

See discussions, stats, and author profiles for this publication at: <https://www.researchgate.net/publication/298725966>

Distribution of megafaunal species in the Southwestern Atlantic: Key ecological areas and opportunities for marine conservation

Article in ICES Journal of Marine Science · March 2016

DOI: 10.1093/icesjms/fsw019

CITATIONS

25

READS

637

12 authors, including:



Victoria González Carman

INIDEP - IIMyC (CONICET - UNMDP)

28 PUBLICATIONS 754 CITATIONS

[SEE PROFILE](#)



Agustina Mandiola

Universidad Nacional de Mar del Plata

34 PUBLICATIONS 144 CITATIONS

[SEE PROFILE](#)



Daniela Alemany

National Scientific and Technical Research Council

17 PUBLICATIONS 307 CITATIONS

[SEE PROFILE](#)



Mariela Dassis

Instituto de Investigaciones Marinas y Costeras (IIMyC), CONICET Universidad Nac...

31 PUBLICATIONS 367 CITATIONS

[SEE PROFILE](#)

Some of the authors of this publication are also working on these related projects:



Histología cardíaca de la Foca Leopardo [View project](#)



South American Consortium for the Census of Antarctic Marine Life (CAML). [View project](#)



Distribution of megafaunal species in the Southwestern Atlantic: key ecological areas and opportunities for marine conservation

V. González Carman^{1,2,3*}, A. Mandiola¹, D. Alemany¹, M. Dassis¹, J. P. Seco Pon¹, L. Prosdocimi⁴, A. Ponce de León⁵, H. Mianzan^{1,2}, E. M. Acha^{1,2}, D. Rodríguez¹, M. Favero¹, and S. Copello¹

¹Instituto de Investigaciones Marinas y Costeras (IIMyC), Consejo Nacional de Investigaciones Científicas y Técnicas (CONICET)—Universidad Nacional de Mar del Plata (UNMDP), Mar del Plata, Argentina

²Instituto Nacional de Investigación y Desarrollo Pesquero (INIDEP), Paseo Victoria Ocampo N° 1 (B7602HSA), Mar del Plata, Argentina

³Aquamarina—CECIM, Programa Regional de Investigación y Conservación de Tortugas Marinas (PRICTMA), Pinamar, Argentina

⁴Coordinación de Gestión de Pesquerías, Dirección Nacional de Planificación Pesquera, Subsecretaría de Pesca y Acuicultura, Buenos Aires, Argentina

⁵Departamento Mamíferos Marinos, Dirección Nacional de Recursos Acuáticos (DINARA), Ministerio de Ganadería, Agricultura y Pesca, Montevideo, Uruguay

*Corresponding author. tel: +54 2234861292; e-mail: vgcarman@inidep.edu.ar; vgcarman@gmail.com.

González Carman, V., Mandiola, A., Alemany, D., Dassis, M., Seco Pon, J.P., Prosdocimi, L., Ponce de León, A., Mianzan, H., Acha, E.M., Rodríguez, D., Favero, M., and Copello, S. Distribution of megafaunal species in the Southwestern Atlantic: key ecological areas and opportunities for marine conservation. – ICES Journal of Marine Science, doi: 10.1093/icesjms/fsw019.

Received 14 July 2015; revised 7 January 2016; accepted 29 January 2016.

During the last centuries, populations of marine megafauna—such as seabirds, turtles, and mammals—were intensively exploited. At present, other threats such as bycatch and pollution affect these species, which play key ecological roles in marine ecosystems as apex consumers and/or nutrient transporters. This study analyses the distribution of six megafaunal species (*Chelonia mydas*, *Caretta caretta*, *Dermochelys coriacea*, *Thalassarche melanophris*, *Otaria flavescens*, and *Arctocephalus australis*) coexisting in the Southwestern Atlantic to discuss their protection in terms of current management strategies in the region. Through the prediction of the species potential distributions and their relation to bathymetry, sea temperature and oceanographic fronts, key ecological areas are defined from a multi-taxa perspective. Information on the distribution of 70 individuals (18 sea turtles, 19 albatrosses, and 33 otariids) was obtained through satellite tracking conducted during 2007–2013 and analysed using a Geographic Information System and maximum entropy models. During the autumn–winter period, megafaunal species were distributed over the continental shelves of Argentina, Uruguay, and Brazil, mainly over the Argentine Exclusive Economic Zone and the Argentina-Uruguay Common Fishing Zone. Despite some differences, all megafaunal species seems to have similar environmental requirements during the autumn–winter period. Mostly waters shallower than 50 m were identified as key ecological areas, with the Río de la Plata as the habitat with the highest suitability for all the species. This area is highly productive and sustains the main coastal fisheries of Uruguay and Argentina, yet its role as a key ecological area for megafaunal species has been underestimated until now. This approach provides a basis to analyse the effect of anthropic activities on megafaunal species through risk maps and, ultimately, to generate knowledge to improve national and bi-national management plans between Argentina and Uruguay.

Keywords: biodiversity conservation, fronts, habitat use, otariids, seabirds, sea turtles, satellite tracking.

Introduction

Since the 16th century and well into the 20th century, populations of marine megafauna—such as sea turtles, birds, and mammals—were intensively exploited in their breeding sites for eggs, feathers, fur, oil, and meat (e.g. Crespo and Pedraza, 1991; Medway, 1998; Rodríguez and Bastida, 1998; Brooke, 2004; Broderick *et al.*, 2006; Croxall *et al.*,

2012). Past levels of exploitation have been reduced significantly, yet some wild populations remain at low levels if compared with historical baselines (Jackson *et al.*, 2001; McClenachan *et al.*, 2006; Grandi *et al.*, 2012). At present, these populations are far from being fully protected because of land-based (e.g. coastal development, introduced predators) and at-sea (e.g. pollution, bycatch and overfishing)

threats (IUCN, 2014). Megafaunal species have been recognized as conservation priorities in marine ecosystems due to their key ecological role as apex consumers and/or nutrient transporters (Jackson, 2001; Heithaus et al., 2008; Baum and Worm, 2009).

This study analyses the distribution of megafaunal species feeding in sympatry in the Southwest Atlantic. It focuses on the Warm Temperate Southwestern Atlantic (WTSA) province (*sensu* Spalding et al., 2007) and on adjacent international waters, comprising part of the exclusive economic zones (EEZs) of Argentina, Brazil, and Uruguay (Figure. 1). In the region, megafaunal species are affected (or have the potential for being affected) by threats like interaction with fisheries and pollution (Fossette et al., 2010; González Carman et al., 2011, 2014a; Rodríguez et al., 2013; Copello et al., 2013, 2014). The region is ground for coastal and high seas fisheries targeting pelagic and bottom-demersal fish and crustaceans species (FAO, 2011) and incidental mortality of megafaunal species has been widely reported in a variety of fishing gear (e.g. Crespo et al., 2007; Bugoni et al., 2008; González Carman et al., 2011; Favero et al., 2013; Seco Pon et al., 2013, 2015; Fossette et al., 2014).

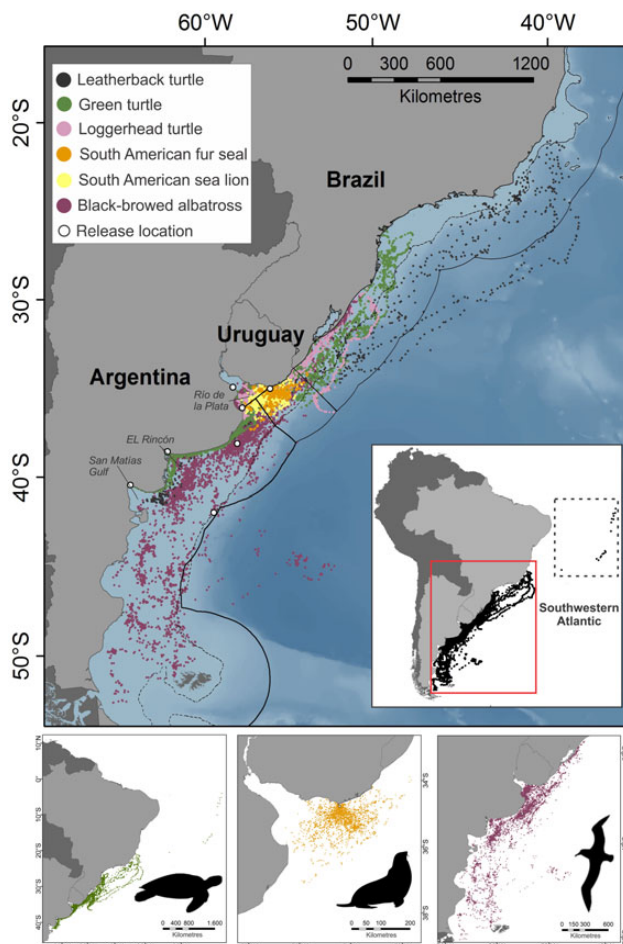


Figure 1. Distribution of tracked marine megafaunal species during autumn–winter (2007–2013) in the WTSA province and adjacent international waters. Upper panel shows locations by species in different colour codes (fixes inside dashed-line square of the inset belong to leatherback turtle). Lower panels show locations of megafaunal species grouped by taxa. EEZs of each country are delimited by solid line and 200 m isobaths by dotted line. This figure is available in black and white in print and in colour at ICES Journal of Marine Science online.

