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**Bin Hu, Yanguo Teng, Yilun Zhang & Chen Zhu**

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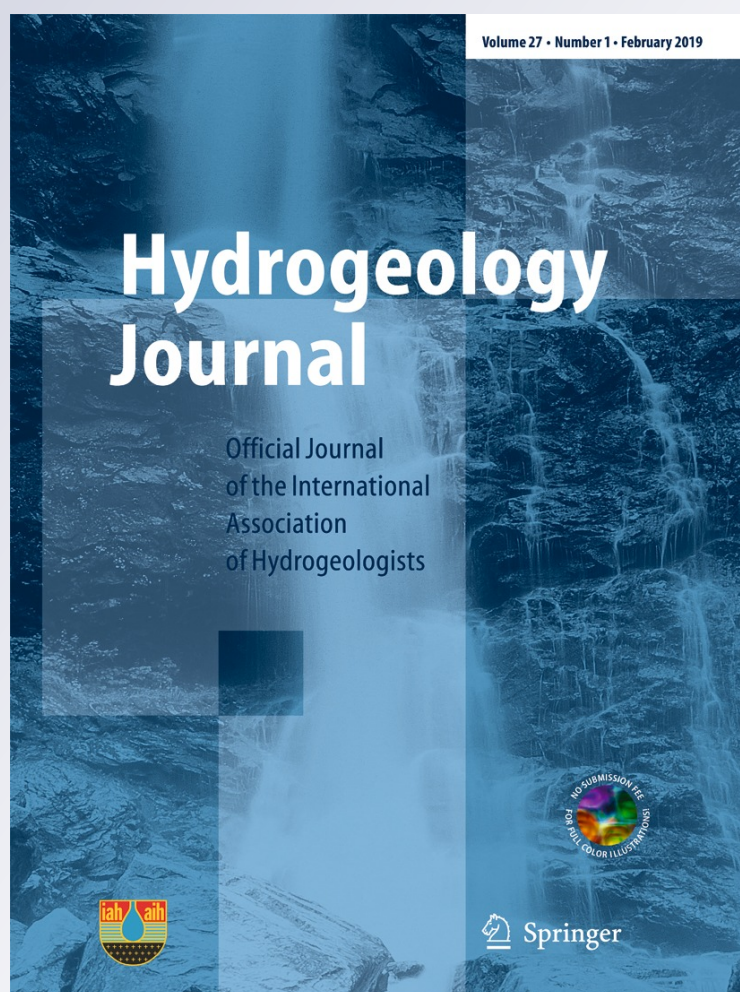
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# Review: The projected hydrologic cycle under the scenario of 936 ppm CO<sub>2</sub> in 2100

Bin Hu<sup>1,2</sup> · Yanguo Teng<sup>1</sup> · Yilun Zhang<sup>2</sup> · Chen Zhu<sup>2</sup>

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## Abstract

A host of environmental consequences will result from global warming, but arguably, the effect on water resources is one of the most consequential. This paper synthesizes results of published modeling studies that examined the groundwater system and hydrologic cycle under the Representative Concentration Pathway (RCP) 8.5 scenario (CO<sub>2</sub> at 936-ppm level at year 2100). Natural replenishment of groundwater occurs predominantly from infiltration of precipitation and surface impoundments. Therefore, the first stage of the study was to review a changed hydrologic cycle under RCP 8.5 in terms of influence from (1) precipitation, evapotranspiration, and soil moisture; (2) surface water and its interaction with groundwater; (3) extreme hydrologic events, and (4) teleconnection patterns. The general pattern of climate impact on groundwater resources follows the precipitation pattern with depletion in tropical and/or subtropical regions and with increase in the high-latitude regions of the Northern and Southern Hemispheres. However, regional variability also corresponds to the heterogenous impact of climate changes on regional distribution of precipitation and evapotranspiration, localized interaction between surface water and groundwater, and distances from the oceans with rising sea level. The decline of the water table in many areas may seriously reduce irrigated crop production and adversely impact groundwater-dependent ecosystems. Relatively fewer studies have been conducted on climate-change impacts on groundwater resources compared to surface-water systems, and the uncertainties on the recharge estimates are large. There appears to be an urgent need for including groundwater systems in climate impact assessment and in climate mitigation strategies.

**Keywords** Climate change · RCP 8.5 · Hydrologic cycle · Groundwater recharge/water budget · Water resources management

## Introduction

Global CO<sub>2</sub> concentration has exceeded 405 ppm since 2016, and the current global mean temperature is 0.83 °C above the pre-industrial level (1850–1900 period; NOAA 2017). In order to project future carbon-climate scenarios, Representative Concentration Pathways (RCPs)—a set of greenhouse gas (GHG) concentration and emissions pathways (Moss et al. 2010; van Vuuren et al. 2011)—are used to support climate change research on potential impacts and to guide policy responses to global warming (Moss et al. 2010; van Vuuren et

al. 2011). Specifically, the RCP 8.5 scenario—similar to Special Report on Emissions Scenarios (SRES) A1FI, a scenario with extreme global warming and CO<sub>2</sub> emission proposed in IPCC 2001—corresponds to the upper bound of the RCPs, known as the “baseline scenario”, which states that GHG emissions and concentrations considerably increase over time, leading to a radiative forcing of 8.5 W/m<sup>2</sup> by the end of the twenty-first century (Riahi et al. 2011). Under this scenario, CO<sub>2</sub> levels will rise to 936 ppm by 2100 (Wittmann and Pörtner 2013), resulting in a global temperature increase of 4–6 °C above the pre-industrial level (IPCC 2013; Parry et al. 2007; Van Vuuren et al. 2011; Wittmann and Pörtner 2013; Fig. 1).

Climate change is projected to change global hydrologic behavior (Hansen et al. 2016; Held and Soden 2006; Huntington 2006; Trenberth 2011). Increased precipitation, high-intensity evapotranspiration, and enhanced glacier melting will alter the global water resource availability and distribution (Betts et al. 2007; Collins et al. 2013; Radić and Hock

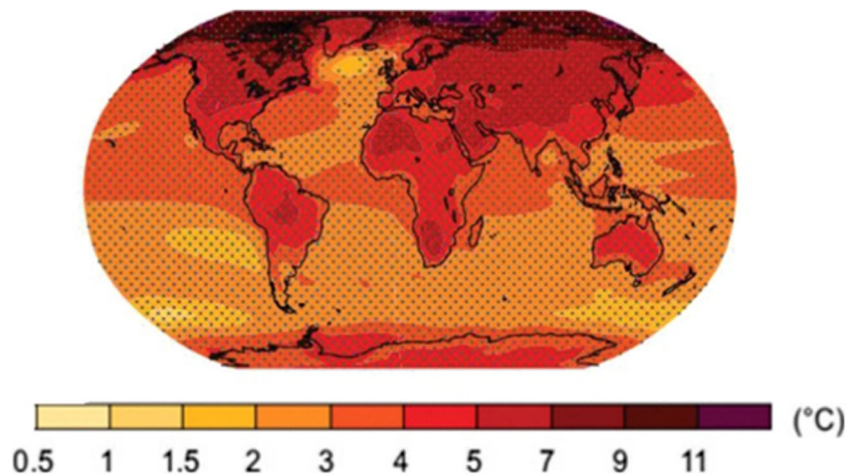
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**Fig. 1** Projected change in average annual temperature over the period 2081–2100 (compared to the period 1986–2005) under the scenario RCP 8.5 (IPCC 2013)



2011). Extreme hydrologic events such as flooding and drought, will increase in frequency in certain regions, threaten inhabited locations, and affect land use (Hirabayashi et al. 2013). Additionally, global temperature and GHG level increases will have a significant effect on seawater pH (Allen and Ingram 2002; Bala et al. 2008; Barnett et al. 2005; Collins et al. 2010; Fu et al. 2008; Hall-Spencer et al. 2008), increase the exceedance frequency of ecosystem tolerance thresholds, and potentially lead to chronic degradation of water quality (Murdoch et al. 2000). Obviously, impacts of climate change on the hydrological cycle cover a wide range of topics. Specifically, this review emphasizes the impacts of climate change on groundwater under RCP 8.5.

In this paper, projected hydrologic changes under the RCP 8.5 scenario are presented in seven sections: (1) precipitation, evapotranspiration, soil moisture, and glaciers; (2) surface water; (3) oceans; (4) groundwater; (5) soil erosion and sediment load; (6) extreme hydrologic events and (7) teleconnection patterns influence. The projected impacts, vulnerabilities, and risks of hydrologic changes under RCP 8.5 are also presented in order to demonstrate hydrologic cycle changes in the most extreme climate conditions by the end of twenty-first century. Based on these changes, impacts, vulnerabilities, and risks, proposed water resources management strategies, responding to such extreme climate conditions and hydrologic changes, are discussed.

