

# Mobility and flexibility enable resilience of human harvesters to environmental perturbation

## Abstract

Characteristics of natural resources that enable sustainable management are often more fully understood than the adaptive behaviors of human harvesters in those same systems. Given increasing environmental variability due to climate change, it is especially critical to understand how human harvesters may respond to environmental perturbation. In this study, we identify characteristics that promoted resilience of one the most valuable fisheries on the west coast of the United States to a record marine heatwave. Using movement telemetry linked to fishery landings records from more than 500 fishing vessels, encompassing 2.2 million geolocations and more than \$2 billion in revenue, we found that vessels employed two, non-mutually exclusive strategies to cope with the anomalous environmental and management conditions imposed by the heatwave: increasing spatial mobility and diversifying fishery participation. The combination of these strategies appeared to be the most adaptive, as it produced the greatest increase in profits. In contrast, participants that specialized in a single fishery and concentrated fishing effort in small spatial areas experienced the greatest losses driven by the heatwave. Our data-driven approach reveals behaviors that can be promoted to improve the adaptive capacity of human harvesters in an era of unprecedented environmental perturbation.

*Key words:* climate change adaptation | environmental perturbation | marine heatwave | fisheries dynamics

## 1. Introduction

Sustainability in social-ecological systems—the continued provision of human and ecological benefits from healthy ecosystems (Leslie et al., 2015)—requires resilience to environmental perturbations. Often, though, people respond to environmental change in diverse and complex ways. Just as multiple species occupying similar ecological niches may react differently to physical changes in their environments (Elmqvist et al., 2003), human actors in a social-ecological system can exhibit diverse behaviors within the constraints imposed by the governance system (McGinnis and Ostrom, 2014). Groups of resource users with distinct livelihood portfolios, available capital, or spatial patterns of resource extraction will not respond the same way to environmental or management changes (Young et al., 2019). In response to change, some users might stick to established knowledge and reliable spatial patterns of exploitation, while others might employ more exploratory strategies that carry higher potential

upsides but also higher risks and costs (Cohen et al., 2007). Understanding the adaptive behaviors of resource users is all the more important given the increasing prevalence of extreme climate events attributable to climate change (Abatzoglou et al., 2019; Cook et al., 2018; Oliver et al., 2018; Townhill et al., 2018), but empirical evidence making the link between climate extremes and contemporaneous human adaptation remains lacking.

Fisheries are a prominent example of a social-ecological system where complex links between resource user (harvester) behavior and natural resource dynamics drive sustainability (Branch et al., 2006). Fisheries represent the last large-scale wild harvest of food on Earth, but also one of the most traditional livelihoods in human history. Difficulties in achieving sustainability in fisheries have often been linked to an inadequate understanding of harvester dynamics (Fulton et al., 2011; Hilborn, 1985). Differences in fisher behaviors, both within and across fisheries, can affect the stability and sustainability of fish populations (Fryxell et al., 2017; Salas and Gaertner, 2004), of other species—for instance, endangered marine mammals or seabirds—and of the fishery itself (Gladics et al., 2017; Hamilton and Baker, 2019).

Additionally, different behavioral segments of fishing fleets may respond in different ways to management measures, or may be differentially vulnerable to environmental perturbations (Salas and Gaertner, 2004). For example, O’Farrell, Sanchirico, et al. (2019) found that more exploratory fishing vessels—those that, on average, traveled further and more often traversed new fishing grounds—were better able to cope with an extended spatial closure. Heterogeneous behavioral responses of fishers can be difficult to study, despite their potential impact on resource dynamics. Partly, this is due to a lack of detailed spatial and economic information on harvester behavior. However, recent years have seen a rise in availability of these types of fishery data, paired with methods to extract behavioral insights from them (Joo et al., 2015; Mendo et al., 2019; Watson and Haynie, 2016). In the following, we apply a range of data-driven methods to ask: how did human harvesters cope with and adapt to a major environmental perturbation in the most valuable fishery on the U.S. west coast?

The Dungeness crab fishery on the west coast of the United States often obtains in excess of \$200 million in revenue from over 1,000 participating vessels each year (Rasmuson, 2013; Richerson et al., 2020). It is a fishery that is central both ecologically and economically (Fuller et al., 2017) to the west coast social-ecological system, making it at once a cornerstone of fishers’ portfolios and a source of complexity in fisheries governance (Holland and Leonard, 2020; Holland et al., 2017). Dungeness crab populations appear able to withstand immense fishing pressure, and although crab abundance can fluctuate markedly from year to year, long term abundance has been relatively stable for more than a half century (Richerson et al., 2020). Harvester characteristics vary widely for an industrialized fishery—Dungeness crab vessels have a large range of sizes (in our data, 21 to 103 feet), and operate out of both large urban and small rural fishing ports across the U.S. west coast.

Recent environmental shocks have challenged the social and economic sustainability of the Dungeness crab fishery. In 2014-2016, a record marine heatwave

(MHW) led to a harmful algal bloom of unprecedented scale (McCabe et al., 2016), causing toxin levels in Dungeness crabs to reach levels dangerous for human consumption and correspondingly lengthy delays in large regions of the coast in the 2015-16 and 2016-17 Dungeness fishing seasons. Concurrently, the MHW caused shoreward compression of the preferred feeding habitat of large whales, contributing to a rise in entanglements of whales in Dungeness crab fishing gear and increasing risk of fishery closure due to marine mammal interactions, effects that continued to directly affect fishery closures through the 2017-18 Dungeness crab season (Feist et al., 2021; Santora et al., 2020). During this period, Dungeness crab fishers had to contend with significant ecological changes and the management measures those changes precipitated. Like with climate extremes in other systems (Loon et al., 2016), the effects of this MHW were complex, reverberated through the social-ecological system, and persisted for years after the anomalous warming dissipated (Fisher et al., 2021; Smale et al., 2019; Suryan et al., 2021). While much recent literature is dedicated to examination of biophysical and ecological impacts of the MHW (Cavole et al., 2016; McCabe et al., 2016; von Biela et al., 2019), to date far less attention has been given to exploring how social systems cope and change with these perturbations (Fisher et al., 2021; Jardine et al., 2020; K. M. Moore et al., 2020).

In this study, we compare the adaptive responses of behavioral groups within the Dungeness crab fishery to the multi-year MHW that directly affected the 2015-16 through 2017-18 Dungeness crab seasons. The 2015-16 Dungeness crab season was the first season to be significantly delayed as a direct result of ecosystem changes, a trend that continued through the 2017-18 season. While previous work has investigated economic impacts (Holland and Leonard, 2020; Jardine et al., 2020; Mao and Jardine, 2020) and changes in fishery participation due to the MHW-associated harmful algal bloom (Fisher et al., 2021), here we explicitly investigate and quantify fishers' adaptive spatial behaviors in response to the MHW more broadly and for the full three-year period over which the MHW impacts manifested. Using a 10-year time-series of more than 2 million satellite-derived fishing vessel location records, linked to fishery revenue and landings data, we derive quantitative behavioral metrics describing space use and mobility of Dungeness crab vessels, then organize these behaviors into characteristic behavioral groups. We explore the overlap of spatial behaviors with profitability, fishing season length, and revenue diversity. We track these behavioral groups over time, and identify key behavioral metrics that promoted adaptation during and after the MHW. This analysis therefore offers insights into the types of adaptive behaviors that may promote sustainable outcomes in other commercial fisheries and perhaps in social-ecological systems more broadly.

## 2. Materials and Methods

### 2.1. Data sources

We used satellite-based Vessel Monitoring System (VMS) data and port level fishery landings data to define most of the behavioral variables used in

the study. The VMS database is maintained by the National Marine Fisheries Service’s Office of Law Enforcement, and records the positions of vessels at approximately one hour intervals. Similar VMS data has been used in other studies of fishery spatial dynamics (Feist et al., 2021; Joo et al., 2015; O’Farrell, Chollett, et al., 2019; Watson and Haynie, 2016). A subset of the vessels that participate in the Dungeness crab fishery are equipped with VMS transponders (primarily vessels that also participate in the west coast groundfish fishery, where VMS transponders are mandatory). This subset varies between 19 and 26 percent of all vessels recording landings for Dungeness crab between the 2008-2009 and 2018-2019 seasons, representing between 10 and 57 percent of all Dungeness crab landings by weight, and between 15 and 42 percent of Dungeness revenue, depending on the year and state (California, Oregon, or Washington). Oregon has the highest relative VMS representation, followed by California, then Washington.

Fish ticket information was obtained through the Pacific Fisheries Information Network (PacFIN). These data represent 1949 vessels targeting Dungeness crab in California, across more than 300,000 fish tickets (i.e., fishing trips). Fishing trips were defined as targeting Dungeness crab if the total landings of Dungeness on the individual fish ticket were at least 10 percent greater than the landed weight of the next greatest species.

We joined the fish ticket data to the VMS data through unique vessel identification numbers and timestamps. VMS geolocations comprising a fishing trip were defined as all of the geolocations between a landed fish ticket and the one immediately preceding it (i.e., the previous ticket landed by the same vessel). After joining the VMS and fish ticket data, we removed the small number of trips in which the final VMS data point for a trip was greater than 50km from the port of landing recorded on the ticket, reasoning that these are unreliable records. Finally, we removed VMS records from vessels sitting idle in port. To do so, we truncated all but the first and last VMS records for each trip that fell within a small buffer zone (1.5 to 3 km) around each port of landing and with an average calculated speed of less than 0.75 m/s.

