

Metacomunity dynamics of complex life-cycles in heavily-disturbed landscapes

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1 Abstract

Human sprawl is all over the globe, affecting most of species on Earth. Metacommunity models predicts that populations can survive in a fragmented landscape consisting of several patches. Structured metacommunity models have been extensively used to determine dynamics of species in future environmental scenarios. We herein developed a metacommunity dynamics model for highly-fragmented species in areas with high anthropogenic pressure. Four models were developed considering the main trait-based drivers concerning mating, dispersal, biotic and environmental variables.

2 Keywords

Metacommunities, dispersal dynamics, individual-based model, human-driven perturbations, Approximate Bayesian Computational

3 Introduction

Movement is a pervasive trait among species on Earth. All species disperse to some extent, because resources become limited as population grow, but a few have a worldwide distribution. Mobility occurs during the whole life cycle or limited to dispersal stages, such as seeds in plants or eggs in fish. Most of the species move over short distances and, to a lesser extent, to greater distances (Davidson et al., 2004). Dispersal, defined as the movement of individuals

away from their source (Nathan et al., 2003), is a key strategy to increase fitness in dynamic landscapes by moving to different environments. Hence, ecological and evolutionary processes that occur in one location may drive changes at far sites or in other ecosystems through the ecological coupling of long-distance dispersal (Gaines et al., 2007). Fitness variability between habitat patches is the driving force for dispersal evolution, comprising pivotal ecological processes such as, interactions, habitat quality and competition (Bowler and Benton, 2005). Yet, highly-fragmented populations may not experience sufficient levels of dispersal, with a subsequent increase of inbreeding which lead to decreased fitness in many species (Fath, 2018). Small populations with high dispersal rates may be prone to extinction under such situations whilst marginal populations with immigration rates may experience a rescue effect where individuals may persist (Eriksson et al., 2014). Unfortunately, most models assumed a fixed dispersal strategy (McCallum et al., 2001; Levin et al., 2003), but understanding the link between dispersal and population dynamics is pivotal for predicting population responses to habitat loss and fragmentation, and interaction with alien species (Bowler and Benton, 2005).

Human footprint is everywhere, even in protected areas to safeguard biodiversity (Tournaire, 2014; Jones et al., 2018). Earths ecosystems are becoming increasingly fragmented, with extensive habitat loss and species depletion (Haddad et al., 2015). Fragmentation modifies the structure, diversity, dynamics, species composition and recruitment rates of communities through habitat reduction and subsequent edge effects (Laurance et al., 1998; Short et al., 2011). Natural fragmentation has been associated with processes that maintain or even increase biodiversity (Tilman, 1988), but human-driven fragmentation is a major threat for biodiversity (Pimm and Raven, 2000) and a disruption of ecosystem processes (Achard et al., 2002). The effects of fragmentation have been traditionally studied using the framework of island biogeog-

raphy (Whittaker and Fernández-Palacios, 2007), but recent works have included tools, e.g. population genetics (Young et al., 1996), or approaches such as, metapopulation (Hanski and Gaggiotti, 2004) and metacommunity theories (Mouquet et al., 2011).

