

The legacy of episodic climatic events in shaping temperate, broadleaf forests

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Abstract. In humid, broadleaf-dominated forests where gap dynamics and partial canopy mortality appears to dominate the disturbance regime at local scales, paleoecological evidence shows alteration at regional-scales associated with climatic change. Yet, little evidence of these broad-scale events exists in extant forests. To evaluate the potential for the occurrence of large-scale disturbance, we used 76 tree-ring collections spanning $\sim 840\,000 \text{ km}^2$ and 5327 tree recruitment dates spanning $\sim 1.4 \text{ million km}^2$ across the humid eastern United States. Rotated principal component analysis indicated a common growth pattern of a simultaneous reduction in competition in 22 populations across $61\,000 \text{ km}^2$. Growth-release analysis of these populations reveals an intense and coherent canopy disturbance from 1775 to 1780, peaking in 1776. The resulting time series of canopy disturbance is so poorly described by a Gaussian distribution that it can be described as “heavy tailed,” with most of the years from 1775 to 1780 comprising the heavy-tail portion of the distribution. Historical documents provide no evidence that hurricanes or ice storms triggered the 1775–1780 event. Instead, we identify a significant relationship between prior drought and years with elevated rates of disturbance with an intense drought occurring from 1772 to 1775. We further find that years with high rates of canopy disturbance have a propensity to create larger canopy gaps indicating repeated opportunities for rapid change in species composition beyond the landscape scale. Evidence of elevated, regional-scale disturbance reveals how rare events can potentially alter system trajectory: a substantial portion of old-growth forests examined here originated or were substantially altered more than two centuries ago following events lasting just a few years. Our recruitment data, comprised of at least 21 species and several shade-intolerant species, document a pulse of tree recruitment at the subcontinental scale during the late-1600s suggesting that this event was severe enough to open large canopy gaps. These disturbances and their climatic drivers support the hypothesis that punctuated, episodic, climatic events impart a legacy in broadleaf-dominated forests centuries after their occurrence. Given projections of future drought, these results also reveal the potential for abrupt, meso- to large-scale forest change in broadleaf-dominated forests over future decades.

Key words: *closed-canopy forests; deciduous forests; disturbance; drought; frost; gap dynamics; historical documents; mesoscale dynamics; tree-ring analysis.*

INTRODUCTION

Anthropogenic climate change is altering the structure and function of forested ecosystems and challenging our understanding of how systems react to disturbance.

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Increased warming has complex effects on ecosystem dynamics through the interaction of drought, species tolerances, insect outbreaks, fire regimes and other perturbations (e.g., Allen and Breshears 1998, Taylor et al. 2006, Worrall et al. 2008, van Mantgem et al. 2009, Allen et al. 2010). Climatic variation drives ecosystem dynamics at large scales in semiarid and boreal forests, where tree species richness is relatively low (e.g., Payette et al. 1985, Savage et al. 1996, Villalba and Veblen 1997, 1998, Allen and Breshears 1998, Brown and Wu 2005, Raffa et al. 2008). An extreme example includes massive forest mortality where one species comprised 95% of the

composition (Michaelian et al. 2011). While recent disturbances appear to be partly driven by anthropogenic warming, they generally fit the expected disturbance regime in these forests and provide insight into how climate modulates these systems at regional scales.

In contrast to these systems, many forests, especially in temperate, broadleaf forests, experience gap dynamics and partial canopy mortality as the prevailing disturbance regime at stand to landscape scales (Runkle 1982, Yamamoto 1992, Rebertus and Veblen 1993, McCarthy 2001, Gutierrez et al. 2008). These small-scale events are thought to be largely stochastic and not directly tied to climate. Triggered by individual tree mortality, these dynamics could aid in the maintenance of stand composition and structural complexity and play a role in prolonged forest stability at large scales (Runkle 1982, Yamamoto 1992, McCarthy 2001, Gutierrez et al. 2008), an observation that is supported through long-term, simulated forest dynamics (Smith and Urban 1988). Long-term forest stability is primarily achieved through in-filling by lateral branches in small gaps or filling through recruitment of more shade-tolerant species. Temperate forests in humid regions dominated by broadleaf species often have high species diversity, making them resilient to species-specific disturbances like insect outbreaks. In marked contrast with low-diversity coniferous forests, for example, the broadleaf-dominated forests of the eastern United States have no native forest insects that trigger large-scale dieback (Man 2012). And, unlike semiarid or coniferous systems, fire in this region is typically fine-scaled, spatially heterogeneous (Clark and Royall 1996, Parshall and Foster 2002, Guyette et al. 2006, McEwan et al. 2007) and less likely to occur at regional scales (though see McMurry et al. [2007] for an exception). Thus, dynamics in broadleaf-dominated forests are often characterized as asynchronous in space and time and not seen as strongly influenced by climatic variation like other forest types (Schleeweis et al. 2013, Vanderwel et al. 2013). Given future climate forecasts, it is imperative that we understand how broadleaf forests in humid regions might respond to climate change at large spatial scales.

Paleoecological studies have documented regional-scale changes in humid forests driven by climate. Specifically, drought shaped the structure and composition of these systems during the Holocene (Jackson and Booth 2002, Shuman et al. 2004, 2009, Foster et al. 2006, Booth et al. 2012). For example, Booth et al. (2012) recently showed how the decline of American beech during the Medieval Climate Anomaly was triggered by drought and climatic variability in the moist, lake-effect-influenced forests in the Great Lakes region. While paleoecological studies are critical in identifying long-term, regional-scale forest dynamics, they typically can only identify the timing of forest dynamics at decadal to multi-decadal scales and at broad compositional (i.e., genera) levels. How this translates to contemporary systems is less certain. To

date, disturbance histories of extant old-growth forests in humid regions primarily identify stand-scale (Runkle 1982, Yamamoto 1992, McCarthy 2001) to moderate-scale dynamics (White et al. 1999, Frelich 2002, Woods 2004, Baker et al. 2005, Worrall et al. 2005, Stueve et al. 2011), although large-scale dynamics have been postulated (e.g., Cho and Boerner 1995). Thus, there is a disparity between paleoecological records of regional-scale forest dynamics and recent forest history reconstructions that primarily document gap dynamics. Bridging this divide for broadleaf-dominated forests could yield important analogues for how forests will respond to climate change at time scales relevant for climate forecasts and human decision making.

Here we develop and analyze two large data sets to investigate the possibility of synchronous, regional-scale disturbance and tree recruitment over the last 500 years of the Common Era (CE) in the humid, broadleaf-dominated region of the eastern United States. We test a null hypothesis of small-scale (gap) dynamics under which disturbance and tree recruitment resemble a stochastic or white-noise process. Explicitly, forest dynamics would not be regionally synchronous and time-series of disturbance and recruitment would not deviate significantly above the long-term background rate. Under the alternative hypothesis, forest dynamics would be regionally synchronous and time-series of disturbance and recruitment would be punctuated by extreme events. We know of no study that has identified severe, synchronous, regional-scale disturbance in temperate humid regions outside of studies covering decadal to centennial time scales (sediment core records). In fact, two tree-ring-based investigations designed to identify regional-scale disturbance did not reveal evidence for synchronous disturbance over the last 300+ years (Retch 2003, D'Amato and Orwig 2008). The first data set for our investigation, 76 tree-ring records over a $\sim 840\,000\text{ km}^2$ area, is examined for temporal clustering of large-scale canopy disturbance. The second data set, >5000 tree recruitment dates from multiple sources that cover ~ 1.4 million km^2 , are examined for evidence of synchronous, large-scale recruitment events.

Identifying the exact triggers of historical disturbance from tree rings or other types of ecological proxies can be difficult given that many factors drive forest dynamics. We thus turned to historical documents to assist in identifying potential drivers of disturbance. Historical documents have been used to clarify the potential impact of climate on human societies (e.g., Endfield 2012) and in physical geographical and ecological studies (e.g., Hooke and Kain 1982). Here, we use historical documents to conduct a “ground-truthing” of events in tree-ring records. Several sources of historical documents were used including a compilation of Atlantic hurricanes (Chenoweth 2006), diaries of frontier settlers, Moravians, in western North Carolina (Fries 1922, 1925, 1926), Thomas Jefferson's Garden Book (1774) and 18th-century newspaper and other

accounts (e.g., Ludlum 1963). While many of the Moravian records contain information about births, marriages, deaths, civil unrest, and smallpox, they are also a rich source of information for historical ecology, documenting events that impacted agriculture (like hail storms, oppressive heat, and individual rain events; see *Examples of weather observation from the Moravian Diaries* in Appendix A for examples). We note, however, that their records are of limited spatial coverage and cannot account for possible impacts across our study region. Given their detailed observations of important weather events, spatial coverage of the records, and multiple observers at potentially daily temporal resolution, it is possible that significant, large-scale, weather events were documented, which could complement our understanding of forest dynamics from tree-ring records.

METHODS

We investigate the potential for regional scale-disturbance using a network of 76 tree-ring chronologies from the International Tree-Ring Databank (ITRDB) covering $\sim 840\,000 \text{ km}^2$ in the broadleaf-dominated region of the humid eastern United States (also called the Eastern Deciduous Forest; Fig. 1a; see Table A1). Tree-ring records from sites in southeastern Kentucky, central Tennessee, and North Carolina indicated the possibility of large-scale disturbance in the late 1770s (see *Temporal and geographic disturbance detection methods* in Appendix A: Fig. A1, and two of these forests in Appendix B). With a larger data set, we sought to test the hypothesis that a regional-scale disturbance event occurred during this period. If there is evidence of a regional event, the larger network will allow us to determine its spatial extent. The network is composed of at least 11 species collected by 15 investigators to study climate, ecology, and fire history. Most trees were targeted for maximum age, drought sensitivity, or fire history, although some were collected via random sampling (Appendix A: Table A1). Most of the collections for climatic reconstruction were not derived from the canonical “open-grown” trees on highly stressed sites (see Fritts 1976), but from closed-canopy forests; 44 of the 49 sites were either visited by a co-author (N. Pederson) or confirmed to be closed-canopy from the collector (E. Cook, *personal communication*). To avoid inflating results, the Blanton Forest populations were combined during analysis, as two *Quercus* species were intermingled throughout the forest.

Tree-ring chronologies were selected if they met several criteria, including species composition (trees from broadleaf-dominated forests), chronology length (inner ring date pre-1750), geographic location, and the likelihood that they represented old-growth forest. We targeted broadleaf-dominated forest to increase the likelihood that small-scale dynamics was the predominant disturbance regime; we did not, for example, include species such as *Pinus resinosa* or *Pinus rigida*, which would be expected to have more episodic

recruitment because of a fire-dominated disturbance regime. We were not able to meet these criteria in all cases. For example, 12 of the 76 chronologies came from conifer-dominated forests with possible episodic recruitment (*Picea rubens*, *Pinus echinata*, *Pinus* spp.; Appendix A: Table A1); *Tsuga canadensis* is an important component of this network, but generally has more of a gap-dynamics life history trait in reaching the forest canopy. To increase the likelihood that potential collections for analysis were not from logged forests, we excluded collections if (1) one-third or more of trees with rapid, early growth were recruited in a single cohort after ca. 1850 or (2) a substantial number of old trees showed major growth releases during this same period ($>100\%$ change in growth; see Lorimer [1985], Lorimer and Frelich [1989]). Even though little stand-scale logging occurred in the mountainous and plateau areas of our study region prior to 1850 (Williams 1992: Fig. 6.4), there is the possibility that stands with only natural disturbance were omitted from this analysis. Since many of the collections are from extant old-growth forests, it is less likely that they have been logged because of the common relationship between low site productivity and old trees (Stahle and Chaney 1994). For finer details on discerning expected radial growth patterns and other considerations of our approach, see *Temporal and geographic disturbance detection methods* in Appendix A. The ultimate goal in omitting some trees and collections was to avoid false-positive results that could occur if all data were included in the initial analysis.

To objectively detect if step-change increases in radial increment might resemble a regional-scale canopy disturbance, raw ring widths were standardized using a straight-line fit to remove differences in mean growth rate. The resulting standard chronologies were entered into rotated varimax principal component analysis (RPCA). RPCA identifies the highest loadings of each variable on a single eigenvector while maintaining orthogonality and maximizing variance of retained eigenvectors (Richman 1986). The Monte Carlo “Rule-N” technique was used to determine the number of eigenvectors to use for analysis (Preisendorfer et al. 1981). RPCA retained eight significant eigenvectors, accounting for 62.9% of the common variation (see Table A2 and *Temporal and geographic disturbance detection methods* in Appendix A). We report on the first three for this study. Eigenvector 1 (EV1), accounting for 20.8% of the common variation, represents the temporal decline in ring widths as constrained by allometry (Appendix A: Fig. A2a). Eigenvector 2 (EV2), accounting for 11.1% of the common variance, has a large, step-change in radial increment around 1780 (Fig. 1b). Eigenvector 3 (EV3), accounting for 8.3% of the common variance, reveals an abrupt increase in ring widths in 1840 and 1857 (Appendix A: Fig. A2b, Table A2). As EV1 reflects the commonly observed pattern of large rings when trees are small and there is a good

chance that EV3 likely represents the era of European settlement, we focus on investigating the potential for a regional disturbance event among the populations loading strongly on EV2. A complete analysis and discussion of EV3 is beyond the goals of this study.

To investigate whether the step-change in radial increment of EV2 was a regional-scale disturbance event, disturbance history was reconstructed from the 22 chronologies loading positively and significantly onto EV2 (≥ 0.224 , $P \leq 0.05$; following Koutsoyiannis [1977]). Evidence of canopy disturbance in individual raw ring-width series of the 22 chronologies was investigated using conservative methods adopted from a traditional approach. Here, a major canopy disturbance is an increase in radial growth of $>99.9\%$ over a 15-year period relative to the prior 15 years; a minor release is an increase in radial growth of 50–99.9% over 15-year periods ([Lorimer and Frelich 1989]; for a deeper discussion on potential pitfalls for these methods, see *Tree and population-level disturbance history analysis* in Appendix A). A release $>99.9\%$ is considered an opportunity for an understory tree to reach the canopy (cf. canopy accession; Lorimer and Frelich 1989). While these methods were primarily developed for shade-tolerant species, previous work indicates that these methods are effective for reconstructing disturbance histories using shade-intolerant species (Orwig and Abrams 1994), even in a “gappy forest” (McGuire et al. 2001) dominated by shade-intolerant species (see Pederson et al. 2008). Our method might not detect multiple disturbances in some trees because trees lose some sensitivity to changes in competition as they gain canopy status (Nowacki and Abrams 1997, Renth et al. 2002, Druckenbrod et al. 2013). However, many trees in our data set recorded more than one disturbance. Further, a test of lower detection thresholds does not alter our findings, but does seem to be sensitive to climate in ways that could result in false positives (see *Detection sensitivity analysis* in Appendix A). Because of the methodology used here, we do not have a complete record of disturbance history. And, given that most ITRDB data are composed of only 20–30 canopy trees per stand, we would not expect to detect all possible disturbances in a particular forest, although increased core replication of the ITRDB collections likely increases the chance of detecting disturbance vs. single-core studies (Copenheaver et al. 2009). Ultimately, our final time-series of canopy disturbance should reflect a lower number of false positives and, more importantly, large canopy gap formation, which should have a more meaningful impact on forest composition and structure than smaller gaps.