Dungeness crab fishing seasons on the west coast typically begin in the middle of November (for Central California) or beginning of December (for Northern California, Oregon, and Washington), but can be variable in their starting dates, depending on state (California, Oregon, or Washington), harmful algal bloom closures, price and market conditions, crab condition and meat quality, and potential interactions with protected species like humpback whales. Therefore, we used a data-driven approach to define the start date for each crab season in each of the 20 fishing port groups on the west coast. Port groups are defined by PacFIN and include clusters of small, neighboring fishing ports. For each port group in each season, we found the date after October 31 of each season that the total Dungeness crab landings into that port reached 1 percent of the eventual, season-long landings. This approach identifies the realized start date of the crab fishery in each portion of the coast in each year.

The maximum length of a Dungeness fishing trip was defined as seven days. That is, if there was a gap of greater than seven days between consecutive tickets,

155 the VMS geolocations greater than seven days prior to the landed ticket were  
156 discarded. The final dataset comprises a clean record of geolocations associated  
157 with each Dungeness crab fishing trip.

158 The only other data source used in the calculation of behavioral metrics is a  
159 measure of average daily wind speeds, from AVHRR Pathfinder satellite-derived  
160 measurements (<https://data.nodc.noaa.gov>; <https://doi.org/10.7289/v52j68xx>).  
161 The data are modelled daily on a 0.04 degree grid (approximately 5 km at the  
162 equator) and are available from 1981-present.

## 163 *2.2. Construction of Fishing Behavioral Metrics*

164 Fishing behavioral metrics were calculated from fish ticket, VMS, and wind  
165 speed data. Our choice of behavioral variables to calculate was driven by previous  
166 evidence of the importance of each variable in describing fisher behavioral patterns  
167 (Fuller et al., 2017; Kasperski and Holland, 2013; O’Farrell, Chollett, et al., 2019;  
168 O’Farrell, Sanchirico, et al., 2019; Pfeiffer and Gratz, 2016). Each of the fisher  
169 behavioral variables described one characteristic of a vessel’s apparent behavior  
170 over the course of a fishing season—a vessel-season. Cluster analysis (see next  
171 section) was performed on these vessel-seasons, and individual vessels could be  
172 clustered into different behavioral groups in different seasons. To determine  
173 whether a vessel would be included in the analysis, we calculated the total  
174 Dungeness crab revenue for each vessel in each season from 2008-09 to 2018-19.  
175 The 5th percentile for annual Dungeness revenue per vessel was \$5828. We  
176 retained all vessel-seasons with greater than \$5828 in revenue in any season (i.e.,  
177 we retain the top 95 percent of all vessel-seasons as measured by revenue).

178 Our behavioral metrics fall into five general categories: port use, fishing  
179 trip characteristics, participation in other fisheries, risk-taking behavior, and  
180 exploration and mobility (see Table A.1 for full technical definitions of metrics).  
181 Port use metrics include the number of ports visited per fishing trip, ports visited  
182 per month, diversity of port use (calculated as a Shannon diversity index on the  
183 proportions of trips landed in each port), and the total number of ports visited  
184 across the entire season. The trip metrics are the mean and standard deviation  
185 of trip distance (in km) and duration (in days). Vessel size has been used as  
186 a proxy for fleet segments in other studies (Jardine et al., 2020). We did not  
187 include vessel size as a metric, since vessel size alone is not a behavioral variable,  
188 but we explored its relationship to included metrics as a point of comparison  
189 (Figs. A.5, A.6).

190 Fishery participation metrics include season length, proportion of revenue  
191 and fish tickets from other (non-Dungeness) fisheries, and revenue diversity. The  
192 Dungeness fishery operates as a derby, where the majority of the landings and  
193 profits are obtained in the first few months of each season (Fig. A.4). Our  
194 season length metric captures this phenomenon and indicates the day of the crab  
195 season that each vessel reaches 90 percent of its cumulative landings for that  
196 season. To calculate the proportion of revenue and tickets from other fisheries,  
197 and revenue diversity, we use a version of the fish ticket data that includes all  
198 fishery targets (not just Dungeness crab). Using these tickets, the proportion  
199 of non-Dungeness revenue is calculated, as well as the proportion of fish tickets

submitted by that vessel with a target other than Dungeness crab. Revenue diversity for each vessel-season is an inverse Simpson index calculated on the proportion of revenue obtained from each species in a vessel’s fishing portfolio.

Risk-taking behavior is modelled after the definition in Pfeiffer and Gratz (2016), who also studied west-coast fisheries, as propensity to fish in high-wind conditions. Using the Pathfinder winds data, we extracted the wind speed at each VMS location, then calculated the 95th percentile of wind speed experienced by each vessel on each trip. Finally, the risk-taking metric was defined as the proportion of trips in a season where the 95th percentile of experienced wind speed was greater than 7.5 m/s (Pfeiffer and Gratz, 2016).

Exploration and mobility were measured with home range and location choice entropy, adopting the definitions in O’Farrell et al. (2019). Home range was calculated as the area of the minimum convex polygon encompassing all VMS locations in a vessel-season, after removing the five percent of locations that were the furthest from other points (i.e., spatial outliers). Location choice entropy measures the propensity of vessels to explore new locations versus returning to the same locations, and is calculated cumulatively across each vessel’s fishing season (O’Farrell, Sanchirico, et al., 2019). Spatial locations were defined as individual cells on a 5x5km grid. As a season progresses, entropy increases as vessels explore novel locations and decreases as the same locations are revisited repeatedly. The season-long metric for exploration for each vessel is defined as the 90th percentile of maximum location choice entropy in that season.

Definitions of all metrics used in the clustering analysis are provided in the Appendix.

### 2.3. Cluster Analysis

All metrics were checked for collinearity, and thinned such that no two metrics had a Pearson correlation greater than 0.7. This thinning removed mean and standard deviation of trip distance, total number of visited ports, and proportion of non-Dungeness tickets from the analysis. The remaining 11 metrics were scaled to range from zero to one by dividing each metric by its maximum value (across all seasons). Clustering was performed using Euclidean distances and Ward aggregation that minimizes total within-cluster variance. The number of clusters was determined using the Nbclust package in R (Charrad et al., 2014), which calculates 22 clustering indices before recommending an optimal number of clusters via majority vote amongst indices. Adopting the optimal clusters defined by NbClust, we visualized results graphically using principal component analysis. After vessel-seasons were assigned to groups, we tested for differences between groups along specific behavioral metrics using Tukey’s HSD.

The importance of individual metrics in discriminating between clusters was calculated using random forest analysis, utilizing the randomForest package in R (Liaw and Wiener, 2002). Random forests were grown on subsamples of the data to classify vessel-seasons according to their defined clusters from the previous step. Then, these random forests were used to predict withheld data. Variable importance was defined as the increase in the rate of mis-classification of vessel-seasons into clusters when the particular variable was randomly permuted.

245 *2.4. Dungeness Fishing Profitability*

246 We used fish ticket data to assess the per-trip, per-week, and per-season  
 247 landings and revenue of vessels in each fisher behavioral group over time. Addi-  
 248 tionally, we modeled fishing costs following the approach of Dewees et al. (2004)  
 249 to assign an estimated profit to each fishing trip. The cost of a fishing trip  $C_t$  is  
 250 assumed to be a function of fuel  $C_f$  and bait costs, and the costs of labor (i.e.,  
 251 crew)  $C_c$ :

$$C_t = C_f + C_c$$

252 Fuel and bait cost is a function of vessel size  $L$  and number of days fished  $d$ , as  
 253 well as trip year  $y$  to adjust for an assumed 2 percent inflation rate.

$$C_f = f(L, d, y)$$

254 Crew cost is a function of vessel size (because larger vessels require more crew  
 255 members) and total trip revenue  $R$  (since crew members receive a proportion of  
 256 revenue).

$$C_c = f(L, R)$$

257 The above cost relationships were parameterized using data from Dewees  
 258 et al. (2004), who administered a survey to 243 Dungeness crab fishers and  
 259 compiled estimates of fishing costs by vessel size. The survey estimated costs  
 260 associated with bait, fuel, and labor (crew) for small (less than 9.1m), medium  
 261 (9.1-15.2 m) and large (greater than 15.2 m) fishing vessels. Using the means  
 262 and standard deviations of these costs reported in Dewees et al. (2004), we  
 263 simulated 10,000 trip costs for vessels ranging in length from 6.4 to 31.4 m,  
 264 which is the range of vessel sizes in our data. Then, linear relationships between  
 265 vessel size and both types of costs were estimated with simple linear regression.  
 266 The resulting relationships,

$$C_f = d(150 + 3.5L) * 1.02^{y-2004}$$

$$C_c = R(0.17 + 0.0018L)$$

267 were used to deterministically assign a cost to each Dungeness fishing trip in  
 268 our data. From there, a profit for each trip could be estimated by subtracting  
 269 costs from revenue. Using trip-level profits, we calculated mean profits per week—  
 270 across seasons—for vessels in each behavioral group, as well as season-long  
 271 profits.

272 Using the fish ticket revenue data, we also calculated total revenue from all  
 273 non-Dungeness fisheries for each vessel-season in the analysis. We constrained  
 274 the calculation of non-Dungeness revenue to only those fishing trips that occurred  
 275 within each vessel's apparent Dungeness season (that is, within the time period  
 276 where the vessel was also landing Dungeness crab).

## 277 2.5. Adaptation to the Marine Heatwave

278 Using the results of cluster analyses, we compared key characteristics of  
279 behavioral groups in MHW versus non-MHW crab seasons. We defined the MHW  
280 as encompassing the crab fishing seasons from 2015-16 to 2017-18. Although  
281 there is evidence that the MHW began affecting west coast ecosystems as early  
282 as late 2014 (Cavole et al., 2016; McCabe et al., 2016), the 2015-16 Dungeness  
283 crab season was the first to be significantly delayed as a direct result of ecosystem  
284 changes (Jardine et al., 2020), a trend that continued through the 2017-18 season.

