A four-step Bayesian workflow for improving ecological science

Keywords: big data; scientific workflow; data simulation; forecasting; null hypothesis testing

Example & Dataavailability: No new data, but full code and raw data are provided in the supplement for an example workflow. We provide an example of the workflow with complete code available eventually on its GitHub site, but for double-blind review we provide it here as a compressed set of files. The data used for the example is provided and full metadata on it is available via the Knowledge Network for Biocomplexity: doi:10.5063/F12J69B2.

1 Abstract

Growing anthropogenic pressures have increased the need for robust predictive models. Meeting this demand requires approaches that can handle bigger data to yield forecasts that capture the variability and underlying uncertainty of ecological systems. Bayesian models are especially adept at this and are growing in use in ecology. Yet many ecologists are not trained in current methods to build flexible robust models. Here we describe a broadly generalizable workflow for statistical analyses and show how it can enhance training in ecology. Building on the increasingly computational toolkit of many ecologists, this approach leverages simulation to integrate model building and testing for empirical data more fully with ecological theory. In turn this workflow can fit models that are more robust and well-suited to provide new ecological insights—allowing us to refine where to put resources for better estimates, better models, and better forecasts.

Introduction

12

13

14

15

16

17

18

19

20

21

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

In recent years, as ecologists have worked to develop global predictive models, they have developed ever larger datasets (Hampton et al., 2013). These bigger data, however, are also messier data. Such data generally requires a model of both the underlying biological processes and how the measurements were made. Some fields have long used these types of models (generally in fields focused on inferring population sizes of things people want to eat or manage, Muthukumarana et a (generally in fields focused on inferring population sizes for management, Muthukumarana et al., 2008; Zheng et al., . Most, however, have not. This has left many researchers to try to adapt what they were trained in—traditional statistical methods (e.g. F and t tests) and a strong focus on null hypothesis testing—to increasingly complex datasets. Often they have done this by fitting multi-way interaction terms, using random effects to correct for group level factors, or comparing across a large suite of models. Common embedded Yet many common statistical approaches do not often align with ecology's aims today. Beyond the reality that most traditional methods are fragile when used beyond the cleaner, simpler experiments these methods assume (e.g. spatial, temporal and phylogenetic correlations often violate common independence assumptions), they will usually fail to produce robust, reproducible results. Multi-way interaction terms can make main effects hard to interpret and require much larger sample sizes to estimate reliably. Null hypotheses are rarely true and can lead to confusion over what is scientifically important versus 'significant' (Gelman and Hill, 2009; Muff et al., 2022). Many model comparison and machine learning approaches prefer models whose inferences match the idiosyncrasics and biases in the available data, but don't generalize For example, an overly zealous focus on p-values has led to a replication crises in several fields, where results seem most likely the outcome of noisy data combined with a search for statistical significance through many models (effectively a garden of forking paths, Halsey et al., 2015; Loken and G . Some model selection approaches, including new machine learning methods, try to avoid this by comparing across models, but may not generalize to provide useful forecasts. This is especially true when forecasts have to adapt to changes in the underlying biology. The disconnect between

the traditional statistical tools available in ecology and the field's fundamental theory is being made ever-clearer as the need for robust ecological forecasts grows.

Bayesian approaches provide a pathway to build powerful models that can transform how we understand our systems as largescale ecological data become increasingly available. Recognizing this, many in ecology are increasingly using Bayesian methods (Anderson et al., 2021). New algorithms (e.g. Hamiltonian Monte Carlo, Hoffman and Gelman, 2014; Betancourt, 2019) that have made fitting and implementing Bayesian models faster, more robust and—in many ways—easier (Carpenter et al., 2017). Fitting larger and sometimes more complex models, however, presents challenges that are frequently not addressed in traditional ecological training. We suggest that many of these challenges can be overcome by approaching analyses through specific workflows (Betancourt, 2020; Grinsztajn et al., 2021; van de Schoot et al., 2021), which themselves are built on a process of how to do not just statistics, but how to do science (Box, 1976).

Such approaches move away from a focus on null hypothesis testing, towards estimating effect sizes, using models calibrated (see Table 1) and better understood through simulating data at multiple steps—using a number of skills often reserved in ecology more for 'theorists' than empirical ecologists. But this theoretical-vs-empirical divide ignores that the average modern ecologist is computational, and thus already has many of the basic skills to build be poke models.

Here we provide steps. Here we outline a simplified—but powerful—workflow that builds on new insights from statistics (Betancourt, 2020; van de Schoot et al., 2021) and the increasingly computational nature of ecology today. By integrating bespoke model building more fully with ecological theory and understanding—and vice versa—this approach can fit models that are more robust and better-suited to providing new ecological insights and improved predictionsOur aim is to provide an approachable rubric for those new to fitting complex models, and is not intended to be a comprehensive overview (see 'Next steps' in the Supplement). We introduce our workflow assuming a Bayesian statistical framework (with an example in the supplement shown in R and Stan); however, it can also be applied to other statistical inference methods.

A basic Bayesian workflow

67

68

69

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

Statistical analyses are designed for inference—to learn about some process, effect or behavior from data. Robust analyses yield inferences consistent with the underlying truth more often than not (quantifying this consistency is called calibration, and is a critical part of using models for inference, see Table 1). Because we do not know the 'truth' how we approach our analyses—and our inference—is a critical component of how we do science. Fragile statistical analyses can lead science away from the type of repeatable and generalizable findings expected when inferences align with the truth. For example, an overly zealous focus on p-values has led to a replication crises in several fields, where results seem most likely the outcome of noisy data combined with a search for statistical significance through many models (effectively a garden of forking paths, Halsey et al., 2015; Loken . Some model selection approaches, including new machine learning methods, try to avoid this by comparing across models, but may not generalize to provide useful forecasts. This is especially true when forecasts have to adapt to changes in the underlying ecology. We argue that robust analyses comes from explicitly building and challenging a model (or set of models) in an organized sequence of steps—a workflow. We outline one such workflow using a Bayesian approach (with an example in the supplement shown in R and Stan). In ecology, Bayesian methods often shift focus away from null hypothesis testing and point estimates and towards evaluating more fully a model's output, including uncertainty often using a specific model designed for the system and question at hand. These bespoke Our workflow outlines what we consider the major steps for building bespoke models (Fig. 1). Such models can be fit by applying Bayes' theorem, which generates a posterior distribution from a combination of a likelihood and a prior distribution (an initial uncertainty estimate derived from basic ecological knowledge), and using iterative algorithms (e.g. MCMC, Markov Chain Monte Carlo) that provide samples that can be used to extract information from the posterior distribution (for more, see A brief review of statistical inference using Bayesian approaches in the Supplement).