Statistical analysis of the disturbance record indicated a heavy tail (i.e., large-disturbance events) that strongly deviated from a Gaussian distribution (see Fig. 2c, d). Therefore, we applied tools from extreme value theory (Davison and Smith 1990, Coles 2001), to analyze the statistical properties of extremes in the disturbance

record (i.e., years with many recorded disturbances). In this framework, we use a peak-over threshold approach, based on the generalized Pareto distribution (GPD), to investigate the tail properties of the time-series of canopy disturbance. We fit a GPD to disturbance events above a disturbance rate of 1%. The determination of a suitable threshold for which the asymptotic GPD approximation holds is an essential step that requires the consideration of a trade-off between bias and variance (e.g., Coles 2001). Note that if a threshold is chosen too low, the GPD will fit the exceedances poorly and introduce a bias in the estimates, while if a threshold is chosen too high it will reduce the number of exceedances and thus increase the estimation variance. In practice, threshold choice involves comparing the theoretical behavior of the GPD with the empirical behavior of the data. Tools like the mean residual life plot assist in the threshold choice, and if the observations follow a GPD with a shape parameter <1 , the mean exceedance should vary linearly with the threshold. For our application we choose the threshold as a disturbance rate of 1%. The rationale behind this threshold choice is that (1) it fulfills the statistical criteria described here and (2) it allows us to consider more moderate disturbances (that lie clearly above the internal variability of the data record; as a disturbance rate of 1% falls at about the 80th percentile of the record) together with the “high impact” extremes. This model described the high tail much better than the Gaussian distribution. In addition, return intervals (in years T) can be described from the probability of exceeding a disturbance rate x within a time window T directly from the fitted GPD.

We tallied the number of major canopy disturbances per year from the 22 collections and created an index of disturbance magnitude. The magnitude index, based on a method to compare climatic events of differing lengths and intensities (Biondi et al. 2005, Gray et al. 2011), is calculated as

$$M_I = D \times DR \times MR$$

where M_I is the magnitude index, D is the duration in years, DR is the anomaly of the disturbance rate as the departure from the mean of the percentage of trees disturbed per year from 1685 to 1880, and MR is the percentage of releases qualifying as a “major release” during each event. The magnitude index describes the intensity of each release event. Canopy disturbance analysis is limited to 1685–1880 because it represents the period when tree replication ≥ 100 (1685) and precedes large-scale logging (ca. 1880; Williams 1992) when we hypothesize that trees from ITRDB collections would have reduced sensitivity to changes in competition. As many collections from the ITRDB contain trees >200 years of age at the time of sampling, it is not unreasonable to expect that most of these trees, especially species like *Quercus alba* or *Liriodendron tulipifera*, would have reached canopy status within 100–150 years

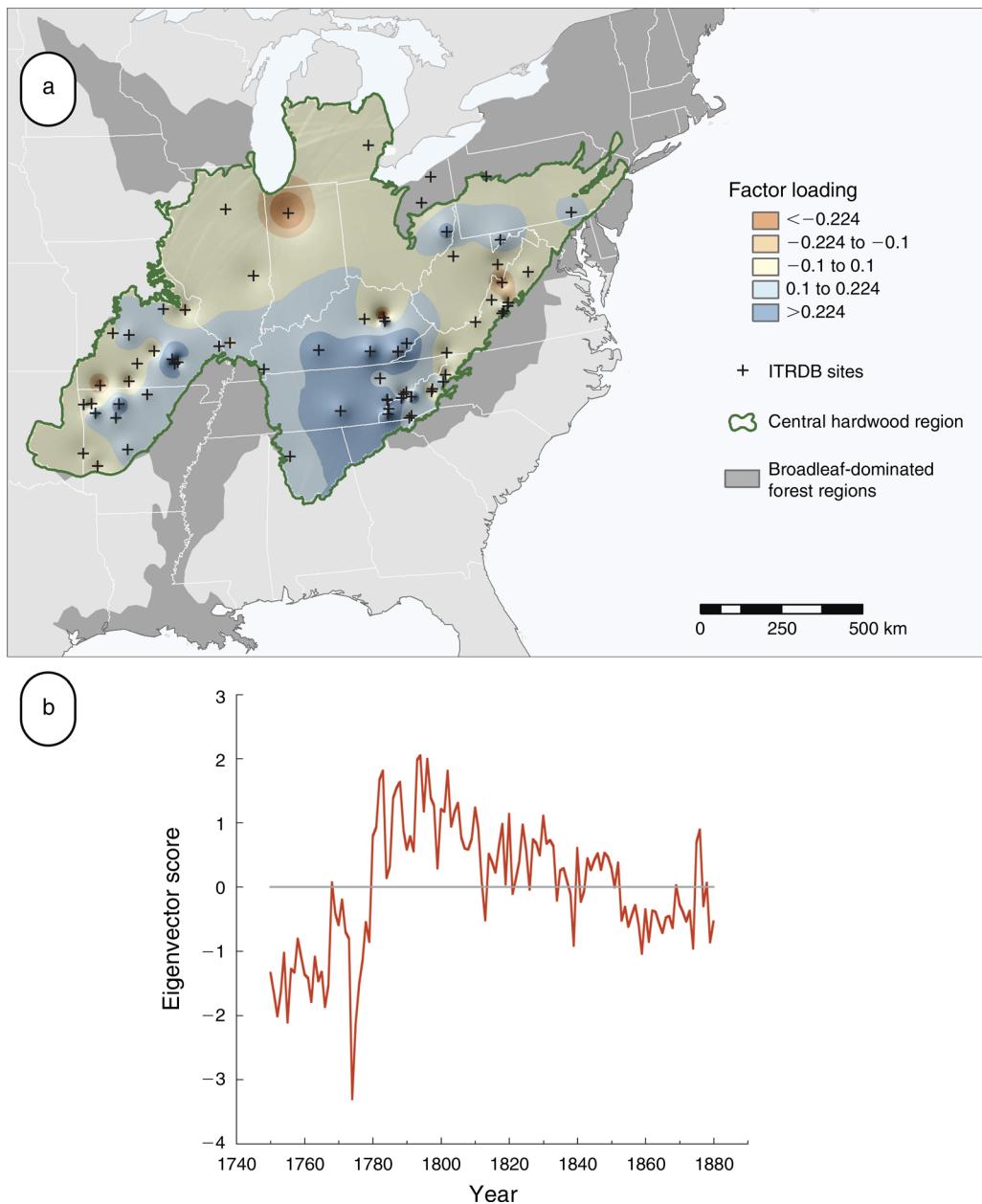


FIG. 1. Spatial loading of populations onto eigenvector 2 (EV2). The central hardwood region is adapted from Fralish and Franklin (2002), and the U.S. Environmental Protection Agency (2003); the broadleaf-dominated forest regions are adapted from Dyer (2006). Plus signs show International Tree-Ring Databank (ITRDB) sites. (b) Time-series of ring width variation of EV2 from 1750–1880 of populations within the dark blue areas in panel a. The gray line in panel b represents the 1750–1880 mean.

prior to sampling; tree-ring sampling in the eastern United States began in earnest during the late 1970s (e.g., Cook and Jacoby 1977, 1983, Cook 1982, Stahle et al. 1985, Stahle and Cleaveland 1988). Reliance on the oldest trees in a reconstruction of disturbance history has the potential to be biased as one moves closer to the period of sampling (McEwan et al. 2014). Thus, we expect the mid-1800s to be the beginning of reduced tree sensitivity to release in our data set, a trend observable in our record.

We then investigated the relationship between canopy disturbance and drought in two ways. First, we made composite drought maps from the North American Drought Atlas (NADA; Cook and Krusic 2004, Cook et al. 2004) for the years prior to the 23 years with elevated disturbance (years with disturbance ≥ 1.0 SD above the 1680–1880 mean) after we noticed that elevated disturbance often followed regional marker rings. Second, we used superposed epoch analysis (SEA; Swetnam 1993) to examine moisture conditions before,

during, and after years with elevated disturbance. Because some of the tree-ring records used for disturbance analysis are used in the NADA, we developed an independent drought proxy (IDP) to test for a relationship between drought and disturbance. IDP is a tree-ring-based proxy of drought using records not used for disturbance analysis from within and around the periphery of the late-1770s disturbance region (see *Creation of the independent drought proxy for superposed epoch analysis* in Appendix A).

We mined published and unpublished data sets of tree establishment dates from old-growth forests dominated by broadleaf species to reconstruct regional-scale recruitment history across the eastern United States (see Supplement). Like our disturbance-detection analysis, forests that would be expected to have episodic recruitment, i.e., pine-dominated forests, were avoided. While some scattered conifers within broadleaf-dominated forests are included in this analysis, recruitment dates for these trees were drawn from broadleaf-dominated forests or conifers with more of a gap-phase life history trait like *Tsuga canadensis* and *Tsuga caroliniana*. We examined a larger area than that of the 76 chronologies for disturbance detection because a review of independent and geographically dispersed studies explicitly discussed a recruitment event, broad compositional change, or stand initiation dates in the late 1600s (Huntington 1914, Haasis 1923, Hough and Forbes 1943, Henry and Swan 1974, Grimm 1983, Guyette et al. 1994, Renth 2003). Thus, this data set would be another test of regional-scale disturbance in forests dominated by gap dynamics. We only examined recruitment dates prior to 1850 to reduce the influence of widespread regional land-use change associated with land clearing and cutting (Williams 1992). The final data set includes 49 published studies from 56 different stands comprised of 5327 individual tree establishment dates (Appendix A: Fig. A6). These studies used a variety of methods to investigate long-term development of old-growth forests at local scales, although Renth (2003) is the exception with five sites distributed across $\sim 30\,000\text{ km}^2$. Recruitment dates were tallied from 34 species, not counting “other” and “unknown” categories as other species. The most common species are *Tsuga canadensis*

(25.6%) and *Quercus alba* (21.3%). Eight *Quercus* species accounted for 37.6% of the recruitment dates while four *Pinus* species combined for only 6.4%. Recruitment dates are estimates of tree age at stump or coring height. Because these data varied in precision, methodology, recruitment dates, and dates when regeneration reached stump or coring height, dates were placed into four categories: Category 1, from randomized or representative sampling; Category 2, from studies targeting the oldest trees or historical timbers; Category 3, from studies that do not include post-1700 recruitment dates; and Category 4, from studies that have recruitment dates binned at >10 years. We compiled dates by decade because of associated uncertainties in methodology (see *Subcontinental-scale recruitment data* in Appendix A). Raw Category 1 recruitment increases through time (see Appendix A: Fig. A7). This trend was removed using segmented regression (R package segmented; Muggeo 2008) allowing us to detect individual recruitment events from the residuals of this trend as well as an objective assessment of breakpoints or changes in the trajectory of recruitment over time.

Multiple sources of observed weather events were used to conduct ground-truthing of the disturbance events embedded in our tree ring network. Due to the close proximity of the region experiencing the 1770s disturbance event, we relied upon the Moravian records from western North Carolina more than other sources. One value of the Moravian observations is that multiple resident diarists for each year are distributed over an area of at least 400 km^2 ; visitors to the Moravian settlement would occasionally extend the scale of observations in the diaries. In addition, Moravian records occasionally revealed the intensity and scale of impact. The wind event of 17 March 1776, as one example, was recorded by three observers, but it was noted by one of those observers from a neighboring village to have caused little damage (Fries 1926).

The Moravian records were used in two ways. First, they were used to determine the potential cause of the 1774 “white ring” seen in increment cores across genera (*Liriodendron*, *Carya*, and *Quercus*) collected in Kentucky and Virginia (Fig. A8). White rings are rings with low lignification and have been produced following

FIG. 2. (a) Frequency of canopy disturbance through time of populations loading >0.224 onto eigenvector 2. The orange filling represents the percentage of major canopy disturbance per year per tree. The blue line with triangles represents tree replication per year. The short dashed line represents one standard deviation (SD) above the 1685–1880 mean, while the long dashed line represents 2 SD above the mean. (b) Map of the spatial extent of the regional-scale 1775–1780 canopy disturbance event. Mapping of the four quartiles of the total accumulation of trees recording disturbance from 1775–1780 indicates severe damage over an $\sim 61\,000\text{ km}^2$ area. Specific site names corresponding to the numbers on this figure are listed in Appendix A: Table A5. Panels (c) and (d) compare the observed percentage of disturbed trees with a Gaussian distribution (least-mean-square fitted to the observations) for 1685–1880. (c) Density (measured as a proportion) plot of the observed percentage of disturbed trees (orange) and theoretical Gaussian distribution (gray). The labels q95, q99, and q99.9 mark the 95%, 99%, and 99.9% quantiles (blue lines), respectively; red line denotes maximum percentage of disturbed trees (6.4%); σ represents the standard deviation. (d) Quantile–quantile plot comparing the observed percentage of disturbed trees with the corresponding Gaussian distribution from panel (a). Black solid line represents the identity line. For convenient reference, the seven years with the highest rate of disturbance are marked to emphasize the uniqueness of the 1775–1778 era.

