



Estimating agricultural water productivity using remote sensing derived data

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Abstract

The 2030 Agenda aims at ending extreme poverty, inequality, injustice and climate change. Progress is evaluated through a set of Sustainable Development Goals (SDGs), targets and indicators. However, there are various challenges affecting regular and timely reporting. Remote sensing (RS) derived data has been shown to provide a valuable complementary data source in reporting SDGs. This study focuses on how RS derived data could support SDG 6 related to water, and in particular SDG indicator 6.4.1 - change in Water Use Efficiency (WUE) over time presented in USD per m³ of water withdrawn. Although water withdrawals cannot be monitored through RS, water use in agriculture, globally withdrawing the largest amount of water, can be monitored through RS based evapotranspiration.

Two approaches were modelled to compute the progress of SDG 6.4.1 in the agricultural sector. The first approach uses the standard equation of SDG 6.4.1, replacing water withdrawal with blue evapotranspiration in the irrigation sector. The second approach distributes the gross value added to the gross domestic product by irrigated agriculture according to the land productivity in irrigated agriculture as observed by RS. The results of these two approaches were compared to the standard way SDG 6.4.1 is calculated. The analyses were carried out for Lebanon, which faces critical water challenges while experiencing a difficult economic and political situation.

The results for Lebanon show that the different approaches to estimate A_{wp} show similar trends as A_{we} , initially showing an increasing trend followed by a sharp decline in 2019 due to the deteriorating economic situation in the country. However, the absolute values differ substantially, largely due to discrepancies between the estimated irrigated area from RS data and the static data reported in AQUASTAT. The results illustrate the spatial variability of A_{wp} in Lebanon, with the area that contributes significantly to the agricultural production nationally (Bekaa and Baalbek) shows lower land and water productivity compared to irrigated areas in other governorates. The contribution of agriculture to the overall SDG 6.4.1 indicator was relatively small, although agriculture is a major consumer of water.

Keywords Water productivity · Remote sensing · SDG · WaPOR · Earth Observation

Introduction

In 2015, the United Nations (UN) together with the heads of the state of 193 countries agreed to adopt the 2030 Agenda as a response to global challenges such as population growth, changing lifestyles, and climate change which deplete resources and stress the earth system (Bhaduri et al.

2016). They were built on the achievements of the Millennium Development Goals (MDGs) and are far more ambitious in scope of global development (Purvis et al. 2019). It consists of 17 Sustainable Development Goals (SDGs), 169 targets and 231 indicators (UN 2017; WWAP and UN-Water 2018; Eisenmenger et al. 2020). The 2030 Agenda now also incorporates a specific target focussing on achieving sustainable management of clean water and sanitation (SDG 6) (UN 2015). This is in recognition of the fact that water availability supports economic growth and the wellbeing of people and has been identified as one of the main challenges for human survival (Steffen et al. 2015).

SDG 6 consists of two indicators, one focusses on overall water scarcity (SDG indicator 6.4.2) and the other one

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focusses on efficient use of water from an economic perspective (SDG indicator 6.4.1). These indicators take into consideration water withdrawals of three main economic sectors – agriculture, industries and services and compare them to their economic contribution to the economy (6.4.1) and the total water available (6.4.2) (UN-Water 2021). SDG indicator 6.4.1 is a composite of the water use efficiency (WUE) of the three sectors, expressed into economic value added per unit of water withdrawn (in USD/m³). As a composite, this indicator is dominated by the high value, low water use sectors such as industry and therefore to be able to provide sector specific recommendations, literature suggests to evaluate each sector's water use efficiency separately (Hellegers and van Halsema 2021).

Monitoring WUE for the agricultural sector, and thus Agricultural Water Use Efficiency (A_{we}), is also important for SDG indicator 6.4.2 as it is responsible for over 70% of the globally withdrawn water (FAO 2015). Even though the methodology for monitoring SDG indicator 6.4.1 has been clearly defined (FAO 2019), there is an ongoing debate on the rationale for using water withdrawn as it doesn't take into consideration return flows which could be used elsewhere or is returned to the environment (Hellegers and van Halsema 2021; Vanham 2018). An alternative recommended approach therefore is to evaluate the agricultural productivity against the water consumed in the form of evapotranspiration (ET) (Hellegers and van Halsema 2021), which is also known as agricultural water productivity. One of the discrepancies between considering water withdrawals compared to water consumed is that the latter includes water stemming from rainfall. It is therefore important to consider the water consumed from irrigation only (also called blue ET (ET_b)) (Chukalla et al. 2015).

The progress of the SDGs is monitored by the countries themselves through comprehensive data collection (Connor 2015; Reyers et al. 2017). However, this is a costly effort and most countries do not collect this information on a regular basis (Roopnarine et al. 2019). In 2019, OECD (2019) reported that on average countries reported on just 55% of the SDG indicators since 2015, with only few countries providing data for all years. The low level of progress compared to the MDG monitoring, which was 68% in 2015, can be explained by the fact that the SDGs has more indicators (231 vs. 48) and more complex computing workflows (MacFeely 2018).