Our workflow describes what we consider the major steps for Bayesian model fitting (Fig. 1).

This includes model development (Step 1) and calibration (Step 2), inference (Step 3), and then model improvement (Step 4). Many of these steps will be familiar to statistical ecologists, but are often overlooked, whereas other steps may appear particular to Bayesian methods (e.g. prior predictive checks), but are actually useful for anyone—using Bayesian models or not—to challenge their models of how the world works. Parts of this workflow could be dropped, or expanded as workflows in themselves, given other aims (see Supplement: Which workflow?). For simplicity, many of the smaller but still critical steps are omitted, including visualization, which is required at every step (and for which there are many good resources, e.g. Gabry et al., 2019)

Step 1: Develop your model(s)

We start the workflow with what can feel like the biggest step—build a model (or potentially, models) that you want to fit your data to based on your aims. By developing a model designed for your biological question, data and aims, your statistical workflow naturally becomes a scientific workflow. You will more clearly see the assumptions and mechanisms in your model, which is especially valuable given how often our intuition of how models 'work' is wrong (Kokko, 2005). You likely already have a model, though it may be only verbal or conceptual. For this workflow, however, you'll need to convert such models into mathematical versions (Servedio et al., 2014). This step is better

Though it can feel challenging at first, this step is best approached before you collect your data. After data collection, it becomes far more tempting to focus on the particular details in the data and not the latent processes and biological models from which the data were generated.

Getting to the point where this step is part of your data design and collection, however, requires starting somewhere—with some model (or models) in hand, any data. A suite of resources for 'generative' or 'narratively generative' modeling can help (MacElreath, 2016; Betancourt,

2021b), along with two points. First, know that you can and will improve on this skill. Second,

as you. As you start, ask lots of questions—and push yourself on your answers—about what

you expect and what's reasonable biologically from your model. For example, instead of simply identifying which distribution your observed response variable looks most similar to, ask yourself what ecological processes generate that distribution and how they influence its properties (e.g. its mean, minimum and maximum). Do you expect data below zero? Up to infinity? If not, why not? Effective model building is about efficient brainstorming and this is a critical part of the process.

As you do this, you'll be generating your model—including its priors, which are important for Bayesian analysis. Assigning priors generally forces you to think about your model with regards to your study system, and interrogate what's probable, possible or actually unreasonable. While many packages (e.g. brms, rstanarm, which fit a suite of pre-defined models) will automatically set default priors, assigning them yourself unreasonable—and can quickly disabuse users of their prejudices prejudices regarding priors. For example, you may not think you have a prior on how sunlight affects plant growth, until you realize your 'agnostic prior' actually allows plants to grow hundreds of meters per day.

Step 2: Check your model on simulated data

Once you have your model and its priors jotted down, you need to write up your model in a particular modeling language and check it. As with all code: just because it runs, does not mean it does what you think it does. The worst errors often still permit code to run. Whether writing it out in Stan, where you need to be able to write out the full likelihood and set all your own priors, or using a package that writes much of the model for you (e.g. rstanarm), you need a way to verify the code is correct: test data.

Test data (aka 'simulated data', or 'fake data,' etc.), and the skills required to build it, are central to this workflow. With 'test data' you simulate data from your model in such a way that you can use the resulting data to test if your model code is correct (i.e., you fix values for your model parameters, then test how well your model recovers them, see the Supplement for an example). While there's no guarantee that inferences will always recover the parameter

values you set, even when using the correct model, extreme disagreement is often an indicator that something is amiss in the implementation of the model. At the same time these simulation studies can help understand how often a model might lead to the correct inference (see Fig. 2). As you do this, you will also be calibrating your model—seeing how accurately and precisely it estimates parameters you set and under what conditions. This very basic model checking step is uncommon for many ecologists, but critical in our view. If you can simulate data from your model, then you can powerfully—and easily—answer questions related to statistical power, what effect sizes are reasonable, and—most likely—have new insights into how your model suggests the world works, all before looking at any real data. 'All models are wrong; some models are useful, 'becomes much clearer when you have the power to generate data from your model under any parameter set and sample size you want. Conversely, if you cannot complete this step, you'll struggle to understand if the model fits well, and struggle further to meaningfully interpret the model output, making an Thus, this apparently simple programmatic task actually encapsulate encapsulates a far deeper understanding of your model. You can learn only so much, however, from data simulated from a particular parameter set. Simulation studies across multiple parameter sets allow you to investigate how robust your inferential performance might be. Prior predictive checks (Betancourt, 2021a; Wesner and Pomeranz, 2021; Winter and Depaoli, use the Bayesian prior model to set this scope of such simulations. For these, you draw values from your prior distribution (usually randomly in your code) and then explore how your model performs under those draws. Seeing how this influences your resulting output reveals the extent to which your model can capture known variation in your data, and gives insight into whether your model is capable of distinguishing among competing hypotheses. It also serves as a check on the priors you're using (addressing one of the common concerns of those inexperienced with Bayesian models). How exactly to do this depends on your question, model and aims, but many

guides can help you think through this (Betancourt, 2021a; Wesner and Pomeranz, 2021; Winter and Depaoli, 2023)

146

147

148

149

150

151

152

153

154

155

156

157

158

159

160

161

162

163

164

165

166

167

168

169

170

171

Step 3: Run your model on your empirical data

The next step is to run the model—you've now validated valuated, test-run and have ready to go—on your exciting new empirical data. Check diagnostics so you know it's running well and adjust until it is (this includes a suite of convergence and efficiency metrics that are well-discussed elsewhere, Betancourt, 2020; Gelman et al., 2020; van de Schoot et al., 2021; Gabry et al., 2019). A model that doesn't converge, or seems to suggest coefficients that are completely at odds with the data, can result from a model that was mis-specified and could never capture real-world variation. For serious problems you may need to reevaluate your model entirely and return to Steps 1 and 2.