285 Adopting this definition of the MHW period, we compared mean Dungeness  
286 profit, non-Dungeness revenue (i.e., external fishery revenue), and home range  
287 size over time among behavioral groups to assess potential adaptive strategies.  
288 For each of these three comparisons, we performed a two-way ANOVA to test for  
289 significant differences in mean profits, revenue, and home range by behavioral  
290 group and period (non-MHW or MHW).

291 All analyses in the study were performed in R (R Core Team, 2021).

## 292 3. Results

### 293 3.1. Describing Fisher Behavior

294 The combined vessel telemetry and fisheries landings dataset captured the  
295 behaviors of 596 different vessels spanning 11 fishing seasons (2008-2019), with  
296 ~2.2 million satellite-derived Vessel Monitoring System (VMS) geolocations, and  
297 315,000 fishery landing records. Using these combined data, we analyzed 11  
298 behavioral variables in five general behavioral categories: fishing port use, fishing  
299 trip characteristics, participation in other fisheries, risk-taking behavior, and  
300 exploration and mobility (definitions of all metrics are provided in Table A.1).

301 The 3391 vessel-seasons in our data fell into four behavioral cluster groups  
302 (Figs. 1a, A.1). The most important discriminating variables driving the  
303 clustering according to random forest analysis were proportion of revenue from  
304 non-Dungeness crab fisheries, followed by diversity of port use, revenue diversity,  
305 and mean trip duration (Fig. 1b). These analyses suggest that the behavior of  
306 the four groups can be conceptualized as varying along two major axes (Fig. 1c):  
307 (1) spatial mobility (principal component 1 in Fig. 1a) and (2) propensity to  
308 fish in non-Dungeness crab fisheries (fishery flexibility, principal component 2 in  
309 Fig. 1a).

310 Vessels with higher spatial mobility, which we term Roving groups, move  
311 between ports throughout a fishing season and have large fishing ranges, while  
312 those with lower mobility—Local groups—show greater fidelity to a single  
313 port. Vessels with greater fishery flexibility, deemed Generalist groups, have  
314 high revenue diversity and derive a relatively greater portion of their total  
315 fishery revenue from fisheries other than Dungeness crab. Vessels exhibiting  
316 less flexibility—Specialists—concentrate fishing effort within the Dungeness crab  
317 fishery. A vessel-season is therefore defined as either Roving or Local, and  
318 either Specialist or Generalist. As an example, for crab vessels fishing out of  
319 Newport, Oregon, Local Specialists have the smallest fishing grounds, followed by



Category	Variable	Definition
Port Use	Ports per Trip	Average ports visited per trip
	Ports per Month	Number of ports visited per month
	Port Diversity	Inverse Simpson diversity index of port use across the entire season
	Total Ports*	Total number of ports visited across the entire season
Trip Length	Mean Trip Distance*	Mean distance per fishing trip
	Mean Trip Duration	Mean number of days per fishing trip
	SD Trip Distance*	Standard deviation of distance traveled per trip
	SD Trip Duration	Standard deviation of days per fishing trip
Participation in Other Fisheries	Season Length	Day-of-season on which fisher reached 90% of eventual, cumulative catch
	Proportion Non-Dungeness Revenue	Proportion of revenue from non-Dungeness crab fisheries
	Proportion Non-Dungeness Tickets*	Proportion of all fish tickets from non-Dungeness crab fisheries
	Revenue Diversity	Inverse Simpson diversity index of revenue by fished species
Risk-Taking	Risk Taking	Propensity to fish in high winds. Proportion of trip pursued where the 95% quantile of wind speed was greater than 7.5 m/s
Exploration & Mobility	Location Entropy	Cumulative choice entropy, measuring how likely a vessel is to fish in new versus past locations. The metric used is the 90th percentile of maximum choice entropy per vessel per season
	Home Range Size	Home range defined as the area of the convex hull surrounding all of a vessel's VMS pings during the season, excluding the top 5% spatial outliers
Vessel Size	Vessel Length in Feet	Registered length of the fishing vessel

Table 1: Fisher behavioral and demographic variables derived and used in the clustering and random forest analyses. Variables with asterisks were removed from the final clustering analysis due to high collinearity with other variables.

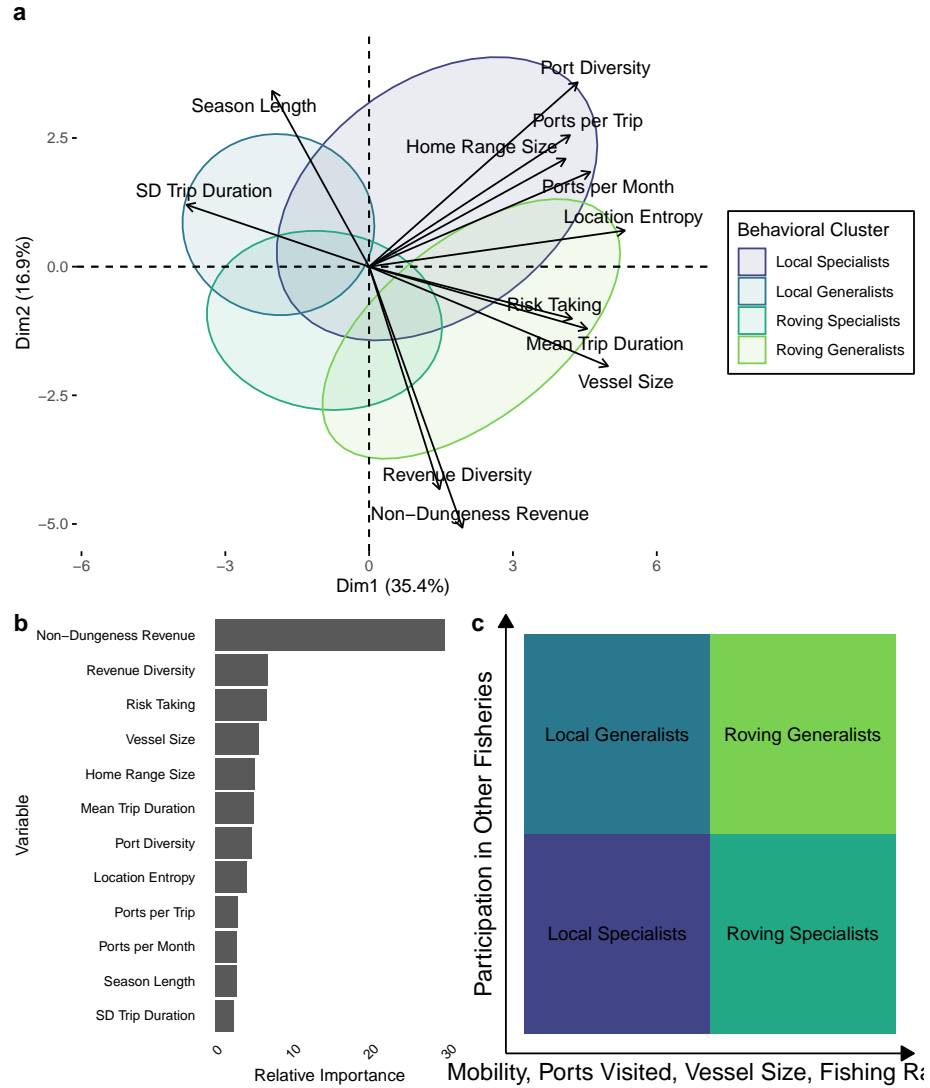


Figure 1: Data-driven formation of fishing behavioral groups. (a) Principal component analysis of vessel-seasons. Clusters of vessel-seasons, which determine behavioral groups, are enclosed by ellipses. Arrows represent the association between metrics in the cluster analysis relative to the placement of vessel-seasons. (b) Ranked importance of top variables used to classify vessel-seasons into behavioral groups, as determined by random forest analysis. (c) Conceptual visualization of the major axes defining behavioral groups.

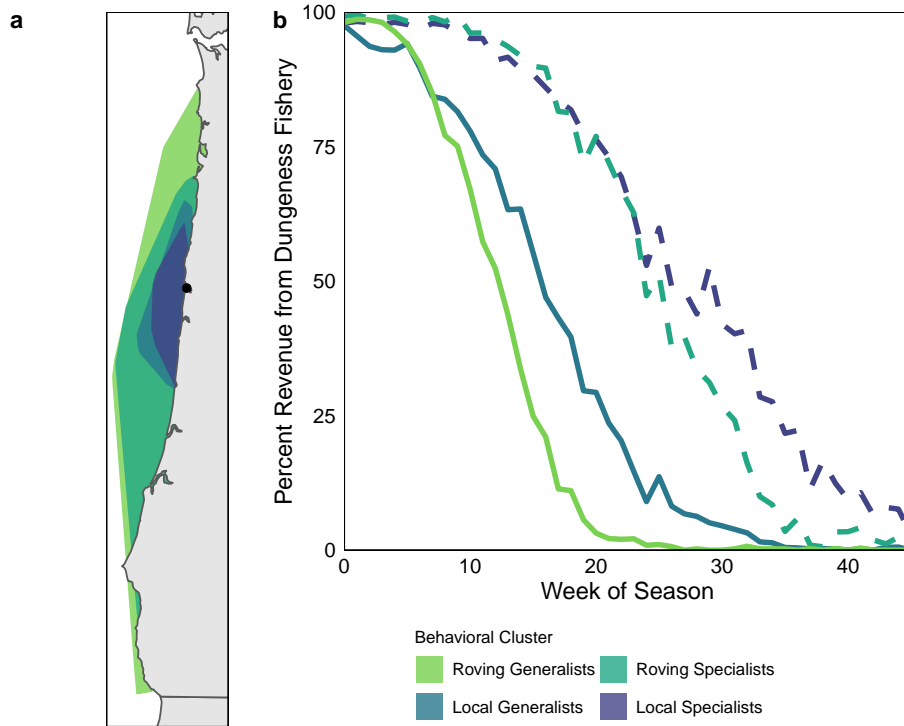


Figure 2: Characteristic patterns in spatial mobility and fishery flexibility across behavioral groups in the west coast Dungeness crab fishery, exemplified by an Oregon port. (a) Fishing footprints of each behavioral group across all seasons for vessels originating from the Port of Newport, Oregon, USA. Shaded polygons are 95 percent convex hulls of all VMS locations for each group. (b) Fishery flexibility, displayed as the mean percent of total weekly revenue obtained from the Dungeness crab fishery (relative to all other fisheries) by vessels in each behavioral group. Weekly revenues are averaged across crab seasons and across all vessels in each group. Generalist groups are represented with solid lines, while Specialist groups are represented with dashed lines.

