

# PLOS ONE

## Entanglement rates and population trends of Steller (*Eumetopias jubatus*) and California (*Zalophus californianus*) sea lions on the north coast of Washington state

--Manuscript Draft--

<b>Manuscript Number:</b>	
<b>Article Type:</b>	Research Article
<b>Full Title:</b>	Entanglement rates and population trends of Steller ( <i>Eumetopias jubatus</i> ) and California ( <i>Zalophus californianus</i> ) sea lions on the north coast of Washington state
<b>Short Title:</b>	Sea lion entanglement rates and population trends on the north coast of Washington state
<b>Corresponding Author:</b>	Elizabeth Allyn Makah Tribe Neah Bay, WA UNITED STATES
<b>Keywords:</b>	Sea lion; Pinniped; <i>Zalophus californianus</i> ; <i>Eumetopias jubatus</i> ; Entanglement; Fisheries; Marine Debris; Population trends; Marine Pollution
<b>Abstract:</b>	<p>Entanglements affect marine mammal species around the globe, and for some, those impacts are great enough to cause population declines. This study aimed to document rates and causes of entanglement in Steller and California sea lions on the north coast of Washington from 2010-2018 and to determine if entanglements caused population impacts. We conducted small boat surveys to count sea lions and document entangled individuals. Rates of entanglement and entangling material occurrence were compared with records of stranded individuals on the Washington and Oregon coast and with packing bands recorded during beach debris surveys. California sea lions experienced a higher rate of entanglement than Steller sea lions (2.13% and 0.41%, respectively). The age composition of entangled Steller sea lions was 77% adults (32.4% male, 63.3% female), 17.1% juveniles, 5.9% unknown age, and no pups. All entangled California sea lions were adult males except for one juvenile. Steller sea lion entanglements showed no seasonality, but California sea lions experienced an entanglement rate peak in June and July. The majority of identifiable entanglements were packing bands, followed by salmon flashers, which only occurred in June – September during the peak of the local ocean salmon troll fishery, and monofilament line. Counts of packing bands in debris surveys correlated with entanglements caused by packing bands observed on haulouts (Pearson's <math>R=0.81</math>). However, no packing band entanglements were observed in the stranding record, and the overall proportion of stranded animals exhibiting evidence of entanglement was lower than expected. During the study period, Steller sea lions exhibited a <math>7.9\% \pm 3.2</math> rate of increase at the study haulouts, which was similar to that seen in California sea lions (<math>7.8\% \pm 4.2</math>), suggesting that the high observed entanglement rates did not have population level consequences, though they are still a welfare issue for individual sea lions.</p>
<b>Order of Authors:</b>	Elizabeth Allyn Jonathan Scordino
<b>Opposed Reviewers:</b>	
<b>Additional Information:</b>	
<b>Question</b>	<b>Response</b>
<b>Financial Disclosure</b>  Enter a financial disclosure statement that describes the sources of funding for the work included in this submission. Review the <a href="#">submission guidelines</a> for detailed	Field research was funded by three grants awarded to author JS from the National Marine Fisheries Service through their Species Recovery Grants to Tribes ( <a href="https://www.fisheries.noaa.gov/grant/species-recovery-grants-tribes">https://www.fisheries.noaa.gov/grant/species-recovery-grants-tribes</a> ). Award numbers were: NA10NMF4720372, NA13NMF4720121, and NA16NMF4720059. Award NA16NMF4720059 funded analysis and writing of the manuscript. The National Marine Fisheries Service had no role in study design, data collection and analysis, decision to

<p>requirements. View published research articles from <a href="#">PLOS ONE</a> for specific examples.</p> <p>This statement is required for submission and <b>will appear in the published article</b> if the submission is accepted. Please make sure it is accurate.</p> <p><b>Unfunded studies</b> Enter: <i>The author(s) received no specific funding for this work.</i></p> <p><b>Funded studies</b> Enter a statement with the following details:</p> <ul style="list-style-type: none"> <li>• Initials of the authors who received each award</li> <li>• Grant numbers awarded to each author</li> <li>• The full name of each funder</li> <li>• URL of each funder website</li> <li>• Did the sponsors or funders play any role in the study design, data collection and analysis, decision to publish, or preparation of the manuscript?</li> <li>• <b>NO</b> - Include this sentence at the end of your statement: <i>The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.</i></li> <li>• <b>YES</b> - Specify the role(s) played.</li> </ul> <p>* typeset</p>	<p>publish, or preparation of the manuscript.</p>
<p><b>Competing Interests</b></p> <p>Use the instructions below to enter a competing interest statement for this submission. On behalf of all authors, disclose any <a href="#">competing interests</a> that could be perceived to bias this work—acknowledging all financial support and any other relevant financial or non-financial competing interests.</p> <p>This statement <b>will appear in the published article</b> if the submission is accepted. Please make sure it is accurate. View published research articles from <a href="#">PLOS ONE</a> for specific examples.</p>	<p>EA and JS both work for the Makah Fisheries Management Department of the Makah Tribe. Funding for this project was provided by the Species Recovery Grant to Tribes that is administered by the National Marine Fisheries Service. EA and JS also receive funding from the Bonneville Power Administration Tribal Capacity Building Grant, the National Science Foundation through a subaward granted to the University of Chicago, and the John H. Prescott Marine Mammal Rescue Assistance Grant administered by the National Marine Fisheries Service. Author JS has also received funding support through the Saltonstall-Kennedy Grant Program, the Bycatch Reduction Engineering Program, the Preserve America Grant, and a NOAA Cooperative Research Grant which are all administered by the National Marine Fisheries Service; the Bureau of Indian Affairs; Washington SeaGrant; and the North Pacific Coast Marine Resource Committee. Funders played no role in the development, implementation, or analysis of the work completed for this manuscript.</p>

### NO authors have competing interests

Enter: *The authors have declared that no competing interests exist.*

### Authors with competing interests

Enter competing interest details beginning with this statement:

*I have read the journal's policy and the authors of this manuscript have the following competing interests: [insert competing interests here]*

\* typeset

### Ethics Statement

Enter an ethics statement for this submission. This statement is required if the study involved:

- Human participants
- Human specimens or tissue
- Vertebrate animals or cephalopods
- Vertebrate embryos or tissues
- Field research

Write "N/A" if the submission does not require an ethics statement.

General guidance is provided below.

Consult the [submission guidelines](#) for detailed instructions. **Make sure that all information entered here is included in the Methods section of the manuscript.**

The National Marine Fisheries Service reviewed and approved our research methodologies and granted Marine Mammal Protection Act research permits 14326, 13430, and 19430. We also obtained Special Use Permits from the United States Fish and Wildlife Service for all land-based survey activities conducted on haulouts within the Flattery Rocks National Wildlife Refuge.

### Format for specific study types

#### Human Subject Research (involving human participants and/or tissue)

- Give the name of the institutional review board or ethics committee that approved the study
- Include the approval number and/or a statement indicating approval of this research
- Indicate the form of consent obtained (written/oral) or the reason that consent was not obtained (e.g. the data were analyzed anonymously)

#### Animal Research (involving vertebrate animals, embryos or tissues)

- Provide the name of the Institutional Animal Care and Use Committee (IACUC) or other relevant ethics board that reviewed the study protocol, and indicate whether they approved this research or granted a formal waiver of ethical approval
- Include an approval number if one was obtained
- If the study involved *non-human primates*, add *additional details* about animal welfare and steps taken to ameliorate suffering
- If anesthesia, euthanasia, or any kind of animal sacrifice is part of the study, include briefly which substances and/or methods were applied

#### Field Research

Include the following details if this study involves the collection of plant, animal, or other materials from a natural setting:

- Field permit number
- Name of the institution or relevant body that granted permission

#### Data Availability

Authors are required to make all data underlying the findings described fully available, without restriction, and from the time of publication. PLOS allows rare exceptions to address legal and ethical concerns. See the [PLOS Data Policy](#) and [FAQ](#) for detailed information.

Yes - all data are fully available without restriction

A Data Availability Statement describing where the data can be found is required at submission. Your answers to this question constitute the Data Availability Statement and **will be published in the article**, if accepted.

**Important:** Stating 'data available on request from the author' is not sufficient. If your data are only available upon request, select 'No' for the first question and explain your exceptional situation in the text box.

Do the authors confirm that all data underlying the findings described in their manuscript are fully available without restriction?

**Describe where the data may be found in full sentences. If you are copying our sample text, replace any instances of XXX with the appropriate details.**

- If the data are **held or will be held in a public repository**, include URLs, accession numbers or DOIs. If this information will only be available after acceptance, indicate this by ticking the box below. For example: *All XXX files are available from the XXX database (accession number(s) XXX, XXX).*
- If the data are all contained **within the manuscript and/or Supporting Information files**, enter the following:  
*All relevant data are within the manuscript and its Supporting Information files.*
- If neither of these applies but you are able to provide **details of access elsewhere**, with or without limitations, please do so. For example:

*Data cannot be shared publicly because of [XXX]. Data are available from the XXX Institutional Data Access / Ethics Committee (contact via XXX) for researchers who meet the criteria for access to confidential data.*

*The data underlying the results presented in the study are available from (include the name of the third party*

All data underlying the results presented in the study are available from Mendeley Data (DOI: 10.17632/447sm2rwrk.1).

<p><i>and contact information or URL).</i></p> <ul style="list-style-type: none"><li>• This text is appropriate if the data are owned by a third party and authors do not have permission to share the data.</li></ul> <p>* typeset</p>	
Additional data availability information:	

To Whom It May Concern,

Jonathan Scordino and I are pleased to provide for your review for publication in PLOS ONE a research article on the population trends and entanglement rates of sea lions in northern Washington titled, "Entanglement rates and population trends of Steller (*Eumetopias jubatus*) and California (*Zalophus californianus*) sea lions on the north coast of Washington state." This study presents population trends, entanglement rates, and entangling material proportions for two pinniped species in a region with no previous studies of entanglements. Entanglement rates are contextualized through comparison to trends in counts conducted at sea lion haulouts and through comparison to the stranding record to understand the impacts of entanglement on sea lion health and survival. Entanglement rates and material proportions are also compared to beach debris surveys conducted along the Washington coast to analyze patterns in material occurrence and entanglement susceptibility. We also situate our entanglement rates and material proportions in existing literature through a literature review of otariid entanglement rates globally. The entanglement rates observed in this study are the highest recorded rates for Steller sea lions and the second highest recorded rates for California sea lions globally. The work in this manuscript has not been submitted or published elsewhere and is the original work of the authors.

Appropriate Academic Editors to handle this manuscript in order of relevance would be Robert Schick, Songhai Li, or Andrew Hoskins.

Thank you for your consideration of this research article. We think it will make an important contribution to the current knowledge on pinniped entanglement and fishery interactions, as well as aid in the development of more targeted mitigation strategies. Furthermore, I am very excited to submit this manuscript as it is my first time submitting my work for peer-reviewed publication.