This is the step many ecologists skip straight to, ourselves included. It's easy to see the appeal: this is the inference step and where you have the opportunity to see might gain new ecological insights. Fitting new data to the model can feel like the moment when you'll learn something new. But, at least in our experience, this is not always the case. When we rush to this step, that first model we fit is often followed by another, and another—perhaps because one does not converge, or the results of another do not make immediate sense. And with the excitement of getting a model to run we can get distracted from what we are actually most interested in—the inference into ecology.

Following this workflow can make this step much more satisfying. Here the benefits of the workflow may become excitedly apparent: you have estimates in useful units with uncertainty you can understand. You can use this information to draw new conclusions, design new experiments and more—but this is also a point to stop and check your model.

Step 4: Check your model on data simulated from your empirical model output (also known as posterior retrodictive checks)

Once you have your posterior based on your model and new empirical data, it's time to remember that it's wrong $\frac{1}{1}$ (as all models are wrong' after all, $\frac{1}{1}$) and ask how useful it is. You can do some of this through common model-fit diagnostics, such as R^2 , which compares point predictions

to the observed data. With a posterior, however, you can compare an entire distribution of predictions to the observed data.

This is where simulating from your model can be especially insightful. It will not only indicate that the model isn't adequately fitting the data but also can suggest what the problems might be. Steps 1-2 have set you up well for this, as you have a sense of what different parameter estimates do to the model, and test data provide a sense of how it works on data similar to yours. Now with the Using the parameter estimates from your posterior you can to simulate new data from them and (Held et al., 2010; Gelman et al., 2000; Conn et al., 2018) lets you see how that new world compares to the observed data—called posterior retrodictive checks (or posterior predictive checks, Fig. 3). Exactly how to do these depend—again—on your question, model and aims (see Held et al., 2010; Gelman et al., 2000; Conn et al., 2018).

Often here you may find big differences from your empirical data, and can start to generate hypotheses for why. For example, you may find patterns that suggest missing grouping factors (e.g. site or biome) through visual posterior retrodictive checks, or you may quickly realize your model predicts impossible numbers for your biological reality. You may begin to see inadequacies in your model, or even potentially your data. This is one of the main benefits of the workflow: models don't fail silently, they fail with a wealth of context the that helps to generate new models and better experiments. experiments.

Feedbacks & workflows

A key feature of this workflow is that it can be iterated. If you find that you want to tweak your model then you just return to the beginning, adjust your model, and repeat the rest of the workflow. In this way, fitting multiple models is encouraged, but this is distinct from the quest for a minimum adequate model or one 'best' fit. Feedbacks in this workflow are focused far more on what is biologically reasonable, and understanding the utility—and limits—of inference from your data for your model. And there are big benefits to it.

This process more fully integrates mathematical modeling into statistical modeling. To complete Steps 1-2, you have to understand the underlying math of your model enough to simulate data from it. This can be challenging at first (e.g. recalling how to simulate y data for a simple linear regression is not straightforward when you rarely do it), but is immensely beneficial to forcing you to understand your model and its consequences. Indeed, we have found the greatest insights come not from the step we all know best—fitting the model with empirical data (Step 3)—but from every other step in this workflow.

How this workflow changed our science

As we have used this workflow, how we approach our statistical models has changed. These changes have generally been similar for each of us. We suspect they are not unique to us, our study systems, or our questions. Instead, we think they represent common approaches to statistical modeling that could help ecology advance, much as we believe they have helped our science advance.

Looking at parameters, not p-values

Before this workflow, not all of us commonly discussed the values that parameters in our model took—things like the slope and intercept (two common model parameters) were sometimes reported, but we did not know them as well as we knew whether the p-value for the slope was < 0.05. This changes quickly when you need to build simulated data (Step 2); for . For example, when modeling phenological events (observations of biological events on numbered days within the calendar year: 1-365 most years) it is not uncommon to find seemingly-reasonable models generating predictions of events on the non-existent calendar day of 1000. This focus on the value of parameters scales up through this and other modeling workflows. Having a better sense of parameter values across different biological contexts, model parameterizations, and time periods gives a better sense of how the biological world works, including what's reasonable, possible or wildly unrealistic.

Thinking about priors

This workflow also can quickly ease the common concern of those unfamiliar with Bayesian approaches: priors. Often treated as the big bad wolf of Bayesian, or the unseen hand producing biased model fits, according to some. In reality, how much priors influence your model fit is up to your model and your data. Depending on those two parts, the likelihood (influenced by your data) can easily overwhelm your priors (Fig. 2). Indeed, most work on the dangers of priors and 'prior misspecification' focuses on cases where you have very little data for the model you're trying to fit. Priors, however, can only matter more than you know when you fail to think through and check them—that is, you skip Steps 1-2.

Understanding nonidentifiability

This workflow also A closer inspection of our parameters also taught us a lot about identifiability and nonidentifiability, which refers to when all parameters in a model ean be can—or cannot—be uniquely identified with infinite data. More common in our experience is nonidentifiability. Models can be nonidentifiable in several ways, including when mathematically some parameters cannot be distinguished from each other no matter how much data we have. A statistical kinof non-identifiability is degeneracy—, and a statistical kin: degeneracy (see Table 1). Degeneracy concerns the kinds of complex uncertainties that can arise from finite data sets (Gelman and Hill, 2009), and something we have often found in Steps 2-3 of our workflow.

