

Supply of Research Services to establish MSY proxies for
data-limited stocks (2017-18) to the Marine Institute, Rinville,
Oranmore, Co. Galway. (Ref: ITT17-015)

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SUMMARY

This report summarises the main activities and outcomes of MyDas (MSY proxies for DAta-limited Stocks) with respect to the proposed tasks: i) Stock prioritisation, ii) Data Collation, iii) Framework Development; iv) Performance Appraisal, v) Reference Point Comparison, vi) Liaison and vii) Linkage with other Projects.

The main outcomes were i) R packages with documentation, ii) vignettes with reproducible examples, iii) manuscripts and peer review papers, and iv) evaluations of Management procedures for data poor stocks.

Initially a number of stocks were proposed as potential case studies and then fisheries dependent and independent data and life history characteristics were collated. After discussion with the Marine Institute (MI) seven case study stocks, i.e. sprat, brill, turbot, pollack, ray, razor clams and lobster, were selected, based on ecological and economic importance.

Data-poor stock assessment models in use worldwide were reviewed and simulation tested. The best performing methods were identified and then implemented in a Management Strategy Evaluation framework based on R. Two R packages were developed: **FLife** for simulating stocks based on life history theory and **mydas** for running and testing data poor harvest control rules. The packages include a number of live vignettes with examples (links provided herein). The framework was tested at an MI workshop and results presented at the ICES Workshop On The Development Of Quantitative Assessment Methodologies Based On Life-History Traits, Exploitation Characteristics, And Other Relevant Parameters For Data-Limited Stocks (WKLIFE).

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1. Introduction

The goal of MyDas was to develop and test a range of assessment methods to establish Maximum Sustainable Yield (MSY) or proxy MSY reference points, across the spectrum of data-limited stocks.

Case studies were identified based on their economic and ecological importance and then appropriate datasets on life histories, commercial catches and surveys collated. Candidate data poor assessment methods were identified and their performance evaluated using simulation. A Management Strategy Evaluation (MSE) framework for data poor stocks was developed in R, using [FLR](#), a collection of tools for quantitative fisheries science, developed in the R language [[Kell et al., 2007](#)]. This required extending existing FLR packages; for example [FLife](#) was extended to include methods to build simulation models based on life history theory. A new package [mydas](#) was developed to allow the assessment methods to be run with FLR and to provide tools for simulation of data advice frameworks.

As well as liaison with the Marine Institute and connected projects, MyDas supported participation at the ICES Workshop On The Development Of Quantitative Assessment Methodologies Based On Life-History Traits, Exploitation Characteristics, And Other Relevant Parameters For Data-Limited Stocks ([WKLIFE](#)).

Outputs from MyDas included

- R packages with assessment models for data-limited stocks implemented in an MSE framework.
- R Vignettes describing the methods and findings.
- Publication in peer-reviewed journals on data poor approaches methods.

Each of these outputs is detailed in the following sections and mapped against the project tasks, which are described next.

2. Tasks

There were seven tasks i.e. i) Stock prioritisation, ii) Data Collation, iii) Framework Development; iv) Performance Appraisal, v) Reference Point Comparison, vi) Liaison with Marine Institute, and vii) Linkage with other Projects. Here we summarise the main activities and outcomes, and provide links to appropriate reports and cloud based resources.

2.1 Stock Prioritisation

- **Objective** A number of potential stocks were identified at the start of the project from which the final choice of case studies were selected based on economic value; importance of the species to the ecosystem (i.e. key-stone species); sensitivity to the impacts of fishing; and availability of data [include hyperlink to PSA work here?].
- **Outcome** The final case studies chosen were sprat, brill, turbot, thornback ray, pollack, lobster, and razor clams, and reflect a range of valuable stocks, ecological roles and [life histories](#).

2.2 Data collation

- **Objective** Data collation ran in parallel with other tasks, and relied on existing data sets which were either available from the Marine Institute, other European labs/agencies or publicly available.
 - The main datasets were fishery independent surveys, commercial catches, stock assessment inputs and outputs and life history parameters.
- **Outcome** The data were collated into a PostgreSQL database [need to populate this appendix] and R code [make link live] developed for processing.
 - At the end of the project the data were archived on [google drive](#).

2.3 Assessment Methods

- **Objective** A number of data limited methods already exist, in order to allow them to be simulation tested and their performance compared they were implemented within a common R framework.
- **Outcome** Data limited methods in use worldwide were summarised in a [google spreadsheet](#). Following a review five catch-only and three length-only assessment methods were then compared, and the best performing methods identified.
 - Two peer review papers were produced that evaluated length based methods [Pons et al., 2019] and compared length and catch based methods [Pons et al., submitted].
 - The best performing methods were Length Based Spawning Potential Ratio [LBSPR Hordyk et al., 2014] and Depletion Based Stock Reduction Analysis [DBSRA Dick and MacCall, 2011].
 - These were then implemented as in the mydas package, see the vignettes `lbspr` and `bdsra` [echo = TRUE these vignettes for reproducibility?].

2.4 Performance Appraisal

- **Objective** Method performance appraisal was based on the development of a set of diagnostics that can be applied across a range of models to assess the stability of the model, sensitivity to assumptions and bias.
- **Outcome** A simulation framework based on R. This was based on [FLR](#), a collection of tools for quantitative fisheries science developed in the R language, that facilitates the construction of simulation models of fisheries systems. Development focussed on two main R packages, [FLife](#) which uses life history theory to build simulation models and [mydas](#) which provides an interface to the data poor methods for simulation testing. [Vignettes](#) document the approach and provide worked examples.
 - Operating Models conditioned on life history theory were developed for each of the [case studies](#).
 - An [Observation Error Model](#) was developed to simulate pseudo data under a range of assumptions to allow the bias and sensitivity of stock estimates and reference points to be estimated.
 - Vignettes include examples of conducting [MSE](#) and the use of [parallel computing](#), estimation of [proxy reference points](#), [OM conditioning](#), calculation of [performance metrics](#), and simulation testing [length](#) and running [catch](#) [typo in vignette title] based methods.

2.5 Simulation

- **Objective** Once reference points are identified, their performance was evaluated through simple Management Strategy Evaluations.
- **Outcome**
 - Simulation testing was conducted for the ICES control rule using the `FLife` package for data limited stocks [[Fischer et al., submitted](#)].
 - A comparison of [length based indicators](#) was conducted and a peer review paper is currently in preparation.
 - An [empirical control rule](#) was simulation tested using MSE, and a peer review paper is in preparation.

2.6 Liason

- **Objective** Meetings on a regular basis with the MI.
- **Outcome** Meetings at the Marine Institute ensured that the over-arching goals of the project were achieved.
 - Regular [presentations](#) were made at the MI and a Workshop was held where the methods were tested.
 - In addition to ensure good communication between the consortium and the Marine Institute and that the project keeps focus and delivers as well as the scheduled meetings code, data and results were made available on the cloud and in a github repository during the life of the project. This has now been turned into the `mydas` package [wiki](#)

2.7 Linkage

- **Objective** Linkage with other projects
- **Outcome**
 - The MyDas framework was developed through case studies in collaboration with the MI.
 - Direct links to Cullen fellowship project *Monkfish management strategy evaluation* (CF/16/03) were established. In particular, Ph.D. candidate Luke Batts is testing stage-based stock assessment methods using monkfish stocks simulated with `FLife`. This project has also made extensive use of the MSE frameworks established within MyDas and FLR to test key uncertainties in advice for monkfish stocks.
 - Attendance at Workshop on the Development of Quantitative Assessment Methodologies based on LIFE-history traits, exploitation characteristics, and other relevant parameters for data-limited stocks ([WKLIFE](#)) and the Working Group for the Celtic Seas Ecoregion ([WGCSE](#)), see presentations.
 - * [WKLIFE VIII](#)
 - * [WKLIFE IX](#)
 - Attendance at the [first Irish Environmetrics Forum](#) where methods for policy-relevant advice were presented and discussed.

3. Case Studies

A number of example stocks were identified in the original call, i.e.

- Sprat in the Celtic Sea and West of Scotland Sprat (Sub-area VI & Divisions VIIa-c and f-k)
- Grey gurnard VI & VII (excl. VIId)
- Ling IIIa, IVa, VI, VII, VIII, IX, XII, and XIV
- Rays, primarily in areas VIIa,f,g
- John dory in ICES Sub-area VII and Divisions VIIa,b and d (Northeast Atlantic)
- In collaboration with Newport STO:
 - Saithe VII, VIII, IX, X
 - Pollock VII
- Turbot VIIe,f,j,h and sub area VIII and IXa
- Brill VII (or suitably defined)

Available data sources and relevant publications were reviewed (Table 1). Following which the final choice of case studies was then made in liaison with the Marine Institute based on economic value of the stock; importance of the species to the ecosystem (key-stone species); sensitivity to the impacts of fishing and available data.

| Species | TAC | Commercial Data | Data | Comments |
|-------------------------|-----|---|---|--|
| Sprat | No | Targeted species for small fleet | Poor; mainly landings weights | Key-stone prey fish |
| Gurnards | No | Nearly 100% discarded & Reasonable discard and survey data. No age data | Key-stone prey – widely distributed and abundant | |
| Saithe, Pollock Ling | Yes | Mixed fishery | Some port sampling, observer and survey data. Very limited age data | Key-stone predator |
| Rays, Skates | Yes | Targeted and mixed fishery | Some port sampling, observer and survey data. No age data | Sensitive species – slow reproduction |
| John Dory | No | Mixed fishery but can be targeted to an extent | Some port sampling, observer and survey data. No age data | Sensitive species – valuable non-TAC species (not protected by fisheries management) |
| Turbot, Brill | No | Mixed fishery but can be targeted to an extent | Some port sampling, observer and survey data. Very limited age data | Sensitive species – valuable non-TAC species (not protected by fisheries management) |

Table 1: Summary of potential study stocks.

The final case studies were [sprat](#), [pollack](#), [brill](#), [turbot](#), [thornback ray](#), [lobster](#) and [razor clams](#) as these represent a range of life histories, and have a variety of roles in maintaining the structure of pelagic and demersal communities. The data on vertebrate life histories were accessed directly from [fishnets](#).

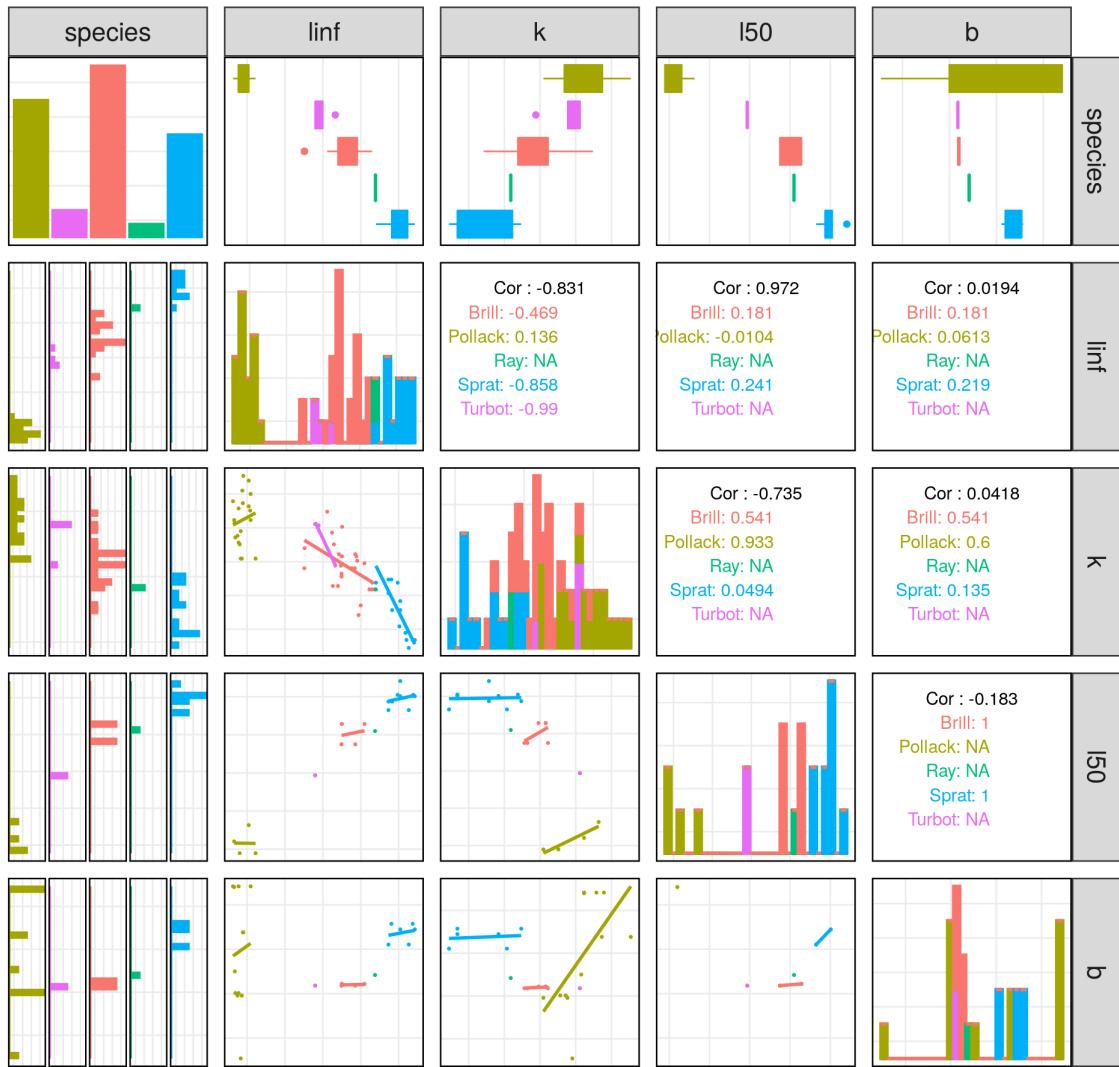


Figure 1: Life history parameters for case study vertebrate stocks.

4. Data Collation

The project relied on existing datasets rather than collecting new data and collating these data ran in parallel with other tasks. As well as Marine Institute data datasets were also obtained from ICES¹ and the JRC². While life history parameters were obtained from Fishbase³ and fishnets⁴, and stock assessment inputs and outputs from stockassessment.org⁵.

A database was designed see the [database summary](#) in the Appendix. These data were [summarised](#) to help identify the case studies. To select the case studies a key dataset were the life history parameters. For data poor stocks where information on key processes is lacking life history theory can be used allowing simulation models to be developed. since many studies have shown that relationships between life history traits exists for processes such as growth, maturity and natural mortality. See the vignette on [conditioning Operating Models](#) on life history parameters.

¹<https://www.ices.dk/marine-data/dataset-collections/Pages/default.aspx>

²<https://stecf.jrc.ec.europa.eu/dd/medbs/ram>

³<http://www.fishbase.org/search.php>

⁴<https://github.com/fishnets/fishnets>

⁵<https://www.stockassessment.org>

The various preliminary analyses based are archived on [google drive](#)

5. Assessment Methods

A number of data-limited methods already exist, in order to compare their performance they need to be run in a common framework. Therefore methods were interfaced to R, using the Fishery Library in R (**FLR**) [Kell et al., 2007] a toolset that is composed of a variety of packages, covering the various steps in the fisheries advice and simulation workflow. Using **FLR** means that advantage can be taken of existing methods, and that dissemination and support is easier and will be maintained after the life of the project.

To identify the appropriate advice given the quality of data rules ICES classifies stock assessments into six [categories](#) on the basis of the available data, e.g. total landings, indices of abundance, length frequency and age data.

Category 1: stocks with quantitative assessments Full analytical assessments and forecasts as well as stocks with quantitative assessments based on production models.

Category 2: stocks with analytical assessments that are only treated qualitatively Quantitative assessments and forecasts which for a variety of reasons are considered indicative of trends in fishing mortality, recruitment, and biomass.

Category 3: stocks for which survey based assessments indicate trends Survey or other indices are available that provide reliable indications of trends in stock metrics, such as total mortality, recruitment, and biomass.

Category 4: stocks for which only reliable catch data are available Time series of catch can be used to approximate MSY.

Category 5: landings only Only landings data are available.

Category 6: negligible landings Landings are negligible in comparison to discards and stocks that are primarily caught as bycatch species in other targeted fisheries

For data limited stocks a number of methods already exist, for example the ICES Working Group on Category 3 and 4 stocks ([WGMSYC AT34](#)) has been working on rules for survey-based assessments which indicate trends in stock status (Category 3) or for which only catch data are available (Category 4).

Data limited methods in use worldwide were summarised in a [google spreadsheet](#). Following which five catch and three length based methods, were then selected for simulation testing. The choice of methods was based on the availability of code, whether the method had been peer reviewed and the appropriateness of estimated quantities for use as proxy reference points. The methods were then simulation tested [Pons et al., submitted] in order to select the best catch and length-only methods.

Catch Only

Catch-only assessment methods evaluated were Catch-Maximum Sustainable Yield [Catch-MSY Martell and Froese, 2013], Depletion Based Stock Reduction Analysis [DBSRA Dick and MacCall, 2011], Simple Stock Synthesis [SSS Cope, 2013], an extension of Catch-MSY (CMSY), and State Space Catch Only Method [SSCOM Thorson and Cope, 2015]. Length based methods evaluated were Length Based Spawning Potential Ratio [LBSPR Hordyk et al., 2014, 2015], Length-Based Integrated Mixed Effects [LIME Rudd and Thorson, 2017], and Length-Based Bayesian [LBB Froese et al., 2018].

Catch-MSY is based on stock reduction analysis (SRA) that assume a Schaefer biomass dynamic model. Inputs are a time series of removals, priors for the population rate of increase at low population

size (r), carrying capacity (K), and a probability distribution of stock depletion in the final year. CMSY, extends Catch-MSY by using a Monte-Carlo filter to fix systematic biases in the Catch-MSY.

DBSRA further modifies the SRA approach by using Monte Carlo draws from the parameter distributions for M , F_{MSY}/M , B_{MSY}/B_0 , depletion, and age at maturity (A_{mat}) to separate total biomass into immature and mature biomass using a delay-difference production model.

SSS is based on Stock Synthesis [Methot Jr and Wetzel, 2013], and fixes all parameters in a Stock Synthesis model except for initial recruitment ($\ln R_0$). An artificial index of abundance that represents the relative stock biomass is used for fitting where the first value of the index is 1, and the value in the final year represents depletion (i.e. the proportion of the population left in the final year). Values of steepness (h) and depletion are randomly drawn from a specified distribution using a Monte Carlo approach and $\ln R_0$ is then estimated.

SSCOM is a Bayesian state-space model that integrates across three stochastic functional forms, variation in effort, population dynamics and fishing efficiency.

Length Only

LBSPR, estimates the proportion of the unfished reproductive potential per recruit (SPR) under a given level of fishing pressure. In an exploited population is calculated as a function of the ratio of fishing mortality to natural mortality (F/M), selectivity, and the two life-history ratios M/k and L_m/L_∞ ; where k is the von Bertalanffy growth coefficient, L_m is the size of maturity and L_∞ is asymptotic size. The inputs are length at maturity specified in terms of L_{50} and L_{95} (the size at which 50% and 95% of a population matures) and length frequency data. LBSPR is an equilibrium based method and assumes asymptotic selectivity, von Bertalanffy growth, length at-age is normally distributed, rates of natural mortality are constant across adult age classes, recruitment is constant over time, and growth rates remain constant across the cohorts within a stock.

LIME like LBSPR uses biological information and the length composition of the catch to estimate F and SPR. Unlike LBSPR, however, it does not assume equilibrium conditions and uses mixed effects to estimate changes in recruitment and fishing mortality separately over time.

LBB is a simple and fast method for estimating relative stock size that uses a Bayesian Monte Carlo Markov Chain (MCMC) approach. LBB uses pre-specified priors on parameters, and thus, technically does not require any inputs other than to length frequency data. It does provides the user the option to specify priors for L_∞ , length at first capture (L_c), and relative natural mortality (M/k). F/M is estimated over the age range represented in the length-frequency sample.

Evaluation

The methods were simulation tested for three life history types (short, medium and long-lived), three historical exploitation scenarios, and three depleted levels [Pons et al., submitted]. The results of the simulations are shown in Figure 5.. Of the five catch based methods SSCOM performs poorly in terms of precision and bias, while SSS is the least bias DBSRA the most precise. The length based methods tend to perform less well than the catch based methods, while LIME is the least biased while LBSPR is the most precise.

Bias and precision are both important factors to consider when assessing fish stocks. Bias reflects how close an estimate is to a known value; precision reflects reproducibility of the estimate. For example, if an assessment is to be re-conducted every year to monitor the impact of a management measure, a precise but biased method would be able to detect a trend better than an unbiased but imprecise method. Like a scientific instruments, this trade-off require calibration to correct for the bias, and such calibration can be explored using closed-loop simulations (i.e. MSE), where the choice of parameters and reference points in a Management Procedure are tuned (i.e. calibrated) to meet the desired management objectives

as represented by the Operating Model. Thus, a biased method (e.g., DBSRA) may be preferable to one that is less biased, but more imprecise (e.g., LIME). Alternatively, imprecision can be addressed through the choice of the percentile (e.g., median being the 50% percentile value) for the derived model output used by management (e.g., catch or SPR); assuming that the true value is contained within the parameter distribution. For example, instead of taking the median value, one could instead use the derived model output associated with the 40th percentile to incorporate risk tolerance as reflected in the calculated imprecision. Such an approach [Ralston et al., 2011] is used in fisheries management systems to directly incorporate scientific uncertainty (both bias and imprecision), and can also be explored and tuned using MSE.

The best performing catch-only methods were SSS and DBSRA, SSS was the least biased method while DBSRA was the most precise. A problem with SSS is that it is based on SS3 and so is a complicated method that requires a lot of assumptions in the form of fixed parameters. DBSRA in contrast is based on a biomass based stock production function and so requires fewer assumptions, it also allows data poor and data rich assessments to be compared, i.e. if an index of abundance could be developed then an ICES Category 3 assessment could be converted into a Category 1 assessment. This allows the value of information to be evaluated. Therefore DBSRA was selected for further evaluation.

Both LBSPR and LIME use the same inputs, the main difference is that LIME is a non-equilibrium method, takes longer to run than LBSPR and often failed to converge, so although LIME estimates are less biased than LBSPR it is likely to perform poorly in closed loop simulations. LBB performed poorly in terms of both bias and precision. Therefore LBSPR was selected for the next stage of simulation evaluation.

6. Framework

The majority of fish stocks worldwide lack sufficient data on which to base quantitative assessments. In some cases Management Strategy Evaluation [MSE [Geromont and Butterworth, 2014](#)] has been used to develop harvest control rules (HCRs) based on empirical indicators that follow trends in survey estimates of abundance, catch rates or size composition.

A reason for the use of MSE is because it is recognised that the robustness of advice depends on the combination of data, estimation method, choice of reference points as well as the rules used to set management action, i.e. the Management Procedure (MP). When conducting MSE an Operating Model (OM) is used to represent the dynamics of system being managed, and control actions from an MP are fed back into the OM so that its influence on the stock and hence on future fisheries data is propagated through the stock and fishery dynamics.

Conducting MSE requires six steps; namely i) identification of management objectives; ii) selection of hypotheses for the OM; iii) conditioning the OM based on data and knowledge, and possible weighting and rejection of hypotheses; iv) identifying candidate management strategies; v) running the Management Procedure (MP) as a feedback control in order to simulate the long-term impact of management; and then vi) identifying the MPs that robustly meet management objectives.

The MSE framework was based on the Fishery Library in R [[FLR Kell et al., 2007](#)], which comprises a variety of packages that cover the various steps in the fisheries advice and simulation workflow. Using **FLR** means that advantage can be taken of existing methods, and that dissemination and support is easier and will be maintained after the life of the project. Under MyDas development focused on two main packages **FLife** and **mydas**. **FLife** is a package for modelling life history relationships and **mydas** provides a set of tools for simulation and conducting Management Strategy Evaluation (MSE) by providing wrappers for the various assessment methods, implementing Observation Error Models (OEMs) to simulate empirical indicators and other datasets, and to model harvest control rules (HCRs).

FLife was used to condition OMs based on life history relationships and ecological theory, MP were then implemented using **mydas**. The link between the OM and the MP is the **OEM**, which generates fishery-dependent or independent resource monitoring data. The OEM models the uncertainties due to sampling and limited data and so mimics the types of data currently required for each assessment method. In addition the types of data that could be made available in the future,

Simulation tests were also performed without feedback where the OEM was used to generate datasets from the OM. This allows the performance of the candidate methods to be compared, since if there is little correlation between the estimates of reference points and status from a candidate method and the OM then there is little point in running that method in the MSE. Tests were performed for a [category 1 assessment](#) (a biomass dynamic assessment with catch and an index of abundance that estimates stock status and reference points), a [category 3 assessment](#) that used a length based indicator, and a [category 4 assessment](#) for which only catch data are available.

Performance measures

Management objectives under the [Marine Directive](#) of the European Union) that stocks should be exploited sustainably consistent with high long-term yields, have full reproductive capacity in order to maintain stock biomass and that the proportion of older and larger individuals should be maintained (or increased) as they are indicators of a healthy stock. These general objectives can be mapped to performance measures, so that alternative management strategies can be compared. For example

A range of summary statistics are required to illustrate trade-offs between multiple potentially conflicting objectives. Although there are many potential summary statistics so that decision makers can choose between tangible options on the basis of actual projections rather than abstract concepts and

performance statistics, however, should ideally be few, informative and based axes such as stock status, 'safety', 'stability' and 'yield'. It is also necessary to distinguish between technical summary statistics (i.e. those required to evaluate model fits and performance) and those required to evaluate management objectives. Examples of summary statistics include

Safety Probability of avoiding a limit such as B_{lim} where recruitment is impaired

Status Probability of achieving targets related to MSY, e.g. B_{MSY} and F_{MSY}

Yield Yield:MSY

Variability Inter-annual variation in catches or stock status

7. Simulations

ICES is in the process of developing methods to identify MSY proxy reference points for data-limited stocks at the ICES Workshop On The Development Of Quantitative Assessment Methodologies Based On Life-History Traits, Exploitation Characteristics, And Other Relevant Parameters For Data-Limited Stocks ([WKLIFE](#)). **MyDas** contributed to this process by proposing and testing new assessment models and methods of establishing reference points and attended both WKLIFE VIII and IX. There are key differences with the ICES approach, however, since **MyDas** stocks are not currently assessed by ICES and **MyDas** focused on the available data for each stock first and then on methods, while the ICES approach focuses on the methods first and then applies a limited number of methods to a large number of stocks.

The importance of the work of **MyDas** was recognised at WKLIFEVIII where recommendations for [future directions](#) were made. Following this several examples based on the work of **MyDas** were made at WKLIFEIX. These included i) the use of life history theory to condition the OM used to evaluate the ICES advice rule proposed for [category 3 and 4 stocks](#); ii) Receiver Operating Characteristic (ROC) curves to explore the setting of appropriate [reference levels](#) in the the catch rule; and iii) the use of trends in an index [without a reference level](#). It was also recommended to use the methods of **MyDas** to evaluate the robustness of SPiCT based upon the development of Operating Models developed under [MyDas](#).

For stocks with analytical assessments (i.e. category 1 and 2 stocks), the advice rules applied by ICES are consistent with the objective of achieving MSY. For stocks in categories 3 and 4 ICES uses MSY proxy reference points as part of a Precautionary Approach to provide advice on the status of the stock and exploitation. The F_{MSY} proxy corresponds to the exploitation rate that will provide maximum long-term yield, while the $MSY_{Btrigger}$ proxy corresponds to the stock size that triggers a cautious response; i.e. advises a reduced fishing mortality relative to F_{MSY} proxy in order to allow the stock to rebuild.

Advice Rule

For stocks with MSY proxy reference points not derived from an assessment model, ICES uses a generic rule of the form

$$C_{y+1} = C_{current} r f b \quad (1)$$

where $C_{current}$ is the catch either in the most recent year available (typically $y - 1$) or the average over a number of recent years (e.g. $y - n, \dots, y - 1$), r should account for the trend in stock biomass ($r > 0$; with $r = 1$ if there is no trend), f is a proxy for the ratio $F_{MSY}/(\text{current exploitation})$ and $b = \min(1, \text{proxy for the ratio (current stock size)}/(MSY_{Btrigger}))$. The later is in effect a hockey stick type HCR.

f is derived from length based indicators and a number of potential length based reference point and indicators⁶ have been identified by ICES for stocks in categories 3 and 4 (see Table 2)

The **mydas** package includes methods to simulate length frequency data and estimate indicators; Figure 3 shows an OM based on pollack that was used to simulate indicators (Figure 4).

To ensure that advice based on such indicators is precautionary, Management Strategy Evaluation was used to evaluate an empirical rule of the form $C_{y+1} = C_{current} r f b$ that bases catch advice on recent catches, information from a biomass survey index (r), length based indicators (f) and proxy MSY reference point (b) [[Fischer et al., submitted](#)]. The OM was conditioned on life histories using the method developed by **MyDas**. The performance of the rule varied substantially between stocks, and the risk

⁶ 'http://ices.dk/sites/pub/PublicationReports/GuidelinesandPolicies/16.04.03.02_Category_3-4_Reference_Points.pdf'

Table 2 Selected indicators for LBI screening plots. Indicator ratios in bold used for stock status assessment with traffic light system.

| Ind | Select table column | Calculation | Reference point | Indicator ratio | Expected value | Property |
|--------------|---------------------|---|-------------------------------------|------------------------|----------------|----------------------------------|
| $L_{max5\%}$ | | Mean length of largest 5% | L_{inf} | $L_{max5\%} / L_{inf}$ | > 0.8 | Conservation (large individuals) |
| $L_{95\%}$ | | 95 th percentile | | $L_{95\%} / L_{inf}$ | | |
| P_{mega} | | Proportion of individuals above $L_{opt} + 10\%$. (L_{opt} is estimated from L_{inf}). | 0.3 – 0.4 | P_{mega} | > 0.3 | |
| $L_{25\%}$ | | 25 th percentile of length distribution | L_{mat} | $L_{25\%} / L_{mat}$ | > 1 | Conservation (immature) |
| L_c | | Length at 50% of modal abundance* | L_{mat} | L_c / L_{mat} | > 1 | |
| L_{mean} | | Mean length of individuals > L_c | $L_{opt} = \frac{2}{3} L_{inf}$ | L_{mean} / L_{opt} | ≈ 1 | Optimal yield |
| L_{maxy} | | Length class with maximum biomass in catch | $L_{opt} = \frac{2}{3} L_{inf}$ | L_{maxy} / L_{opt} | ≈ 1 | |
| L_{mean} | | Mean length of individuals > L_c | $L_{F=M} = (0.75L_c + 0.25L_{inf})$ | $L_{mean} / L_{F=M}$ | ≥ 1 | MSY |

*Note this definition is different from the L_c used for the Mean-length Z estimator.

Table 2: Length based indicators.

of breaching limit reference points was inversely correlated to the von Bertalanffy growth parameter k . Stocks with $k < 0.32$ had a high probability of stock collapse. It was shown that a single generic catch rule cannot be applied across all life-histories, and management should instead be linked to life-history traits, and in particular, the nature of the time series. This supported the premise of **MyDas** that to develop robust management it is important to consider the nature of the stock dynamics, knowledge and data. Counter-intuitively the catch rule performed poorly for the more productive stocks (those with higher k) compared to the less productive stocks (with lower k). The performance of the catch rule, however, is an emergent property of the interaction between the operating model and the catch rule. Therefore this is an important results which would not have been apparent without the work of our manuscript.

The advised catch was mainly influenced by one component of the rule (the trend in the relative index of abundance) and biomass trends for stocks with higher are inherently more variable, which in turn leads to higher fluctuations in catch. Therefore when managed by the ICES catch rule, the stocks with higher were more likely to collapse during simulation. This behaviour can be attributed to an initial rapid recovery, which resulted in an increase in catch. Once the stocks started to decline again, however, catch was not reduced quickly enough to avoid stock collapse. This undesirable feature is caused by the design of the catch rule, which bases the newly advised catch on the previous catch and observed data with a time-lag. Since the less productive stocks (those with low k) were also less variable, the catch rule was sufficiently reactive to avoid stock collapse.

Therefore two papers are currently being drafted. The first used Receiver Operating Characteristic (ROC) curves used to explore the setting of appropriate reference levels in the f component of the catch rule for the **MyDas** case studies. In particular to explore how these depend on life-history traits and the nature of the time-series. The second evaluated the use of trends in an index without a reference level for pollack. To do this, a Management Strategy Evaluation (MSE) was conducted to evaluate an empirical harvest control rule (HCR) based on a trend in an index of abundance. The Operating Model (OM) was conditioning on pollack life-history characteristics and the HCR was based on that used by the Commission for the Conservation of Southern Bluefin Tuna (CCSBT). The HCR has several parameters that require tuning [Hillary et al., 2015], i.e. the parameters are found by choosing values that best meet the objectives of asset and stakeholders; i.e. optimises the outcomes modelled as a reward function.

8. Liason and Linkage

Liaison with Marine Institute](<https://github.com/laurieKell/mydas/wiki/6-Liaison-with-Marine-Institute>)
The service provider is expected to meet on a regular basis with MI staff involved in the project: Monthly update meetings at the Marine Institute premises in Oranmore Galway with researcher providing the research services, 6-monthly progress reports and meetings at the Marine Institute with the researcher providing the research services and contract manager.

The service provider meet on a regular basis with Marine Institute staff involved in the project. The proposal is ambitious but achievable. However there needs to be good communication between the consortium and the Marine Institute to ensure the project keeps focused and delivers. Therefore we will arrange monthly face-to-face meetings, make all code, data and results available on the cloud and in a suitable repository (e.g. github) and provide a web based interface for model results.

Wider project 6-monthly progress reports and meetings at the Marine Institute will ensure the overarching goals of the project are achieved. Attendance at [wklifeix: 4 Length-based approaches](#) and ICES WKLIFE & Working Group for the Celtic Seas Ecoregion ([WGCSE](#))

Linkage with other projects](<https://github.com/laurieKell/mydas/wiki/7-Linkage-with-other-projects>)
The service provider is required to link research output to the following projects:

The MyDas framework was developed through case studies in collaboration with a range of partners, as well as the MI and ICES, these included the tuna Regional Management Fisheries Organisations (tRFMOs), the University of Washington and the JRC. These case studies resulted in a number of peer review manuscripts.

- Monkfish

This project will develop in close collaboration with the Cullen Fellowship of Mr Luke Batts, co-supervised by Dr Hans Gerritsen (MI) and Dr Cóilín Minto (GMIT). Active collaboration will occur with Tasks 3–5, as these are similarly proposed in the Cullen Fellowship where they are applied specifically to *Lophius budegassa* and *Lophius piscatorius* stocks in ICES areas VII–VIII.

- Pollock

Active collaboration exists between GMIT and the Newport Research Cluster (e.g., *Unlocking the Archive* project). Further collaboration and linkages will be built around data-poor assessment of pollock (liaising with the dedicated Scientific and Technical Officer working on pollock at the Furnace research facility). Both visits to Newport and group attendance at the monthly meetings will facilitate collaboration and crossover.

- DRuMFISH project

The project will also link with the DGMARE project: “Study on approaches to management for data-poor stocks in mixed fisheries (DRuMFISH)” to which GMIT is a partner in the consortium. Methodological development from DRuMFISH (e.g., hierarchical methods) will be directly relevant to the present proposal.

- CPV codes 71354500-9 Marine survey services 73112000-0 Marine research services 90712300-4 Marine conservation strategy planning 98360000-4 Marine services 77700000-7 Services incidental to fishing 73000000-2 Research and development services and related consultancy services 73110000-6 Research services 73200000-4 Research and development consultancy services 73210000-7 Research consultancy services.

- There are also other projects worldwide that can be link to e.g. the global group on stock assessment methods and the tRFMO MSE WG.

- Achievements
- Promises & Differences
- What & Why

wklifeix: 4 Length-based approaches

wklifeix: ROC

wklifeix: hake

wklifeix: hake

Evaluate further improvements to the performance of the WKMSYCat34 catch rule 3.2.1. Focus on improving the catch rule for stocks with von Bertalanffy growth parameter $k_t < 0.32$, investigate more extensively the definition of the catch rule components and their impact on performance, and investigate the possibility of alternative catch rules. Explore the operating model set-up for data-limited simulations, including sensitivity analyses based on the Jacobian; e.g. elasticity analysis, on how the different life-history and fishery parameters affect the simulated stock behaviour under exploitation, an analysis of the nature of time-series and trends of observable stock characteristics (such as fishery-dependent and -independent metrics) and how the knowledge gained can be used to further improve the performance of catch rules

The International Council of the Exploration of the Sea (ICES) is in the process of developing methods to identify MSY proxy reference points for data-limited stocks ([WKLIFE](#)). The service provider is required to contribute to this process by proposing and testing new assessment models and methods of establishing reference points and will be expected to attend up to 4 one-week meetings at ICES headquarters in Copenhagen. However there are key differences with the ICES approach. Since this research contract will include stocks not currently assessed by ICES; focusing on the available data for each stock first and on the methods second; the ICES approach focuses on the methods first and then applies a limited number of methods to a large number of stocks.

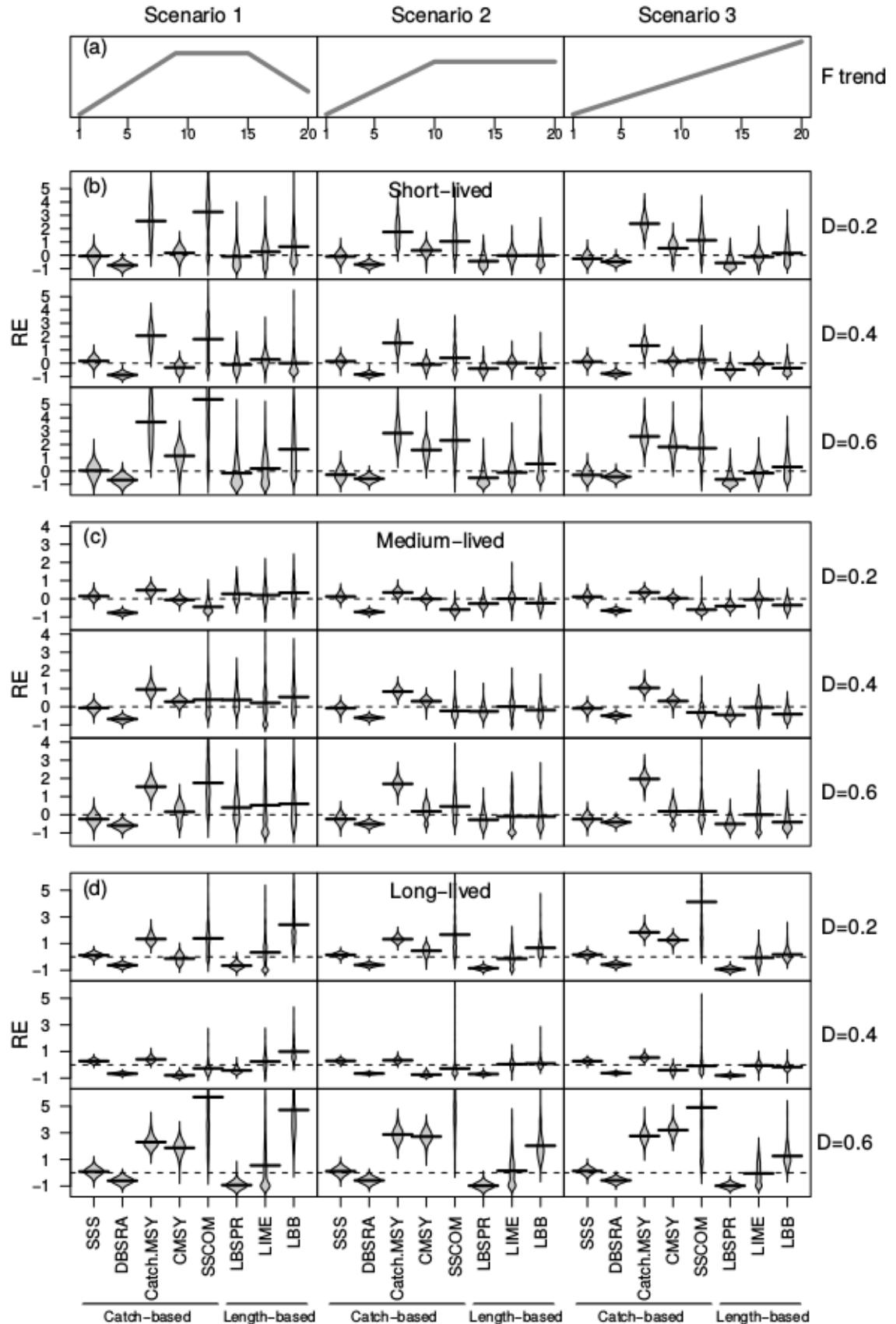


Figure 2: Relative error (RE) in exploitation rate for all the catch-based and length-based models considered under the different harvest scenarios (a) and depletion scenarios for differing life histories, (b) short-lived, (c) medium-lived, and (d) long-lived.

