

**1Letters**

**2Mismatches in thermal and nutrient physiology predict competitive outcomes among  
3phytoplankton**

**4Short running title:** Physiological mismatches predict competition

5Elvire Bestion<sup>1\*</sup>, Bernardo Garcia-Carreras<sup>2</sup>, Charlotte-Elisa Schaum<sup>1</sup>, Samraat Pawar<sup>2</sup>,  
6Gabriel Yvon-Durocher<sup>1\*</sup>

7<sup>1</sup>Environment and Sustainability Institute, University of Exeter, Penryn, Cornwall TR10 9EZ,  
8UK

9<sup>2</sup> Department of Life Sciences, Imperial College London, Silwood Park Campus, Ascot,  
10Berkshire, SL5 7PY, UK

11 [e.bestion@exeter.ac.uk](mailto:e.bestion@exeter.ac.uk), [bernardo.garcia-carreras08@imperial.ac.uk](mailto:bernardo.garcia-carreras08@imperial.ac.uk),

12 [C.L.Schaum@exeter.ac.uk](mailto:C.L.Schaum@exeter.ac.uk), [s.pawar@imperial.ac.uk](mailto:s.pawar@imperial.ac.uk), [g.yvon-durocher@exeter.ac.uk](mailto:g.yvon-durocher@exeter.ac.uk)

**13\* Corresponding authors:**

14Elvire Bestion

15Environment and Sustainability Institute, University of Exeter, Penryn, Cornwall TR10 9EZ,  
16UK **Add phone + fax** Email: [e.bestion@exeter.ac.uk](mailto:e.bestion@exeter.ac.uk)

17Gabriel Yvon-Durocher

18Environment and Sustainability Institute, University of Exeter, Penryn, Cornwall TR10 9EZ,  
19UK **Add phone + fax** Email: [g.yvon-durocher@exeter.ac.uk](mailto:g.yvon-durocher@exeter.ac.uk)

20**Statement of authorship:** GYD and SP conceived the study. EB and GYD designed the experiment, EB  
21performed the experiment, SP and BGC wrote the theory. EB and BGC analysed the data, EB wrote  
22the first draft and all authors contributed to writing.

23**Data accessibility:** data will be available on dryad upon publication

24**Keywords:** nutrients, phosphate, global changes, climate change, interspecific competition,  
25trait-based ecology

26

27**Number of words in the abstract:**

28**Number of words in the main text:**

29**Number of references:**

30**Number of figures, tables and text boxes:**

31

## 32Abstract

33

34Current climate change affects species through both direct effects of temperature on species  
35physiology and indirect effects of temperature on species interactions. To better predict the  
36consequences of future climate change, it is thus crucial to understand how increased  
37temperatures affect species interactions. Recent theoretical studies have demonstrated the  
38potential for mismatches between prey and predators' thermal physiology to alter consumer-  
39resource dynamics. However fewer resources have been devoted to explaining interspecific  
40competition, and, to our knowledge, no large experimental study has tackled this issue to  
41build a bridge between theory and experiments. Here we investigated how mismatches in  
42competing species' thermal and nutrient physiology affected the outcome of the competition  
43in phytoplankton. We developed a theoretical model based on the Monod model of nutrient  
44physiology to investigate competition between species, and tested the predictions of this  
45model against a large scale competition experiment of six species of freshwater  
46phytoplankton at two temperatures and two nutrient conditions. We show that competitive  
47outcomes are driven by mismatches in species maximum growth rates  $\mu_{\max}$  and half-  
48saturation constant  $K_S$ . Further, reversals in competitive outcomes with temperature were  
49linked to temperature-driven reversals in nutrient physiology traits  $\mu_{\max}$  and  $K_S$ .

50

## 51Introduction

52Climate change is predicted to be a major cause of species extinctions over the next century  
53(Field *et al.* 2014), and a considerable threat to biodiversity (Thomas *et al.* 2004; Bellard *et*  
54*al.* 2012). Susceptibility to climate change will depend on species' environmental tolerances  
55(Pacifici *et al.* 2015), with those occupying narrower thermal niches expected to be more  
56vulnerable to climate warming (Magozzi & Calosi 2015). However, recent studies have  
57highlighted that species interactions may play a greater role in mediating the impacts of  
58climate change on populations than physiological tolerance limits (Dunn *et al.* 2009; Bellard  
59*et al.* 2012; Cahill *et al.* 2013; Field *et al.* 2014). Indeed the key drivers of global change  
60(warming, CO<sub>2</sub> and nutrient enrichment) are known to affect various types of species  
61interactions, including competition (Tylianakis *et al.* 2008). To better predict the  
62consequences of future climate change, it is therefore crucial to understand how increased  
63temperatures affect species interactions (Bestion & Cote 2017).

64Metabolism sets the pace of life (Brown *et al.* 2004) and dictates a host of life-history traits  
65and attributes that determine fitness, including population growth rate (Savage *et al.* 2004),  
66abundance, mortality and interspecific interactions (Dell *et al.* 2011). Species vary widely in  
67the way in which their metabolism and associated traits respond to temperature (Kingsolver  
682009; Dell *et al.* 2011), and these differences in thermal physiology can greatly impact  
69species interactions (Reuman *et al.* 2014; Dell *et al.* 2014). Mismatches can arise when  
70species' metabolic traits differ in their magnitude (the elevated of thermal performance  
71curve), sensitivity to temperature (the slope of their temperature-performance relationship)  
72and/or thermal optima (the temperature at which the performance is maximised) (Kordas *et*  
73*al.* 2011). Recent theory suggests that mismatches in the thermal responses of body velocity  
74between interacting species can play a key role in shaping the effects of temperature on  
75consumer-resource dynamics (Dell *et al.* 2014). Mismatches in the temperature-dependence  
76of metabolic rate, nutrient supply rate, consumer consumption efficiency and mortality rates  
77all have the potential to affect biomass fluxes between consumers and resources, and in turn,  
78the stability of food webs (Gilbert *et al.* 2014). In plant-herbivore interactions, higher  
79temperature-dependence of heterotroph respiration compared to photosynthesis has been  
80predicted to increase the strength of top down control in aquatic ecosystems (O'Connor *et al.*  
812011). However, while there have major advances in ecological theory, linking the effects of  
82temperature to metabolism and species interactions (O'Connor *et al.* 2011; Dell *et al.* 2014;  
83Gilbert *et al.* 2014; Amarasekare 2015; Uszko *et al.* 2017), there have been very few  
84empirical tests of this theory, and to our knowledge, no large scale experimental study has  
85confronted recent theoretical developments to test how mismatches in thermal physiology  
86drive the outcome of species interactions.

