

A four-step simulation-based workflow for improving ecological science

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Example & Data: We provide an example of the workflow with complete code available eventually on its GitHub site, but for double-blind review we provide it here as a compressed set of files. The data used for the example is provided and full metadata on it is available via the Knowledge Network for Biocomplexity: `doi:10.5063/F12J69B2`.

Abstract

Growing anthropogenic pressures have increased the need for robust predictive models. Meeting this demand requires approaches that can handle bigger data to yield forecasts that capture the variability and underlying uncertainty of ecological systems. Bayesian models are especially adept at this and are growing in use in ecology. Yet many ecologists are not trained in current methods to build flexible robust models. Here we describe a broadly generalizable workflow for statistical analyses and show how it can enhance training in ecology. Building on the increasingly computational toolkit of many ecologists, this approach leverages simulation to integrate model building and testing for empirical data more fully with ecological theory. In turn this workflow can fit models that are more robust and well-suited to provide new ecological insights—allowing us to refine where to put resources for better estimates, better models, and better forecasts.

Introduction

In recent decades, the field of ecology has become more policy-relevant as growing anthropogenic pressures increase the need to manage and understand ecologists systems (Hák et al., 2016; Lindenmayer and Likens, 2010). As ecologists have worked to develop global predictive models to meet these demands they have developed ever larger datasets (Hampton et al., 2013). These bigger data, however, are also messier data. Such data generally requires more complex models, such as models that accommodate both the underlying biological processes and how the measurements were made. Some fields have long used these types of models (generally in fields focused on inferring population sizes for management, Muthukumarana et al., 2008; Zheng et al., 2007; Trijoulet et al., 2018; Strinella et al., 2020). Most, however, have not. This has left many researchers to try to adapt what they were trained in—traditional statistical methods (e.g. F and t tests) and a strong focus on null hypothesis testing—to increasingly complex datasets.

Yet many common statistical approaches do not align with the demands on ecology today. Beyond the reality that most traditional methods are fragile when used beyond the cleaner, simpler experiments these methods assume (e.g. spatial, temporal and phylogenetic correlations often violate independence assumptions), they will usually fail to produce robust, reproducible results. For example, an overly zealous focus on p -values has led to a replication crises in several fields, where results seem most likely the outcome of noisy data combined with a search for statistical significance through many models (effectively a garden of forking paths, Halsey et al., 2015; Loken and Gelman, 2017). Some model selection approaches, including new machine learning methods, try to avoid this by comparing across models, but may not generalize to provide useful forecasts (Boettiger, 2022). This is especially true when forecasts have to adapt to changes in the underlying biology. These leaves ecology in a predicament shared across other fields—concerns of a looming replication crisis (Filazzola and Cahill Jr, 2021; Fraser et al., 2020) and overly confident forecasts with the potential to erode public trust in science (Leroux, 2019; Boettiger, 2022).

Many ecologists recognize these issues and have turned to methods better designed for forecasting from complex systems and messy data. Many in ecology are increasingly using Bayesian methods (Anderson et al., 2021), which can many of the complexities of ecological data, including models of both the measurement and underlying biological process (Hobbs and Hilborn, 2006). Because Bayesian methods allow a huge diversity of bespoke models and increasingly can accommodate large datasets they provide a potential pathway to transform how we understand our systems as more ecological data become increasingly available. New algorithms (e.g. Hamiltonian Monte Carlo, Hoffman and Gelman, 2014; Betancourt, 2019) that have made fitting and implementing Bayesian models faster, more robust and—in many ways—easier (Carpenter et al., 2017). At the same time, machine learning methods are also increasing in use (Pichler and Hartig, 2023). These methods benefit from large datasets—often with many predictors—to make predictive models via test and training datasets (Breiman, 2001) have revolutionized image detection in ecology, but are increasingly used to forecast ecological processes (e.g., Zwart et al., 2023). Machine learning methods often build complex, opaque models (Breiman, 2001; Shmueli, 2010), thus providing opaque inference into ecological processes, but efforts underway aim to change this (e.g., Kutz, 2023).