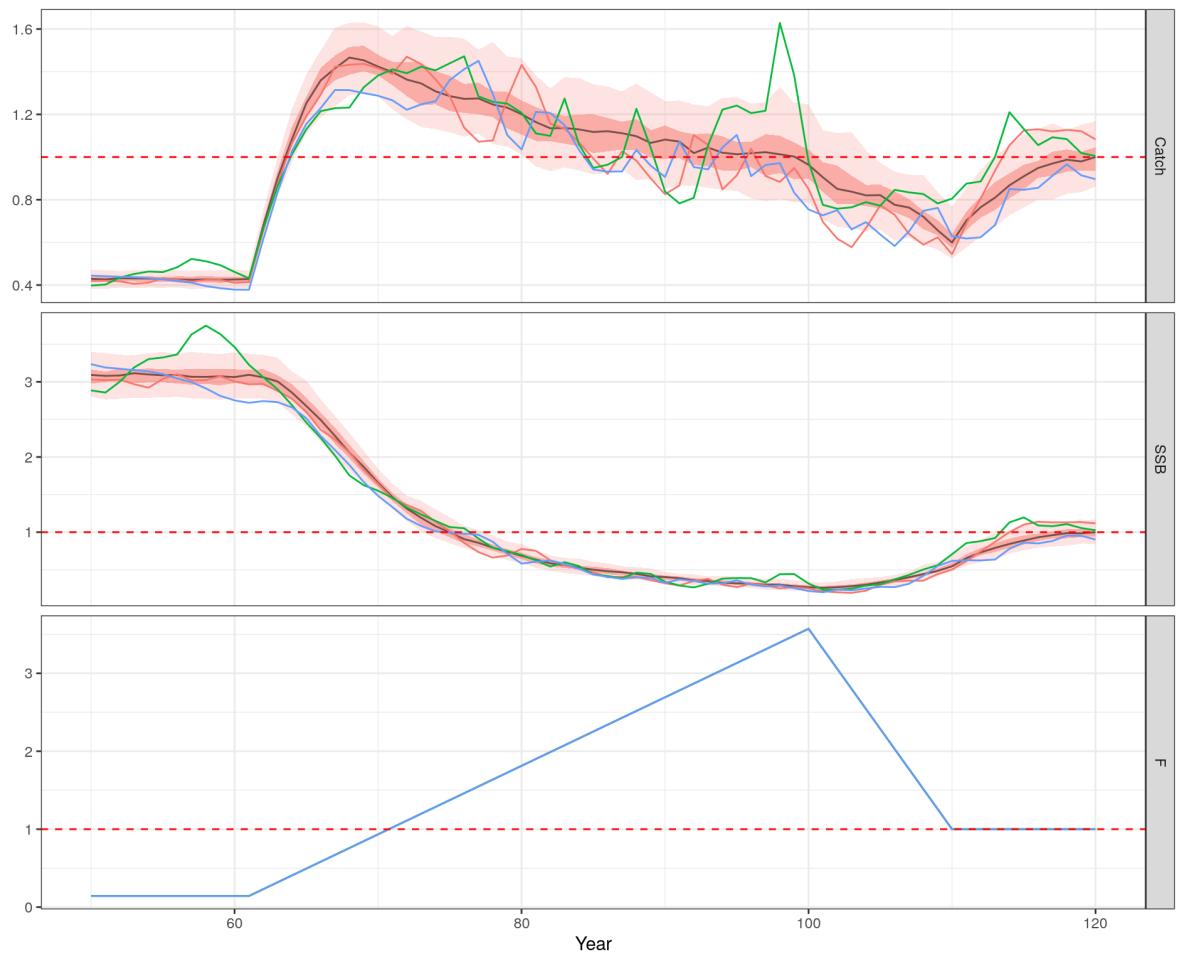


Figure 3: Time series of simulated length based indicators, compared to F/F_{MSY}

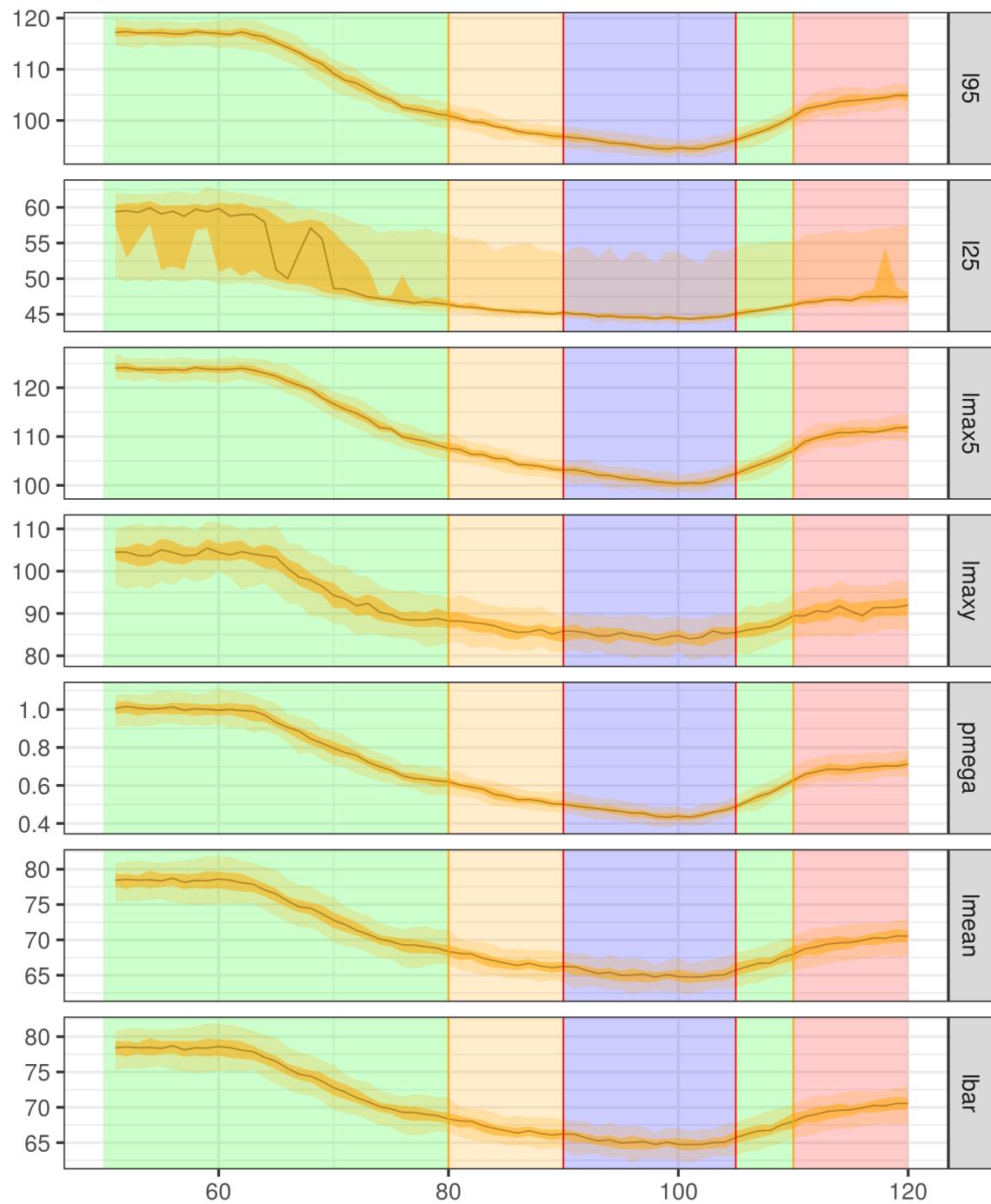


Figure 4: Time series of simulated length based indicators, compared to F/F_{MSY}

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A

A1 *Database*

PostgreSQL database mydas_dev

data_faocodes (includes price – source Hans Gerritsen Marine Institute)

| | Variable name | Type | Description |
|----------|-----------------|-----------|--|
| 1 | species_fao | Character | e.g. BLL, SPR, LIN, GUG, SKA, POK, POL,TUR, JOD. |
| 2 | scientific_name | Character | Latin names e.g. <i>Sprattus sprattus</i> . |
| 3 | english_name | Character | Sprat. |
| 4 | speciesgroup | Character | e.g. ANG, MON, ANF = Monkfish. |
| 5 | priceperkg | Double | Irish market prices 2015. |

data_icesrects (source ICES web site)

| | Variable name | Type | Description |
|----------|----------------|-----------|-------------------------|
| 1 | longitude | Double | Decimal degrees |
| 2 | latitude | Double | Decimal degrees |
| 3 | ices_rectangle | Character | ICES Rectangle code |
| 4 | area_km2 | Integer | Area of ICES rectangle. |
| 5 | ecoregion | Character | Ecological region |

data_icesrects_div (source ICES website)

| | Variable name | Type | Description |
|----------|----------------|-----------|--|
| 1 | area_27 | Character | FAO area code and division notation, i.e. 10.a.1 |
| 2 | ices_rectangle | Character | ICES Rectangle code |
| 3 | ices_division | Character | ICES division. |

data_lhistories (life histories of MYDAS species sourced from literature)

| | Variable name | Type | Description |
|-----------|---------------|-----------|--|
| 1 | species | Character | Latin name. |
| 2 | speciesgp | Character | e.g. grouping blue ling and common ling into 1 var BLI, LIN = LIN. |
| 3 | sex | Character | Code for sex of individual fish i.e. M = Male, F=Female, U=Unsexed |
| 4 | t.0 | Double | Time at which fish is at 0 length |
| 5 | linf | Double | Asymptotic maximum length (cm) |
| 6 | k | Double | Growth rate. |
| 7 | tm | Double | Age at maturity |
| 8 | lmat | Double | Length at maturity (cm) |
| 9 | area | Character | Area experiment was performed i.e. North Sea etc. |
| 10 | source | Character | Reference from literature source (table on shiny) |

data_psa_default (life histories (above) of MYDAS species together with this table produce interactive page for shiny app for year 2015)

| | Variable name | Type | Description |
|-----------|---------------|-----------|---|
| 1 | stock | Character | i.e. Speciesgroup BLL.7 for area 7 catching BLL. |
| 2 | gear | Character | Beam, Otter etc.. |
| 3 | speciesgp | Character | e.g. grouping blue ling and common ling into 1 variable BLI, LIN = LIN. |
| 4 | totland | Double | Total landings in tonnes. |
| 5 | totfleetarea | Double | Total area gear went km ² |
| 6 | stockarea | Double | Total area for the stock |
| 7 | olap_percent | Double | Totfleetarea/ stockarea x 100 |
| 8 | price | Double | Price per kg of the species |
| 9 | score_olap | Integer | 1 to 3 overlap score 3 is high susceptibility. |
| 10 | score_price | Integer | 1 to 3 price score 3 is high susceptibility. |
| 11 | score_catch | Integer | 1 to 3 catchability 3 is high susceptibility. |
| 12 | score_postc | Integer | 1 to 3 post capture mortality 3 is high susceptibility. |
| 13 | tm_score | Integer | 1 to 3 ages at maturity 1 is high productivity. |
| 14 | fec_score | Integer | 1 to 3 fecundity score 1 is high productivity. |
| 15 | repro_score | Integer | 1 to 3 reproductive score 1 is high productivity. |
| 16 | troph_score | Integer | 1 to 3 trophic level score 1 is high productivity. |
| 17 | lmat_score | Integer | 1 to 3 length at maturity score 1 is high productivity. |
| 18 | linf_score | Integer | 1 to 3 asymptotic maximum length 1 is high productivity. |

data_speciescodes (look up table)

| | Variable name | Type | Description |
|----------|-----------------|-----------|--|
| 1 | species_fao | Character | Species FAO codes for MYDAS species |
| 2 | scientific_name | Character | Latin names e.g. <i>Sprattus sprattus</i> |
| 3 | english_name | Character | Sprat. |
| 4 | speccode | Integer | WORMS code |
| 5 | speciesgp | Character | e.g. grouping blue ling and common ling into 1 var BLI, LIN = LIN. |

data_stecf_aer_cpuedays (summary table to calculate cpue for shiny app from STECF annual economic report)

| | Variable name | Type | Description |
|----------|---------------|-----------|--|
| 1 | year | Integer | Year of fishing activity |
| 2 | speciesgp | Character | e.g. grouping blue ling and common ling into 1 var BLI, LIN = LIN. |
| 3 | country_code | Character | ESP = Spain etc. See appendix |
| 4 | gear_type | Character | Fishing activity etc BEAM = beam trawl see appendix. |
| 5 | vessel_length | Character | Length of fishing boat. |
| 6 | totval | Double | Total value in Euros (€) |
| | totctch | Double | Total catch in tonnes (t) |
| | totdays | Double | Total days fished. |

data_stecf_aer_econ (summary table to calculate cpue for shiny app from STECF annual economic report)

| | Variable name | Type | Description |
|----------|---------------|-----------|---|
| 1 | year | Integer | Year of fishing activity |
| 2 | speciesgp | Character | e.g. grouping blue ling and common ling into 1 var BLI, LIN = LIN. |
| 3 | country_code | Character | ESP = Spain etc. See appendix |
| | country_name | Character | Spain |
| 4 | fishing_tech | Character | Fishing activity cluster constructed fleet segment. See AER report. |
| 5 | variable_name | Character | e.g. vessel tonnage. |
| 6 | variable_code | Character | Total vessel tonnage = totgt |
| | value | Double | Value of variable code |

data_stecflandings (just for mydas species/areas) and (data_otherstecflandings – all species)

| | Variable name | Type | Description |
|-----------|----------------|-----------|--|
| 1 | annex | Character | ?? |
| 2 | area | Character | ICES area see appendix below |
| 3 | country | Character | ESP = Spain etc. See appendix |
| 4 | fishery | Character | ?? |
| 5 | gear | Character | Fishing gear |
| 6 | landings | Double | Landings in tonnes |
| 7 | mesh | Character | Meshsize range e.g. 100-119 |
| 8 | quarter | Integer | Given as 1-4 |
| 9 | ices_rectangle | Character | ICES Rectangle code |
| 10 | regulated_area | Character | ?? |
| 11 | regulated_gear | Character | ??? |
| 12 | species | Character | 3 letter FAO code |
| 13 | speccon | Character | ??? |
| 14 | length | Character | Vessel length |
| 15 | year | Integer | Year of fishing |
| 16 | division | Character | ICES division |
| 17 | stock | Character | Mydas stock grouping |
| 18 | speciesgp | Character | e.g. grouping blue ling and common ling into 1 var BLI, LIN = LIN. |
| 19 | longitude | Double | Midpoint of ices rectangle |
| 20 | latitude | Double | Midpoint of ices rectangle |
| 21 | area_km2 | Integer | Area of ICES rectangle |
| 22 | ecoregion | Character | Ecological region |

data_stockprior (summary table resulting from PSA and economic weighting – see stock prioritisation code on mydas wiki)

| | Variable name | Type | Description |
|----------|---------------|-----------|------------------------|
| 1 | year | Character | Year of fishing |
| 2 | country | Character | Country i.e ESP |
| 3 | gear | Integer | Fishing gear i.e BEAM. |

| | | | |
|----------|-----------|-----------|---|
| 4 | rank | Double | Weighted ranked vulnerability |
| 5 | value | Double | Total value of catch in € |
| 6 | speciesgp | Character | e.g. grouping blue ling and common ling into 1 var BLI, LIN = LIN. |

div_area (also geometric table for shiny called ices areas shapefile)

| | Variable name | Type | Description |
|----------|---------------|-----------|---|
| 1 | subarea | Integer | 3, 4, 5 etc.. |
| 2 | subdiv | Character | d, e, f, g... etc. |
| 3 | area_full | Character | Full code for ICES area |
| 4 | area_27 | Character | Full code for ICES area 27 (North Atlantic) |
| 5 | area_km2 | Double | Area of sub region in km2 |
| 6 | division | Character | ICES division. |

For tables data_surveybio/data_surveystns (see DATRAS @ ICES)

Appendix 1 Country coding

| COUNTRY | CODE |
|-------------------------------------|------|
| Belgium | BEL |
| Denmark | DEN |
| Estonia | EST |
| Finland | FIN |
| France | FRA |
| Germany | GER |
| Ireland | IRL |
| Latvia | LAT |
| Lithuania | LIT |
| Netherlands | NED |
| Norway | NOR |
| Poland | POL |
| Portugal | POR |
| Spain | SPN |
| Sweden | SWE |
| United Kingdom (Jersey) | GBJ |
| United Kingdom (Guernsey) | GBG |
| United Kingdom (Alderney/Sark/Herm) | GBC |
| United Kingdom (England and Wales) | ENG |
| United Kingdom (Isle of Man) | IOM |
| United Kingdom (Northern Ireland) | NIR |
| United Kingdom (Scotland) | SCO |
| Other countries | OTH |

Appendix 2 Gear coding

| TYPES OF FISHING TECHNIQUES | | | Gear code |
|-----------------------------|---------------------------------|---------------------------|-----------|
| Mobile gears | Beam trawl | | BEAM |
| | Demersal trawl & demersal seine | Bottom trawl | OTTER |
| | | Danish & Scottish seiners | DEM_SEINE |
| | Pelagic trawl & Seiners | Pelagic Trawl | PEL_TRAWL |

| | | |
|---------------|--|-----------|
| | Pelagic seiner & purse seiner | PEL_SEINE |
| Dredges | | DREDGE |
| Passive gears | Longlines | ONGLINE |
| | Drift & fixed Nets except Trammel Nets | GILL |
| | Trammel Nets | TRAMMEL |
| | Pots & traps | POTS |

Appendix 3 Mesh size coding

| Gear type | Mesh size range |
|---------------|-----------------|
| Mobile gears | <16 |
| | 16-31 |
| | 32-54 |
| | 55-69 |
| | 70-79 |
| | 80-89 |
| | 90-99 |
| | 100-119 |
| | >=120 |
| | |
| Passive gears | 10-30 |
| | 50-59 |
| | 60-69 |
| | 70-79 |
| | 80-89 |
| | 90-99 |
| | 100-109 |
| | 110-119 |
| | 120-219 |
| | >=220 |

A2 *Dataset Summary*

Dataset for use in data poor methods

Laurence Kell and Alex Tidd

25 December, 2018

MYDAS

Task 5 of the MYDAS project <https://github.com/laurieKell/mydas> requires a review of data available on length composition, mean size, catch and indices of abundance (commercial or survey).

Data are available from the MI or from collective sources such as DATRAS <http://www.ices.dk/marine-data/data-portals/Pages/DATRAS.aspx> for survey data and the STECF <https://steclf.jrc.ec.europa.eu/data-dissemination> for commercial data.

Work plan

- Condition the operating model using life histories (from table above) from brill, turbot, skate and Pollack as per the Mydas project (see the quick start). Will have to use MI estimates for sprat. Use estimates of length (brill, sprat, skate). Obtain length frequencies for turbot and pollack from MI.

The life histories were reviewed by Hans, using data from the MI databases.

- With further investigation look at predicted abundance estimates for brill, sprat (although on high side) and skate (abundance was for all species of skates and rays, will have to make consistent with commercial species... maybe use thornback only). Pollack and turbot catches are greater than abundance so cannot use these.

If the absolute abundance estimates are wrong relative indices can be used, as normal practice.

- With such good time series of survey length data for brill, sprat and skate it would be good to simulate the mean size by using the mlz package <https://cran.r-project.org/web/packages/MLZ/vignettes/MLZ.html#introduction>. This would give an estimate of Z (assuming constant M) and thus look at changes in F.

MLZ uses mean size, it is unlikely to provide good estimates of Z for sprat due to high M and high recruitment variability, the mean size is therefore determined by recruitment not F.

- It may also be worth pursuing the LBSR for length-based composition for the above (and compare) with life-history parameters M/K ratio and Linf to estimate F/M and F/FMSY. With SPR being the biological reference point. <https://cran.r-project.org/web/packages/LBSPR/vignettes/LBSPR.html>

LBSPR is better at estimating F.

- For turbot and pollack it maybe beneficially to use an empirical approach (as survey info pretty poor) based on the commercial cpue (biomass index from MI lengths or estimated by weights) such as the ICES 2/3 rule (as per Simon Fischer).

The surveys could be used in the

For the observation model commercial cpues could be used.

Further: Need to discuss with MI what data we can get for Lobsters and Razors in terms of lengths. We have 1 set of priors from Iyes.

Survey data extraction

Extract via sql the case study stocks from Mydas survey database and get numbers at length

For more in depth detail of the data see <https://data.marine.gov.scot/sites/default/files//SMFS%200816.pdf>

Calculate weighted mean length for case study species

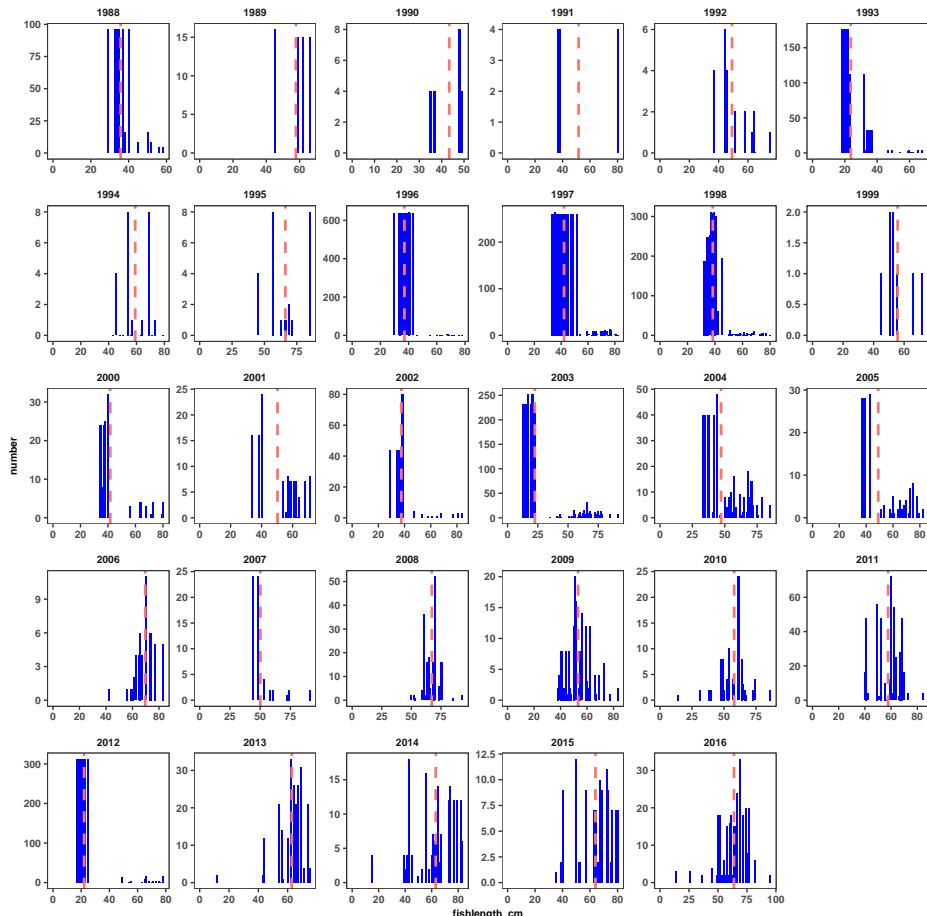


Figure 1 Plot of numbers at length for pollock with weighted mean length represented by dashed red line.

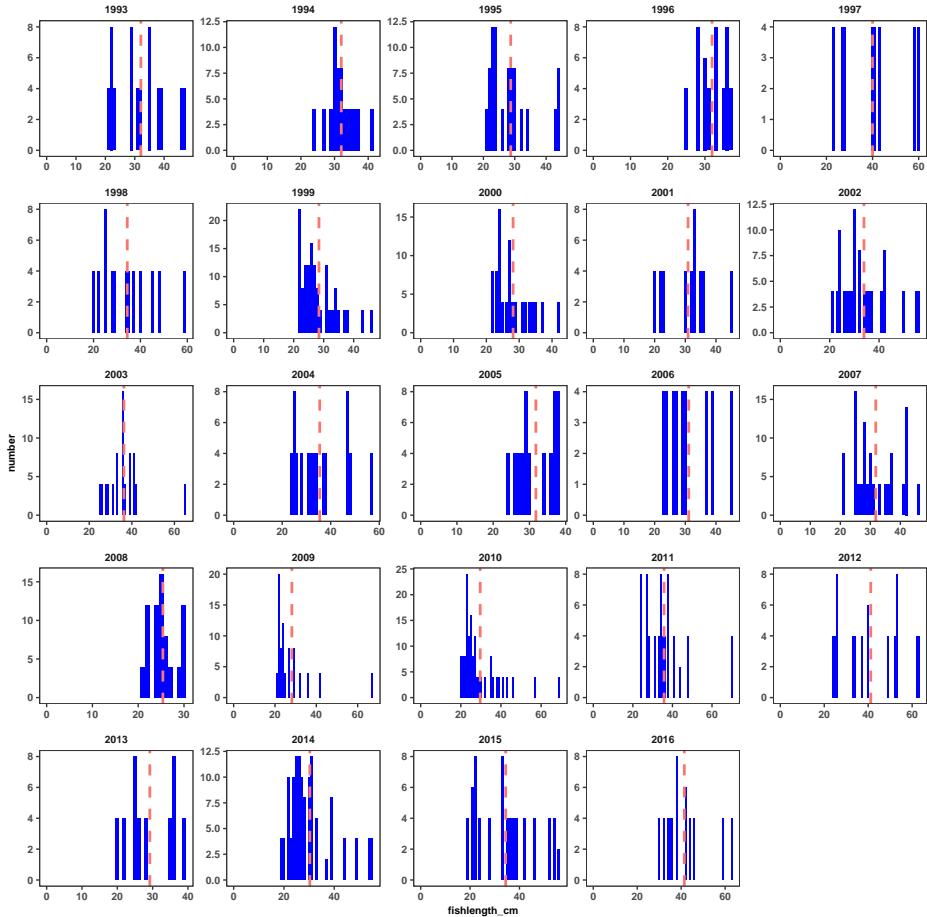


Figure 2 Plot of numbers at length for turbot with weighted mean length represented by dashed red line.

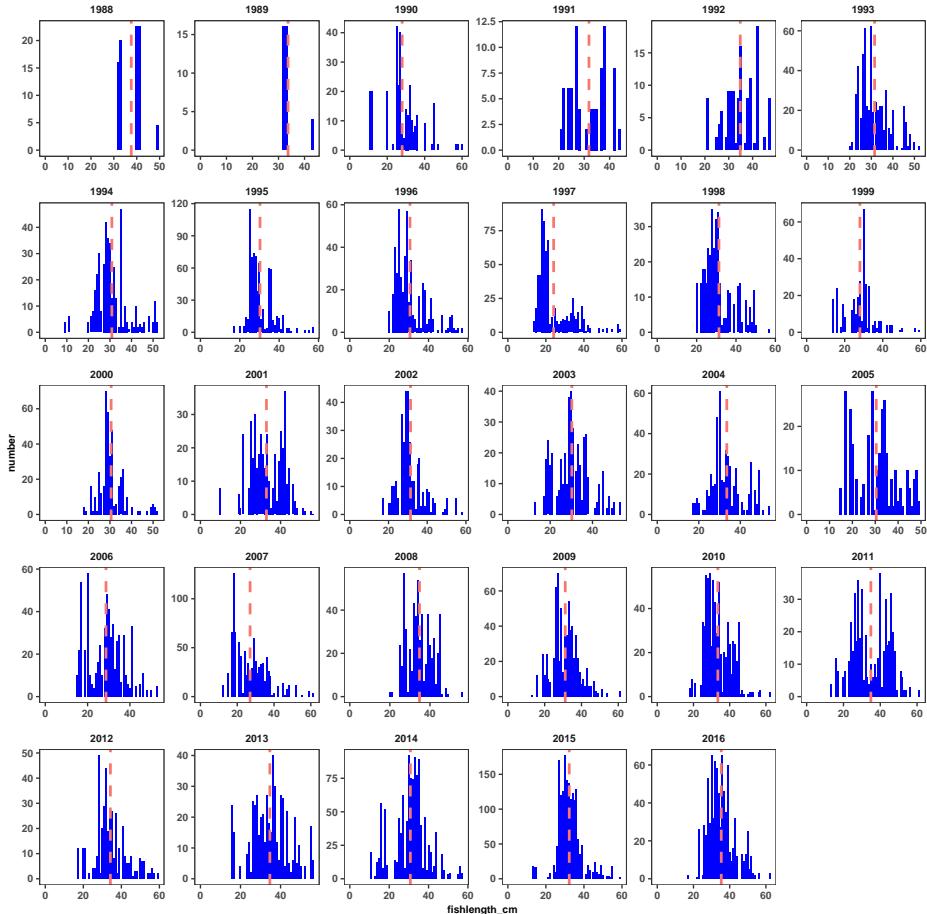


Figure 3 Plot of numbers at length for brill with weighted mean length represented by dashed red line.

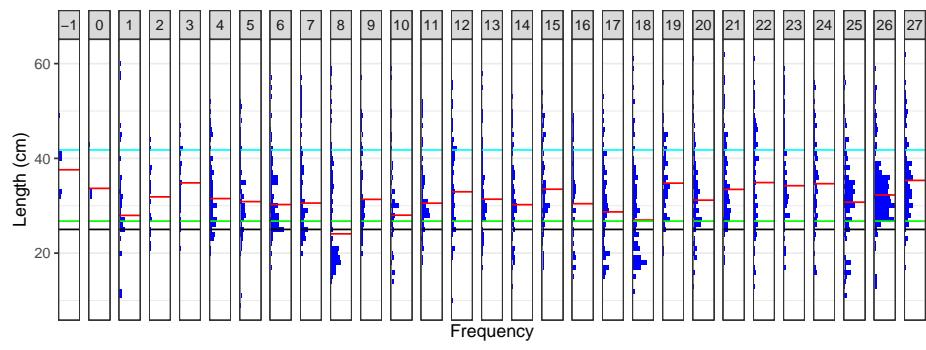


Figure 4 Plot of numbers at length for brill with weighted mean length represented by dashed red line.

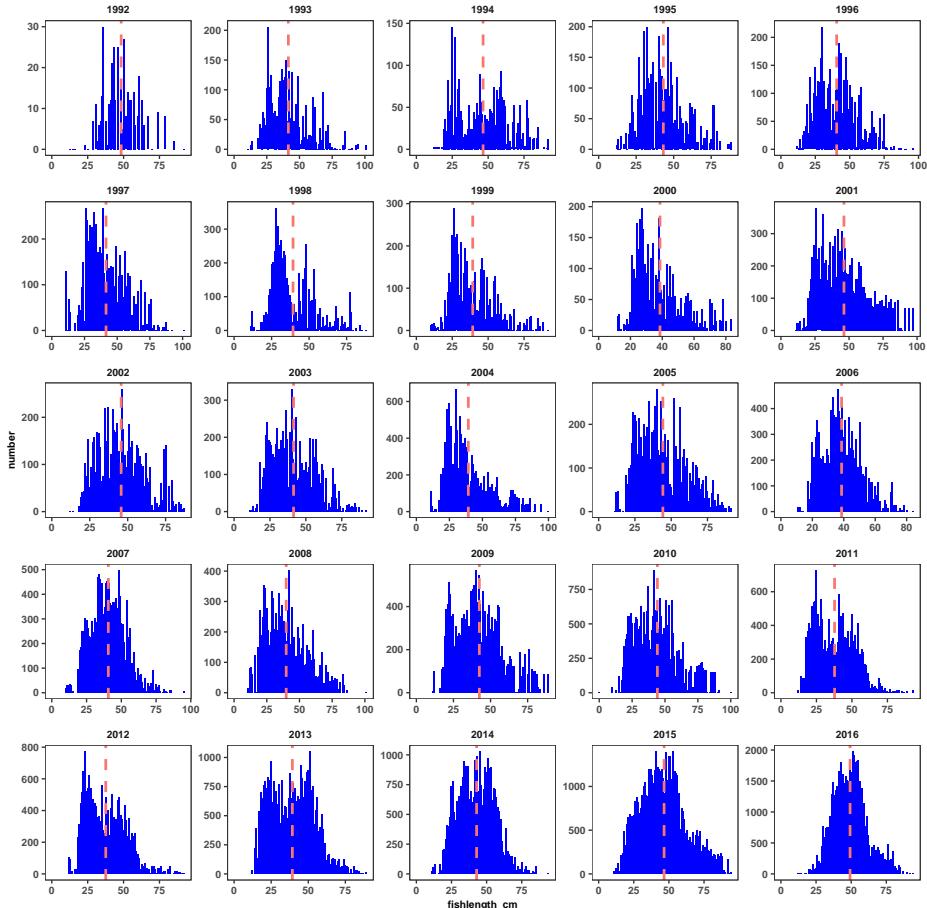


Figure 5 Plot of numbers at length for skate with weighted mean length represented by dashed red line.

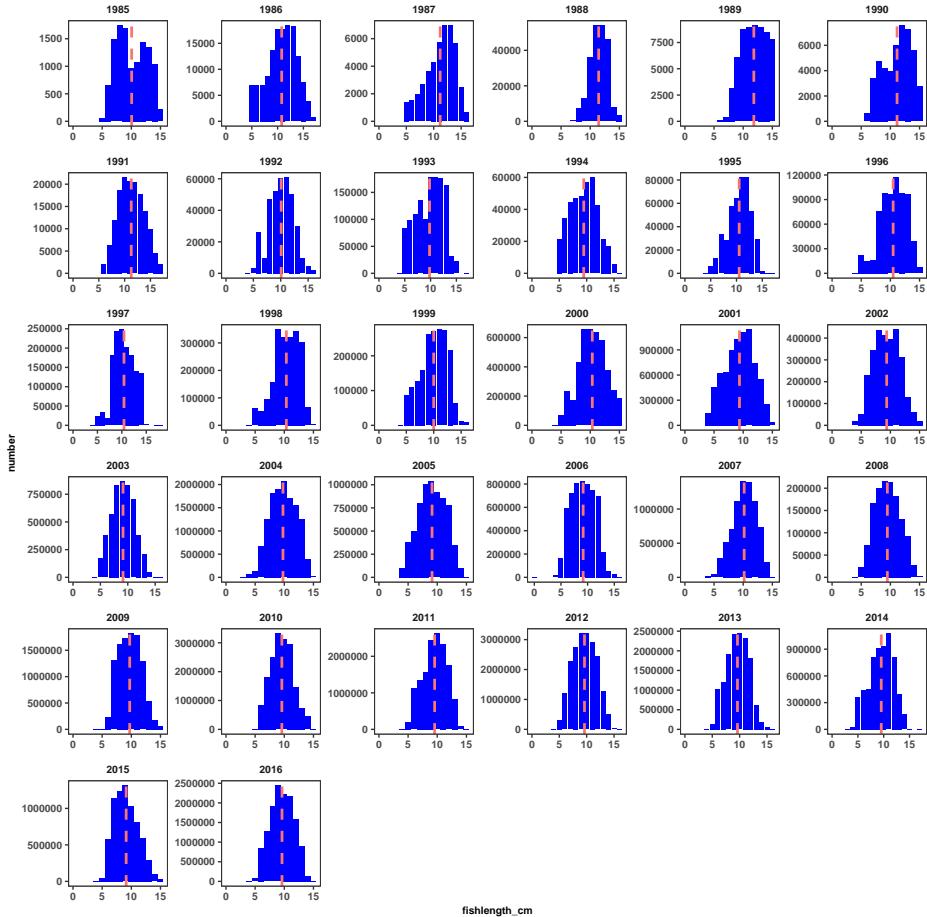


Figure 6 Plot of numbers at length for sprat with weighted mean length represented by dashed red line.

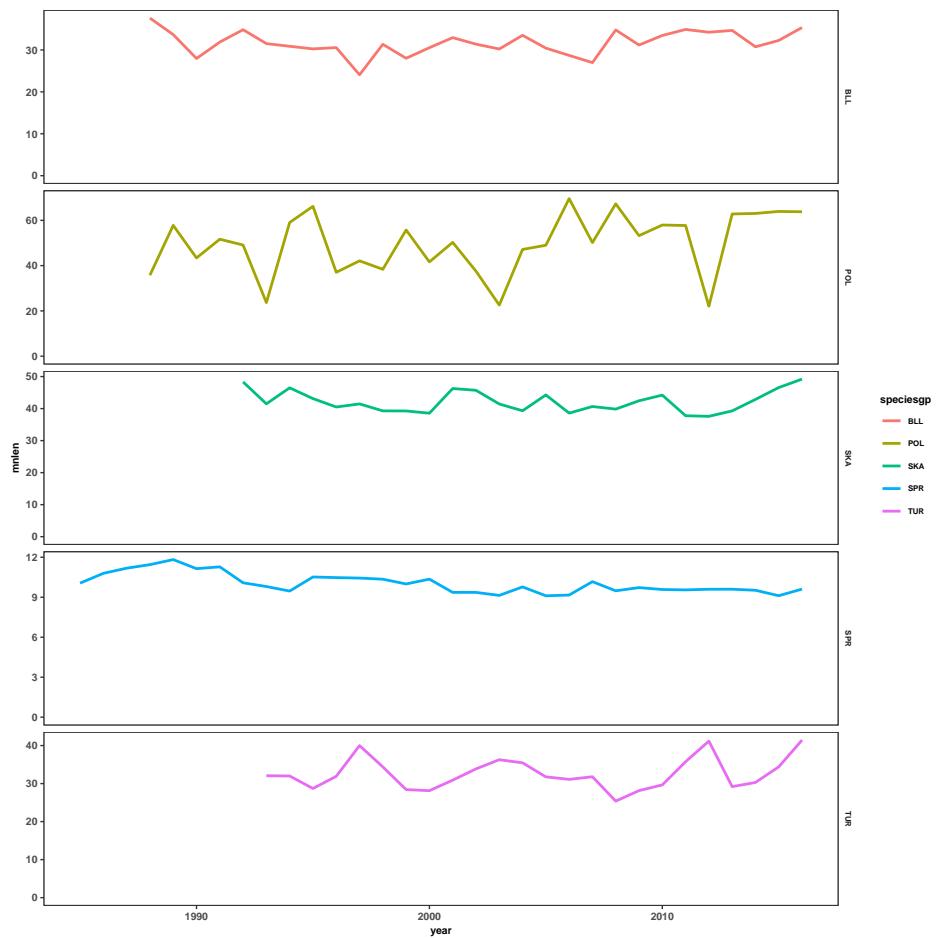


Figure 7 Time series of weighted mean length.

Get all survey data from sql query

Estimate fish weight in kgs

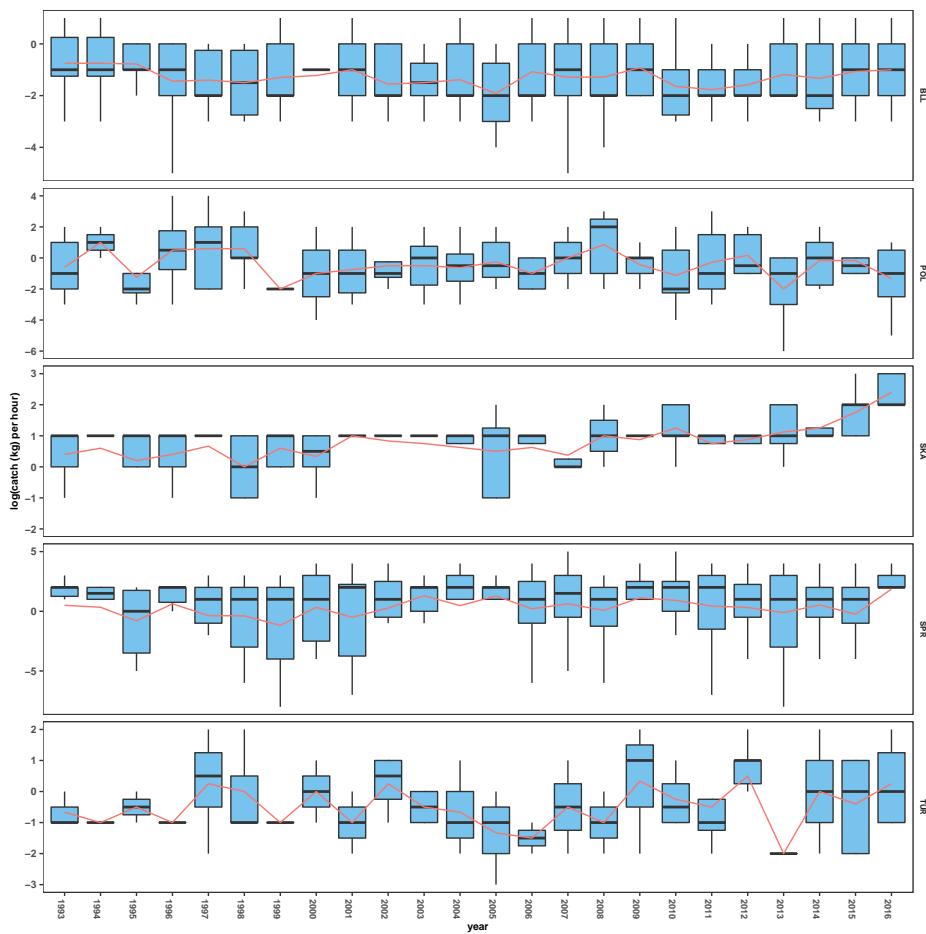


Figure 8 Catch per unit effort in kgs (CPUE) time series with red line depicting mean cpue.

Fish numbers

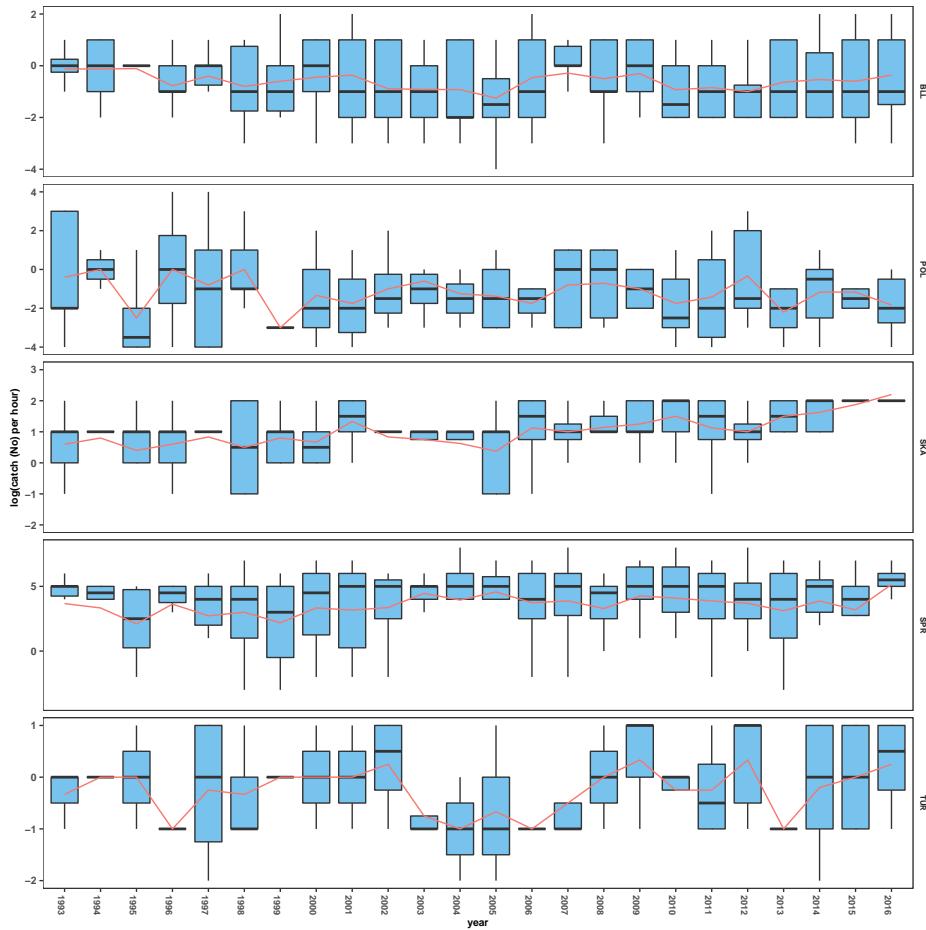


Figure 9 Catch per unit effort in numbers (CPUE) time series with red line depicting mean cpue.

Commercial data

Part of the issues with the main catch data in the stecf database is that the effort does not match up by gear and area with the effort estimations. Here the STECF Annual Economic Report database is used.

Extract the data via sql

Estimate catch in kgs per day

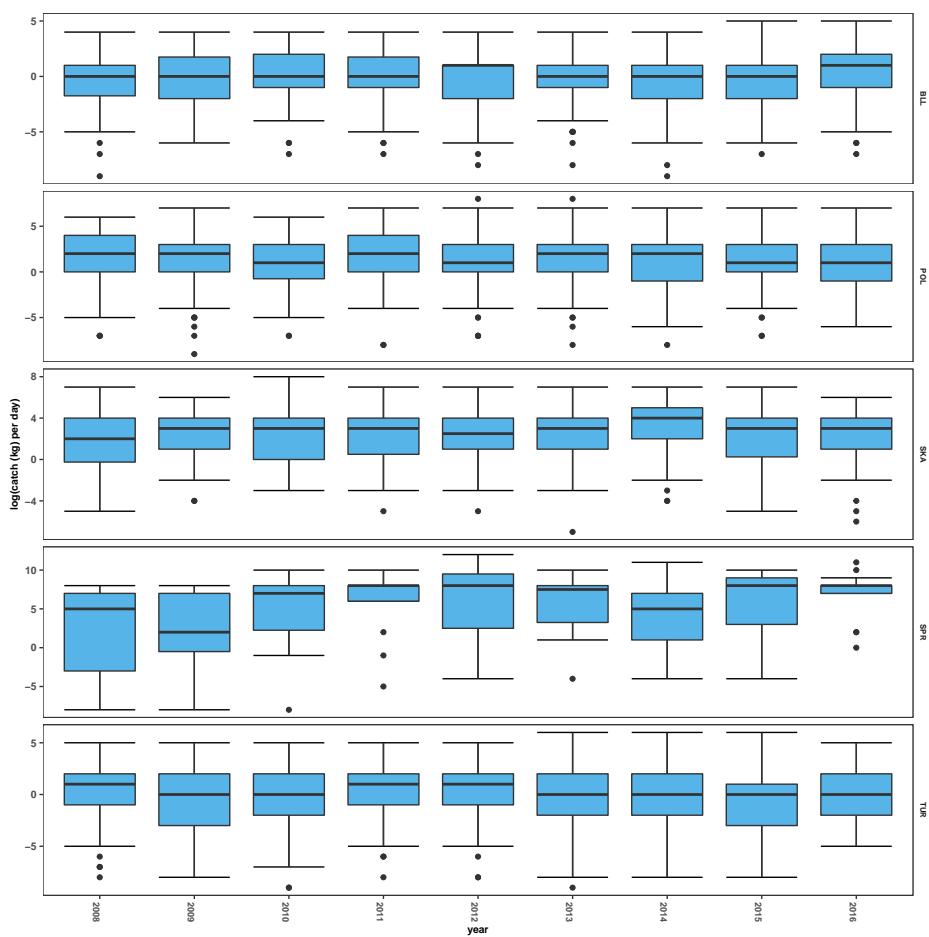


Figure 10 Time series of commercial cpue.

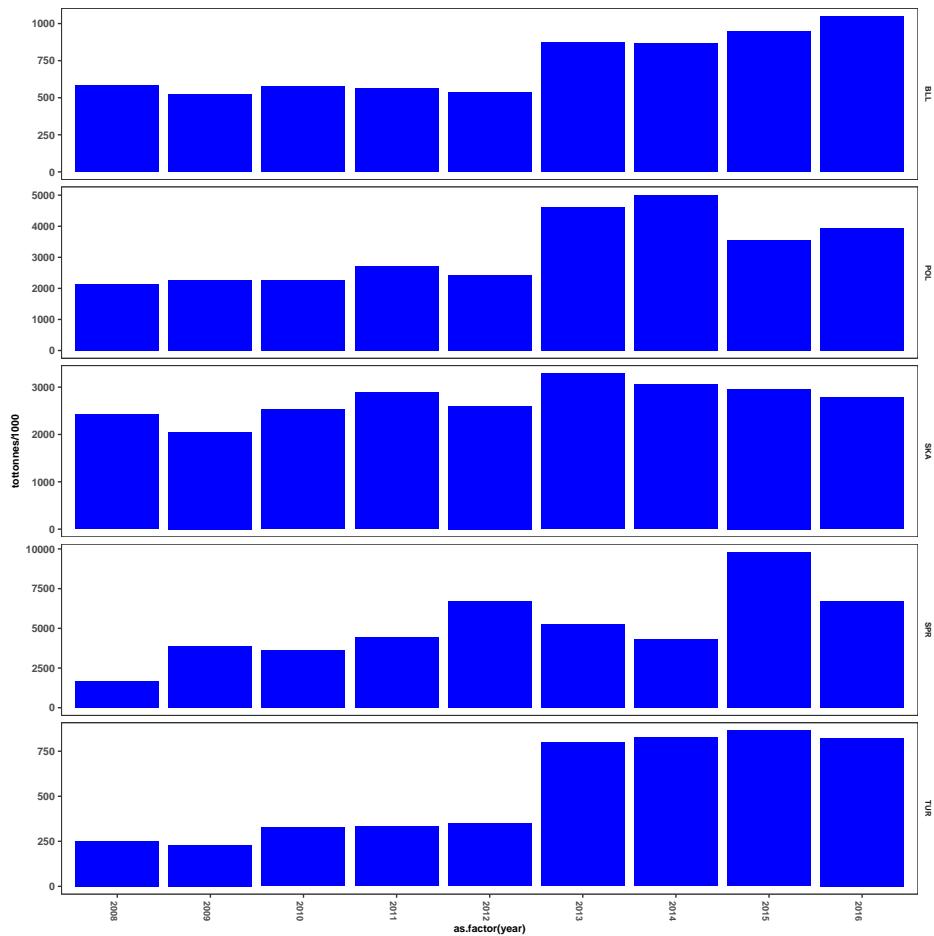


Figure 11 Time series of total catch in tonnes.

