Childhood exposure to non-persistent endocrine disruptors, glucocorticosteroids, and attentional function: A cross-sectional study based on the parametric g-formula

# List of Acronyms

ADHD

Attention-Deficit / Hyperactivity Disorder

ANT

Attention Network Test

BPA

Bisphenol A

CI

Confidence interval

DAG

Directed acyclic graph

DAP

Dialkylphosphate

EDC

Endocrine disrupting chemical

HELIX

Human Early-Life Exposome

HPA

Hypothalamic-pituitary-adrenocortical

HRT-SE

Hit reaction time standard error

IQ

Intelligence quotient

IQR

Interquartile range

**LC**

Liquid chromatography

LLOQ

Lower limit of quantification

LOD

Limit of detection

LOQ

Limit of quantification

MC

Marginal contrast

**MS**

Mass spectrometry

NPSEM

Non-parametric structural equation model

OP pesticide

Organophosphate pesticide

**TOFMS**

Time-of-flight mass spectrometry

**UHPLC**

[Ultra-high performance liquid chromatography](https://www.sciencedirect.com/science/article/pii/B9780128053935000294)

# Abstract

**Background**

Evidence suggests that endocrine disrupting chemicals (EDCs) may perturb the hypothalamic-pituitary-adrenocortical (HPA) axis, which has a major role in brain development. We aimed to evaluate the effects of childhood exposure to organophosphate pesticides, phenols, and phthalate metabolites, on urinary glucocorticosteroids and inattention in childhood.

**Methods**

We used data from the Human Early-Life Exposome (HELIX) cohort (2013-2016) and the parametric g-formula to estimate associations between EDCs, glucocorticosteroids, and hit reaction time standard error (HRT-SE), a measure of inattention, and tested for possible effect modification by sex.

**Results**

We observed a positive marginal contrast (MC) for exposure increases from the 10th to the 90th percentile for methyl-paraben (MC: 0.042 and confidence interval (CI): (0.013, 0.071)), and the phthalate metabolites oxo-MiNP (MC: 0.023 and CI: (0.003, 0.044)), oh-MiNP (MC: 0.039 and CI: (0.001, 0.076)), and MEHP (MC: 0.036 and CI: (0.008, 0.063)), on HRT-SE, indicating lower attention. Several EDCs were also associated with a positive MC for cortisone, cortisol, and corticosterone production. Increased levels of the glucocorticosteroids were not associated with HRT-SE, although we found a possible effect modification by sex.

**Conclusions**

Our results suggest that multiple EDCs might interfere with inattention and with the homeostasis of the HPA axis.

**Keywords:** Endocrine Disruptors, Cortisone, Hydrocortisone, Neuropsychological Tests, Causality.

# 1. Introduction

The prevalence of several neurodevelopmental disorders has increased in the pediatric population ([Grandjean and Landrigan, 2014](#ref-GrandjeanLandrigan:2014)), and multiple environmental pollutants may play a role in the increased rates of these disorders ([Ramírez et al., 2022](#Xd81cf38a3b251ec377f8aa13b097ea9d0c190e3)). Multiple endocrine disrupting chemicals (EDCs), ubiquitous chemicals present in many every-day products and diet, are capable of interfering with the endocrine system, and have shown associations with childhood neurodevelopment and behavior ([Bouchard et al., 2010](#ref-BouchardBellingerWright:2010); [Braun, 2017](#ref-Braun:2017); [Cartier et al., 2016](#Xa0cc0fdd8206e54b0a9b4bdcd9d2102efd24cb7); [Furlong et al., 2017](#ref-FurlongHerringBuckley:2017); [González-Alzaga et al., 2015](#X2a5c5c17453fffe785b7f0bcffd564283892c6c); [Huang et al., 2015](#ref-HuangChenSu:2015); [Huang et al., 2017](#ref-HuangTsaiChen:2017); [Kim et al., 2017](#ref-KimHongShin:2017); [Li et al., 2018](#ref-LiZhangKuang:2018); [Oh et al., 2023](#ref-OhKimKannan:2023); [Rodríguez-Carrillo et al., 2019](#X71fca68020aad3c79232160399444b9923ab76e); [Shoaff et al., 2020](#ref-ShoaffCoullWeuve:2020); [Tewar et al., 2016](#ref-TewarAuingerBraun:2016); [Vilmand et al., 2023](#ref-VilmandBeckBilenberg:2023); [Yu et al., 2016](#ref-YuDuChiou:2016)). Phthalates, phenols, and organophosphate pesticides (OP pesticides) are non-persistent EDCs that have raised concerns about potential risks to human health (Braun, 2017). Similarly, the few studies assessing exposure to OP pesticides during childhood and through the use of biomarkers suffered from a series of limitations, including a small sample size ([Ramírez et al., 2022](#Xd81cf38a3b251ec377f8aa13b097ea9d0c190e3)).

Moreover, little is known about the biological mechanisms of action ([Ramírez et al., 2022](#Xd81cf38a3b251ec377f8aa13b097ea9d0c190e3)). There is some toxicological evidence, however, that exposure to certain EDCs, specifically phthalates, might interfere with the hypothalamic-pituitary-adrenocortical (HPA) axis, which is responsible for the production of glucocorticosteroids, and might interact with the glucocorticoid receptor ([Kim et al., 2018](#ref-KimLeeMoon:2018); [Sears et al., 2023](#ref-SearsLiuLanphear:2023); [Sun et al., 2018](#ref-SunLiJin:2018)). This potential mechanism is corroborated by the presence of receptors for these hormones in the brain ([Lupien et al., 2009](#ref-LupienMcEwenGunnar:2009); [Sun et al., 2018](#ref-SunLiJin:2018)). While glucocorticosteroids are necessary for brain maturation, their under- or over-production might interfere with its normal development and ultimately lead to long-term impaired functioning ([Lupien et al., 2009](#ref-LupienMcEwenGunnar:2009); [Sears et al., 2023](#ref-SearsLiuLanphear:2023)).

Taken together, these results suggest that disruption of the HPA axis’ homeostasis may be one biological mechanism by which EDCs may affect neurodevelopment.

While most of the available literature is focused on the effects of prenatal exposure to EDCs on neurodevelopment in infancy and childhood ([Ramírez et al., 2022](#Xd81cf38a3b251ec377f8aa13b097ea9d0c190e3)), mid-childhood is also a crucial stage of (neuro)development (Mah and Ford-Jones, 2012) and further studies are necessary to understand the impact of exposure to EDCs in this life stage. Our goal in this study was twofold. First, we examined the impact of childhood exposure to non-persistent EDCs on glucocorticosteroids and attentional function in childhood. Second, we examined the impact of changes in glucocorticosteroid levels on attentional function in childhood. To do so, we used the parametric g-formula, a causal inference technique, and marginal contrasts (MCs) to estimate cross-sectional associations between non-persistent EDCs, glucocorticosteroids, and attentional function in children of a multi-country cohort in Europe.

# 2. Material and methods

## 2.1 Study population and design

The Human Early-Life Exposome (HELIX) project aims to characterize early-life exposures and their potential association with endogenous biomarkers and health outcomes ([Vrijheid et al., 2014](#ref-VrijheidSlamaRobinson:2014)). It consists of six existing population-based birth cohort studies across Europe: BiB (Born in Bradford, UK) ([Wright et al., 2013](#ref-WrightSmallRaynor:2013)), EDEN (Study of determinants of pre- and postnatal developmental, France) ([Heude et al., 2016](#ref-HeudeForhanSlama:2016)), INMA (Environment and Childhood, Spain) ([Guxens et al., 2012](#ref-GuxensBallesterEspada:2012)), KANC (Kaunas Cohort, Lithuania) ([Grazuleviciene et al., 2009](#Xd30c40380c9e99bac70b7fa3b0ada5ae8dec3e4)), MoBa (The Norwegian Mother and Child Cohort Study, Norway) ([Magnus et al., 2006](#ref-MagnusIrgensHaug:2006)), and Rhea (Mother–Child Cohort in Crete, Greece) ([Chatzi et al., 2009](#ref-ChatziPlanaDaraki:2009)). The HELIX subcohort of 1,301 mother-child pairs was fully characterized for the external and internal exposome, including exposure and omics biomarkers during childhood ([Maitre et al., 2018](#ref-MaitreBontCasas:2018)). Eligibility criteria for inclusion in the HELIX subcohort included: a) age 6-11 years, with a preference for 7-9 years; b) availability of sufficient stored pregnancy blood and urine samples; c) availability of complete address history from first to last follow-up; d) no serious health problems, which might affect the results of the clinical testing or the volunteer’s safety (e.g., acute respiratory infection). Ethical permission was obtained from the relevant authorities in the corresponding country.

## 2.2 Variables

### 2.2.1 Endocrine disrupting chemicals

Children were assessed between December 2013 and February 2016, and assessments included neurological testing and urine collection. Urine samples of the night before and the first morning void on the day of the visit were combined to provide a more reliable exposure assessment. Non-persistent EDCs assessed in the urine samples included phthalate metabolites, phenols, and organophosphate (OP) pesticide metabolites. A list of the environmental chemicals determined in urine samples and used for the present study is given in [Table S1](#supptbl-info-chems). Briefly, we analyzed a total of 7 phenols (bisphenol A (BPA), ethyl-paraben (ETPA), methyl-paraben (MEPA), n‑butyl‑paraben (BUPA), oxybenzone (OXBE), propyl-paraben (PRPA), triclosan (TRCS)), 6 non-specific organophosphate pesticide metabolites (diethyl dithiophosphate (DEDTP), diethyl phosphate (DEP), diethyl thiophosphate (DETP), dimethyl dithiophosphate (DMDTP), dimethyl phosphate (DMP), dimethyl thiophosphate (DMTP)), and 10 phthalate metabolites (mono benzyl phthalate (MBzP), monoethyl phthalate (MEP), mono‑2‑ethyl 5‑carboxypentyl phthalate (MECPP), mono‑2‑ethylhexyl phthalate (MEHP), mono‑2‑ethyl‑5‑hydroxyhexyl phthalate (MEHHP), mono‑2‑ethyl‑5‑oxohexyl phthalate (MEOHP), mono‑4‑methyl‑7‑hydroxyoctyl phthalate (oh-MiNP), mono‑4‑methyl‑7‑oxooctyl phthalate (oxo-MiNP), mono‑iso‑butyl phthalate (MiBP), mono‑n‑butyl phthalate (MnBP)) originating from 6 distinct phthalate parent compounds ([Table S1](#supptbl-info-chems)). Laboratory protocols for the analysis are described in Haug et al. ([2018](#ref-HaugSakhiCequier:2018)), while quality control and interlaboratory comparison are described in Maitre et al. ([2018](#ref-MaitreBontCasas:2018)). Briefly, the urine samples were collected in high-quality polypropylene tubes during the visit, and the aliquots were stored at -80°C. Concentrations of the phthalate metabolites, phenols, and OP pesticide metabolites were determined using online column-switching LC-MS/MS, online column-switching UHPLC-MS/MS, and UHPLC-TOFMS, respectively. Procedure blanks and internal quality control samples were analyzed along with each batch of samples.

### 2.2.2 Glucocorticosteroids

Urine samples of the night before the day of the visit were used to measure levels of the glucocorticosteroids. These included glucocorticosteroids, glucocorticosteroid metabolites, glucocorticosteroid precursors, glucocorticosteroid precursor metabolites, androgens, and androgen metabolites. A list of the glucocorticosteroids determined in urine samples and used for the present study is given in [Table S2](#supptbl-info-mets).

To assess the levels of glucocorticosteroids and their metabolites, LC-MS/MS analysis was applied at the Applied Metabolomics Research Group, IMIM (Hospital del Mar Medical Research Institute). The laboratory protocols for the analysis are described elsewhere ([Gomez-Gomez and Pozo, 2020](#ref-Gomez-GomezPozo:2020); [Marcos et al., 2014](#ref-MarcosRenauCasals:2014)).

Three additional markers, total cortisol production, total cortisone production, and total corticosterone production, were computed based on the following: cortisol production as the sum of cortisol and its metabolites (20α-dihydrocortisol (20aDHF), 20β-dihydrocortisol (20bDHF), 5α,20α-cortol (5a20acortol), 5α,20β-cortol (5a20bcortol), 5α-tetrahydrocortisol (5aTHF), 5β,20α-cortol (5b20acortol), 5β,20β-cortol (5b20bcortol), 5β-dihydrocortisol (5bDHF), 5β-tetrahydrocortisol (5bTHF), 6β-hydroxycortisol (6OHF)), cortisone production as the sum of cortisone and its metabolites (20α-dihydrocortisone (20aDHE), 20β-dihydrocortisone (20bDHE), 5α-tetrahydrocortisone (5aTHE), 5β,20α-cortolone (5b20acortolone), 5β,20β-cortolone (5b20bcortolone), 5β-tetrahydrocortisone (5bTHE), 6β-hydroxycortisone (6OHE)), and corticosterone production as the sum of 11-dehydrocorticosterone (A), 17-deoxycortolone (17-DO-cortolone), 5α-tetrahydrocorticosterone (5aTHB), 5β-tetrahydrocorticosterone (5bTHB).

### 2.2.3 Attentional function

Cognitive and motor function outcomes were assessed during the visit with standardized, non-linguistic, and culturally blind computer tests, including the Attention Network Test (ANT) ([Rueda et al., 2004](#ref-RuedaFanMcCandliss:2004)), which provides a measure of efficiency of attentional function. The tests were administered in a standardized way, and with minimal interference from the field workers. Further information can be found elsewhere ([Forns et al., 2014](#ref-FornsEsnaolaLopez-Vicente:2014); [Maitre et al., 2018](#ref-MaitreBontCasas:2018); [Rueda et al., 2004](#ref-RuedaFanMcCandliss:2004)). The outcome of interest for the present study is the hit reaction time standard error (HRT-SE) ([Sunyer et al., 2015](#ref-SunyerEsnaolaAlvarez-Pedrerol:2015)), a measure of response speed consistency throughout the test. A high HRT-SE indicates highly variable reaction times and is considered a measure of inattentiveness.

### 2.2.4 Confounders

For each research question, defined by a specific type of exposure and outcome, the minimal set of covariates for inclusion in the analyses was selected on the basis of a directed acyclic graph (DAG) (Greenland et al., 1999) built with DAGitty ([Textor et al., 2016](#ref-TextorvanderZanderGilthorpe:2016)) and ggdag ([Barrett, 2023](#ref-Barrett:2023)). The sets of covariates were selected to estimate the total effect of the exposure on the outcome. For effect estimation of the EDCs on glucocorticosteroids and of glucocorticosteroids on HRT-SE, these sets were also sufficient to estimate direct effects. Sample-specific creatinine values were included in the models to adjust for possible dilution effects. Further, each minimal adjustment set was *augmented* with precision covariates, defined as the set of parent variables of the outcome that are not parents of the exposure. Common confounders were cohort, ethnicity, sex, age, height, weight, and head circumference of the child, urine creatinine , consumption of fish, fruit, vegetables, organic food, and fast food with a food frequency questionnaire, maternal tobacco consumption, family financial situation and affluence scale (FAS). Models for estimating the effects of EDCs on HRT-SE were further adjusted for child breastfeeding, prenatal maternal active and passive smoking, child mood and rest before assessment, child neuropsychological diagnosis, maternal marital status, season, and fasting time before assessment. Models for estimating the effects of EDCs on glucocorticosteroids were further adjusted for season, and fasting time before assessment. Models for estimating the effects of glucocorticosteroids on HRT-SE were further adjusted for child breastfeeding, prenatal maternal active and passive smoking, marital status, EDCs, child mood and rest before assessment, and child neuropsychological diagnosis. The adjustment sets are provided in the Supplementary Material as text files compatible with DAGitty. A simplified DAG is provided in [Figure 1](#fig-dag). Codebooks for the used covariates are provided in [Table S3](#supptbl-codebooks).

## 2.3 Statistical methods

### 2.3.1 Data pre-processing

Concentrations of the glucocorticosteroids were classified as quantifiable, below the limit of quantification (LOQ), possible interference or out of range, and not detected. For each metabolite, we computed the fraction of values below the LOQ and not detected, both within each cohort and overall. We proceeded to impute these values using half the value of the corresponding LOQ, for those metabolites that had less than 30% of non-detected within each cohort and 20% of non-detected overall. Information about the lower limit of quantification (LLOQ) for the glucocorticosteroids is provided in [Table S4](#supptbl-lloq-mets). Missing values were imputed using kNN from the VIM R package ([Kowarik and Templ, 2016](#ref-KowarikTempl:2016)) for those metabolites that had less than 40% of missings within each cohort and 30% of missings overall. For each metabolite, imputation was based on the median of the 5 nearest neighbors and the distance metric was computed based on all remaining metabolites. We natural log-transformed them to improve model fit, assessed with posterior predictive checks. To do so, replicated data were simulated with the fitted models and compared to the observed data. We used the check\_predictions function from the performance R package using the default arguments ([Lüdecke et al., 2021](#ref-LudeckeBen-ShacharPatil:2021)). Values of total cortisol, cortisone, and corticosterone production were expressed in nanograms per millilitre (ng/ml).

Concentrations of the non-persistent EDCs were classified as quantifiable, below the limit of detection (LOD), possible interference or out of range, and not analysed. Concentrations of each chemical below the LOD were singly imputed using a quantile regression approach for the imputation of left-censored missing data, as implemented in the impute.QRILC function from the imputeLCMD R package ([Lazar, 2015](#ref-lazar2015imputelcmd)). Information about the lower limits of detection can be found in Haug et al. ([2018](#ref-HaugSakhiCequier:2018)). Chemicals with more than 70% of observations below the LOD were excluded from the present study. Missing values were imputed similarly using kNN. Values of the chemicals were expressed in grams per litre (g/L).

Missing values in the clinical outcome were imputed similarly using kNN. We natural log-transformed these to improve model fit, assessed with posterior predictive checks. Values of the clinical outcome were expressed in milliseconds (ms).

Missing values in the covariates were imputed similarly using kNN. Categorical covariates were imputed using the maxCat function, which chooses the level with the most occurrences. Creatinine values were expressed in grams per litre (g/L).

### 2.3.2 Estimation of balancing weights

To reduce the effect of measured confounders on the exposure-outcome association, stabilized balancing weights were estimated using the energy method available in the WeightIt R package ([Greifer, 2023a](#ref-Greifer:2023)). This method estimates weights by minimizing an energy statistic related to covariate balance ([Huling et al., 2023](#ref-HulingGreiferChen:2023)), thus avoiding the need to specify a parametric model. Weights below the 0.1 and above the 0.9 quantiles were trimmed. Trimming might lead to decreased covariate balance and potentially change the estimand, but can also decrease the variability of the weights. Covariate balance was assessed using functionalities provided by the cobalt R package ([Greifer, 2023b](#ref-Greifer:2023a)). Specifically, we used *Love* plots to visualize covariate balance before and after weighting.

### 2.3.3 G-computation

We estimated MCs with the parametric g-formula, a method of standardization. The parametric g-formula involves the following steps: 1) fit a outcome model including both exposure and covariates, and balancing weights (Smith et al., 2022); 2) create two counterfactual datasets identical to the original one but with the exposures shifted according to a user-specified dynamic intervention set by a deterministic function of the observed exposure levels; 3) use the outcome model to compute adjusted predictions in the two counterfactual datasets; 4) compute the difference between the means of the adjusted predictions in the counterfactual datasets. The causal parameter of interest was thus specified as the difference between the expected value of the outcomes in the counterfactual datasets under the shifted exposure levels and specified below: . For this parameter to be identified, the usual causal identifiability conditions (no unmeasured confounding, positivity, and consistency) are required. Since these conditions are likely not satisfied, we focused on the estimation of a statistical estimand that is as close as possible to the causal parameter of interest.

We fit the outcome model using the glm function and a Gaussian family with identity link from base R. The exposure variable was modeled using natural cubic splines with 3 degrees of freedom, to more flexibly capture the average dose-response function (ADRF).

To estimate the MCs, we used the avg\_comparisons function from the marginaleffects R package ([Arel-Bundock, 2023](#ref-Arel-Bundock:2023)). The two counterfactual datasets were obtained by setting the exposures levels to 90th percentile () and the 10th percentile (), for each cohort separately to account for cross-cohort differences in the levels of the exposures ([Table S6](#supptbl-chems-desc) and [Table S7](#supptbl-met-desc)). The MCs were computed using the estimated balancing weights above. Robust standard errors were computed with the sandwich R package, using cohort as variable indicating clustering of observations ([Zeileis et al., 2020](#ref-ZeileisKollGraham:2020); [Zeileis, 2004](#ref-Zeileis:2004)). For each outcome, we report the results as differences between MCs with 95% confidence intervals (CIs).

### 2.3.4 Effect-modification analysis

We further estimated separate MCs for possible effect-modification by sex. To do so, balancing weights were estimated separately for each level of the sex variable, and an interaction term between the exposure and sex was included in the outcome model. Similarly, the MCs were estimated with the avg\_comparisons function. For each outcome, we report the results as pairwise differences between female MCs and male MCs, with 95% confidence intervals (CIs).

# 3. Results

[Table 1](#tbl-pop-desc) and [Table S5](#supptbl-pop-desc) provide descriptive statistics for the outcome and covariates for the HELIX subcohort and for each cohort, respectively. Of the 1,301 children of the HELIX subcohort, 1,297 had measurements of the non-persistent EDCs. Measurements of the glucocorticosteroids were available for 1,004 children, of which 980 were matched to the HELIX subcohort. Measurements of both non-persistent EDCs and glucocorticosteroids were available for 976 children of the subcohort. A flowchart describing the sample size for each research question is presented in [Figure S1](#suppfig-flow-pop). The sample consisted of 55% males. The median HRT-SE was 300 ms (interquartile range (IQR), 231-368), with lower median values for EDEN, MOBA, and INMA, corresponding to the cohorts with older children. At the time of visit, the median age of the children was 8.06 years. The children were mostly Caucasian (90%), and the largest minority were of Pakistani origin (6.2%).

Levels of unprocessed non-persistent EDCs, after imputation of values below the LOD, and glucocorticosteroids, are presented in [Table 2](#tbl-edc-desc), [Table 3](#tbl-met-new-desc), [Table S6](#supptbl-chems-desc), and [Table S7](#supptbl-met-desc). Supplementary Figures [2](#suppfig-desc-chems) and [3](#suppfig-desc-mets) provide information on the measurement classification of the EDCs and glucocorticosteroids by cohort, respectively.