Metapopulation models predict the survival of highly-disperse populations regardless local depletion of species (Hanski, 1999; Akçakaya et al., 2007). However, habitat loss and fragmentation may be so extensive that result in a massive species extinction (Montoya, 2008; Rybicki and Hanski, 2013; Haddad et al., 2015). Model predictions have shown that a decrease in connectivity among assemblages, from continuous to sparsely-distributed populations is accompanied by species loss (Metzger et al., 2009; Niebuhr et al., 2015). The role of natural (e.g. stochastic events) and human-driven (e.g. pollution, habitat loss and exploitation) perturbations have been extensively studied in the last decades (Dornelas, 2010) and references therein) and it has been frequently studied in ecological theory (Volkov et al., 2007; Gardner and Engelhardt, 2008). The degree of anthropogenic pressure may be a capital factor for landscape connectivity, since directly affects the persistence or decrease of assemblages (Supp and Ernest, 2014). Besides perturbations, dispersal rates in disturbed landscapes need to be high in order to maintain viable populations (Provan et al., 2008). Thus, individual-based models (IBM) are needed to predict dispersal rates in species where small specimens, i.e. juveniles, are not reproductively active and even larger-sized adults harbour the highest reproductive potential (Hendriks and Mulder, 2008; Werner and Griebeler, 2011). IBM simulate populations as being composed of discrete individual organisms (DeAngelis and Grimm, 2014; van der Vaart et al., 2016). In IBMs the actions of single individuals are simulated, and they interact with others and the landscape they live in (DeAngelis and Mooij, 2005). They incorporate attributes that vary among individuals and can change through time such as, growth, foraging, dispersal and reproduction, among others (Mar-

tin et al., 2013; DeAngelis and Grimm, 2014). These models have been used as size-structured methods to integrate a high variety of data which output are pivotal for management purposes (Punt et al., 2013) and conservation strategies (Nabe-Nielsen et al., 2014). Metapopulation models have been primarily used to certain vagile species such as insects (Harrison et al., 1995; Hilker et al., 2006) or mammals (Krohne, 1997; Lawes et al., 2000), but sessile or semi-sessile species have been neglected though a high proportion have a larval dispersal phase (Sale et al., 2006; Gaggiotti, 2017). This strategy is extensively used in most of reduced-mobility marine species, with a subsequent increase of populations connectivity to nearby areas. However, assemblages from distant places are prone to isolation, as genetic analysis showed in semisessile mollusks from oceanic archipelagos (Côte-Real et al., 1996; Gaggiotti, 2017; Faria et al., 2017). Oceanographic conditions, i.e. currents and eddies, and large geographic distances showed to be pivotal environmental factors to decrease the dispersal capacity (Palumbi, 1994). High larvae mortality plays also a crucial role, mainly due to predation and larvae traits, i.e. growth, reproductive and recruitment strategies, and mobility capacity (Cowen et al., 2000; Cowen and Sponaugle, 2009) in a limited connectivity within species. However, genetic analysis has been observed to be insufficient to accurately determine the demographic connectivity among populations of terrestrial (Chapuis et al., 2011), estuarine (Turner et al., 2002) and marine species (Hawkins et al., 2016).

We herein develop metacommunity models based on individuals of two intertidal species, namely the limpets *Patella aspera* and *P. candei*. The first model assumes that dispersal rates between patches are distance-dependent, with low rates between highly-separated assemblages. The second model assumes that dispersal rates are positively correlated to individual density. The third model assumes that larger individuals have higher reproductive potential. The fourth

model considers a low probability of dispersal to peripheral assemblages relative to central ones. We confront the model with long-time series data (1994-2014) of two commercial limpet species (*Patella candei* and *P. aspera*) in two overpopulated oceanic islands (> 500 inhab km^{-2}) with a high harvesting pressure (Riera et al., 2016; Sousa et al., 2019). Former studies highlighted the sharp decrease of sizes of both limpets, a symptom of overexploitation and hence, a major driver for future local extinction of these species in Tenerife (Riera et al., 2016) and Madeira (Sousa et al., 2019). We herein develop a series of models of metacommunity dynamics to predict future trends on the limpet populations of both islands. The predictive power of these models is pivotal to articulate an integrative management plan to guide future conservation actions. A deep understanding of the eco-evolutionary effects of connectivity loss is essential to guide future conservation actions (Travis et al., 2013; ?). Trait evolution is a pivotal topic in metacommunity dynamics, and we herein particularly focused on traits related to mating, dispersal and biological and environmental traits. These traits greatly influence the dispersal potential of species, specifically important in highly-fragmented populations, and complex life-cycles with contrasting dispersal potential. We expected individuals from size-biased populations to have lower dispersal rates than well-structured populations with a good representation of adults. Also, theory predicts that populations from more fragmented habitats are subjected to a lower propensity of dispersal.

