Appendix 1: Study sites, bird assemblages, and trait selections

Complementary information about study areas

For the placement of the 23 focal landscapes with different forest covers, we also evaluated if forest cover varied within radii of 1, 2, or 3 km based on each landscape centroid. Landscape-level forest cover did not vary more than 5% within those radii. It shows that we are avoiding the influence of any outside larger patch on colonization processes (Pasher et al. 2013).

Sampling sites inside forest patches were spaced a minimum of 800 m apart (1591 \pm 621 m) and they were placed within the forest patch at the end of a 100 m transect, always oriented to the center of the forest patch.

Bird assemblages

We restricted our analysis to forest specialist and forest generalist species, excluding open area species eventually detected inside smaller forest patches. We recorded 180 bird species in both regions: 93 specialists and 87 generalists. **Specialist species richness** was higher in the low-quality than in the high-quality matrix region (82 and 68 species respectively), with 57 species common to both regions. However, the median of specialist species richness per landscape and site in the high-quality matrix was higher than in the low-quality matrix (Table S1.1). There were 11 specialist species (16% of total richness) in the high-quality matrix detected only once (singletons), and 15 singletons (18% of total richness) in the low-quality matrix.

Generalist species richness was quite similar between regions, with 77 and 74 species in the high and low-quality matrix, respectively (64 common species). Accordingly, generalist species richness per landscape and site did not vary much between high-quality and low-quality matrices (Table S1,1). There were 9 singleton generalist species in the high-quality matrix (12% of total richness) and 12 in the low-quality matrix (16% of total richness). For more details in biodiversity metrics of both regions see Boesing et al. (2018).

Table S1.1: Species richness for the assemblages in the high and low-quality matrix landscapes for landscape and local spatial scales.

		Landscape	Local
Assemblage	Total richness	Median (min – max)	Median (min – max)
Specialists	93		
Low-quality matrix	82	32 (17 – 61)	16 (4 – 42)
High-quality matrix	68	40 (29 – 48)	24 (12 – 40)
Generalists	87		
Low-quality matrix	74	38 (34 – 44)	19 (10 – 31)
High-quality matrix	77	44.5 (38 – 49)	22.5 (15 – 31)
Total	180		

Species traits selection

Bird traits selection and the associated hypothesis of how species may respond to habitat loss according to its traits are summarized in Table S1.2 and described in detail in the following paragraphs.

Table S1.2: Bird traits with the information of the operational variables, the hypothesis of species habitat loss filtering according to trait values, and the sources of data acquisition.

Trait	Operational variable	Hypothesis	Data source
Body size	Continuous. Log of mean body mass	Abundance decreases more intensely with	Ramirez et al. 2008;
	in grams.	habitat loss for larger species.	Rodrigues et al. 2019
Nest type	Categorical. Nest in cavities; open or	The abundance of species with open/semi-	Del Hoyo et al. 2014
	semi-open nest; closed nest.	open nests decreases more intensely with	
		habitat loss than for species with other nest	
		types.	
Diet	Categorical main diet: omnivorous,	The abundance of frugivorous and	Sick 1997, Del Hoyo
	frugivorous, nectarivorous,	insectivorous species decreases more	et al. 2014; Willman
	insectivorous, granivorous.	intensely with habitat loss than for species	et al. 2014
	Continuous. Percentage of fruits in	with other diets.	
	the diet.		
	Continuous. Percentage of		
	invertebrates in the diet.		

Foraging Categorical. Main foraging stratum: stratum ground and/or understory, midstory and/or canopy, and all strata.

Continuous. Percentage of lower strata (ground and understory) use.

The abundance of species in the ground and Sick 1997, Del Hoyo understory strata decreases more intensely et al. 2014; Willman than species using midstory, canopy, or all strata.

Body size is one of the key attributes of vertebrates in respect of extinction risk, reproduction, and dispersal (Owens and Bennett 2000, Ripple et al. 2017). In birds, species with large mean body sizes are often considered more vulnerable to extinction given their low population densities, large home ranges, slow growth rates, high energetic requirements, and high sensitivity to anthropogenic overexploitation (Cardillo et al. 2005). The risk of local extinction in altered and smaller habitats correlates with mean body size (e.g. Barbaro and Halder 2009, Flynn et al. 2009, Newbold et al. 2013, Bregman et al. 2016, Bovo et al. 2018). However, sometimes this trait is not a good predictor of community changes (e.g. Tscharntke et al. 2008, Angert et al. 2011), probably because large species may also benefit from having higher mobility (Tscharntke et al. 2012). We compiled information on avian body mass in Boesing et al. (2018a), which followed Ramirez et al. (2008) and Rodrigues et al. (2019). Body mass was log-transformed before modeling.

Nest type is often associated with reproduction effort and is most likely to affect recruitment (Bennett and Owens 2002). For example, species that make nests in cavities have higher growth rates (Bellier et al. 2018) probably because it is a safer nest against parasitism and predation (Sibly et al. 2012). Nest predation and parasitism are among the most impacting factors of bird populations' decline in fragmented landscapes (Cavitt and Martin 2002). In addition, the lack of suitable nesting habitats in disturbed environments can have a strong effect on the reproductive success of certain bird species, such as those from Picidae and Psittacidae families, which require old or dead trees to build their nests (Sick 1997). We assigned the species to 3 nesting categories: closed, cavity, and open/semi-open. We collapsed open and semi-open nest types due to the low proportion of specialist species with open nests in our data. Nest type information was collected in Del Hoyo et al. (2014).

Habitat loss, fragmentation, and land-use change affect the structure of the habitat by altering differently the provision of food for birds. For example, nectarivorous, frugivorous, and insectivorous species seem to be more sensitive to habitat loss and fragmentation than omnivorous and granivorous (Şekercioğlu et al. 2004, Newbold et al. 2013, Bovo et al. 2018, Chatterjee and Basu 2018). We assigned species to five **main diet** categories according to information available in the literature (Sick 1997, Del Hoyo 2014): omnivorous, insectivorous, frugivorous, nectarivorous, granivorous. Because of the special relationship found for frugivorous and insectivorous species in

land-use change (Newbold et al. 2014), we also assigned the percentage of each of these components in species' diet as trait variables. The percentages of fruits and insects in the species' diet were extracted from the EltonTrait database (Wilman et al. 2014). For the analysis, we excluded the single granivorous forest specialist species in the high-quality matrix region and two omnivorous forest specialist species in the low-quality matrix region because of issues during model fit.