Moreover, major metropolitan areas located along the coast (e.g. Buenos Aires, Montevideo, Rio de Janeiro) are responsible for the generation of large amounts of waste (Acha et al., 2003; González Carman et al., 2015) that affects megafaunal species due to the risk of ingestion or entanglement (e.g. Tourinho et al., 2010; Denuncio et al., 2011; González Carman et al., 2014a; Jiménez et al., 2015).

The oceanic region of the study area is also characterized by the opposing flows of the Brazil (subtropical) and Malvinas (subantarctic) Currents that meet, in average, at 36°S (Olson et al., 1988; Piola et al., 2000; Lucas et al., 2005). In this area, referred to as the Brazil/Malvinas Confluence, the two flows turn offshore in a series of large amplitude meanders (Figure. 2). The neritic region is characterized by a narrow shelf at the north which widens southward to form the broad Patagonian shelf. The northern shelf waters are of subtropical origin while those in the south are of subantarctic origin, both are modified by continental drainage. Important oceanographic frontal systems (Acha et al., 2004) are present in the area, such as those of the Río de la Plata (RDP), El Rincón (ERF), the mid-continental shelf (MSF), and the continental shelf break (SBF) (Figure. 2; Acha et al., 2004; Piola et al., 2005, 2008). The RDP is a two-layered estuarine system where freshwater flows seaward on the surface while denser and more saline shelf water intrudes along the bottom (Mianzan et al., 2001; Acha et al., 2008). This dynamic generates two salinity fronts separated by c. 150 km and connected by a salt-wedge: a bottom (RDP bottom) and a surface (RDP surface) front at the inner and outer part of the estuary, respectively. The discharge of estuarine water into the continental shelf forms as distinct surface layer of low salinity water that extends northeastward beyond 28–30°S (~850 km from the estuarine area) during the austral autumn–winter period. This is known as the RDP plume, in which the boundary between the fresh and marine waters also forms a front (hereafter RDP plume, Muelbert et al., 2008; Piola et al., 2008). As well as the RDP fronts, the ERF separates relatively freshwaters (influenced by river discharges) from high salinity

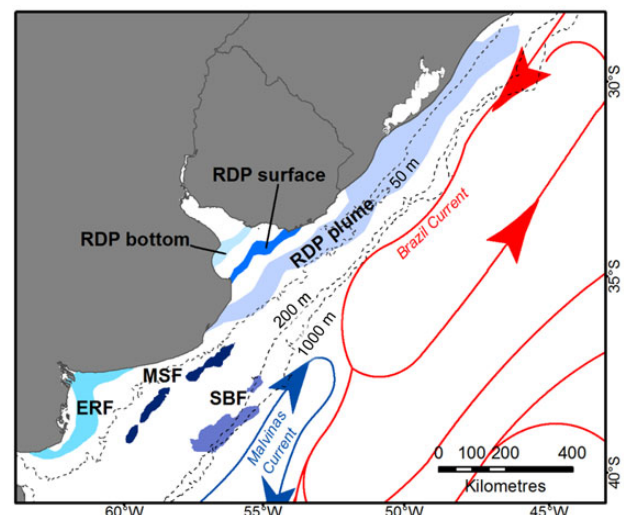


Figure 2. Oceanographic frontal areas in the WTSA province and adjacent international waters. Codes are: RDP bottom: Río de la Plata bottom front, RDP surface: Río de la Plata surface front, RDP plume: Río de la Plata plume front, ERF: El Rincón front, SBF: continental shelf break front, and MSF: mid-continental shelf front. Dotted lines indicate 50, 200, and 1000 m isobaths. This figure is available in black and white in print and in colour at ICES Journal of Marine Science online.

waters generated at the San Matías Gulf (Figure. 1). The MSF holds a thermal front that separates coastal, mixed waters from stratified shelf waters (Lucas *et al.*, 2005; Romero *et al.*, 2006). Similarly, the SBF is characterized by a permanent thermohaline front that separates the shelf waters from the colder and saltier Malvinas waters (Acha *et al.*, 2004; Piola *et al.*, 2008).

These oceanographic frontal areas are known to be crucial habitats for megafaunal species in terms of increased feeding opportunities (Copello *et al.*, 2011, 2013; Rodríguez *et al.*, 2013; Acha *et al.*, 2015). However, they may enhance threats to these species through the concentration of fishing activities (Alemán *et al.*, 2014; Copello *et al.*, 2014) and accumulation of drifting pollutants (e.g. plastics) (Acha *et al.*, 2003; González Carman *et al.*, 2014a). The analysis of the distribution of megafaunal species in relation to oceanographic frontal areas is therefore relevant to their conservation (Ramos *et al.*, 2013; Scales *et al.*, 2014), and the WTSA province and adjacent international waters remains an important region with significant data scarcity, in particular regarding multi-taxa studies.

This study identifies key ecological areas for megafaunal species in the WTSA province and adjacent international waters using a Geographic Information System (GIS) and a modelling tool rooted in maximum entropy. Specifically, it (i) describes the distribution of seabird, turtle, and otariid species; (ii) relates such spatial distribution to bathymetry, sea surface temperature (SST), and main oceanographic frontal areas of the region; and (iii) defines key ecological areas from a multi-taxa perspective. Results are discussed in relation to current management regimes and jurisdictions to identify conservation opportunities.

Methodology

Satellite tracking data

A total of 70 individuals from six megafaunal species were satellite tracked during austral autumn–winter (April to September) from 2007 to 2013: the Green turtle (*Chelonia mydas*), the Loggerhead turtle (*Caretta caretta*), the Leatherback turtle (*Dermochelys coriacea*), the Black-browed albatross (*Thalassarche melanophris*), the South American sea lion (*Otaria flavescens*), and the South American fur seal (*Arctocephalus australis*) (Figure. 1; Table 1). Electronic tags were deployed on the individuals to monitor their distribution after being released at the vicinity of their capture sites. Location information for all taxa were filtered according to different criteria depending on the species (see Supplementary material and Tables S1 and S2 therein).

Environmental data

The oceanographic frontal areas were defined according to those variables showing clear gradients related to the fronts. For the

RDP and ERF, polygons between some selected isohalines were constructed in a GIS with the Feature to Polygon tool of ArcGIS 10.1® (Copyright© ESRI). The isohalines used were 12.5–22.5 for the RDP bottom, 20.0–25.0 for the RDP surface, 30.0–33.0 for the RDP plume, and 33.0–33.7 for ERF; according to Lucas *et al.* (2005), Piola *et al.* (2000, 2008), and Guerrero *et al.* (2010). The SBF and MSF were defined by satellite chlorophyll *a* (Chl *a*) patterns in a GIS. Following Alemán *et al.* (2014), Standard Mapped Images of satellite-derived chlorophyll *a* concentrations (NASA Ocean Colour; <http://oceancolor.gsfc.nasa.gov>, MODIS-Aqua sensor, processing level L3) were used to construct contour lines of 3 and 2 mg m⁻³ Chl *a* mean amplitude that defined the SBF and MSF polygons, respectively. Images corresponded to the 2007–2013 period and were limited to the area where the species co-occurred (197 by 166 pixels, 9 km spatial resolution). Following the same criterion, only the northern zone of the SBF was considered.

Bathymetric information for the study area was obtained from the GEBCO Digital Atlas and ETOPO2 Global 2' Elevations datasets (British Oceanographic Data Centre and NOAA's National Geophysical Data Centre, <http://gebco.net>).

Climatological rasters from Aqua MODIS L3 SST were downloaded from NASA JPL PO.DACC website (ftp://podaac-ftp.jpl.nasa.gov/allData/modis/L3/docs/modis_sst.html, OBPG, 2002) to a GIS using the Marine Geospatial Ecology Tool (Roberts *et al.*, 2010). Data were monthly climatologies of SST images of the study area for the period April 2007–September 2013, with a spatial resolution of 4 km.