4 Material and Methods

4.1 Metapopulation dynamics

Spatial population dynamics are based on eco-evolutionary traits of species. Metacommunity and metapopulation models (REFERENCES HEREIN, INCLUDE SEVERAL STATE-OF-THE-ART PAPER Xu et al. 2006 CHECK IT OUT!!) focused on simple life cycles (EXAMPLE) continuous habitats (EXAMPLES) or in equilibrium between extinctions and colonizations (Hanski, 1994). Also, our metacommunity model is based on stochastic patch occupancy models (Moilanen, 2004) that have been extensively used in metapopulation studies (Ovaskainen and Hanski, 2004). However, species with complex life stages, with patchy distributions in a highly-fragmented landscape have been neglected by these models. Current scenarios show that an extensive number of species are subjected to intense human-driven disturbances directly, e.g. hunting or fishing, or indirectly, e.g. habitat loss and degradation. The connectivity of populations is of utmost importance to prevent their extinction in local and global populations. Reproductive individuals become rather important for the maintenance of future generations; however, extraction activities are mostly primarily targeting larger and older specimens, with the highest reproductive potential.

4.2 Trait dynamics

Four main traits were included in the present metacommunity model, comprising the pivotal ecological, evolutionary and environmental factors concerning the studied metacommunities.

(i) *MATING* The studied species have external fecundation, with random encounter between gametes in the water column. Thus, a lottery model was assumed. A third submodel was conducted to integrate the exponential decay of gametes after their release and the subsequent

decrease of fertilized eggs per time.

(ii) *ABIOTIC* We assumed that the study species are constrained to disperse passively by the large-scale (the Canary Current, (Barton et al., 1998) and meso-scale (eddies) oceanographic conditions. Current data were integrated in a submodel to calculate distances between sites. We also assumed that exposed sites are less affected by harvesting pressure than sites easily-accessible all over the year; thus, a submodel integrating exposition and extraction as trade-off was herein carried out.

(iii) *BIOTIC* Harvesting pressure is the main responsible of the high-fragmented limpet populations in both studied islands (Riera et al., 2016; Sousa et al., 2019). We assumed that this anthropogenic driver remains constant throughout the last 20 years, and with no seasonal variations within a year. We herein used the limpet size as a reproductive proxy, since individuals over 3.5 mm are considered adults, since 50% of individuals are reproductive (Henriques et al., 2012), and the higher the size the higher the reproductive potential (Boaventura et al., 2002; Martins et al., 2017). Several submodels were conducted to integrate the human-driven consequences on the studied species. Management measures were integrated, i.e. closed season (December-March in Madeira); minimal harvesting size (40 mm in Madeira and 45 mm in Canaries for both species), harvesting limitations per fishermen (15 kg per day for professional fishermen in Madeira; 10 kg per day for professional fishermen and 3 kg per day for recreative fishermen). A submodel comprises the dependence of extraction on limpet size, i.e. that larger limpets are subjected to more intense harvesting pressure than shorter individuals. No human-driven factors were also herein included in this model trait, since death probability regardless the individual size was also integrated in a submodel.

(iv) *DISPERSAL* The 5-stage life-cycle of the studied species greatly determines the dispersal

potential of *Patella candei* and *P. aspera*. The limpet size is a proxy of reproductive potential, i.e. the larger the specimen the higher the number of gametes released by the individual. Because of the different dispersal potential of first stages, i.e. egg, trocophore and veliger larvae, three kinds of dispersal were considered in the submodel. Namely, global dispersion was considered for large distances (> 100 kms), belonging to eggs and trocophores, regional dispersion for distances ranging from 1 to 100 km, belonging to trocophores and veligers, and local dispersion for shorter distance (< 1 km), belonging to late-stage veligers. No dispersion was also included in the submodel to represent juveniles and adults, considered sessile organisms with homing behaviour.

