

Tagging Rate at Lower Granite Dam

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1 Introduction

We are interested in determining the relationship between the number of PIT tags implanted in adults at Lower Granite Dam (LGR) and the precision of abundance estimates at the technical recovery team (TRT) population scale for Snake River Basin spring/summer Chinook salmon ESU and steelhead DPS. The number of PIT tags to implant at LGR can then be used to make inferences about trapping and tagging rates. We also consider the precision of sex and total age specific abundance estimates, which are necessary for estimating population productivity (e.g., recruits per spawner/female). Finally, we consider a theoretical scenario in which the detection efficiencies across the basin were increased by 5-20%.

2 Methods

First, we wanted to leverage results based on empirical data from the inception of the PIT tagging program at LGR (2010) to the present. We started by gathering data on all estimates that have been made for population abundance at the TRT spacial scale, and examining the relationship between the number of tags observed in each population, and the coefficient (CV) of the estimate. We fit a model on the semi-log scale:

$$CV \sim e^{a+b*n_{tags}}$$

$$\log(CV) \sim a + b * n_{tags}$$

the log-log scale:

$$CV \sim a * n_{tags}^b$$

$$\log(CV) \sim \log(a) + b * \log(n_{tags})$$

and, in addition, a more flexible version using a loess spline. We were interested in how many tags would need to be detected within each TRT population, and using each model, to predict a CV of ≤ 0.15 . Similarly, we examined the relationship between the proportion of all tags that were deployed at LGR that were detected within each population, and the CV of the population estimate.

The previous provides information on the number of tags, or the proportion of all deployed tags, returning to a population necessary to achieve a precise abundance estimate. We then need to determine how many tags would need to be deployed at LGR to achieve that minimum number of tags to each TRT population to achieve that level of precision in our abundance estimates. To do this, we started by looking at the proportion of tags deployed at LGR that were detected in each TRT population by year, and then averaged across years. Finally, given the minimum number of detections required to achieve a precise abundance estimate, we estimated the number of total tags needed to be deployed at LGR that would be expected to result in the minimum number of detections for that TRT population. We can also predict for a given number of tags deployed at LGR whether there will be sufficient detections in each TRT population to achieve a reasonably precise abundance estimate.

2.1 Populations with No Detections

There are a number of TRT populations for both species that currently lack any PIT tag detection infrastructure, and are therefore not represented in DABOM estimates at all. These are listed in Table 1, although we recognize that IPTDS have been installed in a small group of these recently (e.g., Marsh Creek).

Table 1: Populations, by species, with no PIT tag detection infrastructure and therefore which are not accounted for by DABOM.

Species	MPG_DPS	TRT	Name	Status
Steelhead	Salmon	SRCHA-s	Chamberlain Creek	Extant
Steelhead	Clearwater	CRNFC-s	North Fork Clearwater River	Extirpated
Steelhead	Hells Canyon	SNHCT-s	Hells Canyon tributaries	Extirpated
Chinook	Upper Salmon	SRLMA	Upper Salmon lower main stem	Extant
Chinook	Middle Fork Salmon	MFLMA	Lower Middle Fork main stem	Extant
Chinook	Middle Fork Salmon	MFCAM	Camas Creek	Extant
Chinook	Middle Fork Salmon	MFLOO	Loon Creek	Extant
Chinook	Middle Fork Salmon	MFSUL	Sulphur Creek	Extant
Chinook	Middle Fork Salmon	MFMAR	Marsh Creek	Extant
Chinook	Middle Fork Salmon	MFUMA	Upper Middle Fork main stem	Extant
Chinook	Middle Fork Salmon	SRCHA	Chamberlain Creek	Extant
Chinook	Grande Ronde Imnaha	GRMIN	Minam River	Extant

2.2 Populations with Low Detections

When we initially examined the proportion of all tags deployed at LGR that were detected in each TRT population, by species, we noted a number of populations had very low proportions (e.g., $< 2\%$) of tags observed within them. The low proportion of tags observed within a population can be attributed to three primary reasons:

1. True low spawner abundance.
2. Low detection probability among sites within the population.
3. Detection infrastructure within the population was not intended to monitor natural origin spawner abundance (e.g., hatchery rack detections).

In practice, detection probabilities at instream PIT tag detection systems (IPTDS) intended for natural origin population abundance monitoring have high detection probabilities, so we further examined populations with low proportions of detections and whether that was due to reasons #1 or #3 (or a combination thereof) above. In addition, the proportion of detections within a population was averaged across years, and thus, we

considered populations where infrastructure was installed in a later year. Table 2 and the following narrative provide a summary of populations and years we chose to remove from the analysis due to non-existent IPTDS or infrastructure not intended for natural population monitoring. We additionally note some populations with low proportions of detections that we believe to be due to low population size.

Table 2: Populations, by species, with a low number of total tags deployed at LGR detected within. These populations were further examined as to whether low detections were due to true low population size or infrastructure not intended for population monitoring. Species, TRT population, and year combinations containing a [R] were removed from analysis because IPTDS did not exist or were not intended for population monitoring.

Species	TRT	Sites	2013	2014	2015	2016	2017	2018	2019
Chinook	CRLOC	LRL, LRU, FISTRP	R	R	R	R	R	-	-
Chinook	CRLOL	LC1, LC2	-	-	-	-	-	-	-
Chinook	GRCAT	CCW, CATHEC	R	R	-	-	-	-	-
Chinook	GRLOO	LOOKGC	R	R	R	R	R	R	-
Chinook	GRUMA	UGS, GRANDW	R	R	R	R	R	-	-
Chinook	GRWEN	WEN	R	R	R	R	R	R	-
Chinook	IRBSH	BSC, CMP, LSHEEF	-	-	-	-	-	-	-
Chinook	SEMEA	SW1, SW2	R	R	R	R	R	-	-
Chinook	SNTUC	LTR, MTR, UTR, TUCH	R	R	R	R	R	R	-
Chinook	SREFS	SALEFT	R	R	R	R	R	R	-
Chinook	SRLSR	RAPH	R	R	R	R	R	R	-
Chinook	SRNFS	NFS	R	R	R	R	R	R	-
Chinook	SRPAH	PAHH	R	R	R	R	R	R	-
Chinook	SRPAN	PCA	R	R	R	R	-	-	-
Chinook	SRYFS	YFK	-	-	-	-	-	-	-
Steelhead	CRLOC-s	LRL, LRU, FISTRP	R	R	R	R	R	-	-
Steelhead	CRSEL-s	SW1, SW2	R	R	R	R	R	-	-
Steelhead	SFSEC-s	ZEN, LAKEC	-	-	-	-	-	-	-
Steelhead	SNTUC-s	LTR, MTR, UTR, TUCH	R	R	R	R	R	R	R
Steelhead	SREFS-s	SALEFT	R	R	R	R	R	R	R
Steelhead	SRLSR-s	RAPH	R	R	R	R	R	R	R
Steelhead	SRNFS-s	NFS	R	R	R	R	R	R	-
Steelhead	SRPAH-s	PAHH	R	R	R	R	R	R	R
Steelhead	SRUMA-s	RFL, STL, VC1, VC2, YFK	-	-	-	-	-	-	-

2.2.1 Chinook

Table 3: Reasons for certain Chinook populations being excluded from this analysis.