Species distribution modelling

A maximum entropy (MaxEnt) species distribution modelling (Phillips *et al.*, 2006) was used to relate the distribution of megafaunal species to bathymetry, SST, and main oceanographic frontal areas of the region. MaxEnt predicts the potential distribution of a species from species occurrence data and environmental background data. It generates maps of habitat suitability (HS) scaled from lowest (blue) to highest (red) suitability (Phillips *et al.*, 2006; Elith *et al.*, 2011).

MaxEnt modelling was implemented in MaxEnt version 3.3.3 k for each megafaunal species. To limit autocorrelation, tracking location data were reduced to best daily locations using R (R 3.0.1, R Development Core Team, 2013). Best daily locations were positions with the highest quality location class recorded during a 24-h period (Supplementary Table S2). If more than one location was determined with equal quality within the 24-h period, the first received location was retained (Pikesley *et al.*, 2013). Environmental layers for bathymetry, SST, and frontal systems were created in a GIS ensuring that they have the same geographic coordinate system (datum

Table 1. Metadata for marine megafaunal species satellite tracked in the WTSA province and adjacent international waters.

Common name	Scientific name	No. of individuals	Stage	Sex	Global conservation status (IUCN, 2014)
Green turtle	<i>Chelonia mydas</i>	9	J	U	Endangered
Loggerhead turtle	<i>Caretta caretta</i>	6	J	U	Vulnerable
Leatherback turtle	<i>Dermochelys coriacea</i>	3	A	F	Critically Endangered
Black-browed albatross	<i>Thalassarche melanophris</i>	19	A	16 F 3 M	Near Threatened
South American sea lion	<i>Otaria flavescens</i>	22	6 SA 16 A	F	Least Concern
South American fur seal	<i>Arctocephalus australis</i>	11	A	F	Least Concern

Codes are J, juvenile; SA, subadult; A, adult; U, unknown; F, female; M, male.

WGS 84), resolution (4 km), and spatial extent restricted to the area where all taxa co-occurred (i.e. 30–40°S, 45–65°W) (Supplementary Table S2). For SST data, monthly climatological images were combined into an image representing the mean SST using the Extract Multivalues to Points, Spatial Join and Polygon to Raster tools of ArcGIS 10.1® (Copyright© ESRI).

MaxEnt models were run 100 times, obtaining a mean response model from the 100 runs for each individual species. Each time, a random sample of 30% of the dataset was saved to test the model (Supplementary Table S2). To ensure convergence of the model, the number of iterations was set to 5000 (Young *et al.*, 2011). The receiver-operating characteristic analysis and the area under the curve (AUC) were used to provide a single measure of model performance. An AUC value of 0.5 indicates that the performance of the model is no better than random, while values closer to 1.0 indicate better model performance (Phillips *et al.*, 2006).

To compare the predicted HS maps generated for each megafaunal species, the Niche Overlap function of the Dismo package (Hijsmans *et al.*, 2013) was used in R (R 3.0.1, R Development Core Team, 2013). This function computes niche overlap from pairwise predictions of species distributions with a measure derived from Hellinger distance called “I” (Warren *et al.*, 2008, 2010). This similarity measure is obtained by comparing the estimates of HS calculated for each grid cell of a study area using a Maxent generated distribution model, after normalizing each model so that all suitability scores within the geographic space sum to 1. “I” ranges from 0 (when species predicted environmental tolerances do not overlap at all) to 1 (all grid cells are estimated to be equally suitable for both species) (Warren *et al.*, 2008, 2010).

Multi-taxa key ecological areas were defined overlapping the species maps of HS in a GIS and calculating the weighted average of HS for each cell of the species maps as follows:

$$\text{multi-taxa HS} = \frac{\sum r_i \times HS_{ij}}{N},$$

where r is a coefficient calculated as the ratio of the number of tracking locations used to run the model for species i ($i = 1, \dots, N$), HS is the HS value in map cell j ($j = 1, \dots, J$) for species i , N is the total number of species and J is the total number of map cells. The multi-taxa HS was then represented in a GIS using the Extract Multivalues to Points, Spatial Join, and Polygon to Raster tools of ArcGIS 10.1® (Copyright© ESRI).

Results

Distribution of megafaunal species in the Southwestern Atlantic

A total of 17 233 filtered locations was obtained from 2007 to 2013 for the six analysed species (Supplementary Table S2). Tracked taxa showed an ample distribution in the SW Atlantic, ranging from 2°S to 53°S, from coastal areas to the high seas, and dispersed over Argentinean, Uruguayan, and Brazilian EEZs and international waters (Figure. 1). Although the use of these areas was not homogeneous, tracked individuals spent a large proportion of their time (93%) in shelf waters. Individuals of all taxa spent most of the time in Argentinean waters (48%), followed by Uruguayan (29%), and Brazilian (14%) EEZs, while <9% of the time was spent in international waters.

There were differences in the spatial scale of at-sea distribution by the different taxa (Figure. 1). Sea turtles covered the largest marine area (c. 450 000 km²), with the three species distributed from northern Brazilian waters to 42°S (exceptionally, one leatherback

turtle undertook a long-distance journey north of 3°S into the high seas, see locations inside dashed-line square in Figure. 1). Black-browed albatrosses spread over an area about half the size of sea turtles, although tracked birds distributed further south reaching 53°S. Compared with the other taxa, marine mammals were confined to a smaller area (c. 80 000 km²) between 33–37°S and 52–56°W within the RDP, near their breeding colony.

Distribution of megafaunal species in relation to bathymetry, SST, and frontal areas in the WTSa province and adjacent international waters

A total of 1643 locations were used to run the models. The potential distribution of each megafaunal species during the autumn–winter period in the study area are shown in Figure. 3. Distribution models generated for each species returned AUC values >0.8 indicating a good model performance (Table 2).

Green turtle highest suitable areas (red areas) were in shelf waters of the RDP (associated to the surface and plume fronts) and throughout ERF (Figure. 3), spanning shallow waters of <50 m. Medium suitable areas (orange to yellow areas) occurred in the interior of the RDP, the RDP bottom front and along the 200 m isobath between 30 and 33°S. For loggerheads, highest suitable areas were in shallow (<50 m) waters of the RDP (bottom, surface, and plume fronts) (Figure. 3). Leatherback highest suitable areas were in shelf waters of the RDP (mainly the surface front) and ERF. Medium suitable areas extended to the 50 m isobath and also to the east of the 200 m isobath between 30 and 33°S (Figure. 3).

Albatross highest suitable areas were in waters of the RDP (mainly associated to the RDP bottom front), ERF, MSE, and over the 200 m isobath between 35 and 37°S. Medium suitable areas extended through the continental shelf barely beyond the 50 m isobath and along the 200 m isobath between 37 and 40°S, up to the SBF (Figure. 3).

For otariids, highest suitable areas were in waters of the RDP, mainly associated with the RDP surface and plume fronts (Figure. 3). For the South American fur seal, high suitable areas reach the 200 m isobath at 34–35°S and medium suitable areas extended south along the coast limited by the 50 m isobath (Figure. 3).

Bathymetry had the greatest explanatory power for all species models (>61–85%), whereas frontal areas had the smallest explanatory power (<8%). The mean SST explained between 14 and 30% of the species distribution predicted by the models. For marine turtles and albatross, the mean SST explained 20–30% of the variation observed in the distribution of those species (Table 2). The models predicted that suitable areas for marine turtles occurred in waters >12°C, with maximum probability of occurrence between 18 and 22°C of mean SST. Suitable areas for albatross and otariids included colder waters with 12–16°C of mean SST.