Local Generalists, Roving Specialists, and Roving Generalists (Fig 2a). Across all vessel-seasons, Generalist vessels have shorter crab fishing seasons, exiting the Dungeness crab fishery earlier to pursue other fishing opportunities, while Specialists continue to garner a large percentage of their weekly landed revenue from Dungeness crab over the course of the season (Fig. 2b).

### 3.2. Behavioral Changes During the Marine Heatwave

The four fishing behavioral groups defined by our cluster analysis responded to the social-ecological disruption of the marine heatwave (MHW) by increasing their dependence on other, non-Dungeness fisheries and expanding their fishing ranges. All groups had higher non-Dungeness fishery revenue during the MHW period than during other seasons, indicating a potential fallback to other fisheries during a period of delays and management disruptions in the crab fishery (Fig.

3)(Fisher et al., 2021; Holland et al., 2020). The 2016-17 and 2017-18 seasons had the highest non-Dungeness crab revenue in the time series (Fig. 3a). The Generalist groups in particular more than doubled their revenues from non-Dungeness fisheries (ANOVA  $p < 0.01$ ; Fig. 3b). The Specialist groups also had greater non-Dungeness revenues during the MHW period, but the differences were not as substantial as for the Generalist groups (Table S2, ANOVA  $p = 0.06$  for Roving Specialists,  $p=0.99$  for Local Specialists).

Some Dungeness fishers also expanded their Dungeness crab fishing grounds during the MHW, particularly the two Roving groups (Fig. 4). Prior to the MHW (2008-15), Roving Generalists had the largest mean home range size at more than 4000 square kilometers (Fig. 4a). Roving Specialists had the second-largest ranges on average (around 2500 square kilometers), while the Local groups had much smaller ranges (less than 1000 square kilometers). In the MHW period from 2015-18, the Roving groups fished significantly larger areas, with the Roving Generalist and Roving Specialist groups averaging more than 5500 and 3500 square kilometers fished, respectively ( $p=0.001$  and  $p\ll 0.001$  for Roving Specialists and Roving Generalists). In contrast, the areas fished for the Local groups did not change significantly (Fig. 4b and Table S3,  $p>0.99$  for both Local groups). For all four groups, within the MHW period, the most pronounced change in mobility occurred during the 2016-17 fishing season.

### 3.3. Profitability of Behavioral Groups during the Marine Heatwave

An open question is whether the adaptive responses we detected and quantified—greater spatial mobility and more flexible fishing—allowed fishers to maintain profits in the face of this major environmental perturbation. Our fishing cost model provides an estimation of Dungeness crab profit (reported revenue minus estimated cost) for every fishing trip in the data (i.e., for those vessels that continued to fish), and allowed us to describe how profits within each behavioral group varied over time (Fig. 5).

For all groups, average revenues and estimated costs both increased during the MHW period (Figs. A.7, A.8), but revenue increases outweighed the increases in cost, resulting in increased estimated profits. Dungeness crab profits for all behavioral groups increased during the MHW, significantly so for Local Generalists ( $p=0.05$ ), Roving Generalists ( $p\ll 0.0001$ ) and Roving Specialists ( $p=0.001$ , Table A.4). The Roving Generalist group saw the largest increase in estimated profits in both raw and percent increase in profits (more than a \$63,000 increase per vessel, a 48 percent increase, on average). Local Specialists experienced the smallest increase in profits of all groups (25 percent) during the MHW, while Roving Specialists and Local Generalists experienced a greater than 40 percent increase. In the season after the dissipation of the MHW, estimated profits declined, particularly for the Roving groups.

## 4. Discussion

The pace and magnitude of environmental change in the Anthropocene demand assessment of how social-ecological systems will respond. Ideally, man-

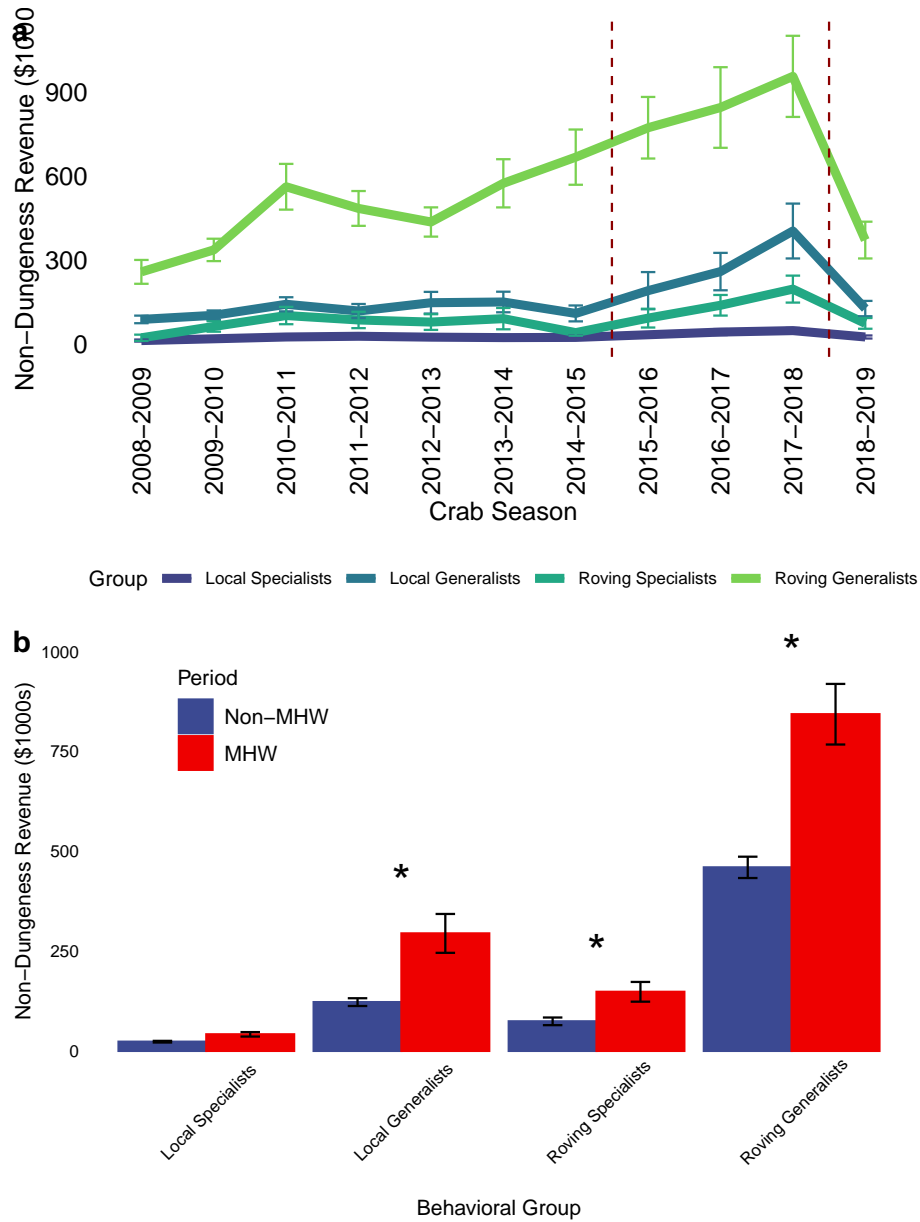


Figure 3: Non-Dungeness revenue for vessels in the analysis. (a) Seasonal mean revenue ( $\pm$  2SE) for vessels in each behavioral group coming from all non-Dungeness fisheries combined. Vertical lines delineate the period of the marine heatwave (MHW). (b) Barplot of mean revenue ( $\pm$  2SE) for vessels in each group during MHW and non-MHW seasons. Stars indicate groups with significantly different non-Dungeness revenue in MHW seasons.

agement approaches can be designed to help humanity adapt by meeting the basic needs of people without compromising ecosystems for future generations (Lubchenco et al., 2016). As one of the last remaining hunter-gatherer activities occurring at scale, commercial fisheries offer an important lens through which to understand human adaptations to novel and extreme conditions. The 2014-2016 marine heatwave on the U.S. west coast stressed the adaptive ability of participants in the highly lucrative Dungeness crab fishery, because an environmental perturbation—the MHW and associated harmful algal bloom and shoreward compression of large whale habitat—led to cascading regulatory actions and market effects (Holland et al., 2020). Our analysis revealed that Dungeness crab fishers that remained in the fishery responded to unprecedented environmental and management changes in multiple ways. Behavioral groups characterized by spatial mobility used expanded fishing grounds in the 2016-17 and 2017-18 seasons to maintain or increase revenues. Similarly, fishers with strategies based around diversified fishing portfolios (Generalists) were able to increase their revenue from other fisheries to bolster their total fishing income. We found that vessels combining greater spatial mobility with higher participation rates in other fisheries were the most profitable, and that these financial benefits were maintained or magnified during the MHW. The behavioral strategies observed in the Dungeness crab fishery suggest that both portfolio and spatial diversification pathways can improve adaptive capacity for human harvesters across industrialized food systems during an era in which the magnitude, frequency, and intensity of environmental perturbations are increasing.

