

Mobility and flexibility enable resilience of human harvesters to environmental perturbation

Abstract

Characteristics of natural resources that enable sustainable management are often more fully understood than the adaptive behaviors of human harvesters in those same systems. Given increasing environmental variability due to climate change, it is especially critical to understand how human harvesters may respond to environmental perturbation. In this study, we identify characteristics that promoted resilience of one the most valuable fisheries on the west coast of the United States to a record marine heatwave. Using movement telemetry linked to fishery landings records from more than 500 fishing vessels, encompassing 2.2 million geolocations and more than \$2 billion in revenue, we found that vessels employed two, non-mutually exclusive strategies to cope with the anomalous environmental and management conditions imposed by the heatwave: increasing spatial mobility and diversifying fishery participation. The combination of these strategies appeared to be the most adaptive, as it produced the greatest increase in profits. In contrast, participants that specialized in a single fishery and concentrated fishing effort in small spatial areas experienced the greatest losses driven by the heatwave. Our data-driven approach reveals behaviors that can be promoted to improve the adaptive capacity of human harvesters in an era of unprecedented environmental perturbation.

1. Introduction

Sustainability in social-ecological systems—the continued provision of human and ecological benefits from healthy ecosystems (Leslie et al., 2015)—requires resilience to environmental perturbations. Often, though, people respond to environmental change in diverse and complex ways. Just as multiple species occupying similar ecological niches may react differently to physical changes in their environments (Elmqvist et al., 2003), human actors in a social-ecological system can exhibit diverse behaviors within the constraints imposed by the governance system (McGinnis and Ostrom, 2014). Groups of resource users with distinct livelihood portfolios, available capital, or spatial patterns of resource extraction will not respond the same way to environmental or management changes (Young et al., 2019). In response to change, some users might stick to established knowledge and reliable spatial patterns of exploitation, while others might employ more exploratory strategies that carry higher potential upsides but also higher risks and costs (Cohen et al., 2007). Understanding the adaptive behaviors of resource users is all the more important given the

20 increasing prevalence of extreme climate events attributable to climate change
21 (Abatzoglou et al., 2019; Cook et al., 2018; Oliver et al., 2018; Townhill et al.,
22 2018), but empirical evidence making the link between climate extremes and
23 contemporaneous human adaptation remains lacking.

24 Fisheries are a prominent example of a social-ecological system where complex
25 links between resource user (harvester) behavior and natural resource dynamics
26 drive sustainability (Branch et al., 2006). Fisheries represent the last large-scale
27 wild harvest of food on Earth, but also one of the most traditional livelihoods in
28 human history. Difficulties in achieving sustainability in fisheries have often been
29 linked to an inadequate understanding of harvester dynamics (Fulton et al., 2011;
30 Hilborn, 1985). Differences in fisher behaviors, both within and across fisheries,
31 can affect the stability and sustainability of fish populations (Fryxell et al., 2017;
32 Salas and Gaertner, 2004), of other species—for instance, endangered marine
33 mammals or seabirds—and of the fishery itself (Gladics et al., 2017; Hamilton
34 and Baker, 2019).

35 Additionally, different behavioral segments of fishing fleets may respond in
36 different ways to management measures, or may be differentially vulnerable to
37 environmental perturbations (Salas and Gaertner, 2004). For example, O’Farrell,
38 Sanchirico, et al. (2019) found that more exploratory fishing vessels—those that,
39 on average, traveled further and more often traversed new fishing grounds—were
40 better able to cope with an extended spatial closure. Heterogeneous behavioral
41 responses of fishers can be difficult to study, despite their potential impact on
42 resource dynamics. Partly, this is due to a lack of detailed spatial and economic
43 information on harvester behavior. However, recent years have seen a rise
44 in availability of these types of fishery data, paired with methods to extract
45 behavioral insights from them (Joo et al., 2015; Mendo et al., 2019; Watson and
46 Haynie, 2016). In the following, we apply a range of data-driven methods to
47 ask: how did human harvesters cope with and adapt to a major environmental
48 perturbation in the most valuable fishery on the U.S. west coast?

49 The Dungeness crab fishery on the west coast of the United States often
50 obtains in excess of \$200 million in revenue from over 1,000 participating vessels
51 each year (Rasmuson, 2013; Richerson et al., 2020). It is a fishery that is
52 central both ecologically and economically (Fuller et al., 2017) to the west coast
53 social-ecological system, making it at once a cornerstone of fishers’ portfolios
54 and a source of complexity in fisheries governance (Holland and Leonard, 2020;
55 Holland et al., 2017). Dungeness crab populations appear able to withstand
56 immense fishing pressure, and although crab abundance can fluctuate markedly
57 from year to year, long term abundance has been relatively stable for more than
58 a half century (Richerson et al., 2020). Harvester characteristics vary widely for
59 an industrialized fishery—Dungeness crab vessels have a large range of sizes (in
60 our data, 21 to 103 feet), and operate out of both large urban and small rural
61 fishing ports across the U.S. west coast.

62 Recent environmental shocks have challenged the social and economic sustain-
63 ability of the Dungeness crab fishery. In 2014-2016, a record marine heatwave
64 (MHW) led to a harmful algal bloom of unprecedented scale (McCabe et al.,
65 2016), causing toxin levels in Dungeness crabs to reach levels dangerous for

human consumption and correspondingly lengthy delays in large regions of the coast in the 2015-16 and 2016-17 Dungeness fishing seasons. Concurrently, the MHW caused shoreward compression of the preferred feeding habitat of large whales, contributing to a rise in entanglements of whales in Dungeness crab fishing gear and increasing risk of fishery closure due to marine mammal interactions, effects that continued to directly affect fishery closures through the 2017-18 Dungeness crab season (Feist et al., 2021; Santora et al., 2020). During this period, Dungeness crab fishers had to contend with significant ecological changes and the management measures those changes precipitated. Like with climate extremes in other systems (Loon et al., 2016), the effects of this MHW were complex, reverberated through the social-ecological system, and persisted for years after the anomalous warming dissipated (Fisher et al., 2021; Smale et al., 2019; Suryan et al., 2021). While much recent literature is dedicated to examination of biophysical and ecological impacts of the MHW (Cavole et al., 2016; McCabe et al., 2016; von Biela et al., 2019), to date far less attention has been given to exploring how social systems cope and change with these perturbations (Fisher et al., 2021; Jardine et al., 2020; K. M. Moore et al., 2020).

In this study, we compare the adaptive responses of behavioral groups within the Dungeness crab fishery to the multi-year MHW that directly affected the 2015-16 through 2017-18 Dungeness crab seasons. The 2015-16 Dungeness crab season was the first season to be significantly delayed as a direct result of ecosystem changes, a trend that continued through the 2017-18 season. While previous work has investigated economic impacts (Holland and Leonard, 2020; Jardine et al., 2020; Mao and Jardine, 2020) and changes in fishery participation due to the MHW-associated harmful algal bloom (Fisher et al., 2021), here we explicitly investigate and quantify fishers' adaptive spatial behaviors in response to the MHW more broadly and for the full three-year period over which the MHW impacts manifested. Using a 10-year time-series of more than 2 million satellite-derived fishing vessel location records, linked to fishery revenue and landings data, we derive quantitative behavioral metrics describing space use and mobility of Dungeness crab vessels, then organize these behaviors into characteristic behavioral groups. We explore the overlap of spatial behaviors with profitability, fishing season length, and revenue diversity. We track these behavioral groups over time, and identify key behavioral metrics that promoted adaptation during and after the MHW. This analysis therefore offers insights into the types of adaptive behaviors that may promote sustainable outcomes in other commercial fisheries and perhaps in social-ecological systems more broadly.

2. Materials and Methods

2.1. Data sources

We used satellite-based Vessel Monitoring System (VMS) data and port level fishery landings data to define most of the behavioral variables used in the study. The VMS database is maintained by the National Marine Fisheries Service's Office of Law Enforcement, and records the positions of vessels at

approximately one hour intervals. Similar VMS data has been used in other studies of fishery spatial dynamics (Feist et al., 2021; Joo et al., 2015; O’Farrell, Chollett, et al., 2019; Watson and Haynie, 2016). A subset of the vessels that participate in the Dungeness crab fishery are equipped with VMS transponders (primarily vessels that also participate in the west coast groundfish fishery, where VMS transponders are mandatory). This subset varies between 19 and 26 percent of all vessels recording landings for Dungeness crab between the 2008-2009 and 2018-2019 seasons, representing between 10 and 57 percent of all Dungeness crab landings by weight, and between 15 and 42 percent of Dungeness revenue, depending on the year and state (California, Oregon, or Washington). Oregon has the highest relative VMS representation, followed by California, then Washington.

Fish ticket information was obtained through the Pacific Fisheries Information Network (PacFIN). These data represent 1949 vessels targeting Dungeness crab in California, across more than 300,000 fish tickets (i.e., fishing trips). Fishing trips were defined as targeting Dungeness crab if the total landings of Dungeness on the individual fish ticket were at least 10 percent greater than the landed weight of the next greatest species.

We joined the fish ticket data to the VMS data through unique vessel identification numbers and timestamps. VMS geolocations comprising a fishing trip were defined as all of the geolocations between a landed fish ticket and the one immediately preceding it (i.e., the previous ticket landed by the same vessel). After joining the VMS and fish ticket data, we removed the small number of trips in which the final VMS data point for a trip was greater than 50km from the port of landing recorded on the ticket, reasoning that these are unreliable records. Finally, we removed VMS records from vessels sitting idle in port. To do so, we truncated all but the first and last VMS records for each trip that fell within a small buffer zone (1.5 to 3 km) around each port of landing and with an average calculated speed of less than 0.75 m/s.