TRT	Notes
CRLOC	LRL installed October 2016. LRU installed December 2017.
CRLOL	Not removed from analysis. Poor precision likely due to small population size.
GRCAT	CCW installed in March 2015.
GRLOO	Only includes ad-intact adults that return to the Lookingglass Hatchery (LOOKGC).
GRUMA	UGS site installed in November 2017. Complete data only for 2018.
GRWEN	Tandem array was installed in 2018, but blew out soon after. Re-installed in 2019. No long term data.
IRBSH	Not removed from analysis. Poor precision likely due to small population size.
SEMEA	SW1 installed December 2016. SW2 installed September 2017.

TRT	Notes
SNTUC	Population occurs below LGR.
SREFS	Array installed in 2016, but blown out soon after and not re-installed.
SRLSR	Only includes ad-intact adults that return to the Rapid River Hatchery weir (RAPH).
SRNFS	Installed in 2016, but performed poorly initially. System upgraded in 2018.
SRPAH	Only includes ad-intact adults that return to the Pahsimeroi Hatchery weir (PAHH).
SRPAN	Site installed in July 2017.
SRYFS	Not removed from analysis. Poor precision likely due to small population size.

Also note that the GRLOS, GRMIN, GRCAT, and GRUMA Chinook salmon TRT populations all occur above the UGR array; however, we don't know which population an adult is destined for after passing UGR unless it is detected again upstream. As a result, Chinook salmon adults that are last observed at the UGR site cannot be used in population abundance estimates. One might consider how final observations at UGR could be leveraged in the future for upriver abundance estimates.

2.2.2 Steelhead

Table 4: Reasons for certain steelhead populations being excluded from this analysis.

TRT	Notes
CRLOC-s	LRL installed October 2016. LRU installed December 2017. Remaining abundance estimates from FISTRP.
CRSEL-s	SW1 installed December 2016. SW2 installed September 2017.
SFSEC-s	Not removed from analysis. Believe poor precision is due to low population size.
SNTUC-s	Population occurs below LGR.
SREFS-s	Array installed in 2016, but blown out soon after and not re-installed.
SRLSR-s	Only includes ad-intact adults that return to the Rapid River Hatchery weir (RAPH).
SRNFS-s	Installed in 2016, but performed poorly initially. System upgraded in 2018.
SRPAH-s	Only includes ad-intact adults that return to the Pahsimeroi Hatchery weir (PAHH).
SRUMA-s	Not removed from analysis. Believe poor precision is due to low population size.

3 Results

3.1 Number of Tags Required in a TRT Population

Figure 1 shows the number of PIT tags detected within a TRT population and the corresponding CV of the population abundance estimate, by species, using results going back to 2013 and after removing population and year combinations in which low/no observations were due to lack of IPTDS or IPTDS not intended for natural origin population abundance (Table 2).