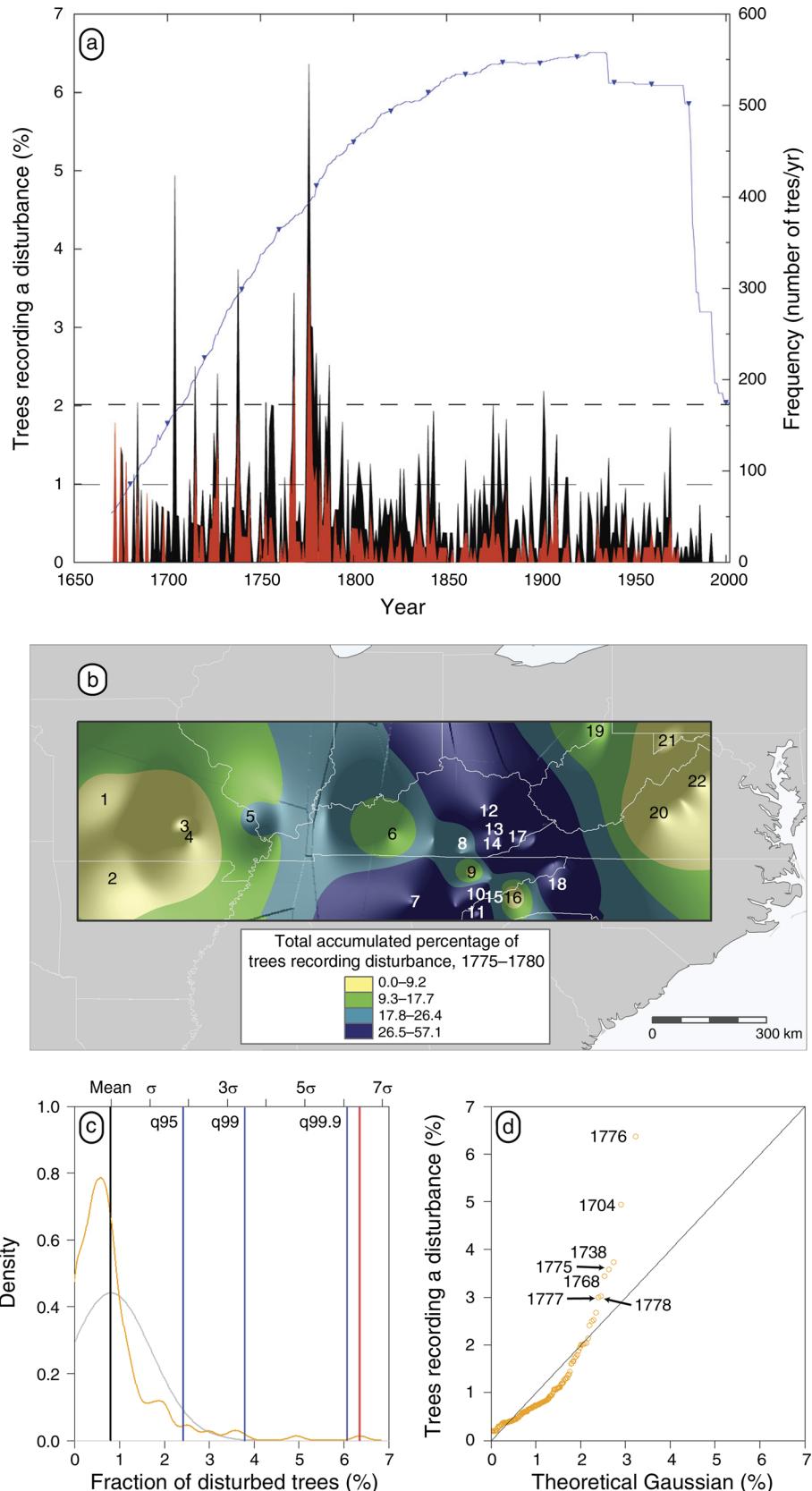


TABLE 1. Characteristics of periods with extended canopy disturbance events from 1685–1880.

Elevated disturbance years	Duration (yr)	Anomaly of disturbance rate (departure from mean/yr)	Proportion major release (%)	Magnitude of disturbance index	Sites recording disturbance
1725–1727	3	0.98	61.1	1.79	14.0
1737–1739	3	1.15	67.2	2.33	14.3
1755–1758	6	0.79	22.7	0.72	16.7
1766–1768	3	1.26	69.6	2.63	22.2
1774–1782	9	1.92	52.6	9.07	33.9
1775–1780	6	2.61	56.8	8.92	41.3
1776	1	5.33	62.5	3.33	47.6
1784–1787	4	0.84	57.1	1.92	26.2
1790–1794	5	0.16	33.3	0.27	18.1
1799–1805	7	0.05	44.4	0.16	16.3
1816–1821	6	-0.10	25.0	-0.15	15.1
1834–1836	3	0.28	37.5	0.31	17.5
1840–1843	4	0.53	40.7	0.86	22.6
1870–1872	3	0.20	38.5	0.02	17.5
1877–1883	4	0.29	24.4	0.50	19.0
1685–1880		0.78	39.6		11.3
Annual mean†		(0.83)			(9.8)
Elevated disturbance	4.6	1.92	44.2	1.57	19.4
Mean	(1.89)	(0.59)	(16.1)	(2.43)	(5.58)

Notes: Characteristics for the peak year, 1776, and peak event within the 1774–1782 event, 1775–1780, are included to display their exceptionality. The event beginning in 1877 continues until 1883. Thus, its characteristics for the full event are included for completeness. See *Methods* for further discussion of this calculation.

† Values are means with SD in parentheses.

defoliation experiments (Hogg et al. 2002). They are also present during years of gypsy moth defoliation (Pederson 2005). Because the 1774 white ring appeared across genera, we hypothesized that defoliation was caused by a frost event. Second, the Moravian records were used as an independent ground-truthing of hurricanes striking the eastern United States in the year prior to or during elevated canopy disturbance events. These strikes were complied from Ludlum (1963), Rappaport and Ruffman (1999), Landsea et al. (2004), and Chenoweth (2006) and are in Appendix A (Table A4). The Moravian records then became vital in determining if known hurricanes impacted our study region because they lived adjacent to the southeast border of our canopy disturbance region and in an area that would likely experience tropical storm wind.

RESULTS

Disturbance history

Twenty-two collections loaded significantly onto EV2. These collections are composed of six species including, mesic, ravine-bottom *Tsuga canadensis* and *Liriodendron tulipifera*, a high-elevation *Picea rubens* collection, and two *Quercus* species growing on a dry, southeast-facing slope (Appendix A: Table A2). The strongest loadings clustered in the southern Appalachian Mountain-Cumberland Plateau region (Fig. 1a; Table A2; a detailed discussion of the RPCA results are in Appendix A: *Rotated principal component analysis results*). EV2's time-series of radial increment (RPCA scores) reveals below average increment prior to 1780 followed by a 201% increase in average radial increment in 1780–1794

vs. 1765–1779 and a linear decline until 1853 that resembles trees following canopy accession (Fig. 1b).

We detected a total of 866 canopy disturbances over the 1570–2000 CE period from the 558 trees and the 916 time-series of radial increment that comprise the 21 populations loading significantly onto EV2 (Fig. 2a; see Fig. A9 for the raw data; please note that the two Blanton Forest collections were combined prior to this step). The peak period of disturbance was from 1775 to 1780, when 81 disturbances were detected, while the peak year was 1776, when 24 disturbances were detected. We detected 588 disturbances between 1665 and 1880. Despite high tree replication throughout this period, 30 years had no evidence for disturbance and 49 years indicated only one disturbance. Twenty-three years had disturbance rates >1 SD above the long-term (1685–1880) mean, hence called “elevated disturbance” (mean = 0.79 disturbance/yr, SD = 0.83; Fig. 2a). From these data we identified 13 “extended disturbance” events, or three consecutive years with ≥ 3 disturbances/year (>2.5 SD above the mean; Table 1).

Of the 588 canopy disturbances from 1685–1880, 60.4% are classified as minor canopy disturbance (Fig. 2a). For the 23 years with elevated disturbance, the minor:major canopy disturbance ratio is closer to 1:1 (51.0% vs. 49.0%, respectively). This ratio swings in favor of major disturbances for years with disturbance rates ≥ 2 SD (48.2% vs. 51.8%). We also found a positive association between severity and disturbance extent. That is, when disturbance was widespread, it was also more severe. Seven of the 13 extended events had a greater proportion of major canopy disturbance than the 1685–1880 mean (Table 1) and individual years and

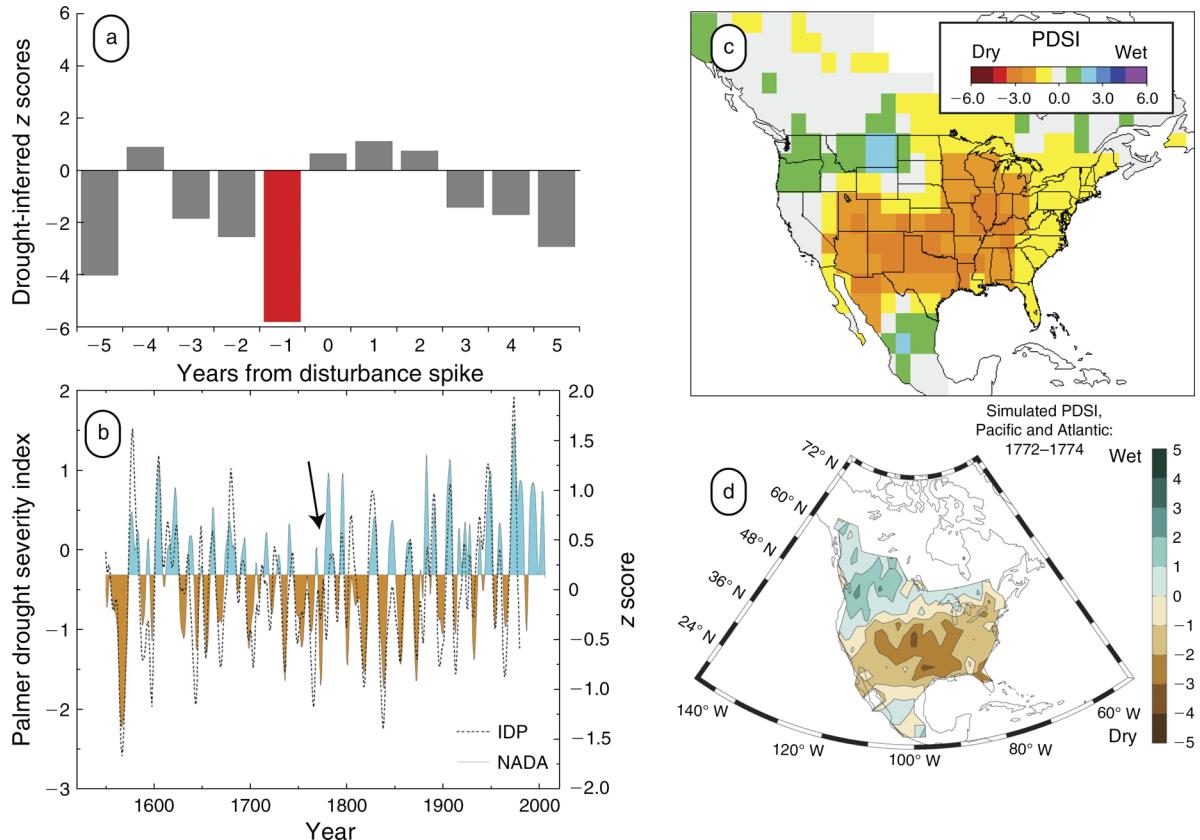


FIG. 3. Relation between climate and disturbance. Superposed epoch analysis of years with disturbance 1 SD or greater above the mean vs. (a) independent drought proxy (IDP). The red bar represents the only year where disturbance is significantly correlated to drought as represented by the IDP ($P < 0.05$). (b) Smoothed reconstruction of drought from 1550–2006 (9-yr spline). The filled curve is derived from the North American Drought Atlas (NADA; Cook et al. 2010) while the dashed line is the IDP. The arrow marks the end of the 1772–1775 drought. (c) NADA composite map of 1772–1775 drought just prior to the 1776 disturbance spike. (d) A 2000 member ensemble simulation for 1772–1775 drought as forced by Pacific and Atlantic sea surface temperatures.

periods with elevated disturbance also appear to be widespread. For example, the annual mean percentage of sites with disturbance from 1685–1880 is 11.3% ($SD = 9.8$), but during periods of elevated disturbance, the mean of sites recording disturbance was 19.4 (range = 14.0–33.9, $SD = 5.58$; Table 1). Disturbance was recorded in 33.9% of all sites from 1774–1782, >3 SD above the mean, and in 41.3% of all sites from 1775–1780 (Table 1). The spatial extent of disturbance peaked in 1776 when it was found in 47.6% of all sites. The total accumulated percentage of trees recording disturbance per site during the 1775–1780 event ranged from 0 to 57.1% (average = 19.7%, $SD = 17.7\%$). Mapped quartiles of the total amount of disturbance recorded from 1775–1780 indicate severe damage over ca 61 000 km² (Fig. 2b). The four collections with less damage during this event include two *Liriodendron tulipifera* and two *Quercus alba* populations. The magnitude index for extended, elevated disturbance events for the 1770s events was 3.4–3.9 times greater than the next two most severe events (1737–1739, 1766–1768) (Table 1).

The time-series of canopy disturbance from 1685–1880 is not well described by a Gaussian distribution. In fact, the distribution of canopy disturbances is heavy tailed (Fig. 2c, d). Notably, all years from 1775–1778 are included in the list of the top seven most-disturbed years (Fig. 2d) and show a disturbance rate $\geq 3\%$, which is \geq the 98th percentile of our data set. These years are well out in the heavy tail of the disturbance rate distribution (Fig. A10). Return intervals for disturbance rates of 2%, 4%, and 6% of disturbed trees per year ranged from 40, 250, and ~ 930 years, respectively. Uncertainty around return intervals greatly flares out beyond the 4% rate because of small sample size (Fig. A11).

We found that elevated canopy disturbance is significantly correlated with a low index of inferred drought during the prior year (Fig. 3a). Drier conditions prevail during four of the five years prior to elevated disturbance. The southern Appalachian Mountain region experienced three intense droughts between 1742 and 1775 with the 1772–1775 drought, spatially expressed across much of the temperate United States, as the most intense (Fig. 3b, c). Composite maps of

TABLE 2. Comparison of tree recruitment categories from 1500–1849.

Era	Category 4											
	Category 1, <i>n</i> = 2277		Category 2, <i>n</i> = 1236†		Category 3, <i>n</i> = 1105		20-yr bins, <i>n</i> = 270		25-yr bins, <i>n</i> = 209		50-yr bins, <i>n</i> = 230‡	
	No. trees	%	No. trees	%	No. trees	%	No. trees	%	No. trees	%	No. trees	%
1450–1499	1	0.04	0	0	0	0	0	0	0	0	0	0
1500–1549	3	0.1	5	0.4	3	0.3	0	0	0	0	1	0.4
1550–1599	20	0.9	27	2.2	4	0.4	3	1.1	0	0	9	3.9
1600–1649	86	3.8	95	7.7	105	9.5	6	2.2	5	2.4	13	5.6
1650–1699	330	14.5	301	24.4	814	73.7	15	5.55	11	5.3	23	10.0
1700–1749	425	18.7	318	25.7			26	9.63	22	10.5	60	26.1
1750–1799	626	27.5	272	22.0			78	28.89	78	37.3	51	22.2
1800–1849	786	34.5	218	17.6			142	52.59	93	44.5	73	31.7
Other eras for comparison												
Pre-1650	110	4.8	127	10.3	112	10.1	13	4.81	5	2.4	23	10.0
1670–1689	168	7.4	127	10.3	413	37.4	4	1.48	6	2.4¶		

Notes: See *Methods* for category details. Total number of trees is shown by *n*. Empty cells indicate that there is no data.

† Dates decline after ca. 1735 because of sampling methods.

‡ Tyrrell and Crowe (1994) age categories crossed some of the time periods used in this table. Thus, there is some uncertainty in total numbers per 50-year period.

§ Gates and Nichols (1930) first recruitment date could have been 1500. To be conservative, we are using it in pre-1500 and 1500–1549 class.

¶ Represents 1675–1699.