The effective sample sizes before and after balancing weights estimation are presented in Supplementary Tables [8](#supptbl-balance-1), [9](#supptbl-balance-2), [10](#supptbl-balance-3), while basic summary statistics of the estimated balancing weights are presented in Supplementary Tables [11](#supptbl-weights-1), [12](#supptbl-weights-2), [13](#supptbl-weights-3). As expected, the median value of the weights for each exposure was close to . The Love plots to assess balance are available in Supplementary Material – Love plots.

[Figure 2](#fig-marginal-1) presents the forest plot for the MCs on the logarithmic scale of the non-persistent EDCs on HRT-SE. For most EDCs, a cohort-specific increase in the levels of the exposures from the 10th to the 90th percentiles was associated with a positive MC, indicating an increase in the values of HRT-SE and thus lower attention. Most of the CIs included the null effect, though. Noteworthy associations were observed for the paraben MEPA (MC: 0.042 and CI: (0.013, 0.071)), and the phthalate metabolites oxo-MiNP (MC: 0.023 and CI: (0.003, 0.044)), oh-MiNP (MC: 0.039 and CI: (0.001, 0.076)), and MEHP (MC: 0.036 and CI: (0.008, 0.063)). The organophosphate pesticide (OP pesticide) DETP was associated with lower HRT-SE (MC: -0.026 and CI: (-0.054, 0.001)).

[Figure 3](#fig-marginal-2) presents the forest plot for the MCs on the logarithmic scale of the non-persistent EDCs on total cortisone, cortisol, and corticosterone production. For most EDCs, a cohort-specific increase in the levels of the exposures from the 10th to the 90th percentiles was associated with a positive MC, indicating an increase in the total production of these metabolites. Exceptions were BUPA, which was associated with negative MCs for all three outcomes, and MiBP, which was associated with a negative MC for total cortisone production only. Most of the effects for the phenols and phthalate metabolites included the null. The phenol BPA showed the largest MCs across all three outcomes (cortisone production, MC: 0.263 and CI: (0.131, 0.394); cortisol production, MC: 0.274 and CI: (0.107, 0.441); corticosterone production, MC: 0.285 and CI: (0.106, 0.464)).

[Figure 4](#fig-marginal-3) presents the forest plot for the MCs on the logarithmic scale of the glucocorticosteroids on HRT-SE. All MCs included the null, with no clear indication of directionality of the effect.

## 3.1 Effect modification by sex

Basic summary statistics of the estimated balancing weights for effect modification are presented in Supplementary Tables [14](#supptbl-weights-1sa), [15](#supptbl-weights-2sa), [16](#supptbl-weights-3sa). As expected, the median value of the weights for each exposure was close to . The Love plots to assess balance are available in Supplementary Material – Love plots.

[Table 4](#tbl-hypothesis-1and2) presents the results of the difference between estimates of the MCs on the logarithmic scale for females and males, for the EDCs on the glucocorticosteroids and HRT-SE. For HRT-SE, significant differences were present for the phenol OXBE (MC: 0.032 and CI: (0.004, 0.061)) and the phthalate metabolite MbZP (MC: -0.066 and CI: (-0.126, -0.007)). For the glucocorticosteroids, significant differences were present across all three classes of EDCs and for all outcomes. The largest differences were attributable to the phenol ETPA (corticosterone production, MC: -0.254 and CI: (-0.416, -0.092)) and the phthalate metabolite MEHP (cortisol production, (MC: -0.221 and CI: (-0.289, -0.153)); cortisone production, (MC: -0.177 and CI: (-0.299, -0.055))). The individual MCs are presented in Supplementary Tables [17](#supptbl-marginal-1sa) and [18](#supptbl-marginal-2sa), and in Supplementary Figures [4](#suppfig-marginal-1sa) and [5](#suppfig-marginal-2sa).

[Table 5](#tbl-hypothesis-3) presents the results of the difference between estimates of the MCs on the logarithmic scale for females and males, for the glucocorticosteroids on HRT-SE. Significant differences were present for cortisone production (MC: 0.14 and CI: (0.019, 0.261)) and corticosterone production (MC: 0.126 and CI: (0.009, 0.243)). Furthermore, for all exposures, the MCs had opposite sign (positive for males and negative for females). The individual MCs are presented in [Figure 4](#fig-marginal-3).

# 4. Discussion

In this study, consisting of 1,297 children from 6 European birth cohorts, we observed that short-term childhood exposure to certain non-persistent EDCs was associated with lower attentional function (MEPA, MEHP, oh-MiNP, and oxo-MiNP), and with an increase in total production of cortisol, cortisone, and corticosterone (DEP, DMP, DMTP, BPA, ETPA, MEPA, MEHP, oh-MiNP, and oxo-MiNP). Some of these associations differed for females and males. Increased production of these glucocorticosteroids was not associated with attentional function in the study population overall, but showed associations with lower attentional function in males and higher attentional function in females.

To the best of our knowledge, no other study has investigated the effects of childhood exposure to multiple classes of non-persistent EDCs in relation to attentional function. More generally, the literature on childhood exposure to non-persistent EDCs and other neurodevelopment outcomes in children has mostly focused on OP pesticides ([Bouchard et al., 2010](#ref-BouchardBellingerWright:2010); [Cartier et al., 2016](#Xa0cc0fdd8206e54b0a9b4bdcd9d2102efd24cb7); [González-Alzaga et al., 2015](#X2a5c5c17453fffe785b7f0bcffd564283892c6c); [Yu et al., 2016](#ref-YuDuChiou:2016)), phthalate metabolites ([Balalian et al., 2019](#ref-BalalianWhyattLiu:2019); [Huang et al., 2015](#ref-HuangChenSu:2015); [Huang et al., 2017](#ref-HuangTsaiChen:2017); [Jankowska et al., 2019](#ref-JankowskaPolanskaHanke:2019); [Kim et al., 2017](#ref-KimHongShin:2017); [Li et al., 2019](#ref-LiPapandonatosCalafat:2019); [Shoaff et al., 2020](#ref-ShoaffCoullWeuve:2020); [Vilmand et al., 2023](#ref-VilmandBeckBilenberg:2023)), and BPA ([Li et al., 2018](#ref-LiZhangKuang:2018); [Rodríguez-Carrillo et al., 2019](#X71fca68020aad3c79232160399444b9923ab76e); [Tewar et al., 2016](#ref-TewarAuingerBraun:2016)).

In children aged 6 to 11 years, higher levels of dialkylphosphate (DAP) metabolites were associated with lower scores of intelligence quotient (IQ) and verbal comprehension, especially in boys ([González-Alzaga et al., 2015](#X2a5c5c17453fffe785b7f0bcffd564283892c6c)), while higher levels of diethylphosphate metabolites were associated with lower working memory scores ([Cartier et al., 2016](#Xa0cc0fdd8206e54b0a9b4bdcd9d2102efd24cb7)). There is also preliminary evidence of a possible association between exposure to certain OP pesticides and Attention-Deficit / Hyperactivity Disorder (ADHD) in children ([Bouchard et al., 2010](#ref-BouchardBellingerWright:2010); [Yu et al., 2016](#ref-YuDuChiou:2016)). In our study we found higher levels of DETP to be associated with better attentional function, with concordant results in the sex-stratified analysis. While we cannot exclude a protective effect of DETP in relation to attentional function, we note that in our study DETP has a higher percentage of missing values compared to the other OP pesticides and that other imputation strategies might lead to different associations.

Previous evidence is also available for several phthalate metabolites in relation to cognitive development in childhood. Higher levels of di(2-ethylhexyl) phthalate metabolites (including MEHP, MEHHP, and MEOHP) were associated with lower intelligence scores in children aged 2 to 12 years ([Huang et al., 2015](#ref-HuangChenSu:2015)), lower scores of IQ and verbal intelligence, more omission errors (a measure of inattention), and higher scores of response time variability (a measure of sustained attention) in 6-year old Korean children ([Kim et al., 2017](#ref-KimHongShin:2017)), poorer fine motor skills in preadolescent boys ([Balalian et al., 2019](#ref-BalalianWhyattLiu:2019)), and lower intelligence scores in 7-year old children ([Vilmand et al., 2023](#ref-VilmandBeckBilenberg:2023)). Further associations were found for higher levels of MEOHP with lower scores of IQ ([Huang et al., 2015](#ref-HuangChenSu:2015)) and verbal intelligence in Taiwanese children aged 6 to 12 years ([Huang et al., 2017](#ref-HuangTsaiChen:2017)), and for higher levels of dibutyl phthalate metabolites (MnBP and MiBP) with impaired verbal intelligence ([Huang et al., 2017](#ref-HuangTsaiChen:2017)). Few studies have investigated different classes of non-persistent EDCs. Shoaff et al. investigated cross-sectional associations between multiple EDCs and ADHD-related behaviors in 15-year old adolescents, finding a higher risk of ADHD-related behavior problems with higher levels of antiandrogenic phthalate metabolites (molar sum of MnBP, MiBP, MBzP, MEHP, MEHHP, MEOHP, MECPP, monocarboxyoctyl phthalate, monohydroxyisobutyl phthalate (MHiBP), monohydroxybutyl phthalate (MHBP), and mono-isononyl phthalate) and the molar sum of di(2-ethylhexyl) phthalate (DEHP) metabolites (MECPP, MEHHP, MEOHP, and MEHP), especially in boys ([Shoaff et al., 2020](#ref-ShoaffCoullWeuve:2020)). Our findings, indicating that short-term childhood exposure to certain phthalate metabolites (MEHP, oh-MiNP, and oxo-MiNP) was associated with attentional function, adds to this growing evidence base suggesting that childhood phthalate exposure may impact child neurodevelopment.

Among phenols, some studies provide preliminary evidence of an association between BPA and ADHD in children aged 8 to 15 years ([Tewar et al., 2016](#ref-TewarAuingerBraun:2016)) and in a case-control study of children aged 6 to 12 years ([Li et al., 2018](#ref-LiZhangKuang:2018)), especially in boys. Except for working memory, there was no evidence of an association between BPA and cognitive abilities in Spanish boys aged 9 to 11 years ([Rodríguez-Carrillo et al., 2019](#X71fca68020aad3c79232160399444b9923ab76e)). We did not observe an association between BPA and attention function in the present study, but this study is the first to suggest that childhood exposure to MEPA may be associated with lower attentional function.

We are not aware of other epidemiological studies investigating childhood exposure to phthalates metabolites, phenols, and OP pesticides, in relation to urinary glucocorticosteroid levels in childhood. However, prior epidemiological research provides preliminary evidence for an association between certain non-persistent EDCs measured at other time points with higher levels of glucocorticoids measured in other biological matrices ([Kim et al., 2018](#ref-KimLeeMoon:2018); [Sears et al., 2023](#ref-SearsLiuLanphear:2023); [Sun et al., 2018](#ref-SunLiJin:2018)). Repeated measures up to 15 months of age of the phthalate metabolites MEHHP, MEOHP, MiBP, and MnBP showed positive associations with free urine cortisol in Korean children ([Kim et al., 2018](#ref-KimLeeMoon:2018)). In a cohort of Chinese pregnant women, phthalate metabolites were measured at 14, 24, and 36 weeks of gestation, and the glucocorticoids cortisol and cortisone were measured in cord blood. Third-trimester levels of MEHP were positively associated with cortisol, while MECPP and MEOHP were negatively associated with cortisone ([Sun et al., 2018](#ref-SunLiJin:2018)). Time- and chemical-dependent sex differences were also found: during the third trimester, MEHHP and MEOHP were positively associated with cortisol in females, while negatively associated in males ([Sun et al., 2018](#ref-SunLiJin:2018)). In a longitudinal study, a mixture of several phthalate metabolites, driven by MEP, MiBP, and MBzP, measured in childhood, showed a positive association with hair cortisol measured at 12 years of age ([Sears et al., 2023](#ref-SearsLiuLanphear:2023)). Our findings also indicate associations between certain phthalate metabolites (MEHP, oh-MiNP, and oxo-MiNP) and glucocorticosteroids, but differences in the exposure assessment time points, the biological matrices used for glucocorticosteroids determinations, and the possible effect of parturition on cord concentrations of cortisol ([Tribe et al., 2018](#ref-TribeTaylorKelly:2018)), make a direct comparison difficult. Adding to these epidemiological studies, previous toxicological research provide evidence for the inhibition by phthalates of human 11-hydroxysteroid dehydrogenase 2 (11-HSD2) activity, responsible for the conversion of active cortisol into inactive cortisone ([Ma et al., 2011](#ref-MaLianDong:2011); [Zhao et al., 2010](#ref-ZhaoChuHuang:2010)).

We are also not aware of prior epidemiological studies specifically investigating the effects of elevated levels of glucocorticosteroids in relation to attentional function, although there is evidence that under- or over-production of glucocorticosteroids interfere with the normal development of the brain, possibly in a sex-specific manner ([Lupien et al., 2009](#ref-LupienMcEwenGunnar:2009)). Our findings support evidence for such sex-specific interference.

Our findings should be interpreted considering the following strengths and limitations. Strengths include its relatively large sample size and its inclusion of multiple classes of non-persistent EDCs. Further, this study used pooled urine samples for chemical assessment to reduce measurement error, since it is known that these specific EDCs have very short half-lives ([Casas et al., 2018](#ref-CasasBasaganaSakhi:2018); [Perrier et al., 2016](#ref-PerrierGiorgis-AllemandSlama:2016)). We decided to model both the *treatment* mechanisms, for the estimation of balancing weights, and the outcomes, with traditional covariates adjustment, to obtain *doubly robust* effect estimates. Finally, we decided not to interpret our results by focusing on the estimated coefficients of possibly misspecified regression models, but by making use of the g-computation procedure.

Limitations include the cross-sectional design of the present study. Importantly, the non-persistent EDCs were measured in a pool of night and morning urine samples before the clinical visit, to represent exposure over the previous day, whereas the glucocorticosteroids were measured in the night urine sample. Cross-sectional designs also have limitations to study the etiologic period of susceptibility of attentional function. Prior studies have investigated associations between short-term effects of air pollution and attentional function, with null results in adolescents but higher response times in adults ([Gignac et al., 2022](#ref-GignacRighiToran:2022), [2021](#ref-GignacBarrera-GomezPersavento:2021)). Finally, longitudinal studies are necessary for a formal causal mediation analysis (Fairchild and McDaniel, 2017), which is why we did not attempt such a analysis even for the sex-stratified case, where glucocorticosteroids were associated with HRT-SE. Although we included a wide range of confounders, there is the possibility, as with other observational studies, of residual confounding, which might lead to a bias away from the null. We decided not to include prenatal levels of the EDCs as confounders due to the *small* correlation between prenatal and childhood levels of these non-persistent chemicals ([Haug et al., 2018](#ref-HaugSakhiCequier:2018)). There is further the possibility of misspecification of the outcome model, although we included a spline of the exposure to relax some of the linearity assumptions and doubly robust estimators are consistent if either the outcome model or the exposure model is specified correctly (Naimi et al., 2023). Finally, we acknowledge that the present study does not consider the possible mixture effect of these chemicals. While methods for causal mixture analysis are available ([Keil et al., 2020](#ref-KeilBuckleyO:2020)), we also emphasize that using these tools without thoughtful consideration may lead to misleading results ([Webster and Weisskopf, 2020](#ref-WebsterWeisskopf:2020)).

## 4.1 Conclusion

In conclusion, in a study of 1,297 children from 6 European birth cohorts, we observed that (i) exposure to certain non-persistent EDCs is associated with higher values of HRT-SE and might disrupt the HPA axis, and (ii) certain glucocorticosteroids are associated with HRT-SE in a sex-specific manner.

While several of our results are *statistically significant* (i.e., the CIs do not include the null effect), the estimates are generally *small* in magnitude. Nevertheless, two important factors should be considered when discussing the public health significance of these findings. First, the reported marginal associations are averages of the conditional associations over the study sample (e.g., confounders). While the marginal associations are *small* in magnitude, individual-level effects might still be clinically significant. Second, EDCs are ubiquitous chemicals present in many every-day products and diet. Thus, even *small* marginal associations might be significant at the population level. For non-persistent EDCs, larger studies with repeated measurements of the exposures at several time points are necessary to fully understand their impact on neurodevelopment.

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# References

Arel-Bundock, V., 2023. [Marginaleffects: Predictions, comparisons, slopes, marginal means, and hypothesis tests](https://marginaleffects.com/).

Balalian, A.A., Whyatt, R.M., Liu, X., Insel, B.J., Rauh, V.A., Herbstman, J., Factor-Litvak, P., 2019. Prenatal and childhood exposure to phthalates and motor skills at age 11 years. Environmental Research 171, 416–427. <https://doi.org/10.1016/j.envres.2019.01.046>

Barrett, M., 2023. [Ggdag: Analyze and Create Elegant Directed Acyclic Graphs](https://github.com/r-causal/ggdag).

Bouchard, M.F., Bellinger, D.C., Wright, R.O., Weisskopf, M.G., 2010. Attention-Deficit/Hyperactivity Disorder and Urinary Metabolites of Organophosphate Pesticides. Pediatrics 125, e1270–e1277. <https://doi.org/10.1542/peds.2009-3058>

Braun, J.M., 2017. Early-life exposure to EDCs: Role in childhood obesity and neurodevelopment. Nat Rev Endocrinol 13, 161–173. <https://doi.org/10.1038/nrendo.2016.186>

Cartier, C., Warembourg, C., Le Maner-Idrissi, G., Lacroix, A., Rouget, F., Monfort, C., Limon, G., Durand, G., Saint-Amour, D., Cordier, S., Chevrier, C., 2016. Organophosphate Insecticide Metabolites in Prenatal and Childhood Urine Samples and Intelligence Scores at 6 Years of Age: Results from the Mother–Child PELAGIE Cohort (France). Environmental Health Perspectives 124, 674–680. <https://doi.org/10.1289/ehp.1409472>

Casas, M., Basagaña, X., Sakhi, A.K., Haug, L.S., Philippat, C., Granum, B., Manzano-Salgado, C.B., Brochot, C., Zeman, F., de Bont, J., Andrusaityte, S., Chatzi, L., Donaire-Gonzalez, D., Giorgis-Allemand, L., Gonzalez, J.R., Gracia-Lavedan, E., Grazuleviciene, R., Kampouri, M., Lyon-Caen, S., Pañella, P., Petraviciene, I., Robinson, O., Urquiza, J., Vafeiadi, M., Vernet, C., Waiblinger, D., Wright, J., Thomsen, C., Slama, R., Vrijheid, M., 2018. Variability of urinary concentrations of non-persistent chemicals in pregnant women and school-aged children. Environ. Int. 121, 561–573. <https://doi.org/10.1016/j.envint.2018.09.046>

Chatzi, L., Plana, E., Daraki, V., Karakosta, P., Alegkakis, D., Tsatsanis, C., Kafatos, A., Koutis, A., Kogevinas, M., 2009. Metabolic Syndrome in Early Pregnancy and Risk of Preterm Birth. American Journal of Epidemiology 170, 829–836. <https://doi.org/10.1093/aje/kwp211>

Fairchild, A.J. and McDaniel, H.L., 2017. Best (but oft-forgotten) practices: mediation analysis. The American journal of clinical nutrition 105, 1259-1271. https://doi.org/10.3945/ajcn.117.152546

Forns, J., Esnaola, M., López-Vicente, M., Suades-González, E., Alvarez-Pedrerol, M., Julvez, J., Grellier, J., Sebastián-Gallés, N., Sunyer, J., 2014. The n-back Test and the Attentional Network Task as measures of child neuropsychological development in epidemiological studies. Neuropsychology 28, 519–529. <https://doi.org/10.1037/neu0000085>

Furlong, M.A., Herring, A., Buckley, J.P., Goldman, B.D., Daniels, J.L., Engel, L.S., Wolff, M.S., Chen, J., Wetmur, J., Barr, D.B., Engel, S.M., 2017. Prenatal exposure to organophosphorus pesticides and childhood neurodevelopmental phenotypes. Environmental Research 158, 737–747. <https://doi.org/10.1016/j.envres.2017.07.023>

Gignac, F., Barrera-Gómez, J., Persavento, C., Solé, C., Tena, È., López-Vicente, M., Foraster, M., Amato, F., Alastuey, A., Querol, X., Llavador, H., Apesteguia, J., Júlvez, J., Couso, D., Sunyer, J., Basagaña, X., 2021. Short-term effect of air pollution on attention function in adolescents (ATENC!Ó): A randomized controlled trial in high schools in Barcelona, Spain. Environment International 156, 106614. <https://doi.org/10.1016/j.envint.2021.106614>

Gignac, F., Righi, V., Toran, R., Paz Errandonea, L., Ortiz, R., Mijling, B., Naranjo, A., Nieuwenhuijsen, M., Creus, J., Basagaña, X., 2022. Short-term NO2 exposure and cognitive and mental health: A panel study based on a citizen science project in Barcelona, Spain. Environment International 164, 107284. <https://doi.org/10.1016/j.envint.2022.107284>

Gomez-Gomez, A., Pozo, O.J., 2020. Determination of steroid profile in hair by liquid chromatography tandem mass spectrometry. Journal of Chromatography A 1624, 461179. <https://doi.org/10.1016/j.chroma.2020.461179>

González-Alzaga, B., Hernández, A.F., Rodríguez-Barranco, M., Gómez, I., Aguilar-Garduño, C., López-Flores, I., Parrón, T., Lacasaña, M., 2015. Pre- and postnatal exposures to pesticides and neurodevelopmental effects in children living in agricultural communities from South-Eastern Spain. Environment International 85, 229–237. <https://doi.org/10.1016/j.envint.2015.09.019>

Grandjean, P., Landrigan, P.J., 2014. Neurobehavioural effects of developmental toxicity. Lancet Neurol 13, 330–338. <https://doi.org/10.1016/S1474-4422(13)70278-3>

Grazuleviciene, R., Danileviciute, A., Nadisauskiene, R., Vencloviene, J., 2009. Maternal Smoking,GSTM1 and GSTT1 Polymorphism and Susceptibility to Adverse Pregnancy Outcomes. International Journal of Environmental Research and Public Health 6, 1282–1297. <https://doi.org/10.3390/ijerph6031282>

Greenland, S., Pearl, J., Robins, J, M., 1999. Causal Diagrams for Epidemiologic Research. Epidemiology 10, 37-48. <http://dx.doi.org/10.1097/00001648-199901000-00008>

Greifer, N., 2023a. Cobalt: Covariate balance tables and plots.

Greifer, N., 2023b. WeightIt: Weighting for covariate balance in observational studies.