Comparison of the distribution of megafaunal species and key multi-taxa areas in the WTSa province

Similarity measure “I” was close to 1 in all pairwise comparison between potential distributions indicating that megafaunal species have similar environmental requirements during the autumn–winter period (Table 3). More similar were the potential distributions among marine turtle species, among otariid species, and between Green turtles, Black-browed albatrosses, and South American fur seals. In particular, waters shallower than 50 m potentially hosted the most suitable environmental conditions for all megafaunal species (Figure. 4). Highest suitable areas were mostly in waters of the RDP (in the bottom, surface, and plume fronts),

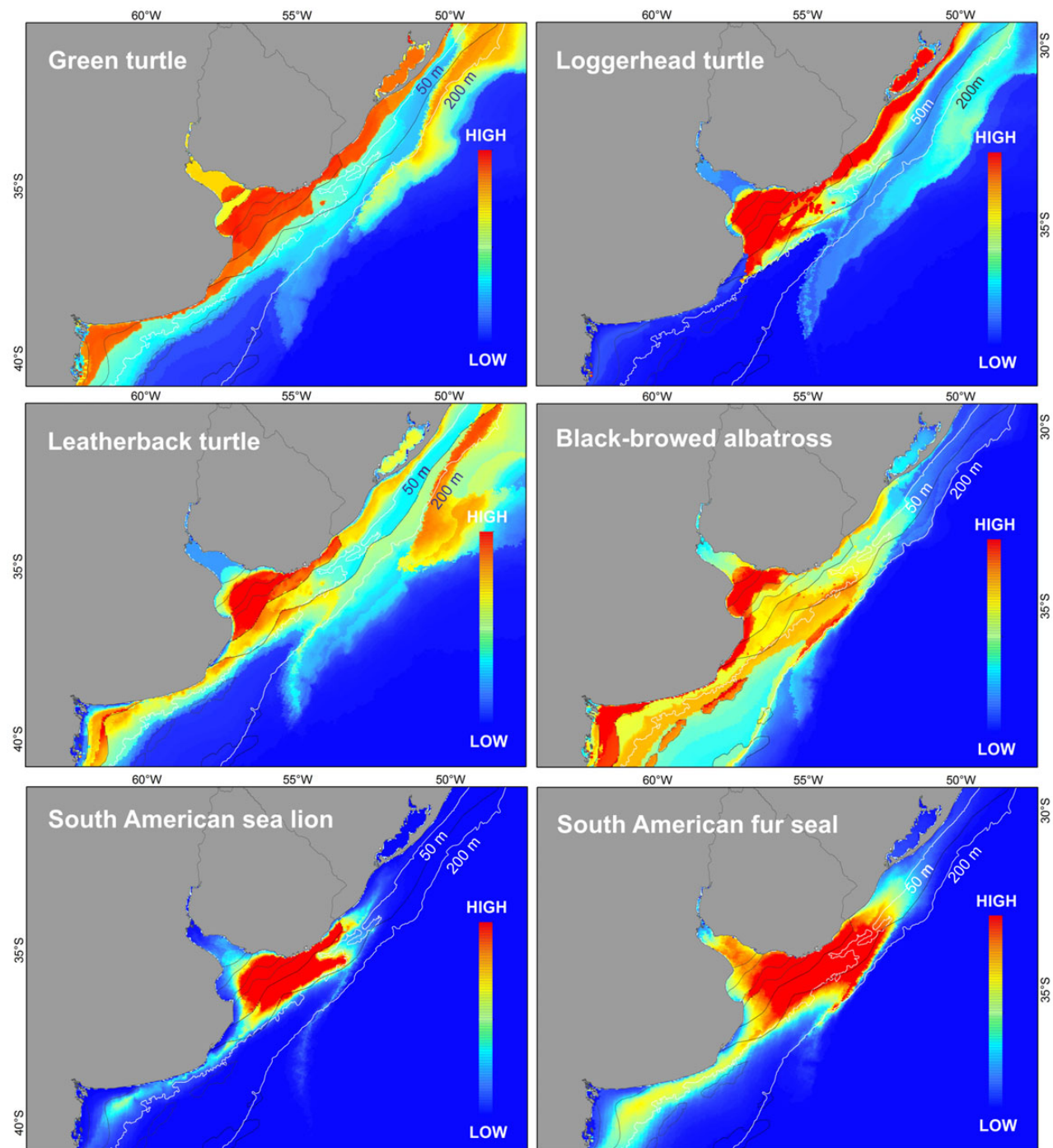


Figure 3. Potential distribution of megafaunal species during the autumn – winter period in the WTSA province and adjacent international waters modelled through maximum entropy. Black lines delimit oceanographic frontal areas and white lines indicate 50 and 200 m isobaths. This figure is available in black and white in print and in colour at *ICES Journal of Marine Science* online.

encompassing an area of ca. 85 000 km². Medium suitable areas extended south along the coast to the ERF and were limited by the 50 m isobath. At the latitude of 34–35°S, high to medium suitable areas extended to the 200 m isobath (Figure. 4).

Discussion

The Green turtle, the Loggerhead turtle, the Leatherback turtle, the Black-browed albatross, the South American sea lion, and the South

American fur seal were widely distributed in waters of the Southwestern Atlantic, mostly over Argentinean, Uruguayan, and Brazilian shelves. For the three taxa—sea turtles, seabirds, and otariids—mostly shallow waters no deeper than 50 m were identified as key areas (i.e. highly suitable). Particularly, the RDP frontal areas were identified as the habitat with the highest suitability for all the species, presumably due to good feeding opportunities. The novel multispecies approach taken in this study expands the bounds of

individual species data sources, aiming to provide a basis for the development of multi-taxa management tools to protect vulnerable marine species and their habitats in the region.

Distribution modelling of megafaunal species

Modelling performance was good judging by the AUC values obtained for all species (Table 2). The potential distributions of the six megafaunal species seem to be defined mostly by the 50 m isobath (Table 2). Only the model generated for the Leatherback turtle exhibited medium suitable areas in the oceanic realm (Figure. 3), a species considered to be oceanic during most part of their lifetime (Bolten, 2003). Marine turtle potential distribution was also influenced by higher mean SST than albatross and otariids, as expected for ectothermic animals that perform seasonal movements from warmer coastal waters of Argentina and Uruguay in summer to coastal and oceanic areas off southern Brazil in winter (López-Mendilaharsu et al., 2009; González Carman et al., 2012, 2016).

Frontal areas had low explanatory power than expected, especially when compared with bathymetry. This can be due to the nature of the environmental layer used to predict the potential distribution of the species, which integrates information from salinity and Chl *a* concentration through the spatial definition of frontal and non-frontal areas. It was not possible to use GIS layers of Chl *a* or salinity values (as for SST) because of restrictions imposed by the modelling procedure. MaxEnt requires all the environmental layers have the exact same extent of data to execute the model (Young et al., 2011). But Chl *a* values in coastal waters, and especially within the RDP, are highly overestimated compared with values in the open sea due to coloured dissolved organic matter altering their optical properties (Piola et al., 2008). In addition to this, *in situ* bottom and surface salinity values were not available for the entire study area, and satellite-derived salinity are only available for surface layers and in a coarse spatial resolution ($1^\circ \times 1^\circ$).

Biological importance of the RDP for megafaunal species

The RDP and neighbouring waters have been recognized as a highly productive area sustaining a range of commercial fisheries (Mianzan

et al., 2001; Chaluleu, 2002; Carozza, 2010; Sánchez et al., 2011). However, its role and relevance from the ecological perspective for megafaunal species has been underestimated. Although there were differences in the time of sampling (i.e. not all individuals were tracked at the same time, see Supplementary Table S1) — and thus our conclusions on multi-taxa key ecological areas should be taken with caution —, information on the feeding ecology of the individual species support the RDP as the main area for all species. They would benefit from high biomasses of their natural prey, but also from anthropogenic or facilitated resources (fishery discards and other by-products of fishery operations, as well as surface or shallow-diving predators during the recovery of the fishing gear). Green and leatherback turtles forage on gelatinous plankton such as medusae (*Liriope tetrphylla* and *Lychnorhiza lucerna*) (Estrades et al., 2007; González Carman et al., 2014b), which are highly abundant in estuarine waters (Mianzan and Guerrero, 2000; Alvarez Colombo et al., 2003). Loggerhead's diet is known to include salps (Martinez Souza, 2009) that reach high biomasses in adjacent shelf waters (Mianzan and Guerrero, 2000; Alvarez Colombo et al., 2003). The Black-browed albatross is known not only to feed on fish, cephalopods, and in some areas crustaceans but also to approach fishing vessels looking for discards, offal, and prey facilitated by fishery operations, in particular trawlers that heavily exploited the area (ACAP, 2010; Copello et al., 2014; Mariano-Jelicich et al., 2014; Seco Pon et al., 2015). Summer diet of otariids from Isla de Lobos (35.1°S, 54.9°W) is dominated by Whitemouth croaker (*Micropogonias furnieri*), Weakfish (*Cynoscion guatucupa*), and Argentine anchovy (*Engraulis anchoita*) (Naya et al., 2000, 2002; Ponce de León and Pin, 2006; Franco-Trecu et al., 2013), two target species for the commercial fisheries operating in the area (Sánchez and de Ciechowski, 1995; Chaluleu, 2002; Jaureguizar et al., 2003).