87In aquatic ecosystems, temperature and nutrients are the two main drivers of phytoplankton  
88productivity (Litchman *et al.* 2010). The effects of temperature on phytoplankton growth  
89typically follow a characteristic left-skewed unimodal function, where rates increase  
90exponentially to an optimum followed by a steeper exponential decline. Phytoplankton  
91exhibit substantial variation among species and functional groups in these thermal response  
92curves (Thomas *et al.* 2016) and interspecific variation in thermal tolerance can be an  
93important driver of community dynamics and seasonal succession in phytoplankton  
94communities (Grover & Chrzanowski 2006). Nutrient availability also has a major impact on  
95phytoplankton growth, with rates typically increasing as a saturating, hyperbolic function of  
96increasing nutrients, characterised by the Monod curve (Monod 1949). Interspecific variation  
97in the functional traits that shape nutrient uptake and growth (e.g. the half saturation constant

98and the maximum growth rate) are widely recognised to be key drivers of competition  
 99(Tilman 1981), community assembly (Bulgakov & Levich 1999) and ultimately the  
 100productivity of phytoplankton communities (Behrenfeld *et al.* 2005). The non-linear effects  
 101of temperature and nutrients also interact multiplicatively. For example, temperature can  
 102influence both the half-saturation constant and the maximum growth rate (Aksnes & Egge  
 1031991; Sterner & Grover 1998; Carter & Lathwell 1967; Mechling & Kilham 1982; Senft *et*  
 104*al.* 1981) and vice-versa, recent work has shown that the optimum temperature for growth  
 105increases as a saturating function of nutrient availability (Thomas *et al.* 2017). Thus changes  
 106in environmental conditions can potentially amplify mismatches between competitors'  
 107functional traits, and this could affect species competition and community assembly  
 108(Litchman *et al.* 2010; Kordas *et al.* 2011). Given that both temperature and nutrient balance  
 109are predicted to shift with global changes (IPCC 2013; Behrenfeld *et al.* 2006; Ye *et al.*  
 1102011), understanding the potential for such climate-driven mismatches is made all the more  
 111urgent.

112Here we investigate how mismatches between species thermal and nutrient traits affect  
 113competition in phytoplankton. We do so by modelling species growth rate as a function of  
 114nutrient physiology through a Monod equation (Monod 1949). Species nutrient physiology is  
 115defined by the two parameters of the model, the maximum growth rate (maximum growth  
 116rate achieved in nutrient replete conditions) and the half-saturation constant (concentration of  
 117limiting nutrient at which the growth is half of the maximum growth rate), which can  
 118themselves vary with temperature. We study how mismatches in these traits can predict  
 119competitive outcomes. We then test our model's predictions against empirical data on 6  
 120species of freshwater phytoplankton in a large-scale experiment with all pairwise  
 121combinations of the six species at two temperature and nutrient levels.

## 122Theory

123We used the Monod equation to characterize the effects of nutrient concentration on  
 124phytoplankton growth rate. The dynamics of a single species limited by a single resource can  
 125be described as

$$126 \quad \frac{1}{N} \frac{dN}{dt} = \mu = \frac{\mu_{\max} S}{K_S + S} \quad (1)$$

$$127 \quad \frac{dS}{dt} = -\alpha N \frac{\mu_{\max} S}{K_S + S} \quad (2)$$

128Where  $N$  is the phytoplankton cell density (cells·mL<sup>-1</sup>),  $\mu$  is the realised growth rate (d<sup>-1</sup>)  
 129of the species,  $\mu_{\max}$  is the maximum growth rate in nutrient saturated conditions (d<sup>-1</sup>),  
 130which reflects a species' performance under nutrient saturated conditions,  $K_S$  is the half-  
 131saturation constant (μmol·L<sup>-1</sup>) which is a measure of performance at low nutrient  
 132concentrations,  $S$  is the nutrient concentration (μmol·L<sup>-1</sup>) and  $\alpha$  is the term that  
 133converts units of phytoplankton density to nutrient concentration ((1000·μmol)·cell<sup>-1</sup>). The  
 134parameters of the Monod equation,  $\mu_{\max}$  and  $K_S$ , can be considered as 'functional  
 135traits' that characterise a species' nutrient physiology. These traits have been shown to vary  
 136among species and play an important role in shaping competitive dynamics in phytoplankton  
 137communities. Furthermore,  $\mu_{\max}$  and  $K_S$ , are likely to exhibit temperature dependence.  
 138We expect maximum growth rate to be tightly coupled to metabolism, and consequently the  
 139temperature dependence of  $\mu_{\max}$  is expected to follow a left-skewed unimodal function of

temperature, where rates increase exponentially to an optimum followed by a steeper exponential decline. The effects of temperature on  $K_S$  are poorly understood and empirical studies have documented a wide range of temperature dependence functions (Aksnes & Egge 1991; Sterner & Grover 1998; Carter & Lathwell 1967; Mechling & Kilham 1982; Senft *et al.* 1981). The joint effects of temperature and nutrient concentrations on phytoplankton growth can be described by,

146

Then need some text emphasising how different spp, might differ in  $\mu$  &  $K_S$  and their respective temperature sensitivities.

149

A model for the dynamics of two species competing for a single limiting resource at a range of temperatures would be

$$152 \quad \frac{1}{N_a} \frac{dN_a}{dt} = \mu_a = \frac{\mu_{\max,a} S}{K_{S,a} + S}$$

$$153 \quad \frac{1}{N_b} \frac{dN_b}{dt} = \mu_b = \frac{\mu_{\max,b} S}{K_{S,b} + S}$$

$$154 \quad \frac{dS}{dt} = -\alpha_a N_b \frac{\mu_{\max,a} S}{K_{S,a} + S} - \alpha_b N_b \frac{\mu_{\max,b} S}{K_{S,b} + S}$$

155

would be great to show here how competitive outcomes depend on mismatches in traits by  $S$  and  $T$ .

158

Where the underscripts  $a$  and  $b$  denote of the identity of each species. In a scenario where species colonise a novel experiment, the competition starts when the two species are rare. In this situation, we hypothesise that differences in exponential growth (and therefore differences in the traits and variables that give rise to this growth rate) are key, whereas other mechanisms that might be more relevant once populations reach a carrying capacity, such as inter and intraspecific competition, might here play a minor role. Further, we assume that species only compete for a common resource. We hypothesise that competition will be driven by differences, or mismatches, in individual species' nutrient physiology traits, and that competitive outcomes are not significantly affected by direct interspecific interference such as production of toxins or competition for light. Finally, we neglect mortality rate. In this scenario, when the concentration of nutrients  $S$  is important, the growth rate of the two species  $\mu_a$  and  $\mu_b$  is close to  $\mu_{\max}$ . Therefore the competitive outcome between the two species is mainly driven by mismatches in the maximum growth rate, and species  $a$  wins when  $\mu_{\max,a} > \mu_{\max,b}$ . However, when  $S$  becomes lower, the importance of the half-saturation constant  $K_S$  becomes higher as it limits growth rate. A species having a lower  $K_S$  will be able to have a higher growth rate at low nutrient concentrations, thus species  $a$  will win when  $K_{S,a} < K_{S,b}$ .