Regardless of the approach, fitting larger and often more complex models presents challenges that are frequently not addressed in traditional ecological training. We suggest that many of these challenges can be overcome by approaching analyses through specific workflows (Betancourt, 2020; Grinsztajn et al., 2021; van de Schoot et al., 2021), which themselves are built

on a process of how to do not just statistics, but how to do science (Box, 1976). While these approaches are slowly gaining traction in other fields (e.g., Esfahani et al., 2021; Schad et al., 2021; Bouman et al., 2024), they are not widely used in ecology today. Such approaches move away from a focus on null hypothesis testing, towards estimating effect sizes, using models calibrated (see Table 1) and better understood through simulating data at multiple steps. These workflows use a number of skills often reserved in ecology more for ‘theorists’ than empirical ecologists. But this theoretical-vs-empirical divide ignores that the average modern ecologist is computational, and thus already has many of the basic skills to build bespoke models (Hilborn and Mangel, 2013).

Here we outline a simplified—but powerful—workflow that builds on new insights from statistics (Betancourt, 2020; van de Schoot et al., 2021) and the increasingly computational nature of ecology today. Our aim is to provide an approachable rubric for those new to fitting complex models or simply those interested in re-considering their current workflow (and is not intended to be a comprehensive overview; see ‘Next steps’ in the Supplement). Because of this aim, we provide examples of relatively simple models, and suggest resources as users build more complex models. Our examples include several statistical inference methods, though we focus more on implementing the workflow through a Bayesian statistical framework (with an example in the supplement shown in R and Stan), because this framework allows integrating bespoke model building more fully with ecological theory and understanding. We argue this approach can fit models that are more robust and better-suited to providing new ecological insights and improved predictions, and may provide a blueprint for other fields similarly challenged by complex systems, growing datasets and limited training in how to best approach them.

A four-step workflow

Our workflow outlines what we consider the major steps for building bespoke models (Fig. 1). Many of these steps will be familiar to statistical ecologists, but are often overlooked, whereas other steps may appear particular to Bayesian methods (e.g. prior predictive checks), but are actually useful for anyone—using Bayesian models or not—to challenge their models of how the world works. We find a Bayesian framework the easiest way to apply this workflow (see *A brief review of statistical inference using Bayesian approaches* in the Supplement), especially for building bespoke models, but we touch on other approaches throughout. We argue this workflow is useful for all approaches, and may be especially important to understand how new techniques to ecology, such as those from machine learning (Pichler and Hartig, 2023), actually apply to help answer ecological questions. Parts of this workflow could be expanded as workflows in themselves, given other aims (see Supplement: Which workflow?).

Step 1: Develop your model(s)

We start the workflow with what can feel like the biggest step—build a model (or potentially, models) based on your aims (Hilborn and Mangel, 2013). By developing a model designed for your biological question, data and aims, your statistical workflow naturally becomes a scientific workflow. You will more clearly see the assumptions and mechanisms in your model, which is especially valuable given how often our intuition of how models ‘work’ is wrong (Kokko, 2005). You likely already have a model, though it may be only verbal or conceptual. For this workflow, however, you’ll need to convert such models into mathematical versions (Servedio et al., 2014).

Though it can feel challenging at first, this step is best approached before you collect any data. A suite of resources for ‘generative’ or ‘narratively generative’ modeling can help (MacElreath, 2016; Betancourt, 2021*b*). As you start, ask lots of questions—and push yourself on your answers—about what you expect and what’s reasonable biologically from your model. As you do this, you’ll be generating your model—including its priors. Priors are important for Bayesian analysis, but the basic idea of them—coming up with a distribution of reasonable values for parameters in your model (see Table 1)—is useful to all analyses. Assigning priors generally forces you to think about your model with regards to your study system, and interrogate what’s probable, possible or actually unreasonable—and can quickly disabuse users of prejudices regarding priors. For example, you may not think you have a prior on how sunlight affects plant growth, until you realize your ‘agnostic prior’ actually allows plants to grow hundreds of meters per day.

Step 2: Check your model on simulated data

Once you have your model and its priors jotted down, you need to write up your model in a particular modeling language and check it. As with all code: just because it runs, does not mean it does what you think it does. The worst errors often still permit code to run.

Test data (aka ‘simulated data’, or ‘fake data,’ etc.), and the skills required to build it, are central to this workflow. With ‘test data’ you simulate data from your model in such a way that you can use the resulting data to test if your model code is correct (i.e., you fix values for your model parameters, then test how well your model recovers them, see the Supplement for an example). While there’s no guarantee that inferences will always recover the parameter values you set, even when using the correct model, extreme disagreement is often an indicator that something is amiss in the implementation of the model. At the same time these simulation studies can help understand how often a model might lead to the correct inference (see Fig. 2). As you do this, you will also be calibrating your model—seeing how accurately and precisely it estimates parameters and under what conditions.

