The Combined Status of Gopher (*Sebastes carnatus*) and Black-and-Yellow Rockfishes (*Sebastes chrysomelas*) in U.S. Waters Off California in 2019

Gopher rockfish (left) and black-and-yellow rockfish (right). Photos by Steve Lonhart.

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The Combined Status of Gopher (*Sebastes carnatus*) and Black-and-Yellow Rockfishes (*Sebastes chrysomelas*) in U.S. Waters Off California in 2019

Contents

Executive Summary

executive-summary

Stock

stock

This assessment reports the status of the GBYR (*Sebastes carnatus/Sebastes chrysomelas*) resource in U.S. waters off the coast of . . . using data through 2018.

Catches

catches

Information on historical landings of GBYR are available back to xxxx. . . (Table a). Commercial landings were small during the years of World War II, ranging between 4 to 28 metric tons (mt) per year.

(Figures a-b)
(Figure c)

Since 2000, annual total landings of GBYR have ranged between 70-169 mt, with landings in 2018 totaling 92 mt.

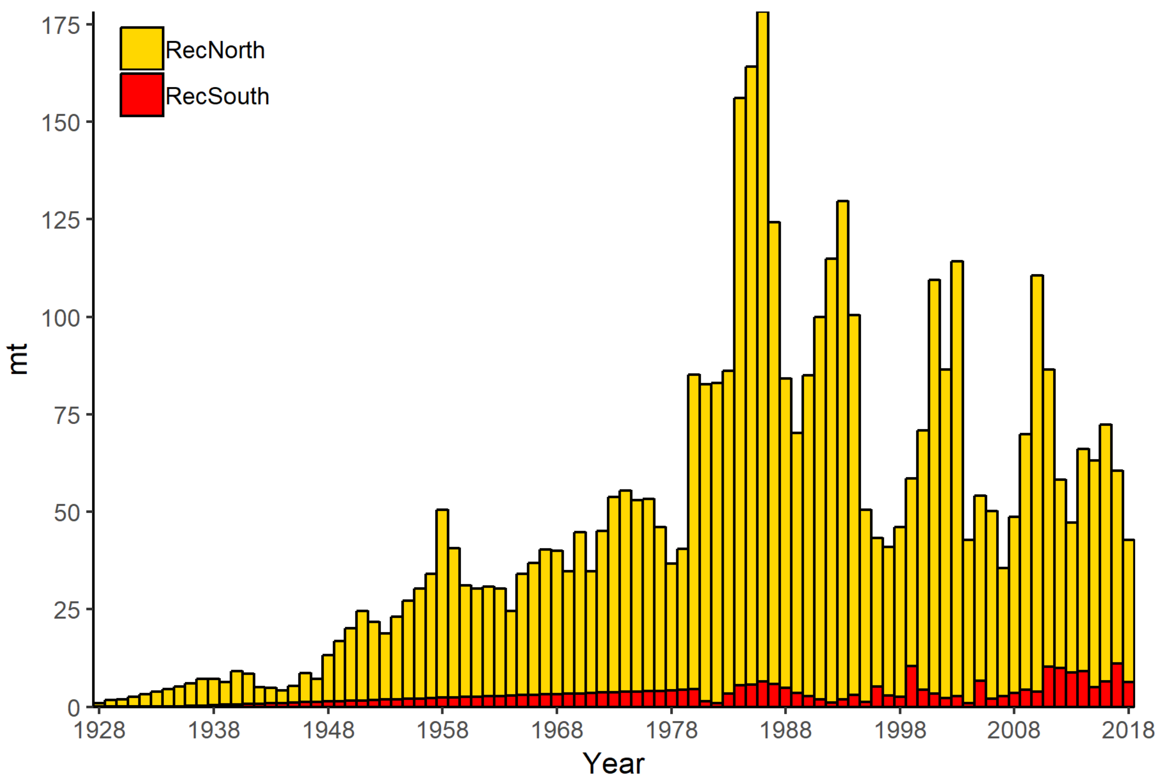

Figure a: Catch history of GBYR for the recreational fleet. fig:Exec_catch1

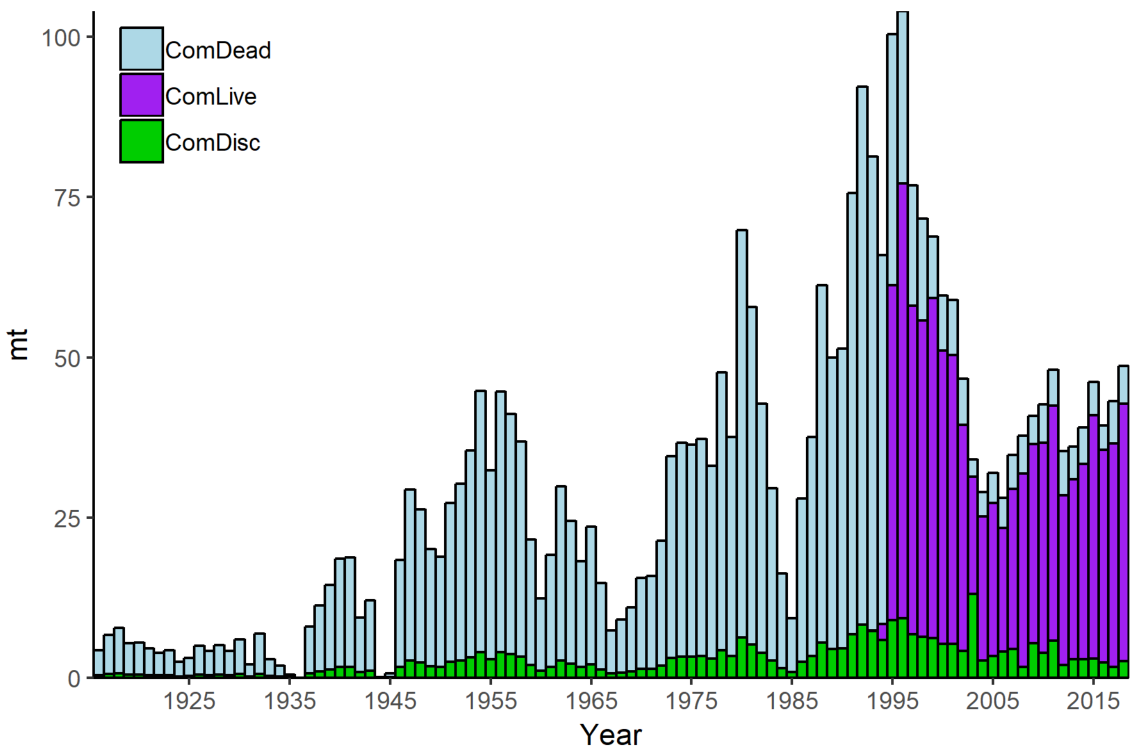

Figure b: Catch history of GBYR for the commercial fleet by dead and live landings, and discards. Catches in 1936 and 1946 were minimal. fig:Exec_catch2

Table a: Recent GBYR landings (mt) by fleet.

tab:Exec_catch

Year	Commercial Retained	Commercial Discard	Recreational North	Recreational South	Total
2009	35.62	5.38	65.64	4.30	110.93
2010	38.83	3.92	106.76	3.90	153.41
2011	42.39	5.72	76.16	10.24	134.52
2012	33.55	1.93	48.25	9.89	93.62
2013	33.45	2.85	38.43	8.86	83.59
2014	36.40	2.85	56.96	9.06	105.27
2015	43.25	2.93	58.09	5.00	109.27
2016	36.96	2.42	65.72	6.57	111.67
2017	42.04	1.65	49.36	11.15	104.19
2018	47.00	2.54	36.48	6.30	92.32

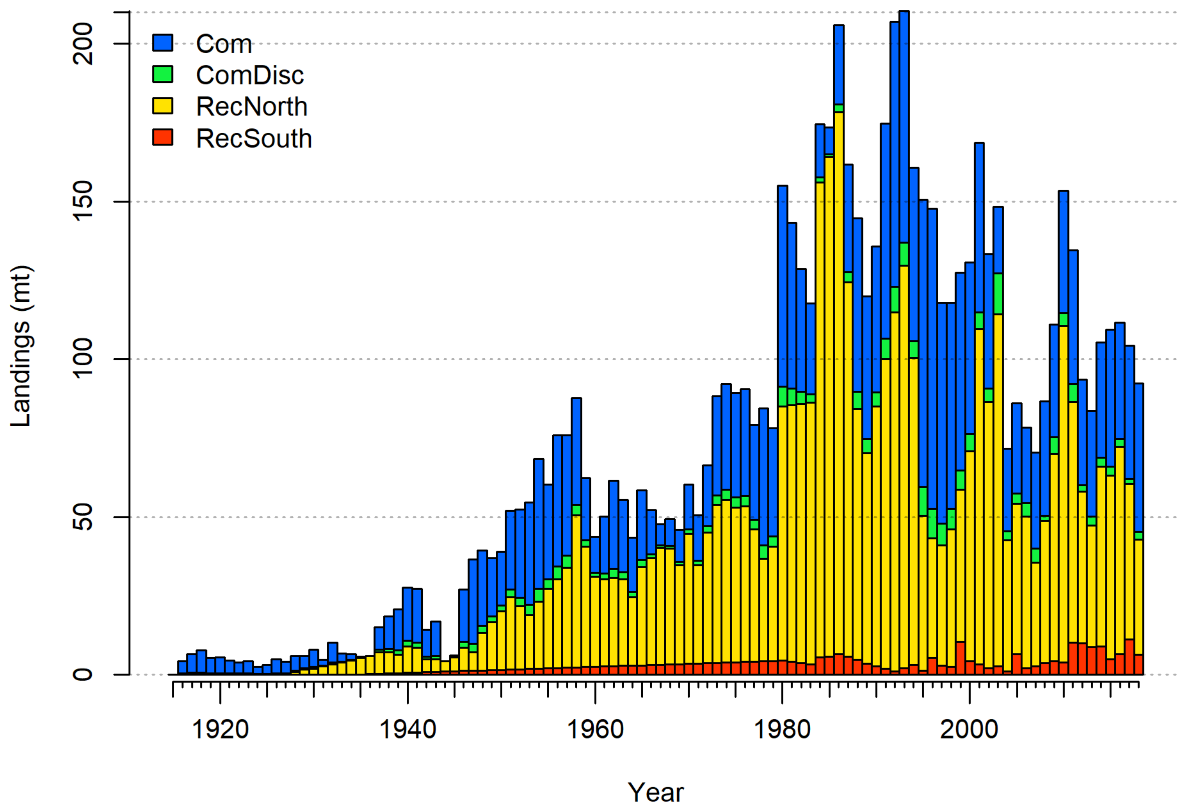

Figure c: Catch history of GBYR in the model. fig:r4ss_catches

Data and Assessment

data-and-assessment

This a new full assessment for GBYR, which was last assessed in ... using Stock Synthesis Version xx. This assessment uses the newest version of Stock Synthesis (3.30.xx). The model begins in 1916, and assumes the stock was at an unfished equilibrium that year.

(Figure d).

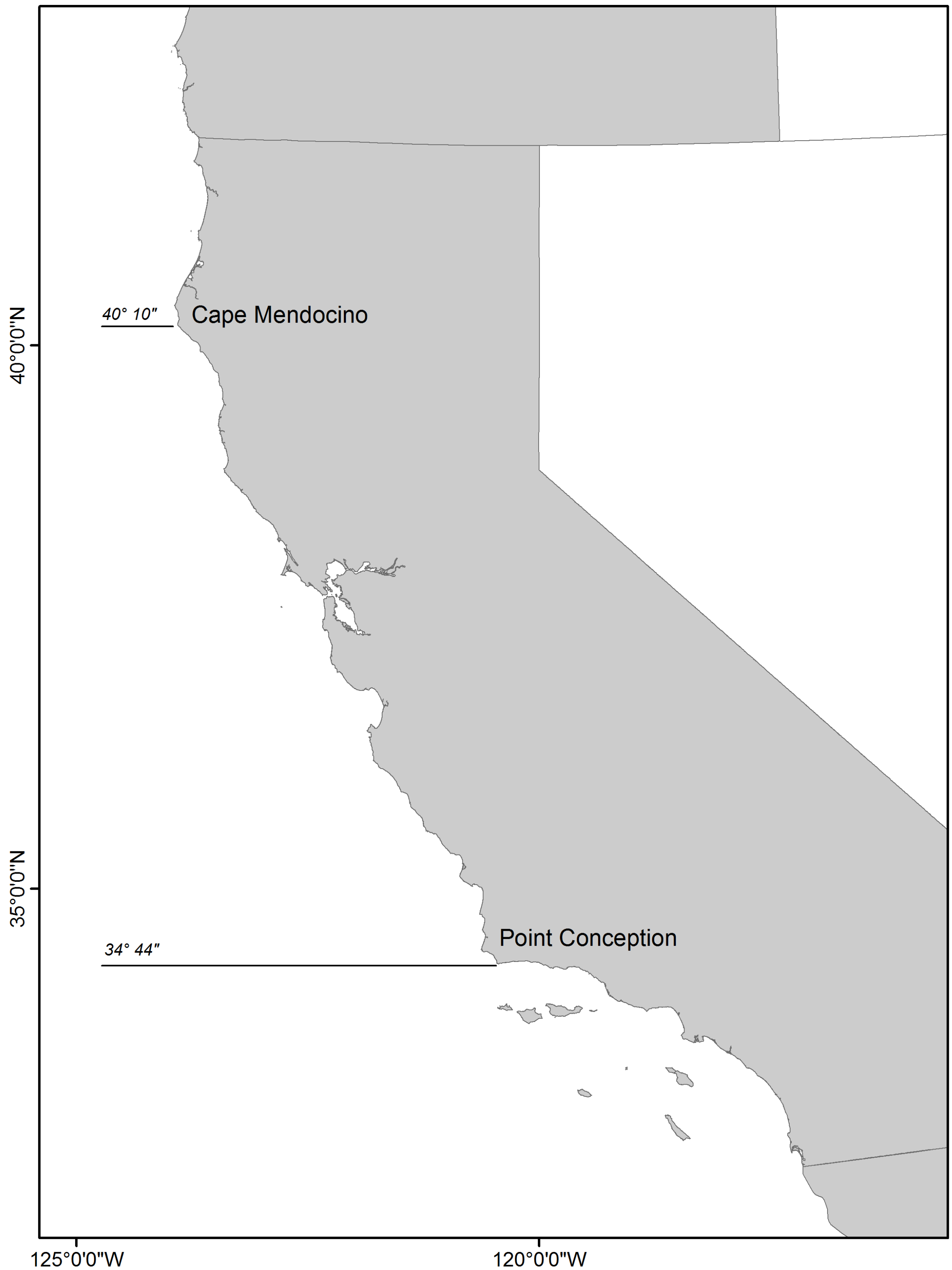

Figure d: Map depicting the core distribution of gopher and black-and-yellow rockfishes. The stock assessment is bounded at Cape Mendocino in the north to the U.S./Mexico border in the south. fig:assess_region_map

Stock Biomass

stock-biomass

(Figure e and Table b).

The 2018 estimated spawning biomass relative to unfished equilibrium spawning biomass is above the target of 40% of unfished spawning biomass at 4 520% (95% asymptotic interval: ± 3 000% - 6 050%) (Figure f). Approximate confidence intervals based on the asymptotic variance estimates show that the uncertainty in the estimated spawning biomass is high.

Table b: Recent trend in beginning of the year spawning output and depletion for the model for GBYR.

tab:SpawningDeplete_mod1

Year	Spawning Output (million eggs)	~ 95% confidence interval	Estimated depletion	~ 95% confidence interval
2010	834	520 - 1149	63.32	45.47 - 81.16
2011	764	469 - 1060	58.03	41.45 - 74.61
2012	707	429 - 985	53.69	38.2 - 69.18
2013	676	410 - 942	51.33	36.74 - 65.91
2014	654	397 - 911	49.64	35.74 - 63.55
2015	625	374 - 877	47.48	33.95 - 61.01
2016	602	352 - 852	45.71	32.24 - 59.18
2017	585	332 - 838	44.41	30.69 - 58.12
2018	580	320 - 841	44.06	29.81 - 58.32
2019	596	320 - 872	45.25	29.99 - 60.5

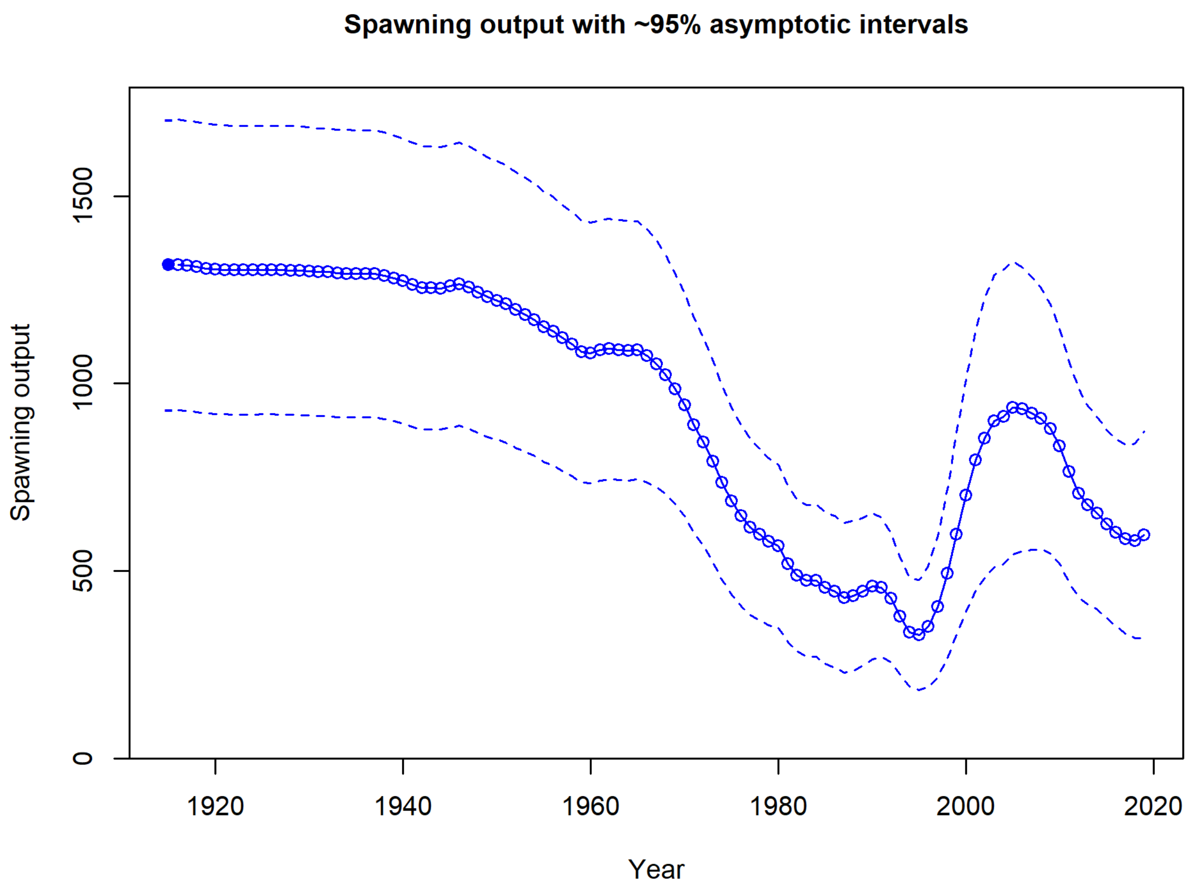

Figure e: Time series of spawning biomass trajectory (circles and line: median; light broken lines: 95% credibility intervals) for the base case assessment model. fig:Spawnbio_all

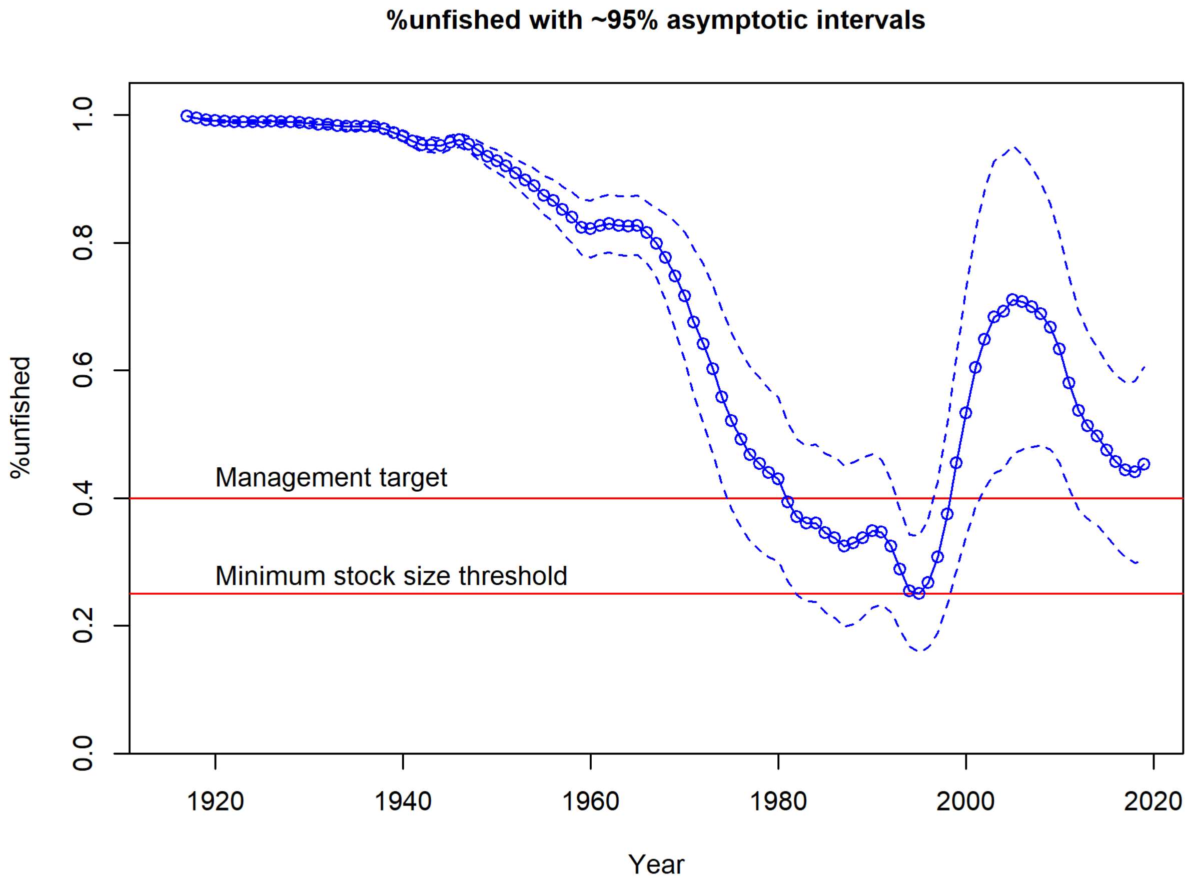

Figure f: Estimated percent depletion with approximate 95% asymptotic confidence intervals (dashed lines) for the base case assessment model. fig:RelDeplete_all

Recruitment

recruitment

Recruitment deviations were estimated from xxxx-xxxx (Figure g and Table c).

Table c: Recent recruitment for the GBYR assessment.

tab:Recruit_mod1

Year	Estimated Recruitment (1,000s)	~ 95% confidence interval
2010	3680	1452 - 9327
2011	3319	1277 - 8625
2012	3393	1279 - 9001
2013	4194	1580 - 11131
2014	6319	2385 - 16738
2015	8445	3154 - 22610
2016	6941	2564 - 18788
2017	5137	1845 - 14301
2018	4177	1469 - 11875
2019	4276	1486 - 12305

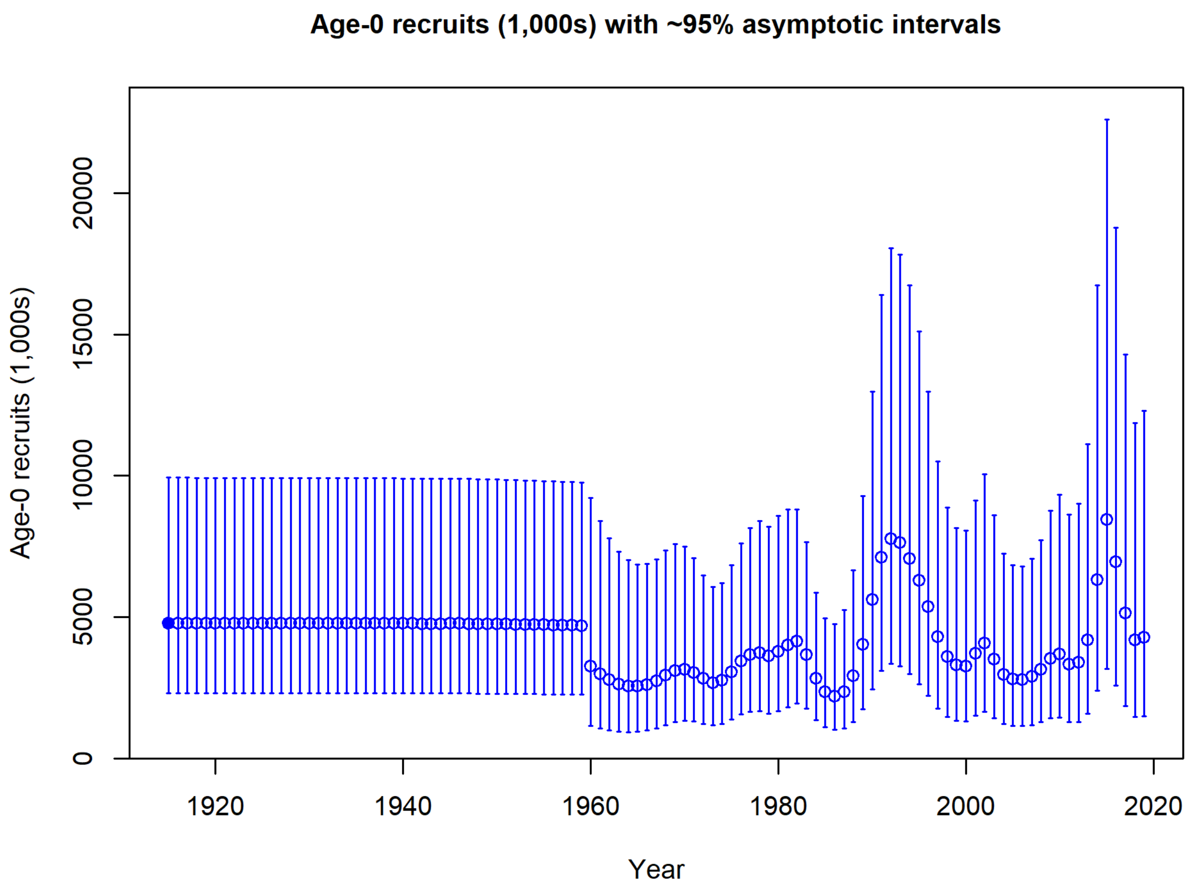

Figure g: Time series of estimated GBYR recruitments for the base-case model with 95% confidence or credibility intervals. fig:Recruits_all

Exploitation status

exploitation-status

Harvest rates estimated by the base model ….. management target levels (Table d and Figure h).

Table d: Recent trend in spawning potential ratio and exploitation for GBYR in the model. Fishing intensity is (1-SPR) divided by 50% (the SPR target) and exploitation is F divided by F_{SPR}.

tab:SPR_Exploit_mod1

Year	Fishing intensity	~ 95% confidence interval	Exploitation rate	~ 95% confidence interval
2009	0.61	0.38 - 0.83	0.08	0.05 - 0.1
2010	0.74	0.5 - 0.99	0.11	0.07 - 0.15
2011	0.74	0.49 - 0.99	0.10	0.06 - 0.14
2012	0.63	0.4 - 0.87	0.08	0.05 - 0.1
2013	0.61	0.38 - 0.84	0.07	0.04 - 0.1
2014	0.71	0.46 - 0.96	0.09	0.05 - 0.12
2015	0.74	0.49 - 0.99	0.09	0.05 - 0.13
2016	0.77	0.51 - 1.04	0.09	0.05 - 0.13
2017	0.77	0.5 - 1.04	0.08	0.04 - 0.12
2018	0.73	0.46 - 1	0.07	0.03 - 0.1

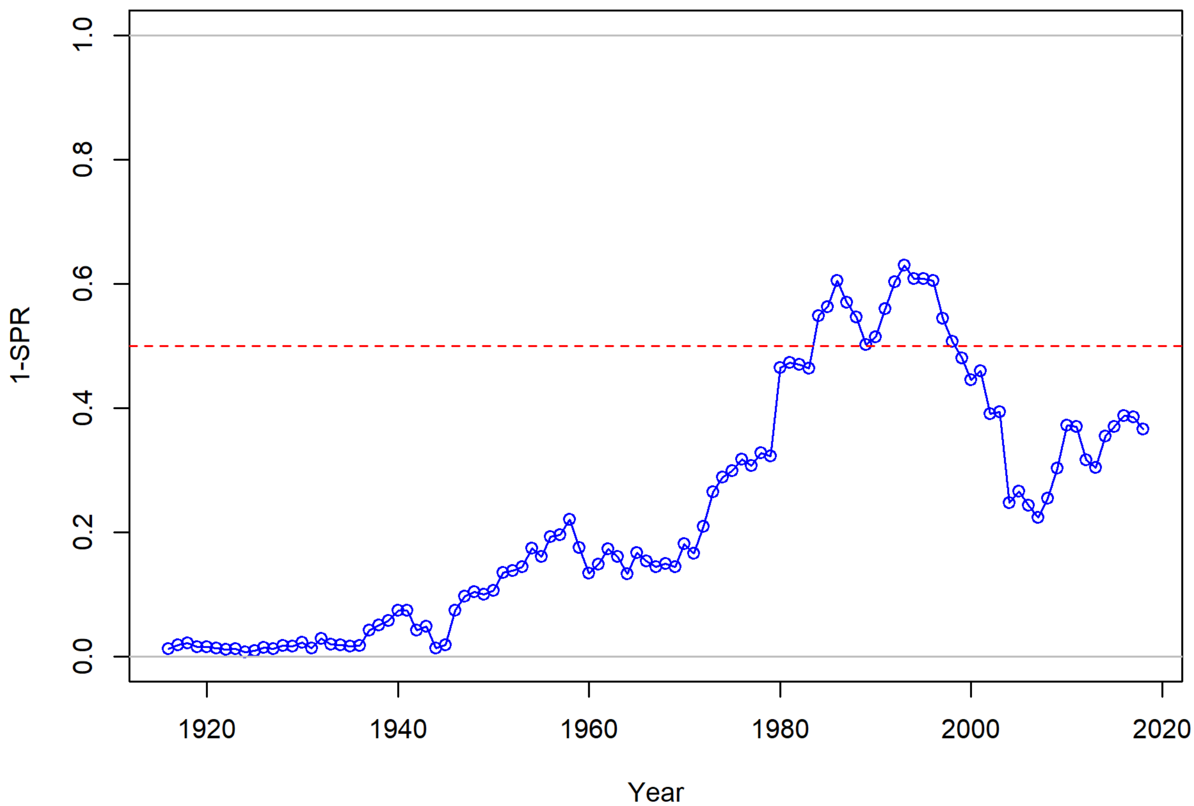

Figure h: Estimated spawning potential ratio (SPR) for the base-case model. One minus SPR is plotted so that higher exploitation rates occur on the upper portion of the y-axis. The management target is plotted as a red horizontal line and values above this reflect harvests in excess of the overfishing proxy based on the $SPR_{50\%}$ harvest rate. The last year in the time series is 2018. fig:SPR_all

Ecosystem Considerations

ecosystem-considerations

In this assessment, ecosystem considerations were not explicitly included in the analysis. This is primarily due to a lack of relevant data and results of analyses (conducted elsewhere) that could contribute ecosystem-related quantitative information for the assessment.

Reference Points

reference-points

This stock assessment estimates that GBYR in the model is above the biomass target ($SB_{40\%}$), and well above the minimum stock size threshold ($SB_{25\%}$). The estimated relative depletion level for the base model in 2019 is 4 520% (95% asymptotic interval: ± 3 000% - 6 050%, corresponding to an unfished spawning biomass of 596 million eggs (95% asymptotic interval: 320 - 872 million eggs) of spawning biomass in the base model (Table e). Unfished age 1+ biomass was estimated to be 2,149 mt in the base case model. The target spawning biomass ($SB_{40\%}$) is 527 million eggs, which corresponds with an equilibrium yield of 179 mt. Equilibrium yield at the proxy F_{MSY} harvest rate corresponding to $SPR_{50\%}$ is 167 mt (Figure i).

Table e: Summary of reference points and management quantities for the base case model.
tab:Ref_pts_mod1

Quantity	**Estimate**	**Low 2.5% limit**	**High 2.5% limit**
Unfished spawning output (million eggs)	1,317	930	1,705
Unfished age 1+ biomass (mt)	2,149	1,664	2,634
Unfished recruitment (R_0)	4,784	1,164	8,404
Spawning output(2018 million eggs)	580	320	841
Depletion (2018)	0.441	0.298	0.583
Reference points based on $\mathbf{SB_{40\%}}$			
Proxy spawning output ($B_{40\%}$)	527	428	626
SPR resulting in $B_{40\%}$ ($SPR_{B40\%}$)	0.458	0.458	0.458
Exploitation rate resulting in $B_{40\%}$	0.151	0.107	0.195
Yield with $SPR_{B40\%}$ at $B_{40\%}$ (mt)	179	108	251
Reference points based on SPR proxy for MSY			
Spawning output	588	477	698
SPR_{proxy}	0.5		
Exploitation rate corresponding to SPR_{proxy}	0.131	0.094	0.169
Yield with SPR_{proxy} at SB_{SPR} (mt)	167	101	233
Reference points based on estimated MSY values			
Spawning output at MSY (SB_{MSY})	279	226	333
SPR_{MSY}	0.289	0.278	0.299
Exploitation rate at MSY	0.266	0.194	0.339
Dead Catch MSY (mt)	210	121	299
Retained Catch MSY (mt)	210	121	299

Management Performance

management-performance

Table f

Unresolved Problems and Major Uncertainties

unresolved-problems-and-major-uncertainties

Table f: Recent trend in total mortality for gopher and black-and-yellow rockfishes (GBYR), combined, relative to the management guidelines for the minor nearshore rockfish south of 40°10′ N. latitude. Total mortality estiamtes are based on annual reports from the NMFS NWFSC

tab:mnmgt_perform

	GBYR	Minor Nearshore Rockfish		
Year	Total mortality	Total mortality	ACL	OFL
2011	122.87	436	1,001	1,156
2012	91.96	445	1,001	1,145
2013	104.53	495	990	1,164
2014	103.63	596	990	1,160
2015	107.95	676	1,114	1,313
2016	111.55	641	1,006	1,288
2017	-	-	1,329	1,163
2018	-	-	1,344	1,179

139 # Decision Table

decision-table

Table g: Projections of potential OFL (mt) for each model, using the base model forecast.

tab:OFL_projection

Year	OFL
2019	180.21
2020	168.28
2021	167.05
2022	174.54
2023	187.31
2024	200.67
2025	210.28
2026	213.76
2027	211.19
2028	204.47
2029	196.15
2030	188.32

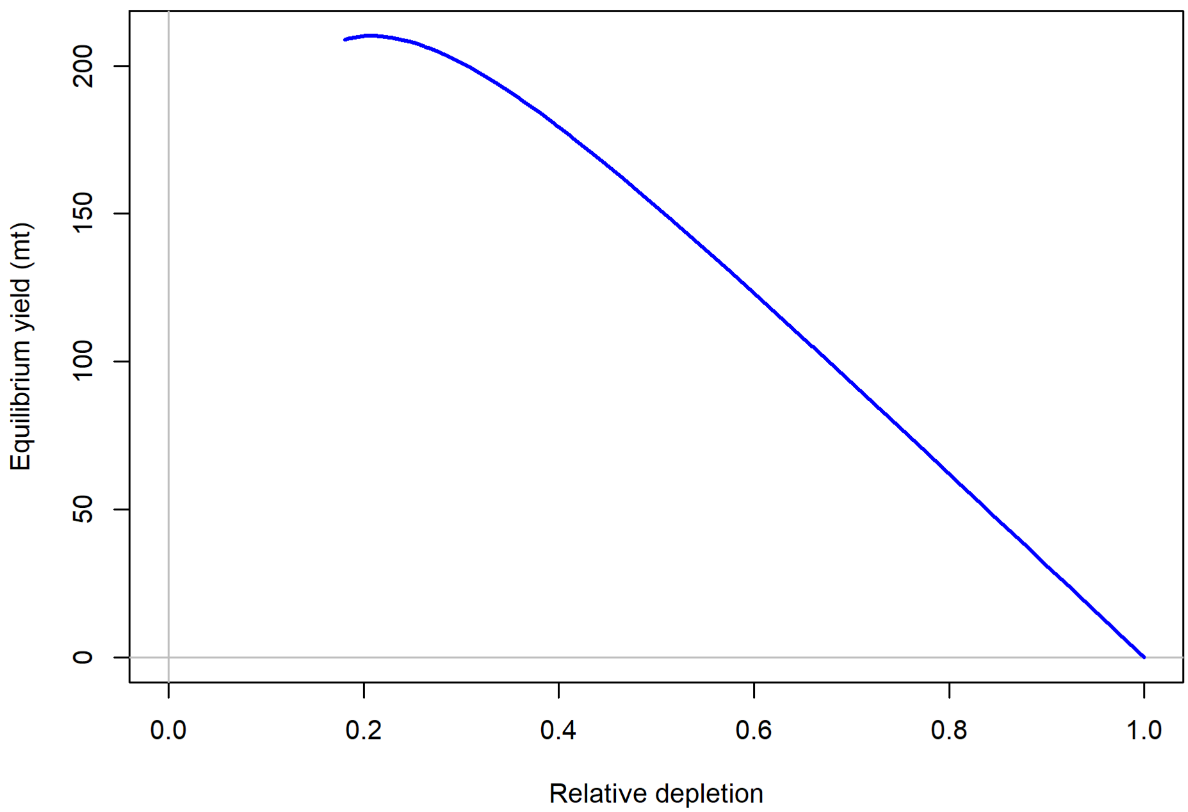

Figure i: Equilibrium yield curve for the base case model. Values are based on the 2018 fishery selectivity and with steepness fixed at 0.718. fig:Yield_all

Table h: Summary of 10-year projections beginning in 2020 for alternate states of nature based on an axis of uncertainty for the model. Columns range over low, mid, and high states of nature, and rows range over different assumptions of catch levels. An entry of "–" indicates that the stock is driven to very low abundance under the particular scenario.

tab:Decision_table_mod1

			States of nature					
			Low M 0.05		Base M 0.07		High M 0.09	
	Year	Catch	Spawning Output	Depletion	Spawning Output	Depletion	Spawning Output	Depletion
	2019	-	-	-	-	-	-	-
	2020	-	-	-	-	-	-	-
	2021	-	-	-	-	-	-	-
40-10 Rule, Low M	2022	-	-	-	-	-	-	-
	2023	-	-	-	-	-	-	-
	2024	-	-	-	-	-	-	-
	2025	-	-	-	-	-	-	-
	2026	-	-	-	-	-	-	-
	2027	-	-	-	-	-	-	-
	2028	-	-	-	-	-	-	-
	2019	-	-	-	-	-	-	-
	2020	-	-	-	-	-	-	-
	2021	-	-	-	-	-	-	-
40-10 Rule	2022	-	-	-	-	-	-	-
	2023	-	-	-	-	-	-	-
	2024	-	-	-	-	-	-	-
	2025	-	-	-	-	-	-	-
	2026	-	-	-	-	-	-	-
	2027	-	-	-	-	-	-	-
	2028	-	-	-	-	-	-	-
	2019	-	-	-	-	-	-	-
	2020	-	-	-	-	-	-	-
	2021	-	-	-	-	-	-	-
40-10 Rule, High M	2022	-	-	-	-	-	-	-
	2023	-	-	-	-	-	-	-
	2024	-	-	-	-	-	-	-
	2025	-	-	-	-	-	-	-
	2026	-	-	-	-	-	-	-
	2027	-	-	-	-	-	-	-
	2028	-	-	-	-	-	-	-
	2019	-	-	-	-	-	-	-
	2020	-	-	-	-	-	-	-
	2021	-	-	-	-	-	-	-
Average Catch	2022	-	-	-	-	-	-	-
	2023	-	-	-	-	-	-	-
	2024	-	-	-	-	-	-	-
	2025	-	-	-	-	-	-	-
	2026	-	-	-	-	-	-	-
	2027	-	-	-	-	-	-	-
	2028	-	-	-	-	-	-	-

Table i: Base case results summary.

tab:base_summary

Quantity	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019
Landings (mt)										
Total Est. Catch (mt)										
OFL (mt)										
ACL (mt)										
(1-SPR)(1-$SPR_{50\%}$)	0.74	0.74	0.63	0.61	0.71	0.74	0.77	0.77	0.73	
Exploitation rate	0.11	0.10	0.08	0.07	0.09	0.09	0.09	0.08	0.07	
Age 1+ biomass (mt)	1454.47	1385.60	1298.92	1236.26	1210.39	1198.97	1188.32	1203.33	1250.27	1329.14
Spawning Output	834	764	707	676	654	625	602	585	580	596
95% CI	520 - 1149	469 - 1060	429 - 985	410 - 942	397 - 911	374 - 877	352 - 852	332 - 838	320 - 841	320 - 872
Depletion	63.3	58.0	53.7	51.3	49.6	47.5	45.7	44.4	44.1	45.2
95% CI	45.47 - 81.16	41.45 - 74.61	38.2 - 69.18	36.74 - 65.91	35.74 - 63.55	33.95 - 61.01	32.24 - 59.18	30.69 - 58.12	29.81 - 58.32	29.99 - 60.5
Recruits	3680	3319	3393	4194	6319	8445	6941	5137	4177	4276
95% CI	1452 - 9327	1277 - 8625	1279 - 9001	1580 - 11131	2385 - 16738	3154 - 22610	2564 - 18788	1845 - 14301	1469 - 11875	1486 - 12305

Research and Data Needs

research-and-data-needs

We recommend the following research be conducted before the next assessment:

1. **xxxx**:
2. **xxxx**:
3. **xxxx**:
4. **xxxx**:
5. **xxxx**:

1 Introduction

introduction

1.1 Basic Information and Life History

basic-information-and-life-history

Population Structure and Complex Assessment Considerations

There have been a number of analyses of the genetic differentiation between gopher rockfish and black-and-yellow rockfish that have yielded some range of results, but have generally concluded that there is unusually low genetic differentiation between the two species. The most frequently used measure of genetic analyses to evaluate evidence for population differentiation is the fixation index (F_{ST}), defined as the proportion of the total genetic variation in one sub-population (subscript S) relative to the total genetic variation (subscript T) (Hauser and Carvalho 2008, Waples et al. 2008). Values of F_{ST} range from 0 to 1 where a zero value implies the populations are panmictic and a value closer to one implies the two populations are genetically independent. Values of F_{ST} thought to be consistent with biologically meaningful genetic differentiation and demographic isolation between populations range from 0.05 to 0.1 (Waples and Gaggiotti 2006). It is also important to note that F_{ST} values are dependent on the study's sample size and it may not necessarily be appropriate to compare them across studies.

Morphologically, gopher and black-and-yellow rockfishes are almost indistinct, except for their color variation (Hubbs and Schultz 1933). Early efforts to evaluate whether the two species were genetically distinct began with an allozyme analysis by Seeb et al. (1986), which did not detect genetic differentiation between gopher and black-and-yellow rockfish. However, as allozymes are proteins that are often conserved due against variation, this early work was not enormously conclusive. In a subsequent study of restriction site polymorphisms, Hunter et al. (1994) found slight but significant differences between species based on restriction fragment length polymorphisms (RFLP's). Following that study, an analysis of the mitochondrial control region by Alesandrini and Bernardi (1999) did not detect differences between the two species, although there were limitations regarding how representative those results were across the genome. Analysis of seven microsatellite loci by Narum et al. (2004) found an F_{ST} of 0.049 across the overlapping range of the two species, which provided some evidence of divergence, although such divergence is relatively low compared to other species within *Sebastes. Those authors characterized their results as suggesting that the two are "reproductively isolated incipient species."
Buonaccorsi et al. (2011) found an even lower F_{ST} of 0.01 using 25 loci, and concluded that gopher and black-and-yellow rockfish "have not completed the speciation process." All of these studies are indicative of low levels of genetic divergence and a high probability of ongoing gene flow between the two nominal species.

Most recently, an analysis of microhaplotypes by Baetscher (2019) observed a higher frequency of mis-assignments of individuals to between gopher and black-and-yellow rockfishes compared to all other pairs of species in the genus *Sebastes*. In addition, comparisons of F_{ST}

values within the study indicated that the level of genetic differentiation observed between gopher and black-and-yellow rockfishes is less than that observed among all other pairwise comparisons of the 54 species in the *Sebastes* genus that were included in their analysis. Baetscher (2019) characterized the results as suggestive of the two species representing "sister species with evidence of ongoing gene flow," noting that a more rigorous evaluation of the level of genetic distinction between these two species would benefit from whole-genome sequencing of representatives from each species group.

In addition to the differences in coloration, the depth distribution and range differ between the two species. The range of both species extends from Cape Blanco Oregon to Baja California. Both species are uncommon north of Fort Bragg, California and black-and-yellow rockfish is uncommon south of Pt. Conception, California. However, gopher rockfish can be found as far south as Punta San Roque on the Baja peninsula.Gopher rockfish are found in rocky reef habitat from the intertidal to depths of 264 ft (80 m) with a predominant depth distribution of 30 to 120 ft (9-37 m), while the black-and-yellow rockfish occupies depths from the intertidal to 120 ft (40 m) and is predominantly observed in depths shallower than 60 ft (18 m) (Eschmeyer et al. 1983, Love et al. 2002).

Both species are solitary, sedentary, territorial species with home ranges of 10-12 square meters, with occasional extended trips (Love et al. 2002). A large percentage (67-71%) of black-and-yellow rockfish returned to the site of capture within two weeks after tranlocated within 50 m (Hallacher 1984). Lea et al. (1999) found that gopher rockfish exhibit minor patterns of movement (<1.5 nm, 2.8 km) with all fish being recaptured on the same reef system where they were tagged. Another study, conducted by (Matthews 1985), reported movements up to 1.2 km by gopher rockfish that traveled from a low-relief natural reef to a high-relief artificial reef. The change in substrate type may have been a factor in the movement in the Matthews(1985) study.

Larson (1980) conducted a study on the territoriality and segregation between gopher and black-and-yellow rockfishes. When one species was removed, the other extended its depth range to areas where the other previously occupied, indicating interspecific competition plays a role in controlling their depth distributions where both species are present. Of the two species, black-and-yellow rockfish are socially dominant and aggressive towards excluding gopher rockfish from shallower waters.

Both species are be feed at night, with similar diets composed primarily of crabs and shrimp,supplemented by fish and cephalopods (Larson 1972, Hallacher and Roberts 1985, Love et al. 2002). Loury et al. (Loury et al. 2015) found no significant differences in the diet of gopher rockfish inside and outside the 35 year old Point Lobos Marine Protected Area (MPA). She did find the diet of gopher rockfish at A˜{n}o Nuevo (shallower and north of Point Lobos) dominated by crabs and dominated by brittle stars at southern, deeper study locations. Zuercher (2019) examined the diets of a suite of nearshore rockfish species including black-and-yellow and found that they relied on hard-bodied benthic invertebrates such as Brachyuran crabs, shrimps, other arthropods, and octopus. The diet of black-and-yellow rockfish remained the same across sampling years, but they occupied a lower trophic level during the upwelling season.