## 176Methods

177

### 178Study design

179We used an experimental approach to test the predictive ability of this simple competition  
180model in predicting competitive outcomes in a context of climate change using 6 species of  
181freshwater phytoplankton. We first determined species nutrient physiology traits  $\mu_{\max}$  and  
182  $K_s$  and their temperature dependence. We then competed the 6 species in all pairwise  
183combinations at two temperature and two nutrient levels. The results from the competition  
184experiment were then matched to the model's predictions, using the empirical data on species  
185nutrient physiology to parametrize the model. The model is simple and explicitly only  
186captures the traits and variables measured experimentally, except for the biomass conversion  
187parameter,  $\alpha$ . We therefore ran models with both a fixed  $\alpha$  for all species (results  
188largely robust to the choice of value), and with  $\alpha_i$  proportional to median cell size,  
189assuming that species with larger cell sizes would equate to a greater amount of phosphate  
190per cell. The results are largely insensitive to the choice of  $\alpha$  (indeed, species were  
191initially chosen to be similar in size). Therefore, the results here are presented for a constant  
192  $\alpha = 1 \cdot 10^{-4}$ . If the model were to predict correctly competitive outcomes, it would show  
193that mismatches between traits have a great importance for species competition. Conversely,  
194if the predictive power was low, it could suggest that mismatches in nutrient physiology alone  
195are not the most important driver of the competition, and that other factors, and more  
196complex models that include factors such as interspecific interference and density dependent  
197growth need to be taken into account.

### 198Species and culture conditions

199The experiment was conducted with six species of freshwater green algae that are known to  
200naturally co-occur, *Ankistrodesmus nannoselene*, *Chlamydomonas moewusii*, *Chlorella*  
201*sorokiniana*, *Monoraphidium minutum*, *Scenedesmus obliquus* and *Selenastrum*  
202*capricornutum* (Fritschie *et al.* 2014). We chose these 6 species because (i) they are similar  
203cell size and (ii) can be cultured on the same media (e.g. standard COMBO culture medium  
204without animal trace elements (Kilham *et al.* 1998)). Strains of each species were ordered in  
205October 2015 from the Culture Collection of Algae and Protozoa ([www.ccap.ac.uk](http://www.ccap.ac.uk), see  
206Supplementary Table 1 for detailed information about the strains). Upon arrival, species were  
207grown on COMBO culture medium, and maintained in semi-continuous culture in an Aralab  
208incubator at 15°C on a 12:12 light-dark cycle with a light intensity of 90  $\mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$ .

209

### 210Nutrient and temperature dependence of growth rate

211We measured growth rates of the 6 species of green algae across gradients in temperature and  
212phosphate concentration. Each of the 6 species was grown in a factorial experiment at 5  
213temperatures and 13 phosphate concentrations, with 3 replicates per combination, amounting  
214to a total of 1170 cultures. We created 13 solutions of different phosphate concentrations  
215ranging from 0.01  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate to 50  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate (original phosphate  
216concentration in the COMBO medium) by mixing different amounts of COMBO medium  
217with and without potassium phosphate dibasic (Table S1B). This range was relevant to  
218phosphate concentrations commonly found in lakes (Downing *et al.* 2001). Small tissue  
219culture flasks (Nunc) filled with 40 mL of each solution were inoculated with each species  
220in monoculture with around 100 cells $\cdot\text{mL}^{-1}$ . Samples were diluted or concentrated by

221filtration to allow for the same inoculation volume, 10  $\mu\text{L}$  (for the very low phosphate  
222concentrations, 0.01, 0.1 and 0.5  $\mu\text{mol}\cdot\text{L}^{-1}$ ) and 50  $\mu\text{L}$  (for all of the other samples), ensuring  
223that the increase in phosphate concentration due to the inoculum was minimal (respectively  
2240.01 and 0.06  $\mu\text{mol}\cdot\text{L}^{-1}$ ). Samples were then grown in Percival incubators at 15, 20, 25, 30,  
225and 35°C on a 12:12 light-dark cycle and with a light intensity of 90  $\mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$  (range: 70-  
226110). Every day, samples were shaken and their position inside of the incubators was  
227randomly changed. Every two days, a 200  $\mu\text{L}$  sample was taken and 10  $\mu\text{L}$  of 1% sorbitol  
228solution was added as a cryoprotectant. After one hour of incubation in the dark, samples  
229were frozen at -80°C until further analysis. Cell density in each sample was determined by  
230flow cytometry (BD Accuri C6). Plates were thawed in a water bath at *ca* 38°C for 10  
231minutes and then run on the flow cytometer on fast flux settings (66  $\mu\text{L}\cdot\text{min}^{-1}$ ), counting 10  
232 $\mu\text{L}$  of each sample. Cleaning fluid was run after each species to avoid contamination of  
233measurements between species. The experiment was run for one month. During the  
234experiment, some samples failed to grow properly and were therefore removed from the  
235subsequent analyses.

### 236Species competition

237To investigate the joint effects of temperature and phosphate availability on competitive  
238outcomes among the 6 species of algae, we competed each of the species in all pairwise  
239combinations (15 pairs) at two temperatures (15 and 25°C; low temperature and a  
240temperature close to the optimum for most species) and two phosphate concentrations  
241(saturating [30  $\mu\text{mol}\cdot\text{L}^{-1}$ ] and limiting [1  $\mu\text{mol}\cdot\text{L}^{-1}$ ] concentrations, chosen from the Monod  
242curves, see Fig. 1, Fig. S1), each replicated 6 times. We also grew the 6 species in  
243monoculture at the two temperature and nutrient levels. The monoculture trials were divided  
244into two subsets, one training subset, used to train the cell discrimination algorithm, which  
245was replicated 3 times per temperature and nutrient levels and inoculated with 200 cells  
246 $\text{cells}\cdot\text{mL}^{-1}$ , and a testing subset used to test the accuracy of the cell discrimination algorithm,  
247which was replicated 6 times per temperature and phosphate level and inoculated with 100  
248 $\text{cells}\cdot\text{mL}^{-1}$ . In total, the design included 576 samples. The competition experiment was  
249done in twenty-four 24 well plates filled with 2 mL of media, and inoculated with 100 or 200  
250 $\text{cells}\cdot\text{mL}^{-1}$  of each species. The position of the species pairs were randomised within the  
251plates, however given the large number of samples and to minimise experimenter error, we  
252separated low-P from high-P plates. Plates were covered with AeraSeal breathable membrane,  
253minimising evaporation and contamination but allowing gas exchange. The competition  
254plates were incubated in the same way as described above for the monoculture growth curves.  
255After 14 days, which was identified from the monoculture experiments as being sufficient  
256time to reach stationary phase, a 200  $\mu\text{L}$  sample was taken and preserved in the same way as  
257described above. Cell density in each sample was determined by flow cytometry (BD Accuri  
258C6) on the slow flux setting (14  $\mu\text{L}\cdot\text{min}$ ), counting 20  $\mu\text{L}$  of each sample. Cleaning fluid was  
259run after each sample to avoid contamination of measurements between samples.

### 260Data analyses

261All statistical analyses were undertaken using R v3.3.2 (R Core Team 2014).