Our work builds on research from the economics (Gordon, 1954; Smith and McKelvey, 1986), evolution (Gallagher et al., 2015), and ecology (Beever et al., 2017) literatures investigating the relative ability of specialists and generalists to cope with environmental change. The cross-disciplinary consensus is that generalists may adapt better to increasingly variable environments. Smith and McKelvey (1986) suggested that specialists and generalists in fisheries use different strategies to cope with variability and uncertainty in income—specialists are efficient and minimize income risk through fishery-specific acumen, while generalists hedge against risk by building diverse portfolios (Kasperski and Holland, 2013; Oken et al., 2021). In a direct ecological analogy, generalist consumers in an ecosystem experiencing novel environmental conditions may be able to gain a competitive advantage over specialists by efficiently switching to alternative prey sources (Beever et al., 2017).

While management dynamics, markets, stochastic resource abundance, and conditions in other fisheries are complicating factors (Holland et al., 2020), the relative performance of specialist versus generalist strategies in the Dungeness crab fishery largely adhere to these existing economic and ecological models. Although some Specialists and Generalists persisted through the MHW period, repeated environmental disruptions in the future that cause further seasonal and spatial restrictions on the Dungeness crab fishery may begin to favor a Generalist, diversified strategy. Within the US west coast context, existing fishery governance systems may constrain this type of generalist adaptation (Kasperski and Holland, 2013; Russell et al., 2018), but there are calls for “climate-ready”

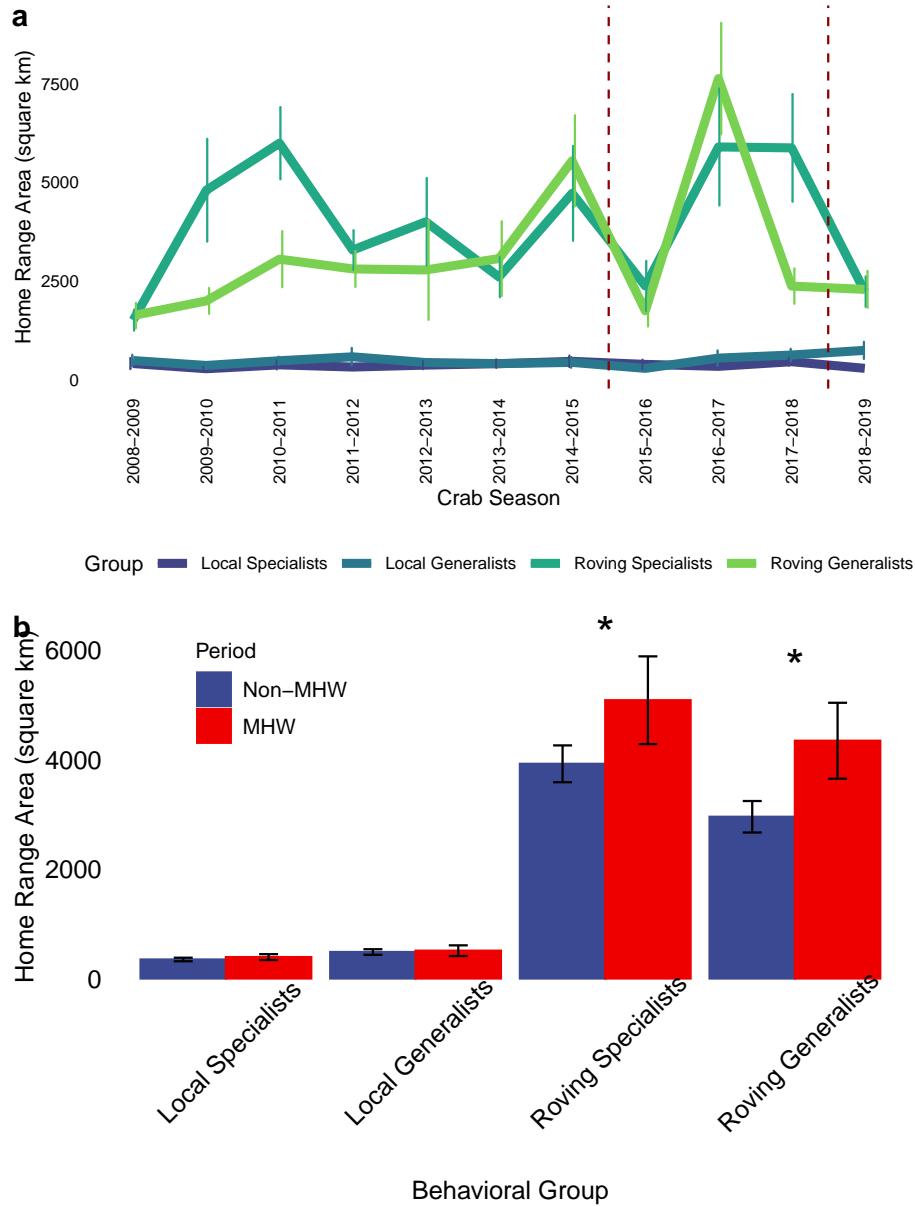


Figure 4: Home range (fishing area) size for vessels in the analysis. (a) Seasonal mean home range area in square kilometers ( $\pm 2SE$ ) for vessels in each behavioral group. Vertical lines delineate the period of the MHW. (b) Barplot of mean home range area ( $\pm 2SE$ ) for vessels in each group during MHW and non-MHW seasons. Stars indicate groups with significantly different home range size during MHW seasons.

421 fisheries that include the flexibility for fishers to move between fisheries (Wilson  
422 et al., 2018). A better understanding of the social, economic, and cultural drivers  
423 of fishers’ decisions to be specialists or generalists is a core component of a  
424 sustainable livelihoods approach to small-scale fisheries management (Allison  
425 and Ellis, 2001; Finkbeiner, 2015). Such an approach can also offer insights for  
426 the design of regulatory approaches that facilitate resilience to environmental  
427 perturbation in larger-scale fisheries and natural resource management contexts  
428 (Salas and Gaertner, 2004).

429 Diversification of fishery revenue was not the only axis of variation associated  
430 with persistence in the face of the MHW. Spatial mobility was also a key  
431 component of the fishing strategies we observed. Following others who have  
432 used recently emerging technologies to understand the sustainability of human  
433 harvester strategies (Brodie and Fragoso, 2020; Frawley et al., 2020; Renner and  
434 Kuletz, 2015), we used satellite data to characterize the spatial behavior of vessels.  
435 Roving groups, whether Specialists or Generalists, were more profitable than their  
436 Local counterparts under all conditions. The benefits of this spatial mobility were  
437 clear during the marine heatwave. We hypothesize that Roving vessels were the  
438 most capable of responding to management actions, market forces, and ecological  
439 factors (e.g., product quantity and quality) that shifted spatially during the  
440 heatwave. The ability of more exploratory fishers to cope during an environmental  
441 disturbance has recently been demonstrated in other commercial fisheries systems  
442 (O’Farrell, Sanchirico, et al., 2019), and our findings confirm that more mobile  
443 vessels performed better during the environmental perturbation. Similar patterns  
444 have been shown among foraging marine mammals, where individual animals  
445 that are more exploratory have greater foraging success during anomalous climate  
446 conditions than more site-faithful conspecifics (Abrahms et al., 2018).

447 Importantly, the nature of the data used in this study means that we studied  
448 the behavior of the ‘survivors’—that is, the fishers who decided or were able to  
449 remain in the Dungeness crab fishery during the MHW period. The MHW acted  
450 as a selective force on Dungeness crab fishery participation. Many Dungeness  
451 crab fishers during the 2016 and 2017 fishery closures chose (or were forced  
452 by circumstance) to not participate in the fishery at all, instead opting to exit  
453 fishing entirely or to re-concentrate all effort in alternative fisheries (Fisher et  
454 al., 2021). Some of the relative success of the Dungeness crab fishers during the  
455 MHW observed in this study, therefore, may be due to reduced competition, as  
456 well as periods of supply shortages and high prices. Although outside the scope  
457 of the current analysis, an important area for further research is to determine  
458 how and why, when faced with an environmental perturbation, fishers choose to  
459 remain or exit a fishery (S. K. Moore et al., 2020).

460 With climate change expected to increase the frequency of extreme environ-  
461 mental perturbations like MHWs (Oliver et al., 2018), established patterns of  
462 natural resource management and human harvester behavior will be challenged.  
463 In our study, following multiple adaptive pathways by both diversifying and  
464 mobilizing appears to be one solution to an extreme environmental event and  
465 rapid management changes in the Dungeness crab fishery. Management mea-  
466 sures that restrict the fishery temporally or spatially—such as spatially-explicit



467 biotoxin-related closures or early termination of the fishing season due to risk  
468 of interactions with protected or bycatch species—will differentially affect dis-  
469 tinct groups of fishers. Single-fishery specialists may thrive when the harvested  
470 resource is stable and productive, but these fishers may struggle to adapt if  
471 management measures restrict fishing season lengths. Likewise, localized fishers  
472 can be successful through intimate knowledge of fishing grounds, but if large-scale  
473 environmental perturbations have spatially-explicit negative effects, fishers with  
474 knowledge of a wider array of fishing grounds and greater mobility will naturally  
475 gain an advantage (O’Farrell, Sanchirico, et al., 2019). Over time, management  
476 context, or failures of management to adapt, can drive changes in the makeup  
477 of fishing fleets as a whole (Frawley et al., 2020). These changes are not in-  
478 herently negative, but in order to maintain the social, economic, and cultural  
479 benefits provided by a fishery, managers should endeavour to anticipate behav-  
480 ioral changes within fleets. More generally, these insights are congruent with  
481 an evolving understanding of adaptation in complex social-ecological systems  
482 (Lubchenco et al., 2016). Because complex systems are an emergent product of  
483 the individual actions of human actors, informed adaptive management requires  
484 an understanding of the drivers of behaviors like those identified in this study  
485 along with well-calibrated and nimble responses within governance systems.