## Projected hydrologic changes

### Precipitation, evapotranspiration, soil moisture, and glaciers

#### Precipitation

Overall, precipitation is expected to increase under high GHG emissions. An increase of 4.5 mm/month of global average

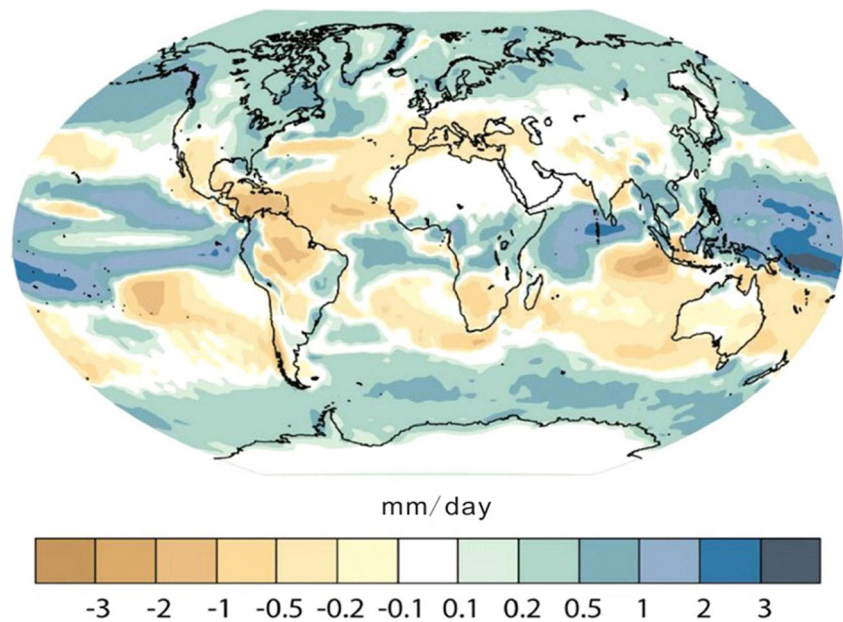
precipitation by the end of the twenty-first century is based on results of simulation of Coupled Model Intercomparison Project Phase 3 (CMIP3) from 15 general circulation models (GCMs; Collins et al. 2013). Precipitation associated with monsoons is projected to increase by 16%, affecting East Asia, South Asia, and Australia. These predictions used monthly mean data from 29 Coupled Model Intercomparison Project Phase 5 (CMIP5) models and daily precipitation data from 21 CMIP5 models (Christensen et al. 2013; Hsu et al. 2013; Kitoh et al. 2013). A GHG emission increase will potentially affect hurricane intensity, frequency, and storm size (Knutson and Tuleya 1999). This might already be occurring in an increasing number of intense storms in the Arabian Sea, including severe cyclones that have killed more than 3,500 people and caused over \$6.5 billion in damage since 1998—in 2011 US\$, Evan et al. (2011).

However, the change in precipitation is uneven across the globe due to internal variability in atmospheric circulation (Fig. 2). Compared to the pre-industrial global precipitation level, the global precipitation under RCP 8.5 has changed significantly. The dry areas have become even drier at mid-latitudes and subtropical dry regions of the Southern Hemisphere and the Northern Hemisphere, while the wet areas have become even wetter at high latitudes of the Northern and Southern Hemispheres as well as at some mid- and low latitude wet regions (Chou and Neelin 2004; Christensen et al. 2013; Giorgetta et al. 2013; Held and Soden 2006; Mitchell et al. 1987).

#### Evapotranspiration and soil moisture

Potential evapotranspiration—defined as the amount of water that would evaporate given unlimited available water (Freeze and Cherry 1979)—is expected to increase with global warming (Kingston et al. 2009). Specifically, global mean evapotranspiration is projected to increase approximately 5%/year due to CO<sub>2</sub> enrichment (Betts et al. 2007). The same

**Fig. 2** Projected change in precipitation by twenty-first century (2081–2100) under RCP 8.5 (Giorgetta et al. 2013)



conclusions were reached through six distinct methodologies, with pronounced increases projected especially in Southern Europe, Central America, Southern Africa, and Siberia (Seneviratne et al. 2010).

Increased evapotranspiration will likely have a serious impact on soil moisture and may induce soil drought conditions. Strong influence of regional soil-moisture dryness is also estimated by the Global Land-Atmosphere Climate Experiment–Coupled Model Intercomparison Project phase 5 (GLACE-CMIP5; Seneviratne et al. 2013). Based on the simulation results of GLACE-CMIP5 for example, in the Mediterranean region soil moisture will decrease about 25% under RCP 8.5 scenario by 2100 (Seneviratne et al. 2013). Droughts that last 4–6 months will double in both extent and frequency under RCP 8.5, and droughts longer than 12 months will be three times more common between the mid-twentieth century and the end of the twenty-first century (Sheffield and Wood 2008).

### Glaciers

Glaciers will be subjected to higher melting rates due to increased temperatures. Global average volume losses range from 12 to 30%, according to ten GCM simulations (Radić and Hock 2011). In Greenland, a projected glaciers' volume reduction of 41–63.8% from 2000 to 2100 has been established by a regional climate model (RCM) and glacier model (Nick et al. 2013). Over roughly the same period (2006–2100) a 68% reduction in the Himalayas is projected by CMIP5 projections from 14 GCMs (Radić and Hock 2014). The approximately 40,800 km<sup>2</sup> of glaciers in the Karakoram and Himalaya (Himalaya: 22,800 km<sup>2</sup>; Karakoram: 18,000 km<sup>2</sup>) contribute irrigation water used to

grow food for 800 million people in Asia, a crucial resource especially in the dry months (Immerzeel et al. 2010; Bolch et al. 2012). Under RCP 8.5, these glaciers would lose about 35 Gt/year from 2070 to 2099, compared to a glacial mass balance in 2000 based on simulation results of 21 CMIP5 models, which is equivalent to 38.8 km<sup>3</sup>/year water resources loss. Moreover, about 27% of the glaciers may face “eventual disappearance” by 2100, straining the water resources available to 800 million people (Chaturvedi et al. 2014).

### Surface water

The current hydrologic cycle consists of continental precipitation of 110,000 km<sup>3</sup>/year, evapotranspiration of 65,000 km<sup>3</sup>/year, and runoff of 40,000–45,000 km<sup>3</sup>/year worldwide (FAO 2005; Oki and Kanae 2006; Oelkers et al. 2011). Under RCP 8.5 at 2100, continental precipitation is projected to increase to 118,158 km<sup>3</sup>/year, global mean evapotranspiration to 68,208 km<sup>3</sup>/year, and runoff to 49,950 km<sup>3</sup>/year (Betts et al. 2007).

However, these changes are not projected to be uniform. A decrease in average annual runoff is predicted for low latitude, inland areas, while an increase is predicted at high latitudes (Arnell and Gosling 2013). By the late twenty-first century (2071–2100), mean streamflow increases are projected in 68.9% of streams worldwide that were included in the model simulation (excluding Greenland and Antarctica) but projected decreases in 31.1% of streams. Increased streamflows were projected in Asia, Africa, Australia, parts of eastern South America, and high latitudes of North America and Eurasia (Koirala et al. 2014). On the other hand, decreased stream-flows were primarily in Europe, the Middle East, Central Asia, northern and southern Africa, the southwestern

United States, and most of Central and South America (Koirala et al. 2014). These results were from simulations using 11 recent atmospheric–oceanic general circulation models (AOGCMs), which, despite differences in details and by independent modeling institutes participating in CMIP5 (Taylor et al. 2012), projected similar trends (Koirala et al. 2014). The magnitude of change varies regionally. In 5.5% of the world, the increase in 2071–2100 is 50% over the 1971–2000 values; 0.2% displays a decrease of the same magnitude (Koirala et al. 2014).

## Groundwater

Groundwater accounts for 98% of accessible freshwater. It is the vital component of the hydrological cycle in the critical zone between the surface and subsurface (Green et al. 2011). There is a marked increase in research that focuses on the direct and indirect effects of climate change and variable projections on groundwater, especially under high GHG emission and extreme global warming scenarios (Holman et al. 2012; Taylor et al. 2013a, b; Meixner et al. 2016). In this article, groundwater resource variability under global warming, as well as feedback between climate and groundwater, are overviewed.

### Groundwater quantity variation

As climate impacts on surface-water quantity and quality are geographically uneven under RCP 8.5, impacts on groundwater resource variability also vary from region to region. Increased precipitation contributes to increased groundwater recharge and replenishes aquifer storage; however, increased evapotranspiration or decreased precipitation can also lead to groundwater depletion (Taylor et al. 2013b). For example, mean annual water tables in Jylland and Sjaelland, Denmark, are projected to increase 0.2 and 0.5 m, respectively, by 2100, assuming a temperature increase of 3.2 °C compared to 1990–2004 (van Roosmalen et al. 2007). The renewable groundwater resources, defined as those which can be recharged or replenished to overcome resource depletion caused by natural processes and anthropogenic activities (Postel et al. 1996), are expected to decrease by 38% under RCP 8.5, especially in arid and semi-arid regions (Portmann et al. 2013). From the regional climate model (RCM) simulations, the total decline of the groundwater level is projected to a range of 6.6–83.5 m in four major South Korean river basins (Han River, Nakdong River, Geum River, and Yeongsan-Sumjin River basins) at 2100 under RCP 8.5 (Jang et al. 2015). Thus, the regional characteristics of precipitation and evapotranspiration directly impact groundwater resource in specific regions. Generally, the groundwater resource variability corresponds to precipitation distribution patterns: (1) groundwater resource depletion increases in tropical

and/or subtropical regions; and (2) groundwater resource quantity increases in the high-latitude regions of the Northern and Southern Hemispheres.