Respectfully,

Elizabeth Allyn

1 Entanglement rates and population trends of Steller (*Eumetopias*  
2 *jubatus*) and California (*Zalophus californianus*) sea lions on the north  
3 coast of Washington state

4

5 Elizabeth Allyn<sup>1\*</sup> and Jonathan Scordino<sup>1</sup>

6 <sup>1</sup>Marine Mammal Program, Fisheries Management Department, Makah Tribe, Neah Bay, Washington,  
7 United States of America

8

9 \* Corresponding author

10 Email: liz.allyn@makah.com (EA)



## Abstract

Entanglements affect marine mammal species around the globe, and for some, those impacts are great enough to cause population declines. This study aimed to document rates and causes of entanglement in Steller and California sea lions on the north coast of Washington from 2010-2018 and to determine if entanglements caused population impacts. We conducted small boat surveys to count sea lions and document entangled individuals. Rates of entanglement and entangling material occurrence were compared with records of stranded individuals on the Washington and Oregon coast and with packing bands recorded during beach debris surveys. California sea lions experienced a higher rate of entanglement than Steller sea lions (2.13% and 0.41%, respectively). The age composition of entangled Steller sea lions was 77% adults (32.4% male, 63.3% female), 17.1% juveniles, 5.9% unknown age, and no pups. All entangled California sea lions were adult males except for one juvenile. Steller sea lion entanglements showed no seasonality, but California sea lions experienced an entanglement rate peak in June and July. The majority of identifiable entanglements were packing bands, followed by salmon flashers, which only occurred in June – September during the peak of the local ocean salmon troll fishery, and monofilament line. Counts of packing bands in debris surveys correlated with entanglements caused by packing bands observed on haulouts (Pearson's  $R=0.81$ ). However, no packing band entanglements were observed in the stranding record, and the overall proportion of stranded animals exhibiting evidence of entanglement was lower than expected. During the study period, Steller sea lions exhibited a  $7.9\% \pm 3.2$  rate of increase at the study haulouts, which was similar to that seen in California sea lions ( $7.8\% \pm 4.2$ ), suggesting that the high observed entanglement rates did not have population level consequences, though they are still a welfare issue for individual sea lions.

## 32 Introduction

33 The prevalence of man-made marine debris is of global concern and has been gaining attention from  
34 media, researchers, and the public in recent decades as the impact to marine life becomes better  
35 understood [1–5]. Many marine organisms are affected by marine debris through entanglement.  
36 Instances of entanglement have been recorded for at least 32 species of marine mammals globally [4],  
37 and for some, like the northern fur seal (*Callorhinus ursinus*) and the endangered Hawaiian monk seal  
38 (*Monachus schauinslandi*), entanglement was thought to have contributed to population declines [6–8].  
39 For pinnipeds specifically, entanglement has been documented for more than half of the existing species  
40 [2,4,9]. In this study, entanglement is defined as the presence of marine debris attached to an animal's  
41 body, including materials that are looped around the appendages, torso, or neck (e.g. netting or packing  
42 bands) and instances where materials are internally or externally embedded (e.g. hooking injuries).

43 The mechanisms by which an animal becomes entangled are almost as varied as the entangling  
44 materials themselves. Entangling debris can come from terrestrial and marine pollution, and from  
45 derelict and active fishing gear. Any marine debris that form loops that can ensnare or sharp objects that  
46 can embed, such as hooks, pose an entanglement risk. The mechanism of entanglement can often be  
47 determined by identifying the entangling material. Packing bands, rubber bands, and monofilament line  
48 are likely encountered passively as debris, while net fragments can be a sign of either passive  
49 encounters with floating derelict gear or a sign of interaction with an actively fished net. Salmon  
50 flashers and other hook and line gear are likely encountered as actively fished gear and are evidence of  
51 fishery depredation behaviors, which cause harm both to the entangled animal and to the fisher's catch  
52 [10,11]. Otariids are especially curious of novel objects, and can become entangled in debris while  
53 attempting to explore or play with them [12,13]. The factors leading to entanglement in any given  
54 location are therefore governed by both local and regional dynamics, as ocean currents, upwelling

patterns, fishing effort and gear types, prey distributions, abundance of pinnipeds, and marine traffic patterns all may contribute to both the distribution of entangling materials and the behavior of pinnipeds in the area [9,14–17].

The objective of this study was to characterize the rates and causes of entanglement in Steller (*Eumetopias jubatus*) and California (*Zalophus californianus*) sea lions in northern Washington state and to evaluate if the observed entanglements were negatively impacting the populations. We described temporal trends in entanglement occurrence and determined the most commonly observed entangling materials. Based on previous studies, we expected to see entanglements caused by mainly packing bands and netting [1,4,9,18–20]. We expected little change in annual entanglement occurrence but anticipated that there would be a peak in entanglements observed in the mid- to late-summer months due to these being the peak months for recreational and commercial fishing effort. We also compared entanglement rates with beach debris survey data to discern patterns in entanglement occurrence due to material availability, and with the stranding record to briefly explore the impacts of entanglement on health and survival. Understanding the patterns behind entanglement occurrence will enable the development of more targeted prevention and response efforts and a more accurate understanding of the impacts of entanglement on local populations.

## Methods

### Data Collection

The National Marine Fisheries Service reviewed and approved our research methodologies and granted Marine Mammal Protection Act research permits 14326, 13430, and 19430. We also obtained Special Use Permits from the United States Fish and Wildlife Service for all land-based survey activities conducted on haulouts within the Flattery Rocks National Wildlife Refuge.

Observations of hauled out Steller and California sea lions were carried out from small boats along the north coast of Washington from 2010 –2018 focusing on four major haulout complexes (Fig 1). Occasionally, surveyors were landed on haulouts to conduct these surveys. Surveys were conducted year-round with more effort from late spring through early fall due to availability of survey days with suitable weather and sea conditions. Surveys often did not include all haulouts during a day due to logistical challenges such as sea conditions and daylight. During surveys, we counted actively entangled individuals and individuals showing evidence of past entanglement (e.g. scarring) and counted the total abundance of the two sea lion species at each haulout. We attempted to photograph all entangled sea lions and those that appeared entangled with a digital SLR camera with a 100-400 mm lens for later assessment. Entangled individuals encountered along the survey route in locations other than the four major haulout complexes were excluded from entanglement rate calculations due to the lack of reliable and regular total counts of hauled individuals but were still photographed to identify the source and nature of the injury. Entanglement and count data are publicly available through Mendeley Data [21].

*Fig 1: Map of the four major Steller and California sea lion haulout complexes surveyed for entangled individuals: Tatoosh Island, the Bodelteh Islands, Carroll Island, and Sea Lion Rock.*

## Population Trends

Population trends were calculated using a three-step process. First, for each species we pooled the counts from the four major haulout complexes on days when all four haulouts were visited. Next, we averaged all complete survey days within a month for a monthly average. Last, we averaged the mean monthly counts for an annual estimate of the average number of Steller and California sea lions using the four major haulout complexes over the study duration. The observed change in annual counts were calculated for each year using the formula

$$r_t = \ln(N_{t+1}/N_t)$$

where  $r_t$  is the realized per capita rate of population change,  $t$  is year, and  $N$  is the average count for the year. The annual rates of change were then averaged over all study years to produce the overall average rate of change in haulout counts for each species. We excluded 2018 data from the analysis because there were no survey days that covered all four haulout sites after June, potentially biasing the counts by not including the full range of seasonal variation (Table 1).

*Table 1: The number of surveys conducted in each month of the study period 2010-2018 with the number of complete surveys where all four major haulout complexes were visited in parentheses. Note that no complete surveys were conducted after June in 2018.*

	Month												Total
	1	2	3	4	5	6	7	8	9	10	11	12	
2010		1	1	1	3(1)	2	3(1)	5(3)	8(2)	2	2(1)		28(8)
2011	1	2(1)	2	4(1)	6(3)	5(2)	4(2)	6(2)	3(2)	4(1)	3(1)		40(15)
2012		2	2(1)	2(1)	3(2)	5(4)	8(1)	4(2)	5(2)	2(1)	3(1)	2(1)	38(16)
2013	2(1)	1(1)	2(1)	2(1)	3(1)	4(3)	3(2)	3(2)	3(1)	3	2		28(13)
2014				2	2(1)	3(2)	4(2)	4(1)	4(3)	2			21(9)
2015	3(2)	2(1)	1	3	2	2	4(1)	5(2)	4(1)	1	1		28(7)
2016	1(1)		4	1	5(3)	1(1)	4(2)	4(2)	3(3)	3(1)		1(1)	27(14)
2017	1(1)	2(1)	1			3(3)	1	4(3)	3(1)	1(1)			16(10)
2018		1	3(1)	2(1)	1(1)	3(2)	3	3	3	1			20(5)
<b>Total</b>	<b>8(5)</b>	<b>11(4)</b>	<b>16(3)</b>	<b>17(4)</b>	<b>25(12)</b>	<b>28(17)</b>	<b>34(11)</b>	<b>38(17)</b>	<b>36(15)</b>	<b>19(4)</b>	<b>11(3)</b>	<b>3(2)</b>	<b>246(97)</b>

## Entanglement Rates

Our goal was to calculate an average annual entanglement rate for California and Steller sea lions for the northern Washington coast. Our survey effort was greatest during the summer and early fall when sea conditions were most predictable (Table 1). In order to ensure that our calculated entanglement rate was representative of the year, and not biased to time periods when we had more surveys, we calculated average yearly entanglement rates using a multistep process. Counts of the total number of

individuals hauled out and counts of entangled individuals, including both active and inactive entanglements taken from photographs and survey notes, were pooled across haulout complexes within survey days, and an entanglement rate was calculated for each survey day by dividing the total number of entangled individuals by the total count. Average entanglement rates were then calculated for each month of the nine-year study period. The mean rates for each month of the study were then averaged across years for each month and across months for each year to discern seasonal and annual patterns, which were analyzed using single-factor ANOVA and Tukey-Kramer post-hoc tests. An overall average entanglement rate was calculated for each species by taking the average of the monthly mean entanglement rates. We used a paired two-tailed t-test to compare the average entanglement rate for Steller and California sea lions with monthly averages ( $n=12$ ) as our sampling unit. We conducted a literature review to catalog published entanglement rates for California and Steller sea lions along with other otariid species to provide a comparison to our calculated rates.

## Photo Analysis

We assessed photographs of sea lions with evidence of entanglement to determine if the entanglement was active or inactive, identify the entangling material, and record the age and sex of the entangled individual. Entangled individuals were assigned to demographic groups by age as adult, juvenile, pup, or unknown, and by sex for adults based on a number of physical characteristics, including body size and shape, whisker length, and presence of secondary sexual features. The proportion of entangled individuals in each sex and age class were calculated.