North American drought for the year prior to elevated disturbance, the 1772–1775 drought, and a statistical model of the 1772–1775 drought reveals pan-continental drying over most of the United States and wetter than average conditions in the Pacific Northwest and northern Great Plains (Figs. 3d, A12). These findings indicate that regional-scale drought is associated with elevated disturbance across the 1775–1780 disturbance region.

Tree recruitment across the eastern deciduous forest

Recruitment data from studies using plot level or representative sampling (Category 1) have 2277 dates from trees that recruited between 1460 and 1850. Of these trees, 14.5% (*n* = 330 trees) recruited between 1650–1699, with more than one-half of these trees recruiting from 1670–1689 (*n* = 168). In sharp contrast, only 86 trees (3.8%) recruited from 1600–1649, or less than one-half of those recruiting between 1670 and 1689. Of the 620 recruitment dates collected through representative or plot-level sampling in old-growth forests before 1944 (e.g., Gates and Nichols 1930, Williams 1936, Hough and Forbes 1943), 21.1% recruited from 1650 to 1699 vs. 8.5% from 1600 to 1649 and 10.8% from 1500 to 1649 (Table 2). Segmented linear regression on Category 1 data indicates a significant break in recruitment around 1599 CE (± 10.6 yr). Residuals from this regression indicate a large and prolonged period of recruitment from 1640 to 1699, followed by below average recruitment from 1700 to 1729, and decadal-scale fluctuation through 1849 (Fig. 4a). The three greatest positive departures in recruitment occur in the 1670s (a residual departure of +1.36), 1680s (+0.79), and 1780s (+0.79). Data from targeted sampling

(Category 2) or collections made during the early-1900s that do not have data after 1699 (Category 3), reveal similar jumps in recruitment during the latter half of the 17th century (Table 2). Within Category 2, the peak in recruitment of the 461 historical timber dates is centered on 1660–1699 (35.4% of total sample) with 15.6% recruited from 1670 to 1689 vs. 18.4% during the preceding 119 years. Despite deliberate attempts by 15 different tree-ring scientists over the last 30 years to core the oldest living trees in various forests, only 5.1% of the 730 trees recruited between 1500 and 1649. In comparison, 7.5% of the 730 trees recruited between 1670–1689.

DISCUSSION

Our records of forest dynamics—two large, species-rich, and geographically extensive data sets—indicate that (1) the dynamics of broadleaf forests in a temperate, humid region occur synchronously across different scales, from the stand to subcontinental level and (2) extended events of canopy disturbance are often severe. These findings more closely resemble our alternative hypothesis where forest dynamics can be regionally synchronous and punctuated by extreme events. Discovering that larger canopy gaps are often formed during synchronous large-scale events in broadleaf-dominated forests is broadly relevant because it provides a mechanism for rapid, large-scale change. That is, a greater number of larger canopy openings in light-limited forests offer increased opportunities for a compositional shift in the canopy at the time of major disturbance. Our large-scale analysis also reveals a greater spatial extent of previously reported disturbances. The large and severe 1775–1780 event is a period of increased disturbance observed at the stand scale in

TABLE 2. Extended.

Category 4, n = 709		Early category 1 studies, 1930–1943, n = 620	
No. trees	%	No. trees	%
0	0	1§	0.2
1	0.1	1	0.2
12	1.7	13	2.1
24	3.4	53	8.5
49	6.9	131	21.1
108	15.2	162	26.1
207	29.1	138	22.2
308	43.4	113	18.2
41	5.8	67	10.8

western North Carolina (Lorimer 1980: Fig. 7) and landscape scale in central Tennessee (Hart et al. 2012: Figs. 7 and 8) and central Pennsylvania (Nowacki and Abrams 1997: Table 3). This large-scale event precedes the highest peak in tree recruitment in our data set from 1700 to 1849 (Fig. 4a). Further, two other periods of extended disturbance discovered in our study, 1737–1739 and 1755–1758 (Table 1), are evident in a landscape-scale study (Hart et al. 2012). While the resulting time-series of disturbance still resembles a white-noise process, we find synchronous disturbance at multi-annual to nearly decadal time scales. Our findings go beyond the limitations of local studies and reveal forest dynamics at both the landscape and mesoscale (from decades to centuries over 100 to 100 000 km²). As such, they are relevant for anthropogenic climate change and have important implications for forest management.

We have also statistically identified a plausible trigger for these disturbance events: drought-induced canopy mortality. These findings support observations of drought-induced forest dynamics and sensitivities conducted at short time scales and local to regional scales (Hough and Forbes 1943, Clinton et al. 1993, Jenkins and Pallardy 1995, Olano and Palmer 2003) across different forest types including humid regions (Allen et al. 2010, Choat et al. 2012). In doing so, we bridge the spatial and temporal gaps between local and sediment studies by providing insights from paleoecological records while revealing broad-scale patterns not seen in stand-scale or landscape-level studies.

Disturbance in humid to wet temperate regions can be characterized by frequent, small, low-severity disturbance events with occasional large-scale, intense disturbance (Lorimer 1989, White et al. 1999). Most studies in extant forests in humid regions have not revealed regional-scale events perhaps because of a predominant focus on local to landscape scales. Experimental forest modeling suggests small-scale analysis reduces the ability to detect large-scale change (Smith and Urban 1988). Our results reveal dynamical processes at small

and large scales over the last 400 years (cf., Jackson 2006) and demonstrate the legacy of large-scale, intense, disturbance events centuries after their occurrence in broadleaf-dominated forests, a finding similar to previous work (e.g., Lorimer 1980, Frelich and Lorimer 1991, Nowacki and Abrams 1994, Hanson and Lorimer 2007). The distinction here is that we document events at regional to subcontinental spatial scales and show that some of these events can occur repeatedly within the maximum longevity of many canopy species. The infrequent, but meso- to large-scale disturbances are important because they can create the “substrate” that the more frequent, but less-intense, small-scale dynamics act upon. While it is known that historical events resonate for centuries and millennia through the structure and dynamics of forested ecosystems (e.g., Lorimer 1989, Sprugel 1991, Swetnam and Betancourt 1998, Foster et al. 1999, Williams and Jackson 2007, Turner 2010), we have identified large-scale events from 230–360 years ago at high resolution that are still reverberating in the structure of today’s old-growth, broadleaf-dominated forests.

Interestingly, some old trees in today’s old-growth forests in the eastern United States are the result of historical events (Tables 1, 2; Figs. 1b, 2a, b) that occurred during a drier era than the more moist conditions that prevailed during the period of repeated measures and field studies (Fig. 5). Dry conditions have the potential to alter other processes, like increased fire or insect outbreaks (e.g., Raffa et al. 2008, Lynch and Hessl 2010), and feed into forest dynamics in direct and indirect ways. The rare, but coherent, spatially broad, and severe events identified here can provide greater opportunities for regeneration as canopy gap formation increases. Alternatively, these species-rich forests, interacting with historical contingencies and a wide range of possible future scenarios, could abruptly change into substantially different types than the current forest (Williams and Jackson 2007).

It is important to note that these results also suggest that local dynamics and other endogenous factors are at play. Four populations within the 1775–1780 event region have low amounts of canopy disturbance (Fig. 2b). Two populations are *Liriodendron tulipifera* whose requirement for relatively large gaps for successful recruitment might have made them less sensitive to changes in competition if they had reached canopy status prior to the 1770s. Although we have evidence for large-scale, synchronous disturbance, not all populations were similarly affected, which could be due to the many factors influencing forest dynamics.

Triggers for elevated canopy disturbance

Triggers of elevated canopy disturbance at large scales across a humid and diverse region are likely complex. Our data support early and more recent research linking tree mortality to drought across forest types and land-use histories (e.g., Lorimer 1984, Olano and Palmer

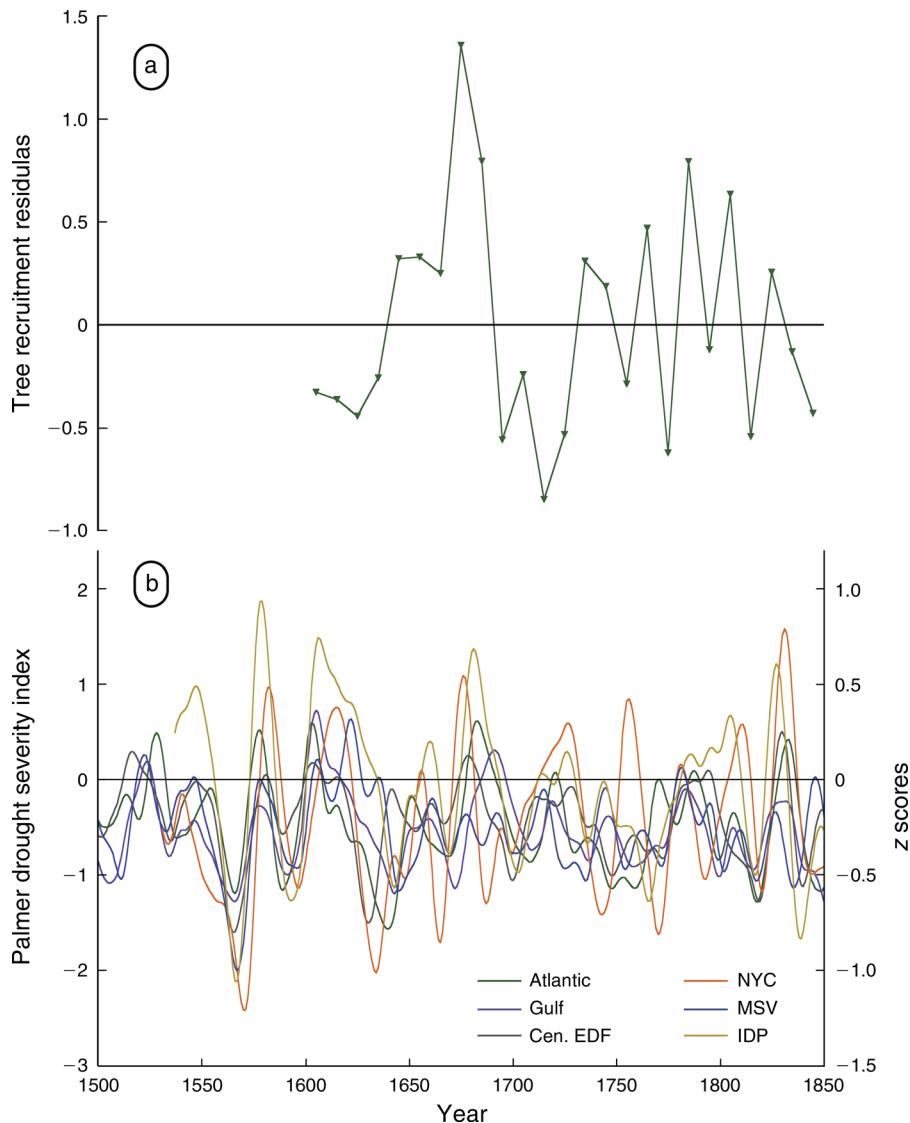


FIG. 4. Forest recruitment and climate across the broadleaf-dominated forests of the eastern United States. (a) Residual of tree initiation dates from 1500–1850 after removing demographic trends (n dates = 2276). (b) Drought proxies for Atlantic Coast (Atlantic; Georgia to Virginia), Gulf Coast (Gulf; Georgia to Louisiana), center of the Central Eastern Deciduous Forest region (Cen. EDF), northeastern United States (NYC; New Jersey to central Massachusetts and western New York State), Mississippi Valley (MSV; Louisiana to Illinois), and the independent drought proxy (IDP; see Appendix A for more information on IDP). The Atlantic, Gulf, Cen. EDF, and MSV records are from the NADA. NYC is from Pederson et al. (2013). Despite the inclusion of trees used in our disturbance analysis, the Cen. EDF record is included to show that drought variation in that record is not dramatically different from records from surrounding areas.

2003, Mueller et al. 2005, Pederson et al. 2008, 2012b, Anderegg et al. 2012). The association between disturbance extent and disturbance intensity indicating the mortality of canopy trees dovetails with observations that tall and large canopy trees are more susceptible to drought-induced mortality (e.g., Hersh and Haasis 1931, Hough and Forbes 1943, Floyd et al. 2009, Hartmann 2011). Trees in closed-canopy forests primarily compete for canopy access and solar radiation (Hartmann 2011). Competition for solar radiation likely pushes tree height near the maximum height possible within the context of

microsite and other prevailing environmental conditions. Because tree height limits leaf-specific hydraulic conductance (McDowell et al. 2008), it is likely that canopy trees in closed-canopy forests live closer to the margin of water balance for survival and are more susceptible to drought-induced mortality. This has been directly observed within our larger study region: the “extreme drought of 1930” in Pennsylvania led to “mortality of the larger or overstory trees both of hemlock and of the subsequently exposed beech” (Hough and Forbes 1943: 311). Further, repeated

drought increases the mortality risk of trees (Pedersen 1998, McDowell et al. 2008). Therefore, it is not too surprising that the 1775–1780 disturbance event is preceded by three intense droughts during the previous three decades (Fig. 3b). Our data suggests that drought plays an important role in canopy dynamics of broadleaf-dominated forests in the eastern United States.

While the 1772–1775 drought was severe, it was not the most severe (Fig. 3b). Therefore, a variety of additional triggers likely interacted with drought to generate the broad-scale patterns of disturbance and canopy dynamics. For instance, colonial-era documents reveal an early onset of leaf-out in 1774 and the commencement of farming two weeks early across the southeastern United States (e.g., Thomas Jefferson's Garden Book 1774, Fries 1925). Heavy frosts after the cold nights of 4 and 5 May 1774 made green leaves look “black and dead” (Fries 1925). Additional accounts confirm cold air, frost, or damage to plant tissue throughout the southeastern United States up to southeastern Pennsylvania (Pennsylvania Gazette [Philadelphia, Pennsylvania] 11 May 1774, page 2; Virginia Gazette [Williamsburg, Virginia] 12 May 1774, page 4; Essex Gazette [Salem, Massachusetts] 7 June 1774, page 176). A frost was observed to have killed “every tender thing” near coastal South Carolina (Rudisill 1993). These observations confirm tree-ring evidence of a frost event in the southern Appalachian Mountain region (Fig. A8) and extend it throughout the southeastern United States.

