

# Closing the gap between statistical and scientific workflows for improved forecasts in ecology

Victor Van der Meersch, J. Regetz, T. J. Davies\* & EM Wolkovich

April 1, 2025

\* Says he is happy to help and give friendly review, but not sure he will reach level of co-author.

**Deadline:** 1 May 2025 (was 1 April 2025)

*For:* Scientific and Statistical Workflow theme issue for *Phil Trans A* as an *Opinion*

## Abstract

Increasing biodiversity loss and climate change have led to greater demands for useful ecological models and forecasts. Relevant datasets to meet these demands have also increased in size and complexity, including in their geographical, temporal and phylogenetic scales. While new research often suggests that accounting for these complexities variously increases, removes or otherwise alters major trends, I argue that the fundamental approach to model fitting in ecology makes it impossible to evaluate and compare models. These problems stem in part from continuing gaps between statistical workflows – where the data processing and model development are often addressed separately from the ecological question and aim – and scientific workflows, where all steps are integrated. Yet, as ecologists become increasingly computational, and new tools make it easier to share data, the opportunity to close this gap has never been greater. I outline how increased data simulation at multiple steps in the scientific workflow could revolutionize our understanding of ecological systems, yielding new insights. Combining these changes with more open model and data sharing – and developing new efforts to race the same data – could be transformative for ecological forecasting.

**Goal:** Increase awareness of how we can merge statistical and scientific workflows in ecology (especially forecasting) and what we would get out of it.

## Introduction

Nature is increasingly threatened by multiple drivers of change, with a largely dominant influence of human activities (Díaz *et al.*, 2019). This ongoing biodiversity crisis is expected to increase in the next decades because of climate change, and will continue to alter ecosystem services and human well-being (IPBES, 2019). To support implementation of sustainable policies among the socioeconomic and environmental dimensions, it is critical to understand trends to date and be able to forecast future dynamics.

Estimation of global biodiversity indicators and current trends depends on large-scale and long-term datasets—across terrestrial, freshwater, and marine ecosystems (e.g. Dornelas *et al.*, 2018). These data, gathered opportunistically and from multiple sources, are often unbalanced and have geographic, temporal and taxonomic biases. Addressing these biases requires the use of appropriate

statistical inference. Forecasting future changes—under different plausible scenarios—generally relies on either correlative models or process-based models (IPBES, 2019). The latter, which focus on a mechanistic representation of ecosystem functioning, are often promoted as the most realistic approach (Urban *et al.*, 2016; Pilowsky *et al.*, 2022).

The urgent need to answer policy-relevant questions has favored the proliferation of diverse methods developed by different researchers, lacking an overall coherence. Though there is no doubt nature is declining globally, significant uncertainty remains. There is no consensus on current species trends, with ongoing debates driven by widely varying reports that sometimes show conflicting trend directions (Dornelas *et al.*, 2014; Leung *et al.*, 2020; Buschke *et al.*, 2021; Johnson *et al.*, 2024). Future projections also diverge considerably, due to a high model uncertainty at the ecological level. Predictive modeling is increasingly relying on overly complex models (with a huge number of parameters), making it less adequate to generate new scientific insights (Franklin *et al.*, 2020).

Current controversies and focus on methodological aspects stem from the lack of coherence between current workflows in ecology (Loreau *et al.*, 2022; Talis & Lynch, 2023; Johnson *et al.*, 2024). Each new model development is added as a separate layer, disconnected from the original research aim, the data stream and the previous scientific insights. Workflows should fully integrate all the steps required to build a model from an ecological question, evaluate its limitations and degeneracies, before estimating its parameters and making projections. Here, we introduce an universal workflow that proposes to iteratively build upon all these steps, harmonizing both trend estimation and forecasting, with the aim of refocusing the debate on ecological questions and increasing the speed of scientific progress.

## Scientific method and workflows

Quantitative science rely on a model-based framework to confront hypothesis and data, making some approximations (). The general scientific method stresses out that the research question should guide both the design of the experiment and the corresponding model building. The experiment should then be conducted—according to this specific design. Finally, the resulting experimental data should be used to inform our model and differentiate between hypotheses—hopefully answering our initial question. However, this is often an idealized view.