Dungeness crab fishing seasons on the west coast typically begin in the middle of November (for Central California) or beginning of December (for Northern California, Oregon, and Washington), but can be variable in their starting dates, depending on state (California, Oregon, or Washington), harmful algal bloom closures, price and market conditions, crab condition and meat quality, and potential interactions with protected species like humpback whales. Therefore, we used a data-driven approach to define the start date for each crab season in each of the 20 fishing port groups on the west coast. Port groups are defined by PacFIN and include clusters of small, neighboring fishing ports. For each port group in each season, we found the date after October 31 of each season that the total Dungeness crab landings into that port reached 1 percent of the eventual, season-long landings. This approach identifies the realized start date of the crab fishery in each portion of the coast in each year.

The maximum length of a Dungeness fishing trip was defined as seven days. That is, if there was a gap of greater than seven days between consecutive tickets, the VMS geolocations greater than seven days prior to the landed ticket were discarded. The final dataset comprises a clean record of geolocations associated

155 with each Dungeness crab fishing trip.

156 The only other data source used in the calculation of behavioral metrics is a
157 measure of average daily wind speeds, from AVHRR Pathfinder satellite-derived
158 measurements (<https://data.nodc.noaa.gov>; <https://doi.org/10.7289/v52j68xx>).
159 The data are modelled daily on a 0.04 degree grid (approximately 5 km at the
160 equator) and are available from 1981-present.

161 2.2. Construction of Fishing Behavioral Metrics

162 Fishing behavioral metrics were calculated from fish ticket, VMS, and wind
163 speed data. Our choice of behavioral variables to calculate was driven by previous
164 evidence of the importance of each variable in describing fisher behavioral patterns
165 (Fuller et al., 2017; Kasperski and Holland, 2013; O’Farrell, Chollett, et al., 2019;
166 O’Farrell, Sanchirico, et al., 2019; Pfeiffer and Gratz, 2016). Each of the fisher
167 behavioral variables described one characteristic of a vessel’s apparent behavior
168 over the course of a fishing season—a vessel-season. Cluster analysis (see next
169 section) was performed on these vessel-seasons, and individual vessels could be
170 clustered into different behavioral groups in different seasons. To determine
171 whether a vessel would be included in the analysis, we calculated the total
172 Dungeness crab revenue for each vessel in each season from 2008-09 to 2018-19.
173 The 5th percentile for annual Dungeness revenue per vessel was \$5828. We
174 retained all vessel-seasons with greater than \$5828 in revenue in any season (i.e.,
175 we retain the top 95 percent of all vessel-seasons as measured by revenue).

176 Our behavioral metrics fall into five general categories: port use, fishing
177 trip characteristics, participation in other fisheries, risk-taking behavior, and
178 exploration and mobility (see Table A.1 for full technical definitions of metrics).
179 Port use metrics include the number of ports visited per fishing trip, ports visited
180 per month, diversity of port use (calculated as a Shannon diversity index on the
181 proportions of trips landed in each port), and the total number of ports visited
182 across the entire season. The trip metrics are the mean and standard deviation
183 of trip distance (in km) and duration (in days). Vessel size has been used as
184 a proxy for fleet segments in other studies (Jardine et al., 2020). We did not
185 include vessel size as a metric, since vessel size alone is not a behavioral variable,
186 but we explored its relationship to included metrics as a point of comparison
187 (Figs. A.5, A.6).

188 Fishery participation metrics include season length, proportion of revenue
189 and fish tickets from other (non-Dungeness) fisheries, and revenue diversity. The
190 Dungeness fishery operates as a derby, where the majority of the landings and
191 profits are obtained in the first few months of each season (Fig. A.4). Our
192 season length metric captures this phenomenon and indicates the day of the crab
193 season that each vessel reaches 90 percent of its cumulative landings for that
194 season. To calculate the proportion of revenue and tickets from other fisheries,
195 and revenue diversity, we use a version of the fish ticket data that includes all
196 fishery targets (not just Dungeness crab). Using these tickets, the proportion
197 of non-Dungeness revenue is calculated, as well as the proportion of fish tickets
198 submitted by that vessel with a target other than Dungeness crab. Revenue

diversity for each vessel-season is an inverse Simpson index calculated on the proportion of revenue obtained from each species in a vessel’s fishing portfolio.

Risk-taking behavior is modelled after the definition in Pfeiffer and Gratz (2016), who also studied west-coast fisheries, as propensity to fish in high-wind conditions. Using the Pathfinder winds data, we extracted the wind speed at each VMS location, then calculated the 95th percentile of wind speed experienced by each vessel on each trip. Finally, the risk-taking metric was defined as the proportion of trips in a season where the 95th percentile of experienced wind speed was greater than 7.5 m/s (Pfeiffer and Gratz, 2016).

Exploration and mobility were measured with home range and location choice entropy, adopting the definitions in O’Farrell et al. (2019). Home range was calculated as the area of the minimum convex polygon encompassing all VMS locations in a vessel-season, after removing the five percent of locations that were the furthest from other points (i.e., spatial outliers). Location choice entropy measures the propensity of vessels to explore new locations versus returning to the same locations, and is calculated cumulatively across each vessel’s fishing season (O’Farrell, Sanchirico, et al., 2019). Spatial locations were defined as individual cells on a 5x5km grid. As a season progresses, entropy increases as vessels explore novel locations and decreases as the same locations are revisited repeatedly. The season-long metric for exploration for each vessel is defined as the 90th percentile of maximum location choice entropy in that season.

Definitions of all metrics used in the clustering analysis are provided in the Appendix.

2.3. Cluster Analysis

All metrics were checked for collinearity, and thinned such that no two metrics had a Pearson correlation greater than 0.7. This thinning removed mean and standard deviation of trip distance, total number of visited ports, and proportion of non-Dungeness tickets from the analysis. The remaining 11 metrics were scaled to range from zero to one by dividing each metric by its maximum value (across all seasons). Clustering was performed using Euclidean distances and Ward aggregation that minimizes total within-cluster variance. The number of clusters was determined using the Nbclust package in R (Charrad et al., 2014), which calculates 22 clustering indices before recommending an optimal number of clusters via majority vote amongst indices. Adopting the optimal clusters defined by NbClust, we visualized results graphically using principal component analysis. After vessel-seasons were assigned to groups, we tested for differences between groups along specific behavioral metrics using Tukey’s HSD.

The importance of individual metrics in discriminating between clusters was calculated using random forest analysis, utilizing the randomForest package in R (Liaw and Wiener, 2002). Random forests were grown on subsamples of the data to classify vessel-seasons according to their defined clusters from the previous step. Then, these random forests were used to predict withheld data. Variable importance was defined as the increase in the rate of mis-classification of vessel-seasons into clusters when the particular variable was randomly permuted.

243 2.4. Dungeness Fishing Profitability

244 We used fish ticket data to assess the per-trip, per-week, and per-season
 245 landings and revenue of vessels in each fisher behavioral group over time. Addi-
 246 tionally, we modeled fishing costs following the approach of Dewees et al. (2004)
 247 to assign an estimated profit to each fishing trip. The cost of a fishing trip C_t is
 248 assumed to be a function of fuel C_f and bait costs, and the costs of labor (i.e.,
 249 crew) C_c :

$$C_t = C_f + C_c$$

250 Fuel and bait cost is a function of vessel size L and number of days fished d , as
 251 well as trip year y to adjust for an assumed 2 percent inflation rate.

$$C_f = f(L, d, y)$$

252 Crew cost is a function of vessel size (because larger vessels require more crew
 253 members) and total trip revenue R (since crew members receive a proportion of
 254 revenue).

$$C_c = f(L, R)$$

255 The above cost relationships were parameterized using data from Dewees
 256 et al. (2004), who administered a survey to 243 Dungeness crab fishers and
 257 compiled estimates of fishing costs by vessel size. The survey estimated costs
 258 associated with bait, fuel, and labor (crew) for small (less than 9.1m), medium
 259 (9.1-15.2 m) and large (greater than 15.2 m) fishing vessels. Using the means
 260 and standard deviations of these costs reported in Dewees et al. (2004), we
 261 simulated 10,000 trip costs for vessels ranging in length from 6.4 to 31.4 m,
 262 which is the range of vessel sizes in our data. Then, linear relationships between
 263 vessel size and both types of costs were estimated with simple linear regression.
 264 The resulting relationships,

$$C_f = d(150 + 3.5L) * 1.02^{y-2004}$$

$$C_c = R(0.17 + 0.0018L)$$

265 were used to deterministically assign a cost to each Dungeness fishing trip in
 266 our data. From there, a profit for each trip could be estimated by subtracting
 267 costs from revenue. Using trip-level profits, we calculated mean profits per week—
 268 across seasons—for vessels in each behavioral group, as well as season-long
 269 profits.

270 Using the fish ticket revenue data, we also calculated total revenue from all
 271 non-Dungeness fisheries for each vessel-season in the analysis. We constrained
 272 the calculation of non-Dungeness revenue to only those fishing trips that occurred
 273 within each vessel's apparent Dungeness season (that is, within the time period
 274 where the vessel was also landing Dungeness crab).