Nonidentifiability and degeneracy can come up in many ways in ecology, and be hard to see if you rush through model fitting. But if you have to write out your model and simulate data (Steps make us think we understand processes we do not. We never noticed them before using this workflow, but since then we have realized (especially in steps 1-2), you may suddenly realize lots of places for nonidentifiability and degenercies to live. For example, when species do not occur across most sites, a model including separate parameters for site and species is often degenerate, but there's often no warning in packages to tell you this. We have become far better at noticing nonidentifiability and degenercies based on this workflow—and live—and we have adjusted how

we collect data and interpret results because of it. For example, we have found fitting both site and species in a model with highly imbalanced data or trying to estimate interaction terms with low sample sizes (Gelman and Hill, 2020, for more details) leads to degenerate models, but there's often no warning in packages to tell us this.

Knowing our limits

data collection for that aim.

274

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

292

293

294

295

296

297

298

Once we noticed how pervasive non-identifiability and degeneracies are how our approach to building complex models also changed. Before using this workflow, we might start with more complex models, then simplify them until their fits converged in our given software package. Often these were hierarchical models with many levels—for example, including every column of site, plot, transect and quadrat in our dataframe, without stopping to check how well sampled they were, or what degeneracies they might introduce. We also often fit a suite of interactions: multiple two-way interactions and the occasional three-way interaction were common fare. But in simulating data, and fitting models to real, messy, imbalanced data using the workflow we came to see how much we were asking of our data and models together. Fitting a two-way interaction with half the effect size of a main effect takes a 16X sample size, compared to fitting the main effects alone (the main effects then average over the interactions, see G . This is sobering. It's more sobering when you see it played out again and again through this workflow. We now both add complexity and simplify based on a more careful reckoning. Often our starting model is not simple. It often includes grouping factors that may be difficult to fit, but only those grouping factors that we see as absolutely critical to the question, model and data at hand. We still add and consider additional grouping factors and interactions, but we do so

with a careful idea of how stable the model given the data likely is with them, and we rarely

fit complex three-way interactions or similar—unless we have carefully designed the model and

Fitting bespoke models

299

300

301

302

303

304

305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

Once we have identified the limits to our data, we can fit a bespoke model to estimate the parameters we are actually interested in rather than the numbers that are convenient to estimate. Models can be specifically designed to estimate and report effects in relevant units (e.g. per degree C of warming)—always with estimated uncertainty. Such flexibility is incredibly powerful in ecology where data are often influenced by complex spatial or temporal patterns, non-linear processes are widespread, and common data types are non-Gaussian (e.g. counts, percent cover, etc.). Bayesian models have many benefits, but an often-mentioned one is that 'you can fit any model you want.' While this is not entirely true (Gelman et al., 2014; Beaumont, 2019), Bayesian modeling options can feel limitless when compared to the modelsecologists can fit in popular modeling packages (e.g. Ime4 package in R). As long as you can write out the likelihood of your desired model (and sometimes even if you can't, Sunnåker et al., 2013) and assign priors to all parameters, you can generally 'fit' the model. However, the flexibility of Bayesian models is sometimes seen as a weakness: you can fit almost whatever you want, but critical parts of your model might be almost entirely unimpacted by your data. In ecological model fitting, we're currently most often interested in parameter estimates strongly informed by our data, making this problem sound especially dangerous. In reality, however, this problem is not related to modeling, but to experimental design—and a flawed experimental design leading to low power

for your model is much easier to see through using our workflow compared to using traditional

null hypothesis testing methods. Low power usually becomes very obvious in Step 2 when you

need to recover your parameter estimates and find you need simulated data very different (e.g.

higher sample size, lower error) than your empirical data to achieve this.

How this workflow intersects with ecological training

322

323

324

325

326

327

328

329

330

331

332

333

334

335

336

337

338

339

340

341

342

343

344

345

346

348

This four-step workflow is a simplified version of the current best practices for Bayesian model fitting (Betancourt, 2020; van de Schoot et al., 2021), but many of the skills required are not part of traditional ecological training. Writing out the math behind most statistical models enough to complete Steps 1-2 bleeds into the skillset usually reserved for those working on theory, where coding and simulating from a model are common tasks. In contrast field, lab and otherwise empirical-data based ecologists often fit models they could not simulate data from. This dichotomy seems short-sighted in our current era of bigger, messier data and greater computational methods poised to handle such messy data. Further, the increasingly computational toolkit of the modern ecologist makes it easier to bridge the gap between ecological models and their underlying math. We argue training in simulating data as part of an organized workflow could speed progress in ecology and is possible given the current skillset of many ecologists. A reasonably competent coder could easily simulate data under a complex model that they might not have the mathematical expertise to solve analytically—if doing so was part of their training and the workflows they regularly use. Simulating data rapidly clarifies underlying assumptions. While training in frequentist methods often includes memorizing assumptions for a particular test, or steps specifically designed to test particular assumptions (e.g. normal quantile plots), this workflow requires no such training. Instead it requires only the skills to identify whatever the assumptions have been encoded in your models. Uptake of new Bayesian R packages highlight that Bayesian methods are no longer the purview of only a few, and these changes come alongside advances in Bayesian workflows, algorithms, and visualizations (e.g. Betancourt, 2020; van de Schoot et al., 2021; Gabry et al., 2019). Ecology must adapt its training to embrace these shifts. While this is an active area, we highlight three major changes. (1) Prior 'beliefs' are changing. Best practices for determining priors is an active area of

statistical research (Gelman et al., 2014; Gelman and Hill, 2020; Betancourt, 2021a), and training