Foraging stratum is of most importance for birds in fragmented landscapes. Ground and understory species are more prone to extinction (Laurance and Gomez 2005), mostly because of higher dispersal limitation and avoidance of open areas (gaps, matrix, and forest edges). We assigned each species to 3 foraging strata categories: ground-understory, midstory-canopy, and all strata. We also used the percentage of use of lower foraging strata (ground and understory) as an alternative operational variable extracted from EltonTrait database (Wilman et al. 2014).

Comparing traits between forest specialists and generalists

All traits were compared between forest specialists and forest generalists using graphical and multivariate analysis to ensure that the selected traits are comparable between groups, i.e., there was not a single trait that could separate specialist from generalist species. The only noticeable difference between generalists and specialists was in the main diet variable, where there were nectarivorous birds only for generalists and insectivorous were more common among specialists.

Below, we present the comparisons of trait values between specialists and generalists. Continuous variables were Z-score scaled and are presented in Figure S1.1, Categorical variables are summarized in Table S1.3.

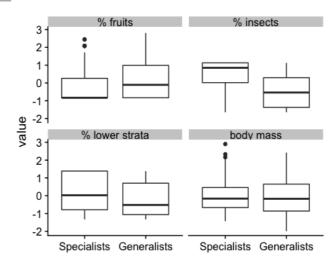


Figure S1.1: Boxplots of values for the traits measured as continuous variables for the specialists and generalists.

 Table S1.3: Percentage of the species in each trait category for specialists and generalists. Numbers

inside brackets are the number of species.

Traits	Specialists	Generalists
Nest type		
Cavities	25% (23)	22% (19)
Closed	27% (25)	21% (18)
Open/semi-open	48% (45)	57% (50)
Main diet		
Frugivorous	17% (16)	23% (20)
Granivorous	2% (2)	6% (5)
Insectivorous	77% (72)	40% (35)
Nectarivorous	0% (0)	15% (13)
Onivorous	3% (3)	16% (14)
Foraging stratum		
All	5% (5)	20% (17)
Ground/Understory	58% (54)	32% (28)
Midstory/Canopy	37% (34)	48% (42)

References on bird traits and environmental change

Below, we list the consulted references of bird traits.

- Barbaro, L. & Halder, I.V. (2009). Linking bird, carabid beetle and butterfly life-history traits to habitat fragmentation in mosaic landscapes. Ecography 32, 321–333.
- Barros, F.M., Martello, F., Peres, C.A., Pizo, M.A. & Ribeiro, M.C. (2019). Matrix type and landscape attributes modulate avian taxonomic and functional spillover across habitat boundaries in the Brazilian Atlantic Forest. *Oikos* **128**, 1600–1612.
- Bellier, E., Kéry, M. & Schaub, M. (2018). Relationships between vital rates and ecological traits in an avian community. Journal of Animal Ecology 87, 1172–1181.
- Boesing, A.L., Nichols, E. & Metzger, J.P. (2018). Biodiversity extinction thresholds are modulated by matrix type. Ecography 41, 1520–1533.
- Bovo, A.A.A., Ferraz, K.M.P.M.B., Magioli, M., Alexandrino, E.R., Hasui, É., Ribeiro, M.C. & Tobias, J.A. (2018). Habitat fragmentation narrows the distribution of avian functional traits associated with seed dispersal in tropical forest. Perspectives in Ecology and Conservation 16, 90–96.

- Bregman, T.P., Lees, A.C., MacGregor, H.E.A., Darski, B., Moura, N.G. de, Aleixo, A., Barlow, J. & Tobias, J.A. (2016). Using avian functional traits to assess the impact of land-cover change on ecosystem processes linked to resilience in tropical forests. *Proc. R. Soc. B* **283**, 20161289.
- Bregman, T.P., Sekercioglu, C.H. & Tobias, J.A. (2014). Global patterns and predictors of bird species responses to forest fragmentation: Implications for ecosystem function and conservation. *Biological Conservation* **169**, 372–383.
- Chatterjee, S. & Basu, P. (2018). Food preferences determine habitat selection at multiple scales: Implication for bird conservation in tropical forests. *Animal Conservation* **21**, 332–342.
- Cleary, D.F.R., Boyle, T.J.B., Setyawati, T., Anggraeni, C.D., Loon, E.E.V. & Menken, S.B.J. (2007). Bird species and traits associated with logged and unlogged forest in borneo. *Ecological Applications* **17**, 1184–1197.
- Cormont, A., Vos, C., Turnhout, C. van, Foppen, R. & Braak, C. ter. (2011). Using life-history traits to explain bird population responses to changing weather variability. *Climate Research* **49**, 59–71.
- De Coster, G., Banks-Leite, C. & Metzger, J.P. (2015). Atlantic forest bird communities provide different but not fewer functions after habitat loss. *Proceedings of the Royal Society B: Biological Sciences* **282**, 20142844.
- Flynn, D.F.B., Gogol-Prokurat, M., Nogeire, T., Molinari, N., Richers, B.T., Lin, B.B., Simpson, N., Mayfield, M.M. & DeClerck, F. (2009). Loss of functional diversity under land use intensification across multiple taxa. *Ecology Letters* **12**, 22–33.
- Kupsch, D., Vendras, E., Ocampo-Ariza, C., Batáry, P., Motombi, F.N., Bobo, K.S. & Waltert, M. (2019). High critical forest habitat thresholds of native bird communities in Afrotropical agroforestry landscapes. *Biological Conservation* **230**, 20–28.
- Luck, G.W., Carter, A. & Smallbone, L. (2013). Changes in Bird Functional Diversity across Multiple Land Uses: Interpretations of Functional Redundancy Depend on Functional Group Identity. PLOS ONE 8, e63671.
- Luck, G.W., Lavorel, S., McIntyre, S. & Lumb, K. (2012). Improving the application of vertebrate trait-based frameworks to the study of ecosystem services. *Journal of Animal Ecology* **81**, 1065–1076.