2.5. *Adaptation to the Marine Heatwave*

Using the results of cluster analyses, we compared key characteristics of behavioral groups in MHW versus non-MHW crab seasons. We defined the MHW as encompassing the crab fishing seasons from 2015-16 to 2017-18. Although there is evidence that the MHW began affecting west coast ecosystems as early as late 2014 (Cavole et al., 2016; McCabe et al., 2016), the 2015-16 Dungeness crab season was the first to be significantly delayed as a direct result of ecosystem changes (Jardine et al., 2020), a trend that continued through the 2017-18 season.

Adopting this definition of the MHW period, we compared mean Dungeness profit, non-Dungeness revenue (i.e., external fishery revenue), and home range size over time among behavioral groups to assess potential adaptive strategies. For each of these three comparisons, we performed a two-way ANOVA to test for significant differences in mean profits, revenue, and home range by behavioral group and period (non-MHW or MHW).

All analyses in the study were performed in R (R Core Team, 2021).

3. Results

3.1. *Describing Fisher Behavior*

The combined vessel telemetry and fisheries landings dataset captured the behaviors of 596 different vessels spanning 11 fishing seasons (2008-2019), with ~2.2 million satellite-derived Vessel Monitoring System (VMS) geolocations, and 315,000 fishery landing records. Using these combined data, we analyzed 11 behavioral variables in five general behavioral categories: fishing port use, fishing trip characteristics, participation in other fisheries, risk-taking behavior, and exploration and mobility (definitions of all metrics are provided in Table A.1).

The 3391 vessel-seasons in our data fell into four behavioral cluster groups (Figs. 1a, A.1). The most important discriminating variables driving the clustering according to random forest analysis were proportion of revenue from non-Dungeness crab fisheries, followed by diversity of port use, revenue diversity, and mean trip duration (Fig. 1b). These analyses suggest that the behavior of the four groups can be conceptualized as varying along two major axes (Fig. 1c): (1) spatial mobility (principal component 1 in Fig. 1a) and (2) propensity to fish in non-Dungeness crab fisheries (fishery flexibility, principal component 2 in Fig. 1a).

Vessels with higher spatial mobility, which we term Roving groups, move between ports throughout a fishing season and have large fishing ranges, while those with lower mobility—Local groups—show greater fidelity to a single port. Vessels with greater fishery flexibility, deemed Generalist groups, have high revenue diversity and derive a relatively greater portion of their total fishery revenue from fisheries other than Dungeness crab. Vessels exhibiting less flexibility—Specialists—concentrate fishing effort within the Dungeness crab fishery. A vessel-season is therefore defined as either Roving or Local, and either Specialist or Generalist. As an example, for crab vessels fishing out of Newport, Oregon, Local Specialists have the smallest fishing grounds, followed by

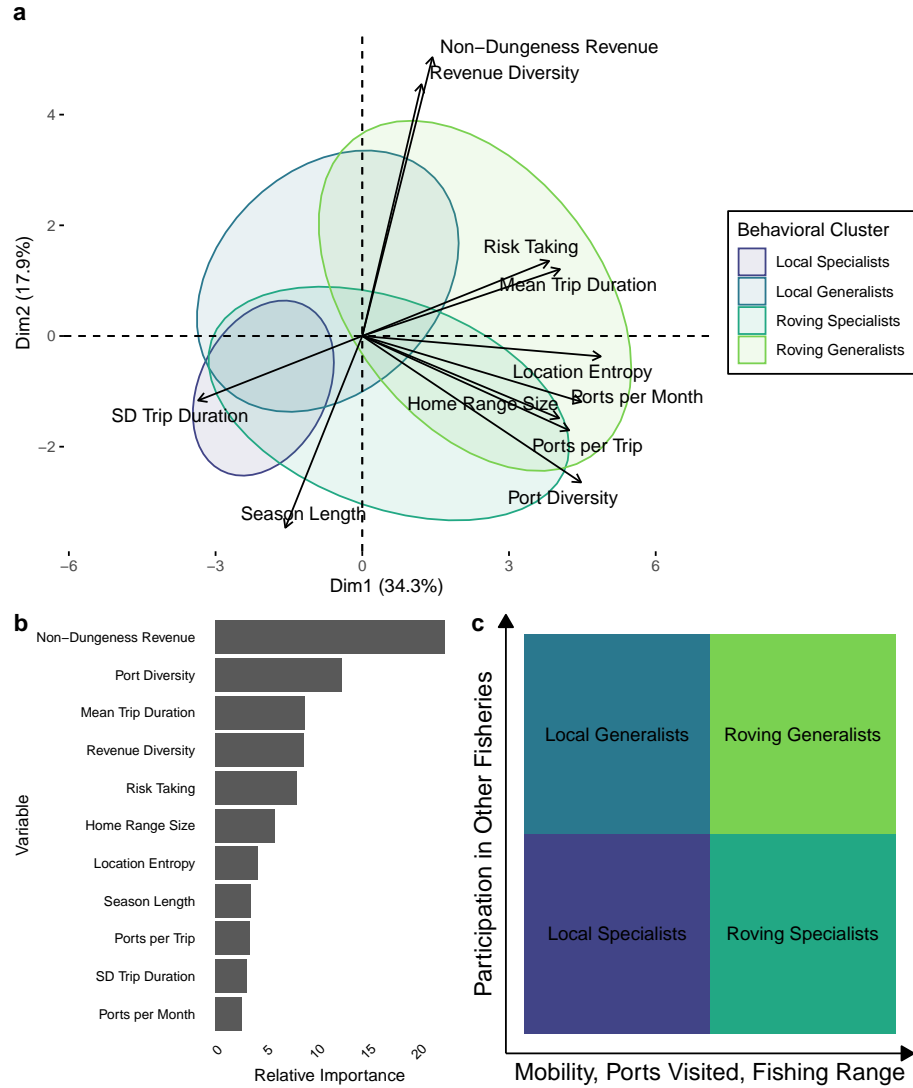


Figure 1: Data-driven formation of fishing behavioral groups. (a) Principal component analysis of vessel-seasons. Clusters of vessel-seasons, which determine behavioral groups, are enclosed by ellipses. Arrows represent the association between metrics in the cluster analysis relative to the placement of vessel-seasons. (b) Ranked importance of top variables used to classify vessel-seasons into behavioral groups, as determined by random forest analysis. (c) Conceptual visualization of the major axes defining behavioral groups.

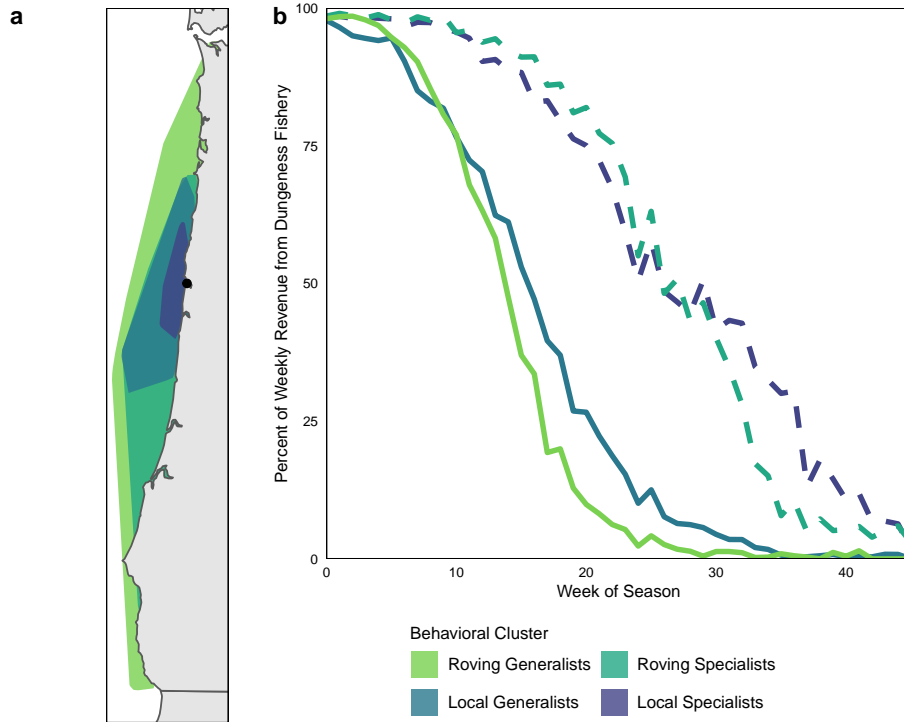


Figure 2: Characteristic patterns in spatial mobility and fishery flexibility across behavioral groups in the west coast Dungeness crab fishery, exemplified by an Oregon port. (a) Fishing footprints of each behavioral group across all seasons for vessels originating from the Port of Newport, Oregon, USA. Shaded polygons are 95 percent convex hulls of all VMS locations for each group. (b) Fishery flexibility, displayed as the mean percent of total weekly revenue obtained from the Dungeness crab fishery (relative to all other fisheries) by vessels in each behavioral group. Weekly revenues are averaged across crab seasons and across all vessels in each group. Generalist groups are represented with solid lines, while Specialist groups are represented with dashed lines.

Local Generalists, Roving Specialists, and Roving Generalists (Fig 2a). Across all vessel-seasons, Generalist vessels have shorter crab fishing seasons, exiting the Dungeness crab fishery earlier to pursue other fishing opportunities, while Specialists continue to garner a large percentage of their weekly landed revenue from Dungeness crab over the course of the season (Fig. 2b).