should reflect current best practices. These include that 'non-informative priors' are a misnomer, 349 as they are often informative (Lemoine, 2019), and priors can easily be 'weakly informative.' 350 Thus a strong focus on the dangers of priors in trainingcan be overkill, and verges on scaremongering. 351 352 Current training often includes a very strong focus on mathematically-convenient priors, because 353 of the importance of conjugate priors in closed form solutions for particular posteriors. Modern 354 algorithms, such as HMC, do not require conjugate priors. Prior Advances in developing 355 Bayesian workflows have come alongside improved algorithms, visualizations (e.g. Betancourt, 2020; van de Schoot et a 356 , perspectives on priors (Gelman et al., 2014; Gelman and Hill, 2020; Betancourt, 2021a) and 357 hierarchical approaches that could also improve training. For example, new work shows that 358 prior predictive checks provide a far more powerful more powerful and intuitive way to under-359 stand how priors work within a particular model, and are more useful than rules about which 360 priors should be fit in certain cases or memorizing which priors are conjugate (Betancourt, 2021a) 361 362 (Betancourt, 2021a), compared to past approaches. Similarly, traditional ecological training in 363 hierarchical models still often refers to grouping factors (2) 'Random effects' are not just random. 364 Hierarchical models contain grouping factors, sometimes referred to as random effects, such as 365 species or individual. The term) as 'random effects', however, ', which is misleading, imprecise 366 and thus no longer recommended (Gelman and Hill, 2009). In ecology, it also carries with it 367 many older 'rules' of what is 'random' versus 'fixed,' including that 'random effects are things 368 you don't care about' (for example the 'block' effect from a randomized block design). Training 369 in posterior retrodictive checks (Step 4) may reshape these views, as hierarchical effects are 370 (by definition) drawn from an underlying distribution—meaning they can predict outside of the 371 specific set sampled (for example, to predict for a new species or individual), whereas the same 372 is not true for most categorical 'fixed' effects. 373 (3) P-values These new best practices have gained traction at the same time many fields have 374 recognized that p-values, and null hypothesis testing in general, are easily misleading, and there's 375 no easy fix for that. The replication crisis, rampant in other fields, is based in part on an overly 376

hopeful belief that p-values will separate the signal from the noise, with one easy number. In reality, small. Small sample sizes, lack of routine reporting of interpretable effect sizes, fitting of many models without adequate explanation, poor data and code reporting habits all increase the chance of finding 'significance' at a level of ≤ 0.05 (Halsey et al., 2015; Loken and Gelman, 2017). This reality means a similar crisis is likely lurking in ecology, especially given small Small sample sizes alongside a tendency to fit complicated models with multiple interactions makes ecological research vulnerable to these problems. The answer to this, however, is not to make p-values smaller (Halsey et al., 2015; Colquhoun, 2017), nor is it Bayesian approaches. The answer is training in workflows designed for careful model building, model fitting and model interrogation informed by ecological theory and understanding—including the one we outline here.

Acknowledgements: Comments from F. Baumgarten, D. Loughnan and N. Pates improved this manuscript and J. Ngo improved Figure 1.

References

391

390

- Anderson, S. C., P. R. Elsen, B. B. Hughes, R. K. Tonietto, M. C. Bletz, D. A. Gill, M. A.
- Holgerson, S. E. Kuebbing, C. McDonough MacKenzie, M. H. Meek, et al. 2021. Trends
- in ecology and conservation over eight decades. Frontiers in Ecology and the Environment
- 395 19:274–282.
- Beaumont, M. A. 2019. Approximate Bayesian computation. Annual review of statistics and
- its application 6:379-403.
- Betancourt, M. 2019. The Convergence of Markov Chain Monte Carlo Methods: From the
- Metropolis Method to Hamiltonian Monte Carlo. Annalen der physik 531.
- 400 . 2020. Towards A Principled Bayesian Workflow. https://betanalpha.github.io/
- assets/case_studies/principled_bayesian_workflow.html.
- 402 . 2021a. Prior modeling. https://betanalpha.github.io/assets/case_studies/
- prior_modeling.htmll.
- 404 ——. 2021b. (what's the probabilistic story) modeling glory? https://betanalpha.github.
- io/assets/case_studies/generative_modeling.html.
- Box, G. E. 1976. Science and statistics. Journal of the American Statistical Association pages
- 407 791–799.
- Carpenter, B., A. Gelman, M. Hoffman, D. Lee, B. Goodrich, M. Betancourt, M. A. Brubaker,
- J. Guo, P. Li, and R. Allen. 2017. Stan: A probabilistic programming language. Journal of
- 410 Statistical Software 76:10.18637/jss.v076.i01.
- Colquhoun, D. 2017. The reproducibility of research and the misinterpretation of p-values.
- Royal Society Open Science 4.
- Conn, P. B., D. S. Johnson, P. J. Williams, S. R. Melin, and M. B. Hooten. 2018. A guide to
- Bayesian model checking for ecologists. Ecological Monographs 88:526–542.

- Gabry, J., D. Simpson, A. Vehtari, M. Betancourt, and A. Gelman. 2019. Visualization in bayesian workflow. Journal of the Royal Statistical Society Series a-Statistics in Society 182:389–402.
- Gelman, A., J. B. Carlin, H. S. Stern, D. B. Dunson, A. Vehtari, and D. B. Rubin. 2014.