1.2 Early Life History

early-life-history

Both gopher and black-and-yellow rockfish are viviparous and release one brood per season between January and July (Echeverria 1987). Larvae are approximately 4 mm in length at birth and have a 1-2 month pelagic stage before recruiting to the kelp forest canopy, i.e., surface fronds of *Macrosystis pyrifera* and *Cysteoseira osmundacea* at around 15-21 mm (Anderson 1983, Wilson et al. 2008). The larvae are transparent until they reach juvenile stage at 22-23 mm. Differences in coloration begin to occur at 25-30 mm and can be used to identify one species from the other. Gopher rockfish become more orange and brown, while black-and-yellow rockfish become more black and yellow. Benthic juveniles associate with *Macrosystis* holdfasts and sporophylls (Anderson 1983). Gopher and black-and-yellow rockfish have similar development

The juveniles undergo ontogenetic migration down the stalks to deeper depths, finally settling on rocky reef habitat in their respective adult depth distribution. Juvenile bocaccio and other fish predate on young of year and other reef dwelling species including cabezon predate on post-settlement juveniles. Individuals avoid rough surge conditions and predators by hiding in the rocky bottom during the daylight hours, then returning to more open water at dusk (Love et al. 2002).

1.3 Map

map

A map showing the scope of the assessment and depicting boundary at Pt. Conception for the recreational fishing fleet (Figure d).

1.4 Ecosystem Considerations

ecosystem-considerations-1

In this assessment, ecosystem considerations were not explicitly included in the analysis. This is primarily due to a lack of relevant data and results of analyses (conducted elsewhere) that could contribute ecosystem-related quantitative information for the assessment.

1.5 Fishery Information

fishery-information

The hook-and-line fishery off California developed in the late 19th century (Love et al. 2002). The rockfish trawl fishery was established in the early 1940s, when the United States became involved in World War II and wartime shortage of red meat created an increased demand for other sources of protein (Harry and Morgan 1961, Alverson et al. 1964).

Gopher and black-and-yellow (referred to from hereon as GBRY when discussing the complex) rockfish have been a minor component of the commercial and recreational rockfish fishery since at least the late 1960s (CFIS and RecFIN). The commercial catch histories of the two species cannot easily be separated (Figure 2). From 1916-1936 only black-and-yellow rockfish were reported in the landings, and an average of 0.04 mt of black-and-yellow rockfish are reported from 1937-1983. Black-and-yellow rockfish reappear in the landings in 1984 with 7.2 mt landed commercially. From 1985-1988 the trend switches and only black-and-yellow rockfish appear in the commercial landings, with gopher rockfish averaging 0.1 mt landed, and 0 mt reported in 1987. From 1988 and on, the landings are dominated by gopher rockfish, and both species are represented in the commercial landings.

The landings from south of Pt. Conception are minor throughout the time period, with peaks in the 1950s and 60s for gopher rockfish. Black-and-yellow rockfish are rare south of Pt. Conception and expected that these catches are minimal.

The live fish fishery began in the early 1990s, with the first commercial landings of live gopher rockfish in 1993, and black-and-yellow rockfish a year later. By 1995 over half (57%; 39 mt) of the commercial landings were from the live fish fishery. This increased quickly over the next few years and has been on average 84% of the landed GBYR since 2000. The majority of the landings are from gopher rockfish north of Pt. Conception. Landings of live GBYR south of Pt. Conception were higher in the late 1990s, (max. 3.2 mt in 1999), and have been averaging 0.4 mt since 2003.

The ex-vessel value of GBYR increased from less than $40,000 in 1984 and peaked at $680,452 in 1996 (Figure 3). The ex-vessel revenue has been fairly stable at around $500,000 a year since 2007. Prior to the live fish fishery in 1994, the average price per pound for either species was around $2 a pound. The live fish fishery increased the value of both species to an average of $6-$8 a pound. The maximum reported value of either a gopher or black-and-yellow rockfish was $20 a pound in 2003.

The recreational GBYR fishery for California is most prominent north of Pt. Conception throughout the entire catch history (Figure a). The recreational landings increased from 1928 to 1980. The sharp increase in the 1980s could be an artifact of the MRFSS sampling program that began in 1980; however, the more recent recreational landings also exhibit a cyclical trend of years with high catches followed by period of decreased recreational landings. The CRFS era recreational total mortality represents the most accurate description of the recreational fleet's catches in terms of area, mode and species (Figure 4).

Recreational catches are dominated by gopher rockfish north of Pt. Conception in the private/rental (PR) and party/charter (PC or CPFV) modes. South of Pt. Conception gopher rockfish are predominately caught by the CPFV fleet, with all other modes being insignificant. The total recreational mortality of black-and-yellow rockfish south of Pt. Conception since 2005 is 3 mt, compared to 106 mt north of Pt. Conception. The total mortality since 2005 for gopher rockfish is 86 mt south of Pt. Conception and 669 mt north of Pt. Conception.

1.6 Summary of Management History

summary-of-management-history

Prior to the adoption of the Pacific Coast Groundfish Fishery Management Plan (FMP) in 1982, GBYR were managed through a regulatory process that included the California Department of Fish and Wildlife (CDFW) along with either the California State Legislature or the Fish and Game Commission (FGC) depending on the sector (recreation or commercial) and fishery. With implementation of the Pacific Coast Groundfish FMP, GBYR came under the management authority of the Pacific Fishery Management Council (PFMC), and were managed as part of the *Sebastes* complex. Because neither species had undergone rigorous stock assessment and did not compose a large fraction of the landings they were classified and managed as part of "Remaining Rockfish" under the larger heading of "Other Rockfish" (PFMC (2002, 2004)).

Since the early 1980s a number of federal regulatory measures have been used to manage the commercial rockfish fishery including cumulative trip limits (generally for two- month periods) and seasons. Starting in 1994 the commercial groundfish fishery sector was divided into two components: limited entry and open access with specific regulations designed for each component. Other regulatory actions for the general rockfish categories have included area closures, gear restrictions, and cumulative bimonthly trip limits set for the four different commercial sectors - limited entry fixed gear, limited entry trawl, open access trawl, and open access non-trawl. Harvest guidelines are also used to regulate the annual harvest for both the recreational and commercial sectors.

In 2000, changes in the PFMC's rockfish management structure resulted in the discontinued use of the *Sebastes* complex, and was replaced with three species groups: nearshore, shelf, and slope rockfishes (January 4, 2000; 65 FR 221), of which GBYR are included in the nearshore group. Within the nearshore group, they are included in the "shallow nearshore rockfish" component.

During the late 1990s and early 2000s, major changes also occurred in the way that California managed its nearshore fishery. The Marine Life Management Act (MLMA), which was passed in 1998 by the California Legislature and enacted in 1999, required that the FGC adopt an FMP for nearshore finfish (Wilson-Vandenberg et al. 2014). It also gave authority to the FGC to regulate commercial and recreational nearshore fisheries through FMPs and provided broad authority to adopt regulations for the nearshore fishery during the time prior to adoption of the nearshore finfish FMP. Within this legislation, the Legislature also included commercial size limits for ten nearshore species including GBYR (10-inch minimum size) and a requirement that commercial fishermen landing these ten nearshore species possess a nearshore permit.

Following adoption of the Nearshore FMP and accompanying regulations by the FGC in fall of 2002, the FGC adopted regulations in November 2002 which established a set of marine reserves around the Channel Islands in southern California (which became effective April 2003). The FGC also adopted a nearshore restricted access program in December 2002 (which included the establishment of a Deeper Nearshore Permit) to be effective starting in the 2003 fishing year.

Also, since the enactment of the MLMA, the Council and State in a coordinated effort developed and adopted various management specifications to keep harvest within the harvest targets, including seasonal and area closures (e.g. the CCAs; a closure of Cordell Banks to specific fishing), depth restrictions, minimum size limits, and bag limits to regulate the recreational fishery and license and permit regulations, finfish trap permits, gear restrictions, seasonal and area closures (e.g. the RCAs and CCAs; a closure of Cordell Banks to specific fishing), depth restrictions, trip limits, and minimum size limits to regulate the commercial fishery.

The state of California has adopts regulatory measures to manage the fishery based on the harvest guidelines set forth by the PFMC. The commercial open access and limited entry fixed gear sectors have undergone three different spatial management changes in since 2000. Since 2005, both have managed the area south of $40°10'$ N. latitude as one area. he open access commercial fishery is managed based on bimonthly allowable catches, that have ranged from 200 pounds to 1800 pounds per two months since 2000.
From 2005 to 2018, the catch limits have doubled and are now set at 1200 pounds per two months (for all months) with March and April remaining closed. The limited entry fixed year sector has followed the same pattern as the open access sector with bi-monthly limits and a doubling of the catch since 2005. The limited entry trawl fleet is managed on monthly limits on an annual basis. Since 2011, the limit has been 300 pounds per month for non-IFQ species. A history of California's commercial regulations from 2000-2018 can be found in Appendix X.

Significant regulatory changed in California's recreational sector began with a change from unlimited number of hooks and lines allowed in 1999 and prior to no more than three hooks and one line per angler in 2000. Since 2001, the limit has been no more than two hooks and one line per angler. There is no size limit in the recreational fishery for gopher or black-and-yellow rockfish. GBYR are part of the nearshore complex which has had a sub-bag limit within the rockfish bag limit since 1999. The nearshore sub-bag limit has been 10 fish since 2005.

California also began spatial management, closures, and depth restrictions for the recreational fleet in 2000. In general, the recreational season north of Pt. Conception extends from April to December, and south of Pt. Conception from March to December. In the area that GBYR are most commonly landed, from Monterey to Morro Bay, the depth restrictions have been between 30 and 40 fathoms until 2017. In 2017 the depth restrictions were eased by 10 fathoms, opening up fishing depths along the central California coast that had not been open consistently since 2002. In both 2017 and 2018, the deepest 10 fathoms was closed prior to the season in December due to high by-catch rates of yelloweye rockfish, one of two rockfish species that are still overfished. A full history of the recreational regulations relating to the spatial management of the fleet can be found in Appendix XXX.

1.7 Management Performance

`management-performance-1`

NEED TO FINISH
The contribution of GBYR to the minor nearshore rockfish OFLs is currently derived from two sources: 1) forecasts from Key et al. (n.d.), from Cape Mendocino to Pt. Conception and 2) a Depletion Corrected Average Catch (DCAC (MacCall 2009)) for the area south of Point Conception. Estimated catch of GBYR has been...... A summary of these values as well as other base case summary results can be found in Table f.

1.8 Fisheries Off Mexico or Canada

`fisheries-off-mexico-or-canada`

The range of GBYR does not extend north to the Canadian border, and they are rarely encountered in Oregon and Washington. The southern end of the gopher rockfish's range extends to Punta San Roque (southern Baja California) while the southern end of the black-and-yellow rockfish's range extends to Isla Natividad (central Baja California) (Love et al. 2002). However, black-and-yellow rockfish are rare south of Pt. Conception, California. This was no available information on the fishery for GBYR at the time of this assessment, nor additional details on the abundance or distribution patterns in Mexican waters.

2 Assessment

`assessment`

2.1 Data

`data`

Data used in the GBYR assessment are summarized in Figure 5. Descriptions of the data sources are in the following sections.

2.1.1 Commercial Fishery Landings

`commercial-fishery-landings`

Overview of gopher and black-and-yellow catch history

Commercial fishery landings for gopher and black-and-yellow rockfishes have not been reported consistently by species throughout the available catch history (Figure 2). The period from 1916-1935 indicates that only black-and-yellow rockfish were landed in the commercial fishery, which then switched to predominately gopher rockfish from 1937-1984. From 1985-1988 the landings data suggest that only black-and-yellow rockfish were landed and not until 1995 are both species well-represented in the catches. There is no way to tease apart the historical catches by species and even across north and south of Pt. Conception prior to about 1995. This precludes the ability to model the catch histories for either species accurately. Given these constraints, all commercial data were combined to represent one commercial fleet in the assessment.

The stock assessment of gopher rockfish in 2005 did not include black-and-yellow rockfish landings. A comparison of recreational and commercial landings from the 2005 assessment to those used in this assessment suggest the 2005 assessment may have included some black-and-yellow rockfish landings (Figure 6). The 2005 assessment estimated recreational landings from 1969-1980 based on a ratio of commercial to recreational landings, where as this assessment makes use of the California Catch Reconstruction landings estimates (Ralston et al. 2010).

Commercial Landings Data Sources

The California Catch Reconstruction (Ralston et al. 2010) contains landings estimates of commercial landings from 1916-1968 and was queried on 4 April 2019 for GBYR. There were no estimated gopher rockfish landings prior to 1937. Landings in this database are divided into trawl and 'non-trawl.' Since the majority of GBYR are caught in the commercial fixed gear fisheries, only estimated catch in the 'non-trawl' was used. A total of 0.154 mt (3.18%) were removed from Eureka commercial landings (based on current proportions of commercial catch from north of Cape Mendocino in Eureka) since the assessment represents the GBYR stock south of Cape Mendocino. The majority of GBYR commercial landings (avg. 83%) are landed in the Monterey and Morro Bay port complexes.

Contemporary landings were extracted from two data sources, the California Cooperative Groundfish Survey, CALCOM) and the Pacific Fisheries Information Network PacFIN landings database.Both databases are based on the same data sources (CALCOM data), but apply a catch expansion based on different algorithms. CALCOM collects information including species composition data (i.e. the proportion of species landed in a sampling stratum), and landing receipts (sometimes called "fish tickets") that are a record of pounds landed in a given stratum. Strata in California are defined by market category, year, quarter, gear group, port complex, and disposition (live or dead). Although many market categories are named after actual species, catch in a given market category can consist of several species. These data form the basis for the "expanded"" landings, i.e., species composition data collected by port samplers were used to allocate pounds recorded on landing receipts to species starting in 1978. Use of the "Gopher Rockfish" or the "Black-and-Yellow Rockfish" categories alone to represent actual landings of GBY would not be accurate.

See Pearson et al. Appendix C (2008) for a simple example of the expansion calculations for the CALCOM database. A description of the landings in PacFIN can be found in Sampson and Crone (1997). Both databases, including species compositions, and expanded landings estimates are stored at the Pacific States Marine Fisheries Commission, a central repository of commercial landings data for the U.S. West Coast. As a note, CALCOM is the only source for landings from 1969-1980.

Commercial landings from 1981-2018 were queried for a final time from the CALCOM database on 4 April 2019 and from PacFIN on 3 June 2019. There are very small differences in commercial landings between CALCOM and PacFIN from 1981-2018 (Figure 7). Landings estimates from PacFIN were used in the assessment (Table 1). Landings were stratified by year, quarter, live/dead, market category, gear group, port complex, and source of species composition data (actual port samples, borrowed samples, or assumed nominal market category). Data from individual quarters were aggregated at the year level. Fish landed live or dead were combined, due to changes over time in the reliability of condition information (D. Pearson, pers. comm.). From 1916-1968, on average, 74% of GBYR were landed north of Point Conception, which rose to 97% from 1978-2018. Given the smaller landings south of Pt. Conception and the similar length composition of GBYR north and south of Pt. Conception, no spatial separation was considered for the commercial fleet.

2.1.2 Commercial Discards

commercial-discards

The West Coast Groundfish Observer Program (WCGOP) provides observer data on discarding across fishery sectors back to 2003. Gopher and black-and-yellow rockfishes have different depth-stratified commercial fishery discard mortality rates (Pacific Fishery Managment Council 2018). In consultation with WCGOP staff, the STAT used estimates of total discard mortality from WCGOP's Groundfish Expanded Mortality Multiyear (GEMM) report. WCGOP observes between 1-5% of nearshore fixed gear landings annually south of $40°10'$ N. latitude (coverage rates available here). The expanded estimates of total discard weight by species is calculated as the ratio of the observed discard weight of the individual species divided by the observed landed weight from PacFIN landing receipts. WCGOP discard estimates for the nearshore fixed gear fishery take into account the depth distribution of landings in order to appropriately apply the depth-stratified discard mortality rates by species (Somers, K.A., J. Jannot, V. Tuttle, K. Richerson and McVeigh 2018). The discard mortality for 2018 was estimated as an average of the discard mortality from 2013-2017. Discard mortality was estimated from the period prior to WCGOP discard estimates (1916-2002) based on the average discard mortality rate from 2003-2016 (2017 was excluded because 2017 discard mortality was disproportionately higher than all other years) (Table 1).

2.1.3 Commercial Fishery Length and Age Data

commercial-fishery-length-and-age-data

Biological data from the commercial fisheries that caught GBYR were extracted from CALCOM on 9 May 2019. The CALCOM length composition data were catch-weighted to "expanded" length the raw length composition data (Table 2). The 2005 assessment used commercial length composition information from CALCOM, but did not include black-and-yellow rockfish and is not directly comparable. The 2005 assessment used 2 cm length bins from 16-40 cm, where this assessment uses 1 cm length bins from 4-40 cm. Sex was not available for the majority (99.5%) of the commercial length, and the assessment did not find sexual dimorphism in growth for either species. We aggregated the commercial length composition among all gears and regions south of Cape Mendocino.

Discard length compositions from WCGOP (2003-2017) were expanded based on the the discard estimates and were aggregated for all regions south of Cape Mendocino and across all fixed gear fisheries.

A total of 46 ages were available for gopher rockfish from the commercial fisheries 2009-2011, 2016, and 2018. Though sparse, the data were included as conditional age-at-length for the commercial fleet.

The input sample sizes for commercial length composition data were calculated via the Stewart Method for fisheries (Ian Stewart, personal communication, IPHC):

$$\text{Input effN} = N_{\text{trips}} + 0.138 * N_{\text{fish}} \text{ if } N_{\text{fish}}/N_{\text{trips}} \text{ is} < 44$$

$$\text{Input effN} = 7.06 * N_{\text{trips}} \text{ if } N_{\text{fish}}/N_{\text{trips}} \text{ is} \geq 44$$

Commercial length composition data are made available from PacFIN and the expanded catch-weight length compositions were provided by Andi Stephens (NWFSC) processed through the PacFIN Utilities package. We compares differences between the catch-weighted length composition expansions from CALCOM and PacFIN. We were unable to reconcile the difference between the two data sets. Sample sizes became more similar if the PacFIN data were restricted to the same market categories used by CALCOM in the expansion. However, both data sets apply other filters that we did not have time to explore. For instance, in the year 2000, 290 more fish were used in the CALCOM expansion than in PacFIN, but in 2002, 150 more fish were used in the PacFIN expansions that were not used in CALCOM. Given these caveats, Figure 8 shows the percent difference in the expanded length comps within a year. The biggest difference is in length bin 32 in 2006. However, the same number of fish and samples were used to expand the 2006 lengths in both databases, indicating there are also fundamental differences in how the data are treated. Full documentation is not available for the PacFIN length composition expansion program. The base model for this assessment uses the CALCOM length composition data as described above, but a sensitivity was conducted using the PacFIN length composition data.

2.1.4 Recreational Fishery Landings and Discards

recreational-fishery-landings-and-discards

Historical recreational landings and discard, 1928-1980

Ralston et al. (2010) reconstructed estimates of recreational rockfish catch and discard in California, 1928-1980. Reported landings of total rockfish were allocated to species based on several sources of species composition data. Estimates of GBYR landings and discard (combined) from 1928-1979 are available from the SWFSC. For this assessment, historical recreational catch was stratified by year and area (north and south of Point Conception). The catches of GBYR reported in Ralston et al. (2010) are higher by an order of magnitude than expected given the more recent catches of GBYR in the MRFSS and CRFS eras south of Pt. Conception (Figure 9). The recreational catches estimated by Ralston et al. (2010) were discussed with the paper's co-authors and also CPFV captains in California. A consensus was reached that the estimated landings did not accurately represent the historical GBYR landings and an alternative catch stream should be developed. One possibility for the inflated catches of GBYR in southern California is that all nearshore shallow species were combined and all of the nearshore deep species were combined and a constant relative fraction between the two was used to assign catches to each combination of CDFW fishing block and year. The fraction of GBYR within the nearshore shallow species group was likely overestimated.

The California Catch Reconstruction applied a linear ramp from from 1928-1936 that was not altered in this assessment. From 1937-1979 a linear ramp was developed from the 1936 estimate to the average recreational landing from 1980 and 1983 (1981-1982 catches interpolated as described in the next section) of 4.3 mt. The recreational catches north of Pt. Conception were not altered from the original catch reconstruction. The resulting alternate recreational catch streams are in (Table 3 and Figure 10).

Marine Recreational Fisheries Statistics Survey (MRFSS), 1980-2003

From 1980-2003, the Marine Recreational Fisheries Statistics Survey (MRFSS) executed a dockside (angler intercept) sampling program in Washington, Oregon, and California (see Holliday et al. (1984) for a description of methods). Data from this survey are available from the Recreational Fisheries Information Network RecFIN. RecFIN serves as a repository for recreational fishery data for California, Oregon, and Washington. Catch estimates for years 1980-2003 were downloaded on 23 March 2019, and are consistent from 1992-2004 with the previous assessment (Meisha Key, Alec D MacCall, Traci Bishop n.d.) (Figure 6).

MRFSS-era recreational removals for California were estimated for two regions: north and south of Point Conception. No finer-scale estimates of landings are available for this period. Catches were downloaded in numbers and weight. Catch in weight is sometimes missing from the database due to missing average weight estimates. We estimated average weights based on adjacent strata as needed, although the effect was relatively minor (7.4 mt over all years for gopher rockfish and 0.6 mt for black-and-yellow rockfish). Data were not available for the CPFVs in Northern California from 1980-1982, and we used the average value from this mode and region from 1983-1987 for these three years. MRFSS sampling was temporarily suspended from 1990-1992, and we used linear interpolation to fill the missing years. Sampling of CPFVs in Northern California was further delayed, and the linear interpolation spans the period 1990-1995 for this boat mode and region. Landings data for the shore-based modes (beach/bank, man-made/jetty and shore) were sparse throughout the MRFSS sampling. All three shore-based modes were combined by region and linear interpolations were applied missing data in 1981 for the Northern California and 1995, 1996-2001, and 2004 in Southern California.

Catches from north of Cape Mendocino were removed based on a CRFS-era average of fraction of recreational landings north of Cape Mendocino by mode (3.3% of shore-based, 0.1% of CPFV, and 0.2% of private/rental were removed). From 1980-1989, San Luis Obispo County was sampled as part of Southern California (personal observation from MRFSS Type 3 sampler examined catch where county is available for 1980-2004). This assessment separates the recreational fleet at Pt. Conception. Recreational landings were re-allocated from southern California from 1980-1992 by fleet based on the average proportion of recreational landings in northern California from 1996-2004 (after sampling of the CPFV fleet in northern California resumed). The average proportion re-allocated from southern to northern California for the CPFV mode was 85%, 97% for the private/rental mode, and 81% for the shore-based modes. Data were pooled over all years and modes to estimate the landings re-allocation for the shore-based modes. Total recreational landings for 1981 and 1982 were 18.8 mt and 18.6 mt, respectively. These landings were >60 mt lower than any of the neighboring years. Landings from 1981-1982 were interpolated from the 1980 and 1983 landings.

California Recreational Fisheries Survey (CRFS), 2004-2016

MRFSS was replaced with the California Recreational Fisheries Survey (CRFS) beginning January 1, 2004. Among other improvements to MRFSS, CRFS provides higher sampling intensity, finer spatial resolution (6 districts vs. 2 regions), and onboard CPFV sampling. Estimates of catch from 2004-2018 were downloaded from the RecFIN database a final time on 4 June 2019, We queried and aggregated CRFS data to match the structure of the MRFSS data, by year, and region (Table 3). Catches in the shore-based modes are small compared to the CPFV and private rental modes. All modes are combined, but separated at Point Conception for two recreational fleets in this assessment, just as was done for the California Catch Reconstruction and MRFSS time series.

Recreational Discard

Recreational discards were only added to the California Catch Reconstruction landings, as Ralston et al. (2010) did not address discards for the recreational reconstruction. Recreational removals from the California Department of Fish and Wildlife MRFSS era (1980-2003) includes catch type A + B1. Catch type A refers to estimates of catch based on sampler-examined catch. Catch type B1 includes mainly angler-reported discard, but also angler-reported retained fish that were unavailable to the sampler during the interview (e.g., fillets). The CRFS era removals account for depth-stratified discard mortality rate and the catch time series includes both retained and discarded catch (total mortality). We calculated the ratio of dead discards to total mortality from the CRFS era by region and mode. The region average across modes was applied to the California Catch Reconstruction as a constant. The result added 4.68% annually to recreational removals north of Pt. Conception and 4.05% annually to the removals South of Pt. Conception). The final time series of landings and discard mortality are in Table 3.

2.1.5 Recreational Fishery Length and Age Data

`recreational-fishery-length-and-age-data`

Recreational length composition samples for California were obtained from several sources, depending on the time period and boat mode (Table 2). This assessment makes use of a much longer time series of length composition data, relative to the previous assessment, as described below. Input sample sizes for recreational length composition data were based on the number of observed trips, when available. Other proxies that were used to estimate the number of trips are described below.

There were no standardized coastwide surveys measure retained or discarded fish from the recreational fleet prior to 1980.

CPFV length composition data, 1959-1978

The earliest available length data for this assessment were described by Karpov et al. (1995), who assembled a time series (1959-1972) of available California CPFV length data (made available courtesy of W. Van Buskirk). For GBYR, data from 1959-1961 and 1966 were available north of Pt. Conception and from 1959-1961 from south of Pt Conception. A total of 716 (680 north of Pt. Conception) unsexed measurements of retained fish (no discards) were included in the assessment (Table 2). Sampling of these length data did not follow a consistent protocol over time and areas (data are unweighted), and therefore may not be representative of total catch. Since the number of trips sampled was not reported by Karpov et al. (1995), we assume the number of sampled trips is proportional to the number of measured fish in each year, and estimated the number of trips using the ratio of fish measured per trip in the MRFSS data (roughly 10 fish per trip).

Collins and Crooke (n.d.) conducted an onboard observer survey of the CPFV fleet in southern California from 1975-1978. A total of 1,308 GBYR lengths were available from the study and were assumed to all be from retained fish. Ally et al. (1991) conducted an onboard observer program of the CPFV fleet from 1985-1987 in southern California. Because MRFSS data were available for this time period as well and represents multiple recreational modes, the Ally et al. (1991) length data were not used in the assessment.

MRFSS Recreational Length Data, 1980-1989 and 1993-2003

Unsexed length data of retained fish were collected by MRFSS dockside samplers and downloaded from the RecFIN website. We identified a subset of lengths that were converted from weight measurements, and these were excluded from the final data set (Table 2). The length measurements from Collins and Crooke (n.d.) from 1975-1978 are assumed to all be from retained fish. As of 2003, the CDFW Onboard Observer program has taken length measurements for discarded fish. The retained catch is measured during the dockside (angler intercept) surveys.

The number of CPFV trips used as initial sample sizes for the MRFSS was based on the number of CPFV trips was determined from the trip-level MRFS CPFV database and the number of private boat trips was determined based on unique combinations of the variables ASSNID ,ID_CODE, MODE_FX, AREA_X, DIST, INTSITE, HRSF, CNTRBTRS, SUB_REG, WAVE, YEAR, and CNTY in the Type 3 (sampler-examined catch) data.

During the recent restructuring of the CRFS data on RecFIN, a “trip” identifier was not carried over for all modes, and trip-level sample sizes could not be extracted from the biological detail table on RecFIN. A proxy for initial sample sizes for 2004-2018 were developed using the 2015 data for which I had access to raw data files by mode from CDFW. In more recent years, sampling of the shore-based modes has declined and were not sampled at all in 2018. Samples sizes were calculated by mode as the number of port-days (or site-days for shore-based modes) during bi-weekly intervals (e.g., Jan 1-15, Jan 16-31, etc). The number of port-days sampled in the bi-weekly intervals was used as the initial sample size for number of trips to calculate initial input sample sizes using Ian Stewart’s method (described above). All length data were re-weighted in the assessment model.

2.1.6 Fishery-Dependent Indices of Abundance

fishery-dependent-indices-of-abundance

A summary of all indices in the assessment can be found in Table 27 and Figure 11.

MRFSS Dockside CPFV Index

From 1980 to 2003 the MRFSS program conducted dockside intercept surveys of recreational fishing fleet. The program was temporarily suspended from 1990-1992 due to lack of funding. For purposes of this assessment, the MRFSS time series was truncated at 1998 due to sampling overlap with the onboard observer program (i.e., the same observer samples the catch while onboard the vessel and also conducts the dockside intercept survey for the same vessel). Each entry in the RecFIN Type 3 database corresponds to a single fish examined by a sampler at a particular survey site. Since only a subset of the catch may be sampled, each record also identifies the total number of that species possessed by the group of anglers being interviewed. The number of anglers and the hours fished are also recorded. The data, as they exist in RecFIN, do not indicate which records belong to the same boat trip. A description of the algorithms and process used to aggregate the RecFIN records to the trip level is outlined Supplemental Materials (“Identifying Trips in RecFIN”).

Initial trip filters included eliminating trips targeting species caught near the surface waters for all or part of the trip, including trips with catch of bluefin tuna, yellowfin tuna, skipjack, and albacore.

The following filtering steps were applied to gopher rockfish, as well as the sum of the two species to represent GBYR. No filtering or indices were developed for black-and-yellow rockfish alone due to the sparseness in the data. In the raw data, unfiltered data, black-and-yellow rockfish only occurred in 48 trips that did not also observer gopher rockfish. There were an additional 65 trips that encountered both species. There was little difference between indices developed for gopher-only and the GBYR complex for both north and south of Pt. Conception (Figure 12). The descriptions of the filtering and data below represent those for the GBYR complex.

The species composition of catch in California varies greatly with latitude. Therefore, Stephens-MacCall filtering was applied independently for north and south of Pt. Conception. Separate indices were also developed to represent two recreational fleets in the model.

Since recreational fishing trips target a wide variety of species, standardization of the catch rates requires selecting trips that are likely to have fished in habitats containing GBYR. The Stephens-MacCall (2004) filtering approach was used to identify trips with a high probability of catching GBYR, based on the species composition of the catch in a given trip. Prior to applying the Stephens-MacCall filter, we identified potentially informative predictor species, i.e., species with sufficient sample sizes and temporal coverage (at least 30 positive trips total) to inform the binomial model. Coefficients from the Stephens-MacCall analysis (a binomial GLM) are positive for species which co-occur with GBYR, and negative for species that are not caught with GBYR. Each of these filtering steps and the resulting number of trips remaining in the sampling frame are provided in Table 16.

MRFSS filtering and index standardization for north of Pt. Conception Prior to the Stephens-MacCall filter, a total of 2,788 trips were retained for the analysis. As expected, positive indicators of GBYR trips include several species of nearshore rockfish, treefish, kelp rockfish, and blue rockfish, and the strongest counter-indicator was striped bass (Figure 13). While the filter is useful in identifying co-occurring or non-occurring species assuming all effort was exerted in pursuit of a single target, the targeting of more than one target species can result in co-occurrence of species in the catch that do not truly co-occur in terms of habitat associations informative for an index of abundance. Stpehens and MacCall (Stephens and MacCall 2004) recommended including all trips above a threshold where the false negatives and false positives are equally balanced. However, this does not have any biological relevance and for this data set, we assume that if a GBYR was landed, the anglers had to have fished in appropriate habitat, especially given how territorial GBYR and both species are strongly associated with rocky habitat.

Two levels of possible filtering were applied using the Stephens-MacCall filter (Table ??tab:Fleet10_11_Filter}). The Stephens-MacCall filtering method identified the probability of occurrence (in this case 0.4) at which the rate of "false positives" equals "false negatives." The trips selected using this criteria were compared to an alternative method including all the "false positive" trips, regardless of the probability of encountering GBYR (Table 19). This assumes that if GBYR were caught, the anglers must have fished in appropriate habitat during the trip. The catch included in this index is "sampler-examined" and the samplers are well trained in species identification. The last filter applied was to exclude years after 1999 due to a number of regulation changes, and years in which there were less than 20 observed trips. The final index is represented by 544 trips, 220 of which encountered GBYR.

Due to the large number of zeros in the data, we modeled catch per angler hour (CPUE; number of fish per angler hour) using maximum likelihood and Bayesian negative binomial regression. Models incorporating temporal (year, 2-month waves) and geographic (region and area_x) factors were evaluated. Counties were grouped into three regions, north of Sonoma county, Sonoma county through Santa Cruz county, and San Luis Obispo county. Based on AIC values from maximum likelihood fits (Table 17), a main effects model including all factors (year, region, area_x, and 2-month waves) was fit in the "rstanarm" R package (version 2.18.2). Diagnostic checks of the Bayesian model fit (Neff, Rhat, and Monte Carlo standard error values) were all reasonable. Predicted means by stratum (Year) were strongly correlated with observed means, suggesting a reasonable fit to the data (Figure 15). The NB model generated data sets with roughly 50-70% zeros, compared to the observed 60% (Figure 16).

The index represents the years 1984-1989, 1995, 1996 and 1999. There is not a lot of contrast in the index, except for a small increase in 1986. The final index values and associated log standard error included in the assessment can be found in Table 6.

MRFSS filtering and index standardization for south of Pt. Conception Prior to the Stephens-MacCall filter, a total of 7,334 trips were available for the analysis. As expected, positive indicators of GBYR trips included several nearshore species, e.g., kelp rockfish, treefish, and black croaker, while the strongest counter-indicator was opaleye (Figure 17). While the filter is useful in identifying co-occurring or non-occurring species assuming all effort was exerted in pursuit of a single target, the targeting of more than one target species can result in co-occurrence of species in the catch that do not truly co-occur in terms of habitat associations informative for an index of abundance.

For consistency with the methods used north of Pt. Conception (Table **??**tab:Fleet10_11_Filter}) the index includes the trips identified as "false positives" from the Stephens-MacCall filtering that had a lower threshold level of 0.22 (Table 20). The last filter applied was to exclude years after 1999 due to a number of regulation changes, and years in which there were less than 20 observed trips. The final index is represented by 475 trips, 342 of which encountered GBYR.

Catch per angler hour (CPUE; number of fish per angler hour) was modelled using the delta-GLM approach (Lo et al. 1992, Stefánsson 1996). A negative binomial model was explored, but the proportion of zeroes was not well estimated in the negative binomial models. This is likely due to the facts that MRFSS sampling effort was higher south of Pt. Conception, and GBYR are also rare south of Pt. Conception, both leading to a higher proportion of zeroes in the trip data than for north of Pt. Conception.

Model selection using Akaike Information Criterion (AIC) supported inclusion of year, region, area_x, and 2-month waves. Counties were grouped into three regions,Santa Barbara to Ventura counties, Los Angeles and Orange counties, and San Diego county for both the positive observation model and the binomial model. Area_x is a measure of distance from shore, a categorical variable indicating whether most of the fishing occurred inside or outside three nautical miles from shore.

The resulting index for south of Pt. Conception represents different years than the index for north of Pt. Conception (Table 6). The index starts in 1980 with continuous data through 1986, and three additional years in 1996, 1998 and 1999.
The index increases through 1983 and a marked decrease to 1986. The index for the three years in the 1990s does not exhibit any significant trend.

CPFV Onboard Observer Surveys

Onboard observer survey data were available from three sources for this assessment, 1) a CDFW survey in central California from 1987-1998 (referred to as the Deb Wilson-Vandenberg survey, (Reilly et al. 1998)), 2) the CDFW survey from 1999-2018, and 3) a Cal Poly survey from 2003-2018. During an onboard observer trip the sampler rides along on a CPFV trip and records location-specific catch and discard information to the species level for a subset of anglers onboard the vessel. The subset of observed anglers is usually a maximum of 15 people that changes during each fishing stop. The catch cannot be linked to an individual, but rather to a specific fishing location. The sampler also records the starting and ending time, number of anglers observed, starting and ending depth, and measures discarded fish. The fine-scale catch and effort data allow us to better filter the data for indices to fishing stops within suitable habitat for the target species.

California implemented a statewide sampling program in 1999 (Monk et al. 2014). California Polytechnic State University (Cal Poly) has conducted an independent onboard sampling program as of 2003 for boats in Port San Luis and Morro Bay (Stephens and MacCall 2004), but follows the protocols established in Reilly et al. (1998), and has been modified to reflect sampling changes that CDFW has also adopted, e.g., observing fish as they are landed instead of at the level of a fisher's bag. Therefore, the Cal Poly data area incorporated in the same index as the CDFW data from 1999-2018. Cal Poly collects lengths of both retained and discarded fish.

We generated separate relative indices of abundance in for the 1987-1999 and 2000-2018 data sets due to the number of regulation changes occurring throughout the time period (see Appendix H, p.**??**). Separate indices were also developed for north and south of Pt. Conception.

Deb Wilson-Vandenberg Onboard Observer Index Filtering and Standardizatiom The specific fishing locations at each fishing stop were recorded at a finer scale than the catch data for this survey. We aggregated the relevant location information (time and number of observed anglers) to match the available catch information. Between April 1987 and July 1992 the number of observed anglers was not recorded for each fishing stop, but the number of anglers aboard the vessel is available. We imputed the number of observed anglers using the number of anglers aboard the vessel and the number of observed anglers at each fishing stop from the August 1992-December 1998 data (see Supplemental materials for details). In 1987, trips were only observed in Monterey, CA and were therefore excluded from the analysis. The years 1990 and 1991 were also removed for low sample sizes. Final data filters included removing reefs that never encountered GBYR, drifts that had fishing times outside 95% of the data, and fishing stops with depths <9 m and >69m. The final data set contained 2,411 fishing stops, with 1,096 of those encountering GBYR (Figure 18).

The index was fit using a delta-GLM model, with a lognormal model (AIC: 1,088) selected over a gamma model (AIC: 1,143) for the positive encounters. Covariates considered in the full model included year, depth, and month (Table 8). The model selected by AIC for both the lognormal and binomial components of the delta-GLM included year, depth and reef. Depth was included in 10 m depth bins and eight reefs were select in the final model (Figure coming). The final index did not indicate an increasing trend that was seen in the 2005 gopher rockfish assessment using the same data set (Figure 19). A number of reasons include that finer-scale location data was keypunched in 2012 for this survey, the index in this assessment includes black-and-yellow rockfish, and different filters were applied to the data. However, the the same peaks and decreases in the two indices are present.

CDFW and Cal Poly Onboard Observer Index Filtering and Standardization As described above the CDFW and Cal Poly onboard observer programs are identical in that the same protocols are followed. The only difference is that Cal Poly measures both retained and discarded fish from the observed anglers. CDFW measures discarded fish only from the observed anglers, and measures retained fish as part of the angler interview at the bag level. Cal Poly has also begun collecting otoliths during the onboard observer trips, which are used as conditional age-at-length data the recreational fishery north of Pt. Conception in this assessment.

A number of filters are applied to these data. All of the Cal Poly data have been through a QA/QC process once key-punched, whereas a number of errors remain in the data from CDFW. Data sheets from CDFW are not longer available prior to 2012 and staff constraints have also prevented a quality control review of the data.

Each drift was assigned to a reef (hard bottom). Hard bottom was extracted from the California Seafloor Mapping Project, with bathymetric data from state waters available at a 2 m resolution. Reefs were developed based on a number of factors described in the supplemental material (xxx).

Initial filters were applied to the entire data set, north and south of Pt. Conception combined. After an initial clean-up of the data, 67,850 drifts remained, with GBYR present in 9,317 (Table 9). This was reduced to 25,427 drifts GBYR present in 7,250 drifts) after filtering the data to remove potential outliers in the time fished and observed anglers, limiting the data to reefs that observed GBYR and were sampled in at least 2/3 of all years, and to drifts with starting locations within 1,000 m of a reef.

Recreational fishing trips north and south of Pt. Conception can be fundamentally different due to differences in habitat structure, target species, and weather.

Filtering and Index Standardization for north of Pt. Conception The number of drifts remaining before region specific filtering was 13,792, with 6,036 drifts encountering GBYR

(Table 9).

Because GBYR are strongly associated with hard bottom habitat, the distance from a reef at the start of a drift was re-examined for drifts encountering GBYR. The maximum distance was 872 m, but the 97% quantile dropped to 42 m and was chosen as a reasonable cutoff value, and only resulted in a reduction of 182 drifts that encountered gopher rockfish. The final data were filtered to ensure all selected reefs were sampled in at least 2/3 of all years, leaving 12,965 drifts for the final index, 5,796 of which encountered GBYR (Figure 20).

The index of abundance was modeled with the a delta-GLM modeling approach, with year, month, 10 m depth bins from 10-59 m, and 12 reefs as possible covariates. A lognormal model (AIC: 12,185) was selected over a gamma (AIC: 12,520) for the positive observations using AIC. The full model was selected by AIC for the lognormal and binomial components of the delta-GLM. The index indicates a relatively stable trend from 2001-2009 and a steady decrease from 2010-2013. The relative index of abundance has increased since 2014.

Filtering and Index Standardization for south of Pt. Conception The bathymetric data is not available at as fine-scale resolution for the Southern California Bight and more of the trips and drifts target mid-water species, including mid-water rockfish (Table 9). Therefore, instead of using distance to reef as a filter, we filtered the data to drifts that encountered 20% or more groundfish. This resulted in the total number of drifts decreasing from 11,635 to 5,495, but only decreased the number of drifts encountering GBYR from 1,277 to 1,171. A final check was made to ensure all reefs were sampled in at least 2/3 of all years, leaving 5,440 drift for the final index, of which 1,132 encountered GBYR (Figure 21).

The index of abundance was modeled with the a delta-GLM modeling approach, with year, month, 10 m depth bins from 10-59 m, and four reefs as possible covariates. A lognormal model (AIC: 162) was selected over a gamma (AIC: 277) for the positive observations using AIC. A model with year, depth and reef was selected by AIC for both the lognormal and binomial components of the delta-GLM. The index indicates a relatively stable trend from 2001-2004 and a steady increase from 2005-2017.

2.1.7 Fishery-Depenent Indices: Availabe Length and Age Data fishery-depenent-indices-availabe-length-and-age-data

Length data associated with the MRFSS dockside CPFV survey and the current onboard observer surveys conducted by CDFW are incorporated into the biological data pulled from the respective data sources, MRFSS and CRFS. The additional length data are not incorporated as separate length composition data as they represent the same portion of the population sampled by the CDFW onboard observer program.

Cal Poly collected ototlihs from the onboard observer program starting in 2017 as part of a special study to correlate fish length before and and after the fish was filleted by the deckhands onboard the CPFV vessels. All fish collected in 2017 only had associated post-fillet lengths and were not used in the assessment since the study has not been finalized nor has the method been endorsed by the SSC. A subset of fish form the 2018 collection included both pre- and post-fillet length and were used in the assessment as conditional age-at-length data associated with the recreational fleet north of Pt. Conception.

Length composition from Deb Wilson-Vandenberg's onboard observer survey are included in the asessment. This program measured both retained and discarded fish, and represent the portion of the population sampled with the spatial extent of the index. This onboard observer program continued during the period from 1990-1992 when MRFSS was on hiatus.

2.1.8 Fishery-Independent Data Sources

fishery-independent-data-sources

Neither of the two fishery-independent surveys described below have previously been used in stock assessments as indices of abundance.

California Collaborative Fisheries Research Project The California Collaborative Fisheries Research Project, CCFRP, is a fishery-independent hook-and-line survey designed to monitor nearshore fish populations at a series of sampling locations both inside and adjacent to MPAs along the central California coast (Wendt and Starr 2009, Starr et al. 2015). The CCFRP survey began in 2007 and was originally designed as a statewide program in collaboration with NMFS scientists and fishermen. From 2007-2016 the CCFRP project was focused on the central California coast, and has monitored four MPAs consistently since then (Figure 22). In 2017, the program was expanded coastwide within California. The index of abundance was developed from the four MPAs sampled consistently (A˜{n}o Nuevo and Point Lobos by Moss Landing Marine Labs; Point Buchon and Piedras Blancas by Cal Poly).

The survey design for CCFRP consists a number 500 x 500 m cells both within and outside each MPA. On any given survey day for a chose site cells are randomly selected within a stratum (MPA and/or reference cells). CPFVs are chartered for the survey and the fishing captain is allowed to search within the cell for a fishing location. During a sampling event, each cell is fished for a total of 45? minutes by volunteer anglers. Each fish encountered is recorded, measured, and can be linked back to a particular angler, and released (or descended to depth). Starting in 2017, a subset of fish have been retained to collect otoliths and fin clips, needed biological information for nearshore species. For the index of abundance, CPUE was modelled at the level of the drift, similar to the fishery-dependent onboard observer survey described above.

The CCFRP data are quality controlled at the time they are key punched and little filtering was needed for the index. Cells not consistently sampled over time were excluded as well as cells that never encountered GBYR. CCFRP samples shallower depths to avoid barotrauma-induced mortality. The index was constrained to 5-39m in 5 m depth bins. The final index included 4,920 drifts, 3,848 of which encountered GBYR.

We modeled catch per angler hour (CPUE; number of fish per angler hour) using maximum likelihood and Bayesian negative binomial regression. The proportion of zeroes in this data was relatively small (22%), and if overdispersion were not present, the regression would innately become Poisson. Models incorporating temporal (year, month) and geographic (MPA site and MPA vs Reference cells) factors were evaluated. Based on AIC values from maximum likelihood fits (Table 15), a main effects model including all factors (year, month, site and MPA/REF) was fit in the "rstanarm" R package (version 2.18.2). Diagnostic checks of the Bayesian model fit (Neff, Rhat, and Monte Carlo standard error values) were all reasonable. Predicted means by stratum (Year) were strongly correlated with observed means, suggesting a reasonable fit to the data (Figure 23). The NB model generated data sets with roughly 18-22% zeros, compared to the observed 22% (Figure 16).

The CCFRP index of abundance closely matches the trend observers in the onboard observer index from 2009-2018 (Figure 11). The index decreases from 2009 to 2013, and then exhibits the same increase through 2018. When both indices are standardized to their means, the values for 2013 and 2018 are the same.

CCFRP Length Measurements and Available Ages

The CCFRP program measures every fish encountered to the nearsest half centimeter. A total of 22,470 GBYR were measured by CCFRP from 2007-2018, of which only 212 were black-and-yellow rockfish. The length distributions for each of the four MPAs used in the index for this assessment show slight difference in their peaks (Figure 25). A˜{n}o Nuevo is the most northern site and Point Buchon the most southern.

Conditional age-at-lenght data were also incorporated into the assessment from the CCFRP program, including two master's theses that are products of the CCFRP. Erin Loury (Loury 2011) collected gopher rockfish otoliths as part of her thesis work from 2007-2009 that included specimens from both inside and outside the MPAs. Natasha Meyers-Cherry (Meyers-Cherry 2014) conducted another thesis focused on the life history of gopher rockfish and collected otoliths from 2011-2012, also both inside and outside the MPAs. Both MLML and Cal Poly began routinely collecting otoliths from a select number of fish in 2017 as part of the CCFRP program. Also included in the conditional age-at-length dat for this fleet are ototliths collected in 2018 by the University of California Davis Bodega Marine Lab CCFRP program.

Partnership for Interdisciplinary Studies of Coastal Oceans The Partnership for Interdisciplinary Studies of Coastal Oceans, PISCO, conducts a number of surveys to monitor the kelp forests, one of which is a kelp forest fish survey. PISCO has monitored fish population in the 0-20 m depth range as part of the Marine Life Protection Act (MLPA) since 1998. Paired sites inside and outside MPAs are surveyed to monitor the long-term dynamics of the kelp forest ecosystem and provide insight into the effect of MPAs on kelp forest species. PISCO conducts the fish surveys from late July through September. At each site, benthic, midwater, and canopy scuba transcets are conducted at 5, 10, 15, and 20 m depth. All divers are trained in species identification. Along each 30 m transect, divers enumerate all identifiable non-cryptic fish, and measure total length to the nearest centimeter. PISCO surveys are conducted by the University of California Santa Cruz (UCSC) in central

California and through the University of California Santa Barbara in southern California. All PISCO data were provided by Dan Malone (UCSC).