Divergences from this scientific method are common. One explanation is the lack of rigor, leading to questionable practices, such as retrospectively crafting hypothesis to explain the results of a model rather than testing a clear pre-defined question (data-driven analysis, Fig. 1). But the reality of ecological research is also a major driver of the divergences from the ideal scientific method. Many important questions often cannot be addressed by conducting experiments and replications, and we must often rely on existing datasets to have a large-scale and long-term perspective (Hilborn & Mangel, 1997). However, this does not explain the persistent flaws and lack of overall coherence in ecological modeling. Trend estimation still often relies on a one-way ‘inference’, and most of the time is spent fitting the model to empirical data (Fig. 1). For forecasting, researchers focus on making predictions with complex models, but the steps of model building and parameterization are not very transparent and not clearly delineated (Fig. 1). The different parts of the model are often calibrated separately rather than as a whole, and some parameter values are just fixed based on experiments and expert knowledge.

To address these flaws while accounting for the realities of working with ecological data, a com-

prehensive workflow is needed to move along the data-model space, following a coherent sequence of steps (Fig 1). In particular, more efforts should be placed on model evaluation before incorporating any real data. This would force the modeler to acknowledge that some parameters might be non-identifiable and to reconsider the model structure. Similarly, it is essential to assess whether model predictions—once parameters are informed by data—are consistent with observations. The strength of such workflow lies in its flexibility, making it applicable to a wide range of modeling approaches, from simple trend analyses to more complex process-based models. At each step, the modeler need to critically examine its understanding of ecological processes, questioning previous assumptions, and explicitly acknowledge sources of uncertainty. This approach has the potential to enhance model interpretability and allow for a more transparent evaluation of model strengths and limitations. It also replace parameters at the core of the modeling process, as fundamental components that shape both inference and forecasting.

The first step of the workflow is to define an explicit research question and formulate some hypotheses (Fig 2). This involves making assumptions about the most influential drivers, within the specific context of our study. This step help us to think about the mechanisms that could generate the data we observe, including the observational error. Naturally, this leads to the development of a model—an ensemble of mathematical equations that encapsulates our knowledge and designed to answer our research question. The general idea is to start with a relatively simple model, that we could refine later. At this stage, prioritizing biologically meaningful parameters is crucial, as it allows us to have a sense of plausible parameter values. The next step is to simulate data that reflect our assumptions about the data structure and biases, by fixing parameter values to some values (which is straightforward if the parameters are interpretable). We then fit our model to this simulated dataset and evaluate its ability to recover the prescribed values. Structural degeneracies in the model might become apparent, revealing that some parameters are highly non-identifiable. This is the first feedback loop: if the model is not able to recover known parameter values, we must reconsider its structure. This way, we ensure the model works well on simulated data before incorporating real observations. We can then fit the model to real data, to obtain parameter estimates constrained by observations. Here, difficulty in fitting the model might indicate an inherent need for more data to address our initial question. This could lead us to either simplify our research question or—ideally—launch new data collection efforts. The next step is a new data simulation step, this time using our fitted model to generate predictions. This retrodictive check allows model output to be compared to observations. Discrepancies may indicate a missing key driver—perhaps an expected outcome if we known our initial model was too simplistic. This is a second feedback loop. We can refine the model to integrate the missing process and restart the workflow. Insights from the retrodictive check can also lead us to introduce additional complexity when simulating fake data, such as phylogenetic structure or observational biases (e.g. unbalanced data). This iterative evaluation of the model moves beyond a simple reliance on goodness-of-fit metrics. At each iteration, we are able to evaluate the model behavior, both with simulated and real data, taking into account our expert knowledge of the ecological processes. Within such a workflow, forecasting emerges as a natural outcome: rather than being a final goal, it only involves jointly modeling new circumstances along with the original data.