This very basic model checking step is uncommon for many ecologists, but critical in our view. If you can simulate data from your model, then you can powerfully—and easily—answer questions related to statistical power, what effect sizes are reasonable, and—most likely—have new insights into how your model suggests the world works, all before looking at any real data. Thus, this apparently simple programmatic task actually encapsulates a far deeper understanding of your model. ‘All models are wrong; some models are useful,’ becomes much clearer when you have the power to generate data from your model under any parameter set and sample size you want.

You can learn only so much, however, from data simulated from a particular parameter set. Simulation studies across multiple parameter sets allow you to investigate how robust your inferential performance might be. Prior predictive checks (Betancourt, 2021*a*; Wesner and Pomeranz, 2021; Winter and Depaoli, 2023) use the Bayesian prior model to set this scope of such simulations, but can be used in any analysis. For these, you draw values from your prior distribution and then explore how your model performs. Seeing how this influences your resulting output reveals the extent to which your model can capture known variation in your data, and gives insight into whether your model is capable of distinguishing among competing hypotheses.

Step 3: Run your model on your empirical data

The next step is to run the model—you’ve now evaluated, test-run and have ready to go—on your exciting new empirical data. Check diagnostics so you know it’s running well and adjust until it is (this includes a suite of convergence and efficiency metrics that are well-discussed elsewhere, Betancourt, 2020; Gelman et al., 2020; van de Schoot et al., 2021; Gabry et al., 2019).

This is the step many ecologists skip straight to, ourselves included. It’s easy to see the appeal: this is the inference step and where you might gain new ecological insights. Fitting new data to the model can feel like the moment when you’ll learn something new. But, at least in our experience, this is not always the case. When we rush to this step, that first model we fit is often followed by another, and another—perhaps because one does not converge, or the results of another do not make immediate sense. And with the excitement of getting a model to run we can get distracted from what we are actually most interested in—the inference into ecology.

Following this workflow can make this step much more satisfying. Here the benefits of the workflow may become excitedly apparent: you have estimates in useful units with uncertainty you can understand. You can use this information to draw new conclusions, design new experiments and more—but this is also a point to stop and check your model.

Step 4: Check your model on data simulated from your empirical model output (also known as posterior retrodictive checks)

Once you have your parameter estimates based on your model and new empirical data, it’s time to remember that your model is wrong (as all models are) and ask how useful it is. You can do some of this through common model-fit diagnostics, such as R^2 , which compares point predictions to the observed data. With a Bayesian posterior (see Table 1), however, you have an added benefit in that you can compare an entire distribution of predictions to the observed data. This is where simulating from your model can be especially insightful. It will not only indicate that the model isn’t adequately fitting the data but also can suggest what the problems might be. Using the parameter estimates from your posterior to simulate new data (Held et al., 2010; Gelman et al., 2000; Conn et al., 2018) lets you see how that new world compares to the observed data—called posterior retrodictive checks (or posterior predictive checks, Fig. 3). This is most easily done in a Bayesian framework where your posterior captures your uncertainty in a useful way, but can be done with estimates of your parameters and their uncertainty from other inferential frameworks.

Often here you may find big differences from your empirical data, and can start to generate hypotheses for why. For example, you may find patterns that suggest missing grouping factors (e.g. site or biome) through visual posterior retrodictive checks, or you may quickly realize your model predicts impossible numbers for your biological reality. You may begin to see inadequacies in your model, or even potentially your data. This is one of the main benefits of the workflow: models don’t fail silently, they fail with a wealth of context that helps to generate new models and experiments.

180 **Feedbacks & workflows**

181 A key feature of this workflow is that it can be iterated. If you find that you want to tweak
182 your model then you return to the beginning, adjust your model, and repeat the rest of the
183 workflow. In this way, fitting multiple models is encouraged, but this is distinct from the quest
184 for a minimum adequate model or one ‘best’ fit. Feedbacks in this workflow are focused far more
185 on what is biologically reasonable, and understanding the utility—and limits—of inference from
186 your data for your model. And there are big benefits to it.