### 262Nutrient and temperature dependence of growth rate

263To characterise the effects of phosphorous availability and temperature on growth we  
264estimated specific growth rate for each of the 1170 combinations of species, phosphate and  
265temperatures from the time-series of cell densities. Population dynamics were fitted to the  
266Buchanan three-phase linear growth model (Buchanan *et al.* 1997) using non-linear least  
267squares regression.



$$N_t = \begin{cases} N_0 & \text{for } t \leq t_{\text{lag}}, \\ N_0 + \mu(t - t_{\text{lag}}) & \text{for } t_{\text{lag}} < t < t_{\text{max}}, \\ N_{\text{max}} & \text{for } t \geq t_{\text{max}}, \end{cases} \quad (x)$$

269

270where  $t_{\text{lag}}$  is the duration of the lag phase (days),  $t_{\text{max}}$  is the time when the maximum  
 271population density is reached (days),  $N_0$  is the  $\log_{10}$  of the initial population density  
 272( $\log_{10}(\text{cells} \cdot \text{mL}^{-1})$ ),  $N_{\text{max}}$  is the  $\log_{10}$  of the maximum population density supported by the  
 273environment ( $\log_{10}(\text{cells} \cdot \text{mL}^{-1})$ ), and  $\mu$  is the specific growth rate ( $\text{day}^{-1}$ ). Fits to the  
 274Buchanan model were determined using the 'nlsLM' function in the 'minpack.lm' package in  
 275R (Elzhov *et al.* 2010), which uses the Levenberg-Marquardt optimisation algorithm.  
 276Parameter estimation was achieved by running 1000 different random combination of starting  
 277parameters picked from uniform distributions and returning the parameter set with the lowest  
 278AICc score.

279The Monod equation (Eq 1, Monod 1949), was fitted to the estimates of  $\mu$  for each  
 280species at each temperature and for each of the three replicates using the 'nlsLM' function in  
 281the 'minpack.lm' package. Parameter estimation was achieved by running 1000 different  
 282random combination of starting parameters picked from a uniform distribution and returning  
 283the parameter set that returned the lowest AICc score.

284We used generalized additive models (GAMs) to describe the thermal variation in  $\mu_{\text{max}}$   
 285and  $K_s$ . For each species, we fitted a gam model of each parameter with temperature as a  
 286smoother term with the number of knots fixed at 3 with the gam function from the mgcv  
 287package.

## 288 **Competition**

289FSC files returned by the flow cytometer were read into R using the Bioconductor package  
 290'FlowCore', returning side scatter (SSC), forward scatter (FSC), green fluorescence (FL1),  
 291orange fluorescence (FL2), red fluorescence (FL3), and blue fluorescence (FL4) values that  
 292could be used to define species morphology and thus discriminate between species in  
 293pairwise competition samples and determine species identity for each cell. We first filtered  
 294the data to remove noise by removing every data point where either  $\ln(\text{FSC.H}) < 10.3$ ,  
 295 $\ln(\text{SSC.H}) < 3$  or  $\ln(\text{FL3.H}) < 1.5$ , which are below minimum values observed for life cells of  
 296all species. We then separated the data set into 3 data frames, one for the isolates inoculated at  
 297100  $\text{cells} \cdot \text{mL}^{-1}$ , and one for the isolates inoculated at 200  $\text{cells} \cdot \text{mL}^{-1}$ , and one for the  
 298competing species. The 200  $\text{cells} \cdot \text{mL}^{-1}$  isolates dataset measured at day 14 was used to  
 299determine pairwise discrimination functions between pair of species. We first removed  
 300outliers from this dataset by manually inspecting FSC.H by FL3.H clustering plots and  
 301choosing visual thresholds for these two values for each species. We then applied 3 different  
 302procedures to discriminate between pairs of species for each temperature and phosphate level:  
 303a linear discriminant analysis with the 'lda' function from the 'MASS' R package, a random  
 304forest analysis with the 'randomForest' function from the 'randomForest' R package and a  
 305recursive partitioning and regression tree analysis with the 'rpart' function from the 'rpart' R  
 306package. These analyses were performed using the natural logarithm of the 10 morphological  
 307variables returned by the flow cytometer (that is FSC.H, FSC.A, SSC.H, SSC.A, FL1.H,  
 308FL1.A, FL2.H, FL2.A, FL3.H, FL3.A, FL4.H and FL4.A, .H standing for height and .A for  
 309area), on each of the 15 pairs of species for each combination of temperature and phosphate



level. These different discriminant functions were then applied to the 100 cells·mL<sup>-1</sup> isolates dataset previously filtered by removing visually determined outliers to test the accuracy of the predictions for the different discriminant methods. We then chose the method that gave the maximum level of accuracy to apply to the competition dataset (Fig. S2A). The best method was the linear discriminant analysis that gave 84 % of accuracy in predicting species identity (Table S2A).

After determining species identity for each sample, we computed cell density and calculated competition coefficients as the proportion of cells from the focal species over the total number of cells. We also computed a binary competition outcome where the competitive dominant was defined as the species that attained more than 50% of the total number of cells.

## Results

### Nutrient and temperature dependence of growth rate

The responses of growth rate to the gradients in phosphate concentration were well fit by the Monod equation (Fig. 1a). The half-saturation constant,  $K_s$ , and the maximum growth rate,  $\mu_{max}$ , varied with temperature, and the temperature response of these traits differed between the 6 species (Table S3A and S3B). Maximum growth rate exhibited a unimodal temperature dependence in *Ankistrodesmus*, *Chlamydomonas*, and *Selenastrum* (Fig 1b, Table S3A). In *Chlorella* and *Monoraphidium*,  $\mu_{max}$  increased monotonically and did not reach their optima by 35°C, while *Scenedesmus* exhibited negligible temperature dependence (Fig 1b, Table S3A).  $K_s$  increased with temperature for *Ankistrodesmus* and *Chlamydomonas*, while *Chlorella* and *Selenastrum* exhibited a unimodal response to temperature and there was no discernible trend for *Monoraphidium* and *Scenedesmus* (Fig. 1c, Table S3B).  $\mu_{max}$  and  $K_s$  were also positively correlated (Pearson  $r = 0.45$  [0.27,0.60],  $t = 4.77$ ,  $df = 88$ ,  $p < 0.001$ ), highlighting a trade-off between maximum growth rate and performance at low nutrient concentrations.

### Species competition

Competition between species varied depending on temperature, nutrient conditions and pair identity (Fig. 2). For instance, for the pair *Chlorella*-*Ankistrodesmus*, *Chlorella* dominated the competition at low temperature, while *Ankistrodesmus* dominated at high temperature under both nutrient conditions. For the pair *Monoraphidium*-*Chlorella*, *Monoraphidium* won in every instance except at high nutrient concentration and high temperature, where *Chlorella* won. For the pair *Scenedesmus*-*Chlamydomonas*, at low temperature, there was no clear winner between the two species regardless of nutrient conditions, while at high temperature the outcome depended of nutrient conditions: at low nutrient conditions, there was no clear winner while at high nutrient conditions *Chlamydomonas* won (Fig. 2).