- Macchi, L., Baumann, M., Bluhm, H., Baker, M., Levers, C., Grau, H.R. & Kuemmerle, T. (2019). Thresholds in forest bird communities along woody vegetation gradients in the South American Dry Chaco. *Journal of Applied Ecology* **56**, 629–639.
- Martin, A.E., Desrochers, A. & Fahrig, L. (2017). Homogenization of dispersal ability across bird species in response to landscape change. *Oikos* **126**, 996–1003.
- Martin, C.A. & Proulx, R. (2016). Habitat geometry, a step toward general bird community assembly rules in mature forests. *Forest Ecology and Management* **361**, 163–169.
- Morante-Filho, J.C., Faria, D., Mariano-Neto, E. & Rhodes, J. (2015). Birds in Anthropogenic Land scapes: The Responses of Ecological Groups to Forest Loss in the Brazilian Atlantic Forest. *PLOS ONE* **10**, e0128923.
- Newbold, T., Scharlemann, J.P.W., Butchart, S.H.M., Şekercioğlu, Ç.H., Alkemade, R., Booth, H. & Purves, D.W. (2013). Ecological traits affect the response of tropical forest bird species to land-use intensity. *Proceedings of the Royal Society B: Biological Sciences* **280**.
- Owens, I.P.F. & Bennett, P.M. (2000). Ecological basis of extinction risk in birds: Habitat loss versus human persecution and introduced predators. *Proceedings of the National Academy of Sciences* **97**, 12144–12148.
- Ripple, W.J., Wolf, C., Newsome, T.M., Hoffmann, M., Wirsing, A.J. & McCauley, D.J. (2017). Extinction risk is most acute for the world's largest and smallest vertebrates. *Proceedings of the National Academy of Sciences* **114**, 10678–10683.
- Smith, Y.C.E., Smith, D.A.E., Seymour, C.L., Thébault, E. & Veen, F.J.F. van. (2015). Response of avian diversity to habitat modification can be predicted from life-history traits and ecological attributes. *Landscape Ecology* **30**, 1225–1239.
- Şekercioğlu, Ç.H., Daily, G.C. & Ehrlich, P.R. (2004). Ecosystem consequences of bird declines. *Pro- ceedings of the National Academy of Sciences* **101**, 18042–18047.
- Tscharntke, T., Sekercioglu, C.H., Dietsch, T.V., Sodhi, N.S., Hoehn, P. & Tylianakis, J.M. (2008). Landscape constraints on functional diversity of birds and insects in tropical agroecosystems. *Ecology* **89**, 944–951.
- Vance, M.D., Fahrig, L. & Flather, C.H. (2003). Effect of reproductive rate on minimum habitat requirements of forest-breeding birds. *Ecology* **84**, 2643–2653.

White, H.J., Montgomery, I.W. & Lennon, J.J. (2018). Contribution of local rarity and climatic suitability to local extinction and colonization varies with species traits. *Journal of Animal Ecology* **87**, 1560–1572.

Appendix 2: Forest cover at local and landscape scale

This appendix is a description of the forest cover variables at both local and landscape scales. We present the baseline models for the selection of the best scale for the local forest cover variable for each dataset. For the local scale, we measured the percentage of forest cover within buffers of 400, 600 and 800 m around each sampling site. For the landscape forest cover, we used the 2 km buffer around the landscape centroid.

1. Relationships among forest cover variables

We calculated Pearson correlation coefficients for forest cover variables in each matrix quality region (Figure S2.1). Also, we plotted the range of local forest cover (400 m) within the landscapes to see how local forest cover varies among landscapes in both regions (Figure S2.2).

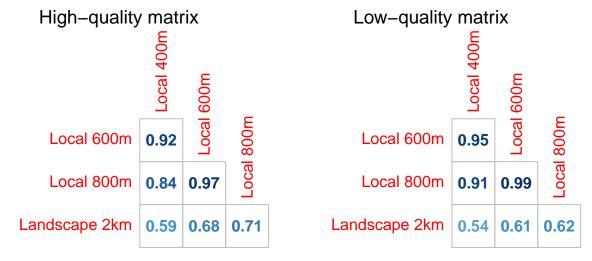


Figure S2.1: Correlations among forest cover variables in the high-quality (left) and low-quality matrix (right) landscapes.

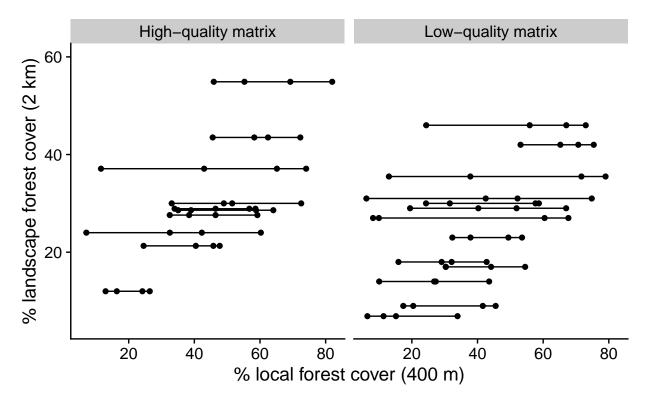


Figure S2.2: Range of local forest cover (buffer 400 m) within each landscape (buffer 2 km) for both landscapes with different matrix quality. Each line represent a landscape and the dots area the local forest cover for each sampling site.

2. Scale of effects for local forest cover

We ran different models with each local forest cover variable and selected the scale of effect using AIC model selection and the R^2 of the models. The models follow the specification presented in the paper (*Modeling* section), except that here we did not include trait variables, i.e. we only modeled the occurrence of species according to forest cover.

We used lme4 package (Bates et al. (2015)) to perform a GLMM with binomial (proportion) distribution. An example of the code for each assemblage is as follows:

In Figure S2.3, we present the occurrence probability predicted for the models with different local forest cover scales for all the assemblages. Predictions were quite similar and decreased with forest cover for the specialists, especially in the low-quality matrix region, and increased or remained flat for the generalists.

Table S2.1: Overal and marginal r-squares and model comparisons with Akaike Information Criterion (AIC) for models with different local forest cover scales as predictor for the specialist and generalists species in both regions with different matrix qualities. For the terms see Table 1 (main text). dAIC is the difference in Akaike Information Criterion to the best model; df are the degrees of freedom.