Guxens, M., Ballester, F., Espada, M., Fernández, M.F., Grimalt, J.O., Ibarluzea, J., Olea, N., Rebagliato, M., Tardón, A., Torrent, M., Vioque, J., Vrijheid, M., Sunyer, J., on behalf of the INMA Project, 2012. Cohort Profile: The INMA—INfancia y Medio Ambiente—(Environment and Childhood) Project. International Journal of Epidemiology 41, 930–940. <https://doi.org/10.1093/ije/dyr054>

Haug, L.S., Sakhi, A.K., Cequier, E., Casas, M., Maitre, L., Basagana, X., Andrusaityte, S., Chalkiadaki, G., Chatzi, L., Coen, M., de Bont, J., Dedele, A., Ferrand, J., Grazuleviciene, R., Gonzalez, J.R., Gutzkow, K.B., Keun, H., McEachan, R., Meltzer, H.M., Petraviciene, I., Robinson, O., Saulnier, P.-J., Slama, R., Sunyer, J., Urquiza, J., Vafeiadi, M., Wright, J., Vrijheid, M., Thomsen, C., 2018. In-utero and childhood chemical exposome in six European mother-child cohorts. Environment International 121, 751–763. <https://doi.org/10.1016/j.envint.2018.09.056>

Heude, B., Forhan, A., Slama, R., Douhaud, L., Bedel, S., Saurel-Cubizolles, M.-J., Hankard, R., Thiebaugeorges, O., De Agostini, M., Annesi-Maesano, I., Kaminski, M., Charles, M.-A., Annesi-Maesano, I., Bernard, J., Botton, J., Charles, M.-A., Dargent-Molina, P., de Lauzon-Guillain, B., Ducimetière, P., de Agostini, M., Foliguet, B., Forhan, A., Fritel, X., Germa, A., Goua, V., Hankard, R., Heude, B., Kaminski, M., Larroque, B., Lelong, N., Lepeule, J., Magnin, G., Marchand, L., Nabet, C., Pierre, F., Slama, R., Saurel-Cubizolles, M., Schweitzer, M., Thiebaugeorges, O., on behalf of the EDEN mother-child cohort study group, 2016. Cohort Profile: The EDEN mother-child cohort on the prenatal and early postnatal determinants of child health and development. International Journal of Epidemiology 45, 353–363. <https://doi.org/10.1093/ije/dyv151>

Huang, H.-B., Chen, H.-Y., Su, P.-H., Huang, P.-C., Sun, C.-W., Wang, C.-J., Chen, H.-Y., Hsiung, C.A., Wang, S.-L., 2015. Fetal and Childhood Exposure to Phthalate Diesters and Cognitive Function in Children Up to 12 Years of Age: Taiwanese Maternal and Infant Cohort Study. PLOS ONE 10, e0131910. <https://doi.org/10.1371/journal.pone.0131910>

Huang, P.-C., Tsai, C.-H., Chen, C.-C., Wu, M.-T., Chen, M.-L., Wang, S.-L., Chen, B.-H., Lee, C.-C., Jaakkola, J.J.K., Wu, W.-C., Chen, M.-K., Hsiung, C.A., Group, R., 2017. Intellectual evaluation of children exposed to phthalate-tainted products after the 2011 Taiwan phthalate episode. Environmental Research 156, 158–166. <https://doi.org/10.1016/j.envres.2017.03.016>

Huling, J.D., Greifer, N., Chen, G., 2023. Independence Weights for Causal Inference with Continuous Treatments. Journal of the American Statistical Association 0, 1–14. <https://doi.org/10.1080/01621459.2023.2213485>

Jankowska, A., Polańska, K., Hanke, W., Wesołowska, E., Ligocka, D., Waszkowska, M., Stańczak, A., Tartaglione, A.M., Mirabella, F., Chiarotti, F., Garí, M., Calamandrei, G., 2019. Prenatal and early postnatal phthalate exposure and child neurodevelopment at age of 7 years – Polish Mother and Child Cohort. Environmental Research 177, 108626. <https://doi.org/10.1016/j.envres.2019.108626>

Keil, A.P., Buckley, J.P., O, ’Brien.K.M., Ferguson, K.K., Zhao, S., White, A.J., 2020. A Quantile-Based g-Computation Approach to Addressing the Effects of Exposure Mixtures. Environmental Health Perspectives 128, 047004. <https://doi.org/10.1289/EHP5838>

Kim, J.H., Lee, J., Moon, H.-B., Park, J., Choi, K., Kim, S.K., Kim, S., 2018. Association of phthalate exposures with urinary free cortisol and 8-hydroxy-2’-deoxyguanosine in early childhood. Science of The Total Environment 627, 506–513. <https://doi.org/10.1016/j.scitotenv.2018.01.125>

Kim, J.I., Hong, Y.-C., Shin, C.H., Lee, Y.A., Lim, Y.-H., Kim, B.-N., 2017. The effects of maternal and children phthalate exposure on the neurocognitive function of 6-year-old children. Environmental Research 156, 519–525. <https://doi.org/10.1016/j.envres.2017.04.003>

Kowarik, A., Templ, M., 2016. Imputation with the R Package VIM. Journal of Statistical Software 74, 1–16. <https://doi.org/10.18637/jss.v074.i07>

Lazar, C., 2015. imputeLCMD: A collection of methods for left-censored missing data imputation. R package, version 2.

Li, N., Papandonatos, G.D., Calafat, A.M., Yolton, K., Lanphear, B.P., Chen, A., Braun, J.M., 2019. Identifying periods of susceptibility to the impact of phthalates on children’s cognitive abilities. Environmental Research 172, 604–614. <https://doi.org/10.1016/j.envres.2019.03.009>

Li, Y., Zhang, H., Kuang, H., Fan, R., Cha, C., Li, G., Luo, Z., Pang, Q., 2018. Relationship between bisphenol A exposure and attention-deficit/ hyperactivity disorder: A case-control study for primary school children in Guangzhou, China. Environmental Pollution 235, 141–149. <https://doi.org/10.1016/j.envpol.2017.12.056>

Lüdecke, D., Ben-Shachar, M.S., Patil, I., Waggoner, P., Makowski, D., 2021. performance: An R package for assessment, comparison and testing of statistical models. Journal of Open Source Software 6, 3139. <https://doi.org/10.21105/joss.03139>

Lupien, S.J., McEwen, B.S., Gunnar, M.R., Heim, C., 2009. Effects of stress throughout the lifespan on the brain, behaviour and cognition. Nat Rev Neurosci 10, 434–445. <https://doi.org/10.1038/nrn2639>

Ma, X., Lian, Q.-Q., Dong, Q., Ge, R.-S., 2011. Environmental inhibitors of 11β-hydroxysteroid dehydrogenase type 2. Toxicology 285, 83–89. <https://doi.org/10.1016/j.tox.2011.04.007>

Magnus, P., Irgens, L.M., Haug, K., Nystad, W., Skjærven, R., Stoltenberg, C., The Moba Study Group, 2006. Cohort profile: The Norwegian Mother and Child Cohort Study (MoBa). International Journal of Epidemiology 35, 1146–1150. <https://doi.org/10.1093/ije/dyl170>

Mah, V.K. and Ford-Jones, E.L., 2012. Spotlight on middle childhood: Rejuvenating the ‘forgotten years’. Paediatrics & child health 17, 81-83. https://doi.org/10.1093/pch/17.2.81

Maitre, L., Bont, J. de, Casas, M., Robinson, O., Aasvang, G.M., Agier, L., Andrušaitytė, S., Ballester, F., Basagaña, X., Borràs, E., Brochot, C., Bustamante, M., Carracedo, A., Castro, M. de, Dedele, A., Donaire-Gonzalez, D., Estivill, X., Evandt, J., Fossati, S., Giorgis-Allemand, L., Gonzalez, J.R., Granum, B., Grazuleviciene, R., Gützkow, K.B., Haug, L.S., Hernandez-Ferrer, C., Heude, B., Ibarluzea, J., Julvez, J., Karachaliou, M., Keun, H.C., Krog, N.H., Lau, C.-H.E., Leventakou, V., Lyon-Caen, S., Manzano, C., Mason, D., McEachan, R., Meltzer, H.M., Petraviciene, I., Quentin, J., Roumeliotaki, T., Sabido, E., Saulnier, P.-J., Siskos, A.P., Siroux, V., Sunyer, J., Tamayo, I., Urquiza, J., Vafeiadi, M., Gent, D. van, Vives-Usano, M., Waiblinger, D., Warembourg, C., Chatzi, L., Coen, M., Hazel, P. van den, Nieuwenhuijsen, M.J., Slama, R., Thomsen, C., Wright, J., Vrijheid, M., 2018. Human Early Life Exposome (HELIX) study: A European population-based exposome cohort. BMJ Open 8, e021311. <https://doi.org/10.1136/bmjopen-2017-021311>

Marcos, J., Renau, N., Casals, G., Segura, J., Ventura, R., Pozo, O.J., 2014. Investigation of endogenous corticosteroids profiles in human urine based on liquid chromatography tandem mass spectrometry. Analytica Chimica Acta 812, 92–104. <https://doi.org/10.1016/j.aca.2013.12.030>

Naimi, A.I., Mishler, A.E., Kennedy, E.H., 2023. Challenges in obtaining valid causal effect estimates with machine learning algorithms. American Journal of Epidemiology 192, 1536-1544. https://doi.org/10.1093/aje/kwab201

Oh, J., Kim, K., Kannan, K., Parsons, P.J., Mlodnicka, A., Schmidt, R.J., Schweitzer, J.B., Hertz-Picciotto, I., Bennett, D.H., 2023. Early childhood exposure to environmental phenols and parabens, phthalates, organophosphate pesticides, and trace elements in association with attention deficit hyperactivity disorder (ADHD) symptoms in the CHARGE study. Res Sq rs.3.rs–2565914. <https://doi.org/10.21203/rs.3.rs-2565914/v1>

Perrier, F., Giorgis-Allemand, L., Slama, R., Philippat, C., 2016. Within-subject Pooling of Biological Samples to Reduce Exposure Misclassification in Biomarker-based Studies. Epidemiology 27, 378. <https://doi.org/10.1097/EDE.0000000000000460>

Rafi, Z., Greenland, S., 2020. Semantic and cognitive tools to aid statistical science: Replace confidence and significance by compatibility and surprise. BMC Medical Research Methodology 20, 244. <https://doi.org/10.1186/s12874-020-01105-9>

Ramírez, V., Gálvez-Ontiveros, Y., González-Domenech, P.J., Baca, M.Á., Rodrigo, L., Rivas, A., 2022. Role of endocrine disrupting chemicals in children’s neurodevelopment. Environmental Research 203, 111890. <https://doi.org/10.1016/j.envres.2021.111890>

Rodríguez-Carrillo, A., Mustieles, V., Pérez-Lobato, R., Molina-Molina, J.M., Reina-Pérez, I., Vela-Soria, F., Rubio, S., Olea, N., Fernández, M.F., 2019. Bisphenol A and cognitive function in school-age boys: Is BPA predominantly related to behavior? NeuroToxicology 74, 162–171. <https://doi.org/10.1016/j.neuro.2019.06.006>

Rueda, M.R., Fan, J., McCandliss, B.D., Halparin, J.D., Gruber, D.B., Lercari, L.P., Posner, M.I., 2004. Development of attentional networks in childhood. Neuropsychologia 42, 1029–1040. <https://doi.org/10.1016/j.neuropsychologia.2003.12.012>

Sears, C.G., Liu, Y., Lanphear, B.P., Buckley, J.P., Meyer, J., Xu, Y., Chen, A., Yolton, K., Braun, J.M., 2023. Evaluating mixtures of urinary phthalate metabolites and serum per-/polyfluoroalkyl substances in relation to adolescent hair cortisol: The HOME Study. American Journal of Epidemiology kwad198. <https://doi.org/10.1093/aje/kwad198>

Shoaff, J.R., Coull, B., Weuve, J., Bellinger, D.C., Calafat, A.M., Schantz, S.L., Korrick, S.A., 2020. Association of Exposure to Endocrine-Disrupting Chemicals During Adolescence With Attention-Deficit/Hyperactivity Disorder–Related Behaviors. JAMA Network Open 3, e2015041. <https://doi.org/10.1001/jamanetworkopen.2020.15041>

Smith, M.J., Mansournia, M.A., Maringe, C., Zivich, P.N., Cole, S.R., Leyrat, C., Belot, A., Rachet, B., Luque‐Fernandez, M.A., 2022. Introduction to computational causal inference using reproducible Stata, R, and Python code: a tutorial. Statistics in Medicine 41, 407-432. https://doi.org/10.1002/sim.9234

Sun, X., Li, J., Jin, S., Li, Y., Liu, W., Zhao, H., Zhou, Y., Jiang, Y., Liu, H., Xia, W., Cai, Z., Xu, S., Shen, X., 2018. Associations between repeated measures of maternal urinary phthalate metabolites during pregnancy and cord blood glucocorticoids. Environment International 121, 471–479. <https://doi.org/10.1016/j.envint.2018.09.037>

Sunyer, J., Esnaola, M., Alvarez-Pedrerol, M., Forns, J., Rivas, I., López-Vicente, M., Suades-González, E., Foraster, M., Garcia-Esteban, R., Basagaña, X., Viana, M., Cirach, M., Moreno, T., Alastuey, A., Sebastian-Galles, N., Nieuwenhuijsen, M., Querol, X., 2015. Association between Traffic-Related Air Pollution in Schools and Cognitive Development in Primary School Children: A Prospective Cohort Study. PLOS Medicine 12, e1001792. <https://doi.org/10.1371/journal.pmed.1001792>

Tewar, S., Auinger, P., Braun, J.M., Lanphear, B., Yolton, K., Epstein, J.N., Ehrlich, S., Froehlich, T.E., 2016. Association of Bisphenol A exposure and Attention-Deficit/Hyperactivity Disorder in a national sample of U.S. children. Environmental Research 150, 112–118. <https://doi.org/10.1016/j.envres.2016.05.040>

Textor, J., van der Zander, B., Gilthorpe, M.S., Liśkiewicz, M., Ellison, G.T., 2016. Robust causal inference using directed acyclic graphs: The R package “dagitty.” International Journal of Epidemiology 45, 1887–1894. <https://doi.org/10.1093/ije/dyw341>

Tribe, R.M., Taylor, P.D., Kelly, N.M., Rees, D., Sandall, J., Kennedy, H.P., 2018. Parturition and the perinatal period: Can mode of delivery impact on the future health of the neonate? The Journal of Physiology 596, 5709–5722. <https://doi.org/10.1113/JP275429>

Vilmand, M., Beck, I.H., Bilenberg, N., Andersson, A.-M., Juul, A., Schoeters, G., Boye, H., Frederiksen, H., Jensen, T.K., 2023. Prenatal and current phthalate exposure and cognitive development in 7-year-old children from the Odense child cohort. Neurotoxicology and Teratology 96, 107161. <https://doi.org/10.1016/j.ntt.2023.107161>

Vrijheid, M., Slama, R., Robinson, O., Chatzi, L., Coen, M., van den Hazel, P., Thomsen, C., Wright, J., Athersuch, T.J., Avellana, N., Basagaña, X., Brochot, C., Bucchini, L., Bustamante, M., Carracedo, A., Casas, M., Estivill, X., Fairley, L., van Gent, D., Gonzalez, J.R., Granum, B., Gražulevičienė, R., Gutzkow, K.B., Julvez, J., Keun, H.C., Kogevinas, M., McEachan, R.R.C., Meltzer, H.M., Sabidó, E., Schwarze, P.E., Siroux, V., Sunyer, J., Want, E.J., Zeman, F., Nieuwenhuijsen, M.J., 2014. The human early-life exposome (HELIX): Project rationale and design. Environ Health Perspect 122, 535–544. <https://doi.org/10.1289/ehp.1307204>

Webster, T.F., Weisskopf, M.G., 2020. Epidemiology of exposure to mixtures: We cant be casual about causality when using or testing methods [WWW Document]. <https://doi.org/10.48550/arXiv.2007.01370>

Wright, J., Small, N., Raynor, P., Tuffnell, D., Bhopal, R., Cameron, N., Fairley, L., Lawlor, D.A., Parslow, R., Petherick, E.S., Pickett, K.E., Waiblinger, D., West, J., on behalf of the Born in Bradford Scientific Collaborators Group, 2013. Cohort Profile: The Born in Bradford multi-ethnic family cohort study. International Journal of Epidemiology 42, 978–991. <https://doi.org/10.1093/ije/dys112>

Yu, C.-J., Du, J.-C., Chiou, H.-C., Chung, M.-Y., Yang, W., Chen, Y.-S., Fuh, M.-R., Chien, L.-C., Hwang, B., Chen, M.-L., 2016. Increased risk of attention-deficit/hyperactivity disorder associated with exposure to organophosphate pesticide in Taiwanese children. Andrology 4, 695–705. <https://doi.org/10.1111/andr.12183>

Zeileis, A., 2004. Econometric computing with HC and HAC covariance matrix estimators. Journal of Statistical Software 11, 1–17. <https://doi.org/10.18637/jss.v011.i10>

Zeileis, A., Köll, S., Graham, N., 2020. Various versatile variances: An object-oriented implementation of clustered covariances in R. Journal of Statistical Software 95, 1–36. <https://doi.org/10.18637/jss.v095.i01>

Zhao, B., Chu, Y., Huang, Y., Hardy, D.O., Lin, S., Ge, R.-S., 2010. Structure-dependent inhibition of human and rat 11β-hydroxysteroid dehydrogenase 2 activities by phthalates. Chemico-Biological Interactions 183, 79–84. <https://doi.org/10.1016/j.cbi.2009.09.014>

# 6. Tables for descriptive data

## 6.1 Study populations

|  |  |
| --- | --- |
| Table 1. **Participant characteristics (HELIX subcohort; 2013-2016).** | |
| **Characteristic** | **N = 1,297***a* | |
| Child age (years) | 8.1 (6.5, 8.9) | |
| Child breastfeeding | 1,093.0 (84.7%) | |
| Unknown | 6 | |
| Child ethnicity |  | |
| Caucasian | 1,157.0 (90.0%) | |
| Pakistani | 80.0 (6.2%) | |
| Asian | 21.0 (1.6%) | |
| Other | 19.0 (1.5%) | |
| African | 7.0 (0.5%) | |
| Native American | 2.0 (0.2%) | |
| White non European | 0.0 (0.0%) | |
| Unknown | 11 | |
| Child head circumference (cm) | 51.8 (50.6, 52.9) | |
| Unknown | 3 | |
| Child height (m) | 1.3 (1.2, 1.4) | |
| Child neuropsychological diagnosis | 95.0 (7.3%) | |
| Child rest before assessment |  | |
| Yes | 1,209.0 (93.3%) | |
| Not as well as usual | 87.0 (6.7%) | |
| Unknown | 1 | |
| Child sex |  | |
| Male | 710.0 (54.7%) | |
| Female | 587.0 (45.3%) | |
| Child weight (kg) | 26.9 (22.9, 32.6) | |
| Chiod mood before assessment |  | |
| Usual | 1,232.0 (95.1%) | |
| Not usual | 64.0 (4.9%) | |
| Unknown | 1 | |
| Cohort |  | |
| MOBA | 272.0 (21.0%) | |
| INMA | 221.0 (17.0%) | |
| BIB | 204.0 (15.7%) | |
| KANC | 203.0 (15.7%) | |
| RHEA | 199.0 (15.3%) | |
| EDEN | 198.0 (15.3%) | |
| Creatinine night sample (g/l) | 1.7 (0.9, 3.0) | |
| Unknown | 321 | |
| Creatinine pooled sample (g/l) | 1.0 (0.8, 1.2) | |
| Date of test (season) |  | |
| Spring | 358.0 (27.7%) | |
| Winter | 339.0 (26.2%) | |
| Autumn | 300.0 (23.2%) | |
| Summer | 297.0 (23.0%) | |
| Unknown | 3 | |
| Family affluence scale |  | |
| 6 | 410.0 (31.7%) | |
| 5 | 325.0 (25.1%) | |
| 7 | 248.0 (19.2%) | |
| 4 | 174.0 (13.4%) | |
| 3 | 92.0 (7.1%) | |
| 2 | 28.0 (2.2%) | |
| 1 | 12.0 (0.9%) | |
| 0 | 6.0 (0.5%) | |
| Unknown | 2 | |
| Fast food/take away (times/week) | 0.1 (0.1, 0.5) | |
| Unknown | 7 | |
| Fasting time before visit (hours) | 3.3 (2.8, 4.0) | |
| Financial situation of the parents |  | |
| Doing alright | 414.0 (32.1%) | |
| Living comfortably | 412.0 (31.9%) | |
| Getting by | 331.0 (25.6%) | |
| Finding it quite difficult | 86.0 (6.7%) | |
| Finding it very difficult | 40.0 (3.1%) | |
| Does not wish to answer | 8.0 (0.6%) | |
| Unknown | 6 | |
| Fish and seafood (times/week) | 2.0 (1.1, 3.5) | |
| Unknown | 5 | |
| Fruits (times/week) | 9.0 (5.9, 18.0) | |
| Unknown | 7 | |
| Hit reaction time standard error (ms) | 299.6 (231.3, 368.2) | |
| Unknown | 18 | |
| Marital status |  | |
| Living with the father | 1,212.0 (94.5%) | |
| Living alone | 39.0 (3.0%) | |
| Other situation | 31.0 (2.4%) | |
| Unknown | 15 | |
| Maternal tobacco consumption |  | |
| Non-smoker and has never smoked | 681.0 (52.6%) | |
| Daily smoker | 200.0 (15.5%) | |
| Non-smoker but previously smoked daily | 186.0 (14.4%) | |
| Non-smoker but previously smoked although not daily | 163.0 (12.6%) | |
| Smoker but not daily | 64.0 (4.9%) | |
| Unknown | 3 | |
| Organic food (times/week) | 0.5 (0.0, 3.0) | |
| Unknown | 7 | |
| Pregnancy maternal active smoking | 190.0 (15.1%) | |
| Unknown | 40 | |
| Pregnancy maternal passive smoking | 514.0 (40.3%) | |
| Unknown | 21 | |
| Vegetables (times/week) | 6.5 (4.0, 10.0) | |
| Unknown | 6 | |
| *a*n (%); Median (IQR) | | |