Following repeated drought during the mid 1700s, the 1774 frost event and subsequent defoliation must have reduced available energy from surviving, but drought-stressed, trees. Widespread frost events in April 2007 (Gu et al. 2008, Augspurger 2009) and May 2010 (Hufkens et al. 2012) lead to tissue and shoot dieback and necrosis in ways similar to 1774 (see Fig. A13a). Refoliation in 2007 did not occur for more than a month in some cases (Augspurger 2009), especially for *Liriodendron tulipifera* (Gu et al. 2008; see Appendix A: Fig. A13b), and was estimated to be 46–99% of normal for seven species, with 40% to ~90% coming from a second flush (Gu et al. 2008, Augspurger 2009). In 2010, net carbon assimilation of *Acer sacharrum* was reduced following leaf necrosis and delayed canopy development (Hufkens et al. 2012). If these frosts were similar in intensity to the 1774 frost, preceding climatic conditions prior to 1774 likely predisposed canopy trees to higher rates of mortality (*sensu* Manion 2003). The mid-18th century was one of the driest periods in the southeastern United States of the last 300 years (Cook et al. 1988, Pederson et al. 2012a). Trees adapt to aridity by (1) shedding leaves, (2) reducing the root/sapwood to leaf area ratio, and (3) experiencing a reduction in height through crown dieback (McDowell et al. 2008). Drought-stressed trees recovering from the severe 1774 frost could struggle with alteration of the carbon sink or

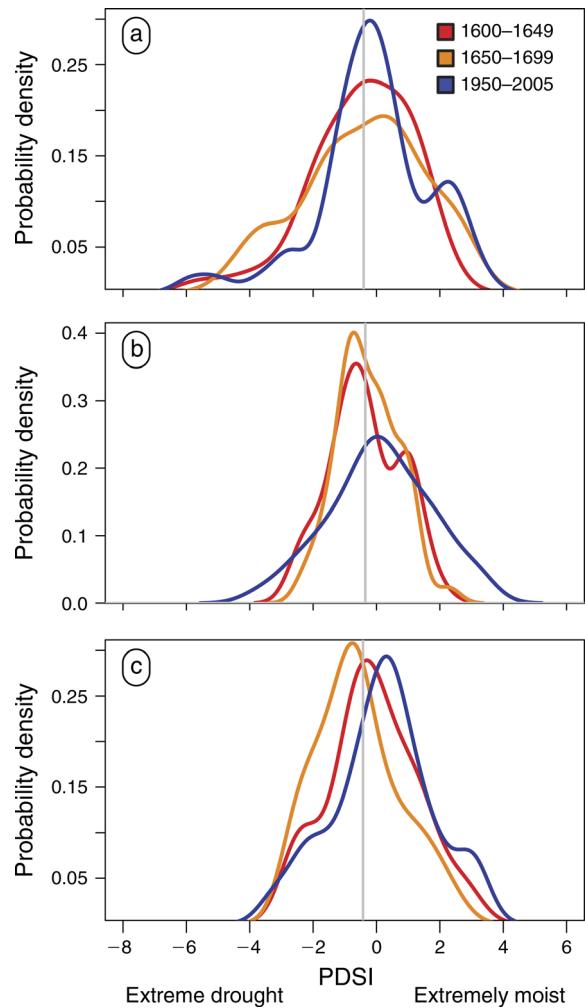


FIG. 5. Probability density functions of reconstructed drought (PDSI) for 1600–1649 (red), 1650–1699 (orange), and 1950–2005 (blue) for three broad regions: (a) northeastern United States, (b) central hardwood region of Fig. 1a, and (c) central Mississippi River Valley. Panels a, b, and c are adapted from the original data used to construct Fig. 4. These plots show that, during the 17th century, more years tended to be dry while, during the period of observation (repeated measures and many ecological field studies), tends to be substantially wetter and close to extremely moist conditions, conditions that are nearly absent during the 17th century.

carbon metabolism (cf., Adams et al. 2013). Lesser, 20th-century droughts are known to have led to tree mortality (Jenkins and Pallardy 1995, Pedersen 1998), suggesting the more severe mid-18th century droughts could have resulted in greater mortality (Fig. 3b). The rate of disturbance in our data during the late 1770s was up to 7.8 times greater than the mean (Fig. 2), a rate even greater than that observed in a tropical ever-wet forest during the strong, El Niño drought of 1998 (Potts 2003). Based upon these observations, we postulate that the 1770s period of elevated disturbance was likely triggered by several interacting factors, including drought and frost.

Following decades of dry conditions, however, fire is another potential contributor to elevated disturbance of the late 1770s. Fire occurs more often and burns a larger area in a humid region during drier conditions (Lynch and Hessl 2010) and in areas with a great range of daily precipitation variability (Lafon and Quiring 2012). Fire was generally rare, spatially restricted, or absent from 1775–1780 in or near the event region (Guyette et al. 1994, Aldrich et al. 2010, Feathers 2010, Flatley et al. 2013, McEwan et al. 2014). As an example, Flatley et al. (2013) reports area-wide fires in 1773 and 1775 in only one of three forests they investigated. These findings suggest a lack of broad-scale fire during the 1770s. Fire is not recorded in the mid 1700s in broadleaf-dominated forests, although it is recorded before and after this period (Guyette et al. 1994, McEwan et al. 2014). Guyette et al. (2006) show an increase in fire during the 1770s, but also note that “Fire frequency was highly variable in both time and space even at regional scales...” (Guyette et al. 2006: 20). The broadest fire in this region occurred in 1780 (Guyette et al. 2002), which is after the peak of the 1770s disturbance event. Due to the low incidence of spatially extensive fire across our study region during the 1770s, a pattern supported by paleo studies (Clark and Royall 1996, Parshall and Foster 2002), and the ability of drought to kill overstory trees, it would seem repeated, intense drought leading up to the mid 1770s was a primary contributor to elevated canopy mortality with frost-induced defoliation and potentially fire as secondary contributors.

Windstorms are another important canopy disturbance that have been well documented across our study region (Lorimer 1977, White 1979, Foster and Boose 1992, Everham and Brokaw 1996, Peterson 2000, 2007, Stueve et al. 2011). Windstorms can leave a lasting impact in forests beyond the stand scale. A squall line increased mortality over roughly 27 km^2 in the Manaus region and an estimated $4.5 \times 10^6 \text{ km}^2$ across the Amazon (Negrón-Juárez et al. 2010), while a conservative estimate of windstorms in the midwestern United States was extrapolated to damage forests over $\sim 1500 \text{ km}^2$ over a 26-year period (Stueve et al. 2011). In contrast with line storms or tornadoes, which are local in scale, hurricanes have the potential to affect large regions. From a compilation of hurricanes striking the eastern United States during each elevated disturbance event in our study (Table A4), we find that five elevated disturbance events coincide with relatively high landfall years (1766–1768, 1834–1836, 1840–1843, 1870–1872, 1877–1883); “relatively high” here is two or more hurricanes per year. However, the simple occurrence of a hurricane making landfall might not translate into an ecological impact because of insufficient intensity or a failure to penetrate inland. Focusing on the most important of these reconstructed events, 1775–1780, the Moravian records hold no evidence of forest damage from wind during the hurricane season. There are three hints that low-pressure systems might have moved

through western North Carolina during this period, however. First, it was noted on 30 August 1775 that, “For some time it has been raining every day” and that by 4 September there has been large-scale clearing of the atmosphere, “It is clear, but the air feels like fall” (Fries 1925: 883). The track of this storm appears to have traveled far to the east of the 1775–1780 event region (Rappaport and Ruffman 1999). Then, in 1778, a year with hurricanes in August and October (Table A4), there are observations of “a hard storm from the northeast” on 11 August (Fries 1926: 1244) and that a “strong wind from the north-east cleared the sky” on 11 October (Fries 1926: 1248). No forest damage is reported with these observations. Observations of winds from the northeast, but with less than tropical-storm strength, matches model predictions of tropical storm decay into this region (Kaplan and DeMaria 1995). And, observations of winds from the northeast causing no damage are in contrast to the windstorms of note between 1774 and 1779. The windstorm of March 1775 caused damage “over a strip about 14 miles long and four wide” (Fries 1925: 873). Similarly, the March 1776 windstorm was reported to be more local: “Br. Praezal returned from Bethabara; the storm was not nearly so severe there, and had done no particular damage” (Fries 1926: 1057). It is beyond the scope of this investigation to quantify the paths of all windstorms and their impacts within our network. While geographically limited, the Moravian records suggest only one wind event at the landscape scale and no late-season windstorms of the years we reviewed (the Moravians diaries contain no evidence of hurricane-like storms during the 1766–1768 event). Nevertheless, windstorms are a potential factor affecting historical canopy disturbance.

Ice storms are another common canopy disturbance in this region that can cause limb breakage, snapped stems, and treefall (Lemon 1961, Irland 2000, Proulx and Greene 2001, Wonkka et al. 2013) over large areas (Millward and Kraft 2004, Vanderwel et al. 2013). While the impact of ice storms can be severe, they also trigger a mixed response, ranging from tree death to a positive growth response in surviving trees (Lafon and Speer 2002). Within the Southern Appalachian region, up to 40% of the trees in a stand were observed to have been killed (Lafon 2006). However, there is no mention of a severe ice storm in the Moravian records preceding or during the 1775–1780 event (Fries 1925, 1926). Ice storms could have been a trigger in our reconstruction of disturbance, but we are lacking evidence of it being an agent for the most severe and widespread event in our records.

Tree recruitment across the eastern deciduous forest

Our recruitment data reveals a subcontinental-scale event in a forest type where this scale of event would be less expected. This finding synthesizes nearly a century of reported stand origin dates scattered across the literature. Given the temporal distribution of these

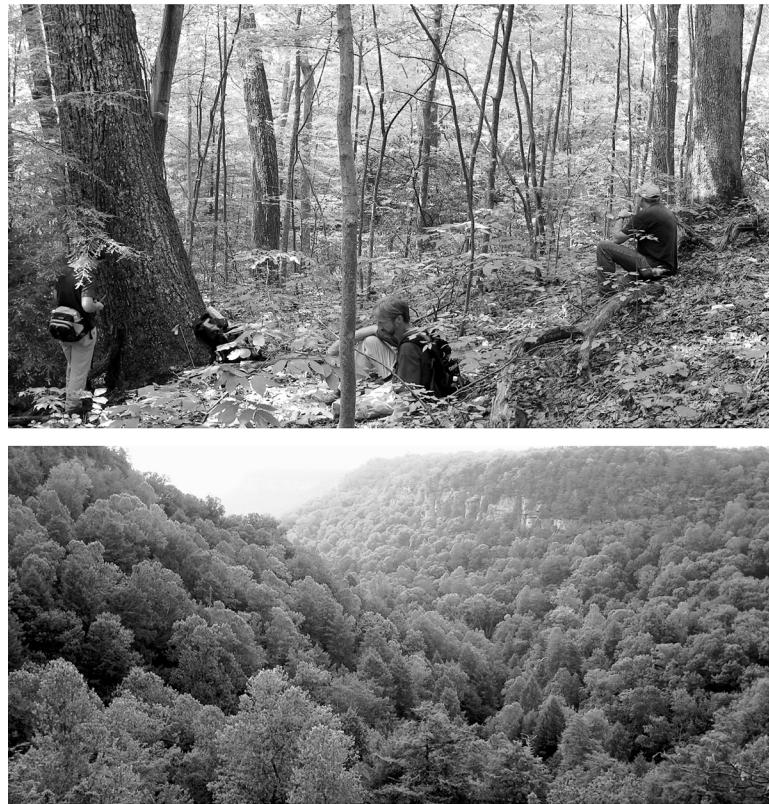


PLATE 1. (Upper) The *Quercus montana* being cored in the Blanton Forest of Kentucky (USA) reflect the subcontinental- and regional-scale dynamics discovered through our investigation. Most of the cored *Quercus* on a dry, southeast-facing slope were recruited to coring height during the late 1600s; none before 1670. We observed this recruitment event in data from across much of the eastern U.S. Many of the cored *Quercus* trees in the Blanton Forest were simultaneously released from competition in the synchronous, large-scale event of the late 1770s in the Southern Appalachian Region. (Lower) Overlooking the old-growth forest at the bottom of Savage Gulf, Tennessee at dusk. Many of the *Tsuga canadensis* cored in this ravine experienced a significant and synchronous reduction in competition at the same time as the *Quercus* trees in the Blanton Forest of Kentucky. The forests on the plateau above this ravine at Savage Gulf also experienced increased canopy disturbance (Hart et al. 2012), which independently helps to confirm the regional-scale disturbance in the late 1770s. These photos and others are available in color in Appendix B. Photo credits: N. Pederson.

studies over the last 90 years and that the forests from which these data are collected can be characterized by continuous, small-scale dynamics, the date of stand origin would be expected to shift back in time according to the time of sample collection. We found, in fact, that a disproportionate amount of recruitment occurs during the mid to late 1600s, even in the oldest studies with recruitment dates. These findings resemble recent evidence of a regional-scale synchrony of understory forest dynamics in a broadleaf-dominated forest (Gravel et al. 2010). Our findings suggest that synchronized tree recruitment at large scales can be an important process in broadleaf-dominated ecosystems.

While it is possible that the late 1600s recruitment pulse is an artifact of tree longevity, four lines of evidence argue against this idea. First, 21 tree species in Category 1 recruited between 1650 and 1699, including shade-intolerant, fast-growing species like *Betula lenta*, *Castanea dentata*, *Liriodendron tulipifera*, and *Pinus strobus* as well as shade-tolerant, slower-growing species

like *Acer saccharum*, *Fagus grandifolia*, and *Tsuga canadensis* (Burns and Honkala 1990). Given variations in longevity, shade tolerance, the spatial extent of our recruitment data set, and light limitations in this forest type, a recruitment pulse seems to require an exogenous factor opening the canopy, not an intrinsic factor such as longevity. Second, conventional wisdom on maximum tree age has proven to be underestimated for many species, sometimes by a century or two (Pederson 2010). Thus, the longevity of 255-year-old trees described in 1923 as comprising most of the recruitment between 1660 and 1674 (Haasis 1923) should not be a factor as most of the species in that study can live longer than 250 years. Third, the late 1600s recruitment pulse is evident in data collected >75 years ago from old-growth forests (Huntington 1914, Haasis 1923, Gates and Nichols 1930, Williams 1936, Hough and Forbes 1943). These early studies identify a slightly higher percentage of trees before 1650 vs. our database (Table 2). But, all authors, save Williams (1936), note a recruitment pulse during

the mid to late 1600s when discussing their results. Fourth, dates from tree-ring investigations targeting old trees during the 20th century or historical timbers cut in the late 1700s to mid 1800s (Category 2) show a similar recruitment pulse (Fig. A7). In fact, there is nearly a tripling of recruitment from 1650–1699 vs. pre-1650 in Category 2 and the number of trees from 1650 to 1699 in the tree-ring data set is five times greater than the prior 50 years (Fig. A7). Although there is uncertainty in these data due to different methods, source material, and potential selection biases between the different sources of data, the striking consensus from these data is that the late 1600s was an important era of subcontinental scale tree recruitment in broadleaf forests of temperate eastern North America.

Potential triggers of the 17th-century recruitment pulse

Though recruitment is often associated with wet conditions, drought could lead to tree recruitment (Shuman et al. 2009). In this scenario, formation of canopy gaps from drought-induced mortality would increase the amount of solar radiation penetrating the understory to stimulate potential recruitment. A return to mesic conditions following drought could aid recruitment. Hydroclimatic records with decadal-scale resolution from Quebec to Ohio and Lake Michigan indicate drier conditions during the early to mid 17th century (Bégin and Payette 1988, Wolin 1996, Lichter 1997, Loope and Arbogast 2000, Greenlee 2006, Argyilan et al. 2010, Hubeny et al. 2011). The dating uncertainties of these records are a limitation, but a record in this region reports a “high probability” of trees growing below current water levels “as early as AD 1663” (Shuman et al. 2009: 2796). These geologic records suggest regional drought from ~1640–1680 that could have caused wide-scale increased canopy tree mortality.