486 For fishers and other human harvesters, future work using mixed methods  
487 from the social sciences like participatory mapping and semi-structured inter-  
488 views (Frawley et al., 2020; S. K. Moore et al., 2020; Pellowe and Leslie, 2019;  
489 Ritzman et al., 2018) will provide complementary insights into the motivations  
490 and social drivers behind adaptive decisions, and could help identify system-  
491 specific metrics of success or performance beyond profitability. Furthermore,  
492 as integrated biophysical and socioeconomic data streams become increasingly  
493 available for environmental management (Bradley et al., 2019), data-driven, inter-  
494 disciplinary studies of resilience and adaptation will enable dynamic management  
495 of natural resources (Hazen et al., 2018; Maxwell et al., 2015). This push for the  
496 incorporation of multiple data streams in environmental management extends  
497 beyond marine fisheries. For example, in wildland fire management in the United  
498 States, integrated data platforms that combine geospatial data with risk models  
499 and fuel treatment scenarios are empowering adaptive fire management plans  
500 (Ager et al., 2011; Krofcheck et al., 2018).

501 This study revealed the elements of behavioral diversity among human  
502 harvesters in a lucrative, keystone commercial fishery, and described how those  
503 elements enabled adaptation during an extreme environmental event attributable  
504 to climate change (Hinder et al., 2012). Just as biological response diversity  
505 can lead to enhanced ecosystem resilience to environmental change (Elmqvist  
506 et al., 2003), behavioral diversity among natural resource users may promote  
507 resilience of social-ecological systems. Given the impending increase in extreme  
508 climatic events such as marine heatwaves (Burge et al., 2014; Smale et al., 2019),  
509 recognition of social and ecological traits that enable resilience now can help to  
510 build toward a more prepared future. As quantitative data become increasingly  
511 available in the United States and far beyond (Bradley et al., 2019), behavioral  
512 analyses like ours can be used in the design of adaptive management measures,

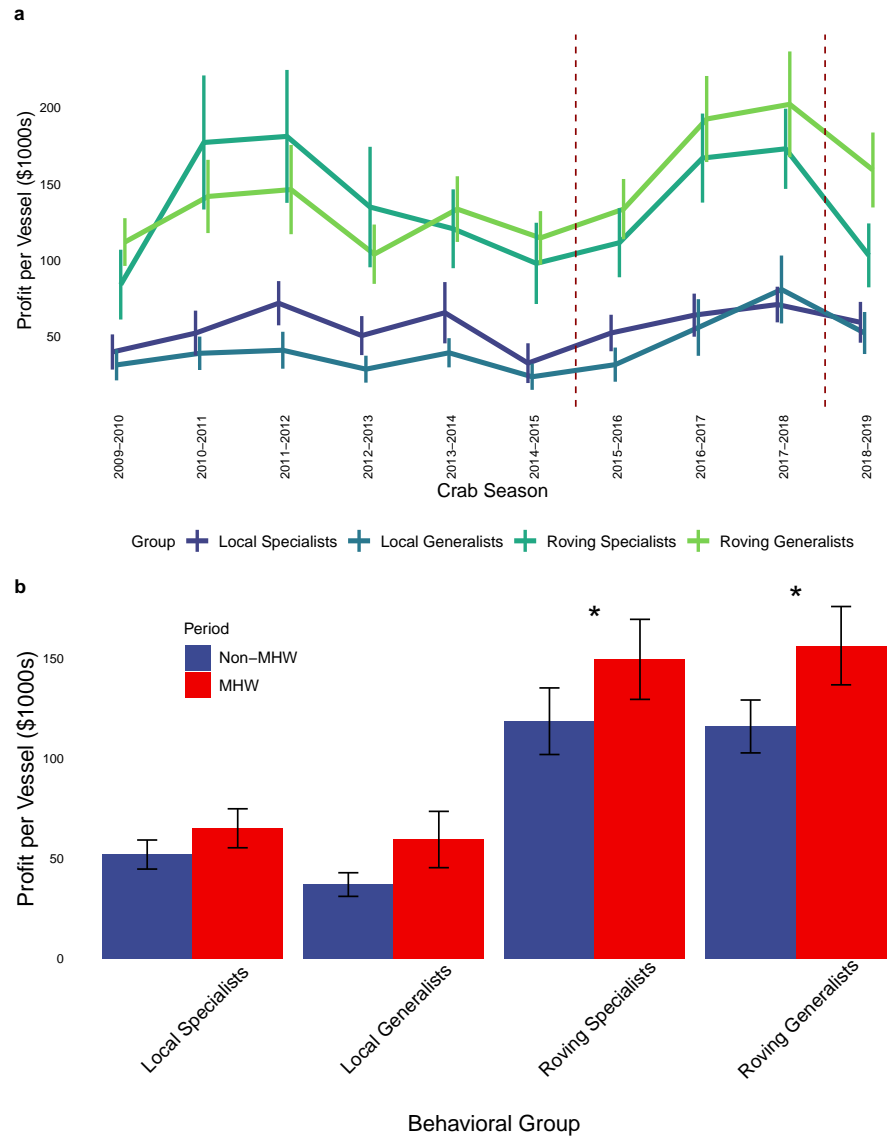


Figure 5: Estimated profits by behavioral group. (a) Mean profit ( $\pm$  2 SE) for vessels in each behavioral group over the full crab season. Vertical lines delineate the period of the marine heatwave. (b) Mean profit ( $\pm$  2 SE) for each group in heatwave (MHW) versus non-MHW seasons. Stars indicate groups with significantly different estimated profits during MHW seasons.

to bolster policy analyses (Cabral et al., 2018), and to inform decision-making under environmental uncertainty.

## References

- Abatzoglou, J.T., Williams, A.P., Barbero, R., 2019. Global emergence of anthropogenic climate change in fire weather indices. *Geophysical Research Letters* 46, 326–336. doi:10.1029/2018GL080959
- Abrahms, B., Hazen, E.L., Bograd, S.J., Brashares, J.S., Robinson, P.W., Scales, K.L., Crocker, D.E., Costa, D.P., 2018. Climate mediates the success of migration strategies in a marine predator. *Ecology Letters* 21, 63–71. doi:10.1111/ele.12871
- Ager, A.A., Vaillant, N.M., Finney, M.A., 2011. Integrating fire behavior models and geospatial analysis for wildland fire risk assessment and fuel management planning. *Journal of Combustion* 2011. doi:10.1155/2011/572452
- Allison, E.H., Ellis, F., 2001. The livelihoods approach and management of small-scale fisheries. *Marine Policy* 25, 377–388.
- Beever, E.A., Hall, L.E., Varner, J., Loosen, A.E., Dunham, J.B., Gahl, M.K., Smith, F.A., Lawler, J.J., 2017. Behavioral flexibility as a mechanism for coping with climate change. *Frontiers in Ecology and the Environment* 15, 299–308. doi:10.1002/fee.1502
- Bradley, D., Merrifield, M., Miller, K.M., Lomonico, S., Wilson, J.R., Gleason, M.G., 2019. Opportunities to improve fisheries management through innovative technology and advanced data systems. *Fish and Fisheries* 20, 564–583. doi:10.1111/faf.12361
- Branch, T.A., Hilborn, R., Haynie, A.C., Fay, G., Flynn, L., Griffiths, J., Marshall, K.N., Randall, J.K., Scheuerell, J.M., Ward, E.J., Young, M., 2006. Fleet dynamics and fishermen behavior: Lessons for fisheries managers. *Canadian Journal of Fisheries and Aquatic Sciences* 63, 1647–1668.
- Brodie, J.F., Fragoso, J.M.V., 2020. Understanding the distribution of bushmeat hunting effort across landscapes by testing hypotheses about human foraging. *Conservation Biology* 0, 1–10. doi:10.1111/cobi.13612
- Burge, C.A., Eakin, C.M., Friedman, C.S., Froelich, B., Hershberger, P.K., Hofmann, E.E., Petes, L.E., Prager, K.C., Weil, E., Willis, B.L., Ford, S.E., Harvell, C.D., 2014. Climate change influences on marine infectious diseases: Implications for management and society. *Annual Review of Marine Science* 6, 249–277. doi:10.1146/annurev-marine-010213-135029
- Cabral, R.B., Mayorga, J., Clemence, M., Lynham, J., Koeshendrajana, S., Muawanah, U., Nugroho, D., Anna, Z., Ghofar, A., Zulfainarni, N., others, 2018. Rapid and lasting gains from solving illegal fishing. *Nature Ecology & Evolution* 2, 650–658.
- Cavole, L.M., Demko, A.M., Diner, R.E., Giddings, A., Koester, I., Pagniello, C.M., Paulsen, M.-L., Ramirez-Valdez, A., Schwenck, S.M., Yen, N.K., others, 2016. Biological impacts of the 2013–2015 warm-water anomaly in the northeast pacific: Winners, losers, and the future. *Oceanography* 29, 273–285.

556 Charrad, M., Ghazzali, N., Boiteau, V., Niknafs, A., 2014. NbClust: An R  
557 package for determining the relevant number of clusters in a data set. *Journal*  
558 *of Statistical Software* 61, 1–36.

559 Cohen, J.D., McClure, S.M., Yu, A.J., 2007. Should i stay or should i  
560 go? How the human brain manages the trade-off between exploitation and  
561 exploration. *Philosophical Transactions of the Royal Society B: Biological*  
562 *Sciences* 362, 933–942.