## 6.2 Endocrine disruptors

|  |  |  |
| --- | --- | --- |
| Table 2. **Participants endocrine disruptors concentrations expressed in grams/L (HELIX subcohort; 2013-2016).** | | |
| **Characteristic** | **N = 1,297***a* | **N = 1,297***b* | |
| **OP pesticide metabolites** | | | |
| DEP | 1.8 (0.4, 4.6) | 2.0 (0.2) | |
| DETP | 0.1 (0.1, 1.7) | 21.0 (1.6) | |
| DMP | 0.4 (0.3, 4.6) | 6.0 (0.5) | |
| DMTP | 2.8 (1.2, 6.3) | 1.0 (0.1) | |
| **Phenols** | | | |
| BPA | 3.8 (2.3, 7.0) | 12.0 (0.9) | |
| BUPA | 0.1 (0.0, 0.1) | 5.0 (0.4) | |
| ETPA | 0.7 (0.4, 1.2) | 3.0 (0.2) | |
| MEPA | 6.3 (3.1, 24.1) | 2.0 (0.2) | |
| OXBE | 2.0 (0.8, 6.6) | 0.0 (0.0) | |
| PRPA | 0.2 (0.0, 1.6) | 17.0 (1.3) | |
| TRCS | 0.6 (0.3, 1.5) | 0.0 (0.0) | |
| **Phthalate metabolites** | | | |
| MBzP | 4.8 (2.7, 8.7) | 1.0 (0.1) | |
| MECPP | 32.8 (19.9, 57.6) | 1.0 (0.1) | |
| MEHHP | 19.3 (11.4, 33.1) | 3.0 (0.2) | |
| MEHP | 2.8 (1.6, 5.1) | 41.0 (3.2) | |
| MEOHP | 12.2 (7.1, 20.4) | 1.0 (0.1) | |
| MEP | 32.5 (15.0, 79.2) | 0.0 (0.0) | |
| MiBP | 40.2 (24.5, 71.1) | 0.0 (0.0) | |
| MnBP | 22.7 (14.5, 38.8) | 0.0 (0.0) | |
| oh-MiNP | 5.0 (3.1, 9.3) | 0.0 (0.0) | |
| oxo-MiNP | 2.7 (1.7, 5.0) | 0.0 (0.0) | |
| *a*Median (IQR) | | | |
| *b*N missing (% missing) | | | |

## 6.3 Glucocorticosteroids

|  |  |
| --- | --- |
| Table 3. **Participants derived glucocorticosteroids concentrations expressed in ng/ml (HELIX subcohort; 2013-2016).** | |
| **Characteristic** | **N = 1,004***a* | **N = 976***a,b* |
| cortisol production | 4,607.9 (2,859.9, 6,789.5); 18.0 (1.79) | 4,559.5 (2,828.4, 6,750.2); 17.0 (1.74) |
| cortisone production | 4,608.1 (2,920.8, 6,843.9); 19.0 (1.89) | 4,580.7 (2,894.4, 6,802.4); 18.0 (1.84) |
| corticosterone production | 257.8 (157.9, 410.5); 3.0 (0.30) | 256.7 (157.5, 409.7); 3.0 (0.31) |
| *a*Median (Q1, Q3); N Missing (% Missing) | | |
| *b*Measurements available for the HELIX subcohort. | | |

# 7. Tables for other analyses

## 7.1 Marginal hypotheses for effect modification

|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| Table 4. **Pairwise differences between marginal contrasts on the logarithmic scale of males and females, for the effect of an increase from the 10th to the 90th percentile of endocrine disrupting chemicals (EDCs) on hit reaction time standard error (HRT-SE), expressed in ms, and on the glucocorticosteroids, expressed in ng/ml (HELIX subcohort; 2013-2016).** | | | | |
|  | HRT-SE*a* | corticosterone production*a* | cortisol production*a* | cortisone production*a* |
| **OP pesticide metabolites** | | | | |
| DEP | 0.019 (-0.022, 0.061) | -0.082 (-0.276, 0.113) | -0.139 (-0.374, 0.096) | -0.104 (-0.311, 0.103) |
| DETP | 0.025 (-0.054, 0.104) | -0.16 (-0.332, 0.012) | -0.071 (-0.264, 0.123) | -0.096 (-0.269, 0.076) |
| DMP | -0.034 (-0.093, 0.025) | 0.007 (-0.217, 0.231) | -0.031 (-0.119, 0.057) | -0.069 (-0.207, 0.07) |
| DMTP | 0.005 (-0.095, 0.106) | -0.014 (-0.165, 0.137) | -0.21 (-0.326, -0.094) | -0.166 (-0.353, 0.022) |
| **Phenols** | | | | |
| BPA | 0.032 (-0.026, 0.09) | -0.153 (-0.291, -0.015) | -0.125 (-0.269, 0.018) | -0.085 (-0.216, 0.047) |
| BUPA | -0.022 (-0.067, 0.024) | -0.117 (-0.247, 0.012) | -0.129 (-0.209, -0.048) | -0.013 (-0.112, 0.085) |
| ETPA | 0.012 (-0.021, 0.045) | -0.254 (-0.416, -0.092) | -0.184 (-0.39, 0.022) | -0.219 (-0.472, 0.034) |
| MEPA | -0.001 (-0.061, 0.058) | -0.129 (-0.271, 0.013) | -0.127 (-0.258, 0.004) | -0.144 (-0.257, -0.03) |
| OXBE | 0.032 (0.004, 0.061) | -0.213 (-0.486, 0.059) | -0.077 (-0.306, 0.153) | -0.064 (-0.274, 0.146) |
| PRPA | 0.015 (-0.045, 0.074) | -0.12 (-0.262, 0.022) | -0.043 (-0.238, 0.151) | -0.102 (-0.223, 0.019) |
| TRCS | -0.017 (-0.076, 0.042) | -0.142 (-0.251, -0.034) | -0.13 (-0.248, -0.012) | -0.152 (-0.207, -0.096) |
| **Phthalate metabolites** | | | | |
| MBzP | -0.066 (-0.126, -0.007) | -0.025 (-0.098, 0.047) | -0.018 (-0.142, 0.107) | -0.079 (-0.174, 0.015) |
| MECPP | 0.008 (-0.077, 0.092) | -0.014 (-0.165, 0.137) | -0.043 (-0.084, -0.001) | 0.017 (-0.055, 0.09) |
| MEHHP | 0.028 (-0.075, 0.131) | -0.052 (-0.264, 0.161) | -0.091 (-0.208, 0.026) | -0.006 (-0.087, 0.075) |
| MEHP | 0.017 (-0.082, 0.115) | -0.165 (-0.26, -0.071) | -0.221 (-0.289, -0.153) | -0.177 (-0.299, -0.055) |
| MEOHP | 0.02 (-0.068, 0.108) | -0.061 (-0.232, 0.111) | -0.075 (-0.157, 0.006) | 0.009 (-0.063, 0.08) |
| MEP | -0.053 (-0.138, 0.033) | -0.05 (-0.408, 0.308) | -0.083 (-0.384, 0.218) | -0.119 (-0.339, 0.1) |
| MiBP | -0.02 (-0.138, 0.098) | 0.037 (-0.175, 0.25) | -0.042 (-0.267, 0.184) | -0.021 (-0.163, 0.12) |
| MnBP | -0.035 (-0.11, 0.041) | 0.029 (-0.186, 0.243) | 0.063 (-0.134, 0.26) | 0.017 (-0.077, 0.111) |
| oh-MiNP | 0.046 (-0.009, 0.102) | -0.127 (-0.335, 0.08) | -0.181 (-0.33, -0.033) | -0.164 (-0.304, -0.024) |
| oxo-MiNP | -0.026 (-0.059, 0.008) | -0.12 (-0.315, 0.076) | -0.146 (-0.303, 0.011) | -0.127 (-0.238, -0.016) |
| *a*Estimate and 95% CI. | | | | |

|  |
| --- |
| Table 5. **Pairwise differences between marginal contrasts on the logarithmic scale of males and females, for the effect of an increase from the 10th to the 90th percentile of the glucocorticosteroids on hit reaction time standard error (HRT-SE) expressed in ms (HELIX subcohort; 2013-2016).** |
|  | HRT-SE*a* |
| **Glucocorticosteroids** | |
| corticosterone production | 0.126 (0.009, 0.243) |
| cortisol production | 0.096 (-0.045, 0.238) |
| cortisone production | 0.14 (0.019, 0.261) |
| *a*Estimate and 95% CI. | |

# 8. DAG

|  |
| --- |
| Figure 1. **Simplified directed acyclic graph (DAG) showing endocrine disrupting chemicals (EDCs), hit reaction time standard error (HRT-SE), and the glucocorticosteroids.** |

# 9. Figures for main results

## 9.1 Marginal contrasts

|  |
| --- |
| Figure 2. **Marginal contrasts on the logarithmic scale for the effect of an increase from the 10th to the 90th percentile of the endocrine disrupting chemicals (EDCs) on hit reaction time standard error (HRT-SE) expressed in ms (HELIX subcohort; 2013-2016).** Circles indicate effect estimates. Solid lines indicate the CI. The size of the circles represents the S value of the effect estimate ([Rafi and Greenland, 2020](#ref-RafiGreenland:2020)). |

|  |
| --- |
| Figure 3. **Marginal contrasts on the logarithmic scale for the effect of an increase from the 10th to the 90th percentile of the endocrine disrupting chemicals (EDCs) on the glucocorticosteroids expressed in ng/ml (HELIX subcohort; 2013-2016).** Circles, triangles, and squares indicate effect estimates. Solid lines indicate the CI. The size of the circles represents the S value of the effect estimate ([Rafi and Greenland, 2020](#ref-RafiGreenland:2020)). |

|  |
| --- |
| Figure 4. **Marginal contrasts on the logarithmic scale for the effect of an increase from the 10th to the 90th percentile of the glucocorticosteroids on hit reaction time standard error (HRT-SE) expressed in ms, overall and by sex (HELIX subcohort; 2013-2016).** Circles indicate effect estimates. Solid lines indicate the CI. The size of the circles represents the S value of the effect estimate ([Rafi and Greenland, 2020](#ref-RafiGreenland:2020)). Abbreviations: cortisone production (cortisone prod.); cortisol production (cortisol prod.); corticost. prod. (corticosterone production). |

# 10. Supplementary information

## 10.1 Directed Acyclic Graphs

dag {  
age\_child  
biomarker  
breastfeeding  
bw  
characteristics\_child  
chemical [exposure]  
child\_diet  
child\_smoking  
cohort  
creatinine  
envFactors\_visit  
ethnicity\_child  
ethnicity\_mother  
familySEP  
gestational\_age  
maternalAlcohol\_preg  
maternalDiet\_preg  
maternalSEP\_preg  
maternalSmoking\_preg  
neuropsychologicalDiagnosis\_child  
outcome [outcome]  
paternalSEP\_preg  
season\_visit  
sex\_child  
time\_lastMeal  
type\_sample  
age\_child -> biomarker  
age\_child -> characteristics\_child  
age\_child -> creatinine  
age\_child -> outcome  
age\_child -> type\_sample  
biomarker -> outcome  
breastfeeding -> neuropsychologicalDiagnosis\_child  
breastfeeding -> outcome  
bw -> characteristics\_child  
bw -> neuropsychologicalDiagnosis\_child  
characteristics\_child -> biomarker  
characteristics\_child -> chemical  
characteristics\_child -> creatinine  
characteristics\_child -> outcome  
chemical -> biomarker  
chemical -> outcome  
child\_diet -> biomarker  
child\_diet -> characteristics\_child  
child\_diet -> chemical  
child\_diet -> outcome  
child\_smoking -> biomarker  
child\_smoking -> characteristics\_child  
child\_smoking -> creatinine  
child\_smoking -> outcome  
cohort -> biomarker  
cohort -> bw  
cohort -> characteristics\_child  
cohort -> chemical  
cohort -> child\_diet  
cohort -> creatinine  
cohort -> outcome  
creatinine -> biomarker  
creatinine -> chemical  
creatinine -> outcome  
envFactors\_visit -> outcome  
ethnicity\_child -> biomarker  
ethnicity\_child -> bw  
ethnicity\_child -> characteristics\_child  
ethnicity\_child -> chemical  
ethnicity\_child -> child\_diet  
ethnicity\_child -> child\_smoking  
ethnicity\_child -> creatinine  
ethnicity\_child -> neuropsychologicalDiagnosis\_child  
ethnicity\_child -> outcome  
ethnicity\_mother -> biomarker  
ethnicity\_mother -> breastfeeding  
ethnicity\_mother -> bw  
ethnicity\_mother -> characteristics\_child  
ethnicity\_mother -> child\_diet  
ethnicity\_mother -> familySEP  
ethnicity\_mother -> maternalAlcohol\_preg  
ethnicity\_mother -> maternalDiet\_preg  
ethnicity\_mother -> maternalSEP\_preg  
ethnicity\_mother -> maternalSmoking\_preg  
ethnicity\_mother -> neuropsychologicalDiagnosis\_child  
ethnicity\_mother -> outcome  
familySEP -> biomarker  
familySEP -> characteristics\_child  
familySEP -> chemical  
familySEP -> child\_diet  
familySEP -> child\_smoking  
familySEP -> creatinine  
familySEP -> outcome  
gestational\_age -> bw  
gestational\_age -> characteristics\_child  
gestational\_age -> neuropsychologicalDiagnosis\_child  
maternalAlcohol\_preg -> bw  
maternalAlcohol\_preg -> characteristics\_child  
maternalAlcohol\_preg -> neuropsychologicalDiagnosis\_child  
maternalAlcohol\_preg -> outcome  
maternalDiet\_preg -> characteristics\_child  
maternalDiet\_preg -> neuropsychologicalDiagnosis\_child  
maternalDiet\_preg -> outcome  
maternalSEP\_preg -> breastfeeding  
maternalSEP\_preg -> bw  
maternalSEP\_preg -> characteristics\_child  
maternalSEP\_preg -> familySEP  
maternalSEP\_preg -> maternalAlcohol\_preg  
maternalSEP\_preg -> maternalDiet\_preg  
maternalSEP\_preg -> maternalSmoking\_preg  
maternalSEP\_preg -> neuropsychologicalDiagnosis\_child  
maternalSEP\_preg -> outcome  
maternalSmoking\_preg -> bw  
maternalSmoking\_preg -> characteristics\_child  
maternalSmoking\_preg -> neuropsychologicalDiagnosis\_child  
maternalSmoking\_preg -> outcome  
neuropsychologicalDiagnosis\_child -> outcome  
paternalSEP\_preg -> breastfeeding  
paternalSEP\_preg -> bw  
paternalSEP\_preg -> characteristics\_child  
paternalSEP\_preg -> familySEP  
paternalSEP\_preg -> maternalAlcohol\_preg  
paternalSEP\_preg -> maternalDiet\_preg  
paternalSEP\_preg -> maternalSmoking\_preg  
paternalSEP\_preg -> neuropsychologicalDiagnosis\_child  
paternalSEP\_preg -> outcome  
season\_visit -> biomarker  
season\_visit -> chemical  
sex\_child -> biomarker  
sex\_child -> characteristics\_child  
sex\_child -> chemical  
sex\_child -> child\_diet  
sex\_child -> child\_smoking  
sex\_child -> creatinine  
sex\_child -> neuropsychologicalDiagnosis\_child  
sex\_child -> outcome  
sex\_child -> type\_sample  
time\_lastMeal -> biomarker  
time\_lastMeal -> chemical  
type\_sample -> chemical  
type\_sample -> creatinine  
}

dag {  
age\_child  
biomarker [outcome]  
breastfeeding  
bw  
characteristics\_child  
chemical [exposure]  
child\_diet  
child\_smoking  
cohort  
creatinine  
envFactors\_visit  
ethnicity\_child  
ethnicity\_mother  
familySEP  
gestational\_age  
maternalAlcohol\_preg  
maternalDiet\_preg  
maternalSEP\_preg  
maternalSmoking\_preg  
neuropsychologicalDiagnosis\_child  
outcome  
paternalSEP\_preg  
season\_visit  
sex\_child  
time\_lastMeal  
type\_sample  
age\_child -> biomarker  
age\_child -> characteristics\_child  
age\_child -> creatinine  
age\_child -> outcome  
age\_child -> type\_sample  
biomarker -> outcome  
breastfeeding -> neuropsychologicalDiagnosis\_child  
breastfeeding -> outcome  
bw -> characteristics\_child  
bw -> neuropsychologicalDiagnosis\_child  
characteristics\_child -> biomarker  
characteristics\_child -> chemical  
characteristics\_child -> creatinine  
characteristics\_child -> outcome  
chemical -> biomarker  
chemical -> outcome  
child\_diet -> biomarker  
child\_diet -> characteristics\_child  
child\_diet -> chemical  
child\_diet -> outcome  
child\_smoking -> biomarker  
child\_smoking -> characteristics\_child  
child\_smoking -> creatinine  
child\_smoking -> outcome  
cohort -> biomarker  
cohort -> bw  
cohort -> characteristics\_child  
cohort -> chemical  
cohort -> child\_diet  
cohort -> creatinine  
cohort -> outcome  
creatinine -> biomarker  
creatinine -> chemical  
creatinine -> outcome  
envFactors\_visit -> outcome  
ethnicity\_child -> biomarker  
ethnicity\_child -> bw  
ethnicity\_child -> characteristics\_child  
ethnicity\_child -> chemical  
ethnicity\_child -> child\_diet  
ethnicity\_child -> child\_smoking  
ethnicity\_child -> creatinine  
ethnicity\_child -> neuropsychologicalDiagnosis\_child  
ethnicity\_child -> outcome  
ethnicity\_mother -> biomarker  
ethnicity\_mother -> breastfeeding  
ethnicity\_mother -> bw  
ethnicity\_mother -> characteristics\_child  
ethnicity\_mother -> child\_diet  
ethnicity\_mother -> familySEP  
ethnicity\_mother -> maternalAlcohol\_preg  
ethnicity\_mother -> maternalDiet\_preg  
ethnicity\_mother -> maternalSEP\_preg  
ethnicity\_mother -> maternalSmoking\_preg  
ethnicity\_mother -> neuropsychologicalDiagnosis\_child  
ethnicity\_mother -> outcome  
familySEP -> biomarker  
familySEP -> characteristics\_child  
familySEP -> chemical  
familySEP -> child\_diet  
familySEP -> child\_smoking  
familySEP -> creatinine  
familySEP -> outcome  
gestational\_age -> bw  
gestational\_age -> characteristics\_child  
gestational\_age -> neuropsychologicalDiagnosis\_child  
maternalAlcohol\_preg -> bw  
maternalAlcohol\_preg -> characteristics\_child  
maternalAlcohol\_preg -> neuropsychologicalDiagnosis\_child  
maternalAlcohol\_preg -> outcome  
maternalDiet\_preg -> characteristics\_child  
maternalDiet\_preg -> neuropsychologicalDiagnosis\_child  
maternalDiet\_preg -> outcome  
maternalSEP\_preg -> breastfeeding  
maternalSEP\_preg -> bw  
maternalSEP\_preg -> characteristics\_child  
maternalSEP\_preg -> familySEP  
maternalSEP\_preg -> maternalAlcohol\_preg  
maternalSEP\_preg -> maternalDiet\_preg  
maternalSEP\_preg -> maternalSmoking\_preg  
maternalSEP\_preg -> neuropsychologicalDiagnosis\_child  
maternalSEP\_preg -> outcome  
maternalSmoking\_preg -> bw  
maternalSmoking\_preg -> characteristics\_child  
maternalSmoking\_preg -> neuropsychologicalDiagnosis\_child  
maternalSmoking\_preg -> outcome  
neuropsychologicalDiagnosis\_child -> outcome  
paternalSEP\_preg -> breastfeeding  
paternalSEP\_preg -> bw  
paternalSEP\_preg -> characteristics\_child  
paternalSEP\_preg -> familySEP  
paternalSEP\_preg -> maternalAlcohol\_preg  
paternalSEP\_preg -> maternalDiet\_preg  
paternalSEP\_preg -> maternalSmoking\_preg  
paternalSEP\_preg -> neuropsychologicalDiagnosis\_child  
paternalSEP\_preg -> outcome  
season\_visit -> biomarker  
season\_visit -> chemical  
sex\_child -> biomarker  
sex\_child -> characteristics\_child  
sex\_child -> chemical  
sex\_child -> child\_diet  
sex\_child -> child\_smoking  
sex\_child -> creatinine  
sex\_child -> neuropsychologicalDiagnosis\_child  
sex\_child -> outcome  
sex\_child -> type\_sample  
time\_lastMeal -> biomarker  
time\_lastMeal -> chemical  
type\_sample -> chemical  
type\_sample -> creatinine  
}

dag {  
age\_child  
biomarker [exposure]  
breastfeeding  
bw  
characteristics\_child  
chemical  
child\_diet  
child\_smoking  
cohort  
creatinine  
envFactors\_visit  
ethnicity\_child  
ethnicity\_mother  
familySEP  
gestational\_age  
maternalAlcohol\_preg  
maternalDiet\_preg  
maternalSEP\_preg  
maternalSmoking\_preg  
neuropsychologicalDiagnosis\_child  
outcome [outcome]  
paternalSEP\_preg  
season\_visit  
sex\_child  
time\_lastMeal  
type\_sample  
age\_child -> biomarker  
age\_child -> characteristics\_child  
age\_child -> creatinine  
age\_child -> outcome  
age\_child -> type\_sample  
biomarker -> outcome  
breastfeeding -> neuropsychologicalDiagnosis\_child  
breastfeeding -> outcome  
bw -> characteristics\_child  
bw -> neuropsychologicalDiagnosis\_child  
characteristics\_child -> biomarker  
characteristics\_child -> chemical  
characteristics\_child -> creatinine  
characteristics\_child -> outcome  
chemical -> biomarker  
chemical -> outcome  
child\_diet -> biomarker  
child\_diet -> characteristics\_child  
child\_diet -> chemical  
child\_diet -> outcome  
child\_smoking -> biomarker  
child\_smoking -> characteristics\_child  
child\_smoking -> creatinine  
child\_smoking -> outcome  
cohort -> biomarker  
cohort -> bw  
cohort -> characteristics\_child  
cohort -> chemical  
cohort -> child\_diet  
cohort -> creatinine  
cohort -> outcome  
creatinine -> biomarker  
creatinine -> chemical  
creatinine -> outcome  
envFactors\_visit -> outcome  
ethnicity\_child -> biomarker  
ethnicity\_child -> bw  
ethnicity\_child -> characteristics\_child  
ethnicity\_child -> chemical  
ethnicity\_child -> child\_diet  
ethnicity\_child -> child\_smoking  
ethnicity\_child -> creatinine  
ethnicity\_child -> neuropsychologicalDiagnosis\_child  
ethnicity\_child -> outcome  
ethnicity\_mother -> biomarker  
ethnicity\_mother -> breastfeeding  
ethnicity\_mother -> bw  
ethnicity\_mother -> characteristics\_child  
ethnicity\_mother -> child\_diet  
ethnicity\_mother -> familySEP  
ethnicity\_mother -> maternalAlcohol\_preg  
ethnicity\_mother -> maternalDiet\_preg  
ethnicity\_mother -> maternalSEP\_preg  
ethnicity\_mother -> maternalSmoking\_preg  
ethnicity\_mother -> neuropsychologicalDiagnosis\_child  
ethnicity\_mother -> outcome  
familySEP -> biomarker  
familySEP -> characteristics\_child  
familySEP -> chemical  
familySEP -> child\_diet  
familySEP -> child\_smoking  
familySEP -> creatinine  
familySEP -> outcome  
gestational\_age -> bw  
gestational\_age -> characteristics\_child  
gestational\_age -> neuropsychologicalDiagnosis\_child  
maternalAlcohol\_preg -> bw  
maternalAlcohol\_preg -> characteristics\_child  
maternalAlcohol\_preg -> neuropsychologicalDiagnosis\_child  
maternalAlcohol\_preg -> outcome  
maternalDiet\_preg -> characteristics\_child  
maternalDiet\_preg -> neuropsychologicalDiagnosis\_child  
maternalDiet\_preg -> outcome  
maternalSEP\_preg -> breastfeeding  
maternalSEP\_preg -> bw  
maternalSEP\_preg -> characteristics\_child  
maternalSEP\_preg -> familySEP  
maternalSEP\_preg -> maternalAlcohol\_preg  
maternalSEP\_preg -> maternalDiet\_preg  
maternalSEP\_preg -> maternalSmoking\_preg  
maternalSEP\_preg -> neuropsychologicalDiagnosis\_child  
maternalSEP\_preg -> outcome  
maternalSmoking\_preg -> bw  
maternalSmoking\_preg -> characteristics\_child  
maternalSmoking\_preg -> neuropsychologicalDiagnosis\_child  
maternalSmoking\_preg -> outcome  
neuropsychologicalDiagnosis\_child -> outcome  
paternalSEP\_preg -> breastfeeding  
paternalSEP\_preg -> bw  
paternalSEP\_preg -> characteristics\_child  
paternalSEP\_preg -> familySEP  
paternalSEP\_preg -> maternalAlcohol\_preg  
paternalSEP\_preg -> maternalDiet\_preg  
paternalSEP\_preg -> maternalSmoking\_preg  
paternalSEP\_preg -> neuropsychologicalDiagnosis\_child  
paternalSEP\_preg -> outcome  
season\_visit -> biomarker  
season\_visit -> chemical  
sex\_child -> biomarker  
sex\_child -> characteristics\_child  
sex\_child -> chemical  
sex\_child -> child\_diet  
sex\_child -> child\_smoking  
sex\_child -> creatinine  
sex\_child -> neuropsychologicalDiagnosis\_child  
sex\_child -> outcome  
sex\_child -> type\_sample  
time\_lastMeal -> biomarker  
time\_lastMeal -> chemical  
type\_sample -> chemical  
type\_sample -> creatinine  
}