3.2. Behavioral Changes During the Marine Heatwave

The four fishing behavioral groups defined by our cluster analysis responded to the social-ecological disruption of the marine heatwave (MHW) by increasing their dependence on other, non-Dungeness fisheries and expanding their fishing ranges. All groups had higher non-Dungeness fishery revenue during the MHW period than during other seasons, indicating a potential fallback to other fisheries during a period of delays and management disruptions in the crab fishery (Fig.

3)(Fisher et al., 2021; Holland et al., 2020). The 2016-17 and 2017-18 seasons had the highest non-Dungeness crab revenue in the time series (Fig. 3a). The Generalist groups in particular more than doubled their revenues from non-Dungeness fisheries (ANOVA $p < 0.01$; Fig. 3b). The Specialist groups also had greater non-Dungeness revenues during the MHW period, but the differences were not as substantial as for the Generalist groups (Table S2, ANOVA $p = 0.06$ for Roving Specialists, $p=0.99$ for Local Specialists).

Some Dungeness fishers also expanded their Dungeness crab fishing grounds during the MHW, particularly the two Roving groups (Fig. 4). Prior to the MHW (2008-15), Roving Generalists had the largest mean home range size at more than 4000 square kilometers (Fig. 4a). Roving Specialists had the second-largest ranges on average (around 2500 square kilometers), while the Local groups had much smaller ranges (less than 1000 square kilometers). In the MHW period from 2015-18, the Roving groups fished significantly larger areas, with the Roving Generalist and Roving Specialist groups averaging more than 5500 and 3500 square kilometers fished, respectively ($p=0.001$ and $p\ll 0.001$ for Roving Specialists and Roving Generalists). In contrast, the areas fished for the Local groups did not change significantly (Fig. 4b and Table S3, $p>0.99$ for both Local groups). For all four groups, within the MHW period, the most pronounced change in mobility occurred during the 2016-17 fishing season.

3.3. Profitability of Behavioral Groups during the Marine Heatwave

An open question is whether the adaptive responses we detected and quantified—greater spatial mobility and more flexible fishing—allowed fishers to maintain profits in the face of this major environmental perturbation. Our fishing cost model provides an estimation of Dungeness crab profit (reported revenue minus estimated cost) for every fishing trip in the data (i.e., for those vessels that continued to fish), and allowed us to describe how profits within each behavioral group varied over time (Fig. 5).

For all groups, average revenues and estimated costs both increased during the MHW period (Figs. A.7, A.8), but revenue increases outweighed the increases in cost, resulting in increased estimated profits. Dungeness crab profits for all behavioral groups increased during the MHW, significantly so for Local Generalists ($p=0.05$), Roving Generalists ($p\ll 0.0001$) and Roving Specialists ($p=0.001$, Table A.4). The Roving Generalist group saw the largest increase in estimated profits in both raw and percent increase in profits (more than a \$63,000 increase per vessel, a 48 percent increase, on average). Local Specialists experienced the smallest increase in profits of all groups (25 percent) during the MHW, while Roving Specialists and Local Generalists experienced a greater than 40 percent increase. In the season after the dissipation of the MHW, estimated profits declined, particularly for the Roving groups.

4. Discussion

The pace and magnitude of environmental change in the Anthropocene demand assessment of how social-ecological systems will respond. Ideally, man-

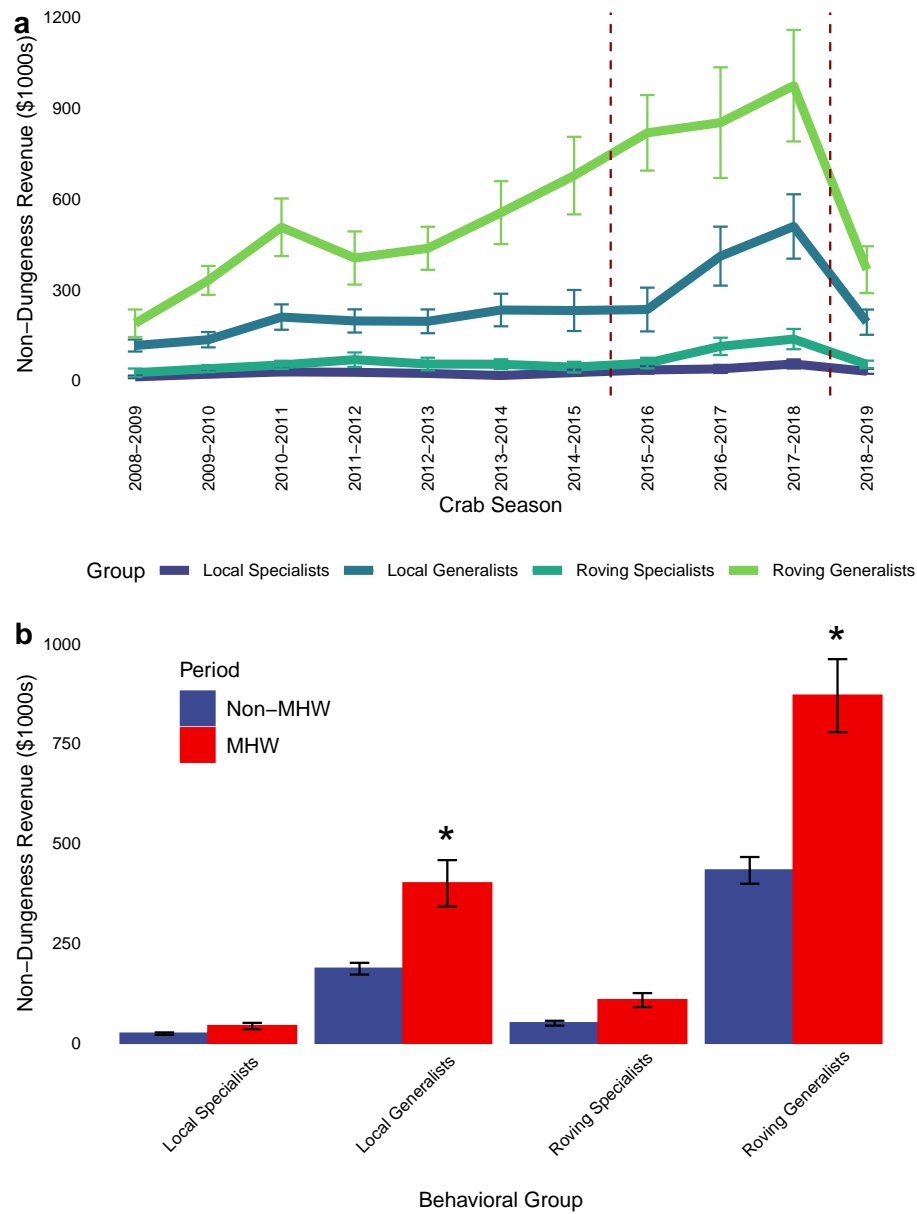


Figure 3: Non-Dungeness revenue for vessels in the analysis. (a) Seasonal mean revenue (\pm 2SE) for vessels in each behavioral group coming from all non-Dungeness fisheries combined. Vertical lines delineate the period of the marine heatwave (MHW). (b) Barplot of mean revenue (\pm 2SE) for vessels in each group during MHW and non-MHW seasons. Stars indicate groups with significantly different non-Dungeness revenue in MHW seasons.

agement approaches can be designed to help humanity adapt by meeting the basic needs of people without compromising ecosystems for future generations (Lubchenco et al., 2016). As one of the last remaining hunter-gatherer activities occurring at scale, commercial fisheries offer an important lens through which to understand human adaptations to novel and extreme conditions. The 2014-2016 marine heatwave on the U.S. west coast stressed the adaptive ability of participants in the highly lucrative Dungeness crab fishery, because an environmental perturbation—the MHW and associated harmful algal bloom and shoreward compression of large whale habitat—led to cascading regulatory actions and market effects (Holland et al., 2020). Our analysis revealed that Dungeness crab fishers that remained in the fishery responded to unprecedented environmental and management changes in multiple ways. Behavioral groups characterized by spatial mobility used expanded fishing grounds in the 2016-17 and 2017-18 seasons to maintain or increase revenues. Similarly, fishers with strategies based around diversified fishing portfolios (Generalists) were able to increase their revenue from other fisheries to bolster their total fishing income. We found that vessels combining greater spatial mobility with higher participation rates in other fisheries were the most profitable, and that these financial benefits were maintained or magnified during the MHW. The behavioral strategies observed in the Dungeness crab fishery suggest that both portfolio and spatial diversification pathways can improve adaptive capacity for human harvesters across industrialized food systems during an era in which the magnitude, frequency, and intensity of environmental perturbations are increasing.

Our work builds on research from the economics (Gordon, 1954; Smith and McKelvey, 1986), evolution (Gallagher et al., 2015), and ecology (Beever et al., 2017) literatures investigating the relative ability of specialists and generalists to cope with environmental change. The cross-disciplinary consensus is that generalists may adapt better to increasingly variable environments. Smith and McKelvey (1986) suggested that specialists and generalists in fisheries use different strategies to cope with variability and uncertainty in income—specialists are efficient and minimize income risk through fishery-specific acumen, while generalists hedge against risk by building diverse portfolios (Kasperski and Holland, 2013; Oken et al., 2021). In a direct ecological analogy, generalist consumers in an ecosystem experiencing novel environmental conditions may be able to gain a competitive advantage over specialists by efficiently switching to alternative prey sources (Beever et al., 2017).