## **The workflow in practice**

Evidence of declining populations of vertebrate species in the latter half of the 20th century, alongside increasing ecosystem health concerns, led to growing public concern about protecting and maintaining the environment, challenging ecology to help predict and prevent further losses (Soulé & Terborgh, 1999; Soulé, 1991). The idea that important taxa were declining was clear from data from certain species and their populations, such as elephants and rhinos (Soule *et al.*, 1979; Leader-Williams *et al.*, 1990). Such trends drove a number of new subfields within ecology—some of which are now complete disciplines within themselves (such as conservation biology, Soulé, 1985)—focused on these problems, and potential ecological solutions to them CITES. Yet, as the magnitude and number of threats to these and other species have increased, with rates of habitat loss, overharvesting, pollution only increase, and anthropogenic climate change now clearly driving species loss (Waller *et al.*, 2017), so has the data and its complexity.

Ecologists have now amassed data on populations and species over the globe, they have also engaged in an increasing number of debates on regional and global trends, with arguments over the magnitude and even direction (Dornelas *et al.*, 2014; Leung *et al.*, 2020; Terry *et al.*, 2022; Müller *et al.*, 2024). The Living Planet Index (LPI), which aims to include long-term data on vertebrate populations of species across the globe is emblematic of these debates. With updated data released semi-annually (??) alongside new estimates of decline, a growing number of high-profile papers have challenged how strong the evidence is for population decline (Dornelas *et al.*, 2014; Gonzalez *et al.*, 2016; Wagner *et al.*, 2021; Müller *et al.*, 2024), with each paper taking a slightly different analytical approach. For example, Leung *et al.* (2020) published a mixture model that suggested most populations were not significantly declining, followed by other alternative modeling approaches (Buschke *et al.*, 2021; Puurtinen *et al.*, 2022) including a recent one suggesting a basic analysis of the dataset should always include three sources of autocorrelation, finding trends that encompassed most previous results (Johnson *et al.*, 2024).

Apparently conflicting results have made it harder for policy-makers to advance initiatives aimed at slowing declines, and has led to a debates within ecology about whether such analyses undermine public confidence in science (Gonzalez *et al.*, 2016). While shifting estimates are part of the process of science—refining our approaches and thus estimates over time—we argue much of the work underlying these debates stems from a poor workflow. In the current workflow for estimating trends over time a new model with a new estimate often leads to a paper (see Fig.) because ecologists spend far too little time interrogating their models with simulated data, or their model performance fit to empirical data. Functionally, research on the LPI has somewhat reverse-engineered the recommended workflow: after a series of papers debating different estimates from different models, more recent papers have focused on simulated data to highlight uncertainty given the model and data together (though I don’t think they link their simulations to the model they use that well, Dove *et al.*, 2023; Toszogyova *et al.*, 2024), but this should have been part of the process for the very first papers.

We argue than an improved workflow that required retrodictive checks and data simulation would lead to larger model advances and a greater recognition of uncertainty—thus highlighting likely consistency in estimates across models—that could better aid policy. Using the workflow would make what now appear as major discrepancies more obviously shifts in point estimates that are generally all in the same uncertainty space (Johnson *et al.*, 2024)—and it would challenge modelers to show major predictive advances, which is not currently part of the process. Explanatory

power in most models of observational data is usually very low (Low-Décarie *et al.*, 2014; Møller & Jennions, 2002) and thus tests of models’ predictions rarely expected. But the workflow highlights that predictions from the model—what we call retrodictive checks (or whatever we call them)—are part of the process of science, and critical to testing for what may be missing in a model. We expect retrodictive checks on most published trend analyses would highlight major missing components in these models (expand here?? ADD example?), and drive changes both in the models themselves and in the simulated data to check the models. While ecologists have started to use simulated data more to understand potential limitations of their models and data combined, this is still extremely rare, and efforts to date often treat simulations as separate from the statistical model (Buschke *et al.*, 2021; Dove *et al.*, 2023), short-circuiting their full utility