Increased evapotranspiration may lead to a dramatic decline of streams, lakes, and other surface-water resources. The traditional source of urban, industrial, and agricultural water will shift from surface water to groundwater, thus increasing groundwater pumping and accelerating its depletion (e.g. Southwestern America, Southern Europe, India; Falloon and Betts 2010; Aeschbach-Hertig and Gleeson 2012; Taylor et al. 2013b).

### Impacts on groundwater quality

In particular, intensified global hydrologic cycle (e.g., increase or decrease in frequency and intensity of precipitation event, increased evapotranspiration intensity, sea-level rise) will directly influence groundwater quality via pollutant infiltration, interaction between groundwater and other water systems, and socio-economic development for meeting increasing population requirements (Cotterman et al. 2018; Dunn et al. 2012; Ferguson and Gleeson 2012; Van Vuuren and Carter 2014). This paper specifically discusses the groundwater quality variation in two categories of regions: (1) agricultural areas and (2) coastal areas.

**Groundwater quality in agricultural areas** The increased demand of future agricultural development will aggravate groundwater depletion and degradation as a result of high-intensity fertilization and land-use changes (Green et al. 2011). Agricultural intensification usually corresponds to the heavy use of fertilization, especially nitrogen (N; Gordon et al. 2008; Johansson et al. 2009), which will accelerate the cycling of N, contributing globally to observed increases in N-loads in surface runoff and subsurface-water systems (Seitzinger et al. 2010). For example, Basso et al. (2016) evaluated impacts of different N fertilizer treatments on  $\text{NO}_3^-$  leaching with the SALUS model in Italy, and the simulation results indicated that urea fertilization treatment might lead to the highest biomass production and  $\text{NO}_3^-$  leaching under RCP 8.5 in 2070. Savard (2016) simulated groundwater flow impact of climate change and agricultural adaptation under RCP 4.5 and 8.5 by year 2050 on Prince Edward Island (PEI, Canada), predicting that it would lead to a 25–32% increase in  $\text{NO}_3\text{-N}$  concentration over the Island aquifer system. Arheimer et al. (2005) modelled N leaching in the Rönneå River in southern Sweden using different climate change scenarios (SRES A2 and B2, scenarios with less GHG emission than in RCP 8.5) and found an average increase in N concentrations of 50% in the root zone on arable land. Similarly, Kaste et al. (2006) used modelling of N-fluxes in a Norwegian catchment to reveal significant growth in nitrate-N fluxes (40–50%) when regional temperature increased approximately 3 °C. Based on the



projection of hydrological N and phosphorus (P) in coastal subsurface-water system of agricultural areas in Sweden, the N and P loads from the subsurface-water system to the sea were projected to increase from 12.9 and 50% in long-term climate changes (2005–2100), respectively, due to the increased non-point source pollution from cropland (Destouni and Darracq 2009).

Furthermore, higher nitrate loads in agricultural regions may lead to higher nitrous oxide ( $\text{N}_2\text{O}$ ) emission because of increased denitrification in the groundwater system (Groffman et al. 2000). Increased subsurface temperature and  $\text{NO}_3^-$  concentration in groundwater can result in motivating denitrification; meanwhile, it can also contribute to the feedback of GHG emission (Jeppesen et al. 2011; Stocker et al. 2013; van Vuuren et al. 2011). Besides, global warming can also influence the danger and transformation of pesticides and therefore affect groundwater quality (Bloomfield et al. 2006; Howarth et al. 2006). In addition, higher temperatures can result in increased toxicity of pesticides (Noyes et al. 2009; Noyes and Lema 2015). Compared to surface water, groundwater receptors are usually minor pesticide receptors due to a combination of major pesticide loss in the soil by sorption, degradation and volatilization, and dilution and dispersion in the saturated zone (Bloomfield et al. 2006). However, future agricultural development will not only increase the demand for fertilization but also enhance the application of pesticides. Moreover, high intense precipitation and flooding will induce pesticide percolation and leaching from adsorbed colloids and sediments in soil into surface-water and groundwater systems. For instance, in the Odense Fjord in Denmark, the integrity of water ecosystems, no matter surface water or groundwater, has been damaged due to urbanization and use of fertilizers and pesticides derived from the industrialization of the agricultural sector, extensive channelization of rivers, summer droughts and groundwater abstraction (Molina-Navarro et al. 2018). Characteristic features of pesticide concentrations in groundwater system are temporal-spatial disparities (Goody et al. 2005). Overall, the impact of climate change on pesticide exposure in groundwater systems is important in future warming. However, projected methodology, field investigation, and data acquisition for pesticide contamination in groundwater systems for various climate change scenarios needs to be further developed.

**Groundwater quality in coastal areas** Coastal aquifers in the interface between the oceanic and terrestrial hydrologic systems supply water resources for more than one billion people living in coastal areas (Small and Nicholls 2003). Likewise, water quality in coastal aquifers is likely to be affected by seawater intrusion or saltwater inundation (Taylor et al. 2013b). Sea level rise will increase the water-table difference between seawater and surface water or groundwater in coastal zones, causing seawater intrusion in coastal areas (Ferguson

and Gleeson 2012). World population along the coasts was estimated by using the digital World Vector Shoreline, and results indicated that approximately 37% (2.07 billion) of the 1994 population (5.62 billion) lived within 100 km of a coastline, and approximately 44% (2.45 billion) lived within 150 km of a coastline (Small and Nicholls 2003). Assuming a similar current population distribution, for 2017 these figures would be 2.8 and 3.3 billion (total 7.69 billion), respectively. Seawater intrusion caused by sea-level rise will inevitably impact fresh groundwater in coastal aquifers (Werner and Simmons 2009). Vandenbohede et al. (2008) simulated a probable 15% increase in recharge of a Belgian coastal aquifer at 2100. Due to a 0.4 m sea-level rise, the interaction between sea water in low-lying inland areas and groundwater in adjacent aquifers will become more frequent, which will result in brackish and saltwater intrusion, thus increasing salinity of groundwater and causing water degradation. Under RCP 8.5, sea-level projections indicate a possible rise of 1.4 m in the Caribbean, which will result in inevitable impact on coastal aquifers and polluted freshwater in aquifers by saline intrusion, even in the absence of human activity such as pumping (Cashman and Nagdee 2017). Along the southwest coast of the Netherlands, due to rising sea levels, salt loads will double by the year 2100. In the deeper inland, increasing salt loads will lead to saline groundwater, and salt loads will also result in shallow groundwater salinization (Oude Essink et al. 2010).

In summary, sea-level rise driven by climate change and intense groundwater withdrawal driven by anthropogenic activity will exacerbate seawater intrusion in coastland regions. Shallow aquifer salinization will increase groundwater salinity and lead to water quality degradation. Moreover, it will bring more challenges to aquifer management in the coastal areas in the future.

### Impacts on groundwater recharge

Due to regional variability of precipitation and evapotranspiration, groundwater recharge also varies from region to region. Compared to groundwater recharge levels during 1971–2000, groundwater recharge in the 2080s is projected to range from –70% to +100%. According to results of five CMIP5 GCMs provided by the ISI-MIP effort to the global hydrological model WaterGAP, more than a 10% decrease of groundwater recharge is projected for large parts of eastern China, southern Australia, the Mediterranean region, the Middle East, Central and South America, and southwestern South Africa. In contrast, a significant increase is projected in the high latitudes of the Northern Hemisphere due to increased precipitation, sea level, or other groundwater recharge sources (Portmann et al. 2013). Moreover, results of GCM hydrological simulations show that groundwater recharge in the San Pedro Basin (Arizona/Sonora, USA) in 2100 would decrease from 17 to 30% with a 3 °C temperature increase compared to



2000 (Serrat-Capdevila et al. 2007). Groundwater recharge in the Dill catchment, which is situated in the southeast of the Rhenish Massif in Germany, is projected to have a significant summertime reduction (more than 50%) by the 2080s. Simulation results of the Soil and Water Assessment Tool (SWAT) indicate that enhanced potential evapotranspiration and decreased precipitation are the main contributors to groundwater recharge decrease under the 4.65 °C annual mean temperature increase scenario (Eckhardt and Ulbrich 2003).

### Surface-water/groundwater interaction

Surface water-groundwater interaction will be further complicated when the individual and combined impacts of climate change and human activities are considered. The projected changes in Denmark are related to increased groundwater–river interaction brought on by an increase of up to 30% in annual net precipitation under SRES A2 for the 2080s (van Roosmalen et al. 2007). According to a simulation with GSFLOW in the Luanhe River Basin (China), Feng et al. (2018) found that when temperature increased by 4 °C (RCP 8.5), precipitation would simultaneously increase 15%, river outflow would increase approximately 2%, and groundwater would decline approximately 4%; therefore, groundwater resources were more sensitive to climate change than surface water in the study area. Through coupling variable infiltration capacity (VIC) and MODFLOW models (VIC-MF), surface-water levels will increase 0.12–0.96 m in most study areas, and water tables will increase 2.6 m under RCP 8.5 due to intense rainfall and frequent surface-water/groundwater interaction by the year 2042. (Sridhar et al. 2017).