The relatively minor use of other oceanographic frontal areas—namely ERF, MSF, and SBF—by megafaunal species could be, at least partially, attributed to season, sex, and life stage of the studied individuals. For example, juvenile Black-browed albatross (not analysed in this study) intensively use the SBF at the latitude of the RDP during autumn and winter (Falabella et al., 2009). Post-breeding males of South American sea lions (in contrast to females studied here) disperse from haul out sites on the coast of Argentina to Uruguay and northern Patagonia (Giardino et al., 2014), at the latitudes of the MSF and ERF. Hence, the association of megafaunal species to such frontal areas could be important under other circumstances or periods in the annual cycle. In fact, the SBF is important for other megafaunal species not included in this study, like the Southern Giant petrel, the Wandering albatross (*Diomedea exulans*), and the Southern Elephant seal (*Mirounga leonina*) (Croxall and Wood, 2002; Campagna et al., 2006; Falabella et al., 2009; Quintana et al., 2010).

Table 2. AUC values and relative contributions of the environmental variables to the Maxent models.

Species	AUC	Variable contribution (%)		
		Frontal areas	Bathymetry	Mean SST
Green turtle	0.878	0.5	85.3	14.2
Loggerhead turtle	0.936	3.4	76.7	19.9
Leatherback turtle	0.867	7.9	67.2	24.9
Black-browed albatross	0.890	5.9	66.5	27.6
South American sea lion	0.984	7.0	66.0	27.0
South American fur seal	0.957	7.4	61.6	31.0

Table 3. Similarity measure “I” showing the overlap between pairwise predictions of megafaunal species potential distributions.

	Green turtle	Loggerhead turtle	Leatherback turtle	Black-browed albatross	South American sea lion	South American fur seal
Green turtle		0.89	0.94	0.84	0.65	0.89
Loggerhead turtle			0.89	0.66	0.71	0.75
Leatherback turtle				0.76	0.62	0.75
Black-browed albatross					0.66	0.87
South American sea lion						0.88
South American fur seal						

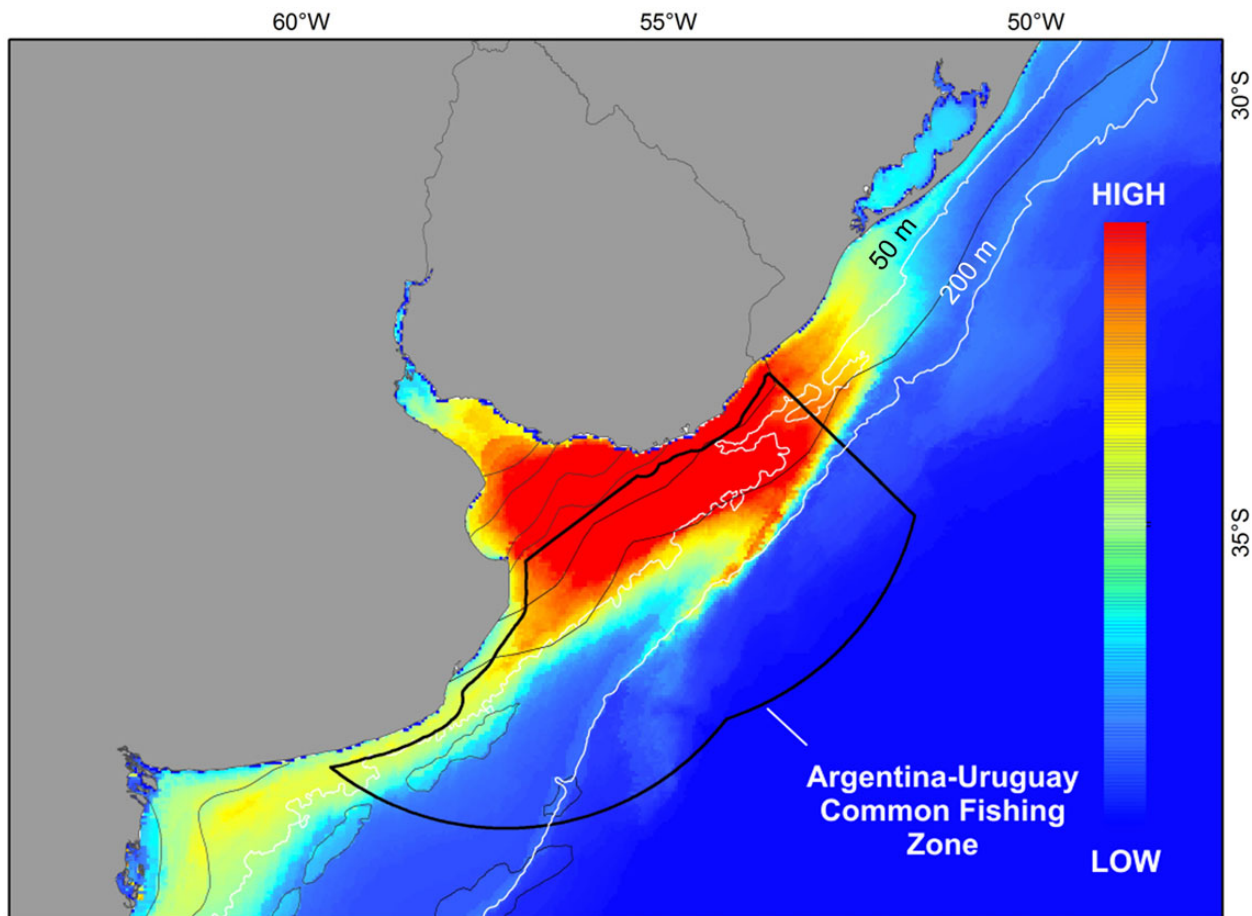


Figure 4. Overlap map of potential distribution of megafaunal species during the autumn–winter period in the WTSa province and adjacent international waters. Black lines delimit oceanographic frontal areas and white lines indicate 50 and 200 m isobaths. This figure is available in black and white in print and in colour at *ICES Journal of Marine Science* online.

Opportunities for conservation of megafaunal species

The RDP is an area under international administration through the RDP Bilateral Treaty (*Tratado del Río de la Plata y su Frente Marítimo*) established between Argentina and Uruguay in 1973. This treaty administrates human activities, such as fishing and coastal development, to ensure sustainability, prevent pollution and promote research, and management to evaluate and preserve resources. Since several megafaunal species intensively use the area and interact with (and may be affected by) human activities, this treaty and its enforcement authority—the Technical Commission of the Maritime Front—are key instruments for their conservation and management. So far, bilateral coordinated actions involving the conservation of marine megafaunal species have been neglected in this forum. But our new understanding of the potential distribution of six megafaunal species of conservation concern improves the probability of success of protection measures in the RDP. The relatively restricted geographic area identified as a multi-taxa key area should be taken into account when zoning of human activities, especially those activities conducted in the Argentina-Uruguay Common Fishing Zone (Figure 4). Future actions to be applied in this area and under this treaty could, for instance, focus on addressing the interactions between megafaunal species and commercial fisheries—such as bycatch and competition for resources—and on preventing and reducing marine pollution in terms of plastic debris disposed from coastal areas as well as from fishing activities.

These actions should also be included into Argentina National Action Plans for seabirds, marine mammals, and sea turtles, in some cases already adopted on either side of the RDP but lacking of any coordination in terms of implementation.

Other threatened megafaunal species inhabiting the RDP area could be beneficiaries of protection actions promoted from these instruments, namely the Franciscana dolphin (*Pontoporia blainvillei*), the Magellanic penguin (*Spheniscus magellanicus*), the Manx (*Puffinus puffinus*), Great (*Ardenna gravis*) and Sooty (*A. grisea*) shearwaters, the Northern Royal albatross (*Diomedea sanfordi*), and the White-chinned (*Procellaria aequinoctialis*), the Spectacled (*P. conspicillata*), and the Southern Giant (*Macronectes giganteus*) petrels (Nicholls *et al.*, 2002; Falabella *et al.*, 2009; Guilford *et al.*, 2009; Ronconi *et al.*, 2010; Secchi, 2010; Hedde *et al.*, 2012; Reid *et al.*, 2014; Blanco and Quintana, 2014). Next steps should focus on more comprehensive analyses with the addition of more species, improvement of models through the inclusion of other variables (e.g. wind, surface currents, and fishing activity) as well as the assessment of the impact of fisheries and pollution on megafaunal species through risk or sensibility maps within the Río de la Plata.

Supplementary data

Supplementary material is available at the *ICESJMS* online version of the manuscript.