Differences in  $\mu$  (the realised growth rate estimated at the specific temperature and nutrient concentration) between the two competitors alone predicted the correct competitive outcome 62% of times (that is, the competitor with the higher  $\mu$  won the competition; Table 2). Mismatches in  $\mu_{max}$  (maximum growth rate at saturating nutrient concentrations from the Monod model) predicted the competitive outcome 60% of times, and differences in  $K_s$  only predicted the correct competitive outcome 37% of times (i.e., the competitor with the lower  $K_s$  won the competition). The difference in predictive power of mismatches in  $\mu_{max}$  and  $K_s$  stands to reason given the positive correlation between individuals'

354  $\mu_{\max}$  and  $K_S$  ; given a high  $\mu_{\max}$  is associated to a high  $K_S$  , both  $\mu_{\max}$  and  
 355  $K_S$  are unlikely to equally be able to predict the outcome of a competition. The  
 356 competition model, which incorporates both mismatches in  $\mu_{\max}$  and  $K_S$  , predicted the  
 357 correct competitive outcome 65% of times. Therefore, the inclusion of both mismatches  
 358 allowed for a marginally greater predictive power. Results remained largely the same when  
 359 looking at the two temperatures and nutrient concentrations separately, but the predictability  
 360 of the competitive outcome was very dependent on the species involved (Table 2).  
 361 Competitions involving *Selenastrum* were considerably more difficult to predict with any of  
 362 the mismatches (Table 2). This could, in part, be due to the lesser power of discrimination  
 363 between cells in pairs involving this species (Table S2A), as well as to the wide confidence  
 364 intervals around  $\mu_{\max}$  and  $K_S$  for this species (Fig 1). Indeed, only removing  
 365 competitions involving *Selenastrum* increased the predictive power of mismatches in  $\mu$  ,  
 366  $\mu_{\max}$  and the competition model (72.5%, 72.5%, and 77.5% of outcomes correctly  
 367 predicted, respectively) but not that of mismatches in  $K_S$  (32.5%). The results were robust  
 368 to the statistical method used to calculate  $\mu_{\max}$  and  $K_S$  , as well as to the statistical  
 369 method used to discriminate between species (Supplementary material S7).

370 The results indicate that  $\mu_{\max}$  is a distinctly more important trait for predicting competitive  
 371 outcomes than is  $K_S$  , suggesting that performance at low nutrient concentrations had little  
 372 bearing in a species' competitive ability. Simulations clearly show that at higher nutrient  
 373 concentrations, mismatches in  $K_S$  have little or no influence on competitions (Fig. 3a, c).  
 374 At the lower nutrient concentration, mismatches in  $K_S$  are of greater importance, but  
 375 mismatches in  $\mu_{\max}$  nonetheless still dominate (Fig. 3b, d).

376 In some cases, the winner of a competition changed depending on the nutrient concentration  
 377 and/or temperature. For example, *Chlorella* won against *Chlamydomonas* at 15°C, but the  
 378 reverse was true at 25°C. These reversals, or flips of the competitive outcome, were far more  
 379 likely to occur between temperatures (in 18 out of 30 competitions; 15 pairs and two nutrient  
 380 concentrations) than between nutrient concentrations (six out of 30). In the 18 reversals due  
 381 to the change in temperature, 14 coincided with similar reversals in the species'  $\mu_{\max}$  (i.e.,  
 382  $\mu_{\max}$  was higher for one species at one temperature, but not at the other), and only six  
 383 coincided with reversals in species'  $K_S$  , thus corroborating the greater significance of  
 384  $\mu_{\max}$  in determining competitive outcomes (Table 2).

## 385 Discussion

386

387 Global change is predicted to affect both the temperature of aquatic ecosystems (IPCC 2013)  
 388 and their nutrient balance (e.g. through an increase in vertical stratifications, reducing  
 389 nutrient supply (Behrenfeld *et al.* 2006), or through an increase in eutrophication, increasing  
 390 nutrient supply (Ye *et al.* 2011)). These shifts could lead to mismatches between competing  
 391 species' physiological traits, and influence the outcome of the competition. Here we showed  
 392 that phytoplankton species varied for their temperature response of nutrient physiology traits,  
 393 and that this variation affected competition between species. Mismatches between species  
 394 maximum growth rate  $\mu_{\max}$  and/or between species half-saturation constant  $K_S$  led to  
 395 reversals in competitive outcome between pairs of species depending on environmental  
 396 conditions.

397 We found that traits governing species nutrient physiology were not fixed values for a species  
 398 but varied plastically with temperature. Growth rate depended both on temperature and  
 399 nutrients non-linearly for each species. Half-saturation constants generally increased with  
 400 temperature. These results are in accordance with previous results showing a positive  
 401 relationship between  $K_s$  and temperature in plants (Carter & Lathwell 1967) or in algae  
 402 for nitrogen (Aksnes & Egge 1991; Sterner & Grover 1998) and silicate (Mechling & Kilham  
 403 1982), and a hump shaped relationship between  $K_s$  and temperature in algae for phosphorus  
 404 (Senft *et al.* 1981). Studies investigating effects of climate change on algal biomass often  
 405 consider the half-saturation constant to be independent of temperature (Goldman & Carpenter  
 406 1974; Ye *et al.* 2011; Thomas *et al.* 2017); our results highlight the temperature-dependence  
 407 of nutrient-limited growth. Further, the relationship between temperature and nutrient  
 408 physiology depended on species identity, with for instance *Selenastrum* having a much higher  
 409 half-saturation constant than *Ankistrodesmus*.

410 Competitive outcomes between pairs of species varied with temperature and nutrient  
 411 conditions. These results match previous studies, where temperature has been shown to  
 412 influence competitive interactions in various groups including phytoplankton, arthropods and  
 413 vertebrates (see (Dunson & Travis 1991) for a review). Mismatches in nutrient uptake traits  
 414 were a good predictor of competitive outcomes between species. Particularly, mismatches in  
 415  $\mu_{\max}$  were clearly a better predictor of competitive ability than performance at low nutrient  
 416 concentrations, although knowledge of mismatches in both traits using the model helped  
 417 improve predictive power. Our results in fact indicate that  $\mu_{\max}$  was the more important  
 418 trait even at low nutrient concentrations, when we would have expected  $K_s$  to be most  
 419 significant (Table 2, Figure 3). This could be due to the relatively large confidence intervals  
 420 around our estimates of  $K_s$ , or to the fact that the lowest concentration of phosphate used  
 421 in the competition experiment,  $1 \mu\text{mol}\cdot\text{L}^{-1}$ , was still relatively high compared to the half-  
 422 saturating constant of most species. Our results on the predictability of competitive outcomes  
 423 (Table 2) should also be interpreted in the context that competitions were very variable across  
 424 replicates (Figure 2), that is to say, both competing species were often observed to win across  
 425 six replicates. Competitive outcomes were highly predictable when excluding competitions  
 426 involving *Selenastrum*, suggesting that the other species predominantly compete for  
 427 resources (and implying that there is little direct interference). On the other hand, the  
 428 predictability of *Selenastrum*'s competitive outcomes was poor. This might have been due to  
 429 the fact that our discriminating power for this species was low (Table S2Ab), but could also  
 430 indicate that competitions with this species might have involved some significant form of  
 431 direct interspecific interference (e.g. production of toxins), as also indicated by the fact that  
 432 competitive interactions involving this species were all strongly negative, leading to a  
 433 strongly diminished yield of the pair of competitors relative to the same species in  
 434 monoculture (Loreau & Hector 2001) deviation from expected yield in pairs involving the  
 435 focal species, mean  $\pm$  SD,  $-0.77 \pm 0.36$  in pairs involving *Selenastrum*, compared to  $-0.37 \pm$   
 436  $0.64$ ,  $-0.13 \pm 0.51$ ,  $-0.19 \pm 0.72$  and  $-0.23 \pm 0.53$  in pairs involving *Ankistrodesmus*,  
 437 *Chlamydomonas*, *Chlorella*, *Monoraphidium* and *Scenedesmus* respectively).

