

Unnecessarily Complicated Research Title

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Abstract

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1. Introduction

- Risk and uncertainty
- One rule for all?
- Impact of life histories
- Comparison of constant catch v changing catch based on trends in an empirical index.
- catch only - need good catch data if you havent got this how can you set catch limits

Sustainability and risks to non target exploited marine fish stock populations requires both estimates of current stock status, the effects of fishing pressure (catchability and fishing effort) and the effects of management measures on target populations, however these data are often lacking. Subsequently there is increasing concern and a growing need for the development of innovative approaches so that management of all marine stocks not just those of high commercial value can be included into the Common Fisheries Policy (CFP [1]) framework. Under the CFP management objectives are to recover stocks and to maintain stocks within safe biological limits to levels that can produce Maximum Sustainable Yield (MSY), including by-catch species by 2015 (Implementation Plan adopted at the World Summit on Sustainable Development, Johannesburg in 2002) and no later than 2020. These conservation objectives are currently being achieved by introducing

22 biological target (can fluctuate around targets) and limit (i.e must not be
 23 exceeded) reference points e.g. population size (stock biomass) and/or yields
 24 (catches) and/or long-term yields and fishing mortality against which the
 25 preservation of stocks within such limits are assessed. These targets or limit
 26 reference points are often referred to as harvesting strategies which include
 27 an operational component called a harvest control rule (HCR) that are based
 28 on indicators (e.g. monitoring data or models) of stock status and to prevent
 29 overfishing.

30 The International Council for the Exploration of the Sea (ICES) cate-
 31 gorises stocks in to classes *data-rich*, (categories 1 and 2) i.e those that have
 32 a quantitative assessment based on conventional methods that require large
 33 amounts of data that include a long historical time series of catches and
 34 sound biological information [2]; or *data-limited* [3](categories 3 and 4) (of-
 35 ten called data poor) those without assessment, forecasts and have limited
 36 funding for research. For data-rich stocks ICES uses two types of reference
 37 points for providing fisheries advice;

- 38 1. Precautionary Approach (PA) reference points (those relating to stock
 39 status and exploitation relative to precautionary objectives) and
- 40 2. MSY reference points (those relating to achieving MSY)

41 In contrast for data limited stocks MSY *proxy* reference points are used to
 42 estimate stock status and exploitation. Often many of the methods used to
 43 estimate MSY proxy reference points require length based inputs as they are
 44 cheap, easy to collect [4] and are related to life history parameters such as
 45 fish size, mortality and fecundity as well as fishery selectivity. For example
 46 many methods are being developed to estimate MSY, but currently only 4 are
 47 approved by ICES, these include, Surplus Production model in Continuous
 48 Time (catch based) (SPiCT; [5], Mean Length Z (MLZ; [6]), Length Based
 49 Spawner Per Recruit (LBSPR; [7]) and Length Based Indicators (LBI; e.g.
 50 [8]). The aforementioned data limited procedures have differing data require-
 51 ments, intended uses and obviously have their own strengths and weaknesses.

52 To test the performance of candidate management procedures often re-
 53 quires evaluation of alternative hypothesis about the dynamics of the system
 54 e.g. population dynamics (life history dynamics such as growth parameters
 55 which are an indication of fishery exploitation levels and management) and
 56 the behaviour of the fishery (e.g range contraction and density dependence)
 57 etc.. Due to the nature of conflicting objectives, stakeholder interests and the
 58 uncertainty in the dynamics of the resource and/ or the plausibility of alter-

native hypotheses can lead to poor decision making and can be problematic when defining management policy.

An intense area of work being researched over the last 2 decades is Management Strategy Evaluation (MSE), which focuses on the broader aspects of fishing (the Ecosystem) whereby different management options are tested against a range of objectives (see [9]). The approach is not to come up with a definitive answer, but to lay-bare the trade offs associated with each management objective, along with identifying and incorporating uncertainties in the evaluation and communicating the results effectively to client groups and decision-makers. MSE is not intended to be complex but to provide a robust framework that account for conflicting poorly defined objectives and uncertainties that have been absent in conventional management [9].