4.3 Case study

Two species of limpets (*Patella candei* and *P. aspera*) were herein used as model system, since they harbour a series of traits that have been uncovered by former metacommunity models. They are characterized by a complex five-staged life cycle, with ontogenic differences in dispersal rate, from large-dispersed eggs to sessile adults. The transition among stages is fast, driven by post-fertilization hours. The size of these species are used as a reproductive proxy, the higher the size the higher the reproductive potential, with individuals over 35 mm considered adults since 0% of individuals are reproductive (Henriques et al., 2012). Hence, a decrease of limpet size is a symptom of local extinctions in patches with low connectivity. Lastly, limpets have been heavily exploited in the study locations, underpinning patchy-dispersed populations within a disturbed landscape. Metacommunity dynamics may reveal the main traits structuring the dispersal, recruitment and settlement of these species in these isolated areas, without input from extensive coastal continental areas.

4.4 Modelling framework

A metacommunity model was developed to explore the probability of occupancy of the two studied species in each oceanic island, i.e. Madeira and Tenerife, based on previous spatially-explicit metacommunity models (Hanski, 1999; Hanski and Ovaskainen, 2000; Ovaskainen and Hanski, 2001; Hanski and Ovaskainen, 2002; Bertuzzo et al., 2015). In the present model, each pixel of the modelled landscape is assumed to be a patch that may be either occupied or not by larvae of the studied limpet species. We assumed that the study species are constrained to disperse passively by the large-scale (the Canary Current, (Barton et al., 1998) and meso-scale (eddies) oceanographic conditions, and also by the active dispersion of the larvae (Henriques et al., 2012). Former studies have demonstrated that the main driver of the sharp decrease of limpet populations is the human harvesting pressure, we assumed that remains constant throughout the last 20 years, and with no seasonal variations within a year. Both species (*Patella candei* and *P. aspera*) are in clear regression in both islands due to their commercial interest (Riera et al., 2016; Sousa et al., 2019).

We consider an empirically sampled landscape consisting of N sites. At each site, there are two competing consumer species each with a complex life-cycle population of different abundances. Resources and natural predators are not considered in the present metacommunity model. To model spatio-temporal changes in population abundances, trait, population and dispersal dynamics need to be defined (see Table 1). We combine the lottery competition model (Chesson and Warner, 1981) with trait-based dynamics accounting for complex life-cycles, exposition to human settlements, Marine Protected Areas (MPAs), spatial heterogeneity and exploitation intensity. The spatial structure of the present model is static, and the site matrix comprises 3 vectors, i.e. size, exposition to environment and extraction by humans. Individuals

are dynamic spatial and temporally, because of their dispersal potential.

Model 1: Infinite sites in homogeneous landscapes This scenario represents a large number of sites with no differentiation between North and South, different rates of human-driven disturbance or asymmetry in dispersal probabilities. Details of the biotic (weak competition), abiotic, mating and dispersal traits

Model 2: Infinite sites in heterogeneous landscapes A large number of sites with differentiation between North and South, exposition to human settlements and asymmetry in dispersal probabilities Details of the biotic (weak competition), abiotic, mating and dispersal traits

Model 3: Finite sites in homogeneous landscapes Details of the biotic (strong competition), abiotic, mating and dispersal traits

Model 4: Finite sites in heterogeneous landscapes Details of the biotic (strong competition), abiotic, mating and dispersal traits

5 Results

(3) OUTPUTS IN-OUT ratio -¿ Source or Sink sites Overlap with Marine Protected Areas (MPAs), Sources-Sinks?