438 We highlighted different competitions where some species always won, while there were  
 439 frequent reversals of competitive outcomes, particularly with temperature and less so with  
 440 nutrients. Reversals in competitive outcomes were often linked to analogous reversals in the  
 441 values of  $\mu_{\max}$ . Mismatches in both  $\mu$  and  $\mu_{\max}$  were themselves linked to  
 442 mismatches in physiological traits. Therefore having a better understanding on the thermal-  
 443 dependency of species nutrient physiology is an important step if we are to understand how  
 444 species competition and community functioning can be affected by climate change (Litchman

445& Klausmeier 2008; Litchman *et al.* 2010). The results of our study contrast with some  
446earlier studies, as (Park 1954) found that higher growth rate of a competitor at higher  
447temperature did not lead to a switch in competitive dominance in *Tribolium* species.

448More generally, our findings stress the importance of considering how species traits will  
449plastically change with temperature to better understand biotic interactions in a context of  
450global climate change. Studying consequences of climate change in terms of mismatches  
451between physiological traits should be a useful approach in understanding how species  
452interactions will be modified by warming climates (Dell *et al.* 2014). Further, because global  
453changes are unlikely to act only through temperature changes but should involve rapid  
454modifications of both nutrient and thermal conditions (Behrenfeld *et al.* 2006; Ye *et al.* 2011;  
455IPCC 2013), it is crucial to better understand how the combination of multiple stressors  
456should affect species and community responses to global changes. We highlight the interest of  
457considering a spectrum of different ecological contexts to predict successful competitors and  
458invaders, and to pinpoint which mismatch in species traits is more important in which  
459ecological context.

460

461**Acknowledgements:** We thank Saskia Johnson and Emily Budd for their help in the experiments. This  
462work was supported by a NERC grant number XXXXXX to GYD and SP.

## 463References

464

465

4661.

467Aksnes, D.L. & Egge, J.K. (1991). A theoretical model for nutrient uptake in phytoplankton. *Marine*  
468*Ecology Progress Series*, 70, 65–72

469

4702.

471Amarasekare, P. (2015). Effects of temperature on consumer–resource interactions. *J Anim Ecol*, 84,  
472665–679

473

4743.

475Behrenfeld, M.J., Boss, E., Siegel, D.A. & Shea, D.M. (2005). Carbon-based ocean productivity and  
476phytoplankton physiology from space. *Global Biogeochem. Cycles*, 19, GB1006

477

4784.

479Behrenfeld, M.J., O'Malley, R.T., Siegel, D.A., McClain, C.R., Sarmiento, J.L., Feldman, G.C., *et al.*  
480(2006). Climate-driven trends in contemporary ocean productivity. *Nature*, 444, 752–755

481

4825.

483Bellard, C., Bertelsmeier, C., Leadley, P., Thuiller, W. & Courchamp, F. (2012). Impacts of climate  
484change on the future of biodiversity. *Ecology Letters*, 15, 365–377

485

4866.

487Bestion, E. & Cote, J. (2017). Species Responses to Climate Change: Integrating Individual-Based  
488Ecology Into Community and Ecosystem Studies. In: *Reference Module in Earth Systems and*  
489*Environmental Sciences*. Elsevier

490  
4917.  
492Brown, J.H., Gillooly, J.F., Allen, A.P., Savage, V.M. & West, G.B. (2004). Toward a metabolic theory of  
493ecology. *Ecology*, 85, 1771–1789  
494  
4958.  
496Buchanan, R.L., Whiting, R.C. & Damert, W.C. (1997). When is simple good enough: a comparison of  
497the Gompertz, Baranyi, and three-phase linear models for fitting bacterial growth curves. *Food*  
498*Microbiology*, 14, 313–326  
499  
5009.  
501Bulgakov, N.G. & Levich, A.P. (1999). The nitrogen : Phosphorus ratio as a factor regulating  
502phytoplankton community structure : Nutrient ratios. *Archiv für Hydrobiologie*, 146, 3–22  
503  
50410.  
505Cahill, A.E., Aiello-Lammens, M.E., Fisher-Reid, M.C., Hua, X., Karanewsky, C.J., Ryu, H.Y., *et al.* (2013).  
506How does climate change cause extinction? *Proc. R. Soc. B*, 280, 20121890  
507  
50811.  
509Carter, O.G. & Lathwell, D.J. (1967). Effects of Temperature on Orthophosphate Absorption by Excised  
510Corn Roots. *Plant Physiol.*, 42, 1407–1412  
511  
51212.  
513Dell, A.I., Pawar, S. & Savage, V.M. (2011). Systematic variation in the temperature dependence of  
514physiological and ecological traits. *PNAS*, 108, 10591–10596  
515  
51613.  
517Dell, A.I., Pawar, S. & Savage, V.M. (2014). Temperature dependence of trophic interactions are  
518driven by asymmetry of species responses and foraging strategy. *J Anim Ecol*, 83, 70–84  
519  
52014.  
521Downing, J.A., Watson, S.B. & McCauley, E. (2001). Predicting Cyanobacteria dominance in lakes.  
522*Can. J. Fish. Aquat. Sci.*, 58, 1905–1908  
523  
52415.  
525Dunn, R.R., Harris, N.C., Colwell, R.K., Koh, L.P. & Sodhi, N.S. (2009). The sixth mass coextinction: are  
526most endangered species parasites and mutualists? *Proc. R. Soc. B*, 276, 3037–3045  
527  
52816.  
529Dunson, W.A. & Travis, J. (1991). The Role of Abiotic Factors in Community Organization. *The*  
530*American Naturalist*, 138, 1067–1091  
531  
53217.  
533Elzhov, T.V., Mullen, K.M. & Bolker, B. (2010). R interface to the Levenberg-Marquardt nonlinear least-  
534squares algorithm found in MINPACK  
535  
53618.  
537Field, C.B., Barros, V.R. & Intergovernmental Panel on Climate Change (Eds.). (2014). *Climate change*  
538*2014: impacts, adaptation, and vulnerability: Working Group II contribution to the fifth assessment*  
539*report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, New York, NY