However, studies that quantify integrated surface-water/groundwater systems in global warming are rarer. As already mentioned, most studies focused separately on global warming impacts on either surface water or groundwater, but a few studies attempted to quantify the interaction process in surface-water/groundwater systems. Although there is some research that explores an integrated hydrologic cycle in a specific region in the current situation (2000–2020) by developing integrated hydrologic model (Essaid and Caldwell 2017; Goderniaux et al. 2009; Huntington and Niswonger 2012; Kløve et al. 2014; Maxwell et al. 2014), there are few studies attempting to project surface-water/groundwater interaction under various future climate change impacts (2080–2100; Li et al. 2015; Pulido-Velazquez et al. 2015). For extreme future climate change scenarios, the relevant research studies are even more rare due to the imperfection of integrated hydrological model development, the discrepancy of temporal-spatial scale in climate model, hydrological model and groundwater model, and the difficulty in accurately generalizing and describing the complicated hyporheic zone. Thus,

extreme climate change impact on surface water-groundwater interaction needs further study.

### Oceans

Although there is a consensus that the global mean sea level will continue increasing under global warming, there is no general agreement on the exact extent. An adopted value of global mean sea level rise for RCP 8.5 by IPCC is 0.63 m in 2100 compared to 1986–2005 (with a range from 0.45 to 0.82 m; Church et al. 2013).

Meanwhile, water chemistry will also change with increased temperature and GHG emissions; however, due to a high level of uncertainty, few projections of the impacts of climate change on water quality are available (Jiménez-Cisneros et al. 2014). Water quality projections depend strongly on (1) local conditions; (2) climatic and environmental assumptions; and (3) the current or reference pollution state (Bonte and Zwolsman 2010; Sahoo et al. 2011; Whitehead et al. 2009a, 2009b). For instance, surface ocean pH is projected to decrease by 0.3–0.5 (Ciais et al. 2014) and dissolved oxygen (DO) is projected to decline significantly (Cocco et al. 2013). Projections of DO decrease range by 6–12  $\mu\text{mol/kg}$  by 2100 (the current level is approximately 175  $\mu\text{mol/kg}$ ; Bopp et al. 2002; Matear et al. 2000; Matear and Hirst 2003; Oschlies et al. 2008; Plattner et al. 2002; Sarmiento et al. 1998; Schmittner et al. 2008; Shaffer et al. 2009). Increased temperatures will also result in expanded ranges of harmful warm-water algal species and algal bloom seasons (Hallegraeff 2010). Blooms of *A. catenella* increase when water temperatures are greater than 13 °C (Nishitani and Chew 1984); the algae already cause shellfish toxicity in Puget Sound (Washington, USA), primarily in the late summer and early fall when water temperatures reach their maxima (Moore et al. 2009). Harmful algal blooms of toxin-producing *Microcystis aeruginosa* were found from 2010 to 2011 in Monterey Bay (California, USA), a time period that was associated with warm water temperatures (Gibble and Kudela 2014). Cyanobacteria blooms usually occur in water temperatures higher than 16–17 °C and extensive accumulate in the Baltic Sea, reaching an area of  $\sim 200,000 \text{ km}^2$  (Kahru and Elmgren 2014). Furthermore, global warming and increasing  $\text{CO}_2$  concentration may benefit the growth of *Synechococcus* (Fu et al. 2007), a cosmopolitan cyanobacterium, and it also leads to harmful blooms in various ecosystems including Florida Bay, USA (Boyer et al. 2006).

### Soil Erosion and sediment load

Predicted changes in precipitation volume and intensity, runoff conditions, and soil moisture will affect soil erosion and sediment load. On the global scale, the mean value of global soil erosion is projected to increase about 14% by the 2090s

compared to the 1980s under RCP 8.5. Of the projected increase in soil erosion, 64% is attributed to climate change and 36% to land use change. As much as 40 to 50% increase of global mean soil erosion is projected in Australia and Africa (Yang et al. 2003). Assuming a typical soil bulk density of 1.3 Mg/m, the estimated erosion corresponds to removal of a surface layer of 0.7–1.0 mm annually (Ito 2007).

By the end of the twenty-first century, the impact of climate change on soil erosion is expected to have twice the impact of land use change (Yang et al. 2003). Increased precipitation in many regions will enhance rainfall erosivity, resulting in intensification of erosion phenomena and stream sediment loads (Jiménez-Cisneros et al. 2014). For instance, a 10% increase in winter rainfall accompanying a projected increase of 4.5 °C would increase annual erosion of arable land by up to 180% in the UK (Favis-Mortlock and Boardman 1995). While erosion rates are likely to rise in locations that are projected to receive more rainfall, rates may abate where precipitation decreases—for example, more than 14% of the projected decline in rainfall in Australia is during erosion-sensitive months for 2070–2099, thus reducing soil erosion over 24% according to results of a revised Morgan Morgan Finney (MMF) model (Scholz et al. 2008). On the other hand, in the tropics, the intensity of cyclones is projected to increase by 2–11% by 2100, which might increase soil erosion and landslides (Knutson et al. 2010).

## Extreme hydrologic events

### Floods

Overall, annual global flooding is projected to dramatically increase under RCP 8.5 in many densely inhabited locations. With Shared Socio-economic Pathways 2 (SSP2, “middle-of-the-road”) and RCP 8.5 by 2050, the climate change will increase water resource stress and river flood risk for approximately 920–3,400 million people and 100 and 580 million people, respectively (Arnell and Lloyd-Hughes 2014). The number of people exposed to 100-year flood events in 2100 is projected to be more than twice the population affected in 2040 (Fig. 3).

High intensity and frequent precipitation is the dominant factor for flood occurrence. Under conditions of global warming, precipitation is more likely to form heavy rainfalls accompanied by an increase of flood risk in different regions (Goswami et al. 2006; Min et al. 2011)—for example, in RCP 8.5, global flood exposure will increase by about  $14 \pm 10$  times from the time-period 1971–2000 to 2071–2100 in the areas of increasing annual precipitation and heavy precipitation, including South Asia, Southeast Asia, Northeast Eurasia, eastern and low-latitude Africa (Hirabayashi et al. 2013).

Likewise, sea-level rise can also impact flood risk, especially in coastal areas. Flood events’ probabilities can be estimated via probabilistic projections of future local sea-level rise (Hinkel et al. 2014; Nicholls and Cazenave 2010). For

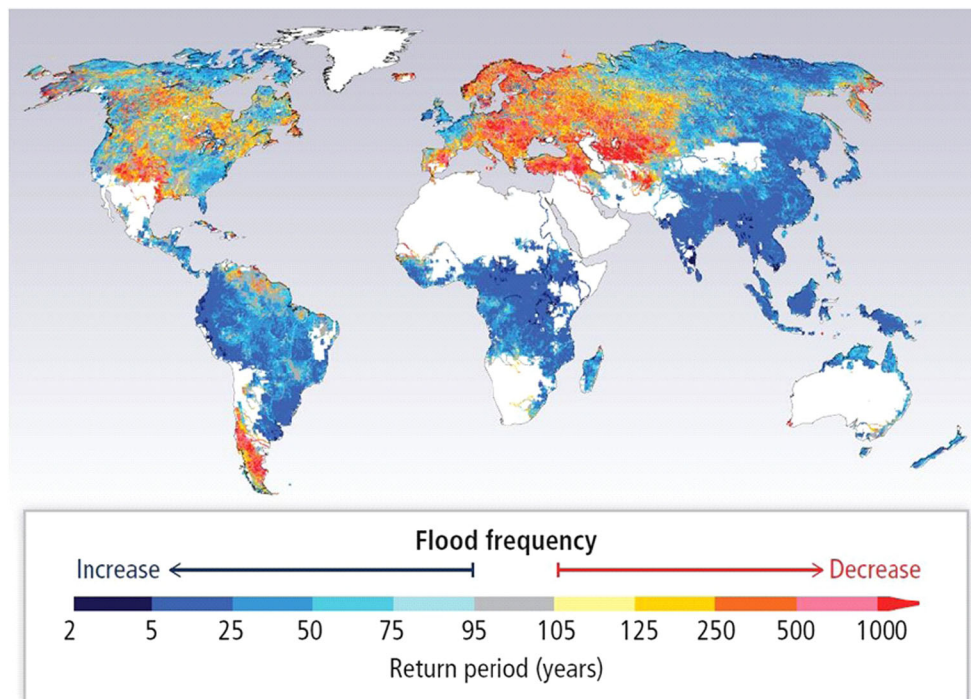
example, when considering local sea-level rise impact, it will likely result in more extreme flood events, i.e., New York is projected nine 100-year-floods over 2001–2100 under RCP 8.5, the same as 11-year events without sea-level rise (Kopp et al. 2014).