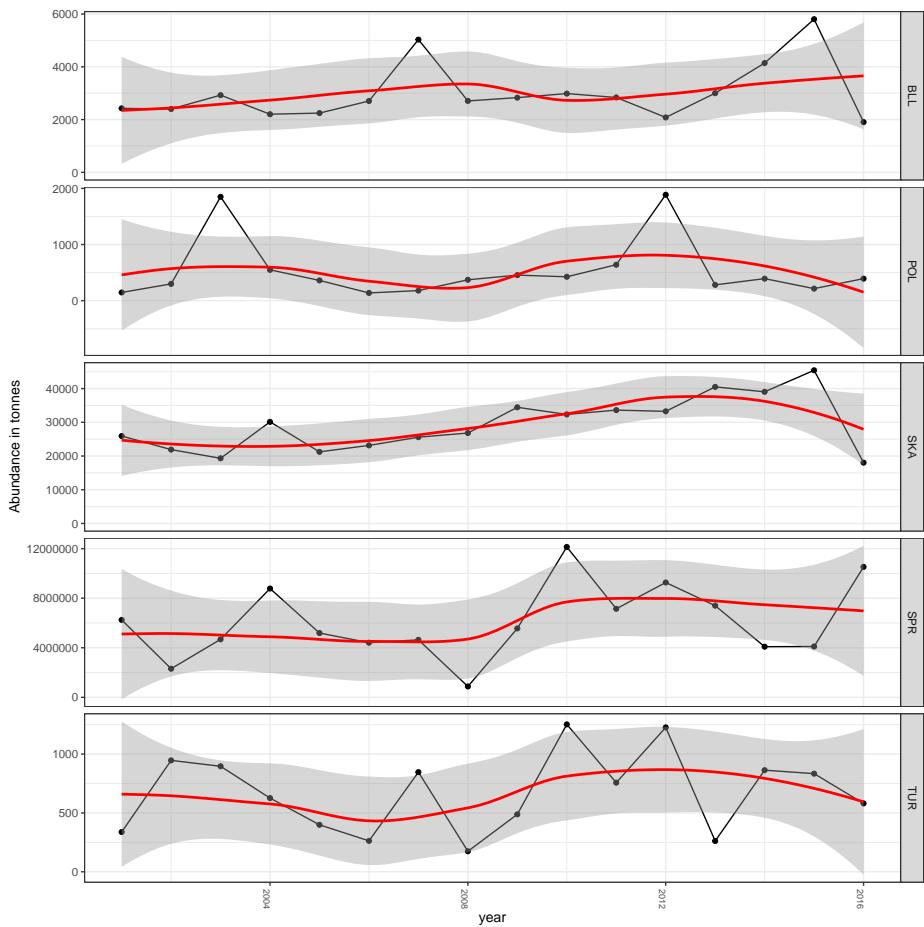
Estimate abundance from survey data

See <http://www.fao.org/docrep/w5449e/w5449e0f.htm> section 13.7

Get survey stations

Calculate total area/total number of stations the survey covered by division and merge with catch tonnes perkm²

Biomass = totalsurveyarea x 1/total number of stations within the survey area x sum of catch



Life history parameters from the literature

Table 1: life history paramters

| id | species | speciesgrp | t.0 | linf | k | lmat | tm | sex | area | source |
|----|-----------------------|------------|-------|-------|--------|------|-----|-----|-----------|--------|
| 1 | Amblyraja radiata | SKA | | 66.00 | 0.2330 | 46.0 | 4.0 | U | North Sea | [12] |
| 7 | Leucoraja circularis | SKA | | 98.80 | 0.1210 | 51.2 | 5.3 | U | North Sea | [12] |
| 8 | Leucoraja fullonica | SKA | | 98.80 | 0.1210 | 51.2 | 5.3 | U | North Sea | [12] |
| 17 | Pollachius pollachius | POL | | 85.60 | 0.1900 | NA | NA | U | North Sea | [18] |
| 18 | Pollachius pollachius | POL | | 85.60 | 0.1860 | 44.8 | 3.7 | U | North Sea | [12] |
| 23 | Psetta maxima | TUR | -0.05 | 64.80 | 0.2600 | 46.0 | NA | F | North Sea | [15] |
| 24 | Psetta maxima | TUR | -0.51 | 49.20 | 0.3700 | NA | NA | M | North Sea | [15] |

| id | species | speciesgp | t.0 | linf | k | lmat | tm | sex | area | source |
|----|----------------------|-----------|--------|--------|--------|-------|------|-----|---------------|-----------------------|
| 25 | Psetta maxima | TUR | | 65.20 | 0.3240 | 49.0 | NA | M | Bay of Biscay | [9], [7] |
| 26 | Psetta maxima | TUR | | 73.60 | 0.2770 | 49.0 | NA | F | Bay of Biscay | [9], [7] |
| 27 | Psetta maxima | TUR | -1.79 | 69.60 | 0.2497 | 49.0 | 4.5 | U | North Sea | [19], [8], [16], [17] |
| 28 | Psetta maxima | TUR | | 57.00 | 0.3200 | 46.0 | 4.5 | U | North Sea | [12] |
| 29 | Psetta maxima | TUR | | 52.50 | 0.3200 | NA | NA | M | North Sea | [17] |
| 30 | Psetta maxima | TUR | -1.79 | 70.00 | 0.1480 | NA | NA | F | North Sea | [17] |
| 31 | Raja batis | SKA | -1.63 | 254.00 | 0.0600 | 130.0 | 11.0 | U | North Sea | [11] |
| 32 | Raja batis | SKA | | 253.70 | 0.0570 | 155.0 | 15.0 | U | North Sea | [12] |
| 33 | Raja brachyura | SKA | | 139.00 | 0.1200 | 100.0 | 9.3 | U | North Sea | [12] |
| 34 | Raja brachyura | SKA | -0.8 | 118.00 | 0.1900 | NA | NA | M | Irish Sea | [13] |
| 35 | Raja brachyura | SKA | -1.52 | 139.00 | 0.1200 | NA | NA | F | Irish Sea | [13] |
| 36 | Raja clavata | SKA | -1.32 | 128.00 | 0.0900 | NA | NA | F | Irish Sea | [13] |
| 37 | Raja clavata | SKA | -0.6 | 85.60 | 0.2100 | NA | NA | M | Irish Sea | [13] |
| 38 | Raja clavata | SKA | | 105.00 | 0.2200 | 65.0 | 5.0 | U | North Sea | [12] |
| 39 | Raja montagui | SKA | -0.37 | 72.80 | 0.1800 | NA | NA | F | Irish Sea | [13] |
| 40 | Raja montagui | SKA | -0.56 | 68.70 | 0.1900 | NA | NA | M | Irish Sea | [13] |
| 41 | Raja montagui | SKA | | 97.80 | 0.1480 | 67.0 | 6.0 | U | North Sea | [12] |
| 42 | Raja naevus | SKA | -0.465 | 92.00 | 0.1100 | 59.0 | 9.0 | U | North Sea | [11] |
| 43 | Raja naevus | SKA | | 91.60 | 0.1090 | 59.0 | 9.0 | U | North Sea | [12] |
| 44 | Raja radiata | SKA | | 66.00 | 0.2300 | 46.0 | 4.0 | U | North Sea | [22] |
| 45 | Raja undulata | SKA | | 112.00 | 0.1000 | 57.5 | 6.3 | U | North Sea | [12] |
| 46 | Scophthalmus rhombus | BLL | | 74.88 | 0.1400 | 37.0 | NA | M | Bay of Biscay | [9], [7] |
| 47 | Scophthalmus rhombus | BLL | | 85.23 | 0.1470 | 37.0 | NA | F | Bay of Biscay | [9], [7] |
| 48 | Scophthalmus rhombus | BLL | | 50.00 | 0.2700 | 37.0 | 4.5 | U | North Sea | [12] |
| 49 | Sprattus sprattus | SPR | * | 16.00 | 0.6500 | 13.0 | NA | U | North Sea | [20] |

Conclusion from the above analysis for 5 of the 7 case studies

- Pollack and turbot have poor coverage in terms of numbers at length from the survey data and hence potentially poor abundance estimates.
- Brill, sprat and skates have long time series of numbers at length. Brill has adequate numbers from 1993 to 2016 as has skate, while sprat contains a longer time series from 1985-to 2016.
- Observations from the commercial time-series show that for all species there are 9 years of data.
- There are enough life history parameters from the literature to condition an operating model for all case-study species apart from sprat.

Software Versions

- R version 3.5.1 (2018-07-02)
- plyr: 1.8.4
- dplyr: 0.7.8
- ggplot2: 3.1.0
- DBI: 1.0.0
- RPostgreSQL: 0.6.2
- reshape 0.8.8
- **Compiled:** Tue Dec 25 20:37:56 2018

Author information

Alex Tidd. emperorfish@gmail.com

Acknowledgements

This vignette and many of the methods documented in it were developed under the MyDas project funded by the Irish exchequer and EMFF 2014-2020. The overall aim of MyDas is to develop and test a range of assessment models and methods to establish Maximum Sustainable Yield (MSY) reference points (or proxy MSY reference points) across the spectrum of data-limited stocks.

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Session Info

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Running under: Ubuntu 16.04.2 LTS

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LAPACK: /usr/lib/lapack/liblapack.so.3.6.0

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[3] LC_TIME=en_GB.UTF-8      LC_COLLATE=en_US.UTF-8
[5] LC_MONETARY=en_GB.UTF-8   LC_MESSAGES=en_US.UTF-8
[7] LC_PAPER=en_GB.UTF-8     LC_NAME=C
[9] LC_ADDRESS=C             LC_TELEPHONE=C
[11] LC_MEASUREMENT=en_GB.UTF-8 LC_IDENTIFICATION=C

attached base packages:
[1] splines    stats      graphics   grDevices utils      datasets   methods
[8] base

other attached packages:
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B Peer Review Papers

B1 Linking the performance of a data-limited empirical catch rule to life-history traits

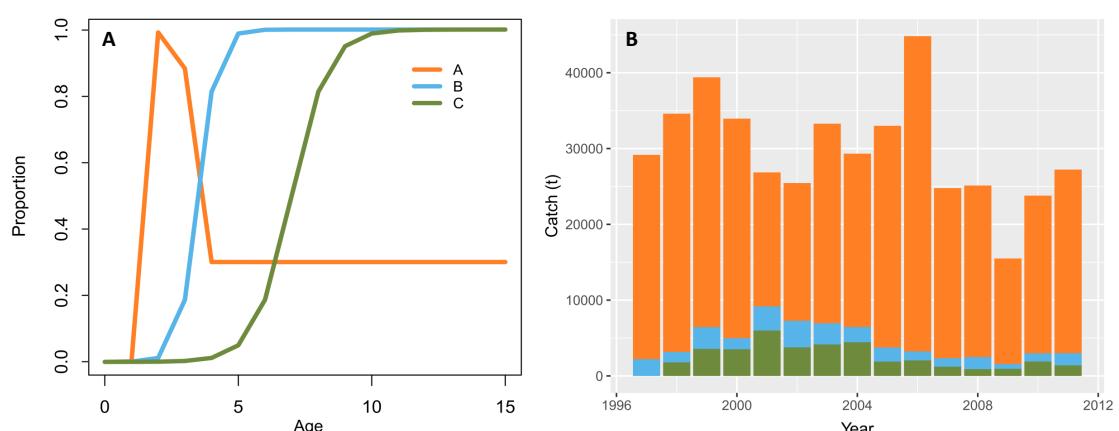


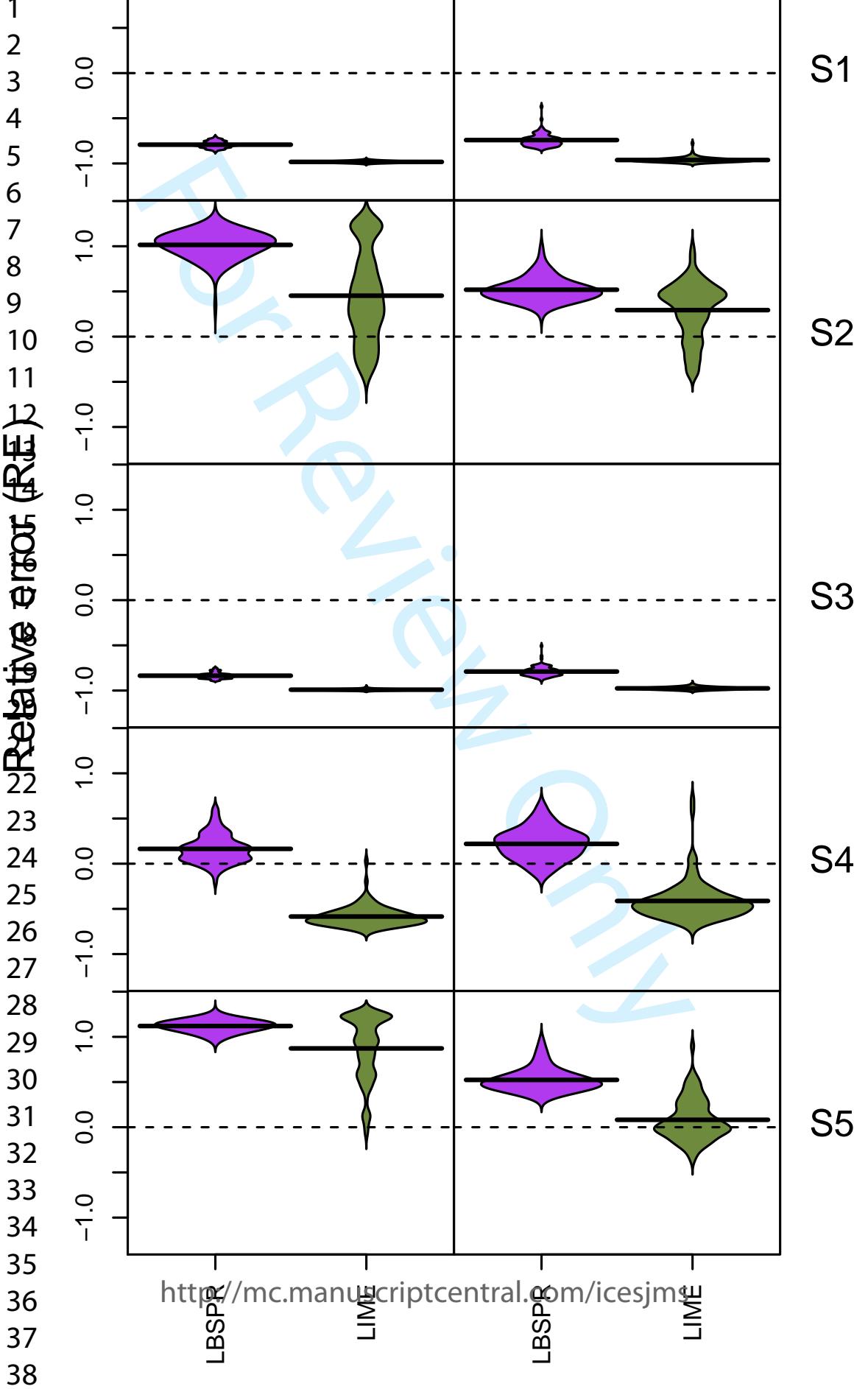
Performance of length-based data-limited methods in a multi-fleet context: application to small tunas, mackerels and bonitos in the Atlantic Ocean

| | |
|-------------------------------|--|
| Journal: | <i>ICES Journal of Marine Science</i> |
| Manuscript ID | ICESJMS-2018-485.R1 |
| Manuscript Types: | Original Article |
| Date Submitted by the Author: | 20-Dec-2018 |
| Complete List of Authors: | Pons, Maite; University of Washington, School of Aquatic and Fishery Sciences Kell, Laurence; Imperial College London, Centre for Environmental Policy Rudd, Merrill; University of Washington, School of Aquatic and Fishery Sciences; National Marine Fisheries Service, Northwest Fisheries Science Center Cope, Jason; Northwest Fisheries Science Center, National Oceanic and Atmospheric Administration Fisheries Lucena Frédou , Flávia; Universidade Federal Rural de Pernambuco, Departamento de Pesca e Aquicultura |
| Keyword: | length based assessments, spawning potential ratio, Scombrids, management |
| | |

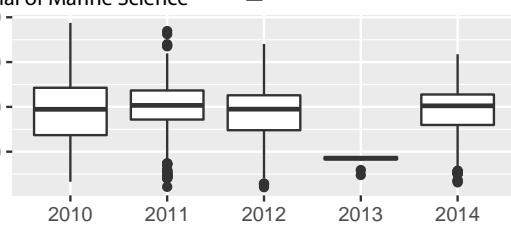
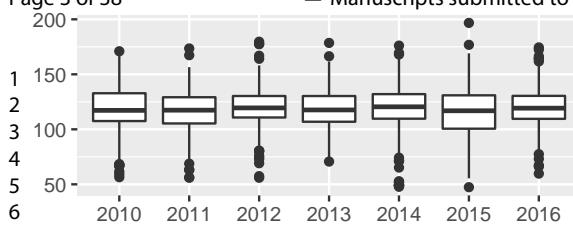
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Manuscripts

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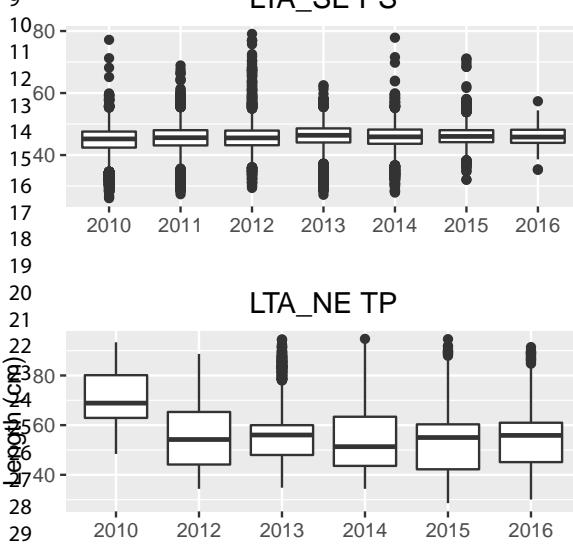




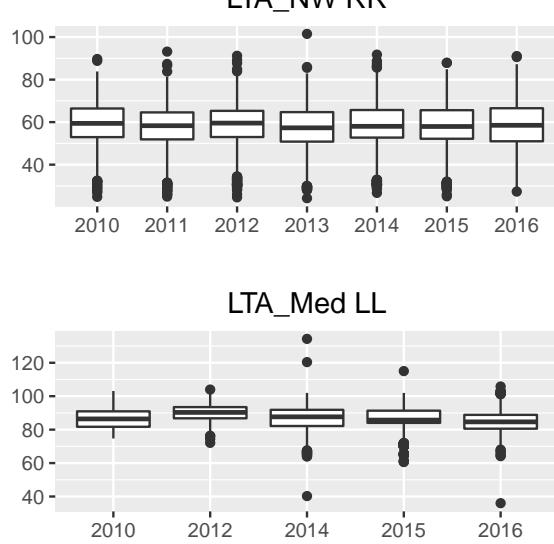
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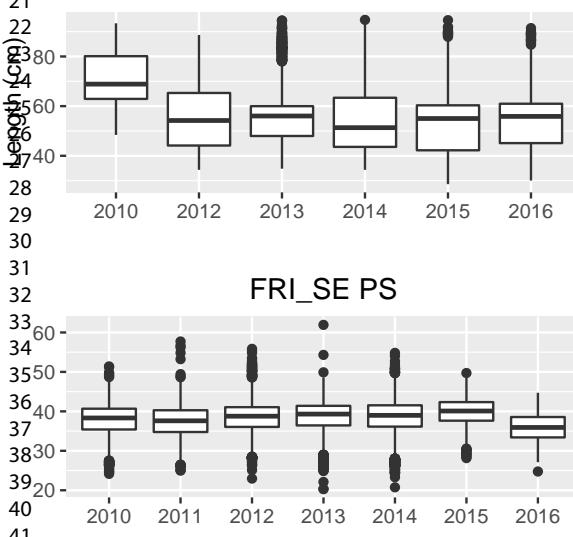
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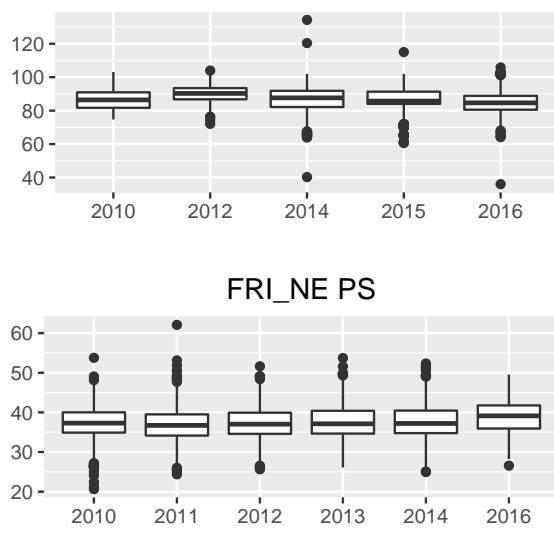
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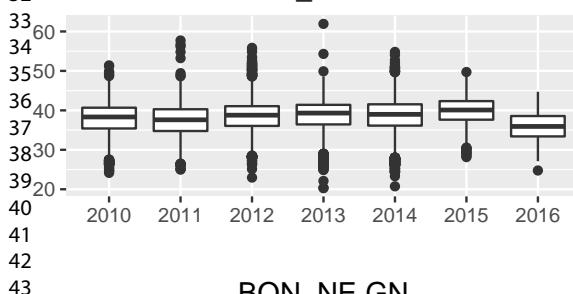
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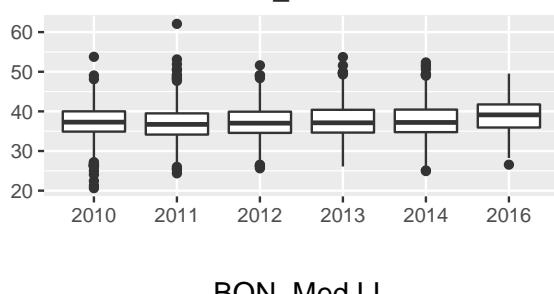
LTA_Med LL



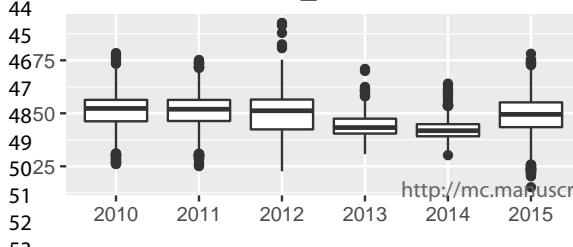
FRI_SE PS



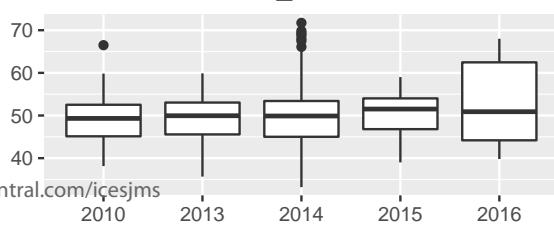
FRI_NE PS

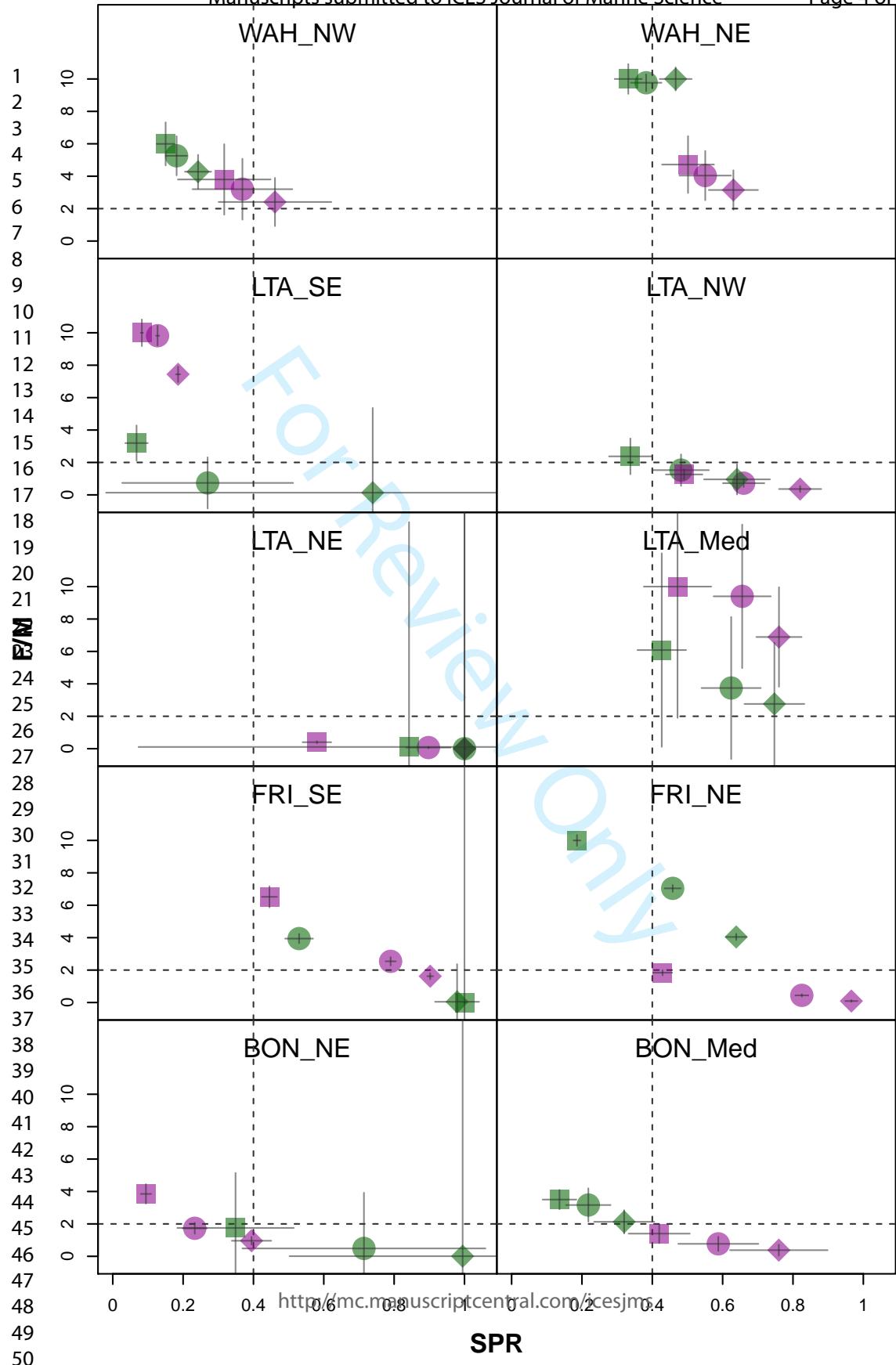


BON_NE GN



BON_Med LL





Supplementary information: ICES Journal of Marine Science, 76**Performance of length-based data-limited methods in a multi-fleet context: application to small tunas, mackerels, and bonitos in the Atlantic Ocean**

Maite Pons, Laurence Kell, Merrill B. Rudd, Jason M. Cope, and Flávia Lucena Frédou

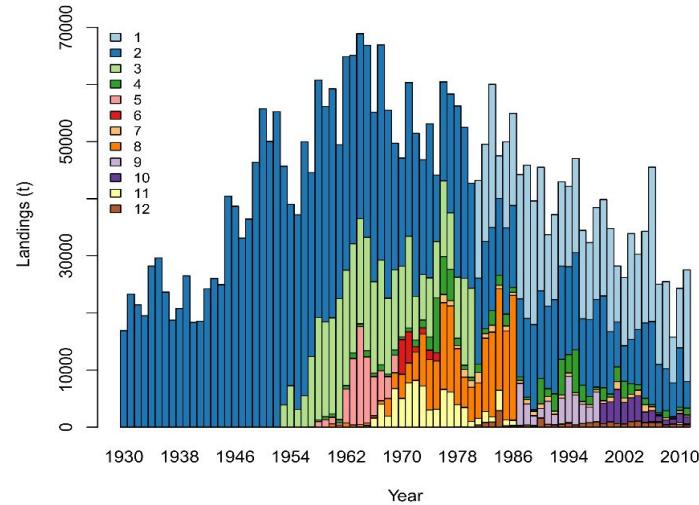


Figure S1. Catch of North Atlantic albacore by fleet from 1930 to 2011. Fleets 1–12 were defined in the 2013 stock assessment performed by ICCAT (see the ICCAT, 2014 report for more details).

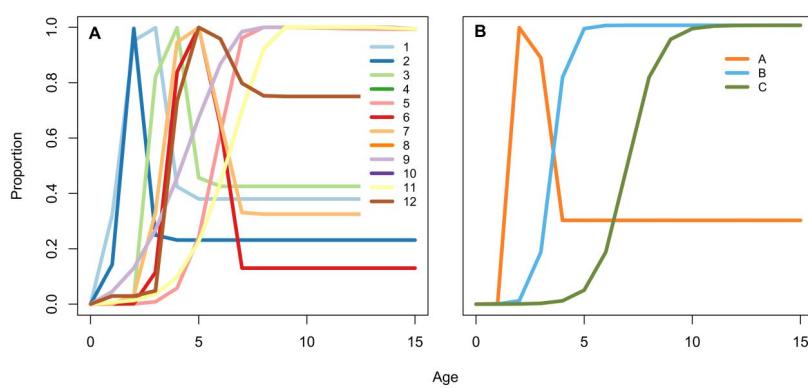


Figure S2. Selectivity curves. (a) The 12 selectivity curves used in the 2013 North Atlantic albacore assessment (fleets 10 and 11, 8 and 9, and 4 and 12 have the same selectivity pattern). (b) Combined selectivity curves to test under the different scenarios for the fleets that operated in the last 15 years (fleets A–C). Label the two curves as (a) and (b), not A and B.

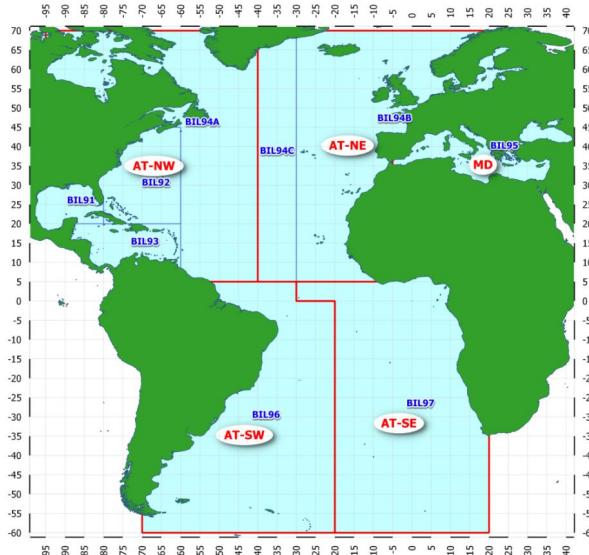


Figure S3. ICCAT geographical definitions (Version: 2016.02) for small scombrids. Taken from: http://iccat.es/Data/ICCAT_maps.pdf.

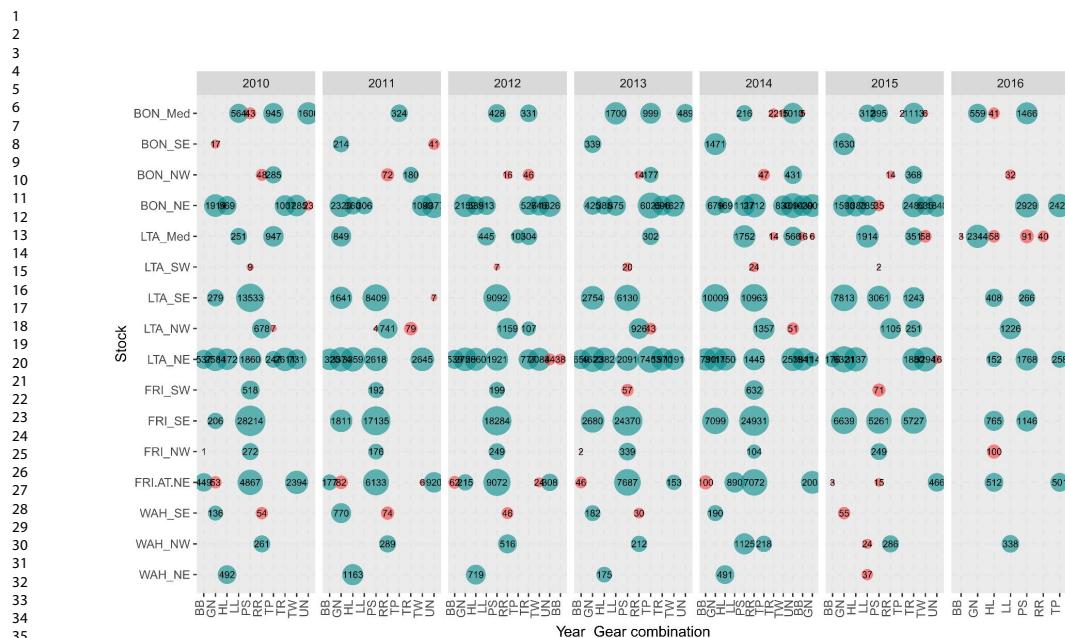


Figure S4. Number of fish measured by stock, year, and gear during 2010–2016. The size of the bubbles is proportional to the logarithm of the number of fish measured, and the number is inside the bubble. Bubbles where the number of fish measured is less than 100 are in red. Gears: gillnet (GN); handline (HL); longline (LL); purse-seine (PS); trap (TP); trolling (TR); sport (SP); baitboat (BB); rod and reel (RR); haulseine (HS); trammelnet (TN); unclassified (UN).

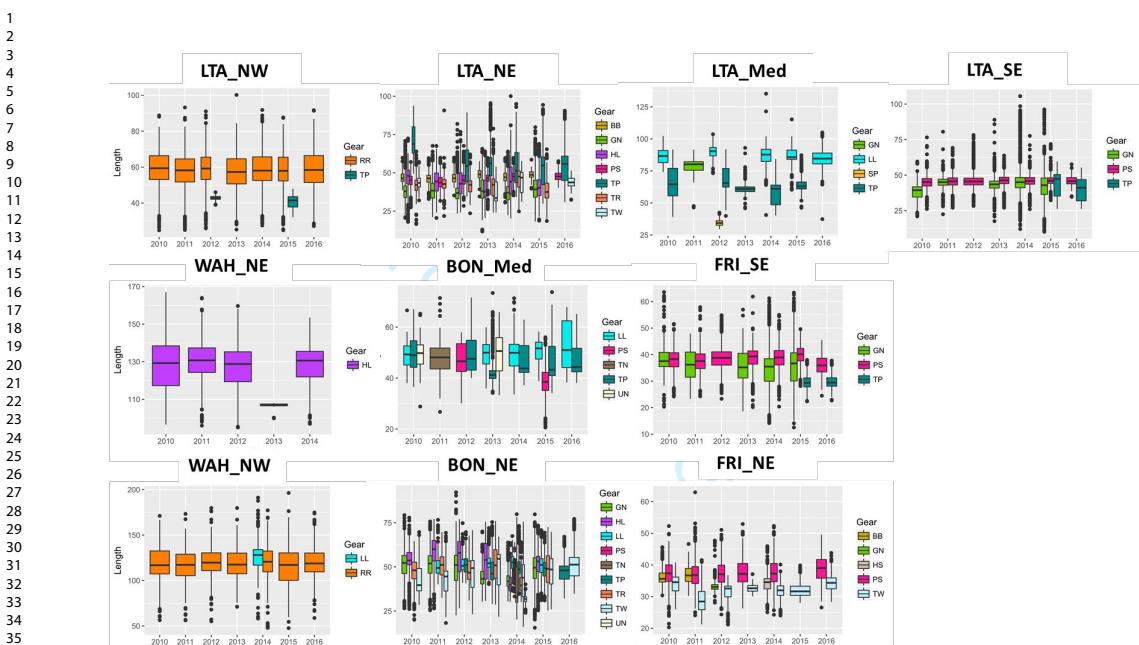


Figure S5. Length composition data for the main stocks of small scombrids by gear in the Atlantic Ocean available in Task2sz database of ICCAT. LTA, little tunny; WAH, wahoo; BON, bonito; FRI, frigate tuna. BB, bait boat; GN, gillnet; HL, handline; HS, haulseine; LL, longline; PS, purse-seine; RR, rod and reel; SP, sport; TN, trammelnet; TP, trap; TR, trolling; TW, trawl; UN, unknown. NE, Northeast; SE, Southeast; NW, Northwest, and Med, Mediterranean Sea.

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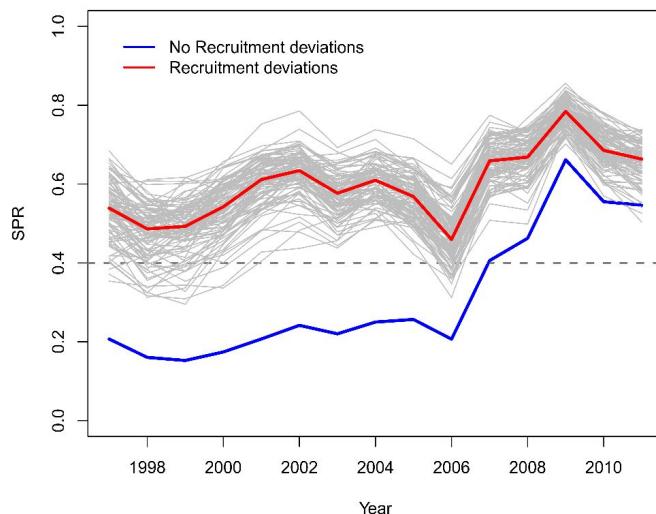


Figure S6. Time-series of the true SPR for the two OM (with and without recruitment deviations). Grey lines are the SPR time-series for each of the 100 runs with random recruitment deviations.

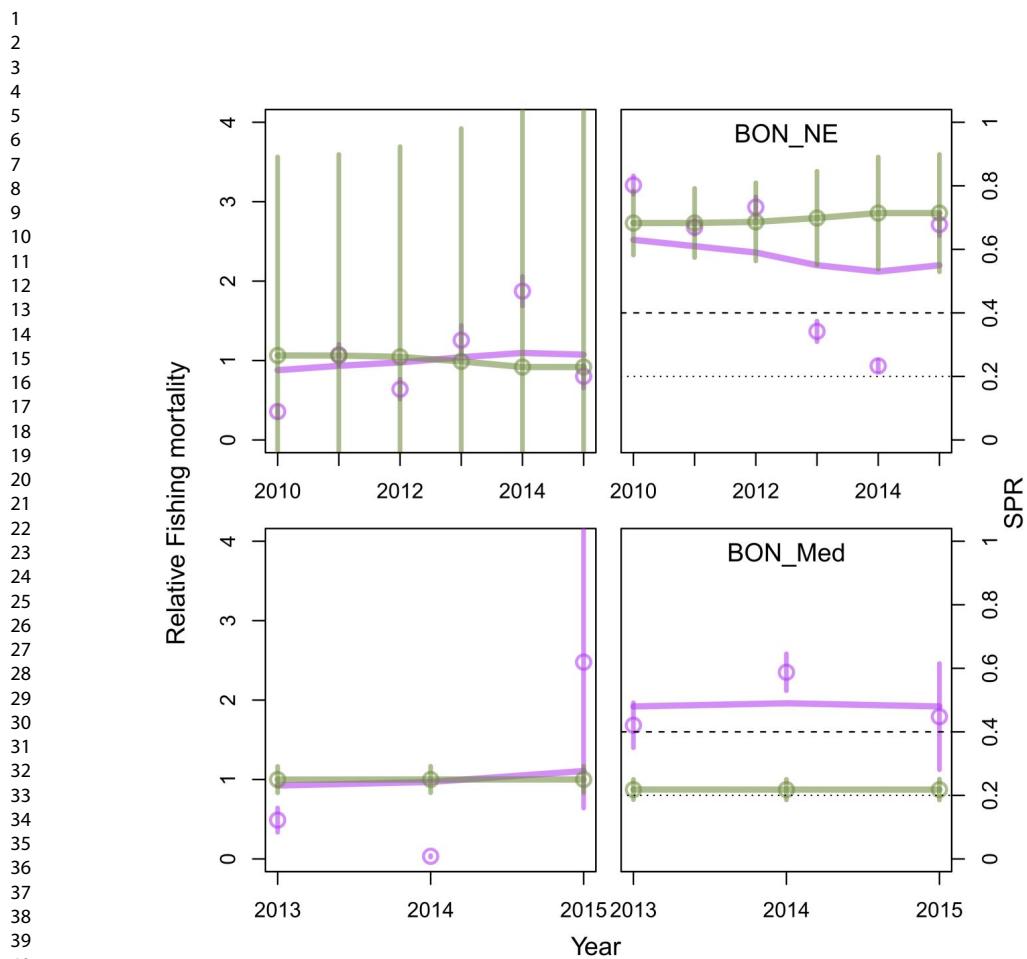


Figure S7. Bonito (BON) estimates of fishing mortality and SPR in the Northeast (NE) Atlantic and in the Mediterranean Sea (Med). Estimates from LIME are in green, and estimates from LBSPR are in purple. Bars represent standard deviation of the estimates.

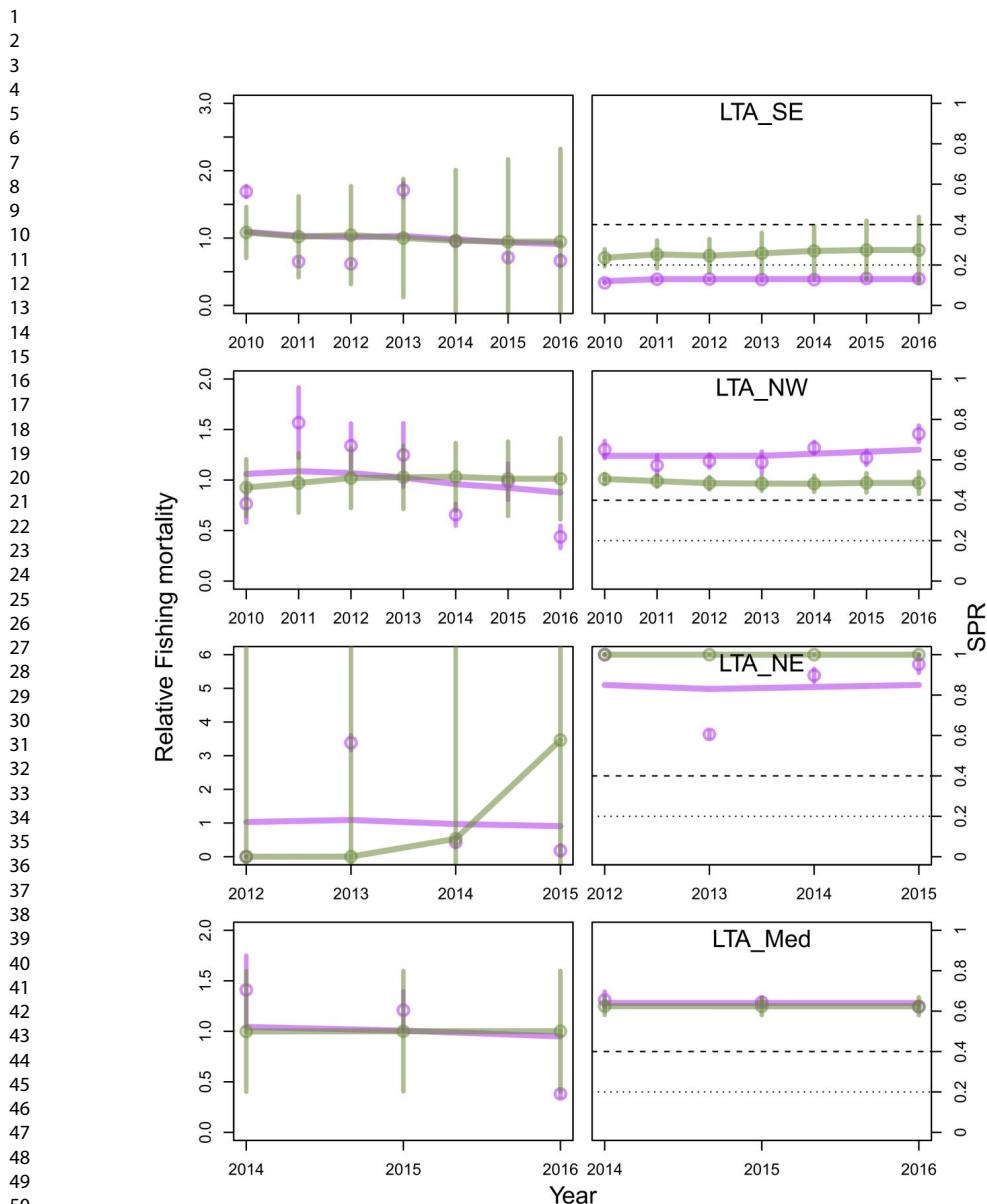


Figure S8. Little tunny (LTA) estimates of fishing mortality and SPR in the Southeast (SE), Northwest (NW), and Northeast (NE) Atlantic Ocean and in the Mediterranean Sea (Med). Estimates from LIME are in green, and estimates from LBSPR are in purple. Bars represent standard deviation of the estimates.

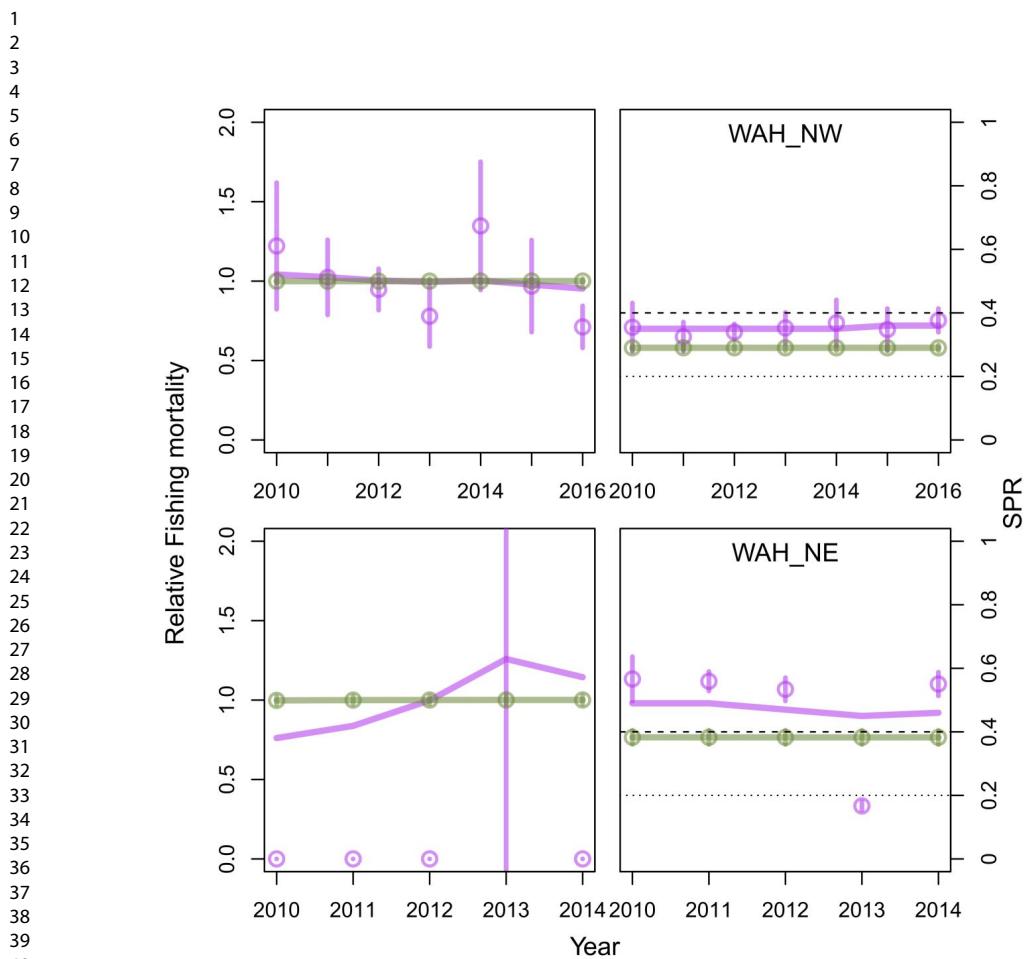


Figure S9. Wahoo (WAH) estimates of fishing mortality and SPR in the Northwest (NW) and Northeast (NE) Atlantic Ocean and in the Mediterranean Sea (Med). Estimates from LIME are in green, and estimates from LBSPR are in purple. Bars represent standard deviation of the estimates.