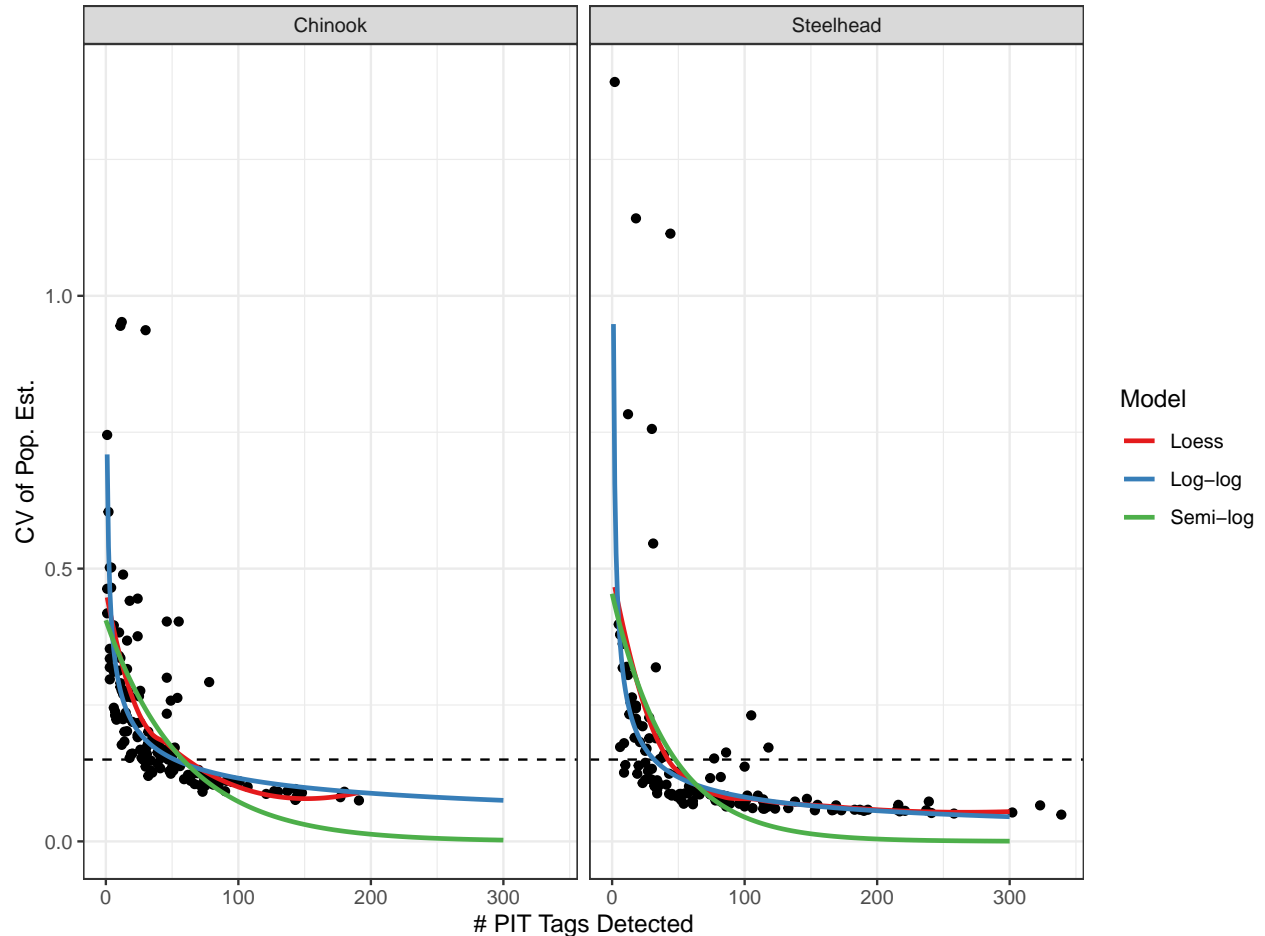


Figure 1: Scatterplot of the number of PIT tags detected in each population and the CV of the abundance estimates for that population. The lines of semi-log, log-log and loess fits to the data. The dashed horizontal line indicates a CV of 15%.

Additionally, using the same data, we examined the proportion of total tags deployed at LGR that were observed within a TRT population and the CV of the abundance estimate (Figure 2).

Ultimately, we were interested in the number of PIT tags that need to be observed within a population or the proportion of total PIT tags deployed at LGR detected within a population necessary to achieve a predicted CV of ≤ 0.15 for the abundance estimate. To do this, we tried fitting a separate model for each species, and after fitting those models (semi-log, log-log and loess), we examined the minimum number and proportion of total tags that would need to be detected to achieve that CV level. Those results are shown in Tables 5 and 6.

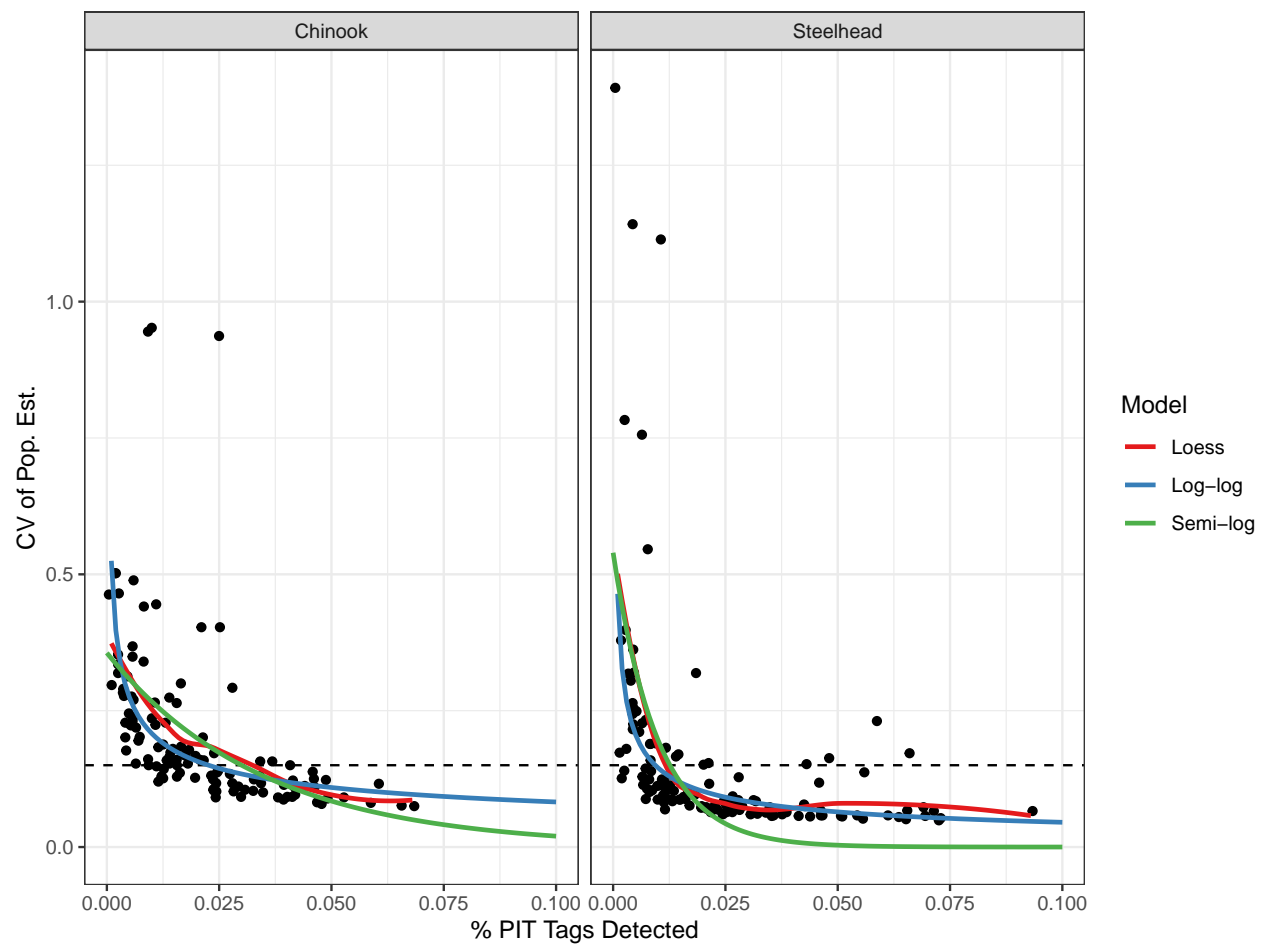


Figure 2: Scatterplot of the proportion of PIT tags detected in each population and the CV of the abundance estimate for that population. The lines are semi-log, log-log and Loess fits to that data. The dashed horizontal line denotes a CV of 15%.

Table 5: Estimated number of PIT tags that need to be detected within a TRT population to achieve a CV of 15% or less for abundance estimates.

Species	Model	# Tags	Pred. CV
Chinook	Loess	61	0.150
Chinook	Log-log	52	0.150
Chinook	Semi-log	58	0.150
Steelhead	Loess	43	0.147
Steelhead	Log-log	32	0.149
Steelhead	Semi-log	48	0.149

Table 6: The estimated proportion of total tags deployed at LGR that need to be detected within a TRT population to achieve a CV of 15% or less for abundance estimates.