To assess case specific harvest strategies (via simulation) within the MSE, we will implement a management procedure based on a empirical HCR that adjusts yield depending on stock status for a given range set of hyper-parameters for each of the harvest strategies and to test their robustness to uncertainty. This approach could also help identify similar conditions across species where particular advice rules are likely to work well, and where they perform poorly for a given a set of hyper-parameters. Often empirical harvest control rules require extensive exhaustive parameter searches to tune or optimise hyper-parameters that aren't directly learnt from estimators. This requires a technique known as a grid search that extensively searches for all combinations of all parameters. In contrast, and some what less time consuming alternative and efficient parameter search strategies can be considered for a given range of parameter space and a known distribution. As such a random sample can be obtained and used to perform the different experiments for parameter optimisation. Assessment is made as to the performance of each HCR via a set of utilities: safety ($B/B_{MSY} > 1$), yield ($yield/MSY$), kobe proportion (proportion of years that stay in the green zone of kobe plot ($B/B_{MSY} > 1$), and Yield Annual Variation (yield changes by 10% year on year).

2. Material and Methods

2.1. Materials

Fishnets

93 *2.2. Methods*

94 FLife and MSE

95 *2.2.1. Operating Model*

96 Age based.

97 *2.2.2. Management Procedure*

98 The management procedure was based on an empirical MP, where an
99 increase in an index of abundance resulted in an increase in the TAC, while
100 a decrease in the index results in a decrease in the TAC.

101 *2.2.3. Random Search*

102 When running an MSE commonly a set of MP scenarios are run to tune
103 the MP, this requires running the MSE for each OM scenario for a range of
104 fixed values in the HCR and then choosing the rule that best meets manage-
105 ment objectives. If there are a lot of parameters to tune then a grid search
106 may become unfeasible. An alternative is random search [10] as randomly
107 chosen trials are more efficient for parameter optimisation than trials based
108 on a grid.

109 **3. Results**

110 Estimates of the simulated life history parameters obtained from Fishbase
111 are presented in Fig.1. These show that for fast growing species which are
112 small in size l_{∞} (asymptotic length parameter) species such as sprat, the
113 growth parameter k is high. The sprats age-at-50%-maturity $a50$ are low,
114 in contrast to a slower growing larger longer lived species l_{∞} such as rays or
115 pollack.

116 Observations resulting from Fig.2 shows that maturity is related to se-
117 lectivity and that the faster growing species are more susceptible to fishing,
118 although the slower growing larger (by mass and length) species (e.g. pollack
119 has a higher natural mortality rate at lower ages) with the most significant
120 rate increases associated with turbot. A levelling off in the mortality rate is
121 evident for ray just prior to age 4.5. In contrast, there have been less steep
122 declines in natural mortality estimates for brill, but most notably for sprat.

123 Fig.3 displays the simulated time series resulting from the operating
124 model in relation to reference points by species with the trends in annual
125 stock estimates in relation to (bmsy). The relationships were compared via

126 their statistical significance (see Table 3), using scatter plots. To highlight
 127 the changes in catchability over the 20-year period the data were divided
 128 equally into two 10-year time periods (Fig. 3, Fig. 4), and Fig. 5 reflects
 129 the mean indices over time, which is a summary of the indices presented in
 130 Fig. 3 and Fig. 4. Several spatial indices of fisher behaviour were corre-
 131 lated with estimates of catchability for both tuna species but, as expected,
 132 the nature and strength of these relationships differed between associated
 133 and unassociated sets and the timescales of the fishing behaviour (trip-level
 134 versus annually averaged).

135 **[EXAMPLES TO BE UPDATED]**

- 136 • Figure 3 shows the life history parameters
- 137 • Figure 2 shows the vectors
- 138 • Figure 4 shows the time series relative to reference points
- 139 • Figure ?? shows the performance statistics; points are
 - 140 1.
 - 141 2.
 - 142 3.
 - 143 4.
 - 144 5. Figure 5 shows the utility functions for the seven study stocks
 - 145 points area
- 146 1.
- 147 2.
- 148 3.
- 149 4.