Despite some asynchrony, tree-ring records across the eastern United States indicate drier conditions during the mid 1600s. Each record contains severe drought between 1630 and 1650 (Fig. 4b). A mid-Mississippi River Valley record indicates the mid-17th century to be the driest since 1600 CE (Cook et al. 2010) while a northeastern U.S. record indicates six severe droughts between 1629 and 1700 (Pederson et al. 2013). Of these six droughts, five rank as the 10 most severe droughts since 1531 CE, with 1661–1667 and 1630–1636 ranked second and third, respectively. Other annually resolved proxies support these records over a large region (St. George et al. 2009, Maxwell et al. 2011). Interestingly, all of these tree-ring proxies generally substantiate some geologic evidence for a mid-1600s drought sandwiched between two pluvials, prolonged periods of above-average moisture (Wolin 1996, Lichter 1997, Loope and Arbogast 2000, Argyilan et al. 2010). The switch from pluvial conditions following drought during the 17th century likely has far-reaching ecological consequences. If trees adapt to aridity by root and shoot

dieback (McDowell et al. 2008), then pluvial conditions likely stimulate the opposite. Therefore, an abrupt, severe drought following a pluvial could exacerbate drought stress by making it difficult to maintain pluvial-level biomass, elevating mortality, increasing solar radiation to the forest floor, and possibly increasing opportunities for tree recruitment.

Like many aspects of macroecology, the exact cause of regional-scale disturbance cannot be ascribed to a single trigger (McEwan et al. 2011). Low tree replication prior to the mid 1600s forces us to consider circumstantial evidence. The combination of fire and drought cannot be ruled out. Several studies found charcoal or asserted that drought and fire led to stand origin during the mid-1600s (Huntington 1914, Hough and Forbes 1943, Henry and Swan 1974, Foster 1988). The most consistent fires in northwestern Vermont, for example, occurred in 1586, 1595, 1635, and 1670 (Mann et al. 1994), which coincides with some of the driest periods in the northeastern United States (Pederson et al. 2013). While sample replication is low in the heart of our study region, fire is more often recorded from 1660–1680 than during the mid 1700s (Mann et al. 1994, Guyette and Dey 1995, Dey and Guyette 2000, Guyette et al. 2002, 2006, Guyette and Spetich 2003, McMurry et al. 2007, Stambaugh et al. 2011).

Given that forests in the Western Hemisphere were in a managed landscape prior to European settlement (Crosby 1986, Mann 2005, Krumhardt 2010), it is possible that the catastrophic decline in indigenous populations contributed to the recruitment pulse. However, the locations of recruitment data and resettlement patterns post-population collapse are in opposition to one another. Recruitment data used here are primarily from old-growth forests in mountainous areas (Fig. A6; see raw recruitment data in the Supplement). Most of today’s old-growth forests grow on sites with low productivity and commercial viability (Therrell and Stahle 1998). Most indigenous populations were observed in large and fertile river valleys at the time of European settlement (Cronon 1983, Williams 1992, Davis 2000). In the southern Appalachian Mountain region, people moved into flatter, moister, river valleys following the population collapse (Davis 2000). Supporting this, an estimation of natural vegetation inferred from human demographics indicates forests to be broadly recovered by 1600 CE over much of our study region (Appendix A; Fig. A14; Krumhardt 2010). The uncertainties regarding land-use and human demographics are large at this time (Milner and Chaplin 2010), but they suggest broad forest recovery at least 50 years prior to our reconstructed recruitment pulse.

Climate change implications

Secular trends in anthropogenic climate change are expected to have significant impacts on eastern U.S. forests (Iverson and Prasad 2001, McKenney et al. 2011), although change could occur relatively slowly due

to hysteresis and canopy persistence (e.g., Loehle 2000). However, as observed here, rare or low probability events could have immediate and long-term ramifications for ecosystem structure and function. The severe frost event of 1774 following repeated droughts over large spatial scales altered the trajectory of forest structure through synchronous canopy disturbance. These changes in forest structure generated historical contingencies that can be observed today. Severe droughts in the eastern United States, like the 1960s drought that increased *Acer rubrum* mortality (Lorimer 1984), can occur as a result of internal variability of the climate system (Seager et al. 2012) making stochastic climate dynamics an important aspect of forest dynamics. Further, Pacific Ocean sea surface temperatures seems to have been a trigger for the 1772–1775 drought (Fig. 3d) and, thus, ocean–atmosphere dynamics may have created a legacy in eastern U.S forests where this part of the climate system has less of an influence on the average climatology. These events illustrate how hard-to-predict, singular scenarios can shape forests for decades to centuries in regions where the average scenario might not.

Understanding how the effects of long-term climate trends and extreme events interact is a major challenge for ecological forecasting, not the least because extreme events such as droughts have often been perceived as a minor risk in humid regions such as eastern North America (e.g., McMahon et al. 2010, Schleeweis et al. 2013, Vanderwel et al. 2013). Future droughts could turn out to be analogous to the 16th-century megadrought (Stahle et al. 2000), the most synchronous drought across our study region over the last 450 years (Fig. 4b). If future warming exacerbates drought stress in trees and other stressors, including elevated air pollution (Dietze and Moorcroft 2011), nonnative pests, and pathogens (Aukema et al. 2010), or a combinations of stressors (e.g., Waller 2013), widespread tree mortality and subsequent canopy turnover could drive rapid rates of change in temperate forests across humid eastern North America.

CONCLUSIONS

We have identified two synchronous disturbance events at regional to subcontinental scales in old-growth, broadleaf-dominated forests in a humid region. Not only do these records reveal a severe, regional-scale event from 1775–1780, but also disturbances at other spatial scales in the broadleaved-dominated forest of the eastern United States. Importantly, we find elevated canopy disturbance at moderate frequency beyond the stand scale. Thus, our analysis supports the alternative hypothesis that dynamics in broadleaf-dominated forests can be regionally synchronous and punctuated by extreme events.

Our results can aid future simulations attempting to forecast carbon sinks or rates of ecological change under a changing climate (e.g., Millar et al. 2007) at the scales relevant to the management of forests across a region

with high ecosystem functionality (Freudenberger et al. 2012). Large-scale, synchronous, and climatically influenced disturbances over the last four centuries are still detectable, and thus, important in today's old-growth forests. These events point to the possibility that severe events can push ecosystems to new structures and/or compositions (Jackson 2006, Williams and Jackson 2007, Jackson et al. 2009, Frelich and Reich 2010), even by distal climate systems that do not typically have an influence on a particular region. As local and regional climate and ecologies are influenced by anthropogenic climatic change, historical events and species diversity will interact in complex ways as the future forest develops. Evidence here indicates that broadleaf-dominated forests could change abruptly from small to subcontinental scales in the coming decades.

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LITERATURE CITED

- Adams, H. D., M. J. Germino, D. D. Breshears, G. A. Barron-Gafford, M. Guardiola-Claramonte, C. B. Zou, and T. E. Huxman. 2013. Nonstructural leaf carbohydrate dynamics of *Pinus edulis* during drought-induced tree mortality reveal role for carbon metabolism in mortality mechanism. *New Phytologist* 197:1142–1151.
- Aldrich, S. R., C. W. Lafon, H. D. Grissino-Mayer, G. G. Deweese, and J. A. Hoss. 2010. Three centuries of fire in montane pine–oak stands on a temperate forest landscape. *Applied Vegetation Science* 13:36–46.
- Allen, C. D., and D. D. Breshears. 1998. Drought-induced shift of a forest–woodland ecotone: Rapid landscape response to climate variation. *Proceedings of the National Academy of Sciences USA* 95:14839–14842.
- Allen, C. D., A. K. Macalady, H. Chenchouni, D. Bachelet, N. McDowell, M. Vennetier, T. Kitzberger, A. Rigling, D. D. Breshears, and E. H. Hogg. 2010. A global overview of drought and heat-induced tree mortality reveals emerging climate change risks for forests. *Forest Ecology and Management* 259:660–684.
- Anderegg, W. R. L., J. A. Berry, D. D. Smith, J. S. Sperry, L. D. L. Anderegg, and C. B. Field. 2012. The roles of hydraulic and carbon stress in a widespread climate-induced

- forest die-off. *Proceedings of the National Academy of Sciences USA* 109:233–237.
- Argyilan, E. P., S. L. Forman, and T. A. Thompson. 2010. The Holocene variability of Lake Michigan water level during the past 1000 years. *Holocene* 20:723–731.
- Augspurger, C. K. 2009. Spring 2007 warmth and frost: phenology, damage and refoliation in a temperate deciduous forest. *Functional Ecology* 23:1031–1039.
- Aukema, J. E., D. G. McCullough, B. Von Holle, A. M. Liebhold, K. Britton, and S. J. Frankel. 2010. Historical accumulation of nonindigenous forest pests in the continental United States. *BioScience* 60:886–897.
- Baker, P., S. Bunyavejchewin, C. D. Oliver, and P. S. Ashton. 2005. Disturbance history and historical stand dynamics of a seasonal tropical forest in western Thailand. *Ecological Monographs* 75:317–343.
- Bégin, Y., and S. Payette. 1988. Dendroecological evidence of lake-level changes during the last three centuries in subarctic Quebec. *Quaternary Research* 30:210–220.
- Biondi, F., T. J. Kozubowski, and A. K. Panorska. 2005. A new model for quantifying climate episodes. *International Journal of Climatology* 25:1253–1264.
- Booth, R. K., S. T. Jackson, V. A. Sousa, M. E. Sullivan, T. A. Minckley, and M. J. Clifford. 2012. Multi-decadal drought and amplified moisture variability drove rapid forest community change in a humid region. *Ecology* 93:219–226.
- Brown, P. M., and R. Wu. 2005. Climate and disturbance forcing of episodic tree recruitment in a southwestern ponderosa pine landscape. *Ecology* 86:3030–3038.
- Burns, R. M., and B. H. Honkala. 1990. *Silvics of North America: 1. Conifers; 2. Hardwoods*. Handbook 654. USDA Forest Service Agriculture, Washington, D.C., USA.
- Chenoweth, M. 2006. A reassessment of historical Atlantic Basin tropical cyclone activity, 1700–1855. *Climatic Change* 76:169–240.
- Cho, D.-S., and R. E. J. Boerner. 1995. Dendrochronological analysis of the canopy history of two Ohio old-growth forests. *Vegetatio* 120:173–183.
- Choat, B., et al. 2012. Global convergence in the vulnerability of forests to drought. *Nature* 491:752–755.
- Clark, J., and P. Royall. 1996. Local and regional sediment charcoal evidence for fire regimes in presettlement northeastern North America. *Journal of Ecology* 84:365–382.
- Clinton, B. D., and W. T. Swank. 1993. Canopy gap characteristics and drought influences in oak forests of the Coweeta Basin. *Ecology* 74:1551–1558.
- Coles, S. 2001. An introduction to statistical modeling of extreme values. Springer-Verlag, London, UK.
- Cook, E., and G. Jacoby. 1977. Tree-ring-drought relationships in the Hudson Valley, New York. *Science* 198:399–401.
- Cook, E., M. Kablack, and G. Jacoby. 1988. The 1986 drought in the southeastern United States: How rare an event was it? *Journal of Geophysical Research* 93:14,257–14,260.
- Cook, E., R. Seager, R. R. J. Heim, R. S. Vose, C. Herweijer, and C. A. Woodhouse. 2010. Megadroughts in North America: placing IPCC projections of hydroclimatic change in a long-term palaeoclimate context. *Journal of Quaternary Science* 25:48–61.
- Cook, E. R. 1982. Tree-ring data networks: comment. Pages 6–7 in M. K. Hughes, P. M. Kelly, J. R. Pilcher, and V. C. LaMarche, Jr., editors. *Climate from tree rings*. Cambridge University Press, London, UK.
- Cook, E. R., and G. C. Jacoby. 1983. Potomac River streamflow since 1730 as reconstructed by tree rings. *Journal of Climate and Applied Meteorology* 22:1659–1672.
- Cook, E. R., and P. J. Krusic. 2004. The North American drought atlas. <http://iridl.ldeo.columbia.edu/SOURCES/.LDEO/.TRL/.NADA2004/.pdsi-atlas.html>
- Cook, E. R., C. A. Woodhouse, C. M. Eakin, D. M. Meko, and D. W. Stahle. 2004. Long-term aridity changes in the western United States. *Science* 306:1015–1018.
- Copenheaver, C. A., B. A. Black, M. B. Stine, R. H. McManamay, and J. Bartens. 2009. Identifying dendroecological growth releases in American beech, jack pine, and white oak: within-tree sampling strategy. *Forest Ecology and Management* 257:2235–2240.
- Cronon, W. 1983. Changes in the land: Indians, colonists, and the ecology of New England. Hill and Wagner, New York, New York, USA.
- Crosby, A. W. 1986. *Ecological imperialism: the biological expansion of Europe, 900–1900*. Cambridge University Press, Cambridge, UK.
- D'Amato, A. W., and D. A. Orwig. 2008. Stand and landscape-level disturbance dynamics in old-growth forests in western Massachusetts. *Ecological Monographs* 78:507–522.
- Davis, D. E. 2000. *An environmental history of the southern Appalachians*. University of Georgia Press, Athens, Georgia, USA.
- Davison, A. C., and R. L. Smith. 1990. Models for exceedances over high thresholds. *Journal of the Royal Statistical Society B* 52:393–442.
- Dey, D. C., and R. P. Guyette. 2000. Anthropogenic fire history and red oak forests in south-central Ontario. *Forestry Chronicle* 76:339–347.
- Dietze, M. C., and P. R. Moorcroft. 2011. Tree mortality in the eastern and central United States: patterns and drivers. *Global Change Biology* 17:3312–3326.
- Druckenbrod, D. L., N. Pederson, J. Rentch, and E. R. Cook. 2013. A comparison of times series approaches for dendroecological reconstructions of past canopy disturbance events. *Forest Ecology and Management* 302:23–33.
- Dyer, J. M. 2006. Revisiting the deciduous forests of eastern North America. *BioScience* 56:341–352.
- Endfield, G. H. 2012. Complexity, causality and collapse: social discontinuity in history and prehistory. Pages 355–374 in J. Matthews, editor. *Handbook of environmental change*. Volume 2. Sage Publications, London, UK.
- Essex Gazette. 1774. Extract of a letter from Baltimore, dated May 21. 7 June 1774, page 176. Essex Gazette, Salem, Massachusetts, USA.
- Everham, E., and N. Brokaw. 1996. Forest damage and recovery from catastrophic wind. *Botanical Review* 62:113–185.
- Feathers, I. C. 2010. Fire history from dendrochronological analyses at two sites near Cades Cove, Great Smoky Mountains National Park, U.S.A. Thesis. University of Tennessee, Knoxville, Tennessee, USA.
- Flatley, W. T., C. W. Lafon, H. D. Grissino-Mayer, and L. B. Laforest. 2013. Fire history related to climate and land use in three southern Appalachian landscapes in the eastern United States. *Ecological Applications* 23:1250–1266.
- Floyd, M. L., M. Clifford, N. S. Cobb, D. Hanna, R. Delph, P. Ford, and D. Turner. 2009. Relationship of stand characteristics to drought-induced mortality in three southwestern piñon-juniper woodlands. *Ecological Applications* 19:1223–1230.
- Foster, D. R. 1988. Disturbance history, community organization and vegetation dynamics of the old-growth Pisgah Forest, South-western New Hampshire, U.S.A. *Journal of Ecology* 76:105–134.
- Foster, D., and E. Boose. 1992. Patterns of forest damage resulting from catastrophic patterns of forest wind in central New England, USA. *Journal of Ecology* 80:79–98.
- Foster, D., M. Fluet, and E. Boose. 1999. Human or natural disturbance: landscape-scale dynamics of the tropical forests of Puerto Rico. *Ecological Applications* 9:555–572.