Entangling materials were identified to one of nine categories: packing band, salmon flasher, rubber band, monofilament line, hook and line, netting, rope, scar, or unknown. Salmon flashers are plastic or metal attractants attached to a line with a 60 – 200cm leader ahead of the lure or baited hook. The hook from the lure or baited hook is often swallowed leaving the flasher to dangle out of the mouth by the

leader. The hook and line category included fishing lures (not attached to flashers) and longline gear, both of which are found hooked externally on entangled individuals. Rubber bands are thick black bands cut from truck tire inner tubes that are often used in crab fisheries to secure trap doors. Packing bands are thin plastic strips attached at the ends to form loops that are used to increase the integrity of containers generally made of cardboard. The netting category included both gillnets made of monofilament line and trawl netting made of nylon or synthetic lines. Monofilament lines are commonly used in recreational fisheries and for leaders in commercial salmon fisheries and were differentiated from gillnets by the absence of knotted webbing. Active entanglements where the material could not be identified were recorded as 'Unknown'. Animals with evidence of a previous entanglement but where no debris was observed on the sea lion were recorded as 'Scar'. The proportion of entanglements that were active or inactive and the proportion exhibiting each entangling material were summarized and reported over months and years to observe trends in material occurrence.

### Packing Band Analysis

Annual packing band entanglement occurrence was compared to data from marine debris surveys conducted by the Olympic Coast National Marine Sanctuary (OCNMS) to discern patterns in material availability in the environment. The year 2018 was excluded from annual trend analysis due to low sea lion survey effort after the month of June. OCNMS conducted 1,548 beach debris surveys in the Olympic Coast region from 2012-2017, covering 17 beaches in Washington State, from Roosevelt Beach (47.1770° N, 124.1972° W) to Wa'atch Beach (48.3441° N, 124.6792° W). Surveys were conducted by volunteers in an OCNMS citizen science program adhering to standardized debris monitoring procedures developed by NOAA's Marine Debris Program [22]. The number of packing bands encountered each year in beach debris surveys was divided by the total number of surveys conducted in that year to correct for variation in survey effort. The annual proportion of entanglements caused by packing bands observed during surveys was analyzed for correlation with the number of packing bands per beach debris survey.

## Stranding Analysis

The West Coast Marine Mammal Stranding Network, overseen by the West Coast Regional Office of NOAA's Protected Resources Division, has recorded opportunistic sightings of marine mammal strandings since the early 1980's. Data on Steller and California sea lions that stranded dead on the Washington and Oregon coast from 2010-2018 were analyzed to determine the occurrence of stranded individuals bearing evidence of entanglement. Entanglements were assigned to three categories depending on the nature of the entanglement evidence: animals that stranded with the entangling material still present were marked as 'Active', animals with evidence of entanglement-related injuries without entangling material present were marked 'Scar', and animals showing probable but inconclusive evidence of entanglement were marked 'Possible'. For active entanglements, the entangling material was determined using notes and comments accompanying the stranding record and assigned to one of the categories used during our live surveys (e.g. packing band or flasher). Entanglements marked "Possible" were excluded from summary statistics due to inconsistencies in reporting suspicious lesions as potential entanglement evidence.

## Statistical Analysis

All statistical analyses were conducted in Microsoft Excel. Figures 2-6 were created with R Statistical Program version 3.6.1 using ggplot2 [23,24].

## Results

### Population Trends

There were 92 survey days from 2010-2017 where counts were recorded at all four major complexes. The average annual rate of change at the haulout complexes in northern Washington for Steller sea lion counts was  $7.9\% \pm 3.2$  (95% CI), and for California sea lion counts was  $7.8\% \pm 4.2$  (95% CI; Fig 2).



*Fig 2: Trends in average annual counts of Steller and California sea lions present at four major haulout complexes on the north coast of Washington from 2010-2017.*

## Entanglement Rates

There were 648 active and inactive entanglements observed in the survey area from 2010-2018, 611 (433 Steller, 178 California) of which were documented at the four major haulout complexes. The average entanglement rate for California sea lions (2.13%) was greater than for Steller sea lions (0.41%), but the difference was not statistically significant (Paired t-test,  $df = 11$ ,  $t = 1.41$ ,  $p = 0.19$ ). There were no annual or seasonal trends of statistical significance in entanglement rates for Steller or California sea lions (Fig 3). However, California sea lions experienced high rates of entanglement in 2014, and both species experienced somewhat elevated rates of entanglement in 2015. California sea lions also exhibited some seasonal variability with a peak in entanglement rates in the summer, coinciding with the lowest months for haulout counts (Fig 4). While other months exhibited elevated rates of entanglement (November: 1.5%) or comparatively low average haulout counts (February: 168, March: 218), June and July were the only months to exhibit both low average haulout counts and high entanglement rates (June: 167, 10.2%; July: 35, 12.1%).

*Fig 3: Average entanglement rates (expressed as entanglements per individual) and entangling material proportions for California and Steller sea lions in northern Washington from 2010-2018 by year. Entanglement rate calculations only included entangled individuals observed at one of four major haulout complexes. Entangling materials were only analyzed for individuals with photos of sufficient quality observed hauled out anywhere along the survey route.*

*Fig 4: Average pooled counts at the four major haulouts, average entanglement rates (expressed as entanglements per individual), and entangling material proportions for California and Steller sea lions in northern Washington from 2010-2018 by month. Entanglement rate calculations only included entangled individuals observed at one of four major haulout complexes. Entangling materials were analyzed for any entangled individuals with photos of sufficient quality observed hauled out anywhere along the survey route.*

## Material Analysis

There were 502 (357 Steller, 145 California) sightings of entanglements with photos of a quality sufficient for analysis. For Steller sea lions, active entanglements comprised 77.9% of all entanglements and of those only 55.4% ( $n = 154$ ) were identifiable. The majority of identifiable entanglements were caused by packing bands (67.5%) and salmon flashers (13.6%). Other materials comprised less than 10% of identifiable entanglements: rubber bands (7.8%), monofilament line (6.5%), netting (1.9%), hook and line (1.9%), and rope (0.6%). For California sea lions, 80.0% of all entanglements were active, and 41.4% of active entanglements were identifiable. Packing bands made up the majority of entanglements (70.8%), followed by monofilament line (12.5%), rope (10.4%), and salmon flashers (6.3%). For both species salmon flashers were only observed in the months of June – September coinciding with the local recreational and commercial ocean salmon troll fishery (Fig 4). In all cases where the entangling material could not be identified or was no longer present the entanglement scar or wound was located on the neck, indicating that those entanglements were caused by an encircling material, such as a packing band, rubber band, monofilament line, or netting.

## Sex and Age

For Steller sea lions both the sex and age could be identified for 74.5% of entanglements, and either the sex or the age could be identified for an additional 19.9% for the 357 Steller sea lions analyzed. The age composition of entangled Steller sea lions was 77% adults (32.4% male, 63.3% female), 17.1% juveniles, 5.9% unknown age, and no pups. For the most part, entangling materials were evenly distributed among sex and age classes, but 16.4% of entangled juveniles exhibited a flasher and 11.5% exhibited rubber bands, higher percentages than any other sex or age class grouping (Fig 5). The sex and age could be identified for 98.6% ( $n = 143$ ) of entangled California sea lions, 142 of which were adult males, with one juvenile male. The single juvenile male was entangled in a packing band, and the remaining adult males

exhibited entanglements in the same proportions as what was seen for California sea lions overall in the survey area.

*Fig 5: The proportion of entanglements caused by each material type for Steller sea lion juveniles, adult females, and adult males in northern Washington, 2010-2018.*

## Packing Band Analysis

Annual trends in the proportion of entanglements caused by packing bands from 2012-2017 positively correlated with the annual occurrence of packing bands observed during OCNMS beach debris surveys (Pearson's  $R=0.81$ ; Fig 6).

*Fig 6: The proportion of entanglements caused by packing bands for sea lions at haulouts in northern Washington (primary axis) and the number of packing bands per survey recorded in beach debris surveys along the north Pacific coast of Washington conducted by the Olympic Coast National Marine Sanctuary (secondary axis).*

## Stranding Analysis

There were confirmed stranding records of 551 dead Steller sea lions and 1,048 dead California sea lions on the outer coast of Washington and Oregon from 2010-2018. The proportion of dead strandings exhibiting evidence of entanglement was 1.6% for Steller sea lions and 0.38% for California sea lions. All four entangled California sea lions that stranded dead were adult males. Of the nine dead stranded entangled Steller sea lions, 7 were adults (4 females, 3 males), one subadult, and one unknown. Of the 13 total entanglements observed, five were entangled in salmon flashers and other assorted hook and line gear. There was also a single Steller sea lion entangled in rope, and another exhibiting scars indicative of entanglement. The remaining six records did not have enough detail to determine the status of the entanglement or the entangling material. No sea lions stranded dead were observed entangled in packing bands.

## Discussion

Despite exhibiting high rates of entanglement, populations of both California sea lions and Steller sea lions exhibited high rates of growth in northern Washington, suggesting that entanglements did not affect the population dynamics of either species in this area enough to cause concern. The California sea lions in this study experienced the second highest rate for any otariid population in the published literature and the highest otariid entanglement rate documented in the United States (Table 2). Our observed count increase rate for California sea lions (7.8%) was similar to the range-wide population growth estimate for 1975-2014 (7%) [25]. The entanglement rate observed in this study for Steller sea lions was almost double other published rates [18,26], and the count increase rate calculated for Steller sea lions in this study (7.9%) was close to double the population growth rate observed by Pitcher et al. [27] and the National Marine Fisheries Service [28] for the eastern distinct population segment of Steller sea lions.

*Table 2: A review of pinniped entanglement rates in the published literature in ascending order of entanglement rate.*

*Entanglement rates were calculated using many different methodologies based on many different data collection methods and are not meant to be directly comparable without caution. Species are listed using the first letters of their genus and species: Af - Arctocephalus forsteri, Ag - Arctocephalus gazella, Ap - Arctocephalus pusillus, Apd - Arctocephalus pusillus doriferus, App - Arctocephalus pusillus pusillus, At - Arctocephalus tropicalis, Cu - Callorhinus ursinus, Ej - Eumetopias jubatus, Nc - Neophoca cinerea, Zc - Zalophus californianus.*

Year	Reference	Location	Species	Rate (%)
1988-1989	[20]	Channel Islands, CA	Cu	0.00
1997-2013	[29]	Bass Strait, Australia	App	0.002-0.019
1988-1997	[30]	Livingston Island, Antarctica	Ag	0.024
1996-2002	[31]	Bouvetøya	Ag	0.024-0.059
2001-2005	[1]	Point Reyes, CA	Zc	0.03