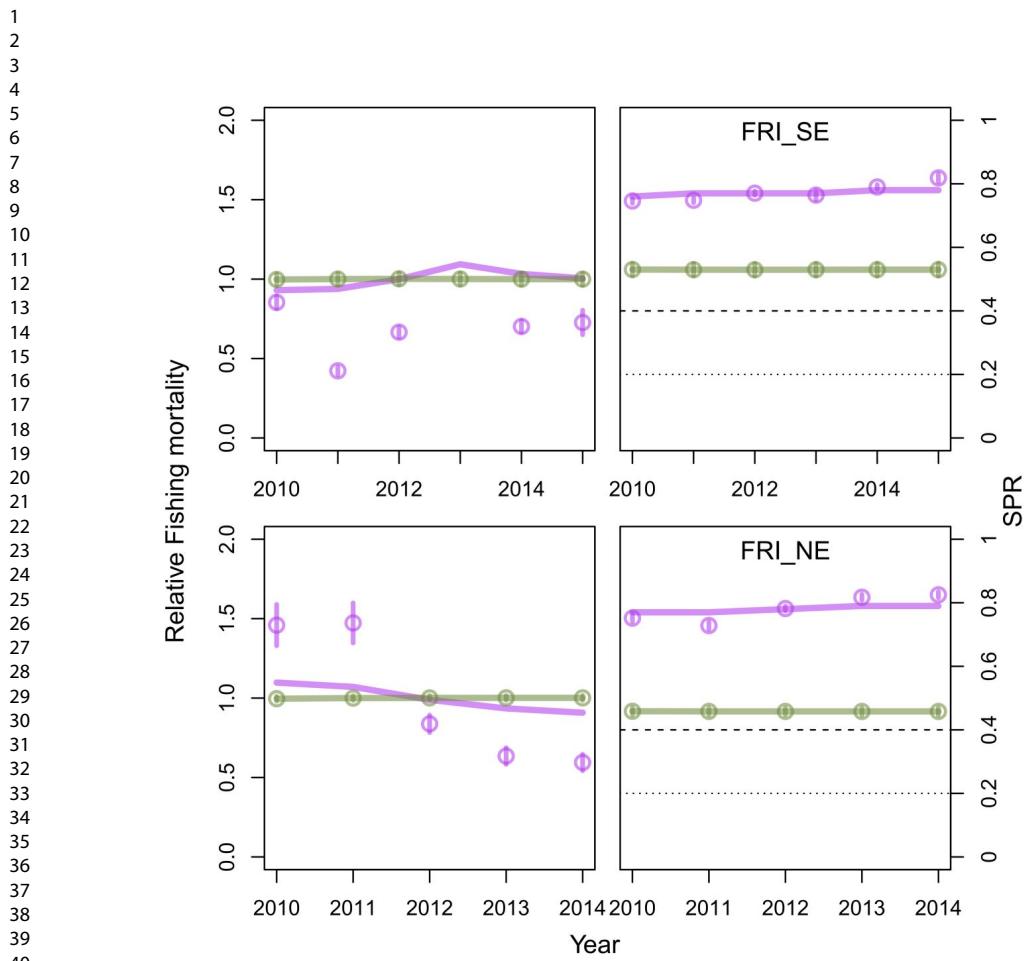


Figure S10. Frigate tuna (FRI) estimates of fishing mortality and SPR in the Southeast (SE) and Northeast (NE) Atlantic Ocean. Estimates from LIME are in green, and estimates from LBSPR are in purple. Bars represent standard deviation of the estimates.

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5**Table S1.** North Atlantic albacore age-length conversion matrix extracted from SS. Columns
are the ages and rows are length bins.6
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| Length/Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | |
|------------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|
| 150 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 6.05E-14 | 4.10E-09 | 8.20E-09 | 2.40E-09 | 3.20E-09 | 1.09E-09 | 2.32E-09 | 7.47E-09 | 1.27E-09 | 2.16E-09 | 3.07E-09 | 5.26E-09 | |
| 148 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 2.79E-13 | 1.14E-09 | 1.59E-09 | 3.56E-09 | 2.75E-05 | 1.11E-04 | 2.97E-04 | 6.02E-04 | 1.01E-03 | 1.50E-03 | 2.01E-03 | 3.19E-03 | |
| 146 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 1.47E-12 | 4.07E-09 | 4.42E-09 | 8.32E-09 | 5.70E-05 | 2.11E-04 | 5.28E-04 | 1.02E-03 | 1.65E-03 | 2.37E-03 | 3.11E-03 | 4.75E-03 | |
| 144 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 7.33E-12 | 1.39E-08 | 1.17E-08 | 1.87E-09 | 1.14E-05 | 3.87E-04 | 9.08E-04 | 1.67E-03 | 2.61E-03 | 3.63E-03 | 4.67E-03 | 6.89E-03 | |
| 142 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 1.11E-16 | 3.46E-11 | 4.50E-08 | 3.00E-08 | 4.05E-05 | 2.20E-04 | 6.87E-04 | 1.51E-03 | 2.66E-03 | 4.00E-03 | 5.42E-03 | 6.81E-03 | 9.70E-03 | |
| 140 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 4.44E-16 | 1.55E-10 | 1.40E-07 | 7.34E-06 | 8.45E-05 | 4.10E-04 | 1.18E-03 | 2.45E-03 | 4.10E-03 | 5.95E-03 | 7.85E-03 | 9.65E-03 | 1.33E-02 | |
| 138 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 4.00E-15 | 6.59E-10 | 4.13E-07 | 1.72E-05 | 1.70E-04 | 7.37E-04 | 1.96E-03 | 3.83E-03 | 6.13E-03 | 8.61E-02 | 1.10E-02 | 1.33E-02 | 1.77E-02 | |
| 136 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 2.90E-14 | 3.89E-09 | 3.08E-06 | 8.30E-06 | 6.12E-05 | 2.15E-05 | 4.88E-03 | 8.51E-03 | 1.25E-02 | 1.64E-02 | 2.00E-02 | 2.32E-02 | 2.88E-02 | |
| 134 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 1.97E-13 | 1.81E-08 | 3.10E-06 | 8.03E-06 | 5.70E-05 | 1.79E-05 | 4.19E-02 | 7.00E-02 | 1.21E-02 | 1.51E-02 | 1.78E-02 | 2.29E-02 | 2.40E-02 | |
| 132 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 1.24E-12 | 3.65E-09 | 8.44E-06 | 1.79E-05 | 4.01E-05 | 1.04E-05 | 2.44E-02 | 4.84E-02 | 1.22E-02 | 1.71E-02 | 2.10E-02 | 2.50E-02 | 2.90E-02 | |
| 130 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 7.49E-12 | 1.25E-07 | 2.01E-05 | 3.48E-04 | 1.90E-03 | 5.43E-03 | 1.07E-02 | 1.67E-02 | 2.36E-02 | 2.78E-02 | 3.21E-02 | 3.59E-02 | 4.20E-02 | |
| 128 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 4.18E-11 | 4.06E-07 | 4.72E-05 | 6.67E-05 | 3.16E-03 | 8.20E-03 | 1.50E-02 | 2.22E-02 | 2.89E-02 | 3.46E-02 | 3.92E-02 | 4.29E-02 | 4.87E-02 | |
| 126 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 2.19E-10 | 1.25E-06 | 1.06E-04 | 1.22E-03 | 5.07E-03 | 1.19E-02 | 2.04E-02 | 2.87E-02 | 3.60E-02 | 4.18E-02 | 4.63E-02 | 4.97E-02 | 5.48E-02 | |
| 124 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 1.07E-09 | 3.64E-06 | 2.27E-04 | 2.16E-03 | 7.82E-03 | 1.68E-02 | 2.68E-02 | 3.60E-02 | 4.34E-02 | 4.90E-02 | 5.31E-02 | 5.60E-02 | 6.00E-02 | |
| 122 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 3.33E-16 | 4.90E-09 | 1.01E-05 | 4.65E-05 | 3.65E-03 | 1.16E-02 | 2.28E-02 | 3.41E-02 | 4.36E-02 | 5.07E-02 | 5.57E-02 | 5.90E-02 | 6.13E-02 | 6.39E-02 |
| 120 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 2.78E-15 | 2.10E-08 | 2.64E-05 | 9.10E-04 | 5.92E-03 | 1.66E-02 | 2.99E-02 | 4.20E-02 | 5.12E-02 | 5.75E-02 | 6.14E-02 | 6.38E-02 | 6.51E-02 | 6.62E-02 |
| 118 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 2.76E-14 | 8.45E-08 | 6.55E-04 | 1.70E-03 | 3.20E-02 | 7.39E-02 | 5.00E-02 | 5.83E-02 | 6.32E-02 | 6.58E-02 | 6.70E-02 | 6.73E-02 | 6.66E-02 | |
| 116 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 2.52E-16 | 3.18E-07 | 1.54E-04 | 3.07E-04 | 3.05E-02 | 4.63E-02 | 5.76E-02 | 6.42E-02 | 6.74E-02 | 6.84E-02 | 6.83E-02 | 6.76E-02 | 6.52E-02 | |
| 114 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 2.18E-15 | 1.81E-06 | 5.31E-03 | 1.02E-02 | 3.00E-02 | 5.47E-02 | 6.42E-02 | 6.75E-02 | 6.97E-02 | 7.03E-02 | 6.95E-02 | 6.83E-02 | 6.73E-02 | |
| 112 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 1.62E-11 | 3.68E-05 | 7.26E-04 | 8.39E-03 | 5.41E-02 | 1.40E-02 | 2.71E-02 | 4.05E-02 | 5.09E-02 | 6.08E-02 | 6.42E-02 | 6.64E-02 | 6.75E-02 | |
| 110 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 1.13E-10 | 1.14E-05 | 1.45E-03 | 1.30E-02 | 3.60E-02 | 5.70E-02 | 6.86E-02 | 7.21E-02 | 7.10E-02 | 6.78E-02 | 6.42E-02 | 5.66E-02 | 5.18E-02 | |
| 108 | 0.00E+00 | 0.00E+00 | 0.00E+00 | 7.30E-10 | 3.30E-05 | 2.76E-03 | 1.93E-02 | 4.05E-02 | 6.56E-02 | 6.51E-02 | 7.28E-02 | 7.27E-02 | 6.89E-02 | 6.39E-02 | 5.50E-02 | 5.14E-02 | 4.54E-02 |
| 106 | 0.00E+00 | 0.00E+00 | 4.30E-09 | 8.93E-05 | 4.96E-03 | 2.73E-02 | 5.56E-02 | 7.16E-02 | 7.47E-02 | 7.09E-02 | 6.47E-02 | 5.84E-02 | 5.29E-02 | 4.84E-02 | 4.47E-02 | 3.86E-02 | |
| 104 | 0.00E+00 | 0.00E+00 | 2.32E-08 | 2.26E-04 | 8.46E-03 | 7.03E-02 | 6.49E-02 | 7.58E-02 | 7.40E-02 | 6.69E-02 | 5.89E-02 | 5.18E-02 | 4.60E-02 | 4.14E-02 | 3.78E-02 | 3.20E-02 | |
| 102 | 0.00E+00 | 1.15E-07 | 5.37E-07 | 1.37E-02 | 4.77E-02 | 7.27E-02 | 7.74E-02 | 7.07E-02 | 6.10E-02 | 5.19E-02 | 4.45E-02 | 3.87E-02 | 3.43E-02 | 3.10E-02 | 2.58E-02 | | |
| 100 | 0.00E+00 | 1.11E-16 | 5.21E-07 | 1.19E-03 | 2.09E-02 | 5.89E-02 | 7.82E-02 | 6.71E-02 | 5.53E-02 | 4.58E-02 | 4.32E-02 | 3.71E-02 | 3.17E-02 | 2.77E-02 | 2.47E-02 | 2.02E-02 | |
| 98 | 0.00E+00 | 9.99E-07 | 2.16E-06 | 2.48E-03 | 3.04E-02 | 6.93E-02 | 8.08E-02 | 7.20E-02 | 5.82E-02 | 4.60E-02 | 3.67E-02 | 3.00E-02 | 2.52E-02 | 2.17E-02 | 1.91E-02 | 1.54E-02 | |
| 96 | 0.00E+00 | 1.92E-14 | 8.27E-06 | 4.84E-03 | 4.18E-02 | 7.79E-02 | 8.00E-02 | 6.56E-02 | 5.00E-02 | 3.80E-02 | 3.35E-02 | 1.94E-02 | 1.65E-02 | 1.44E-02 | 1.14E-02 | | |
| 94 | 0.00E+00 | 2.95E-06 | 2.80E-05 | 8.82E-05 | 8.82E-02 | 8.00E-02 | 7.95E-02 | 7.56E-02 | 4.16E-02 | 3.03E-02 | 2.88E-02 | 1.78E-02 | 1.45E-02 | 1.22E-02 | 1.04E-02 | | |
| 92 | 0.00E+00 | 4.00E-06 | 4.00E-05 | 1.39E-05 | 1.39E-02 | 8.00E-02 | 8.00E-02 | 7.95E-02 | 4.16E-02 | 3.03E-02 | 2.88E-02 | 1.78E-02 | 1.45E-02 | 1.22E-02 | 1.04E-02 | | |
| 90 | 0.00E+00 | 4.73E-11 | 2.69E-04 | 2.41E-02 | 7.27E-02 | 8.00E-02 | 8.00E-02 | 7.95E-02 | 4.16E-02 | 3.03E-02 | 2.88E-02 | 1.78E-02 | 1.45E-02 | 1.22E-02 | 1.04E-02 | | |
| 88 | 0.00E+00 | 4.91E-10 | 7.23E-04 | 3.61E-02 | 8.74E-02 | 7.84E-02 | 5.11E-02 | 1.93E-02 | 1.31E-02 | 1.27E-02 | 8.82E-03 | 6.51E-03 | 5.06E-03 | 4.11E-03 | 3.47E-03 | 2.59E-03 | |
| 86 | 0.00E+00 | 4.46E-09 | 1.78E-03 | 5.05E-02 | 9.20E-02 | 7.00E-02 | 4.12E-02 | 2.35E-02 | 1.39E-02 | 1.28E-02 | 6.85E-03 | 6.02E-03 | 4.37E-03 | 3.86E-03 | 2.70E-03 | 2.26E-03 | 1.66E-03 |
| 84 | 0.00E+00 | 3.55E-08 | 4.01E-03 | 6.63E-02 | 9.17E-02 | 5.96E-02 | 3.19E-02 | 1.71E-02 | 9.71E-03 | 5.98E-03 | 3.99E-03 | 2.85E-03 | 2.16E-03 | 1.72E-03 | 1.43E-03 | 1.04E-03 | |
| 82 | 0.00E+00 | 2.47E-07 | 8.28E-03 | 8.15E-02 | 8.66E-02 | 4.85E-02 | 2.37E-02 | 1.19E-02 | 6.53E-03 | 3.92E-03 | 2.56E-03 | 1.80E-03 | 1.35E-03 | 1.07E-03 | 8.81E-04 | 6.35E-04 | |
| 80 | 0.00E+00 | 1.51E-06 | 1.57E-02 | 9.38E-02 | 7.75E-02 | 3.77E-02 | 1.69E-02 | 8.06E-03 | 4.24E-03 | 2.40E-03 | 1.52E-03 | 9.55E-04 | 6.55E-04 | 4.81E-04 | 3.75E-04 | 3.06E-04 | 2.16E-04 |
| 78 | 0.00E+00 | 8.08E-07 | 2.72E-02 | 1.01E-01 | 6.58E-02 | 2.08E-02 | 1.16E-02 | 5.23E-03 | 2.66E-03 | 1.52E-03 | 9.55E-04 | 6.55E-04 | 4.81E-04 | 3.75E-04 | 3.06E-04 | 2.17E-04 | |
| 76 | 0.00E+00 | 3.79E-05 | 4.31E-02 | 1.02E-01 | 5.29E-02 | 1.98E-02 | 7.60E-03 | 3.27E-03 | 1.61E-03 | 8.98E-04 | 5.57E-04 | 3.77E-04 | 2.75E-04 | 2.13E-04 | 1.73E-04 | 1.21E-04 | |
| 74 | 0.00E+00 | 1.55E-04 | 6.23E-02 | 9.62E-02 | 4.03E-02 | 1.34E-02 | 4.79E-03 | 1.97E-03 | 9.43E-04 | 5.14E-04 | 3.14E-04 | 2.11E-04 | 1.52E-04 | 1.17E-04 | 9.47E-05 | 6.60E-05 | |
| 72 | 0.00E+00 | 3.52E-04 | 8.48E-02 | 8.81E-02 | 8.00E-02 | 2.00E-02 | 1.20E-02 | 6.00E-03 | 3.00E-03 | 1.70E-03 | 1.10E-04 | 8.10E-04 | 5.30E-04 | 3.50E-04 | 2.50E-04 | 1.75E-04 | |
| 70 | 0.00E+00 | 1.76E-03 | 1.02E-01 | 7.04E-02 | 1.99E-02 | 5.35E-03 | 1.68E-03 | 6.40E-04 | 3.40E-04 | 1.55E-04 | 9.08E-05 | 5.99E-05 | 4.27E-05 | 2.31E-05 | 1.80E-05 | | |
| 68 | 0.00E+00 | 4.85E-03 | 1.15E-01 | 5.46E-02 | 1.29E-02 | 3.15E-03 | 9.37E-04 | 3.44E-04 | 1.53E-04 | 7.92E-05 | 4.67E-05 | 3.05E-05 | 2.17E-05 | 1.64E-05 | 1.21E-05 | 9.01E-06 | |
| 66 | 0.00E+00 | 1.17E-02 | 1.18E-01 | 3.96E-02 | 7.94E-02 | 1.77E-03 | 5.01E-04 | 1.78E-04 | 7.75E-05 | 3.97E-05 | 2.32E-05 | 1.51E-05 | 1.06E-05 | 8.05E-06 | 6.43E-06 | 4.39E-06 | |
| 64 | 0.00E+00 | 2.48E-02 | 1.11E-01 | 2.69E-02 | 6.43E-03 | 9.53E-04 | 2.57E-03 | 8.90E-05 | 1.93E-05 | 1.26E-05 | 7.21E-06 | 5.08E-06 | 3.83E-06 | 2.05E-06 | 2.08E-06 | | |
| 62 | 0.00E+00 | 4.60E-02 | 9.54E-02 | 1.71E-02 | 2.55E-03 | 4.89E-04 | 1.27E-03 | 3.09E-04 | 1.08E-04 | 4.45E-04 | 2.23E-04 | 1.30E-04 | 8.43E-05 | 5.21E-06 | 3.55E-06 | 2.35E-06 | 1.77E-06 |
| 60 | 0.00E+00 | 7.48E-02 | 7.53E-02 | 1.02E-02 | 1.34E-03 | 2.40E-04 | 5.98E-05 | 1.24E-05 | 7.64E-07 | 4.36E-07 | 2.79E-07 | 1.95E-07 | 1.47E-07 | 1.16E-07 | 7.91E-08 | | |
| 58 | 0.00E+00 | 1.60E-01 | 5.45E-02 | 5.69E-03 | 6.63E-04 | 1.12E-04 | 8.83E-06 | 3.64E-06 | 1.80E-06 | 1.03E-06 | 6.59E-07 | 4.61E-07 | 3.46E-07 | 2.75E-07 | 1.87E-07 | | |
| 56 | 0.00E+00 | 1.33E-01 | 3.61E-02 | 2.97E-03 | 3.12E-04 | 5.01E-05 | 1.18E- | | | | | | | | | | |

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Table S2. Life history information and references available for small scombrids in the Atlantic Ocean (ICCAT, 2018).

| Species | Parameter | Northeast | Source | Southeast | Source | Mediterranean | Source | Northwest | Source | Southwest | Source |
|-------------------------------------|-----------|-----------|--|-----------|--------|------------------|---|------------|---|-----------|--------------------|
| <i>Sardinops sagax (BON)</i> | Lmax | 91.4 | Collette and Nauen, 1983 | | | 69.56 | Collette and Nauen, 1983 | | | | |
| | Linf | 73.0 | Baibat et al., 2016 | | | 0.439 | Kahraman et al., 2014 | | | | |
| | R | 0.3075 | Baibat et al., 2016 | | | -1.327 | Kahraman et al., 2014 | | | | |
| | to | -2.4469 | Baibat et al., 2016 | | | 5.04 | Collette et al., 1993 | | | | |
| | Tmax | | Baibat et al., 2016 | | | 36.6-39.93 | Hattour, 2000; Saber et al., 2017 | | | | |
| | Lm50 | 42.6 | Baibat et al., 2016 | | | 0.71 | Kahraman et al., 2014 | | | | |
| | Tm50 | | | | | 0.006321-0.00822 | Sabre et al., 2004 | | | 0.0135 | Hansen, 1987 |
| | WL_a | 5.00E-05 | Baibat et al., 2016 | | | | Sabre et al., 2017; | | | | |
| | WL_b | 2.7852 | Baibat et al., 2016 | | | 3.21-3.13 | Sinoviel et al., 2004 | | | 2.952 | Hansen, 1987 |
| <i>Euthynnus alletteratus (TA)</i> | Lmax | 82.6 | Cayré and Diouf, 1980 (spines) | | | 122 | Claro, 1994 | 106.68 | IGFA, 2011 | | |
| | Linf | | | | | 117-130.8 | Hattour, 2009; Hajjej et al., 2012 | 86 | Cabrera et al., 2005 | | |
| | R | | | | | 0.19-0.133 | Hattour, 2009; Hajjej et al., 2012 | 0.26 | Cabrera et al., 2005 | | |
| | to | | | | | -1.13 - -2.22 | Hattour, 2009; Hajjej et al., 2012 | -0.32 | Cabrera et al., 2005 | | |
| | Tmax | | Cayré and Diouf, 1980 (spines) | | | 10-7.4 | Hattour, 2009; Hajjej et al., 2012 | | | | |
| | Lm50 | 42 | Diouf, 1980 | | | 44.8-51.13 | Hajjej et al., 2012 | 39.7 | Ramirez-Arredondo, 1993 | | |
| | Tm50 | | | | | 1.89 | Hattour, 2009 | | | | |
| | WL_a | 0.0138 | Diouf, 1980 | | | 0.0329-0.01242 | Hajjej et al., 2011; Saber et al., 2017 | 0.00000205 | Ramirez-Arredondo et al., 1996; W in g and Fl in mm | | |
| | WL_b | 3.035 | Diouf, 1980 | | | 2.8101-3.050 | Hajjej et al., 2011; Saber et al., 2017 | | Ramirez-Arredondo et al., 1996; W in g and Fl in mm | | |
| <i>Acanthocybium solandri (WAM)</i> | Lmax | | | | | | | 2.96 | | | |
| | Linf | | | | | | | 200 | Hogarth, 1986 | | 197 |
| | R | | | | | | | 179.5 | Jenkins and McBride et al., 2008 | | Viana et al., 2013 |
| | to | | | | | | | 0.74 | McBride et al., 2008 | | |
| | Tmax | | | | | | | -1.91 | McBride et al., 2008 | | |
| | Lm50 | | | | | | | 9 | McBride et al., 2008 | | |
| | Tm50 | | | | | | | 0.25 | Jenkins and McBride, 2009 | | |
| | WL_a | 0.02749 | Santana et al., 1993 (Size distribution) | | | | | 0.64 | Jenkins and McBride, 2009 | | |
| | WL_b | 2.72252 | Santana et al., 1993 (Size distribution) | | | | | 3.242 | Beerthcker2005 | 0.0016 | Frota et al., 2004 |
| | | | | | | | | | | 3.275 | Frota et al., 2004 |
| <i>Alosa thunbergii (FIL)</i> | Lmax | | | | | | | | | | |
| | Linf | | | | | | | | | | |
| | R | | | | | | | | | | |
| | to | | | | | | | | | | |
| | Tmax | | | | | | | | | | |
| | Lm50 | | | | | | | | | | |
| | Tm50 | | | | | | | | | | |
| | WL_a | | | | | | | | | | |
| | WL_b | | | | | | | | | | |

Table S3. We run the OMs under perfect equilibrium conditions and three different scenarios using one fleet: (i) fish selected before maturing; (ii) fish selected at the length at maturity; and (iii) fish selected after maturing. A large sample size of 12 000 individuals was used. It seems that a persistent bias exists for both assessment methods compared with the OM, particularly for LIME in all cases and LBSPR except when fish are selected before maturing. LIME underestimates SPR for this species (albacore tuna), particularly when individuals are selected by the fishery before maturing. LIME also overestimates selectivity. LBSPR, however, is unbiased when fish are selected by the fishing gear before maturing and positively biased when fish are caught after maturing. For LBSPR, selectivity values are very close to the real values. These results under perfect equilibrium conditions help to understand biases in performance by both methods when different OMs are used.

| Scenario | Year | SPR_OM | SPR_LBSPR | SPR_LIME | S ₅₀ _OM | S ₉₅ _OM | S ₅₀ _LBSPR | S ₉₅ _LBSPR | S ₅₀ _LIME | S ₉₅ _LIME |
|----------|-----------|--------|-----------|----------|---------------------|---------------------|------------------------|------------------------|-----------------------|-----------------------|
| i | Last year | 0.42 | 0.41 | 0.09 | 78 | 86 | 77 | 91 | 81 | 89 |
| ii | Last year | 0.42 | 0.56 | 0.16 | 90 | 96 | 89 | 103 | 95 | 99 |
| iii | Last year | 0.42 | 0.68 | 0.21 | 101 | 108 | 96 | 112 | 113 | 125 |

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ICESJMS-2018-485 (EA)**Performance of length-based data-limited methods in a multi-fleet context: application to small tunas, mackerels, and bonitos in the Atlantic Ocean**

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Large scombrids, commercial tuna species, are regularly assessed and managed. However, most of the small scombrids, many mackerels and bonitos, lack accurate catch data to implement traditional stock assessments despite their economic importance in many small-scale fisheries. In this study, we analyzed different approaches using length composition data from multiple fleets with different gear selectivity to assess small scombrids in the Atlantic Ocean. Using simulated populations, we compared two length-based methods [length-based spawning potential ratio (LBSPR) and length-based integrated mixed effects (LIME)], under different length data grouping scenarios. We found that using length data from the fleet targeting the broadest range of sizes resulted in the lowest bias in spawning potential ratio of all options tested. Based on these results, we used biological and length data to estimate a quantitative proxy of current stock status for ten small scombrid stocks in the Atlantic Ocean. We found that some small scombrid stocks are likely to be overfished, such as little tunny (*Euthynnus alletteratus*) in the Southeast Atlantic and wahoo (*Acanthocybium solandri*) in the Northwest Atlantic. This is a starting point in the estimation of stock status for these species, but should not be thought of as a replacement for other more data-intensive assessment techniques. Instead, the framework developed should be used to help identify the data and knowledge needed to improve the sustainable utilization of these species.

Introduction

For the “principal market tunas”, like Southern (*Thunnus maccoyii*), Pacific (*T. orientalis*), and Atlantic bluefin (*T. thynnus*), bigeye (*T. obesus*), yellowfin (*T. albacares*), albacore (*T. alalunga*), and skipjack (*Katsuwonus pelamis*), stock assessments are performed regularly, and a variety of management procedures are in place to protect these stocks from overfishing (Pons *et al.*, 2017). However, there are also other scombrid species, commonly referred to as “small

tunas, mackerels, Spanish mackerels and bonitos" (from now on "small scombrids"), that account for a notable proportion of the total tuna and tuna-like species catch and that are mostly unassessed and unmanaged (Juan-Jordá *et al.*, 2015; Pons *et al.*, 2018). Small scombrids are generally coastal and associated with continental shelves and islands (Collette and Nauen, 1983). Although their economic value is lower than that of the principal market tunas (Collette *et al.*, 2011), they sustain important regional commercial fisheries in many coastal communities throughout their distributions (Majkowski, 2007). Juan-Jordá *et al.* (2011) showed that within the Scombridae family, the fastest declines in biomass are exhibited not only for the largest, longest-lived, most valuable tunas, but also for a few smaller, short-lived mackerels. Also, some small Scombridae stocks in the Atlantic Ocean were assigned as "moderate to high risk" of being overfished or subject to overfishing in a recent qualitative risk assessment, even if they have not been formally assessed in recent years (Lucena-Frédu^o *et al.*, 2017a).

Total catch is one of the main data sources required for most classical stock assessment methods, particularly when estimating absolute estimates of spawning or total biomass. Stock assessment methods used for principal market tunas use catch data, but obtaining accurate landings and discards for small scombrids is generally challenging (Pitcher *et al.*, 2002). Small scombrids are targeted by multiple fleets, particularly medium- and small-scale fisheries, and caught as bycatch in many industrial fisheries targeting commercial tuna species. Available catch data usually consist of incomplete catch time-series from tuna regional fisheries management organizations (tRFMO) statistics and from catch time-series that might be highly aggregated by species from the Food and Agriculture Organization (FAO) database (FAO, 2016). While quantifying total catch is difficult, there is a wide-ranging toolbox of qualitative and quantitative assessment approaches for data-limited fisheries to infer the exploitation status of the stocks (Chrysafi and Kuparinen, 2016; Dowling *et al.*, 2016). In 2017, Lucena-Frédu^o *et al.* (2017b) performed a qualitative risk assessment for small scombrids in the Atlantic Ocean. They identified five of 13 species as priority for evaluation and implementation of future management actions: the low productivity and susceptible *Euthynnus alletteratus* (little tunny), *Acanthocybium solandri* (wahoo), and *Scomberomorus cavalla* (king mackerel), and the highly targeted *Sarda sarda* (bonito) and *Auxis thazard* (frigate tuna). This study served to identify priority species, but does not estimate population processes, productivity, or stock status that would be required for more specific management advice. In addition, qualitative risk assessment methods have been questioned recently since their performance is poor under a wide range of conditions (Hordyk and Carruthers, 2018).

The International Commission for the Conservation of Atlantic Tunas (ICCAT) suggested that length composition of the catch available in the ICCAT database should be used to quantitatively assess the status of these species and inform management advice (ICCAT, 2017). In fisheries without total catch data or information on relative or absolute abundance, stock assessments typically use the spawning potential ratio (SPR) as an alternative reference point to the biomass at maximum sustainable yield (B_{MSY}). SPR can be expressed in terms of spawning stock biomass per recruit (SSBR), which is often defined as the expected lifetime reproductive potential of an average recruit. SPR then is the ratio of the fished SSBR to the unfished SSBR under equilibrium conditions (Goodyear, 1993). SPR has been recommended for data-limited assessment because it can be estimated using only biological information and length data (Brooks *et al.*, 2010).

There are several methods that use life history information and length composition from the catch to estimate fishing intensity and derive values of SPR that can be used as a proxy for

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3 stock status. One of them is length-based spawning potential ratio (LBSPR, Hordyk *et al.*, 2015).
4 This method uses the Beverton–Holt life history ratios in an equilibrium-based population model
5 applying the shape of the length composition data compared with the expected unfished length
6 structure to estimate the ratio of fishing mortality and natural mortality (F/M) and derive SPR.
7 Another new method is the length-based integrated mixed effects model (LIME, Rudd and
8 Thorson, 2018), that also requires biological information and length composition data to derive
9 SPR, but relaxes the equilibrium conditions by treating recruitment as a random effect over time
10 and estimating annual F as fixed effects (Rudd and Thorson, 2018). Both methods can be
11 implemented in *R* (Hordyk, 2017; R Core Team, 2017; Rudd, 2018).

12 Data-limited, length-based stock assessment methods typically assume selectivity is
13 asymptotic by default (Hordyk *et al.*, 2015; Rudd and Thorson, 2018). If large fish are absent
14 from the catch, logistic selectivity models assume that they do not exist in the population (as
15 opposed to being present in the population, but evading the fishing gear). The logistic selectivity
16 assumption is usually violated in highly size-selective fisheries (i.e. gillnets), and it could be
17 problematic in multifleet fisheries where stocks are caught in different proportions by multiple
18 gears with different selectivity patterns. As an example, the majority of the catch of the North
19 Atlantic albacore stock comes from pole and line fisheries which have dome-shaped selectivity,
20 catching mainly juvenile albacores. In addition, a smaller proportion of the catch comes from
21 longline fisheries targeting larger individuals, but with different selectivity patterns depending on
22 the fishery (ICCAT, 2014). These different selectivity patterns, catches, and indices of
23 abundance are included in complex assessment models that allow for multiple fleet interactions
24 in the formal assessment performed regularly by ICCAT. When fitting only to the length
25 composition of a proportion of the catch, assumptions regarding fishery dynamics, particularly
26 the shape of the selectivity curve, need to be carefully analyzed.

27 The main objective of this study is to determine the current stock status of small
28 scombrids in the Atlantic Ocean. However, the only data available to evaluate these populations
29 are limited biological information and length composition of the catch coming from different
30 fleets with different selectivity patterns. Therefore, before applying any length-based assessment,
31 we need to determine how to combine these length data to obtain an unbiased estimate of fishing
32 intensity. Developing best practices for combining length data across multiple fleets for length-
33 based assessments of small scombrids in the Atlantic Ocean thus becomes a critical challenge to
34 face in order to estimate stock status. To address this challenge and meet the main objective, we
35 used simulation testing to evaluate the performance of LBSPR and LIME by combining length
36 composition data of the catch from multiple fleets with different selectivity patterns. Using
37 conclusions from the simulation, we applied both length-based approaches to estimate a proxy of
38 stock status for the priority small scombrid species determined by ICCAT (ICCAT, 2018).

45 Methods

46 First, we compared the performance of two length-based methods using a simulation study based
47 on North Atlantic albacore. Then, based on the insight about the robustness of the methods, we
48 estimated stock status for the small tuna stocks.

51 Simulation study

52 We chose the North Atlantic albacore stock on which to develop an operating model (OM) to
53 simulate resource dynamics in order to evaluate the performance of the different assessment
54 methods. ICCAT, in the 2017 report of the small tunas species group intersessional meeting

(ICCAT, 2017), suggested the North Atlantic albacore stock as a good example of a multifleet fishery to use for simulation purposes, where the selectivity patterns are considered well estimated for 12 different fleets (ICCAT, 2014). This stock is targeted by pole and line, troll, longline, and other surface gears in the Atlantic Ocean.

In the next sections, we describe the data and specifications used in the OM, the estimation models (EM), namely LBSPR and LIME, and how we measured their performance under different scenarios.

Operating models

Input data

We extracted the catch time-series (Figure S1 in the Supplementary material), selectivity patterns from 12 different fleets (Figure S2 in the Supplementary material), and the biological parameters (Table 1) from the formal assessment performed by ICCAT in 2013 for North Atlantic albacore (ICCAT, 2014) to use for the OM. For simplicity and based on similarity of selectivity patterns (Figure S2 in the Supplementary material), we combined some fleets using an average age-specific selectivity: (A) fleets 1 and 2 (bait boat and troll fisheries) which target small individuals and have a dome-shaped selectivity curve; (B) fleets 4 and 12 which are other surface gears targeting a broader range of sizes; and (C) fleets 10 and 11 which are longline fisheries targeting mainly adults with an asymptotic selectivity curve (Figure S2 in the Supplementary material). During the last 15 years only, the three fleet combinations mentioned earlier were still operating plus fleet 7, which is a longline fleet that captures less than 1% of the total catch. Eighty-seven percent of the total catch corresponds to the fleets' bait boat and troll fisheries (A), 6% longline (C) and 7% other surface gears with same selectivity (B) (Figure 1).

Model specifications

We simulated an age-structured population using stock synthesis (SS) Version 3.30.10 (Methot and Wetzel, 2013; Methot *et al.*, 2018). We specified a final depletion fitting to an artificial abundance survey index equal to 1 at the beginning of the time-series (1930) and $0.4 B_0$ in the last year (2011). The depletion value was arbitrarily selected to mimic the current depletion of the North Atlantic albacore population. All parameters were fixed, except the average recruitment in the unfished state (R_0). We assumed a Beverton–Holt spawner-recruit function (Beverton and Holt, 1957; Methot and Wetzel, 2013). We simulated two populations, one with and one without recruitment deviations in order to verify if the outputs were different between the estimations models since LIME specifically estimates recruitment.

Fishing intensity in SS was estimated to match the observed North Atlantic albacore catch. SS assumes that the absolute level of catch is known, using the catch time-series to calculate the level of fishing intensity needed to obtain that level of catch, conditioned on the model's current estimate of age-specific population abundance and age-specific selectivity (Methot and Wetzel, 2013). Single estimates of fishing mortality rates (F) were calculated for all gears combined in stock synthesis. LBSPR and LIME (see the estimation models section) assume a single gear, so the F estimates represent the F as covered by the stock sampled. SS calculates the SPR as the equilibrium level of spawning biomass per recruit that would occur with the current year's level of fishing intensity relative to the unfished level of spawning biomass per recruit (Goodyear, 1993).

After running SS to generate the OM, we extracted the expected catch at age by year and fleet from the SS report. We converted this catch at age in biomass into catch at age in numbers

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3 using the mean weight at age. We used the age-length transition matrix output from SS to assign
4 a distribution of length at each age (Table S1 in the Supplementary material). Summing across
5 each length bin by gear gave us the length distribution of the catch. We used a 2-cm length bin,
6 which corresponds to the bin structure applied in the formal ICCAT assessment (ICCAT, 2014).
7 In order to analyze different length sampling scenarios, we sampled 100 individuals from the
8 catch by year and fleet with a multinomial distribution using the probability of being harvested at
9 each length bin for each year. We repeated this process of simulating a population and generating
10 data for 100 replicates for each scenario. We chose to sample only 100 individuals because no
11 big differences in the results were found when using a larger sample size (i.e. 1000 and 10 000
12 individuals), just a small reduction in variance (see Pons, 2018).
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15 Scenarios

16 A common question that arises with length data from multifleet fisheries with different
17 selectivity patterns is which fleets to use and how to combine data when applying length-based,
18 data-limited methods that only estimate selectivity and fishing mortality for one fishing gear. We
19 explored the performance of each estimation method under different approaches combining
20 length data into one common “fleet”. The single fleet could actually be one selected fleet or
21 multiple fleets combined in some way. In all scenarios, the selectivity for this one fleet was
22 estimated and starting values were the same for each model run. We explored five possible
23 scenarios:
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- 25 • Scenario 1: Length composition sampled proportional to the catch of each fleet (Figure
26 1). This means that fish measured from the fleet with the highest catch would be more
27 represented in the length composition data than other fleets.
- 28 • Scenario 2: Length composition sampled with equal weight from each fleet. This means
29 that the same number of individuals were measured from each fleet and combined in one
30 length-sample. All fleets are equally represented in the length composition data.
- 31 • Scenario 3: Only use length data from the fleet that targets small individuals (fleet A).
32 Fleet A has a dome-shaped selectivity (Figure 1a) where the true S_{50} is 57 cm (~ age 1.5)
33 and S_{95} is 61 cm (~ age 2). It was modeled in SS with a double normal distribution with 6
34 parameters. This fishery catches mainly juveniles and it is the main fishery for North
35 Atlantic albacore in terms of catch (Figure 1b).
- 36 • Scenario 4: Only use length data from the fleet that targets a broad range of lengths (fleet
37 B). Fleet B has an asymptotic selectivity harvesting both juveniles and adults, with a true
38 S_{50} of 78 cm (~ age 3.5) and S_{95} of 90 cm (~ age 5, Figure 1a). In terms of catch, this fleet
39 resents a small proportion of the total (Figure 1b).
- 40 • Scenario 5: Only use length data from the fleets that target adults (fleet C). Fleet C also
41 catches a small fraction of the total catch (Figure 1b), but is a longline fishery that targets
42 mainly adults with a true S_{50} of 100 cm (~ age 7) and a S_{95} of 108 cm (~ age 9) (Figure
43 1a).

44 Estimation models

45 In LBSPR, SPR in an exploited population is a function of the ratio of fishing mortality to
46 natural mortality (F/M), selectivity and the two life history ratios M/k and L_m/L_∞ ; k is the von
47 Bertalanffy growth coefficient, L_m is the size of maturity, and L_∞ is asymptotic size (Hordyk *et*
48 *al.*, 2015). The inputs to the LBSPR are: M/k , L_∞ , the variability of length-at-age (CVL_∞), which
49 was set as 10% in the OMs; and size of maturity specified in terms of L_{50} and L_{95} , the size at
50

which 50 and 95% of a population matures (Table 1). Given the assumed values for the M/k and L_∞ parameters and length composition data from an exploited stock, the LBSPR model uses maximum likelihood methods to estimate the selectivity ogive, which is assumed to be a logistic curve defined by the selectivity-at-length parameters S_{50} and S_{95} and the relative fishing mortality (F/M), which are then used to calculate the SPR (Hordyk *et al.*, 2015). LBSPR estimates a selectivity curve for each length sample. Estimates of SPR are primarily determined by the length of the fish in a sample, relative to the maturity and L_∞ . For example, if a reasonable proportion of fish in a sample attain lengths approaching L_∞ , estimates of F/M will be relatively low leading to a high estimate of SPR. However, the proportion of length samples near L_∞ will vary with the life history parameters such as fecundity-at-age/length and selectivity. LBSPR is an equilibrium-based method with some underlying assumptions including: (i) asymptotic selectivity, (ii) growth adequately described by the von Bertalanffy equation, (iii) a single growth curve can be used to describe both sexes which have equal catchability, (iv) length-at-age is normally distributed, (v) rates of natural mortality are constant across adult age classes, (vi) growth rates remain constant across the cohorts within a stock, and (vii) constant recruitment (Hordyk *et al.*, 2015). In this study, we used LBSPR package version 0.1.2 in R (Hordyk, 2017). The LBSPR package uses the Rauch–Tung–Striebel smoother function to smooth out the multiyear estimates of SPR (Hordyk, 2017), and these smoothed values were used for comparisons with the OM.

LIME is an integrated, age-structured model which requires, as input, biological information such as growth, natural mortality, and maturity, and, at minimum, one year of length composition data. LIME estimates, as fixed effects, annual fishing mortality rates F , S_{50} , and S_{95} , the recruitment standard deviation, and a Dirichlet-multinomial parameter governing the relationship between the nominal and effective sample size of length measurements. LIME has most of the same assumptions as LBSPR, but LIME does not assume equilibrium conditions when recruitments can be estimated (i.e. more than one year of length data). LIME extends length-based methods by deriving time-varying recruitment deviations (Rudd and Thorson, 2018) using automatic differentiation and Laplace approximations (TMB) (Kristensen *et al.*, 2015) to calculate the marginal likelihood for the random effect on recruitment. Using the assumed biological information, recruitment deviates, estimated F , and estimated selectivity, LIME calculates the predicted abundance-at-age over time. To predict length composition of the catch, LIME calculates the predicted probability of being captured at age over time, the probability of being in a length bin given age, and then the probability of being captured in each length bin. LIME fits the predicted length composition to the observed length composition using the Dirichlet-multinomial likelihood function. In addition to the length composition marginal likelihood, the joint negative log likelihood also includes a penalty on the fishing mortality rate to avoid varying unrealistically between years and the likelihood of the random effect of annual recruitment deviations. LIME can also accommodate catch and/or abundance data if available (Rudd and Thorson, 2018), although this feature was not used in this study. We used the LIME package version 1.0.5 (Rudd, 2018).

Performance measures

The performance of the EMs under different scenarios was compared with the simulated “truth” from the OM using relative error (RE) calculated as $(\text{estimated}-\text{true})/\text{true}$, where *estimated* comes from the EM and *true* from the OM. This is a measure of uncertainty, in both bias and precision, of the EM under each scenario, and it is commonly used as a standardized metric of

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3 model performance. We used SPR as the performance measure for all scenarios estimated by
4 both LIME and LBSPR. We presented the relative error of the last year of the time-series of SPR
5 in all cases for the 100 simulation replicates for each scenario.
6

7 **Assessment of small scombrids in the Atlantic Ocean**

8 Little tunny, bonito, wahoo, king mackerel, and frigate tuna have been identified as priority to be
9 evaluated by ICCAT in 2017 (ICCAT, 2017). In the present study, the only species that we did
10 not evaluate was king mackerel. In the Southwest Atlantic, there is no good information on
11 length data to evaluate this stock; in the Northwest Atlantic, it is regularly assessed by the US as
12 two independent stocks: one in the Gulf of Mexico and the other off the southeast coast of the
13 US (SEDAR, 2014a, 2014b). According to these reports, neither stock of king mackerel in the
14 Northwest Atlantic is currently overfished nor undergoing overfishing.
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16 None of the other four species of small scombrids have studies defining stocks
17 boundaries in the Atlantic Ocean. So, for management purposes, ICCAT uses five sampling or
18 statistical areas for small scombrids: Northwest Atlantic, Southwest Atlantic, Northeast Atlantic,
19 Southeast Atlantic, and Mediterranean Sea (Figure S3 in the Supplementary material). Hence, we
20 decided to use these areas as proxies for stock boundaries to assess these putative “stocks”.
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22 The ICCAT database (<http://iccat.es/en/accesingdb.htm>) has length data from 1975 to
23 2016 for the four priority species assessed in this study. The length composition data available
24 for each stock come from different regions and different gear types. To estimate current stock
25 status, we used only data from 2010 to the present where there is a better representation of the
26 length composition of the catch by year and gear (Figure S4 in the Supplementary material). We
27 used the length data reported in 1- and 2-cm bins and pooled them into 2-cm length bins for the
28 analysis. The number of fish measured by year for the priority species varies between 17 429
29 individuals measured in 2016 to 98 173 in 2014, all species combined (Figure S4 in the
30 Supplementary material). We presented the stock status for the year 2014 where there are more
31 length data and are consistent among species and representative of different gears.
32

33 For some stocks, the length data available were limited, so we removed samples
34 numbering fewer than 100 fish per year and gear combination (Figure S4 in the Supplementary
35 material). Some stocks, such as wahoo in the Southwest Atlantic, were excluded from the
36 analysis because they are targeted by multiple fleets, but length data are available only for one
37 gear (gillnets) and would bias the results. This filtering process reduced the number of stocks
38 with enough information to run the length-based models. We did not run these models for bonito
39 in the Southeast, Northwest, and Southwest, little tunny in the Southwest, wahoo in the
40 Mediterranean, Southeast, and Southwest (stock not present in the Mediterranean), and frigate
41 tuna in the Mediterranean, Southwest, and Southeast, resulting in 10 stocks with representative
42 information of length composition data of catch by gear (Figure S5 in the Supplementary
43 material, Table 2).
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45 Both LBSPR and LIME require life history information on growth, maturity, and length-
46 weight relationships as input parameters. These methods are very sensitive to these parameters
47 (Hordyk *et al.*, 2015; Rudd and Thorson, 2018). In 2018, the ICCAT small tunas working group
48 met, and a set of life history parameters were agreed among scientists from each region in the
49 Atlantic Ocean for each stock to run data-limited methods (ICCAT, 2018, Table S2 in the
50 Supplementary material). There are many gaps in the life history information available for these
51 species. In cases where there was missing information, we borrowed information from the
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nearest stock of that species (i.e. when missing information existed for the Southeast Atlantic, we borrowed the information from the Northeast Atlantic) to run the length-based models.