(4) Approximate Bayesian Computational (ABC) and Bayes Networks (BN)

6 Discussion

Hihgly fragmented populations Complex life-history (5 phases), with different dispersal rates Trait-based model,based on empirical data Exploring gradient of complexity in the model

This strategy is extensively used in most of reduced-mobility marine species, with a subse-

quent increase of populations connectivity to nearby areas. However, assemblages from distant places are prone to isolation, as genetic analysis showed in semisessile mollusks from oceanic archipelagos (Corte-Real et al. 1996; Bird et al. 2011; Faria et al. 2017). Oceanographic conditions, i.e. currents and eddies, and large geographic distances showed to be pivotal environmental factors to decrease the dispersal capacity (e.g. Palumbi et al. 1994). High larvae mortality plays also a crucial role, mainly due to predation and larvae traits, i.e. growth, reproductive and recruitment strategies, and mobility capacity (Cowen et al. 2000; Cowen and Sponaugle 2009) in a limited connectivity within species. However, genetic analysis have been observed to be insufficient to accurately determine the demographic connectivity among populations of terrestrial (Chapuis et al. 2011), estuarine (Turner et al. 2002) and marine species (Hawkins et al. 2016).

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References

- Achard, F., H. D. Eva, H.-J. Stibig, P. Mayaux, J. Gallego, T. Richards, and J.-P. Malingreau, 2002. Determination of deforestation rates of the world’s humid tropical forests. *Science* 297:999–1002.
- Akçakaya, H. R., G. Mills, and C. P. Doncaster, 2007. The role of metapopulations in conservation. *Key topics in conservation biology* Pp. 64–84.
- Barton, E. D., J. Arístegui, P. Tett, M. Cantón, J. García-Braun, S. Hernández-León, L. Nykjaer,

- C. Almeida, J. Almunia, S. Ballesteros, and Others, 1998. The transition zone of the Canary Current upwelling region. *Progress in Oceanography* 41:455–504.
- Bertuzzo, E., I. Rodriguez-Iturbe, and A. Rinaldo, 2015. Metapopulation capacity of evolving fluvial landscapes. *Water Resources Research* 51:2696–2706.
- Boaventura, D., L. C. da Fonseca, and S. J. Hawkins, 2002. Analysis of competitive interactions between the limpets *Patella depressa* Pennant and *Patella vulgata* L. on the northern coast of Portugal. *Journal of Experimental Marine Biology and Ecology* 271:171–188.
- Bowler, D. E. and T. G. Benton, 2005. Causes and consequences of animal dispersal strategies: relating individual behaviour to spatial dynamics. *Biological Reviews* 80:205–225.
- Chapuis, M.-P., J.-A. M. Popple, K. Berthier, S. J. Simpson, E. Deveson, P. Spurgin, M. J. Steinbauer, and G. A. Sword, 2011. Challenges to assessing connectivity between massive populations of the Australian plague locust. *Proceedings of the Royal Society B: Biological Sciences* 278:3152–3160.
- Chesson, P. L. and R. R. Warner, 1981. Environmental variability promotes coexistence in lottery competitive systems. *The American Naturalist* 117:923–943.
- Côrte-Real, H., S. J. Hawkins, and J. P. Thorpe, 1996. Population differentiation and taxonomic status of the exploited limpet *Patella candei* in the Macaronesian islands (Azores, Madeira, Canaries). *Marine Biology* 125:141–152.
- Cowen, R. K., K. M. M. Lwiza, S. Sponaugle, C. B. Paris, and D. B. Olson, 2000. Connectivity of marine populations: open or closed? *Science* 287:857–859.