The majority of filtering for the PISCO data set was to determine which sites to keep for the final index. After initial filtering the data for GBYR in southern California were too sparse to be considered for the index of abundance. Gopher and black-and-yellow rockfish were also rarely observed in the midwater and canopy transects, and therefore the index is based only on the benthic transects. Only sites sampled consistently throughout the time period 2001-2018 were kept for the index. Multiple transects can be conducted along the same line within a sampling event. All transects within a site were combined and effort was modeled as the number of transects represented in the number of fish observed. The final index included 3,231 transects, of which 1,729 observed GBYR (Figure 26).

We modeled number of fish observed per transect(s) using maximum likelihood and Bayesian negative binomial regression. Models incorporating temporal (year, month) and geographic (site and zone) factors were evaluated. The zone is a factor indicating the depth stratification at a site, i.e., 5 m,10 m,15 m, or 20 m targetted bottom depth.. Based on AIC values from maximum likelihood fits (Table 13), a main effects model including all factors (year, month, site and zone) was fit in the "rstanarm" R package (version 2.18.2). Diagnostic checks of the Bayesian model fit (Neff, Rhat, and Monte Carlo standard error values) were all reasonable. Predicted means by stratum (Year) were strongly correlated with observed means, suggesting a reasonable fit to the data (Figure 27). The NB model generated data sets with roughly 16-25% zeros, compared to the observed 23% (Figure 28).

The final index decreases from 2001 to the late 2000s, with lower estiamtes of relative abundance from 2005-2010. From 2010 to 2015, the index increaes and peaks in 2015, before the decreasing trends from 2016-2018. The pattern observed in this index indicates almost an counter trend to that observed in the onboard observer and CCFRP indices for north of Pt. Conception (Figure 11). The PISCO survey is sampling different habitat types than the other two surveys, and covers much shallower depths.

PISCO Length Measurements

Every GBYR observed by PISCO divers except one, had an associated length measurement (N = 11,965). Divers measure fish to the nearest centimeter, and are trained to measure fish underwater and be aware of possible biases, e.g., ambient light, body color, visibility, and body shape. Both juvenile and adult GBYR were observed in the PISCO kelp forest fish survey data (Figure **??**). Of note is the similarity in length distibutions both between the species and for the two species combined across sites. Fish in the 10-17 cm size range (approximately) are not observed in this survey. There is significant post-settlement mortality for both species, which is thought to be due to density-dependent predation (Johnson 2006, 2007). Secondly, both species are cryptic and observed more often at night than during the day (Mark Carr, PISCO-UCSC, per comm.).

2.1.9 Biological Parameters and Data

biological-parameters-and-data

Neither gophr nor black-and-yellow rockfish have forked tails, therefore total legnth and fork length are equal. All of the data provided for this assessment were either in fork length or total length.

(Table 4)

Length and Age Compositions

Length compositions were provided from the following sources:

- CALCOM (*commercial retained dead fish*, 1987, 1992-2018)
- WCGOP (*commercial discarded fish*, 2004-2018)
- Deb Wilson-Vandenber's onboard observer survey (*recreational charter retained and discarded catch*, 1987-1998)
- California recreational sources combined (*recreational charter retained catch*)
  - Miller and Gotshall dockside survey (1959-1966)
  - Ally et al. onboard observer survey (1985-1987)
  - Collins and Crooke onboard observer survey (1975-1978)
  - MRFSS dockside survey (1980-2003)
  - CRFS onboard and dockside survey (2004-2018)
- PISCO dive survey (*research*, 2001-2018)
- CCFRP hook-and-line survey (*research*, 2007-2018)

The length composition of all fisheries aggregated across time by fleet is in Figure 30. Descriptions and details of the length composition data are in the above section for each fleet or survey.

Age Structures

A total of 2,467 otoliths were incorporated in this assessment and a summary by source can be found in Table 21. Gopher rockfish comprised 80% of the samples (946 females, 901 males, 121 unknown sex), and all but a few black-and-yellow rockfish (247 females, 232 males, 20 unknown sex) came from a directed study by Jody Zaitlin (1986) (Figure 31).

Of the available ages, 91% were collected during fishery-independent surveys.
An additional 36 otoliths were collected by Cal Poly during their CPFV onboard observer survey in 2018. The remaining 7.5% were from commercial port samples or recreational dockside surveys. Black-and-yellows represent 20% of the samples collected, and are mainly derived from Ralph Larson's work in Monterey Bay.

All otoliths were read by Don Pearson (NMFS SWFSC, now retired) and ages ranged from 1-28. The aged black-and-yellow rockfish ranged in length from 7-32 cm with a mean of 24 cm and gopher rockfish ranged in length from 11-36 cm, with a mean of 26. In terms of ages, the black-and-yellow rockfish ranged from 2-19 and gophers from 2-28. Fits to the von Bertalanffy growth curve (Bertalanffy 1938), $L_i = L_\infty e^{(-k[t-t_0])}$, where L_i is the length (cm) at age i, t is age in years, k is rate of increase in growth, t_0 is the intercept, and L_∞ is the asymptotic length, were explore by species and sex.

No significant differences were found in growth between males and females, or between gopher and black-and-yellow rockfishes.

Aging Precision and Bias

Uncertainty in ageing error was estimated using a collection of xxx gopher and black-and-yellow rockfish scorpionfish otoliths with two age reads (Figure **??**). Age-composition data used in the model were from a number of sources described above. All otoliths were read by Don Pearson (NMFS SWFSC, no retired) who also conducted all blink double reads.

Ageing error was estimated using publicly available software (Thorson et al. 2012). The software setting for bias and standard deviation were the same for both readers, (unbiased and curvilinear increase in standard deviation) with age, respectively (Figure **??**).

The resulting estimate indicated a standard deviation in age readings increasing from xxx years to a standard deviation of xxx years at age 28.

Weight-Length

The weight-length relationship is based on the standard power function: $W = \alpha(L^\beta)$ where W is individual weight (kg), L is length (cm), and α and β are coefficients used as constants.

The weight-length relationships was estimated from the three studies, Loury (2011), Meyers-Cherry (2014) (both gopher rockfish only from CCFRP) and Zaitlin (Zaitlin 1986) (black-and-yellow rockfish only). Only one weight-length relationship was estimated for the GBYR complex. The estimated parameters are $\alpha = 8.84e^{-006}$ and$\beta = 3.25584$. The estimated relationship is similar to that estimated by Lea et al. (1999) for gopher rockfish (Figure 33). The weight-length relationship estimated here was used in the assessment model to best represent the GBYR complex.

Sex Ratio, Maturity, and Fecundity

The sex ratio for GBYR is assumed to be 50:50 as there is no evidence to suggest otherwise.

Zaitlin (1986) found that females reached 50% maturity at 17.5 cm or 4 years of age in Central California and were 100% mature by age 6, with the same age of maturity found in southern California though individuals were smaller at age; all ages from this study were surface ages and have been re-read for this assessment. Echeverria (1987) estimated maturity for 17 rockfish species in central California. She found the size at first maturity and the size at 50% maturity for male and female gopher rockfish to be 17 cm total length, and 100% mature by 21 cm. Black-and-yellow rockfish males and females were first mature at 14 cm, 50% of females were mature at 15 cm, 50% of males mature at 16 cm. Male black-and-yellow rockfish were 100% mature at 20 cm and females at 19 cm. In southern California waters, both males and females were found to reach first maturity at 13 cm total length (Larson 1980). We did not have any samples from southern California to re-analyze the maturiy ogive for that portion of the population.

The maturity data from Zaitlin (1986) (422 black-and-yellow rockfish) were re-analyzed along with samples from Meyers-Cherry (2014) (115 gopher rockfish). Combining the two dataset provided an updated maturity ogive for the GBYR complex females (Figure **??**). The first observed mature fish was 19 cm and the length at 50% was 21.66 cm, larger than suggested from the studies in the 1980s. Using only the gopher data, 50% was estimated at 23.33 cm and for black-and-yellow rockfish at 21.26 cm. An important note is that the smaller fish from these studies were black-and-yellow rockfish and the larger fish were gopher rockfish. Alhtough not used in this assessment, the estimate of 50% maturity for 23 GBYR from these studies was 21.88 cm. The age at 50% mature increased in this assessment to 21.66 cm, which is 3.96 cm larger than the value used int eh 2005 assessment.

Mature females in central California release larvae between January and July, peaking in February, March, and May (Larson 1980, Lea et al. 1999, Love et al. 2002).Both species of GBYR release one brood per season (Love et al. 2002). Black-and-yellow rockfish females can produce 25,000 - 450,000 eggs spawning from January to May. Gopher rockfish females ranging between 176 and 307 grams carry approximately 249 eggs per gram of body weight (MacGregor 1970). The larval stage lasts one to two months (Moser 1996) and it may take up to 90 days before the larvae settle out of the plankton at around 20 mm total length (Lea et al. 1999). The fecundity estimates used in this assessment were provided by E.J. Dick (NMFS SWFSC) from a meta-analysis of fecundity in the genus *Sebastes* (Dick et al. 2017).

Natural Mortality Hamel (2015) developed a method for combining meta-analytic approaches to relating the natural mortality rate M to other life-history parameters such as longevity, size, growth rate and reproductive effort, to provide a prior on M. In that same issue of ICESJMS, Then et al. (2015), provided an updated data set of estimates of M and related life history parameters across a large number of fish species, from which to develop an M estimator for fish species in general. They concluded by recommending M estimates be based on maximum age alone, based on an updated Hoenig non-linear least squares (nls) estimator $M = 4.899 * {A_{max}}^{-.916}$. The approach of basing M priors on maximum age alone was one that was already being used for west coast rockfish assessments. However, in fitting the alternative model forms relating $-.916M$ to A_{max}, Then et al. (2015) did not consistently apply their transformation. In particular, in real space, one would expect substantial heteroscedasticity in both the observation and process error associated with the observed relationship of M to A_{max}. Therefore, it would be reasonable to fit all models under a log transformation. This was not done. Revaluating the data used in Then et al. (2015) by fitting the one-parameter A_{max} model under a log-log transformation (such that the slope is forced to be -1 in the transformed space (as in Hamel (2015)), the point estimate for M is:

$$M = \frac{5.4}{A_{max}} \tag{1}$$

The above is also the median of the prior. The prior is defined as a lognormal with mean $ln\frac{5.4}{A_{max}}$ and SE = 0.4384343 (Owen Hamel, personal communication, NMFS). Using a maximum age of 28 the point estimate and median of the prior is 0.211138, which is used as a prior for in the assessment model.

2.1.10 Environmental or Ecosystem Data Included in the Assessment

`environmental-or-ecosystem-data-included-in-the-assessment`

In this assessment, neither environmental nor ecosystem considerations were explicitly included in the analysis. This is primarily due to a lack of relevant data and results of analyses (conducted elsewhere) that could contribute ecosystem-related quantitative information for the assessment.

2.2 Previous Assessments

`previous-assessments`

2.2.1 History of Modeling Approaches Used for this Stock

`history-of-modeling-approaches-used-for-this-stock`

2.2.2 yyyy Assessment Recommendations

`yyyy-assessment-recommendations`

Recommendation 1:

STAT response: xxxxx

Recommendation 2:

STAT response: xxxxx

Recommendation 3:

STAT response: xxxx

2.3 Model Description

model-description

2.3.1 Transition to the Current Stock Assessment

transition-to-the-current-stock-assessment

2.3.2 Summary of Data for Fleets and Areas

summary-of-data-for-fleets-and-areas

There are xxx fleets in the base model. They include:

Commercial: The commercial fleets include ...

Recreational: The recreational fleets include ...

Research: There are xx sources of fishery-independent data available ...

2.3.3 Other Specifications

other-specifications

2.3.4 Modeling Software

modeling-software

The STAT team used Stock Synthesis 3 version 3.30.05.03 by Dr. Richard Methot at the NWFSC. This most recent version was used, since it included improvements and corrections to older versions. The r4SS package (GitHub release number v1.27.0) was used to post-processing output data from Stock Synthesis.

2.3.5 Data Weighting

data-weighting

2.3.6 Priors

priors

The log-normal prior for female natural mortality were based on a meta-analysis completed by Hamel (2015), as described under "Natural Mortality." Female natural mortality was fixed at the median of the prior, 0.xxx for an assumed maximum age of xx. An uninformative prior was used for the male offset natural mortality, which was estimated.

The prior for steepness (h) assumes a beta distribution with parameters based on an update for the Thorson-Dorn rockfish prior (Dorn, M. and Thorson, J., pers. comm.), which was endorsed by the Science and Statistical Committee in 2018. The prior is a beta distribution with mu=0.xxx and $sigma$=0.xxx. Steepness is fixed in the base model at the mean of the prior. The priors were applied in sensitivity analyses where these parameters were estimated.

2.3.7 Estimated and Fixed Parameters

`estimated-and-fixed-parameters`

A full list of all estimated and fixed parameters is provided in Tables 23.

The base model has a total of xxx estimated parameters in the following categories:

- xxx,
- xxx
- xxx, and
- xxx selectivity parameters

The estimated parameters are described in greater detail below and a full list of all estimated and parameters is provided in Table 23.

Growth.

Natural Mortality.

Selectivity.

Other Estimated Parameters.

Other Fixed Parameters.

2.4 Model Selection and Evaluation

`model-selection-and-evaluation`

2.4.1 Key Assumptions and Structural Choices

`key-assumptions-and-structural-choices`

2.4.2 Alternate Models Considered

`alternate-models-considered`

2.4.3 Convergence

`convergence`

2.5 Response to the Current STAR Panel Requests

`response-to-the-current-star-panel-requests`

Request No. 1:

Rationale: xxx

STAT Response: xxx

Request No. 2:

Rationale: xxx

STAT Response: xxx

Request No. 3:

Rationale: x.

STAT Response: xxx

Request No. 4:

Rationale: xxx

STAT Response: xxx

Request No. 5:

Rationale: xxx

STAT Response: xxx

2.6 Base Case Model Results

`base-case-model-results`

The following description of the model results reflects a base model that incorporates all of the changes made during the STAR panel (see previous section). The base model parameter estimates and their approximate asymptotic standard errors are shown in Table 23 and the likelihood components are in Table 24. Estimates of derived reference points and approximate 95% asymptotic confidence intervals are shown in Table e. Time-series of estimated stock size over time are shown in Table 25.

2.6.1 Parameter Estimates

`parameter-estimates`

The additional survey variability (process error added directly to each year's input variability) for all surveys was estimated within the model.

(Figure **??**).

The stock-recruit curve . . . Figure 36 with estimated recruitments also shown.

2.6.2 Fits to the Data

fits-to-the-data

Model fits to the indices of abundance, fishery length composition, survey length composition, and conditional age-at-length observations are all discussed below.

2.6.3 Uncertainty and Sensitivity Analyses

uncertainty-and-sensitivity-analyses

A number of sensitivity analyses were conducted, including:

1. Sensitivity 1
2. Sensitivity 2
3. Sensitivity 3
4. Sensitivity 4
5. Sensitivity 5, etc/

2.6.4 Retrospective Analysis

retrospective-analysis

2.6.5 Likelihood Profiles

likelihood-profiles

2.6.6 Reference Points

reference-points-1

Reference points were calculated using the estimated selectivities and catch distribution among fleets in the most recent year of the model, (2017). Sustainable total yield (landings plus discards) were 167 mt when using an $SPR_{50\%}$ reference harvest rate and with a 95% confidence interval of 101 mt based on estimates of uncertainty. The spawning biomass equivalent to 40% of the unfished level ($SB_{40\%}$) was 527 mt.

(Figure 46

The 2018 spawning biomass relative to unfished equilibrium spawning biomass is above/below the target of 40% of unfished levels (Figure **??**). The relative fishing intensity, $(1 - SPR)/(1 - SPR_{50\%})$, has been xxx the management target for the entire time series of the model.

Table e shows the full suite of estimated reference points for the base model and Figure 48 shows the equilibrium curve based on a steepness value xxx.

3 Harvest Projections and Decision Tables

harvest-projections-and-decision-tables

This section will be completed at or after the STAR panel.

4 Regional Management Considerations

regional-management-considerations

While the proportion of the stock residing within U.S. waters is unknown, the assessment provides an adequate geographic representation of the portion assessed for management purposes. There is little evidence that black-and-yellow rockfish extend into Mexico, and the proportion of gopher rockfish residing south of Pt. Conception is small. while there has been work on the genetic structure of the two species, there has not been work done within species to inform whether or not there is genetic spatial structure to the population. Given the relatively small area in the waters of California where these species occur, there is relatively little concern regarding exploitation in proportion to the regional distribution of abundance in the area assessed in this study.

The state of California implements regional management for the recreational fleet in the form of five regions. Neither gopher nor black-and-yellow rockfish are a large componenet of the total recreational landings and are managed as part of a shallow nearshore rockfish complex. Current regional management appears appropriate for these species.

5 Research Needs

research-needs

This section will be completed at or after the STAR panel.

6 Acknowledgments

acknowledgments

This section will be completed at or after the STAR panel.

7 Tables

tables

Table 1: Commercial landings and discards (mt) from the commercial fisheries. Data sources are the California Catch Reconstruction, CALCOM, PacFIN, and WCGOP GEMM report.

Year	Landings	Discards	Total Commercial Removals	Source
1916	3.88	0.38	4.27	Catch Reconstruction
1917	6.03	0.59	6.63	Catch Reconstruction
1918	7.06	0.69	7.75	Catch Reconstruction
1919	4.91	0.48	5.39	Catch Reconstruction
1920	5.01	0.49	5.50	Catch Reconstruction
1921	4.13	0.41	4.54	Catch Reconstruction
1922	3.56	0.35	3.90	Catch Reconstruction
1923	3.84	0.38	4.22	Catch Reconstruction
1924	2.22	0.22	2.44	Catch Reconstruction
1925	2.78	0.27	3.05	Catch Reconstruction
1926	4.48	0.44	4.92	Catch Reconstruction
1927	3.81	0.37	4.18	Catch Reconstruction
1928	4.60	0.45	5.06	Catch Reconstruction
1929	3.81	0.37	4.18	Catch Reconstruction
1930	5.40	0.53	5.93	Catch Reconstruction
1931	1.93	0.19	2.11	Catch Reconstruction
1932	6.24	0.61	6.85	Catch Reconstruction
1933	2.58	0.25	2.84	Catch Reconstruction
1934	1.75	0.17	1.92	Catch Reconstruction
1935	0.43	0.04	0.47	Catch Reconstruction
1936	0.01	0.00	0.01	Catch Reconstruction
1937	7.27	0.71	7.98	Catch Reconstruction
1938	10.29	1.01	11.30	Catch Reconstruction
1939	13.13	1.29	14.42	Catch Reconstruction
1940	16.90	1.66	18.56	Catch Reconstruction
1941	17.06	1.67	18.73	Catch Reconstruction
1942	8.55	0.84	9.38	Catch Reconstruction
1943	11.00	1.08	12.08	Catch Reconstruction
1944	0.05	0.00	0.05	Catch Reconstruction
1945	0.59	0.06	0.65	Catch Reconstruction
1946	16.71	1.64	18.35	Catch Reconstruction
1947	26.71	2.62	29.33	Catch Reconstruction
1948	23.95	2.35	26.30	Catch Reconstruction
1949	18.29	1.79	20.09	Catch Reconstruction
1950	17.15	1.68	18.83	Catch Reconstruction
1951	24.83	2.44	27.26	Catch Reconstruction

Continues next page

tab:CommCatches

Table 1: Commercial landings and discards (mt) from the commercial fisheries. Data sources are the California Catch Reconstruction, CALCOM, PacFIN, and WCGOP GEMM report.

Year	Landings	Discards	Total Commercial Removals	Source
1952	27.59	2.71	30.29	Catch Reconstruction
1953	32.30	3.17	35.47	Catch Reconstruction
1954	40.75	4.00	44.74	Catch Reconstruction
1955	29.49	2.89	32.38	Catch Reconstruction
1956	40.66	3.99	44.65	Catch Reconstruction
1957	37.52	3.68	41.20	Catch Reconstruction
1958	33.56	3.29	36.86	Catch Reconstruction
1959	19.62	1.92	21.54	Catch Reconstruction
1960	11.30	1.11	12.41	Catch Reconstruction
1961	17.49	1.72	19.20	Catch Reconstruction
1962	27.18	2.67	29.85	Catch Reconstruction
1963	22.29	2.19	24.48	Catch Reconstruction
1964	16.55	1.62	18.17	Catch Reconstruction
1965	21.50	2.11	23.61	Catch Reconstruction
1966	13.44	1.32	14.76	Catch Reconstruction
1967	6.70	0.66	7.36	Catch Reconstruction
1968	8.29	0.81	9.10	Catch Reconstruction
1969	9.99	0.98	10.97	CALCOM
1970	14.21	1.39	15.60	CALCOM
1971	14.41	1.41	15.83	CALCOM
1972	19.42	1.91	21.33	CALCOM
1973	31.43	3.08	34.51	CALCOM
1974	33.41	3.28	36.69	CALCOM
1975	33.08	3.25	36.33	CALCOM
1976	33.90	3.33	37.23	CALCOM
1977	30.13	2.96	33.09	CALCOM
1978	43.41	4.26	47.67	CALCOM
1979	34.24	3.36	37.60	CALCOM
1980	63.65	6.24	69.89	CALCOM
1981	52.71	5.17	57.87	PacFIN
1982	38.97	3.82	42.79	PacFIN
1983	28.67	2.64	31.30	PacFIN
1984	16.74	1.45	18.20	PacFIN
1985	8.54	0.83	9.37	PacFIN
1986	25.16	2.50	27.66	PacFIN
1987	34.05	3.36	37.40	PacFIN
1988	54.98	5.47	60.44	PacFIN
1989	45.22	4.46	49.68	PacFIN

Continues next page

tab:CommCatches
Table 1: Commercial landings and discards (mt) from the commercial fisheries. Data sources are the California Catch Reconstruction, CALCOM, PacFIN, and WCGOP GEMM report.

Year	Landings	Discards	Total Commercial Removals	Source
1990	46.08	4.59	50.67	PacFIN
1991	67.98	6.75	74.73	PacFIN
1992	83.91	8.24	92.15	PacFIN
1993	73.43	7.27	80.70	PacFIN
1994	54.84	5.89	60.74	PacFIN
1995	91.10	8.97	100.07	PacFIN
1996	95.08	9.29	104.37	PacFIN
1997	69.99	6.81	76.80	PacFIN
1998	65.29	6.40	71.70	PacFIN
1999	62.65	6.15	68.80	PacFIN
2000	54.44	5.29	59.72	PacFIN
2001	53.76	5.24	59.00	PacFIN
2002	42.64	4.15	46.79	PacFIN
2003	21.08	13.04	34.12	PacFIN & WCGOP
2004	26.25	2.66	28.91	PacFIN & WCGOP
2005	28.67	3.33	31.99	PacFIN & WCGOP
2006	24.05	4.10	28.15	PacFIN & WCGOP
2007	30.36	4.50	34.87	PacFIN & WCGOP
2008	36.22	1.63	37.85	PacFIN & WCGOP
2009	35.62	5.38	40.99	PacFIN & WCGOP
2010	38.83	3.92	42.75	PacFIN & WCGOP
2011	42.39	5.72	48.12	PacFIN & WCGOP
2012	33.55	1.93	35.48	PacFIN & WCGOP
2013	33.45	2.85	36.31	PacFIN & WCGOP
2014	36.40	2.85	39.24	PacFIN & WCGOP
2015	43.25	2.93	46.18	PacFIN & WCGOP
2016	36.96	2.42	39.38	PacFIN & WCGOP
2017	42.04	1.65	43.68	PacFIN & WCGOP
2018	47.00	2.54	49.54	PacFIN & WCGOP

Table 2: Length composition sample sizes for fishery dependent data. Continuous years begin in 1975. Recreational north samples include Karpov et al., MRFSS, and CRFS data. Recreational south samples include Karpov et al., Collins and Crooke unpub., Ally et al. 1991, MRFSS, and CRFS data.

tab:length_samples_fishery

	CALCOM		WCGOP		Rec North		Rec South		Deb VW	
Year	Trips	Lengths	Trips	Lengths	Trips	Lengths	Trips	Lengths	Trips	Lengths
1959					27	271	2.10	21		
1960					39	394	1.40	14		
1961					1	8	0.10	1		
1966					1	7				
1975							50.00	159		
1976							73.00	224		
1977							96.00	392		
1978							91.00	533		
1979										
1980					4	164	21.00	53		
1981					1	19	30.00	100		
1982					1	50	17.00	58		
1983					6	323	60.00	170		
1984					14	849	42.00	150		
1985					35	1027	34.00	180		
1986					36	826	28.00	86		
1987	2	82			28	392	5.00	7	14	73
1988					30	303	10.00	30	54	664
1989					19	303	7.00	11	70	727
1990									17	109
1991									38	722
1992	56	671							55	838
1993	148	1648			14	1094	8.00	24	75	614
1994	170	1379			12	608	1.00	15	86	735
1995	174	1523							90	1171
1996	256	3270			74	607	14.00	32	100	1364
1997	140	1319			95	1424	7.00	23	107	1415
1998	206	2549			89	614	19.00	66	83	1048
1999	251	3283			49	1112	33.00	301		
2000	384	4918			21	695	12.00	58		
2001	142	2179			46	929	14.00	35		
2002	59	870			58	1656	22.00	65		
2003	55	625			72	1690	15.00	100		
2004	63	770	72	572	19	2023	3.00	42		
2005	72	700	42	260	30	3217	8.00	93		
2006	31	478	42	266	35	3737	9.00	106		
2007	80	1165	37	268	30	3200	10.00	126		
2008	46	503	12	46	39	4165	11.00	132		
2009	73	854	22	263	43	4612	15.00	184		
2010	75	925	37	344	47	4992	16.00	192		
2011	61	858	68	366	44	4692	22.00	270		
2012	57	709	69	302	46	4904	89.00	1081		
2013	48	581	56	348	40	4339	77.00	930		
2014	15	184	62	388	44	4746	49.00	595		
2015	48	578	93	521	54	5789	36.00	436		
2016	77	928	56	317	58	6265	37.00	444		
2017	67	1581	49	226	44	4691	39.00	478		
2018	67	1210			33	3563	26.00	317		

Table 3: Recreational removals (mt) of GBYR. Data sources are the California Catch Reconstruction (modified for south of Pt. Conception), MRFSS (modified for 1981-1982), and CRFS.

Year	North of Pt. Conception	South of Pt. Conception	Total Recreational Removals	Source
1928	0.84	0.02	0.85	Catch Reconstruction
1929	1.67	0.03	1.70	Catch Reconstruction
1930	1.92	0.05	1.97	Catch Reconstruction
1931	2.56	0.06	2.62	Catch Reconstruction
1932	3.20	0.08	3.28	Catch Reconstruction
1933	3.84	0.09	3.93	Catch Reconstruction
1934	4.48	0.11	4.59	Catch Reconstruction
1935	5.12	0.12	5.24	Catch Reconstruction
1936	5.76	0.22	5.98	Catch Reconstruction
1937	6.82	0.31	7.14	Catch Reconstruction
1938	6.71	0.41	7.12	Catch Reconstruction
1939	5.87	0.50	6.37	Catch Reconstruction
1940	8.45	0.60	9.05	Catch Reconstruction
1941	7.81	0.69	8.51	Catch Reconstruction
1942	4.15	0.79	4.94	Catch Reconstruction
1943	3.97	0.88	4.85	Catch Reconstruction
1944	3.26	0.98	4.24	Catch Reconstruction
1945	4.35	1.07	5.42	Catch Reconstruction
1946	7.48	1.17	8.65	Catch Reconstruction
1947	5.92	1.26	7.18	Catch Reconstruction
1948	11.81	1.36	13.17	Catch Reconstruction
1949	15.30	1.45	16.76	Catch Reconstruction
1950	18.65	1.55	20.20	Catch Reconstruction
1951	22.97	1.64	24.61	Catch Reconstruction
1952	19.99	1.74	21.73	Catch Reconstruction
1953	17.02	1.83	18.85	Catch Reconstruction
1954	21.16	1.93	23.09	Catch Reconstruction
1955	25.23	2.02	27.25	Catch Reconstruction
1956	28.17	2.12	30.28	Catch Reconstruction
1957	31.80	2.21	34.01	Catch Reconstruction
1958	48.15	2.31	50.46	Catch Reconstruction
1959	38.25	2.40	40.65	Catch Reconstruction
1960	28.66	2.50	31.15	Catch Reconstruction
1961	27.74	2.59	30.33	Catch Reconstruction
1962	28.04	2.69	30.73	Catch Reconstruction
1963	27.53	2.78	30.32	Catch Reconstruction
1964	21.73	2.88	24.61	Catch Reconstruction

Continues next page

tab:Rec_removal

Table 3: Recreational removals (mt) of GBYR. Data sources are the California Catch Reconstruction (modified for south of Pt. Conception), MRFSS (modified for 1981-1982), and CRFS.

Year	North of Pt. Conception	South of Pt. Conception	Total Recreational Removals	Source
1965	31.10	2.97	34.07	Catch Reconstruction
1966	33.85	3.07	36.91	Catch Reconstruction
1967	37.08	3.16	40.25	Catch Reconstruction
1968	36.78	3.26	40.03	Catch Reconstruction
1969	31.46	3.35	34.81	Catch Reconstruction
1970	41.25	3.45	44.70	Catch Reconstruction
1971	31.18	3.54	34.72	Catch Reconstruction
1972	41.50	3.64	45.13	Catch Reconstruction
1973	50.02	3.73	53.75	Catch Reconstruction
1974	51.60	3.83	55.43	Catch Reconstruction
1975	49.01	3.92	52.93	Catch Reconstruction
1976	49.30	4.02	53.32	Catch Reconstruction
1977	41.99	4.11	46.10	Catch Reconstruction
1978	32.57	4.21	36.77	Catch Reconstruction
1979	36.23	4.30	40.53	Catch Reconstruction
1980	80.56	4.54	85.10	MRFSS
1981	81.32	1.42	82.74	Estimated
1982	82.08	0.90	82.99	Estimated
1983	82.85	3.29	86.14	MRFSS
1984	150.47	5.58	156.05	MRFSS
1985	158.34	5.74	164.08	MRFSS
1986	171.81	6.52	178.33	MRFSS
1987	118.51	5.78	124.29	MRFSS
1988	79.43	4.80	84.23	MRFSS
1989	66.61	3.57	70.19	MRFSS
1990	82.33	2.73	85.06	MRFSS
1991	98.04	1.89	99.93	MRFSS
1992	113.76	1.04	114.80	MRFSS
1993	127.71	1.97	129.68	MRFSS
1994	97.39	3.03	100.42	MRFSS
1995	49.25	1.19	50.44	MRFSS
1996	38.06	5.23	43.28	MRFSS
1997	38.15	2.84	40.99	MRFSS
1998	43.55	2.52	46.07	MRFSS
1999	48.17	10.45	58.61	MRFSS
2000	66.53	4.39	70.92	MRFSS
2001	106.23	3.29	109.53	MRFSS

Continues next page

tab:Rec_removal

Table 3: Recreational removals (mt) of GBYR. Data sources are the California Catch Reconstruction (modified for south of Pt. Conception), MRFSS (modified for 1981-1982), and CRFS.

Year	North of Pt. Conception	South of Pt. Conception	Total Recreational Removals	Source
2002	84.28	2.15	86.43	MRFSS
2003	111.50	2.70	114.20	MRFSS
2004	41.75	0.98	42.73	CRFS
2005	47.51	6.59	54.10	CRFS
2006	48.10	2.13	50.22	CRFS
2007	32.88	2.70	35.58	CRFS
2008	45.14	3.61	48.74	CRFS
2009	65.64	4.30	69.94	CRFS
2010	106.76	3.90	110.67	CRFS
2011	76.16	10.24	86.40	CRFS
2012	48.25	9.89	58.14	CRFS
2013	38.43	8.86	47.28	CRFS
2014	56.96	9.06	66.02	CRFS
2015	58.09	5.00	63.09	CRFS
2016	65.72	6.57	72.29	CRFS
2017	49.36	11.15	60.51	CRFS
2018	36.48	6.30	42.78	CRFS

Table 4: Length composition sample sizes for survey data.
tab:length_samples_survey

	CCFRP		PISCO	
Year	Trips	Lengths	Trips	Lengths
2001			55	222
2002			56	438
2003			64	473
2004			64	312
2005			65	241
2006			68	220
2007	35	2147	68	156
2008	52	3143	67	198
2009	35	1579	68	154
2010	32	2201	58	144
2011	32	1727	68	260
2012	32	1820	40	183
2013	32	685	61	258
2014	32	1655	61	313
2015	18	1121	64	622
2016	32	2015	56	346
2017	58	2402	58	317
2018	29	1975	60	264

¡!– *** –¿

Table 5: Summary of indices used in this assessment.

tab:Index_summary

Fleet	Years	Name	Type	Area	Method	Endorsed
5	1988-1998	Deb Wilson-Vandenberg's Onboard Observer Survey	Fishery-dependent	Central California	Delta lognormal	SSC
6	2001-2018	CRFS CPFV Onboard Observer Survey	Fishery-dependent	North of Pt. Conception	Delta lognormal	SSC
7	2001-2018	CRFS CPFV Onboard Observer Survey	Fishery-dependent	South of Pt. Conception	Delta lognormal	SSC
8	2001-2018	PISCO Dive Survey	Fishery-independent	North of Pt. Conception	Negative Binomial	First use in stock assessment
9	2007-2018	CCFRP Hook-and-Line Survey	Fishery-independent	Central California	Negative Binomial	First use in stock assessment
10	1984-1999	MRFSS Dockside Survey	Fishery-dependent	North of Pt. Conception	Negative Binomial	SSC
11	1980-1999	MRFSS Dockside Survey	Fishery-dependent	South of Pt. Conception	Negative Binomial	SSC

Table 6: Index inpus.

tab:Indices

	Deb WV		MRFSS N		MRFSS S		Onboard N		Onboard S		CCFRP		PISCO	
Year	Obs	se_log	Obs	se_log	Obs	se_log	Obs	se_log	Obs	se_log	Obs	se_log	Obs	se_log
1980					0.08	0.21								
1981					0.05	0.24								
1982					0.07	0.25								
1983					0.13	0.13								
1984			0.04	0.60	0.09	0.17								
1985			0.03	0.55	0.09	0.21								
1986			0.09	0.58	0.03	0.19								
1987			0.02	0.66										
1988	0.22	0.17	0.03	0.61										
1989	0.34	0.15	0.02	0.66										
1990														
1991														
1992	0.30	0.17												
1993	0.20	0.14												
1994	0.23	0.12												
1995	0.25	0.10	0.04	0.64										
1996	0.28	0.10	0.04	0.52	0.04	0.28								
1997	0.21	0.09												
1998	0.24	0.11			0.05	0.26								
1999			0.03	0.53	0.05	0.22								
2000														
2001							0.32	0.12	0.01	0.52			1.66	0.23
2002							0.19	0.14	0.01	0.37			2.05	0.21
2003							0.28	0.07	0.03	0.33			2.53	0.19
2004							0.27	0.06	0.01	0.37			1.29	0.22
2005							0.26	0.08	0.02	0.24			0.91	0.24
2006							0.34	0.08	0.04	0.21			0.87	0.23
2007							0.33	0.08	0.08	0.16	1.20	0.15	0.69	0.24
2008							0.33	0.08	0.06	0.16	1.14	0.16	0.92	0.22
2009							0.27	0.08	0.07	0.16	1.13	0.16	0.59	0.22
2010							0.26	0.07	0.08	0.15	1.32	0.16	0.67	0.21
2011							0.24	0.07	0.15	0.11	0.97	0.16	1.24	0.19
2012							0.18	0.08	0.09	0.11	1.00	0.15	1.34	0.23
2013							0.09	0.09	0.07	0.12	0.38	0.16	1.45	0.22
2014							0.10	0.10	0.09	0.13	0.81	0.15	1.43	0.23
2015							0.17	0.10	0.06	0.17	1.03	0.16	2.55	0.22
2016							0.18	0.08	0.09	0.14	0.96	0.16	2.17	0.22
2017							0.15	0.12	0.08	0.17	1.18	0.16	1.80	0.23
2018							0.30	0.10	0.08	0.18	1.33	0.16	1.24	0.19

Table 7: Data filtering steps for Deb Wilson-Vandenberg's CPFV onboard observer index of abundance

tab:Fleet5_Filter

Filter	Drifts	Positive Drifts
Remove errors, missing data	6691	1470
Remove 1987 (sampled only MNT), 1990-1991 low sample sizes	4283	1372
Remove reefs that never encountered GBY	4022	1372
Remove lower and upper 2.5% of time fished	3762	1300
Remove depth less than 9 m and greater than 69 m	3515	1279
Remove reefs with low sample rates	2411	1096

Table 8: Model selection for Deb Wilson-Vandenberg's CPFV onboard observer index of abundance. Bold values indicate the model selected.

tab:Fleet5_AIC

Model	Lognormal	Binomial
Year	2834	3330
Year + Depth	2781	2906
Year + Reef	2716	2880
Year + Month	2839	3286
Year + Depth + Reef	**2625**	**2488**
Year + Month+ Reef	2725	2844
Year + Depth + Month	2780	2902
Year+ Depth+Month+Reef	2632	2479

Table 9: Data filtering steps for the CRFS CPFV onboard observer index of abundance for north and south of Pt. Conception.

tab:Fleet6_7_Filter

Filter	Drifts	Positive Drifts
Data from SQL filtered for missing data	67850	9317
Remove years prior to 2001 and north of Cape Mendocino	64448	9129
Depth, remove 1% data on each tail of positive catches	50846	8955
Time fished, remove 1% data on each tail	50100	8903
Observed anglers, remove 1% data on each tail	48089	8774
Limit to reefs observering gopher/byel in at least 20 drifts	29639	8025
Limit to reefs sampled in at least 2/3 of all years	32672	7517
Limit to drifs within 1000 m of a reef	27355	7358
Put depth in 10m depth bins, remove 0-9 and 60-69 m bins	25427	7250
Start of north filtering	13792	6036
Filter to drifts within 43 m of a reef, 97% quantile	13145	5854
Make sure reefs still sampled at least 2/3 of years	12965	5796
Start of south filtering	11635	1277
Filter to drifts with >=20% groundfish and recheck reefs	5495	1171
Make sure reefs still sampled at least 2/3 of years	5440	1132

Table 10: Model selection for the CRFS CPFV onboard observer index of abundance for north of Pt. Conception. Bold values indicate the model selected.

tab:Fleet6_AIC

Model	Lognormal	Binomial
Year	14135	17531
Year + Month	14120	17529
Year + Depth	13953	17025
Year + Reef	14126	17293
Year + Month + Depth	13951	17027
Year + Month + Depth + Reef	**13921**	**16674**

Table 11: Model selection for the CRFS CPFV onboard observer index of abundance for south of Pt. Conception. Bold values indicate the model selected.

tab:Fleet7_AIC

Model	Lognormal	Binomial
Year	2798	5490
Year + Month	2799	5487
Year + Depth	2744	5159
Year + Reef	2653	5390
Year + Depth + Reef	**2652**	**5071**
Year + Depth + Reef + Month	2663	5072

Table 12: Data filtering steps for the PISCO dive survey.

tab:Fleet8_Filter

Filter	Transects	Positive Transects
Remove missing data and retain only bottom transects	22,055	6,330
Remove month of June - few samples	21,941	6,318
Remove dives earlier than 2004 for UCSB and 2001 for UCSC	20,659	6,165
Keep sites sampled in at least half of all years (UCSC and UCSB separate)	14,721	4,097
Keep sites observing GBYR in at least half of all years	12,139	4,002
Remove transects denoted as old, no longer sampled	10,712	3,268
Subset to just UCSC sites	5,686	2,939
Use only consistently sampled sites	3,231	1,729
Collapse repeated transects	1,928	1,487

Table 13: Model selection for the PISCO dive survey data.

tab:Fleet8_AIC

Model	AIC
Year	5,687
Year + Month	5,672
Year + Month + Site	5,623
Year + Month + Site + Zone	**5,512**

Table 14: Data filtering steps for the fishery-independent CCFRP hook-and-line survey.

tab:Fleet9_Filter

Filter	Drifts	Positive Drifts
All data	5,886	Drift and catch data not merged
Remove missing data and cells not sampled consistently at Piedras Blancas	4,942	3,857
Remove cells that never encountered GBYR	4,934	3,857
Remove depth bins with little or no sampling (keep 5-39 m)	4,920	3,848

Table 15: Model selection for the fishery-independent CCFRP hook-and-line survey.

tab:Fleet9_AIC

Model	AIC
Year	23,212
Year + Month	23,214
Year + Depth	22,901
Year + Depth + Site	22,642
Year + Depth + Site + MPA/REF	**22,341**

Table 16: Data filtering steps for the MRFSS dockside intercept survey index of abundance for north and south of Pt. Conception.

tab:Fleet10_11_Filter

Filter	Trips	Positive Trips
All data	10,392	1,061
Remove north of Cape Mendocino	10,327	1,061
Remove trips targetting offshore species	10,122	1,061
Start northern filtering	2,788	620
Remove species that never co-occurand not present in at least 1% of all	2,788	620
Stephens-MacCall filter (keep all positives - selected filter)	806	620
Alternate Stephens-MacCall filter (keep only above threshold)	623	437
Remove years after 1999 due to regulation changes and with fewer than 20 trips	544	220
Start southern filtering	7,334	441
Remove species that never co-occurand not present in at least 1% of all	7,334	441
Stephens-MacCall filter (keep all positives - selected filter)	687	441
Alternate Stephens-MacCall filter (keep only above threshold)	430	184
Remove years after 1999 due to regulation changes and with fewer than 20 trips	475	342

Table 17: Model selection for the MRFSS dockside intercept survey north of Pt. Conception. Bold values indicate the model selected.

tab:Fleet10_AIC

Model	AIC
Year	1,481
Year + Region	1,429
Year + Region + Area_X	1,403
Year + Region + Area_X + Wave	1,397

Table 18: Model selection for the MRFSS dockside intercept survey south of Pt. Conception. Bold values indicate the model selected.

tab:Fleet11_AIC

Model	Lognormal	Binomial
Year	911	552
Year+ Wave	908	538
Year + Wave + Area_X	905	540
Year + Wave + Area_X + SubRegion	**903**	**537**
Year + Wave + SubRegion	908	536

Table 19: Contingency table for the Stephens-MacCall filtering for the MRFSS dockside CPFV index for GBYR north of Pt. Conception.

tab:Fleet10_contingency

	GBYR absent	GBYR present
Above 0.4	186	437
Below 0.4	1982	183

Table 20: Contingency table for the Stephens-MacCall filtering for the MRFSS dockside CPFV index for GBYR south of Pt. Conception.

tab:Fleet11_contingency

	GBYR absent	GBYR present
Above 0.22	246	184
Below 0.22	6647	257

Table 21: Summary of age data used in the assessment.

tab:Age_data

Project	Source	Years	Region	Gear	Black.and.yellow	Gopher
Port sampling	Commercial	2009-2010; 2018	Bodega; Morro Bay	hook-and-line	0	46
CDFW sampling	Recreational	1978; 1980; 1982-1986	Morro Bay; San Francisco	hook-and-line	0	138
Cal Poly onboard observer	Recreational	2018	Morro Bay	hook-and-line	0	36
E.J.'s trap survey	Research	2012	Monterey	trap	1	25
Zaitlin thesis	Research	1983-1986	Monterey	spear	491	0
Pearson groundfish cruise	Research	2002-2005	Monterey	hook-and-line	0	118
Hanan CPFV survey	Research	2003-2004	Morro Bay; Santa Barbara	hook-and-line	0	189
Juv. rockfish cruise special study	Research	2004-2005	Monterey	hook-and-line	0	79
CCFRP	Research	2007-2013	Central CA	hook-and-line	7	1,191
CCFRP trap	Research	2008-2009	Central CA	trap	0	87
Abrams thesis	Research	2010-2011	Fort Bragg	hook-and-line	0	59
Total					**499**	**1,968**

Table 22: Results from 100 jitters from the base case model.

tab:jitter

Description	Value	NA	NA
Returned to base case	-	-	-
Found local minimum	-	-	-
Found better solution	-	-	-
Error in likelihood	-	-	-
Total	100	100	100

Table 23: List of parameters used in the base model, including estimated values and standard deviations (SD), bounds (minimum and maximum), estimation phase (negative values indicate not estimated), status (indicates if parameters are near bounds, and prior type information (mean, SD).

No.	Parameter	Value	Phase	Bounds	Status	SD	Prior (Exp.Val, SD)
1	NatM_p_1_Fem_GP_1	0.211	2	(0.05, 0.4)	OK	0.029	Log_Norm (-1.6458, 0.4384)
2	L_at_Amin_Fem_GP_1	9.590	3	(4, 50)	OK	0.725	None
3	L_at_Amax_Fem_GP_1	28.549	3	(20, 60)	OK	0.883	None
4	VonBert_K_Fem_GP_1	0.129	3	(0.01, 0.3)	OK	0.028	None
5	CV_young_Fem_GP_1	0.258	3	(0.05, 0.5)	OK	0.032	None
6	CV_old_Fem_GP_1	0.117	3	(0.03, 0.3)	OK	0.013	None
7	Wtlen_1_Fem_GP_1	0.000	-3	(-3, 3)			None
8	Wtlen_2_Fem_GP_1	3.256	-3	(2, 4)			None
9	Mat50%_Fem_GP_1	21.666	-3	(-3, 3)			None
10	Mat_slope_Fem_GP_1	-0.906	-3	(-6, 3)			None
11	Eggs/kg_inter_Fem_GP_1	1.000	-3	(-3, 3)			None
12	Eggs/kg_slope_wt_Fem_GP_1	0.000	-3	(-3, 3)			None
13	CohortGrowDev	1.000	-1	(0.1, 10)			None
14	FracFemale_GP_1	0.500	-4	(0.000001, 0.999999)			None
15	SR_LN(R0)	8.473	1	(2, 15)	OK	0.386	None
16	SR_BH_steep	0.720	-1	(0.2, 1)			Full_Beta (0.72, 0.16)
17	SR_sigmaR	0.400	-2	(0, 2)			None
18	SR_regime	0.000	-4	(-5, 5)			None
19	SR_autocorr	0.678	4	(-1, 1)	OK	0.106	None
103	LnQ_base_DebCPFV(5)	-6.982	-1	(-15, 15)			None
104	Q_extraSD_DebCPFV(5)	0.071	4	(0.0001, 2)	OK	0.048	None
105	LnQ_base_RecOnboardNorth(6)	-7.718	-1	(-15, 15)			None
106	Q_extraSD_RecOnboardNorth(6)	0.230	4	(0.0001, 2)	OK	0.056	None
107	LnQ_base_RecOnboardSouth(7)	-10.300	-1	(-15, 15)			None
108	Q_extraSD_RecOnboardSouth(7)	0.597	4	(0.0001, 2)	OK	0.148	None
109	LnQ_base_PISCO(8)	-7.605	-1	(-15, 15)			None

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Table 23: List of parameters used in the base model, including estimated values and standard deviations (SD), bounds (minimum and maximum), estimation phase (negative values indicate not estimated), status (indicates if parameters are near bounds, and prior type information (mean, SD).