187 **How this workflow changed our science**

188 Before this workflow, not all of us commonly discussed the values that parameters in our model
189 took—things like the slope and intercept (two common model parameters) were sometimes
190 reported, but we did not know them as well as we knew whether the p -value for the slope was
191 < 0.05 . This changes quickly when you need to build simulated data (Step 2). For example,
192 when modeling phenological events (observations of biological events on numbered days within
193 the calendar year: 1-365 most years) it is not uncommon to find seemingly-reasonable models
194 generating predictions of events on the non-existent calendar day of 1000.

195 A closer inspection of our parameters also taught us a lot about identifiability and nonidentifi-
196 ability, which refers to when all parameters in a model can—or cannot—be uniquely identified
197 with infinite data, and a statistical kin: degeneracy (see Table 1). Degeneracy concerns the
198 kinds of complex uncertainties that can arise from finite data sets (Gelman and Hill, 2009), and
199 something we have often found in Steps 2-3 of our workflow. Nonidentifiability and degeneracy
200 can come up in many ways in ecology, and make us think we understand processes we do not.
201 We never noticed them before using this workflow, but since then we have realized (especially
202 in steps 1-2) lots of places for nonidentifiability and degeneracies to live—and we have adjusted
203 how we collect data and interpret results because of it. For example, we have found fitting both
204 site and species in a model with highly imbalanced data or trying to estimate interaction terms
205 with low sample sizes (Gelman and Hill, 2020, for more details) leads to degenerate models, but
206 there’s often no warning in packages to tell us this.

207 **How this workflow intersects with ecological training**

208 This four-step workflow is a simplified version of the current best practices for model fitting
209 (Betancourt, 2020; van de Schoot et al., 2021), but many of the skills required are not part
210 of traditional ecological training. Writing out the math behind most statistical models enough
211 to complete Steps 1-2 bleeds into the skillset usually reserved for those working on theory,
212 where coding and simulating from a model are common tasks. In contrast field, lab and other-
213 wise empirical-data based ecologists often fit models they could not simulate data from. This
214 dichotomy seems short-sighted in our current era of bigger, messier data and greater compu-
215 tational methods poised to handle such messy data. Further, the increasingly computational
216 toolkit of the modern ecologist makes it easier to bridge the gap between ecological models and
217 their underlying math.

218 We argue training in simulating data as part of an organized workflow could speed progress in
219 ecology and is possible given the current skillset of many ecologists. A reasonably competent

coder could easily simulate data under a complex model that they might not have the mathematical expertise to solve analytically—if doing so was part of their training and the workflows they regularly use. While training in frequentist methods often includes memorizing assumptions for a particular test, or steps specifically designed to test particular assumptions (e.g. normal quantile plots), this workflow requires no such training. Instead it requires only the skills to identify whatever the assumptions have been encoded in your models. As such it moves away from some modeling paradigms in ecology, which focus on fewer underlying assumptions (e.g. random forests, non-parametric), to building models where the assumptions are transparent and motivated by the specific domain expertise available in each application.

Advances in developing workflows have come alongside improved algorithms, visualizations (e.g. Betancourt, 2020; van de Schoot et al., 2021; Gabry et al., 2019), perspectives on priors (Gelman et al., 2014; Gelman and Hill, 2020; Betancourt, 2021a) and hierarchical approaches that could also improve training. For example, new work shows that prior predictive checks provide a more powerful and intuitive way to understand how priors work within a particular Bayesian model (Betancourt, 2021a), compared to past approaches. Similarly, traditional ecological training in hierarchical models still often refers to grouping factors (such as species or individual) as ‘random effects,’ which is misleading, imprecise and thus no longer recommended (Gelman and Hill, 2009). In ecology, it also carries with it many older ‘rules’ of what is ‘random’ versus ‘fixed,’ including that ‘random effects are things you don’t care about’ (for example the ‘block’ effect from a randomized block design). Training in posterior retrodictive checks (Step 4) may reshape these views, as hierarchical effects are (by definition) drawn from an underlying distribution—meaning they can predict outside of the specific set sampled (for example, to predict for a new species or individual), whereas the same is not true for most categorical ‘fixed’ effects.