								AIC	7	
Model	Total	fixed	env.sp	lands.sp	site.sp	lands	site	dAIC	df	
Forest spe	ecialist	species	5							
Low-quality matrix										
$400 \mathrm{m}$	64.3	7.5	42.9	6.2	3.3	1.0	3.4	0.00	9	
$600 \mathrm{m}$	63.9	6.2	43.1	6.2	3.4	0.9	4.1	10.03	9	
$800 \mathrm{m}$	63.9	6.0	43.0	6.3	3.5	1.0	4.1	18.08	9	
High-qua	lity ma	trix								
$400 \mathrm{m}$	56.6	1.3	44.3	7.5	1.8	0.7	1.1	0.00	9	
$600 \mathrm{m}$	56.8	1.0	44.2	7.5	1.8	0.9	1.3	8.57	9	
$800 \mathrm{m}$	56.6	0.6	44.3	7.4	1.9	0.9	1.4	12.40	9	
Forest ger	neralist	specie	s							
Low-qual	ity mat	trix								
$400\mathrm{m}$	46.6	0.1	39.7	3.5	2.5	0.0	0.9	0.00	9	
$600 \mathrm{m}$	46.6	0.0	39.1	3.6	3.0	0.0	0.9	20.27	9	
$800 \mathrm{m}$	46.5	0.0	39.0	3.6	3.0	0.0	0.9	22.54	9	
High-qua	lity ma	trix								
$400 \mathrm{m}$	44.1	0.0	37.0	2.2	3.6	0.6	0.7	11.24	9	
$600 \mathrm{m}$	44.3	0.0	37.5	2.0	3.4	0.6	0.7	2.45	9	
800m	44.3	0.1	37.5	2.0	3.5	0.6	0.7	0.00	9	

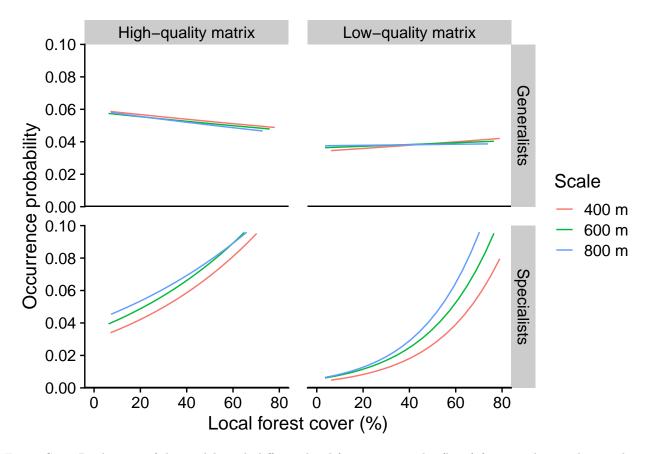


Figure S2.3: Predictions of the models with different local forest cover scales (lines) for specialists and generalists in both regions.

Residual correlations among species

We evaluated the residuals by Kendall correlations among species and among sites for the 400 m models using the predictions for site:sp random effects (Observation Level Random Effect), following the code provided by Miller, Damschen & Ives (2018). Codes for the species names are presented in the dataset available.

Below we show models residual correlation plots for the specialists in the high-quality matrix landscape. All the other assemblages presented similar results.

Range of species correlations: -0.41, 0.46.

Range of sites correlations: -0.25, 0.31.

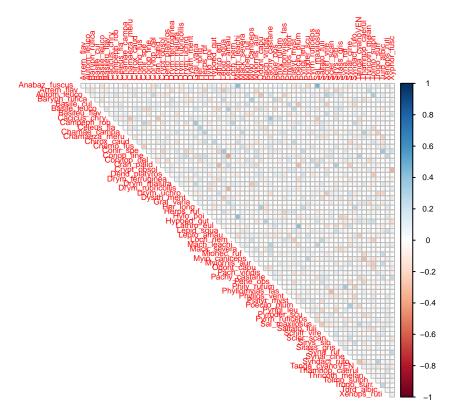


Figure S2.4: Species residual Kendall correlations for the specialist species in the coffee matrix region.

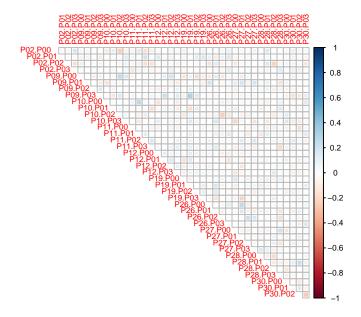


Figure S2.5: Sites residual Kendall correlations for the specialist species in the coffee matrix region.

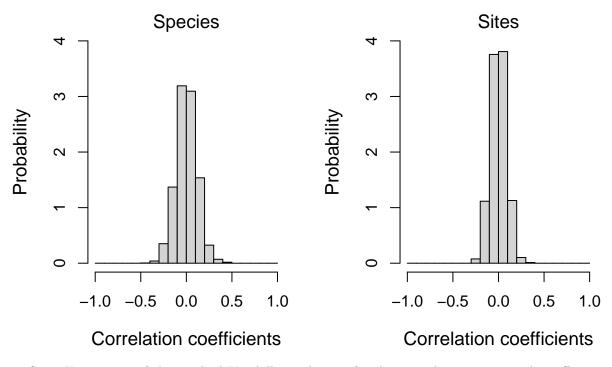


Figure S2.6: Histograms of the residual Kendall correlations for the specialists species in the coffee matrix region.

3. Including landscape forest cover

After selecting the local forest cover of 400 m radius buffer around each site for all datasets, we included the landscape forest cover (2 km radius buffer around the centroid of the landscape) in the model.

The R syntax example of this model area as follows:

Table S2.2: Variance Inflation Factor index for the variables of local forest cover and landscape forest cover.