Applying the workflow to current trend estimates could importantly highlight the best way to improve data collection for more reliable estimates. Returning to the example of a global estimate of trends in vertebrate populations of species over time and applying our proposed workflow would mean more efforts to define the goal and question—is it a simple global estimate? Or a need to also find which species are declining most, including those that may have poor or no data? From there a generative model using simulated data for testing could incorporate many aspects of the populations, and data, that are often only included in ‘null’ or ‘synthetic data generation’ currently (Buschke *et al.*, 2021; McRae *et al.*, 2025) but could be built into the models fit to the empirical data. Eventually fitting the empirical data and performing retrodictive checks would likely highlight major missing components of the generative model. For example, certain populations are recovering for very specific reasons (e.g., elephants in regions where the ivory trade drove declines in the past) that perhaps should be modeled. From this model, what data are most critically needed to address the updated aims would become clearer and could drive new data collection (Toszygova *et al.*, 2024).

Ecological forecasting is a broad field with a diverse range of methods. Process-based modeling is often considered the gold standard in ecology (Urban *et al.*, 2016; Pilowsky *et al.*, 2022) and beyond. Process-based models are built on explicit mathematical equations to describe (supposedly causal) relationships between environmental drivers and ecological responses. They also often incorporate empirical relationships, particularly when knowledge is incomplete or when some processes are intentionally omitted. Processes are often represented at different nested spatiotemporal scales, depending on the underlying assumptions. Model development is the central step, typically requiring several years, yet it often remains opaque from an external perspective. The step of designing the model—translating knowledge and hypotheses into mathematical equations and parameters—is often blurred with the step of model calibration (or tuning), where parameter values are inferred. Models are often treated as an accumulation of multiple submodels, each governing one or several ecological processes. Rather than being fitted as a whole, submodels are calibrated separately against specific subsets of data, and some parameters are simply prescribed (i.e. fix to a value found in the literature) or tuned to reproduce some observations or theory. The way models are currently calibrated is not a coincidence, but rather an inappropriate way to accommodate their complexity, where many parameters compensate for one another.

These limitations are central to current debates. In climate modeling, increasing model complexity has not necessarily led to reduced uncertainty. For instance, the uncertainty range on the effect of increasing CO<sub>2</sub> concentration on temperature have remained largely unchanged (Zelinka

*et al.*, 2020). This has driven calls for more rigorous and transparent calibration processes (Balaji *et al.*, 2022). Similar concerns arise in ecology, where strong discrepancies exist between model projections (Cheaib *et al.*, 2012). Some researchers now advocate for the simplification of models, to avoid over-parametrization when the data provide little information to constrain some parameters (Wang *et al.*, 2017; Harrison *et al.*, 2021). If a model becomes too complex, understanding the sources of uncertainty and how they propagate through the model may become nearly impossible. Each additional process and parameter can increase overall uncertainty to the point where model projections lose their usefulness for decision makers (Saltelli *et al.*, 2020).

Applying the workflow to process-based models is a key for opening the black box. It would serve as a guide through the successive steps of model development. In particular, incorporating data simulation would introduce a crucial step between model building and data fitting, ensuring a clear delineation between the two and exposing strong degeneracies in the model design. This approach would force researchers to begin with a simpler version of the model, providing a clear pathway to support—or reject—the additional complexity and new parameters along the iterative development of the model. Model calibration would no longer be just a hidden aspect of model building but a step as crucial as forecasting to gain new ecological insights. The workflow could also refocus attention on the research question, defining a clear and limited context in which the model should apply. Process-based model would once again be a way to answer a research question—whereas today, model simulations have increasingly become a subject of study on their own. An universal workflow provides an opportunity to merge statistical and process-based approaches, integrating mechanistic knowledge and leveraging robust statistical approaches.

Across the different fields of ecology—for both parameter estimation and forecasting—a systematic application of a coherent workflow holds the promise to highlight the opportunities to best reduce uncertainties through new scientific insights, toward the most critical steps. This will help refocus the debate on designing new hypothesis, formulating new questions—and guiding efforts to collect new data. In a world where machine learning is rapidly advancing, there is no point of sticking to traditional methods if no changes are made. Machine learning may surpass process-based models if the latter lack a robust estimation of their parameters and fall in a complexity trap, at the cost of their interpretability. Similarly, trend analysis, when the focus is on methodological controversies (due to the lack of an iterative workflow) rather than on a robust mechanistic foundation, offers no clear advantage over machine learning.