540  
54119.  
542Fritschie, K.J., Cardinale, B.J., Alexandrou, M.A. & Oakley, T.H. (2014). Evolutionary history and the  
543strength of species interactions: testing the phylogenetic limiting similarity hypothesis. *Ecology*, 95,  
5441407–1417  
545  
54620.  
547Gilbert, B., Tunney, T.D., McCann, K.S., DeLong, J.P., Vasseur, D.A., Savage, V., et al. (2014). A  
548bioenergetic framework for the temperature dependence of trophic interactions. *Ecol Lett*, 17, 902–  
549914  
550  
55121.  
552Goldman, J.C. & Carpenter, E.J. (1974). A kinetic approach to the effect of temperature on algal  
553growth. *Limnology and Oceanography*, 19, 756–766  
554  
55522.  
556Grover, J.P. & Chrzanowski, T.H. (2006). Seasonal dynamics of phytoplankton in two warm temperate  
557reservoirs: association of taxonomic composition with temperature. *J Plankton Res*, 28, 1–17  
558  
55923.  
560IPCC. (2013). *Climate change 2013: the physical science basis : Working Group I contribution to the*  
561*fifth assessment report of the Intergovernmental Panel on Climate Change*. Cambridge University  
562Press, Cambridge, United Kingdom and New York, NY, USA  
563  
56424.  
565Kilham, S.S., Kreeger, D.A., Lynn, S.G., Goulden, C.E. & Herrera, L. (1998). COMBO: a defined  
566freshwater culture medium for algae and zooplankton. *Hydrobiologia*, 377, 147–159  
567  
56825.  
569Kingsolver, J.G. (2009). The Well-Tempered Biologist. *The American Naturalist*, 174, 755–768  
570  
57126.  
572Kordas, R.L., Harley, C.D.G. & O'Connor, M.I. (2011). Community ecology in a warming world: The  
573influence of temperature on interspecific interactions in marine systems. *Journal of Experimental*  
574*Marine Biology and Ecology*, Global change in marine ecosystems, 400, 218–226  
575  
57627.  
577Litchman, E. & Klausmeier, C.A. (2008). Trait-Based Community Ecology of Phytoplankton. *Annual*  
578*Review of Ecology, Evolution, and Systematics*, 39, 615–639  
579  
58028.  
581Litchman, E., Pinto, P. de T., Klausmeier, C.A., Thomas, M.K. & Yoshiyama, K. (2010). Linking traits to  
582species diversity and community structure in phytoplankton. *Hydrobiologia*, 653, 15–28  
583  
58429.  
585Loreau, M. & Hector, A. (2001). Partitioning selection and complementarity in biodiversity  
586experiments. *Nature*, 412, 72–76  
587  
58830.  
589Magozzi, S. & Calosi, P. (2015). Integrating metabolic performance, thermal tolerance, and plasticity

590 enables for more accurate predictions on species vulnerability to acute and chronic effects of global  
 591 warming. *Global Change Biology*, 21, 181–194  
 592  
 593 31.  
 594 Mechling, J.A. & Kilham, S.S. (1982). Temperature Effects on Silicon Limited Growth of the Lake  
 595 Michigan Diatom *Stephanodiscus Minutus* (bacillariophyceae)1. *Journal of Phycology*, 18, 199–205  
 596  
 597 32.  
 598 Monod, J. (1949). The growth of bacterial cultures. *Annual Reviews in Microbiology*, 3, 371–394  
 599  
 600 33.  
 601 O'Connor, M.I., Gilbert, B. & Brown, C.J. (2011). Theoretical Predictions for How Temperature Affects  
 602 the Dynamics of Interacting Herbivores and Plants. *The American Naturalist*, 178, 626–638  
 603  
 604 34.  
 605 Pacifici, M., Foden, W.B., Visconti, P., Watson, J.E.M., Butchart, S.H.M., Kovacs, K.M., *et al.* (2015).  
 606 Assessing species vulnerability to climate change. *Nature Clim. Change*, 5, 215–224  
 607  
 608 35.  
 609 Park, T. (1954). Experimental Studies of Interspecies Competition II. Temperature, Humidity, and  
 610 Competition in Two Species of *Tribolium*. *Physiological Zoology*, 27, 177–238  
 611  
 612 36.  
 613 R Core Team. (2014). *R: A Language and Environment for Statistical Computing*. R Foundation for  
 614 Statistical Computing, Vienna, Austria  
 615  
 616 37.  
 617 Reuman, D.C., Holt, R.D. & Yvon-Durocher, G. (2014). A metabolic perspective on competition and  
 618 body size reductions with warming. *J Anim Ecol*, 83, 59–69  
 619  
 620 38.  
 621 Savage, V.M., Gillooly, J.F., Brown, J.H., West, G.B. & Charnov, E.L. (2004). Effects of Body Size and  
 622 Temperature on Population Growth. *The American Naturalist*, 163, 429–441  
 623  
 624 39.  
 625 Senft, W.H., Hunchberger, R.A. & Roberts, K.E. (1981). Temperature Dependence of Growth and  
 626 Phosphorus Uptake in Two Species of *Volvox* (volvocales, Chlorophyta)1. *Journal of Phycology*, 17,  
 627 323–329  
 628  
 629 40.  
 630 Sterner, R.W. & Grover, J.P. (1998). Algal growth in warm temperate reservoirs: kinetic examination of  
 631 nitrogen, temperature, light, and other nutrients. *Water Research*, 32, 3539–3548  
 632  
 633 41.  
 634 Thomas, C.D., Cameron, A., Green, R.E., Bakkenes, M., Beaumont, L.J., Collingham, Y.C., *et al.* (2004).  
 635 Extinction risk from climate change. *Nature*, 427, 145–148  
 636  
 637 42.  
 638 Thomas, M.K., Aranguren-Gassis, M., Kremer, C.T., Gould, M.R., Anderson, K., Klausmeier, C.A., *et al.*  
 639 (2017). Temperature–nutrient interactions exacerbate sensitivity to warming in phytoplankton. *Glob*  
 640 *Change Biol*, n/a–n/a



641

64243.

643Thomas, M.K., Kremer, C.T. & Litchman, E. (2016). Environment and evolutionary history determine  
644the global biogeography of phytoplankton temperature traits. *Global Ecology and Biogeography*, 25,  
64575–86

646

64744.

648Tilman, D. (1981). Tests of Resource Competition Theory Using Four Species of Lake Michigan Algae.  
649*Ecology*, 62, 802–815

650

65145.

652Tylianakis, J.M., Didham, R.K., Bascompte, J. & Wardle, D.A. (2008). Global change and species  
653interactions in terrestrial ecosystems. *Ecology Letters*, 11, 1351–1363

654

65546.

656Uszko, W., Diehl, S., Englund, G. & Amarasekare, P. (2017). Effects of warming on predator–prey  
657interactions – a resource-based approach and a theoretical synthesis. *Ecol Lett*, n/a–n/a

658

65947.