1982-1984	[32]	St Paul Island, AK	Cu	0.04~
1985	[26]	Aleutian Islands, AK	Ej	0.07
1978	[33]	Namibia & South Africa	Ap	0.11*
1977	[33]	Namibia & South Africa	Ap	0.12*
1979	[33]	Namibia & South Africa	Ap	0.12*
1991-1996	[34]	Marion Island, Australia	Ag & At	0.15
2006	[35]	Pribilof Islands, AK	Cu	0.17*
1983-1984	[20]	Channel Islands, CA	Zc	0.18
2005	[35]	Pribilof Islands, AK	Cu	0.18*
2006	[35]	St Paul Island, AK	Cu	0.20
1988-2000	[36]	Kangaroo Island, Australia	Nc	0.20
1988-1989	[20]	Channel Islands, CA	Zc	0.22
1985-1986	[20]	Channel Islands, CA	Cu	0.24
1996-1999	[34]	Marion Island, Australia	Ag & At	0.24
2001-2007	[18]	SEAK & northern BC	Ej	0.26
1985-1986	[20]	Channel Islands, CA	Zc	0.27
1986-1988	[20]	Channel Islands, CA	Zc	0.27
1986-1988	[20]	Channel Islands, CA	Cu	0.28
1988-1989	[37]	Bird Island, South Georgia	Ag	0.4
1989-2000	[36]	Kangaroo Island, Australia	Af	0.4
2010-2018	[21]	Northwest Coast, WA	Ej	0.43
1991-1995	[38]	Gulf of California, Mexico	Zc	0.49
1995-2005	[39]	Kaikoura, New Zealand	Af	0.6-2.84
1983	[40]	St Paul Island, AK	Cu	0.75*
1984	[41]	St Paul Island, AK	Cu	0.78*

2001-2002	[36]	Kangaroo Island, Australia	Af	0.9
2001	[36]	Kangaroo Island, Australia	Nc	1
2002	[36]	Kangaroo Island, Australia	Nc	1.3
1989-1991	[42]	Bass Strait, Australia	Apd	1.9
2010-2018	[21]	Northwest Coast, WA	Zc	2.86
1992	[16]	Los Islotes, Mexico	Zc	3.9-7.9
2000	[43]	Los Islotes, Mexico	Zc	8.75
1998	[43]	Los Islotes, Mexico	Zc	9.9
1992	[43]	Los Islotes, Mexico	Zc	10.4

\* Harvest data, only subadult males

~ Rookery data during breeding season

While the entanglement rates we observed were high, the low number of recorded mortalities from entanglement in the literature and in the local stranding record highlights our poor understanding of the effects of entanglement on sea lion health and survival. In the stranding record for the Washington and Oregon coast only thirteen California and Steller sea lions were found dead with signs of entanglement from 2010-2018 out of 1,599 total strandings. The proportion of dead stranded sea lions that exhibited evidence of entanglement (0.81%) was of a similar order of magnitude to the proportion of live sea lions observed with signs of entanglement from survey effort (0.41% Steller, 2.13% California). In the literature there are also very few records of animals observed dead with signs of entanglement [19,44]. Since dead stranded animals are a subset of the mortality experienced by a population, it is logical that if entanglement significantly affected the sea lion's health and survival, the proportion of dead individuals with evidence of entanglement would be greater than for the live population at large. Since recorded mortality due to entanglement was lower than expected, it suggests that this was not the case.

285 There are two plausible explanations for the lack of recorded mortalities from entanglement. The first is  
286 that entanglement has a much smaller effect on the health of the affected animals than is assumed [45].  
287 Studies on tagged subadult male northern fur seals on St. Paul Island, Alaska found that entangled  
288 individuals had a similar return rate the following year as the general harvest population, suggesting  
289 that entanglement, at least for the harvestable segment of the population, had little to no impact on  
290 survival [41,46]. However, the probability of survival might be largely dependent on the animal's ability  
291 to shed the entangling material [41]. There are records of animals shedding entangling materials in the  
292 wild, including an adult female Antarctic fur seal (*Arctocephalus gazella*) that removed a tied loop of  
293 rope [47], a female Hawaiian monk seal with a nursing pup who freed herself from a tangle of  
294 monofilament and polypropylene line [48], nursing female northern fur seals who freed themselves  
295 from 200g trawl net fragments [49], multiple Hawaiian monk seals who seemed to entangle and  
296 disentangle themselves in beached netting [44], and several Steller sea lions, including a few branded  
297 individuals, observed shedding salmon flashers in Alaska (Alaska Department of Fish and Game,  
298 unpublished data). The likelihood of successfully shedding entangling materials may depend on the type  
299 of material. Packing bands were the most common entangling material in all study years for both Steller  
300 and California sea lions from live observations, similar to what was seen in other studies in the North  
301 Pacific [18,20,35]. However, not a single sea lion stranded dead on the Washington or Oregon coast  
302 from 2010-2018 while entangled in a packing band. This could indicate that sea lions are able to shed  
303 packing bands at a higher rate than other materials. Flashers, on the other hand, made up one third of  
304 strandings with an identifiable entanglement, a much higher proportion than what was seen in live  
305 observations (13.6% Steller, 6.3% California), indicating that individuals with entanglements caused by a  
306 swallowed hook could have a higher mortality rate. The presence of flasher entanglements on live  
307 individuals only during June – September reinforces that sea lions either quickly shed the gear or die.  
308 Most sea lions were in good body condition when observed, suggesting it is more likely that they quickly

309 shed the gear, though it is likely that some animals retain the hook internally after losing the visible  
310 flasher. The large proportion of individuals exhibiting entanglement-related scarring in our record  
311 (21.5% of all documented entanglements) and in other studies [16,20,41] is another testament both to  
312 the ability of animals to self-shed entangling materials and to survive even severely wounding  
313 entanglements.

314 The second plausible explanation of the lack of recorded mortalities due to entanglement is that  
315 affected animals are dying at sea or otherwise away from areas where they might be detected  
316 [4,9,50,51]. Entanglement in a large entangling material, such as a trawl netting fragment, has been  
317 proven to increase the energy expenditure of affected animals, increase the time they spend at sea, and  
318 decrease the depth and duration of foraging dives, all of which could lead to reductions in health or  
319 survival and cause them to perish away from the scientific eye [49,51,52]. Internal entanglement injuries  
320 from swallowed and embedded hooks are also likely to go undetected and unrecorded. Three animals in  
321 the Oregon stranding record had hooks in their stomach and esophagus, but no external signs of  
322 entanglement, and one individual was found with a hook in the stomach and the attached flasher  
323 wedged in the esophagus, demonstrating that animals impacted by embedded hooks may have  
324 sustained severe injuries without showing any observable evidence of entanglement until necropsy [53].  
325 Likewise, animals entangled in derelict fishing gear are unlikely to be discovered until the gear is  
326 recovered, so the impact of these entanglement mortalities is likely underestimated [54]. At-sea  
327 mortality, internal injuries, and derelict gear are just a few types of entanglement-related mortality  
328 unlikely to be accurately documented and included in published entanglement rates.

329 The age, size, and foraging experience of the sea lion may dictate the materials they become entangled  
330 in, and therefore the outcome of the entanglement [9,55,56]. The high proportion of entangled Steller  
331 juveniles exhibiting flashers and rubber bands may be a function of their age: rubber bands may be too  
332 small to entangle a large adult, and flasher entanglement is a sign of a risky foraging behavior -



333 depredating salmon troll fisheries. The small number of unidentifiable entangling materials on juveniles  
334 may be because of their smaller size, which causes the material to sit on the surface of the skin where it  
335 can be easily identified. This may also explain the large number of unidentifiable entangling materials on  
336 adult males, whose considerable seasonal growth [57] could have caused entanglements to bury deep  
337 into the flesh where they are not readily observed [49]. Age and body size therefore impact both the  
338 entangling materials an individual is likely to encounter, and the severity of the wound caused by that  
339 entanglement.

340 Entanglement may also have an impact on pinniped life history and population dynamics. Most  
341 California sea lions migrate away from our survey area to their breeding grounds to the south during  
342 June and July, but the few animals that stayed during those months exhibited a much higher  
343 entanglement rate than in other months. It is therefore possible that entangled individuals were  
344 prevented from migrating because of restrictions imposed by the greater energy expenditure associated  
345 with entanglement [51,52] or compromised body condition. Even for individuals that did arrive at their  
346 breeding grounds, entanglement could impact their reproductive success. Entangled nursing female  
347 northern fur seals spent longer at sea, weaned smaller pups, and abandoned their pups more frequently  
348 than unentangled females [49,58]. However, records of three entangled female California sea lions  
349 successfully weaning pups in Los Islotes, Baja California [16] demonstrate that the impacts of  
350 entanglement on all aspects of pinniped population dynamics are poorly understood.

351 Entanglement rates also appear to be impacted by the availability and distribution of entangling  
352 materials in the immediate environment [4,9]. In our survey area, the occurrence of packing bands in  
353 beach surveys was positively correlated with the proportion of entangled individuals exhibiting packing  
354 bands. A similar relationship has been observed in Hawaiian monk seals, which frequently haul out on  
355 top of beached debris and therefore experience higher entanglement risk when more debris is present  
356 on the beach [59], and with northern fur seal pups which show higher rates of entanglement in areas on

357 St. Paul Island, Alaska with higher concentrations of debris in the nearshore [56]. It is likely that both  
358 basin-wide circulation patterns and nearshore currents play a role in the concentration of entangling  
359 materials and therefore the distribution of entanglement hot spots. Studies have shown that warm  
360 anomaly ocean conditions, usually associated with an El Niño event, can cause changes to the  
361 distribution of marine debris, fishing effort, and pinniped prey items, all of which can impact rates of  
362 entanglement [14,15,38]. In summer 2014, high sea surface temperatures associated with the warm  
363 anomaly referred to as “the Blob” reached the coast, causing the shortest upwelling season for the  
364 northern California Current on record [60], the impacts of which were seen well into 2016 [61]. 2014  
365 and 2015 were years of high entanglement rates for California and Steller sea lions in our study area,  
366 and 2014 - 2016 were years of elevated large whale entanglements [17,62]. It is possible that these  
367 anomalous ocean conditions changed the distribution of fishing effort, entangling debris, and prey items  
368 important to cetaceans and pinnipeds, contributing to the high levels of entanglement seen for both  
369 taxa. Entanglement rates therefore seem to be driven somewhat by debris circulation conditions  
370 created both by normal ocean currents and abnormal ocean conditions.

371 While entanglement may not currently cause population-level concerns in Steller or California sea lions  
372 in Washington, it is still a significant welfare issue, especially considering that most entanglements are  
373 caused by humans, either through the creation of marine debris or through direct fishery interactions  
374 [4] (except for animals collared by penguin skins [30,31,47]). The good news is that human-caused  
375 entanglements can be addressed through changes in human behavior. For entanglements caused by  
376 actively fished gear, outreach and education paired with deterrence strategies may prove effective,  
377 while marine debris requires tackling pollution sources or redesigning offending materials. In New  
378 Zealand and South Georgia, campaigns to encourage fishermen to cut packing bands before disposal led  
379 to declines in packing band entanglements [16,63]. However, in Australia, large-scale efforts by the  
380 government and local fishermen to reduce entanglement failed to prevent entanglement rates from

continuing to increase [36]. Page et al. (2004) proposed that the debris could originate from areas outside of Australian waters and away from local fishing grounds, making national legislation ineffective at addressing the trans-boundary issue. A similar situation could complicate entanglement prevention efforts in northern Washington because of the close proximity to the Canadian border and the presence of large basin-wide currents just offshore. Page et al. (2004) also commented that laws that fall short of mandating the use of redesigned materials to prevent entanglement risk, such as biodegradable packing bands, may fail to cause an effective change in observed entanglement rates. Similarly, while deterrents exist or are in development that could prevent animals from interacting with various types of actively fished gear [64,65], it can be a challenge to find a solution that balances effectiveness, cost, and reducing potential harm to the ecosystem [66–68]. While preventing entanglements altogether is likely an impossible task, small actions such as encouraging fishers to cut packing bands could decrease the impact of entanglement on the welfare of local pinniped species.