- Foster, D. R., W. W. Oswald, E. K. Faison, E. D. Doughty, and B. C. S. Hansen. 2006. A climatic driver for abrupt mid-Holocene vegetation dynamics and the hemlock decline in New England. *Ecology* 87:2959–2966.
- Fralish, J. S., and S. B. Franklin. 2002. *Taxonomy and ecology of woody plants in North American forests*. John Wiley and Sons, New York, New York, USA.
- Frelich, L. E. 2002. Forest dynamics and disturbance regimes studies from temperate evergreen-deciduous forests. Cambridge University Press, Cambridge, UK.
- Frelich, L., and C. Lorimer. 1991. Natural disturbance regimes in hemlock-hardwood forests of the Upper Great Lakes Region. *Ecological Monographs* 61:145–164.
- Frelich, L. E., and P. B. Reich. 2010. Will environmental changes reinforce the impact of global warming on the prairie-forest border of central North America? *Frontiers in Ecology and the Environment* 8:371–378.
- Freudenberger, L., P. R. Hobson, M. Schluck, and P. L. Ibisch. 2012. A global map of the functionality of terrestrial ecosystems. *Ecological Complexity* 12:13–22.
- Fries, A. L., editor. 1922. *Records of the Moravians in North Carolina. Volume 1. 1752–1771*. North Carolina Historical Commission, Raleigh, North Carolina, USA.
- Fries, A. L., editor. 1925. *Records of the Moravians in North Carolina. Volume 2. 1752–1775*. North Carolina Historical Commission, Raleigh, North Carolina, USA.
- Fries, A. L., editor. 1926. *Records of the Moravians in North Carolina. Volume 3. 1776–1779*. North Carolina Historical Commission, Raleigh, North Carolina, USA.
- Fritts, H. C. 1976. *Tree rings and climate*. Academic Press, New York, New York, USA.
- Gates, F. C., and G. E. Nichols. 1930. Relation between age and diameter of the trees in the primeval northern hardwood forest. *Journal of Forestry* 28:395–398.
- Gravel, D., M. Beaudet, and C. Messier. 2010. Large-scale synchrony of gap dynamics and the distribution of understory tree species in maple-beech forests. *Oecologia* 162:153–161.
- Gray, S. T., J. J. Lukas, and C. A. Woodhouse. 2011. Millennial-length records of streamflow from three major Upper Colorado River tributaries. *Journal of the American Water Resources Association* 47:702–712.
- Greenlee, D. M. 2006. Dietary variation and prehistoric maize farming in the middle Ohio Valley. Pages 215–233 in J. E. Staller, R. H. Tykot, and B. E. Benz, editors. *Histories of Maize: multidisciplinary approaches to the prehistory, linguistics, biogeography, domestication, and evolution of maize*. Academic Press, Waltham, Massachusetts, USA.
- Grimm, E. 1983. Chronology and dynamics of vegetation change in the prairie-woodland region of southern Minnesota, U.S.A. *New Phytologist* 93:311–350.
- Gu, L., P. J. Hanson, W. M. Post, D. P. Kaiser, B. Yang, R. Nemani, S. G. Pallardy, and T. Meyers. 2008. The 2007 eastern US spring freeze: increased cold damage in a warming world? *BioScience* 58:253–262.
- Gutierrez, A. G., J. C. Aravena, N. V. Carrasco-Farias, D. A. Christie, M. Fuentes, and J. J. Armesto. 2008. Gap-phase dynamics and coexistence of a long-lived pioneer and shade-tolerant tree species in the canopy of an old-growth coastal temperate rain forest of Chiloe' Island, Chile. *Journal of Biogeography* 35:1674–1687.
- Guyette, R. P., and D. C. Dey. 1995. A presettlement fire history of an oak-pine forest near Basin Lake, Algonquin Park, Ontario. *Forest Research Report No. 132*. Ontario Forest Research Institute, Sault Ste. Marie, Ontario, Canada.
- Guyette, R. P., D. C. Dey, and M. C. Stambaugh. 1994. Fire and human history of a barren-forest in southern Indiana. *American Midland Naturalist* 149:21–34.
- Guyette, R. P., D. C. Dey, M. C. Stambaugh, and R.-M. Muzika. 2006. Fire scars reveal variability and dynamics of eastern fire regimes. Pages 20–39 in M. B. Dickinson, editor. *Fire in eastern oak forests: delivering science to land managers*. General Technical Report NRS-P-1. U.S. Department of Agriculture, Forest Service, Northern Research Station, Newtown Square, Pennsylvania, USA.
- Guyette, R. P., R. M. Muzika, and D. C. Dey. 2002. Dynamics of an anthropogenic fire regime. *Ecosystems* 5:472–486.
- Guyette, R. P., and M. A. Spetich. 2003. Fire history of oak-pine forests in the lower Boston Mountains, Arkansas, USA. *Forest Ecology and Management* 180:463–474.
- Haasis, F. W. 1923. Significance of a 255-year age class in an eastern Kentucky forest. *Journal of Forestry* 21:700–704.
- Hanson, J. J., and C. G. Lorimer. 2007. Forest structure and light regimes following moderate wind storms: Implications for multi-cohort management. *Ecological Applications* 17:1325–1340.
- Hart, J. L., S. L. Clark, S. J. Torreano, and M. L. Buchanan. 2012. Composition, structure, and dendroecology of an old-growth *Quercus* forest on the tablelands of the Cumberland Plateau, USA. *Forest Ecology and Management* 266:11–24.
- Hartmann, H. 2011. Will a 385 million year-struggle for light become a struggle for water and for carbon? How trees may cope with more frequent climate change-type drought events. *Global Change Biology* 17:642–655.
- Henry, J. D., and J. M. A. Swan. 1974. Reconstructing forest history from live and dead plant material: an approach to the study of forest succession in southwest New Hampshire. *Ecology* 55:772–783.
- Hogg, E. H., M. Hart, and V. J. Lieffers. 2002. White tree rings formed in trembling aspen saplings following experimental defoliation. *Canadian Journal of Forest Research* 32:1929–1934.
- Hooke, J. M., and R. J. P. Kain. 1982. *Historical change in the physical environment*. Butterworth Scientific, London, UK.
- Hough, A. F., and R. D. Forbes. 1943. The ecology and silvics of forests in the High Plateau of Pennsylvania. *Ecological Monographs* 13:299–320.
- Hubeny, J. B., J. W. King, and M. Reddin. 2011. Northeast US precipitation variability and North American climate teleconnections interpreted from late Holocene varved sediments. *Proceedings of the National Academy of Sciences USA* 108:17895–17900.
- Hufkens, K., M. A. Friedl, T. F. Keenan, O. Sonnentag, A. Bailey, J. O'Keefe, and A. D. Richardson. 2012. Ecological impacts of a widespread frost event following early spring leaf-out. *Global Change Biology* 18:2365–2377.
- Huntington, E. 1914. The climatic factor as illustrated in arid America. Carnegie Institution of Washington, Washington, D.C., USA.
- Hursh, C. R., and F. W. Haasis. 1931. Effects of 1925 summer drought on southern Appalachian hardwoods. *Ecology* 12:380–386.
- Irland, L. C. 2000. Ice storms and forest impacts. *Science of the Total Environment* 262:231–242.
- Iverson, L. R., and A. M. Prasad. 2001. Potential changes in tree species richness and forest community types following climate change. *Ecosystems* 4:186–199.
- Jackson, S. T. 2006. Vegetation, environment, and time: the origination and termination of ecosystems. *Journal of Vegetation Science* 17:549–557.
- Jackson, S. T., J. L. Betancourt, R. K. Booth, and S. T. Gray. 2009. Ecology and the ratchet of events: climate variability, niche dimensions, and species distributions. *Proceedings of the National Academy of Sciences USA* 106:19685–19692.
- Jackson, S. T., and R. K. Booth. 2002. The role of Late Holocene climate variability in the expansion of yellow birch

- in the western Great Lakes region. *Diversity and Distributions* 8:275–284.
- Jenkins, M. A., and S. G. Pallardy. 1995. The influence of drought on red oak group species growth and mortality in the Missouri Ozarks. *Canadian Journal of Forest Research* 25:1119–1127.
- Kaplan, J., and M. DeMaria. 1995. A simple empirical model for predicting the decay of tropical cyclone winds after landfall. *Journal of Applied Meteorology* 34:2499–2512.
- Koutsoyiannis, A. 1977. Theory of econometrics: an introductory exposition of econometric methods. MacMillan Press, New York, New York, USA.
- Krumhardt, K. M. 2010. Methodology for worldwide population estimates: 1000 BC to 1850. Technical Report 3. l'Environnement Université de Lausanne Géopolis, Lusanne, Switzerland.
- Lafon, C. W. 2006. Forest disturbance by ice storms in *Quercus* forests of the southern Appalachian Mountains, USA. *Ecoscience* 13:30–43.
- Lafon, C. W., and S. M. Quiring. 2012. Relationships of fire and precipitation regimes in temperate forests of the eastern United States. *Earth Interactions* 16:1–15.
- Lafon, C. W., and J. H. Speer. 2002. Using dendrochronology to identify major ice storm events in oak forests of southwestern Virginia. *Climate Research* 20:41–54.
- Landsea, C. W., C. Anderson, N. Charles, G. Clark, J. Dunion, J. Fernandez-Partgas, P. Hungerford, C. Neumann, and M. Zimmer. 2004. The Atlantic Hurricane Database Re-analysis Project: documentation for 1851–1910 alterations and additions to the HURDAT database. Pages 178–221 in R. Murnane and K. Liu, editors. *Hurricanes and typhoons, past, present and future*. Columbia University Press, New York, New York, USA.
- Lemon, P. C. 1961. Forest ecology of ice storms. *Bulletin of the Torrey Botanical Club* 88:21–29.
- Lichter, J. 1997. Radiocarbon dating of Lake Michigan beachridge and dune development. *Quaternary Research* 48:137–140.
- Loehle, C. 2000. Forest ecotone response to climate change: sensitivity to temperature response functional forms. *Canadian Journal of Forest Research* 30:1632–1645.
- Loope, W. L., and A. F. Arbogast. 2000. Dominance of an ~150-year cycle of sand-supply change in Late Holocene dune-building along the eastern shore of Lake Michigan. *Quaternary Research* 54:414–422.
- Lorimer, C. 1977. The presettlement forest and natural disturbance cycle of northeastern Maine. *Ecology* 58:139–148.
- Lorimer, C. G. 1980. Age structure and disturbance history of a Southern Appalachian virgin forest. *Ecology* 61:1169–1184.
- Lorimer, C. G. 1984. Development of the red maple understory in northeastern oak forests. *Forest Science* 1:3–22.
- Lorimer, C. G. 1985. Methodological considerations in the analysis of forest disturbance history. *Canadian Journal of Forest Research* 15:200–213.
- Lorimer, C. G. 1989. Relative effects of small and large disturbances on temperate hardwood forest structure. *Ecology* 70:565–567.
- Lorimer, C. G., and L. E. Frelich. 1989. A methodology for estimating canopy disturbance frequency and intensity in dense temperate forests. *Canadian Journal of Forest Research* 19:651–663.
- Ludlum, D. W. 1963. Early American Hurricanes, 1492–1870. American Meteorological Society, Boston, Massachusetts, USA.
- Lynch, C., and A. Hessl. 2010. Climatic controls on historical wildfires in West Virginia, 1939–2008. *Physical Geography* 31:254–269.
- Man, G. 2012. Major forest insect and disease conditions in the United States, 2011. FS-1000. USDA Forest Service, Washington, D.C., USA.
- Manion, P. 2003. Evolution of concepts in forest pathology. *Phytopathology* 98:1052–1055.
- Mann, C. C. 2005. 1491: New revelations of the Americas before Columbus. Knopf, New York, New York, USA.
- Mann, D. H., F. B. Engstrom, and J. L. Bubier. 1994. Fire history and tree recruitment in an uncut New England forest. *Quaternary Research* 42:206–215.
- Maxwell, R. S., A. E. Hessl, E. R. Cook, and N. Pederson. 2011. A multispecies tree ring reconstruction of Potomac River streamflow (950–2001). *Water Resources Research* 47:1–12.
- McCarthy, J. 2001. Gap dynamics of forest trees: a review with particular attention to boreal forests. *Environmental Reviews* 9:1–59.
- McDowell, N., W. T. Pockman, C. D. Allen, D. D. Breshears, N. Cobb, T. Kolb, J. Plaut, J. Sperry, A. West, D. G. Williams, and E. A. Yepez. 2008. Mechanisms of plant survival and mortality during drought: why do some plants survive while others succumb to drought? *New Phytologist* 178:719–739.
- McEwan, R. W., J. M. Dyer, and N. Pederson. 2011. Multiple interacting ecosystem drivers: toward an encompassing hypothesis of oak forest dynamics across eastern North America. *Ecography* 34:244–256.
- McEwan, R. W., T. F. Hutchinson, R. P. Long, D. R. Ford, and C. B. McCarthy. 2007. Temporal and spatial patterns in fire occurrence during the establishment of mixed-oak forests in eastern North America. *Journal of Vegetation Science* 18:655–664.
- McEwan, R. W., N. Pederson, A. Cooper, J. Taylor, R. Watts, and A. Hruska. 2014. Fire and gap dynamics over 300 years in an old-growth temperate forest. *Applied Vegetation Science* 17:312–322.
- McGuire, J. P., R. J. Mitchell, E. B. Moser, S. D. Pecot, D. H. Gjerstad, and C. W. Hedman. 2001. Gaps in a gappy forest: plant resources, longleaf pine regeneration, and understory response to tree removal in longleaf pine savannas. *Canadian Journal of Forest Research* 31:765–778.
- McKenney, D. W., J. H. Pedlar, R. B. Rood, and D. Price. 2011. Revisiting projected shifts in the climate envelopes of North American trees using updated general circulation models. *Global Change Biology* 17:2720–2730.
- McMahon, S. M., G. G. Parker, and D. R. Miller. 2010. Evidence for a recent increase in forest growth. *Proceedings of the National Academy of Sciences USA* 107:3611–3615.
- McMurtry, E. R., M. C. Stambaugh, R. P. Guyette, and D. C. Dey. 2007. Fire scars reveal source of New England's 1780 Dark Day. *International Journal of Wildland Fire* 16:266–270.
- Michaelian, M., E. H. Hogg, R. J. Hall, and E. Arsenault. 2011. Massive mortality of aspen following severe drought along the southern edge of the Canadian boreal forest. *Global Change Biology* 17:2084–2094.
- Millar, C. I., N. L. Stephenson, and S. L. Stephens. 2007. Climate change and forests of the future: managing in the face of uncertainty. *Ecological Applications* 17:2145–2151.
- Millward, A. A., and C. E. Kraft. 2004. Physical influences of landscape on a large-extent ecological disturbance: the northeastern North American ice storm of 1998. *Landscape Ecology* 19:99–111.
- Milner, G., and G. Chaplin. 2010. Eastern North American population at ca A.D. 1500. *American Antiquity* 75:707–726.
- Mueller, R. C., C. M. Scudder, M. E. Porter, R. T. Trotter, C. A. Gehring, and T. G. Whitham. 2005. Differential tree mortality in response to severe drought: evidence for long-term vegetation shifts. *Journal of Ecology* 93:1085–1093.