### **Wrap up: how to make it happen?**

The misuse of models and misinterpretation of their outcomes (including by the authors) have multiple origins. One cause, well known, is the pressure to publish academic work, which can lower research standards. This challenge extends beyond the scope of this paper, yet we believe that applying an intelligible workflow can promote better research ethics. This a growing concern, leading to increase in reproducibility and data sharing practices. However, model development in ecology is still rarely transparent, which limits model understandability and prevents peers from properly identifying potential issues. A significant portion of scientific debate thus becomes lost in methodological considerations rather than advancing our ecological understanding.

## References

- Balaji, V., Couvreur, F., Deshayes, J., Gautrais, J., Hourdin, F. & Rio, C. (2022) Are general circulation models obsolete? *Proceedings of the National Academy of Sciences* **119**, e2202075119.
- Buschke, F.T., Hagan, J.G., Santini, L. & Coetzee, B.W.T. (2021) Random population fluctuations bias the living planet index. *Nature Ecology & Evolution* **5**, 1145–1152.
- Cheaib, A., Badeau, V., Boe, J., Chuine, I., Delire, C., Dufrêne, E., François, C., Gritti, E.S., Legay, M., Pagé, C., Thuiller, W., Viovy, N. & Leadley, P. (2012) Climate change impacts on tree ranges: model intercomparison facilitates understanding and quantification of uncertainty. *Ecology Letters* **15**, 533–544.
- Díaz, S., Settele, J., Brondízio, E.S., Ngo, H.T., Agard, J., Arneth, A., Balvanera, P., Brauman, K.A., Butchart, S.H.M., Chan, K.M.A., Garibaldi, L.A., Ichii, K., Liu, J., Subramanian, S.M., Midgley, G.F., Miloslavich, P., Molnár, Z., Obura, D., Pfaff, A., Polasky, S., Purvis, A., Razzaque, J., Reyers, B., Chowdhury, R.R., Shin, Y.J., Visseren-Hamakers, I., Willis, K.J. & Zayas, C.N. (2019) Pervasive human-driven decline of life on earth points to the need for transformative change. *Science* **366**.
- Dornelas, M., Antão, L.H., Moyes, F., Bates, A.E., Magurran, A.E., Adam, D., Akhmetzhanova, A.A., Appeltans, W., Arcos, J.M., Arnold, H., Ayyappan, N., Badihi, G., Baird, A.H., Barbosa, M., Barreto, T.E., Bässler, C., Bellgrove, A., Belmaker, J., Benedetti-Cecchi, L., Bett, B.J., Bjorkman, A.D., Błażewicz, M., Blowes, S.A., Bloch, C.P., Bonebrake, T.C., Boyd, S., Bradford, M., Brooks, A.J., Brown, J.H., Bruelheide, H., Budy, P., Carvalho, F., Castañeda-Moya, E., Chen, C.A., Chamblee, J.F., Chase, T.J., Siegwart Collier, L., Collinge, S.K., Condit, R., Cooper, E.J., Cornelissen, J.H.C., Cotano, U., Kyle Crow, S., Damasceno, G., Davies, C.H., Davis, R.A., Day, F.P., Degraer, S., Doherty, T.S., Dunn, T.E., Durigan, G., Duffy, J.E., Edelist, D., Edgar, G.J., Elahi, R., Elmendorf, S.C., Enemar, A., Ernest, S.K.M., Escribano, R., Estiarte, M., Evans, B.S., Fan, T., Turini Farah, F., Loureiro Fernandes, L., Farneda, F.Z., Fidelis, A., Fitt, R., Fosaa, A.M., Daher Correa Franco, G.A., Frank, G.E., Fraser, W.R., García, H., Cazzolla Gatti, R., Givan, O., Gorgone-Barbosa, E., Gould, W.A., Gries, C., Grossman, G.D., Gutierrez, J.R., Hale, S., Harmon, M.E., Harte, J., Haskins, G., Henshaw, D.L., Hermanutz, L., Hidalgo, P., Higuchi, P., Hoey, A., Van Hoey, G., Hofgaard, A., Holeck, K., Hollister, R.D., Holmes, R., Hoogenboom, M., Hsieh, C., Hubbell, S.P., Huettmann, F., Huffard, C.L., Hurlbert, A.H., Macedo Ivanauskas, N., Janík, D., Jandt, U., Jażdżewska, A., Johannessen, T., Johnstone, J., Jones, J., Jones, F.A.M., Kang, J., Kartawijaya, T., Keeley, E.C., Kelt, D.A., Kinneer, R., Klanderud, K., Knutsen, H., Koenig, C.C., Kortz, A.R., Král, K., Kuhn, L.A., Kuo, C., Kushner, D.J., Laguionie-Marchais, C., Lancaster, L.T., Min Lee, C., Lefcheck, J.S., Lévesque, E., Lightfoot, D., Lloret, F., Lloyd, J.D., López-Baucells, A., Louzao, M., Madin, J.S., Magnússon, B., Malamud, S., Matthews, I., McFarland, K.P., McGill, B., McKnight, D., McLarney, W.O., Meador, J., Meserve, P.L., Metcalfe, D.J., Meyer, C.F.J., Michelsen, A., Milchakova, N., Moens, T., Moland, E., Moore, J., Mathias Moreira, C., Müller, J., Murphy, G., Myers-Smith, I.H., Myster, R.W., Naumov, A., Neat, F., Nelson, J.A., Paul Nelson, M., Newton, S.F., Norden, N., Oliver, J.C., Olsen, E.M., Onipchenko, V.G., Pabis, K., Pabst, R.J., Paquette, A., Pardede, S., Paterson, D.M., Péliissier,