Table 2 shows the final parameters used for each stock to run LBSPR and LIME. Natural mortality (M) was calculated using a suite of empirical life history-based methods (Cope, 2017, see http://barefootecologist.com.au/shiny_m). We used nine empirical methods that used growth parameters (L_∞ , k , t_0 , and maximum age). Four of these methods are described in Then *et al.* (2015), two in Jensen (1996), and the other three in Alverson and Carney (1975), Chen and Watanabe (1989), and Jensen (1997), respectively. Table 2 shows the median and first and third quantile of the distribution of M estimated for each stock. LBSPR and LIME were run with these three M values to test their sensitivity to these parameter estimations.

Reference points for small scombrids

We used SPR as a biological reference point. In general, it is used as a proxy for B_{MSY} when information on the scale population size is not available. A harvest strategy that targets a fishing mortality rate that is expected to result in 40% of the unfished spawning output ($SPR_{40\%}$) is considered a reasonable proxy even for stocks with very low resiliency (Clark, 2002). Moreover, 30% SPR is sometimes considered a threshold beyond which overfishing would be occurring (Clark, 2002; Nadon *et al.*, 2015; Rudd and Thorson, 2018). In addition, we presented the estimated ratio F/M for each stock.

Results

Simulation testing: length data in multifleet fisheries

Based on the observed catch data for North Atlantic albacore used in the OM, the true SPR value in the terminal year was 0.55 for the OM without recruitment deviations; for the OM that includes random recruitment deviations, the median was 0.66, with a range 0.50–0.74 (Figure S6 in the Supplementary material). LBSPR was least biased when using length data from the fleet with asymptotic selectivity catching a broad range of lengths from juveniles to adults (Scenario 4; Figure 2). LIME was least biased with length data from the fleet that targets only adults when considering recruitment variability (Scenario 5; Figure 2). In the Supplementary information, we presented potential biases in the estimation models compared with our OMs under perfect equilibrium conditions. It seems that a persistent bias exists for both assessment methods compared with the OM, except for LBSPR when fish are selected before reaching the length at maturity. In general, LIME underestimates SPR for this species (North Atlantic albacore), particularly when individuals are selected by the fishery before maturing. LIME also overestimates selectivity most of the time, but is less biased when adults are present in the samples. LBSPR is slightly positively biased only when the sampled individuals are above the size at maturity (Table S3 in the Supplementary material).

In Scenario 1, length composition data was weighted by the fleet's proportional catch, meaning that more weight was given to the fleet with dome-shaped selectivity. In this scenario, both LBSPR and LIME underestimated SPR on average in both recruitment scenarios (Figure 2). Under an asymptotic selectivity assumption, if large individuals are absent from the catch, both assessment methods estimate F to be higher than the truth and SPR lower than the truth. LIME estimated SPR to be almost zero. Results from Scenario 3 were similar to Scenario 1 since both scenarios put higher weight on length compositions consisting of mainly juveniles or smaller individuals than the full span of vulnerable fish.

Under Scenario 2, sampling the same number of individuals by gear type, LBSPR and LIME estimated SPR higher than the truth, particularly when the OM did not consider recruitment variability. When considering recruitment variability, LBSPR was positively biased, although LIME was less biased, but less precise. Under these scenarios, the proportion of large individuals in the catch was overrepresented, leading to the EMs estimating higher SPR values than expected. The same overestimation of SPR occurred in Scenario 5 using the fleet that targets adults when no recruitment variability was included in the OM due to the proportions of large individuals in the catch. However, as was shown under perfect equilibrium conditions (Table S3 in the Supplementary material), LIME tends to be less biased when large individuals are in the samples.

LBSPR was less biased in Scenario 4 when considering only the fleet with an asymptotic selectivity that captures a broad range of sizes (from juveniles to adults), while LIME was less biased under the scenarios with recruitment variability when considering the fleets with gears that selected mainly adults in Scenario 5 (Figure 2). We observed that, in many cases, even under equilibrium conditions (Table S3 in the Supplementary material), LIME estimated higher selectivity parameter values, S_{50} and S_{95} , and lower SPR values than LBSPR. This is probably the reason why LIME performs better when using fleets that target large fish and why LIME SPR estimates are lower than LBSPR SPR estimates when using the same data.

Assessments of small scombrids in the Atlantic Ocean

Based on simulation testing, none of the scenarios produced the best performance for both estimation models (LIME and LBSPR) simultaneously. LBSPR performed best in Scenario 4, which used the length data coming from the fleet with an asymptotic selectivity targeting a broad range of lengths. LIME, however, performed better in Scenarios 5, where mainly adults were represented in the catch. So, based on these results, we decided to apply both LIME and LBSPR using the length composition data from small scombrids from the fleet that has a broader range of sizes including adults, but not restricted to the adult portion of the catch. The gears used then varied among small tuna stocks (Figure 3).

The length composition data for each stock by gear, filtered by year–gear combinations with at least 100 length measurements, varies among areas. These differences likely stem from variable fleets operations in each region. Length composition data for little tunny are available for two gears in the Northwest Atlantic, but rod and reel has more representative sampling by year and length range compared with traps (Figure S5 in the Supplementary material). In this case, we used length data only from rod and reel to assess this stock (Figure 3). For little tunny in the Northeast Atlantic, we selected the length data coming from traps since they cover a broader range of ages including adults, despite the fact that there are no data in 2011. For little tunny in the Mediterranean, we used length data from longlines; for the Southeast Atlantic, we used data from gillnets (Figure S5 in the Supplementary material, Figure 3). For wahoo in the Northeast, we used the length composition from handlines since it was the only information available; we used rod and reel data for the Northwest. For bonito in the Mediterranean, we used length data coming from longlines, such as for little tunny in the same area. Finally, for frigate tuna in the Northeast and Southeast, we selected the length data coming from purse-seine fisheries (Figure S5 in the Supplementary material, Figure 3).

For some small scombrid stocks, the SPR estimates were below the target of 40%, but the results varied between assumptions about M and the estimation method considered (Figure 4). LBSPR and LIME predicted different values of SPR; the estimated values were sometimes far

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3 apart, such as for bonito in the Mediterranean and in the Northeast (e.g. 0.2 with LIME and 0.6
4 with LBSPR). LIME estimated a higher F and lower SPR than LBSPR (Figure S7 in the
5 Supplementary material). In contrast, for bonito in the Northeast, LIME estimated a higher
6 selectivity ogive, lower F , and higher SPR than LBSPR. For 2014, LIME estimated high
7 recruitment; the small individuals in the catch were attributed to the recruitment spike, as
8 opposed to LBSPR, which interpreted the small individuals as a high F (Figure S7 in the
9 Supplementary material). As expected, when M was assumed to be lower than in the base case
10 scenario (median M), the SPR estimations were lower (Figure 4).

11 SPR for little tunny in the Southeast was below 0.40 in all cases, except when LIME
12 assumed a high M , which resulted in a SPR estimate of 0.70 (Figure 4). Assuming the median
13 value of M , LBSPR predicted a very high F and SPR values below 0.20 for the entire time-series.
14 LIME also predicted low SPR values, around 0.30, and a much lower S_{50} than LBSPR (Figure S8
15 in the Supplementary material). The results of LBSPR and LIME were more similar for little
16 tunny in the Northwest, Mediterranean, and Northeast Atlantic (Figure S8 in the Supplementary
17 material). SPR for these stocks were above the 40% target reference point. For little tunny in the
18 Mediterranean, the ratio between F/M is high (above 2), even though SPR is above 40%.

19 SPR estimates for both LBSPR and LIME for wahoo in the Northwest were below 40%
20 (except in the high M scenario with LIME). LIME predicted very high F/M and SPR below 40%,
21 except when assuming a higher M for wahoo in the Northeast. In both the Northwest and
22 Northeast, LIME predicted a higher S_{50} and a lower SPR than LBSPR (Figure S9 in the
23 Supplementary material). None of the frigate tuna stock assessments estimated SPR below 40%,
24 except with LIME in the low M scenario where SPR was estimated at ca. 20% for the Northeast
25 and Southeast stocks (Figure 4, Figure S10 in the Supplementary material). Table 3 summarizes
26 the status of the ten small Scombrids stocks assessed under the base-case scenario, considering
27 the median of the M distribution.

33 Discussion

34 Length data in multifleet fisheries

35 We showed how stock assessment results could be highly biased when using only one gear that
36 is not representative of the length of the exploited population, particularly when the assumptions
37 of asymptotic selectivity are violated (e.g. albacore length data coming from bait boat and troll
38 fisheries targeting juveniles with dome-shaped selectivity). In this case, high catches of smaller
39 individuals resulted in an underrepresentation of the proportion of adults in the population,
40 estimating a lower SPR value than the actual value. Even if the asymptotic selectivity
41 assumption is met (i.e. albacore length data coming from longline fleets targeting adults), SPR
42 was overestimated. Hordyk *et al.* (2015) suggested that when there are multiple fleets targeting
43 the same stock, the LBSPR model should be applied to the data from the fleet that targets the
44 adult portion of the stock. However, we found that SPR estimates were positively biased when
45 fish are caught after reaching the size at maturity L_{50} (Table S3 in the Supplementary material).
46 In all the scenarios analyzed by Hordyk *et al.* (2015), the S_{50} was lower than the L_{50} , but in our
47 scenarios 5 and 6, the S_{50} was higher than L_{50} , explaining why they did not find this bias in their
48 results. SPR estimates are primarily determined by the size of the fish in a sample relative to both
49 size at maturity L_{50} and L_{∞} . In our Scenario 4, where the S_{50} was lower than the L_{50} , LBSPR was
50 less biased, as we also showed assuming perfect equilibrium conditions (Table S3 in the
51 Supplementary material).

Rudd and Thorson (2018) tested the performance of LIME under LBSPR's own OM (Hordyk *et al.*, 2015), with relative ages based on the M/k ratio. They found that LBSPR performs well across all life history types, but LIME underestimated SPR for the medium- and longer-lived life history types and overestimated SPR for the short-lived life history type. However, in most of the non-equilibrium scenarios, LIME performed better than LBSPR. We also found in most of the scenarios considered that LIME estimated a lower SPR than LBSPR for this medium-lived albacore tuna. Also, we showed how, even under equilibrium conditions, that different OMs structures can show biases in the performance of the estimation models (Table S3 in the Supplementary material).

Based on our results, we recommend that, when there are multiple fleets with different selectivity patterns targeting one stock, and the length-based estimator assumes asymptotic selectivity, length-based models should be applied to the length data coming from the fleet that targets the broadest range of sizes including adults, but not restricted only to the adult portion of the catch. In particular, it is important to include juveniles because SPR estimates improve when the catch length sample is representative of the length composition of the entire exploited population.

Small scombrids stock status

LBSPR estimates of SPR, selectivity parameters, and the ratio of fishing mortality to natural mortality (F/M) are independently determined each year. However, LIME includes length composition data available for multiple years in the same model to estimate a single selectivity curve for all years, but fishing mortality and recruitment can vary among years. Therefore, assumptions and model structure are different between LIME and LBSPR, leading, unsurprisingly, to different results on the proxy of the stock status for small scombrids.

We did not find a specific pattern in exploitation status among regions (Figure 4) as regions in the Atlantic Ocean showed comparable stock status. Although some combinations of stock assessment model and natural mortality rate resulted in differing estimates of stock status, the approaches agreed under the base scenario with median M that two stocks out of ten are overfished: the little tunny in the Southeast and wahoo in the Northwest.

Little tunny

The length composition data for little tunny in the Northeast and Northwest Atlantic from purse-seiners were very similar and both assessment methods indicated that these stocks are above stock status targets ($SPR > 0.4$). On the other hand, for little tunny in the Southeast, most of the fish caught were below the length at maturity and, in most of the scenarios, this stock was estimated to be overfished. These results agree with a preliminary qualitative risk assessment analysis performed for small scombrids in the Atlantic Ocean considering two populations, north and south. The southern stock was found at high risk, while the northern population was found at moderate risk (Lucena-Frédu^o *et al.*, 2017b). This species has an estimated maximum age between 8 and 10 years (Cayre and Diouf, 1980) and an estimate of L_{∞} between 86 and 117 cm. Adults of this species (> 60 cm) in the Southeast are scarce in the length composition leading to low estimates of SPR. Along with bonito and frigate tuna, this species is one of the most captured among all small scombrids in the Atlantic Ocean (ICCAT, 2018).

Bonito

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In the base-case scenario, bonito in the Northeast was estimated to have a SPR below target reference points ($\text{SPR} < 0.4$) with LBSPR, but not with LIME. The opposite was observed in the Mediterranean, where LIME estimated a lower SPR than LBSPR. Rudd and Thorson (2018) found that LIME generally estimated a higher SPR than the truth for short-lived fish in a yearly time-step. A monthly time-step could be considered in the future for this species to test for sensitivity to this assumption since the life span for this species is 5 years (Baibat *et al.*, 2016).

Previous data-limited assessment methods were applied for bonito in the Northeast using Morocco landings data between 2012 and 2014. A Powell–Wetherall plot approach was used to explore changes in total mortality based on length samples and catch-curve analysis using lengths converted to age and cohort slicing (Ahmed *et al.*, 2015). Assuming $M = 0.2$, they found that fishing mortality is twice this value and suggested that this stock might be fully exploited. The M values used in the present study were > 0.2 in all cases, so using such a low value for M could give similar results as in Ahmed *et al.* (2015). However, for this short-lived species, we assume natural mortality should be > 0.2 (as is typical for other short-lived species).

This species is the most captured among all small scombrids (ICCAT, 2018), but the biological information as well as the length composition data available are highly fragmented and variable. Our results should be analyzed with caution; as better data becomes available, these stocks should be re-evaluated. This should be a high priority item for ICCAT.

Frigate tuna

In almost all scenarios, the stocks were estimated to be $> 40\%$ SPR. However, assessments for the Northeast stock always estimated lower SPR values than for the one in the Southeast. Again, these results matched the preliminary risk assessment for small scombrids in the Atlantic Ocean, where stocks in the south are at lower risk than the ones in the north (Lucena-Frédon *et al.*, 2017b). However, both F and SPR estimates in the Southeast should be considered with caution, since some of the results are in the low-right quadrant at F , close to 0 and SPR close to 1 with very high uncertainty. In particular, for LIME, this could indicate an unconverged model (Figure S8 in the Supplementary material). If F was estimated to be close to 0, it is likely that the life history information is inaccurate because we know the F is not 0 since a fishery is occurring. L_∞ might be too low, so both models would estimate no fishing if the observed lengths are very close to the asymptotic length. The growth parameters should be discussed again at the next small tuna group meeting in order to consider different life history values for this stock.

Wahoo

Both LIME and LBSPR estimated low SPR values for the Northwest stock, suggesting that this stock is overfished. In the Northeast, only LIME in the base case and low M scenarios estimated that this stock is overfished, but not LBSPR. In the South Atlantic, wahoo has been categorized as high risk, and no assessments are available for this stock. This species, along with bonito, should be considered as a priority to assess by ICCAT.

Future directions

We estimated for the first time a proxy of current stock status for ten stocks of the small scombrid group of species in the Atlantic Ocean. This is a very important starting point in the estimation of stock status for these species, but the wide uncertainty in estimates combined with differences in results between LBSPR and LIME demonstrate that data-poor methods are not substitutes for more data-intensive assessment techniques. ICCAT still needs to support the

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3 collection of improved life history information, length data from all fishing gears, and total
4 removals for these stocks and associated fishing data. While collecting such information in small
5 coastal fisheries is challenging, particularly for catch and effort data, it is important to develop
6 indices of abundance. A full assessment might return estimates with higher precision, although it
7 does not mean that it would be more accurate than the estimates provided here. Data collection
8 should focus on filling gaps and improving existing biological information, particularly for
9 stocks such as little tunny in the Southeast Atlantic, where data were borrowed from the
10 Northwest Atlantic stock. Also, ICCAT should emphasize the importance of obtaining length
11 data across the different gears, particularly those that cover a broad range of sizes. For example,
12 wahoo in the Southwest Atlantic were excluded from the analysis because they are targeted by
13 multiple fleets, but length data are available only for one gear (gillnets) and using only this data
14 could bias the results.

15 LBSPR and LIME, like all age- or length-based methods, are sensitive to
16 misspecifications of the inputs of life information (Hordyk *et al.*, 2015; Kokkalis *et al.*, 2017;
17 Rudd and Thorson, 2018). Sensitivity tests in these studies demonstrated the impact of the
18 misspecification of biological parameters. Quantification of uncertainty is one of the next steps
19 in the evaluation of these small scombrid stocks, not only for M , but for other growth and
20 maturity parameters, to provide support for local biological studies of these species. To account
21 for the uncertainty in the biological parameters with the current information available in the
22 Scombridae database (Juan-Jordá *et al.*, 2016), a Monte Carlo algorithm could be applied in
23 specifying prior distributions for life history parameters (Prince *et al.*, 2015) for this group of
24 species.

25 Small scombrid fisheries in the Atlantic Ocean are medium- to small-scale, data-limited,
26 and generally unassessed, with a lack of management and enforcement, with the exception of
27 some regions in the Northwest Atlantic, such as in the US. Determining stock status is the first
28 step to protect these stocks from being overfished and to apply management measures to rebuild
29 overfished stocks. Since stock status for these species is highly uncertain, simulation testing is
30 needed to evaluate different management procedures accounting for data and model uncertainty
31 in a management strategy framework. Management strategy evaluation (MSE) could be used to
32 develop robust management frameworks for such stocks, e.g. using the data-limited-methods
33 toolkit (DLMtool, Carruthers and Hordyk, 2018).

34 **Conclusions**

35 The present study analyzed different approaches using length-based, data-limited assessments
36 when length composition data come from fisheries with multiple gears with different selectivity
37 patterns. An aim of this study was to evaluate how to use length composition data from
38 multifleet fisheries to estimate stock status for small scombrids. We recommend that length-
39 based models should be applied to the length data coming from the fleet that targets the broadest
40 range of sizes, including juveniles and adults when the data are available. Even though the
41 results observed here can be applied to other multifleet fisheries, the results show biases under
42 different selectivity assumptions for the simulated albacore population, so further simulation
43 testing for data-poor, multifleet fisheries with variable life history, selectivity, and exploitation
44 patterns should be explored.

45 Small tunas are an important social and economic resource for many coastal
46 communities; however, most of these stocks have not been assessed, and their status is unknown.
47 This work is, therefore, an important first step in developing management plans, particularly as

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the evaluation of uncertainty recognizes where data are needed to identify stock status and reduce risks of overfishing. Little tunny in the Southeast and wahoo in the Northwest are undergoing overfishing, despite the method and M used, which confirms the previous perception of ICCAT. These species have already been assigned priority, given their social-economic importance and also considering that they were two of the top three stocks at risk in the Atlantic Ocean, hence deserving most of the managers' attention (SCRS, 2018). For the Southeast little tunny, life history parameters for the given "stock" are not available, and data used within this study were borrowed from other areas, which may hamper our analysis. This species should be certainly considered as a priority for data collection within the small tuna group within ICCAT. For the Northwest wahoo, although life history parameters are available, length composition is only available for rod and reel, which may not include the majority of size classes. For both species, ICCAT has already developed a research programme to address knowledge gaps regarding size data and biological parameters (both from biological sampling and tagging programmes).

Supplementary information

The Supplementary material available at the *ICESJMS* online version of this paper includes Figures S1–S10 and Tables S1–S3.

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Table 1. OM biological inputs parameters for North Atlantic albacore (ICCAT, 2014).

| Biological information | Symbol | Value |
|--|--------------|-----------------------|
| Maximum age (years) | T_{max} | 15 |
| Length where 50% of the fish are mature (cm) | L_{50} | 90 |
| Length where 95% of the fish are mature (cm) | L_{95} | 100 |
| Length-weight scaling parameter (g) | a | 1.34×10^{-5} |
| Length-weight allometric parameter (g) | b | 3.107 |
| von Bertalanffy Brody growth coefficient | k | 0.209 |
| von Bertalanffy asymptotic length (cm) | L_∞ | 122 |
| Theoretical age at length 0 | t_0 | -1.3 |
| Variability of length at age | CVL_∞ | 0.1 |
| Recruitment deviations | σ_R | 0.4 |
| Steepness | h | 0.9 |

Table 2. Life history parameters used as inputs to assess stock status of small scombrids in the Atlantic Ocean using length-based data limited methods. * M was estimated empirically through different methods. The first quantile, median, and third quantile are presented. For the same species: ^Ainformation borrowed from the Northwest stock, ^Binformation borrowed from the Northeast stock, ^Cinformation borrowed from the Southeast stock, and ^Dinformation borrowed from the Southwest stock. Please provide this table in an editable format, i.e. not as a figure.

| Species | Parameter | Northeast | Southeast | Mediterranean | Northwest | Southwest |
|------------------------------------|-------------------------------|---|---|--|---|-----------|
| <i>Sarda sarda</i> (RON) | L_{∞} (cm) | 73 (Babai et al., 2016) | | 70 (Kahraman et al., 2014) | | |
| | k (years^{-1}) | 0.31 (Babai et al., 2016) | | 0.44 (Kahraman et al., 2014) | | |
| | t_0 (years) | -2.45 (Babai et al., 2016) | | -1.33 (Kahraman et al., 2014) | | |
| | T_{max} (years) | 5 (Babai et al., 2016) | | 5 (Cayré et al., 1993) | | |
| | L_{50} (cm) | 42.6 (Babai et al., 2016) | | 39.9 (Saber et al., 2017) | | |
| | M^* (years^{-1}) | 0.43; 0.78; 1.11 | | 0.60; 0.83; 1.09 | | |
| | WL_a (g) | 5.0×10^6 (Babai et al., 2016) | | 6.3×10^6 (Saber et al., 2017) | | |
| | WL_b | 2.79 (Babai et al., 2016) | | 3.21 (Saber et al., 2017) | | |
| <i>Insufficient length data</i> | | | | | | |
| <i>Euthynnus alleteratus</i> (LTA) | L_{∞} (cm) | 86 (Cabrera et al., 2005) ^A | 86 (Cabrera et al., 2005) ^A | 117 (Iattouer, 2009) | 86 (Cabrera et al., 2005) | |
| | k (years^{-1}) | 0.26 (Cabrera et al., 2005) ^A | 0.26 (Cabrera et al., 2005) ^A | 0.19 (Iattouer, 2009) | 0.26 (Cabrera et al., 2005) | |
| | t_0 (years) | -0.32 (Cabrera et al., 2005) ^A | -0.32 (Cabrera et al., 2005) ^A | -1.13 (Iattouer, 2009) | -0.32 (Cabrera et al., 2005) ^B | |
| | T_{max} (years) | 8 (Cayré and Douf, 1980) | 8 (Cayré and Douf, 1980) | 10 (Iattouer, 2009) | 8 (Cayré and Douf, 1980) ^B | |
| | L_{50} (cm) | 42.0 (Douf, 1980) | 42.0 (Douf, 1980) ^B | 51.1 (El Hajji et al., 2012) | 39.7 (Ramírez-Arredondo, 1993) | |
| | M^* (years^{-1}) | 0.4; 0.53; 0.68 | 0.4; 0.53; 0.68 | 0.29; 0.43; 0.54 | 0.4; 0.53; 0.68 | |
| | WL_a (g) | 1.4×10^5 (Douf, 1980) | 1.4×10^5 (Douf, 1980) ^B | 1.2×10^6 (Saber et al., 2017) | 2.1×10^5 (Ramírez-Arredondo, 1996) | |
| | WL_b | 3.04 (Douf, 1980) | 3.04 (Douf, 1980) ^B | 3.06 (Saber et al., 2017) | 2.96 (Ramírez-Arredondo, 1996) | |
| <i>Insufficient length data</i> | | | | | | |
| <i>Scadichthys stellifer</i> (WHT) | L_{∞} (cm) | 179.7 (Viana et al., 2013) ^A | | | 179.7 (Viana et al., 2013) | |
| | k (years^{-1}) | 0.32 (McBride et al., 2008) ^A | | | 0.32 (McBride et al., 2008) | |
| | t_0 (years) | -1.91 (McBride et al., 2008) ^A | | | -1.91 (McBride et al., 2008) ^B | |
| | T_{max} (years) | 9 (McBride et al., 2008) ^A | | | 9 (McBride et al., 2008) ^B | |
| | L_{50} (cm) | 92.5 (Jenkins and McBride, 2009) ^A | | | 92.5 (Jenkins and McBride, 2009) | |
| | M^* (years^{-1}) | 0.43; 0.49; 0.60 | | | 0.43; 0.49; 0.60 | |
| | WL_a (g) | 2.8×10^4 (Santana et al., 1993) | | | 2.0×10^6 (Beerkircher 2005) | |
| | WL_b | 2.72 (Santana et al., 1993) | | | 3.24 (Beerkircher 2005) | |
| <i>Stock not defined</i> | | | | | | |
| <i>Anisotremus albus</i> (FRU) | L_{∞} (cm) | 52 (Gradushev and Korolevich, 1986) | 52 (Gradushev and Korolevich, 1986) | | | |
| | k (years^{-1}) | 0.32 (Gradushev and Korolevich, 1986) ^C | 0.32 (Gradushev and Korolevich, 1986) | | | |
| | t_0 (years) | -0.83 (Gradushev and Korolevich, 1986) ^C | -0.83 (Gradushev and Korolevich, 1986) ^C | | | |
| | T_{max} (years) | 4 (Gradushev and Korolevich, 1986) ^C | 4 (Gradushev and Korolevich, 1986) ^C | | | |
| | L_{50} (cm) | 30 (Cayré et al., 1993) ^C | 30 (Cayré et al., 1993) ^C | | | |
| | M^* (years^{-1}) | 0.48; 1.01; 1.37 | 0.48; 1.01; 1.37 | | | |
| | WL_a (g) | 8.9×10^4 (Frota et al., 2004) ^D | 8.9×10^4 (Frota et al., 2004) ^D | | | |
| | WL_b | 3.17 (Frota et al., 2004) ^D | 3.17 (Frota et al., 2004) ^D | | | |
| <i>Insufficient length data</i> | | | | | | |
| <i>Insufficient length data</i> | | | | | | |

Table 3. Summary of stock status for small scombrids in the Atlantic Ocean. SPR is shown for both models, LBSPR and LIME, with the lower (CI_low) and upper (CI_up) confidence interval for the base model (median M).

| Species | Stock | LBSPR | | | LIME | | |
|--------------|---------------|-------|--------|-------|------|--------|-------|
| | | SPR | CI_low | CI_up | SPR | CI_low | CI_up |
| Little tunny | Southeast | 0.13 | 0.12 | 0.13 | 0.27 | 0.03 | 0.51 |
| Bonito | Northeast | 0.23 | 0.20 | 0.27 | 0.71 | 0.37 | 1.06 |
| Wahoo | Northwest | 0.37 | 0.23 | 0.51 | 0.29 | 0.27 | 0.31 |
| Wahoo | Northeast | 0.55 | 0.48 | 0.62 | 0.38 | 0.34 | 0.43 |
| Bonito | Mediterranean | 0.59 | 0.47 | 0.70 | 0.22 | 0.15 | 0.28 |
| Little tunny | Mediterranean | 0.66 | 0.57 | 0.74 | 0.62 | 0.54 | 0.71 |
| Little tunny | Northwest | 0.66 | 0.60 | 0.72 | 0.48 | 0.40 | 0.56 |
| Frigate tuna | Southeast | 0.79 | 0.78 | 0.80 | 0.53 | 0.49 | 0.57 |
| Frigate tuna | Northeast | 0.83 | 0.81 | 0.84 | 0.46 | 0.44 | 0.48 |
| Little tunny | Northeast | 0.90 | 0.83 | 0.96 | 1.00 | 1.00 | 1.00 |

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Figure captions

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Figure 1. (a) Selectivity patterns included in the operating model (OM) for each fleet. (b) Catch time-series included in the OM by fleet in the last 15 years (1997–2011) taken from the North Atlantic albacore assessment (ICCAT, 2013). Fleets A, B, and C are combination of fleets described in the main text. **Label to two graphs as (a) and (b) instead of as A and B. On graph (a), rotate the y-axis numbers 90° clockwise. On graph (b), change the y-axis numbers to 10, 20, 30, and 40 and change the caption to “Catch (‘000 t)”.**

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B2 Performance of length-based data-limited methods in a multi-fleet context



Canadian Journal of Fisheries and Aquatic Sciences

Comparing performance of catch-based and length-based stock assessment methods in data-limited fisheries

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Manuscripts

1 Comparing performance of catch-based and length-based stock
2 assessment methods in data-limited fisheries

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13

14 **Abstract**

15 The quantity of data from many small-scale fisheries is insufficient to allow for the
16 application of conventional assessment methods. Even though in many countries they are
17 moving to close-loop simulations to assess the performance of different management
18 procedures in data limited situations, managers in most developing countries are still
19 demanding information on stock status. In this study we use the common metric of harvest
20 rate to evaluate and compare the performance of the following catch-only and length-only
21 assessment models: Catch-Maximum Sustainable Yield (Catch-MSY), Depletion Based
22 Stock Reduction Analysis (DBSRA), Simple Stock Synthesis (SSS), an extension of Catch-
23 MSY (CMSY), Length Based Spawning Potential Ratio (LBSPR), Length-Based Integrated
24 Mixed Effects (LIME), and Length-Based Bayesian (LBB). In general, results were more
25 biased for slightly depleted than for highly depleted stocks, and for long-lived than for short-

26 lived species. Length-based models, such as LIME, performed as well as catch-based
27 methods in many scenarios and, among the catch-base models the one with the best
28 performance was SSS.

29

30 Keywords: data-limited assessment methods, depletion, life-history, harvest rates

31

Draft

32 INTRODUCTION

33 Major commercial fish species usually have substantial sets of data that can be
34 integrated by complex stock assessment models (e.g., Methot and Wetzel 2013); these data
35 may include time series of total removals, catch-at-length or -age, relative or absolute
36 abundance indices, fishing effort, and information on life-history parameters. Such datasets
37 are typically unavailable for most of the small-scale fisheries and by-catch species around the
38 world. Fisheries and stocks lacking comprehensive datasets are commonly known as “data-
39 poor” or “data-limited” fisheries (Costello et al. 2012; Dowling et al. 2015). Recently, many
40 data-limited approaches have been developed to meet an increasing demand for science-
41 based fisheries management of unassessed fisheries where data and resources are limited
42 (Wetzel and Punt 2011; Costello et al. 2012; Dowling et al. 2015, 2016; Chrysafi and
43 Kuparinen 2016; Rosenberg et al. 2017).

44 Assessing stocks using only catch and life-history data started many years ago with
45 the development of Stock Reduction Analysis (SRA; Kimura and Tagart 1982; Kimura et al.
46 1984). Since then, SRA has been extended to estimate productivity and reconstruct historical
47 abundance trends by making assumptions about final biomass relative to unfished or initial
48 biomass (i.e., stock depletion; Thorson and Cope 2015). SRA has been further extended to
49 incorporate stochastic variability in population dynamics (Stochastic-SRA; Walters et al.
50 2006), a flexible shape for the production function (Depletion-Based SRA; Dick and MacCall
51 2011), prior information regarding resilience and population abundance at the start of the
52 catch time series (Catch-Maximum Sustainable Yield, Catch-MSY; Martell and Froese
53 2013), bayesian approaches (CMSY, Froese et al. 2017), and age-structured population
54 dynamics (Simple Stock Synthesis, SSS; Cope 2013). Despite these differences, this family
55 of catch-only models share a common dependence on prior assumptions about final stock
56 depletion. Simulation testing has previously indicated that these methods perform well only

57 when assumptions regarding final relative abundance are met (Wetzel and Punt 2015).
58 Unsurprisingly, because final stock depletion is a prior assumption, the methods perform
59 differently under different stock depletion levels (i.e., highly depleted or slightly depleted
60 stocks, Walters et al. 2006) or under different harvest history or catch trends.

61 For many small-scale fisheries, obtaining reliable time series on historical total catch
62 is difficult, whereas sampling lengths from the catch is easier. Mean-length mortality
63 estimators (Beverton and Holt 1957) assume that fishing mortality directly influences the
64 mean length of the catch under equilibrium conditions. This basic method has been extended
65 by length-based spawning potential ratio (LBSPR, Hordyk et al. 2015a), length-based
66 Integrated Mixed Effects (LIME, Rudd and Thorson 2017) and Length-Based Bayesian
67 approach (LBB, Froese et al. 2018) models, among others. These allow for the estimation of
68 instantaneous fishing mortality (F) and spawning potential ratio (SPR) when basic biological
69 parameters are known. In contrast to LBSPR and LBB, LIME does not assume equilibrium
70 conditions. The mixed-effects aspect of LIME extends length-based methods by estimating
71 changes in recruitment and separating them from fishing mortality over time (Rudd and
72 Thorson 2018).

73 It is good practice to simulation test the performance of assessment methods before
74 applying them in practice (Cope 2008). This can be done using a variety of approaches,
75 though it is most often accomplished using an Operating Model (OM) to generating pseudo-
76 data with error to fit an assessment model (Punt et al. 2016). Simulation can either be an open
77 loop or a closed loop with feedback. Carruthers (2016), using closed loop simulations, found
78 that data-limited methods using observations of stock depletion offer the best overall
79 performance across life history types, data quality and autocorrelation in recruitment strength.
80 However, these management procedures are based on setting catch limits and were designed
81 for use in data-limited fisheries for which annual catch data are available, sometimes together

82 with a relative abundance index (Carruthers et al. 2014). In many data poor fisheries,
83 measuring total removals is difficult, as is enforcing catch limits. Recently, Hordyk et al.
84 (2015b) tested some harvest strategies using a simulation approach to assess the utility of
85 LBSPR as a tool for management in data-limited fisheries using an effort-based harvest
86 control rule. They found that the LBSPR assessment model with an iterative effort-based
87 harvest control rule can be used to rebuild an overfished stock to sustainable levels or fish
88 down a stock to the target SPR. So far, these two approaches, output-based limits versus
89 input/effort-based limits, have not been directly tested.

90 In addition, there are some studies comparing the performance in determining stock
91 status of catch-based (Wetzel and Punt, 2015; Rosenberg et al. 2017) and length-based
92 assessment models (Chong et al. 2019). However, a comparison of the performance of both
93 length-based and catch-based methods to estimate stock status is needed. Unfortunately,
94 finding a common metric between catch-based and length-based stock status metrics is
95 difficult; the former measures overfishing via fishing rates and the latter biomass-based catch
96 limits.

97 Even though many countries are moving to close-loop simulations or Management
98 Strategy Evaluation (MSE) to assess the performance of different management procedures
99 (i.e., the combination of analytical method and control rule) for data-limited fisheries
100 (Harford and Carruthers, 2017), it remains a need to understand how the individual analytical
101 methods perform. Therefore, this study used a simulation approach to better estimate relative
102 bias and precision of a range of data-limited methods so researchers can choose which one
103 could be more appropriate to use for the fishery they need to assess, with the results
104 providing general information needed to construct an appropriate control rule for a given
105 method's performance.

106 We used OMs (here considered just as a simulated population model) to represent
107 the main sources of uncertainty and to generate data for use in data-limited stock assessment
108 methods and to evaluate how well the methods perform when the data and knowledge
109 requirements are met. A common metric is then used for comparison across model types,
110 namely exploitation or harvest rates, to evaluate the performance of both catch-based and
111 length-based models in a simulation context. We evaluate performance considering three fish
112 populations with contrasting life-history strategies, under three different harvest trends, and
113 three different levels of final stock depletion.

114 METHODS

115 Simulation studies commonly make different operating model assumptions from those
116 of the methods being tested to allow the evaluation of robustness; in some cases, however,
117 the same population model is used for both simulation and estimation, i.e. self-testing
118 (Deroba et al., 2015). Using the same model structure for simulation and estimation can result
119 in more optimistic results that might not be true under many scenarios (Francis, 2012) since it
120 is not possible to explore the impact of model assumptions when the model used for
121 simulation and estimation is the same. If a method performs poorly, however, when the
122 assumptions in the OM are the same as the assessment model, it is unlikely to perform well in
123 practice. To evaluate robustness to model structure our approach evaluated multiple data-
124 limited assessment methods with a range of assumptions about population and fishery
125 dynamics using an OM decoupled from the tested methods.

126 Twenty-seven different OMs were created using a factorial design comprising 3
127 harvest rates, 3 life-history types, and 3 depletion scenarios. The different harvest rates
128 scenarios, each considered a 20-year time series of fishing, correspond to fishing mortality
129 histories commonly seen in many fisheries. In harvest rate scenario 1, fishing mortality

130 increases until it reaches a maximum and starts declining afterwards. This is commonly seen
 131 once management measures are implemented to reduce fishing pressure. Harvest rate
 132 scenario 2 assumes that fishing mortality increases and remains constant after reaching a
 133 maximum. This harvest rate profile could result from the implementation of catch or effort
 134 management limits. Harvest rate scenario 3 has constantly increasing fishing mortality, which
 135 would occur for fisheries that are still developing (Figure 1a).

136 Three life history types of varying longevity and somatic growth rate were simulated,
 137 namely (i) a short-lived fast-growing species, Pacific chub mackerel (*Scomber japonicus*), (ii)
 138 a medium-lived medium-growing fish, albacore tuna (*Thunnus alalunga*), and (iii) a longer-
 139 lived slow-growing species, canary rockfish (*Sebastodes pinniger*) (Table 1). Finally, the
 140 following three final depletion levels (D) were considered: (i) heavily fished ($D = 0.2$; *i.e.*
 141 stock biomass in the last year is 20% of virgin biomass), (ii) moderately fished ($D = 0.4$) and
 142 (iii) lightly fished ($D = 0.6$).

143 *Operating model specifications*

144 The OM was developed to simulate resource dynamics under the different fishing
 145 scenarios, life histories and final depletion levels. The OMs consist of an age-structured
 146 population with numbers at age over time modelled as follows:

147

$$148 N_{a,t} = \begin{cases} R_t, & a = 0 \text{ and } t = 0 \\ \frac{N_{a-1,t} e^{(-M - F_t S_{a-1})}}{1 - e^{(-M - F_t S_{a-1})}}, & 0 < a < A \text{ and } t = 1 \\ \frac{N_{a-1,t-1} e^{(-M - F_t S_{a-1})}}{(N_{a-1,t-1} + N_{a,t-1}) e^{(-M - F_{t-1} S_{a-1})}}, & 0 < a < A \text{ and } t > 1 \\ & a = A \text{ and } t > 1 \end{cases}$$

149 R_t is the number of age-0 animals at the start of year t , $N_{a,t}$ is the number of fish of
150 age a at the start of the year t , S_a is the selectivity at age, F_t is the instantaneous fishing
151 mortality rate during year t , M is the instantaneous rate of natural mortality, and A is the age
152 of the plus group. Fishing mortality deviations were included as $F_t \sim \text{lognormal}(F_{t-1}, \sigma_F^2)$. A
153 Beverton–Holt spawner–recruitment function (Beverton and Holt 1957) and annual normally
154 distributed recruitment deviations $N(0, \sigma_R)$ were assumed (Table 1).

155 The biomass in each year t was calculated as $B_t = \sum_{a=1}^A N_{a,t} w_a$ where $w_a = \alpha L_a^\beta$
156 (parameters in Table 1). In addition, the predicted total catch by year (C_t) was calculated as C_t
157 $= \sum_{a=0}^A C_{a,t}$ with:

$$C_{a,t} = \frac{F_t S_a}{M + F_t S_a} N_{a,t} (1 - e^{(-M - F_t S_a)})$$

159 For each OM (N=27), 100 time series of fishing mortality were simulated, and the
160 harvest rate per year (U_t) as C_t/B_t calculated (Figures S1 to S3). Each simulated population
161 began at the unfished biomass level and all fishing trend scenarios terminate at the specified
162 depletion level (Appendix Figures A1 to A27).

163 To condition the OM, published life-history parameter values (Table 1) reported in
164 formal stock assessments were used (Crone and Hill 2015 for the short-lived Pacific chub
165 mackerel; Anon 2014 for the medium-lived albacore tuna; and Thorson and Wetzel 2015 for
166 long-lived canary rockfish). Each population was assumed to be targeted in a single area, by
167 one fleet with a selectivity pattern (Table 1) that was logistic and constant through time.

168 Length bins were defined as they were in their respective assessments; every 2 cm
169 (Crone and Hill 2015; Anon 2014; Thorson and Wetzel 2015). To obtain the catch length
170 frequency, the probability (p) of being in a length bin (j) given age (a) was calculated as:

$$171 \quad p_{j,a} = \begin{cases} \emptyset \left(\frac{j - L_a}{L_a CV_L} \right), & j = 1 \\ \emptyset \left(\frac{j - L_a}{L_a CV_L} \right) - \emptyset \left(\frac{j - 1 - L_a}{L_a CV_L} \right), & 1 < j < J \\ 1 - \emptyset \left(\frac{j - 1 - L_a}{L_a CV_L} \right), & j = J \end{cases}$$

172 With the predicted probability of harvest by length bin being:

$$173 \quad \pi_j = p_{j,a} \frac{\sum_{a=0}^A N_{a,t} S_a}{N_t}$$

174 One thousand fish per year were drawn using a multinomial distribution with a π_j

175 probability (Rudd and Thorson 2018).

176 *Comparing methods outputs*

177 One of the challenges when comparing catch-based and length-based methods is that
 178 they produce different model outputs. Catch-only models calculate total and/or spawning
 179 stock biomass and MSY, whereas length-based models estimate exploitation and transient
 180 SPR, which can be used to infer stock status. These are fundamentally different measures of
 181 the population status. As a result, our performance metric is defined as the error relative (RE)
 182 to the OM, where $RE = (U_{Method} - U_{OM}) / U_{OM}$. This allows for a measure of uncertainty, in
 183 both bias and precision, for all methods under each scenario, and is used as a standardized
 184 metric of model performance. Bias in this study is how far, on average, the performance
 185 measure from each estimation model is from the OM. Imprecision is related to the variability
 186 (variance) around the central tendency.

187 We used U as a common measure for comparisons between each data limited method
 188 and the OM. For catch-only approaches it is defined as the ratio catch/biomass; while for the
 189 length-based models, the estimated F was transformed to an exploitation rate via $U = 1 - exp$

190 (- F). In addition, we present the average RE across the last five years of the time series, not
191 along the entire time series of data, because we are interested in the estimation of the current
192 exploitation rates. Also, multiple studies have shown that catch-based methods might be
193 appropriate to predict sustainable catch at the end of the time series, but not to reconstruct a
194 biomass time series (Carruthers et al. 2012; Wetzel and Punt 2015).

195 *Estimation models*

196 All simulations, and data-limited model calculations, were conducted using the open-
197 source statistical software R (R Core Team 2018). Each catch-based and length-based method
198 evaluated here are summarized below.

199 *Catch-based data-limited methods*

200 **Catch-MSY** (Martell and Froese 2013) is a SRA approach that assume a Schaefer
201 biomass dynamic model. Inputs are a time series of removals, priors for the population rate of
202 increase at low population size (r), carrying capacity (K), and a range of stock depletion in
203 the final year (Table 1). Values of r and K are randomly sampled using a Monte Carlo
204 approach to detect ‘viable’ r - K pairs. A parameter pair is considered ‘viable’ if the
205 corresponding biomass trajectories calculated from a production model are compatible with
206 the observed catches, so that the population abundance never falls below 0, and is compatible
207 with the prior assumption of relative biomass (i.e., stock depletion; Martell and Froese 2013).
208 r - K pairs are drawn from uniform prior distributions and the Bernoulli distribution is used as
209 the likelihood function for accepting each r - K pair. Catch-MSY uses catch and productivity
210 to estimate MSY. Here we use the modified version of Catch-MSY (Rosenberg et al. 2017) to
211 extract biomass trends from all viable r - K pairs using the R package *datalimited* version 0.1.0
212 (Anderson et al. 2016). The biomass trajectory is calculated as the median of all viable
213 biomass trajectories generated under the Monte Carlo process. We decided to include Catch-

214 MSY in this performance comparison analysis since it is a method that has become very
215 popular in developing countries because of its easy implementation through R libraries.

216 **CMSY** (Froese et al. 2017) extends Catch-MSY by using a Monte-Carlo filter
217 (instead of the SIR algorithm) that fixes systematic biases in the Catch-MSY method. It also,
218 explicitly incorporates process error and estimates target reference points (MSY , F_{MSY} , B_{MSY})
219 as well as relative stock size (B/B_{MSY}) and exploitation (F/F_{MSY}) from catch data and priors
220 for r and depletion at the beginning and the end of the time series. CMSY has an inbuilt
221 piecewise "hockey-stick" to prevent over-estimating of rebuilding potential at very low
222 abundance $B < 0.25B_0$. The CMSY package implements a Bayesian state-space
223 implementation of the Schaefer surplus production model (Winker 2019). CMSY was
224 included in this paper because it has been described as an unbiased version of Catch-MSY
225 and it has been already implemented in different parts of Europe (Froese et al. 2018) and
226 explored by Regional Fisheries Management Organizations such as the International
227 Commission for the Conservation of Atlantic Tunas, ICCAT (Winker et al. 2017, ICCAT
228 2017). We included a prior for depletion in this method (see Table 1).

229 **DBSRA** (Dick and MacCall 2011) modifies the SRA approach by using Monte Carlo
230 draws from four parameter distributions (M , F_{MSY}/M , B_{MSY}/B_0 , and depletion) and age at
231 maturity (A_{mat}) to separate the total biomass into immature and mature biomass (fishery
232 selectivity is also assumed to have an identical pattern to the age-at-maturity ogive). It uses a
233 delay-difference production model with a time lag for recruitment and mortality as:

$$B_{t+1} = B_t + P(B_{t-A_{mat}}) - C_t$$

234 where B_t is the biomass at the start of year t , $P(B_{t-A_{mat}})$ is the latent annual production
235 based on a function of adult biomass in year $t-A_{mat}$ and C_t is the catch in year t . Biomass in
236 the first year (B_0) is assumed to be equal to K . The package *fishmethods* version 1.10-3 was
237 used to perform this analysis (Nelson 2017). For DBSRA we used the A_{mat} and M as fixed

239 inputs and three priors: final stock depletion, F_{MSY}/M , and B_{MSY}/B_0 (distributions in Table 1).
240 Each of these is assigned a distribution from which the Monte Carlo draws are taken. We
241 chose DBSRA to be included in our analysis because it is currently used for providing
242 fisheries management advice on the US West Coast.