563 Cook, B.I., Mankin, J.S., Anchukaitis, K.J., 2018. Climate change and  
564 drought: From past to future. *Current Climate Change Reports* 4, 164–179.  
565 doi:10.1007/s40641-018-0093-2

566 Dewees, C.M., Sortais, K., Krachey, M.J., Hackett, S.C., Hankin, D.G., 2004.  
567 Racing for crabs... costs and management options evaluated in dungeness crab  
568 fishery. *California Agriculture* 58, 186–189. doi:10.3733/ca.v058n04p186

569 Elmqvist, T., Folke, C., Nyström, M., Peterson, G., Bengtsson, J., Walker, B.,  
570 Norberg, J., 2003. Response diversity, ecosystem change, and resilience. *Frontiers*  
571 *in Ecology and the Environment* 1, 488–494. doi:10.1890/1540-9295(2003)001[0488:RDECAR]2.0.CO;2

572 Feist, B.E., Samhouri, J.F., Forney, K.A., Saez, L.E., 2021. Footprints of  
573 fixed-gear fisheries in relation to rising whale entanglements on the u.s. West  
574 coast. *Fisheries Management and Ecology* 28, 283–294. doi:10.1111/fme.12478

575 Finkbeiner, E.M., 2015. The role of diversification in dynamic small-scale  
576 fisheries: Lessons from baja california sur, mexico. *Global Environmental Change*  
577 32, 139–152. doi:10.1016/j.gloenvcha.2015.03.009

578 Fisher, M.C., Moore, S.K., Jardine, S.L., Watson, J.R., Samhouri, J.F., 2021.  
579 Climate shock effects and mediation in fisheries. *Proceedings of the National*  
580 *Academy of Sciences of the United States of America* 118, 1–8. doi:10.1073/pnas.2014379117

581 Frawley, T.H., Muhling, B.A., Brodie, S., Fisher, M.C., Tommasi, D., Fol,  
582 G.L., Hazen, E.L., Stohs, S.S., Finkbeiner, E.M., Jacox, M.G., 2020. Changes to  
583 the structure and function of an albacore fishery reveal shifting social-ecological  
584 realities for pacific northwest fishermen 1–18. doi:10.1111/faf.12519

585 Fryxell, J.M., Hilborn, R., Bieg, C., Turgeon, K., Caskenette, A., McCann,  
586 K.S., 2017. Supply and demand drive a critical transition to dysfunctional  
587 fisheries. *Proceedings of the National Academy of Sciences of the United States*  
588 *of America* 114, 12333–12337. doi:10.1073/pnas.1705525114

589 Fuller, E.C., Samhouri, J.F., Stoll, J.S., Levin, S.A., Watson, J.R., 2017.  
590 Characterizing fisheries connectivity in marine social-ecological systems. *ICES*  
591 *Journal of Marine Science* 74, 2087–2096. doi:10.1093/icesjms/fsx128

592 Fulton, E.A., Smith, A.D.M., Smith, D.C., Putten, I.E.V., 2011. Human  
593 behaviour: The key source of uncertainty in fisheries management. *Fish and*  
594 *Fisheries* 12, 2–17. doi:10.1111/j.1467-2979.2010.00371.x

595 Gallagher, A.J., Hammerschlag, N., Cooke, S.J., Costa, D.P., Irschick, D.J.,  
596 2015. Evolutionary theory as a tool for predicting extinction risk. *Trends in*  
597 *Ecology and Evolution* 30, 61–65. doi:10.1016/j.tree.2014.12.001

598 Gladics, A.J., Melvin, E.F., Suryan, R.M., Good, T.P., Jannot, J.E., Guy,  
599 T.J., 2017. Fishery-specific solutions to seabird bycatch in the u.s. West coast  
600 sablefish fishery. *Fisheries Research* 196, 85–95. doi:10.1016/j.fishres.2017.08.015

601 Gordon, H.S., 1954. The economic theory of a common-property resource:  
602 The fishery. *The Journal of Political Economy* 124–142.

603 Hamilton, S., Baker, G.B., 2019. Technical mitigation to reduce marine  
604 mammal bycatch and entanglement in commercial fishing gear: Lessons learnt  
605 and future directions. *Reviews in Fish Biology and Fisheries* 29, 223–247.  
606 doi:10.1007/s11160-019-09550-6

607 Hazen, E.L., Scales, K.L., Maxwell, S.M., Briscoe, D.K., Welch, H., Bograd,  
608 S.J., Bailey, H., Benson, S.R., Eguchi, T., Dewar, H., Kohin, S., Costa, D.P.,  
609 Crowder, L.B., Lewison, R.L., 2018. A dynamic ocean management tool to  
610 reduce bycatch and support sustainable fisheries. *Science Advances* 4, eaar3001.  
611 doi:10.1126/sciadv.aar3001

612 Hilborn, R., 1985. Fleet dynamics and individual variation: Why some people  
613 catch more fish than others. *Canadian Journal of Fisheries and Aquatic Sciences*  
614 42, 2–13. doi:10.1139/f85-001

615 Hinder, S.L., Hays, G.C., Edwards, M., Roberts, E.C., Walne, A.W., Gravenor,  
616 M.B., 2012. Changes in marine dinoflagellate and diatom abundance under cli-  
617 mate change. *Nature Climate Change* 2, 271–275. doi:10.1038/nclimate1388

618 Holland, D.S., Abbott, J.K., Norman, K.E., 2020. Fishing to live or living to  
619 fish: Job satisfaction and identity of west coast fishermen. *Ambio* 49, 628–639.  
620 doi:10.1007/s13280-019-01206-w

621 Holland, D.S., Leonard, J., 2020. Is a delay a disaster? Economic impacts of  
622 the delay of the california dungeness crab fishery due to a harmful algal bloom.  
623 *Harmful Algae* 98, 101904. doi:10.1016/j.hal.2020.101904

624 Holland, D.S., Speir, C., Agar, J., Crosson, S., Depiper, G., Kasperski, S.,  
625 Kitts, A.W., Perruso, L., 2017. Impact of catch shares on diversification of  
626 fishers’ income and risk. *Proceedings of the National Academy of Sciences of*  
627 *the United States of America* 114, 9302–9307. doi:10.1073/pnas.1702382114

628 Jardine, S.L., Fisher, M.C., Moore, S.K., Samhouri, J.F., 2020. Inequality in  
629 the economic impacts from climate shocks in fisheries: The case of harmful algal  
630 blooms. *Ecological Economics* 176, 106691. doi:10.1016/j.ecolecon.2020.106691

631 Joo, R., Salcedo, O., Gutierrez, M., Fablet, R., Bertrand, S., 2015. Defin-  
632 ing fishing spatial strategies from vms data: Insights from the world’s largest  
633 monospecific fishery. *Fisheries Research* 164, 223–230. doi:10.1016/j.fishres.2014.12.004

634 Kasperski, S., Holland, D.S., 2013. Income diversification and risk for  
635 fishermen. *Proceedings of the National Academy of Sciences* 110, 2076–2081.

636 Krofcheck, D.J., Hurteau, M.D., Scheller, R.M., Loudermilk, E.L., 2018.  
637 Prioritizing forest fuels treatments based on the probability of high-severity fire  
638 restores adaptive capacity in sierran forests. *Global Change Biology* 24, 729–737.  
639 doi:10.1111/gcb.13913

640 Leslie, H.M., Basurto, X., Nenadovic, M., Sievanen, L., Cavanaugh, K.C.,  
641 Cota-Nieto, J.J., Erisman, B.E., Finkbeiner, E., Hinojosa-Arango, G., Moreno-  
642 Báez, M., Nagavarapu, S., Reddy, S.M.W., Sánchez-Rodríguez, A., Siegel, K.,  
643 Ulibarria-Valenzuela, J.J., Weaver, A.H., Aburto-Oropeza, O., 2015. Opera-  
644 tionalizing the social-ecological systems framework to assess sustainability. *Pro-*  
645 *ceedings of the National Academy of Sciences of the United States of America*  
646 112, 5979–5984. doi:10.1073/pnas.1414640112

Liaw, A., Wiener, M., 2002. Classification and regression by randomForest. *R News* 3, 18–22.

Loon, A.F.V., Gleeson, T., Clark, J., Dijk, A.I.J.V., Stahl, K., Hannaford, J., Baldassarre, G.D., Teuling, A.J., Tallaksen, L.M., Uijlenhoet, R., Hannah, D.M., Sheffield, J., Svoboda, M., Verbeiren, B., Wagener, T., Rangelcroft, S., Wanders, N., Lanen, H.A.J.V., 2016. Drought in the anthropocene. *Nature Geoscience* 9, 89–91. doi:10.1038/ngeo2646

Lubchenco, J., Cerny-Chipman, E.B., Reimer, J.N., Levin, S.A., 2016. The right incentives enable ocean sustainability successes and provide hope for the future. *Proceedings of the National Academy of Sciences of the United States of America* 113, 14507–14514. doi:10.1073/pnas.1604982113

Mao, J., Jardine, S.L., 2020. Market impacts of a toxic algae event: The case of california dungeness crab. *Marine Resource Economics* 35, 1–20. doi:10.1086/707643

Maxwell, S.M., Hazen, E.L., Lewison, R.L., Dunn, D.C., Bailey, H., Bograd, S.J., Briscoe, D.K., Fossette, S., Hobday, A.J., Bennett, M., Benson, S., Caldwell, M.R., Costa, D.P., Dewar, H., Eguchi, T., Hazen, L., Kohin, S., Sippel, T., Crowder, L.B., 2015. Dynamic ocean management: Defining and conceptualizing real-time management of the ocean. *Marine Policy* 58, 42–50. doi:10.1016/j.marpol.2015.03.014

McCabe, R.M., Hickey, B.M., Kudela, R.M., Lefebvre, K.A., Adams, N.G., Bill, B.D., Gulland, F.M.D., Thomson, R.E., Cochlan, W.P., Trainer, V.L., 2016. An unprecedented coastwide toxic algal bloom linked to anomalous ocean conditions. *Geophysical Research Letters* 43, 10, 366–10, 376. doi:10.1002/2016GL070023

Mcginnis, M.D., Ostrom, E., 2014. Social-ecological system framework : Initial changes and continuing challenges. *Ecology and Society* 19.