Acknowledgements

This study adhered to the legal requirements of the countries in which the work was carried out, and to all institutional guidelines established by the wildlife agencies of Buenos Aires and Río Negro provinces, and the National Wildlife Agency of Argentina and Uruguay. We thank to PRICTMA (Acuario de Buenos Aires, Aquamarina—CECIM, Fundación Mundo Marino, Instituto de Biología Marina y Pesquera “Alte. Storni”, Reserva Natural de Usos Múltiples Bahía Blanca, Bahía Falsa y Bahía Verde, and Mar del Plata Aquarium) and the fieldwork assistance of DINARA personnel (Cesar Barreiro, Nelson Veiga, Leonardo Olivera, Miguel Casella, and Fernando Area) during capturing and handling of otariids at Isla de Lobos (Uruguay). We thank the anaesthesia and veterinary control performed by DMVs Bruce Heath, Eduardo Mateos, Valeria Ruopollo, and Diego Albareda. We are also grateful to PhD. Silvia Romero and Lic. Graciela N. Molinari for their assistance during the definition of oceanographic frontal areas and to PhD. Santiago Barbini and PhD. Federico Cortés for their advices on Maxent modelling and R procedures, respectively. A special thanks to PhD. Manjula Tiwari of NOAA Southwest Fisheries Science Center, and PhD. Alberto Piola for their financial support. Funding was provided by the Buenos Aires Zoo to Diego Albareda, the Wildlife Conservation Society, Fondo para la Conservación Ambiental from Banco Galicia and the Cleveland Metropark Zoo—Scott Neotropical Fund, and Agencia Nacional de Promoción Científica y Tecnológica FONCyT PICT 2013–2099 to VGC, from Inter-American Institute for Global Change Research (IAI) grant CRN 3070 sponsored by the US National Science Foundation Grant GEO-1128040 to HM and EMA, PIP 2011-070, PICT 2012-295 and PICT 2013-711 to SC, MF and JPSP, the Alaska SeaLife Center (Contracts ASLC # R-1972-01 and R-2972-01), DINARA (Exp.1136/2006, 503/2007 and 1378/2008), FONCyT (Projects PICT 2007–01763; PICT 2011–1834), Universidad Nacional de Mar del Plata, Argentina (Projects 15/E471 and 15/E335), and an NSF-CONICET Cooperation Grant (CONICET Resolution 1340/10) to DR and APL. AM is supported by a scholarship from CONICET. This is INIDEP contribution no. 1958.

References

- Acha, E. M., Mianzan, H., Guerrero, R., Carreto, J., Giberto, D., Montoya, N., and Carignan, M. 2008. An overview of physical and ecological processes in the Río de la Plata Estuary. *Continental Shelf Research*, 28: 1579–1588.
- Acha, E. M., Mianzan, H. W., Guerrero, R. A., Favero, M., and Bava, J. 2004. Marine fronts at the continental shelves of austral South America Physical and ecological processes. *Journal of Marine Systems*, 44: 83–105.
- Acha, E. M., Mianzan, H. W., Iribarne, O., Gagliardini, D. A., Lasta, C., and Daleo, P. 2003. The role of the Río de la Plata bottom salinity front in accumulating debris. *Marine Pollution Bulletin*, 46: 197–202.
- Acha, E. M., Piola, A., Iribarne, O., and Mianzan, H. 2015. Ecological processes at marine fronts. Oases in the ocean. Cham, Springer International Publishing. 68 pp.
- Agreement on the Conservation of Albatrosses and Petrels. 2010. ACAP Species assessment: Black-browed albatross *Thalassarche melanophrys*. <http://acap.aq/en/acap-species/238-black-browed-albatross/file> (last accessed 13 July 2015).
- Aleman, D., Acha, E. M., and Iribarne, O. O. 2014. Marine fronts are important fishing areas for demersal species at the Argentine Sea (Southwest Atlantic Ocean). *Journal of Sea Research*, 87: 56–67.
- Alvarez Colombo, G., Mianzan, H., and Madirolas, A. 2003. Acoustic characterization of gelatinous-plankton aggregations: four case studies from the Argentine continental shelf. *ICES Journal of Marine Science*, 60: 650–657.
- Baum, J. K., and Worm, B. 2009. Cascading top-down effects of changing oceanic predator abundances. *Journal of Animal Ecology*, 78: 699–714.
- Blanco, G. S., and Quintana, F. 2014. Differential use of the Argentine shelf by wintering adults and juveniles southern giant petrels, *Macronectes giganteus*, from Patagonia. *Estuarine, Coastal and Shelf Science*, 149: 151–159.
- Bolten, A. B. 2003. Variation in sea turtle life history patterns: neritic versus oceanic developmental stages. In *The Biology of Sea Turtles*, pp. 243–257. Ed. by P. L. Lutz, J. A. Musick, and J. Wyneken. CRC Press, Boca Raton, FL.
- Broderick, A. C., Frauenstein, R., Glen, F., Hays, G. C., Jackson, A. L., Pelembe, T., Ruxton, G. D., et al. 2006. Are green turtles globally endangered? *Global Ecology and Biogeography*, 15: 21–26.
- de Brooke, M. L. 2004. Albatrosses and Petrels across the World. Oxford University Press, Oxford.
- Bugoni, L., Mancini, P. L., Monteiro, D. S., Nascimento, L., and Neves, T. S. 2008. Seabird bycatch in the Brazilian pelagic longline fishery and a review of capture rates in the southwestern Atlantic Ocean. *Endangered Species Research*, 5: 137–147.
- Campagna, C., Piola, A. R., Rosa Marin, M., Lewis, M., and Fernández, T. 2006. Southern elephant seal trajectories, fronts and eddies in the Brazil/Malvinas Confluence. *Deep Sea Research Part I: Oceanographic Research Papers*, 53: 1907–1924.
- Carozza, C. 2010. Pesquería comercial de corvina rubia (*Micropogonias furnieri*) en Argentina. *Frente Marítimo*, 21: 15–22.
- Chaluleu, J. D. 2002. Shared fishery Argentine-Uruguayan common fishing zone. FAO, Fisheries Report, 695: 86–104. <ftp://193.43.36.93/docrep/fao/005/y4652e/Y4652e03.pdf> (last accessed 13 July 2015).
- Copello, S., Dogliotti, A., Gagliardini, D., and Quintana, F. 2011. Oceanographic and biological landscapes used by the Southern Giant Petrel during the breeding season at the Patagonian Shelf. *Marine Biology*, 158: 1247–1257.
- Copello, S., Seco Pon, J. P., and Favero, M. 2013. Use of marine space by Black-browed albatrosses during the non-breeding season in the Southwest Atlantic Ocean. *Estuarine, Coastal and Shelf Science*, 123: 34–38.
- Copello, S., Seco Pon, J. P., and Favero, M. 2014. Spatial overlap of Black-browed albatrosses with longline and trawl fisheries in the Patagonian Shelf during the non-breeding season. *Journal of Sea Research*, 89: 44–51.
- Crespo, E. A., Dans, S., Koen Alonso, M., and Pedraza, S. 2007. Interacciones entre mamíferos marinos y pesquerías. In *El mar argentino y sus recursos pesqueros*, volumen 5, pp. 51–169. Ed. by J. I. Carreto, and C. Bremec. Instituto Nacional de Investigación y Desarrollo Pesquero INIDEP, Mar del Plata. 169 pp.
- Crespo, E. A., and Pedraza, S. N. 1991. Estado actual y tendencia de la población de lobos marinos de un pelo (*Otaria flavescens*) en el litoral norpatagónico. *Ecología Austral*, 1: 87–95.
- Croxall, J. P., Butchart, S. H., Lascelles, B., Stattersfield, A. J., Sullivan, B., Symes, A., and Taylor, P. 2012. Seabird conservation status, threats and priority actions: a global assessment. *Bird Conservation International*, 22: 1–34.
- Croxall, J. P., and Wood, A. 2002. The importance of the Patagonian Shelf for top predator species breeding at South Georgia. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 12: 101–118.
- Denuncio, P., Bastida, R., Dassis, M., Giardino, G., Gerpe, M., and Rodríguez, D. 2011. Plastic ingestion in Franciscana dolphins, *Pontoporia blainvillei* (Gervais and d’Orbigny, 1844), from Argentina. *Marine Pollution Bulletin*, 62: 1836–1841.