243 **SSS** is based on the Stock Synthesis age-structured stock assessment model (Methot
244 and Wetzel 2013). SSS fix all parameters in the Stock Synthesis model except for initial
245 recruitment ($\ln R_0$). It also sets up an artificial index of abundance that represents the relative
246 stock biomass. The first value of the index is always 1, and the value in the final year
247 represents the percent of the population left in that year (final depletion). The values of
248 steepness (h) and the final year of the abundance survey are all randomly drawn from a
249 specified distribution using a Monte Carlo approach (Cope 2013) and $\ln R_0$ is then estimated.
250 Benefits of this approach are that it retains the same modelling framework as the data-rich
251 stock assessments, but still allows for flexibility in a variety of parameter and model
252 specifications, if desired. The input priors used for SSS were relative stock status and
253 steepness and selectivity was matched to the OM (Cope 2019). We chose SSS to be included
254 in our analysis because it is gaining use as a flexible platform to incorporate more data as it is
255 collected, as well as providing an age-based alternative to the other catch-only models.

256 *Length-based data-limited methods*

257 SPR is the proportion of the unfished reproductive potential per recruit under a given
258 level of fishing pressure (Goodyear 1993). In **LBSPR**, SPR in an exploited population is
259 calculated as a function of the ratio of fishing mortality to natural mortality (F/M), selectivity,
260 and the two life-history ratios M/k and L_m/L_∞ ; k is the von Bertalanffy growth coefficient, L_m
261 is the size of maturity and L_∞ is asymptotic size (Hordyk et al. 2015a). The inputs of LBSPR
262 are M/k , L_∞ , the variability of length-at-age (CVL_∞), which is normally assumed to be around

263 10%, and the length at maturity specified in terms of L_{50} and L_{95} (the size at which 50% and
264 95% of a population matures). Given the assumed values for M/k and L_∞ and that length
265 composition data come from an exploited stock, the LBSPR model uses maximum likelihood
266 methods to estimate the selectivity ogive, which is assumed to be of a logistic form defined
267 by the selectivity-at-length parameters S_{50} and S_{95} (the size at which 50% and 95% of a
268 population is retained by the fishing gear), and F/M . The selectivity ogive and relative
269 fishing mortality are then used to calculate SPR (Hordyk et al. 2015a, 2015b). Estimates of
270 SPR are primarily determined by the length of fish relative to L_{50} and L_∞ , but it also depends
271 on life history parameters such as fecundity-at-age/length and selectivity. LBSPR is an
272 equilibrium based method with the following assumptions: (i) asymptotic selectivity, (ii)
273 growth is adequately described by the von Bertalanffy equation, (iii) a single growth curve
274 can be used to describe both sexes which have equal catchability, (iv) length at-age is
275 normally distributed, (v) rates of natural mortality are constant across adult age classes, (vi)
276 recruitment is constant over time, and (vii) growth rates remain constant across the cohorts
277 within a stock (Hordyk et al. 2015a). Analyses were conducted using the LBSPR package,
278 version 0.1.3 in R (Hordyk 2018). We used the Rauch-Tung-Striebel smoother function
279 available in the LBSPR package to smooth out the multi-year estimates of F . We decided to
280 include LBSPR in the analysis as its easy implementation through R libraries and online apps
281 have made it a popular, widespread choice.

282 **LIME** uses length composition of the catch and biological information to estimate F
283 and SPR. LIME has the same data requirements as LBSPR, but does not assume equilibrium
284 conditions. The mixed effects aspect of LIME extends length-based methods by estimating
285 changes in recruitment and fishing mortality separately over time (Rudd and Thorson 2018).
286 LIME uses automatic differentiation and Laplace approximations as implemented in
287 Template Model Builder (TMB; Kristensen et al. 2016) to calculate the marginal likelihood

288 for the mixed-effects. All other assumptions are the same as LBSPR but LIME estimates one
289 selectivity curve for the entire time series of length data while LBSPR estimates one
290 selectivity curve for each year since each time step estimation in LBSPR is independent
291 (Hordyk et al. 2015a). The inputs to LIME (Rudd 2019) are: M , k , L_∞ , t_0 , CVL_∞ , L_{50} , L_{95} , h ,
292 and the parameters of the length-weight relationship a and b (Table 1). We decided to include
293 LIME in our analysis because it is the only length-based method that allows for recruitment
294 deviations and it has not been widely tested.

295 **LBB** is a simple and fast method for estimating relative stock size that uses a
296 Bayesian Monte Carlo Markov Chain (MCMC) approach (Froese et al. 2018). In contrast to
297 other length-based methods, LBB uses pre-specified priors on parameters, and thus,
298 technically does not require further inputs in addition to length frequency data, if the user is
299 willing to accept the life history defaults. However, it provides the user the option to specify
300 priors for the inputs L_∞ , length at first capture (L_c), and relative natural mortality (M/k). We
301 specified the true M/k value and we let the model calculate L_c (length where 50% of the
302 individuals are retained by the gear) and L_∞ , which is approximated by the maximum
303 observed length L_{max} . In addition, F/M is estimated as means over the age range represented
304 in the length-frequency sample. We decided to include LBB because it has not been widely
305 tested and compared with other length-based assessments methods. As well as CMSY, LBB
306 is gaining consideration as a plausible method in some international commissions such as
307 ICCAT (Anon, 2019).

308 During our simulation testing, we assumed that length models had the correct values
309 for the von Bertalanffy length-at-age relationship, L_∞ , k and t_0 , length-weight parameters α
310 and β , M , and the parameters L_{50} and L_{95} from a logistic maturity-at-length curve. We did not
311 evaluate misspecifications in life history parameters inputs. The objective is only to
312 compare the performance of different data-limited methods under different scenarios, not to

313 evaluate each of them under different parameters misspecifications, which has already been
314 done in the original publications of each method.

315 RESULTS

316 Large variability is seen in model performance across harvest scenarios, life-history
317 types, and depletion levels (Figure 1b-d, Figure 2). Figures 1 and 2 can be used to search for
318 the best model, i.e. to identify the least biased and least imprecise method based on the life
319 history of the species, harvest trends, and knowledge of final depletion. The imprecision of
320 each method is shown in the variability around those estimations in Figure 1 and the specific
321 values in Figure 2. A robust method would show low bias and high precision for all life
322 history and harvest scenarios; for example, among all catch-based methods, SSS appears to
323 be the most robust method (Figure 2).

324 In general, catch-based methods tended to be more biased at relatively higher stock
325 sizes ($D=0.6$) and for long-lived species (Figures 1 and 2). Catch-MSY tended to perform
326 poorly across a range of scenarios. Among the length-based models, all were less biased and
327 more precise for the medium-lived species (Figure 2) without a clear different performance
328 among harvest scenarios. The most biased length-based method was LBSPR although more
329 precise than LIME and LBB (Figure 2). Overall, LIME and SSS were the most robust
330 methods.

331 *Catch-only methods*

332 It was to be expected that the various catch-based models considered in this study
333 would perform differently among them because they have different model structure and
334 assumptions. SSS performed best in most cases estimating unbiased exploitation rates across
335 different scenarios of harvest trends, final stock depletion and life histories (Table 2).
336 However, it tended to underestimate harvest rates by 30% for medium and short-lived species

337 at relatively high stock sizes ($D=0.6$). For long-lived species the estimations were slightly
338 overestimated for moderately fished stock sizes ($D=0.4$), but SSS was always the model that
339 was least biased. DBSRA was the most precise but it underestimated harvest rates in general,
340 and in contrast to the other catch-based methods, it was less biased for stocks at relatively
341 high stock sizes ($D=0.6$, Figure 2). Catch-MSY was the most biased of the catch-based
342 models tested, overestimating harvest rates in particular for stocks lightly fished ($D=0.6$). It
343 was less biased to itself for medium-lived species and produced non-biased estimates of MSY
344 for highly depleted ($D=0.2$) medium-lived or short-lived species at moderate stock sizes
345 ($D=0.4$) (Table A1). CMSY was less biased in the relatively medium to high stock sizes
346 ($D=0.4$ and $D=0.6$) and for medium-lived species (Table 2). In general, catch-based models
347 were less biased and more precise when stocks were at relatively low stock sizes (i.e. using a
348 prior centered around 0.2, Figure 1). In general, estimation models in harvest scenario 1
349 (ramp shaped) produced the most variable estimations in harvest rates (Figure 1, Table 2).
350 The catch-based methods performed the best for the medium-lived species.

351

352 *Length-based methods*

353 In many cases, length-based models gave a less biased estimation of U than catch-
354 based models (Figure 1). LIME was the least biased length-based method (Table 3).
355 However, in general, LIME did not converge around 10% of the time. The three length-based
356 methods used here (LIME, LBSPR and LBB) produced more variable estimations in harvest
357 scenario 1, where the fishing intensity decreased at the end of the time series (Figure 1) but
358 no clear pattern was observed in terms of harvest scenarios in bias (Figure 2). LBSPR in
359 general underestimated harvest rates. Compared to itself, LBB performed better in scenarios
360 where the stocks have relatively low to medium stock sizes ($D=0.2$ and $D=0.4$) and for
361 medium-lived species (Figure 2). Both, LBB and LIME were highly variable for long-lived

362 species and harvest scenario 1 (Figure 1). Multiple modes for harvest scenario 1 may suggest
363 poor convergence. Overall, length-based models were less biased for the medium-lived
364 species.

365 *Short-lived species*

366 For the short-lived life-history strategy, SSS was the least biased and most precise
367 among the catch-based methods. The second most precise method was CMSY, but this
368 method was positively biased for lightly fished or relatively high stock sizes ($D=0.6$). Among
369 the three length-based models, LIME had better overall performance than LBSPR and LBB.
370 However, LBSPR was less biased in harvest scenario 1. LBB was the one model that
371 presented the most variability in harvest-rate estimations for the short-lived species.

372 *Medium-lived species*

373 Both catch-based and length-based models gained more precision as the stocks were
374 more depleted. Among the catch-based methods, CMSY was the less biased followed by
375 SSS, although SSS was more precise. DBSRA was highly negatively biased particularly for
376 relatively low stock sizes ($D=0.2$). Catch-MSY was highly positively biased but less so for
377 highly depleted stocks (Figure 1c). Between the length-based models, LIME showed very
378 good performance with regards to bias for harvest scenarios 2 and 3 under the three depletion
379 levels, although it was slightly biased for harvest scenario 1. LBSPR and LBB showed
380 similar performance but LBSPR was more precise (Figure 1c).

381 *Long-lived species*

382 Both catch-based and length-based methods were less precise (more variability in RE)
383 as the long-lived stocks had relatively higher stock sizes ($D=0.6$), as they did for the other
384 life-history strategies. Among the catch-based methods, SSS was the most precise and least

385 biased method except in harvest scenario 1 and D=0.2 where CMSY was the least biased.
386 SSS underestimated harvest rates in scenarios 2 and 3, where the catch history is constant or
387 increases at the end of the time series. Among assessment methods, a less biased and more
388 precise estimation was observed for medium depleted stocks (D=0.4) than for other depletion
389 levels. LBSPR and DBSRA was negatively biased in all cases (Figure 1d). LIME was highly
390 imprecise in harvest scenario 1, but the least biased among the other length-based methods
391 (Figure 1c, Figure 2).

392

393 DISCUSSION

394 It is to be expected that the various methods would perform differently. Rosenberg et
395 al. (2017) used four catch-based data-limited models and found that models frequently
396 disagreed about population status estimations, with no model showing overall good
397 performance, i.e. high precision and low bias across all case studies. When scenarios
398 represent specific resource dynamics or particular stocks or fisheries, it may be difficult to
399 draw any overall conclusions. Therefore, we chose scenarios that represented different
400 fishing intensity trends, depletion levels, and life histories. We found that model performance
401 is highly dependent on all these factors. More imprecision occurred where fishing pressure
402 decreases at the end of the time series (harvest scenario 1), in comparison with harvest
403 scenarios where F was either stable or increasing (harvest scenarios 2 and 3) that performed
404 similarly in terms of bias and imprecision. In addition, most bias was found when relative
405 stock abundance is high and/or for the slow-grow long-lived life history.

406 In particular, catch-based models performed better (i.e. were less biased and more
407 precise) for stocks that were medium to highly depleted than for lightly depleted stocks.
408 Walters et al. (2006) suggested that for SRA, stocks that have experienced extensive

409 historical depletion gain precision due to a high rate of rejected parameter draws. The SSS
410 approach, which tracks age-structure population dynamics, performed better than the models
411 that are based on lumped biomass production functions such as Catch-MSY, CMSY, and
412 DBSRA, even when priors for depletion were centered on the true values for all methods.
413 Although SSS seems to be the least biased catch-based model, unlike other catch-based
414 models, more detailed life-history information (e.g., age and growth estimates) are required
415 by SSS to define age structure and remove catch according to age-/size-based selectivity
416 patterns (Cope, 2013). So, when this information is highly uncertain, sensitivity analysis to
417 parameter misspecifications is strongly recommended.

418 Catch-MSY performed poorly in all scenarios, overestimating harvest rates even
419 when given a prior for depletion close to the true value. A key point of Catch-MSY is the
420 ability to define a reasonable prior range for the parameters of the Schaefer model, in
421 particular K . In our case, we have arbitrarily chosen 100 times the maximum catch as the
422 upper bound for K based on Martell and Froese (2013). Other priors for K could be explored
423 to see if this improves the outcome, but it remains a difficult parameter to specify under most
424 conditions.

425 CMSY on the other hand performed particularly well with respect to bias and
426 precision for medium-lived species, even better than SSS for medium to low depleted stocks
427 (Figure 2). Also, CMSY was more accurate than the original Catch-MSY method (Martell
428 and Froese, 2013). The difference is that Catch-MSY was designed to select the most
429 probable r - K pair as the geometric mean of this distribution, but CMSY searches not in the
430 center of the distribution but rather near the right tip of viable pairs. According to Froese et
431 al. (2017) since r is defined as the maximum net productivity, the right tip of the distribution
432 of r - K pairs is where these parameters should be found. So, between Catch-MSY and CMSY,
433 CMSY is preferred, though its overall poor performance is noted.

434 Hordyk et al. (2015a) explained how LBSPR relies on detecting the signal of fishing
435 mortality in the right-hand side of the length composition. Consequently, fishing is not likely
436 to have a visible impact on the length composition until fishing mortality is very high and
437 stocks are highly depleted. This is why LBSPR was less biased for more depleted populations
438 and in fishing scenario 1.

439 Our study found that LIME was highly imprecise for long-lived species. Rudd and
440 Thorson (2017) also showed that LIME is more imprecise for long lived species. The model
441 is trying to track cohorts through the length data to estimate recruitment deviations and this is
442 likely difficult for long-lived species when time series of length data are short or much of the
443 population is found near the asymptotic size (Rudd and Thorson 2017).

444 Hordyk et al. (2019) suggested that LBB has not been sufficiently simulated tested
445 and it can produce biased estimates of fishing mortality. We found that LBB was the most
446 biased and imprecise length-based method, although for the less depleted stocks it generally
447 performed better than LBSPR. One of criticisms of LBB by Hordyk et al. (2019) is that it *a*
448 *priori* assumes that $M/k = 1.5$, however here, we specified the true M/k value. So, the main
449 bias was associated with the estimations of L_∞ due to the approximation using maximum
450 observed length L_{max} . There are many reasons why the maximum length may be a biased
451 measure of L_∞ . In addition to L_∞ , L_c was always overestimated by the LBB assessment model
452 (see Table A2), likely due to the bias in the L_∞ estimates.

453 In general, all catch-based and length-based methods seem to perform worse for long-
454 lived life-history types, where there is likely to be less contrast in the dynamics over time
455 (few years relative to their life span), than for medium and short-lived. The length of the time
456 series for the long-lived canary rockfish is therefore probably too short in comparison to the
457 age they reach the maximum length (64 years), to capture the true dynamics of the population
458 and the response to different harvest rates.

459 The present study did not look at parameter misspecification but correctly specified
460 (i.e., unbiased values) the life-history parameters. As we mentioned before, parameters
461 misspecification testing was performed in the original publications for each method. With
462 accurate prior information, length-based models showed better performance in many cases
463 than some catch-based models, as the latter were more sensitive to the catch history scenarios
464 and depletion levels.

465 We note that bias is not symmetrically interpretable, as it could, depending on
466 management objectives, be better or worse to be biased high or low. In data-limited
467 management, maximum sustainable yield is often hard to pinpoint, thus a more precautionary
468 pretty good yield (Hilborn 2010) is likely a better target. When measuring exploitation rates,
469 this would translate to being biased high as a preferable way of being wrong rather than
470 underestimating the exploitation rate and possibly increasing the fishing rate above the true
471 target. The consideration of asymmetric management significance when measuring bias is an
472 important consideration when using the results presented here to develop control rules.

473 Bias and precision are both important factors to consider when assessing fish stocks.
474 Bias reflects how close an estimate is to a known value; precision reflects reproducibility of
475 the estimate. For example, if an assessment is to be re-conducted every year to monitor the
476 impact of a management measure, a precise but biased method would be able to detect a trend
477 better than an unbiased but imprecise method. Like a scientific instrument, this trade-off
478 requires calibration to correct for the bias, and such calibration can be explored using closed-
479 loop simulations such as management strategy evaluation, where the choice of parameters
480 and reference points in a management procedure are tuned (i.e. calibrated) to meet the desired
481 management objectives as represented by the operating model. Thus, a biased method (e.g.,
482 DBSRA) may be preferable to one that is less biased, but more imprecise (e.g., LIME).
483 Alternatively, imprecision can be addressed through the choice of the percentile (e.g., median

484 being the 50% percentile value) for the derived model output used by management (e.g.,
485 catch or SPR); assuming that the true value is contained within the parameter distribution.
486 For example, instead of taking the median value, one could instead use the derived model
487 output associated with the 40th percentile to incorporate risk tolerance as reflected in the
488 calculated imprecision. Such an approach (Ralston et al. 2011) is used in fisheries
489 management systems to directly incorporate scientific uncertainty (both bias and
490 imprecision), and can also be explored and tuned using MSE.

491 *Recommendations*

492 To provide estimates of stock status for unassessed fisheries where data are limited,
493 but reconstructing time series of catch is possible, SSS is recommended. The performance of
494 SSS hinges on the correct specification of the input parameters such as stock depletion,
495 productivity, maturity, and growth parameters. Knowledge about some of these parameters,
496 especially stock depletion, is likely to be poor for data-limited stocks, resulting in
497 misspecification of these parameters. Meta-analyses may offer some starting values for life
498 history parameters (e.g., Myers 2001, Thorson et al. 2012, and Zhou et al. 2012, 2017),
499 however other inputs remain difficult to specify (Chrysafi and Cope *in press*).

500 For fisheries where the time series of catch are unavailable or catches are not
501 consistently monitored and managed, using length-composition data can provide good
502 approximations of the status of the stock, in particular for medium-lived species. It has been
503 shown here that in some cases, length-based models such as LIME can provide the same or
504 less biased estimates of exploitation status than catch-based models. However, growth
505 parameters are even more important for length-based than for any catch-based method, so it
506 is important to have good estimates of those parameters before using any length-based
507 assessment.

508 Making recommendations on which models should be applied to estimate exploitation
509 intensity in different fisheries is challenging because model choice is dependent on data
510 availability, trends in fishing intensity, and the biology of the species. If possible, simulation
511 studies testing different data-limited methods with OMs based on the focus species and the
512 dynamics of the fishery can greatly inform which method is most appropriate. Likewise,
513 decision support tools such as FishPath (Dowling et al. 2015) can also help one weigh the
514 input requirements and assumptions to identify the most appropriate methods given data and
515 life history. Based on the OMs used in this study, we conclude that when only catch data is
516 available, SSS should be considered. When only length data is available, LIME may be less
517 biased than LBSPR and LBB if recruitment variability is an important consideration.
518 However, Pons et al. (2019) found that neither LBSPR or LIME are good in all situations,
519 and thus, both should be considered and compared for inconsistent results. LIME sometimes
520 undergoes convergence issues and has difficulties separating changes in recruitment from
521 changes in fishing mortality (Pons 2018; Pons et al. 2019).

522 For long-lived species it is necessary to have longer time series of data to draw more
523 conclusions. However, Pons (2018) recommended SSS and LBSPR when long time series
524 (i.e. 80 years) of data are available for a species that lives more than 60 years to evaluate
525 changes in fishing intensity.

526 If both catch and length data are available, models that integrate both data types
527 should be considered. LIME, although primarily length based, allows for the inclusion of
528 catch data as well as an index of abundance if one is available. Moreover, integrated
529 assessment models (that use catch as well as length information) like Stock Synthesis could
530 also be considered (Methot and Wetzel 2013). Length information can therefore be added to
531 the SSS data file, with the possibility of freeing up the stock status assumption input, and

532 running the model more like a traditional statistical-catch-at-age model (Cope 2013; Thorson
533 and Cope 2015).

534 For the scenarios analyzed here, including the specific life-histories considered, we do
535 not recommend Catch-MSY for estimating exploitation rates, even with a good estimate of
536 stock depletion. This method will however, as it was originally created to do, produce
537 unbiased estimates of MSY, in particular for short and medium-lived and highly depleted
538 species (Table A1).

539 *Future directions*

540 Dowling et al. (2019) in a review of data limited methods, noted the dangers in the
541 indiscriminate use of generic methods (i.e., methods that are either imported as solutions or
542 considered just because no others are understood, but are not rigorously considered for
543 appropriateness given data and assumptions) and recommended obtaining better data, using
544 care in acknowledging and interpreting uncertainties, developing harvest strategies (including
545 control rules) that are robust to these higher levels of uncertainty and tailoring them to the
546 species and fisheries specific data and context. Management actions to regulate fishing can be
547 based on changes in harvest rates, mainly when catch limits are not an option. The steps, for
548 each fishery, would include i) developing reference points, ii) identifying monitoring and
549 assessments options, and iii) a qualitative and possibly quantitative (e.g., closed-loop
550 simulation) evaluation of how to adjust harvest strategies and paired control rules to meet
551 management objectives (Dowling et al. 2008). Control rules linked to methods explored in
552 this work can be tailored and tested based on the bias and imprecision found in this study.

553 In addition, these data-limited methods could be tested using a fully-specified MSE
554 with stakeholder input to specify the management objectives in order to determine
555 management procedures to help ensure robust and sustainable fisheries management. This

556 evaluation includes the benefits of adaptive improvement of the harvest strategy and
557 management of the fishery.

558 This study provides steps to the above closed-loop simulation by way of conditioning
559 OMs and generating pseudo data for use by the management procedure. Having a common
560 metric (exploitation rate) to compare methods that are often decoupled in performance testing
561 allows for a comparison of the components of uncertainty, bias and imprecision.
562 Understanding the comparative degree each of these methods express component uncertainty
563 under control simulation testing benefits the next steps of closed-loop performance testing if
564 management procedures. The importance of considering assessment methods as part of a
565 management procedure is that a method that provides biased, yet precise results could, if
566 calibrated correctly, provide useful advice, whereas unbiased estimates with high imprecision
567 would need specific consideration on how to manage the risk inherent in imprecision.

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578

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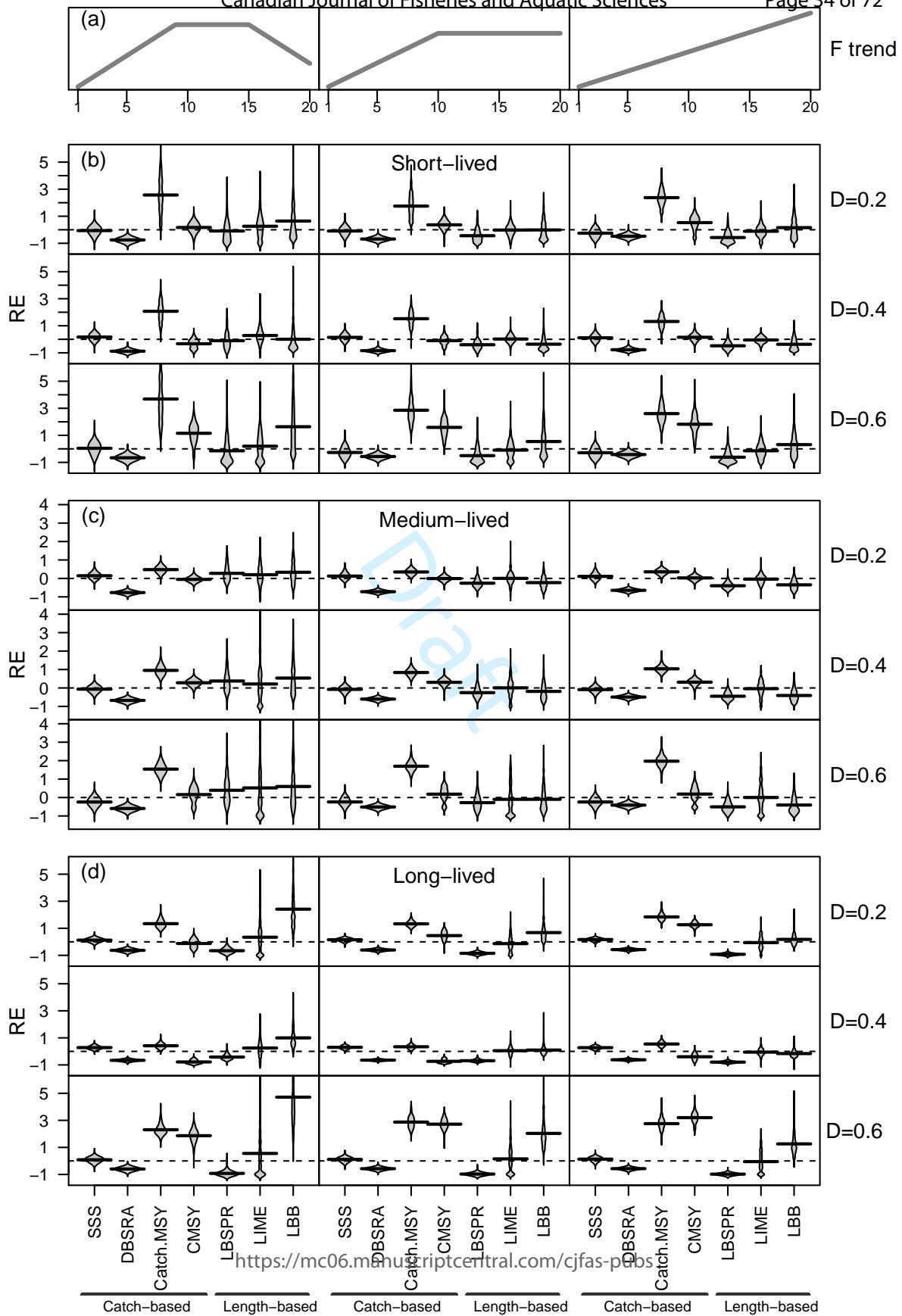
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- 1 Table 1. Life-history information and priors for the three species used in the study. Notation is as follows: *Lognormal* (μ , σ^2); Uniform $U(a, b)$.
- 2 Priors for K were Uniform between the maximum catch in the time series and 100 times the maximum catch. For the Catch-MSY method, the
- 3 depletion priors were Uniform centered on the true value with a minimum of *true* - 0.1 and a maximum of *true* + 0.1.

| Operating model inputs | Symbol | Short-lived | Medium-lived | Long-lived |
|--|---------------|--|--|--|
| Maximum age | Age_{max} | 10 | 15 | 64 |
| Age at 50% maturity (years) | A_{mat} | 3 | 5 | 16 |
| Length where 50% of the fish are mature (FL cm) | L_{50} | 29 | 90 | 55 |
| Length where 95% of the fish are mature (FL cm) | L_{95} | 34 | 100 | 57 |
| Length-weight scaling parameter | α | 2.73×10^{-6} | 1.34×10^{-5} | 1.80×10^{-5} |
| Length-weight allometric parameter | β | 3.444 | 3.107 | 3.094 |
| Von Bertalanffy Brody growth coefficient (1/years) | k | 0.40 | 0.21 | 0.14 |
| Von Bertalanffy asymptotic length (cm) | L_∞ | 38.2 | 122.2 | 60.0 |
| Theoretical age at length=0 | t_0 | -0.6 | -1.3 | -1.9 |
| Coefficient of variation of length at age for all ages | CVL | 0.1 | 0.1 | 0.1 |
| Natural mortality (1/years) | M | 0.60 | 0.30 | 0.05 |
| Steepness | h | 0.5 | 0.9 | 0.8 |
| Selectivity at 50% (cm) | S_{50} | 25 | 60 | 45 |
| Selectivity at 95% (cm) | S_{95} | 30 | 75 | 50 |
| Recruitment deviations | σ_R | 0.3 | 0.4 | 0.5 |
| Fishing mortality deviations | σ_F | 0.2 | 0.2 | 0.2 |
| Estimation models prior distributions | | | | |
| Depletion (used for all catch-based models) | XB_0 | <i>Lognormal</i> (true, 0.1) $U(\max(\text{catch}), \max(\text{catch}) \times 100)$ | <i>Lognormal</i> (true, 0.1) $U(\max(\text{catch}), \max(\text{catch}) \times 100)$ | <i>Lognormal</i> (true, 0.1) $U(\max(\text{catch}), \max(\text{catch}) \times 100)$ |
| Carrying capacity (used for Catch-MSY, CMSY and SSCOM) | K | | | |
| Population rate of increase (used for Catch-MSY, CMSY and SSCOM) | r | $U(0.8, 1.2)$ | $U(0.2, 0.6)$ | $U(0.05, 0.4)$ |
| Steepness (used for SSS) | h | <i>Normal</i> (0.5, 0.1) | <i>Normal</i> (0.9, 0.1) | <i>Normal</i> (0.8, 0.1) |
| Vulnerability (used for DBSRA) | F_{MSY}/M | $U(0, 2)$ | $U(0, 2)$ | $U(0, 2)$ |
| Compensation (used for DBSRA) | B_{MSY}/B_0 | $U(0, 1)$ | $U(0, 1)$ | $U(0, 1)$ |

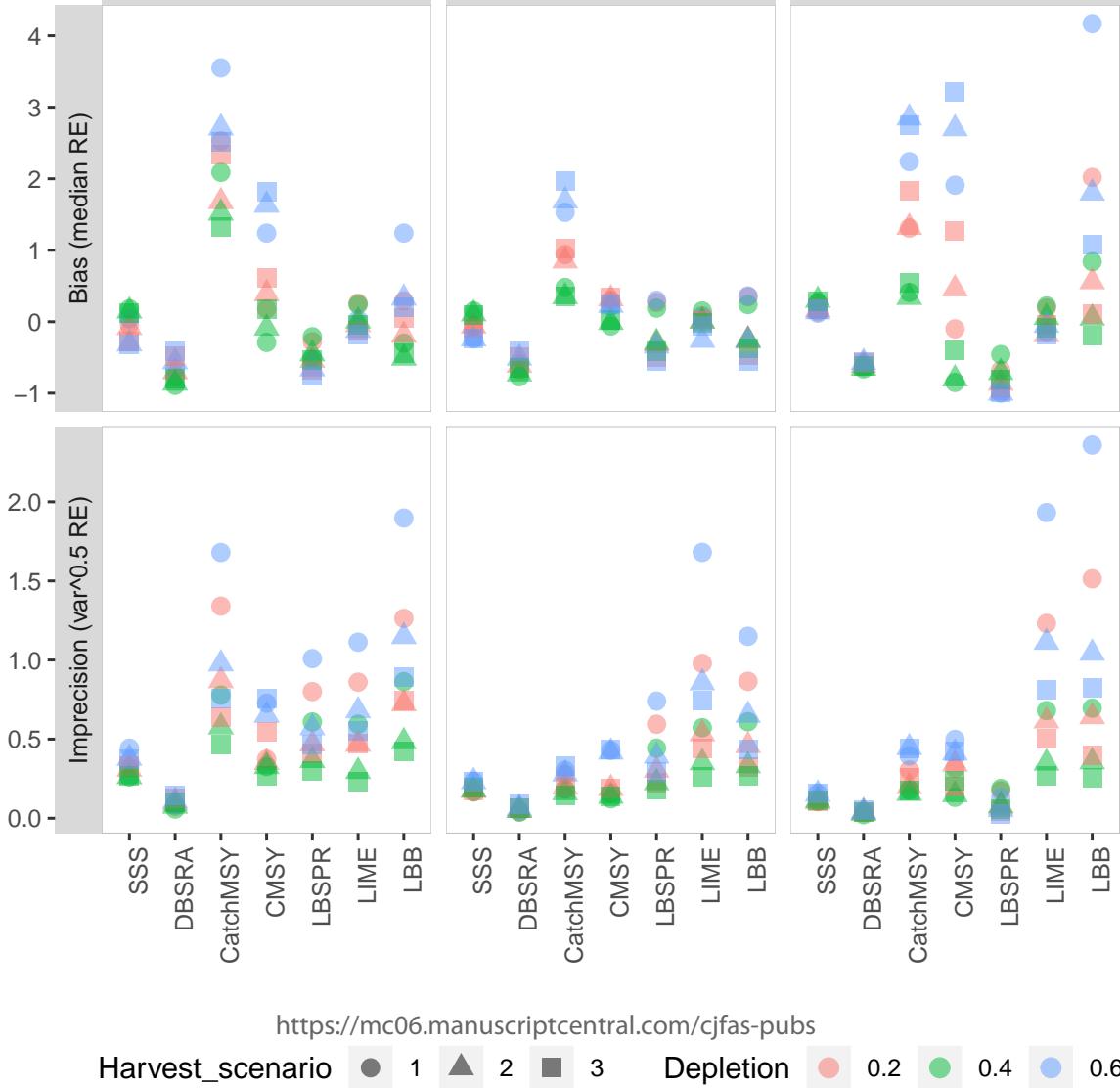


Table 2 Bias in performance for the catch-based models. The values in between brackets are the

| Scenarios | Life-history | Harvest trend | Final depletion | SSS (RE = 0.17) | DBSRA (RE = 0.63) |
|-----------|--------------------------|------------------------|-----------------|--------------------|----------------------|
| 1 | Short-lived (RE = 1.15) | Scenario 1 (RE = 1.53) | 0.2 | -0.06 | -0.75 |
| 2 | | | 0.4 | 0.17 | -0.88 |
| 3 | | | 0.6 | 0.04 | -0.66 |
| 4 | | Scenario 2 (RE = 0.97) | 0.2 | -0.09 | -0.68 |
| 5 | | | 0.4 | 0.14 | -0.84 |
| 6 | | | 0.6 | -0.27 | -0.57 |
| 7 | | Scenario 3 (RE = 0.95) | 0.2 | -0.25 | -0.48 |
| 8 | | | 0.4 | 0.11 | -0.77 |
| 9 | | | 0.6 | -0.29 | -0.42 |
| 10 | Medium-lived (RE = 0.50) | Scenario 1 (RE = 0.57) | 0.2 | -0.06 | -0.67 |
| 11 | | | 0.4 | 0.15 | -0.77 |
| 12 | | | 0.6 | -0.24 | -0.59 |
| 13 | | Scenario 2 (RE = 0.46) | 0.2 | -0.08 | -0.60 |
| 14 | | | 0.4 | 0.13 | -0.72 |
| 15 | | | 0.6 | -0.24 | -0.51 |
| 16 | | Scenario 3 (RE = 0.47) | 0.2 | -0.09 | -0.49 |
| 17 | | | 0.4 | 0.11 | -0.65 |
| 18 | | | 0.6 | -0.24 | -0.41 |
| 19 | Long-lived (RE = 1.33) | Scenario 1 (RE = 1.10) | 0.2 | 0.13 | -0.63 |
| 20 | | | 0.4 | 0.28 | -0.66 |
| 21 | | | 0.6 | 0.09 | -0.60 |
| 22 | | Scenario 2 (RE = 1.44) | 0.2 | 0.15 | -0.60 |
| 23 | | | 0.4 | 0.30 | -0.65 |
| 24 | | | 0.6 | 0.11 | -0.57 |
| 25 | | Scenario 3 (RE = 1.43) | 0.2 | 0.17 | -0.58 |
| 26 | | | 0.4 | 0.28 | -0.62 |
| 27 | | | 0.6 | 0.13 | -0.57 |

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| mean absolute relative error per factor: harvest rates scenarios, species and final de | | |
|--|---------------------|---------------------|
| Catch-MSY (RE = 1.62) | CMSY (RE = 0.71) | Absolute mean RE |
| 2.57 | 0.17 | 1.36 |
| 2.07 | -0.33 | 1.05 |
| 3.69 | 1.15 | 2.19 |
| 1.75 | 0.36 | 0.79 |
| 1.52 | -0.10 | 0.60 |
| 2.85 | 1.59 | 1.52 |
| 2.37 | 0.53 | 0.95 |
| 1.32 | 0.15 | 0.52 |
| 2.61 | 1.82 | 1.37 |
| 0.95 | 0.28 | 0.47 |
| 0.48 | -0.05 | 0.38 |
| 1.54 | 0.16 | 0.86 |
| 0.84 | 0.31 | 0.41 |
| 0.35 | -0.01 | 0.36 |
| 1.70 | 0.18 | 0.62 |
| 1.04 | 0.32 | 0.45 |
| 0.36 | 0.03 | 0.35 |
| 1.97 | 0.19 | 0.60 |
| 1.34 | -0.12 | 0.72 |
| 0.41 | -0.78 | 0.48 |
| 2.31 | 1.87 | 2.11 |
| 1.33 | 0.46 | 0.85 |
| 0.34 | -0.74 | 0.46 |
| 2.87 | 2.72 | 3.03 |
| 1.84 | 1.27 | 1.60 |
| 0.55 | -0.40 | 0.38 |
| 2.76 | 3.20 | 2.31 |

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Table 2 Bias in performance for the length-based methods. The values in between brackets are t

| Scenarios | Life-history | Harvest trend | Final depletion | LBSPR (RE = 0.51) | LIME (RE = 0.15) |
|-----------|--------------------------|------------------------|-----------------|----------------------|---------------------|
| 1 | Short-lived (RE = 0.80) | Scenario 1 (RE = 0.39) | 0.2 | -0.42 | 0.25 |
| 2 | | | 0.4 | -0.69 | 0.05 |
| 3 | | | 0.6 | -0.80 | -0.06 |
| 4 | | Scenario 2 (RE = 0.69) | 0.2 | -0.66 | 0.34 |
| 5 | | | 0.4 | -0.85 | -0.13 |
| 6 | | | 0.6 | -0.92 | -0.06 |
| 7 | | Scenario 3 (RE = 1.29) | 0.2 | -0.92 | 0.55 |
| 8 | | | 0.4 | -0.97 | 0.15 |
| 9 | | | 0.6 | -0.98 | -0.05 |
| 10 | Medium-lived (RE = 0.31) | Scenario 1 (RE = 0.23) | 0.2 | -0.10 | 0.28 |
| 11 | | | 0.4 | -0.40 | 0.02 |
| 12 | | | 0.6 | -0.48 | -0.05 |
| 13 | | Scenario 2 (RE = 0.26) | 0.2 | -0.09 | 0.26 |
| 14 | | | 0.4 | -0.44 | -0.02 |
| 15 | | | 0.6 | -0.57 | -0.11 |
| 16 | | Scenario 3 (RE = 0.46) | 0.2 | -0.14 | 0.20 |
| 17 | | | 0.4 | -0.50 | -0.09 |
| 18 | | | 0.6 | -0.62 | -0.14 |
| 19 | Long-lived (RE = 0.28) | Scenario 1 (RE = 0.23) | 0.2 | 0.27 | 0.20 |
| 20 | | | 0.4 | -0.26 | 0.01 |
| 21 | | | 0.6 | -0.40 | -0.04 |
| 22 | | Scenario 2 (RE = 0.27) | 0.2 | 0.38 | 0.22 |
| 23 | | | 0.4 | -0.25 | 0.01 |
| 24 | | | 0.6 | -0.45 | -0.04 |
| 25 | | Scenario 3 (RE = 0.32) | 0.2 | 0.39 | 0.52 |
| 26 | | | 0.4 | -0.28 | -0.01 |
| 27 | | | 0.6 | -0.51 | 0.01 |

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the mean absolute relativ

| LBB (RE = 0.73) | Absolute mean RE |
|--------------------|---------------------|
| 1.00 | 0.56 |
| 0.10 | 0.28 |
| -0.16 | 0.34 |
| 2.41 | 1.14 |
| 0.68 | 0.56 |
| 0.18 | 0.39 |
| 4.71 | 2.06 |
| 2.04 | 1.05 |
| 1.26 | 0.76 |
| 0.01 | 0.13 |
| -0.36 | 0.26 |
| -0.37 | 0.30 |
| 0.64 | 0.33 |
| -0.01 | 0.16 |
| 0.16 | 0.28 |
| 1.64 | 0.66 |
| 0.54 | 0.38 |
| 0.31 | 0.36 |
| 0.33 | 0.27 |
| -0.23 | 0.17 |
| -0.35 | 0.26 |
| 0.54 | 0.38 |
| -0.19 | 0.15 |
| -0.41 | 0.30 |
| 0.60 | 0.50 |
| -0.10 | 0.13 |
| -0.41 | 0.31 |

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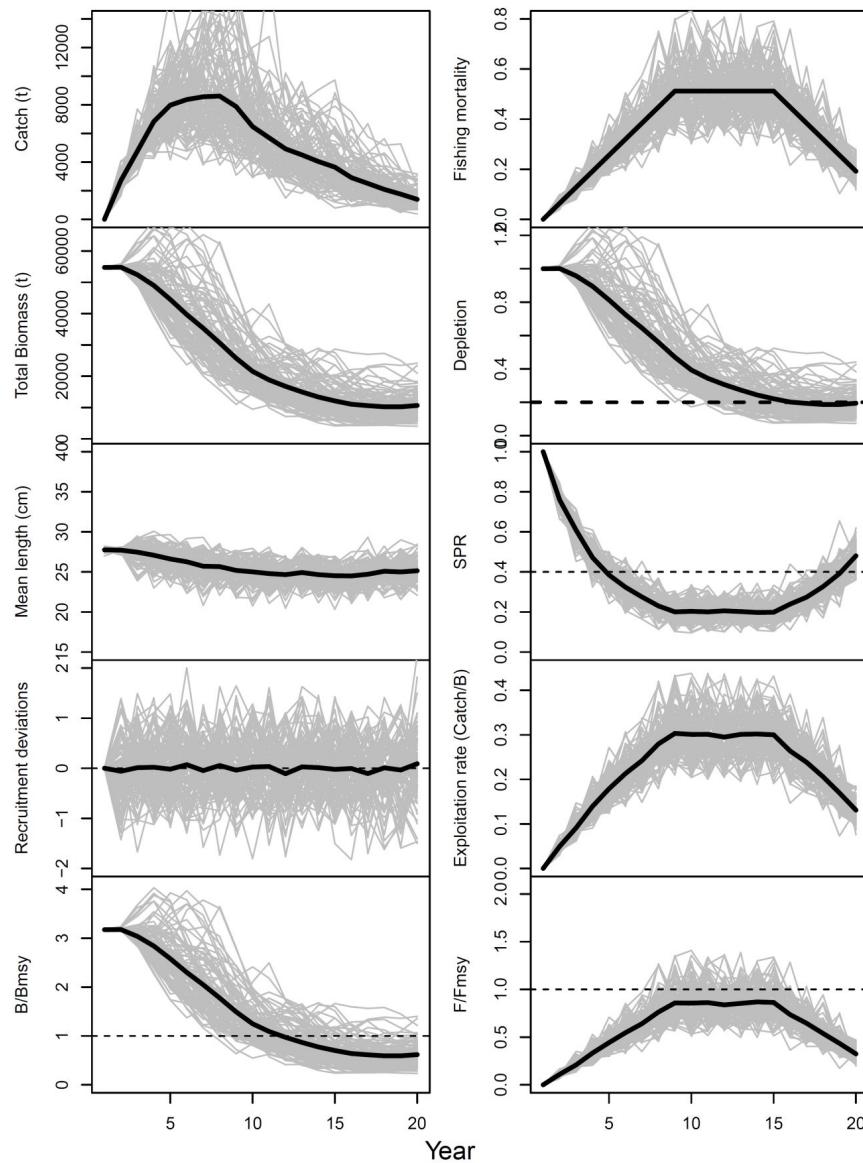
Appendix

Figure A1. Time series for each simulated fast-grow chub mackerel population for harvest scenario 1 and depletion = 0.2. The black solid lines represent the mean value for all runs.

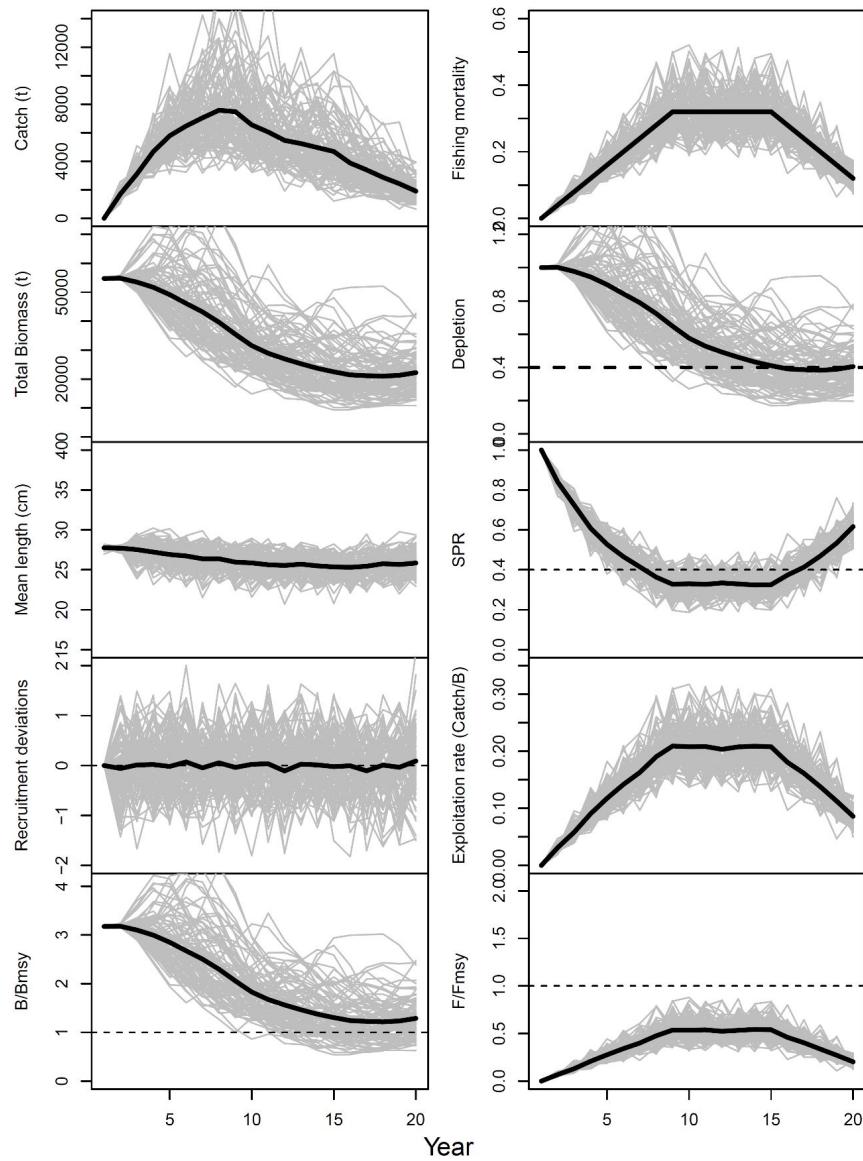


Figure A2. Time series for each simulated fast-grow chub mackerel population for harvest scenario 1 and depletion = 0.4. The black solid lines represent the mean value for all runs.