- Cowen, R. K. and S. Sponaugle, 2009. Larval dispersal and marine population connectivity. *Annual review of marine science* 1:443–466.
- Davidson, I. C., A. C. Crook, and D. K. Barnes, 2004. Quantifying spatial patterns of intertidal biodiversity: is movement important? *Marine Ecology* 25:15–34.
- DeAngelis, D. L. and V. Grimm, 2014. Individual-based models in ecology after four decades. *F1000prime reports* 6.
- DeAngelis, D. L. and W. M. Mooij, 2005. Individual-based modeling of ecological and evolutionary processes. *Annu. Rev. Ecol. Evol. Syst.* 36:147–168.
- Dornelas, M., 2010. Disturbance and change in biodiversity. *Philosophical Transactions of the Royal Society B: Biological Sciences* 365:3719–3727.
- Eriksson, A., F. Elías-Wolff, B. Mehlig, and A. Manica, 2014. The emergence of the rescue effect from explicit within-and between-patch dynamics in a metapopulation. *Proceedings of the royal society B: biological sciences* 281:20133127.
- Faria, J., G. M. Martins, A. Pita, P. A. Ribeiro, S. J. Hawkins, P. Presa, and A. I. Neto, 2017. Disentangling the genetic and morphological structure of *Patella candei* complex in Macaronesia (NE Atlantic). *Ecology and evolution* 7:6125–6140.
- Fath, B. D., 2018. *Encyclopedia of ecology*. Elsevier.
- Gaggiotti, O. E., 2017. Metapopulations of marine species with larval dispersal: A counterpoint to Ilkka’s Glanville fritillary metapopulations. *in* *Annales Zoologici Fennici*, vol. 54, Pp. 97–113. BioOne.

- Gaines, S. D., B. Gaylor, L. R. Gerber, A. Hastings, and B. P. Kinlan, 2007. Connecting places: the ecological consequences of dispersal in the sea. *Oceanography* 20:90–99.
- Gardner, R. H. and K. A. M. Engelhardt, 2008. Spatial processes that maintain biodiversity in plant communities. *Perspectives in Plant Ecology, Evolution and Systematics* 9:211–228.
- Haddad, N. M., L. A. Brudvig, J. Clobert, K. F. Davies, A. Gonzalez, R. D. Holt, T. E. Lovejoy, J. O. Sexton, M. P. Austin, C. D. Collins, and Others, 2015. Habitat fragmentation and its lasting impact on Earth’s ecosystems. *Science advances* 1:e1500052.
- Hanski, I., 1994. A practical model of metapopulation dynamics. *Journal of animal ecology* Pp. 151–162.
- , 1999. Habitat connectivity, habitat continuity, and metapopulations in dynamic landscapes. *Oikos* 87:209–219.
- Hanski, I. and O. Ovaskainen, 2000. The metapopulation capacity of a fragmented landscape. *Nature* 404:755–758.
- , 2002. Extinction debt at extinction threshold. *Conservation biology* 16:666–673.
- Hanski, I. A. and O. E. Gaggiotti, 2004. *Ecology, genetics and evolution of metapopulations*. Academic Press.
- Harrison, S., C. D. Thomas, and T. M. Lewinsohn, 1995. Testing a metapopulation model of coexistence in the insect community on ragwort (*Senecio jacobaea*). *The American Naturalist* 145:546–562.

- Hawkins, S. J., K. Bohn, D. W. Sims, P. Ribeiro, J. Faria, P. Presa, A. Pita, G. M. Martins, A. I. Neto, M. T. Burrows, and Others, 2016. Fisheries stocks from an ecological perspective: Disentangling ecological connectivity from genetic interchange. *Fisheries Research* 179:333–341.
- Hendriks, A. J. and C. Mulder, 2008. Scaling of offspring number and mass to plant and animal size: model and meta-analysis. *Oecologia* 155:705–716.
- Henriques, P., R. Sousa, A. R. Pinto, J. Delgado, G. Faria, A. Alves, and M. Khadem, 2012. Life history traits of the exploited limpet *Patella candei* (Mollusca: Patellogastropoda) of the north-eastern Atlantic. *Journal of the Marine Biological Association of the United Kingdom* 92:1379–1387.
- Hilker, F. M., M. Hinsch, and H. J. Poethke, 2006. Parameterizing, evaluating and comparing metapopulation models with data from individual-based simulations. *Ecological Modelling* 199:476–485.
- Jones, K. R., O. Venter, R. A. Fuller, J. R. Allan, S. L. Maxwell, P. J. Negret, and J. E. M. Watson, 2018. One-third of global protected land is under intense human pressure. *Science* 360:788–791.
- Krohne, D. T., 1997. Dynamics of metapopulations of small mammals. *Journal of Mammalogy* 78:1014–1026.
- Laurance, W. F., L. V. Ferreira, J. M. R.-D. Merona, S. G. Laurance, R. W. Hutchings, and T. E. Lovejoy, 1998. Effects of forest fragmentation on recruitment patterns in Amazonian tree communities. *Conservation biology* 12:460–464.