No.	Parameter	Value	Phase	Bounds	Status	SD	Prior (Exp.Val, SD)
110	Q_extraSD_PISCO(8)	0.209	4	(0.0001, 2)	OK	0.074	None
111	LnQ_base_CCFRP(9)	-6.449	-1	(-15, 15)			None
112	Q_extraSD_CCFRP(9)	0.185	4	(0.0001, 2)	OK	0.074	None
113	LnQ_base_RecDocksideNorth(10)	-8.798	-1	(-15, 15)			None
114	Q_extraSD_RecDocksideNorth(10)	0.000	-4	(0.0001, 2)			None
115	LnQ_base_RecDocksideSouth(11)	-9.771	-1	(-15, 15)			None
116	Q_extraSD_RecDocksideSouth(11)	0.276	4	(0.0001, 2)	OK	0.108	None
117	Size_DblN_peak_Com(1)	32.339	1	(19, 38)	OK	0.716	None
118	Size_DblN_top_logit_Com(1)	8.000	-5	(-5, 10)			None
119	Size_DblN_ascend_se_Com(1)	3.124	5	(-9, 10)	OK	0.126	None
120	Size_DblN_descend_se_Com(1)	5.000	-5	(-9, 9)			None
121	Size_DblN_start_logit_Com(1)	-11.708	5	(-15, -5)	OK	1.749	None
122	Size_DblN_end_logit_Com(1)	10.000	-5	(-5, 15)			None
123	Size_DblN_peak_ComDisc(2)	25.041	2	(19, 38)	OK	0.449	None
124	Size_DblN_top_logit_ComDisc(2)	-9.378	5	(-15, 10)	OK	78.881	None
125	Size_DblN_ascend_se_ComDisc(2)	2.035	5	(-9, 10)	OK	0.220	None
126	Size_DblN_descend_se_ComDisc(2)	5.469	5	(-9, 9)	OK	1.896	None
127	Size_DblN_start_logit_ComDisc(2)	-14.117	5	(-15, -5)	OK	20.152	None
128	Size_DblN_end_logit_ComDisc(2)	-999.000	-5	(-5, 10)			None
129	Size_DblN_peak_RecNorth(3)	32.379	3	(19, 39)	OK	0.404	None
130	Size_DblN_top_logit_RecNorth(3)	8.000	-5	(-5, 10)			None
131	Size_DblN_ascend_se_RecNorth(3)	3.264	5	(-9, 10)	OK	0.072	None
132	Size_DblN_descend_se_RecNorth(3)	5.000	-5	(-9, 9)			None
133	Size_DblN_start_logit_RecNorth(3)	-11.972	5	(-15, -5)	OK	1.526	None
134	Size_DblN_end_logit_RecNorth(3)	10.000	-5	(-5, 15)			None
135	Size_DblN_peak_RecSouth(4)	27.732	4	(19, 38)	OK	1.212	None
136	Size_DblN_top_logit_RecSouth(4)	8.000	-5	(-5, 10)			None

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Table 23: List of parameters used in the base model, including estimated values and standard deviations (SD), bounds (minimum and maximum), estimation phase (negative values indicate not estimated), status (indicates if parameters are near bounds, and prior type information (mean, SD).

No.	Parameter	Value	Phase	Bounds	Status	SD	Prior (Exp.Val, SD)
137	Size_DblN_ascend_se_RecSouth(4)	3.212	5	(-9, 10)	OK	0.265	None
138	Size_DblN_descend_se_RecSouth(4)	5.000	-5	(-9, 9)			None
139	Size_DblN_start_logit_RecSouth(4)	-8.820	5	(-15, -5)	OK	2.760	None
140	Size_DblN_end_logit_RecSouth(4)	10.000	-5	(-5, 15)			None
141	Size_DblN_peak_DebCPFV(5)	30.894	5	(19, 38)	OK	0.617	None
142	Size_DblN_top_logit_DebCPFV(5)	8.000	-5	(-5, 10)			None
143	Size_DblN_ascend_se_DebCPFV(5)	2.999	5	(-9, 10)	OK	0.117	None
144	Size_DblN_descend_se_DebCPFV(5)	5.000	-5	(-9, 9)			None
145	Size_DblN_start_logit_DebCPFV(5)	-17.851	5	(-20, -5)	OK	38.084	None
146	Size_DblN_end_logit_DebCPFV(5)	10.000	-5	(-5, 15)			None
147	SizeSel_P1_RecOnboardNorth(6)	-1.000	-5	(-1, 10)			None
148	SizeSel_P2_RecOnboardNorth(6)	-1.000	-5	(-1, 10)			None
149	SizeSel_P1_RecOnboardSouth(7)	-1.000	-5	(-1, 10)			None
150	SizeSel_P2_RecOnboardSouth(7)	-1.000	-5	(-1, 10)			None
151	Size_DblN_peak_PISCO(8)	30.452	5	(19, 38)	OK	2.149	None
152	Size_DblN_top_logit_PISCO(8)	8.000	-5	(-15, 10)			None
153	Size_DblN_ascend_se_PISCO(8)	3.907	5	(-9, 10)	OK	0.362	None
154	Size_DblN_descend_se_PISCO(8)	5.000	-5	(-9, 9)			None
155	Size_DblN_start_logit_PISCO(8)	-2.801	5	(-15, 15)	OK	0.604	None
156	Size_DblN_end_logit_PISCO(8)	10.000	-5	(-5, 15)			None
157	Size_DblN_peak_CCFRP(9)	31.050	5	(19, 38)	OK	0.623	None
158	Size_DblN_top_logit_CCFRP(9)	-10.648	5	(-15, 10)	OK	65.029	None
159	Size_DblN_ascend_se_CCFRP(9)	3.135	5	(-9, 10)	OK	0.149	None
160	Size_DblN_descend_se_CCFRP(9)	1.643	5	(-15, 9)	OK	0.802	None
161	Size_DblN_start_logit_CCFRP(9)	-999.000	-5	(-15, -5)			None
162	Size_DblN_end_logit_CCFRP(9)	-999.000	-5	(-5, 10)			None
163	SizeSel_P1_RecDocksideNorth(10)	-1.000	-5	(-1, 10)			None

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Table 23: List of parameters used in the base model, including estimated values and standard deviations (SD), bounds (minimum and maximum), estimation phase (negative values indicate not estimated), status (indicates if parameters are near bounds, and prior type information (mean, SD).

No.	Parameter	Value	Phase	Bounds	Status	SD	Prior (Exp.Val, SD)
164	SizeSel_P2_RecDocksideNorth(10)	-1.000	-5	(-1, 10)			None
165	SizeSel_P1_RecDocksideSouth(11)	-1.000	-5	(-1, 10)			None
166	SizeSel_P2_RecDocksideSouth(11)	-1.000	-5	(-1, 10)			None
167	Size_DblN_peak_Com(1)_BLK1repl_1999	28.891	6	(19, 38)	OK	0.328	None
168	Size_DblN_ascend_se_Com(1)_BLK1repl_1999	1.585	6	(-9, 10)	OK	0.167	None
169	Size_DblN_start_logit_Com(1)_BLK1repl_1999	-11.860	6	(-15, -5)	OK	3.470	None

tab:model_params

Table 24: Likelihood components from the base model.

tab:like_components

Likelihood component	Value
TOTAL	1097.30
Catch	0.00
Survey	-98.12
Length composition	763.02
Age composition	421.52
Recruitment	10.88
Forecast recruitment	0.00
Parameter priors	0.00
Parmeter soft bounds	0.01

Table 25: Time-series of population estimates from the base-case model. Relative exploitation rate is $(1 - SPR)/(1 - SPR_{50\%})$.

Year	Total biomass (mt)	Spawning biomass (mt)	Depletion	Age-0 recruits	Total catch (mt)	Relative exploitation rate	SPR
1916	2156	1317	0.000	4784	4	0.00	0.99
1917	2154	1315	0.998	4783	7	0.00	0.98
1918	2150	1311	0.995	4782	8	0.00	0.98
1919	2146	1307	0.993	4780	5	0.00	0.98
1920	2144	1306	0.991	4780	5	0.00	0.98
1921	2142	1304	0.990	4779	5	0.00	0.99
1922	2141	1303	0.989	4779	4	0.00	0.99
1923	2141	1303	0.989	4779	4	0.00	0.99
1924	2140	1303	0.989	4779	2	0.00	0.99
1925	2141	1303	0.989	4779	3	0.00	0.99
1926	2141	1304	0.990	4779	5	0.00	0.99
1927	2140	1303	0.989	4779	4	0.00	0.99
1928	2140	1302	0.989	4779	6	0.00	0.98
1929	2138	1301	0.988	4778	6	0.00	0.98
1930	2137	1300	0.987	4778	8	0.00	0.98
1931	2135	1298	0.985	4777	5	0.00	0.99
1932	2135	1298	0.985	4777	10	0.00	0.97
1933	2131	1294	0.983	4776	7	0.00	0.98
1934	2130	1294	0.982	4775	7	0.00	0.98
1935	2130	1293	0.982	4775	6	0.00	0.98
1936	2130	1293	0.982	4775	6	0.00	0.98
1937	2130	1293	0.982	4775	15	0.01	0.96
1938	2124	1288	0.978	4773	18	0.01	0.95
1939	2117	1281	0.973	4771	21	0.01	0.94
1940	2109	1274	0.967	4768	28	0.01	0.93
1941	2098	1264	0.959	4764	27	0.01	0.93
1942	2088	1255	0.953	4761	14	0.01	0.96
1943	2089	1255	0.953	4761	17	0.01	0.95
1944	2088	1254	0.952	4761	4	0.00	0.99
1945	2094	1261	0.957	4763	6	0.00	0.98
1946	2099	1265	0.961	4765	27	0.01	0.93
1947	2090	1257	0.954	4762	37	0.02	0.90
1948	2076	1244	0.945	4757	39	0.02	0.90
1949	2061	1231	0.935	4752	37	0.02	0.90
1950	2051	1222	0.927	4748	39	0.02	0.89
1951	2041	1212	0.920	4744	52	0.03	0.86
1952	2024	1197	0.908	4738	52	0.03	0.86

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Table 25: Time-series of population estimates from the base-case model. Relative exploitation rate is $(1 - SPR)/(1 - SPR_{50\%})$. tab:Timeseries_mo

Year	Total biomass (mt)	Spawning biomass (mt)	Depletion	Age-0 recruits	Total catch (mt)	Relative exploitation rate	SPR
1953	2009	1183	0.898	4732	55	0.03	0.86
1954	1995	1170	0.888	4726	68	0.03	0.83
1955	1974	1152	0.874	4718	60	0.03	0.84
1956	1961	1140	0.865	4713	76	0.04	0.81
1957	1940	1122	0.851	4704	76	0.04	0.80
1958	1922	1106	0.839	4696	88	0.05	0.78
1959	1899	1085	0.824	4687	62	0.03	0.82
1960	1893	1082	0.821	3255	44	0.02	0.87
1961	1897	1090	0.827	2975	50	0.03	0.85
1962	1888	1093	0.830	2772	61	0.03	0.83
1963	1859	1089	0.827	2630	56	0.03	0.84
1964	1820	1088	0.826	2552	43	0.02	0.87
1965	1775	1089	0.827	2542	58	0.03	0.83
1966	1710	1074	0.815	2600	52	0.03	0.85
1967	1643	1052	0.799	2731	48	0.03	0.86
1968	1575	1023	0.777	2930	49	0.03	0.85
1969	1508	985	0.748	3104	46	0.03	0.86
1970	1447	943	0.716	3149	60	0.04	0.82
1971	1384	890	0.676	3036	51	0.04	0.83
1972	1337	844	0.641	2816	66	0.05	0.79
1973	1288	793	0.602	2664	88	0.07	0.74
1974	1233	735	0.558	2747	92	0.07	0.71
1975	1184	686	0.521	3062	89	0.08	0.70
1976	1144	647	0.492	3430	91	0.08	0.68
1977	1111	616	0.468	3674	79	0.07	0.69
1978	1094	597	0.454	3732	84	0.08	0.67
1979	1083	579	0.439	3610	78	0.07	0.68
1980	1086	566	0.430	3787	155	0.14	0.53
1981	1052	519	0.394	3993	143	0.14	0.53
1982	1037	488	0.370	4134	129	0.12	0.53
1983	1041	474	0.360	3670	118	0.11	0.54
1984	1058	475	0.360	2819	174	0.17	0.45
1985	1045	455	0.346	2339	173	0.17	0.44
1986	1032	445	0.338	2191	206	0.20	0.39
1987	995	428	0.325	2350	162	0.16	0.43
1988	975	433	0.329	2918	145	0.15	0.45
1989	957	445	0.337	4015	120	0.13	0.50

Continues next page

tab:Timeseries_mo

Table 25: Time-series of population estimates from the base-case model. Relative exploitation rate is $(1 - SPR)/(1 - SPR_{50\%})$.

Year	Total biomass (mt)	Spawning biomass (mt)	Depletion	Age-0 recruits	Total catch (mt)	Relative exploitation rate	SPR
1990	949	459	0.348	5623	136	0.14	0.49
1991	939	456	0.346	7114	175	0.19	0.44
1992	927	427	0.324	7774	207	0.23	0.40
1993	931	379	0.288	7619	210	0.23	0.37
1994	972	336	0.255	7063	161	0.17	0.39
1995	1075	329	0.250	6297	150	0.14	0.39
1996	1205	352	0.267	5358	148	0.12	0.39
1997	1342	404	0.307	4289	118	0.09	0.46
1998	1485	494	0.375	3604	118	0.08	0.49
1999	1601	598	0.454	3302	127	0.08	0.52
2000	1678	702	0.533	3250	131	0.08	0.55
2001	1717	795	0.604	3702	169	0.10	0.54
2002	1703	853	0.648	4080	133	0.08	0.61
2003	1683	900	0.683	3497	148	0.09	0.61
2004	1637	912	0.692	2963	72	0.04	0.75
2005	1625	935	0.710	2793	86	0.05	0.73
2006	1592	932	0.708	2782	78	0.05	0.76
2007	1555	921	0.699	2887	70	0.05	0.78
2008	1514	907	0.688	3137	87	0.06	0.75
2009	1460	879	0.667	3534	111	0.08	0.70
2010	1391	834	0.633	3680	153	0.11	0.63
2011	1304	764	0.580	3319	135	0.10	0.63
2012	1242	707	0.537	3393	94	0.08	0.68
2013	1217	676	0.513	4194	84	0.07	0.70
2014	1209	654	0.496	6319	105	0.09	0.65
2015	1202	625	0.475	8445	109	0.09	0.63
2016	1214	602	0.457	6941	112	0.09	0.61
2017	1258	585	0.444	5137	104	0.08	0.61
2018	1336	580	0.441	4177	92	0.07	0.63
2019	1433	596	0.452	4276			

Table 26: Sensitivity of the base model to dropping or down-weighting data sources and alternative assumptions about growth.

`tab:Sensitivity_model1`

Label	Base (Francis weights)	Default weights	Harmonic mean weights	Estimate equal M	Estimate equal M and h	Drop PR data	Drop PC data	Drop RecDD data
TOTAL_like	-	-	-	-	-	-	-	-
Catch_like	-	-	-	-	-	-	-	-
Equil_catch_like	-	-	-	-	-	-	-	-
Survey_like	-	-	-	-	-	-	-	-
Length_comp_like	-	-	-	-	-	-	-	-
Age_comp_like	-	-	-	-	-	-	-	-
Parm_priors_like	-	-	-	-	-	-	-	-
SSB_Unfished_thousand_mt	-	-	-	-	-	-	-	-
TotBio_Unfished	-	-	-	-	-	-	-	-
SmryBio_Unfished	-	-	-	-	-	-	-	-
Recr_Unfished_billions	-	-	-	-	-	-	-	-
SSB_Btgt_thousand_mt	-	-	-	-	-	-	-	-
SPR_Btgt	-	-	-	-	-	-	-	-
Fstd_Btgt	-	-	-	-	-	-	-	-
TotYield_Btgt_thousand_mt	-	-	-	-	-	-	-	-
SSB_SPRtgt_thousand_mt	-	-	-	-	-	-	-	-
Fstd_SPRtgt	-	-	-	-	-	-	-	-
TotYield_SPRtgt_thousand_mt	-	-	-	-	-	-	-	-
SSB_MSY_thousand_mt	-	-	-	-	-	-	-	-
SPR_MSY	-	-	-	-	-	-	-	-
Fstd_MSY	-	-	-	-	-	-	-	-
TotYield_MSY_thousand_mt	-	-	-	-	-	-	-	-
RetYield_MSY	-	-	-	-	-	-	-	-
Bratio_2015	-	-	-	-	-	-	-	-
F_2015	-	-	-	-	-	-	-	-
SPRratio_2015	-	-	-	-	-	-	-	-
Recr_2015	-	-	-	-	-	-	-	-
Recr_Virgin_billions	-	-	-	-	-	-	-	-
L_at_Amin_Fem_GP_1	-	-	-	-	-	-	-	-
L_at_Amax_Fem_GP_1	-	-	-	-	-	-	-	-
VonBert_K_Fem_GP_1	-	-	-	-	-	-	-	-
CV_young_Fem_GP_1	-	-	-	-	-	-	-	-
CV_old_Fem_GP_1	-	-	-	-	-	-	-	-

Table 27: Summary of the biomass/abundance time series used in the stock assessment.

tab:Index_summary

Fleet	Years	Name	Fishery ind.	Filtering	Method	Endorsed
5	1988-1998	Deb Wilson-Vandenberg's Onboard Observer Survey	Fishery-dependent	Central California	Delta lognormal	SSC
6	2001-2018	CRFS CPFV Onboard Observer Survey	Fishery-dependent	North of Pt. Conception	Delta lognormal	SSC
7	2001-2018	CRFS CPFV Onboard Observer Survey	Fishery-dependent	South of Pt. Conception	Delta lognormal	SSC
8	2001-2018	PISCO Dive Survey	Fishery-independent	North of Pt. Conception	Negative Binomial	First use in stock assessment
9	2007-2018	CCFRP Hook-and-Line Survey	Fishery-independent	Central California	Negative Binomial	First use in stock assessment
10	1984-1999	MRFSS Dockside Survey	Fishery-dependent	North of Pt. Conception	Negative Binomial	SSC
11	1980-1999	MRFSS Dockside Survey	Fishery-dependent	South of Pt. Conception	Negative Binomial	SSC

Table 28: Summaries of key assessment outputs and likelihood values from the retrospective analysis. Note that male growth parameters are exponential offsets from female parameters, and depletion and SPR ratio are for the year of 2017. The base model includes all of the data. Retro1 removes the last year of data (2016), Retro2 removes the last two years of data, Retro3 removes three years and Retro4 removes four years.

tab:retro

Label	Base	Retro1	Retro2	Retro3	Retro4
Female natural mortality	0.26	0.26	0.26	0.26	0.26
Steepness	0.72	0.72	0.72	0.72	0.72
lnR0	8.16	8.09	8.07	8.04	8.08
Total Biomass (mt)	2796.86	2593.78	2568.77	2498.07	2650.36
Depletion	57.41	53.57	50.74	50.72	54.78
SPR ratio	0.72	0.76	0.79	0.80	0.74
Female Lmin	12.43	12.45	12.90	12.63	13.03
Female Lmax	33.31	33.50	33.39	33.37	33.46
Female K	0.25	0.24	0.24	0.25	0.23
Male Lmin (offset)	0.00	0.00	0.00	0.00	0.00
Male Lmax (offset)	-0.16	-0.16	-0.15	-0.16	-0.15
Male K (offset)	-0.29	-0.30	-0.43	-0.41	-0.56
Negative log-likelihood	1097.30	1047.56	1009.37	961.81	897.04
No. parameters	0.00	0.00	0.00	0.00	0.00
TOTAL	0.00	0.00	0.00	0.00	0.00
Equililibrium catch	-98.12	-92.00	-89.12	-81.75	-80.59
Survey	763.02	739.90	720.39	700.10	670.66
Length composition	421.52	390.56	369.97	336.26	299.84
Age composition	10.88	9.09	8.12	7.20	7.12
Recruitment	0.00	0.00	0.00	0.00	0.00
Forecast Recruitment	0.00	0.00	0.00	0.00	0.00
Parameter priors	0.01	0.01	0.01	0.01	0.01

Table 29: Summaries of key assessment outputs and likelihood values from selected likelihood profile runs on virgin recruitment (lnR0) and steepness. Note that male growth parameters are exponential offsets from female parameters, and depletion and SPR ratio are for the year of 2017.

tab:like_profiles

Label	R07400	R07800	R08200	R08600	R09000	h0410	h0570	h0710	h0870	h0990
Female M	0.26	0.26	0.26	0.26	0.26	0.26	0.26	0.26	0.26	0.26
Steepness	0.72	0.72	0.72	0.72	0.72	0.41	0.57	0.71	0.87	0.99
lnR0	7.40	7.80	8.20	8.60	9.00	8.34	8.21	8.16	8.13	8.11
Total biomass (m)	1623.19	2113.03	2894.72	4173.95	6142.97	3313.42	2943.85	2802.69	2712.12	2667.97
Depletion (%)	46.83	49.83	58.31	66.23	71.80	51.20	55.27	57.32	58.81	59.60
SPR ratio	1.05	0.91	0.70	0.49	0.34	0.68	0.71	0.72	0.72	0.73
Female Lmin	12.16	12.41	12.43	12.39	12.36	12.43	12.44	12.43	12.43	12.43
Female Lmax	34.29	33.83	33.26	32.76	32.42	33.19	33.28	33.31	33.33	33.34
Female K	0.24	0.25	0.25	0.26	0.26	0.25	0.25	0.25	0.25	0.25
Male Lmin (offset)	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Male Lmax (offset)	-0.18	-0.17	-0.16	-0.15	-0.15	-0.16	-0.16	-0.16	-0.16	-0.16
Male K (offset)	-0.22	-0.31	-0.29	-0.24	-0.21	-0.27	-0.29	-0.29	-0.30	-0.30
Negative log-likelihood										
TOTAL	1117.15	1101.02	1097.33	1099.69	1102.95	1101.35	1098.58	1097.35	1096.72	1100.21
Catch	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Equil_catch	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Survey	-100.10	-99.20	-97.99	-97.00	-96.37	-98.27	-98.18	-98.12	-98.06	-98.03
Length_comp	761.18	760.12	763.44	767.61	770.76	765.11	763.69	763.05	762.58	762.33
Age_comp	437.32	427.37	421.09	418.57	417.98	420.58	421.24	421.51	421.68	421.77
Recruitment	18.74	12.72	10.80	10.50	10.58	12.55	11.40	10.90	10.56	10.38
Forecast_Recruitment	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Parm_priors	0.00	0.00	0.00	0.00	0.00	1.38	0.42	0.01	-0.04	3.76
Parm_softbounds	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01	0.01
Parm_devs	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Crash_Pen	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00

Table 30: Summaries of key assessment outputs and likelihood values from selected likelihood profile runs on female natural mortality. Note that male growth parameters are exponential offsets from female parameters, and depletion and SPR ratio are for the year of 2017.

tab:like_profiles

Label	M0220	M0260	M0300	M0350	M0400
Female M	0.22	0.26	0.30	0.35	0.40
Steepness	0.72	0.72	0.72	0.72	0.72
lnR0	7.67	8.20	8.95	12.21	31.00
Total biomass (m)	2259.39	2861.79	4632.81	89473.50	9753570000000.00
Depletion (%)	47.72	58.15	68.08	79.27	79.74
SPR ratio	0.97	0.70	0.41	0.02	0.00
Female Lmin	12.39	12.44	12.43	12.39	12.24
Female Lmax	33.23	33.31	33.31	33.25	33.73
Female K	0.25	0.25	0.25	0.25	0.24
Male Lmin (offset)	0.00	0.00	0.00	0.00	0.00
Male Lmax (offset)	-0.16	-0.16	-0.15	-0.15	-0.15
Male K (offset)	-0.27	-0.30	-0.31	-0.32	-0.36
Negative log-likelihood					
TOTAL	1102.66	1096.96	1092.96	1089.92	1091.52
Catch	0.00	0.00	0.00	0.00	0.00
Equil_catch	0.00	0.00	0.00	0.00	0.00
Survey	-97.79	-98.14	-98.33	-98.33	-98.95
Length_comp	765.50	762.85	760.88	759.19	755.26
Age_comp	422.97	421.41	420.05	418.75	425.16
Recruitment	11.91	10.82	10.30	10.05	9.54
Forecast_Recruitment	0.00	0.00	0.00	0.00	0.00
Parm_priors	0.06	0.00	0.06	0.25	0.51
Parm_softbounds	0.01	0.01	0.01	0.00	0.00
Parm_devs	0.00	0.00	0.00	0.00	0.00
Crash_Pen	0.00	0.00	0.00	0.00	0.00

Table 31: Projection of potential OFL, spawning biomass, and depletion for the base case model.

tab:Forecast_mod1

Yr	OFL contribution (mt)	ACL landings (mt)	Age 5+ biomass (mt)	Spawning Biomass (mt)	Depletion
2019	180.214	180.214	1425.960	596.045	0.452
2020	168.285	168.285	1466.470	596.045	0.452
2021	167.052	167.052	1506.630	596.045	0.452
2022	174.540	174.540	1535.160	596.045	0.452
2023	187.307	187.307	1546.310	596.045	0.452
2024	200.668	200.668	1538.850	596.045	0.452
2025	210.285	210.285	1515.250	596.045	0.452
2026	213.765	213.765	1480.850	596.045	0.452
2027	211.189	211.190	1442.300	596.045	0.452
2028	204.471	204.472	1405.670	596.045	0.452
2029	196.145	196.145	1375.030	596.045	0.452
2030	188.320	188.320	1351.970	596.045	0.452

8 Figures

figures

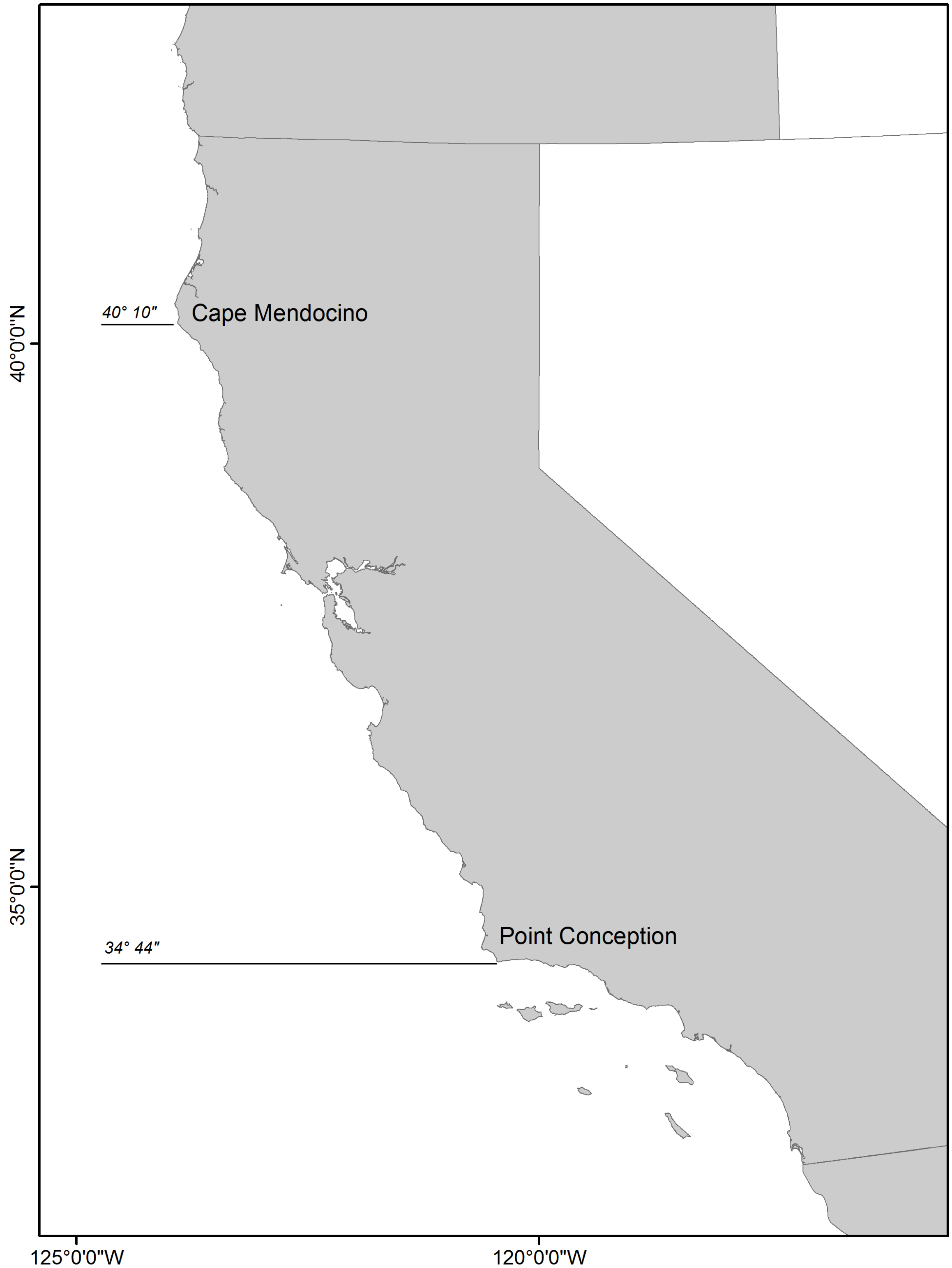

Figure 1: Map showing the management area for gopher and black-and-yellow rockfish from Cape Mendocino to the U.S. Mexico border.{fig:assess_reagion_map}

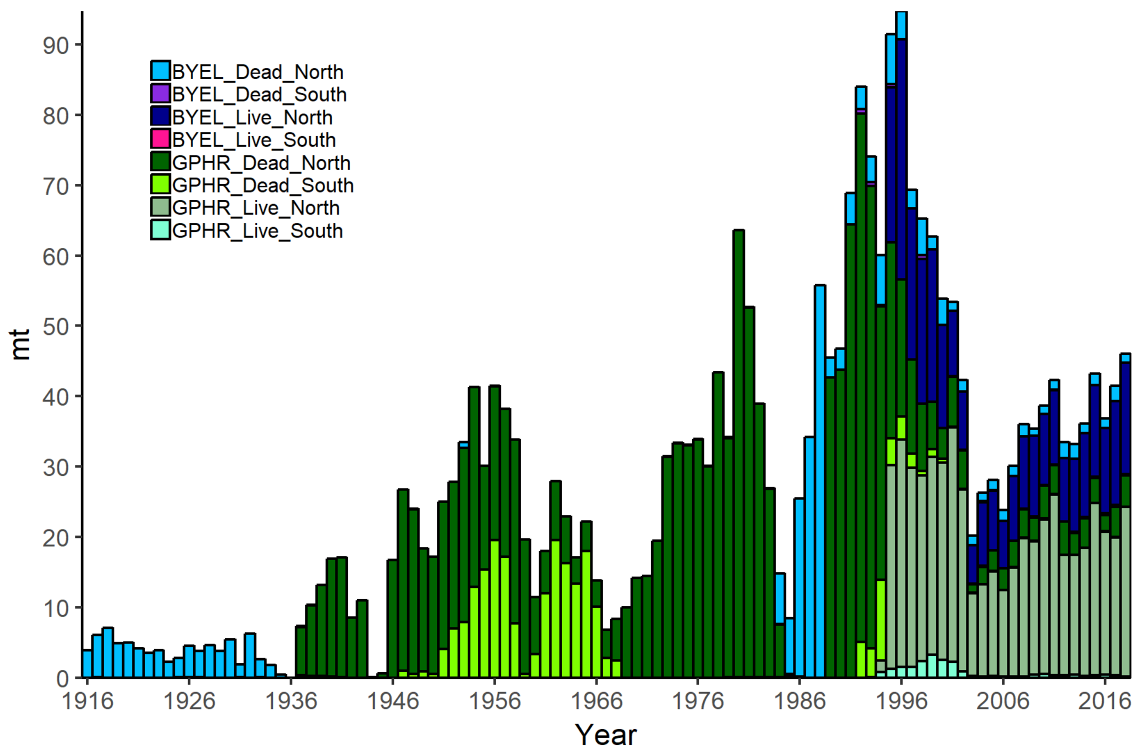

Figure 2: Commercial landings for gopher (GPHR) and black-and-yellow (BYEL) rockfishes landed live and dead north and south of Pt. Conception. All catch time series were combined for the assessment into one commercial fleet. fig:Catches_livedeadNS_gby

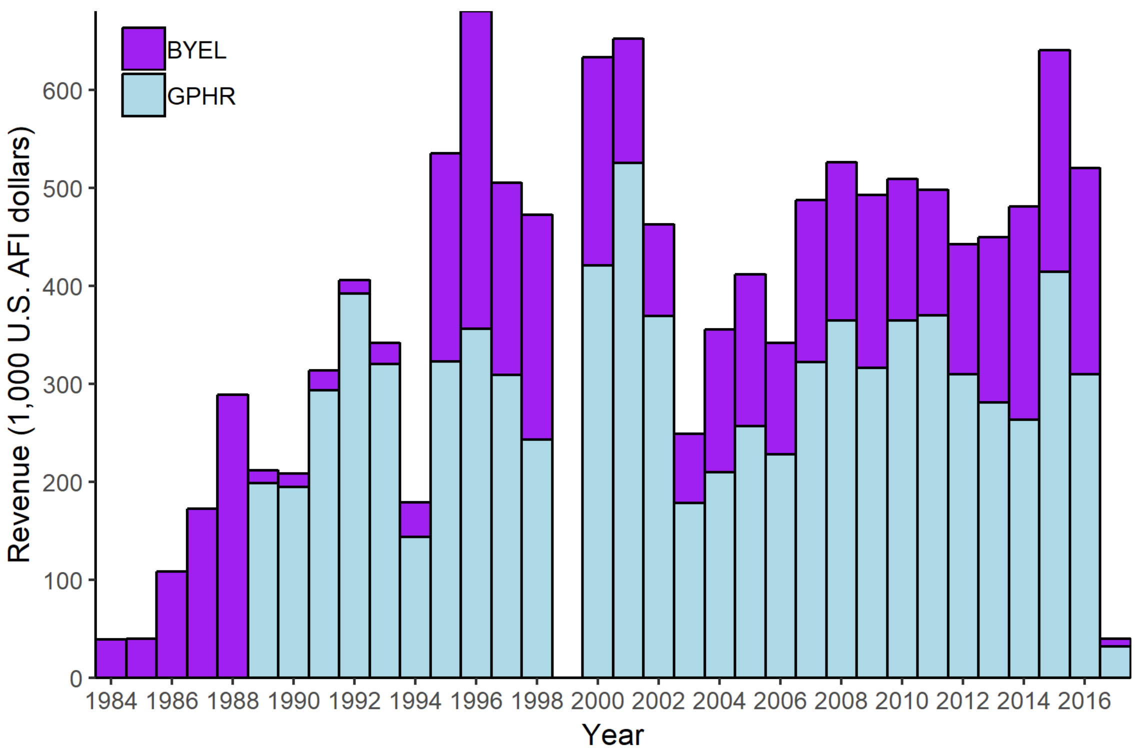

Figure 3: Annual ex-vessel revenue, adjusted for inflation (AFI) in thousands of dollars for gopher and black-and-yellow rockfish. fig:GBY_revenue

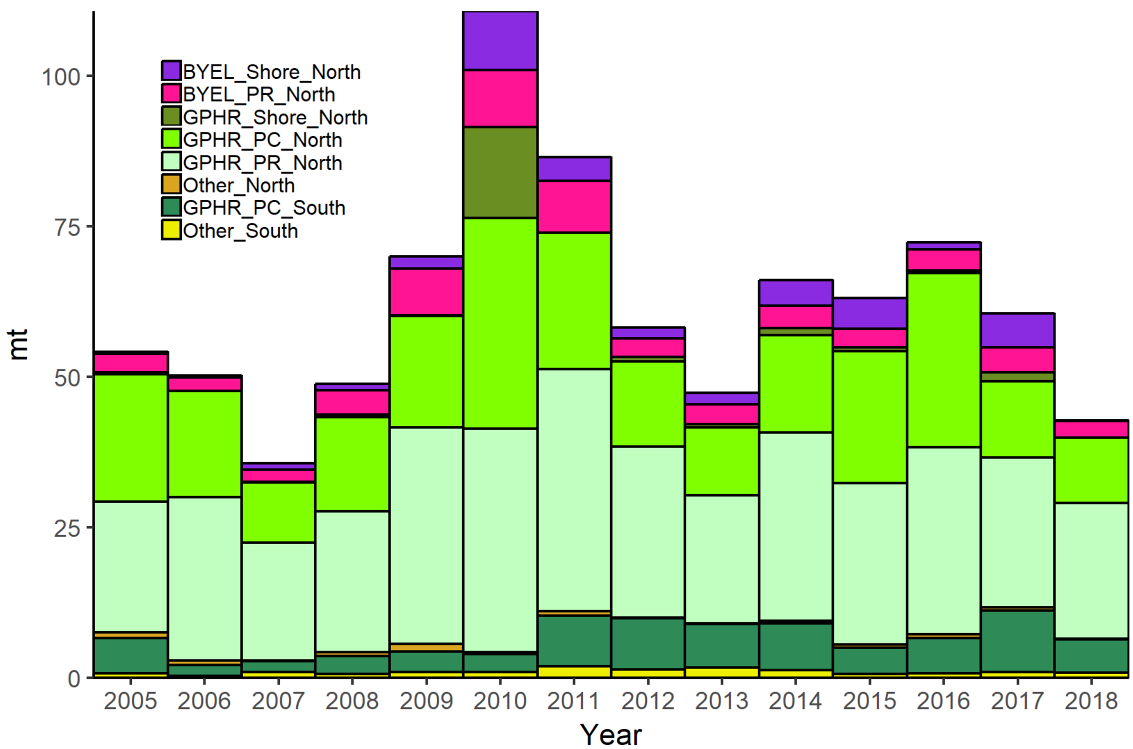

Figure 4: Recreational total mortality for gopher rockfish (GPHR) and black-and-yellow (BYEL) rockfish from the CDFW CRFS sampling era by mode and split north and south of Pt. Conception. fig:CFRS_catches

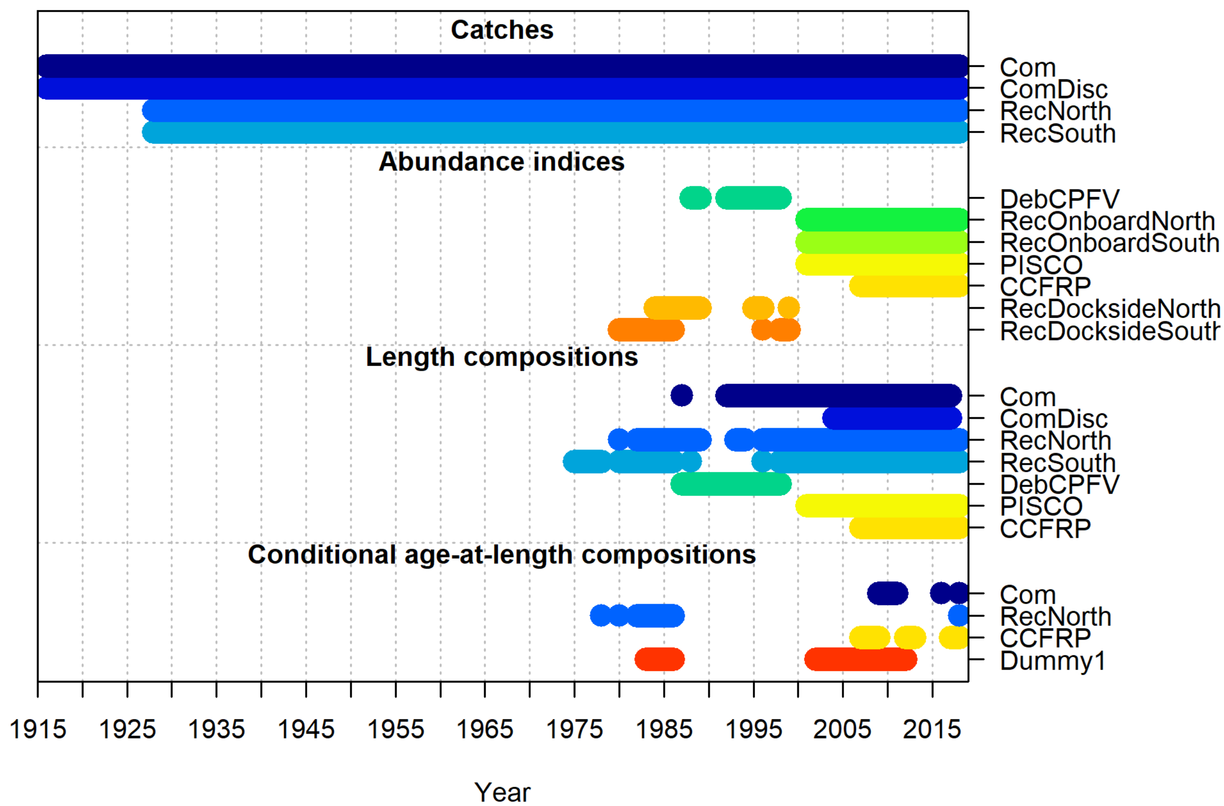

Figure 5: Summary of data sources used in the model. fig:data_plot

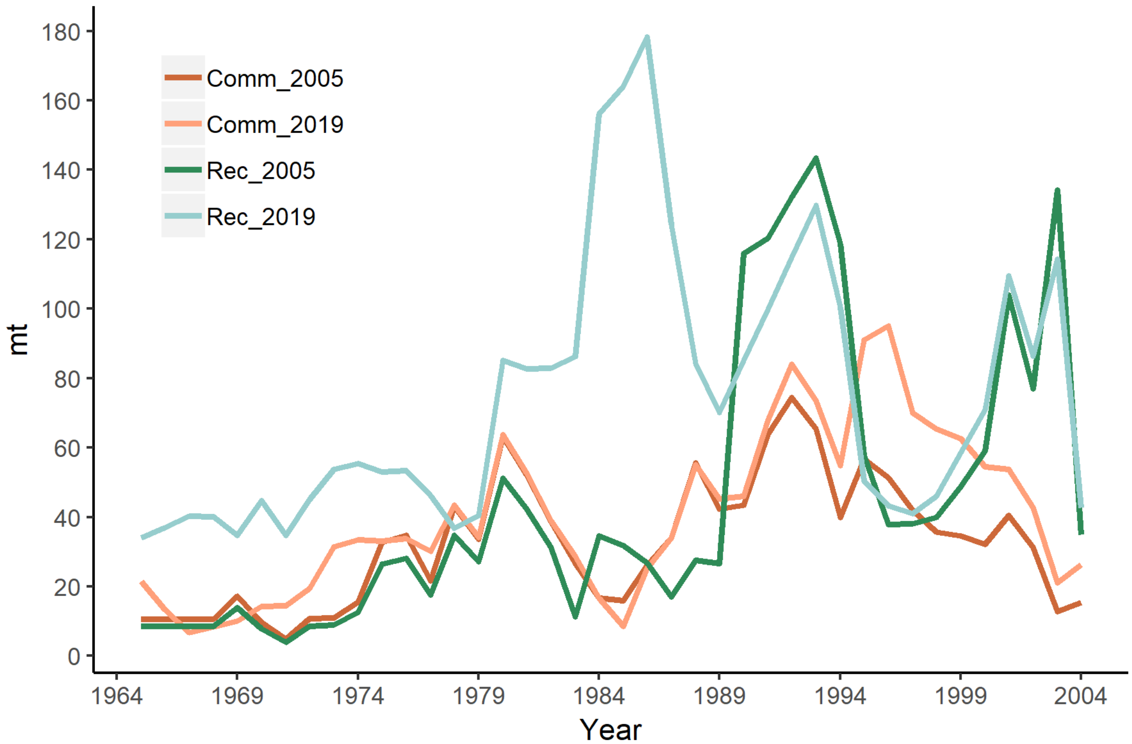

Figure 6: Comparison of the recreational and commercial fishery landings from the 2005 assessment to this 2019 assessment. Note that the 2019 assessment includes both gopher and black-and-yellow rockfish where the 2005 assessment represents gopher rockfish only. The 2005 assessment also did not include landings from south of Pt. Conception. fig:Assessment_compar

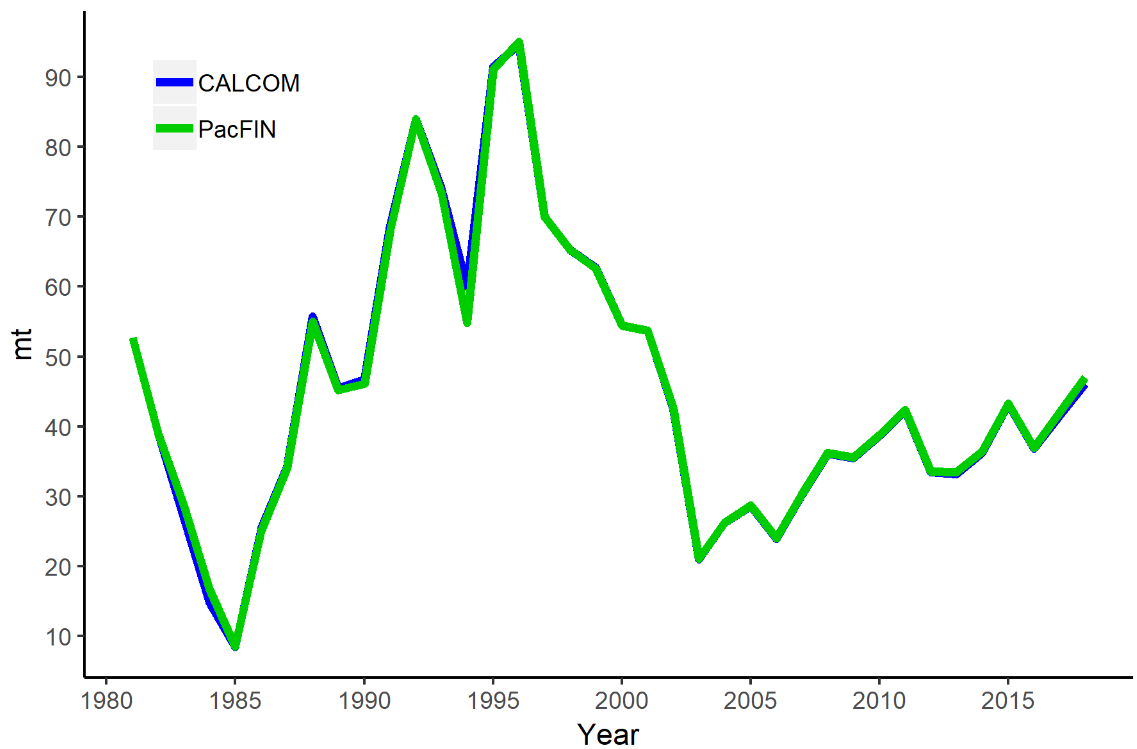

Figure 7: Commercial landings estimates from CALCOM add PacFIN. fig:Calcom_vs_Pacfin

Figure 8: Percent differences in the expanded length compositions by year from CALCOM and PacFIN. The same market categories were used for each dataset, but each database was subject to further independent filtering criteria and expansion algorithms. fig:Calcom_vs_pacfin_length