Species	Model	% Tags Detected	Pred. CV
Chinook	Loess	3.3 %	0.148
Chinook	Log-log	2.3 %	0.149
Chinook	Semi-log	3 %	0.150
Steelhead	Loess	1.2 %	0.144
Steelhead	Log-log	1 %	0.145
Steelhead	Semi-log	1.3 %	0.145

Table 5 shows that we need to detect about 45-55 Chinook salmon and 30-46 steelhead to achieve a CV of 15% or less for natural origin abundance estimates. Further, Table 6 shows that, using past results, we need to detect about 2-3% of Chinook salmon or about 1% of steelhead PIT tags deployed at LGR within a population to achieve reasonably precise estimates of abundance. Looking at model diagnostics (not shown), the log-log model appears to meet the linear model assumptions the best. For the sake of simplicity, let's assume that we need to detect 45 PIT tags within a population, for both species, to achieve a precise abundance estimate (Figure 3).

3.1.1 Sex and Age Estimates

We want to note that reasonably precise estimates of abundance, by sex and total age, are also a desired outcome as they are necessary to construct brood tables and estimate population productivity. However, note that it will not require 45 female tags detected within a population to achieve reasonable precision of female abundance. Rather, we will take the total abundance estimate to a TRT population and multiply it by a binomial (female vs. male) proportion to generate abundance by females and males. The binomial proportions will be estimated from all of the tags detected in that TRT population. If we have about 45 total tags detected in a population for a good abundance estimate, the proportion female or male will be fairly precise because 45 is a good sample size for proportions, especially when the proportion is somewhere near 50%. Therefore, the CV for the number of females in that population will be slightly larger than the CV of the total abundance, but probably not by much.

Similar arguments hold for abundance by total age (multinomial proportions), although as we attempt to account for more age classes, the precision will suffer. Therefore, for a Chinook salmon population with 3 total age classes, we expect to get reasonable estimates with a sample size close to 45. However, for a steelhead population with up to 6 or so total age classes, the precision will suffer, especially since some of the total age classes may have very few fish. However, sex and age classes with proportions outside the tails (e.g., > 10%) would likely have good precision.

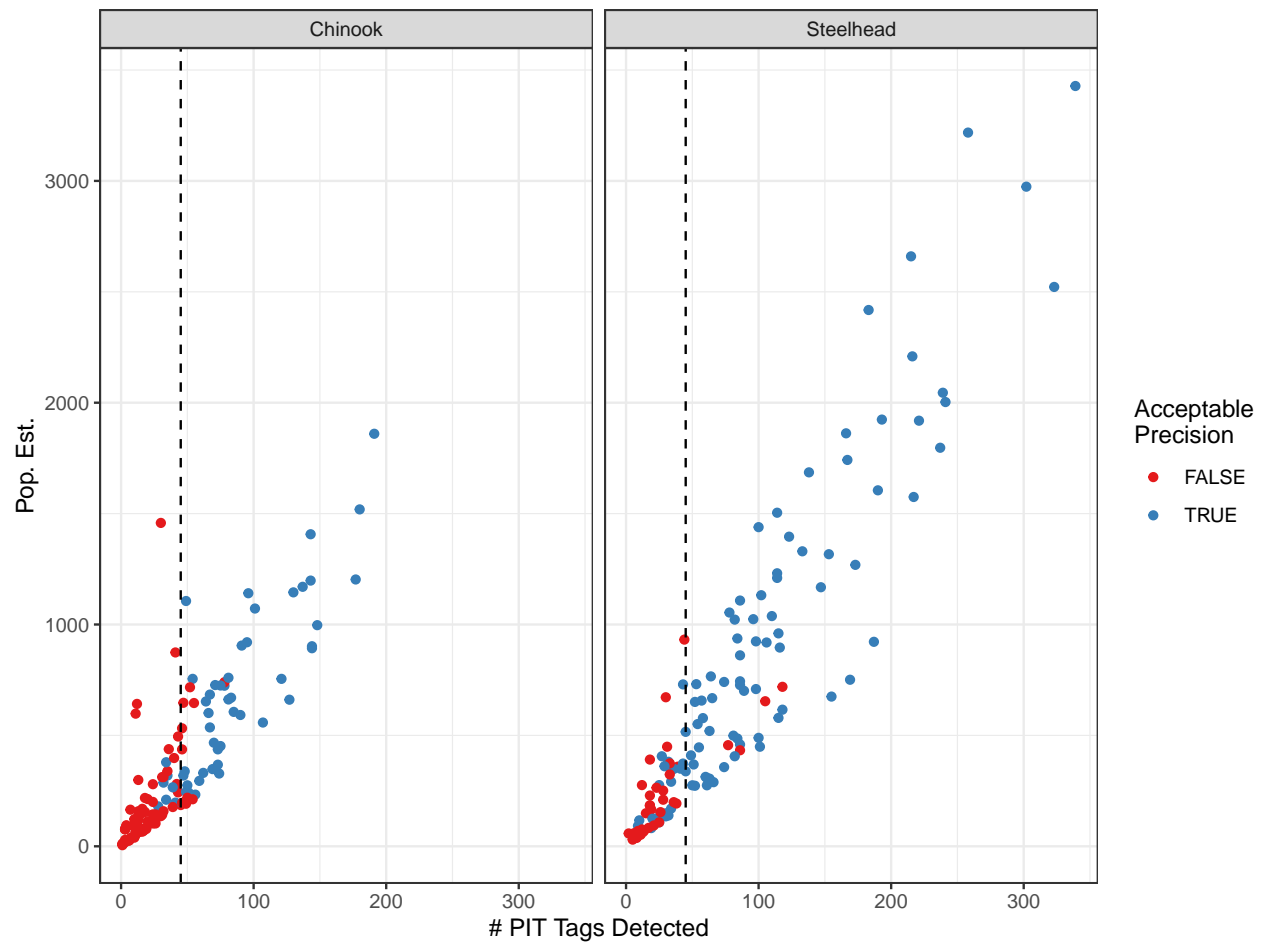


Figure 3: The number of PIT tags detected in a TRT population and the corresponding population abundance estimate, by species, using results back to spawn year 2013. Estimates shown in blue had a CV of 15% or less. The dashed vertical line denotes 45 PIT tags detected.