 Bayesian Data Analysis. 3rd ed. CRC Press, New York.
- Gelman, A., Y. Goegebeur, F. Tuerlinckx, and I. Van Mechelen. 2000. Diagnostic checks for discrete data regression models using posterior predictive simulations. Journal of the Royal Statistical Society Series C-Applied Statistics 49:247–268.
- Gelman, A., and J. Hill. 2009. Data Analysis Using Regression and Multilevel/Hierarchical
 Models. Cambridge, New York.
- Gelman, A., A. Vehtari, D. Simpson, C. C. Margossian, B. Carpenter, Y. Yao, L. Kennedy,
 J. Gabry, P.-C. Bürkner, and M. Modrák. 2020. Bayesian workflow. arXiv.
- Grinsztajn, L., E. Semenova, C. C. Margossian, and J. Riou. 2021. Bayesian workflow for disease transmission modeling in stan. Statistics in Medicine 40:6209–6234.
- Halsey, L. G., D. Curran-Everett, S. L. Vowler, and G. B. Drummond. 2015. The fickle p value generates irreproducible results. Nature Methods 12:179–185.
- Hampton, S. E., C. A. Strasser, J. J. Tewksbury, W. K. Gram, A. E. Budden, A. L. Batcheller,
 C. S. Duke, and J. H. Porter. 2013. Big data and the future of ecology. Frontiers in Ecology
 and the Environment 11:156–162.
- Held, L., B. Schroedle, and H. Rue. 2010. Posterior and Cross-validatory Predictive Checks: A

 Comparison of MCMC and INLA. Pages 91–110 in T. Kneib and G. Tutz, eds. Statistical

 Modelling and Regression Structures: Festschrift in Honour of Ludwig Fahrmeir.
- Hoffman, M. D., and A. Gelman. 2014. The No-U-Turn Sampler: Adaptively Setting Path Lengths in Hamiltonian Monte Carlo. Journal of Machine Learning Research 15:1593–1623.

- Kokko, H. 2005. Useful ways of being wrong. Journal of evolutionary biology 18:1155–1157.
- Lemoine, N. P. 2019. Moving beyond noninformative priors: why and how to choose weakly informative priors in bayesian analyses. Oikos 128:912–928.
- Loken, E., and A. Gelman. 2017. Measurement error and the replication crisis. Science 355:584–585.
- MacElreath, R. 2016. Statistical Rethinking, vol. 469 pp. CRC Press, New York.
- Muff, S., E. B. Nilsen, R. B. O'Hara, and C. R. Nater. 2022. Rewriting results sections in the language of evidence. Trends in ecology & evolution 37:203–210.
- Muthukumarana, S., C. J. Schwarz, and T. B. Swartz. 2008. Bayesian analysis of mark-recapture data with travel time-dependent survival probabilities. Canadian Journal of Statistics 36:5–21.
- Servedio, M. R., Y. Brandvain, S. Dhole, C. L. Fitzpatrick, E. E. Goldberg, C. A. Stern,
 J. Van Cleve, and D. J. Yeh. 2014. Not just a theory—the utility of mathematical models in
 evolutionary biology. PLoS biology 12:e1002017.
- Strinella, E., D. Scridel, M. Brambilla, C. Schano, and F. Korner-Nievergelt. 2020. Potential sex-dependent effects of weather on apparent survival of a high-elevation specialist. Scientific Reports 10:8386.
- Sunnåker, M., A. G. Busetto, E. Numminen, J. Corander, M. Foll, and C. Dessimoz. 2013.

 Approximate bayesian computation. PLoS computational biology 9:e1002803.
- Trijoulet, V., S. J. Holmes, and R. M. Cook. 2018. Grey seal predation mortality on three depleted stocks in the West of Scotland: What are the implications for stock assessments?

 Canadian Journal of Fisheries and Aquatic Sciences 75:723–732.
- van de Schoot, R., S. Depaoli, R. King, B. Kramer, K. Maertens, M. C. Tadesse, M. Vannucci,
 A. Gelman, D. Veen, J. Willemsen, and C. Yau. 2021. Bayesian statistics and modelling.

 Nature Reviews Methods Primers 1.

- Wesner, J. S., and J. P. F. Pomeranz. 2021. Choosing priors in Bayesian ecological models by
- simulating from the prior predictive distribution. Ecosphere 12.
- Winter, S. D. D., and S. Depaoli. 2023. Illustrating the value of prior predictive checking for
- 467 bayesian structural equation modeling. Structural Equation Modeling-a multidisciplinary
- journal .
- Zheng, C., O. Ovaskainen, M. Saastamoinen, and I. Hanski. 2007. Age-dependent survival
- analyzed with Bayesian models of mark-recapture data. Ecology 88:1970–1976.
- 471 Anderson, S.C., Elsen, P.R., Hughes, B.B., Tonietto, R.K., Bletz, M.C., Gill, D.A., Holgerson,
- 472 M.A., Kuebbing, S.E., McDonough MacKenzie, C., Meek, M.H. et al. (2021) Trends in ecology
- and conservation over eight decades. Frontiers in Ecology and the Environment 19, 274–282.
- Beaumont, M.A. (2019) Approximate Bayesian computation. Annual review of statistics and
- *its application* **6**, 379–403.
- Betancourt, M. (2019) The Convergence of Markov Chain Monte Carlo Methods: From the
- 477 Metropolis Method to Hamiltonian Monte Carlo. Annalen der physik 531.
- Betancourt, M. (2020) Towards A Principled Bayesian Workflow.
- Betancourt, M. (2021) Prior modeling. .
- Betancourt, M. (2021) (what's the probabilistic story) modeling glory?...
- Box, G.E. (1976) Science and statistics. Journal of the American Statistical Association pp.
- 482 791-799.
- Goodrich, Betancourt, Brubaker, Guo, Li & Allen Carpenter: 2017stan Carpenter, B., Gelman,
- 484 A., Hoffman, M., Lee, D., Goodrich, B., Betancourt, M., Brubaker, M.A., Guo, J., Li, P. &
- Allen, R. (2017) Stan: A probabilistic programming language. Journal of Statistical Software
- 486 **76**, 10.18637/jss.v076.i01.
- Colquhoun, D. (2017) The reproducibility of research and the misinterpretation of p-values.
- 488 Royal Society Open Science 4.