While management dynamics, markets, stochastic resource abundance, and conditions in other fisheries are complicating factors (Holland et al., 2020), the relative performance of specialist versus generalist strategies in the Dungeness crab fishery largely adhere to these existing economic and ecological models. Although some Specialists and Generalists persisted through the MHW period, repeated environmental disruptions in the future that cause further seasonal and spatial restrictions on the Dungeness crab fishery may begin to favor a Generalist, diversified strategy. Within the US west coast context, existing fishery governance systems may constrain this type of generalist adaptation (Kasperski and Holland, 2013; Russell et al., 2018), but there are calls for “climate-ready”

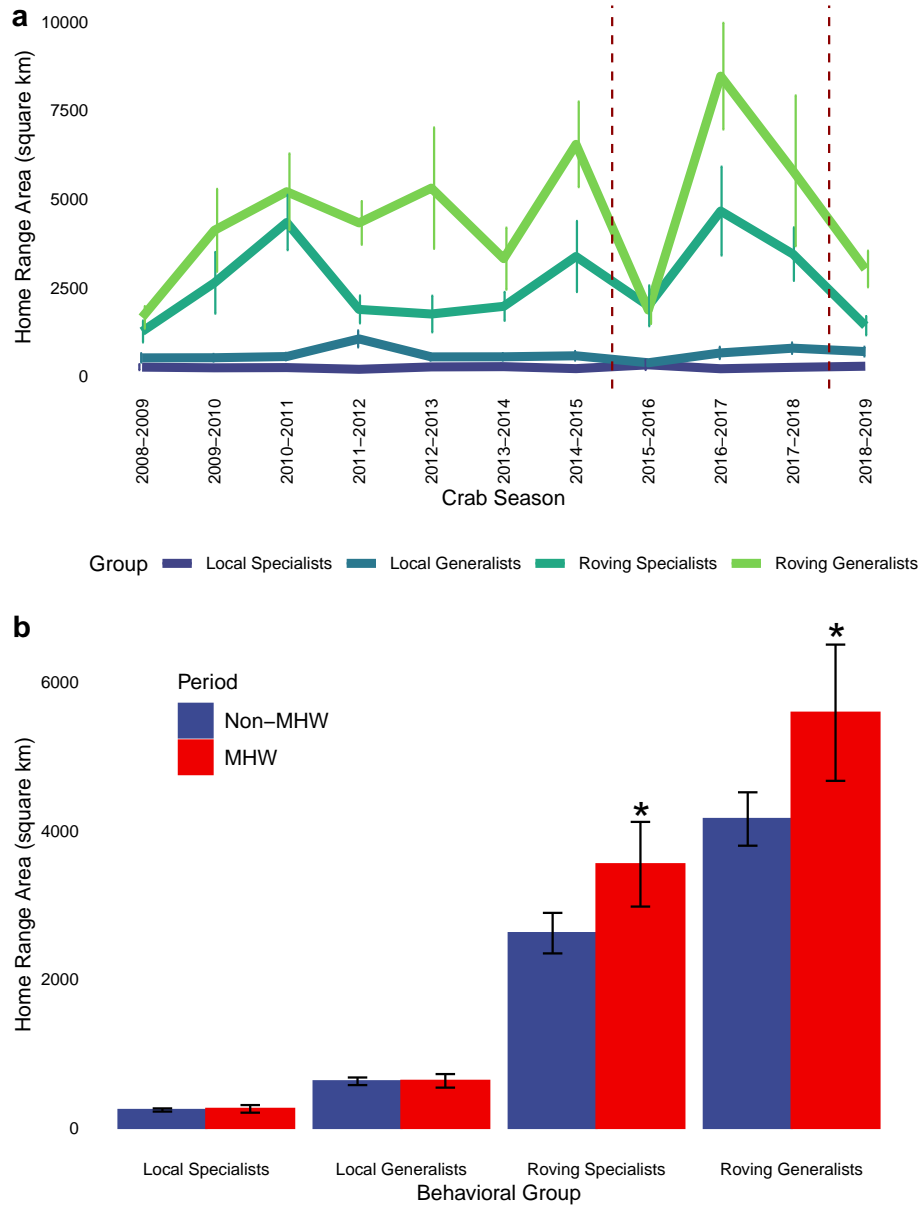


Figure 4: Home range (fishing area) size for vessels in the analysis. (a) Seasonal mean home range area in square kilometers ($\pm 2SE$) for vessels in each behavioral group. Vertical lines delineate the period of the MHW. (b) Barplot of mean home range area ($\pm 2SE$) for vessels in each group during MHW and non-MHW seasons. Stars indicate groups with significantly different home range size during MHW seasons.

419 fisheries that include the flexibility for fishers to move between fisheries (Wilson
420 et al., 2018). A better understanding of the social, economic, and cultural drivers
421 of fishers’ decisions to be specialists or generalists is a core component of a
422 sustainable livelihoods approach to small-scale fisheries management (Allison
423 and Ellis, 2001; Finkbeiner, 2015). Such an approach can also offer insights for
424 the design of regulatory approaches that facilitate resilience to environmental
425 perturbation in larger-scale fisheries and natural resource management contexts
426 (Salas and Gaertner, 2004).

427 Diversification of fishery revenue was not the only axis of variation associated
428 with persistence in the face of the MHW. Spatial mobility was also a key
429 component of the fishing strategies we observed. Following others who have
430 used recently emerging technologies to understand the sustainability of human
431 harvester strategies (Brodie and Fragoso, 2020; Frawley et al., 2020; Renner and
432 Kuletz, 2015), we used satellite data to characterize the spatial behavior of vessels.
433 Roving groups, whether Specialists or Generalists, were more profitable than their
434 Local counterparts under all conditions. The benefits of this spatial mobility were
435 clear during the marine heatwave. We hypothesize that Roving vessels were the
436 most capable of responding to management actions, market forces, and ecological
437 factors (e.g., product quantity and quality) that shifted spatially during the
438 heatwave. The ability of more exploratory fishers to cope during an environmental
439 disturbance has recently been demonstrated in other commercial fisheries systems
440 (O’Farrell, Sanchirico, et al., 2019), and our findings confirm that more mobile
441 vessels performed better during the environmental perturbation. Similar patterns
442 have been shown among foraging marine mammals, where individual animals
443 that are more exploratory have greater foraging success during anomalous climate
444 conditions than more site-faithful conspecifics (Abrahms et al., 2018).

445 Importantly, the nature of the data used in this study means that we studied
446 the behavior of the ‘survivors’—that is, the fishers who decided or were able to
447 remain in the Dungeness crab fishery during the MHW period. The MHW acted
448 as a selective force on Dungeness crab fishery participation. Many Dungeness
449 crab fishers during the 2016 and 2017 fishery closures chose (or were forced
450 by circumstance) to not participate in the fishery at all, instead opting to exit
451 fishing entirely or to re-concentrate all effort in alternative fisheries (Fisher et
452 al., 2021). Some of the relative success of the Dungeness crab fishers during the
453 MHW observed in this study, therefore, may be due to reduced competition, as
454 well as periods of supply shortages and high prices. Although outside the scope
455 of the current analysis, an important area for further research is to determine
456 how and why, when faced with an environmental perturbation, fishers choose to
457 remain or exit a fishery (S. K. Moore et al., 2020).

458 With climate change expected to increase the frequency of extreme environ-
459 mental perturbations like MHWs (Oliver et al., 2018), established patterns of
460 natural resource management and human harvester behavior will be challenged.
461 In our study, following multiple adaptive pathways by both diversifying and
462 mobilizing appears to be one solution to an extreme environmental event and
463 rapid management changes in the Dungeness crab fishery. Management mea-
464 sures that restrict the fishery temporally or spatially—such as spatially-explicit

465 biotoxin-related closures or early termination of the fishing season due to risk
 466 of interactions with protected or bycatch species—will differentially affect dis-
 467 tinct groups of fishers. Single-fishery specialists may thrive when the harvested
 468 resource is stable and productive, but these fishers may struggle to adapt if
 469 management measures restrict fishing season lengths. Likewise, localized fishers
 470 can be successful through intimate knowledge of fishing grounds, but if large-scale
 471 environmental perturbations have spatially-explicit negative effects, fishers with
 472 knowledge of a wider array of fishing grounds and greater mobility will naturally
 473 gain an advantage (O’Farrell, Sanchirico, et al., 2019). Over time, management
 474 context, or failures of management to adapt, can drive changes in the makeup
 475 of fishing fleets as a whole (Frawley et al., 2020). These changes are not in-
 476 herently negative, but in order to maintain the social, economic, and cultural
 477 benefits provided by a fishery, managers should endeavour to anticipate behav-
 478 ioral changes within fleets. More generally, these insights are congruent with
 479 an evolving understanding of adaptation in complex social-ecological systems
 480 (Lubchenco et al., 2016). Because complex systems are an emergent product of
 481 the individual actions of human actors, informed adaptive management requires
 482 an understanding of the drivers of behaviors like those identified in this study
 483 along with well-calibrated and nimble responses within governance systems.

484 For fishers and other human harvesters, future work using mixed methods
 485 from the social sciences like participatory mapping and semi-structured inter-
 486 views (Frawley et al., 2020; S. K. Moore et al., 2020; Pellowe and Leslie, 2019;
 487 Ritzman et al., 2018) will provide complementary insights into the motivations
 488 and social drivers behind adaptive decisions, and could help identify system-
 489 specific metrics of success or performance beyond profitability. Furthermore,
 490 as integrated biophysical and socioeconomic data streams become increasingly
 491 available for environmental management (Bradley et al., 2019), data-driven, inter-
 492 disciplinary studies of resilience and adaptation will enable dynamic management
 493 of natural resources (Hazen et al., 2018; Maxwell et al., 2015). This push for the
 494 incorporation of multiple data streams in environmental management extends
 495 beyond marine fisheries. For example, in wildland fire management in the United
 496 States, integrated data platforms that combine geospatial data with risk models
 497 and fuel treatment scenarios are empowering adaptive fire management plans
 498 (Ager et al., 2011; Krofcheck et al., 2018).