3.2 Total Tags at Lower Granite

We now have an estimate of the number of PIT tags that need to be detected within a TRT population (45) to achieve an abundance estimate with reasonable precision. But now we want to know how many PIT tags would need to be deployed at LGR to observe 45 PIT tags within any population. The proportion of all PIT tags deployed at LGR that were detected in any TRT population has changed over time, due to more IPTDS infrastructure being installed across the Snake River Basin (Figure 4), and thus, we chose to focus on the period from 2013 on, since the infrastructure has been more stable since then. Since 2013, the average proportion of all tags deployed from LGR that were detected in a TRT population, by species, is shown in Table 7.

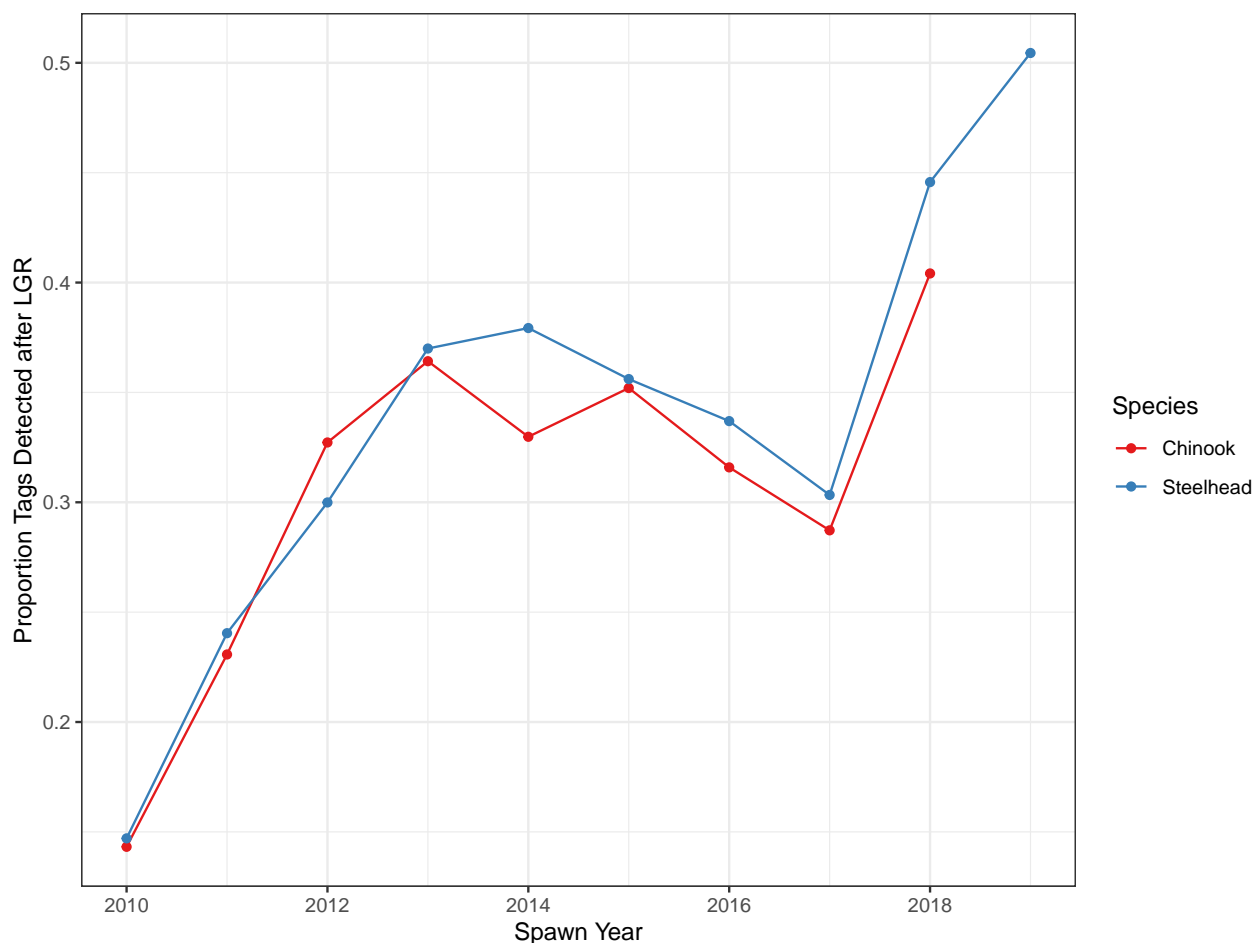


Figure 4: The proportion of all PIT tags deployed at LGR that were later detected within a TRT population each year, by species.

Table 7: The average proportion of tags deployed at LGR detected in a TRT population since 2013, by species.

Species	Avg. % Total Tags Detected
Chinook	34.2 %
Steelhead	38.5 %

Next, Table 8 summarizes the total number of tags that would need to be deployed at LGR to achieve a precise abundance estimate for any given TRT population. Again, Table 8 only contains TRT populations that

contain infrastructure intended for natural origin population abundance monitoring, and within populations, years in which IPTDS were not installed were excluded (Table 2).

Table 8: The expected number of tags needed to be deployed at LGR to achieve a minimum number of detections in each TRT population leading to reasonably precise estimate of abundance.

Species	TRT	Name	% Tags Detected	Min. # Tags	Total Tags N
Chinook	IRBSH	Big Sheep Creek	0.003	45	
Chinook	CRLOL	Lolo Creek	0.008	45	
Chinook	SRYFS	Yankee Fork	0.008	45	
Chinook	SRUMA	Salmon River upper mainstem above Redfish Lake	0.016	45	
Chinook	SRVAL	Valley Creek	0.018	45	
Chinook	SCUMA	Upper South Fork Clearwater	0.021	45	
Chinook	SRLEM	Lemhi River	0.022	45	
Chinook	MFBEA	Bear Valley Creek	0.032	45	
Chinook	IRMAI	Imnaha River mainstem	0.039	45	
Chinook	MFBIG	Big Creek	0.039	45	
Chinook	SFEFS	East Fork South Fork Salmon River	0.040	45	
Chinook	SFSEC	Secesh River	0.043	45	
Chinook	GRLOS	Lostine River	-	45	

Species	TRT	Name	% Tags Detected	Min. # Tags	Total Tags N
Steelhead	SFSEC-s	Secesh River	0.004	45	
Steelhead	SRUMA-s	Salmon River upper mainstem	0.006	45	
Steelhead	MFBIG-s	Big, Camas, and Loon Creek	0.010	45	
Steelhead	SRLEM-s	Lemhi River	0.011	45	
Steelhead	SRPAN-s	Panther Creek	0.011	45	
Steelhead	CRLOL-s	Lolo Creek	0.011	45	
Steelhead	CRSFC-s	South Fork Clearwater River	0.019	45	
Steelhead	SFMAI-s	South Fork Salmon River	0.023	45	
Steelhead	CRLMA-s	Clearwater River lower mainstem	0.023	45	
Steelhead	SNASO-s	Asotin Creek	0.027	45	
Steelhead	GRWAL-s	Wallowa River	0.030	45	
Steelhead	GRUMA-s	Grande Ronde River upper mainstem	0.044	45	
Steelhead	GRLMT-s	Grande Ronde River lower mainstem tributaries	0.048	45	
Steelhead	GRJOS-s	Joseph Creek	0.060	45	
Steelhead	IRMAI-s	Imnaha River	0.066	45	