# 11. Supplementary tables

## 11.1 Tables for descriptive data

### 11.1.1 Information about the endocrine disruptors

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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| |  |  |  |  |  | | --- | --- | --- | --- | --- | | Compound | Symbol | Variable name | PubChem CID | Parental compound | | **OP pesticide metabolites** | | | | | | diethyl dithiophosphate | DEDTP | dedtp | 9274 |  | | diethyl phosphate | DEP | dep | 654 |  | | diethyl thiophosphate | DETP | detp | 3683036 |  | | dimethyl dithiophosphate | DMDTP | dmdtp | 36158 |  | | dimethyl phosphate | DMP | dmp | 13134 |  | | dimethyl thiophosphate | DMTP | dmtp | 168140 |  | | **Phenols** | | | | | | bisphenol A | BPA | bpa | 6623 |  | | n‑butyl‑paraben | BUPA | bupa | 7184 |  | | ethyl-paraben | ETPA | etpa | 8434 |  | | methyl-paraben | MEPA | mepa | 7456 |  | | oxybenzone | OXBE | oxbe | 4632 |  | | propyl-paraben | PRPA | prpa | 7175 |  | | triclosan | TRCS | trcs | 5564 |  | | **Phthalate metabolites** | | | | | | mono benzyl phthalate | MBzP | mbzp | 31736 | BBzP | | mono‑2‑ethyl 5‑carboxypentyl phthalate | MECPP | mecpp | 148386 | DEHP | | mono‑2‑ethyl‑5‑hydroxyhexyl phthalate | MEHHP | mehhp | 170295 | DEHP | | mono‑2‑ethylhexyl phthalate | MEHP | mehp | 21924291 | DEHP | | mono‑2‑ethyl‑5‑oxohexyl phthalate | MEOHP | meohp | 119096 | DEHP | | monoethyl phthalate | MEP | mep | 75318 | DEP | | mono‑iso‑butyl phthalate | MiBP | mibp | 92272 | DiBP | | mono‑n‑butyl phthalate | MnBP | mnbp | 8575 | DnBP | | mono‑4‑methyl‑7‑hydroxyoctyl phthalate | oh-MiNP | ohminp | 102401880 | MiNP | | mono‑4‑methyl‑7‑oxooctyl phthalate | oxo-MiNP | oxominp | 102401881 | MiNP |   Table S1. **Information about non-persistent endocrine disrupting chemicals (EDCs), including the full compound name, the standard symbol, the used variable name, the identifier from PubChem, and the parental compound.** |

### 11.1.2 Information about the glucocorticosteroids

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  | | --- | --- | --- | --- | | Metabolite | Symbol | HMDB ID | CAS number | | **Androgen** | | | | | Androsternedione | AED | HMDB0000053 | 63-05-8 | | Testosterone | T | HMDB0000234 | 58-22-0 | | **Androgen metabolite** | | | | | Androsterone | Andros | HMDB0000031 | 53-41-8 | | Etiocholanolone | Etio | HMDB0000490 | 53-42-9 | | **Glucocorticosteroid** | | | | | 11-dehydrocorticosterone | A | HMDB0004029 | 72-23-1 | | Corticosterone | B | HMDB0001547 | 50-22-6 | | Cortisol | F | HMDB0000063 | 50-23-7 | | Cortisone | E | HMDB0002802 | 53-06-5 | | **Glucocorticosteroid metabolite** | | | | | 11β-hydroxyandrosterone | 11OHAndros | HMDB0002984 | 57-61-4 | | 17-deoxycortolone | 17-DO-cortolone | NA | NA | | 20α-dihydrocortisol | 20aDHF | NA | NA | | 20α-dihydrocortisone | 20aDHE | NA | NA | | 20β-dihydrocortisol | 20bDHF | NA | NA | | 20β-dihydrocortisone | 20bDHE | NA | NA | | 5α,20α-cortol | 5a20acortol | HMDB0003180 | 516-38-1 | | 5α,20β-cortol | 5a20bcortol | HMDB0005821 | 667-65-2 | | 5α-tetrahydrocorticosterone | 5aTHB | HMDB0000449 | 600-63-5 | | 5α-tetrahydrocortisol | 5aTHF | HMDB0000526 | 302-91-0 | | 5α-tetrahydrocortisone | 5aTHE | NA | NA | | 5β,20α-cortol | 5b20acortol | HMDB0003180 | 516-38-1 | | 5β,20α-cortolone | 5b20acortolone | HMDB0003128 | 516-42-7 | | 5β,20β-cortol | 5b20bcortol | HMDB0005821 | 667-65-2 | | 5β,20β-cortolone | 5b20bcortolone | NA | NA | | 5β-dihydrocortisol | 5bDHF | HMDB0003259 | 1482-50-4 | | 5β-tetrahydrocorticosterone | 5bTHB | HMDB0000268 | 68-42-8 | | 5β-tetrahydrocortisol | 5bTHF | HMDB0000949 | 1953-02-01 | | 5β-tetrahydrocortisone | 5bTHE | NA | NA | | 6β-hydroxycortisol | 6OHF | HMDB0247074 |  | | 6β-hydroxycortisone | 6OHE | NA | NA | | **Glucocorticosteroid precursor** | | | | | 17-hydroxyprogesterone | 17OHP | HMDB0000374 | 68-96-2 | | Cortexolone | S | HMDB0000015 | 152-58-9 | | Deoxycorticosterone | DOC | HMDB0000016 | 64-85-7 | | **Glucocorticosteroid precursor metabolite** | | | | | 17-hydroxypregnanolone | 17HP | HMDB0000363 | 387-79-1 | | 5β-dihydrocortexolone | 5bDHS | NA | NA | | 5β-tetrahydrocortexolone | 5bTHS | NA | NA | | Pregnantriol | PT | NA | 1098-45-9 | | Tetrahydrocortexolone | THS | HMDB0005972 | 68-60-0 | | Abbreviations: Human Metabolome Database (HMDB); Chemical Abstracts Service (CAS). | | | |   Table S2. **Information about the glucocorticosteroids, including the full metabolite name, the standard symbol, the identifier from the HMDB, and the CAS number.** |

### 11.1.3 Codebooks

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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| |  |  |  |  |  |  |  |  |  |  |  | | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | |  | type | description | coding | labels | remarks | comments | assessment | q1 | q2 | q3 | | **age\_child** | | | | | | | | | | | | hs\_age\_years | numerical | Child age |  |  |  | years | postnatal | Yes | Yes | Yes | | **breastfeeding** | | | | | | | | | | | | hs\_bf | categorical | Child breastfeeding | 0,1 | No, Yes |  |  | prenatal | Yes | No | Yes | | **characteristics\_child** | | | | | | | | | | | | hs\_c\_height | numerical | Child height |  |  |  | m | postnatal | Yes | Yes | Yes | | hs\_c\_weight | numerical | Child weight |  |  |  | kg | postnatal | Yes | Yes | Yes | | hs\_head\_circ | numerical | Child head circumference |  |  |  | cm | postnatal | Yes | Yes | Yes | | **child\_diet** | | | | | | | | | | | | hs\_fastfood | numerical | Fast food/take away |  |  |  | Times / week | postnatal | Yes | Yes | Yes | | hs\_org\_food | numerical | Organic food |  |  |  | Times / week | postnatal | Yes | Yes | Yes | | hs\_total\_fish | numerical | Fish and seafood |  |  |  | Times / week | postnatal | Yes | Yes | Yes | | hs\_total\_fruits | numerical | Fruits |  |  |  | Times / week | postnatal | Yes | Yes | Yes | | hs\_total\_veg | numerical | Vegetables |  |  |  | Times / week | postnatal | Yes | Yes | Yes | | **child\_smoking** | | | | | | | | | | | | hs\_tob | categorical | Maternal tobacco consumption | 1,2,3,4,5 | Non-smoker and has never smoked, Non-smoker but previously smoked although not daily, Non-smoker but previously smoked daily, Smoker but not daily, Daily smoker |  |  | postnatal | Yes | Yes | Yes | | **cohort** | | | | | | | | | | | | cohort | character | Cohort | SAB,EDEN,BIB,RHEA,KANC,MOBA | SAB, EDEN, BIB, RHEA, KANC, MOBA |  |  |  | Yes | Yes | Yes | | **creatinine** | | | | | | | | | | | | hs\_creatinine\_cg | numerical | Creatinine pooled sample |  |  | Values below the limit of detection imputed | G / L | postnatal | Yes | Yes | No | | creatinine\_to\_helix | numerical | Creatinine night sample |  |  |  | G / L | postnatal | No | Yes | Yes | | **envFactors\_visit** | | | | | | | | | | | | hs\_mood | categorical | Chiod mood before assessment | 1,2 | Usual, Not usual |  |  | postnatal | Yes | No | Yes | | hs\_rest\_nth | categorical | Child rest before assessment | 1,2 | Yes, Not as well as usual |  |  | postnatal | Yes | No | Yes | | **ethnicity\_child** | | | | | | | | | | | | h\_ethnicity\_c | character | Child ethnicity | 1,2,3,4,5,6,7 | African, Asian, Caucasian, Native American, Other, Pakistani, White non European |  |  | postnatal | Yes | Yes | Yes | | **ethnicity\_mother** | | | | | | | | | | | | h\_ethnicity\_m | integer | Mother ethnicity | 1,2,3,4,5,6,7 | White European, Pakistani, Asian, African, Other, Native American, White non European |  |  | postnatal | No | No | No | | **familySEP** | | | | | | | | | | | | FAS\_score | numerical | Family Affluence Scale |  |  |  |  | postnatal | Yes | Yes | Yes | | hs\_finance | categorical | Financial situation of the parents | 1,2,3,4,5,6 | Living comfortably, Doing alright, Getting by, Finding it quite difficult, Finding it very difficult, Does not wish to answer |  |  | postnatal | Yes | Yes | Yes | | **maternalAlcohol\_preg** | | | | | | | | | | | | e3\_alcpreg\_g | numerical | Alcool during pregnancy |  |  |  | Glasses / week | prenatal | No | No | No | | **maternalDiet\_preg** | | | | | | | | | | | | h\_cereal\_preg | numerical | Cereal consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | h\_dairy\_preg | numerical | Dairy consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | h\_fastfood\_preg | numerical | Fast food consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | h\_fish\_preg | numerical | Fish consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | h\_fruit\_preg | numerical | Fruit consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | h\_legume\_preg | numerical | Legume consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | h\_meat\_preg | numerical | Meat consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | h\_veg\_preg | numerical | Vegetables consumption during pregnancy |  |  |  | Times / week | prenatal | No | No | No | | **maternalSEP\_preg** | | | | | | | | | | | | e3\_edum | categorical | Maternal education | 0,1,2 | Primary school, Secondary school, University degree or higher |  |  | prenatal | No | No | No | | e3\_marital | categorical | Marital status | 0,1,2 | Living with the father, Living alone, Other situation |  |  | prenatal | Yes | No | Yes | | e3\_ses | categorical | Socioeconomic status of the parents | 1,2,3 | Low income, Medium income, High income |  |  | prenatal | No | No | No | | **maternalSmoking\_preg** | | | | | | | | | | | | e3\_asmokyn\_p | categorical | Pregnancy maternal active smoking | 0,1 | No, Yes |  |  | prenatal | Yes | No | Yes | | e3\_psmokanyt | categorical | Pregnancy maternal passive smoking | 0,1 | No, Yes |  |  | prenatal | Yes | No | Yes | | **neuropsychologicalDiagnosis\_child** | | | | | | | | | | | | hs\_neuro\_diag | categorical | Child neuropsychological diagnosis | 1,2 | No, Yes |  |  | postnatal | Yes | No | Yes | | **paternalSEP\_preg** | | | | | | | | | | | | e3\_eduf | categorical | Paternal education | 0,1,2 | Primary school, Secondary school, University degree or higher |  |  | prenatal | No | No | No | | **season\_visit** | | | | | | | | | | | | hs\_date\_neu | date | Date of test |  |  |  | season | postnatal | Yes | Yes | No | | **sex\_child** | | | | | | | | | | | | e3\_sex | categorical | Child sex | 0,1 | Male, Female |  |  | postnatal | Yes | Yes | Yes | | **time\_lastMeal** | | | | | | | | | | | | hs\_dift\_mealblood\_imp | numerical | Fasting time before visit |  |  |  | hours | postnatal | Yes | Yes | No | | **chemical** | | | | | | | | | | | | hs\_bpa\_c | numerical | Bisphenol A (BPA) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_bupa\_c | numerical | N-Butyl paraben (BUPA) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_dedtp\_cadj | numerical | Diethyl dithiophosphate (DEDTP) adjusted for creatinine |  |  | Values below the limit of detection imputed | microg / g | postnatal | No | No | No | | hs\_dep\_c | numerical | Diethyl phosphate (DEP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_detp\_c | numerical | Diethyl thiophosphate (DETP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_dmdtp\_craw | numerical | Dimethyl dithiophosphate (DMDTP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | No | | hs\_dmp\_c | numerical | Dimethyl phosphate (DMP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_dmtp\_c | numerical | Dimethyl thiophosphate (DMTP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_etpa\_c | numerical | Ethyl paraben (ETPA) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mbzp\_c | numerical | Mono benzyl phthalate (MbzP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mecpp\_c | numerical | Mono-2-ethyl 5-carboxypentyl phthalate (MECPP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mehhp\_c | numerical | Mono-2-ethyl-5-hydroxyhexyl phthalate (MEHHP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mehp\_c | numerical | Mono-2-ethylhexyl phthalate (MEHP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_meohp\_c | numerical | Mono-2-ethyl-5-oxohexyl phthalate (MEOHP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mep\_c | numerical | Monoethyl phthalate (MEP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mepa\_c | numerical | Methyl paraben (MEPA) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mibp\_c | numerical | Mono-iso-butyl phthalate (MiBP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_mnbp\_c | numerical | Mono-n-butyl phthalate (MnBP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_ohminp\_c | numerical | Mono-4-methyl-7-hydroxyoctyl phthalate (OHMiNP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_oxbe\_c | numerical | Oxybenzone (OXBE) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_oxominp\_c | numerical | Mono-4-methyl-7-oxooctyl phthalate (OXOMiNP) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_prpa\_c | numerical | Propyl paraben (PRPA) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes | | hs\_trcs\_c | numerical | Triclosan (TRCS) |  |  | Values below the limit of detection imputed | microg / L | postnatal | No | No | Yes |   Table S3. **Codebook for the covariates used in the estimation of the marginal comparisons of endocrine disrupting chemicals (EDCs) on hit reaction time standard error (HRT-SE) (q1), endocrine disrupting chemicals (EDCs) on the glucocorticosteroids (q2), and glucocorticosteroids on hit reaction time standard error (HRT-SE) (q3).** |

### 11.1.4 Lower limits of quantification of the glucocorticosteroids

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| |  |  | | --- | --- | | Metabolite | LLOQ | | 5aTHF | 5.00 | | 5bTHE | 5.00 | | 5b20acortolone | 5.00 | | 5b20bcortolone | 5.00 | | 5a20acortol | 2.50 | | 5a20bcortol | 2.50 | | 5b20acortol | 2.50 | | 5b20bcortol | 2.50 | | 11OHAndros | 2.00 | | 17HP | 2.00 | | PT | 2.00 | | 20bDHF | 0.50 | | 5bTHF | 0.50 | | 6OHF | 0.50 | | E | 0.50 | | 20aDHE | 0.50 | | 20bDHE | 0.50 | | 5aTHE | 0.50 | | 6OHE | 0.50 | | 5aTHB | 0.50 | | 5bTHB | 0.50 | | 17DOcortolone | 0.50 | | 5bTHS | 0.50 | | Andros | 0.50 | | Etio | 0.50 | | F | 0.25 | | 20aDHF | 0.25 | | 5bDHF | 0.10 | | A | 0.10 | | S | 0.10 | | 5bDHS | 0.10 | | T | 0.10 | | AED | 0.10 | | Abbreviations: lower limit of quantification (LLOQ). | |   Table S4. **Lower limits of quantification expressed in ng/ml for the glucocorticosteroids (HELIX subcohort; 2013-2016).** |