### Droughts

A comparison of the proportion of land exhibiting significant projected changes in hydrological drought frequency (reduction in the availability of surface and subsurface water) to the proportion exhibiting significant changes in meteorological drought (reduction in precipitation relative to the mean for a particular location) and agricultural drought frequency (results from an insufficient supply of water for plant growth and includes soil moisture deficit) was made under SRES A1fi. Excluding cold areas, 18–30% of the land surface was projected to experience a significant increase in the frequency of 3-month droughts, while about 15–45% will see a decrease from the baseline period 1961–1990 to the 2080s (Taylor et al. 2013a).

Duration of drought, as the length of a drought event in months (Touma et al. 2015), is one of the predominant characteristics of a drought (Giorgi and Bi 2005). It has a significant variability in spatial scales due to the heterogeneous distribution of precipitation and/or evapotranspiration in climate change (Samaniego et al. 2017). For example, according to 15 global climate models available in the CMIP5, there are longer drought duration events in low-latitude regions (subtropical and tropical regions); in contrast, higher latitudes regions (e.g. South Africa in the southern hemisphere, and Northeastern Europe in the northern hemisphere) tend to show little change under RCP 8.5 for the period 2050–2099 (Touma et al. 2015). Meanwhile, longer droughts may result in urban, industrial, and agricultural water shifting from surface water to predominantly groundwater, thus accelerating the depletion of groundwater (e.g. Southwestern America, Taylor et al. 2013a). Furthermore, increased groundwater depletion may also trigger secondary disaster such as land subsidence and seawater intrusion (Taylor et al. 2013a).

Several drought indicators have been applied to evaluate meteorological drought severity, such as the Palmer drought severity index (PDSI, Palmer 1965), the Standardized Precipitation Index (SPI, McKee et al. 1993), the Streamflow Drought Index (SDI, Nalbantis and Tsakiris 2009), the Standardized Precipitation Evaporation Index (SPEI, Vicente-Serrano et al. 2010), the Standardized Groundwater Level Index (SGI, Bloomfield and Marchant 2013), the Groundwater Resource Index (GRI, Mendicino et al. 2008), and the Groundwater Drought Index (GDI, Thomas et al. 2015). Among these indices, PDSI is the most widely available for drought evaluation under different climate change scenarios (Dai 2011, 2013; Trenberth et al. 2014; Zhao and Running 2010). It is a locally normalized index of



**Fig. 3** Multi-model median return period (years) in the 2080s for the twentieth century 100-year flood (Hirabayashi et al. 2013), based on one hydrological model driven by 11 Coupled Model Inter-comparison Project Phase 5 (CMIP5) general circulation models (GCMs) under RCP 8.5. At each location, the magnitude of the 100-year flood was estimated by fitting a Gumbel distribution function to time series of simulated

annual maximum daily discharge in 1971–2000, and the return period of that flood in 2071–2100 was estimated by fitting the same distribution to discharges simulated for that period. Regions with mean runoff less than 0.01 mm/day, Antarctica, Greenland, and small islands are excluded from the analysis and indicated in white (Jiménez-Cisneros et al. 2014)

soil moisture availability calculated from the balance of moisture supply (precipitation) and demand (evapotranspiration; Cook et al. 2015). Relative to a defined baseline, negative PDSI values ( $PDSI < 0$ ) represent drier than average conditions (droughts), and positive PDSI values ( $PDSI > 0$ ) represent wetter conditions (pluvials; Cook et al. 2015). Moreover, the severity of groundwater drought can also be estimated by PDSI (Dai 2011; von Freyberg et al. 2015)—for example, von Freyberg et al. (2015) applied PDSI into drought characteristics analysis in monthly time-scale of precipitation and groundwater recharge in northeastern Switzerland.

In addition to PDSI, the SGI builds on SPI and can efficiently reveal and quantify characteristics (intensity, severity etc.) of groundwater drought under precipitation variation in the future due to the high correlation between SPI and SGI indices (Table 1; Bloomfield and Marchant 2013). The SPI has been successfully applied into groundwater drought assessment in the eastern Netherlands, southern Germany, southern Sweden, the UK, and Iran (Bloomfield et al. 2015; Jönsson 2017; Motlagh et al. 2017; Van Loon et al. 2017).

Moreover, based on the literature, the trend of drought under global warming tendency in the twenty-first century demonstrates a significant variability over spatial and temporal scales due to reduced precipitation and/or increased evapotranspiration (Cook et al. 2015; Park et al. 2015; Seneviratne

et al. 2012). A decreasing trend of drought has been detected in East Asia (Dai 2013) and in most parts of China (Chen et al. 2013). In contrast, an increasing trend of the frequency and intensity of drought was identified in the United Kingdom and France (Vidal et al. 2012; Vidal and Wade 2009), East Asia (Manabe et al. 2004; Wang 2005), Southwestern United States (Cook et al. 2015), Europe (Beniston et al. 2007) and globally (Dai 2013).

### Teleconnection patterns influence

Teleconnection patterns such as the Atlantic Multidecadal Oscillation (AMO), the Pacific Decadal Oscillation (PDO), and the El Niño-Southern Oscillation (ENSO), are closely

**Table 1** The interval value of SPI and SGI (Bloomfield and Marchant 2013)

SPI values	SGI values	Drought degree
$SPI < -2$	$SGI < 0$	Extreme drought
$-2 < SPI < -1.5$	$SGI < 0$	Severe drought
$-1.5 < SPI < -1$	$SGI < 0$	Moderate drought
$-1 < SPI < 0$	$SGI < 0$	Minor drought
$0 < SPI$	$SGI > 0$	No drought



associated with changes in the amount and temporal-spatial distribution of precipitation, air temperature, and evapotranspiration (De Vita et al. 2012; Ropelewski and Halpert 1986; Shang et al. 2011). These teleconnection patterns therefore are also essential to recharge rates and changes in groundwater storage (Fleming and Quilty 2006; Gurdak et al. 2007; Kuss and Gurdak 2014). Holman et al. (2011) applied wavelet analysis to indicate that large-scale climate patterns such as the North Atlantic Oscillation (NAO), may affect groundwater recharge rates and mechanisms in aquifers across the UK. Kuss and Gurdak (2014) quantified the effects of ENSO, PDO, AMO, and NAO on precipitation and groundwater levels of North Atlantic Coastal Plain principal aquifers (PAs) in the US by singular spectrum analysis (SSA), wavelet coherence analysis, and lag correlation. The results demonstrated that ENSO and PDO had greater control than NAO and AMO on variability in groundwater levels across the US, particularly in the western and central PAs. Velasco et al. (2017) used SSA and found that PDO, ENSO, and Pacific/North American Oscillation (PNA) significantly impacted on precipitation and groundwater fluctuations across a north–south gradient of the West Coast in the US, and low frequency signals (PDO) seemed to be better preserved in groundwater fluctuations than high frequency signals (ENSO and PNA).

With increasing GHG emission in the future, the characteristics (duration, frequency, amplitude) of teleconnection patterns will be affected, resulting in the variability of precipitation, temperature, and evapotranspiration (Sabziparvar et al. 2011; Tremblay et al. 2011). Therefore, groundwater storage, groundwater recharge, and groundwater level fluctuation will have more complexity by 2100 (Kuss and Gurdak 2014).

## Uncertainty analysis of the groundwater recharge

Synthesizing hydrologic changes with multiple GCMs allows uncertainty analysis (Crosbie et al. 2013; Deser et al. 2012; Kurylyk and MacQuarrie 2013; Meixner et al. 2016; Rogelj et al. 2012; Smerdon 2017; Taylor et al. 2013a, b). Three sources drive the uncertainty of the hydrologic cycle under climate change: (1) type of GCM; (2) type of catchment scale; and (3) GHG emission scenario (Friedlingstein et al. 2014; Hawkins and Sutton 2009; Tebaldi and Knutti 2007). Groundwater recharge is an essential driver of many hydrologic processes as well as an important variable in the water cycle (Bakker et al. 2013). Arguably, the predicted impacts of climate change on groundwater system are quite uncertain, especially for groundwater recharge, due to the two dominant sources of uncertainty in climate change predictions: GCM types and downscaling (Green et al. 2011; Taylor et al. 2013b; Smerdon 2017).

The process of predicting recharge under climate change in the future by various GCMs involves high uncertainty (Holman et al. 2012; Jyrkama and Sykes 2007; Meixner et al. 2016). Crosbie et al. (2010) fitted recharge projections derived from 15 GCMs with low, medium, and high global warming scenarios to a weighted Pearson Type III distribution, and there was an 80% of the variability range between the wet (10%) and dry (90%) future recharge at 2030. Jackson et al. (2011) simulated groundwater recharge and flow variation at an ensemble of catchment scale by applying 13 GCMs under A2 emission scenario (second highest emission scenario in IPCC 2001), and there was −26% to +31% range in potential groundwater recharge. When only applying a single GCM and emission scenario (A2) and adopting an average decrease recharge of 5% by 2100, Hanson et al. (2012) found that the variability range was still large: −60 to +20%. The potential impact of climate change on groundwater recharge across the western US was analyzed by synthesizing existing studies, and the uncertainty range in these areas demonstrated regional discrepancy from southern (−100 to +30%) to northern regions (little change; Meixner et al. 2016).