Finally, we summarised results by examining the percentage of TRT population abundance estimates we would expect to be reasonably precise (i.e. $CV \leq 0.15$) given a certain number of tags deployed at LGR (Figure 5). Results were similar for each species. As an example, if 2,000 PIT tags were deployed for each species at LGR, we might expect that approximately 40% of TRT population abundance estimates would have a CV of 15% or less (Figures 5 and 6); with 4,000 PIT tags per species, we'd expect that about 75% of monitored Chinook salmon and 62% of monitored steelhead population abundance estimates would be reasonably precise.

Another way to interpret Figure 5 is to see which TRT populations would or would not be expected to have a reasonably precise abundance estimate for a given number of tags. For example, if 4,000 spring/summer Chinook were tagged, we could expect reasonable abundance estimates for all populations except SRYFS, CRLOL and IRBSH. Figure 6 provides the same information as Figure 5, except is zoomed into the lower-left

portion of the plot to focus on lower numbers of tags deployed.

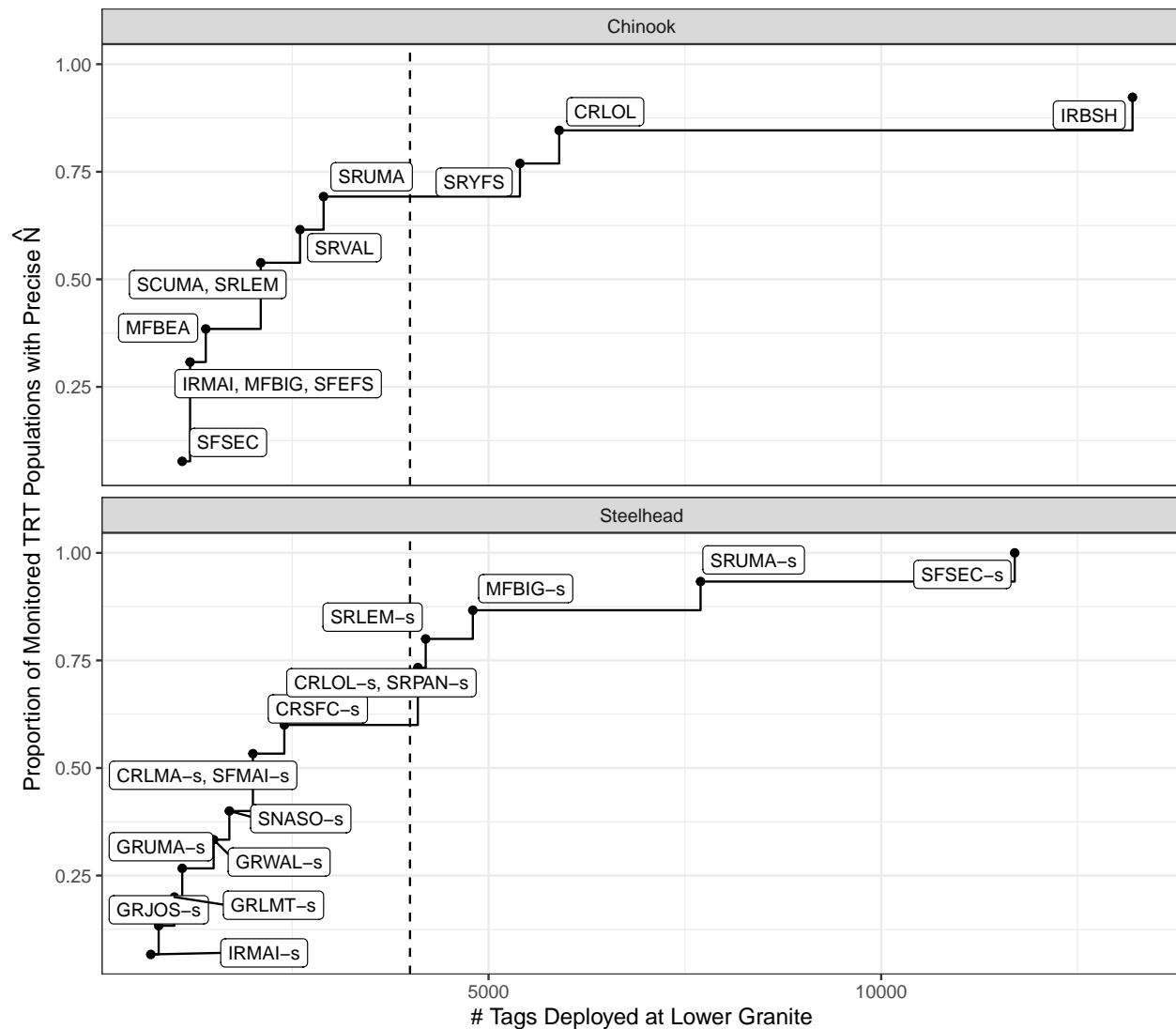


Figure 5: Expected proportion of TRT populations with a good CV of abundance estimates for a given number of PIT tags deployed at LGR, faceted by species. Dashed line shows 4,000 tags. Labels depict which additional TRT populations are expected to have a good CV of abundance as the number of tags deployed increases.

One way to increase the number of tags detected in a given population is to increase the total number of tags deployed at LGR, but another could be to increase the probability of detecting a tag assuming it does arrive in that population. This could be done by upgrading older in-stream PIT tag detection equipment, re-deploying it more effectively, or adding detection sites within the population. Without exploring specific ways that detection efficiency could be increased within each population (or even whether it could be increased by the amounts suggested) we evaluated the potential effect on the number of total tags needed to be deployed at Lower Granite if the detection probability was increased by 5, 10 or 20% across the basin.