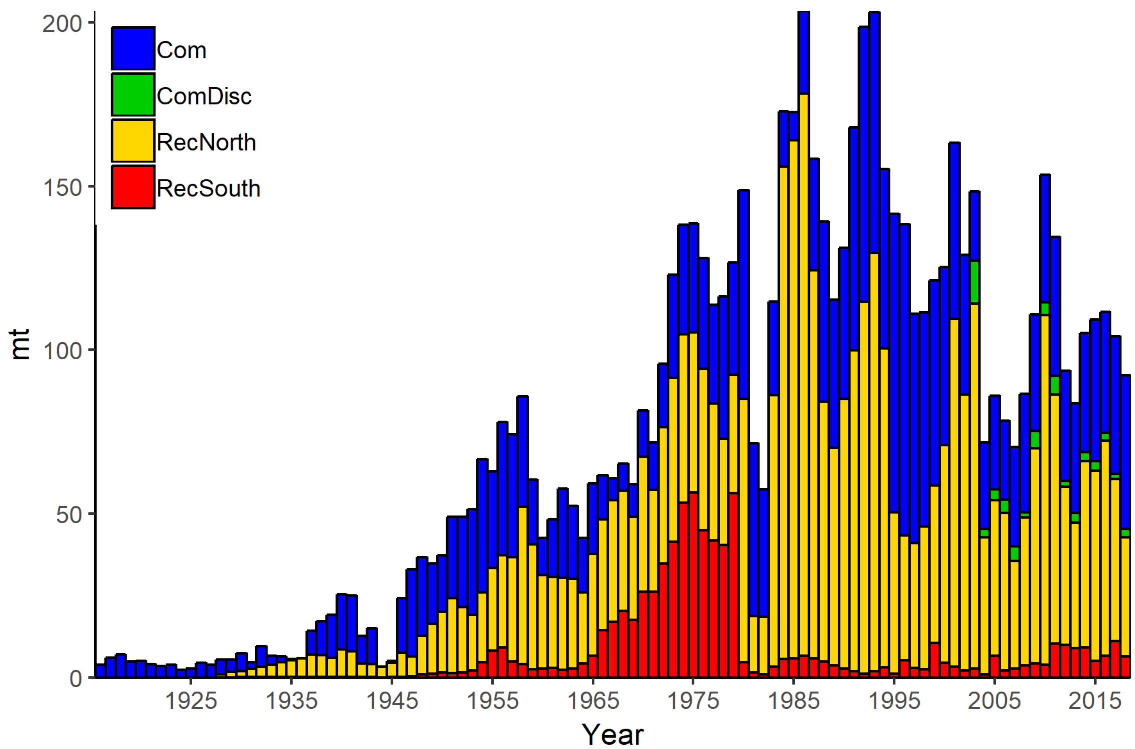

Figure 9: Commercial and recreational landings estimates prior to any data modification or interpolation to the recreational catches or hindcasting of commercial discards. fig:Catches_original

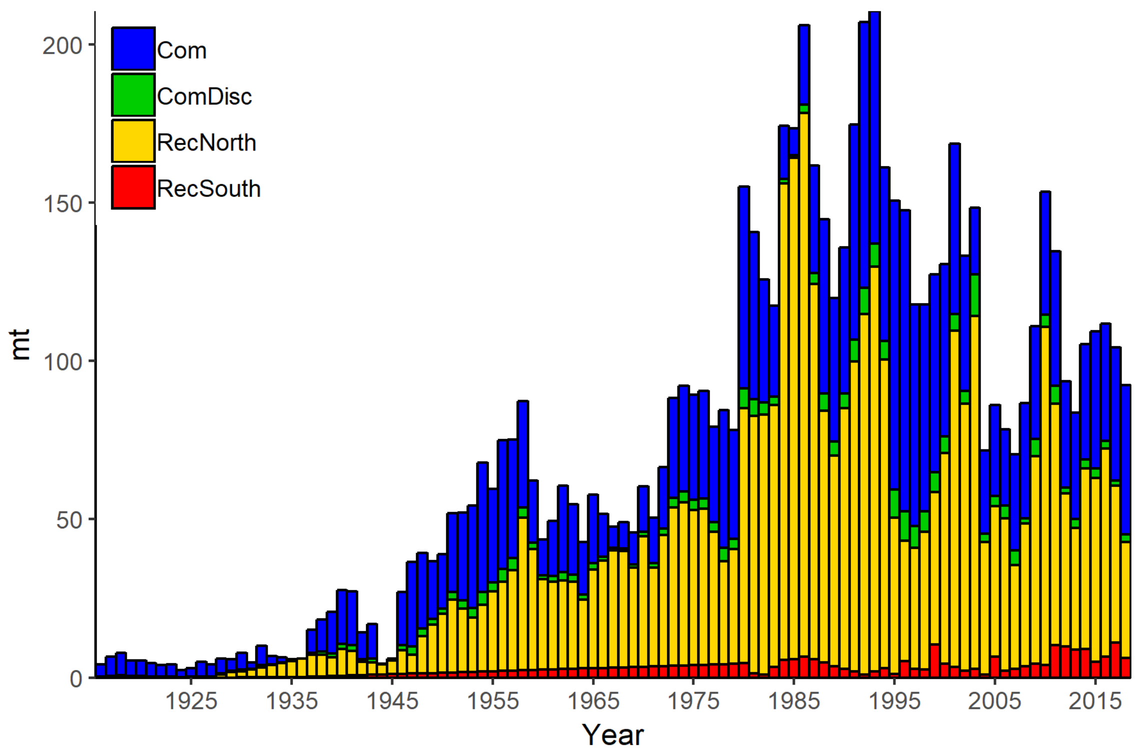

Figure 10: Commercial and recreational landings estimates after data modification and interpolations were made to the recreational catches and commercial discards. fig:Catches_alternate

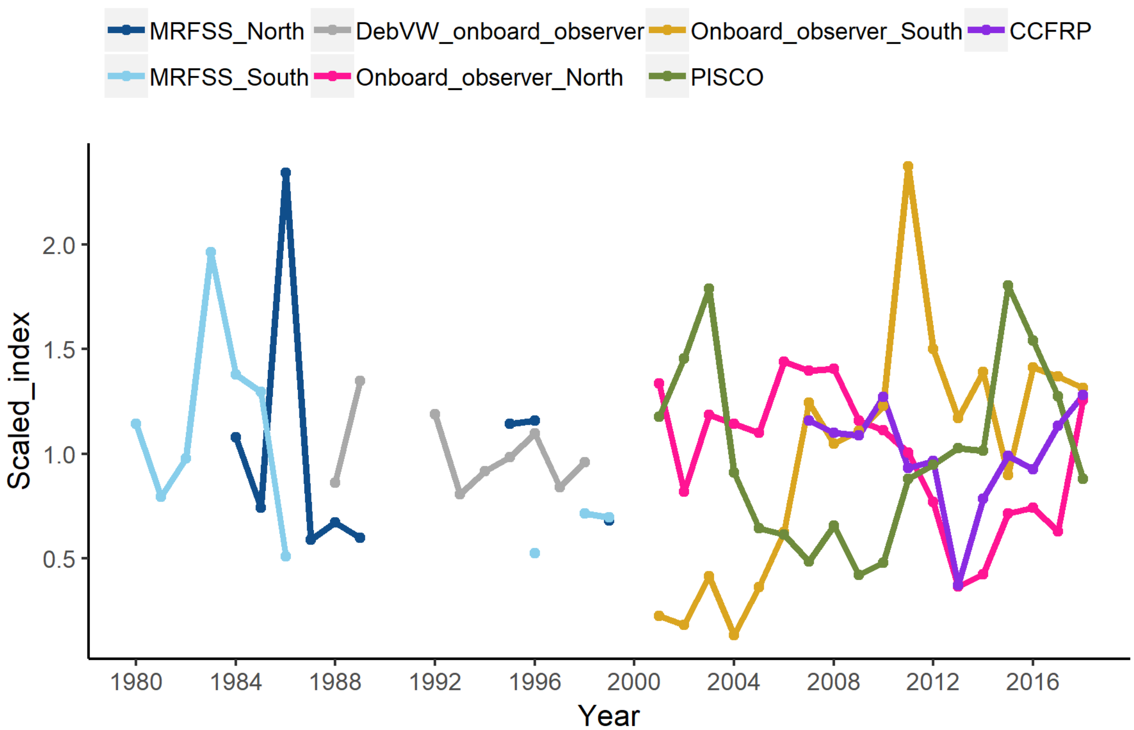

Figure 11: Comparison of all indices of abundance (scaled to their means). fig:All_index_compar

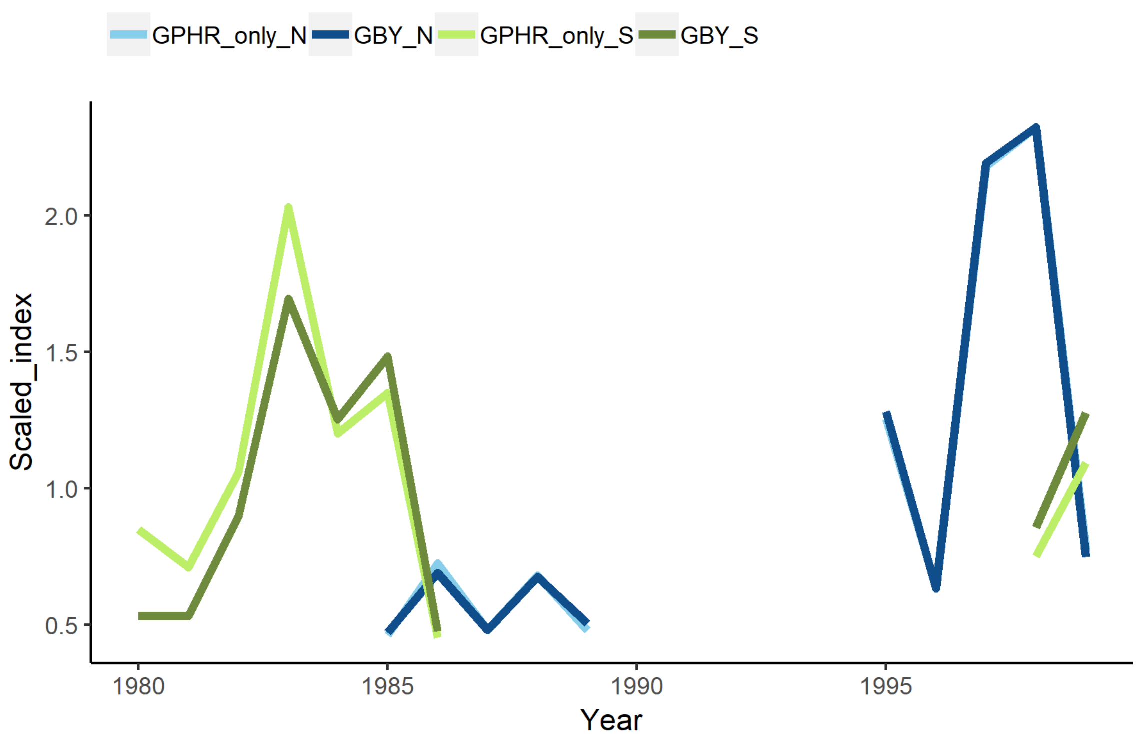

Figure 12: Comparison of indices of abundance (scaled to their means) for the MRFSS dockside CPFV survey between a gopher-only and GBYR complex index for north and south of Pt. Conception. fig:MRFSS_index_compare

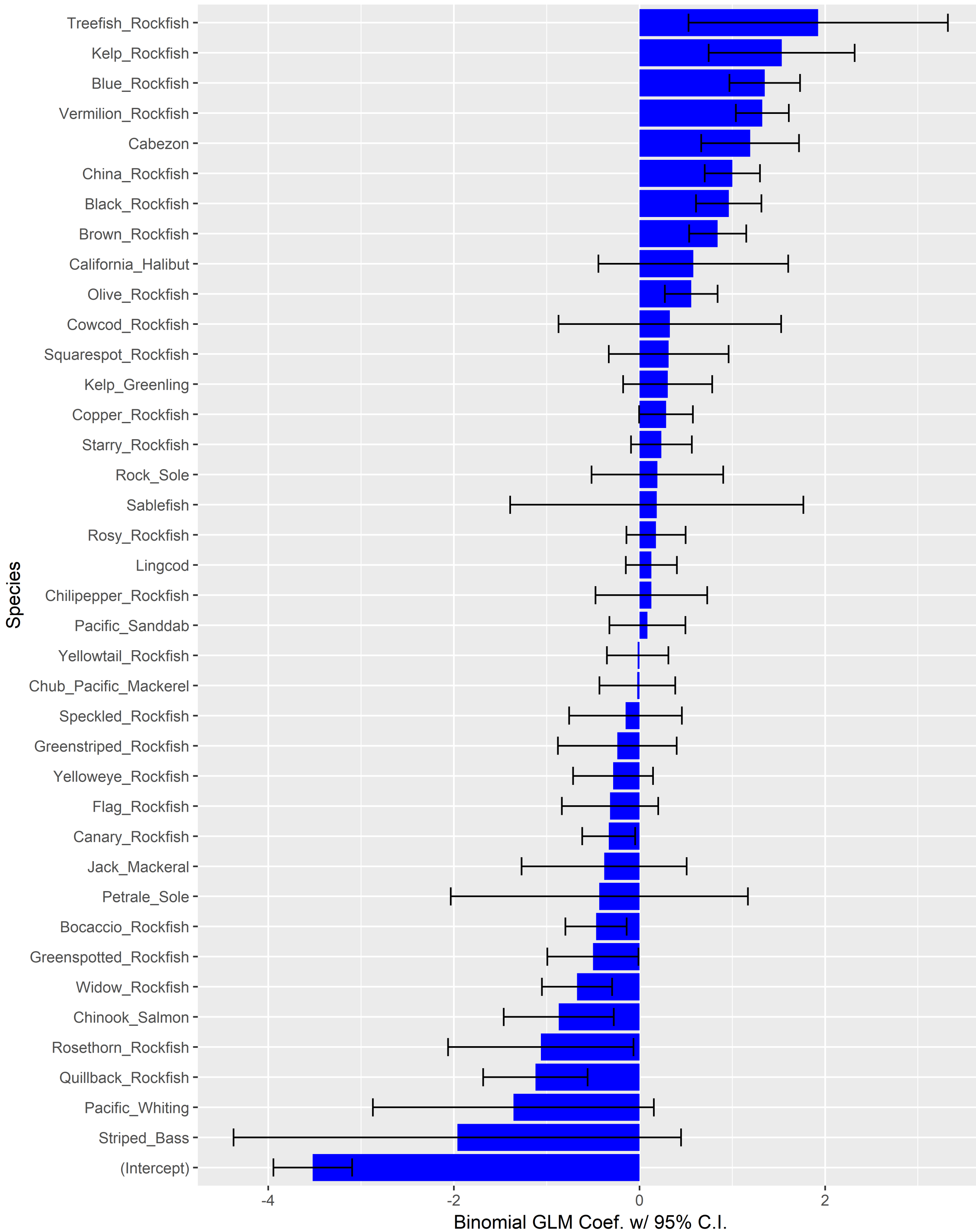

Figure 13: Species coefficients from the binomial GLM for presence/absence of GBYR in the Marine Recreational Fisheries Statistics Survey (MRFSS) CPFV mode dockside survey data set north of Pt. Conception. Horizontal bars are 95% confidence intervals. fig:Fleet10_SM_filter

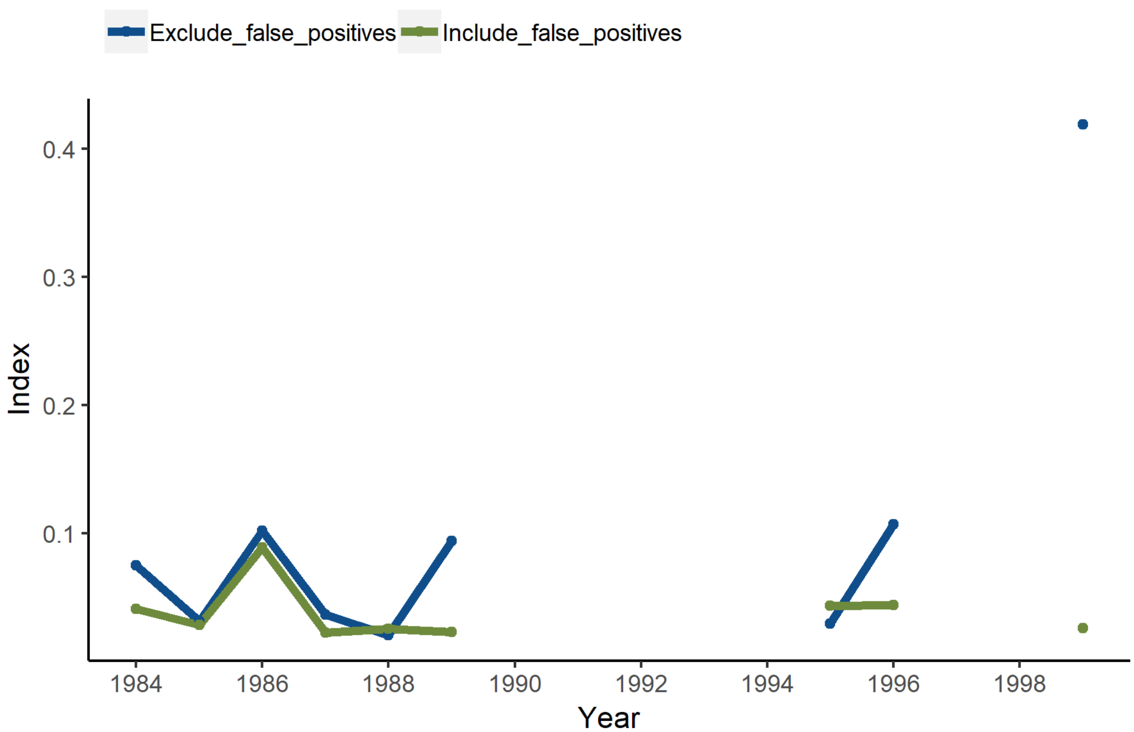

Figure 14: Comparisons of the indices of abundance for GBYR north of Pt. Conception from the MRFSS dockside CPVS survey that either include of exclude the trips identified as false positives from the Stephens-MacCall filter. fig:MRFSS_index_N_SM_falsepos

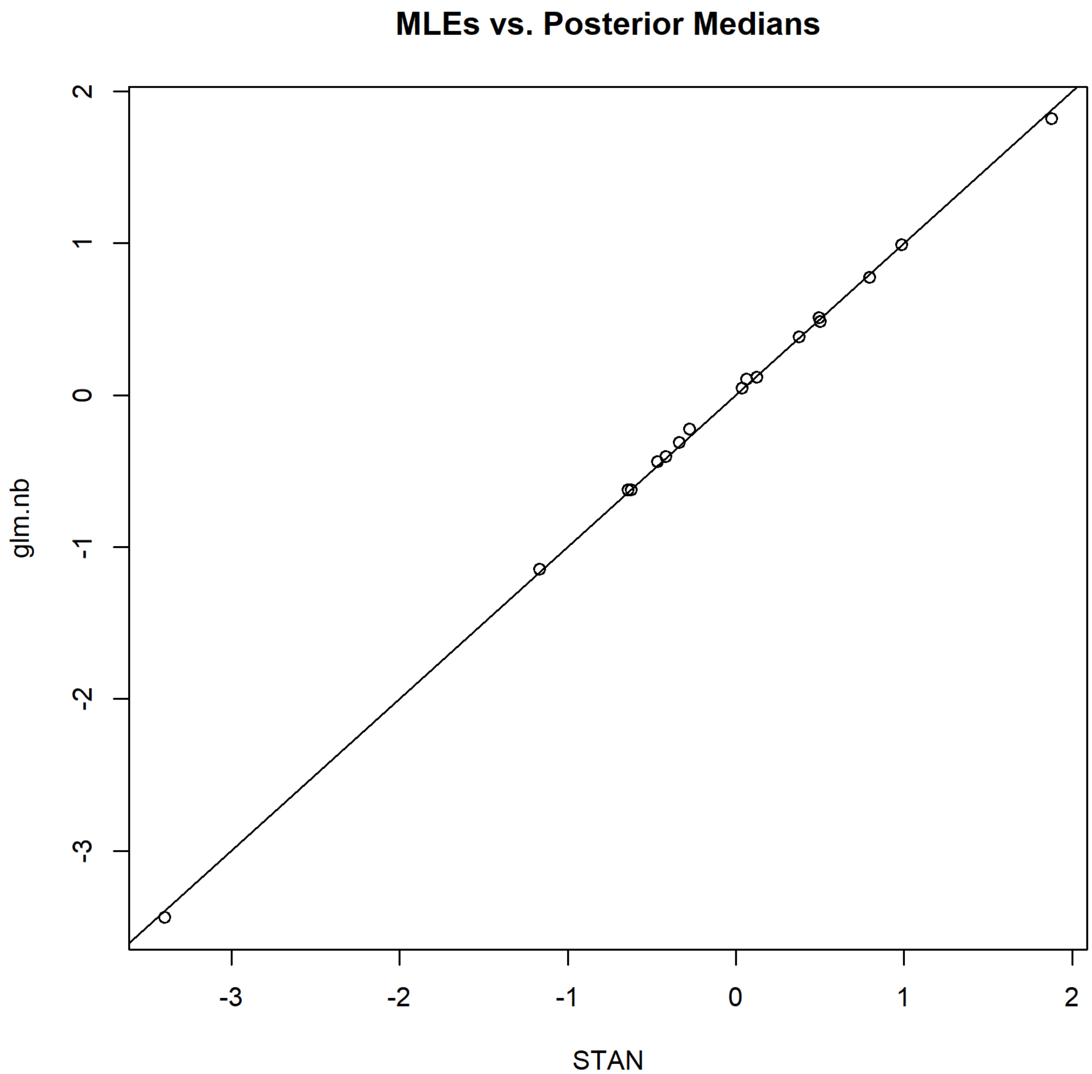

Figure 15: Comparison of negative binomial predictions (CPUE) to observed means in each stratum (year) MRFSS CPFV dockside index north of Pt. Conception. The 1:1 plot is for reference. fig:Fleet10_MLE_stan

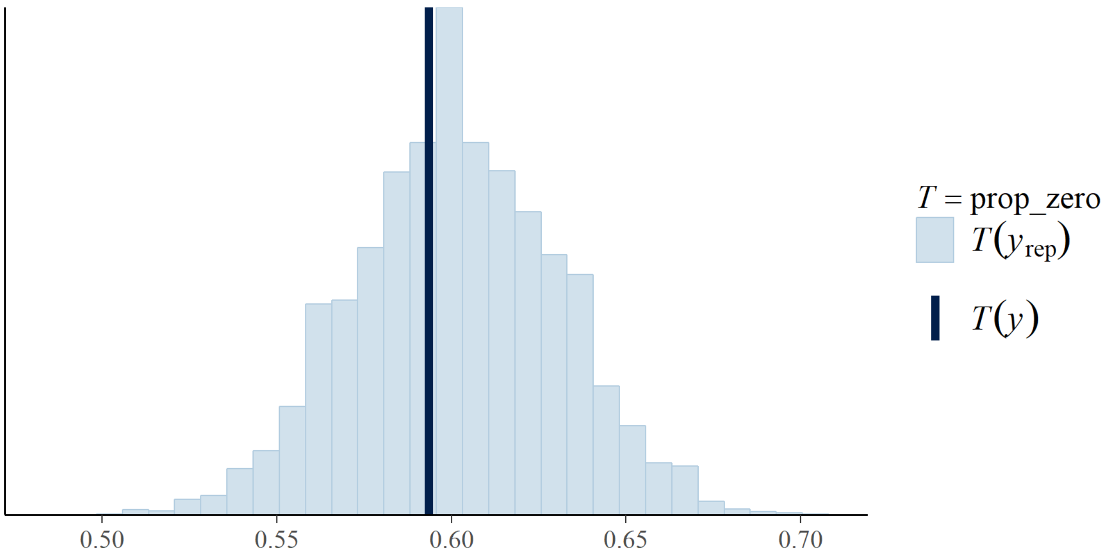

Figure 16: Posterior predictive distribution of the proportion of zero observations in replicate data sets generated by the negative binomial model for MRFSS dockside CPFV index north of Pt. Conception. fig:Fleet10_prop_zero_STAN

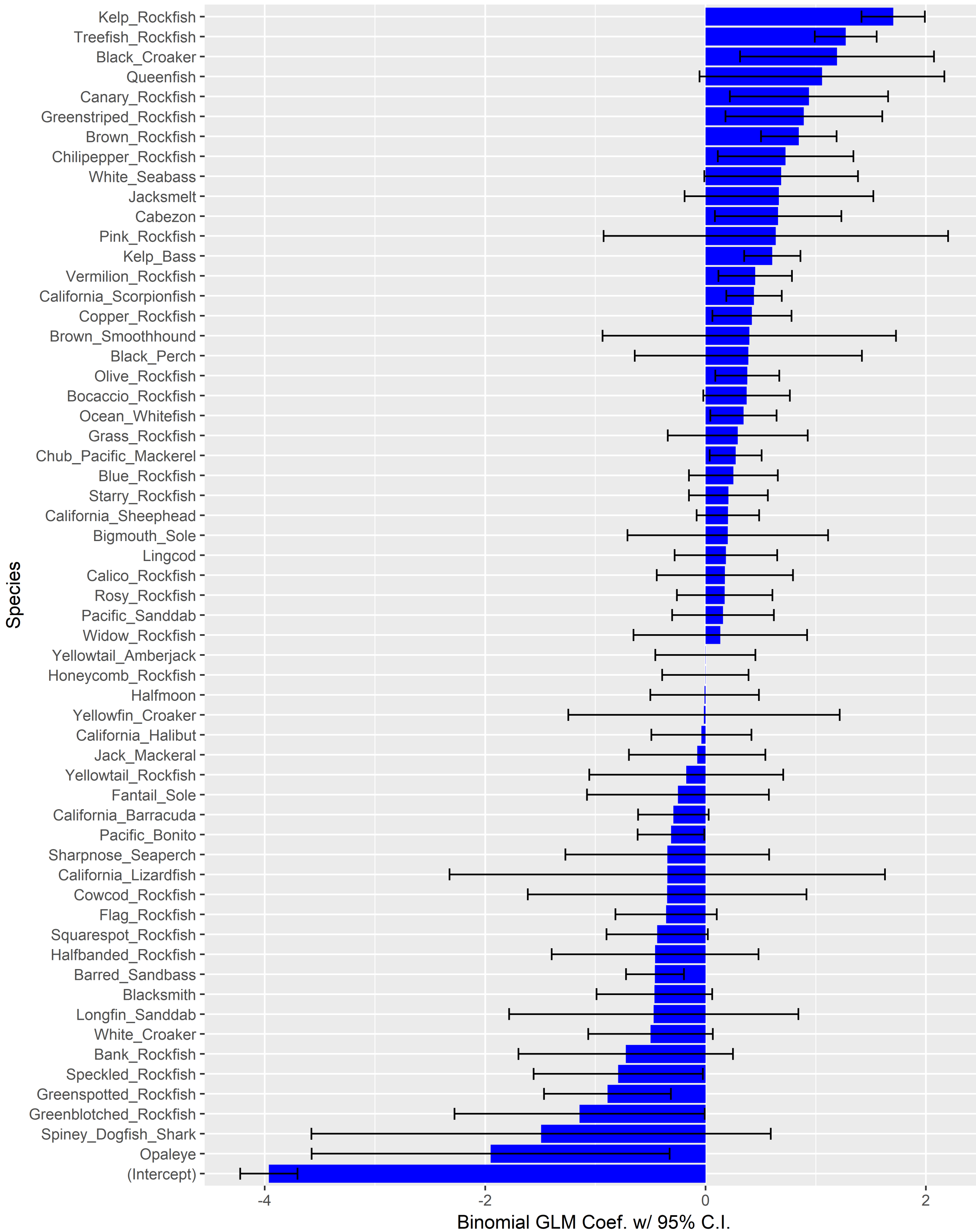

Figure 17: Species coefficients from the binomial GLM for presence/absence of GBYR in the Marine Recreational Fisheries Statistics Survey (MRFSS) CPFV mode dockside survey data set north of Pt. Conception. Horizontal bars are 95% confidence intervals. fig:Fleet11_SM_filter

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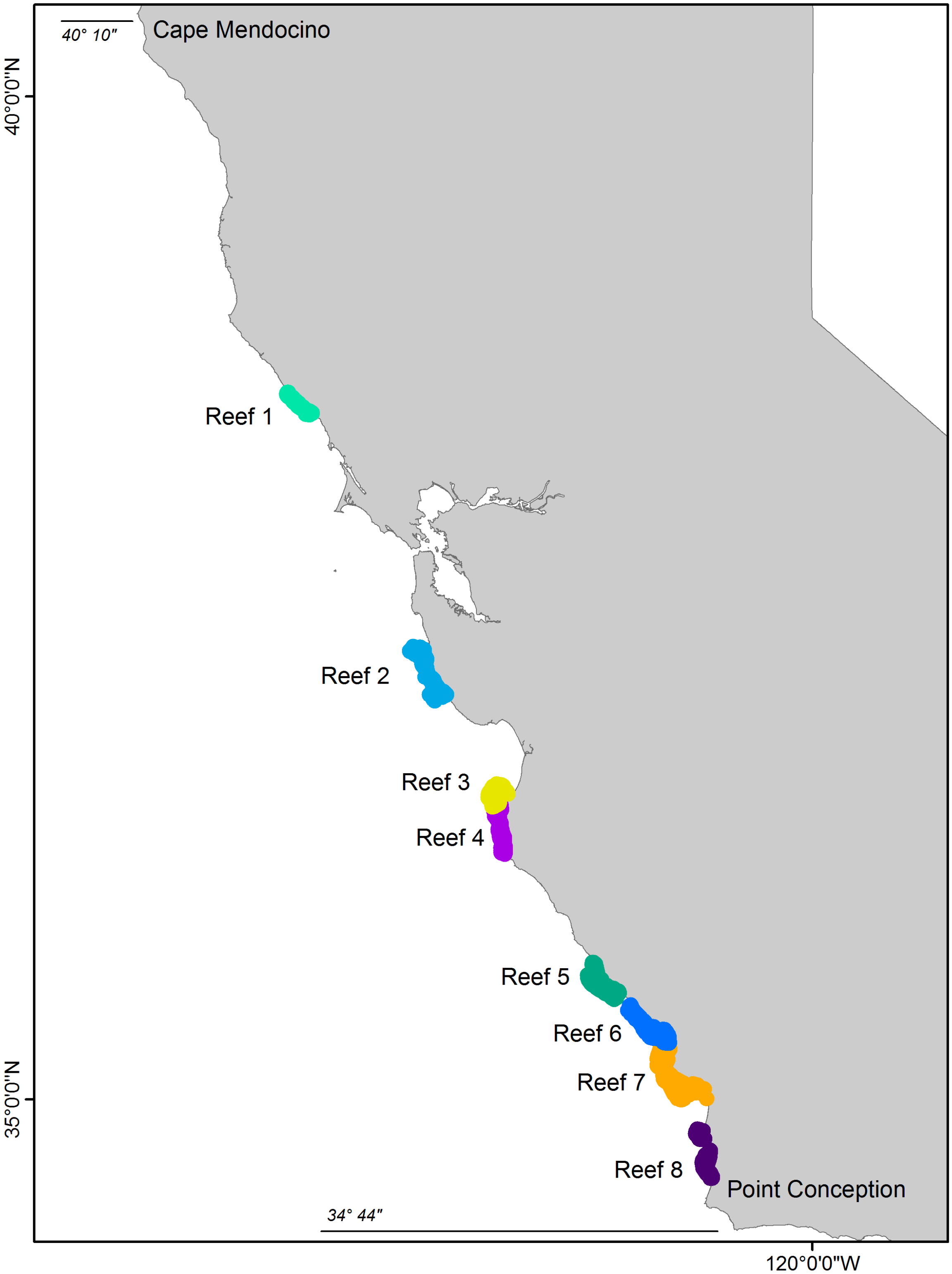


Figure 18: Map of the reefs used in the Deb Wilson-Vandenbery CPFV onboard observer survey index of abundance. fig:DebWV_sites

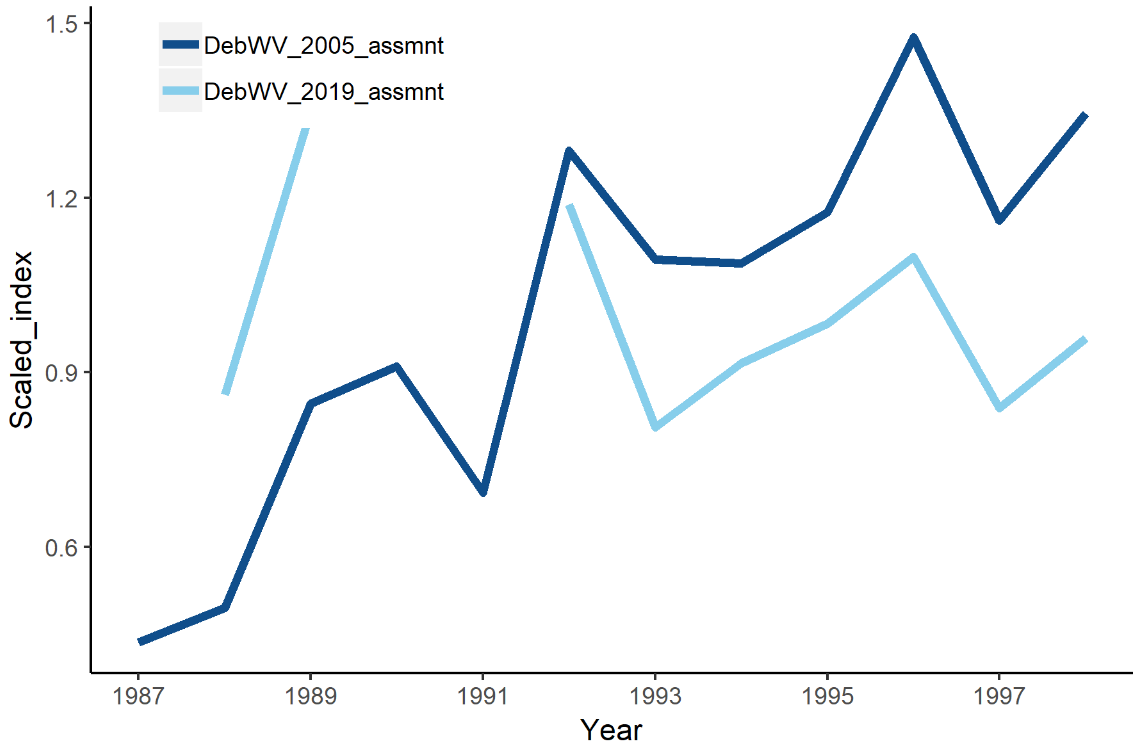


Figure 19: Comparison of the index developed for the Deb Wilson-Vandenbery CPFV on-board observer survey from the 2005 assessment and forthe 2019 assessment. fig:DebWV_index_compare

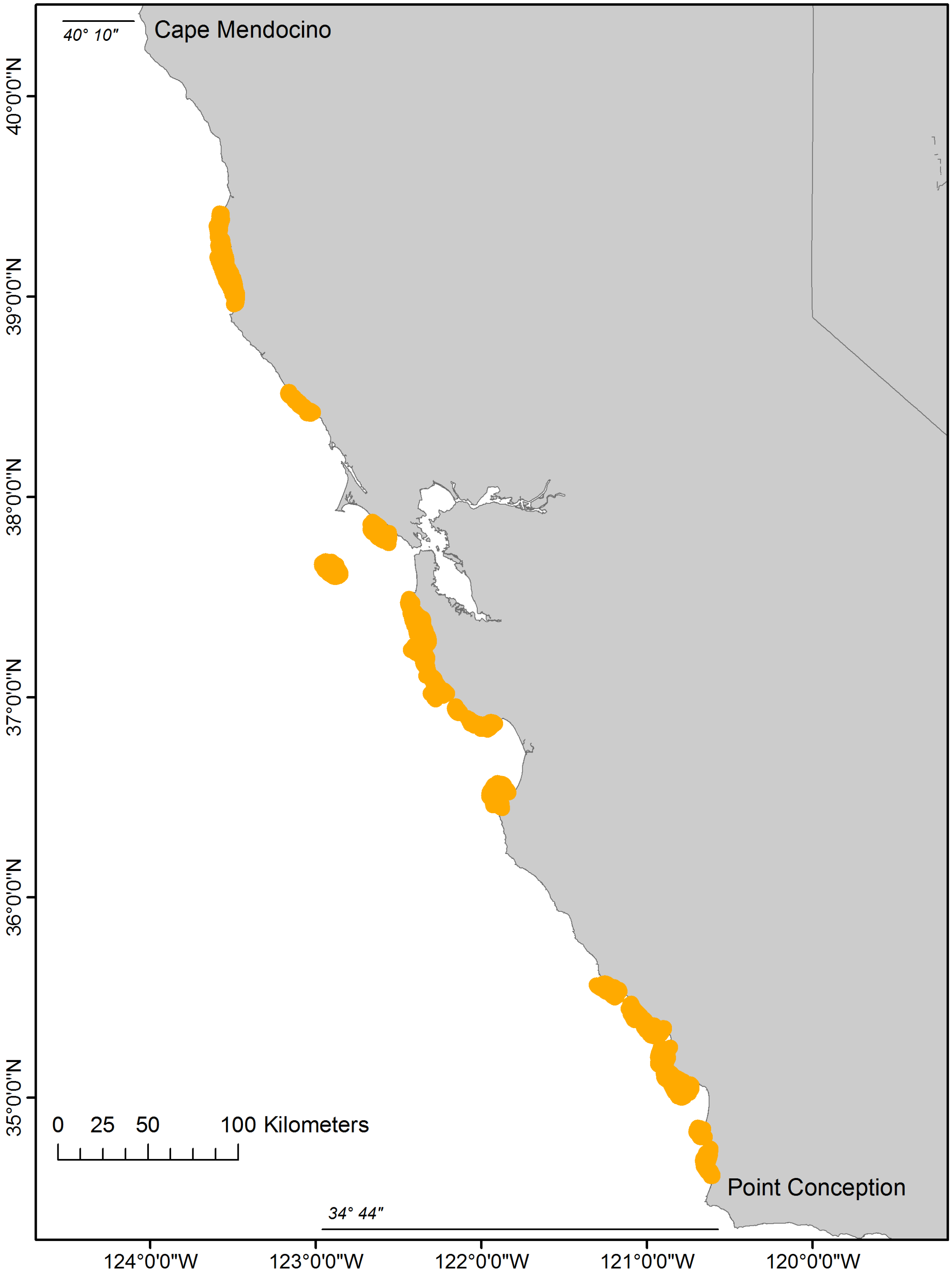


Figure 20: Map of the reefs selected for the final index for the onboard observer surveys (CDFW and Cal Poly) north of Pt. Conception fig:Onboard_observer_north_sites

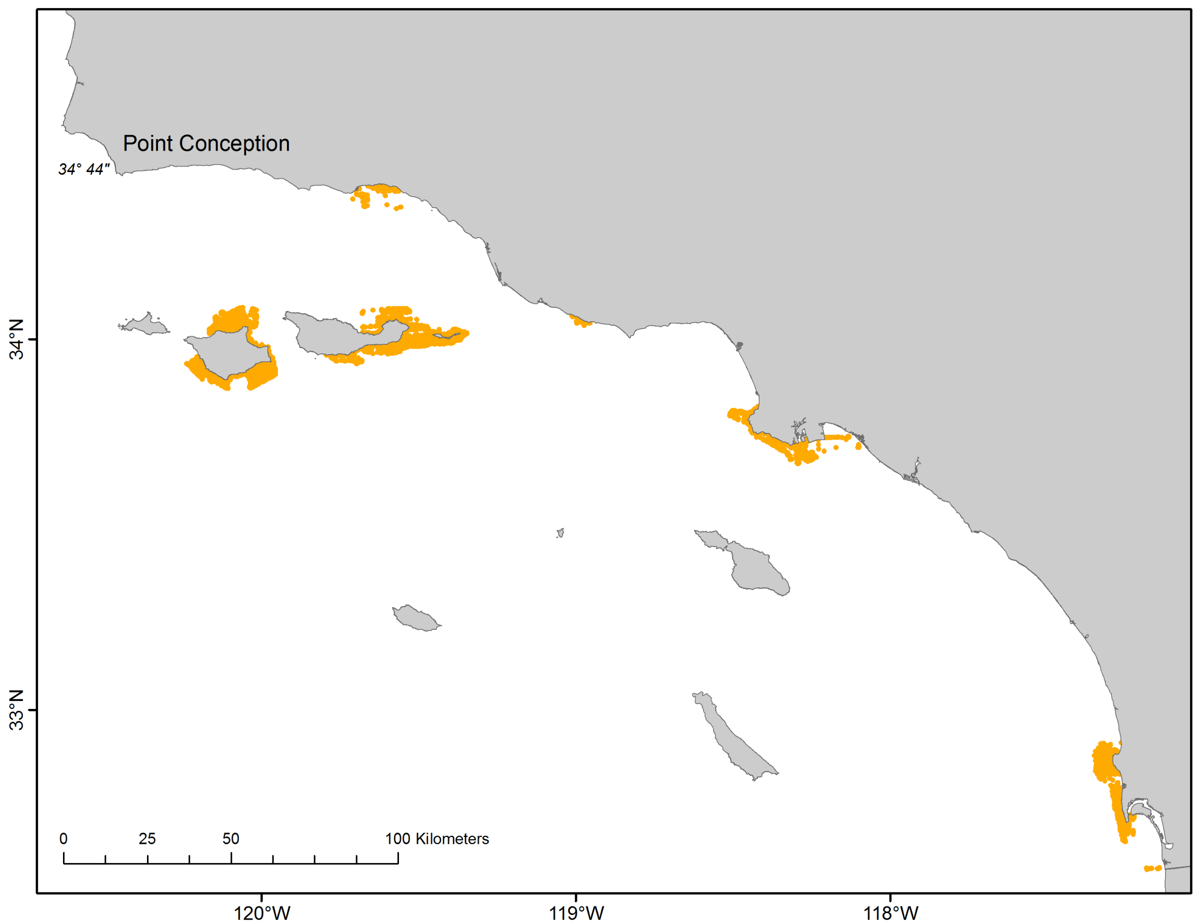


Figure 21: Map of the reefs selected for the final index for the CDFW onboard observer survey south of Pt. Conception fig:Onboard_observer_south_sites

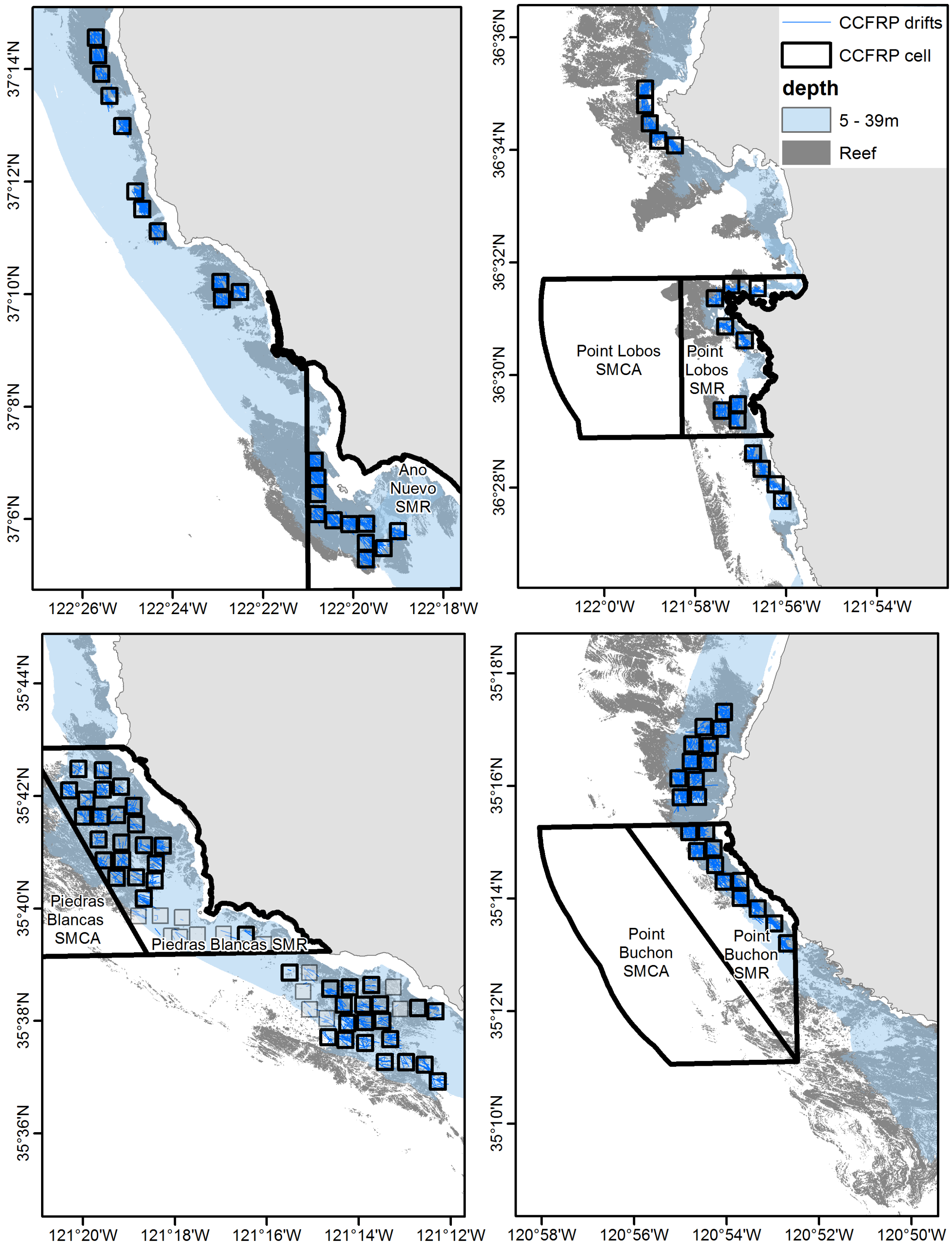


Figure 22: Map of the four MPAs sample consistently through time for the CCFRP fishery-independent survey. fig:CCFRP_sites