499 This study revealed the elements of behavioral diversity among human
 500 harvesters in a lucrative, keystone commercial fishery, and described how those
 501 elements enabled adaptation during an extreme environmental event attributable
 502 to climate change (Hinder et al., 2012). Just as biological response diversity
 503 can lead to enhanced ecosystem resilience to environmental change (Elmqvist
 504 et al., 2003), behavioral diversity among natural resource users may promote
 505 resilience of social-ecological systems. Given the impending increase in extreme
 506 climatic events such as marine heatwaves (Burge et al., 2014; Smale et al., 2019),
 507 recognition of social and ecological traits that enable resilience now can help to
 508 build toward a more prepared future. As quantitative data become increasingly
 509 available in the United States and far beyond (Bradley et al., 2019), behavioral
 510 analyses like ours can be used in the design of adaptive management measures,

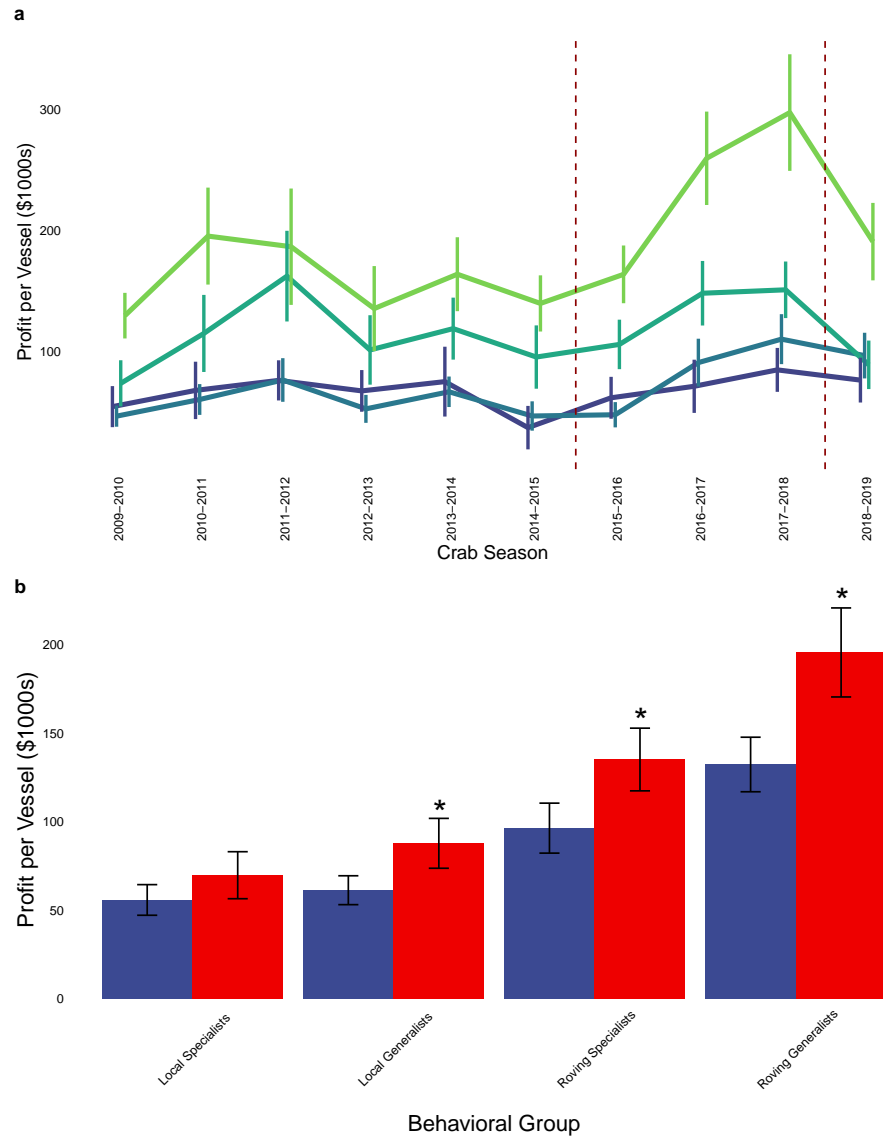


Figure 5: Estimated profits by behavioral group. (a) Mean profit (± 2 SE) for vessels in each behavioral group over the full crab season. Vertical lines delineate the period of the marine heatwave. (b) Mean profit (± 2 SE) for each group in heatwave (MHW) versus non-MHW seasons. Stars indicate groups with significantly different estimated profits during MHW seasons.

511 to bolster policy analyses (Cabral et al., 2018), and to inform decision-making
512 under environmental uncertainty.

513 References

- 514 Abatzoglou, J.T., Williams, A.P., Barbero, R., 2019. Global emergence of
515 anthropogenic climate change in fire weather indices. *Geophysical Research*
516 *Letters* 46, 326–336. doi:10.1029/2018GL080959
- 517 Abrahams, B., Hazen, E.L., Bograd, S.J., Brashares, J.S., Robinson, P.W.,
518 Scales, K.L., Crocker, D.E., Costa, D.P., 2018. Climate mediates the suc-
519 cess of migration strategies in a marine predator. *Ecology Letters* 21, 63–71.
520 doi:10.1111/ele.12871
- 521 Ager, A.A., Vaillant, N.M., Finney, M.A., 2011. Integrating fire behav-
522 ior models and geospatial analysis for wildland fire risk assessment and fuel
523 management planning. *Journal of Combustion* 2011. doi:10.1155/2011/572452
- 524 Allison, E.H., Ellis, F., 2001. The livelihoods approach and management of
525 small-scale fisheries. *Marine Policy* 25, 377–388.
- 526 Beever, E.A., Hall, L.E., Varner, J., Loosen, A.E., Dunham, J.B., Gahl, M.K.,
527 Smith, F.A., Lawler, J.J., 2017. Behavioral flexibility as a mechanism for coping
528 with climate change. *Frontiers in Ecology and the Environment* 15, 299–308.
529 doi:10.1002/fee.1502
- 530 Bradley, D., Merrifield, M., Miller, K.M., Lomonico, S., Wilson, J.R., Gleason,
531 M.G., 2019. Opportunities to improve fisheries management through innova-
532 tive technology and advanced data systems. *Fish and Fisheries* 20, 564–583.
533 doi:10.1111/faf.12361
- 534 Branch, T.A., Hilborn, R., Haynie, A.C., Fay, G., Flynn, L., Griffiths, J.,
535 Marshall, K.N., Randall, J.K., Scheuerell, J.M., Ward, E.J., Young, M., 2006.
536 Fleet dynamics and fishermen behavior: Lessons for fisheries managers. *Canadian*
537 *Journal of Fisheries and Aquatic Sciences* 63, 1647–1668.
- 538 Brodie, J.F., Fragoso, J.M.V., 2020. Understanding the distribution of
539 bushmeat hunting effort across landscapes by testing hypotheses about human
540 foraging. *Conservation Biology* 0, 1–10. doi:10.1111/cobi.13612
- 541 Burge, C.A., Eakin, C.M., Friedman, C.S., Froelich, B., Hershberger, P.K.,
542 Hofmann, E.E., Petes, L.E., Prager, K.C., Weil, E., Willis, B.L., Ford, S.E.,
543 Harvell, C.D., 2014. Climate change influences on marine infectious diseases:
544 Implications for management and society. *Annual Review of Marine Science* 6,
545 249–277. doi:10.1146/annurev-marine-010213-135029
- 546 Cabral, R.B., Mayorga, J., Clemence, M., Lynham, J., Koeshendrajana, S.,
547 Muawanah, U., Nugroho, D., Anna, Z., Ghofar, A., Zulfainarni, N., others, 2018.
548 Rapid and lasting gains from solving illegal fishing. *Nature Ecology & Evolution*
549 2, 650–658.
- 550 Cavole, L.M., Demko, A.M., Diner, R.E., Giddings, A., Koester, I., Pagniello,
551 C.M., Paulsen, M.-L., Ramirez-Valdez, A., Schwenck, S.M., Yen, N.K., others,
552 2016. Biological impacts of the 2013–2015 warm-water anomaly in the northeast
553 pacific: Winners, losers, and the future. *Oceanography* 29, 273–285.

554 Charrad, M., Ghazzali, N., Boiteau, V., Niknafs, A., 2014. NbClust: An R
555 package for determining the relevant number of clusters in a data set. *Journal*
556 *of Statistical Software* 61, 1–36.

557 Cohen, J.D., McClure, S.M., Yu, A.J., 2007. Should i stay or should i
558 go? How the human brain manages the trade-off between exploitation and
559 exploration. *Philosophical Transactions of the Royal Society B: Biological*
560 *Sciences* 362, 933–942.