Figure 7 shows how such increases would impact the number of tags needed to be deployed. The main message from this analysis is that improving detection efficiency, even by as much as 20% (which is actually impossible in most populations, as that would increase it to above 100%), does not reduce the number of tags needed very much. This is partly due to the fact that most populations already have high detection efficiency. However, there may be certain populations where improving the detection infrastructure could lead to gains

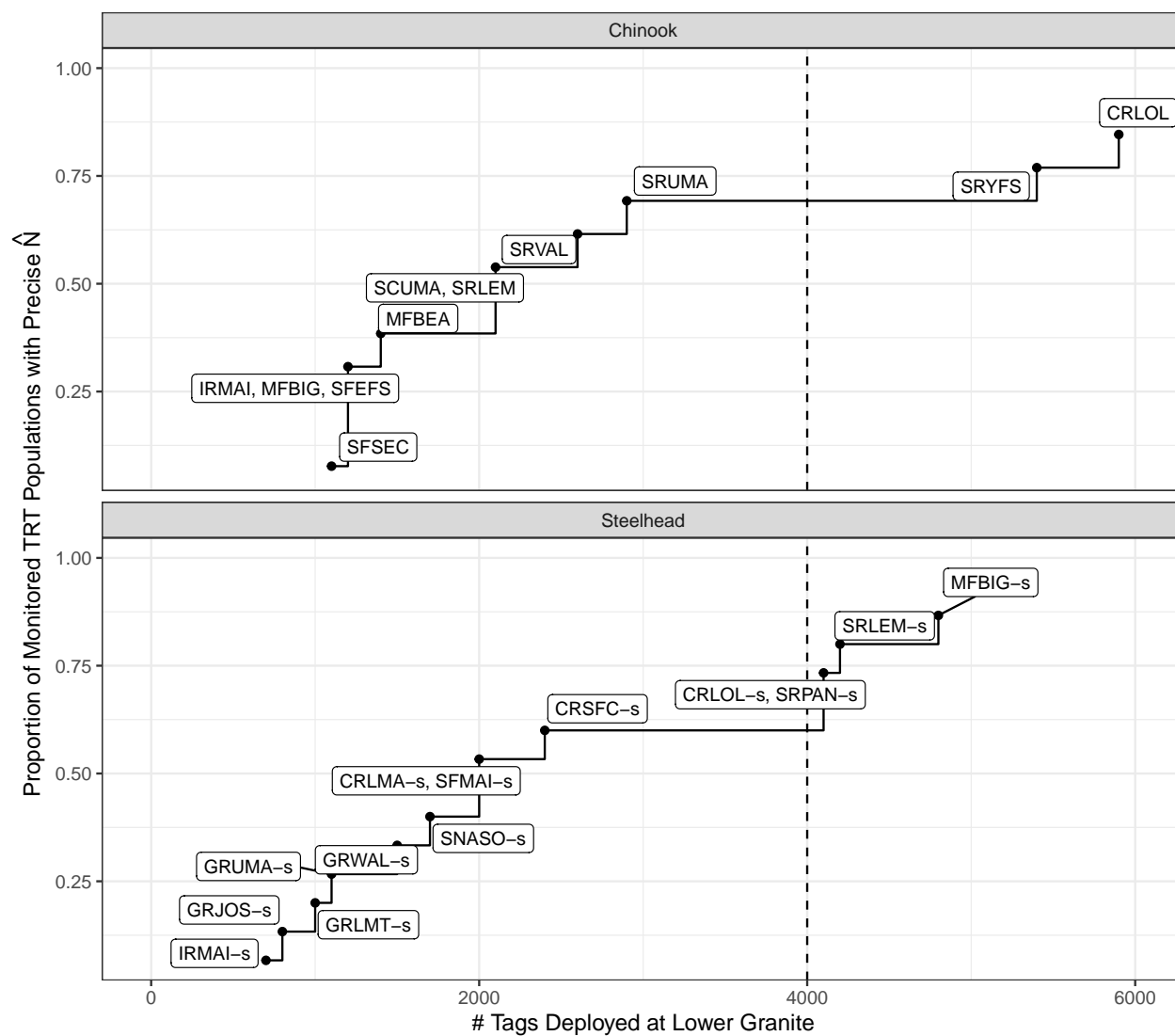


Figure 6: Same as figure above, but zoomed into the lower left corner of that plot.

in the precision of the abundance estimate. Figure 8 summarizes detection probabilities, by population and MPG/DPS, and could be used to provide guidance on populations/areas where IPTDS infrastructure could be best increased or improved.