- R., Peñuelas, J., Pérez-Matus, A., Pizarro, O., Pomati, F., Post, E., Prins, H.H.T., Priscu, J.C., Provoost, P., Prudic, K.L., Pulliainen, E., Ramesh, B.R., Mendivil Ramos, O., Rassweiler, A., Rebelo, J.E., Reed, D.C., Reich, P.B., Remillard, S.M., Richardson, A.J., Richardson, J.P., van Rijn, I., Rocha, R., Rivera-Monroy, V.H., Rixen, C., Robinson, K.P., Ribeiro Rodrigues, R., de Cerqueira Rossa-Feres, D., Rudstam, L., Ruhl, H., Ruz, C.S., Sampaio, E.M., Rybicki, N., Rypel, A., Sal, S., Salgado, B., Santos, F.A.M., Savassi-Coutinho, A.P., Scanga, S., Schmidt, J., Schooley, R., Setiawan, F., Shao, K., Shaver, G.R., Sherman, S., Sherry, T.W., Siciński, J., Sievers, C., da Silva, A.C., Rodrigues da Silva, F., Silveira, F.L., Slingsby, J., Smart, T., Snell, S.J., Soudzilovskaia, N.A., Souza, G.B.G., Maluf Souza, F., Castro Souza, V., Stallings, C.D., Stanforth, R., Stanley, E.H., Mauro Sterza, J., Stevens, M., Stuart-Smith, R., Rondon Suarez, Y., Supp, S., Yoshio Tamashiro, J., Tarigan, S., Thiede, G.P., Thorn, S., Tolvanen, A., Teresa Zugliani Toniato, M., Totland, Ø., Twilley, R.R., Vaitkus, G., Valdivia, N., Vallejo, M.I., Valone, T.J., Van Colen, C., Vanaverbeke, J., Venturoli, F., Verhey, H.M., Vianna, M., Vieira, R.P., Vrška, T., Quang Vu, C., Van Vu, L., Waide, R.B., Waldock, C., Watts, D., Webb, S., Wesolowski, T., White, E.P., Widdicombe, C.E., Wilgers, D., Williams, R., Williams, S.B., Williamson, M., Willig, M.R., Willis, T.J., Wipf, S., Woods, K.D., Woehler, E.J., Zawada, K. & Zettler, M.L. (2018) Biotime: A database of biodiversity time series for the anthropocene. *Global Ecology and Biogeography* **27**, 760–786.
- Dornelas, M., Gotelli, N.J., McGill, B., Shimadzu, H., Moyes, F., Sievers, C. & Magurran, A.E. (2014) Assemblage time series reveal biodiversity change but not systematic loss. *Science* **344**, 296–299.
- Dove, S., Böhm, M., Freeman, R., McRae, L. & Murrell, D.J. (2023) Quantifying reliability and data deficiency in global vertebrate population trends using the living planet index. *Global Change Biology* **29**, 4966–4982.
- Franklin, O., Harrison, S.P., Dewar, R., Farrior, C.E., Brännström, A., Dieckmann, U., Pietsch, S., Falster, D., Cramer, W., Loreau, M., Wang, H., Mäkelä, A., Rebel, K.T., Meron, E., Schymanski, S.J., Rovenskaya, E., Stocker, B.D., Zaehle, S., Manzoni, S., van Oijen, M., Wright, I.J., Ciais, P., van Bodegom, P.M., Peñuelas, J., Hofhansl, F., Terrer, C., Soudzilovskaia, N.A., Midgley, G. & Prentice, I.C. (2020) Organizing principles for vegetation dynamics. *Nature Plants* **6**, 444–453.
- Gonzalez, A., Cardinale, B.J., Allington, G.R., Byrnes, J., Arthur Endsley, K., Brown, D.G., Hooper, D.U., Isbell, F., O'Connor, M.I. & Loreau, M. (2016) Estimating local biodiversity change: a critique of papers claiming no net loss of local diversity. *Ecology* **97**, 1949–1960.
- Harrison, S.P., Cramer, W., Franklin, O., Prentice, I.C., Wang, H., Brännström, A., de Boer, H., Dieckmann, U., Joshi, J., Keenan, T.F., Lavergne, A., Manzoni, S., Mengoli, G., Morfopoulos, C., Peñuelas, J., Pietsch, S., Rebel, K.T., Ryu, Y., Smith, N.G., Stocker, B.D. & Wright, I.J. (2021) Eco-evolutionary optimality as a means to improve vegetation and land-surface models. *New Phytologist* **231**, 2125–2141.
- Hilborn, R. & Mangel, M. (1997) *The Ecological Detective: Confronting Models with Data*. Princeton University Press.