Direct application of GCMs may lead to biased hydrological simulations due to the uncertainty of predicted precipitation and runoff in global scale (Prudhomme and Davies 2009). Thus, downscaling the results of global climate models are appropriate for climate change impact assessments and overcome the limitations previously discussed (Beniston et al. 2007; Moeck et al. 2016). Van Vuuren et al. (2011) suggested that using more-emission scenarios would increase uncertainty of global climate models and that combining some downscaling methods could lower it; however, the uncertainty related to the global-regional climate model was inevitable. Crosbie et al. (2011) indicated that the largest uncertainty of global climate models is higher than the uncertainty of the downscaling method. Dynamical downscaling and statistical downscaling are two major downscaling methods for global-scale or large-scale models (Fowler et al. 2007, Prudhomme et al. 2002; Crosbie et al. 2011).

However, downscaling the projection from a global climate model to the scale of a hydrologic or hydrogeologic model can also introduce additional error and uncertainty into the analysis (Jyrkama and Sykes 2007). Mileham et al. (2009) indicated that multiple downscaling methods could result in 50% increase or decrease in recharge using the same GCM and hydrological model. Allen et al. (2010) applied state-of-the-art downscaling method to four global climate models and showed the variability of groundwater recharge was −10.5 to +23.2% in coastal British Columbia, Canada, and Washington State, USA. Kurylyk and MacQuarrie (2013) found out that the variability of average annual groundwater recharge is −6 to +58% with combinations of global climate models, emission scenarios, and downscaling approaches in eastern Canada. In addition to GCM types and downscaling,



hydrological model (Crosbie et al. 2011), limited data (Ordens et al. 2014), emission scenarios (Holman et al. 2012), land types (Shanafield and Cook 2014), heavy precipitation (Seiller et al. 2012), and heatwave (Merz et al. 2011) can also affect the uncertainty of groundwater recharge projection. Comparing the estimated uncertainty at different scales (local and global), there is no significant correlation among these studies (Table 2). The number of global climate models, emission scenarios, and downscaling methods are different in each case and may result in the lack of correlation between scales and uncertainty of groundwater recharge estimation.

Feedback between water-table dynamics, land-surface energy fluxes, and atmospheric response have been investigated in recent studies, which impart a new perspective to understanding the groundwater recharge process in diverse climate change scenarios (Ferguson and Maxwell 2010; Gilbert et al. 2017; Maxwell and Kollet 2008). The strong positive correlation between groundwater depth and land-surface response is usually investigated in a “critical zone”, which is 2–5 m below the surface (Leung et al. 2011; Maxwell and Kollet 2008). With the case studies in the southern Great Plains (USA), the San Joaquin River watershed in central California (USA), North Rhine in Westphalia region (Germany), Ballona Creek Watershed (Los Angeles, USA), various researchers found that groundwater depth dynamics in a “critical zone”—which is driven by lateral water flow—might not only determine the relative sensitivity of regions to temperature and precipitation change but also in turn affect atmospheric response, especially on moisture transportation and precipitation

(Ferguson and Maxwell 2010; Gilbert et al. 2017; Jiang et al. 2009; Maxwell and Condon 2016; Maxwell and Kollet 2008; Reyes et al. 2016; Sulis et al. 2017). Moreover, lateral groundwater flow processes may be vital to characterizing evapotranspiration and plant transpiration at large scales (Maxwell and Condon 2016) and could result in variation of water table, soil moisture, and atmospheric moisture. Meanwhile, land-energy fluxes will be affected by water-table changes due to the positive correlation relationship between land-energy fluxes and water table, and finally, it would have important implications to atmospheric simulation (Maxwell and Condon 2016). However, comparing to the groundwater recharge process, there is a lack of uncertainty estimation in atmospheric response caused by groundwater depth dynamics and lateral water flow fluxes. Thus, additional work is needed to estimate uncertainties in groundwater feedbacks between water table dynamics and land-surface energy fluxes and atmospheric response at different scales.

## Projected impacts, vulnerabilities, and risks

Under climate change, reliable surface-water supply is expected to decrease due to increased variability of river flow, increased precipitation variability, and decreased snow and ice storage (Barnett et al. 2005; De Wit and Stankiewicz 2006; Vörösmarty et al. 2000). Meanwhile, renewable groundwater resources also decrease remarkably under high GHG emission levels. The percentage of the global population living in river

**Table 2** An overview of uncertainty evaluation in several studies about the impact of projected climate change on groundwater recharge

Scale	Number of GCMs	Number of ES	Selection of DM	Uncertainty	References
Global	2	2	NA	−30% ~ +100%	Döll 2009
7,617,930 km <sup>2</sup>	16	3	Delta change method <sup>a</sup>	+45 ~ +283%	Crosbie et al. 2013
451,000 km <sup>2</sup>	19	4	Delta change method <sup>a</sup>	−37% ~ +27%	Meixner et al. 2016
106,000 km <sup>2</sup>	15	3	Daily downscaling approach <sup>a</sup>	+10% ~ +90%	Crosbie et al. 2010
52,000 km <sup>2</sup>	3	8	Bias corrected statistical downscaling <sup>a</sup>	+20% ~ +60%	Meixner et al. 2016
20,000 km <sup>2</sup>	15	3	Daily downscaling approach <sup>a</sup>	−33% ~ +28%	Ali et al. 2012
7,560 km <sup>2</sup>	17	4	Delta change method and CMIP3 GCM <sup>a</sup>	−100% ~ +35%	Serrat-Capdevila et al. 2007
2,600 km <sup>2</sup>	13	1	Change factor method <sup>b</sup>	−26% ~ +31%	Jackson et al. 2011
2,098 km <sup>2</sup>	1	1	Providing regional climates for impact studies <sup>b</sup>	−50% or + 50%	Mileham et al. 2009
1,600 km <sup>2</sup>	1	2	Stochastic weather generator <sup>a</sup> and change factor method <sup>b</sup>	−14% ~ −37%	Holman et al. 2009
581 km <sup>2</sup>	5	2	Change factor method <sup>b</sup>	−20% ~ +7%	Dams et al. 2012
161 km <sup>2</sup>	4	1	state-of-the-art downscaling method <sup>a</sup>	−10.5% ~ +23.2%	Allen et al. 2010
9.5 km <sup>2</sup>	4	4	Daily translation method <sup>a</sup>	−6% ~ +58%	Kurylyk and MacQuarrie 2013

GCMs global climate models; ES emission scenarios; DM downscaling methods; NA not applicable

<sup>a</sup> Statistical downscaling method

<sup>b</sup> Dynamic downscaling method

basins with new or aggravated water scarcity is projected to increase with global warming, from 8% at 2 °C to 13% at 5 °C (Gerten et al. 2013).

## Water resources

Water scarcity is the lack of sufficient available water resources to meet water needs within a region, and it is referred to, in increasing order of severity, as “water shortage,” “water stress or deficits,” and “water crisis” (Gleick 1998; Rijsberman 2006). Water scarcity is an important index for evaluating regional water-resource crises. Models have shown that mean global warming of 5 °C would expose an additional 13% of the world population to new or aggravated water scarcity (Gerten et al. 2013).

Projected water scarcity varies by region (Bernstein et al. 2008). Northern China and North Africa are expected to suffer less water stress while Southern Europe and most of the Middle East are expected to experience more intense water stress based on simulation with agro-land use model and water supply–demand model (Hayashi et al. 2013).

According to statistical data, there were 1.2 billion people living in areas of water scarcity in 2003 (Rijsberman 2006). By 2050, under RCP 8.5, an additional 0.5–3.1 billion people are projected to be exposed to water scarcity (Gosling and Arnell 2016). That means the population affected by water scarcity will increase by 42 to 258%.

In a study specifically about groundwater, 27–50% of the projected global population (SSP2 population scenario) is projected to suffer from a decrease of renewable groundwater resources of more than 10% by the 2080s (as compared to 1980s levels) under RCP 8.5 (Portmann et al. 2013). The land area affected by the decrease in groundwater resources is projected to increase linearly with global mean temperature rise between 0 and 3 °C; for each degree of global mean temperature increase, an additional 4% of the global land area is projected to suffer a decrease in groundwater resources of more than 30%, and an additional 1% to suffer a decrease of more than 70% (Portmann et al. 2013).

## Agricultural irrigation

Under RCP 8.5, global irrigation water demand (IWD) is projected to increase by 6.7 to 10% (Hanasaki et al. 2013). Future IWD is projected to increase by more than 20% by 2100 during the summer in the Northern Hemisphere (Döll 2002; Konzmann et al. 2013; Wada et al. 2013). Due to the increased water demands of intensified agriculture in 2100 (Schewe et al. 2014), more than 17% of the global population will suffer from a severe reduction (more than 20%) in water resources.