Remote Sensing (RS) is considered as a reliable source of data which can support the monitoring and reporting of the SDG indicators on a regular and near-real time basis (Calera et al. 2017; Hakimdavar et al. 2020; Kavvada et al. 2020). FAO, one of the custodian agencies for reporting several SDGs is therefore actively promoting the uptake of amongst other data, the use of Earth Observation (Gennari

et al. 2019). Open data platforms such as Global surface water explorer (Pekel et al. 2016) and the Freshwater Ecosystems Explorer (Lu et al. 2021) contribute significantly to the periodical estimation of indicators related to SDG 6. Recent publications demonstrate the use of satellite data in mapping of wetlands contributing to SDG 6.6.1 (Weise et al. 2020), mapping land degradation to support SDG 15 (Giuliani et al. 2020), and collect evidence on slavery from space supporting SDG 8 (Foody et al. 2019; Boyd et al. 2018). RS based ET has been used to monitor various aspects of water management such as estimation of water allocation to various sectors (Droogers et al. 2010) and irrigation performance assessment (Karimi et al. 2019), however, the use of RS based ET for monitoring the SDGs is still relatively new (Biancalani and Marinelli 2021). There are various RS based ET products available which can be utilised to support reporting of SDG 6 indicators (e.g. SSE-Bop: Senay et al. 2013; CMRSET: Guerschman et al. 2009; SEBS: Chen et al. 2013; MOD16: Mu et al. 2011).

There are multiple global initiatives globally to coordinate the efforts of monitoring SDG's using RS data. One of the major initiatives is Group on Earth Observations (GEO) which is a partnership of more than 100 national governments and organizations with major objective being main streaming Earth observation data for decisions and actions towards sustainable development (Anderson et al. 2017). GEO initiated the development of Global Earth Observation System of Systems (GEOSS: Nativi et al. 2020) with the Committee on Earth Observation Satellites (CEOS) contributing to the technical side of the development including initiatives to promote RS for SDG monitoring such as EO4SDG (Kavvada et al. 2022).

In this paper, we investigate how RS can support monitoring of SDG 6.4.1 on water use efficiency. We modelled the agricultural component of SDG 6.4.1. with RS data using two different approaches: (i) using volume of consumptive water use (actual evapotranspiration) and (ii) using the water productivity concept. We applied both approaches to the country Lebanon as a case study and compared the results to the standard calculations using AQUASTAT data.

Methodology

Case study: Lebanon

Lebanon has an area of 10,452 km² and is divided into 7 governorates (Fig. 1). The annual rainfall ranges between 700 and 1,500 mm/year, averaging 910 mm/year. Lebanon is considered rich in water resources compared to its surrounding countries, with 12 permanent watercourses with a