How this workflow extends to other fields

These new best practices have gained traction at the same time that ecology alongside many other fields have recognized that p -values, and null hypothesis testing in general, are easily misleading (Gelman and Geurts, 2017; Ferraro and Shukla, 2020; Filazzola and Cahill Jr, 2021; Fraser et al., 2020). Small sample sizes, lack of routine reporting of interpretable effect sizes, fitting of many models without adequate explanation, poor data and code reporting habits all increase the chance of finding ‘significance’ at a level of ≤ 0.05 (Halsey et al., 2015; Loken and Gelman, 2017). Small sample sizes alongside a tendency to fit complicated models with multiple interactions makes ecological research, alongside other fields, vulnerable to these problems (Gelman, 2015).

The answer to these problems is not to make p -values smaller (Halsey et al., 2015; Colquhoun, 2017), nor is it Bayesian, machine learning or ‘new’ causal inference approaches, despite assertions to the contrary, which echo previous promised revolutions through a new method (e.g., Mitchell, 1992; Burnham and Anderson, 2004; Byrnes and Dee, 2025). Many fields are increasingly using machine learning models to guide science, but they can easily lead to poor models that do not match underlying realities of the system (Efron, 2020; Pichler and Hartig, 2023). Ecologists have readily taken up such approaches, in hopes they will help them fit better models. Similarly the field readily took up path analysis, model comparison with AIC, and a suite of other approaches, that promised better inference given mastery of one new approach that led eventually to hundreds of papers reporting poor models, and resulting policy recommendations based on such models (Petratis et al., 1996; Leroux, 2019). This fad approach to statistics is

not unique to ecology, but the cure for it is also not a new statistical approach.

The answer is training in workflows designed for careful model building, model fitting and model interrogation informed by underlying theory and understanding (Hilborn and Mangel, 2013)—including the one we outline here. This workflow depends strongly on simulating data—for testing your model (Step 2), and understanding your model results (Step 4)—an area we actively under-train in many research fields that depend on increasingly complex statistical methods. Simulation approaches encourage interactive learning, build intuition, and stress exploring a model in its relevant context. Ecologists, similar to researchers in any domain-specific field, are much better at thinking about ecological problems than statistical ones. Grounding statistical approaches in theory and understanding will likely bring the best out of statistical modeling (similar to how the best applied statisticians focus on the science CITES). While this idea is not new we argue the need for it is especially high, as the line between estimation and prediction becomes more blurred for trainees (Shmueli, 2010). At the same time, however, computation is increasingly part of a researcher’s toolkit making the opportunity to widely adopt this workflow more possible.

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Table 1: Glossary: We provide below simplified definitions of the major terms we use (many of these terms, such as calibration, may be used differently depending on the particular literature).

<i>Term</i>	<i>Definition</i>
calibration	analyzing how often an estimate is close to the true value over an ensemble of hypothetical observations. An exact calibration would requires simulating from the true data generating process which is impossible in practice. We can, however, calibrate to data simulated from the configurations of models we plan use to fit to our data (<i>Steps 1-2</i>) so we understand the models better, including their limits given data similar to ours. We emphasize simulations to calibrate model behaviors consistent with our ecological systems and understanding (e.g. working within a limited set of parameter ranges through prior predictive checks). In contrast to this approach, frequentist method are calibrated against all possible behaviors, which is not only impractical for complicated models it's also irrelevant given that the most extreme behaviors are unlikely to manifest in reality.
degeneracy	complex uncertainties that come from a mix of sources, including, non-identified models and cases where the data cannot well inform model parameters. When the data are not informing the parameters that we care about, this highlights a measurement issue. Identifying these problems in simulation studies can highlight when we need a better experimental design (e.g. sampling for more overlapping species across sites, or changing what we measure, etc.).
non-identifiability	when all parameters in a model cannot be uniquely identified with infinite data
prior	an distribution of reasonable values for a parameter based on fundamental biological and ecological understanding, previous research, or other sources
statistical model	Mathematical approximations of the true data generating process labeled with numerical parameters. Evaluating a statistical model on observed data gives a likelihood function that quantifies how compatible different parameters are with the observed data, and hence can be used to ‘fit’ the best parameters. In this article, we often simplify to ‘model.’ See also the Supplement: What’s a model?
posterior	product of the likelihood and prior; that is, a probability distribution that quantifies how compatible different model parameters are with both the observed data and the domain expertise encoded in the prior model.
workflow	a set of steps to achieve a goal, with those steps designed to help organize the process, and ideally make it more systematic