The present peak IWD is projected to shift 1 month or more in  $\geq 80\%$  of current global irrigated areas, currently home to 4

billion people—for example, peak IWD is projected to occur approximately a month later in Eastern Asia (June–July), the Arabian Peninsula (May–July), and Northern Africa (May–June). However, these peaks may occur about 1 month earlier over Central America (April–March) and Central Asia (July–June; Wada et al. 2013).

## Energy production

Climate change affects hydropower generation through changes in mean annual streamflow, shifts of seasonal flows, increases of streamflow variability (including floods and droughts), increases in evaporation from reservoirs, and changes in sediment flux (Bergström et al. 2001; Hamududu and Killingtveit 2012; van Vliet et al. 2016). Global hydropower potential is expected to increase 6.3% (RCP 8.5) by the 2080s compared to 1971–2000, according to GCM-global hydrological models (GCM-GHMs; van Vliet et al. 2016). The strongest increases in hydropower potential are expected for Central Africa, India, Central Asia and the northern high-latitudes, with 18–33% of the world population living in these areas by the 2080s (van Vliet et al. 2016). Hydropower projections correspond with a pattern of precipitation under RCP 8.5; high latitudes of the Northern Hemisphere have more precipitation.

In addition to the effects of precipitation, changes in temperature, evapotranspiration, and snowmelt, as well as warmer ambient temperatures will also affect the hydrologic regime and hydropower production (Schaeffli et al. 2007; van Vliet et al. 2016). By the 2050s, global warming effects will be felt in US states in the lower Colorado River Basin where hydropower generation will decline by as much as 40% due to decreasing reservoir levels and flows (Barnett et al. 2004). By the 2080s, warmer ambient temperatures leading to warmer cooling water are projected to lower thermal power plant efficiency and thus decrease electricity production by 1.5–3% in European countries under emissions scenario SRES A1fi (Golombek et al. 2012).

## Municipal services

Rising surface-water temperatures will not only affect hydropower efficiency. They will also affect basic parameters such as pH, DO, and dissolved organic carbon (DOC; Evans et al. 2005) as well as increase pollutant concentrations in drier places (Van Vliet and Zwolsman 2008). For instance, results of the PAYSER model have identified an increased risk of water-treatment-plant “failure” that would result in high nitrate pollution levels in the 2050s through 2090s under A1fi at Grafham (located in the east of England). Additionally, DYRESM-CAEDYM modeling has been applied to project deterioration of water quality by the increased frequency of stratification and increased phytoplankton and DOC levels

caused by decreasing precipitation in the reservoir (Thorne and Fenner 2011).

On the other hand, increased precipitation may cause its own problems in the form of increased pathogen loads. Five simulated rainfall events illustrated that under these events, runoff release of staphylococci, enterococci, and *Clostridium perfringens* (*C. perfringens*) was ~3–6 log units greater in the litter because of aggravated leaching pollution (Brooks et al. 2009). This would increase water treatment costs in farming areas; nevertheless, on the positive side, biological water and wastewater treatment are more efficient when the water is warmer (Tchobanoglous et al. 2003).

### Freshwater ecosystems

Increased air temperature will affect the temperature of surface water and groundwater and result in variation in lentic systems, especially the quality of fish and amphibian habitat (Ficke et al. 2007). In the northeastern United States, there would be a nearly 50% decrease of fish habitat in cold and cool water under a hypothetical scenario of a projected air temperature increase of 4.5 °C (Eaton and Scheller 1996). Similarly, warming of just 3.8 °C is projected to drastically reduce the range of brook trout in the southeastern United States (Flebbe 1993) and in southern Canada (Meisner 1990).

Moreover, the vulnerabilities of groundwater-dependent ecosystems (GDEs, including groundwater-dependent springs, lakes, rivers, wetlands, peatland, wet forest and species) will also be more prominent under such extreme CO<sub>2</sub> concentration (Kløve et al. 2014). CO<sub>2</sub> not only impacts most important climate forcing factors for RCP 8.5 but also affects the productivity of many terrestrial plants and aquatic organisms (Van Vuuren and Carter 2014)—for example, through 26 GCMs simulation based on RCP 8.5, the number of ecoregions that are impacted by habitat loss and fragmentation–climate change (HLF–CC) will increase to 170 in 2090 (Segan et al. 2016). Due to the high density of CO<sub>2</sub>, the acidification of the superficial zone in GDEs may be dominant in the regions of rainfall increase (Bertrand et al. 2008). Meanwhile, it will also alter the sulfur dynamic, organic acid production, and soil acid-buffering ability in GDEs (Kløve et al. 2014).

Increased soil erosion caused by additional precipitation will also impact water quality, especially in agricultural countries such as China, India, and much of Latin America (Lal 1998). Non-point source pollution of nitrogen, phosphorus, and pesticides may affect river water quality along with surface runoff inflow into rivers, thus leading to eutrophication. In eutrophic freshwater systems, 700 ppm of atmospheric CO<sub>2</sub> may result in an increase in phytoplankton productivity of more than 50%; therefore, nuisance phytoplankton blooms may be further aggravated by elevated atmospheric CO<sub>2</sub> concentrations (Schippers et al. 2004).

Climate change also has a connection with eutrophication in lakes (Straile and Adrian 2000). Higher temperatures can reduce water viscosity, enhance nutrient diffusion, and induce vertical gradient variation of pH and DO (Peperzak 2003). Meanwhile, the decreased viscosity caused by increased temperature will promote the sinking of larger, non-motile phytoplankton that have weak buoyancy regulation mechanisms (e.g. diatoms, Paerl and Huisman 2009). Higher temperatures also foster cyanobacteria blooms, increasing the potential of eutrophication in lakes (Paul 2008).

### Links to other sectors and services

Adaptation in other sectors such as agriculture, forestry, and industry can have impacts on the freshwater system and should therefore be considered as part of adaptation planning in the water sector (Jiang et al. 2005). Climate change parameters include higher temperatures, changes in precipitation, and higher atmospheric CO<sub>2</sub> concentrations. Although rising temperature has both positive and negative effects on crop yields, in general it has been found to reduce yields and quality of many crops, most importantly cereal and feed grains (Adams et al. 1988, 1998). An atmosphere with higher CO<sub>2</sub> concentrations would result in higher net photosynthetic rates (Allen et al. 1987; Cure and Acock 1986). It may also reduce transpiration as plants reduce their stomatal apertures, the small openings in the leaves through which CO<sub>2</sub> and water vapor are exchanged with the atmosphere (Adams et al. 1998)—for example, in Amazonia, stomatal closure caused by increasing CO<sub>2</sub> concentration and temperature contributes approximately 20% to the precipitation decrease, suppressing local evaporative water recycling (Betts et al. 2004).

Under RCP 8.5, the increased surface-water temperature will encourage algal blooms not only in the marine system but also in interior lakes (van Vliet et al. 2016). For example, there was a cyanobacterial bloom outbreak event at Taihu (the third largest freshwater lake in China) because of an unusually warm and dry spring in 2007, which induced a drinking water crisis that affected approximately 2 million people for at least 1 week (Zhu and Schwartz 2011). According to five projected model results, cyanobacterial blooms will likely occur with more frequency in the coming decades (Qin et al. 2010).

### Perspective of water resources management

In general, available water resources will decline by 2100 under RCP 8.5 due to increased urban, agricultural, industrial water consumption; increased evapotranspiration in low latitudes; and additional water consumption by vegetation expansion in areas where human pressure will decrease and cultivation will disappear (García-Ruiz et al. 2011). Climate and land use changes indicate an increase in water stress and a shortage

of water availability, especially in the mid-latitude and sub-tropical dry regions. However, due to an increase in precipitation, streamflow and groundwater are projected to increase significantly and flood frequency is also estimated to rise in the high latitudes of both the Northern and Southern Hemispheres (Christensen et al. 2013; Hirabayashi et al. 2013; Portmann et al. 2013). Thus, water resource management needs to be improved to respond to future challenges.

Comprehensive management approaches to water resources may greatly reduce human vulnerability to extreme GHG emissions and climate change and thereby promote global water and food security (Taylor et al. 2013b). However, current strategies and policies cannot cope with such dramatic increase of global water crisis and extreme GHG emission scenarios. Thus, water resource management needs to be adapted to respond to future challenges.