## Acknowledgements

The authors would like to acknowledge all the individuals who assisted with data collection, including Patrick Gearin, Merrill Gosho, and Jeff Harris from NOAA MML, and past technicians for the Makah Tribe, including Adrienne Akmajian, Maria Roberts, Joshua Monette, and Quinton Thompson. Kristin Wilkinson and Lauren De Maio assisted with compiling stranding data, which was collected by the many dedicated organizations that make up the West Coast Marine Mammal Stranding Network. Thanks also to Chris Butler-Minor and the OCNMS staff and volunteers who collect and organize the beach debris survey data. We would also like to thank Wendy Szaniszló for assistance with identifying entangling materials, and Justin Jenniges and Hyejoo Ro for providing reviews of the manuscript.

## References

1. Moore E, Lyday S, Roletto J, Litle K, Parrish JK, Nevins H, et al. Entanglements of marine mammals and seabirds in central California and the north-west coast of the United States 2001–2005. *Mar Pollut Bull.* 2009;58: 1045–1051. doi:10.1016/j.marpolbul.2009.02.006
2. National Oceanic and Atmospheric Administration Marine Debris Program. 2014 Report on the Entanglement of Marine Species in Marine Debris with an Emphasis on Species in the United States. Silver Spring, MD; 2014.
3. Dau BK, Gilardi KVK, Gulland FM, Higgins A, Holcomb JB, Leger JS, et al. Fishing gear-related injury in California marine wildlife. *J Wildl Dis.* 2009;45: 355–362. doi:10.7589/0090-3558-45.2.355
4. Laist DW. Impacts of Marine Debris: Entanglement of Marine Life in Marine Debris Including a Comprehensive List of Species with Entanglement and Ingestion Records. Coe J, Rogers D, editors. New York: Springer; 1997. doi:10.1007/978-1-4613-8486-1
5. Hofman RJ. The changing focus of marine mammal conservation. *Trends Ecol Evol.* 1995;10: 462–465. doi:10.1016/S0169-5347(00)89184-3
6. Fowler C. Marine Debris and Northern Fur Seals: a Case Study. *Mar Pollut Bull.* 1987;18: 326–335. doi:10.1016/S0025-326X(87)80020-6
7. Henderson JR. A Pre- and Post-MARPOL Annex V Summary of Hawaiian Monk Seal Entanglements and Marine Debris Accumulation in the Northwestern Hawaiian Islands, 1982–1998. *Mar Pollut Bull.* 2001;42: 584–589. doi:10.1016/S0025-326X(00)00204-6
8. French DP, Reed M. Potential impact of entanglement in marine debris on the population dynamics of the northern fur seal, *Callorhinus ursinus*. In: Shomura RS., Godfrey ML, editors. Proceedings of the Second International Conference on Marine Debris, 2-7 April 1989, Honolulu,

- 424 HI. Department of Commerce, NOAA Technical Memorandum, NMFS, NOAA-TM-NMFS-SWFC-  
425 154.; 1990. pp. 431–452. doi:10.1051/0004-6361/201116485
- 426 9. Fowler CW. A review of seal and sea lion entanglement in marine fishing debris. Proceedings of  
427 the North Pacific Rim Fishermen’s Conference on Marine Debris 1987. 1988. pp. 16–63.
- 428 10. Read AJ. The looming crisis: interactions between marine mammals and fisheries. J Mammal.  
429 2008;89: 541–548. doi:10.1644/07-mamm-s-315r1.1
- 430 11. Weise MJ, Harvey JT. Impact of the California sea lion (*Zalophus californianus*) on salmon  
431 fisheries in Monterey Bay, California. Fish Bull. 2005;103: 685–696.
- 432 12. Yoshida K, Baba N. The problem with fur seal entanglement in marine debris. In: Shomura RS.,  
433 Yoshida HO, editors. Proceedings of the Workshop on the Fate and Impact of Marine Debris, 27-  
434 29 November 1984, Honolulu, HI. 1985. pp. 448–452.
- 435 13. Cawthorn MW. Entanglement in, and ingestion of, plastic litter in marine mammals, sharks, and  
436 turtles in New Zealand waters. In: Shomura RS, Yoshida HO, editors. Proceedings of the  
437 Workshop on the Fate and Impact of Marine Debris, 27-29 November 1984, Honolulu, HI. 1985.  
438 pp. 336–343.
- 439 14. Donohue MJ, Foley DG. Remote sensing reveals links among the endangered Hawaiian monk  
440 seal, marine debris, and El Niño. Mar Mammal Sci. 2007;23: 468–473. doi:10.1111/j.1748-  
441 7692.2007.00114.x
- 442 15. Keledjian AJ, Mesnick S. The Impacts of El Niño Conditions on California Sea Lion (*Zalophus*  
443 *californianus*) Fisheries Interactions: Predicting Spatial and Temporal Hotspots Along the  
444 California Coast. Aquat Mamm. 2013;39: 221–232. doi:10.1578/AM.39.3.2013.221
- 445 16. Harcourt R, Auriolles D, Sanchez J. Entanglement of California sea lions at Los Islotes, Baja

446 California Sur, Mexico. Mar Mammal Sci. 1994;10: 122–125. doi:10.1111/j.1748-  
 447 7692.1994.tb00399.x

448 17. Santora JA, Mantua NJ, Schroeder ID, Field JC, Hazen EL, Bograd SJ, et al. Habitat compression  
 449 and ecosystem shifts as potential links between marine heatwave and record whale  
 450 entanglements. Nat Commun. 2020;11. doi:10.1038/s41467-019-14215-w

451 18. Raum-Suryan KL, Jemison LA, Pitcher KW. Entanglement of Steller sea lions (*Eumetopias jubatus*)  
 452 in marine debris: Identifying causes and finding solutions. Mar Pollut Bull. 2009;58: 1487–1495.  
 453 doi:10.1016/j.marpolbul.2009.06.004

454 19. Hanni KD, Pyle P. Entanglement of pinnipeds in synthetic materials at South-east Farallon Island,  
 455 California, 1976-1998. Mar Pollut Bull. 2000;40: 1076–1081. doi:10.1016/S0025-326X(00)00050-3

456 20. Stewart BS, Yochem PK. Pinniped entanglement in synthetic materials in the Southern California  
 457 Bight. In: Shomura RS., Godfrey ML, editors. Proceedings of the Second International Conference  
 458 on Marine Debris, 2-7 April 1989, Honolulu, Hawaii. 1990. pp. 554–561.

459 21. Scordino J, Allyn E. Entanglement and count data for Steller sea lions and California sea lions in  
 460 northwest Washington. Mendeley Data; 2020.  
 461 doi:http://dx.doi.org/10.17632/447sm2rwrk.1#file-1ecfafdc-0796-472b-ab7c-cd454b164228

462 22. Opfer S, Arthur C, Lippiatt S. NOAA Marine Debris Shoreline Survey Field Guide. NOAA Mar Debris  
 463 Progr. 2012. Available: [www.MarineDebris.noaa.gov](http://www.MarineDebris.noaa.gov)

464 23. R Core Team. R Statistical Program. 2019.

465 24. Wickham H. ggplot2: Elegant Graphics for Data Analysis. New York: Springer-Verlag; 2016.  
 466 Available: <https://ggplot2.tidyverse.org>

- 467 25. Laake JL, Lowry MS, DeLong RL, Melin SR, Carretta J V. Population Growth and Status of California  
468 Sea Lions. *J Wildl Manage.* 2018;82: 583–595. doi:10.1002/jwmg.21405
- 469 26. Loughlin TR. Assessment of Net Entanglement on Northern Sea Lions in the Aleutian Islands, 25  
470 June - 15 July 1985. NWAFC Process Rep 86-02. 1986.
- 471 27. Pitcher KW, Olesiuk PF, Brown RF, Lowry MS, Jeffries SJ, Sease JL, et al. Abundance and  
472 distribution of the eastern North Pacific Steller sea lion (*Eumetopias jubatus*) population. *Fish*  
473 *Bull.* 2007;107: 102–115.
- 474 28. National Marine Fisheries Service. Steller Sea Lion (*Eumetopias jubatus*): Eastern U.S. Stock. US  
475 Pacific Mar Mammal Stock Assessments. 2018.
- 476 29. McIntosh RR, Kirkwood R, Sutherland DR, Dann P. Drivers and annual estimates of marine wildlife  
477 entanglement rates: A long-term case study with Australian fur seals. *Mar Pollut Bull.* 2015;101:  
478 716–725. doi:10.1016/j.marpolbul.2015.10.007
- 479 30. Huckle-Gaete R, Torres D, Vallejos V. Entanglement of Antarctic fur seals *Arctocephalus gazella* in  
480 marine debris at Cape Shirreff and San Telmo Islets, Livingston Island, Antarctica: 1988-1997. *Ser*  
481 *Científica Ina.* 1997;47: 123–135.
- 482 31. Hofmeyr GJ, Bester MN, Kirkman SP, Lydersen C, Kovacs KM. Entanglement of Antarctic fur seals  
483 at Bouvetøya, Southern Ocean. *Mar Pollut Bull.* 2006;52: 1077–1080.  
484 doi:10.1016/j.marpolbul.2006.05.003
- 485 32. Scordino J. Studies on fur seal entanglement, 1981-1984, St. Paul Island, Alaska. In: Shomura RS.,  
486 Yoshida HO, editors. *Proceedings of the Workshop on the Fate and Impact of Marine Debris*, 27-  
487 29 November 1984, Honolulu, HI. 1985. pp. 278–290.
- 488 33. Shaughnessy P. Entanglement of Cape Fur Seals with Man-made Objects. *Mar Pollut Bull.*