### 11.1.5 Study populations

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  |  |  |  |  | | --- | --- | --- | --- | --- | --- | --- | --- | | **Characteristic** | **Overall**, N = 1,297*a* | **BIB**, N = 204*a* | **EDEN**, N = 198*a* | **INMA**, N = 221*a* | **KANC**, N = 203*a* | **MOBA**, N = 272*a* | **RHEA**, N = 199*a* | | Child age (years) | 8.1 (6.5, 8.9) | 6.6 (6.5, 6.8) | 10.9 (10.4, 11.2) | 8.8 (8.4, 9.3) | 6.4 (6.1, 6.9) | 8.5 (8.2, 8.8) | 6.5 (6.4, 6.6) | | Child breastfeeding | 1,093.0 (84.7%) | 147.0 (72.4%) | 128.0 (65.0%) | 195.0 (88.6%) | 187.0 (92.6%) | 260.0 (96.3%) | 176.0 (88.4%) | | Unknown | 6 | 1 | 1 | 1 | 1 | 2 | 0 | | Child ethnicity |  |  |  |  |  |  |  | | Caucasian | 1,157.0 (90.0%) | 87.0 (42.6%) | 196.0 (99.5%) | 221.0 (100.0%) | 200.0 (100.0%) | 254.0 (95.8%) | 199.0 (100.0%) | | Pakistani | 80.0 (6.2%) | 80.0 (39.2%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | | Asian | 21.0 (1.6%) | 13.0 (6.4%) | 1.0 (0.5%) | 0.0 (0.0%) | 0.0 (0.0%) | 7.0 (2.6%) | 0.0 (0.0%) | | Other | 19.0 (1.5%) | 17.0 (8.3%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 2.0 (0.8%) | 0.0 (0.0%) | | African | 7.0 (0.5%) | 7.0 (3.4%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | | Native American | 2.0 (0.2%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 2.0 (0.8%) | 0.0 (0.0%) | | White non European | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | | Unknown | 11 | 0 | 1 | 0 | 3 | 7 | 0 | | Child head circumference (cm) | 51.8 (50.6, 52.9) | 51.4 (50.3, 52.3) | 50.5 (49.5, 52.0) | 52.3 (51.3, 53.3) | 52.0 (51.0, 53.0) | 52.5 (51.5, 53.6) | 51.2 (50.2, 52.0) | | Unknown | 3 | 0 | 0 | 0 | 0 | 0 | 3 | | Child height (m) | 1.3 (1.2, 1.4) | 1.2 (1.2, 1.2) | 1.4 (1.4, 1.5) | 1.3 (1.3, 1.4) | 1.2 (1.2, 1.3) | 1.3 (1.3, 1.4) | 1.2 (1.2, 1.2) | | Child neuropsychological diagnosis | 95.0 (7.3%) | 3.0 (1.5%) | 58.0 (29.3%) | 24.0 (10.9%) | 1.0 (0.5%) | 1.0 (0.4%) | 8.0 (4.0%) | | Child rest before assessment |  |  |  |  |  |  |  | | Yes | 1,209.0 (93.3%) | 192.0 (94.1%) | 170.0 (86.3%) | 206.0 (93.2%) | 200.0 (98.5%) | 259.0 (95.2%) | 182.0 (91.5%) | | Not as well as usual | 87.0 (6.7%) | 12.0 (5.9%) | 27.0 (13.7%) | 15.0 (6.8%) | 3.0 (1.5%) | 13.0 (4.8%) | 17.0 (8.5%) | | Unknown | 1 | 0 | 1 | 0 | 0 | 0 | 0 | | Child sex |  |  |  |  |  |  |  | | Male | 710.0 (54.7%) | 112.0 (54.9%) | 113.0 (57.1%) | 120.0 (54.3%) | 111.0 (54.7%) | 143.0 (52.6%) | 111.0 (55.8%) | | Female | 587.0 (45.3%) | 92.0 (45.1%) | 85.0 (42.9%) | 101.0 (45.7%) | 92.0 (45.3%) | 129.0 (47.4%) | 88.0 (44.2%) | | Child weight (kg) | 26.9 (22.9, 32.6) | 22.3 (20.3, 25.0) | 35.7 (32.4, 41.2) | 30.7 (26.8, 36.5) | 23.6 (21.4, 27.1) | 28.5 (25.7, 31.6) | 23.3 (21.2, 27.2) | | Chiod mood before assessment |  |  |  |  |  |  |  | | Usual | 1,232.0 (95.1%) | 198.0 (97.1%) | 176.0 (89.3%) | 214.0 (96.8%) | 187.0 (92.1%) | 262.0 (96.3%) | 195.0 (98.0%) | | Not usual | 64.0 (4.9%) | 6.0 (2.9%) | 21.0 (10.7%) | 7.0 (3.2%) | 16.0 (7.9%) | 10.0 (3.7%) | 4.0 (2.0%) | | Unknown | 1 | 0 | 1 | 0 | 0 | 0 | 0 | | Creatinine night sample (g/l) | 1.7 (0.9, 3.0) | 0.8 (0.6, 1.1) | 3.3 (2.0, 4.3) | 2.5 (1.5, 3.8) | 1.7 (0.9, 2.7) | 2.0 (1.2, 3.0) | 0.8 (0.4, 1.3) | | Unknown | 321 | 72 | 64 | 19 | 23 | 72 | 71 | | Creatinine pooled sample (g/l) | 1.0 (0.8, 1.2) | 1.0 (0.8, 1.2) | 1.2 (1.0, 1.5) | 1.0 (0.8, 1.3) | 0.9 (0.7, 1.1) | 0.9 (0.7, 1.1) | 0.9 (0.7, 1.1) | | Date of test (season) |  |  |  |  |  |  |  | | Spring | 358.0 (27.7%) | 48.0 (23.5%) | 64.0 (32.3%) | 71.0 (32.4%) | 61.0 (30.0%) | 37.0 (13.6%) | 77.0 (38.9%) | | Winter | 339.0 (26.2%) | 40.0 (19.6%) | 61.0 (30.8%) | 97.0 (44.3%) | 38.0 (18.7%) | 73.0 (26.8%) | 30.0 (15.2%) | | Autumn | 300.0 (23.2%) | 49.0 (24.0%) | 1.0 (0.5%) | 30.0 (13.7%) | 77.0 (37.9%) | 105.0 (38.6%) | 38.0 (19.2%) | | Summer | 297.0 (23.0%) | 67.0 (32.8%) | 72.0 (36.4%) | 21.0 (9.6%) | 27.0 (13.3%) | 57.0 (21.0%) | 53.0 (26.8%) | | Unknown | 3 | 0 | 0 | 2 | 0 | 0 | 1 | | Family affluence scale |  |  |  |  |  |  |  | | 6 | 410.0 (31.7%) | 34.0 (16.7%) | 64.0 (32.3%) | 75.0 (34.1%) | 50.0 (24.8%) | 142.0 (52.2%) | 45.0 (22.6%) | | 5 | 325.0 (25.1%) | 48.0 (23.5%) | 29.0 (14.6%) | 65.0 (29.5%) | 69.0 (34.2%) | 57.0 (21.0%) | 57.0 (28.6%) | | 7 | 248.0 (19.2%) | 26.0 (12.7%) | 90.0 (45.5%) | 43.0 (19.5%) | 14.0 (6.9%) | 53.0 (19.5%) | 22.0 (11.1%) | | 4 | 174.0 (13.4%) | 40.0 (19.6%) | 13.0 (6.6%) | 22.0 (10.0%) | 38.0 (18.8%) | 16.0 (5.9%) | 45.0 (22.6%) | | 3 | 92.0 (7.1%) | 34.0 (16.7%) | 2.0 (1.0%) | 11.0 (5.0%) | 22.0 (10.9%) | 3.0 (1.1%) | 20.0 (10.1%) | | 2 | 28.0 (2.2%) | 16.0 (7.8%) | 0.0 (0.0%) | 1.0 (0.5%) | 4.0 (2.0%) | 0.0 (0.0%) | 7.0 (3.5%) | | 1 | 12.0 (0.9%) | 4.0 (2.0%) | 0.0 (0.0%) | 2.0 (0.9%) | 4.0 (2.0%) | 1.0 (0.4%) | 1.0 (0.5%) | | 0 | 6.0 (0.5%) | 2.0 (1.0%) | 0.0 (0.0%) | 1.0 (0.5%) | 1.0 (0.5%) | 0.0 (0.0%) | 2.0 (1.0%) | | Unknown | 2 | 0 | 0 | 1 | 1 | 0 | 0 | | Fast food/take away (times/week) | 0.1 (0.1, 0.5) | 0.5 (0.1, 1.0) | 0.1 (0.1, 0.5) | 0.1 (0.1, 0.5) | 0.1 (0.0, 0.1) | 0.1 (0.1, 0.5) | 0.5 (0.1, 0.5) | | Unknown | 7 | 0 | 0 | 5 | 2 | 0 | 0 | | Fasting time before visit (hours) | 3.3 (2.8, 4.0) | 3.3 (2.8, 4.1) | 3.2 (2.8, 3.7) | 3.0 (2.6, 3.8) | 3.3 (2.8, 3.8) | 3.4 (2.8, 3.8) | 4.0 (3.3, 4.8) | | Financial situation of the parents |  |  |  |  |  |  |  | | Doing alright | 414.0 (32.1%) | 73.0 (35.8%) | 94.0 (47.5%) | 64.0 (29.2%) | 61.0 (30.5%) | 64.0 (23.5%) | 58.0 (29.3%) | | Living comfortably | 412.0 (31.9%) | 59.0 (28.9%) | 49.0 (24.7%) | 29.0 (13.2%) | 48.0 (24.0%) | 202.0 (74.3%) | 25.0 (12.6%) | | Getting by | 331.0 (25.6%) | 59.0 (28.9%) | 36.0 (18.2%) | 82.0 (37.4%) | 70.0 (35.0%) | 4.0 (1.5%) | 80.0 (40.4%) | | Finding it quite difficult | 86.0 (6.7%) | 8.0 (3.9%) | 9.0 (4.5%) | 29.0 (13.2%) | 12.0 (6.0%) | 1.0 (0.4%) | 27.0 (13.6%) | | Finding it very difficult | 40.0 (3.1%) | 5.0 (2.5%) | 10.0 (5.1%) | 15.0 (6.8%) | 2.0 (1.0%) | 0.0 (0.0%) | 8.0 (4.0%) | | Does not wish to answer | 8.0 (0.6%) | 0.0 (0.0%) | 0.0 (0.0%) | 0.0 (0.0%) | 7.0 (3.5%) | 1.0 (0.4%) | 0.0 (0.0%) | | Unknown | 6 | 0 | 0 | 2 | 3 | 0 | 1 | | Fish and seafood (times/week) | 2.0 (1.1, 3.5) | 2.0 (1.0, 3.1) | 2.1 (1.4, 3.0) | 3.5 (2.1, 5.0) | 1.0 (0.4, 1.6) | 2.6 (1.6, 5.0) | 1.5 (1.0, 2.0) | | Unknown | 5 | 1 | 0 | 2 | 2 | 0 | 0 | | Fruits (times/week) | 9.0 (5.9, 18.0) | 15.5 (10.0, 21.0) | 6.6 (3.3, 13.5) | 7.5 (3.6, 12.6) | 7.3 (3.8, 9.6) | 14.1 (8.6, 21.0) | 8.5 (6.2, 13.5) | | Unknown | 7 | 2 | 0 | 2 | 2 | 1 | 0 | | Hit reaction time standard error (ms) | 299.6 (231.3, 368.2) | 355.1 (292.1, 397.5) | 237.7 (184.7, 307.0) | 256.0 (197.4, 313.8) | 368.4 (324.2, 406.6) | 248.7 (193.0, 300.9) | 340.9 (281.1, 399.2) | | Unknown | 18 | 3 | 11 | 3 | 0 | 0 | 1 | | Marital status |  |  |  |  |  |  |  | | Living with the father | 1,212.0 (94.5%) | 178.0 (87.3%) | 193.0 (98.0%) | 219.0 (99.1%) | 168.0 (84.4%) | 260.0 (98.5%) | 194.0 (98.5%) | | Living alone | 39.0 (3.0%) | 0.0 (0.0%) | 2.0 (1.0%) | 0.0 (0.0%) | 31.0 (15.6%) | 3.0 (1.1%) | 3.0 (1.5%) | | Other situation | 31.0 (2.4%) | 26.0 (12.7%) | 2.0 (1.0%) | 2.0 (0.9%) | 0.0 (0.0%) | 1.0 (0.4%) | 0.0 (0.0%) | | Unknown | 15 | 0 | 1 | 0 | 4 | 8 | 2 | | Maternal tobacco consumption |  |  |  |  |  |  |  | | Non-smoker and has never smoked | 681.0 (52.6%) | 148.0 (72.5%) | 87.0 (43.9%) | 103.0 (46.8%) | 104.0 (51.7%) | 138.0 (50.7%) | 101.0 (50.8%) | | Daily smoker | 200.0 (15.5%) | 27.0 (13.2%) | 45.0 (22.7%) | 45.0 (20.5%) | 24.0 (11.9%) | 6.0 (2.2%) | 53.0 (26.6%) | | Non-smoker but previously smoked daily | 186.0 (14.4%) | 11.0 (5.4%) | 37.0 (18.7%) | 42.0 (19.1%) | 21.0 (10.4%) | 53.0 (19.5%) | 22.0 (11.1%) | | Non-smoker but previously smoked although not daily | 163.0 (12.6%) | 12.0 (5.9%) | 19.0 (9.6%) | 23.0 (10.5%) | 32.0 (15.9%) | 63.0 (23.2%) | 14.0 (7.0%) | | Smoker but not daily | 64.0 (4.9%) | 6.0 (2.9%) | 10.0 (5.1%) | 7.0 (3.2%) | 20.0 (10.0%) | 12.0 (4.4%) | 9.0 (4.5%) | | Unknown | 3 | 0 | 0 | 1 | 2 | 0 | 0 | | Organic food (times/week) | 0.5 (0.0, 3.0) | 0.0 (0.0, 0.5) | 0.5 (0.1, 3.0) | 0.0 (0.0, 0.5) | 1.0 (0.1, 3.0) | 1.0 (0.5, 3.0) | 0.0 (0.0, 1.0) | | Unknown | 7 | 0 | 0 | 5 | 2 | 0 | 0 | | Pregnancy maternal active smoking | 190.0 (15.1%) | 25.0 (13.7%) | 47.0 (23.7%) | 55.0 (25.1%) | 12.0 (6.0%) | 9.0 (3.4%) | 42.0 (21.2%) | | Unknown | 40 | 22 | 0 | 2 | 4 | 11 | 1 | | Pregnancy maternal passive smoking | 514.0 (40.3%) | 55.0 (27.5%) | 43.0 (21.8%) | 126.0 (57.8%) | 97.0 (48.7%) | 14.0 (5.3%) | 179.0 (90.4%) | | Unknown | 21 | 4 | 1 | 3 | 4 | 8 | 1 | | Vegetables (times/week) | 6.5 (4.0, 10.0) | 6.0 (4.0, 10.0) | 8.3 (4.4, 11.0) | 6.0 (3.0, 8.5) | 6.0 (3.5, 8.5) | 8.5 (6.0, 14.0) | 6.5 (4.0, 10.0) | | Unknown | 6 | 1 | 0 | 2 | 2 | 1 | 0 | | *a*Median (IQR); n (%) | | | | | | | |   Table S5. **Participant characteristics, by cohort and overall (HELIX subcohort; 2013-2016).** |

### 11.1.6 Concentrations of the endocrine disruptors

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  |  |  |  |  | | --- | --- | --- | --- | --- | --- | --- | --- | | **Characteristic** | **Overall**, N = 1,297*a* | **BIB**, N = 204*a* | **EDEN**, N = 198*a* | **INMA**, N = 221*a* | **KANC**, N = 203*a* | **MOBA**, N = 272*a* | **RHEA**, N = 199*a* | | **OP pesticide metabolites** | | | | | | | | | DEP | 1.8 (0.1, 0.4, 4.6, 9.8) | 3.5 (0.1, 1.1, 8.8, 38.5) | 2.0 (0.1, 0.4, 4.2, 7.5) | 1.6 (0.1, 0.5, 3.7, 7.6) | 0.7 (0.1, 0.1, 2.1, 4.0) | 2.0 (0.1, 0.6, 5.7, 10.0) | 2.0 (0.1, 0.6, 5.4, 10.7) | | DETP | 0.1 (0.1, 0.1, 1.7, 5.0) | 0.1 (0.1, 0.1, 1.5, 3.6) | 0.1 (0.1, 0.1, 1.8, 5.7) | 0.1 (0.1, 0.1, 1.8, 6.3) | 0.1 (0.1, 0.1, 0.4, 2.0) | 0.1 (0.1, 0.1, 1.5, 3.6) | 0.7 (0.1, 0.1, 2.9, 6.4) | | DMP | 0.4 (0.2, 0.3, 4.6, 9.5) | 0.3 (0.2, 0.3, 3.5, 9.2) | 0.6 (0.3, 0.3, 6.8, 12.1) | 0.4 (0.2, 0.3, 5.9, 9.0) | 0.3 (0.2, 0.3, 2.3, 4.7) | 1.9 (0.2, 0.3, 4.4, 9.2) | 1.6 (0.2, 0.3, 6.5, 11.5) | | DMTP | 2.8 (0.2, 1.2, 6.3, 13.1) | 2.2 (0.1, 0.9, 5.5, 11.3) | 4.0 (0.3, 1.6, 9.0, 15.1) | 3.5 (0.6, 1.6, 7.9, 14.4) | 2.4 (0.1, 0.9, 4.6, 10.2) | 2.4 (0.1, 1.0, 5.1, 9.9) | 3.4 (0.8, 1.7, 6.6, 13.3) | | **Phenols** | | | | | | | | | BPA | 3.8 (1.5, 2.3, 7.0, 14.4) | 5.2 (1.7, 2.5, 9.7, 20.7) | 2.9 (1.2, 1.8, 4.6, 7.2) | 4.0 (1.6, 2.3, 7.0, 13.1) | 3.4 (1.2, 2.1, 5.9, 11.8) | 4.0 (1.8, 2.7, 7.5, 15.3) | 4.4 (1.5, 2.6, 9.2, 21.3) | | BUPA | 0.1 (0.0, 0.0, 0.1, 0.3) | 0.1 (0.0, 0.0, 0.1, 0.3) | 0.0 (0.0, 0.0, 0.1, 0.2) | 0.1 (0.0, 0.1, 0.2, 0.4) | 0.1 (0.0, 0.1, 0.2, 0.4) | 0.1 (0.0, 0.0, 0.1, 0.2) | 0.1 (0.0, 0.0, 0.1, 0.3) | | ETPA | 0.7 (0.3, 0.4, 1.2, 3.3) | 0.7 (0.3, 0.4, 1.6, 6.0) | 0.7 (0.3, 0.5, 1.1, 2.4) | 0.8 (0.3, 0.5, 1.5, 5.6) | 0.6 (0.2, 0.3, 1.5, 2.7) | 0.6 (0.2, 0.4, 1.0, 2.8) | 0.6 (0.2, 0.4, 1.1, 2.1) | | MEPA | 6.3 (1.9, 3.1, 24.1, 122.3) | 9.9 (2.8, 5.0, 47.0, 217.6) | 5.3 (2.0, 3.0, 22.2, 116.9) | 14.0 (3.0, 5.1, 65.8, 293.9) | 6.4 (1.9, 3.1, 29.7, 129.8) | 3.7 (1.0, 1.9, 8.7, 26.3) | 5.2 (1.9, 2.6, 12.5, 59.7) | | OXBE | 2.0 (0.4, 0.8, 6.6, 23.2) | 4.0 (0.8, 1.4, 12.8, 61.8) | 0.7 (0.2, 0.4, 1.8, 5.1) | 2.2 (0.4, 1.0, 6.2, 17.6) | 2.1 (0.5, 0.9, 6.3, 32.7) | 3.0 (0.6, 1.1, 7.6, 22.4) | 1.6 (0.2, 0.6, 8.3, 23.8) | | PRPA | 0.2 (0.0, 0.0, 1.6, 14.3) | 0.7 (0.0, 0.2, 4.4, 39.2) | 0.1 (0.0, 0.0, 0.5, 5.2) | 1.4 (0.0, 0.2, 8.3, 65.4) | 0.1 (0.0, 0.0, 1.2, 13.2) | 0.0 (0.0, 0.0, 0.4, 1.7) | 0.1 (0.0, 0.0, 0.9, 6.1) | | TRCS | 0.6 (0.2, 0.3, 1.5, 4.2) | 1.2 (0.4, 0.6, 2.8, 10.9) | 0.7 (0.2, 0.4, 1.5, 5.1) | 1.0 (0.3, 0.5, 2.5, 7.6) | 0.7 (0.3, 0.4, 1.5, 2.6) | 0.2 (0.1, 0.1, 0.4, 1.0) | 0.4 (0.1, 0.2, 1.1, 3.4) | | **Phthalate metabolites** | | | | | | | | | MBzP | 4.8 (1.7, 2.7, 8.7, 16.9) | 2.5 (0.9, 1.5, 4.9, 8.9) | 7.3 (3.2, 4.9, 12.3, 25.1) | 5.2 (1.8, 3.1, 8.9, 16.2) | 6.1 (2.0, 3.5, 10.5, 20.4) | 3.6 (1.7, 2.4, 6.3, 11.2) | 6.0 (2.2, 3.5, 10.6, 21.3) | | MECPP | 32.8 (13.4, 19.9, 57.6, 98.6) | 41.3 (15.1, 22.7, 67.4, 115.0) | 21.4 (11.0, 15.3, 32.4, 45.7) | 34.4 (17.1, 24.2, 60.5, 101.6) | 45.3 (22.3, 33.6, 70.7, 103.8) | 19.0 (10.2, 13.5, 29.9, 45.8) | 62.3 (26.0, 35.3, 96.9, 146.7) | | MEHHP | 19.3 (7.1, 11.4, 33.1, 56.0) | 22.3 (8.7, 13.2, 38.7, 71.3) | 13.5 (6.6, 9.3, 20.9, 30.8) | 21.3 (10.1, 14.1, 36.8, 62.7) | 26.1 (11.7, 17.5, 37.9, 55.7) | 10.8 (5.3, 6.8, 17.6, 26.7) | 34.5 (12.1, 20.7, 54.5, 80.9) | | MEHP | 2.8 (1.0, 1.6, 5.1, 8.8) | 3.3 (0.9, 1.7, 6.1, 9.9) | 2.3 (0.9, 1.3, 3.7, 5.6) | 3.3 (1.3, 1.9, 6.0, 10.1) | 3.2 (1.1, 1.8, 4.8, 8.2) | 1.7 (0.8, 1.1, 2.8, 4.4) | 4.7 (1.5, 2.7, 8.3, 15.0) | | MEOHP | 12.2 (4.5, 7.1, 20.4, 34.0) | 13.7 (4.7, 7.9, 24.1, 44.0) | 8.3 (3.9, 5.4, 13.5, 18.4) | 12.7 (6.0, 8.5, 21.4, 37.7) | 16.7 (7.7, 11.3, 24.5, 36.4) | 6.8 (3.4, 4.5, 10.7, 16.6) | 21.1 (8.2, 12.9, 33.8, 50.4) | | MEP | 32.5 (8.8, 15.0, 79.2, 173.3) | 45.2 (12.6, 20.4, 113.8, 249.0) | 50.0 (13.1, 26.4, 97.6, 225.8) | 86.7 (23.1, 41.4, 169.5, 345.1) | 26.1 (10.1, 14.0, 54.4, 84.7) | 12.8 (5.8, 8.3, 20.9, 38.8) | 28.1 (9.3, 14.2, 62.1, 109.6) | | MiBP | 40.2 (14.6, 24.5, 71.1, 123.1) | 61.9 (24.6, 38.8, 103.2, 168.4) | 47.4 (16.3, 27.5, 82.4, 159.2) | 28.5 (13.1, 17.7, 44.6, 61.4) | 69.9 (28.7, 42.4, 108.6, 164.6) | 23.5 (9.5, 14.2, 36.5, 53.9) | 43.2 (22.0, 30.0, 71.8, 117.6) | | MnBP | 22.7 (9.5, 14.5, 38.8, 66.7) | 25.4 (8.6, 16.8, 42.0, 65.3) | 21.7 (10.1, 13.8, 37.5, 55.1) | 15.0 (6.6, 10.2, 23.2, 38.0) | 39.9 (17.7, 25.6, 60.5, 90.7) | 20.4 (10.8, 15.0, 32.6, 55.9) | 22.5 (9.7, 14.0, 38.9, 78.4) | | oh-MiNP | 5.0 (2.1, 3.1, 9.3, 17.4) | 8.2 (2.8, 4.5, 14.4, 27.6) | 4.5 (1.9, 2.9, 8.4, 14.2) | 5.4 (2.2, 3.4, 10.2, 18.0) | 4.5 (1.8, 2.6, 7.5, 14.1) | 4.4 (2.1, 2.9, 7.7, 15.3) | 5.3 (2.3, 3.3, 8.6, 14.8) | | oxo-MiNP | 2.7 (1.1, 1.7, 5.0, 9.1) | 3.1 (1.3, 1.9, 5.7, 8.4) | 2.3 (1.0, 1.5, 3.9, 6.4) | 3.1 (1.4, 2.0, 5.3, 11.7) | 2.4 (1.1, 1.5, 4.3, 8.1) | 2.2 (1.0, 1.4, 4.0, 9.3) | 3.5 (1.4, 2.1, 6.7, 9.5) | | *a*Median (10%, IQR, 90%) | | | | | | | |   Table S6. **Participants non-persistent endocrine disrupting chemicals (EDCs) concentrations, by cohort and overall (HELIX subcohort; 2013-2016).** |