- Lawes, M. J., P. E. Mealin, and S. E. Piper, 2000. Patch occupancy and potential metapopulation dynamics of three forest mammals in fragmented afro-montane forest in South Africa. *Conservation Biology* 14:1088–1098.
- Levin, S. A., H. C. Muller-Landau, R. Nathan, and J. Chave, 2003. The ecology and evolution of seed dispersal: a theoretical perspective. *Annual Review of Ecology, Evolution, and Systematics* 34:575–604.
- Martin, B. T., T. Jager, R. M. Nisbet, T. G. Preuss, and V. Grimm, 2013. Predicting population dynamics from the properties of individuals: a cross-level test of dynamic energy budget theory. *The American Naturalist* 181:506–519.
- Martins, G. M., C. D. G. Borges, M. Vale, P. A. Ribeiro, R. R. Ferraz, H. R. Martins, R. S. Santos, and S. J. Hawkins, 2017. Exploitation promotes earlier sex change in a protandrous patellid limpet, *Patella aspera* Röding, 1798. *Ecology and evolution* 7:3616–3622.
- McCallum, H., N. Barlow, and J. Hone, 2001. How should pathogen transmission be modelled? *Trends in ecology & evolution* 16:295–300.
- Metzger, J. P., A. C. Martensen, M. Dixo, L. C. Bernacci, M. C. Ribeiro, A. M. G. Teixeira, and R. Pardini, 2009. Time-lag in biological responses to landscape changes in a highly dynamic Atlantic forest region. *Biological conservation* 142:1166–1177.
- Moilanen, A., 2004. SPOMSIM: software for stochastic patch occupancy models of metapopulation dynamics. *Ecological modelling* 179:533–550.
- Montoya, D., 2008. Habitat loss, dispersal, and the probability of extinction of tree species. *Communicative & integrative biology* 1:146–147.

- Mouquet, N., B. Matthiessen, T. Miller, and A. Gonzalez, 2011. Extinction debt in source-sink metacommunities. *PLoS One* 6:e17567.
- Nabe-Nielsen, J., R. M. Sibly, J. Tougaard, J. Teilmann, and S. Sveegaard, 2014. Effects of noise and by-catch on a Danish harbour porpoise population. *Ecological Modelling* 272:242–251.
- Nathan, R., G. Perry, J. T. Cronin, A. E. Strand, and M. L. Cain, 2003. Methods for estimating long-distance dispersal. *Oikos* 103:261–273.
- Niebuhr, B. B. S., M. E. Wosniack, M. C. Santos, E. P. Raposo, G. M. Viswanathan, M. G. E. Da Luz, and M. R. Pie, 2015. Survival in patchy landscapes: the interplay between dispersal, habitat loss and fragmentation. *Scientific reports* 5:11898.
- Ovaskainen, O. and I. Hanski, 2001. Spatially structured metapopulation models: global and local assessment of metapopulation capacity. *Theoretical population biology* 60:281–302.
- , 2004. From individual behavior to metapopulation dynamics: unifying the patchy population and classic metapopulation models. *The American Naturalist* 164:364–377.
- Palumbi, S. R., 1994. Genetic divergence, reproductive isolation, and marine speciation. *Annual review of ecology and systematics* 25:547–572.
- Pimm, S. L. and P. Raven, 2000. Biodiversity: extinction by numbers. *Nature* 403:843.
- Provan, J., G. E. Beatty, S. L. Keating, C. A. Maggs, and G. Savidge, 2008. High dispersal potential has maintained long-term population stability in the North Atlantic copepod *Calanus finmarchicus*. *Proceedings of the Royal Society B: Biological Sciences* 276:301–307.