	Local	Landscape
Specialists	3	
Coffee	1.26	1.26
Pasture	1.04	1.04
Generalist	\mathbf{s}	
Coffe	1.13	1.13
Pasture	1.15	1.15

```
(local.400 + landscape.2k |sp) +
  (1|landscape:sp) + (1|site:sp) +
  (1|landscape) + (1|site),
family=binomial, data=high.spe)
```

Before analysing results, we evaluated possible colinearity between local and landscape forest cover using the Variance Inflation Factor with the code provided by John Lefcheck (https://jonlefcheck.net/2012/12/28/dealing-with-multicollinearity-using-variance-inflation-factors/). With VIF we found no evidence of collinearity between the forest cover scales (Table S2.2).

Predictions of the models are present in Figure S2.7. It is important to notice the differences in 20 and 40% landscape forest cover predictions for the specialists in the low-quality matrix.

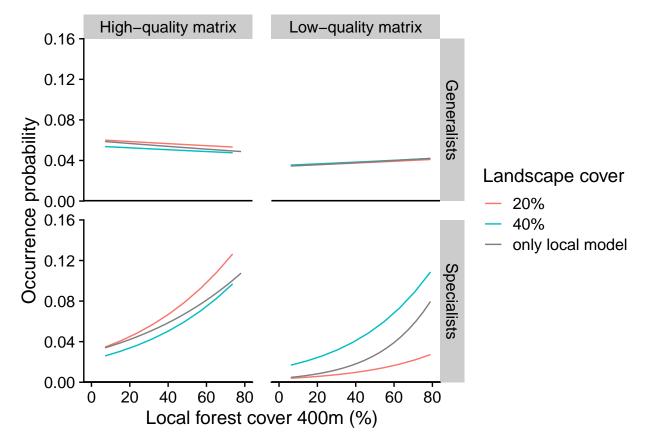


Figure S2.7: Predictions of the models without (gray lines) and with landscape forest cover scales (20 percent cover in red and 40 percent cover in blue lines) for specialists and generalists in both regions.

References

- Bates, D., Mächler, M., Bolker, B. & Walker, S. (2015). Fitting Linear Mixed-Effects Models Using lme4. *Journal of Statistical Software* **67**, 1–48.
- Miller, J.E.D., Damschen, E.I. & Ives, A.R. (2018). Functional traits and community composition: A comparison among community-weighted means, weighted correlations, and multilevel models. *Methods in Ecology and Evolution* **0**.

Appendix 3: Additional models results

1. Models with traits

Specification of the models: We used lme4 package to perform a GLMM with binomial (proportion) distribution. An example of the code for each dataset are as follows:

```
mhigh.spe <- glmer(cbind(occor, n.visit-occor) ~</pre>
                    forest_site400*lbody_size +
                    forest site400*nest +
                    forest site400*diet +
                    forest_site400*lower_stratum +
                    forest land*lbody size +
                    forest_land*nest +
                    forest_land*diet +
                    forest land*lower stratum +
                    (forest site400 + forest land|sp) +
                    (1|landscape:sp) + (1|site:sp) +
                    (lbody size + nest + diet + lower stratum|landscape) +
                    (lbody size + nest + diet + lower stratum|site),
                    family=binomial, data=high.spe,
                    nAGQ = 1, control = glmerControl(optimizer = "bobyqa",
                                     optCtrl = list(maxfun = 500000)))
```

We ran separate models for each assemblage and trait. Afterwards, we ran one model with the combination of the traits body mass, diet, nest type and % of lower strata use. Table S3.1 shows the marginal \mathbb{R}^2 of all models terms.

2. Models coefficients

Tables S3.2, S3.3, S3.4, and S3.5 show the coefficients for each model.

3. Models diagnostic

Variance Inflation Factor of the model parameters for each dataset in Table S3.6.

Table S3.1: Overall and marginal R-squared of trait models in each dataset. For the marginal R-squared terms see Table 2 (main text).

Model	Total	trait*env	env sp	lands:sp	site:sp	trait lands	trait site
Specialists							
High quality							
Combined	57.6	10.0	36.8	6.7	0.7	1.3	2.2
body mass	56.8	7.7	38.9	6.7	1.6	0.8	1.1
nest type	56.8	4.6	41.9	6.8	1.5	0.6	1.4
main diet	57.4	4.5	42.5	6.5	1.6	1.0	1.3
% frugivory	56.8	4.4	42.4	6.8	1.6	0.5	1.1
% insetivory	56.7	3.8	42.9	6.8	1.6	0.5	1.1
% lower strata	56.7	1.5	45.0	6.8	1.4	0.7	1.4
foraging stratum	56.7	2.0	44.5	6.5	1.5	0.8	1.3
Low quality							
Combined	71.9	22.2	39.9	3.8	1.0	0.9	4.0
body mass	70.5	15.8	45.6	3.8	2.4	0.1	2.7
nest type	71.0	17.3	44.5	3.9	2.2	0.0	3.1
main diet	71.4	18.4	43.8	3.7	2.5	0.4	2.5
% frugivory	71.3	17.2	45.1	3.8	2.6	0.0	2.6
% insetivory	71.3	18.4	44.0	3.8	2.6	0.1	2.5
% lower strata	71.1	15.5	46.2	3.8	1.9	0.4	3.2
foraging stratum	70.7	17.0	44.5	3.9	2.3	0.3	2.6
Generalists							
High quality							
Combined	46.2	7.4	31.9	1.3	0.9	1.6	3.1
body mass	44.7	0.2	37.8	1.6	2.9	0.6	1.5
nest type	44.6	0.6	37.4	1.7	3.1	0.6	1.1
main diet	45.8	5.2	33.7	1.5	1.7	1.1	2.7
% frugivory	44.5	0.5	37.4	1.5	3.6	0.7	0.8
% insetivory	44.4	3.2	34.8	1.2	3.7	0.9	0.7
% lower strata	44.5	1.3	36.7	1.5	3.6	0.7	0.7
foraging stratum	44.4	1.9	35.8	1.7	3.2	0.7	1.0
Low quality							
Combined	47.4	7.5	33.3	2.3	0.9	1.0	2.4
body mass	47.1	0.8	39.7	2.7	2.2	0.4	1.2
nest type	46.7	1.0	39.3	3.1	2.1	0.0	1.1
main diet	47.1	5.3	35.2	2.8	1.9	0.5	1.5
% frugivory	47.0	0.4	40.0	2.9	2.4	0.3	0.9
% insetivory	46.7	1.6	38.7	3.0	2.4	0.2	0.9
% lower strata	47.0	0.9	39.5	3.1	2.1	0.1	1.3
foraging stratum	47.0	1.2	39.2	2.9	2.4	0.1	1.1