### 11.1.7 Concentrations of the glucocorticosteroids

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  |  |  |  |  | | --- | --- | --- | --- | --- | --- | --- | --- | | **Characteristic** | **Overall**, N = 1,004*a* | **BIB**, N = 154*a* | **EDEN**, N = 137*a* | **INMA**, N = 205*a* | **KANC**, N = 180*a* | **MOBA**, N = 200*a* | **RHEA**, N = 128*a* | | **Glucocorticosteroid** | | | | | | | | | A | 4.3 (1.4, 2.4, 8.2, 16.4) | 4.8 (1.5, 2.8, 9.0, 18.6) | 5.1 (1.7, 2.6, 9.1, 18.0) | 3.0 (0.7, 1.6, 5.6, 9.9) | 3.8 (1.1, 2.0, 7.3, 15.1) | 4.3 (2.0, 2.7, 8.4, 14.7) | 5.9 (2.1, 3.5, 14.9, 23.6) | | Unknown | 1 | 0 | 0 | 1 | 0 | 0 | 0 | | E | 22.9 (7.7, 13.1, 38.5, 62.9) | 25.7 (8.0, 14.5, 41.4, 61.2) | 28.6 (9.1, 14.1, 42.0, 73.6) | 17.1 (6.6, 10.3, 27.4, 40.9) | 21.4 (7.4, 12.0, 33.7, 51.5) | 23.3 (8.5, 14.1, 38.1, 62.2) | 28.9 (10.0, 19.3, 59.4, 98.8) | | F | 5.5 (1.9, 3.2, 9.5, 16.4) | 6.3 (2.2, 4.2, 10.4, 16.1) | 7.8 (2.8, 4.2, 11.4, 19.8) | 4.6 (1.6, 2.9, 7.1, 11.1) | 4.9 (1.7, 2.7, 8.2, 14.9) | 5.2 (1.6, 3.0, 9.1, 15.7) | 6.2 (2.2, 3.4, 13.1, 18.7) | | Unknown | 2 | 0 | 0 | 0 | 1 | 1 | 0 | | **Glucocorticosteroid metabolite** | | | | | | | | | 11OHAndros | 234.2 (72.3, 130.3, 390.5, 565.2) | 259.7 (86.5, 151.9, 375.0, 475.1) | 413.0 (161.6, 221.7, 617.0, 786.7) | 256.7 (80.1, 142.9, 365.1, 512.1) | 163.3 (38.9, 80.7, 298.5, 421.0) | 254.4 (89.3, 151.5, 408.4, 531.4) | 165.4 (62.3, 95.9, 304.2, 470.7) | | Unknown | 3 | 0 | 0 | 0 | 3 | 0 | 0 | | 17-DO-cortolone | 57.5 (10.7, 29.1, 101.7, 158.8) | 56.1 (11.2, 32.8, 100.6, 170.9) | 76.5 (22.9, 46.0, 137.6, 236.6) | 61.3 (17.1, 32.5, 102.1, 145.7) | 43.7 (4.7, 15.1, 93.4, 133.7) | 56.4 (11.7, 26.4, 92.0, 136.1) | 51.2 (9.4, 28.5, 94.3, 148.8) | | Unknown | 2 | 0 | 0 | 0 | 1 | 0 | 1 | | 20aDHE | 16.6 (5.9, 9.7, 27.5, 40.8) | 14.2 (3.2, 7.0, 25.8, 38.1) | 25.8 (10.2, 15.1, 37.8, 49.8) | 15.6 (6.0, 10.2, 23.0, 34.0) | 14.8 (4.3, 7.7, 25.6, 40.8) | 17.5 (8.1, 11.7, 26.1, 38.1) | 14.8 (5.3, 8.7, 27.6, 41.9) | | Unknown | 11 | 7 | 0 | 0 | 4 | 0 | 0 | | 20aDHF | 6.6 (1.8, 3.3, 13.3, 23.0) | 7.2 (1.8, 3.8, 14.0, 24.0) | 10.0 (2.9, 5.7, 19.5, 31.5) | 5.5 (1.8, 3.0, 9.4, 15.7) | 4.8 (1.1, 2.2, 11.4, 21.4) | 7.4 (2.4, 4.2, 14.0, 24.1) | 6.5 (1.8, 2.9, 13.8, 23.2) | | Unknown | 7 | 4 | 0 | 0 | 3 | 0 | 0 | | 20bDHE | 9.5 (3.7, 6.2, 14.3, 20.0) | 8.7 (2.3, 4.8, 14.8, 22.2) | 13.2 (7.5, 9.7, 17.3, 22.7) | 9.0 (4.7, 6.6, 11.7, 15.6) | 8.9 (3.2, 5.1, 13.7, 19.4) | 9.0 (3.4, 5.9, 14.3, 19.1) | 8.7 (3.3, 5.3, 15.2, 21.9) | | Unknown | 17 | 14 | 0 | 0 | 3 | 0 | 0 | | 20bDHF | 15.2 (5.3, 9.1, 24.8, 37.3) | 16.5 (6.7, 10.8, 26.5, 38.5) | 19.9 (7.8, 12.0, 32.0, 44.0) | 13.0 (4.4, 8.0, 18.1, 26.9) | 14.0 (4.9, 8.5, 24.5, 36.5) | 14.2 (4.7, 8.4, 23.5, 34.2) | 14.3 (5.5, 7.9, 27.5, 42.4) | | 5a20acortol | 88.9 (31.1, 52.1, 141.6, 208.0) | 109.8 (36.4, 61.7, 177.3, 283.2) | 103.0 (35.1, 58.0, 153.8, 213.2) | 83.0 (22.8, 45.9, 118.7, 171.4) | 84.7 (28.3, 46.9, 145.9, 221.9) | 88.6 (35.7, 53.7, 138.2, 182.3) | 72.4 (29.0, 47.2, 130.2, 211.4) | | Unknown | 9 | 9 | 0 | 0 | 0 | 0 | 0 | | 5a20bcortol | 122.4 (43.8, 70.4, 185.0, 262.5) | 131.0 (41.7, 66.3, 182.3, 279.0) | 148.8 (64.2, 108.8, 226.1, 291.5) | 124.3 (41.4, 68.9, 178.8, 241.6) | 115.2 (40.5, 62.9, 189.2, 272.4) | 114.7 (44.3, 67.8, 172.7, 236.0) | 105.3 (43.3, 72.6, 175.0, 239.8) | | Unknown | 5 | 5 | 0 | 0 | 0 | 0 | 0 | | 5aTHB | 133.1 (43.3, 76.1, 222.4, 359.7) | 159.8 (59.2, 101.7, 241.3, 428.4) | 144.2 (58.7, 87.9, 255.3, 365.2) | 115.7 (34.2, 73.3, 171.7, 256.8) | 148.0 (42.6, 82.6, 245.6, 411.5) | 106.1 (38.0, 61.1, 184.9, 293.4) | 139.9 (48.8, 74.6, 260.5, 385.0) | | 5aTHE | 73.9 (22.3, 39.7, 124.0, 195.9) | 82.0 (25.7, 52.1, 145.7, 214.4) | 83.9 (27.0, 41.5, 132.7, 203.9) | 62.2 (15.5, 32.3, 97.3, 141.2) | 71.3 (23.6, 40.3, 121.7, 197.7) | 64.5 (20.9, 36.4, 103.9, 156.8) | 107.9 (28.0, 51.2, 183.2, 355.7) | | Unknown | 1 | 0 | 0 | 0 | 0 | 0 | 1 | | 5aTHF | 2,870.0 (979.4, 1,663.7, 4,389.0, 6,150.4) | 3,394.6 (1,364.9, 2,288.1, 5,308.1, 7,169.4) | 3,474.2 (1,301.5, 1,856.1, 5,253.4, 6,625.7) | 2,756.9 (804.1, 1,565.6, 3,758.3, 5,438.7) | 2,907.3 (1,017.8, 1,656.1, 4,621.2, 6,464.2) | 2,283.3 (840.7, 1,259.8, 3,454.6, 4,906.8) | 3,001.9 (1,040.7, 1,652.3, 4,613.6, 6,203.4) | | 5b20acortol | 147.7 (51.7, 83.5, 225.8, 320.2) | 177.4 (46.4, 98.9, 302.3, 430.2) | 169.7 (64.2, 91.1, 252.9, 331.8) | 141.9 (37.8, 76.6, 187.6, 271.1) | 143.0 (49.8, 80.2, 229.8, 310.1) | 143.7 (60.6, 86.6, 204.2, 289.0) | 137.7 (51.2, 79.6, 220.5, 310.0) | | Unknown | 11 | 11 | 0 | 0 | 0 | 0 | 0 | | 5b20acortolone | 641.9 (206.0, 366.0, 983.1, 1,402.2) | 638.3 (213.4, 385.0, 1,028.2, 1,432.2) | 903.7 (401.7, 574.5, 1,296.1, 1,775.1) | 654.6 (189.4, 398.7, 890.7, 1,143.2) | 518.0 (157.9, 261.2, 870.2, 1,338.1) | 580.6 (200.2, 318.0, 901.5, 1,298.0) | 629.3 (265.8, 400.9, 962.4, 1,418.5) | | 5b20bcortol | 195.7 (67.8, 120.1, 302.4, 438.7) | 242.7 (87.6, 152.0, 356.8, 502.2) | 225.2 (97.8, 142.1, 371.5, 504.5) | 199.9 (58.8, 130.5, 289.3, 407.2) | 155.8 (45.4, 88.0, 270.4, 343.5) | 186.3 (80.6, 115.5, 269.4, 391.8) | 177.5 (66.2, 113.7, 301.7, 450.8) | | Unknown | 3 | 3 | 0 | 0 | 0 | 0 | 0 | | 5b20bcortolone | 546.9 (184.8, 336.3, 837.1, 1,224.3) | 561.3 (230.0, 331.3, 889.9, 1,372.6) | 682.3 (298.5, 452.0, 1,031.1, 1,363.8) | 534.1 (163.5, 372.6, 792.7, 1,071.7) | 505.0 (144.0, 272.3, 769.3, 1,147.1) | 496.1 (170.7, 289.2, 761.3, 1,150.2) | 563.5 (209.8, 328.4, 881.5, 1,370.4) | | 5bDHF | 1.4 (0.6, 0.9, 2.0, 2.9) | 1.4 (0.7, 0.9, 2.2, 2.9) | 1.8 (1.0, 1.3, 2.6, 3.6) | 1.1 (0.4, 0.6, 1.8, 2.6) | 1.5 (0.9, 1.1, 1.9, 2.4) | 1.1 (0.4, 0.6, 1.7, 2.9) | 1.5 (0.7, 1.0, 2.1, 2.7) | | Unknown | 2 | 0 | 0 | 1 | 0 | 1 | 0 | | 5bTHB | 49.3 (17.5, 28.0, 82.7, 120.8) | 53.3 (19.4, 27.5, 98.3, 139.7) | 60.9 (22.5, 34.9, 94.5, 122.4) | 50.0 (18.2, 29.7, 73.1, 103.8) | 43.8 (15.0, 27.5, 89.7, 134.0) | 40.0 (13.6, 24.7, 65.7, 102.1) | 53.5 (18.2, 28.4, 76.7, 123.5) | | Unknown | 1 | 0 | 0 | 0 | 1 | 0 | 0 | | 5bTHE | 3,138.3 (1,060.9, 1,889.5, 4,694.0, 6,288.2) | 3,552.8 (1,452.7, 2,335.3, 4,797.4, 5,636.6) | 3,649.6 (1,594.3, 2,293.5, 5,317.1, 7,827.6) | 2,911.6 (814.5, 1,615.2, 4,050.7, 5,844.7) | 2,754.6 (904.4, 1,448.0, 3,989.3, 5,963.6) | 3,070.1 (1,099.8, 1,785.5, 4,637.7, 5,686.3) | 3,541.6 (1,127.5, 2,010.1, 5,901.3, 7,481.9) | | 5bTHF | 906.5 (303.5, 548.0, 1,416.1, 1,939.7) | 1,116.2 (384.5, 660.8, 1,644.8, 2,389.0) | 1,238.6 (527.4, 743.1, 1,578.3, 2,089.2) | 882.9 (274.3, 542.6, 1,199.8, 1,619.9) | 753.9 (218.1, 389.4, 1,258.7, 1,860.0) | 859.7 (301.2, 492.9, 1,261.3, 1,788.5) | 881.5 (402.8, 565.0, 1,441.1, 1,962.0) | | Unknown | 2 | 2 | 0 | 0 | 0 | 0 | 0 | | 6OHE | 11.9 (3.8, 6.5, 18.4, 27.7) | 13.2 (4.2, 7.6, 20.6, 33.0) | 12.2 (4.3, 6.1, 17.4, 26.2) | 9.2 (3.1, 5.3, 14.1, 20.5) | 13.1 (3.9, 7.1, 19.6, 28.6) | 11.2 (3.7, 6.4, 18.1, 25.5) | 14.3 (4.6, 8.7, 24.3, 32.5) | | 6OHF | 42.8 (11.9, 22.5, 76.7, 127.8) | 51.9 (13.4, 29.8, 93.9, 167.7) | 55.8 (15.3, 29.8, 82.3, 127.7) | 32.3 (8.8, 18.5, 53.3, 77.8) | 36.6 (11.0, 19.7, 68.7, 107.8) | 46.0 (12.3, 27.9, 82.9, 136.6) | 42.0 (13.0, 21.1, 93.2, 128.0) | | **Glucocorticosteroid precursor** | | | | | | | | | S | 0.4 (0.2, 0.3, 0.8, 1.2) | 0.5 (0.2, 0.3, 0.9, 1.4) | 0.4 (0.2, 0.3, 0.7, 1.6) | 0.6 (0.2, 0.4, 0.9, 1.2) | 0.3 (0.2, 0.2, 0.5, 0.7) | 0.4 (0.2, 0.3, 0.7, 1.2) | 0.4 (0.2, 0.2, 0.8, 1.2) | | Unknown | 94 | 6 | 5 | 12 | 9 | 51 | 11 | | **Glucocorticosteroid precursor metabolite** | | | | | | | | | 17HP | 22.3 (10.1, 15.1, 33.5, 49.3) | 17.0 (9.4, 11.1, 27.6, 38.5) | 33.2 (18.6, 23.5, 44.0, 66.0) | 20.3 (8.6, 13.2, 32.2, 48.8) | 20.3 (6.7, 10.8, 33.1, 48.2) | 23.0 (15.3, 17.5, 31.2, 43.8) | 21.8 (11.9, 15.7, 32.2, 50.8) | | Unknown | 1 | 0 | 0 | 0 | 0 | 0 | 1 | | 5bDHS | 0.3 (0.1, 0.2, 0.4, 0.6) | 0.3 (0.1, 0.2, 0.4, 0.6) | 0.3 (0.1, 0.2, 0.5, 0.7) | 0.3 (0.1, 0.2, 0.3, 0.6) | 0.2 (0.1, 0.2, 0.3, 0.4) | 0.3 (0.1, 0.2, 0.4, 0.7) | 0.3 (0.1, 0.2, 0.5, 0.8) | | Unknown | 132 | 5 | 20 | 43 | 0 | 57 | 7 | | 5bTHS | 30.7 (10.1, 18.5, 50.5, 74.3) | 35.7 (14.9, 20.7, 59.2, 98.9) | 34.5 (13.1, 19.8, 52.1, 71.4) | 27.7 (8.8, 17.6, 43.0, 60.0) | 31.3 (9.8, 18.6, 55.1, 84.1) | 26.2 (8.6, 14.2, 40.8, 59.9) | 33.7 (15.7, 20.0, 58.2, 77.9) | | Unknown | 2 | 0 | 0 | 1 | 0 | 1 | 0 | | PT | 200.6 (63.0, 112.8, 342.0, 521.9) | 149.1 (47.9, 87.6, 246.3, 379.7) | 378.8 (148.3, 230.8, 542.8, 813.5) | 253.4 (81.8, 150.0, 404.4, 648.3) | 142.2 (47.3, 82.4, 273.7, 440.3) | 176.4 (57.2, 112.9, 283.3, 378.8) | 189.4 (68.3, 104.9, 306.3, 447.8) | | **Androgen** | | | | | | | | | AED | 0.2 (0.1, 0.2, 0.3, 0.6) | 0.2 (0.2, 0.2, 0.3, 0.3) | 0.3 (0.1, 0.2, 0.5, 0.7) | 0.2 (0.1, 0.1, 0.4, 0.6) | 0.2 (0.1, 0.1, 0.3, 0.4) | 0.2 (0.1, 0.1, 0.3, 0.4) | 0.2 (0.1, 0.1, 1.1, 1.7) | | Unknown | 407 | 0 | 34 | 73 | 117 | 106 | 77 | | T | 0.5 (0.2, 0.3, 1.0, 1.6) | 0.7 (0.5, 0.5, 1.0, 1.3) | 1.0 (0.3, 0.5, 1.9, 3.3) | 0.6 (0.2, 0.3, 1.0, 1.7) | 0.3 (0.1, 0.2, 0.6, 1.0) | 0.4 (0.2, 0.3, 0.7, 1.3) | 0.4 (0.2, 0.3, 0.7, 1.2) | | Unknown | 75 | 0 | 5 | 3 | 29 | 24 | 14 | | **Androgen metabolite** | | | | | | | | | Andros | 186.0 (35.2, 78.1, 394.0, 741.4) | 148.4 (32.8, 72.0, 267.9, 446.0) | 552.2 (203.0, 308.7, 980.2, 1,491.1) | 295.4 (58.8, 129.1, 513.8, 894.7) | 98.4 (19.5, 39.6, 227.5, 388.3) | 134.7 (30.9, 63.4, 293.1, 489.2) | 110.0 (33.1, 61.6, 226.5, 406.1) | | Unknown | 1 | 0 | 0 | 0 | 1 | 0 | 0 | | Etio | 110.9 (24.6, 50.7, 237.8, 451.4) | 75.1 (19.6, 32.6, 151.0, 266.5) | 369.7 (122.7, 231.8, 561.0, 910.0) | 169.7 (44.0, 84.0, 306.1, 560.3) | 74.8 (17.6, 37.6, 122.6, 239.7) | 91.4 (19.2, 45.8, 184.0, 258.4) | 76.2 (24.6, 41.2, 147.0, 234.5) | | Unknown | 1 | 0 | 0 | 0 | 1 | 0 | 0 | | *a*Median (10%, IQR, 90%) | | | | | | | |   Table S7. **Participants glucocorticosteroids concentrations, by cohort and overall (HELIX subcohort; 2013-2016).** |

## 11.2 Tables for main results

### 11.2.1 Balancing weights: sample sizes

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  | | --- | --- | --- | | Exposure | Unadjusted | Adjusted*a* | | **Phenols** | | | | ETPA | 1,297 | 1,289 | | OXBE | 1,297 | 1,277 | | BUPA | 1,297 | 1,276 | | PRPA | 1,297 | 1,275 | | MEPA | 1,297 | 1,266 | | TRCS | 1,297 | 1,255 | | BPA | 1,297 | 1,137 | | **OP pesticide metabolites** | | | | DETP | 1,297 | 1,222 | | DEP | 1,297 | 1,222 | | DMTP | 1,297 | 1,219 | | DMP | 1,297 | 1,172 | | **Phthalate metabolites** | | | | oxo-MiNP | 1,297 | 1,199 | | oh-MiNP | 1,297 | 1,171 | | MBzP | 1,297 | 1,114 | | MEHP | 1,297 | 1,090 | | MEP | 1,297 | 1,054 | | MnBP | 1,297 | 1,035 | | MEHHP | 1,297 | 1,010 | | MEOHP | 1,297 | 1,000 | | MECPP | 1,297 | 980.7 | | MiBP | 1,297 | 927.3 | | *a*Truncated weights. | | |   Table S8. **Effective sample size before and after balancing weights estimation (exposures: endocrine disrupting chemicals (EDCs); outcome: hit reaction time standard error (HRT-SE)) (HELIX subcohort; 2013-2016).** |

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  | | --- | --- | --- | | Exposure | Unadjusted | Adjusted*a* | | **Phenols** | | | | OXBE | 976.0 | 960.1 | | PRPA | 976.0 | 956.0 | | MEPA | 976.0 | 953.7 | | BUPA | 976.0 | 952.3 | | ETPA | 976.0 | 951.7 | | TRCS | 976.0 | 942.4 | | BPA | 976.0 | 856.4 | | **OP pesticide metabolites** | | | | DEP | 976.0 | 922.1 | | DETP | 976.0 | 922.1 | | DMTP | 976.0 | 907.3 | | DMP | 976.0 | 893.3 | | **Phthalate metabolites** | | | | oh-MiNP | 976.0 | 877.9 | | oxo-MiNP | 976.0 | 873.6 | | MBzP | 976.0 | 828.8 | | MEHP | 976.0 | 827.3 | | MEP | 976.0 | 796.3 | | MEHHP | 976.0 | 784.8 | | MECPP | 976.0 | 768.1 | | MEOHP | 976.0 | 761.5 | | MnBP | 976.0 | 745.7 | | MiBP | 976.0 | 690.9 | | *a*Truncated weights. | | |   Table S9. **Effective sample size before and after balancing weights estimation (exposures: endocrine disrupting chemicals (EDCs); outcomes: glucocorticosteroids) (HELIX subcohort; 2013-2016).** |

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  | | --- | --- | --- | | Exposure | Unadjusted | Adjusted*a* | | cortisone production | 976.0 | 777.2 | | corticosterone production | 976.0 | 757.5 | | cortisol production | 976.0 | 751.5 | | *a*Truncated weights. | | |   Table S10. **Effective sample size before and after balancing weights estimation (exposures: glucocorticosteroids; outcome: hit reaction time standard error (HRT-SE)) (HELIX subcohort; 2013-2016).** |

### 11.2.2 Balancing weights: summary statistics

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  | | --- | --- | --- | |  | Median (IQR) | Range | | **Characteristic***a* | **N = 1,297***a* | **N = 1,297***a* | | **OP pesticide metabolites** | | | | DMP | 0.99 (0.73, 1.25) | 0.49, 1.50 | | DMTP | 1.00 (0.81, 1.20) | 0.59, 1.39 | | DEP | 1.01 (0.81, 1.19) | 0.59, 1.39 | | DETP | 0.99 (0.81, 1.18) | 0.61, 1.41 | | **Phenols** | | | | MEPA | 1.01 (0.90, 1.13) | 0.74, 1.25 | | ETPA | 1.01 (0.96, 1.07) | 0.88, 1.14 | | PRPA | 1.01 (0.92, 1.12) | 0.80, 1.23 | | BPA | 0.99 (0.70, 1.27) | 0.38, 1.57 | | BUPA | 1.01 (0.91, 1.11) | 0.81, 1.22 | | OXBE | 1.01 (0.92, 1.09) | 0.79, 1.21 | | TRCS | 1.01 (0.87, 1.13) | 0.68, 1.28 | | **Phthalate metabolites** | | | | MEP | 0.93 (0.61, 1.27) | 0.27, 1.77 | | MiBP | 0.91 (0.46, 1.38) | 0.05, 1.92 | | MnBP | 0.98 (0.59, 1.33) | 0.20, 1.74 | | MBzP | 0.98 (0.66, 1.27) | 0.35, 1.62 | | MEHP | 0.98 (0.64, 1.28) | 0.31, 1.68 | | MEHHP | 0.96 (0.54, 1.35) | 0.16, 1.76 | | MEOHP | 0.96 (0.52, 1.35) | 0.16, 1.78 | | MECPP | 0.95 (0.50, 1.34) | 0.14, 1.84 | | oh-MiNP | 1.01 (0.74, 1.24) | 0.47, 1.51 | | oxo-MiNP | 1.01 (0.78, 1.20) | 0.52, 1.43 | | *a*Truncated weights. | | |   Table S11. **Summary statistics of the estimated balancing weights (exposures: endocrine disrupting chemicals (EDCs); outcome: hit reaction time standard error (HRT-SE)) (HELIX subcohort; 2013-2016).** |