Mendo, T., Smout, S., Photopoulou, T., James, M., 2019. Identifying fishing grounds from vessel tracks: Model-based inference for small scale fisheries. *Royal Society Open Science* 6. doi:10.1098/rsos.191161

Moore, K.M., Allison, E.H., Dreyer, S.J., Ekstrom, J.A., Jardine, S.L., Klinger, T., Moore, S.K., Norman, K.C., 2020. Harmful algal blooms: Identifying effective adaptive actions used in fishery-dependent communities in response to a protracted event. *Frontiers in Marine Science* 6, 803.

Moore, S.K., Dreyer, S.J., Ekstrom, J.A., Moore, K., Norman, K., Klinger, T., Allison, E.H., Jardine, S.L., 2020. Harmful algal blooms and coastal communities: Socioeconomic impacts and actions taken to cope with the 2015 u.s. West coast domoic acid event. *Harmful Algae* 96, 101799. doi:10.1016/j.hal.2020.101799

O’Farrell, S., Chollett, I., Sanchirico, J.N., Perruso, L., 2019. Classifying fishing behavioral diversity using high-frequency movement data. *Proceedings of the National Academy of Sciences* 116, 16811–16816. doi:10.1073/pnas.1906766116

O’Farrell, S., Sanchirico, J.N., Spiegel, O., Depalle, M., Haynie, A.C., Murawski, S.A., Perruso, L., Strelcheck, A., 2019. Disturbance modifies payoffs in the explore-exploit trade-off. *Nature Communications* 10, 1–9. doi:10.1038/s41467-019-11106-y

Oken, K.L., Holland, D.S., Andr´, A., Punt, A.E., 2021. The effects of population synchrony, life history, and access constraints on benefits from fishing portfolios.

693 Oliver, E.C.J., Donat, M.G., Burrows, M.T., Moore, P.J., Smale, D.A.,  
 694 Alexander, L.V., Benthuisen, J.A., Feng, M., Gupta, A.S., Hobday, A.J., Hol-  
 695 brook, N.J., Perkins-Kirkpatrick, S.E., Scannell, H.A., Straub, S.C., Wernberg,  
 696 T., 2018. Longer and more frequent marine heatwaves over the past century.  
 697 *Nature Communications* 9, 1–12. doi:10.1038/s41467-018-03732-9  
 698 Pellowe, K.E., Leslie, H.M., 2019. Heterogeneity among clam harvesters in  
 699 northwest mexico shapes individual adaptive capacity. *Ecology and Society* 24.  
 700 doi:10.5751/ES-11297-240425  
 701 Pfeiffer, L., Gratz, T., 2016. The effect of rights-based fisheries management  
 702 on risk taking and fishing safety. *Proceedings of the National Academy of Sciences*  
 703 of the United States of America 113, 2615–2620. doi:10.1073/pnas.1509456113  
 704 Rasmuson, L.K., 2013. The biology, ecology and fishery of the dungeness  
 705 crab, cancer magister, 1st ed, *Advances in Marine Biology*. Elsevier Ltd.  
 706 doi:10.1016/B978-0-12-410498-3.00003-3  
 707 R Core Team, 2021. R: A language and environment for statistical computing.  
 708 R Foundation for Statistical Computing, Vienna, Austria.  
 709 Renner, M., Kuletz, K.J., 2015. A spatial-seasonal analysis of the oiling risk  
 710 from shipping traffic to seabirds in the aleutian archipelago. *Marine Pollution*  
 711 *Bulletin* 101, 127–136. doi:10.1016/j.marpolbul.2015.11.007  
 712 Richerson, K., Punt, A.E., Holland, D.S., 2020. Nearly a half century of  
 713 high but sustainable exploitation in the dungeness crab (cancer magister) fishery.  
 714 *Fisheries Research* 226, 105528. doi:10.1016/j.fishres.2020.105528  
 715 Ritzman, J., Brodbeck, A., Brostrom, S., McGrew, S., Dreyer, S., Klinger,  
 716 T., Moore, S.K., 2018. Economic and sociocultural impacts of fisheries closures  
 717 in two fishing-dependent communities following the massive 2015 u.s. West coast  
 718 harmful algal bloom. *Harmful Algae* 80, 35–45. doi:10.1016/j.hal.2018.09.002  
 719 Russell, S.M., Oostenburg, M.V., Vizek, A., 2018. Adapting to catch shares:  
 720 Perspectives of west coast groundfish trawl participants. *Coastal Management*  
 721 46, 603–620. doi:10.1080/08920753.2018.1522491  
 722 Salas, S., Gaertner, D., 2004. The behavioural dynamics of fishers: Man-  
 723 agement implications. *Fish and Fisheries* 5, 153–167. doi:10.1111/j.1467-  
 724 2979.2004.00146.x  
 725 Santora, J.A., Mantua, N.J., Schroeder, I.D., Field, J.C., Hazen, E.L., Bograd,  
 726 S.J., Sydeman, W.J., Wells, B.K., Calambokidis, J., Saez, L., Lawson, D., Forney,  
 727 K.A., 2020. Habitat compression and ecosystem shifts as potential links between  
 728 marine heatwave and record whale entanglements. *Nature Communications* 2020  
 729 11:1 11, 1–12. doi:10.1038/s41467-019-14215-w  
 730 Smale, D.A., Wernberg, T., Oliver, E.C.J., Thomsen, M., Harvey, B.P.,  
 731 Straub, S.C., Burrows, M.T., Alexander, L.V., Benthuisen, J.A., Donat, M.G.,  
 732 Feng, M., Hobday, A.J., Holbrook, N.J., Perkins-Kirkpatrick, S.E., Scannell,  
 733 H.A., Gupta, A.S., Payne, B.L., Moore, P.J., 2019. Marine heatwaves threaten  
 734 global biodiversity and the provision of ecosystem services. *Nature Climate*  
 735 *Change* 9, 306–312. doi:10.1038/s41558-019-0412-1  
 736 Smith, C.L., McKelvey, R., 1986. Specialist and generalist: Roles for coping  
 737 with variability. *North American Journal of Fisheries Management* 6, 88–99.  
 738 doi:10.1577/1548-8659(1986)6<88:sag>2.0.co;2

739 Suryan, R.M., Arimitsu, M.L., Coletti, H.A., Hopcroft, R.R., Lindeberg,  
 740 M.R., Barbeaux, S.J., Batten, S.D., Burt, W.J., Bishop, M.A., Bodkin, J.L.,  
 741 Brenner, R., Campbell, R.W., Cushing, D.A., Danielson, S.L., Dorn, M.W.,  
 742 Drummond, B., Esler, D., Gelatt, T., Hanselman, D.H., Hatch, S.A., Haught,  
 743 S., Holderied, K., Iken, K., Irons, D.B., Kettle, A.B., Kimmel, D.G., Konar, B.,  
 744 Kuletz, K.J., Laurel, B.J., Maniscalco, J.M., Matkin, C., McKinstry, C.A.E.,  
 745 Monson, D.H., Moran, J.R., Olsen, D., Palsson, W.A., Pegau, W.S., Piatt, J.F.,  
 746 Rogers, L.A., Rojek, N.A., Schaefer, A., Spies, I.B., Straley, J.M., Strom, S.L.,  
 747 Sweeney, K.L., Szymkowiak, M., Weitzman, B.P., Yasumiishi, E.M., Zador, S.G.,  
 748 2021. Ecosystem response persists after a prolonged marine heatwave. *Scientific*  
 749 *Reports* 11, 1–17. doi:10.1038/s41598-021-83818-5  
 750 Townhill, B.L., Tinker, J., Jones, M., Pitois, S., Creach, V., Simpson, S.D.,  
 751 Dye, S., Bear, E., Pinnegar, J.K., 2018. Harmful algal blooms and climate  
 752 change: Exploring future distribution changes. *ICES Journal of Marine Science*  
 753 75, 1882–1893. doi:10.1093/icesjms/fsy113  
 754 von Biela, V., Arimitsu, M.L., Piatt, J.F., Heflin, B.M., Schoen, S., 2019.  
 755 Extreme reduction in condition of a key forage fish during the pacific marine  
 756 heatwave of 2014–2016. *Marine Ecology Progress Series* 613, 171–182.  
 757 Watson, J.T., Haynie, A.C., 2016. Using vessel monitoring system data to  
 758 identify and characterize trips made by fishing vessels in the united states north  
 759 pacific. *PLoS ONE* 11, 1–20. doi:10.1371/journal.pone.0165173  
 760 Wilson, J.R., Lomonico, S., Bradley, D., Sievanen, L., Dempsey, T., Bell,  
 761 M., McAfee, S., Costello, C., Szuwalski, C., McGonigal, H., Fitzgerald, S.,  
 762 Gleason, M., 2018. Adaptive comanagement to achieve climate-ready fisheries.  
 763 *Conservation Letters* 11, 1–7. doi:10.1111/conl.12452  
 764 Young, T., Fuller, E.C., Provost, M.M., Coleman, K.E., Martin, K.S., McCay,  
 765 B.J., Pinsky, M.L., 2019. Adaptation strategies of coastal fishing communi-  
 766 ties as species shift poleward. *ICES Journal of Marine Science* 76, 93–103.  
 767 doi:10.1093/icesjms/fsy140