561 Cook, B.I., Mankin, J.S., Anchukaitis, K.J., 2018. Climate change and
562 drought: From past to future. *Current Climate Change Reports* 4, 164–179.
563 doi:10.1007/s40641-018-0093-2

564 Dewees, C.M., Sortais, K., Krachey, M.J., Hackett, S.C., Hankin, D.G., 2004.
565 Racing for crabs... costs and management options evaluated in dungeness crab
566 fishery. *California Agriculture* 58, 186–189. doi:10.3733/ca.v058n04p186

567 Elmqvist, T., Folke, C., Nyström, M., Peterson, G., Bengtsson, J., Walker, B.,
568 Norberg, J., 2003. Response diversity, ecosystem change, and resilience. *Frontiers*
569 *in Ecology and the Environment* 1, 488–494. doi:10.1890/1540-9295(2003)001[0488:RDECAR]2.0.CO;2

570 Feist, B.E., Samhouri, J.F., Forney, K.A., Saez, L.E., 2021. Footprints of
571 fixed-gear fisheries in relation to rising whale entanglements on the u.s. West
572 coast. *Fisheries Management and Ecology* 28, 283–294. doi:10.1111/fme.12478

573 Finkbeiner, E.M., 2015. The role of diversification in dynamic small-scale
574 fisheries: Lessons from baja california sur, mexico. *Global Environmental Change*
575 32, 139–152. doi:10.1016/j.gloenvcha.2015.03.009

576 Fisher, M.C., Moore, S.K., Jardine, S.L., Watson, J.R., Samhouri, J.F., 2021.
577 Climate shock effects and mediation in fisheries. *Proceedings of the National*
578 *Academy of Sciences of the United States of America* 118, 1–8. doi:10.1073/pnas.2014379117

579 Frawley, T.H., Muhling, B.A., Brodie, S., Fisher, M.C., Tommasi, D., Fol,
580 G.L., Hazen, E.L., Stohs, S.S., Finkbeiner, E.M., Jacox, M.G., 2020. Changes to
581 the structure and function of an albacore fishery reveal shifting social-ecological
582 realities for pacific northwest fishermen 1–18. doi:10.1111/faf.12519

583 Fryxell, J.M., Hilborn, R., Bieg, C., Turgeon, K., Caskenette, A., McCann,
584 K.S., 2017. Supply and demand drive a critical transition to dysfunctional
585 fisheries. *Proceedings of the National Academy of Sciences of the United States*
586 *of America* 114, 12333–12337. doi:10.1073/pnas.1705525114

587 Fuller, E.C., Samhouri, J.F., Stoll, J.S., Levin, S.A., Watson, J.R., 2017.
588 Characterizing fisheries connectivity in marine social-ecological systems. *ICES*
589 *Journal of Marine Science* 74, 2087–2096. doi:10.1093/icesjms/fsx128

590 Fulton, E.A., Smith, A.D.M., Smith, D.C., Putten, I.E.V., 2011. Human
591 behaviour: The key source of uncertainty in fisheries management. *Fish and*
592 *Fisheries* 12, 2–17. doi:10.1111/j.1467-2979.2010.00371.x

593 Gallagher, A.J., Hammerschlag, N., Cooke, S.J., Costa, D.P., Irschick, D.J.,
594 2015. Evolutionary theory as a tool for predicting extinction risk. *Trends in*
595 *Ecology and Evolution* 30, 61–65. doi:10.1016/j.tree.2014.12.001

596 Gladics, A.J., Melvin, E.F., Suryan, R.M., Good, T.P., Jannot, J.E., Guy,
597 T.J., 2017. Fishery-specific solutions to seabird bycatch in the u.s. West coast
598 sablefish fishery. *Fisheries Research* 196, 85–95. doi:10.1016/j.fishres.2017.08.015

599 Gordon, H.S., 1954. The economic theory of a common-property resource:
600 The fishery. *The Journal of Political Economy* 124–142.

601 Hamilton, S., Baker, G.B., 2019. Technical mitigation to reduce marine
602 mammal bycatch and entanglement in commercial fishing gear: Lessons learnt
603 and future directions. *Reviews in Fish Biology and Fisheries* 29, 223–247.
604 doi:10.1007/s11160-019-09550-6

605 Hazen, E.L., Scales, K.L., Maxwell, S.M., Briscoe, D.K., Welch, H., Bograd,
606 S.J., Bailey, H., Benson, S.R., Eguchi, T., Dewar, H., Kohin, S., Costa, D.P.,
607 Crowder, L.B., Lewison, R.L., 2018. A dynamic ocean management tool to
608 reduce bycatch and support sustainable fisheries. *Science Advances* 4, eaar3001.
609 doi:10.1126/sciadv.aar3001

610 Hilborn, R., 1985. Fleet dynamics and individual variation: Why some people
611 catch more fish than others. *Canadian Journal of Fisheries and Aquatic Sciences*
612 42, 2–13. doi:10.1139/f85-001

613 Hinder, S.L., Hays, G.C., Edwards, M., Roberts, E.C., Walne, A.W., Gravenor,
614 M.B., 2012. Changes in marine dinoflagellate and diatom abundance under cli-
615 mate change. *Nature Climate Change* 2, 271–275. doi:10.1038/nclimate1388

616 Holland, D.S., Abbott, J.K., Norman, K.E., 2020. Fishing to live or living to
617 fish: Job satisfaction and identity of west coast fishermen. *Ambio* 49, 628–639.
618 doi:10.1007/s13280-019-01206-w

619 Holland, D.S., Leonard, J., 2020. Is a delay a disaster? Economic impacts of
620 the delay of the california dungeness crab fishery due to a harmful algal bloom.
621 *Harmful Algae* 98, 101904. doi:10.1016/j.hal.2020.101904

622 Holland, D.S., Speir, C., Agar, J., Crosson, S., Depiper, G., Kasperski, S.,
623 Kitts, A.W., Perruso, L., 2017. Impact of catch shares on diversification of
624 fishers’ income and risk. *Proceedings of the National Academy of Sciences of*
625 *the United States of America* 114, 9302–9307. doi:10.1073/pnas.1702382114

626 Jardine, S.L., Fisher, M.C., Moore, S.K., Samhouri, J.F., 2020. Inequality in
627 the economic impacts from climate shocks in fisheries: The case of harmful algal
628 blooms. *Ecological Economics* 176, 106691. doi:10.1016/j.ecolecon.2020.106691

629 Joo, R., Salcedo, O., Gutierrez, M., Fablet, R., Bertrand, S., 2015. Defin-
630 ing fishing spatial strategies from vms data: Insights from the world’s largest
631 monospecific fishery. *Fisheries Research* 164, 223–230. doi:10.1016/j.fishres.2014.12.004

632 Kasperski, S., Holland, D.S., 2013. Income diversification and risk for
633 fishermen. *Proceedings of the National Academy of Sciences* 110, 2076–2081.

634 Krofcheck, D.J., Hurteau, M.D., Scheller, R.M., Loudermilk, E.L., 2018.
635 Prioritizing forest fuels treatments based on the probability of high-severity fire
636 restores adaptive capacity in sierran forests. *Global Change Biology* 24, 729–737.
637 doi:10.1111/gcb.13913

638 Leslie, H.M., Basurto, X., Nenadovic, M., Sievanen, L., Cavanaugh, K.C.,
639 Cota-Nieto, J.J., Erisman, B.E., Finkbeiner, E., Hinojosa-Arango, G., Moreno-
640 Báez, M., Nagavarapu, S., Reddy, S.M.W., Sánchez-Rodríguez, A., Siegel, K.,
641 Ulibarria-Valenzuela, J.J., Weaver, A.H., Aburto-Oropeza, O., 2015. Opera-
642 tionalizing the social-ecological systems framework to assess sustainability. *Pro-*
643 *ceedings of the National Academy of Sciences of the United States of America*
644 112, 5979–5984. doi:10.1073/pnas.1414640112

Liaw, A., Wiener, M., 2002. Classification and regression by randomForest. *R News* 3, 18–22.

Loon, A.F.V., Gleeson, T., Clark, J., Dijk, A.I.J.V., Stahl, K., Hannaford, J., Baldassarre, G.D., Teuling, A.J., Tallaksen, L.M., Uijlenhoet, R., Hannah, D.M., Sheffield, J., Svoboda, M., Verbeiren, B., Wagener, T., Rangecroft, S., Wanders, N., Lanen, H.A.J.V., 2016. Drought in the anthropocene. *Nature Geoscience* 9, 89–91. doi:10.1038/ngeo2646

Lubchenco, J., Cerny-Chipman, E.B., Reimer, J.N., Levin, S.A., 2016. The right incentives enable ocean sustainability successes and provide hope for the future. *Proceedings of the National Academy of Sciences of the United States of America* 113, 14507–14514. doi:10.1073/pnas.1604982113

Mao, J., Jardine, S.L., 2020. Market impacts of a toxic algae event: The case of california dungeness crab. *Marine Resource Economics* 35, 1–20. doi:10.1086/707643

Maxwell, S.M., Hazen, E.L., Lewison, R.L., Dunn, D.C., Bailey, H., Bograd, S.J., Briscoe, D.K., Fossette, S., Hobday, A.J., Bennett, M., Benson, S., Caldwell, M.R., Costa, D.P., Dewar, H., Eguchi, T., Hazen, L., Kohin, S., Sippel, T., Crowder, L.B., 2015. Dynamic ocean management: Defining and conceptualizing real-time management of the ocean. *Marine Policy* 58, 42–50. doi:10.1016/j.marpol.2015.03.014

McCabe, R.M., Hickey, B.M., Kudela, R.M., Lefebvre, K.A., Adams, N.G., Bill, B.D., Gulland, F.M.D., Thomson, R.E., Cochlan, W.P., Trainer, V.L., 2016. An unprecedented coastwide toxic algal bloom linked to anomalous ocean conditions. *Geophysical Research Letters* 43, 10, 366–10, 376. doi:10.1002/2016GL070023

Mcginnis, M.D., Ostrom, E., 2014. Social-ecological system framework : Initial changes and continuing challenges. *Ecology and Society* 19.