Table S3.2: Fixed effects oefficients for the model of specialists in high-quality matrix landscapes.

effect	term	estimate	std.error	statistic	p.value
fixed	(Intercept)	-2.72	0.98	-2.78	0.01
fixed	forest_site400	0.83	0.33	2.49	0.01
fixed	lbody_size	-0.62	0.28	-2.22	0.03
fixed	nestclosed	0.32	0.67	0.48	0.63
fixed	nestopen_semi	-0.28	0.56	-0.50	0.62
fixed	dietinsectivorous	0.07	0.83	0.09	0.93
fixed	dietonivorous	-1.00	1.25	-0.81	0.42
fixed	lower_stratum	0.18	0.23	0.78	0.44
fixed	forest_land	-0.36	0.41	-0.87	0.38
fixed	$forest_site 400: lbody_size$	-0.03	0.09	-0.30	0.76
fixed	$forest_site400:nestclosed$	-0.38	0.19	-1.98	0.05
fixed	$forest_site 400:nestopen_semi$	-0.24	0.17	-1.36	0.17
fixed	$forest_site 400: diet in sectivo rous$	-0.28	0.29	-0.97	0.33
fixed	$forest_site 400: dietonivorous$	-0.35	0.44	-0.80	0.42
fixed	$forest_site 400: lower_stratum$	0.01	0.07	0.10	0.92
fixed	lbody_size:forest_land	0.00	0.10	0.04	0.97
fixed	$nestclosed:forest_land$	0.30	0.22	1.35	0.18
fixed	$nestopen_semi:forest_land$	0.31	0.19	1.61	0.11
fixed	$dietinsectivorous:forest_land$	-0.04	0.33	-0.11	0.91
fixed	${\it dietonivorous:} forest_land$	0.45	0.47	0.95	0.34
fixed	$lower_stratum:forest_land$	0.00	0.08	-0.03	0.98

Table S3.3: Fixed effects oefficients for the model of specialists in low-quality matrix landscapes.

effect	term	estimate	std.error	statistic	p.value
fixed	(Intercept)	-5.46	1.03	-5.33	0.00
fixed	forest_site400	1.06	0.36	2.94	0.00
fixed	lbody_size	-0.21	0.32	-0.67	0.50
fixed	nestclosed	0.70	0.78	0.90	0.37
fixed	nestopen_semi	-0.12	0.69	-0.18	0.86
fixed	dietgranivorous	0.13	1.82	0.07	0.95
fixed	dietinsectivorous	1.27	0.82	1.54	0.12
fixed	lower_stratum	0.36	0.28	1.28	0.20
fixed	forest_land	1.42	0.44	3.22	0.00
fixed	$forest_site 400: lbody_size$	0.05	0.09	0.51	0.61
fixed	forest_site400:nestclosed	-0.39	0.22	-1.78	0.07
fixed	$forest_site 400:nestopen_semi$	-0.06	0.20	-0.29	0.77
fixed	forest_site400:dietgranivorous	-0.41	0.61	-0.67	0.50
fixed	$forest_site 400: diet in sectivo rous$	-0.40	0.28	-1.45	0.15
fixed	$forest_site 400: lower_stratum$	0.02	0.09	0.18	0.86
fixed	lbody_size:forest_land	0.01	0.12	0.04	0.96
fixed	$nestclosed:forest_land$	-0.59	0.28	-2.10	0.04
fixed	nestopen_semi:forest_land	-0.16	0.27	-0.62	0.54
fixed	dietgranivorous:forest_land	-0.03	0.73	-0.04	0.96
fixed	$diet in sectivo rous: for est_land$	-0.38	0.34	-1.10	0.27
fixed	lower_stratum:forest_land	-0.15	0.12	-1.24	0.22

Table S3.4: Fixed effects oefficients for the model of generalists in high-quality matrix landscapes.

effect	term	estimate	std.error	statistic	p.value
fixed	(Intercept)	-3.45	0.55	-6.29	0.00
fixed	$forest_site400$	0.37	0.19	2.01	0.04
fixed	lbody_size	0.08	0.27	0.29	0.77
fixed	nestclosed	0.20	0.61	0.32	0.75
fixed	nestopen_semi	0.20	0.53	0.38	0.71
fixed	dietgranivorous	-0.91	0.86	-1.06	0.29
fixed	dietinsectivorous	1.00	0.50	2.00	0.05
fixed	dietnectarivorous	-0.12	0.75	-0.16	0.88
fixed	dietonivorous	0.14	0.67	0.21	0.83
fixed	lower_stratum	0.36	0.17	2.06	0.04
fixed	forest_land	-0.11	0.22	-0.50	0.62
fixed	$forest_site 400: lbody_size$	0.05	0.09	0.61	0.54
fixed	$forest_site 400:nest closed$	-0.43	0.20	-2.12	0.03
fixed	forest_site400:nestopen_semi	-0.22	0.17	-1.28	0.20
fixed	$forest_site 400: diet granivorous$	-0.61	0.35	-1.75	0.08
fixed	$forest_site 400: diet in sectivo rous$	-0.13	0.18	-0.75	0.46
fixed	$forest_site 400: diet nectari vorous$	-0.16	0.27	-0.61	0.54
fixed	forest_site400:dietonivorous	-0.50	0.25	-2.04	0.04
fixed	$forest_site 400: lower_stratum$	0.04	0.05	0.71	0.48
fixed	$lbody_size:forest_land$	-0.01	0.10	-0.09	0.93
fixed	nestclosed:forest_land	0.18	0.21	0.88	0.38
fixed	nestopen_semi:forest_land	0.07	0.17	0.38	0.71
fixed	dietgranivorous:forest_land	0.68	0.36	1.87	0.06
fixed	dietinsectivorous:forest_land	-0.13	0.19	-0.65	0.51
fixed	${\it dietnectarivorous:} {\it forest_land}$	-0.11	0.28	-0.40	0.69
fixed	dietonivorous:forest_land	0.06	0.27	0.22	0.83
fixed	$lower_stratum: forest_land$	-0.03	0.06	-0.58	0.56

Table S3.5: Fixed effects oefficients for the model of generalists in low-quality matrix landscapes.

effect	term	estimate	std.error	statistic	p.value
fixed	(Intercept)	-4.23	0.49	-8.66	0.00
fixed	$forest_site400$	0.26	0.20	1.32	0.19
fixed	lbody_size	0.22	0.26	0.83	0.41
fixed	nestclosed	0.79	0.59	1.33	0.18
fixed	nestopen_semi	0.55	0.50	1.10	0.27
fixed	dietgranivorous	0.87	1.12	0.78	0.44
fixed	dietinsectivorous	1.06	0.48	2.20	0.03
fixed	dietnectarivorous	-0.19	0.76	-0.25	0.80
fixed	dietonivorous	0.39	0.64	0.62	0.54
fixed	lower_stratum	0.24	0.18	1.33	0.18
fixed	forest_land	0.04	0.18	0.25	0.81
fixed	forest_site400:lbody_size	0.03	0.10	0.32	0.75
fixed	$forest_site 400:nest closed$	-0.52	0.21	-2.45	0.01
fixed	forest_site400:nestopen_semi	-0.26	0.17	-1.49	0.14
fixed	forest_site400:dietgranivorous	0.55	0.39	1.41	0.16
fixed	forest_site400:dietinsectivorous	0.02	0.19	0.12	0.90
fixed	forest_site400:dietnectarivorous	0.09	0.30	0.30	0.77
fixed	forest_site400:dietonivorous	-0.05	0.23	-0.21	0.84
fixed	forest_site400:lower_stratum	-0.10	0.06	-1.62	0.11
fixed	$lbody_size:forest_land$	0.00	0.09	-0.05	0.96
fixed	nestclosed:forest_land	0.27	0.18	1.46	0.14
fixed	nestopen_semi:forest_land	0.20	0.15	1.37	0.17
fixed	dietgranivorous:forest_land	-0.48	0.34	-1.42	0.15
fixed	dietinsectivorous:forest_land	-0.21	0.17	-1.23	0.22
fixed	${\it dietnectarivorous:} {\it forest_land}$	-0.26	0.28	-0.94	0.35
fixed	dietonivorous:forest_land	-0.25	0.20	-1.27	0.21
fixed	$lower_stratum:forest_land$	0.07	0.06	1.17	0.24

Table S3.6: Variance Inflation Factor index for combined traits models in each dataset.

	Spec	ialists	$Gen \epsilon$	eralists
parameter	Coffee	lowture	Coffee	lowture
forest.local	1.24	1.04	1.17	1.17
$body_mass$	1.23	1.08	1.15	1.15
$\operatorname{nest_closed}$	1.78	1.43	2.37	2.37
nest_open_semi	2.15	1.99	1.96	1.96
$\operatorname{diet_insectivorous}$	1.83	1.94	2.16	2.16
diet_onivorous	1.83	1.49	2.04	2.04
$lower_strata$	1.37	NA	2.22	2.22
diet_granivorous	1.18	1.19	1.15	1.15
forest.landscape	NA	1.16	1.20	1.20
diet_nectarivorous	NA	NA	2.46	2.46

Example of the residual diagnostic of the model with the combined traits (main diet, body mass, nest type and % of lower strata use) for the forest specialists in high-quality matrix landscapes. The models' diagnostics for the other assemblages were all similar and can be checked in this Rmd file.

Residual correlations among species and sites

Below we present the Kendall correlations for the residuals among species and sites for the models using the predictions for site:sp random effect (Observation Level Random Effect). For the residual correlations we followed the code provided by Miller, Damschen & Ives (2018).

Range of species correlations: -0.4, 0.43. Range of sites correlations: -0.3, 0.27.

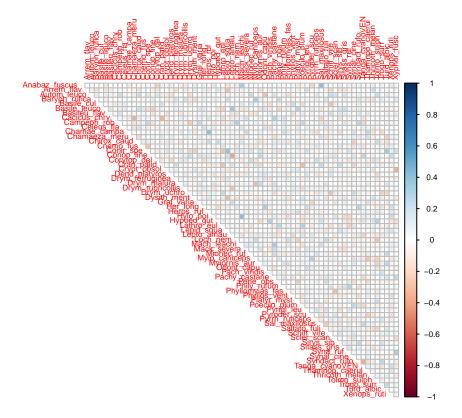


Figure S3.1: Species residual Kendall correlations for the specialist species in high-quality matrix landscapes.

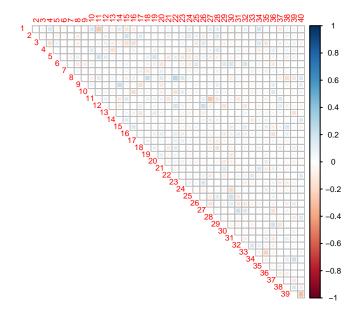


Figure S3.2: Sites residual Kendall correlations for the specialist species in high-quality matrixlandscapes.

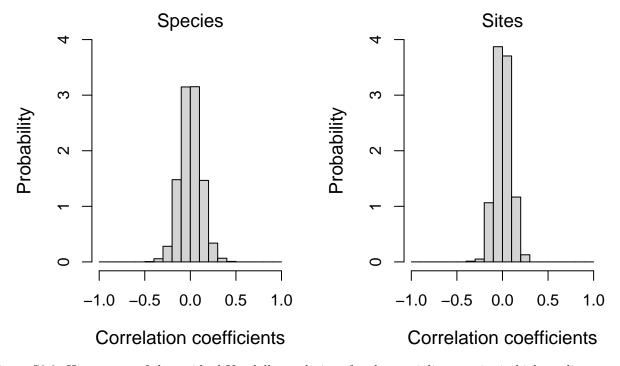
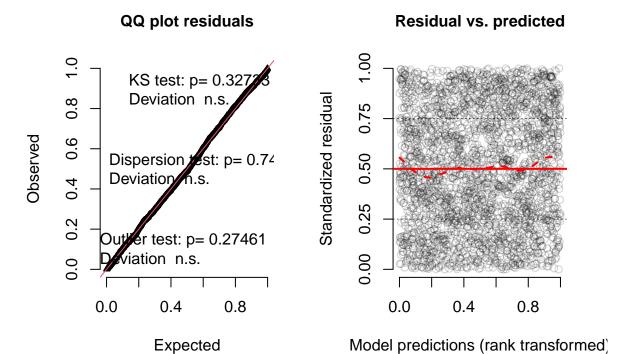


Figure S3.3: Histograms of the residual Kendall correlations for the specialists species in high-quality matrix landscapes.