150 4. Discussion

- 151 • Bullet point one
- 152 • Bullet point two

153 5. Conclusions

- 154 • Bullet point one
- 155 • Bullet point two

6. References

References

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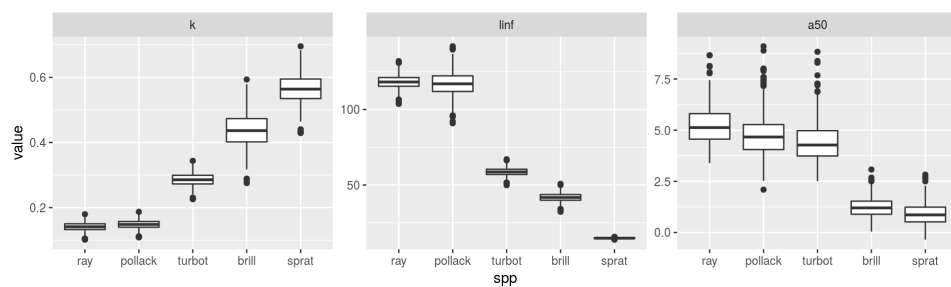


Figure 1:

- 188 [10] J. Bergstra, Y. Bengio, Random search for hyper-parameter optimiza-
 189 tion, *Journal of Machine Learning Research* 13 (2012) 281–305.

190 7. Figures

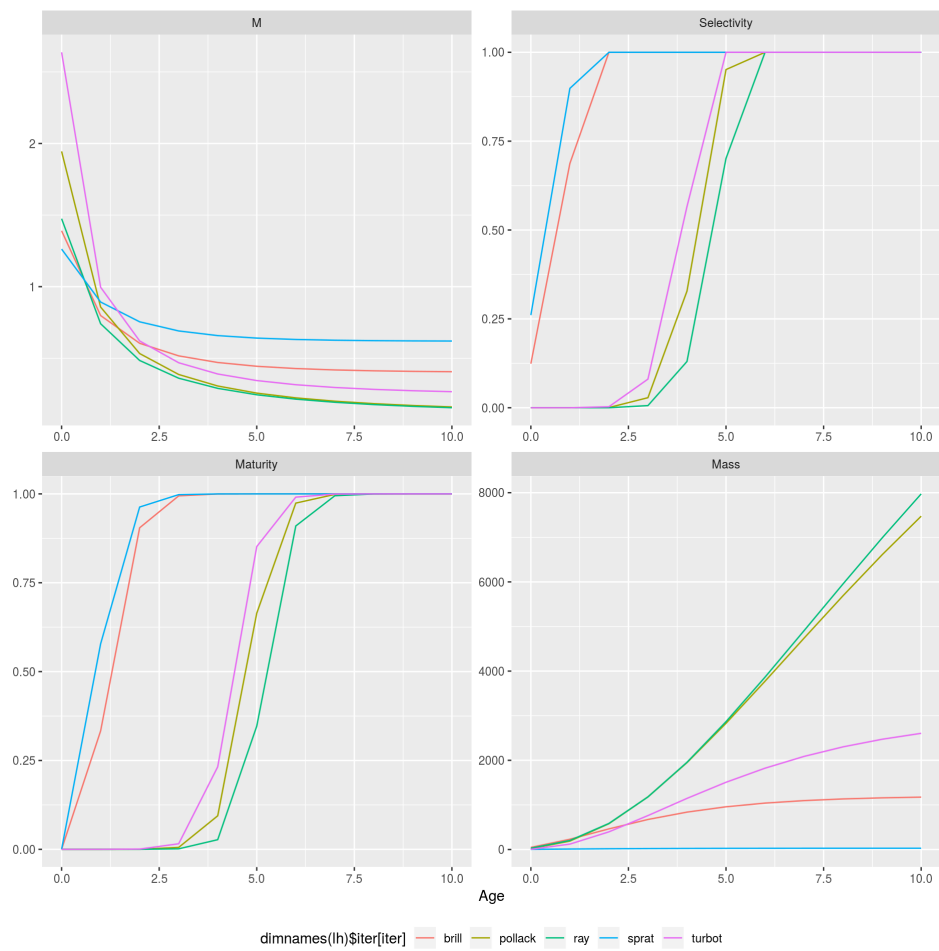
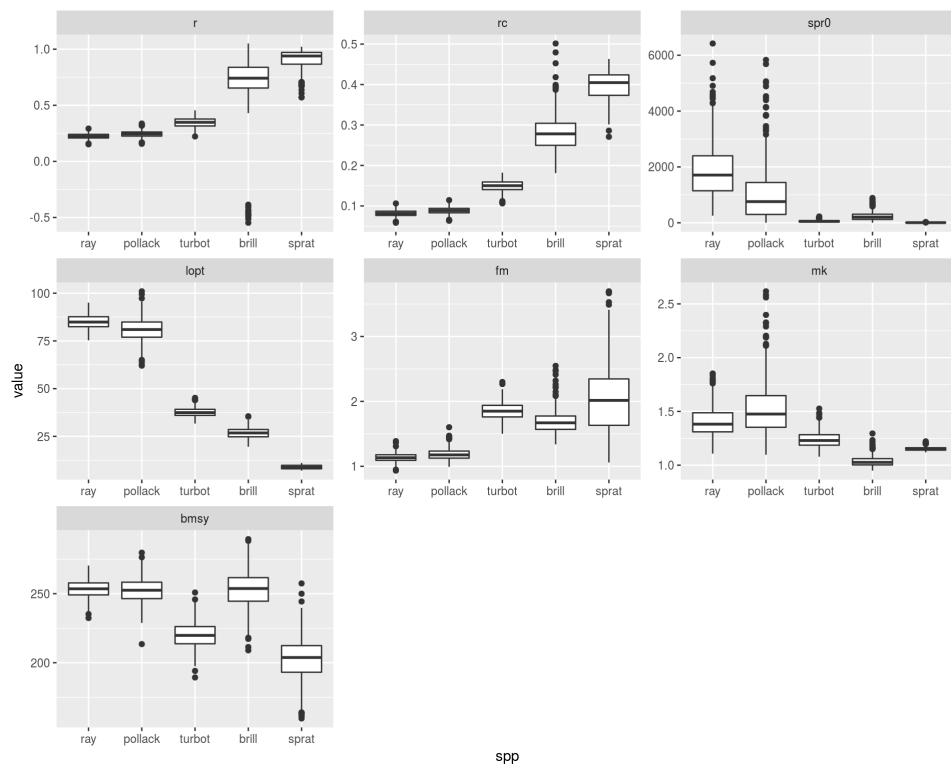