- Elith, J., Phillips, S. J., Hastie, T., Dudík, M., Chee, Y. E., and Yates, C. J. 2011. A statistical explanation of MaxEnt for ecologists. *Diversity and Distributions*, 17: 43–57.
- Estrades, A., López-Mendilaharsu, M., and Fallabrino, A. 2007. *Dermochelys coriacea* (Leatherback Sea turtle). *Diet. Herpetological Review*, 38: 330.
- Falabella, V., Campagna, C., and Croxall, J. P. 2009. Atlas del Mar Patagónico. Especies y espacios. Wildlife Conservation Society and BirdLife International, Buenos Aires. 204 pp. <http://atlas-marpatagonico.org/atlas.html> (last accessed 13 July 2015).
- FAO Fisheries Department (FAO-FI). Review of the state of world marine fishery resources. 2011. Marine resources – Southwest Atlantic, 2009. FIRMS Reports. In *Fishery Resources Monitoring System (FIRMS)*. Rome. <http://firms.fao.org/firms/resource/13329/en> (last accessed 13 July 2015).
- Favero, M., Blanco, G., Copello, S., Pon, J. P. S., Patterlini, C., Mariano-Jelicich, R., García, G., et al. 2013. Seabird bycatch in the Argentinean demersal longline fishery, 2001–2010. *Endangered Species Research*, 19: 187–199.
- Fossette, S., Girard, C., López-Mendilaharsu, M., Miller, P., Domingo, A., Evans, D., Kelle, L., et al. 2010. Atlantic leatherback migratory paths and temporary residence areas. *PLoS ONE*, 5: e13908.
- Fossette, S., Witt, M., Miller, P., Nalovic, M., Albareda, D., Almeida, A., Broderick, A., et al. 2014. Pan-Atlantic analysis of the overlap of a highly migratory species, the leatherback turtle, with pelagic longline fisheries. *Proceedings of the Royal Society B: Biological Sciences*, 281: 20133065.
- Franco-Trecu, V., Drago, M., Riet-Sapirza, F. G., Parnell, A., Frau, R., and Inchausti, P. 2013. Bias in diet determination: incorporating traditional methods in Bayesian mixing models. *PLoS ONE*, 8: e80019.
- Giardino, G. V., Mandiola, M. A., Bastida, J., Denuncio, P. E., Bastida, R. O., and Rodríguez, D. H. 2014. Travel for sex: long-range breeding dispersal and winter haulout fidelity in southern sea lion males. *Mammalian Biology*, 81: 89–95.
- González Carman, V., Acha, E. M., Maxwell, S. M., Albareda, D., Campagna, C., and Mianzan, H. 2014a. Young green turtles, *Chelonia mydas*, exposed to plastic in a frontal area of the SW Atlantic. *Marine Pollution Bulletin*, 78: 56–65.
- González Carman, V., Álvarez, K., Prosdociimi, L., Inchaurrega, M. C., Dellacasa, R. F., Faiella, A., Echenique, C., et al. 2011. Argentinian coastal waters: a temperate habitat for three species of threatened sea turtles. *Marine Biology Research*, 7: 500–508.
- González Carman, V., Botto, F., Gaitán, E., Albareda, D., Campagna, C., and Mianzan, H. 2014b. A jellyfish diet for the herbivorous green turtle *Chelonia mydas* in the temperate SW Atlantic. *Marine Biology*, 161: 339–349.
- González Carman, V., Bruno, I., Maxwell, S., Álvarez, K., Albareda, D., Acha, E. M., and Campagna, C. 2016. Habitat use, site fidelity and conservation opportunities for juvenile loggerhead sea turtles in the Río de la Plata, Argentina. *Marine Biology*, 163: 20.
- González Carman, V., Falabella, V., Maxwell, S., Albareda, D., Campagna, C., and Mianzan, H. 2012. Revisiting the ontogenetic shift paradigm: the case of juvenile green turtles in the SW Atlantic. *Journal of Experimental Marine Biology and Ecology*, 429: 64–72.
- González Carman, V., Machain, N., and Campagna, C. 2015. Legal and institutional tools to mitigate plastic pollution affecting marine species: Argentina as a case study. *Marine Pollution Bulletin*, 92: 125–133.
- Grandi, M. F., Oliveira, L., Dans, S. L., and Crespo, E. A. 2012. A hunted population in recovery: effective population size for South American sea lions from Patagonia. *Animal Biology*, 62: 433–450.
- Guerrero, R. A., Piola, A., Molinari, G., and Osiroff, A. 2010. Climatología de temperatura y salinidad en el Río de la Plata y su Frente Marítimo, Argentina-Uruguay. Instituto Nacional de Investigación y desarrollo Pesquero INIDEP, Mar del Plata. 95 pp.
- Guilford, T., Meade, J., Willis, J., Phillips, R. A., Boyle, D., Roberts, S., Collett, M., et al. 2009. Migration and stopover in a small pelagic seabird, the Manx shearwater *Puffinus puffinus*: insights from machine learning. *Proceedings of the Royal Society of London, Series B: Biological Sciences*, 276: 1215–1223.
- Hedd, A., Montevecchi, W. A., Otley, H., Phillips, R. A., and Fifield, D. A. 2012. Trans-equatorial migration and habitat use by sooty shearwaters *Puffinus griseus* from the South Atlantic during the nonbreeding season. *Marine Ecology Progress Series*, 449: 277–290.
- Heithaus, M. R., Frid, A., Wirsing, A. J., and Worm, B. 2008. Predicting ecological consequences of marine top predator declines. *Trends in Ecology & Evolution*, 23: 202–210.
- Hijsmans, R. J., Phillips, S., Leathwick, J., and Elith, J. 2013. dismo: Species distribution modeling version 0.8–5. <http://CRAN.Rproject.org/package=dismo> (last accessed 12 December 2015).
- IUCN. 2014. IUCN Red List of Threatened Species. Version 2014.2. <http://www.iucnredlist.org/> (last accessed 13 July 2015).
- Jackson, J. 2001. What was natural in the coastal oceans? *Academy of Sciences of the United States of America. Proceedings of the National Academy of Sciences of the United States of America*, 98: 5411–5418.
- Jackson, J. B. C., Kirby, M. X., Berger, W. H., Bjørndal, K. A., Botsford, L. W., Bourque, B. J., Bradbury, R. H., et al. 2001. Historical overfishing and the recent collapse of coastal ecosystems. *Science*, 293: 629–638.
- Jaureguizar, A. J., Bava, J., Carroza, C. R., and Lasta, C. 2003. Distribution of whitemouth croaker *Micropogonias furnieri* in relation to environmental factors at the Río de la Plata estuary, South America. *Marine Ecology Progress Series*, 255: 271–282.
- Jiménez, S., Domingo, A., Brazeiro, A., Defeo, O., and Phillips, R. A. 2015. Marine debris ingestion by albatrosses in the southwest Atlantic Ocean. *Marine Pollution Bulletin*, 96: 149–154.
- López-Mendilaharsu, M., Rocha, C. F. D., Miller, P., Domingo, A., and Prosdociimi, L. 2009. Insights on leatherback turtle movements and high use areas in the Southwest Atlantic Ocean. *Journal of Experimental Marine Biology and Ecology*, 378: 31–39.
- Lucas, A. J., Guerrero, R. A., Mianzan, H. W., Acha, E. M., and Lasta, C. A. 2005. Coastal oceanographic regimes of the Northern Argentine Continental Shelf (34–43°S). *Estuarine, Coastal and Shelf Science*, 65: 405–420.
- Mariano-Jelicich, R., Copello, S., Seco Pon, J., and Favero, M. 2014. Contribution of fishery discards to the diet of the Black-browed albatross (*Thalassarche melanophris*) during the non-breeding season: an assessment through stable isotope analysis. *Marine Biology*, 161: 119–129.
- Martinez Souza, G. 2009. Ecología alimentar da tartaruga marinha cabeçuda (*Caretta caretta*) no Oceano Atlântico Sul Ocidental, Uruguai. Universidade Federal do Rio Grande—FURG, Rio Grande.
- McClenachan, L., Jackson, J. B. C., and Newman, M. J. H. 2006. Conservation implications of historic sea turtle nesting beach loss. *Frontiers in Ecology and the Environment*, 4: 290–296.
- Medway, D. G. 1998. Human induced mortality of Southern Ocean Albatrosses at sea in the 19th century: a historical review. In *Albatross Biology and Conservation*, pp. 189–198. Ed. by G. Robertson, and R. Gales. Surrey Beatty and sons Pty Limited, Chipping Norton.
- Mianzan, H. W., and Guerrero, R. A. 2000. Environmental patterns and biomass distribution of gelatinous macrozooplankton. Three Study Cases in the South-Western Atlantic Ocean. *Scientia Marina*, 64: 215–224.
- Mianzan, H. W., Lasta, C., Acha, E., Guerrero, R., Macchi, G., and Bremec, C. 2001. The Río de la Plata Estuary, Argentina-Uruguay. *Ecological Studies*, 144: 185–204.
- Muelbert, J. H., Acha, M., Mianzan, H., Guerrero, R., Reta, R., Braga, E. S., García, V. M. T., et al. 2008. Biological, physical and chemical properties at the Subtropical Shelf Front Zone in the SW Atlantic Continental Shelf. *Continental Shelf Research*, 28: 1662–1673.