To deal with projected hydrologic cycle changes and potential risks, several strategies for water resource management under scenario RCP 8.5 are feasible:

1. *Increase development of hydraulic infrastructures.* Developing hydraulic infrastructure in the high latitudes, mid-latitudes, and wet tropical basin, while also developing some indispensable hydraulic infrastructure in the mid- and low latitudes of the Northern and Southern Hemispheres would help precipitation to dramatically increase in these regions in both the Northern and Southern Hemispheres and enhance suitable conditions for hydropower generation. The development of reservoirs and hydropower dams would not only adjust streamflow and achieve flood control (Schwanenberg et al. 2015) but also would meet increasing future electricity demands (Hamududu and Killington 2012). Decreased precipitation will challenge the affected countries where water supply already depends mainly on irregular stream flows, thwarting increasing crop production with its increasing irrigation demand. Indispensable reservoir development for rainfall and surface runoff storage can help adjust water resource distribution for regions with a lack of sufficient available water resources to meet demand (García-Ruiz et al. 2011). Moreover, adjustment of water resources and reservoir management cannot be ignored but must rather adapt to changing river regimes and water resource availability.
2. *Optimize conjunctive use of surface water and groundwater.* An effective option to reduce groundwater depletion under RCP 8.5 scenario is conjunctive usage of surface water and groundwater. Although there are several methods for conjunctive use, the optimization of alternative water-resource-development modes still needs further study (Giordano 2009; Iglesias et al. 2007; Scanlon et al. 2012). Identification of inherent uncertainty in water resource projections and the residence time of freshwater

in both surface-water and groundwater systems is essential to set sustainability goals of conjunctive use (Aeschbach-Hertig and Gleeson 2012). Further development of conjunctive usage of surface water and groundwater such as riverbank filtration technology, is also vital to mitigate future water crises (Hiscock and Grischek 2002; Hoppe-Jones et al. 2010; Hu et al. 2016).

3. *Enhance groundwater recharge from supply-side management via technical development.* Aquifer recharge management is essential to mitigate groundwater depletion issue when domestic, agricultural, and industrial water habits alter or occur during droughts (Aeschbach-Hertig and Gleeson 2012). Artificial recharge of groundwater with excess surface water, desalinated water, and treated wastewater is an effective method for aquifer recovery and has been applied in the North China Plain, Western United States, and India (Foster et al. 2004; Shah 2009; Scanlon et al. 2012). However, the efficiency of artificial recharge is unsatisfactory, and technologies of seawater desalination and wastewater treatment are immature (Elimelech and Phillip 2011). Thus, development of seawater desalination and sewage purification with emerging technologies will be beneficial to artificial recharge of groundwater and in the face of droughts.
3. *Improve water treatment technologies.* Technological improvements in sewage treatment, seawater desalination, and so forth are essential to sustainable water resource development in water stressed regions. Increasing freshwater resources by seawater desalination and sewage purification with emerging technologies will help global-water-crisis mitigation. Limited efficiency and capital operation costs are common problems for both sewage treatment and recycling and seawater desalination with current technological level (Missimer et al. 2013; Muga and Mihelcic 2008; Peñate and García-Rodríguez 2012; Van der Bruggen et al. 2003). Moreover, for seawater desalination, the high specific energy consumption, membranes selection, harmful effects of scaling and fouling on membranes may further limit desalination efficiency (Peñate and García-Rodríguez 2012). Although rainwater harvesting (RWH) is an integral part of the sustainable rainwater management for coping with future climate change, especially in the areas where precipitation will significantly increase (Kahinda et al. 2010; Pandey et al. 2003), the capital costs, materials for manufacturing systems, as well as the costs and resource requirements related to system installation and maintenance are also neglected (Ward et al. 2012). Thus, to promote essential water treatment technologies, further research is required in materialogy, microbiology, environmental science, toxicology, and relevant other subjects.
4. *Improve integrated hydrologic modeling.* The purpose for establishing an integrated hydrologic model is to better



understand and predict the hydrologic cycle with climate changes at both the global and regional scale and to develop measures to cope with the projected hydrologic change (Green et al. 2011; Maxwell et al. 2014). However, as already discussed (see section “[Surface-water/groundwater interaction](#)” and section “[Uncertainty analysis of the groundwater recharge](#)”), the limitation of current integrated hydrologic modeling is highlighted such as the discrepancy of temporal-spatial scale in sub-models (e.g. climate model, hydrological model and groundwater model) and hyporheic zone generalization. In the temporal-scale, most of existing integrated hydrologic models prefer to focus on the current situation rather than make a projection for the future with different climate change impacts (Kløve et al. 2014; Pulido-Velazquez et al. 2015). Furthermore, inherent uncertainties in models also affect the reliability of the projection by integrated models (Crosbie et al. 2013; Smerdon 2017). Thus, integrated hydrologic models need to be further developed, not only in computing methods but also in fundamental theories, especially for description and generalization of hyporheic zones. Meanwhile, models also need to shift the simulation target from the current (2000–2020) to the future (2080–2100) under multiple climate change and GHG emission scenarios. Moreover, lowering the uncertainties in integrated hydrologic models is essential to develop possible adaptation strategies for policymakers.

5. *Establish new legal and policy frameworks for water resource management.* It is critical to propose regionally adapted strategies from this range of options and generally strengthen regulation, policy and management for water, energy and agriculture (Giordano 2009; Sophocleous 2010; Theesfeld 2010). There is no uniform solution for water resource management due to climatic, hydrologic, political, social-economic conditions that vary significantly between different regions (Aeschbach-Hertig and Gleeson 2012). Future research needs to identify appropriate, adaptable, and sustainable long-term strategies for each region no matter whether it is arid, semi-arid, or wet, and this research must try to transfer knowledge and measures between regions; only then can change be promoted to meet the challenges of water resource management at a global scale under extreme GHG emission and global warming scenario.

Moreover, mitigation of CO<sub>2</sub> emissions is vital to slowing global warming and mitigating future water crises. Strong institutional frameworks will need to be established socially, economically, and environmentally in order to achieve CO<sub>2</sub> limit goals and relieve water crises caused by global warming. Carbon capture and storage (CCS) and development of renewable clean energy are feasible to mitigate CO<sub>2</sub> emissions and achieve the target. Recently, bioenergy carbon capture and storage

(BECCS) have been promoted as an effective mitigation option for GHG negative emissions (Fuss et al. 2014) due to their ability to achieve negative CO<sub>2</sub> emissions (Obersteiner et al. 2001). However, water consumption of BECCS may itself jeopardize the sustainability of global water resources. Therefore, the water balance of the widespread application of BECCS needs to be assessed.

## Conclusions

This paper synthesizes model predictions of the changed groundwater system and hydrologic cycle in the literature under the global warming scenario of CO<sub>2</sub> at 936 ppm in 2100 (RCP 8.5). These model predictions describe a global hydrologic cycle that is dramatically different from that of today. The impacts of climate change on groundwater resources vary regionally due to the regional distribution of precipitation and evapotranspiration, localized interaction between surface water and groundwater, distances from the oceans and impacts from sea-level rise, and teleconnection pattern effects. Generally, the groundwater resource variability corresponds to precipitation distribution patterns in RCP 8.5 scenario: (1) groundwater resource depletion will increase in tropical and/or subtropical regions, and (2) groundwater resource quantity will increase in the high-latitude regions of the Northern and Southern Hemispheres. Meanwhile, increased evapotranspiration may lead to a dramatic decline of streams, lakes, and other surface-water resources and result in a shift from traditional source for urban, industrial, and agricultural water use from surface water to groundwater, thereby accelerating its depletion (e.g. Southwestern America, Southern Europe, India). An increase in demand of agricultural development and sea-level rise will lead to water degradation in agricultural and coastal areas by nonpoint pollution and seawater intrusion, respectively.

However, predicted groundwater recharge variations as a result of global warming carry large uncertainties. GCMs types and downscaling methods are two dominating factors affecting the reliability of groundwater recharge prediction. Studies of feedbacks among water-table dynamics, land-surface energy fluxes, and atmospheric provide a new perspective to understanding the groundwater recharge processes in diverse climate change scenarios, but there is a lack of uncertainty estimation in atmospheric response caused by groundwater depth dynamics and lateral water flow fluxes.

Besides, for global hydrologic cycle under RCP 8.5, a 4–6 °C increase in global mean temperature will increase the global mean precipitation 54.75 mm/year, evapotranspiration by 5%, and available fresh water by 68.9%, while decreasing groundwater resources by 38% by the end of the twenty-first century compared to the pre-industrial period. However, these

global average values mask a startling reality: water scarcity will be exacerbated in areas already experiencing water stress. Under RCP 8.5, the global mean sea level will rise 0.63 m by 2100 and soil erosion will increase 9% by the 2090s, while the frequency of extreme hydrologic events, i.e., floods and droughts, will increase for 18–30% of the land area. These changes in the hydrologic cycle will seriously threaten water resource security. Thirteen percent of the world's population will suffer new or aggravated water scarcity, whereas due to increased demand of intensified agriculture, more than 17% of the global population may suffer from a severe reduction (more than 20%) in water resources, further complicating water resource distribution. An increase in surface-water temperature will contribute to ecosystem degradation, a reduction of hydropower efficiency and overall generation (affecting more than 46% of the world population), and increased water treatment costs.

Moreover, water security is one important impetus among many for human societies to slow down global warming. In addition to hydraulic infrastructure construction and institutional framework establishment, conjunctive usage of surface water and groundwater, water-treatment-technology improvement, integrated hydrologic modeling improvement, groundwater recharge enhancement by human intervention, and mitigation of CO<sub>2</sub> emission is essential to water resource management. Such an extremely complicated problem will require interdisciplinary cooperation to respond to climate change, river regimes, reservoir management, and CO<sub>2</sub> emission mitigation, necessitating the participation of engineers, climatologists, hydrologists, human and physical geographers, general environmentalists and stakeholders.

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