- 1980;11: 332–336.
34. Hofmeyr GJ, De Maine M, Bester MN, Kirkman SP, Pistorius PA, Makhado AB. Entanglement of pinnipeds at Marion Island, Southern Ocean: 1991–2001. *Aust Mammal.* 2002;24: 141–146.
35. Zavadil PA, Robson BW, Lestenkof AD, Holser R, Malavansky A. Northern Fur Seal Entanglement Studies on the Pribilof Islands in 2006. 2007. Available: [https://www.researchgate.net/publication/237478662\\_Northern\\_Fur\\_Seal\\_Entanglement\\_Studies\\_on\\_the\\_Pribilof\\_Islands\\_in\\_2006](https://www.researchgate.net/publication/237478662_Northern_Fur_Seal_Entanglement_Studies_on_the_Pribilof_Islands_in_2006)
36. Page B, McKenzie J, McIntosh R, Baylis A, Morrissey A, Calvert N, et al. Entanglement of Australian sea lions and New Zealand fur seals in lost fishing gear and other marine debris before and after Government and industry attempts to reduce the problem. *Mar Pollut Bull.* 2004;49: 33–42. doi:10.1016/j.marpolbul.2004.01.006
37. Croxall JP, Rodwell S, Boyd IL. Entanglement in Man-Made Debris of Antarctic Fur Seals At Bird Island, South Georgia. *Mar Mammal Sci.* 1990;6: 221–233. Available: [http://www.ncbi.nlm.nih.gov/entrez/query.fcgi?db=pubmed&cmd=Retrieve&dopt=AbstractPlus&list\\_uids=8369427653589658751related:f-yKmRsuJnQJ](http://www.ncbi.nlm.nih.gov/entrez/query.fcgi?db=pubmed&cmd=Retrieve&dopt=AbstractPlus&list_uids=8369427653589658751related:f-yKmRsuJnQJ)
38. Zavala-González A, Mellink E. Entanglement of California sea lions, *Zalophus californianus californianus*, in fishing gear in the central-northern part of the Gulf of California, Mexico. *Fish Bull.* 1997;95: 180–184.
39. Boren LJ, Morrissey M, Muller CG, Gemmell NJ. Entanglement of New Zealand fur seals in man-made debris at Kaikoura, New Zealand. *Mar Pollut Bull.* 2006;52: 442–446. doi:10.1016/j.marpolbul.2005.12.003
40. Scordino J, Beekman G, Kajimura H, Yoshida K, Fujimaki Y, Tomita M. Investigations on fur seal



511 entanglement in 1983 and comparisons with 1981 and 1982 entanglement data, St. Paul Island,  
512 Alaska. Background paper submitted to the 27th Annual meeting of the Stranding Scientific  
513 Committee, North Pacific Fur Seal Commission March 29-April 6, 1984 Moscow, Russia. 1984.

514 41. Scordino J, Kajimura H, Furuta A. Fur Seal Entanglement Studies in 1984, St. Paul Island, Alaska.  
515 Kozloff P, Kajimura H, editors. Fur Seal Investig 1985. 1988.

516 42. Pemberton D, Brothers NP, Kirkwood R. Entanglement of Australian fur seals in man-made debris  
517 in Tasmanian waters. Wildl Res. 1992;19: 151–159. doi:10.1071/WR9920151

518 43. Aurióles-Gamboa D, García-Rodríguez F, Ramírez-Rodríguez M, Hernández-Camacho C.  
519 Interaction between the California sea lion and the artisanal fishery in La Paz Bay, Gulf of  
520 California, Mexico. Ciencias Mar. 2003;29: 357–370. doi:10.7773/cm.v29i3.151

521 44. Kenyon KW. No man is benign: The endangered monk seal. Oceans. 1980;13: 48–54.

522 45. Andersen MS, Forney KA, Cole TVN, Eagle TC, Angliss RP, Long K, et al. Differentiating Serious and  
523 Non-Serious Injury of Marine Mammals: Report of the Serious Injury Technical Workshop, 10-13  
524 September 2007, Seattle, WA. US Dep Commer, NOAA Tech Memo. 2008.

525 46. Stewart BS, Bengtson JL, Baba N. Northern Fur Seals Tagged and Observed During Entanglement  
526 Studies St. Paul Island, Alaska. Kozloff P, Kajimura H, editors. Fur Seal Investig 1986. 1989.

527 47. Bonner WN, McCann T. Neck collars on fur seals, *Arctocephalus gazella*, at South Georgia. Br  
528 Antarct Surv Bull. 1982;57: 73–77.

529 48. Henderson JR. Encounters of Hawaiian monk seals with fishing gear at Lisianski Island, 1982. Mar  
530 Fish Rev. 1984;46: 59–61.

531 49. DeLong RL, Gearin PJ, Bengtson JL, Dawson P, Feldkamp SD. Studies on the Effects of

532 Entanglement on Individual Northern Fur Seals. In: Shomura RS., Godfrey ML, editors.  
533 Proceedings of the Second International Conference on Marine Debris, 2-7 April 1989, Honolulu,  
534 HI. 1990. pp. 492–493.

535 50. Williams R, Gero S, Bejder L, Calambokidis J, Kraus SD, Lusseau D, et al. Underestimating the  
536 damage: Interpreting cetacean carcass recoveries in the context of the Deepwater Horizon/BP  
537 incident. *Conserv Lett.* 2011;4: 228–233. doi:10.1111/j.1755-263X.2011.00168.x

538 51. Feldkamp SD, Costa DP, DeKrey GK. Energetic and behavioral effects of net entanglement on  
539 juvenile northern fur seals, *Callorhinus ursinus*. *Fish Bull.* 1989;87: 85–94.

540 52. Bengtson JL, Stewart BS, Ferm LM, DeLong RL. The Influence of Entanglement in Marine Debris  
541 on the Diving Behavior of Subadult Male Northern Fur Seals. Kozloff P, Kajimura H, editors. *Fur*  
542 *Seal Investig* 1986. 1989.

543 53. Franco-Trecu V, Drago M, Katz H, Machín E, Marín Y. With the noose around the neck: Marine  
544 debris entangling otariid species. *Environ Pollut.* 2017;220: 985–989.  
545 doi:10.1016/j.envpol.2016.11.057

546 54. Good TP, June JA, Etnier MA, Broadhurst G. Derelict fishing nets in Puget Sound and the  
547 Northwest Straits: Patterns and threats to marine fauna. *Mar Pollut Bull.* 2010;60: 39–50.  
548 doi:10.1016/j.marpolbul.2009.09.005

549 55. Gearin PJ, Stewart BS, DeLong RL. Late Season Surveys for Entangled Northern Fur Seal Females  
550 and Pups St. Paul Island, Alaska. Kozloff P, Kajimura H, editors. *Fur Seal Investig* 1986. 1989.

551 56. Stewart BS, Baba N, Gearin PJ, Baker J. Observations of Beach Debris and Net Entanglement on  
552 St. Paul Island, Alaska. Kozloff P, Kajimura H, editors. *Fur Seal Investig* 1986. 1989.

553 57. Winship AJ, Trites AW, Calkins DG. Growth in Body Size of the Steller Sea Lion (*Eumetopias*

554 *jubatus*). J Mammal. 2001;82: 500–519. doi:10.1644/1545-1542(2001)082<0500:gibsot>2.0.co;2

555 58. DeLong RL, Dawson P, Gearin P. Incidence and Impact of Entanglement in Netting Debris on  
556 Northern Fur Seal Pups and Adult Females, St. Paul Island, Alaska. Kozloff P, Kajimura H, editors.  
557 Fur Seal Investig 1985. 1988.

558 59. Donohue MJ, Boland RC, Sramek CM, Antonelis GA. Derelict fishing gear in the Northwestern  
559 Hawaiian Islands: Diving surveys and debris removal in 1999 confirm threat to coral reef  
560 ecosystems. Mar Pollut Bull. 2001;42: 1301–1312. doi:10.1016/S0025-326X(01)00139-4

561 60. Peterson W, Robert M, Bond NA. The warm Blob continues to dominate the ecosystem of the  
562 northern California Current. PICES. 2015;23: 44–46.

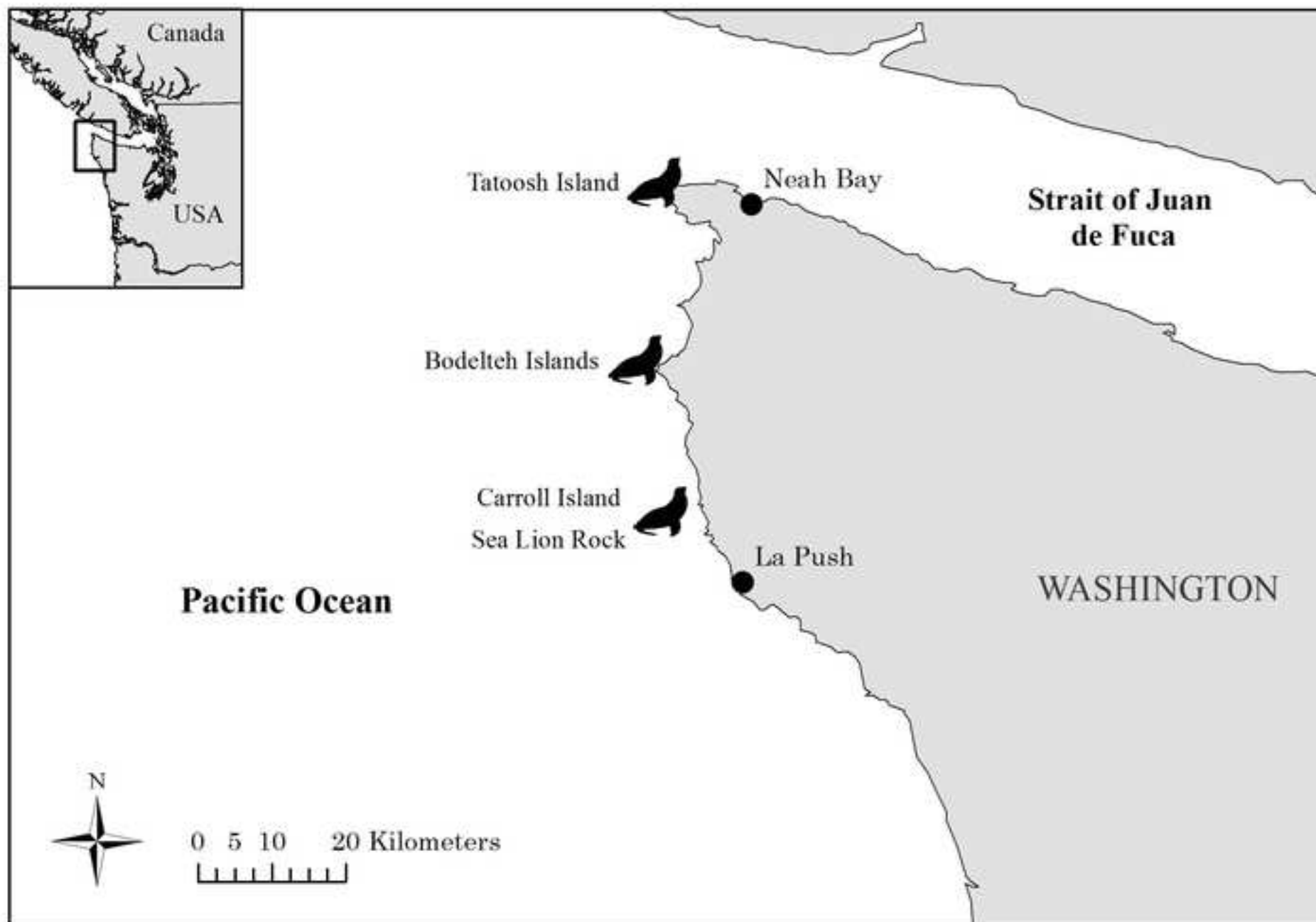
563 61. Gentemann CL, Fewings MR, García-Reyes M. Satellite sea surface temperatures along the West  
564 Coast of the United States during the 2014–2016 northeast Pacific marine heat wave. Geophys  
565 Res Lett. 2017;44: 312–319. doi:10.1002/2016GL071039