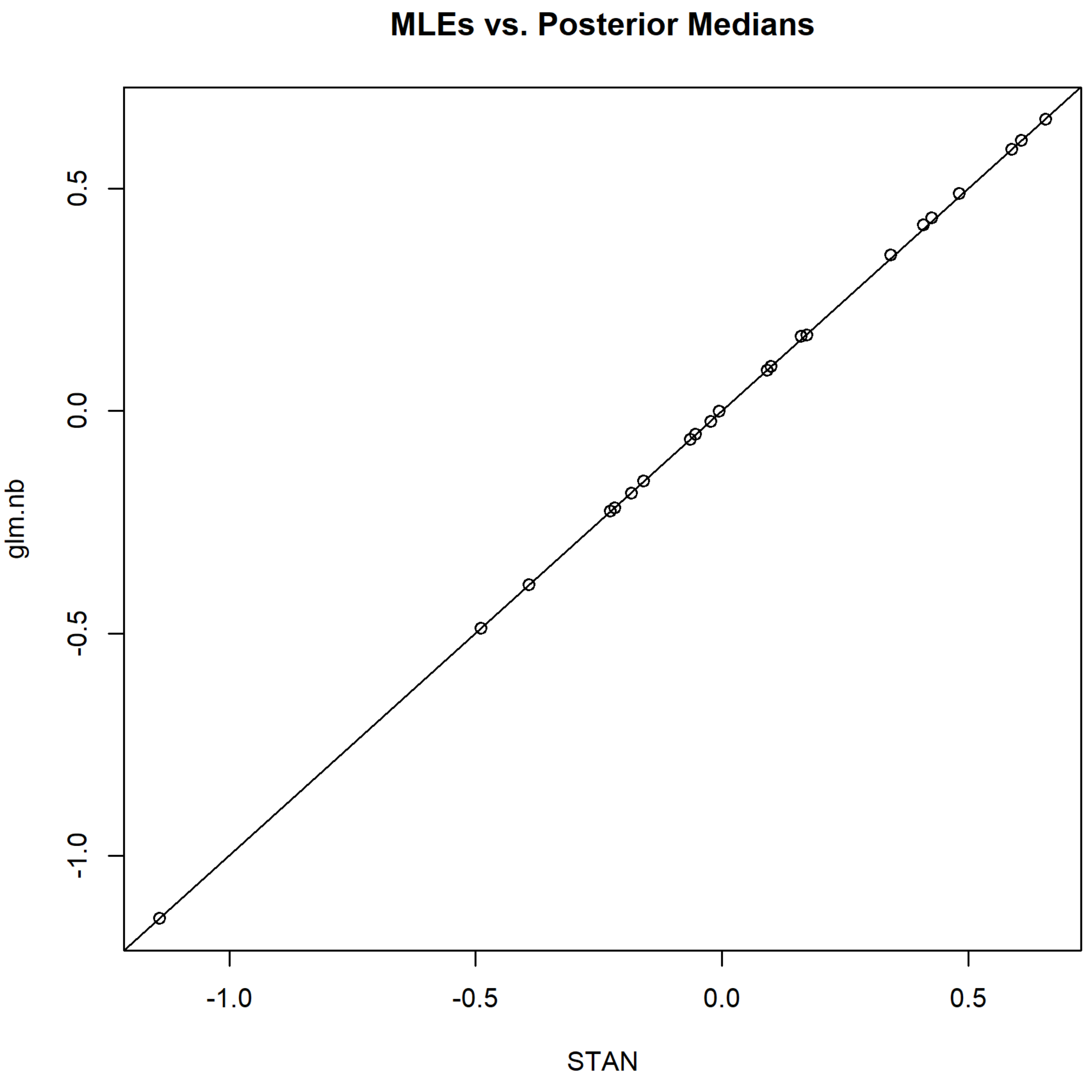


Figure 23: Comparison of negative bionimial predictions (CPUE) to observed means in each stratum (year) for the CCFRP index. The 1:1 plot is for reference. fig:Fleet9_MLE_stan

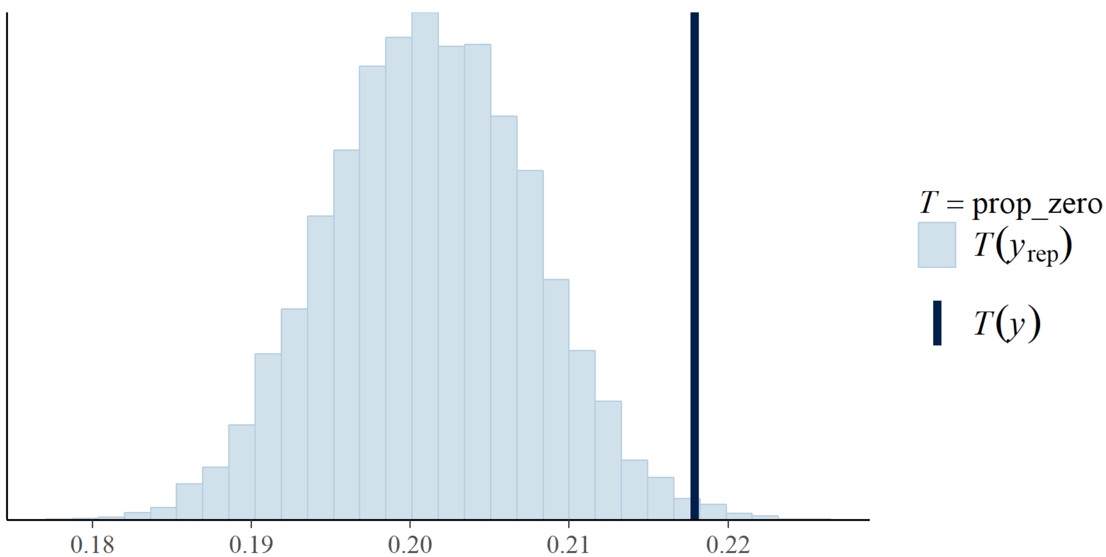


Figure 24: Posterior predictive distribution of the proportion of zero observations in replicate data sets generated by the negative binomial model for the CCFRP index. fig:Fleet9_prop_zero_STAN

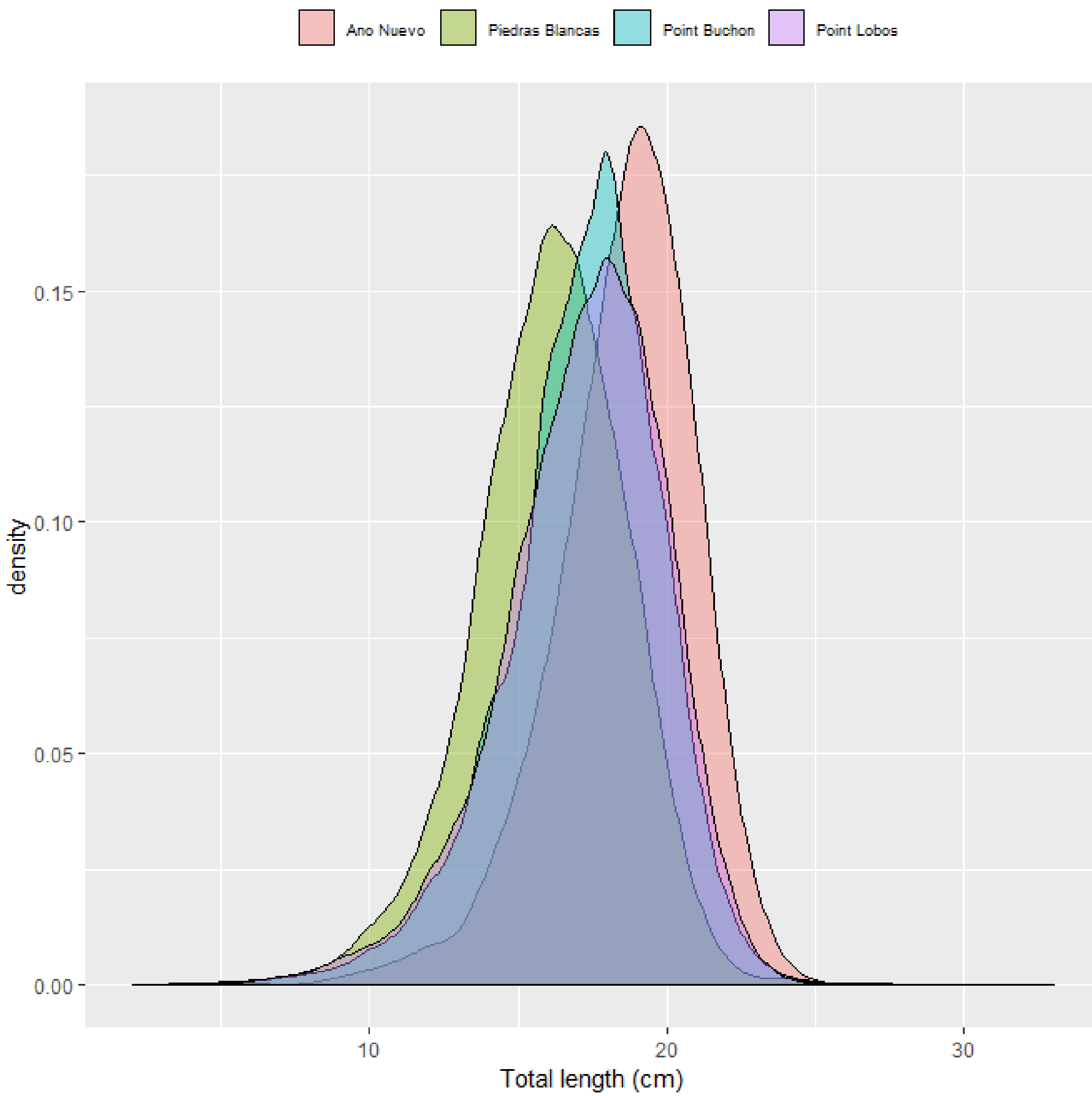


Figure 25: Length distributions of GBYR for the four MPAs sampled by the CCFRP survey used in this assessment. | CCFRP_lengths_by_site

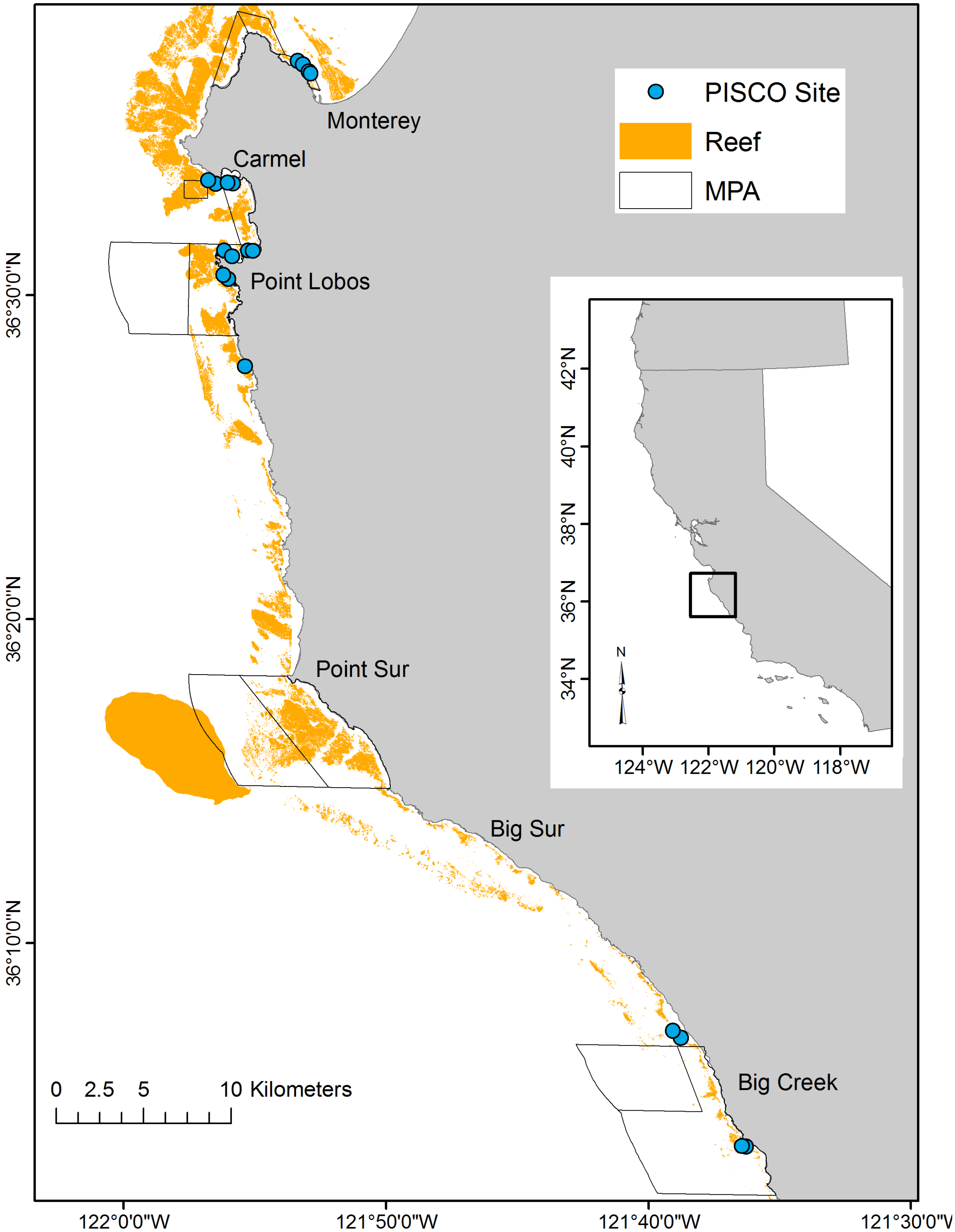


Figure 26: Map of the sites sampled consistently through time for the PISCO kelp forest fish survey. fig:PISCO_sites

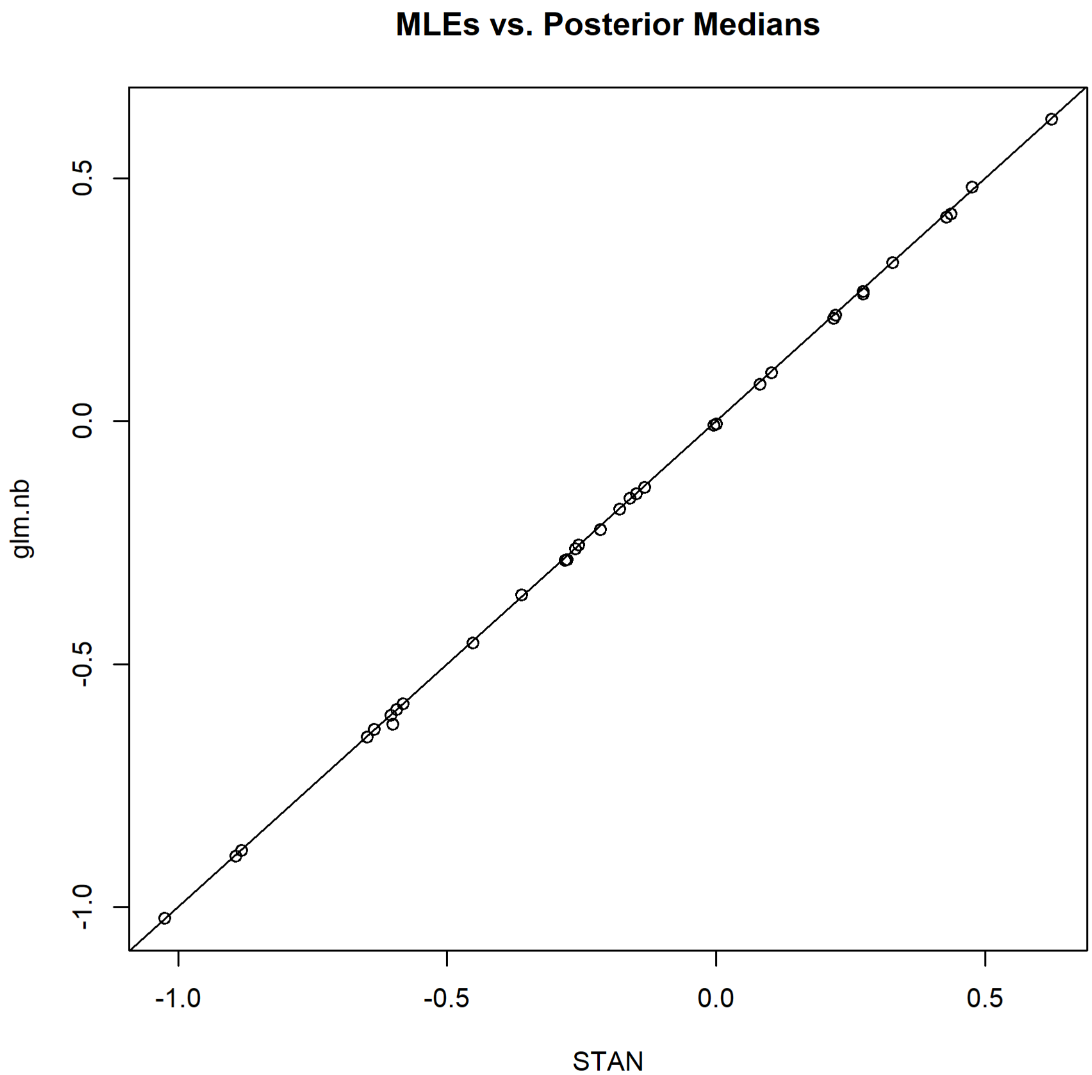


Figure 27: Comparison of negative bionimial predictions (CPUE) to observed means in each stratum (year) for the PISCO kelp forest fish survey index. The 1:1 plot is for reference. fig:Fleet8_MLE_

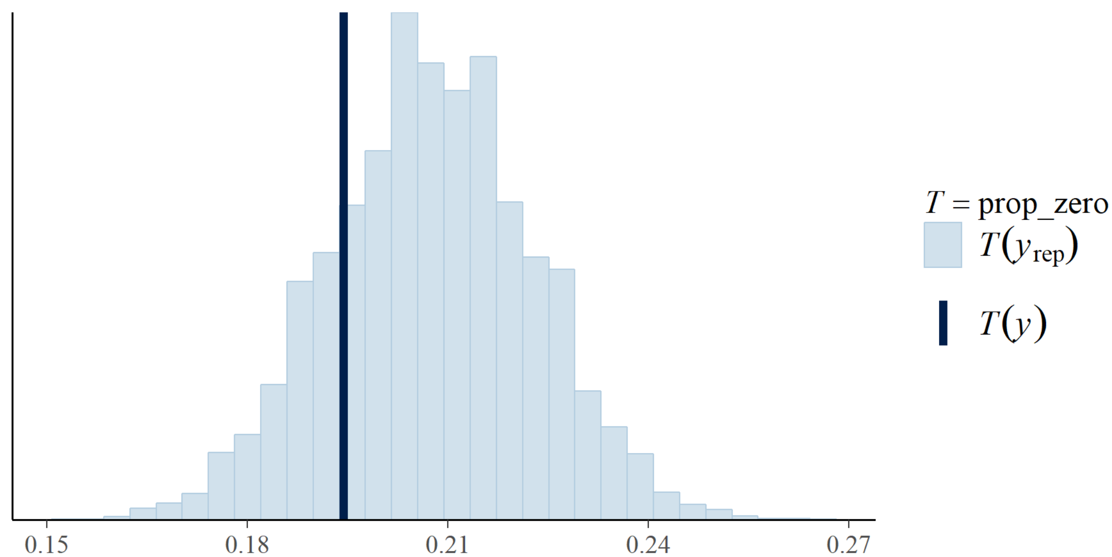


Figure 28: Posterior predictive distribution of the proportion of zero observations in replicate data sets generated by the negative binomial model for the PISCO kelp forest fish survey. fig:Fleet8_pr

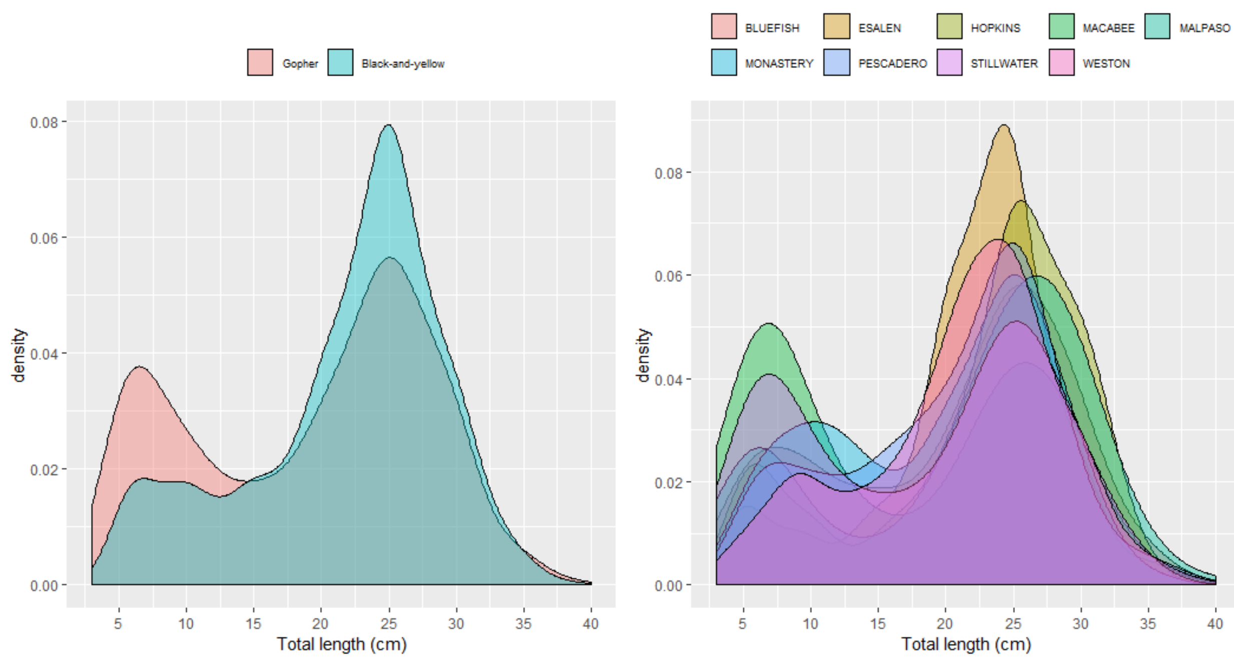


Figure 29: Plots of the length distributions from the PISCO kelp forest fish survey by species (left) and for compbined species by site (right) for sites included in the final index of abundance. fig:PISCO_lenghts

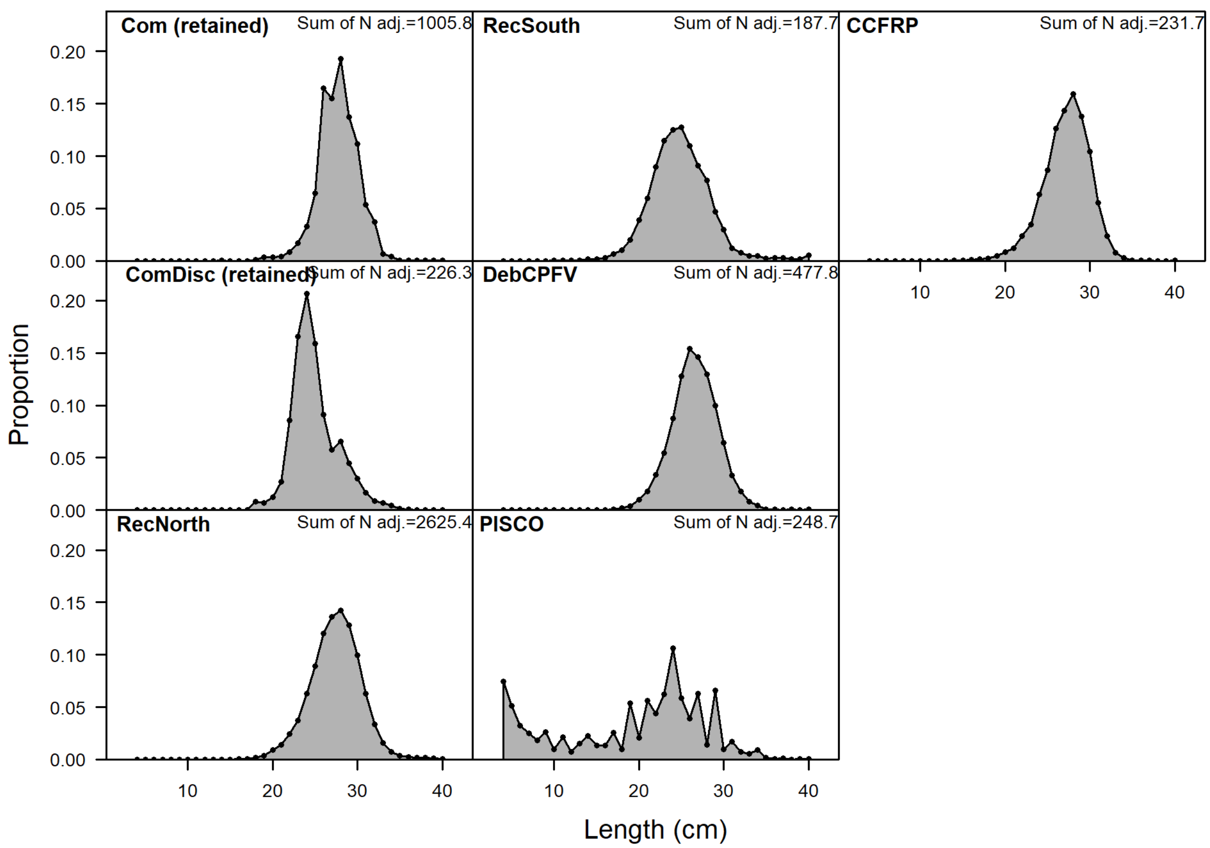


Figure 30: Length comp data, aggregated across time by fleet. Labels 'retained' and 'discard' indicate discarded or retained sampled for each fleet. Panels without this designation represent the whole catch. fig:comp_lendat_aggregated_across_time

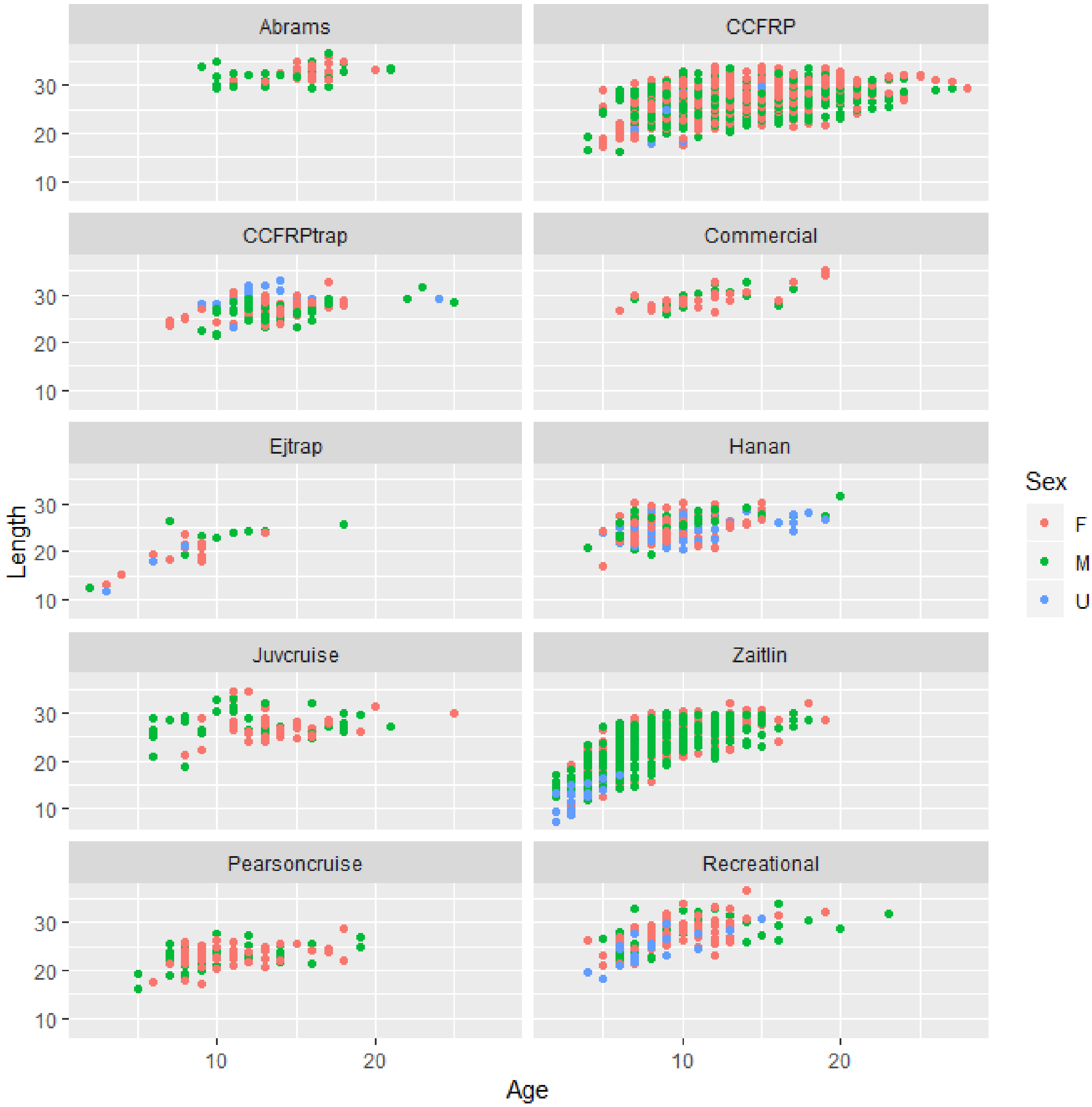


Figure 31: Available length-at-age data for gopher and black-and-yellow rockfish by sex and data source. The Zaitlin study is all black-and-yellow rockfish. The remaining plots represent gopher rockfish fig:growth_samples

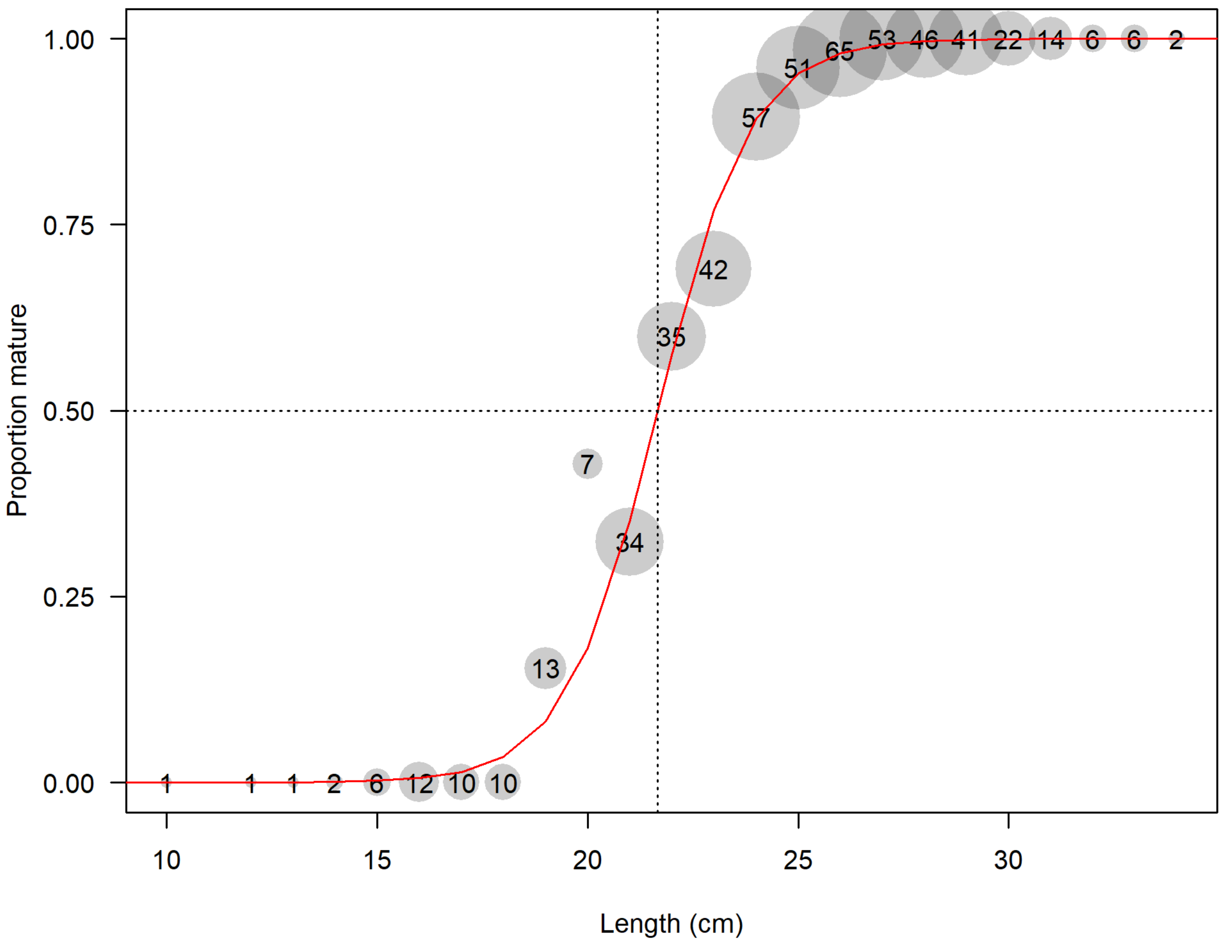


Figure 32: Maturity ogive for females estimated from black-and-yellow rockfish from Zaitlin (1986) and gopher rockfish from Meyers-Cherry (2014). Sample sizes at a given length are shown in the circles. fig:GBY_maturity_ogive

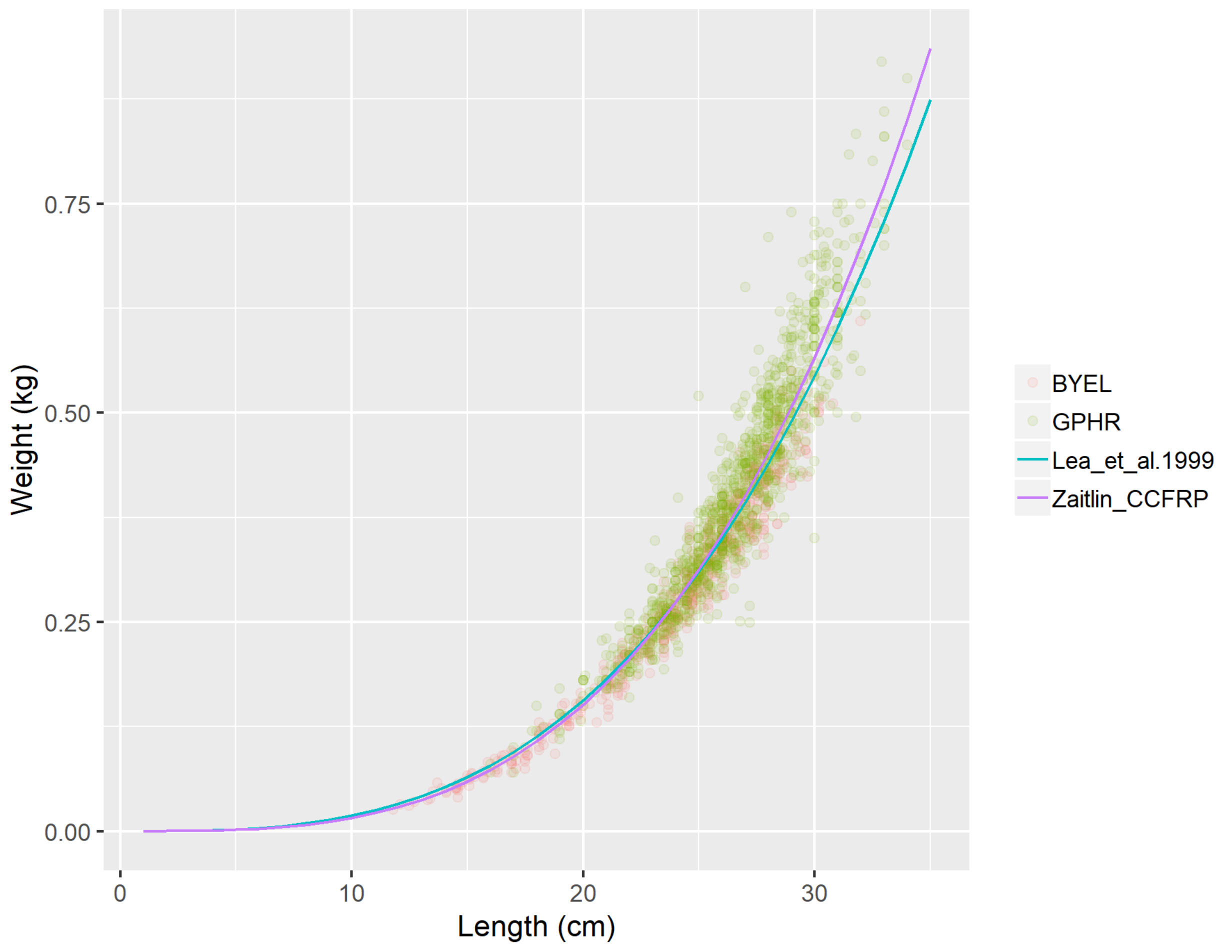


Figure 33: Comparison of the gopher rockfish weight-length curves from Lee et al. (1999) and tha estimated from black-and-yellow rockfishes from Zaitlin (1986), and gopher rockfishes from Loury (2011) and Meyers-Cherry(2014). The estimated curve from the current data is used in this assessment. fig:GBY_weight_length

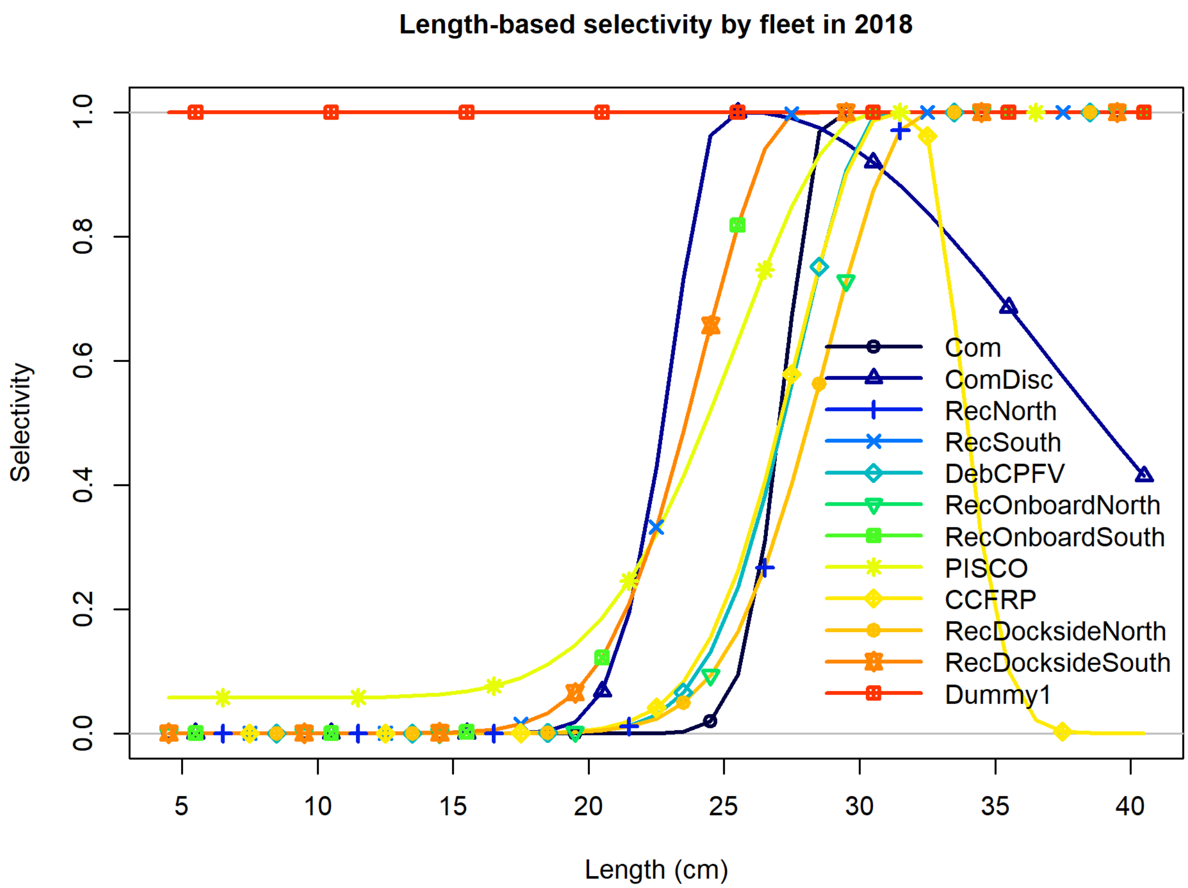


Figure 34: Selectivity at length for all of the fleets in the base model. fig:sel01_multiple_fle

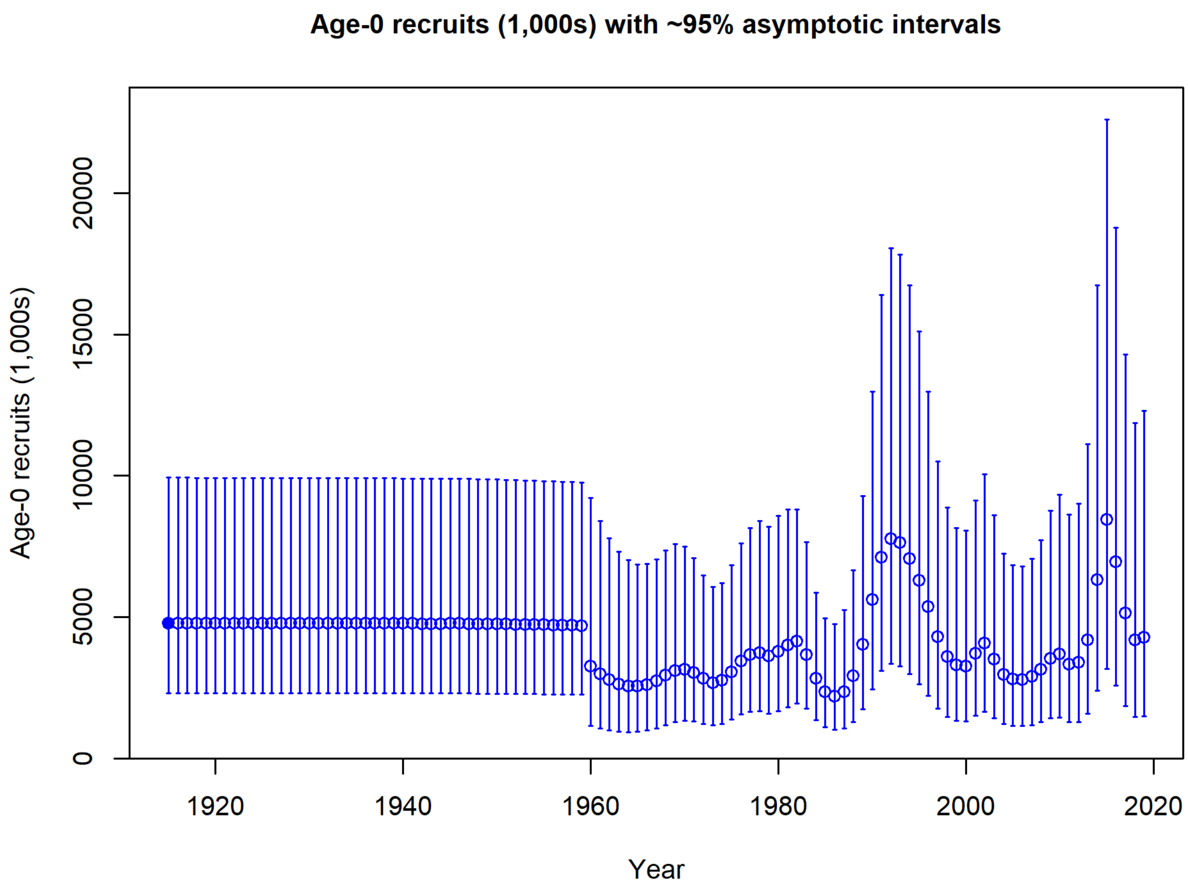


Figure 35: Time series of estimated GBYR recruitments for the base-case model with 95% confidence or credibility intervals. fig:Recruit_mod1

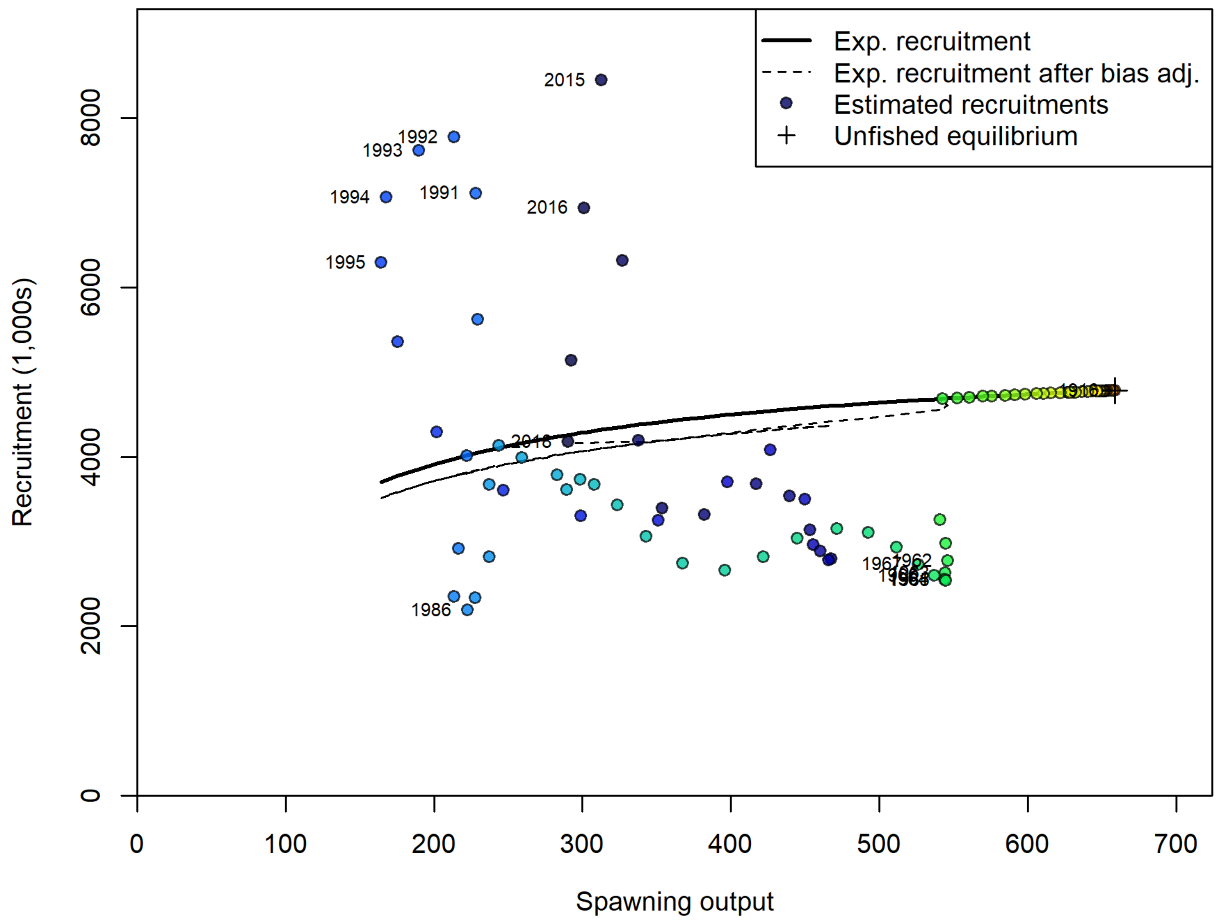
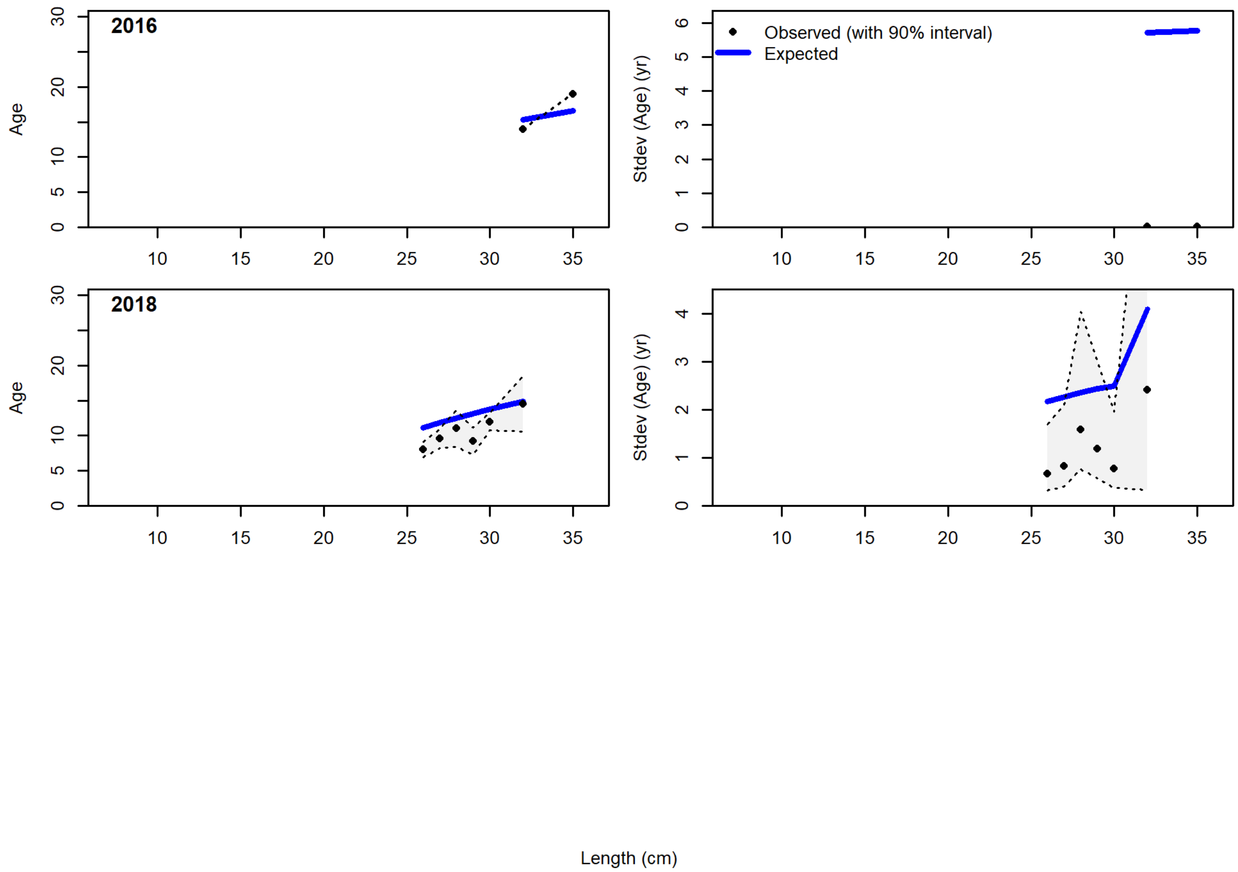