- Hooten conn 2018 Conn, P.B., Johnson, D.S., Williams, P.J., Melin, S.R. & Hooten, M.B. (2018)
- 490 A guide to Bayesian model checking for ecologists. Ecological Monographs 88, 526-542.
- 491 Gelman gabryvis Gabry, J., Simpson, D., Vehtari, A., Betancourt, M. & Gelman, A. (2019)
- 492 Visualization in bayesian workflow. Journal of the Royal Statistical Society Series a-Statistics
- *in Society* **182**, 389–402.
- Rubin BDA Gelman, A., Carlin, J.B., Stern, H.S., Dunson, D.B., Vehtari, A. & Rubin, D.B.
- 495 (2014) Bayesian Data Analysis. CRC Press, New York, 3rd edn.
- Van Mechelen]gelman200ppc Gelman, A., Goegebeur, Y., Tuerlinckx, F. & Van Mechelen, I.
- 497 (2000) Diagnostic checks for discrete data regression models using posterior predictive simulations.
- 498 Journal of the Royal Statistical Society Series C-Applied Statistics 49, 247–268.
- 499 Gelman, A. & Hill, J. (2009) Data Analysis Using Regression and Multilevel/Hierarchical Models.
- 500 Cambridge, New York.
- Gelman, A. & Hill, J. (2020) Regression and Other Stories. Cambridge University Press.
- Carpenter, Yao, Kennedy, Gabry, Bürkner & Modrák gelman 2020 bayesian Gelman, A., Vehtari,
- A., Simpson, D., Margossian, C.C., Carpenter, B., Yao, Y., Kennedy, L., Gabry, J., Bürkner,
- P.C. & Modrák, M. (2020) Bayesian workflow. arXiv.
- Riou grinsztajn 2021 Grinsztajn, L., Semenova, E., Margossian, C.C. & Riou, J. (2021) Bayesian
- workflow for disease transmission modeling in stan. Statistics in Medicine 40, 6209–6234.
- Drummond]halsey2015 Halsey, L.G., Curran-Everett, D., Vowler, S.L. & Drummond, G.B.
- 508 (2015) The fickle p value generates irreproducible results. Nature Methods 12, 179–185.
- Budden, Batcheller, Duke & Porter Hampton 2013 Hampton, S.E., Strasser, C.A., Tewksbury,
- J.J., Gram, W.K., Budden, A.E., Batcheller, A.L., Duke, C.S. & Porter, J.H. (2013) Big data
- and the future of ecology. Frontiers in Ecology and the Environment 11, 156–162.
- Held, L., Schroedle, B. & Rue, H. (2010) Posterior and Cross-validatory Predictive Checks: A
- 513 Comparison of MCMC and INLA. Statistical Modelling and Regression Structures: Festschrift

```
in Honour of Ludwig Fahrmeir (eds. T. Kneib & G. Tutz), pp. 91–110.
```

- Hoffman, M.D. & Gelman, A. (2014) The No-U-Turn Sampler: Adaptively Setting Path Lengths
- in Hamiltonian Monte Carlo. Journal of Machine Learning Research 15, 1593–1623.
- Kokko, H. (2005) Useful ways of being wrong. Journal of evolutionary biology 18, 1155–1157.
- Lemoine, N.P. (2019) Moving beyond noninformative priors: why and how to choose weakly
- informative priors in bayesian analyses. Oikos 128, 912–928.
- Loken, E. & Gelman, A. (2017) Measurement error and the replication crisis. Science 355,
- 521 584 585.
- MacElreath, R. (2016) Statistical Rethinking, vol. 469 pp. CRC Press, New York.
- Nater]muff2022rewriting Muff, S., Nilsen, E.B., O'Hara, R.B. & Nater, C.R. (2022) Rewriting
- results sections in the language of evidence. Trends in ecology & evolution 37, 203-210.
- Swartz]muthuku2008 Muthukumarana, S., Schwarz, C.J. & Swartz, T.B. (2008) Bayesian analysis
- of mark-recapture data with travel time-dependent survival probabilities. Canadian Journal of
- 527 Statistics **36**, 5–21.

531

- Goldberg, Stern, Van Cleve & Yeh]servedio2014not Servedio, M.R., Brandvain, Y., Dhole, S.,
- Fitzpatrick, C.L., Goldberg, E.E., Stern, C.A., Van Cleve, J. & Yeh, D.J. (2014) Not just a
- theory—the utility of mathematical models in evolutionary biology. *PLoS biology* **12**, e1002017.
- Korner-Nievergelt|strinella2020potential Strinella, E., Seridel, D., Brambilla, M., Schano, C. &
- Korner-Nievergelt, F. (2020) Potential sex-dependent effects of weather on apparent survival of
- a high-elevation specialist. Scientific Reports 10, 8386.
- Corander, Foll & Dessimoz]Sunnaaker2013 Sunnaker, M., Busetto, A.G., Numminen, E., Corander,
- J., Foll, M. & Dessimoz, C. (2013) Approximate bayesian computation. PLoS computational
- 537 biology **9**, e1002803.
- 538 Cook trijoulet 2018 Trijoulet, V., Holmes, S.J. & Cook, R.M. (2018) Grey seal predation mortality

on three depleted stocks in the West of Scotland: What are the implications for stock assessments? Canadian 539 Journal of Fisheries and Aquatic Sciences 75, 723-732. 540 Kramer, Maertens, Tadesse, Vannucci, Gelman, Veen, Willemsen & Yaulvandeschoot2021 van de 541 Schoot, R., Depaoli, S., King, R., Kramer, B., Maertens, K., Tadesse, M.C., Vannucci, M., 542 Gelman, A., Veen, D., Willemsen, J. & Yau, C. (2021) Bayesian statistics and modelling. Nature 543 Reviews Methods Primers 1. 544 Wesner, J.S. & Pomeranz, J.P.F. (2021) Choosing priors in Bayesian ecological models by 545 simulating from the prior predictive distribution. Ecosphere 12. 546 Winter, S.D.D. & Depaoli, S. (2023) Illustrating the value of prior predictive checking for 547 bayesian structural equation modeling. Structural Equation Modeling-a multidisciplinary journal 548 549 Hanski]zheng2007 Zheng, C., Ovaskainen, O., Saastamoinen, M. & Hanski, I. (2007) Age-dependent 550 survival analyzed with Bayesian models of mark-recapture data. Ecology 88, 1970–1976. 551

Tables & Figures

Table 1: A set Glossary: We provide below simplified definitions of the major terms we use (many of these terms, such as calibration, may be used with their simplified definitions differently depending on the particular literature).