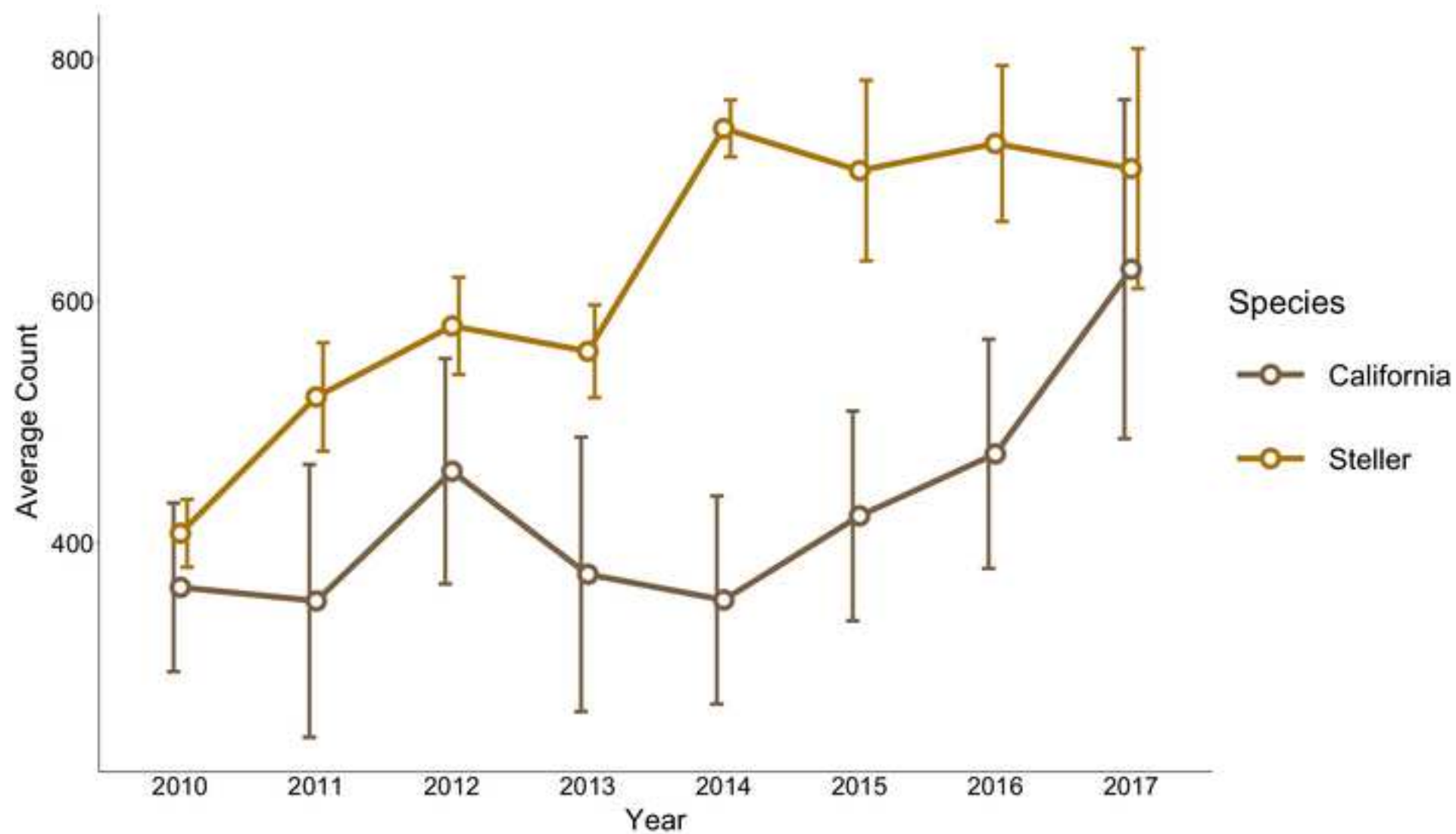
566 62. National Marine Fisheries Service. 2018 West Coast Whale Entanglement Summary. NOAA Fish.  
567 2019. Available:  
568 [https://seagrant.oregonstate.edu/sites/seagrant.oregonstate.edu/files/wcr\\_2018\\_entanglement](https://seagrant.oregonstate.edu/sites/seagrant.oregonstate.edu/files/wcr_2018_entanglement_report_508.pdf)  
569 [\\_report\\_508.pdf](https://seagrant.oregonstate.edu/sites/seagrant.oregonstate.edu/files/wcr_2018_entanglement_report_508.pdf)

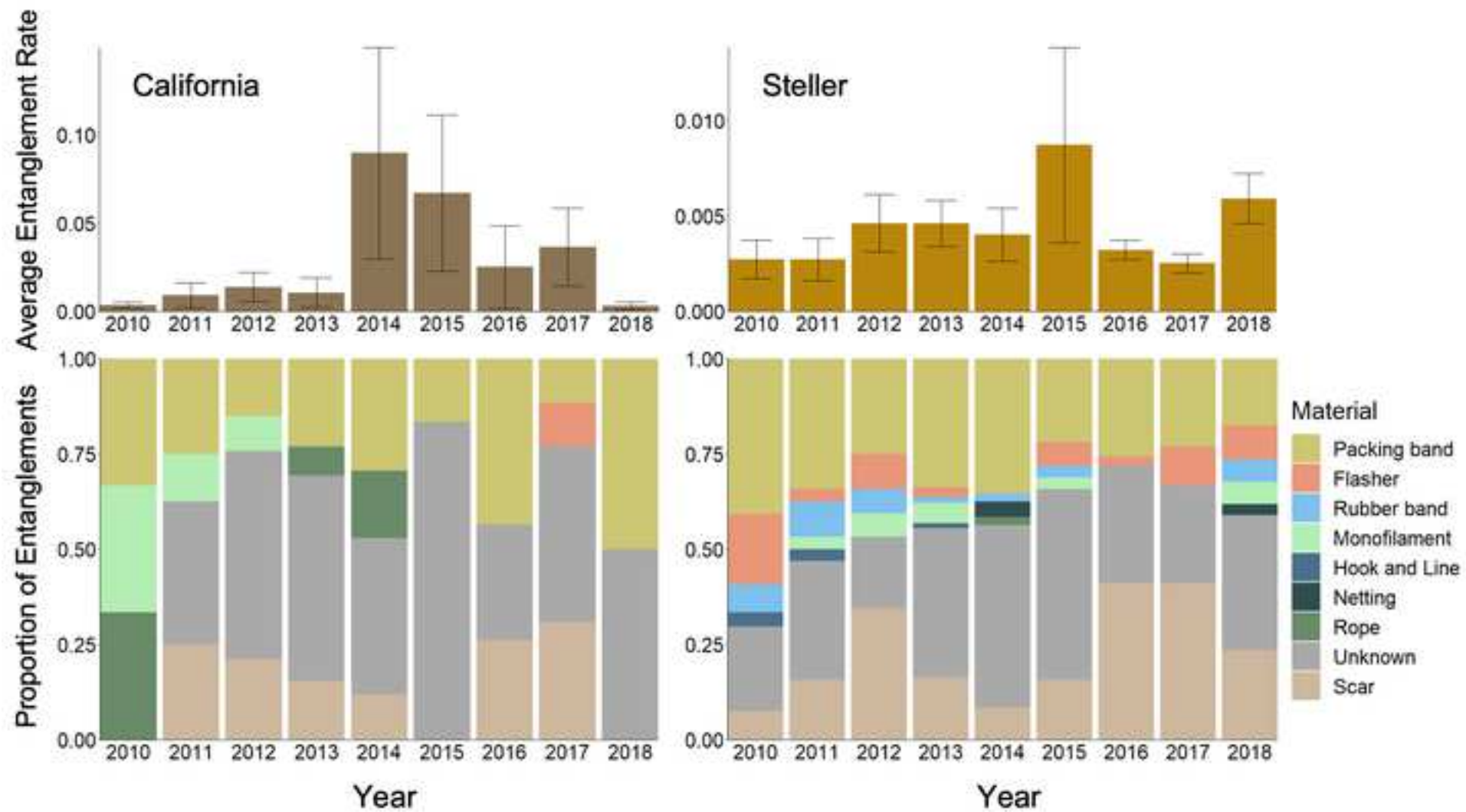
570 63. Arnould JPY, Croxall JP. Trends in entanglement of Antarctic fur seals (*Arctocephalus gazella*) in  
571 man-made debris at South Georgia. Mar Pollut Bull. 1995;30: 707–712. doi:10.1016/0025-  
572 326X(95)00054-Q