- IPBES (2019) Global assessment report on biodiversity and ecosystem services of the intergovernmental science-policy platform on biodiversity and ecosystem services. Tech. rep.
- Johnson, T.F., Beckerman, A.P., Childs, D.Z., Webb, T.J., Evans, K.L., Griffiths, C.A., Capdevila, P., Clements, C.F., Besson, M., Gregory, R.D., Thomas, G.H., Delmas, E. & Freckleton, R.P. (2024) Revealing uncertainty in the status of biodiversity change. *Nature* **628**, 788–794.
- Leader-Williams, N., Albon, S. & Berry, P. (1990) Illegal exploitation of black rhinoceros and elephant populations: patterns of decline, law enforcement and patrol effort in luangwa valley, zambia. *Journal of applied ecology* pp. 1055–1087.
- Leung, B., Hargreaves, A.L., Greenberg, D.A., McGill, B., Dornelas, M. & Freeman, R. (2020) Clustered versus catastrophic global vertebrate declines. *Nature* **588**, 267–271.
- Loreau, M., Cardinale, B.J., Isbell, F., Newbold, T., O’Connor, M.I. & de Mazancourt, C. (2022) Do not downplay biodiversity loss. *Nature* **601**, E27–E28.
- Low-Décarie, E., Chivers, C. & Granados, M. (2014) Rising complexity and falling explanatory power in ecology. *Frontiers in Ecology and the Environment* **12**, 412–418.
- McRae, L., Cornford, R., Marconi, V., Puleston, H., Ledger, S.E., Deinet, S., Oppenheimer, P., Hoffmann, M. & Freeman, R. (2025) The utility of the living planet index as a policy tool and for measuring nature recovery. *Philosophical Transactions B* **380**, 20230207.
- Møller, A. & Jennions, M.D. (2002) How much variance can be explained by ecologists and evolutionary biologists? *Oecologia* **132**, 492–500.
- Müller, J., Hothorn, T., Yuan, Y., Seibold, S., Mitesser, O., Rothacher, J., Freund, J., Wild, C., Wolz, M. & Menzel, A. (2024) Weather explains the decline and rise of insect biomass over 34 years. *Nature* **628**, 349–354.
- Pilowsky, J.A., Colwell, R.K., Rahbek, C. & Fordham, D.A. (2022) Process-explicit models reveal the structure and dynamics of biodiversity patterns. *Science Advances* **8**, eabj2271.
- Puurtinen, M., Elo, M. & Kotiaho, J.S. (2022) The living planet index does not measure abundance. *Nature* **601**, E14–E15.
- Saltelli, A., Bammer, G., Bruno, I., Charters, E., Di Fiore, M., Didier, E., Nelson Espeland, W., Kay, J., Lo Piano, S., Mayo, D., Pielke Jr, R., Portaluri, T., Porter, T.M., Puy, A., Rafols, I., Ravetz, J.R., Reinert, E., Sarewitz, D., Stark, P.B., Stirling, A., van der Sluijs, J. & Vineis, P. (2020) Five ways to ensure that models serve society: a manifesto. *Nature* **582**, 482–484.
- Soulé, M.E. (1985) What is conservation biology? *BioScience* **35**, 727–734.
- Soulé, M.E. (1991) Conservation: tactics for a constant crisis. *Science* **253**, 744–750.
- Soulé, M.E. & Terborgh, J. (1999) Conserving nature at regional and continental scales—a scientific program for north america. *BioScience* **49**, 809–817.