Term	Definition
calibration	analyzing how often an estimate is close to the true value over an ensemble
	of hypothetical observations. An exact calibration would require simulating
	from the true data generating process which is impossible in practice. W
	can, however, calibrate to data simulated from the configurations of model
	we plan use to fit to our data $(Steps 1-2)$ so we understand the models better
	including their limits given data similar to ours. We emphasize simulation
	to calibrate model behaviors consistent with our ecological systems and un
	derstanding (e.g. working within a limited set of parameter ranges through
	prior predictive checks). In contrast to this approach, frequentist method ar
	calibrated against all possible behaviors, which is not only impractical fo
	complicated models it's also irrelevant given that the most extreme behavior
	are unlikely to manifest in reality.
degeneracy	complex uncertainties that come from a mix of sources, including, non
	identified models and cases where the data cannot well inform model param
	eters. When the data are not informing the parameters that we care about
	this highlights a measurement issue. Identifying these problems in simulation
	studies can highlight when we need a better experimental design (e.g. sam
	pling for more overlapping species across sites, or changing what we measure
	etc.).
non-identifiability	when all parameters in a model cannot be uniquely identified with infinit
	data
prior	an distribution of reasonable values for a parameter based on fundamenta
	26 biological and ecological understanding, previous research, or other sources
statistical model	Mathematical approximations of the true data generating process labeled wit

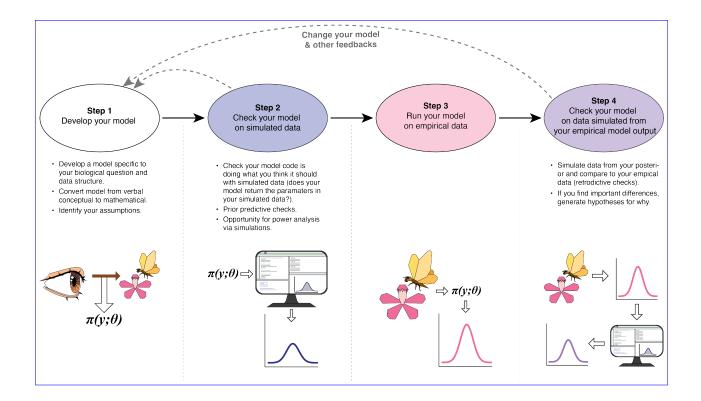


Figure 1: The four-step iterative workflow we outline can help design models for specific ecological questions, data and aims—which makes this a statistical workflow that can naturally become a scientific workflow. It makes the step that many of us focus on—running your model on your empirical data (Step 3)—far more straightforward and insightful by using simulations both before (Step 2) and after (Step 4) it to better understand the model and data together.

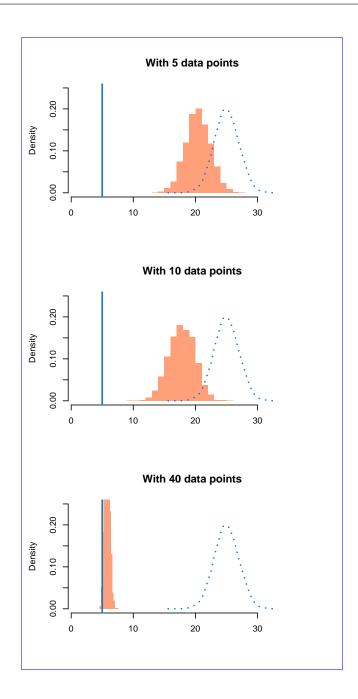


Figure 2: A simple example of how to use simulated data to understand calibration issues in a mis-specified model. Here we know the true model underlying the data is $y = \alpha + \text{normal}(0, \sigma)$ where α is 5 (shown as blue vertical line) and σ is 2. The model, however, is mis-specified by a prior for α of normal(25, 2) (dashed blue line), resulting in a posterior (salmon-colored histogram) not centered on the true value. In our experience it is quite rare to have a prior informed by ecological knowledge be so far off, but this is an example. How mis-calibrated the model will be depends on the data: we show examples with a sample size (N) of 5, 10 and 40 data points. In practice these studies would allow us to determine how much data we would need to be robust to suspect prior models.

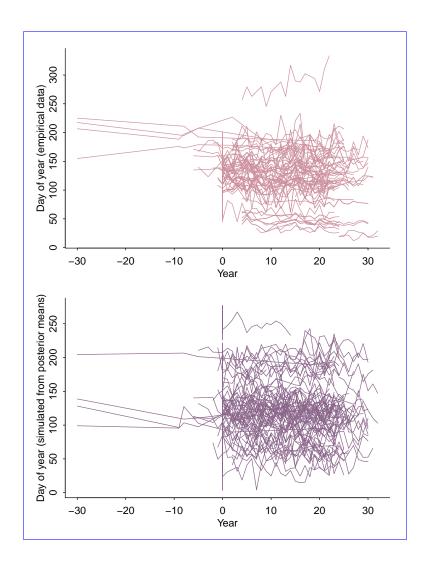


Figure 3: Example of a single retrodictive check from time-series data of phenological events over time. The raw data (top, pink) looks similar to one simulated dataset (bottom, purple), based on existing species number, their respective x data, and simulating from the parameters for each species. See 'An example workflow' in the Supplement for more details.