Figure 2:



spp

Figure 3:

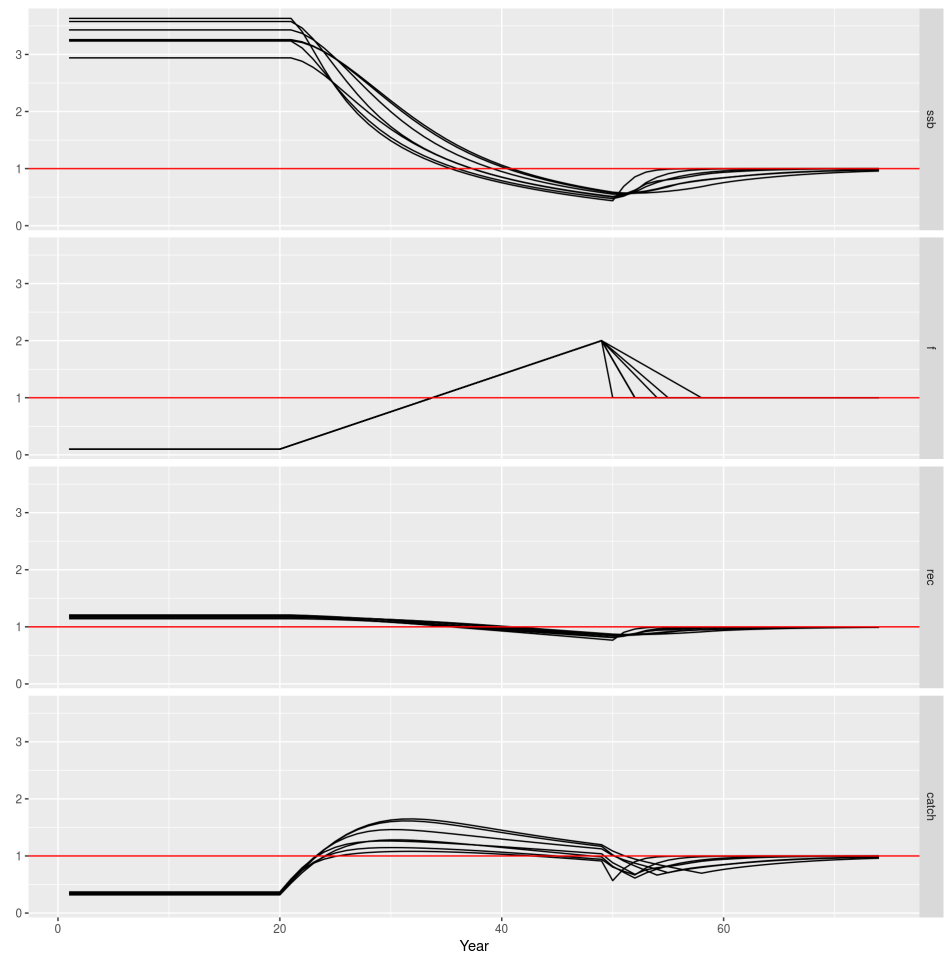


Figure 4:

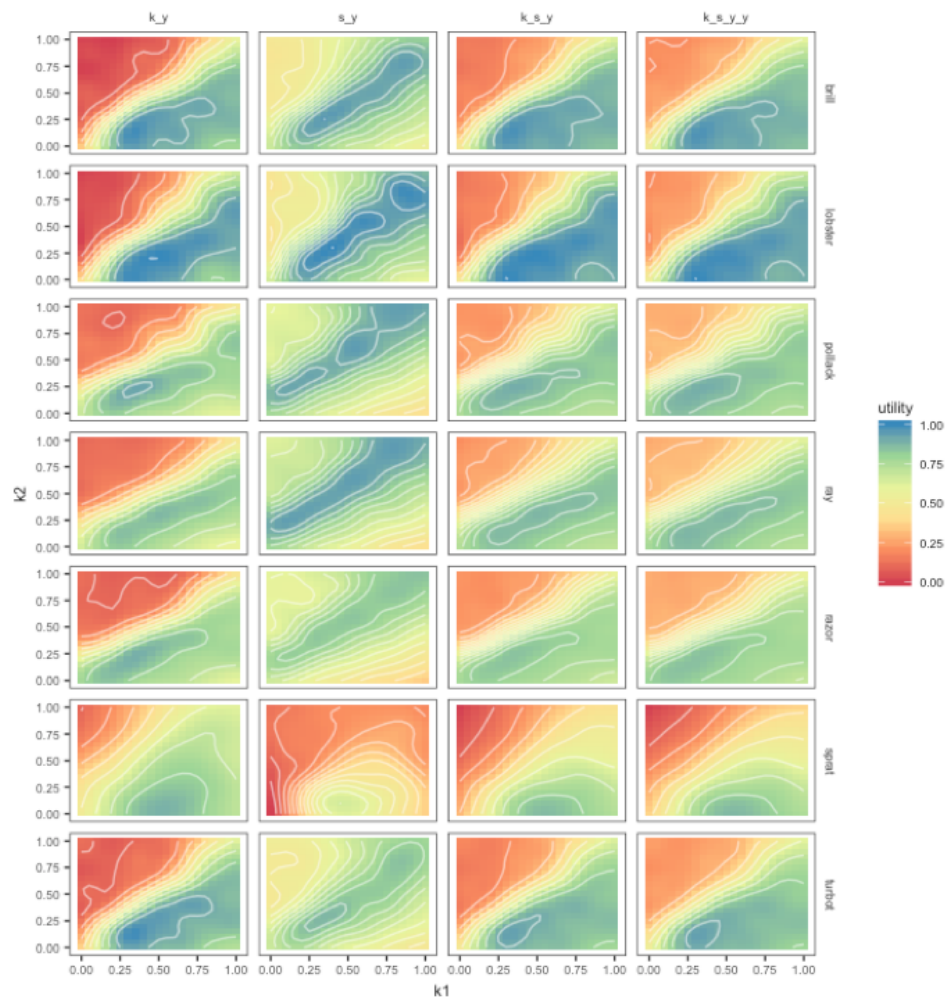


Figure 5: