

¹ Woody plant phenological responses are strongly associated
² with key functional traits

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⁴ August 15, 2024

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²¹ Running title: Budburst responses associated with traits

³¹ **Summary**

³² Species-level variation in phenology—the timing of recurring life history events—can vary seasonally
³³ with changes in climatic risk, light, and nutrients. This favours acquisitive growth strategies early in
³⁴ the spring season and conservative growth strategies under the more climatically benign, but compet-
³⁵ itive, environment later in the season. This framework infers suites of traits that may co-vary with
³⁶ phenologies, but high trait variability across environments makes this challenging to test. Here, we
³⁷ combine a new joint modeling approach to accommodate this variability with global data on plant
³⁸ traits and budburst responses in controlled environment experiments. We find that earlier species—
³⁹ which are generally most responsive to anthropogenic warming—are generally shorter with denser,
⁴⁰ lower nitrogen leaves. These results suggest warming may reshape the trait structure of plant com-
⁴¹ munities, and could help improve predictions of how growth strategies and phenologies together may
⁴² shift with continued climate change.

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⁴⁴ Key Words: Budburst, spring phenology, functional traits, trees, climate change, forest communities

⁴⁵ **Introduction**

⁴⁶ The timing of life history events—phenology—can shape both ecosystem services and community
⁴⁷ dynamics. Spring phenology, for example, defines the start and overall length of the growing season—
⁴⁸ shaping forest carbon storage and species interactions^{3,9,18}. Shifts in phenology with climate change
⁴⁹ across systems^{31,45} have thus led to growing concerns over their possible impacts.

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⁵¹ Predicting these changes requires understanding the drivers of phenology both at a proximate scale—
⁵² the environmental triggers of phenology each year, such as temperature and daylength—and at an
⁵³ ultimate scale, where long-term environmental pressures may select for different phenologies across
⁵⁴ species e.g., certain species are early or late relative to other species each year,^{33,51}. At the proximate
⁵⁵ level, environmental conditions throughout the winter and spring cause species to start growth at dif-
⁵⁶ ferent times. This is well documented for the start of growth in woody plants each year^{15,24}. Similar
⁵⁷ environmental conditions appear to trigger spring phenological events across taxa, including in the
⁵⁸ timing of egg laying in birds^{11,12} and the advance of spawning in amphibians^{23,43}, but current work
⁵⁹ provides limited insights into the drivers of species differences^{6,15,24}.

⁶⁰

⁶¹ At the ultimate level, species phenologies may vary due to changing pressures across the growing season.
⁶² Species that start growth early often risk high tissue loss—due to frost damage^{2,37} or high herbivore
⁶³ apparency⁴⁶—but benefit from higher resource availability^{21,34}. In contrast, later species face greater
⁶⁴ biotic pressures, especially from high competition for resources^{27,52}. For plants, this variation in early
⁶⁵ to late season growth, may mirror the stressors from early to late successional communities, and may
⁶⁶ similarly shape phenology²⁴.

⁶⁷

⁶⁸ Different pressures could shape a number of species attributes related to their growth strategies, includ-
⁶⁹ ing phenology. Species with earlier phenology may produce cheaper tissues that are easily replaced if
⁷⁰ damaged³⁶, while species with later phenology may benefit from investing in tissues that infer greater
⁷¹ resource retention¹⁶. Differences in traits, and trade-offs in allocation of resources to growth and tissue
⁷² quality, can be related to a broader framework of species growth strategies and functional traits⁵² (Fig
⁷³ 1), where species range from acquisitive (fast) to more conservative (slow) growth^{5,53}.

⁷⁴

⁷⁵ Globally, gradients from acquisitive to conservative strategies appear common, and form the foun-
⁷⁶ dation of the leaf economic and the wood economic spectra^{5,13,47,48,53}, but they can make limited
⁷⁷ predictions of trait variability⁴⁴. As a result, highly variable traits like phenology are often excluded
⁷⁸ from trait studies, leaving the relationships between broader trait syndromes and phenology largely

79 unknown. Understanding these relationships is critical to forecasting community dynamics and re-
80 sponses to climate change. To date, studies have generally only examined the relationship between
81 traits and phenology within a single site, as reviewed by Wolkovich and Cleland⁵⁰ and Wolkovich and
82 Donahue⁵¹, where the problem of proximate drivers causing phenological variation can be more easily
83 ignored. Adding phenology to broader trait frameworks becomes more complex when high levels of
84 variation occurs across large spatial and temporal ranges.

85

86 Consistently defining early to late phenology is possible using the underlying cues that predict gra-
87 dients in phenology which do not generally vary strongly across space and time,^{7,15,20}. For many
88 plants, early species generally have responses that are small in magnitude to all three major cues of
89 spring leafout: warm spring temperatures (forcing), cool winter temperatures (chilling) and daylength
90 (photoperiod). In contrast, later species have larger responses to chilling and/or photoperiod^{15,24}, and
91 likely larger forcing responses.

92

93 Studies of spring phenology in temperate forests may provide the best opportunity to integrate phe-
94 nology into functional trait research. In addition to how well we understand the environmental cues
95 that trigger early versus late leafout, spring in many forests includes strong gradients in potential se-
96 lective environments (Fig 1). Based on trade-offs between early and late spring phenologies, we predict
97 acquisitive species to be shorter, with leaf traits favourable to higher light availability and tolerance
98 to late spring frost (high specific leaf area, SLA, and leaf nitrogen content, LNC; Fig 1). Such species
99 should exhibit early phenology, with small cue responses. Canopy species that budburst later via larger
100 cue responses, when competition for soil resources is greater, would then have traits associated with
101 conservative growth—taller with denser wood²⁵—with leaf traits suited for more variable light (low
102 SLA and LNC, Fig 1). Seed size may similarly be predicted from this acquisitive to conservative con-
103 tinuum, as acquisitive species produce smaller seeds and conservative species produce larger—better
104 provisioned—seeds (Fig 1).

105

106 To test our predicted relationships between budburst responses to environmental cues and common
107 functional traits (height, SLA, seed mass, and LNC), we merged available data from trait databases—
108 BIEN²⁹ and TRY²²—with budburst data from the OSPREE database of controlled environment stud-
109 ies¹⁴. We developed a hierarchical Bayesian joint model that predicts phenological responses to forcing,
110 chilling and photoperiod treatments based on species-level trait values, while allowing additional varia-
111 tion due to species. This approach takes a step towards predicting variation via species traits instead of
112 species identity (when traits explain a significant portion of the variation, species identity will explain
113 only a small amount), which could help forecast species phenological responses based on trait values
114 alone.

115

116 Methods

117 We merged three major databases for our analysis. We gathered phenological data from the OSPREE
118 database¹⁴, which contains budburst data for woody, deciduous species from experiments of forcing,
119 chilling and photoperiod. We updated this database since its initial publication, the methods of which
120 are discussed by Morales-Castilla et al.³². We gathered trait data from TRY and BIEN (v. 4.0)^{22,29},
121 both of which are large trait databases that include plant trait data across many individuals, species,
122 and studies (Table S1). We obtained data from both databases on 5 December, 2018, with an updated
123 version of the TRY data obtained 10 April, 2019. We focused our search for trait data on the subset of
124 234 OSPREE species used in Morales-Castilla et al.³². Using the BIEN R package, version 1.2.5²⁹, we
125 downloaded trait data for the 94 species available, for which there were 13 traits. The TRY database
126 included data for 10 traits for 96 of our focal species²². Given our focus on the phenology of adult
127 trees, we included trait data from adult individuals with a minimum height of 1.38 m. We further

removed all data from experiments or from plants growing in non-natural habitats. We also grouped trait values where appropriate, for example categorizing trait values for “SLA”, “SLA with petioles”, and “SLA without petioles” as simply “SLA” in our analysis (see Table S1). Duplicated data in both the TRY and BIEN datasets were also removed ($n = 434905$). Based on our selection criteria, our final dataset included data for 11 traits from 91 of the species also represented in the OSPREE database, with each species differing in the number and types of traits measured.

For our analysis, we only include species for which we had a complete trait profile (i.e., all traits measured for all species). We initially considered six commonly measured traits—SLA, leaf dry matter content (LDMC), height, seed mass, stem specific density (SSD), and LNC—for which 26 species had at least one trait measurement for each trait. We then used a principle component analysis to understand trait correlations and adjusted which traits we included. A PCA of our six initial traits identified high correlations between SLA and LDMC, and between height and SSD (see Supplementary material). The first principal component explained 32% of variation while the second explained 24.2% of the variation (Fig. S1). By excluding one trait from each of these highly correlated trait pairs (specifically LDMC and SSD) we increased the number of species in our dataset from the 26 species with six traits, to 37 species for which we had complete datasets for four traits. The data for these 37 species were from 24 unique studies (height $n = 47781$, seed mass $n = 281$, LNC $n = 3853$, SLA $n = 7656$). We subsampled height measurements to reduce the influence of the 13 most frequently measured tree species on our height model. Since these 13 species were measured 19 times more frequently than other species, for each species, we randomly sampled 3000 height measurements.

149

150 Joint model of trait and phenology

To understand connections between phenology and species traits, we developed and then parameterized a joint model for each trait: height, SLA, LNC, and seed mass. Our model is a joint model insofar as it involves two types of data—trait observations and phenological observations—that arise from shared latent processes. In particular, we assume that species “true” trait values determine observed trait values across different studies (trait sub-model), and separately, that the same “true” trait values interact with phenological cues (forcing, chilling, and photoperiod) to determine observed phenology, specifically day of year of budburst (phenology sub-model). Below we describe the two sub-models, noting which parameters are shared across sub-models and which are independent.

159 Trait sub-model

The trait sub-model describes the processes that determine trait observations for 1 to n species across each of the 1 to m studies in our trait dataset (TRY and BIEN data). We use hierarchical modeling to partition trait variation by measurement error, species identity, and study identity. In particular, we assume that a trait observation for species i from study j , $Y_{\text{trait}_{i,j}}$, has the following normal distribution:

$$Y_{\text{trait}_{i,j}} \sim \mathcal{N}(\mu_{i,j}, \sigma_m^2) \quad (1)$$

with

$$\mu_{i,j} = \alpha_{\text{grand trait}} + \alpha_{\text{trait}_i} + \alpha_{\text{study}_j} \quad (2)$$

where α_{trait_i} and α_{study_j} are elements of the normal random vectors:

$$\begin{aligned} \boldsymbol{\alpha}_{\text{trait}} &= \{\alpha_{\text{trait}_1}, \dots, \alpha_{\text{trait}_n}\}^T \text{ such that } \boldsymbol{\alpha}_{\text{trait}} \sim \mathcal{N}(0, \sigma_{\text{trait}}^2) \\ \boldsymbol{\alpha}_{\text{study}} &= \{\alpha_{\text{study}_1}, \dots, \alpha_{\text{study}_m}\}^T \text{ such that } \boldsymbol{\alpha}_{\text{study}} \sim \mathcal{N}(0, \sigma_{\text{study}}^2) \end{aligned} \quad (3)$$

165 The latent parameter $\alpha_{\text{grand trait}}$ represents a trait value that is independent of study and species,
 166 $\alpha_{\text{species},i}$ and $\alpha_{\text{study},j}$ are species and study-level offsets from that trait value, σ_m^2 is measurement error,
 167 and σ_{trait}^2 and σ_{study}^2 represent species and study-level variances in trait values. Of these parameters,
 168 α_{trait} are shared by the phenology sub-model.

169 Phenology sub-model

170 The phenology sub-model describes the processes that determine phenological observations for 1 to
 171 n species, specifically the timing (day of year) of budburst from the updated OSPREE dataset. We
 172 assume that an observation of budburst day for species k under set g of chilling, forcing, and photoper-
 173 iod treatments, which we z -scored to allow direct comparison of cues (c_g, f_g, p_g) , $Y_{\text{pheno}_{k,g}}$, has the
 174 following normal distribution:

$$Y_{\text{pheno}_{k,g}} \sim \mathcal{N}(\mu_{k,g}, \sigma_d^2) \quad (4)$$

with

$$\mu_{k,g} = \alpha_{\text{pheno},k} + \beta_{\text{chill}_k} \cdot c_g + \beta_{\text{force}_k} \cdot f_g + \beta_{\text{photo}_k} \cdot p_g \quad (5)$$

and

$$\begin{aligned} \beta_{\text{chill}_k} &= \alpha_{\text{chill},k} + \beta_{\text{trait.chill}} \cdot \alpha_{\text{trait},k} \\ \beta_{\text{force}_k} &= \alpha_{\text{force},k} + \beta_{\text{trait.force}} \cdot \alpha_{\text{trait},k} \\ \beta_{\text{photo}_k} &= \alpha_{\text{photo},k} + \beta_{\text{trait.photo}} \cdot \alpha_{\text{trait},k} \end{aligned} \quad (6)$$

175 where α_{pheno_k} , α_{chill_k} , α_{force_k} , and α_{photo_k} are elements of the normal random vectors:

$$\begin{aligned} \boldsymbol{\alpha}_{\text{pheno}} &= \{\alpha_{\text{pheno}_1}, \dots, \alpha_{\text{pheno}_n}\}^T \text{ such that } \boldsymbol{\alpha}_{\text{pheno}} \sim \mathcal{N}(\mu_{\text{pheno}}, \sigma_{\text{pheno}}^2) \\ \boldsymbol{\alpha}_{\text{chill}} &= \{\alpha_{\text{chill}_1}, \dots, \alpha_{\text{chill}_n}\}^T \text{ such that } \boldsymbol{\alpha}_{\text{chill}} \sim \mathcal{N}(\mu_{\text{chill}}, \sigma_{\text{chill}}^2) \\ \boldsymbol{\alpha}_{\text{force}} &= \{\alpha_{\text{force}_1}, \dots, \alpha_{\text{force}_n}\}^T \text{ such that } \boldsymbol{\alpha}_{\text{force}} \sim \mathcal{N}(\mu_{\text{force}}, \sigma_{\text{force}}^2) \\ \boldsymbol{\alpha}_{\text{photo}} &= \{\alpha_{\text{photo}_1}, \dots, \alpha_{\text{photo}_n}\}^T \text{ such that } \boldsymbol{\alpha}_{\text{photo}} \sim \mathcal{N}(\mu_{\text{photo}}, \sigma_{\text{photo}}^2) \end{aligned} \quad (7)$$

176 Parameter $\alpha_{\text{pheno},k}$ represents the day of budburst for species k without the influence of chilling, forc-
 177 ing, or photoperiod treatments. Average day of budburst (independent of treatments) is μ_{pheno} , and
 178 σ_{pheno}^2 is the variance across species. The parameters $\alpha_{\text{chill},k}$, $\alpha_{\text{force},k}$, and $\alpha_{\text{photo},k}$ represent the trait-
 179 independent responses of species k to chilling, forcing, and photoperiod treatments respectively, each
 180 with an associated mean (μ_{chill} , μ_{force} , μ_{photo}) and variance (σ_{chill}^2 , σ_{force}^2 , σ_{photo}^2) across species. The
 181 effect of the species trait value, $\alpha_{\text{trait},k}$ (parameter shared with trait sub-model above), on its responses
 182 to chilling, forcing, and photoperiod are described by parameters $\beta_{\text{trait.chill}}$, $\beta_{\text{trait.force}}$, $\beta_{\text{trait.photo}}$. Fi-
 183 nally, σ_d^2 is the variance arising from measurement error.

184
 185 We chose weakly informative priors, and validated them using a series of prior predictive checks. The
 186 model was coded in the Stan programming language, fit using the rstan package (version 3.3.6)⁴¹,
 187 with 1,000 iterations per chain across 4 chains (4,000 total sampling iterations), and all models met
 188 basic diagnostic checks, including no divergences, high effective sample size (n_{eff}), and \hat{R} close to 1,
 189 fitting the data well (Fig S2). Here we present our model estimates as the means and 90% posterior
 190 uncertainty intervals.

191 **Results**

192 Across traits, height, SLA, and LNC strongly related to chilling, forcing and photoperiod treatments
193 (β_{chill_k} , β_{force_k} , and β_{photo_k} , Fig 2 a-f & j-l), but the direction of these relationships only showed
194 consistent trends for LNC (Fig. 2 j-l). As we predicted, height was related to chilling (β_{chill_k}) and pho-
195 toperiod (β_{photo_k}), with taller species having larger responses to cues (-0.5 m per standardized chilling;
196 90% uncertainty interval (UI): -1.0, -0.1 and -0.2 m per standardized photoperiod; 90% UI: -0.5, 0.0,
197 Fig 2 a-c, Table S2). As illustrated for one characteristically acquisitive species, *Alnus incana*, and
198 one characteristically conservative species, *Quercus rubra* (Fig S3), the cue relationships with height
199 led to generally later budburst ($\mu_{k,g}$; Fig. 3). In contrast, seed mass had the smallest responses, with
200 no relationship between seed mass and any cue (Fig. 2 g-i, Fig 3 d-f, & Table S3).

201
202 Of our leaf traits, we found that species SLA related to photoperiod (β_{photo_k} , -0.2 mm²/mg per stan-
203 dardized photoperiod; 90% UI: -0.4, 0.0, Fig. 2 f, Table S4), but did not strongly predict responses
204 to chilling (β_{chill_k}) or forcing treatments (β_{force_k} , Fig. 2 d and e). Thus, species with more acquisitive
205 growth strategies (thin leaves and a lower investment in leaf mass that leads to large SLA values),
206 had larger responses to photoperiod, contrary to our predictions (Fig. 2 f). For LNC, we found that
207 species that produce leaves with high nitrogen content, which relates generally to high photosynthetic
208 rates and acquisitive growth, show smaller responses to cues (Fig. 2 j-l). These findings are in line
209 with our predictions that high LNC species (acquisitive) would be less responsive to chilling (0.7 mg/g
210 per standardized chilling; 90% UI: 0.3, 1.2, Table S5), but we also found high LNC species to be less
211 responsive to photoperiod (0.3 mg/g per standardized photoperiod; 90% UI: 0, 0.6) and to forcing (0.5
212 mg/g per standardized forcing; 90% UI: 0.1, 0.9, Fig 2 j-l & Fig S3 d-f).

213
214 Across our models, we found species-level variation across traits (σ_{trait}^2) were comparable to or greater
215 than variation across studies (σ_{study}^2 , Fig 4). The magnitude of study-level variation (σ_{study}^2) that we
216 found, however, suggests that models using large trait databases that fail to separate out study from
217 species-level variation (σ_{trait}^2) may poorly estimate species traits. Variation across studies was greatest
218 for height (with σ_{study}^2 of 7.5 m compared to 5.9 m for σ_{trait}^2 , Fig 4a). For seed mass and LNC study-
219 level variation was less than that of the species-level variation, with estimates of 1 mg for study-level
220 variation versus 1.6 for species-level variation in seed mass and estimates of 3.6 mg g⁻¹ for study-level
221 variation and 5.1 mg g⁻¹ for the species-level variation in LNC (Fig 4c and d). At the lowest end,
222 study-level variation in SLA was approximately half the value of the species-level variation (3.3 mm²
223 mg⁻¹ versus 7.8 mm² mg⁻¹ for σ_{study}^2 and σ_{trait}^2 , respectively, Fig 4b).

224

225 **Discussion**

226 We found species traits influenced the timing of budburst in response to the three primary cues of
227 spring phenology: chilling, forcing and photoperiod. These trait effects were associated with earlier
228 or later phenology following well-established gradients in growth strategies predicted by functional
229 trait frameworks^{5,13,47,48,53}: early species tended to have traits associated with fast and acquisitive
230 strategies while later species had traits associated with conservative, slower strategies. We found the
231 largest budburst responses occurred for traits related to resource acquisition and structure, with SLA,
232 LNC, and height all showing large responses across our three cues. In contrast, our one reproductive
233 trait—seed mass—showed a smaller response. Our results provide a major step forward in integrating
234 phenology into broader trait syndromes that shape species growth strategies, and support previous
235 findings from more local scales that found strong relationships between height and species phenol-
236 ogy^{39,40,42}. Our findings also suggest other traits—such as seed mass—show no relationship with
237 phenology in our more global analysis.

238

239 Effects of phenology-trait relationships on community assembly

240 Our findings suggest the changing pressures across the early growing season may affect the temporal
241 assembly of communities. Strong abiotic pressures alongside weak competition early in the season were
242 associated with early-budbursting species with acquisitive traits (shorter heights and low LNC) that
243 allow faster return on resource investments^{5,17,47}. These traits should allow early species to more easily
244 replace tissue if lost to frost or other abiotic disturbances, and benefit from greater light availability in
245 the open canopy of many temperate forests in the early spring. In contrast, later-budbursting species
246 had traits associated with greater competitive abilities and slower growth^{5,17,47} (Fig 2), which may
247 help them compete for soil and light resources when most other species are already growing. These
248 traits can be linked to other ecological processes and species characteristics, such as species succes-
249 sional position, as illustrated by the differences between early and late successional species (e.g., *Alnus*
250 *incana* and *Quercus rubra*; Fig 2).

251 The traits with cue responses that deviated from our expectations also offer novel insights into the
252 tradeoffs between traits and environmental cues. All of our traits are associated with numerous aspects
253 of species growth, and may be adaptive for reasons other than those we predicted. Contrary to our
254 predictions, we found large responses to forcing for short trees, which could prevent frost damage or
255 xylem cavitation under a late spring frost^{10,30} and influence annual cambial meristem growth²⁶. Simi-
256 larly, the lack of a response to chilling or forcing by high SLA individuals could be driven by other trait
257 attributes and environmental cues—selecting for species relative growth rates or leaf longevity—and
258 not photosynthetic potential^{35,47}. These findings highlight the complexity of determining the drivers
259 of species trait profiles, and further our understanding of how traits affect community dynamics under
260 variable environments.

262

263 Phenology-trait relationships under future climates

264 Incorporating phenology within broader trait syndromes could aid forecasting of species and commu-
265 nity responses to climate change. While decades of research have documented phenological shifts with
266 anthropogenic climate change, increasing research suggests a potential connection between phenolog-
267 ical responses to warming and performance with warming, where species that shift their phenology
268 more also perform better^{8,28}.

269

270 Our results suggest this phenology-performance relationship could be driven in part by a suite of traits
271 that covary with phenological cues to determine how responsive species are to warming. Species with
272 smaller responses to all cues, especially chilling and photoperiod, would tend to advance more with
273 warming, which our results suggest would allow these species to also grow more quickly. These results
274 could further aid in predicting the potential for invasion, as communities with similar phenologies
275 and suites of traits, appear more susceptible to fast growing, phenologically more responsive invasive
276 species^{1,38,49}.

277

278 Our analytical approach and results may be especially useful to help forecast changes in forest dy-
279 namics. Identifying the varying trait syndromes of forest communities over a spring season can aid
280 predictions of how climate change will alter species growth and productivity. For example, our results
281 suggest that, by favoring more phenologically responsive species (i.e., with small chilling and pho-
282 toperiod responses), warming may also favor species with acquisitive growth strategies. In contrast,
283 conservative species, which appear less phenologically responsive to changes in temperature (due to
284 larger chilling and photoperiod responses) could face greater abiotic and biotic stress¹⁹.

285

286 Our results could further help identify which species are most likely to be negatively impacted under
287 future climates, and develop better strategies for climate change mitigation and conservation. Species

288 that fail to advance phenologically with warming might experience more competition^{1,4}, as species that
289 begin growth increasingly earlier with warming have more time to deplete resources. In addition to
290 altering the timing and interactions between species within a season, species trait syndromes have the
291 potential to further redefine the environmental conditions under which growth occurs, and as a result,
292 shape community assembly and productivity within ecological communities. By identifying the species
293 most vulnerable to climate change impacts, we can develop more effective management practices that
294 prevent the loss of critical ecosystem services and preserve community diversity under future conditions.

295

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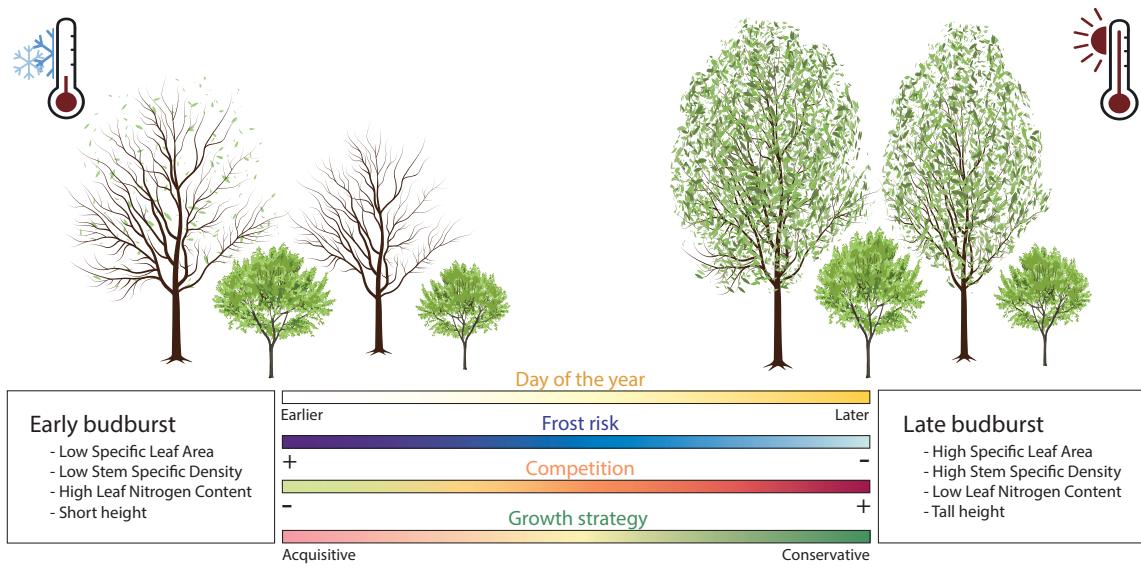


Figure 1: Leaf and wood traits follow a gradient that varies from acquisitive versus conservative growth strategies, which may also include phenology. We predicted early-budbursting species would have traits associated with acquisitive growth, as they are more likely to experience greater risk of frost but reduced competition. In contrast, we expected later-budbursting species would have traits associated with conservative growth, as they experience greater competition but a more climatically benign environment.

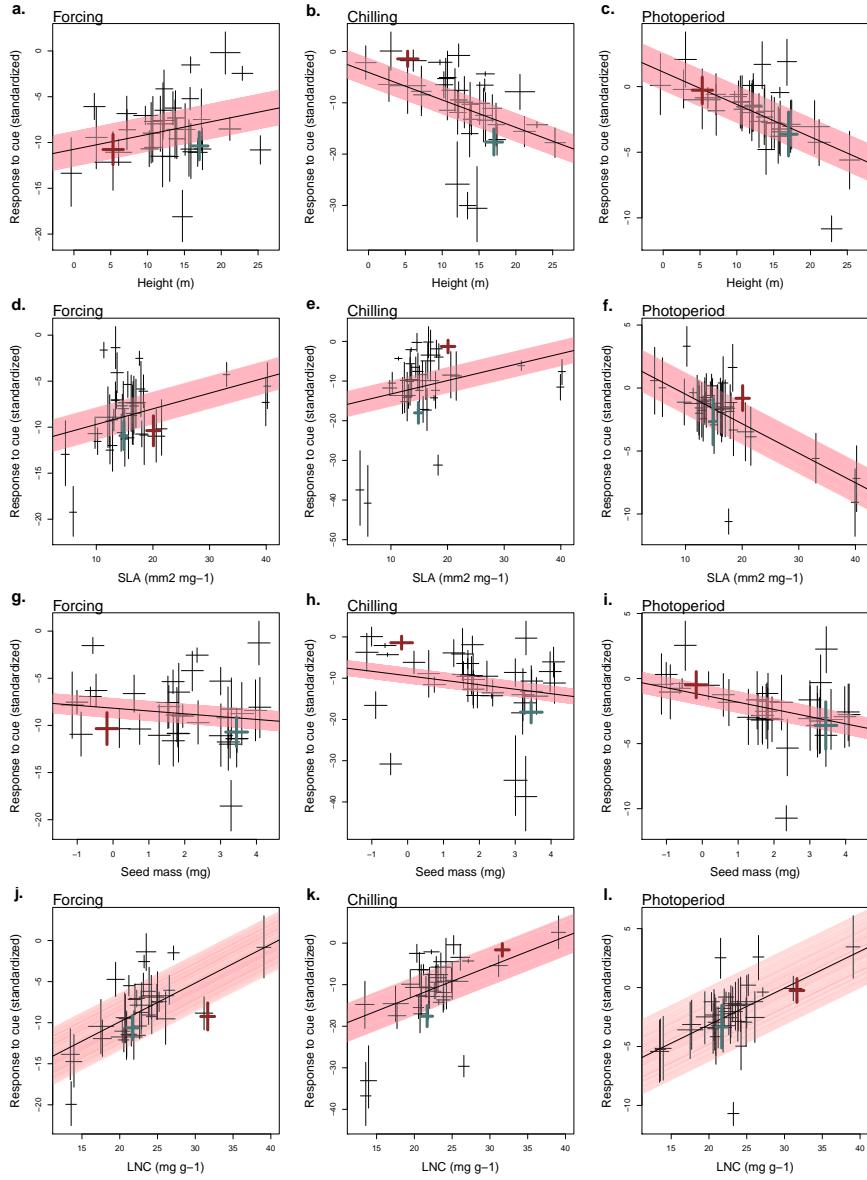


Figure 2: Estimated species-level responses to standardized forcing (left column: a, d, g & j), chilling (middle column: b, e, h & k), and photoperiod treatments (right column: c, f, i & l) predicted by estimated trait values for height (first row: a-c), SLA (second row: d-f), log10 Seed mass (third row: g-i), and LNC (fourth row: j-l). We estimated parameters using a joint trait-phenology model, with the black line depicting the mean linear relationship between estimated trait effects and the slope of the cue response and the pink band the 50% uncertainty interval. Each set of crossed lines represents one species (with each line spanning the 50% uncertainty interval), with the species depicted in Fig 3 colored in each panel, with the acquisitive species (*Alnus incana*) shown in red, and the conservative species (*Quercus rubra*) shown in blue.

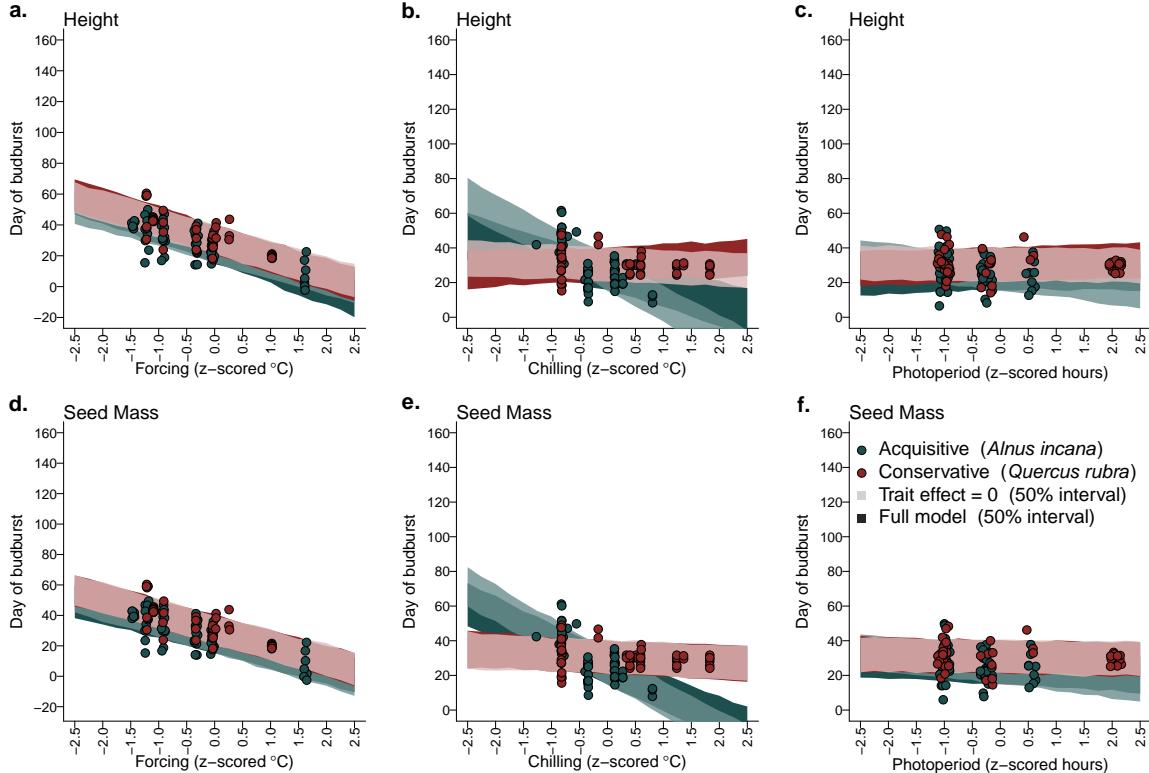


Figure 3: We expected species with traits associated with acquisitive (e.g., smaller heights and small seed mass) versus conservative (e.g., taller with larger seeds) growth strategies would have different budburst responses to phenological cues. Shown here is an example of the cue relationships with height (a-c) and seed mass (d-f) for an acquisitive species, *Alnus incana* shown in red, and a conservative species, *Quercus rubra*, shown in blue. **a**, The effect of height on budburst timing was smaller in response to forcing cues, but larger in response to both **b**, chilling and **c**, photoperiod. In contrast, seed mass had a negligible effect on **d**, forcing and **f**, photoperiod responses, **e**, but a greater response to chilling. Band represent the 50% uncertainty intervals of the model estimates. The coloured bands represent the 50% uncertainty intervals of the model estimates and points individual trait measurements.

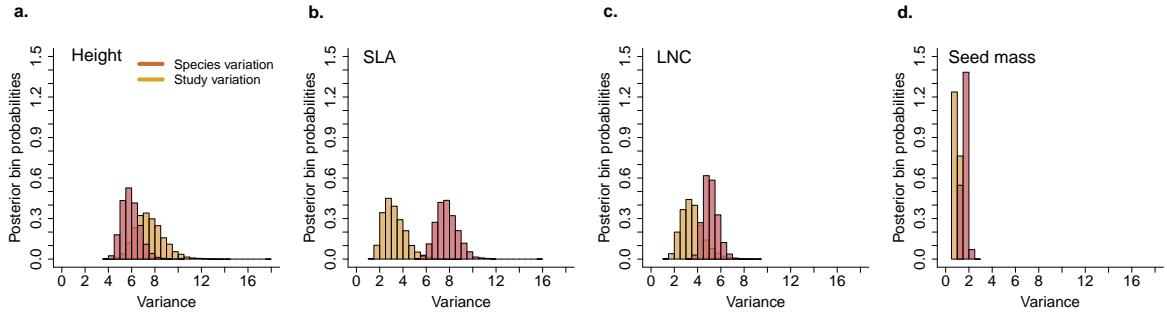


Figure 4: Traits differed in the relative magnitude of their species-level and study-level variation, with only **a**, the height model estimating greater study-level variation than species-level, while our **b**, specific leaf area, **c**, leaf nitrogen content, and **d**, seed mass models all estimated higher species-level variation. Shown here are the posterior densities for the species-level variation, shown in red, and study-level variation, shown in yellow. For comparison, we show all x-axes spanning the minimum to maximum variances across all four traits. The histograms depict the full distribution of the study and species-level variance, where each bin is normalized by the total count of the posterior estimate.