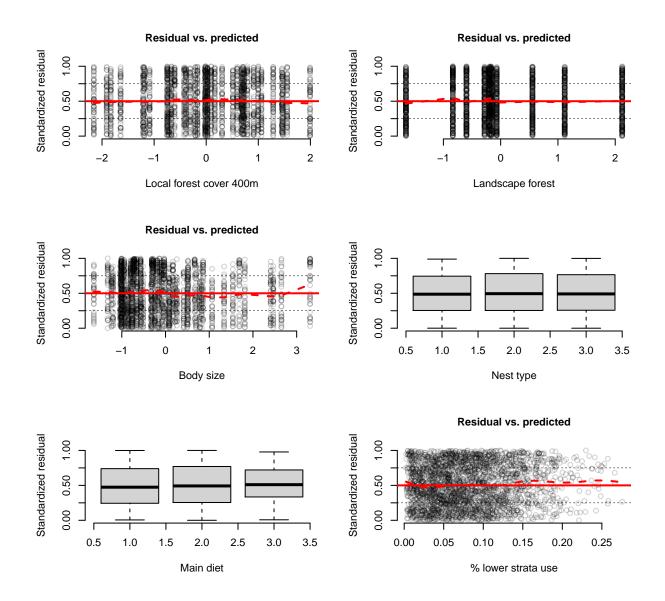
Residual diagnostic

We used DHARMa package (Hartig (2018)) for the diagnostic of quantile residuals.

DHARMa residual diagnostics



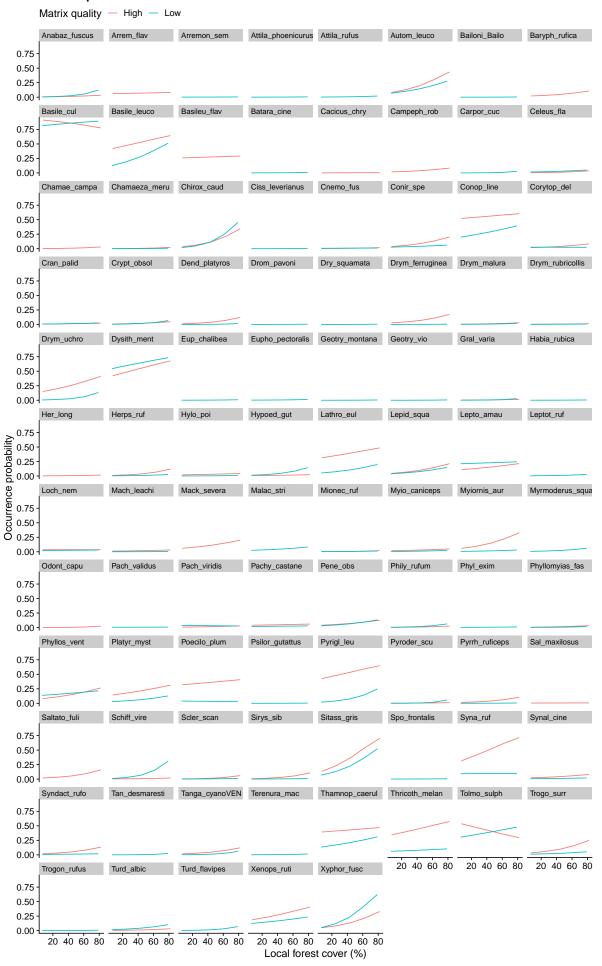
Residuals against predictors:



Predictions for each species local forest cover

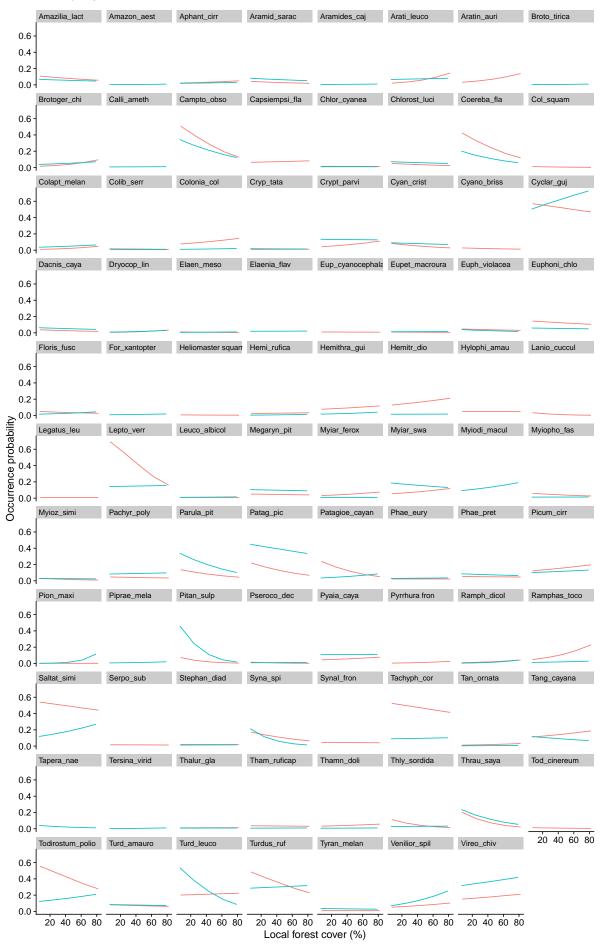
Landscape forest cover was fixed in 30%.

Forest specialists



Forest generalists

Matrix quality - High - Low



References

Hartig, F. (2018). DHARMa: Residual Diagnostics for Hierarchical (Multi-Level / Mixed) Regression Models.

Miller, J.E.D., Damschen, E.I. & Ives, A.R. (2018). Functional traits and community composition: A comparison among community-weighted means, weighted correlations, and multilevel models. *Methods in Ecology and Evolution* **0**.