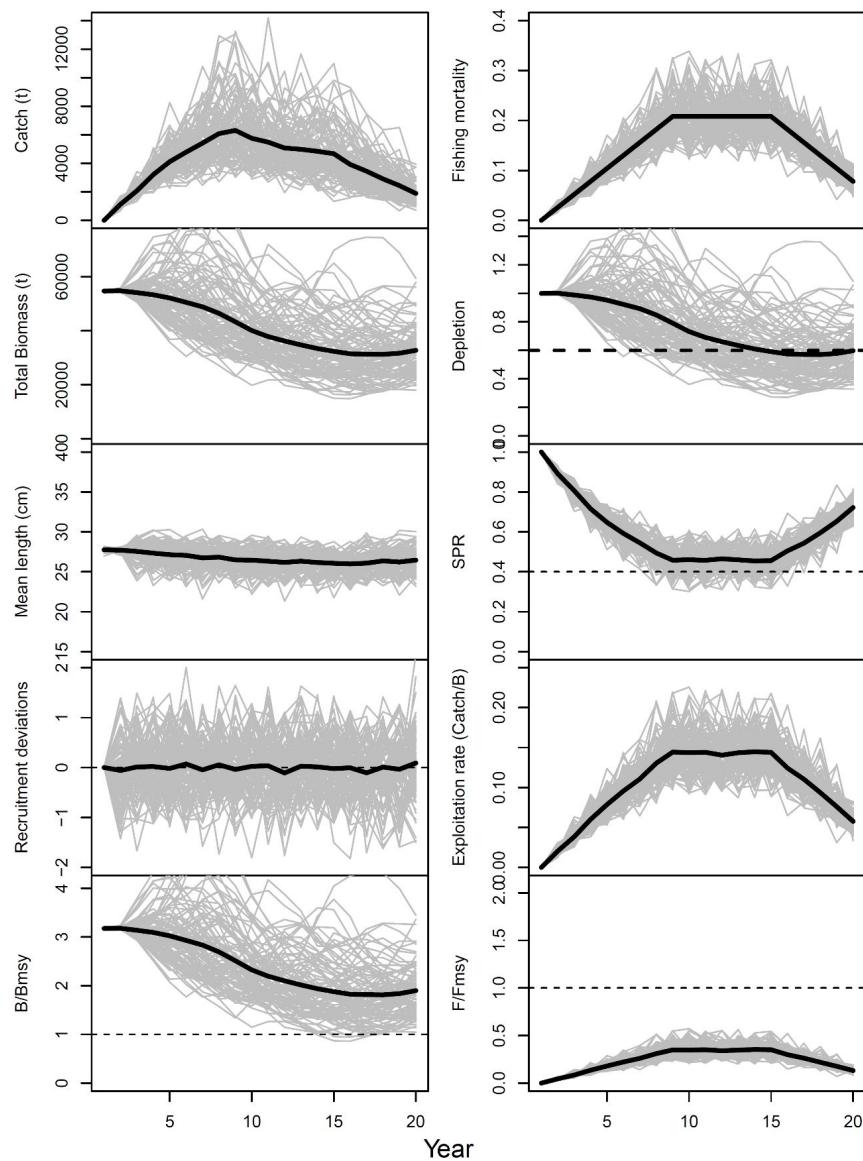


Figure A3. Time series for each simulated fast-grow chub mackerel population for harvest scenario 1 and depletion = 0.6. The black solid lines represent the mean value for all runs.

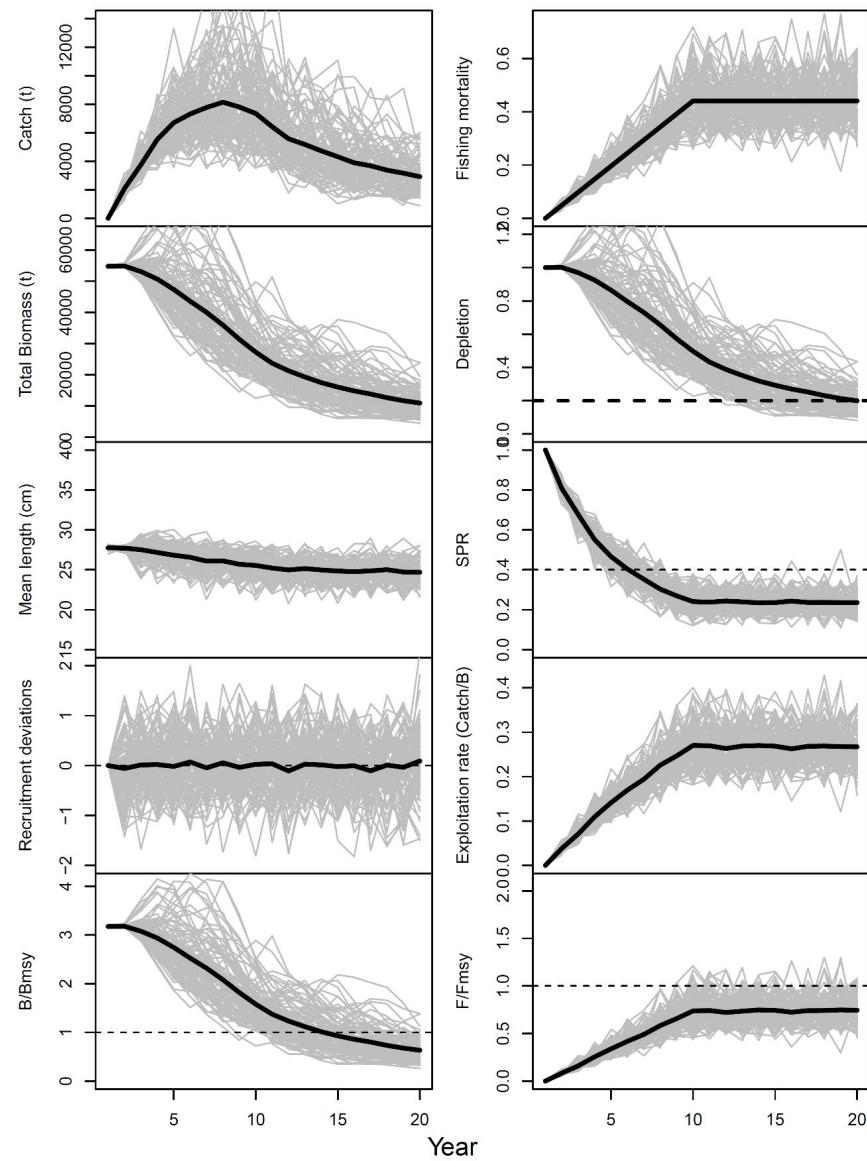


Figure A4. Time series for each simulated fast-grow chub mackerel population for harvest scenario 2 and depletion = 0.2. The black solid lines represent the mean value for all runs.

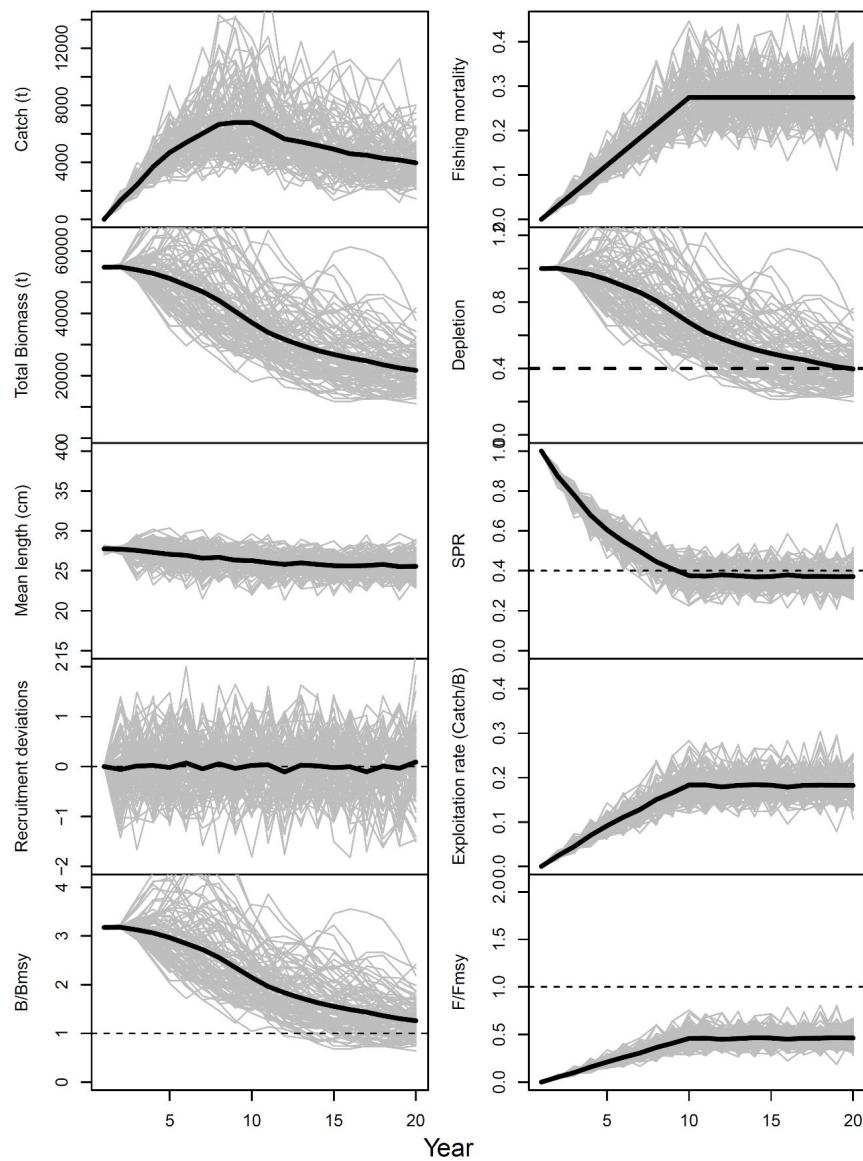


Figure A5. Time series for each simulated fast-grow chub mackerel population for harvest scenario 2 and depletion = 0.4. The black solid lines represent the mean value for all runs.

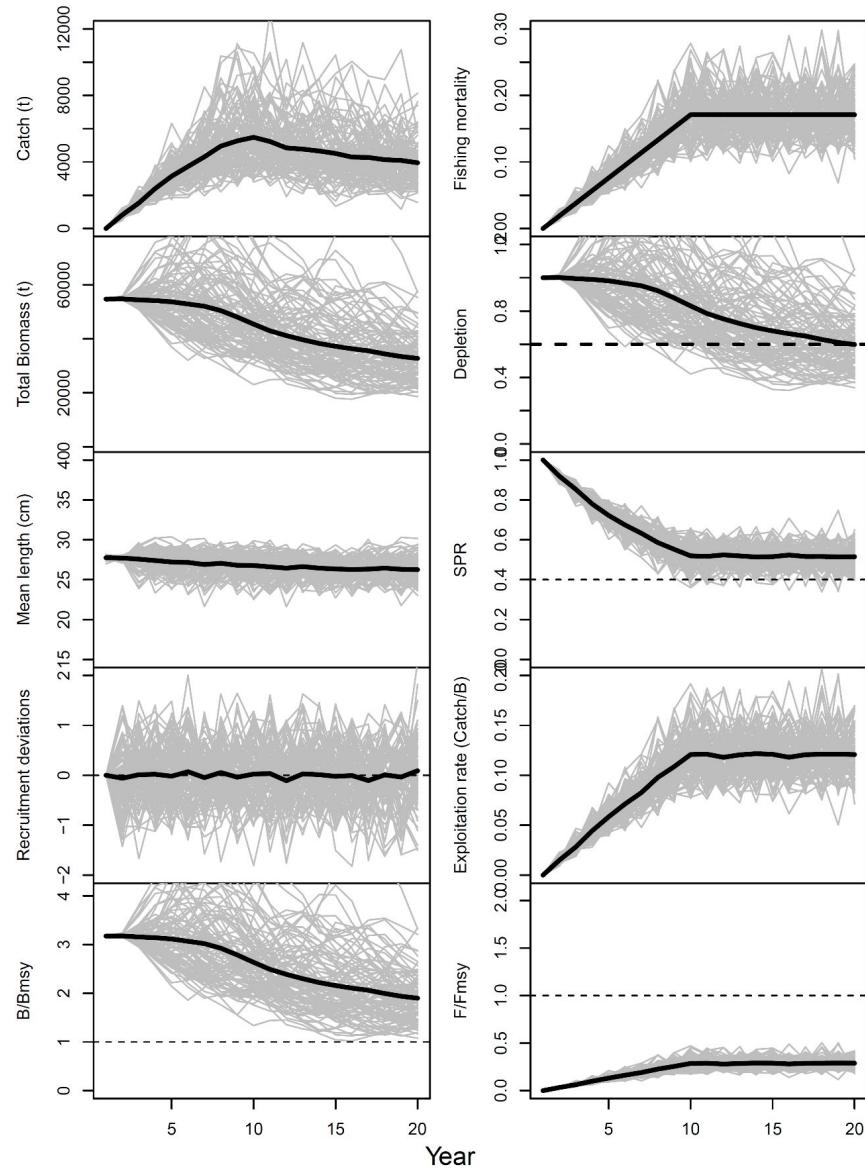


Figure A6. Time series for each simulated fast-grow chub mackerel population for harvest scenario 2 and depletion = 0.6. The black solid lines represent the mean value for all runs.

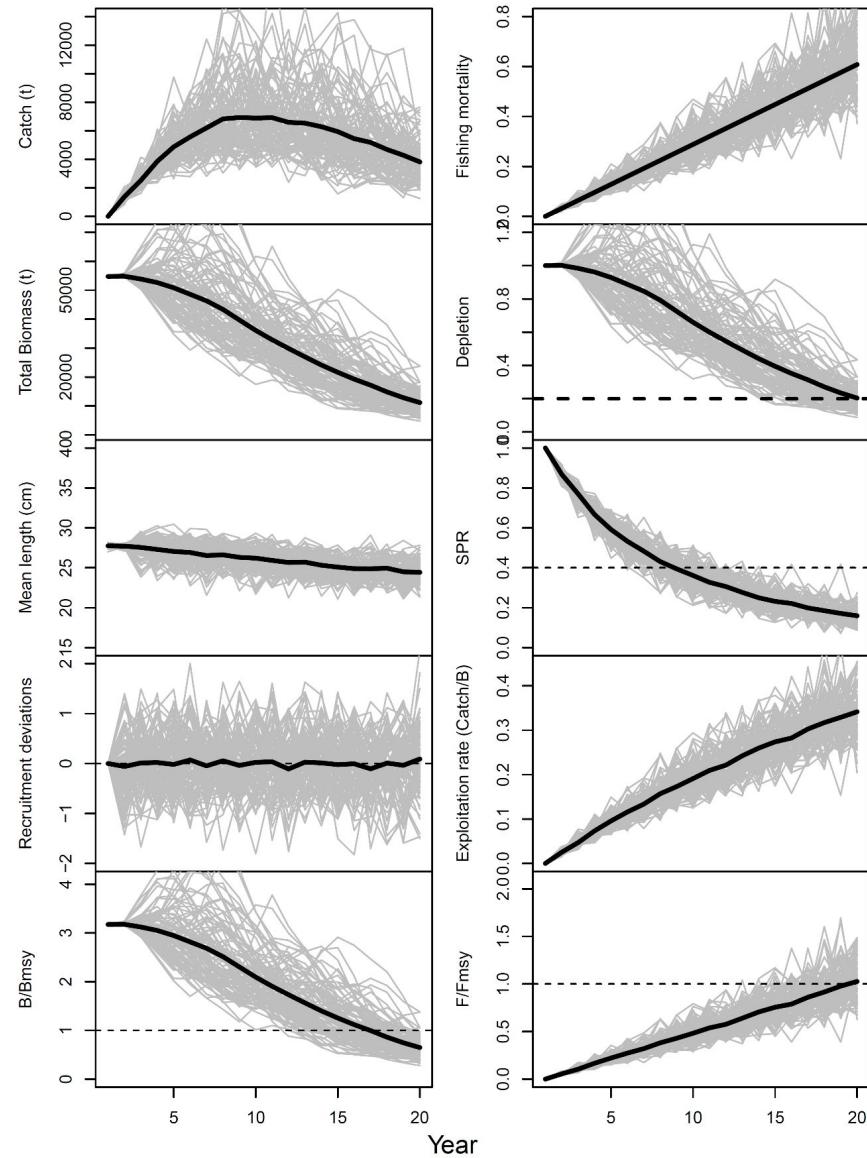


Figure A7. Time series for each simulated fast-grow chub mackerel population for harvest scenario 3 and depletion = 0.2. The black solid lines represent the mean value for all runs.

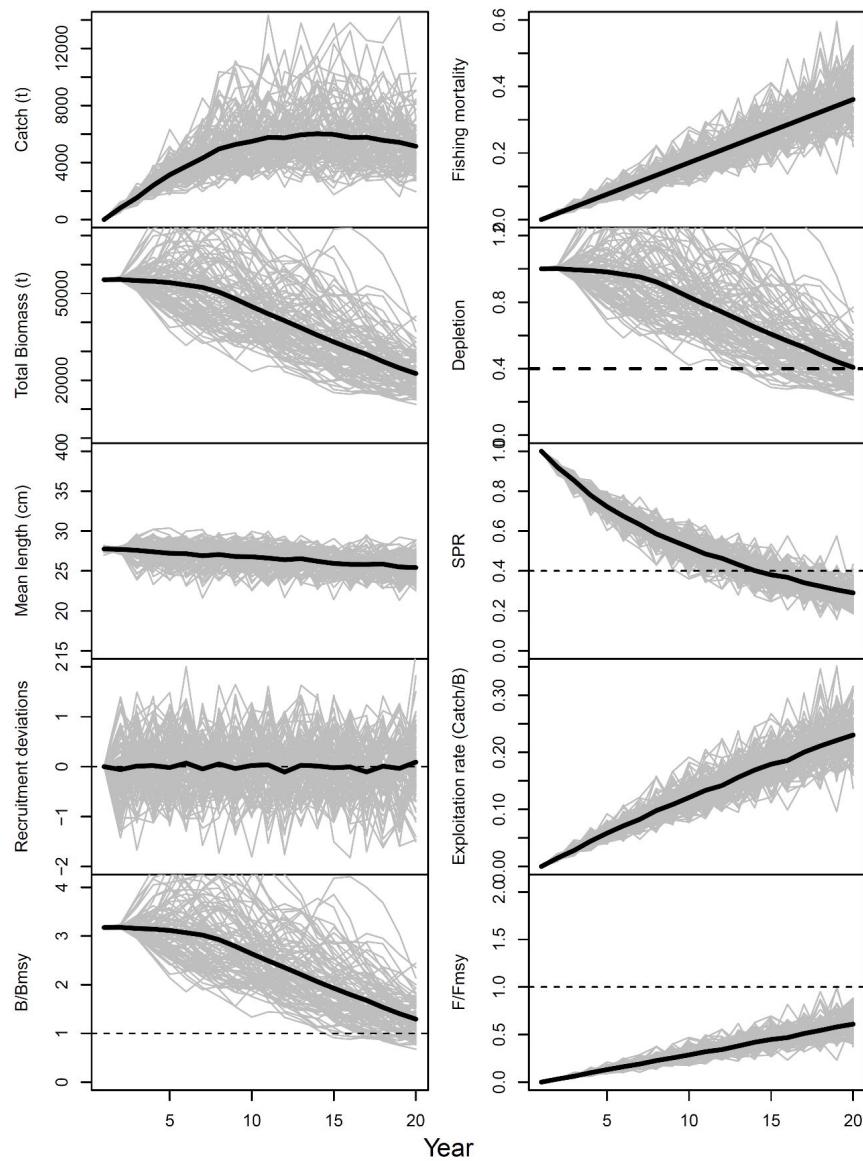


Figure A8. Time series for each simulated fast-grow chub mackerel population for harvest scenario 3 and depletion = 0.4. The black solid lines represent the mean value for all runs.

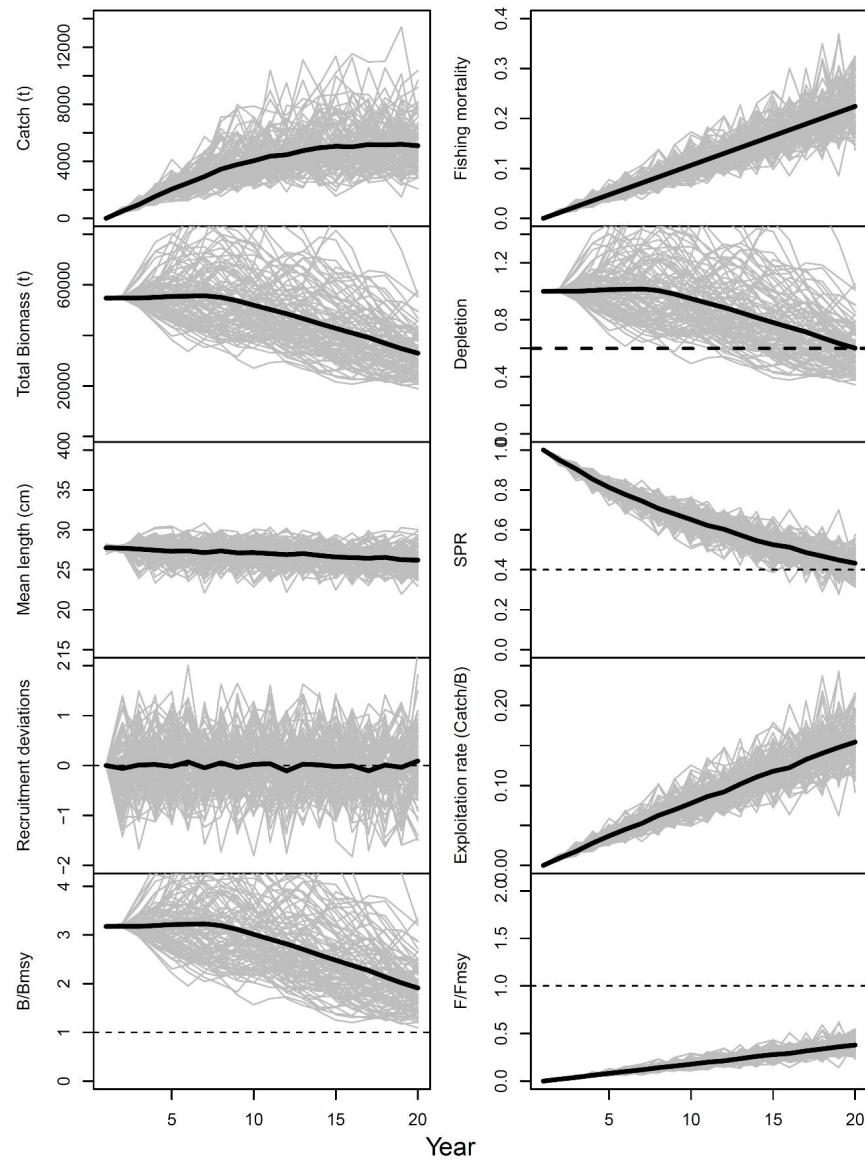


Figure A9. Time series for each simulated fast-grow chub mackerel population for harvest scenario 3 and depletion = 0.6. The black solid lines represent the mean value for all runs.

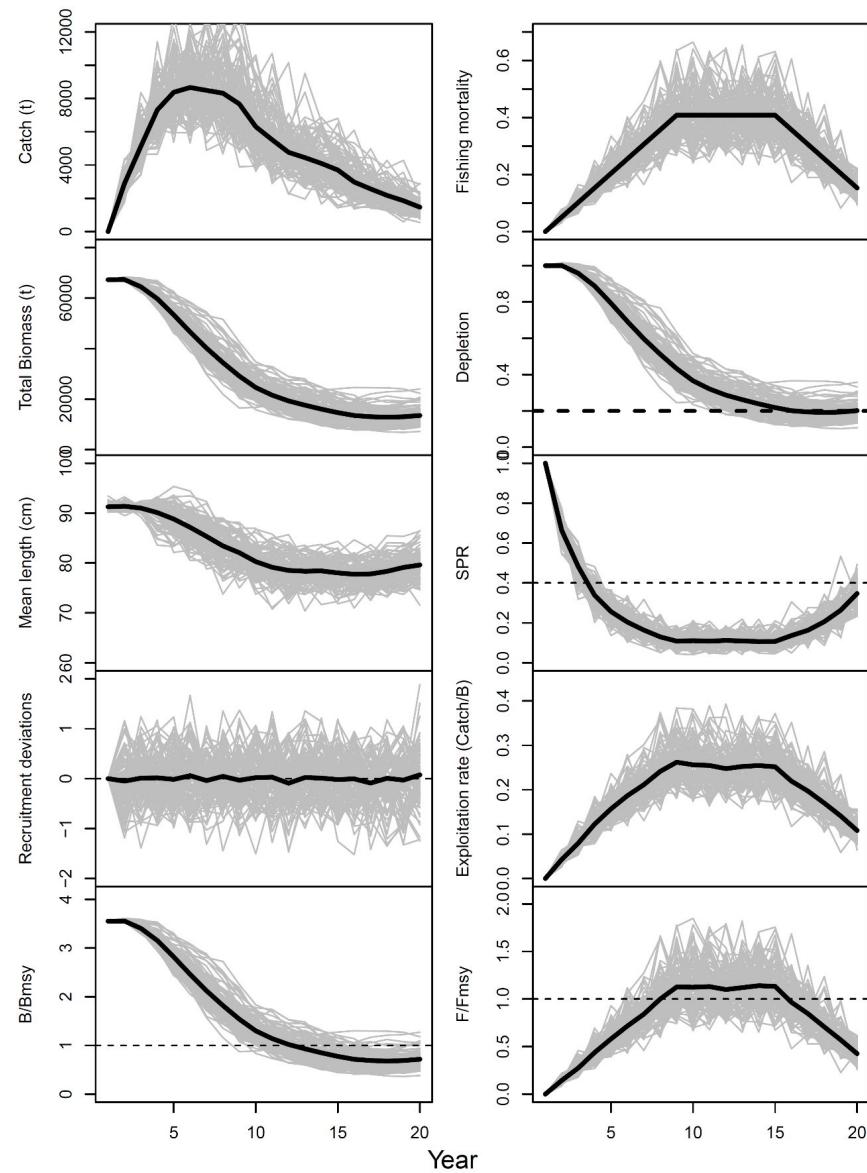


Figure A10. Time series for each simulated medium-grow albacore tuna population for harvest scenario 1 and depletion = 0.2. The black solid lines represent the mean value for all runs.

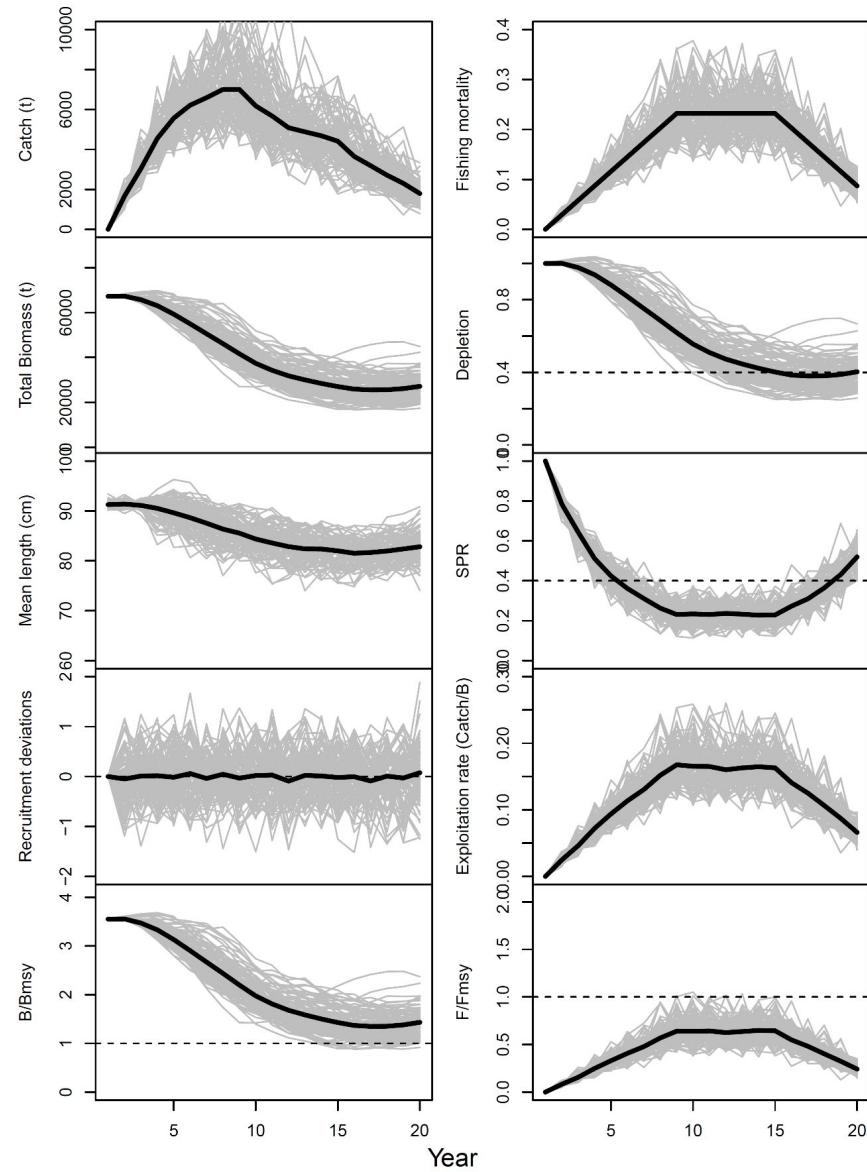


Figure A11. Time series for each simulated medium-grow albacore tuna population for harvest scenario 1 and depletion = 0.4. The black solid lines represent the mean value for all runs.

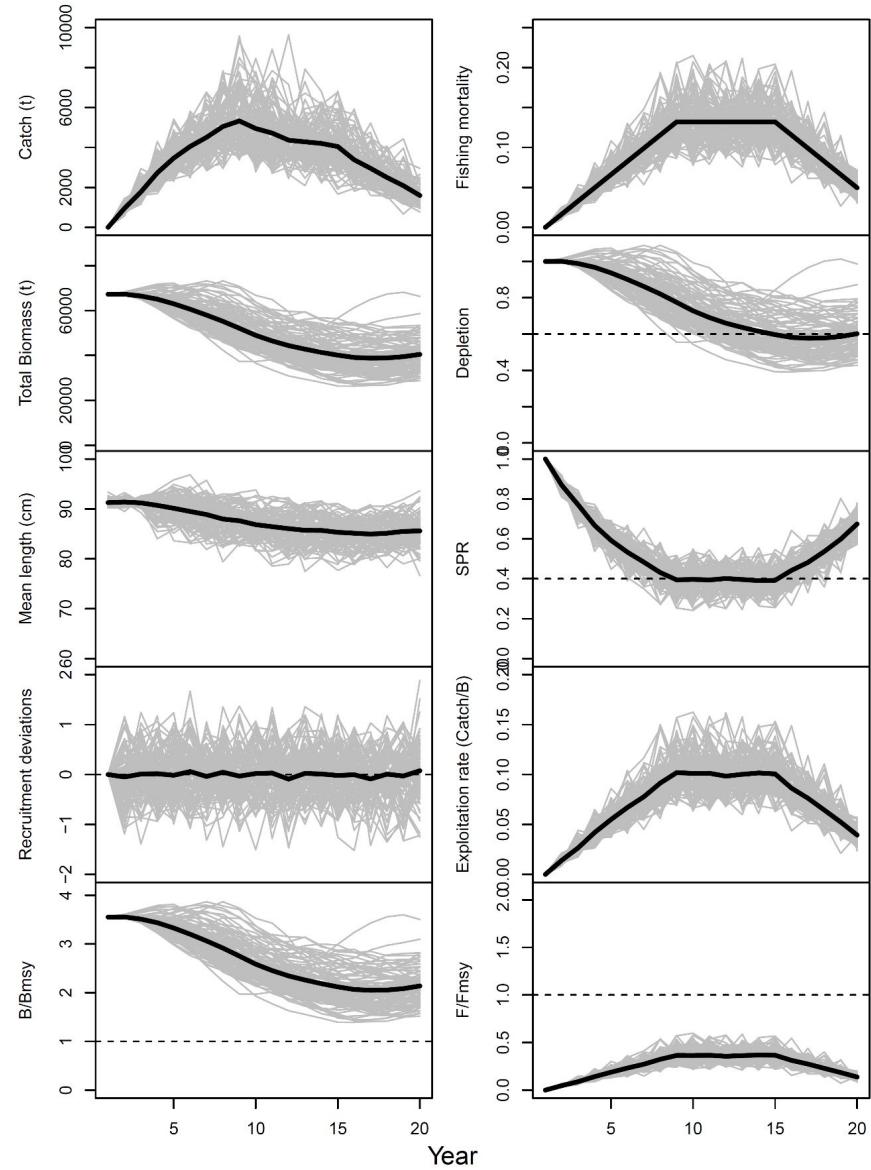


Figure A12. Time series for each simulated medium-grow albacore tuna population for harvest scenario 1 and depletion = 0.6. The black solid lines represent the mean value for all runs.

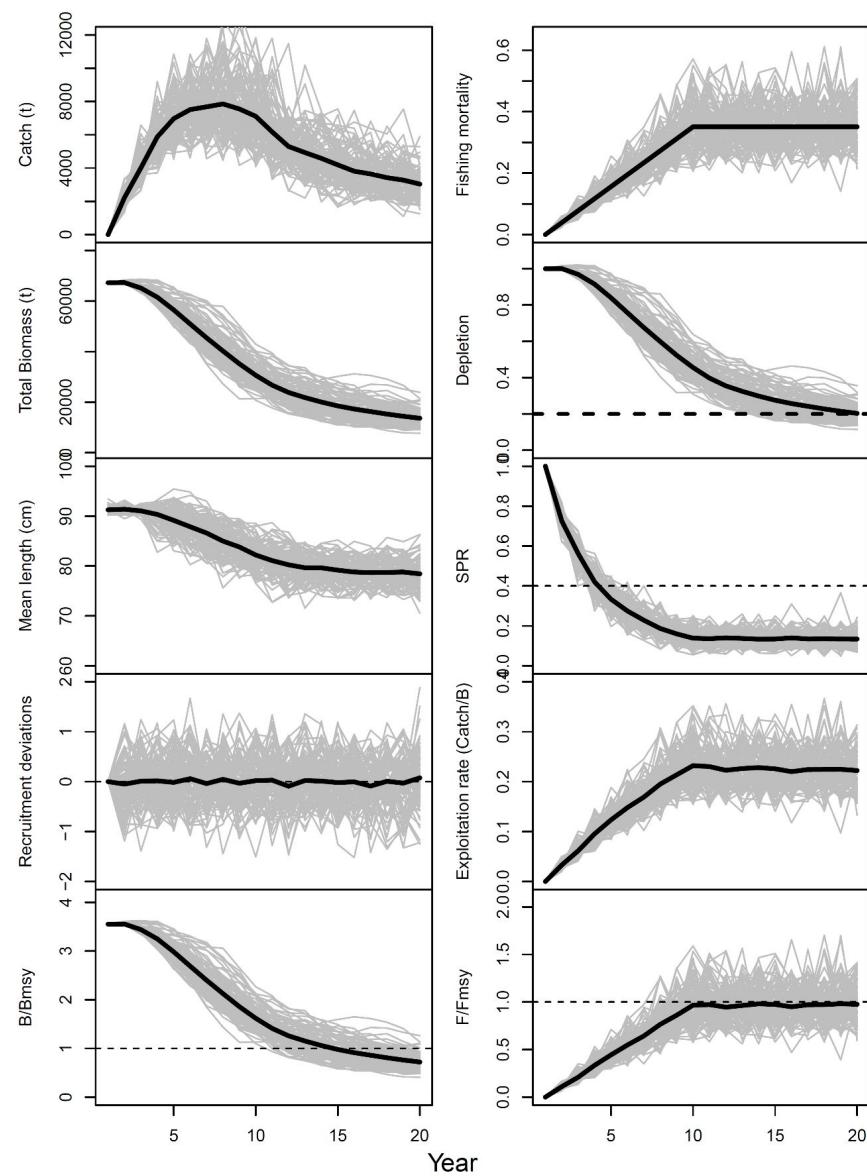


Figure A13. Time series for each simulated medium-grow albacore tuna population for harvest scenario 2 and depletion = 0.2. The black solid lines represent the mean value for all runs.

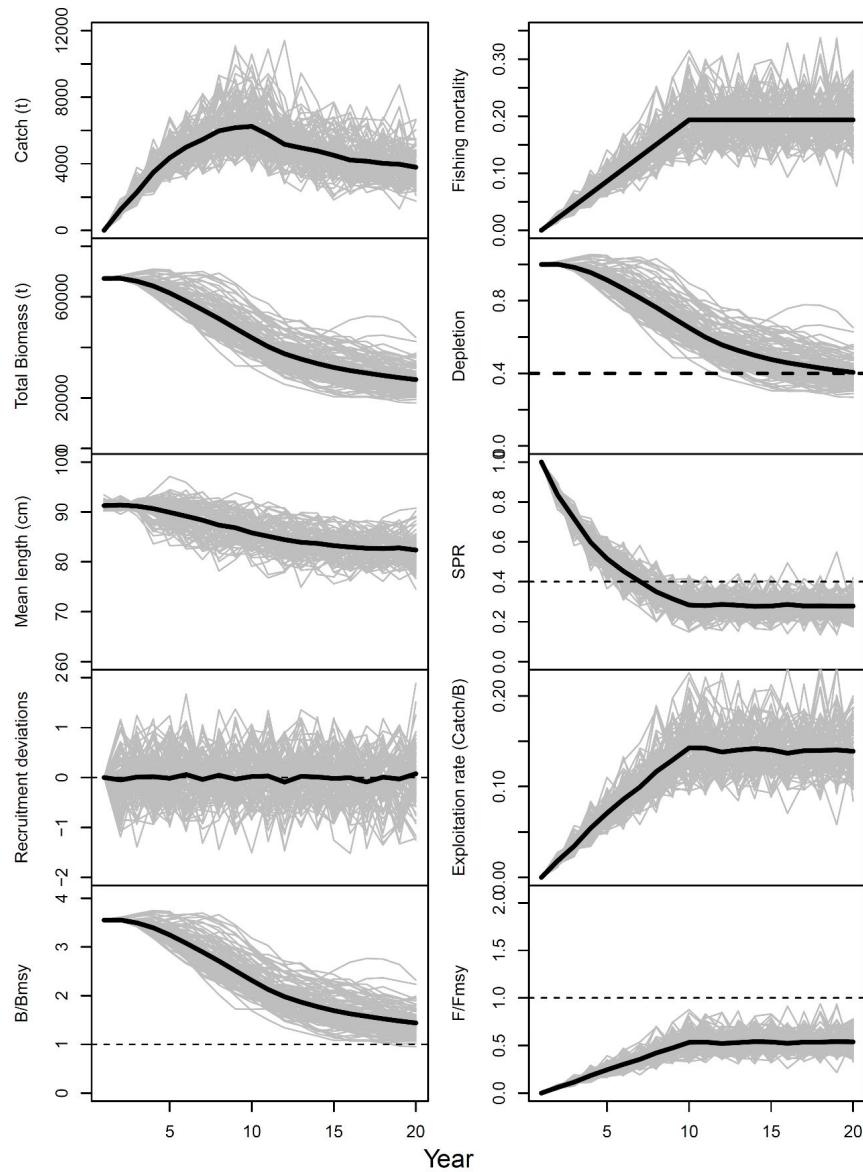


Figure A14. Time series for each simulated medium-grow albacore tuna population for harvest scenario 2 and depletion = 0.4. The black solid lines represent the mean value for all runs.

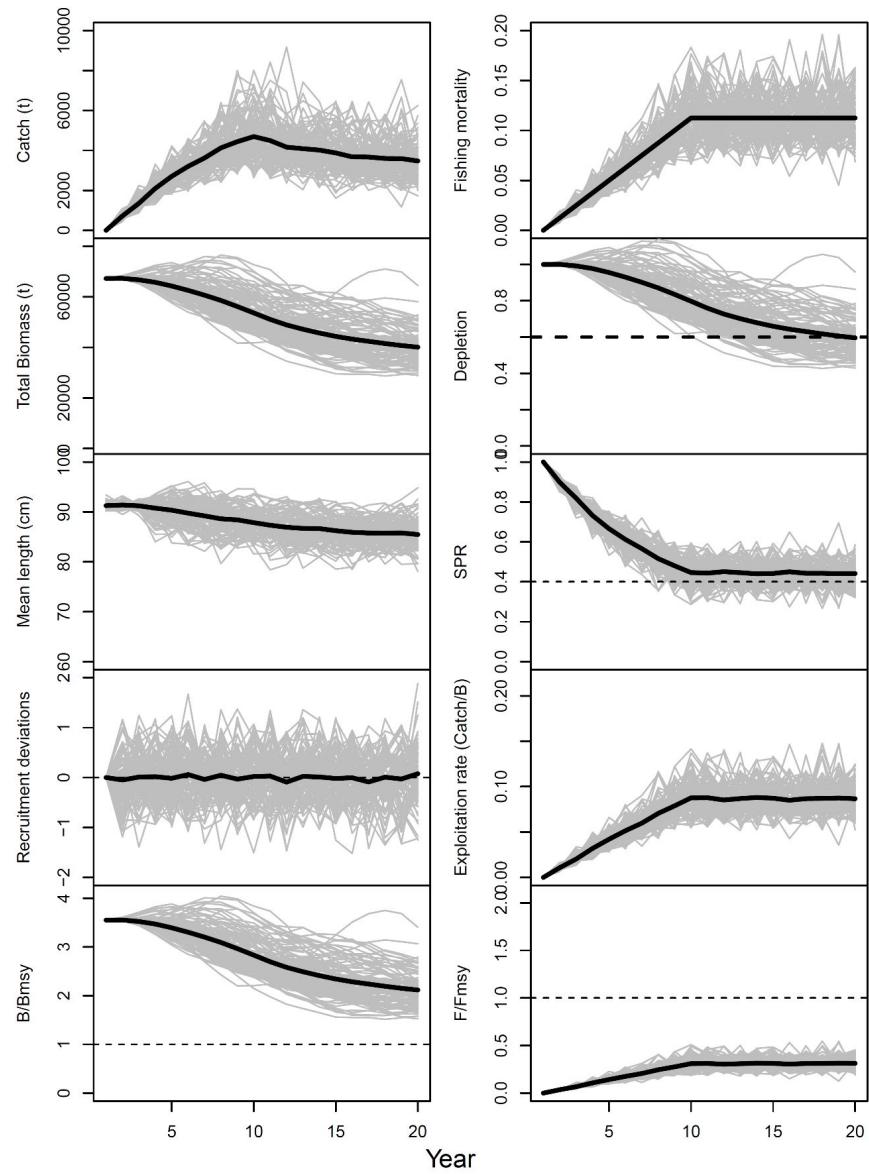


Figure A15. Time series for each simulated medium-grow albacore tuna population for harvest scenario 2 and depletion = 0.6. The black solid lines represent the mean value for all runs.

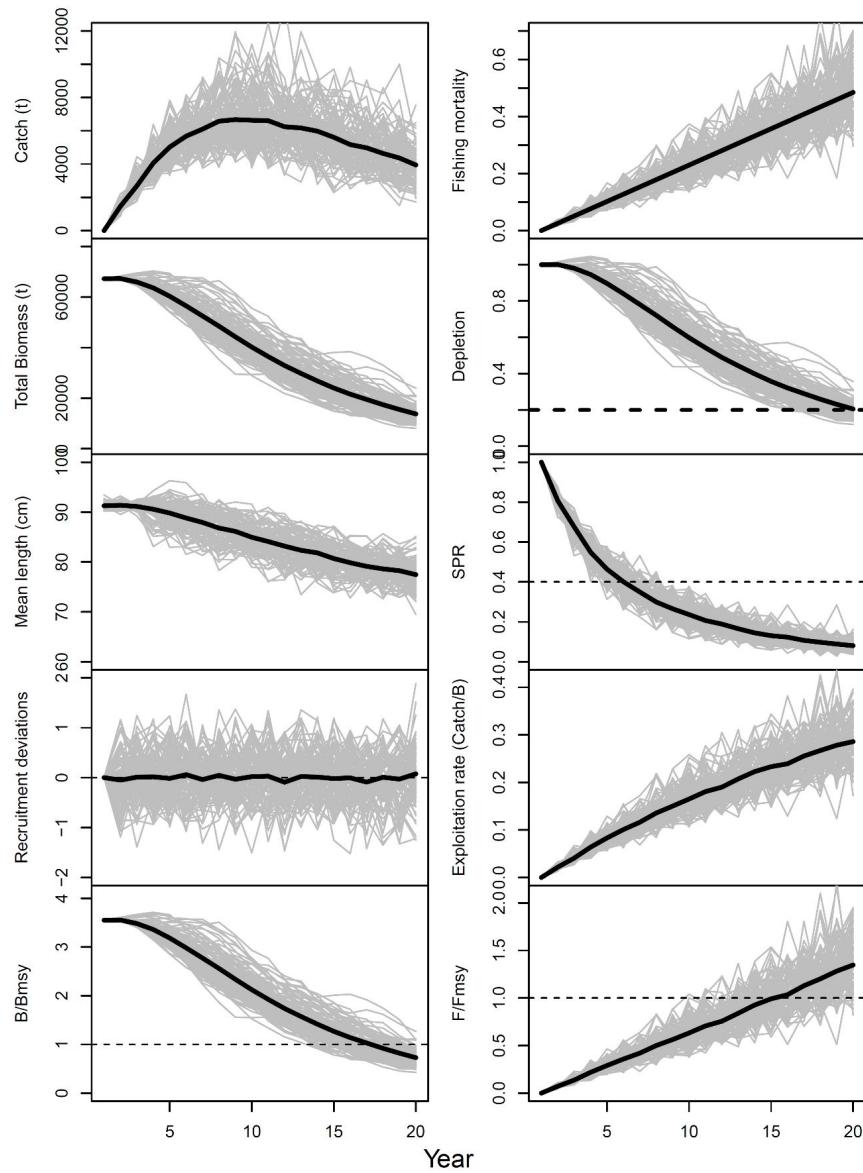


Figure A16. Time series for each simulated medium-grow albacore tuna population for harvest scenario 3 and depletion = 0.2. The black solid lines represent the mean value for all runs.