- Soule, M.E., Wilcox, B.A. & Holtby, C. (1979) Benign neglect: a model of faunal collapse in the game reserves of east africa. *Biological Conservation* **15**, 259–272.
- Talis, E.J. & Lynch, H.J. (2023) Capturing stochasticity properly is key to understanding the nuances of the living planet index. *Nature Ecology & Evolution* **7**, 1194–1195.
- Terry, J.C.D., O’Sullivan, J.D. & Rossberg, A.G. (2022) No pervasive relationship between species size and local abundance trends. *Nature Ecology & Evolution* **6**, 140–144.
- Toszogyova, A., Smyčka, J. & Storch, D. (2024) Mathematical biases in the calculation of the living planet index lead to overestimation of vertebrate population decline. *Nature Communications* **15**, 5295.
- Urban, M.C., Bocedi, G., Hendry, A.P., Mihoub, J.B., Pe’er, G., Singer, A., Bridle, J.R., Crozier, L.G., De Meester, L., Godsoe, W., Gonzalez, A., Hellmann, J.J., Holt, R.D., Huth, A., Johst, K., Krug, C.B., Leadley, P.W., Palmer, S.C.F., Pantel, J.H., Schmitz, A., Zollner, P.A. & Travis, J.M.J. (2016) Improving the forecast for biodiversity under climate change. *Science* **353**, aad8466.
- Wagner, D.L., Grames, E.M., Forister, M.L., Berenbaum, M.R. & Stopak, D. (2021) Insect decline in the anthropocene: Death by a thousand cuts. *Proceedings of the National Academy of Sciences* **118**, e2023989118.
- Waller, N.L., Gynther, I.C., Freeman, A.B., Lavery, T.H. & Leung, L.K.P. (2017) The bramble cay melomys melomys rubicola (rodentia: Muridae): a first mammalian extinction caused by human-induced climate change? *Wildlife Research* **44**, 9–21.
- Wang, H., Prentice, I.C., Keenan, T.F., Davis, T.W., Wright, I.J., Cornwell, W.K., Evans, B.J. & Peng, C. (2017) Towards a universal model for carbon dioxide uptake by plants. *Nature Plants* **3**, 734–741.
- Zelinka, M.D., Myers, T.A., McCoy, D.T., Po-Chedley, S., Caldwell, P.M., Ceppi, P., Klein, S.A. & Taylor, K.E. (2020) Causes of higher climate sensitivity in cmip6 models. *Geophysical Research Letters* **47**.