- Muggeo, V. 2008. Modelling temperature effects on mortality: multiple segmented relationships with common break points. *Biostatistics* 9:613–620.
- Negrón-Juárez, R. I., J. Q. Chambers, G. Guimaraes, H. Zeng, C. F. M. Raupp, D. M. Marra, G. H. P. M. Ribeiro, S. S. Saatchi, B. W. Nelson, and N. Higuchi. 2010. Widespread Amazon forest tree mortality from a single cross-basin squall line event. *Geophysical Research Letters* 37:1–5.
- Nowacki, G. J., and M. D. Abrams. 1994. Forest composition, structure, and disturbance history of the Alan Seeger Natural Area, Huntington, Pennsylvania. *Bulletin of the Torrey Botanical Club* 121:277–291.
- Nowacki, G. J., and M. D. Abrams. 1997. Radial-growth averaging criteria for reconstructing disturbance histories from presettlement-origin oaks. *Ecological Monographs* 67:225–234.
- Olano, J. M., and M. W. Palmer. 2003. Stand dynamics of an Appalachian old-growth forest during a severe drought episode. *Forest Ecology and Management* 174:139–148.
- Orwig, D. A., and M. D. Abrams. 1994. Contrasting radial growth and canopy recruitment patterns in *Liriodendron tulipifera* and *Nyssa sylvatica*: gap-obligate versus gap-facultative tree species. *Canadian Journal of Forest Research* 24:2141–2149.
- Parshall, T., and D. Foster. 2002. Fire on the New England landscape: regional and temporal variation, cultural and environmental controls. *Journal of Biogeography* 29:1305–1317.
- Payette, S., L. Filion, L. Gauthier, and Y. Boutin. 1985. Secular climate change in old-growth tree-line vegetation of northern Quebec. *Nature* 315:135–138.
- Pedersen, B. 1998. The role of stress in the mortality of midwestern oaks as indicated by growth prior to death. *Ecology* 79:79–93.
- Pederson, N. 2005. Climatic sensitivity and growth of southern temperate trees in the eastern US: implications for the carbon cycle. Dissertation. Columbia University, New York, New York, USA.
- Pederson, N. 2010. External characteristics of old trees in the eastern deciduous forest. *Natural Areas Journal* 30:396–407.
- Pederson, N., A. R. Bell, E. R. Cook, U. Lall, N. Devineni, R. Seager, K. Eggleston, and K. P. Vranes. 2013. Is an epic pluvial masking the water insecurity of the greater New York City region? *Journal of Climate* 26:1339–1354.
- Pederson, N., et al. 2012a. Long-term perspective on a modern drought in the American Southeast. *Environmental Research Letters* 7:014034.
- Pederson, N., K. Tackett, R. W. McEwan, S. Clark, A. M. Cooper, G. Brosi, R. Eaton, and R. D. Stockwell. 2012b. Long-term drought sensitivity of trees in second-growth forests in a humid region. *Canadian Journal of Forest Research* 42:1837–1850.
- Pederson, N., J. M. Varner III, and B. J. Palik. 2008. Canopy disturbance and tree recruitment over two centuries in a managed longleaf pine landscape. *Forest Ecology and Management* 254:85–95.
- Pennsylvania Gazette. 1774. May 11, page 2. Pennsylvania Gazette, Philadelphia, Pennsylvania, USA.
- Peterson, C. 2000. Catastrophic wind damage to North American forests and the potential impact of climate change. *Science of the Total Environment* 262:287–311.
- Peterson, C. J. 2007. Consistent influence of tree diameter and species on damage in nine eastern North America tornado blowdowns. *Forest Ecology and Management* 250:96–108.
- Potts, M. 2003. Drought in a Bornean everwet rain forest. *Journal of Ecology* 91:467–474.
- Preisendorfer, R. W., F. W. Zwiers, and T. Barnett. 1981. Foundations of principal components selection rules. Scripps Institution of Oceanography, La Jolla, California, USA.
- Proulx, R. J., and D. F. Greene. 2001. The relationship between ice thickness and northern hardwood tree damage during ice storms. *Canadian Journal of Forest Research* 31:1758–1767.
- Raffa, K. F., B. H. Aukema, B. J. Bentz, A. L. Carroll, J. A. Hicke, M. G. Turner, and W. H. Romme. 2008. Cross-scale drivers of natural disturbances prone to anthropogenic amplification: the dynamics of bark beetle. *BioScience* 58:501–517.
- Rappaport, E., and A. Ruffman. 1999. The catastrophic hurricane(s): the search for data and understanding. Pages 787–790 in A. M. Society, editor. *Preprints, 23rd Conference on Hurricanes and Tropical Meteorology*. American Meteorological Society, Boston, Massachusetts, USA.
- Rebertus, A. J., and T. T. Veblen. 1993. Structure and tree-fall gap dynamics of old-growth *Nothofagus* forests in Tierra del Fuego, Argentina. *Journal of Vegetation Science* 4:641–654.
- Rentch, J. 2003. Oak establishment and canopy accession strategies in five old-growth stands in the central hardwood forest region. *Forest Ecology and Management* 184:285–297.
- Rentch, J. S., F. Desta, and G. W. Miller. 2002. Climate, canopy disturbance, and radial growth averaging in a second-growth mixed-oak forest in West Virginia, U.S.A. *Canadian Journal of Forest Research* 32:915–927.
- Richman, M. B. 1986. Rotation of principal components. *Journal of Climatology* 6:293–335.
- Rudisill, H. F. 1993. The diaries of Evan Pugh (1762–1801). H. F. Rudisill, editor. St David's Society, Florence, South Carolina, USA.
- Runkle, J. R. 1982. Patterns of disturbance in some old-growth mesic forests of eastern North America. *Ecology* 63:1533–1546.
- Savage, M., P. M. Brown, and J. Feddema. 1996. The role of climate in a pine forest regeneration pulse in the southwestern United States. *Ecoscience* 3:310–318.
- Schleeweis, K., S. N. Goward, C. Huang, J. G. Masek, G. Moisen, R. E. Kennedy, and N. E. Thomas. 2013. Regional dynamics of forest canopy change and underlying causal processes in the contiguous U.S. *Journal of Geophysical Research: Biogeosciences* 118:1035–1053.
- Seager, R., N. Pederson, Y. Kushnir, and J. Nakamura. 2012. The 1960s drought and the subsequent shift to a wetter climate in the Catskill Mountains Region of the New York City watershed. *Journal of Climate* 25:6721–6742.
- Shuman, B., A. K. Henderson, C. Plank, I. Stefanova, and S. S. Ziegler. 2009. Woodland-to-forest transition during prolonged drought in Minnesota after ca. AD 1300. *Ecology* 90:2792–2807.
- Shuman, B., P. Newby, Y. Huang, and T. Webb. 2004. Evidence for the close climatic control of New England vegetation history. *Ecology* 85:1297–1310.
- Smith, T., and D. Urban. 1988. Scale and resolution of forest structural pattern. *Vegetatio* 74:143–150.
- Sprugel, D. G. 1991. Disturbance, equilibrium, and environmental variability: What is “natural” vegetation in a changing environment? *Biological Conservation* 58:1–18.
- St. George, S., D. M. Meko, M.-P. Girardin, G. M. MacDonald, E. Nielsen, G. T. Pederson, D. J. Sauchyn, J. C. Tardif, and E. Watson. 2009. The tree-ring record of drought on the Canadian Prairies. *Journal of Climate* 22:689–710.
- Stahle, D. W., and P. L. Chaney. 1994. A predictive model for the location of ancient forests. *Natural Areas Journal* 14:151–158.
- Stahle, D. W., and M. K. Cleaveland. 1988. Texas drought history reconstructed and analyzed from 1698 to 1980. *Journal of Climate* 1:59–74.
- Stahle, D. W., M. K. Cleaveland, and J. Hehr. 1985. A 450-year drought reconstruction for Arkansas, United States. *Nature* 316:530–532.

- Stahle, D. W., E. R. Cook, M. K. Cleaveland, M. D. Therrell, H. D. Grissino-Mayer, E. Watson, and B. H. Luckman. 2000. Tree-ring data document 16th century megadrought over North America. *EOS Transactions* 81:121–125.
- Stambaugh, M. C., J. Sparks, R. P. Guyette, and G. Wilson. 2011. Fire history of a relict oak woodland in northeast Texas. *Rangeland Ecology and Management* 64:419–423.
- Stueve, K. M., C. H. Perry, M. D. Nelson, S. P. Healey, A. D. Hill, G. G. Moisen, W. B. Cohen, D. D. Gormanson, and C. Huang. 2011. Ecological importance of intermediate windstorms rivals large, infrequent disturbances in the northern Great Lakes. *Ecosphere* 2:art2.
- Swetnam, T. W. 1993. Fire history and climate change in giant sequoia groves. *Science* 262:885–889.
- Swetnam, T. W., and J. L. Betancourt. 1998. Mesoscale disturbance and ecological response to decadal climatic variability in the American Southwest. *Journal of Climate* 11:3128–3147.
- Taylor, S. W., A. L. Carroll, R. I. Alfaro, and L. Safranyik. 2006. Forest, climate and mountain pine beetle outbreak dynamics in western Canada. Pages 66–94 in L. Safranyik and B. Wilson, editors. *The mountain pine beetle: a synthesis of biology, management, and impacts on lodgepole pine*. Natural Resources Canada, Canadian Forest Service, Pacific Forestry Centre, Victoria, British Columbia, Canada.
- Therrell, M., and D. Stahle. 1998. A predictive model to locate ancient forests in the Cross Timbers of Osage County, Oklahoma. *Journal of Biogeography* 25:847–854.
- Thomas Jefferson's Garden Book. 1774. Papers of Thomas Jefferson in the collections of the Massachusetts Historical Society. Massachusetts Historical Society, Boston, Massachusetts, USA.
- Turner, M. G. 2010. Disturbance and landscape dynamics in a changing world. *Ecology* 91:2833–2849.
- Tyrrell, L. E., and T. R. Crow. 1994. Structural characteristics of old-growth hemlock–hardwood forests in relation to age. *Ecology* 75:370–386.
- U.S. Environmental Protection Agency. 2003. Level III ecoregions of the conterminous United States. U.S. EPA Office of Research and Development, National Health and Environmental Effects Research Laboratory, Corvallis, Oregon, USA.
- van Mantgem, P. J., et al. 2009. Widespread increase of tree mortality rates in the western United States. *Science* 323:521–524.
- Vanderwel, M. C., D. A. Coomes, and D. W. Purves. 2013. Quantifying variation in forest disturbance, and its effects on aboveground biomass dynamics, across the eastern United States. *Global Change Biology* 19:1504–1517.
- Villalba, R., and T. T. Veblen. 1997. Regional patterns of tree population age structures in northern Patagonia: climatic and disturbance influences. *Journal of Ecology* 85:113–124.
- Villalba, R., and T. Veblen. 1998. Influences of large-scale climatic variability on episodic tree mortality in northern Patagonia. *Ecology* 79:2624–2640.
- Virginia Gazette. 1774. May 12, page 4. Virginia Gazette, Williamsburg, Virginia, USA.
- Waller, M. 2013. Drought, disease, defoliation and death: forest pathogens as agents of past vegetation change. *Journal of Quaternary Science* 28:336–342.
- White, P. 1979. Pattern, process and natural disturbance in vegetation. *Botanical Review* 45:229–299.
- White, P. S., J. Harrod, W. Romme, and J. Betancourt. 1999. The role of disturbance and temporal dynamics. Pages 281–312 in R. C. Szaro, N. C. Johnson, W. T. Sexton, A. J. Malk, editors. *Ecological stewardship*. Volume 2. Elsevier Science, Oxford, UK.
- Williams, A. B. 1936. The composition and dynamics of a beech–maple climax community. *Ecological Monographs* 6:317–408.
- Williams, J. W., and S. T. Jackson. 2007. Novel climates, no-analog communities, and ecological surprises. *Frontiers in Ecology and the Environment* 5:475–482.
- Williams, M. 1992. *Americans and their forests: a historical geography*. Cambridge University Press, Cambridge, UK.
- Wolin, J. A. 1996. Late Holocene lake-level and lake development signals in Lower Herring Lake, Michigan. *Journal of Paleolimnology* 15:19–45.
- Wonkka, C. L., C. W. Lafon, C. M. Hutton, and A. J. Joslin. 2013. A CSR classification of tree life history strategies and implications for ice storm damage. *Oikos* 122:209–222.
- Woods, K. D. 2004. Intermediate disturbance in a late-successional hemlock–northern hardwood forest. *Journal of Ecology* 92:464–476.
- Worrall, J. J., L. Egeland, T. Eager, R. A. Mask, E. W. Johnson, P. A. Kemp, and W. D. Shepperd. 2008. Rapid mortality of *Populus tremuloides* in southwestern Colorado, USA. *Forest Ecology and Management* 255:686–696.
- Worrall, J. J., T. D. Lee, and T. C. Harrington. 2005. Forest dynamics and agents that initiate and expand canopy gaps in *Picea–Abies* forests of Crawford Notch, New Hampshire, USA. *Journal of Ecology* 93:178–190.
- Yamamoto, S.-I. 1992. The gap theory in forest dynamics. *Botanical Magazine Tokyo* 105:375–383.

SUPPLEMENTAL MATERIAL

Ecological Archives

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