Figure 36: Estimated recruitment (red circles) and the assumed stock-recruit relationship (black line) for GBYR. The green line shows the effect of the bias correction for the lognormal distribution. fig:SR_curve2



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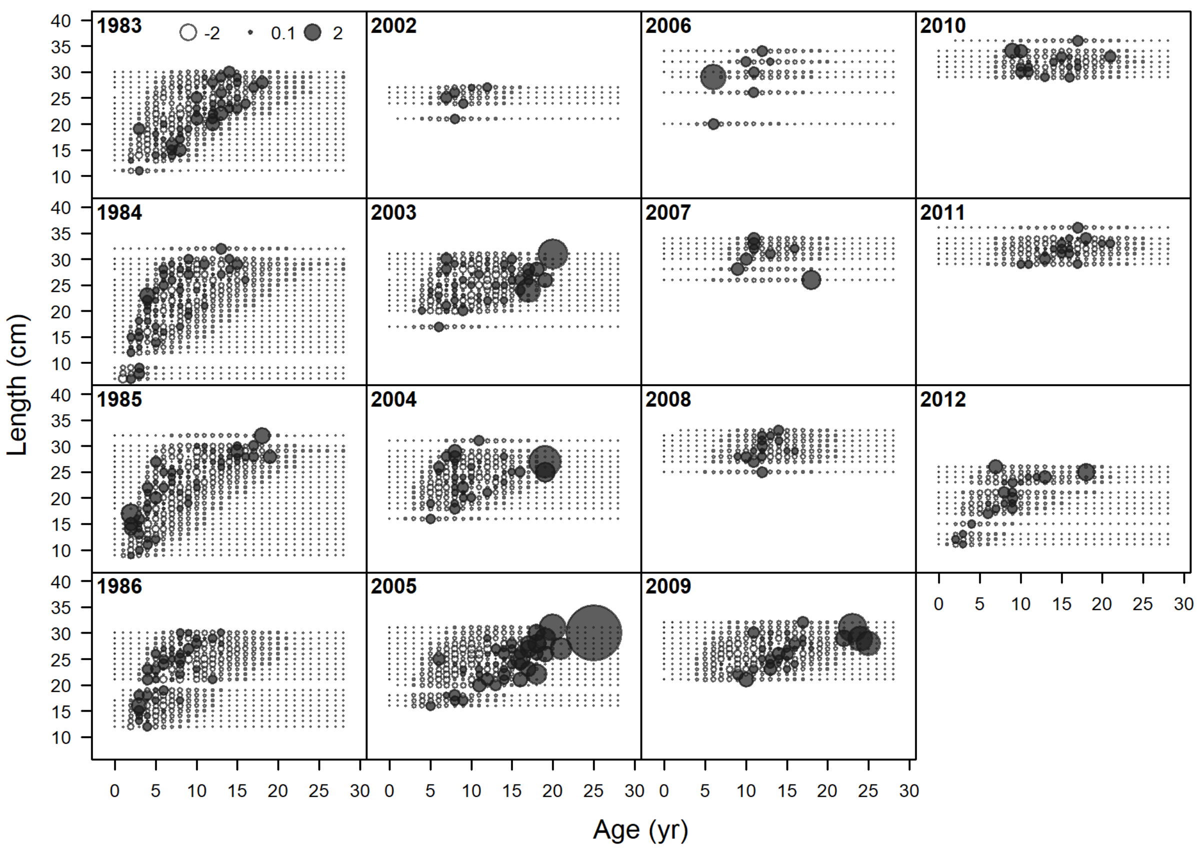


Figure 37: Pearson residuals, whole catch, Dummy1 (max=22.55) fig:mod1_4_comp_condAALf

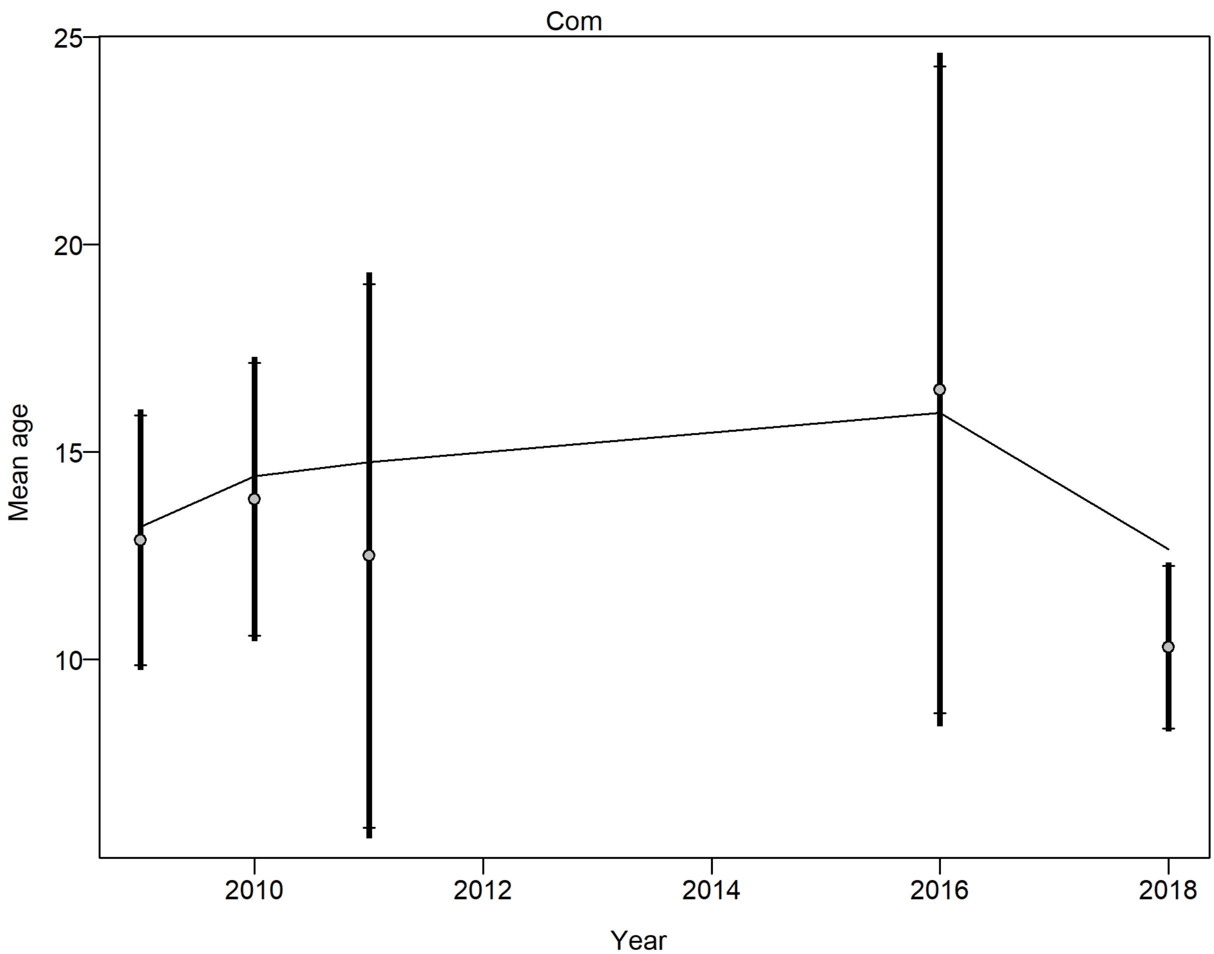


Figure 38: Mean age from conditional data (aggregated across length bins) for Com with 95% confidence intervals based on current samples sizes. Francis data weighting method TA1.8: thinner intervals (with capped ends) show result of further adjusting sample sizes based on suggested multiplier (with 95% interval) for conditional age_at_length data from Com: 1.0859 (0.6193_33.7322) For more info, see Francis, R.I.C.C. (2011). Data weighting in statistical fisheries stock assessment models. Can. J. Fish. Aquat. Sci. 68: 1124_1138. fig:mod1_5_com

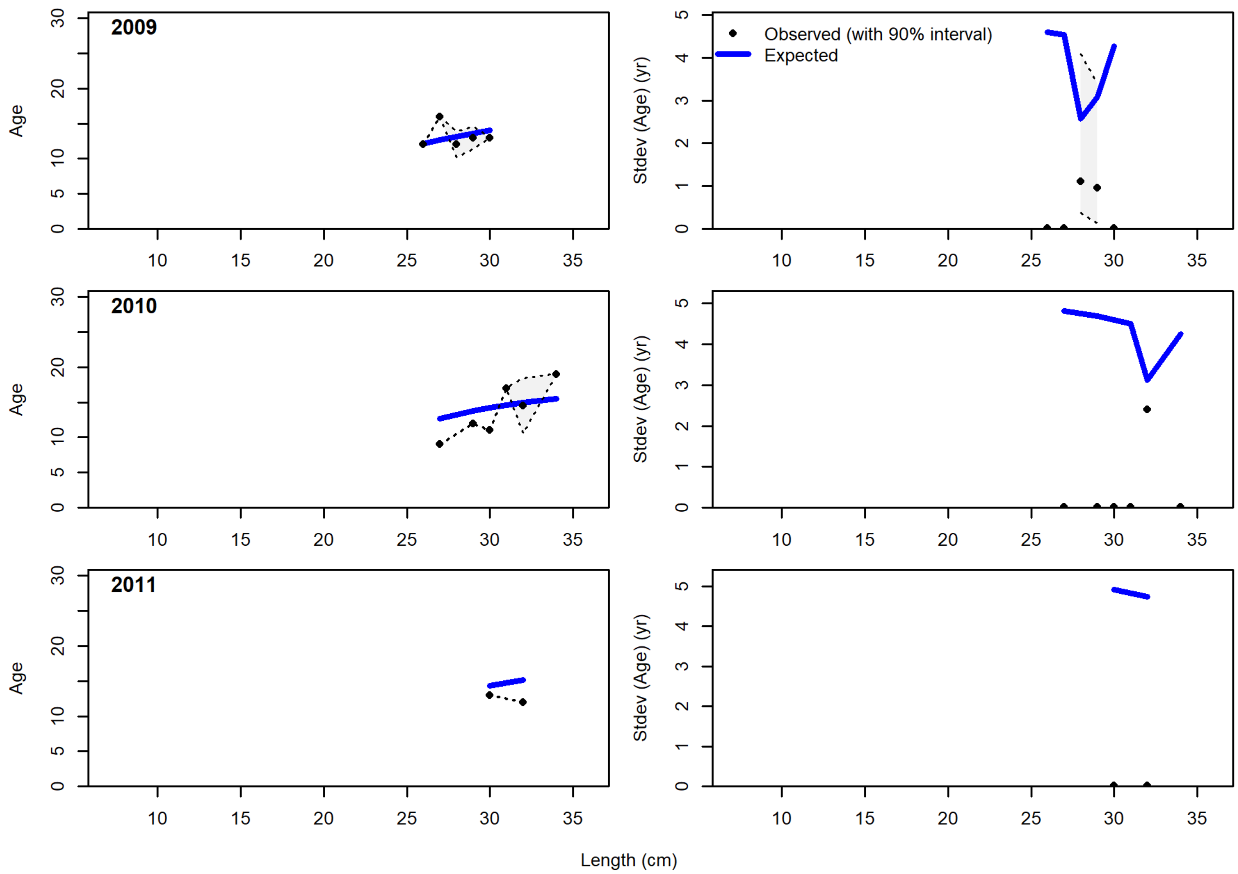


Figure 39: Conditional AAL plot, whole catch, Com (plot 1 of 2) These plots show mean age and std. dev. in conditional AAL. Left plots are mean AAL by size_class (obs. and pred.) with 90% CIs based on adding 1.64 SE of mean to the data. Right plots in each pair are SE of mean AAL (obs. and pred.) with 90% CIs based on the chi_square distribution. fig:mod1_6_comp_cond

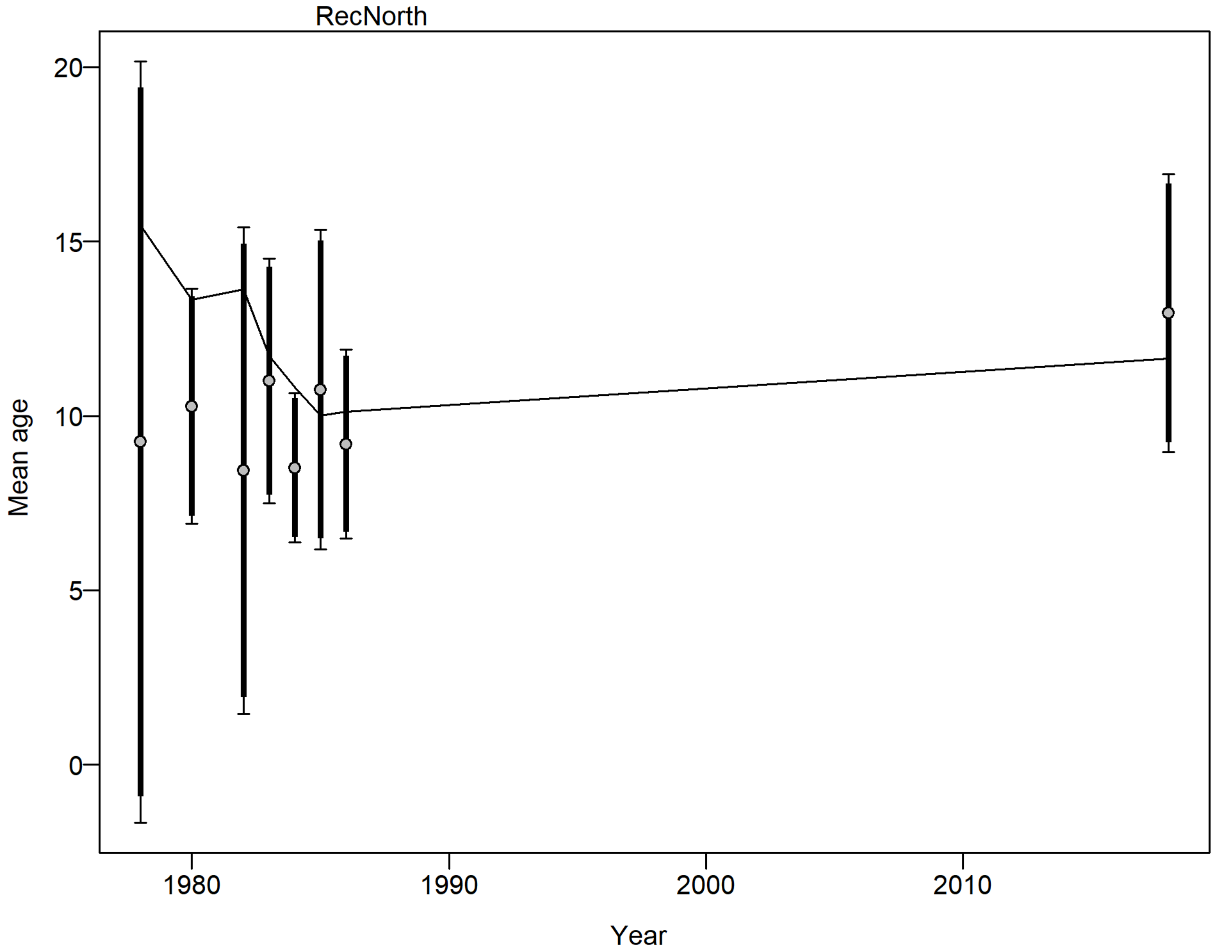


Figure 40: Mean age from conditional data (aggregated across length bins) for RecNorth with 95% confidence intervals based on current samples sizes. Francis data weighting method TA1.8: thinner intervals (with capped ends) show result of further adjusting sample sizes based on suggested multiplier (with 95% interval) for conditional age_at_length data from RecNorth: 0.8677 (0.589_3.1055) For more info, see Francis, R.I.C.C. (2011). Data weighting in statistical fisheries stock assessment models. Can. J. Fish. Aquat. Sci. 68: 1124_1138. fig:mod1_8_com

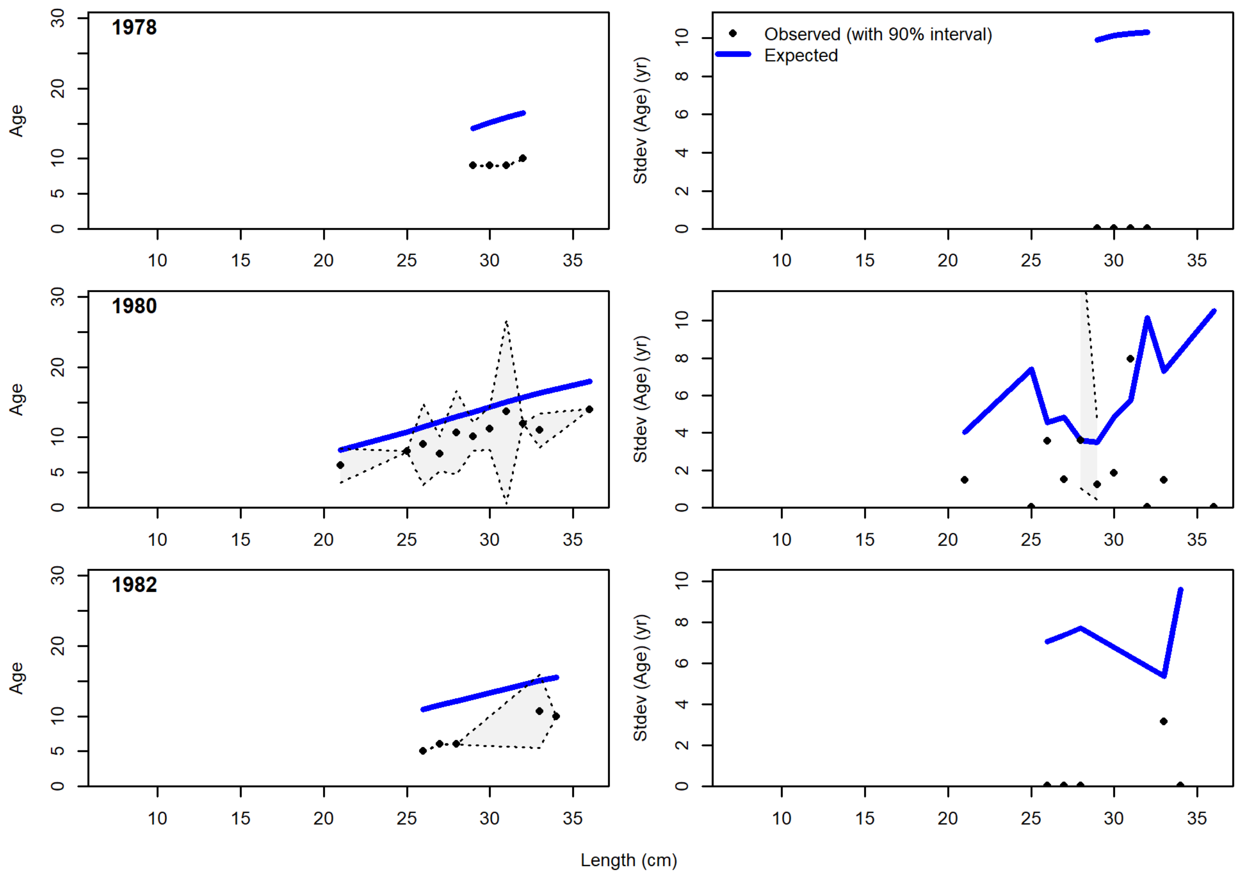
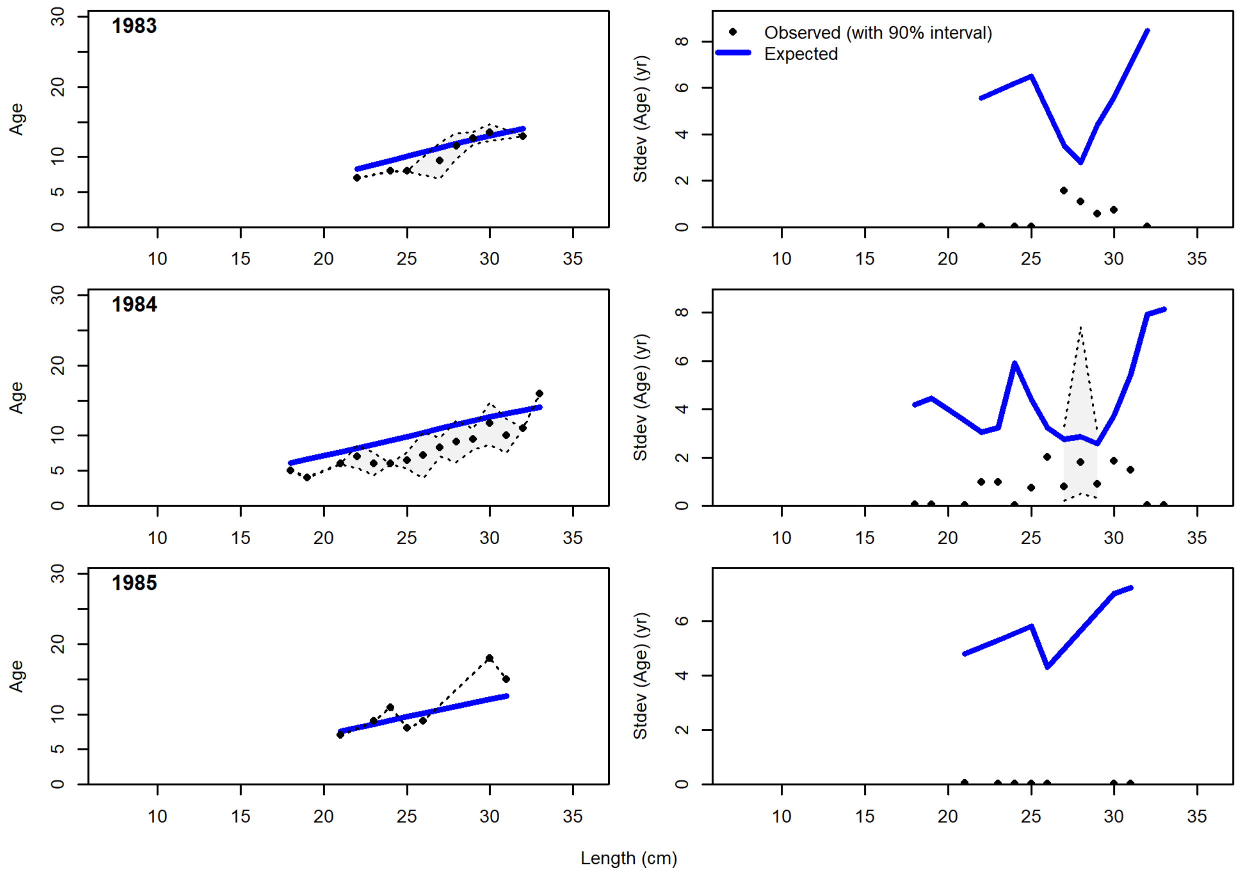


Figure 41: Conditional AAL plot, whole catch, RecNorth (plot 1 of 3) These plots show mean age and std. dev. in conditional AAL. Left plots are mean AAL by size_class (obs. and pred.) with 90% CIs based on adding 1.64 SE of mean to the data. Right plots in each pair are SE of mean AAL (obs. and pred.) with 90% CIs based on the chi_square distribution. fig:mod1_9_comp_condAALfit_Andre_plotsflt3mkt0_page1



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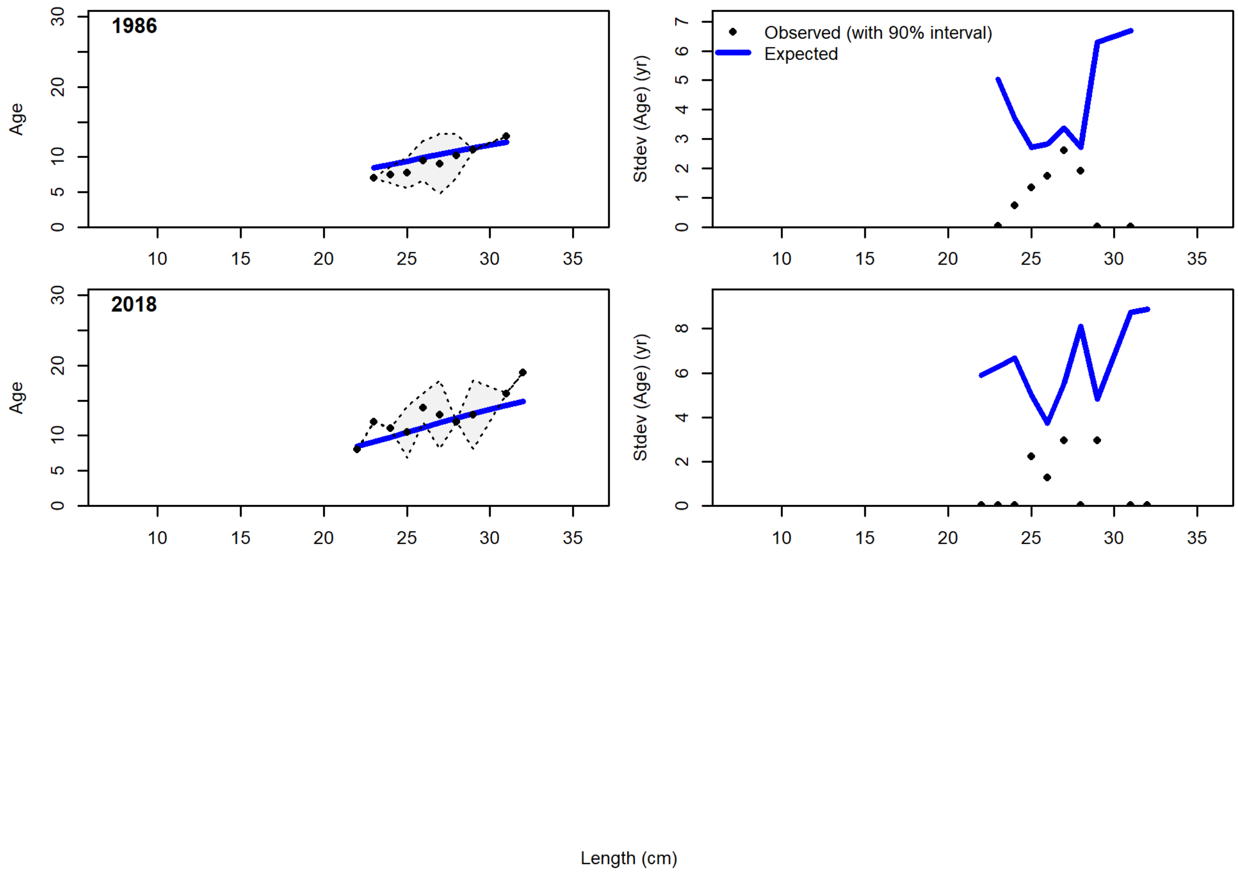

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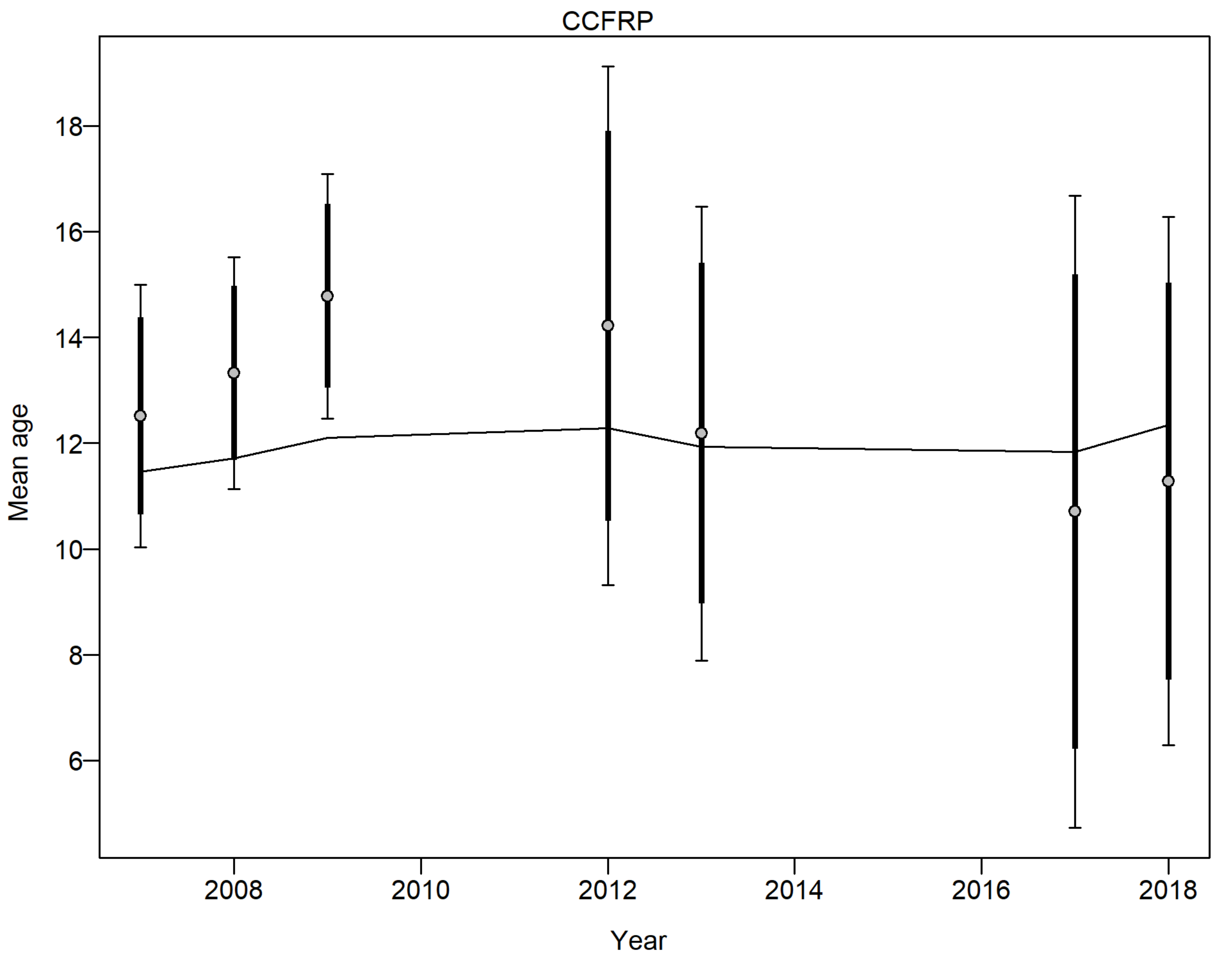


Figure 42: Mean age from conditional data (aggregated across length bins) for CCFRP with 95% confidence intervals based on current samples sizes. Francis data weighting method TA1.8: thinner intervals (with capped ends) show result of further adjusting sample sizes based on suggested multiplier (with 95% interval) for conditional age_at_length data from CCFRP: 0.5645 (0.344_2.4444) For more info, see Francis, R.I.C.C. (2011). Data weighting in statistical fisheries stock assessment models. Can. J. Fish. Aquat. Sci. 68: 1124_1138. fig:mod1_12_co

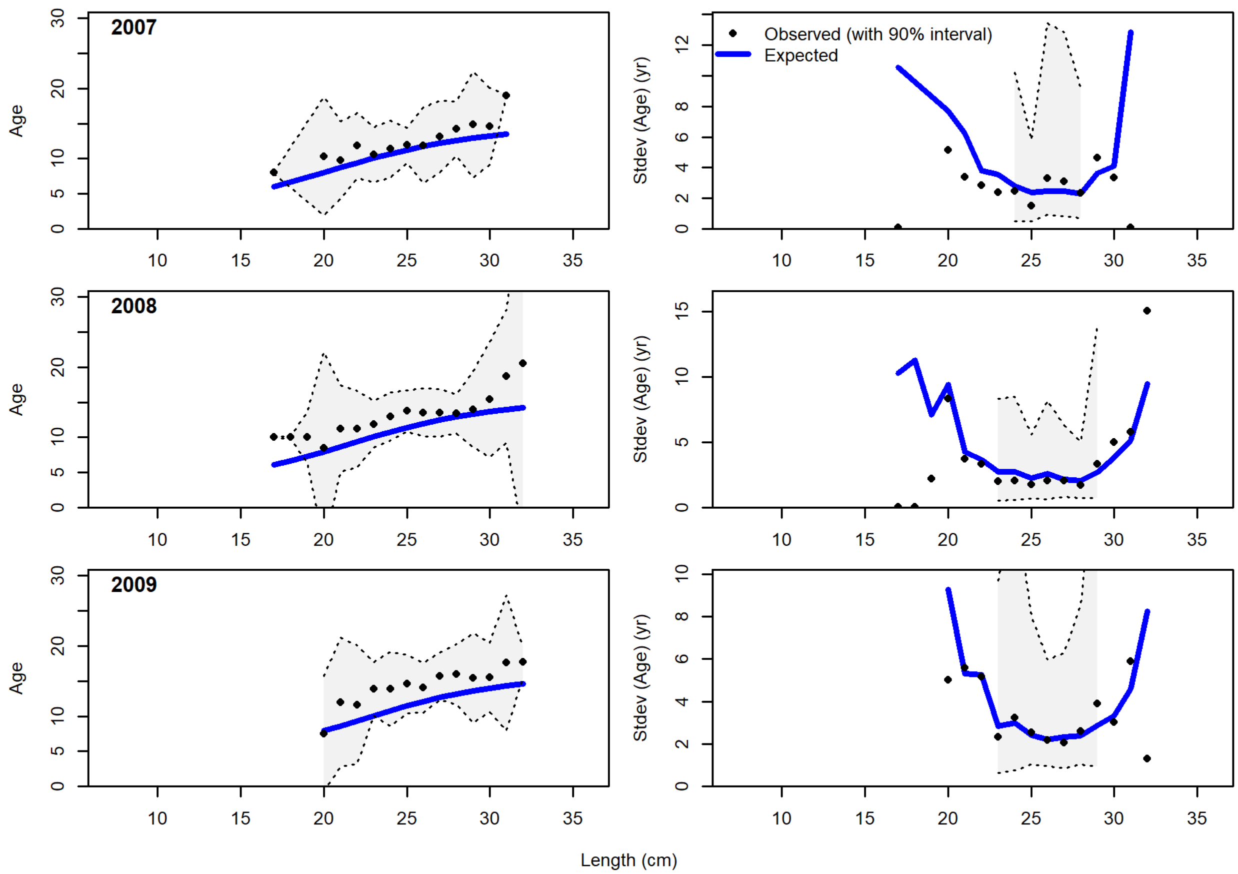
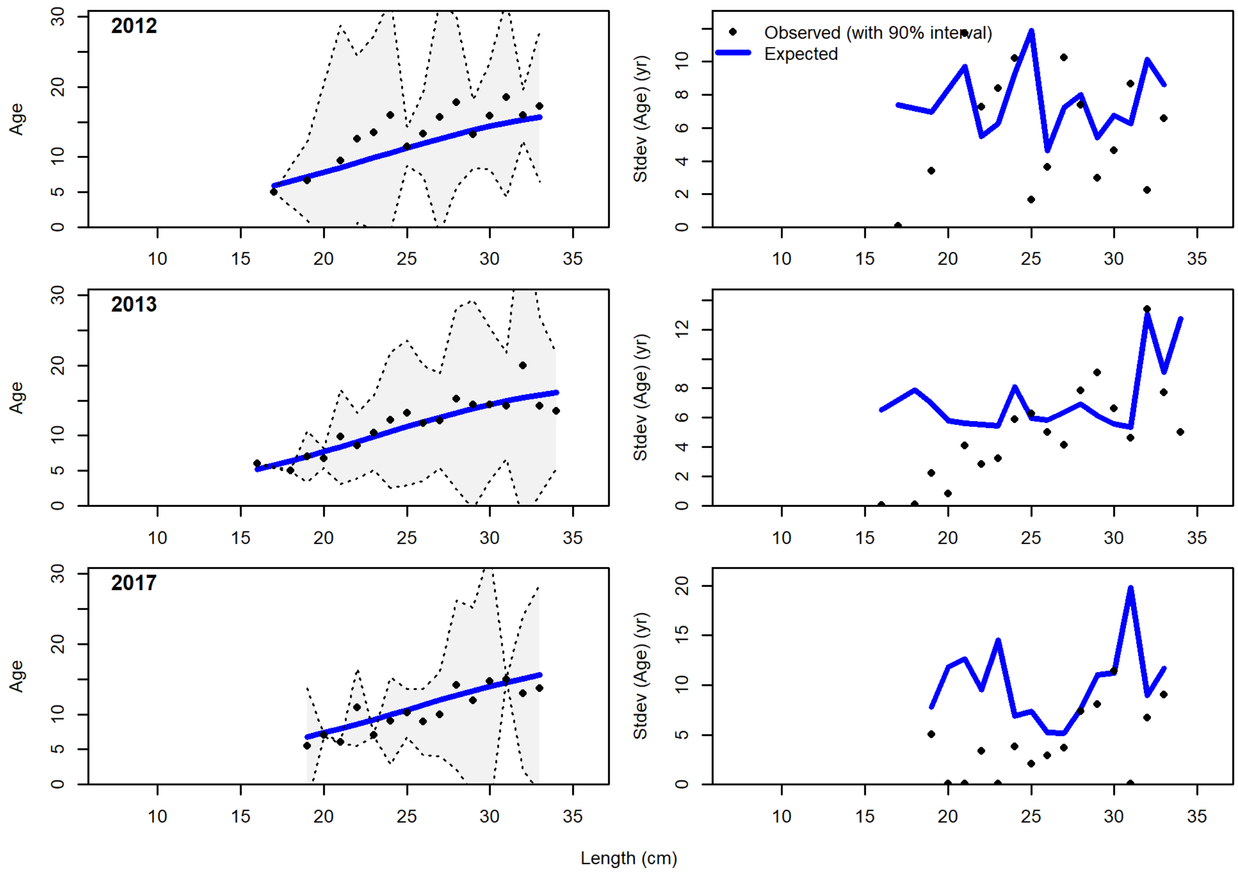


Figure 43: Conditional AAL plot, whole catch, CCFRP (plot 1 of 3) These plots show mean age and std. dev. in conditional AAL. Left plots are mean AAL by size_class (obs. and pred.) with 90% CIs based on adding 1.64 SE of mean to the data. Right plots in each pair are SE of mean AAL (obs. and pred.) with 90% CIs based on the chi_square distribution. fig:mod1_13_c



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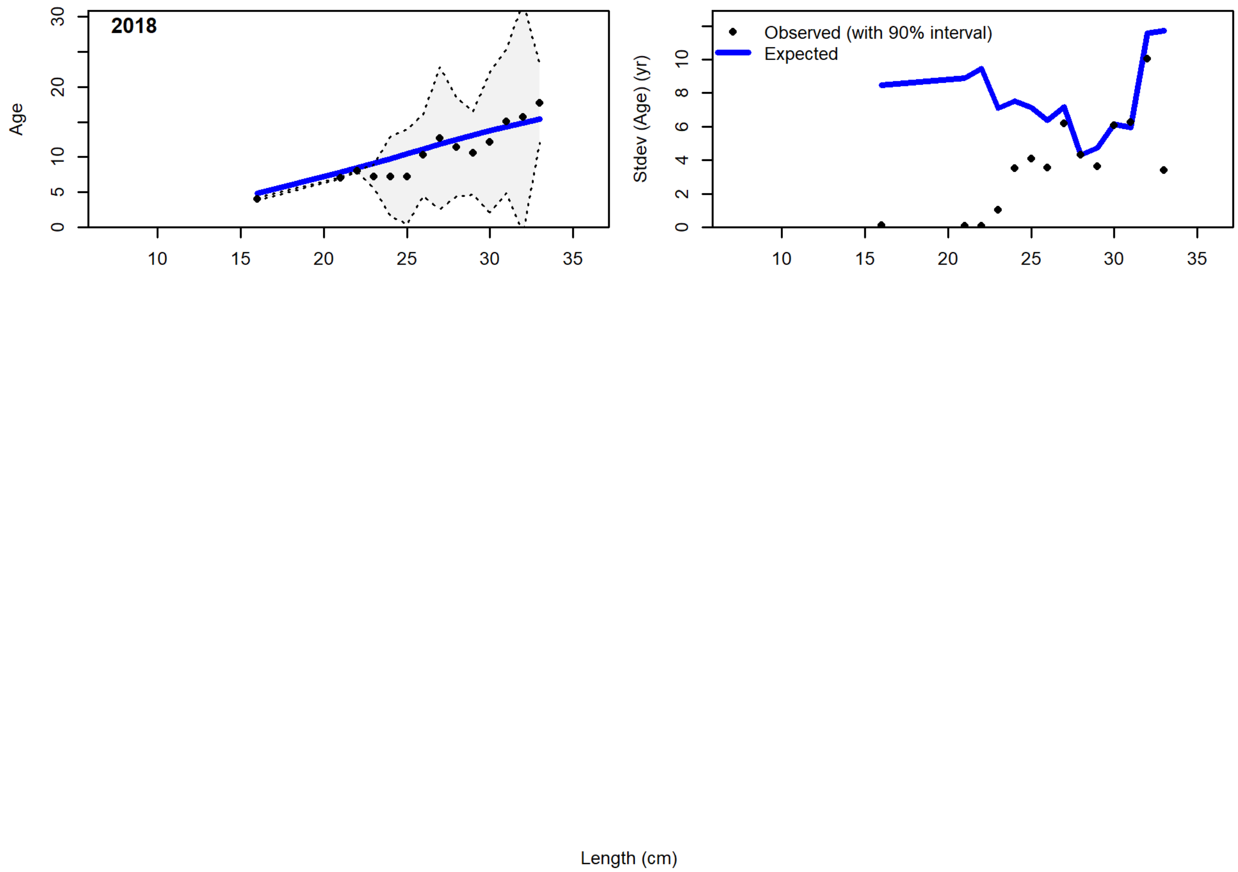

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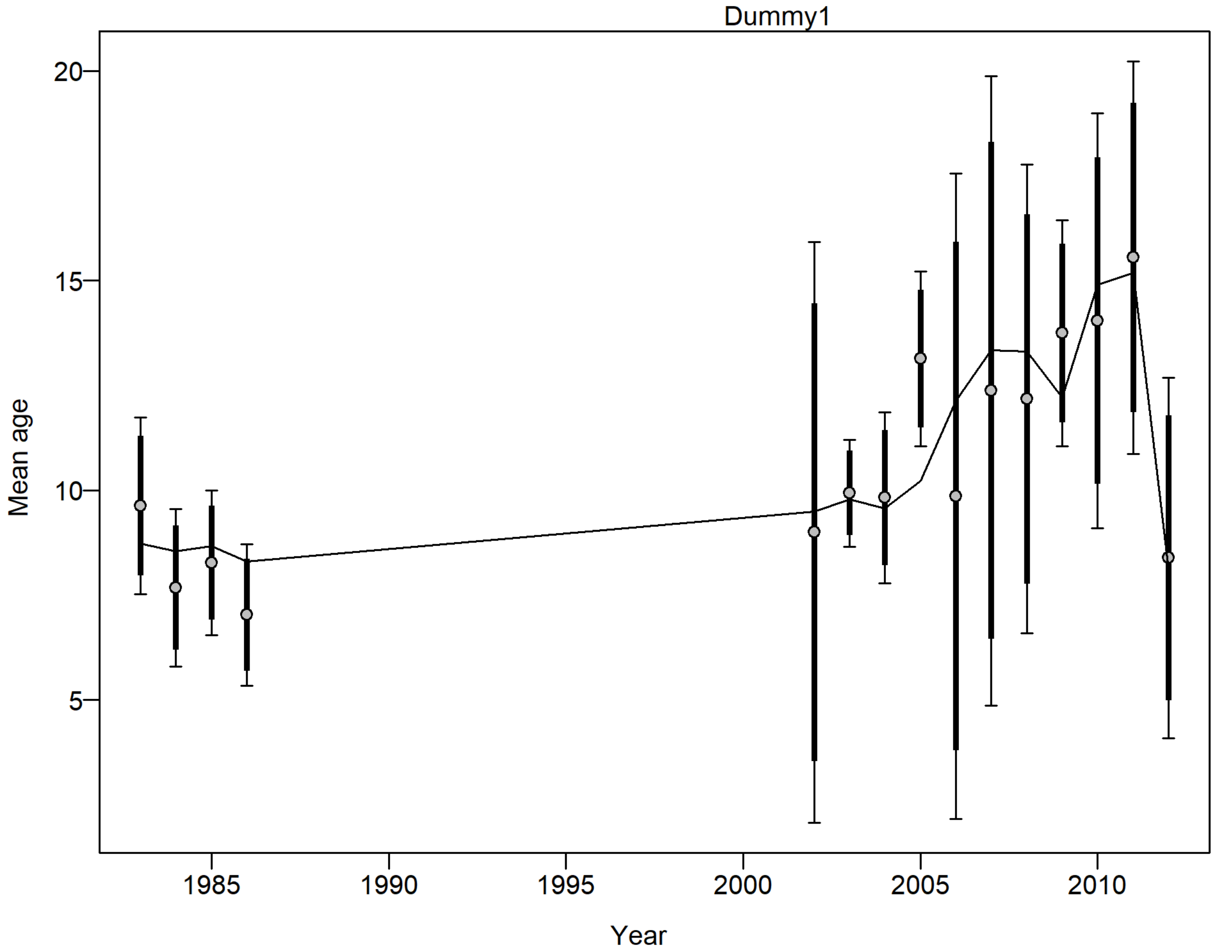