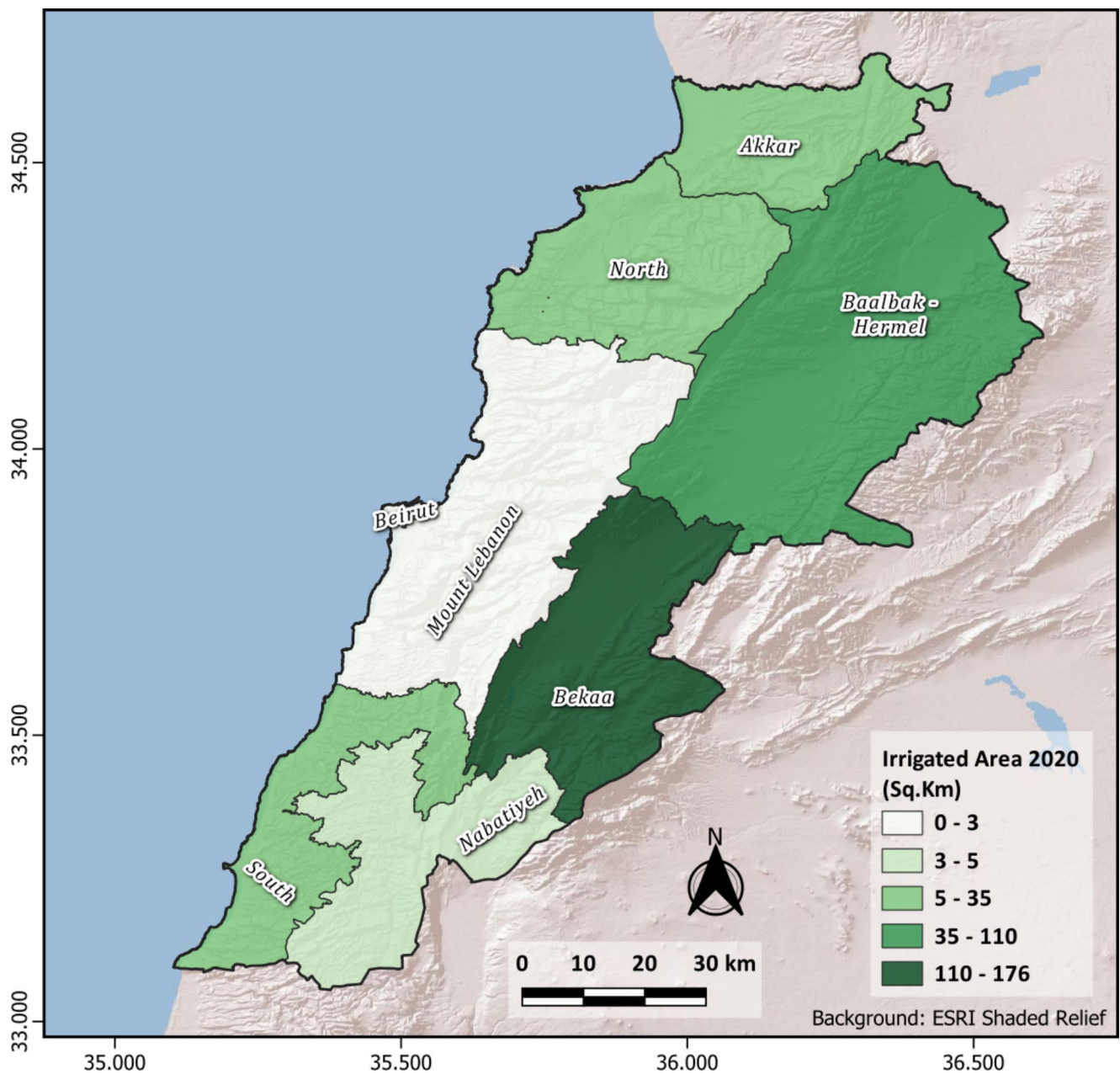


Fig. 1 Map of Lebanon showing distribution of irrigated areas in governorates in 2020. (source of data: WaPOR)

discharge of 3,542 million m^3/year along with 2,000 major springs which provide 630 million m^3/year (Shaban 2020).

Agricultural lands occupy more than one third of the area ($3,600 \text{ km}^2$), of which 1040 km^2 is equipped for irrigation and 900 km^2 was actually irrigated as recorded in the 1998 census (FAO 2008). About 42% this area is located in the Bekaa Valley (FAO 2008). The main irrigated crops are wheat (24%), potatoes (18%), citrus (16%) and vegetables (14%). The source of water for irrigation is evenly distributed between surface water (48%) and groundwater (52%) (Shaban 2014). The agricultural sector contributes 5% of

GDP and employs 8% of the labour force (FAO-Lebanon 2022).

Methods

This study focuses on modelling the agricultural component of the SDG indicator 6.4.1 (A_{we}) using two different RS-based approaches (A_{wp}) and compare the results to the standard equation of A_{we} . The analysis was conducted using annual timesteps for a 5-year period starting with the baseline for the SDG in 2015 to 2020. We used a geospatial modelling approach (Chen et al. 2020; Koubarakis 2023)

to provide a comprehensive analysis of water use efficiency and productivity in agricultural systems.

The standard equation for the agricultural component of SDG indicator 6.4.1 (A_{we}) is defined in FAO (2019) as:

$$A_{we} = \frac{GVA_a(1 - c_r)}{V_a} \quad (1)$$

Where:

A_{we} represents the water use efficiency of irrigated agriculture in USD/m³.

GVA_a is the Gross Value Added to gross domestic product (GDP) by agriculture in USD/year. c_r is the proportion of agricultural GVA_a produced by rainfed agriculture.

V_a is the volume of water used by the agricultural sector in m³/year.

The equation focusses on the gross value added to GDP by irrigated agriculture (represented by $GVA_a^* (1 - c_r)$) as this is where the water withdrawals contribute to. This ensures that the A_{we} calculation focuses solely on the efficiency of the water withdrawals in agriculture.

For the RS-based approaches for modelling the agricultural component of SDG 6.4.1, we developed two approaches. The first approach modifies Eq. 1 by replacing the water withdrawn (V_a) with the total volume of the blue water consumed (V_{ETb}) derived from water supplied through irrigation (Eq. 2).

$$A_{wp1} = \frac{GVA_a(1 - C_r)}{V_{ETb}} \quad (2)$$

Where: A_{wp1} is the water productivity of irrigated agricultural in USD/m³ using approach 1

V_{ETb} is the volume of water consumed by irrigation (blue ET) in m³/year.

To estimate V_{ETb} the first step is to separate ET into water from rainfall (ET green) and irrigation (ET_b) (Chukalla et al. 2015). Various models exist to separate ET, which have different level of complexity and without clear validation data it is not possible to identify the best model (Msigwa et al. 2021). Therefore, for this purpose, we used a rather simple model by considering effective rainfall as ET green with the remainder of ET to be attributed to ET_b. The effective rainfall and ET_b is calculated on a monthly basis and summed to an annual value using Eq. 3:

$$ET_b = \sum_{n=1}^{12} \max(ET - P_e, 0) \quad (3)$$

Where:

ET_b is in mm/year,

ET is in mm/month, and

P_e is effective rainfall in mm/month.

P_e is computed from monthly precipitation by applying the conditions in outlined in the Eq. 4, adapted from Brouwer and Heibloem (1986):

$$P_e = \max(0.8 * P - 25, 0), \text{ if } P > 75 \text{ mm/month} \quad (4)$$

$$P_e = \max(0.6 * P - 10, 0), \text{ if } P \leq 75 \text{ mm/month}$$

Annual ET_b is then converted to volume (V_{ETb}) in m³/year by multiplying the average ET_b of the irrigated areas ($\overline{ET_b}$) with the irrigated area (A_{irr}) as given in Eq. 5.

$$V_{ETb} = \overline{ET_b} \times A_{irr} \quad (5)$$

The second approach starts with calculating Biomass Water Productivity (WP_b) at pixel scale in kg/m³. WP_b is derived from dividing the Total Biomass Production or Land Productivity (LP in kg/ha) by the irrigation water consumed (ET_b). This is represented by the following equation (Eq. 6):

$$WP_b = \frac{LP}{ET_b} \quad (6)$$

WP_b is converted to economic water productivity by multiplying the average price per unit of annual biomass production from irrigated agriculture as presented in the following equation (Eq. 7):

$$A_{wp2} = WP_b \times \bar{p} \quad (7)$$

Where:

A_{wp2} is the agriculture water productivity in USD/m³ using approach 2.