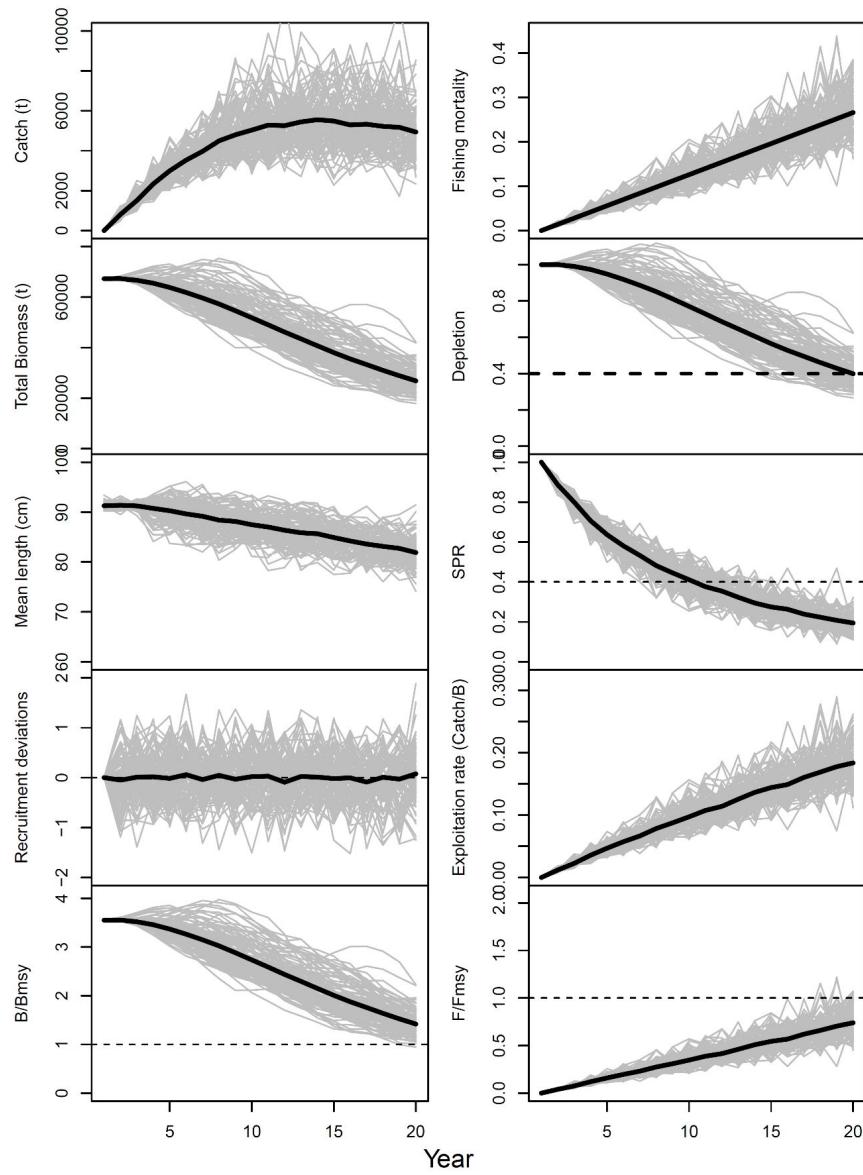


Figure A17. Time series for each simulated medium-grow albacore tuna population for harvest scenario 3 and depletion = 0.4. The black solid lines represent the mean value for all runs.

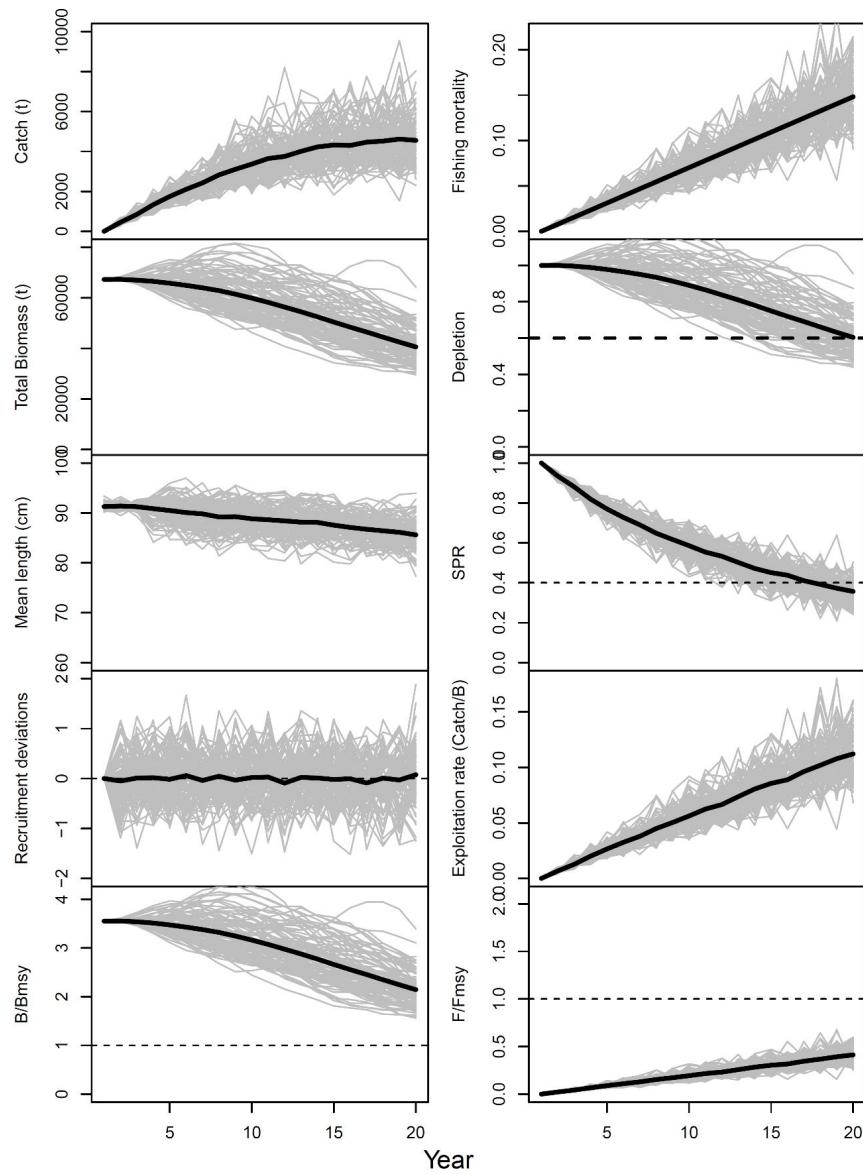


Figure A18. Time series for each simulated medium-grow albacore tuna population for harvest scenario 3 and depletion = 0.6. The black solid lines represent the mean value for all runs.

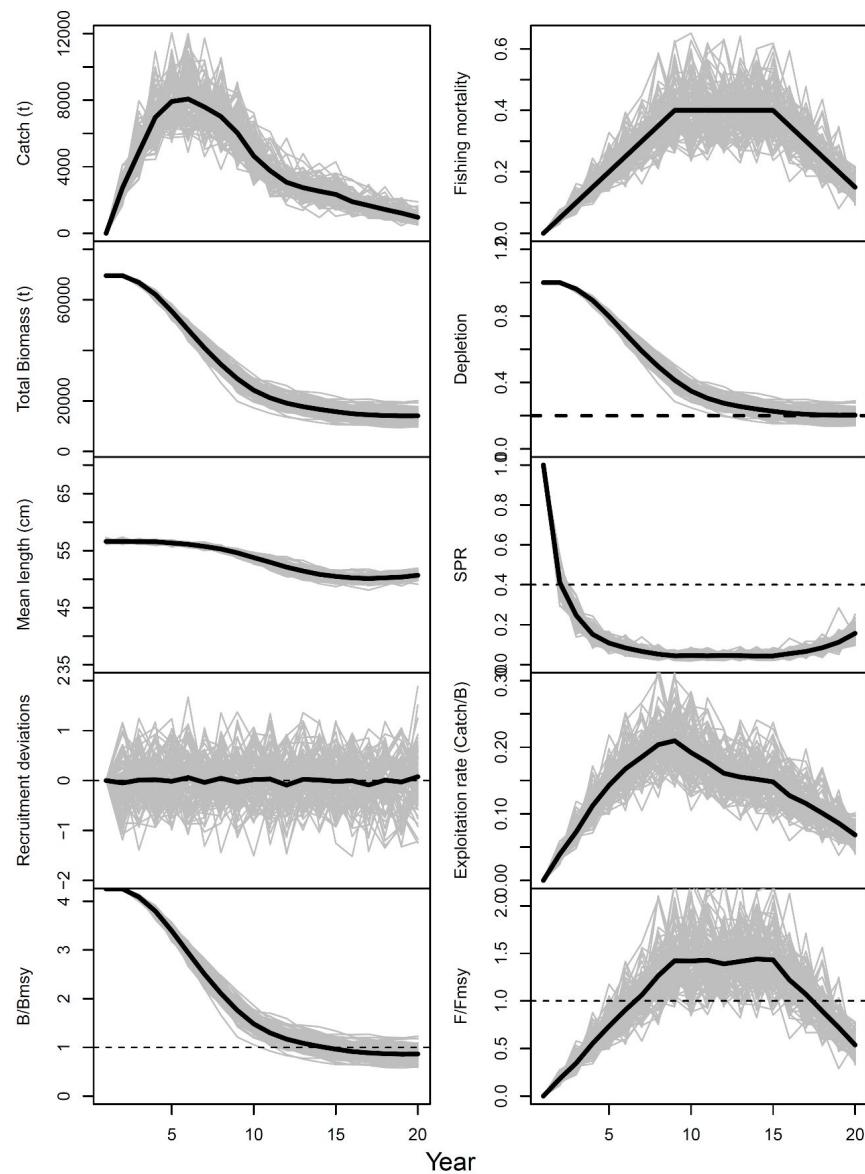


Figure A19. Time series for each simulated slow-grow canary rockfish population for harvest scenario 1 and depletion = 0.2. The black solid lines represent the mean value for all runs.

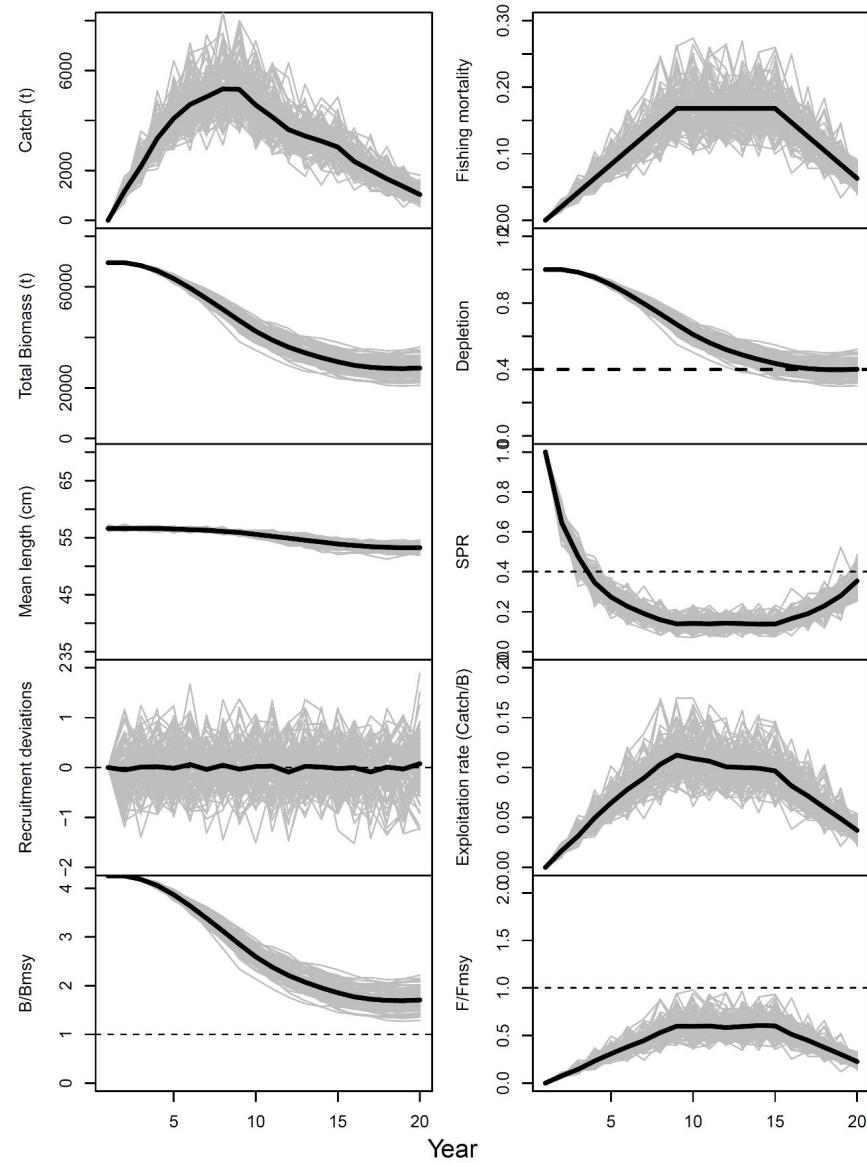


Figure A20. Time series for each simulated slow-grow canary rockfish population for harvest scenario 1 and depletion = 0.4. The black solid lines represent the mean value for all runs.

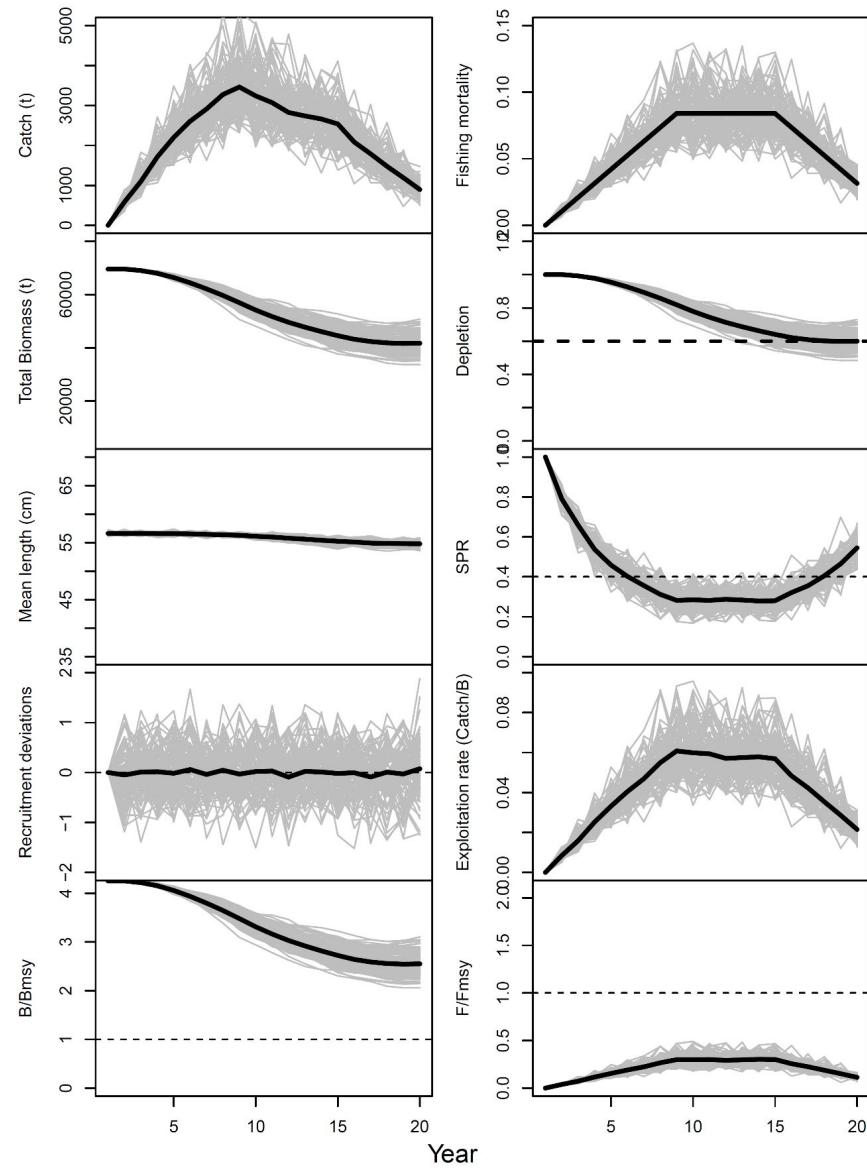


Figure A21. Time series for each simulated slow-grow canary rockfish population for harvest scenario 1 and depletion = 0.6. The black solid lines represent the mean value for all runs.

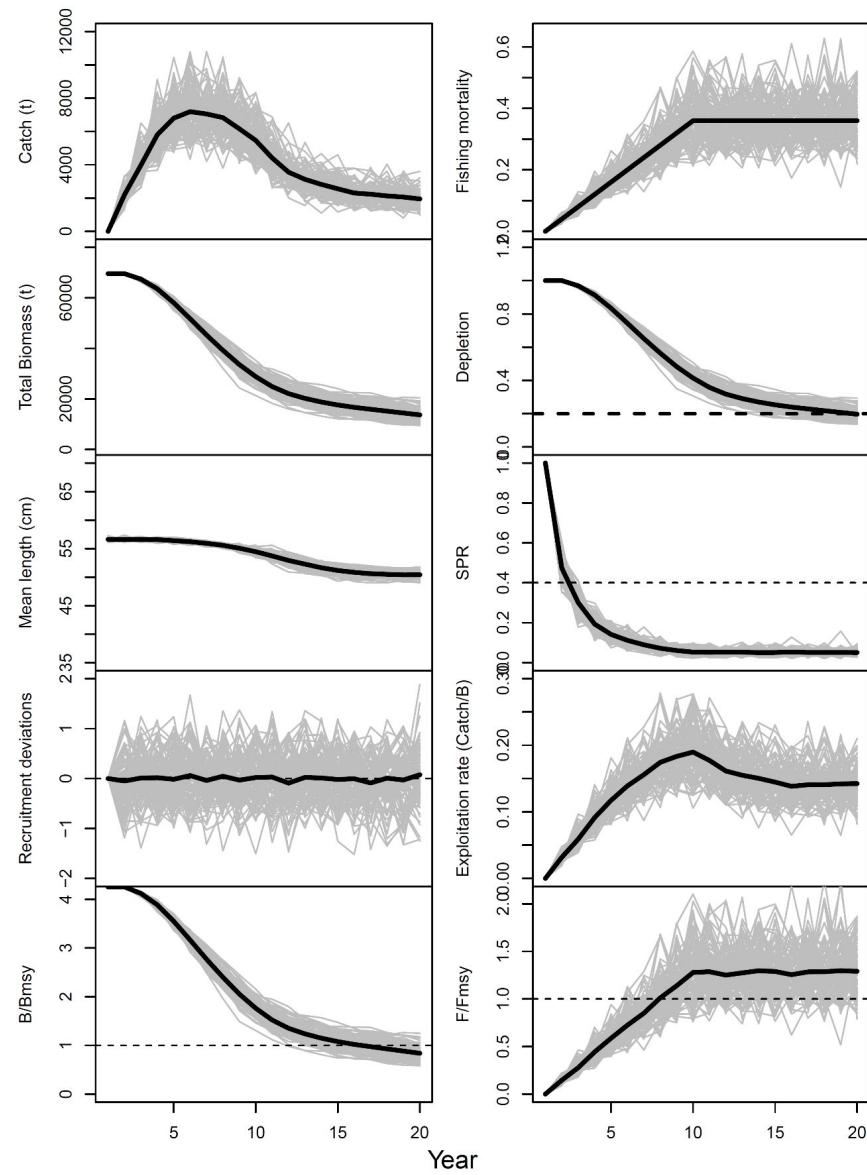


Figure A22. Time series for each simulated slow-grow canary rockfish population for harvest scenario 2 and depletion = 0.2. The black solid lines represent the mean value for all runs.

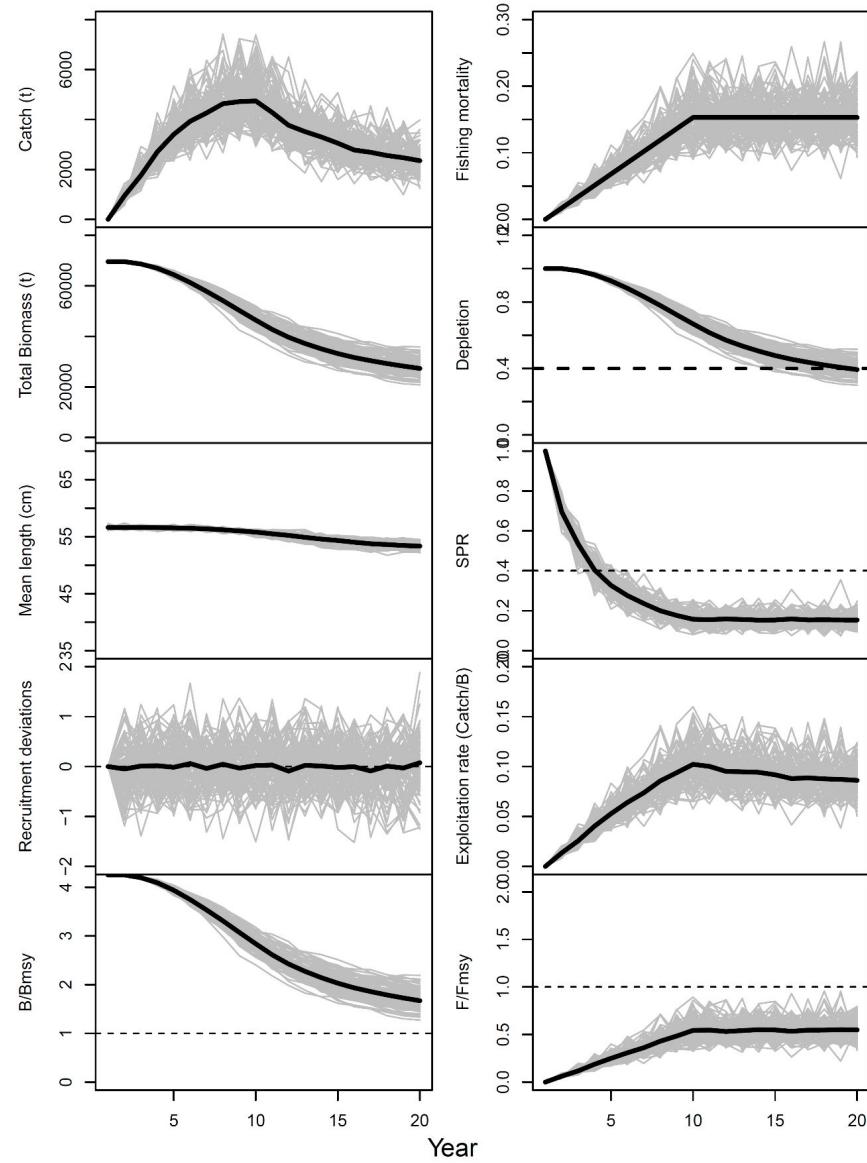


Figure A23. Time series for each simulated slow-grow canary rockfish population for harvest scenario 2 and depletion = 0.4. The black solid lines represent the mean value for all runs.

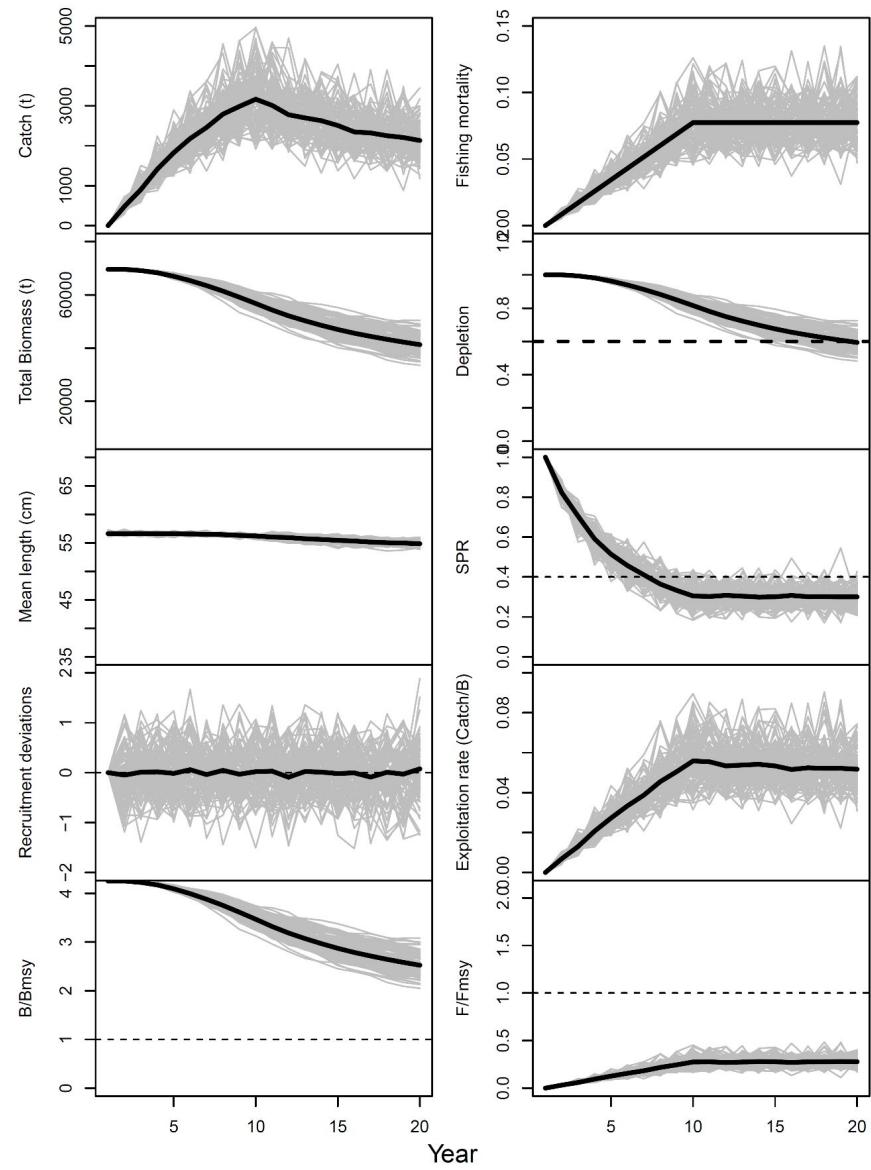


Figure A24. Time series for each simulated slow-grow canary rockfish population for harvest scenario 2 and depletion = 0.6. The black solid lines represent the mean value for all runs.

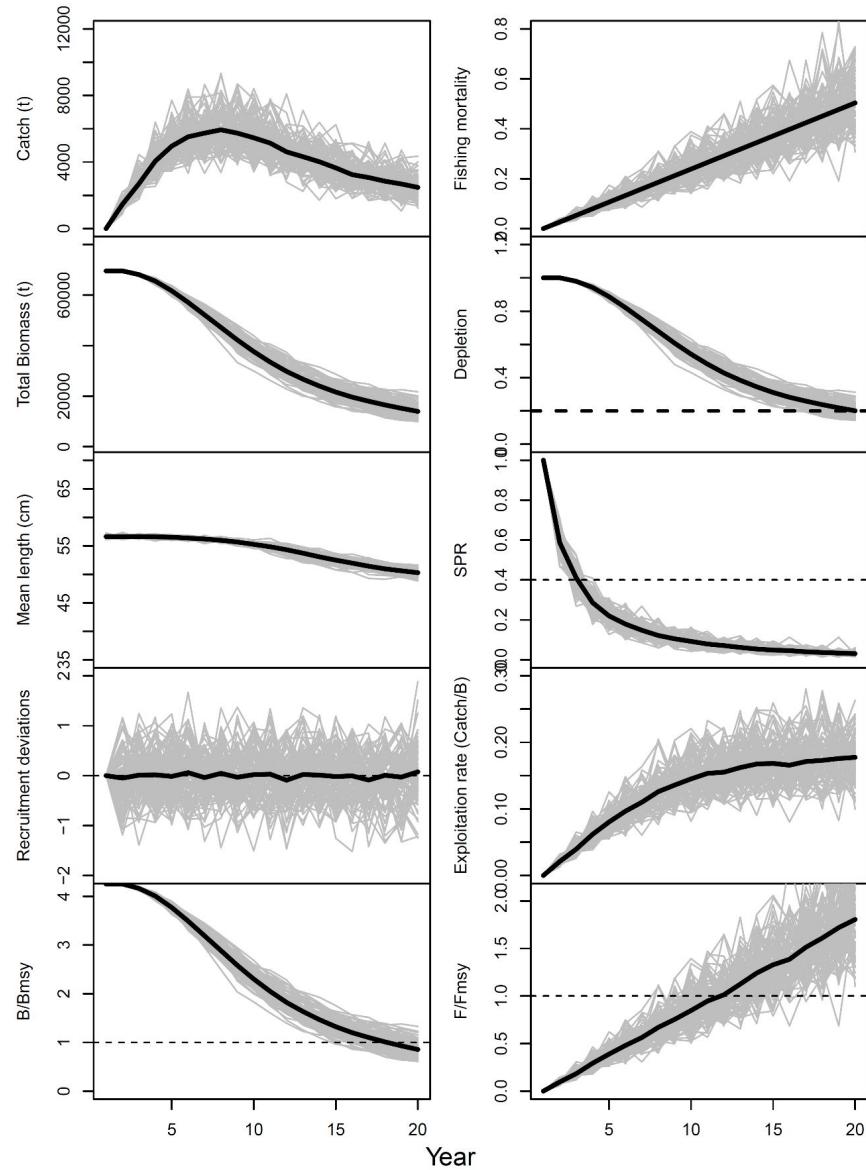


Figure A25. Time series for each simulated slow-grow canary rockfish population for harvest scenario 1 and depletion = 0.2. The black solid lines represent the mean value for all runs.

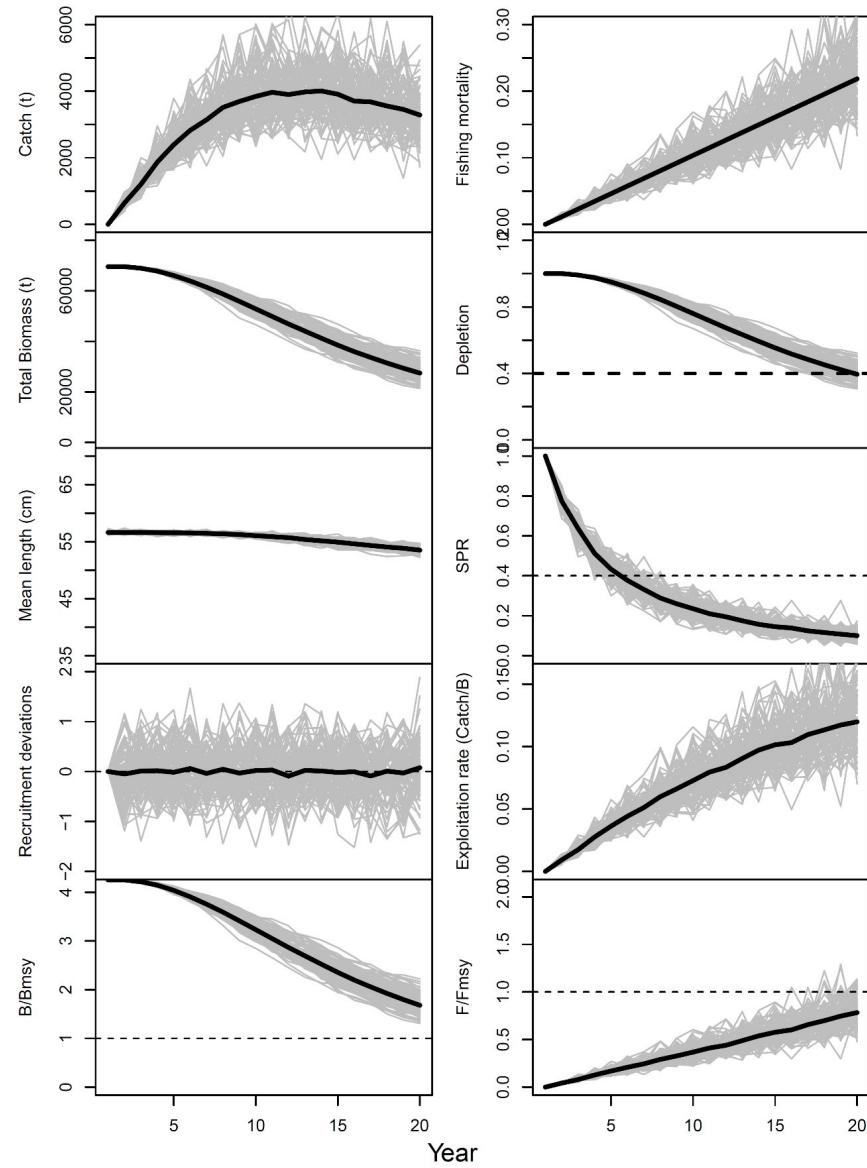


Figure A26. Time series for each simulated slow-grow canary rockfish population for harvest scenario 3 and depletion = 0.4. The black solid lines represent the mean value for all runs.

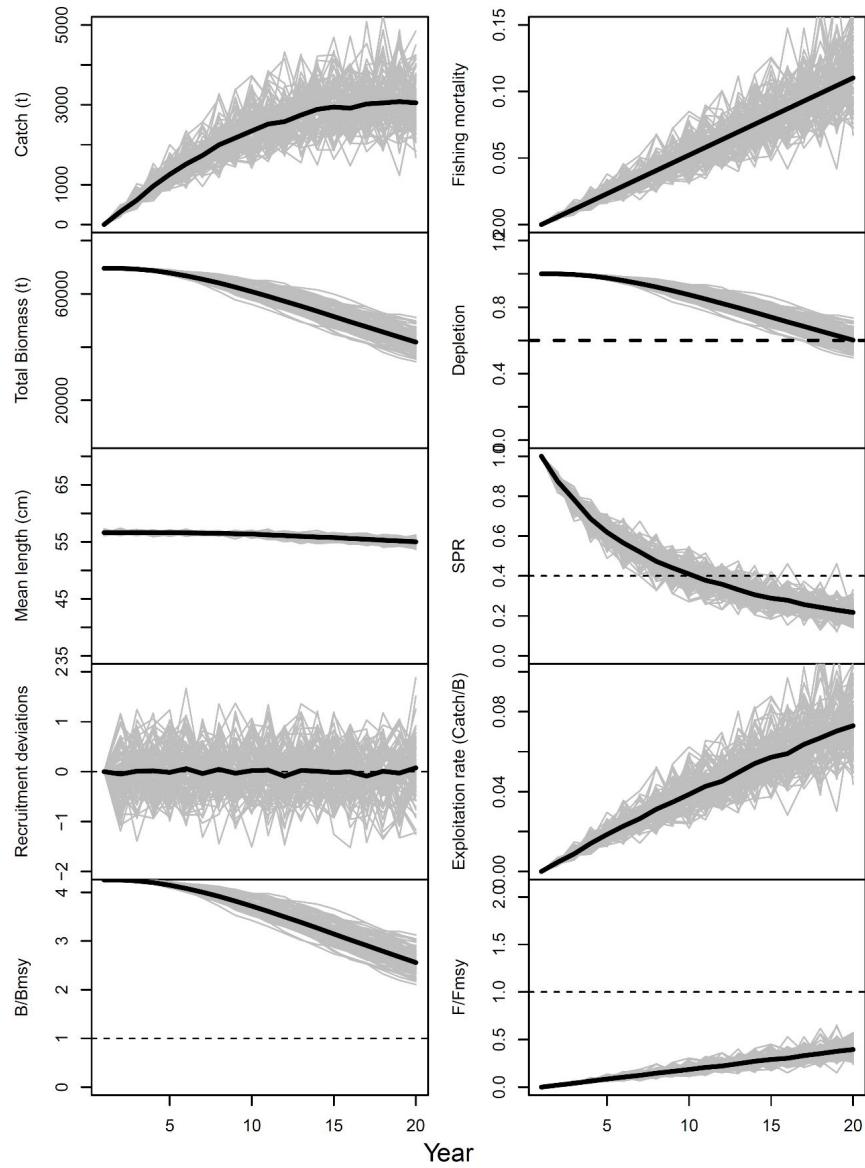


Figure A27. Time series for each simulated slow-grow canary rockfish population for harvest scenario 3 and depletion = 0.6. The black solid lines represent the mean value for all runs.

Table A1. Mean and standard deviation of relative error (RE) between the maximum sustainable yield (MSY) estimated by the operating model (OM) and the MSY estimated by the CMSY method. Values are proportions. Positive values mean that the MSY was overestimated and negative values that it was underestimated.

| Species | Harvest Scenario | Mean ± Standard deviation | | |
|----------|------------------|---------------------------|--------------|--------------|
| | | Depletion level | | |
| | | 0.2 | 0.4 | 0.6 |
| Mackerel | Scenario 1 | 0.37 ± 0.08 | 0.18 ± 0.09 | -0.13 ± 0.15 |
| | Scenario 2 | 0.28 ± 0.08 | 0.03 ± 0.09 | -0.22 ± 0.13 |
| | Scenario 3 | 0.33 ± 0.11 | 0.05 ± 0.11 | -0.26 ± 0.14 |
| Albacore | Scenario 1 | -0.07 ± 0.02 | -0.18 ± 0.03 | -0.46 ± 0.02 |
| | Scenario 2 | -0.25 ± 0.02 | -0.42 ± 0.02 | -0.42 ± 0.02 |
| | Scenario 3 | -0.12 ± 0.02 | -0.24 ± 0.02 | -0.43 ± 0.02 |
| Rockfish | Scenario 1 | 1.61 ± 0.27 | 1.13 ± 0.31 | 0.80 ± 0.32 |
| | Scenario 2 | 1.67 ± 0.26 | 1.27 ± 0.33 | 0.86 ± 0.34 |
| | Scenario 3 | 1.77 ± 0.33 | 1.29 ± 0.45 | 0.94 ± 0.45 |

Table A2. True OM and LBB estimated values for L_{∞} and S_{50} ($\sim L_c$). LCL is the lower confidence limit and UCL the upper confidence limit for the estimated values.

| Scenarios | Life-history | Harvest trend | Final depletion | True L_{∞} | True S_{50} | Estimated L_{∞} | Estimated LCL L_{∞} | Estimated UCL L_{∞} | Estimated L_c | Estimated LCL L_c | Estimated UCL L_c |
|-----------|--------------|---------------|-----------------|-------------------|---------------|------------------------|----------------------------|----------------------------|-----------------|---------------------|---------------------|
| 1 | Short-lived | Scenario 1 | 0.2 | 38.2 | 25 | 44.5 | 44.0 | 45.1 | 16.9 | 16.5 | 17.4 |
| 2 | | | 0.4 | 38.2 | 29 | 43.6 | 43.3 | 44.0 | 15.6 | 15.5 | 15.8 |
| 3 | | | 0.6 | 38.2 | 29 | 45.1 | 44.6 | 45.6 | 20.1 | 19.5 | 20.7 |
| 4 | | Scenario 2 | 0.2 | 38.2 | 29 | 44.7 | 44.2 | 45.2 | 18.0 | 17.4 | 18.6 |
| 5 | | | 0.4 | 38.2 | 29 | 43.7 | 43.4 | 44.1 | 15.7 | 15.5 | 15.8 |
| 6 | | | 0.6 | 38.2 | 29 | 45.3 | 45.0 | 45.8 | 20.3 | 19.7 | 20.9 |
| 7 | | Scenario 3 | 0.2 | 38.2 | 29 | 45.5 | 45.2 | 46.0 | 20.3 | 19.8 | 21.0 |
| 8 | | | 0.4 | 38.2 | 29 | 44.5 | 44.0 | 45.1 | 15.9 | 15.7 | 16.1 |
| 9 | | | 0.6 | 38.2 | 29 | 45.7 | 45.4 | 46.2 | 20.7 | 20.1 | 21.3 |
| 10 | Medium-lived | Scenario 1 | 0.2 | 122.2 | 60 | 142.5 | 141.4 | 143.6 | 62.4 | 61.7 | 63.2 |
| 11 | | | 0.4 | 122.2 | 60 | 141.4 | 140.1 | 142.5 | 60.4 | 59.7 | 61.0 |
| 12 | | | 0.6 | 122.2 | 60 | 143.2 | 142.7 | 143.9 | 64.1 | 63.2 | 64.9 |
| 13 | | Scenario 2 | 0.2 | 122.2 | 60 | 143.0 | 141.8 | 143.9 | 62.8 | 62.0 | 63.6 |
| 14 | | | 0.4 | 122.2 | 60 | 141.9 | 141.1 | 143.3 | 60.7 | 60.0 | 61.4 |
| 15 | | | 0.6 | 122.2 | 60 | 143.3 | 143.0 | 144.0 | 64.6 | 63.8 | 65.4 |
| 16 | | Scenario 3 | 0.2 | 122.2 | 60 | 143.3 | 143.0 | 144.0 | 63.3 | 62.5 | 64.1 |
| 17 | | | 0.4 | 122.2 | 60 | 142.5 | 141.4 | 143.7 | 61.1 | 60.4 | 61.8 |
| 18 | | | 0.6 | 122.2 | 60 | 143.5 | 143.1 | 144.3 | 65.4 | 64.5 | 66.3 |
| 19 | Long-lived | Scenario 1 | 0.2 | 60.0 | 60 | 76.7 | 76.0 | 77.6 | 50.4 | 49.9 | 51.0 |
| 20 | | | 0.4 | 60.0 | 45 | 75.3 | 74.5 | 76.3 | 47.0 | 46.5 | 47.4 |
| 21 | | | 0.6 | 60.0 | 45 | 76.9 | 76.2 | 77.8 | 51.2 | 50.7 | 51.8 |
| 22 | | Scenario 2 | 0.2 | 60.0 | 45 | 76.5 | 75.8 | 77.4 | 50.5 | 50.0 | 51.1 |
| 23 | | | 0.4 | 60.0 | 45 | 75.4 | 74.6 | 76.4 | 47.8 | 47.2 | 48.3 |
| 24 | | | 0.6 | 60.0 | 45 | 77.0 | 76.3 | 78.0 | 51.4 | 50.8 | 52.0 |
| 25 | | Scenario 3 | 0.2 | 60.0 | 45 | 76.8 | 76.1 | 77.7 | 51.0 | 50.4 | 51.6 |
| 26 | | | 0.4 | 60.0 | 45 | 76.0 | 75.3 | 76.9 | 49.5 | 49.0 | 50.1 |
| 27 | | | 0.6 | 60.0 | 45 | 77.1 | 76.3 | 78.0 | 51.7 | 51.1 | 52.2 |

B21 ROC

Proxy MSY Reference Points for data poor stocks

Power of length based indicators to detect overfishing.

L Kell

21 November, 2019

Summary

- The OM has been rerun where F after recovery is $0.7F_{MSY}$.
- The reference period is 61:120 and the period of overfishing is 80:109. this gives 60 years for which 30 are overfished.
- The biomass indicator L_{maxy} has been added.
- Figures aren't final but hopefully show the types of plots we might want to include
- Figure 7 is new and shows the "best performance", i.e. point closest to $TPR=1$ and $FPR=0$, this could be the entry point to the regression tree.

B22 Pareto

Pareto Frontiers - Multiple Objective Trade-offs

Pollack

L Kell

14 November, 2019

Summary of simulations and analyses

- The random MSE has been run to derive the calibration regressions and pareto frontiers, based on these the “final” MSEs have been run.
- The OM is the same as for the ROC paper, however, only one stock was run, pollack, as a single example is interesting enough
- There are factors for CPUE CV with three levels (0.1, 0.2, 0.3) and number of years in the HCR regression (3, 5, 7).
- The MP is an empirical HCR, where K1 is how much you decrease TAC for a -ve trend in CPUE and K2 how much you increase TAC for a +ve trend

Catches are increased when the trend in an index is positive, and decreased if the trend is negative

$$TAC_{y+1}^1 = TAC_y \times \begin{cases} 1 - k_1|\lambda| & \text{for } \lambda < 0 \\ 1 + k_2\lambda & \text{for } \lambda \geq 0 \end{cases}$$

where λ is the slope in the regression of $\ln I_y$ for the most recent n years, k_1 and k_2 are *gain* parameters and γ actions asymmetry so that decreases in the index do not result in the same relative change as as an increase.

Figures

- Fig 1, Example OM
- Fig 2, shows that the stock collapses if K1<K2
- Fig 3,4,5 Calibration regressions, showing K1 and K2 as a function of yield and AAV. These shows that there is a trade-off between yield and AAV, i.e. fix K1 at 1 and keep K2 ~1 if you want high yield, if you want to reduce AAV at the expense of yield then decrease K2.
- Fig 5 for the 7 year regression may look a bit wonky, but I think I can explain it. If the stock varies as a sin wave with a wavelength < 14 year, when you start at the minimum you have <7 years with an upward trend. As you move forward they will be a tipping point and the regression slope will suddenly change sign. If you allow big changes in catch increases (k2) you will get high yield but at the expense of high variability. If you only allow small changes in k2 then you will lose yield.
- Fig 6 are the Pareto frontiers, which shows the trade-offs between yield and AAV
- Fig 7 are the MSE summaries that confirm the calibrations