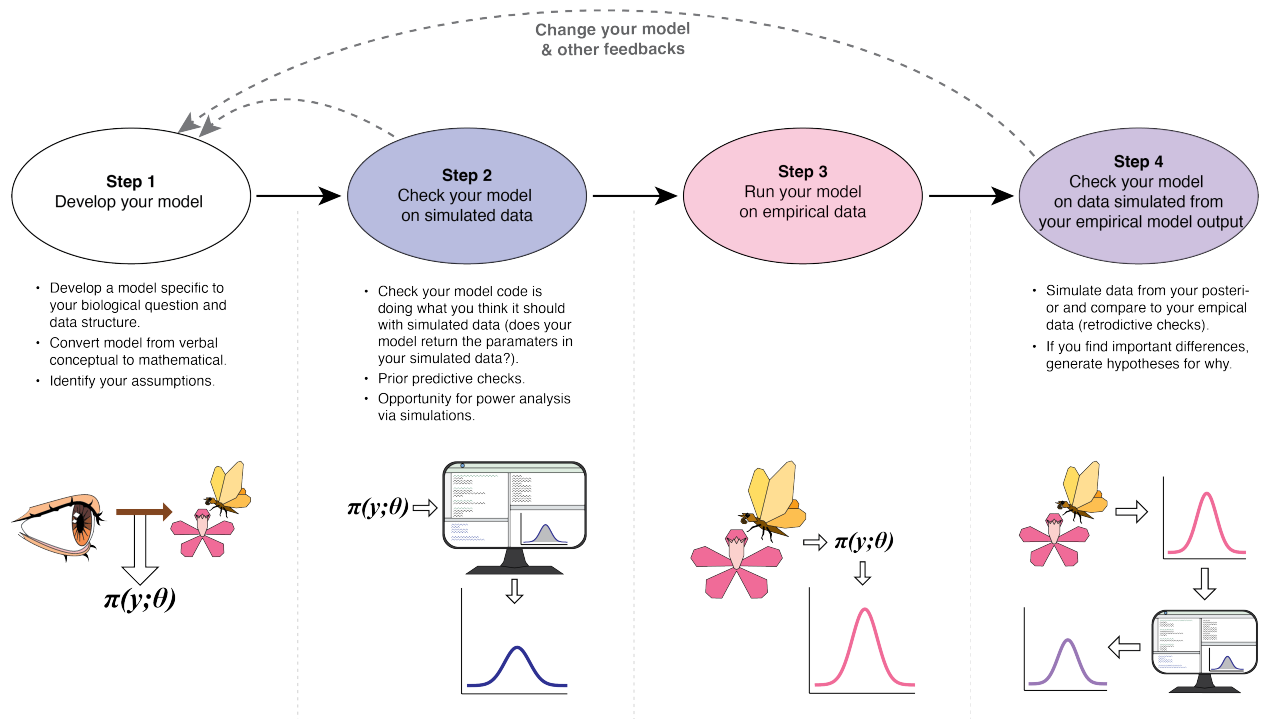


Figure 1: The four-step iterative workflow we outline can help design models for specific ecological questions, data and aims—which makes this a statistical workflow that can naturally become a scientific workflow. It makes the step that many of us focus on—running your model on your empirical data (Step 3)—far more straightforward and insightful by using simulations both before (Step 2) and after (Step 4) it to better understand the model and data together.