WP_b is the Biomass Water Productivity in kg/m³.

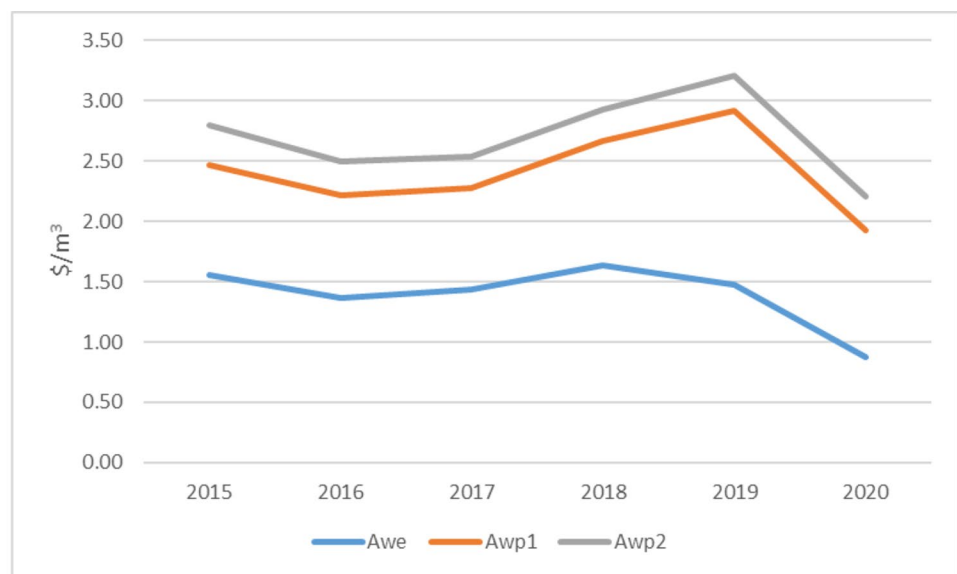
\bar{p} is the average value added in USD per kg of annual biomass production

The average price is obtained by dividing GVA_a of irrigated agriculture with the total LP of the irrigated area. This second approach is modelled at the grid or pixel level, which fills the gap that national level indicators are not able to provide all the required information for decision makers (Giuopponi et al. 2018; Saner et al. 2020). The analyses enables the spatial disaggregation of the results to sub-national levels (governorates and villages), and thereby providing an assessment of spatial variability of A_{wp} within a country.

Table 1 Data and corresponding sources employed in each of the approaches

Variables	Description	Units	Temporal resolution	Spatial resolution	A_{we}	A_{wp1}	A_{wp2}	Source
V_a	Volume of agricultural water withdrawal	m ³ /year	Static	Country	X			AQUASTAT
GVA_a	Gross Value added to GDP from agriculture	USD/year	Yearly	Country	X	X	X	AQUASTAT
c_r	Proportion of GVA_a produced by rainfed agriculture	-	Static	Country	X	X	X	AQUASTAT
LCC	Land cover classification – irrigated & rainfed area	--	Yearly	100 m		X	X	WaPOR
ET	actual EvapoTranspiration and Interception	mm/month	Monthly	100 m		X	X	WaPOR
P	Precipitation	mm/month	Monthly	5 km		X	X	WaPOR
LP	Land Productivity ²	kg/ha/year	Yearly	100 m			X	WaPOR

² In WaPOR the layer is called Total Biomass Production.

Fig. 2 Comparison between A_{we} and the two approaches for estimating A_{wp} 

Data sources

Table 1 shows an overview of the data used in this research and for each approach, it consists of both secondary data and RS derived data. A_{we} is calculated using solely the secondary data from AQUASTAT (FAO 2022). This secondary data consists of annual country-level data on agricultural water withdrawals (V_a), gross value added to GDP from agriculture (GVA_a), and the Proportion of GVA_a produced by rainfed agriculture (c_r). Except for V_a , this data was also used to model the two RS-based approaches. For the RS-based approaches, the data is complemented by RS based ET and LP. This data was obtained from FAO's portal to monitor water productivity through open access RS derived data (WaPOR) (FAO 2023). WaPOR data is available in three resolutions, for Lebanon the highest resolution available covering the entire country of 100 m (WaPOR level 2) was used in this study.

Results and discussion

National level

The results of the two approaches for calculating A_{wp} were compared to A_{we} (Fig. 2). The values of A_{wp1} range between 1.93 USD/m³ in 2020 to 2.46 USD/m³ in 2015, with A_{wp2} showing a slightly higher estimation compared to A_{wp1} . The three curves show similar trends over time, with a decline from 2015 to 2016, a general increase from 2016 to 2019 and a sharp decline in 2020. The results show a high correlation between A_{wp1} and A_{wp2} ($r^2=0.9$), and lower correlation between the two approaches for A_{wp} and A_{we} ($r^2=0.5$).

In absolute values, however, A_{wp1} and A_{wp2} are almost a factor 2 higher than A_{we} . With the GVA_a for A_{we} and A_{wp1} being the same, the difference is in the estimation of the water withdrawn or consumed. Comparing the annual irrigated areas obtained from the WaPOR and those from AQUASTAT, it shows a significant difference. The WaPOR LCC map estimates an area between 354 and 432 km² as irrigated, which is less than half of the area reported by the

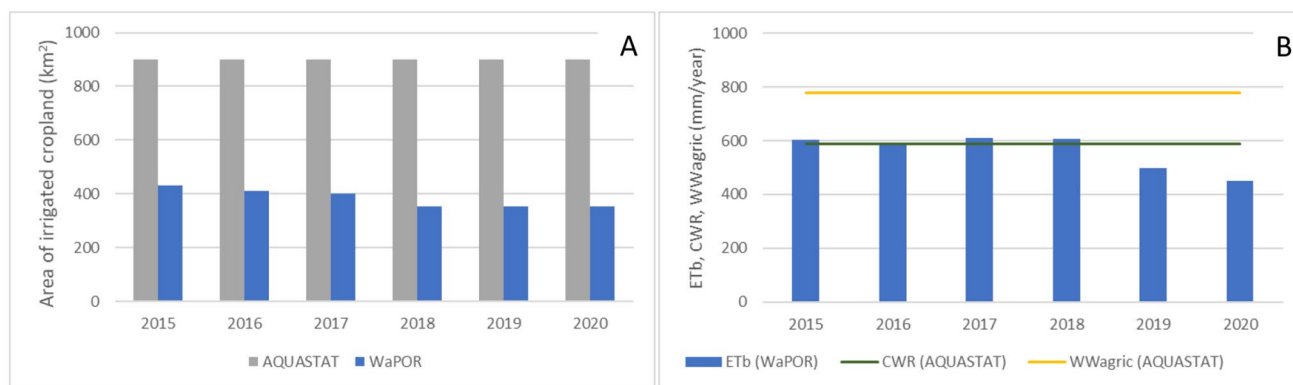
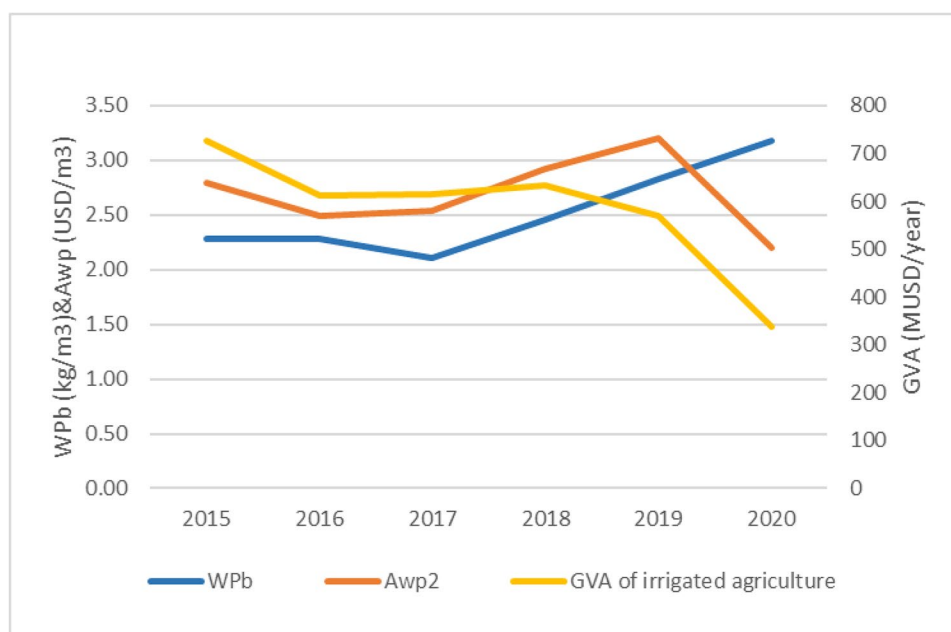


Fig. 3 **A)** Area under irrigation from WaPOR and AQUASTAT and **B)** average ET_b from WaPOR over irrigated areas versus agricultural water withdrawal (WW_{agric}) and estimated crop water requirement (CWR) from AQUASTAT for Lebanon between 2015 and 2020

Fig. 4 Comparison between WP_b , A_{wp2} and GVA_a between 2015–2020, for irrigated agriculture in Lebanon



Lebanese government in AQUASTAT (900 km²) (Fig. 2A). A major distinction between the two is that the AQUASTAT data is based on area equipped for irrigation, which is a static value, whereas WaPOR uses a water deficit index based on precipitation and evapotranspiration to estimate if a cropland is irrigated or not, which varies year by year¹ (FAO 2020).

When comparing the volume of water used for irrigation per area (in mm/year) (Fig. 2B), the values from both sources have a closer match. The WaPOR ET_b data ranges between 449 mm/year in 2020 to 604 mm/year in 2015 (Fig. 2B), while AQUASTAT reports the volume of water withdrawn by agriculture (WW_{agric}) to be 0.7 km³/year (equivalent to 777 mm/year), and crop water requirement (CWR) of 0.529 km³/year (equivalent to 587 mm/year). The

discrepancy in the last two years could be attributed to fields being abandoned because of the economic hardship faced by farmers in Lebanon. Despite these differences, the order of magnitude in estimated water depth estimated using the WaPOR data are reasonable.

As shown in Fig. 2, the results show a consistent drop in A_{we} and both A_{wp} approaches between 2019 and 2020. Comparing WP_b with A_{wp2} shows that the main culprit for the decline in A_{wp2} is not in efficient water use in the agricultural sector as WP_b shows an overall increasing trend from 2015 to 2020, from 2.1 to 2.2 kg/m³ in 2015 to 3.18 kg/m³ in 2020 (Fig. 4). A_{wp2} followed the same trend as WP_b until 2019, starting at 2.79 USD/m³ in 2015 to reach 3.21 USD/m³ in 2019. After 2019, the two indicators start deviating with WP_b increasing to 3.18 kg/m³, and A_{wp2} dropping to 2.2 USD/m³. The drop in A_{wp2} is therefore not the result of a reduced performance of the agricultural sector but is

¹ Because FAO uses a moving window of 5 years, the last two years are not updated.

due to factors affecting the economy in Lebanon 2020, with GVA_a dropping from 634 million USD in 2018 to 339 million USD in 2020. At the same time, the results show that WP_b continues to increase over time.

Sub-national level

The spatial disaggregation of LP, WP_b and A_{wp2} per governorates and villages provides additional information related to the performance of the irrigated areas in the different administrative zones (Fig. 5). The highest LP was found in the North and South with a LP between 20 and 25 ton/ha.

The highest WP_b during these five years was Bekaa governorate in 2020. However, the highest average WP_b of all the governorates was Akkar with 2.8 kg/m^3 and the lowest was recorded in the North Governorate with 2.4 kg/m^3 (Fig. 5D). Overall, WP_b per governorate was increasing over the years (Fig. 6A), however A_{wp2} was decreasing, especially in the last year of the analyses (Fig. 6B).

While Lebanon's average A_{wp2} between 2015 and 2020 was 2.78 USD/m^3 , the spatial disaggregation per governorate or even village shows high variability, with A_{wp2} in the range between 1.8 and $3\text{--}3.5 \text{ USD/m}^3$ (Fig. 6B). These results show that reporting of A_{wp2} (and therefore

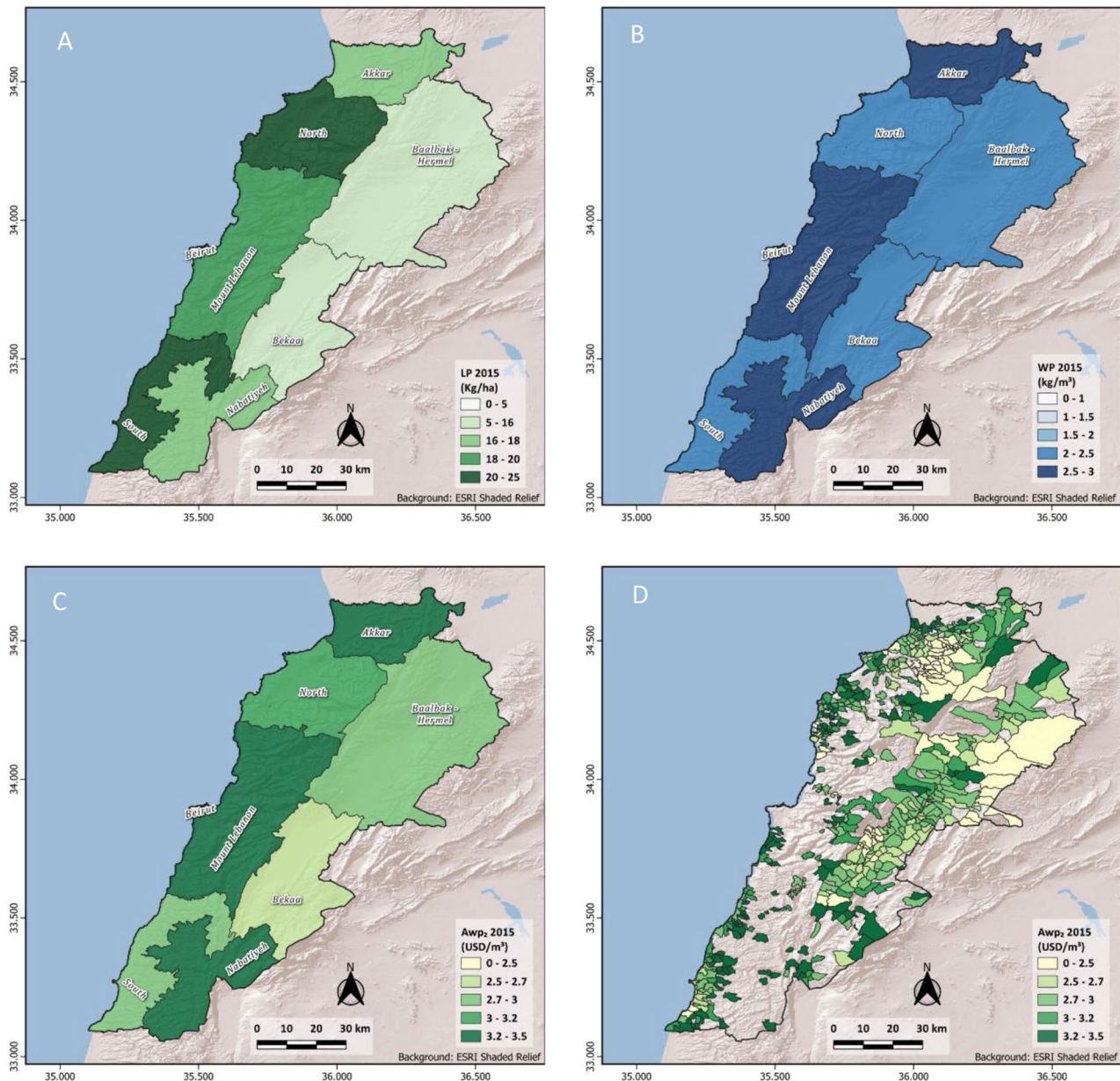


Fig. 5 A) Water Productivity (WP_b), B) Land Productivity (LP) and C) A_{wp2} per governorates in 2015; D) A_{wp2} per village in 2015

Fig. 6 **A)** Annual WP_b and **B)** annual A_{wp2} per governorate in 2015

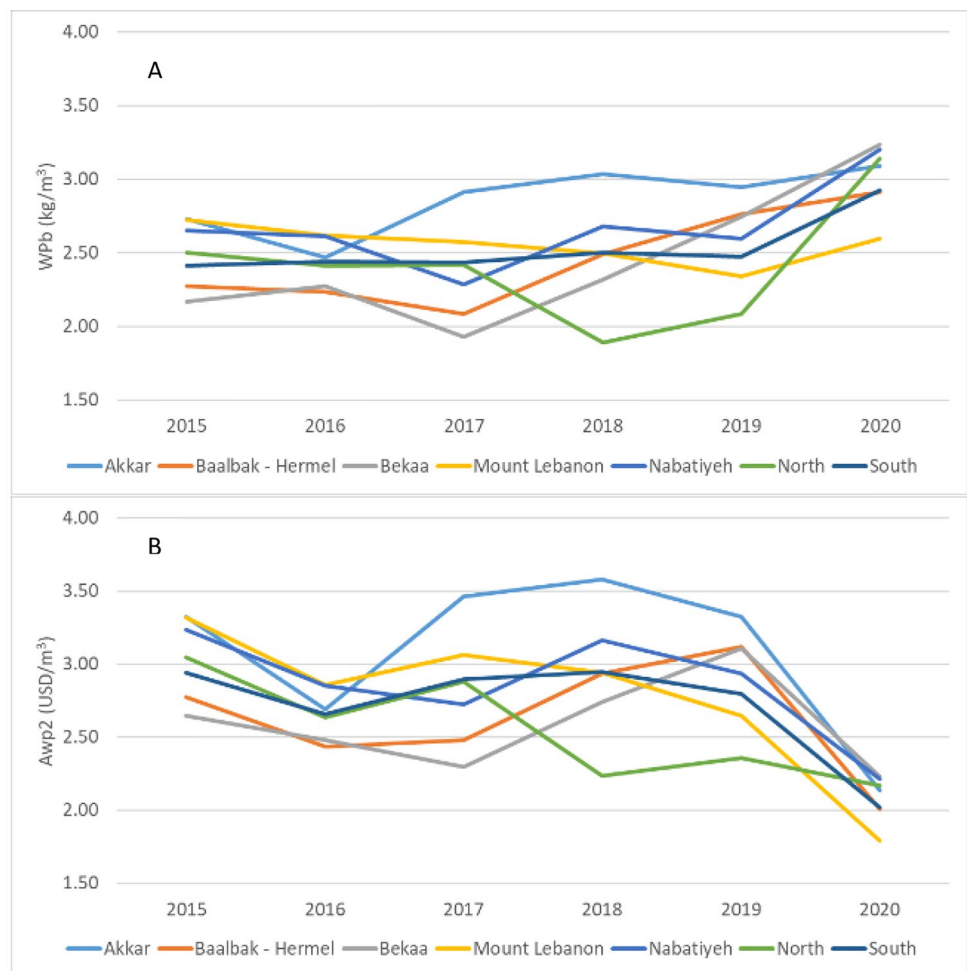
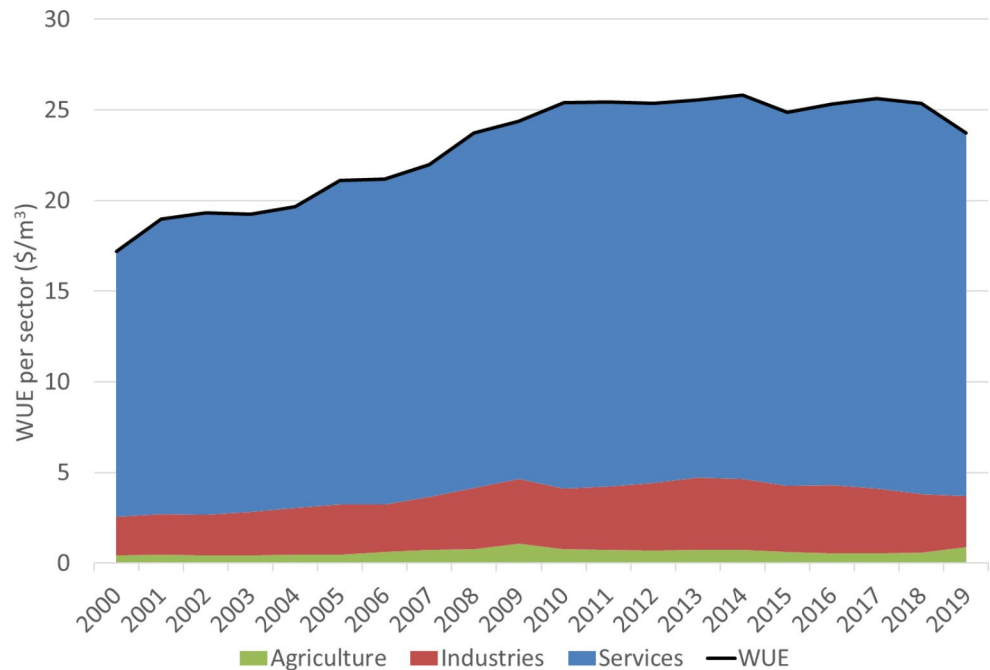


Fig. 7 Contribution of the different sectors to SDG Indicator 6.4.1 water use efficiency for Lebanon over time between 2000 and 2019. (Source of data: UN-Water 2022)



A_{we}) at country level masks the spatial variability within a country and thereby hiding the issues shown when data is

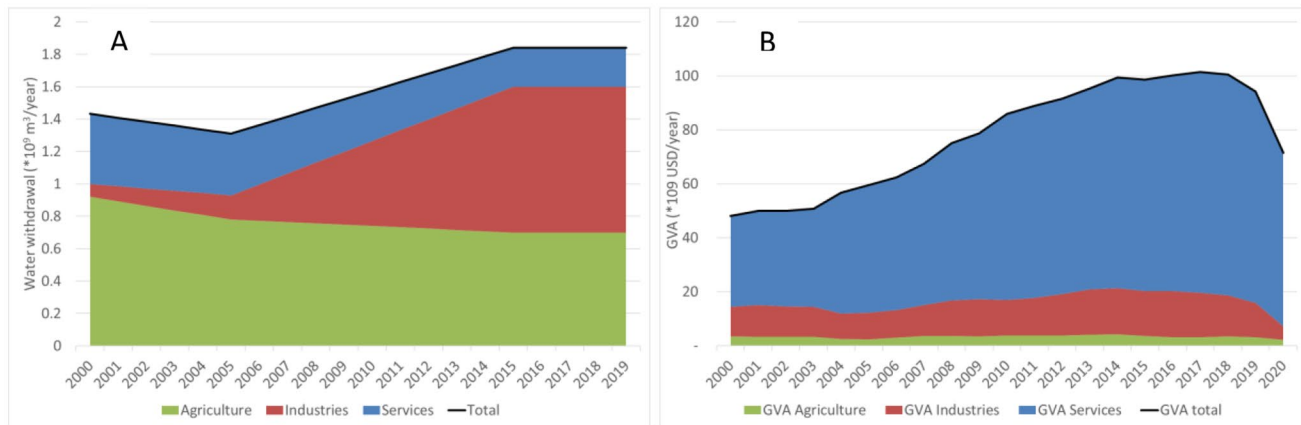


Fig. 8 (a) water withdrawal per sector; (b) GVA per sector over time between 2000 and 2019. (Source of data: UN Water 2022)

disaggregated. Thus, the spatial disaggregation can actually identify hotspots and assist in identifying factors that can explain successes and thus measures for improvement.

Implications for SDG 6 reporting

In the context of reporting SDG indicator 6.4.1 for Lebanon, Fig. 7 shows the indicator increasing from 17.5 USD/m³ in 2000 to 25 USD/m³ in 2019. While the analysis shows that the indicator increased significantly over this period, the changes since 2015 (the baseline of the SDGs) were negligible. The contribution of agriculture to the WUE was relatively small (Fig. 7), although agriculture is a major consumer of water (Fig. 8a).

The data breakdown for SDG indicator 6.4.1 reveals that water withdrawals for the agricultural sector decreases over time while its contribution to the GVA remains relatively constant. The main jump in the overall SDG indicator 6.4.1 between 2000 and 2015 is thus clearly related to the increase in industrial water withdrawals, which has grown nine-fold from 2000 to 2017, surpassing agriculture, while its GVA contribution declines (Fig. 8A and B).

Conclusion

The results of the different approaches show that RS based data can be a valuable source of data for regular reporting and monitoring of SDG indicator 6.4.1. By replacing the water withdrawn by the water consumed, the latter of which can be observed by RS, a proxy to the conventional indicator is estimated. The results of these different approaches to estimate A_{wp} show similar trends as A_{we} . While the overall results on trends are encouraging, the absolute values differ substantially. One main factor is the large inconsistencies in the data on area irrigated in Lebanon between the WaPOR

data and the data reported in AQUASTAT. A possible explanation of this discrepancy can be related to the different definitions used in AQUASTAT where the irrigated area is the area equipped with irrigation, whereas WaPOR uses the water deficit index to identify areas that are actually irrigated (FAO 2020). Arguably, monitoring progress benefits more from information on actually irrigated areas and the water consumed in those areas compared to the potential irrigated areas and the estimated water supply to these areas as reported in AQUASTAT.

A second source of discrepancy is that the A_{wp} approaches only estimate water productivity in irrigated areas and does not consider the aquaculture and livestock components as per the definition of SDG indicator 6.4.1. In the case of Lebanon, the volume of water consumed by irrigation contributes to an important part of the total water withdrawn by agriculture, and therefore will not affect the results.

The methodology used in this paper can be easily extended globally. The combination of spatial and temporal analysis can give a better understanding of the spatial variations and can help incorporate policy changes, decisions, and actions toward Agenda 2030. Before applying the methodology to other countries, it is recommended to validate the RS based estimates to improve the reliability of the analysis and to evaluate if the assumption that irrigated agriculture is the largest consumer of water in agriculture is valid.

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Data Availability The datasets generated during and/or analysed during the current study are available from the corresponding author on reasonable request.

Declarations

Competing interests The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper. The authors declare the following financial interests/personal relationships which may be considered as potential competing interests.

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