Mendo, T., Smout, S., Photopoulou, T., James, M., 2019. Identifying fishing grounds from vessel tracks: Model-based inference for small scale fisheries. *Royal Society Open Science* 6. doi:10.1098/rsos.191161

Moore, K.M., Allison, E.H., Dreyer, S.J., Ekstrom, J.A., Jardine, S.L., Klinger, T., Moore, S.K., Norman, K.C., 2020. Harmful algal blooms: Identifying effective adaptive actions used in fishery-dependent communities in response to a protracted event. *Frontiers in Marine Science* 6, 803.

Moore, S.K., Dreyer, S.J., Ekstrom, J.A., Moore, K., Norman, K., Klinger, T., Allison, E.H., Jardine, S.L., 2020. Harmful algal blooms and coastal communities: Socioeconomic impacts and actions taken to cope with the 2015 u.s. West coast domoic acid event. *Harmful Algae* 96, 101799. doi:10.1016/j.hal.2020.101799

O’Farrell, S., Chollett, I., Sanchirico, J.N., Perruso, L., 2019. Classifying fishing behavioral diversity using high-frequency movement data. *Proceedings of the National Academy of Sciences* 116, 16811–16816. doi:10.1073/pnas.1906766116

O’Farrell, S., Sanchirico, J.N., Spiegel, O., Depalle, M., Haynie, A.C., Murawski, S.A., Perruso, L., Strelcheck, A., 2019. Disturbance modifies payoffs in the explore-exploit trade-off. *Nature Communications* 10, 1–9. doi:10.1038/s41467-019-11106-y

Oken, K.L., Holland, D.S., Andr´, A., Punt, A.E., 2021. The effects of population synchrony, life history, and access constraints on benefits from fishing portfolios.

691 Oliver, E.C.J., Donat, M.G., Burrows, M.T., Moore, P.J., Smale, D.A.,
 692 Alexander, L.V., Benthuisen, J.A., Feng, M., Gupta, A.S., Hobday, A.J., Hol-
 693 brook, N.J., Perkins-Kirkpatrick, S.E., Scannell, H.A., Straub, S.C., Wernberg,
 694 T., 2018. Longer and more frequent marine heatwaves over the past century.
 695 *Nature Communications* 9, 1–12. doi:10.1038/s41467-018-03732-9
 696 Pellowe, K.E., Leslie, H.M., 2019. Heterogeneity among clam harvesters in
 697 northwest mexico shapes individual adaptive capacity. *Ecology and Society* 24.
 698 doi:10.5751/ES-11297-240425
 699 Pfeiffer, L., Gratz, T., 2016. The effect of rights-based fisheries management
 700 on risk taking and fishing safety. *Proceedings of the National Academy of Sciences*
 701 of the United States of America 113, 2615–2620. doi:10.1073/pnas.1509456113
 702 Rasmuson, L.K., 2013. The biology, ecology and fishery of the dungeness
 703 crab, cancer magister, 1st ed, *Advances in Marine Biology*. Elsevier Ltd.
 704 doi:10.1016/B978-0-12-410498-3.00003-3
 705 R Core Team, 2021. R: A language and environment for statistical computing.
 706 R Foundation for Statistical Computing, Vienna, Austria.
 707 Renner, M., Kuletz, K.J., 2015. A spatial-seasonal analysis of the oiling risk
 708 from shipping traffic to seabirds in the aleutian archipelago. *Marine Pollution*
 709 *Bulletin* 101, 127–136. doi:10.1016/j.marpolbul.2015.11.007
 710 Richerson, K., Punt, A.E., Holland, D.S., 2020. Nearly a half century of
 711 high but sustainable exploitation in the dungeness crab (cancer magister) fishery.
 712 *Fisheries Research* 226, 105528. doi:10.1016/j.fishres.2020.105528
 713 Ritzman, J., Brodbeck, A., Brostrom, S., McGrew, S., Dreyer, S., Klinger,
 714 T., Moore, S.K., 2018. Economic and sociocultural impacts of fisheries closures
 715 in two fishing-dependent communities following the massive 2015 u.s. West coast
 716 harmful algal bloom. *Harmful Algae* 80, 35–45. doi:10.1016/j.hal.2018.09.002
 717 Russell, S.M., Oostenburg, M.V., Vizek, A., 2018. Adapting to catch shares:
 718 Perspectives of west coast groundfish trawl participants. *Coastal Management*
 719 46, 603–620. doi:10.1080/08920753.2018.1522491
 720 Salas, S., Gaertner, D., 2004. The behavioural dynamics of fishers: Man-
 721 agement implications. *Fish and Fisheries* 5, 153–167. doi:10.1111/j.1467-
 722 2979.2004.00146.x
 723 Santora, J.A., Mantua, N.J., Schroeder, I.D., Field, J.C., Hazen, E.L., Bograd,
 724 S.J., Sydeman, W.J., Wells, B.K., Calambokidis, J., Saez, L., Lawson, D., Forney,
 725 K.A., 2020. Habitat compression and ecosystem shifts as potential links between
 726 marine heatwave and record whale entanglements. *Nature Communications* 2020
 727 11:1 11, 1–12. doi:10.1038/s41467-019-14215-w
 728 Smale, D.A., Wernberg, T., Oliver, E.C.J., Thomsen, M., Harvey, B.P.,
 729 Straub, S.C., Burrows, M.T., Alexander, L.V., Benthuisen, J.A., Donat, M.G.,
 730 Feng, M., Hobday, A.J., Holbrook, N.J., Perkins-Kirkpatrick, S.E., Scannell,
 731 H.A., Gupta, A.S., Payne, B.L., Moore, P.J., 2019. Marine heatwaves threaten
 732 global biodiversity and the provision of ecosystem services. *Nature Climate*
 733 *Change* 9, 306–312. doi:10.1038/s41558-019-0412-1
 734 Smith, C.L., McKelvey, R., 1986. Specialist and generalist: Roles for coping
 735 with variability. *North American Journal of Fisheries Management* 6, 88–99.
 736 doi:10.1577/1548-8659(1986)6<88:sag>2.0.co;2

737 Suryan, R.M., Arimitsu, M.L., Coletti, H.A., Hopcroft, R.R., Lindeberg,
 738 M.R., Barbeaux, S.J., Batten, S.D., Burt, W.J., Bishop, M.A., Bodkin, J.L.,
 739 Brenner, R., Campbell, R.W., Cushing, D.A., Danielson, S.L., Dorn, M.W.,
 740 Drummond, B., Esler, D., Gelatt, T., Hanselman, D.H., Hatch, S.A., Haught,
 741 S., Holderied, K., Iken, K., Irons, D.B., Kettle, A.B., Kimmel, D.G., Konar, B.,
 742 Kuletz, K.J., Laurel, B.J., Maniscalco, J.M., Matkin, C., McKinstry, C.A.E.,
 743 Monson, D.H., Moran, J.R., Olsen, D., Palsson, W.A., Pegau, W.S., Piatt, J.F.,
 744 Rogers, L.A., Rojek, N.A., Schaefer, A., Spies, I.B., Straley, J.M., Strom, S.L.,
 745 Sweeney, K.L., Szymkowiak, M., Weitzman, B.P., Yasumiishi, E.M., Zador, S.G.,
 746 2021. Ecosystem response persists after a prolonged marine heatwave. *Scientific*
 747 *Reports* 11, 1–17. doi:10.1038/s41598-021-83818-5
 748 Townhill, B.L., Tinker, J., Jones, M., Pitois, S., Creach, V., Simpson, S.D.,
 749 Dye, S., Bear, E., Pinnegar, J.K., 2018. Harmful algal blooms and climate
 750 change: Exploring future distribution changes. *ICES Journal of Marine Science*
 751 75, 1882–1893. doi:10.1093/icesjms/fsy113
 752 von Biela, V., Arimitsu, M.L., Piatt, J.F., Heflin, B.M., Schoen, S., 2019.
 753 Extreme reduction in condition of a key forage fish during the pacific marine
 754 heatwave of 2014–2016. *Marine Ecology Progress Series* 613, 171–182.
 755 Watson, J.T., Haynie, A.C., 2016. Using vessel monitoring system data to
 756 identify and characterize trips made by fishing vessels in the united states north
 757 pacific. *PLoS ONE* 11, 1–20. doi:10.1371/journal.pone.0165173
 758 Wilson, J.R., Lomonico, S., Bradley, D., Sievanen, L., Dempsey, T., Bell,
 759 M., McAfee, S., Costello, C., Szuwalski, C., McGonigal, H., Fitzgerald, S.,
 760 Gleason, M., 2018. Adaptive comanagement to achieve climate-ready fisheries.
 761 *Conservation Letters* 11, 1–7. doi:10.1111/conl.12452
 762 Young, T., Fuller, E.C., Provost, M.M., Coleman, K.E., Martin, K.S., McCay,
 763 B.J., Pinsky, M.L., 2019. Adaptation strategies of coastal fishing communi-
 764 ties as species shift poleward. *ICES Journal of Marine Science* 76, 93–103.
 765 doi:10.1093/icesjms/fsy140