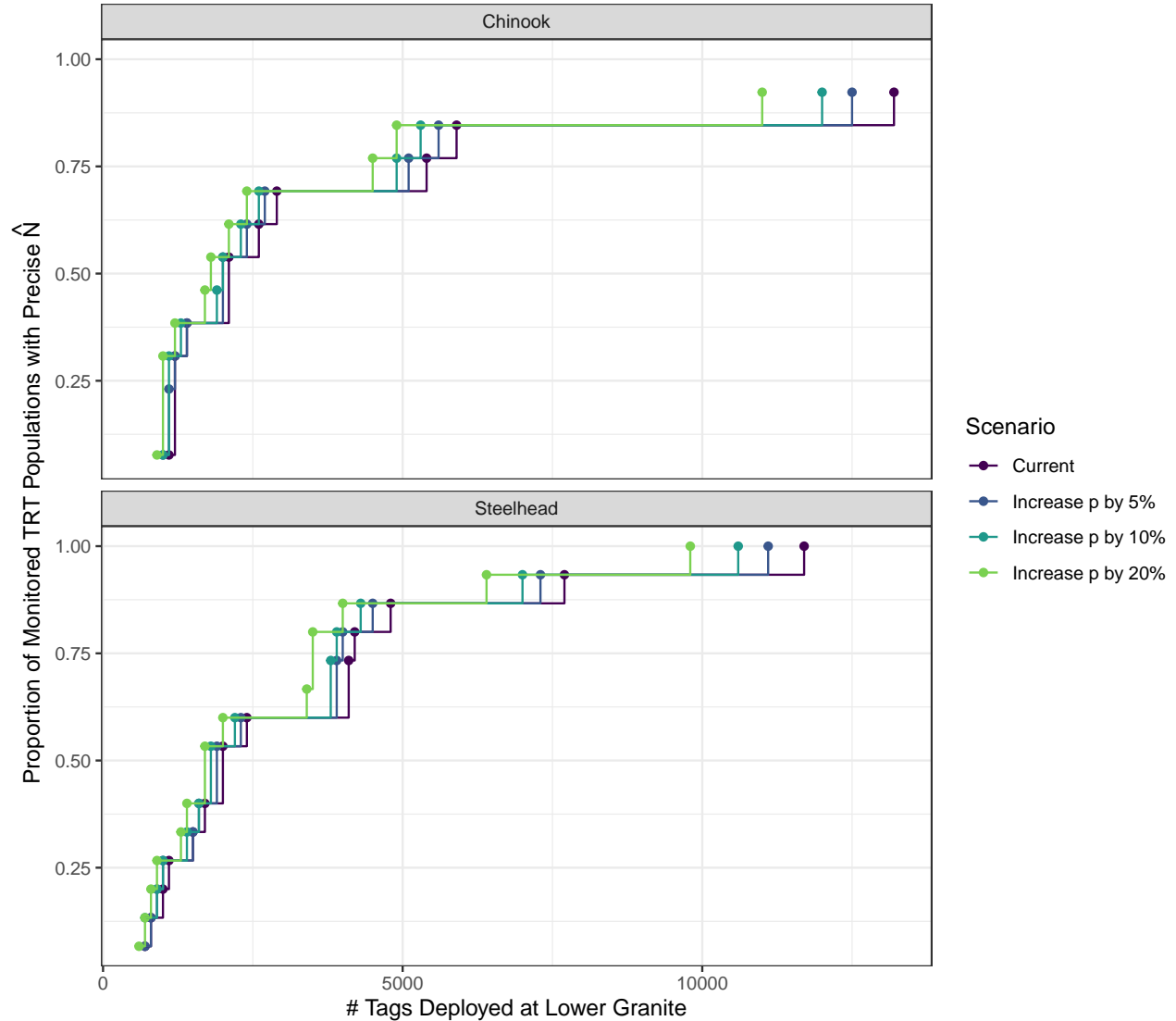


Figure 7: Expected proportion of TRT populations with a good CV of abundance estimates for a given number of PIT tags deployed at LGR, faceted by species, colored by detection probability scenario.

4 Conclusions

There are 28 extant Chinook salmon populations in the Snake River; of those, 13 contain IPTDS used to monitor natural origin population abundance. For steelhead, there are 23 of which 15 contain IPTDS to monitor spawner abundance. Our results suggest we need to detect a minimum of about 45 tags per TRT population to achieve acceptable precision in our abundance estimates at the TRT spatial scale. Figures 5 and 6 can be interpreted to mean that assuming any given TRT population contains IPTDS intended for population monitoring and 4,000 PIT tags are deployed at LGR, we'd expect that about 75% (Chinook) and 60% (steelhead) of TRT populations monitored with PIT tags would have abundance estimates with "good" precision (i.e. $CV \leq 0.15$).

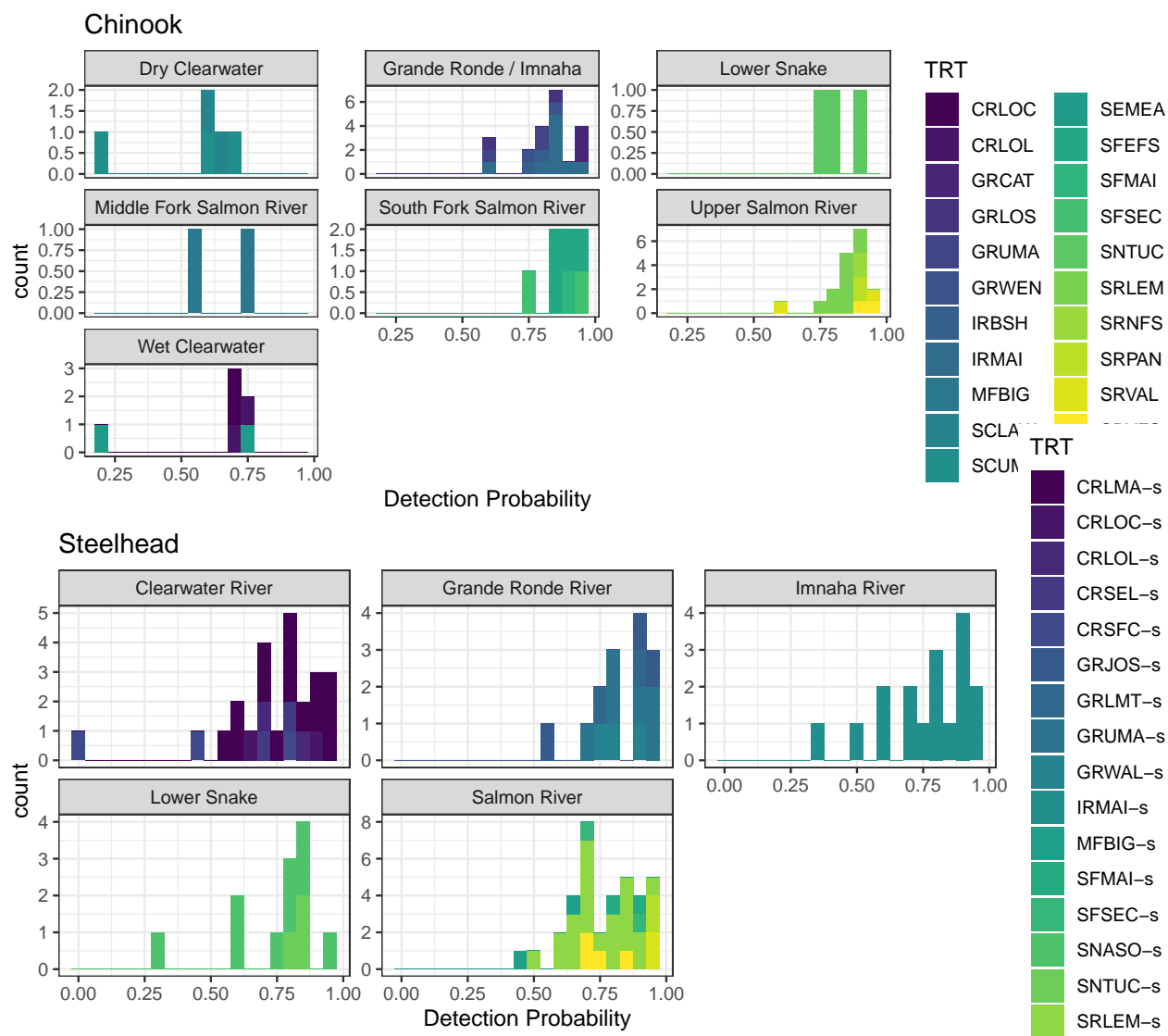


Figure 8: Histograms of detection probabilities at each array, estimated from DABOM, facted by MPG and colored by TRT population.