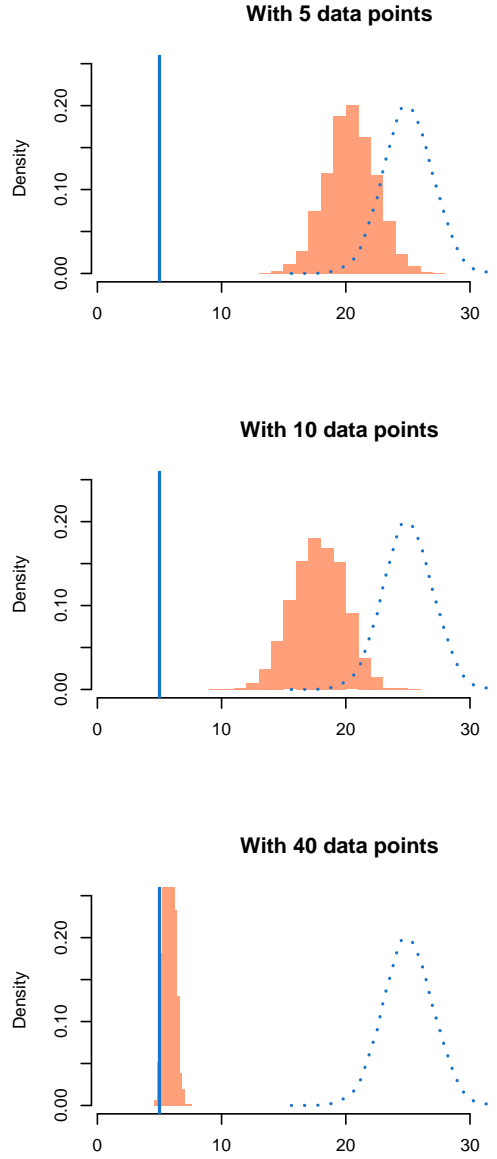


Figure 2: A simple example of how to use simulated data to understand calibration issues in a mis-specified model. Here we know the true model underlying the data is $y = \alpha + \text{normal}(0, \sigma)$ where α is 5 (shown as blue vertical line) and σ is 2. The model, however, is mis-specified by a prior for α of $\text{normal}(25, 2)$ (dashed blue line), resulting in a posterior (salmon-colored histogram) not centered on the true value. In our experience it is quite rare to have a prior informed by ecological knowledge be so far off, but this is an example. How mis-calibrated the model will be depends on the data: we show examples with a sample size (N) of 5, 10 and 40 data points. In practice these studies would allow us to determine how much data we would need to be robust to suspect prior models.

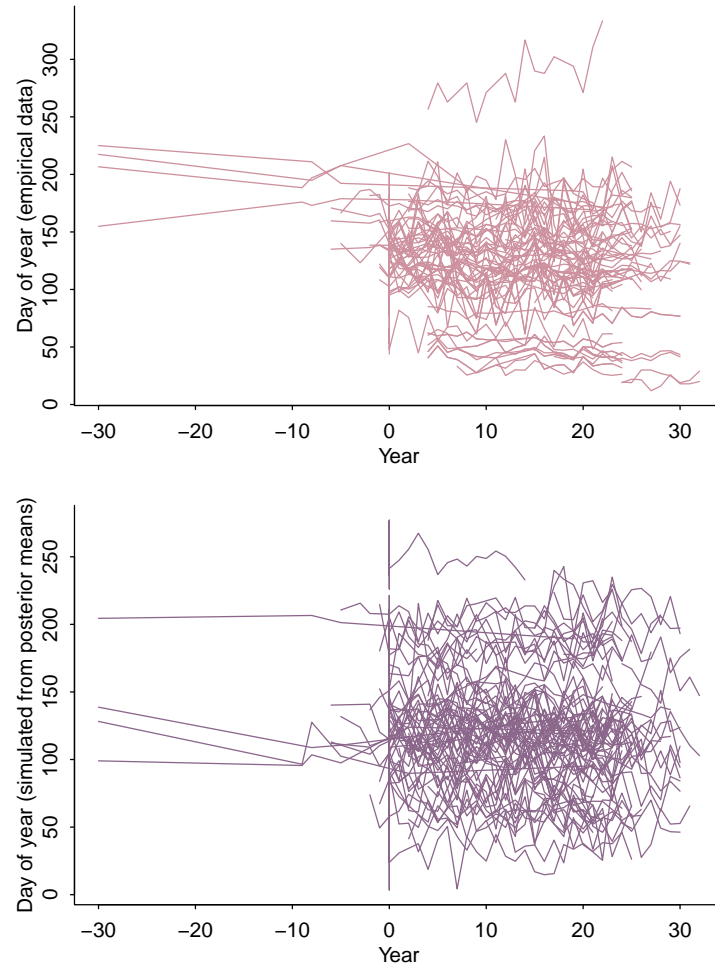


Figure 3: Example of a single retrodictive check from time-series data of phenological events over time. The raw data (top, pink) looks similar to one simulated dataset (bottom, purple), based on existing species number, their respective x data, and simulating from the parameters for each species. See ‘An example workflow’ in the Supplement for more details.