Figure 44: Mean age from conditional data (aggregated across length bins) for Dummy1 with 95% confidence intervals based on current samples sizes. Francis data weighting method TA1.8: thinner intervals (with capped ends) show result of further adjusting sample sizes based on suggested multiplier (with 95% interval) for conditional age_at_length data from Dummy1: 0.6218 (0.324_3.077) For more info, see Francis, R.I.C.C. (2011). Data weighting in statistical fisheries stock assessment models. Can. J. Fish. Aquat. Sci. 68: 1124_1138. fig:mod1_16_co

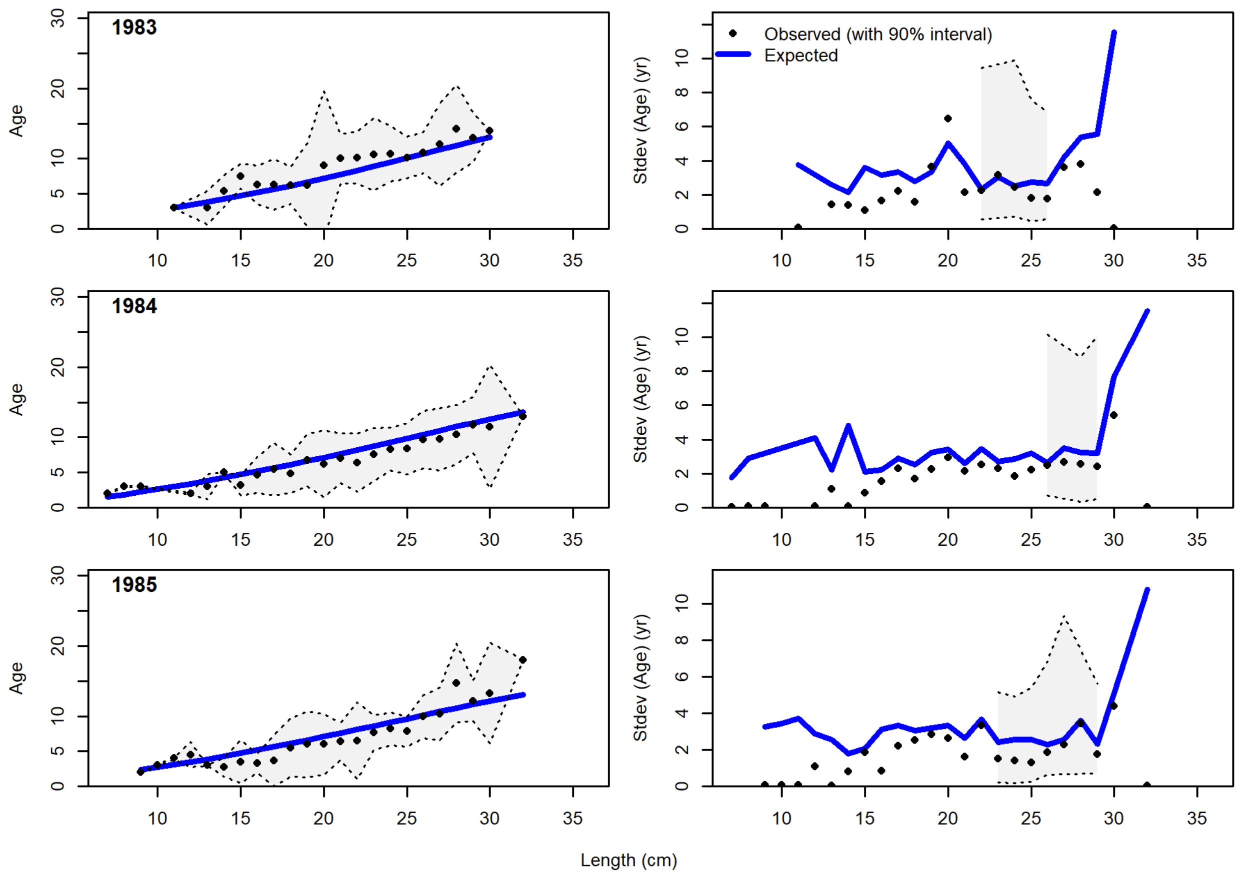
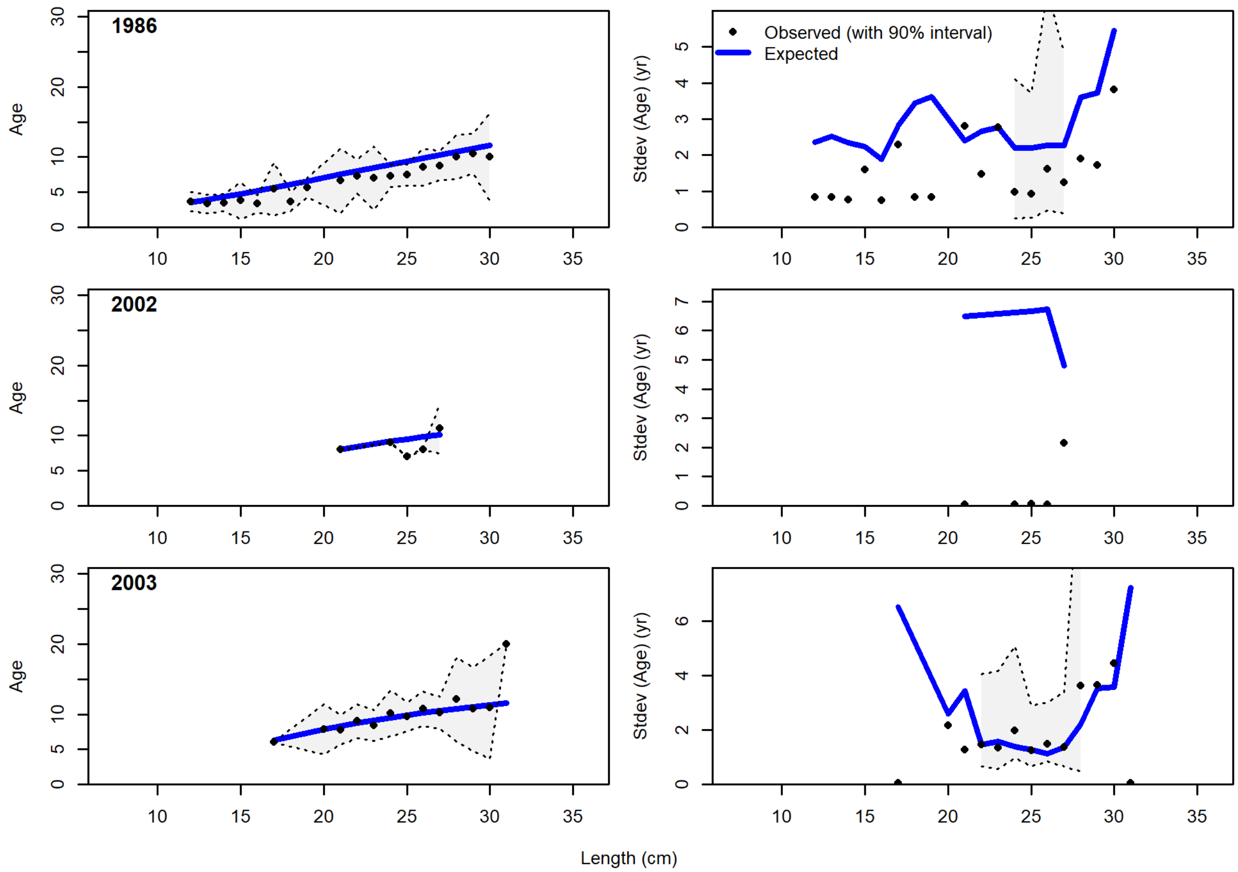


Figure 45: Conditional AAL plot, whole catch, Dummy1 (plot 1 of 5) These plots show mean age and std. dev. in conditional AAL. Left plots are mean AAL by size_class (obs. and pred.) with 90% CIs based on adding 1.64 SE of mean to the data. Right plots in each pair are SE of mean AAL (obs. and pred.) with 90% CIs based on the chi_square distribution. fig:mod1_17_c



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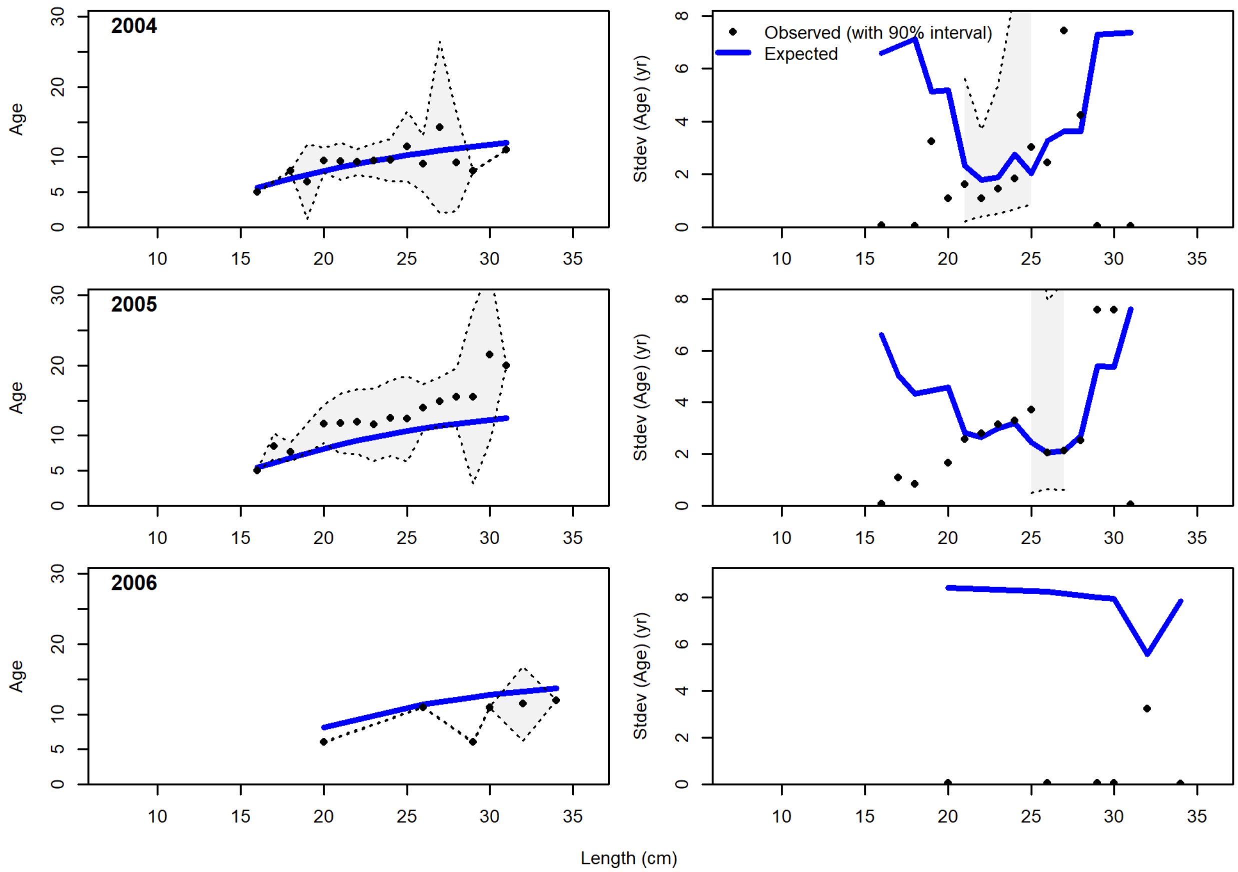

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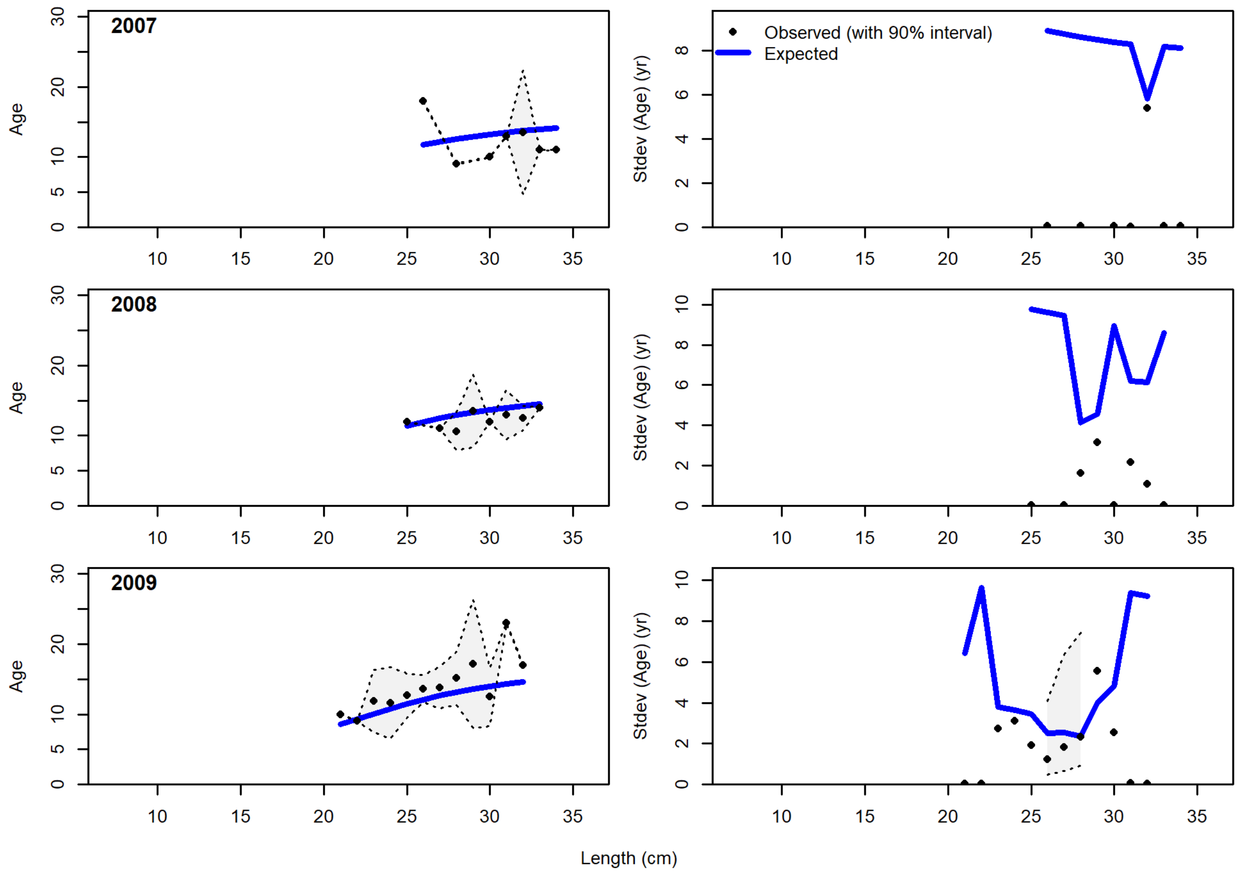

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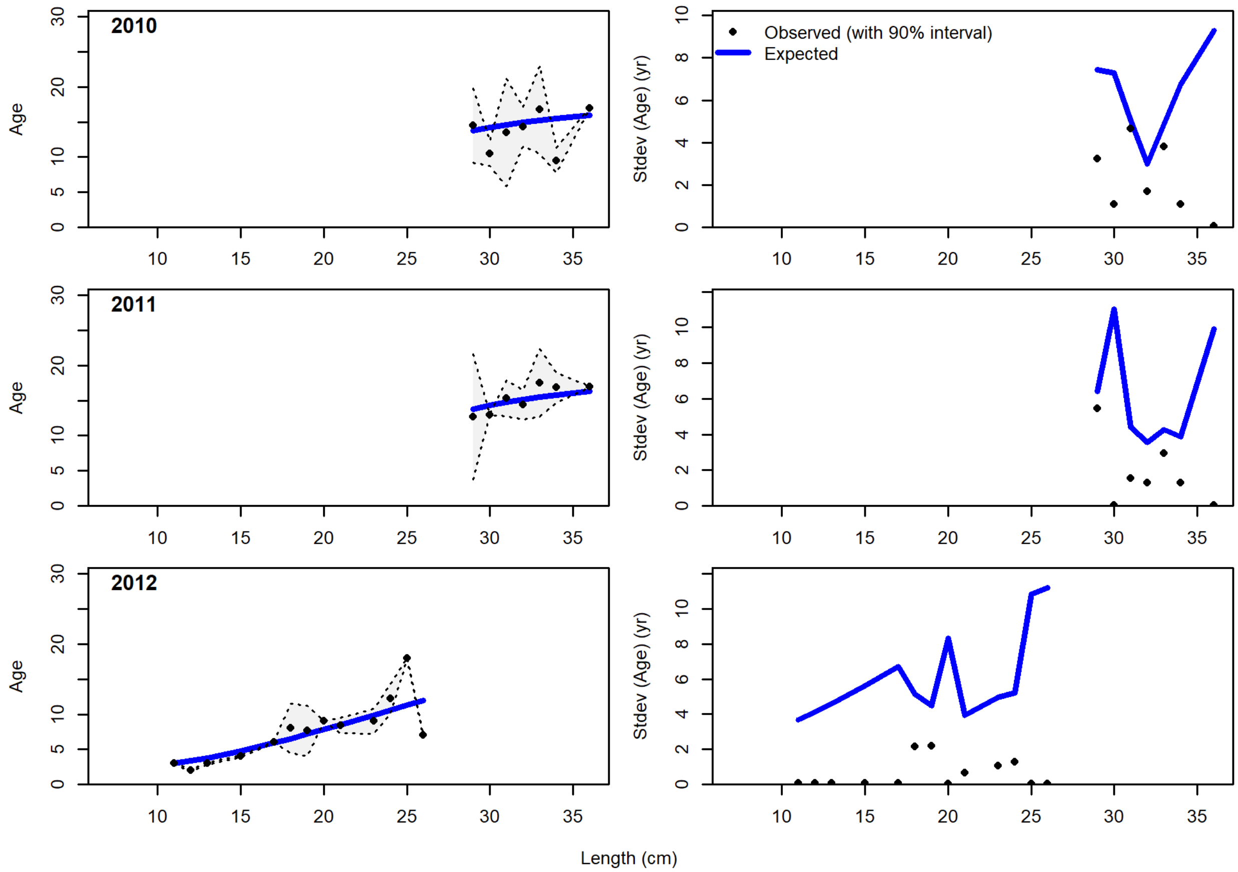

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¡!– ***********h Likelihood profile FIGURES****************************** –¿

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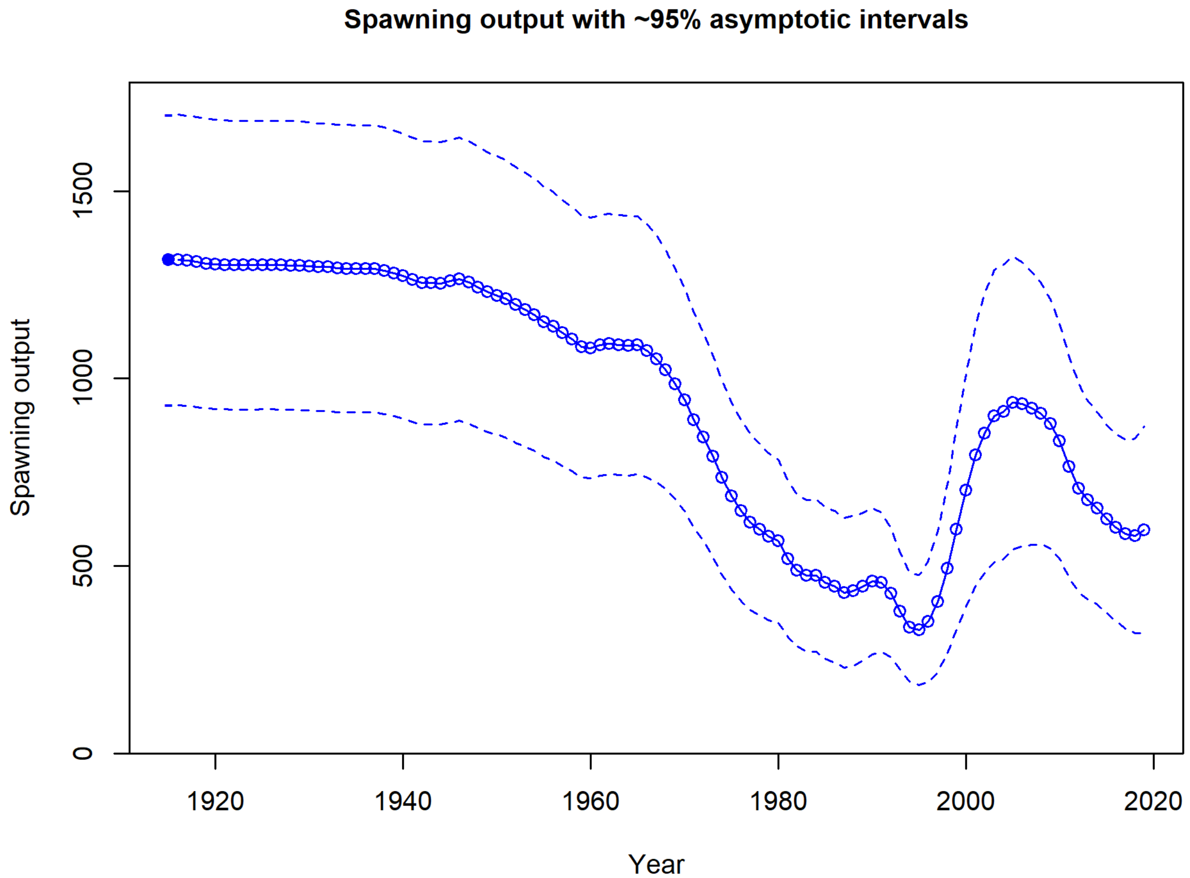

Figure 46: Estimated spawning biomass (mt) with approximate 95% asymptotic intervals. fig:ts7_Spawn

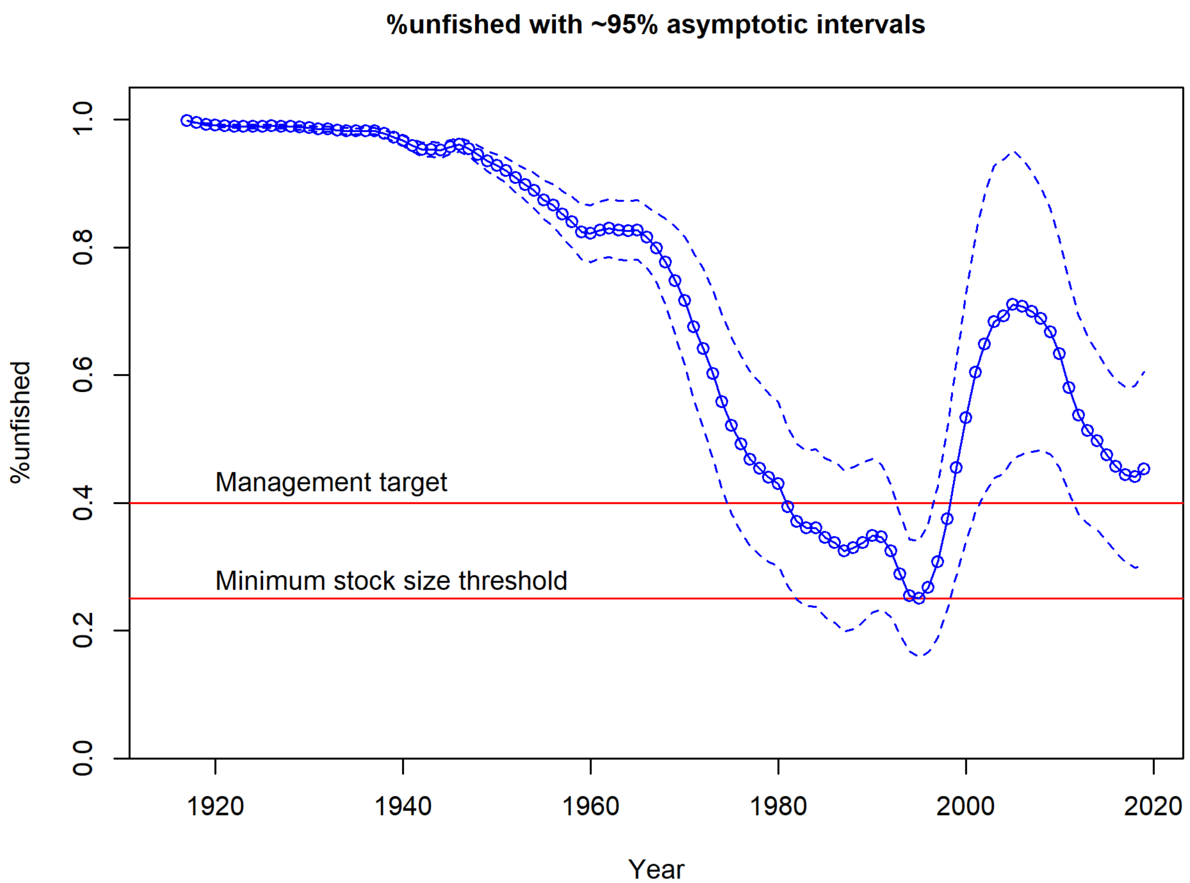

Figure 47: Estimated spawning depletion with approximate 95% asymptotic intervals. fig:ts9_unfishe

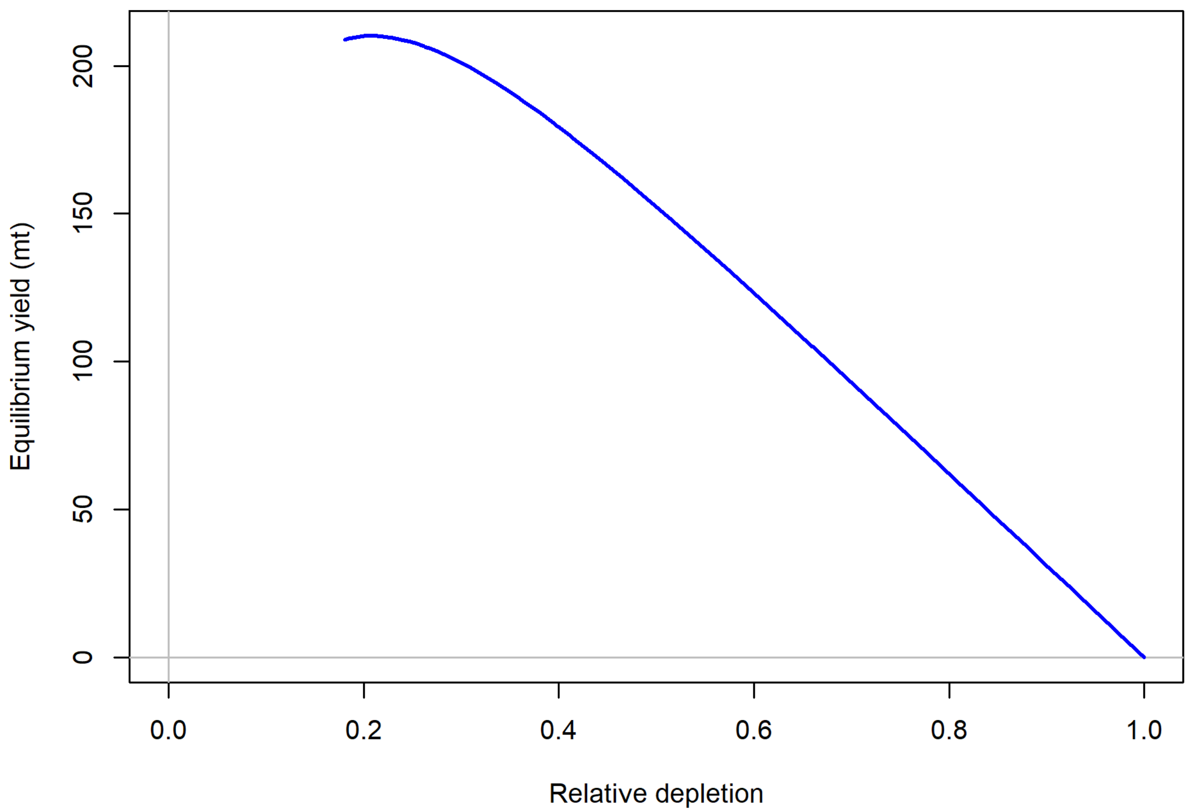

Figure 48: Equilibrium yield curve for the base case model. Values are based on the 2018 fishery selectivity and with steepness fixed at 0.718. fig:yield1_yield_curve

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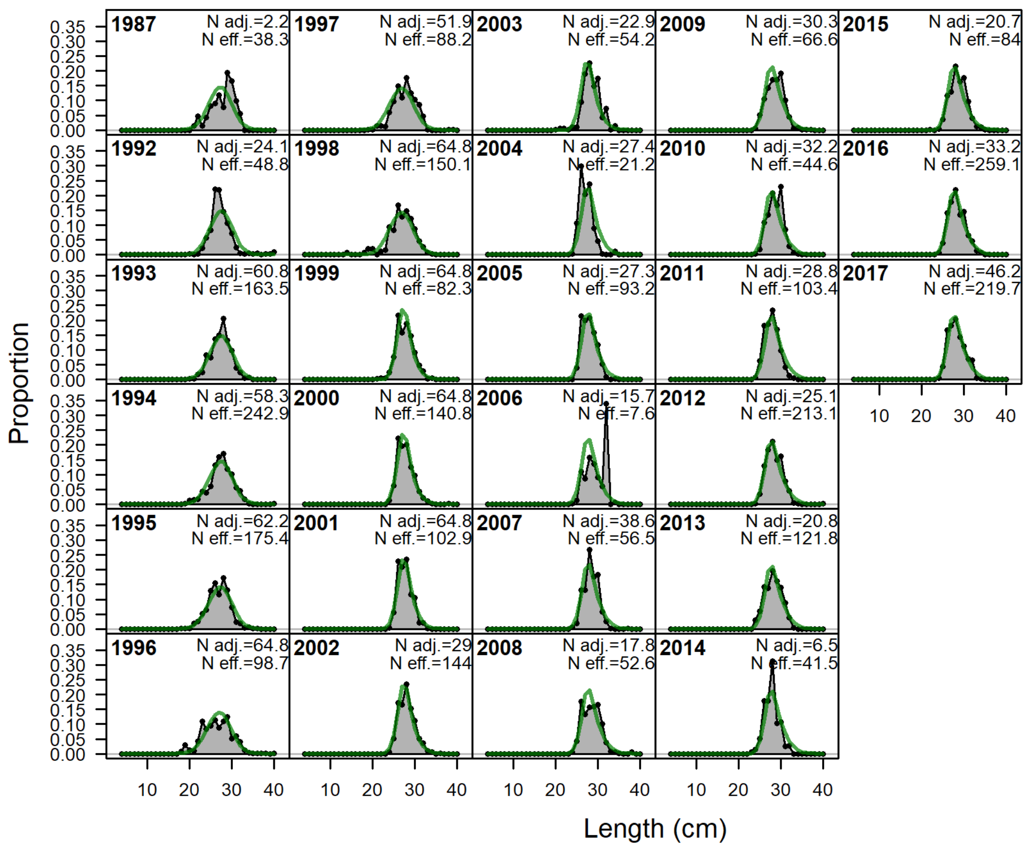

Figure A49: Length comps, retained, Com. 'N adj.' is the input sample size after data_weighting adjustment. N eff. is the calculated effective sample size used in the McAllister_Iannelli tuning method. fig:mod1_1_comp_lenfit_flt1mkt2

Appendix A. Detailed fits to length composition data

appendix-a.-detailed-fits-to-length-composition-data

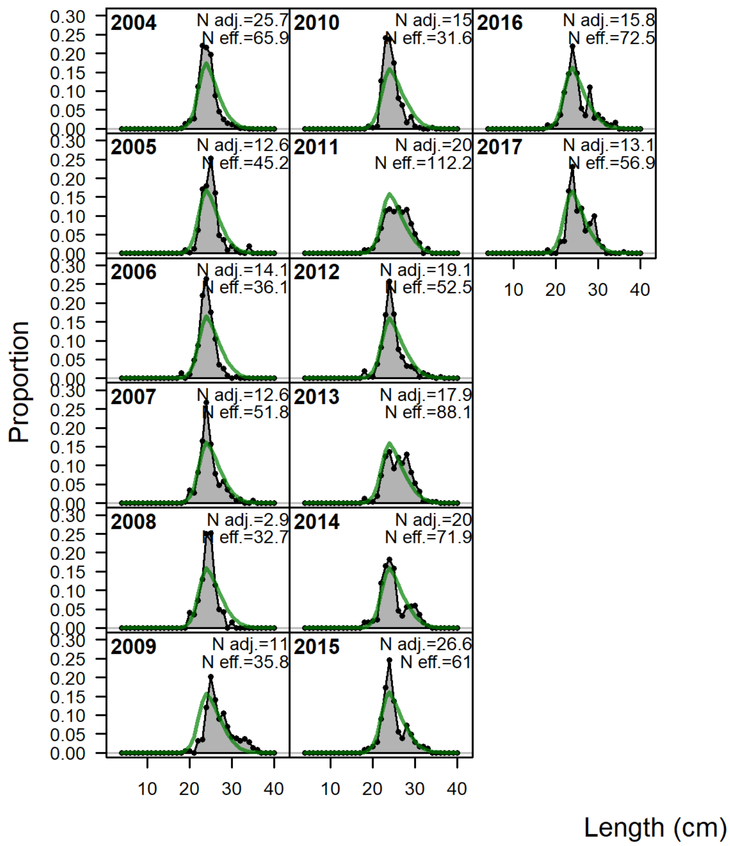

Figure A50: Length comps, retained, ComDisc. 'N adj.' is the input sample size after data_weighting adjustment. N eff. is the calculated effective sample size used in the McAllister_Iannelli tuning method. fig:mod1_2_comp_lenfit_flt2mkt2

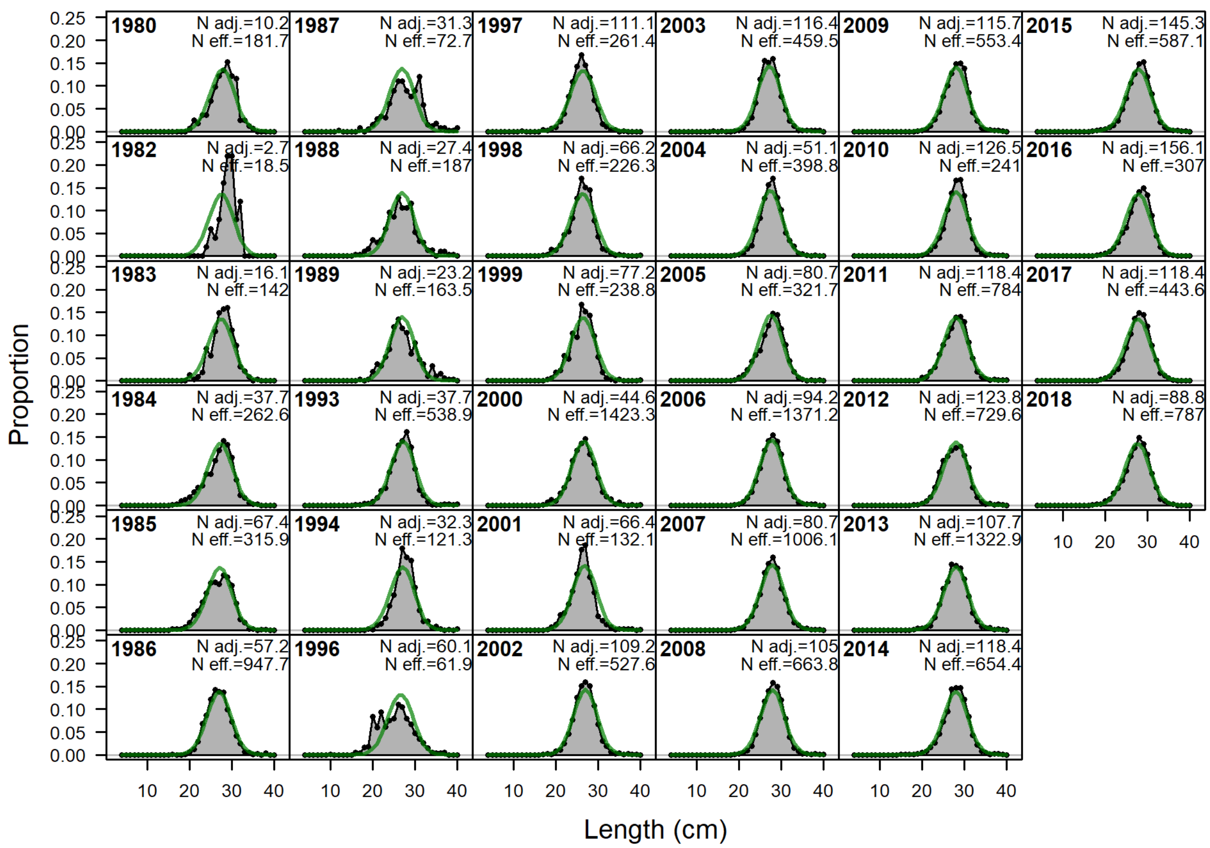

Figure A51: Length comps, whole catch, RecNorth. 'N adj.' is the input sample size after data_weighting adjustment. N eff. is the calculated effective sample size used in the McAllister_Iannelli tuning method. fig:mod1_3_comp_lenfit_flt3mkt0

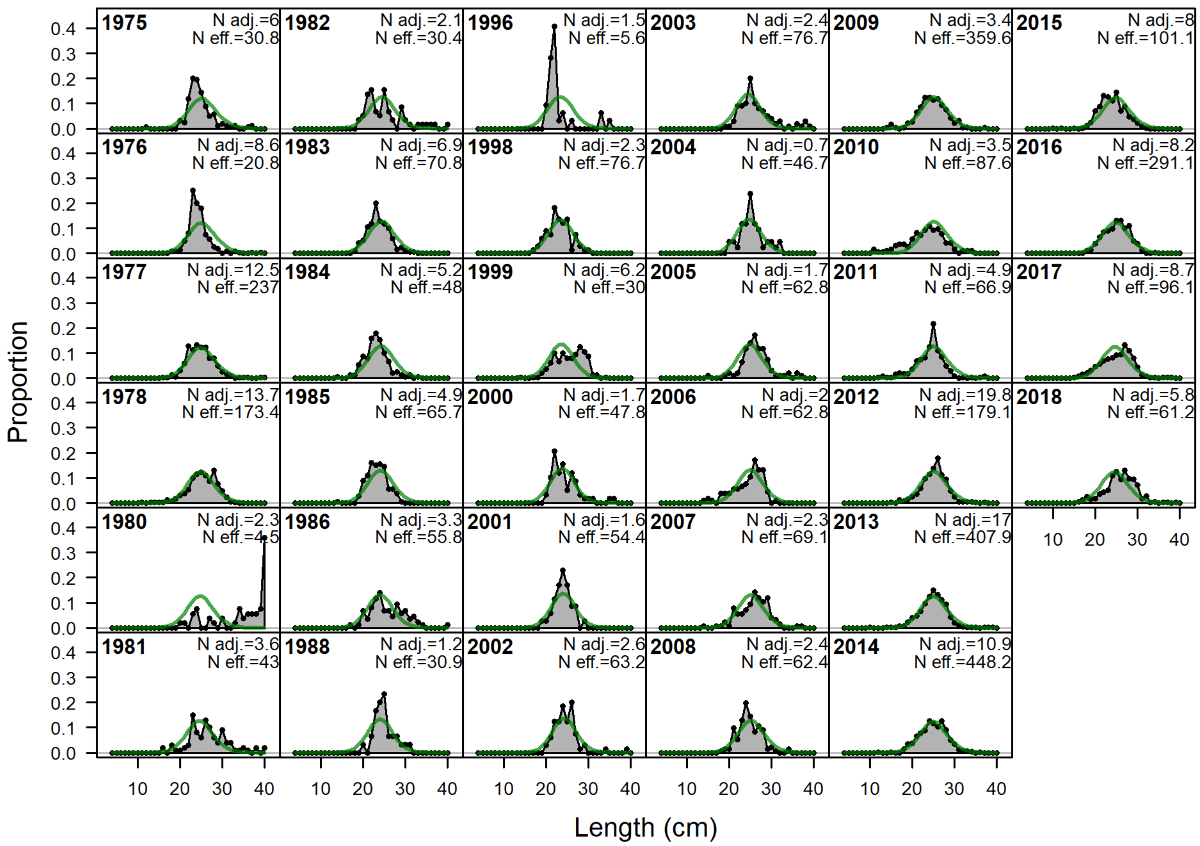

Figure A52: Length comps, whole catch, RecSouth. 'N adj.' is the input sample size after data_weighting adjustment. N eff. is the calculated effective sample size used in the McAllister_Iannelli tuning method. fig:mod1_4_comp_lenfit_flt4mkt0

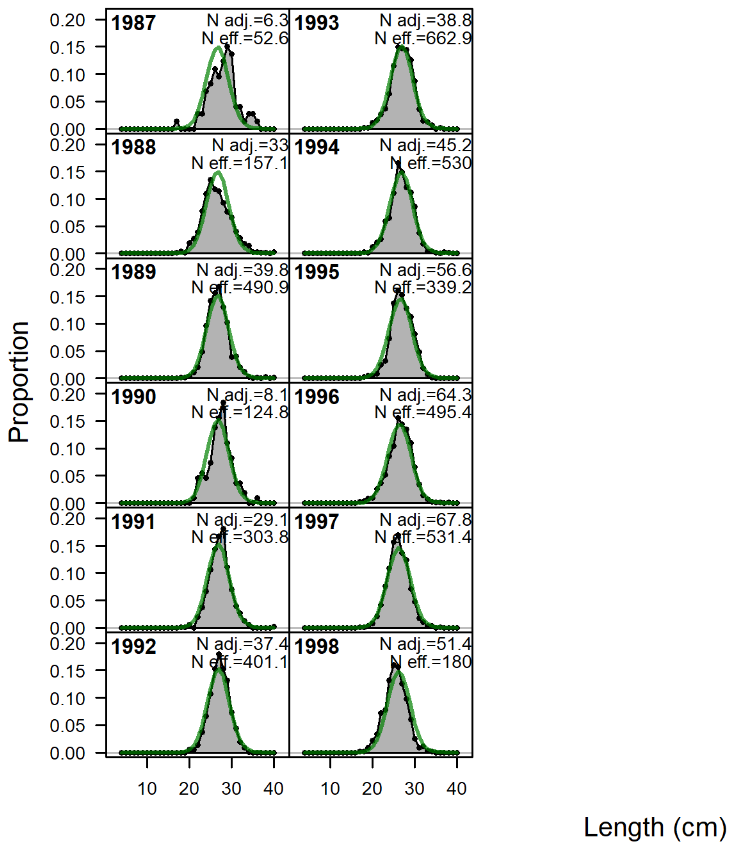

Figure A53: Length comps, whole catch, DebCPFV. 'N adj.' is the input sample size after data_weighting adjustment. N eff. is the calculated effective sample size used in the McAllister_Iannelli tuning method. fig:mod1_5_comp_lenfit_flt5mkt0

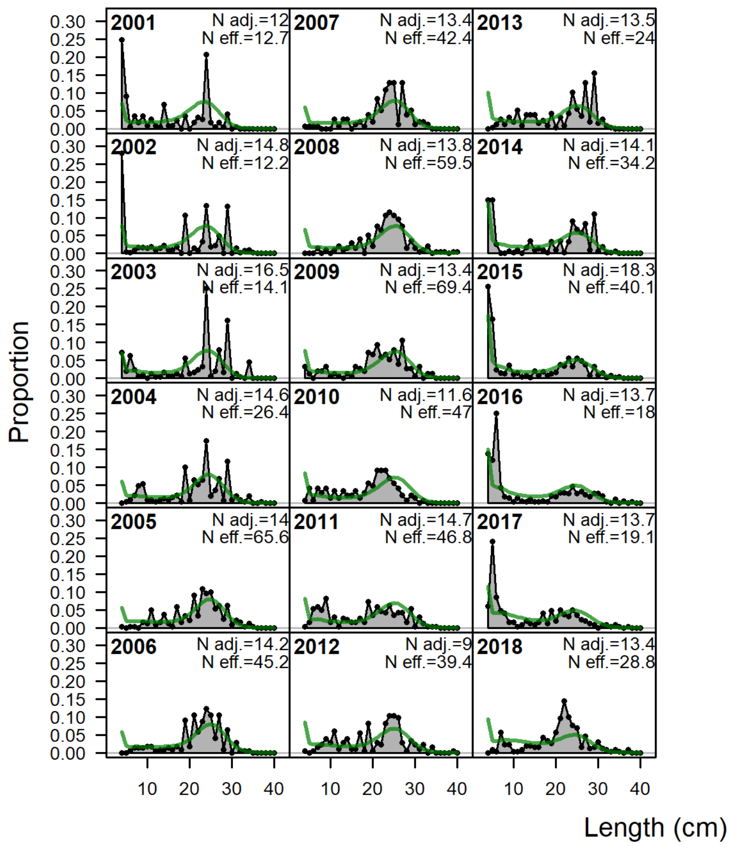

Figure A54: Length comps, whole catch, PISCO. 'N adj.' is the input sample size after data_weighting adjustment. N eff. is the calculated effective sample size used in the McAllister_Iannelli tuning method. fig:mod1_6_comp_lenfit_flt8mkt0

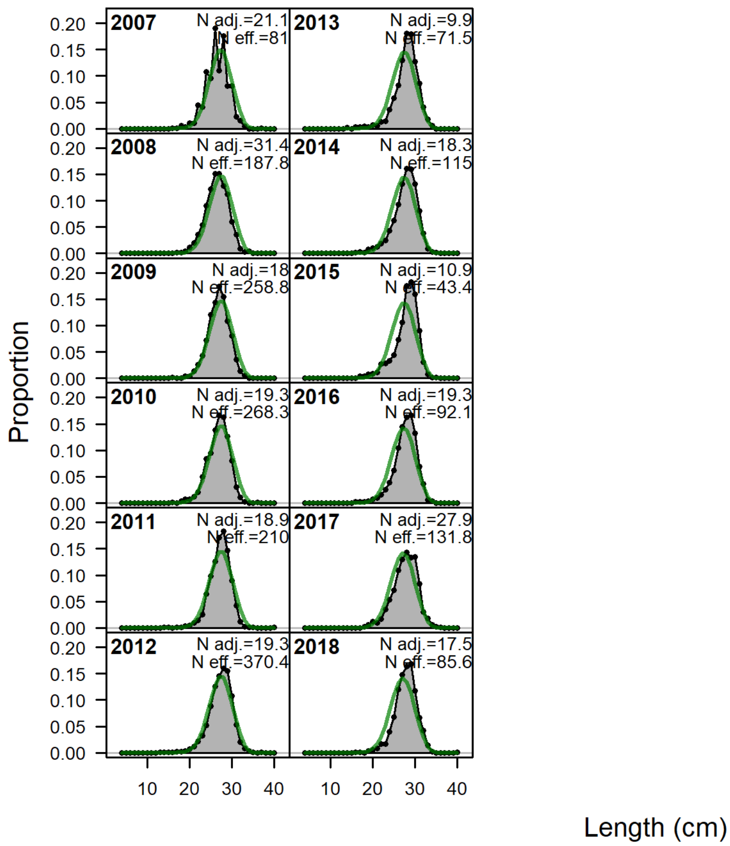

Figure A55: Length comps, whole catch, CCFRP. 'N adj.' is the input sample size after data_weighting adjustment. N eff. is the calculated effective sample size used in the McAllister_Iannelli tuning method. fig:mod1_7_comp_lenfit_flt9mkt0

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