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  | | --- | --- | --- | |  | Median (IQR) | Range | | **Characteristic***a* | **N = 976***a* | **N = 976***a* | | **OP pesticide metabolites** | | | | DMP | 0.99 (0.75, 1.23) | 0.51, 1.46 | | DMTP | 1.00 (0.78, 1.23) | 0.56, 1.41 | | DEP | 0.99 (0.81, 1.20) | 0.64, 1.41 | | DETP | 0.99 (0.82, 1.18) | 0.62, 1.41 | | **Phenols** | | | | MEPA | 1.00 (0.90, 1.13) | 0.75, 1.26 | | ETPA | 1.02 (0.90, 1.14) | 0.72, 1.24 | | PRPA | 1.00 (0.92, 1.12) | 0.76, 1.26 | | BPA | 1.00 (0.70, 1.26) | 0.40, 1.58 | | BUPA | 1.01 (0.90, 1.13) | 0.75, 1.27 | | OXBE | 1.01 (0.92, 1.10) | 0.78, 1.21 | | TRCS | 1.01 (0.86, 1.14) | 0.68, 1.29 | | **Phthalate metabolites** | | | | MEP | 0.92 (0.60, 1.27) | 0.28, 1.74 | | MiBP | 0.88 (0.44, 1.38) | 0.09, 1.98 | | MnBP | 0.97 (0.52, 1.35) | 0.14, 1.84 | | MBzP | 0.94 (0.68, 1.29) | 0.35, 1.68 | | MEHP | 0.98 (0.65, 1.29) | 0.33, 1.64 | | MEHHP | 0.98 (0.56, 1.35) | 0.21, 1.69 | | MEOHP | 0.98 (0.53, 1.35) | 0.18, 1.77 | | MECPP | 0.96 (0.55, 1.36) | 0.19, 1.76 | | oh-MiNP | 0.99 (0.73, 1.26) | 0.45, 1.49 | | oxo-MiNP | 1.01 (0.71, 1.25) | 0.45, 1.52 | | *a*Truncated weights. | | |   Table S12. **Summary statistics of the estimated balancing weights (exposures: endocrine disrupting chemicals (EDCs); outcomes: glucocorticosteroids) (HELIX subcohort; 2013-2016).** |

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  | | --- | --- | --- | |  | Median (IQR) | Range | | **Characteristic***a* | **N = 976***a* | **N = 976***a* | | cortisol production | 1.00 (0.54, 1.39) | 0.14, 1.80 | | cortisone production | 1.00 (0.58, 1.39) | 0.19, 1.73 | | corticosterone production | 0.98 (0.56, 1.39) | 0.15, 1.78 | | *a*Truncated weights. | | |   Table S13. **Summary statistics of the estimated balancing weights (exposures: glucocorticosteroids; outcome: hit reaction time standard error (HRT-SE)) (HELIX subcohort; 2013-2016).** |

## 11.3 Tables for other results

### 11.3.1 Balancing weights for effect modification: summary statistics

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  |  | | --- | --- | --- | --- | --- | |  | Median (IQR) | Range | | | | **Characteristic***a* | **females** N = 587*a* | **males** N = 710*a* | **females** N = 587*a* | **males** N = 710*a* | | **OP pesticide metabolites** | | | | | | DMP | 0.99 (0.74, 1.25) | 1.00 (0.74, 1.25) | 0.53, 1.46 | 0.53, 1.46 | | DMTP | 1.00 (0.79, 1.22) | 1.02 (0.82, 1.20) | 0.58, 1.38 | 0.58, 1.38 | | DEP | 1.01 (0.82, 1.19) | 1.02 (0.84, 1.17) | 0.64, 1.36 | 0.64, 1.36 | | DETP | 1.00 (0.77, 1.22) | 1.01 (0.82, 1.20) | 0.57, 1.39 | 0.57, 1.39 | | **Phenols** | | | | | | MEPA | 1.02 (0.89, 1.15) | 1.02 (0.94, 1.11) | 0.76, 1.23 | 0.76, 1.23 | | ETPA | 1.02 (0.96, 1.08) | 1.01 (0.97, 1.06) | 0.91, 1.12 | 0.91, 1.12 | | PRPA | 1.02 (0.92, 1.13) | 1.02 (0.95, 1.10) | 0.82, 1.21 | 0.82, 1.21 | | BPA | 1.02 (0.73, 1.28) | 1.02 (0.74, 1.25) | 0.41, 1.50 | 0.41, 1.50 | | BUPA | 1.02 (0.94, 1.10) | 1.01 (0.81, 1.20) | 0.67, 1.29 | 0.67, 1.29 | | OXBE | 1.03 (0.92, 1.12) | 1.02 (0.94, 1.09) | 0.81, 1.19 | 0.81, 1.19 | | TRCS | 1.03 (0.92, 1.13) | 1.01 (0.89, 1.12) | 0.73, 1.25 | 0.73, 1.25 | | **Phthalate metabolites** | | | | | | MEP | 0.96 (0.67, 1.26) | 0.93 (0.62, 1.30) | 0.31, 1.67 | 0.31, 1.67 | | MiBP | 0.93 (0.51, 1.39) | 0.96 (0.51, 1.40) | 0.16, 1.85 | 0.16, 1.85 | | MnBP | 1.00 (0.62, 1.33) | 0.98 (0.59, 1.35) | 0.28, 1.69 | 0.28, 1.69 | | MBzP | 1.00 (0.70, 1.27) | 0.99 (0.69, 1.27) | 0.40, 1.57 | 0.40, 1.57 | | MEHP | 1.02 (0.69, 1.28) | 0.98 (0.62, 1.32) | 0.33, 1.62 | 0.33, 1.62 | | MEHHP | 1.01 (0.60, 1.30) | 0.95 (0.55, 1.36) | 0.26, 1.72 | 0.26, 1.72 | | MEOHP | 1.00 (0.63, 1.29) | 0.95 (0.52, 1.40) | 0.23, 1.74 | 0.23, 1.74 | | MECPP | 1.00 (0.59, 1.33) | 0.95 (0.50, 1.37) | 0.23, 1.76 | 0.23, 1.76 | | oh-MiNP | 1.02 (0.78, 1.22) | 1.00 (0.76, 1.23) | 0.51, 1.46 | 0.51, 1.46 | | oxo-MiNP | 1.02 (0.83, 1.17) | 1.01 (0.76, 1.21) | 0.58, 1.39 | 0.58, 1.39 | | *a*Truncated weights. | | | | |   Table S14. **Summary statistics of the estimated balancing weights for effect modification (exposures: endocrine disrupting chemicals (EDCs); outcome: hit reaction time standard error (HRT-SE); modifier: sex) (HELIX subcohort; 2013-2016).** |

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  |  | | --- | --- | --- | --- | --- | |  | Median (IQR) | Range | | | | **Characteristic***a* | **females** N = 434*a* | **males** N = 542*a* | **females** N = 434*a* | **males** N = 542*a* | | **OP pesticide metabolites** | | | | | | DMP | 0.98 (0.77, 1.23) | 1.01 (0.76, 1.21) | 0.57, 1.45 | 0.57, 1.45 | | DMTP | 1.03 (0.78, 1.22) | 1.00 (0.79, 1.23) | 0.56, 1.40 | 0.56, 1.40 | | DEP | 1.01 (0.85, 1.16) | 1.00 (0.84, 1.19) | 0.67, 1.36 | 0.67, 1.36 | | DETP | 1.00 (0.77, 1.22) | 1.01 (0.86, 1.17) | 0.57, 1.40 | 0.57, 1.40 | | **Phenols** | | | | | | MEPA | 1.01 (0.88, 1.17) | 1.03 (0.94, 1.11) | 0.73, 1.26 | 0.73, 1.26 | | ETPA | 1.04 (0.92, 1.12) | 1.02 (0.91, 1.12) | 0.78, 1.22 | 0.78, 1.22 | | PRPA | 1.03 (0.87, 1.16) | 1.02 (0.95, 1.10) | 0.74, 1.24 | 0.74, 1.24 | | BPA | 1.00 (0.71, 1.29) | 1.01 (0.75, 1.24) | 0.44, 1.52 | 0.44, 1.52 | | BUPA | 1.02 (0.95, 1.11) | 1.01 (0.80, 1.20) | 0.64, 1.30 | 0.64, 1.30 | | OXBE | 1.03 (0.86, 1.16) | 1.02 (0.95, 1.09) | 0.76, 1.22 | 0.76, 1.22 | | TRCS | 1.03 (0.92, 1.13) | 1.01 (0.87, 1.14) | 0.73, 1.25 | 0.73, 1.25 | | **Phthalate metabolites** | | | | | | MEP | 0.99 (0.70, 1.24) | 0.95 (0.55, 1.30) | 0.31, 1.68 | 0.31, 1.68 | | MiBP | 0.92 (0.45, 1.40) | 0.92 (0.54, 1.39) | 0.15, 1.85 | 0.15, 1.85 | | MnBP | 0.97 (0.51, 1.41) | 0.98 (0.57, 1.32) | 0.21, 1.78 | 0.21, 1.78 | | MBzP | 0.99 (0.70, 1.26) | 0.98 (0.66, 1.31) | 0.38, 1.58 | 0.38, 1.58 | | MEHP | 1.01 (0.72, 1.29) | 0.98 (0.61, 1.34) | 0.36, 1.58 | 0.36, 1.58 | | MEHHP | 1.02 (0.64, 1.31) | 1.00 (0.59, 1.35) | 0.30, 1.63 | 0.30, 1.63 | | MEOHP | 1.01 (0.62, 1.32) | 1.01 (0.50, 1.41) | 0.24, 1.68 | 0.24, 1.68 | | MECPP | 0.98 (0.62, 1.32) | 0.98 (0.53, 1.40) | 0.29, 1.67 | 0.29, 1.67 | | oh-MiNP | 1.00 (0.73, 1.26) | 1.00 (0.78, 1.24) | 0.49, 1.44 | 0.49, 1.44 | | oxo-MiNP | 1.03 (0.73, 1.27) | 1.02 (0.76, 1.24) | 0.47, 1.45 | 0.47, 1.45 | | *a*Truncated weights. | | | | |   Table S15. **Summary statistics of the estimated balancing weights for effect modification (exposures: endocrine disrupting chemicals (EDCs); outcomes: glucocorticosteroids; modifier: sex) (HELIX subcohort; 2013-2016).** |

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  |  | | --- | --- | --- | --- | --- | |  | Median (IQR) | Range | | | | **Characteristic***a* | **females** N = 434*a* | **males** N = 542*a* | **females** N = 434*a* | **males** N = 542*a* | | cortisol production | 0.97 (0.57, 1.41) | 1.01 (0.58, 1.35) | 0.24, 1.72 | 0.24, 1.72 | | cortisone production | 1.00 (0.61, 1.40) | 1.00 (0.59, 1.38) | 0.26, 1.69 | 0.26, 1.69 | | corticosterone production | 1.00 (0.60, 1.39) | 1.03 (0.56, 1.37) | 0.23, 1.71 | 0.23, 1.71 | | *a*Truncated weights. | | | | |   Table S16. **Summary statistics of the estimated balancing weights for effect modification (exposures: glucocorticosteroids; outcome: hit reaction time standard error (HRT-SE); modifier: sex) (HELIX subcohort; 2013-2016).** |

### 11.3.2 Marginal contrasts for effect modification

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  | | --- | --- | --- | |  | HRT-SE males*a* | HRT-SE females*a* | | **OP pesticide metabolites** | | | | dep | -0.001 (-0.032, 0.031) | -0.02 (-0.066, 0.027) | | detp | -0.016 (-0.064, 0.031) | -0.041 (-0.091, 0.008) | | dmp | -0.009 (-0.069, 0.051) | 0.024 (-0.032, 0.081) | | dmtp | 0.008 (-0.082, 0.099) | 0.003 (-0.083, 0.089) | | **Phenols** | | | | bpa | 0.028 (-0.009, 0.065) | -0.004 (-0.055, 0.047) | | bupa | -0.009 (-0.048, 0.03) | 0.012 (-0.027, 0.052) | | etpa | 0.013 (-0.027, 0.054) | 0.001 (-0.037, 0.039) | | mepa | 0.04 (-0.004, 0.083) | 0.041 (0.01, 0.072) | | oxbe | 0.034 (-0.011, 0.08) | 0.002 (-0.028, 0.031) | | prpa | 0.017 (-0.05, 0.084) | 0.002 (-0.047, 0.052) | | trcs | 0.011 (-0.022, 0.043) | 0.028 (-0.058, 0.113) | | **Phthalate metabolites** | | | | mbzp | -0.01 (-0.037, 0.018) | 0.057 (0.009, 0.104) | | mecpp | 0.034 (-0.037, 0.104) | 0.026 (-0.01, 0.061) | | mehhp | 0.057 (-0.022, 0.136) | 0.028 (-0.017, 0.074) | | mehp | 0.06 (-0.011, 0.131) | 0.043 (0.004, 0.083) | | meohp | 0.067 (-0.016, 0.151) | 0.047 (0.016, 0.079) | | mep | 0.002 (-0.029, 0.033) | 0.054 (-0.005, 0.114) | | mibp | 0.017 (-0.083, 0.117) | 0.037 (-0.029, 0.104) | | mnbp | -0.011 (-0.07, 0.047) | 0.023 (-0.037, 0.083) | | ohminp | 0.058 (-0.003, 0.118) | 0.011 (-0.007, 0.03) | | oxominp | 0.017 (-0.013, 0.046) | 0.042 (0.034, 0.051) | | *a*Estimate and 95% CI. | | |   Table S17. **Marginal contrasts on the logarithmic scale for effect modification by sex of a increase from the 10th to the 90th percentile of the endocrine disrupting chemicals (EDCs) on hit reaction time standard error (HRT-SE) expressed in ms (HELIX subcohort; 2013-2016).** |

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| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| |  |  |  |  |  |  |  | | --- | --- | --- | --- | --- | --- | --- | |  | corticosterone production males*a* | corticosterone production females*a* | cortisol production males*a* | cortisol production females*a* | cortisone production males*a* | cortisone production females*a* | | **OP pesticide metabolites** | | | | | | | | dep | 0.127 (0.003, 0.252) | 0.209 (0.035, 0.383) | 0.105 (-0.049, 0.26) | 0.244 (0.092, 0.396) | 0.11 (-0.024, 0.243) | 0.214 (0.08, 0.348) | | detp | -0.026 (-0.087, 0.036) | 0.135 (-0.053, 0.323) | -0.046 (-0.13, 0.038) | 0.025 (-0.157, 0.207) | -0.024 (-0.117, 0.07) | 0.073 (-0.104, 0.249) | | dmp | 0.07 (-0.051, 0.192) | 0.063 (-0.105, 0.231) | 0.029 (-0.083, 0.14) | 0.059 (-0.021, 0.14) | 0.065 (-0.035, 0.165) | 0.134 (0.025, 0.243) | | dmtp | 0.067 (-0.03, 0.163) | 0.08 (-0.053, 0.214) | -0.026 (-0.084, 0.032) | 0.184 (0.073, 0.296) | 0.063 (-0.024, 0.149) | 0.228 (0.087, 0.37) | | **Phenols** | | | | | | | | bpa | 0.218 (0.027, 0.409) | 0.372 (0.177, 0.566) | 0.22 (0.024, 0.416) | 0.345 (0.176, 0.515) | 0.229 (0.082, 0.376) | 0.314 (0.159, 0.469) | | bupa | -0.109 (-0.22, 0.001) | 0.008 (-0.09, 0.106) | -0.086 (-0.18, 0.008) | 0.043 (-0.062, 0.148) | -0.048 (-0.137, 0.04) | -0.035 (-0.138, 0.068) | | etpa | 0.084 (-0.029, 0.196) | 0.337 (0.18, 0.494) | 0.08 (-0.031, 0.191) | 0.264 (0.105, 0.422) | 0.076 (-0.02, 0.172) | 0.295 (0.091, 0.499) | | mepa | 0.022 (-0.047, 0.092) | 0.151 (0.006, 0.296) | 0.013 (-0.059, 0.085) | 0.14 (0.047, 0.233) | 0.047 (-0.013, 0.108) | 0.191 (0.063, 0.319) | | oxbe | -0.091 (-0.258, 0.076) | 0.122 (-0.034, 0.279) | -0.022 (-0.192, 0.149) | 0.055 (-0.085, 0.194) | -0.023 (-0.188, 0.141) | 0.041 (-0.067, 0.149) | | prpa | -0.013 (-0.109, 0.083) | 0.107 (-0.042, 0.256) | 0.009 (-0.078, 0.097) | 0.053 (-0.112, 0.218) | -0.01 (-0.067, 0.046) | 0.091 (-0.048, 0.231) | | trcs | 0.08 (-0.077, 0.238) | 0.223 (0.06, 0.385) | 0.036 (-0.104, 0.176) | 0.166 (0.012, 0.319) | 0 (-0.094, 0.094) | 0.152 (0.015, 0.289) | | **Phthalate metabolites** | | | | | | | | mbzp | 0.101 (-0.096, 0.297) | 0.126 (-0.104, 0.356) | 0.062 (-0.145, 0.269) | 0.08 (-0.12, 0.279) | 0.019 (-0.186, 0.224) | 0.098 (-0.021, 0.217) | | mecpp | 0.208 (-0.001, 0.418) | 0.222 (-0.035, 0.48) | 0.13 (-0.057, 0.317) | 0.173 (-0.028, 0.373) | 0.148 (-0.042, 0.339) | 0.131 (-0.058, 0.321) | | mehhp | 0.192 (0.014, 0.369) | 0.243 (-0.013, 0.5) | 0.112 (-0.093, 0.318) | 0.203 (-0.011, 0.418) | 0.144 (-0.027, 0.315) | 0.15 (-0.021, 0.321) | | mehp | 0.202 (0.059, 0.345) | 0.367 (0.216, 0.518) | 0.116 (-0.044, 0.277) | 0.337 (0.176, 0.499) | 0.136 (-0.006, 0.278) | 0.313 (0.188, 0.439) | | meohp | 0.174 (-0.004, 0.352) | 0.235 (0.001, 0.469) | 0.1 (-0.092, 0.291) | 0.175 (-0.012, 0.361) | 0.113 (-0.052, 0.278) | 0.104 (-0.044, 0.253) | | mep | 0.03 (-0.246, 0.306) | 0.08 (-0.036, 0.196) | -0.004 (-0.233, 0.225) | 0.079 (-0.048, 0.206) | -0.048 (-0.261, 0.165) | 0.071 (-0.039, 0.181) | | mibp | 0.153 (0.036, 0.27) | 0.115 (-0.081, 0.312) | 0.043 (-0.062, 0.148) | 0.085 (-0.143, 0.313) | -0.041 (-0.145, 0.063) | -0.02 (-0.147, 0.107) | | mnbp | 0.153 (-0.054, 0.359) | 0.124 (-0.182, 0.43) | 0.126 (-0.042, 0.294) | 0.063 (-0.256, 0.382) | 0.1 (-0.07, 0.27) | 0.082 (-0.116, 0.281) | | ohminp | 0.102 (0.045, 0.159) | 0.229 (0.026, 0.432) | 0.051 (-0.065, 0.167) | 0.232 (0.083, 0.382) | 0.046 (-0.027, 0.12) | 0.21 (0.074, 0.347) | | oxominp | 0.073 (0.022, 0.123) | 0.192 (0.021, 0.363) | 0.035 (-0.101, 0.17) | 0.181 (0.097, 0.264) | 0.021 (-0.064, 0.106) | 0.148 (0.069, 0.227) | | *a*Estimate and 95% CI. | | | | | | |   Table S18. **Marginal contrasts on the logarithmic scale for effect modification by sex of a increase from the 10th to the 90th percentile of the endocrine disrupting chemicals (EDCs) on the glucocorticosteroids expressed in ng/ml (HELIX subcohort; 2013-2016).** |

# 12. Supplementary figures

## 12.1 Figures for descriptive data

### 12.1.1 Study populations

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| Figure S1. Flowchart describing the sample size for each research question. |

### 12.1.2 Description of endocrine disruptors

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| Figure S2. **Measurement classification of endocrine disrupting chemicals (EDCs), by cohort (HELIX subcohort; 2013-2016).** Coding: 1, quantifiable; 2, <LOD; 3, interference or out of range; 4. not analysed. |

### 12.1.3 Description of glucocorticosteroids

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| Figure S3. **Measurement classification of the glucocorticosteroids, by cohort (HELIX subcohort; 2013-2016).** Coding: 1, quantifiable; 2, <LOQ; 3, interference or out of range; 4, not detected. |

## 12.2 Figures for other results

### 12.2.1 Marginal contrasts for effect modification

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| Figure S4. **Marginal contrasts on the logarithmic scale for effect modification by sex of a increase from the 10th to the 90th percentile of the endocrine disrupting chemicals (EDCs) on hit reaction time standard error (HRT-SE) expressed in ms (HELIX subcohort; 2013-2016).** Circles and triangles indicate effect estimates. Solid lines indicate the CI. The size of the circles represents the S value of the effect estimate ([Rafi and Greenland, 2020](#ref-RafiGreenland:2020)). |

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| Figure S5. **Marginal contrasts on the logarithmic scale for effect modification by sex of a increase from the 10th to the 90th percentile of the endocrine disrupting chemicals (EDCs) on the glucocorticosteroids expressed in ng/ml (HELIX subcohort; 2013-2016).** Circles and triangles indicate effect estimates. Solid lines indicate the CI. The size of the circles represents the S value of the effect estimate ([Rafi and Greenland, 2020](#ref-RafiGreenland:2020)). |