- Punt, A. E., T. Huang, and M. N. Maunder, 2013. Review of integrated size-structured models for stock assessment of hard-to-age crustacean and mollusc species. *ICES Journal of Marine Science* 70:16–33.
- Riera, R., Ó. Pérez, O. Álvarez, D. Simón, D. Díaz, Ó. Monterroso, and J. Núñez, 2016. Clear regression of harvested intertidal mollusks. A 20-year (1994–2014) comparative study. *Marine Environmental Research* 113:56–61.
- Rybacki, J. and I. Hanski, 2013. Species–area relationships and extinctions caused by habitat loss and fragmentation. *Ecology letters* 16:27–38.
- Sale, P. F., I. Hanski, and J. P. Kritzer, 2006. The merging of metapopulation theory and marine ecology: establishing the historical context. Pp. 3–28, *in* *Marine metapopulations*. Elsevier.
- Short, F. T., B. Polidoro, S. R. Livingstone, K. E. Carpenter, S. Bandeira, J. S. Bujang, H. P. Calumpong, T. J. B. Carruthers, R. G. Coles, W. C. Dennison, and Others, 2011. Extinction risk assessment of the world’s seagrass species. *Biological Conservation* 144:1961–1971.
- Sousa, R., J. Vasconcelos, P. Henriques, A. R. Pinto, J. Delgado, and R. Riera, 2019. Long-term population status of two harvested intertidal grazers (*Patella aspera* and *Patella candei*), before (1996–2006) and after (2007–2017) the implementation of management measures. *Journal of Sea Research* 144:33–38. URL <https://doi.org/10.1016/j.seares.2018.11.002>.
- Supp, S. R. and S. K. M. Ernest, 2014. Species-level and community-level responses to disturbance: a cross-community analysis. *Ecology* 95:1717–1723.
- Tilman, D., 1988. *Plant strategies and the dynamics and structure of plant communities*. Princeton University Press.

- Tournadre, J., 2014. Anthropogenic pressure on the open ocean: The growth of ship traffic revealed by altimeter data analysis. *Geophysical Research Letters* 41:7924–7932.
- Travis, J. M., M. Delgado, G. Bocedi, M. Baguette, K. Bartoń, D. Bonte, I. Boulangeat, J. A. Hodgson, A. Kubisch, V. Penteriani, et al., 2013. Dispersal and species’ responses to climate change. *Oikos* 122:1532–1540.
- Turner, T. F., J. P. Wares, and J. R. Gold, 2002. Genetic effective size is three orders of magnitude smaller than adult census size in an abundant, estuarine-dependent marine fish (*Sciaenops ocellatus*). *Genetics* 162:1329–1339.
- van der Vaart, E., A. S. A. Johnston, and R. M. Sibly, 2016. Predicting how many animals will be where: how to build, calibrate and evaluate individual-based models. *Ecological modelling* 326:113–123.
- Volkov, I., J. R. Banavar, S. P. Hubbell, and A. Maritan, 2007. Patterns of relative species abundance in rainforests and coral reefs. *Nature* 450:45.
- Werner, J. and E. M. Griebeler, 2011. Reproductive biology and its impact on body size: comparative analysis of mammalian, avian and dinosaurian reproduction. *PLoS One* 6:e28442.
- Whittaker, R. J. and J. M. Fernández-Palacios, 2007. *Island biogeography: ecology, evolution, and conservation*. Oxford University Press.
- Young, A., T. Boyle, and T. Brown, 1996. The population genetic consequences of habitat fragmentation for plants. *Trends in ecology & evolution* 11:413–418.