- Naya, D. E., Arim, M., and Vargas, R. 2002. Diet of South American fur seals (*Arctocephalus australis*) in Isla de Lobos, Uruguay. *Marine Mammal Science*, 18: 734–745.
- Naya, D. E., Vargas, R., and Arim, M. 2000. Preliminary analysis of southern sea lion (*Otaria flavescens*) diet in Isla de Lobos, Uruguay. *Boletín de la Sociedad Zoológica del Uruguay* (2da Epoca), 12: 14–21.
- Nicholls, D., Robertson, C., Prince, P., Murray, M., Walker, K., and Elliott, G. 2002. Foraging niches of three Diomedea albatrosses. *Marine Ecology Progress Series*, 231: 269–277.
- OBPG. 2002. MODIS Aqua Level 3 SST Thermal IR Monthly 4 km Daytime. Ver. 1. PO.DAAC, CA, USA. <https://podaac.jpl.nasa.gov/dataset/> (last accessed 11 December 2015).
- Olson, D. B., Podestá, G. P., Evans, R. H., and Brown, O. B. 1988. Temporal variations in the separation of Brazil and Malvinas Currents. *Deep Sea Research Part A. Oceanographic Research Papers*, 35: 1971–1990.
- Phillips, S. J., Anderson, R. P., and Schapire, R. E. 2006. Maximum entropy modeling of species geographic distributions. *Ecological Modelling*, 190: 231–259.
- Pikesley, S. K., Maxwell, S. M., Pendoley, K., Costa, D. P., Coyne, M. S., Formia, A., Godley, B. J., et al. 2013. On the front line: integrated habitat mapping for olive Ridley sea turtles in the southeast Atlantic. *Diversity and Distributions*, 19(12): 1518–1530.
- Piola, A. R., Campos, E. J. D., Möller, O. O., Jr., Charo, M., and Martinez, C. 2000. Subtropical Shelf Front off eastern South America. *Journal of Geophysical Research*, 105: 6565–6578.
- Piola, A. R., Matano, R. P., Palma, E. D., Möller, O. O., Jr., and Campos, E. J. D. 2005. The influence of the Plata River discharge on the western South Atlantic shelf. *Geophysical Research Letters*, 32: L01603.
- Piola, A. R., Romero, S. I., and Zajaczkowski, U. 2008. Space-time variability of the Plata plume inferred from ocean color. *Continental Shelf Research*, 28: 1556–1567.
- Ponce de León, A., and PIN, O. D. 2006. Distribución, reproducción y alimentación del lobo fino *Arctocephalus australis* y del león marino *Otaria flavescens* en Uruguay. In *Bases para la conservación y el manejo de la costa uruguaya*, pp. 305–313. Ed. By R. Menafrá, L. Rodríguez-Gallego, F. Scarabino, and D. Conde. Vida Silvestre Uruguay, Montevideo.
- Quintana, F., Dell’Arciprete, O. P., and Copello, S. 2010. Foraging behavior and habitat use by the Southern Giant Petrel on the Patagonian Shelf. *Marine Biology*, 157: 515–525.
- R Developmental Core Team. 2013. <http://cran.r-project.org> (last accessed 13 July 2015).
- Ramos, R., Granadeiro, J. P., Rodríguez, B., Navarro, J., Paiva, V. H., Bécas, J., Reyes-González, J. M., et al. 2013. Meta-population feeding grounds of Cory’s shearwater in the subtropical Atlantic Ocean: implications for the definition of Marine Protected Areas based on tracking studies. *Diversity and Distributions*, 19: 1284–1298.
- Reid, T. A., Ronconi, R. A., Cuthbert, R. J., and Ryan, P. G. 2014. The summer foraging ranges of adult spectacled petrels *Procellaria conspicillata*. *Antarctic Science*, 26: 23–32.
- Roberts, J. J., Best, B. D., Dunn, D. C., Trembl, E. A., and Halpin, P. N. 2010. Marine Geospatial Ecology Tools: an integrated framework for ecological geoprocessing with ArcGIS, Python, R, MATLAB, and C++. *Environmental Modelling & Software*, 25: 1197–1207.
- Rodríguez, D., and Bastida, R. 1998. Four hundred years in the history of pinniped colonies around Mar del Plata, Argentina. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 8: 721–735.
- Rodríguez, D. H., Dassis, M., Ponce de León, A., Barreiro, C., Farenga, M., Bastida, R. O., and Davis, R. W. 2013. Foraging strategies of Southern sea lion females in the La Plata River Estuary (Argentina-Uruguay). *Deep Sea Research Part II: Topical Studies in Oceanography*, 88: 120–130.
- Romero, S. I., Piola, A. R., Charo, M., and Garcia, C. A. E. 2006. Chlorophyll-a variability off Patagonia based on SeaWiFS data. *Journal of Geophysical Research*, 111: C05021.
- Ronconi, R. A., Ryan, P. G., and Ropert-Coudert, Y. 2010. Diving of great shearwaters (*Puffinus gravis*) in cold and warm water regions of the South Atlantic Ocean. *PLoS ONE*, 5: e1550810.
- Sánchez, R., and de Ciechomski, J. D. 1995. Spawning and nursery grounds of pelagic fish species in the sea-shelf off Argentina and adjacent areas. *Scientia Marina*, 59: 455–478.
- Sánchez, R. P., Navarro, G., Monsalvo, M., and Martínez Puljak, G. 2011. Operatoria de la flota argentina dirigida a los recursos corvina y merluza en la Zona Común de Pesca Argentino-Uruguaya. *Alternativas A la Pesca de los Recursos Objetivo. Frente Marítimo*, 2: 71–129.
- Scales, K. L., Miller, P. I., Hawkes, L. A., Ingram, S. N., Sims, D. W., and Votier, S. C. 2014. On the front line: frontal zones as priority at-sea conservation areas for mobile marine vertebrates. *Journal of Applied Ecology*, 51: 1575–1583.
- Secchi, E. 2010. Review on the threats and conservation status of franciscana, *Pontoporia blainvilliei* (Cetacea, Pontoporiidae). In *Biology, Evolution and Conservation of River Dolphins within South America and Asia*, 1st edn, pp. 323–339. Ed. by M. Ruiz-García, and J. Shostell. Nova Science Publishers Inc, Hauppauge, NY.
- Seco Pon, J. P., Copello, S., Moretinni, A., Lértora, H. P., Bruno, I., Bastida, J., Mauco, L., et al. 2013. Seabird and marine-mammal attendance and by-catch in semi-industrial trawl fisheries in near-shore waters of northern Argentina. *Marine and Freshwater Research*, 64: 237–248.
- Seco Pon, J. P., Copello, S., Tamini, L., Mariano-Jelicich, R., Paz, J., Blanco, G., and Favero, M. 2015. Seabird conservation in fisheries: current state of knowledge and conservation needs for Argentine high-seas fleets. In *Seabirds and Songbirds: Habitat Preference, Conservation and Migratory Behavior*, 1st edn, pp. 45–88. Ed. by G. Mahala. Nova Science Publishers, New York. 172 pp.
- Spalding, M. D., Fox, H. E., Allen, G. R., Davidson, N., Ferdaña, Z. A., Finlayson, M., Halpern, B. S., et al. 2007. Marine ecoregions of the world: a bioregionalization of coastal and shelf areas. *BioScience*, 57: 573–583.
- Tourinho, P. S., Ivar do Sul, J. A., and Fillmann, G. 2010. Is marine debris ingestion still a problem for the coastal marine biota of southern Brazil? *Marine Pollution Bulletin*, 60: 396–401.
- Warren, D. L., Glor, R. E., and Turelli, M. 2008. Environmental niche equivalency versus conservatism: quantitative approaches to niche evolution. *Evolution*, 62: 2868–2883.
- Warren, D. L., Glor, R. E., and Turelli, M. 2010. ENMTools: a toolbox for comparative studies of environmental niche models. *Ecography*, 33: 607–611.
- Young, N., Carter, L., and Evangelista, P. 2011. A MaxEnt Model v3.3.3e Tutorial (ArcGIS v10). http://ibis.colostate.edu/WebContent/WS/ColoradoView/TutorialsDownloads/A_Maxent_Model_v7.pdf (last accessed 24 September 2015).

Handling editor: Marta Coll