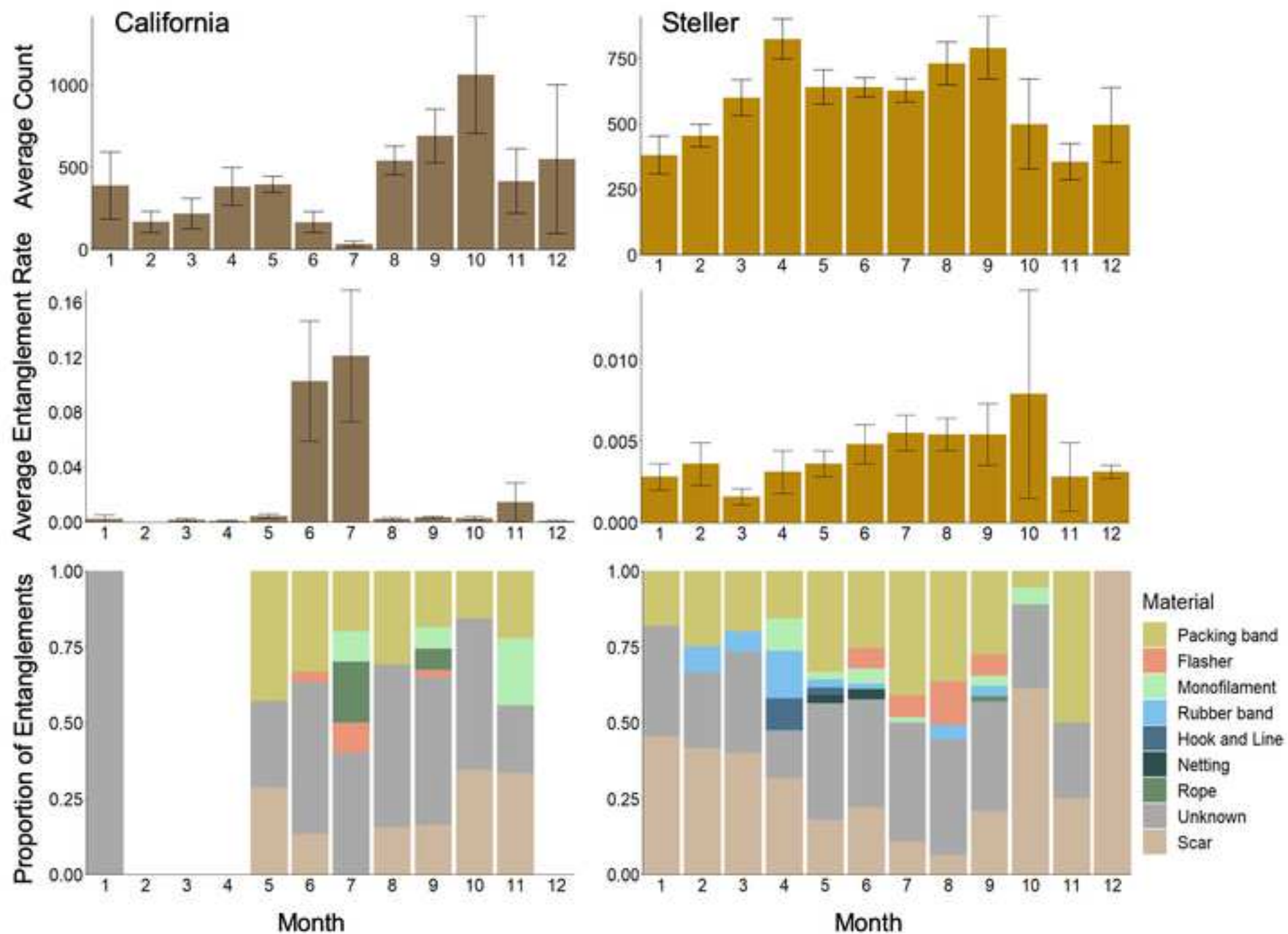
573 64. Forrest KW, Cave JD, Michielsens CGJ, Haulena M, Smith D V. Evaluation of an Electric Gradient to  
574 Deter Seal Predation on Salmon Caught in Gill-Net Test Fisheries. North Am J Fish Manag.  
575 2009;29: 885–894. doi:10.1577/m08-083.1

- 576 65. Barlow J, Cameron GA. Field experiments show that acoustic pingers reduce marine mammal  
577 bycatch in the California drift gill net fishery. *Mar Mammal Sci.* 2003;19: 265–283.  
578 doi:10.1111/j.1748-7692.2003.tb01108.x
- 579 66. Götz T, Janik VM. Acoustic deterrent devices to prevent pinniped depredation: Efficiency,  
580 conservation concerns and possible solutions. *Mar Ecol Prog Ser.* 2013;492: 285–302.  
581 doi:10.3354/meps10482
- 582 67. Jefferson TA, Curry BE. Acoustic methods of reducing or eliminating marine mammal-fishery  
583 interactions: Do they work? *Ocean Coast Manag.* 1996;31: 41–70. doi:10.1016/0964-  
584 5691(95)00049-6
- 585 68. Götz T, Janik VM. Target-specific acoustic predator deterrence in the marine environment. *Anim*  
586 *Conserv.* 2015;18: 102–111. doi:10.1111/acv.12141
- 587

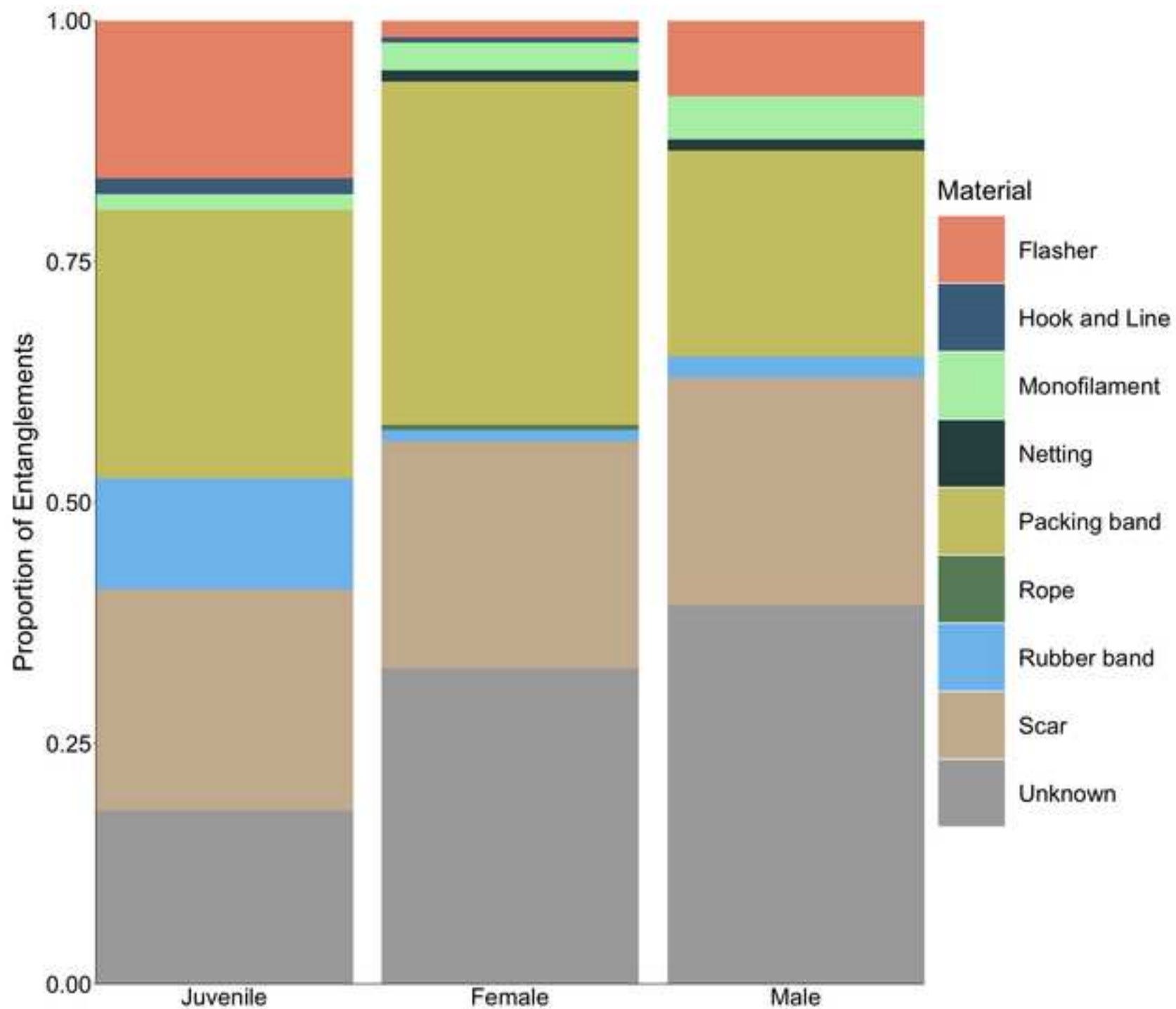


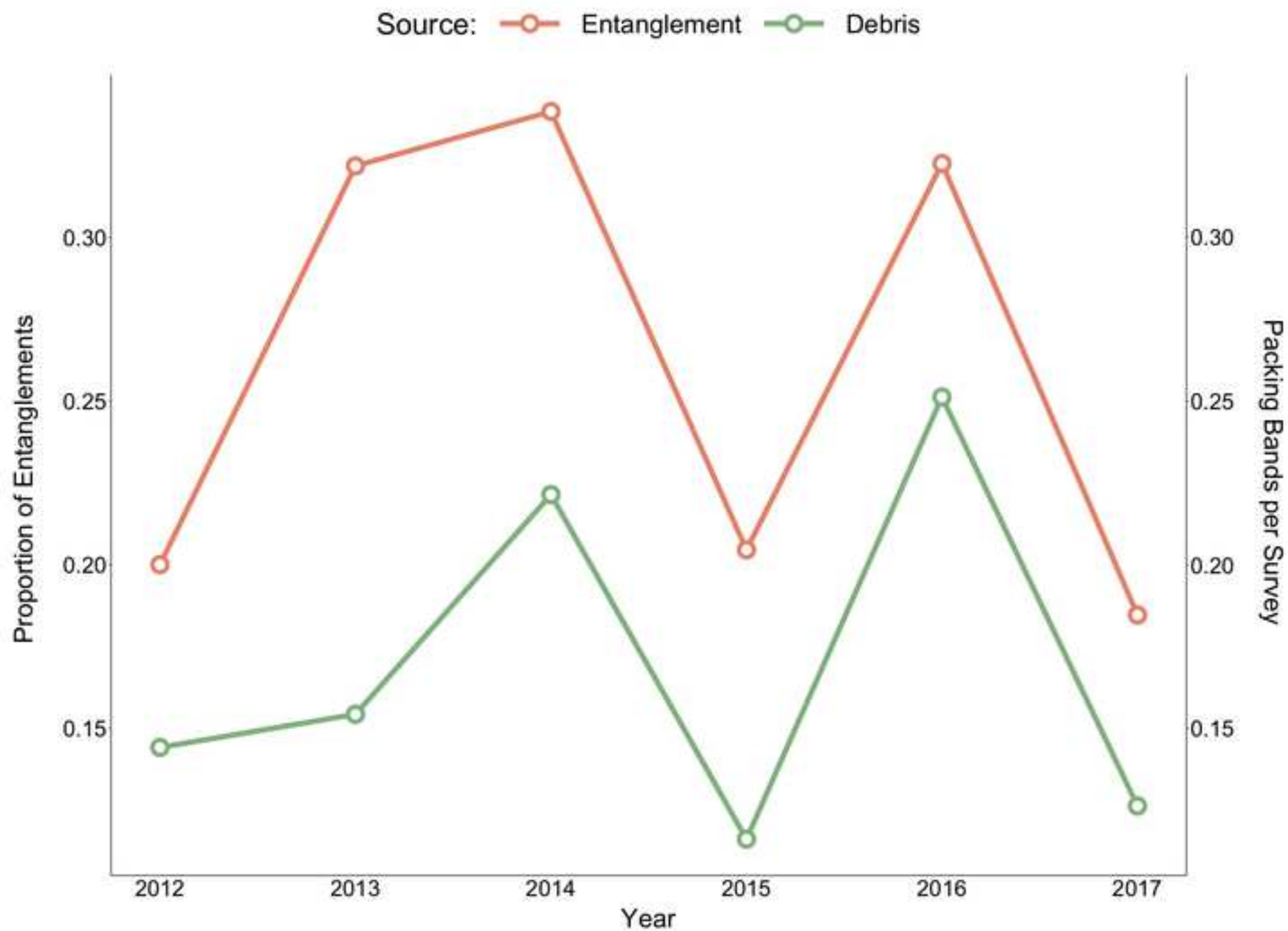














Click here to download Data Review URL  
<http://dx.doi.org/10.17632/447sm2rwrk.1>