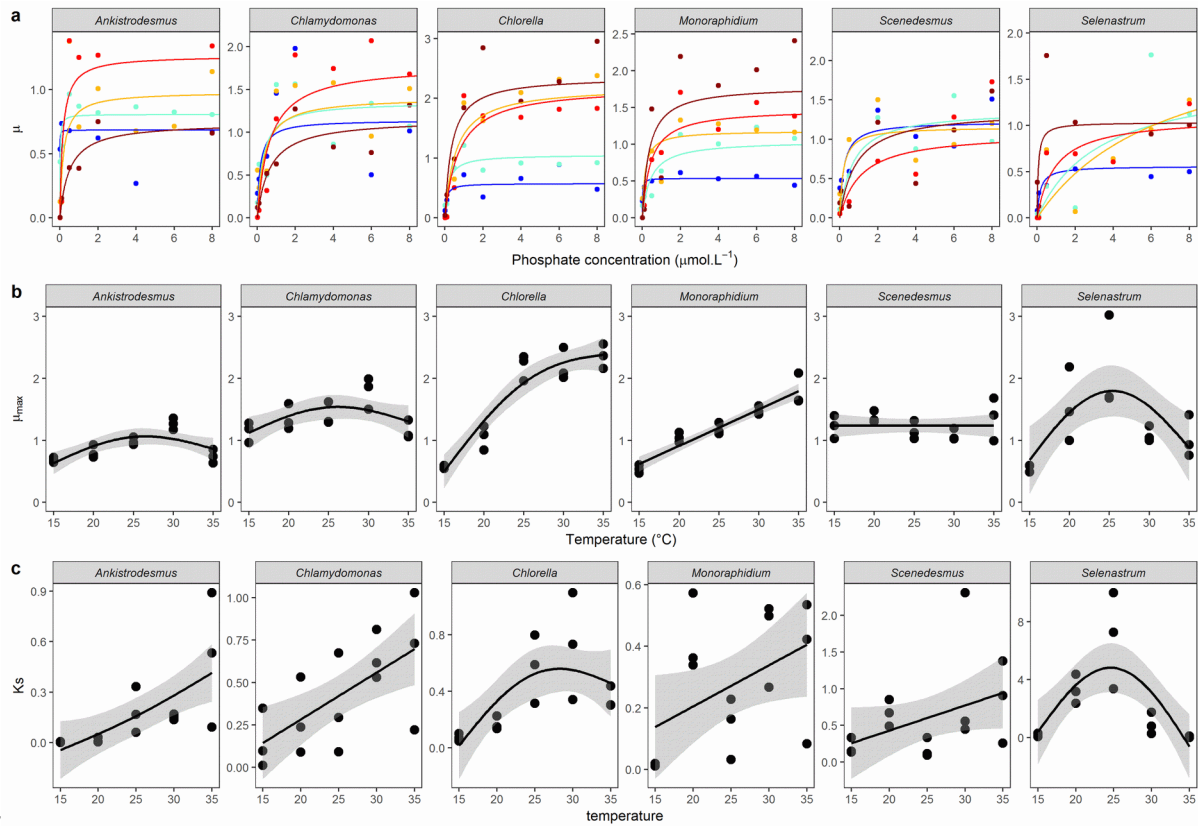
660Ye, C., Shen, Z., Zhang, T., Fan, M., Lei, Y. & Zhang, J. (2011). Long-term joint effect of nutrients and  
661temperature increase on algal growth in Lake Taihu, China. *Journal of Environmental Sciences*, 23,  
662222–227

663

664

665**Figures:**

666



667

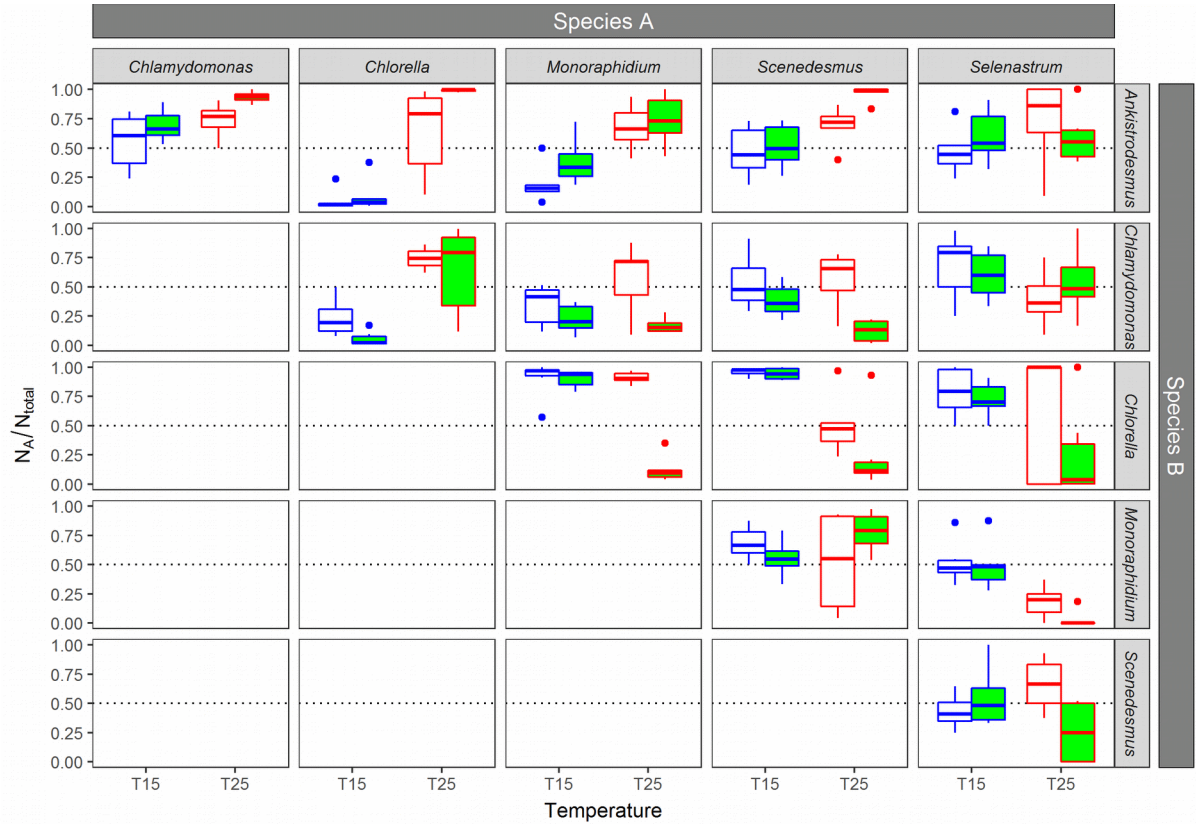
668**Fig 1:**

669 **(a)** Mean Monod curves for each species growth rate. Growth rate  $\mu$  as a function of  
670 phosphate concentration in the medium ( $\mu\text{mol}\cdot\text{L}^{-1}$ ) and temperature (from blue: 15°C to dark  
671 red: 35°C). Points represent the mean of the 3 replicates, and the Monod curve is drawn from  
672 the mean of the rate  $\mu_{\text{max}}$  and  $K_s$  parameters from the 3 replicates. Note that the  
673 phosphate concentration levels in the experiment go from 0.01 to 50  $\mu\text{mol}\cdot\text{L}^{-1}$  but the x-axis  
674 was cut at 8  $\mu\text{mol}\cdot\text{L}^{-1}$  for clarity. **(b)** Half-saturation coefficient  $K_s$  and **(b)** Maximum  
675 growth rate  $\mu_{\text{max}}$  as a function of temperature. **(c)** Half-saturation coefficient  $K_s$  Lines  
676 represent the fit of the GAM models investigating the temperature dependence of each  
677 parameter. See Tables S3A and S3B for more details about the temperature-dependence of the  
678 estimates from the Monod model.

679

680

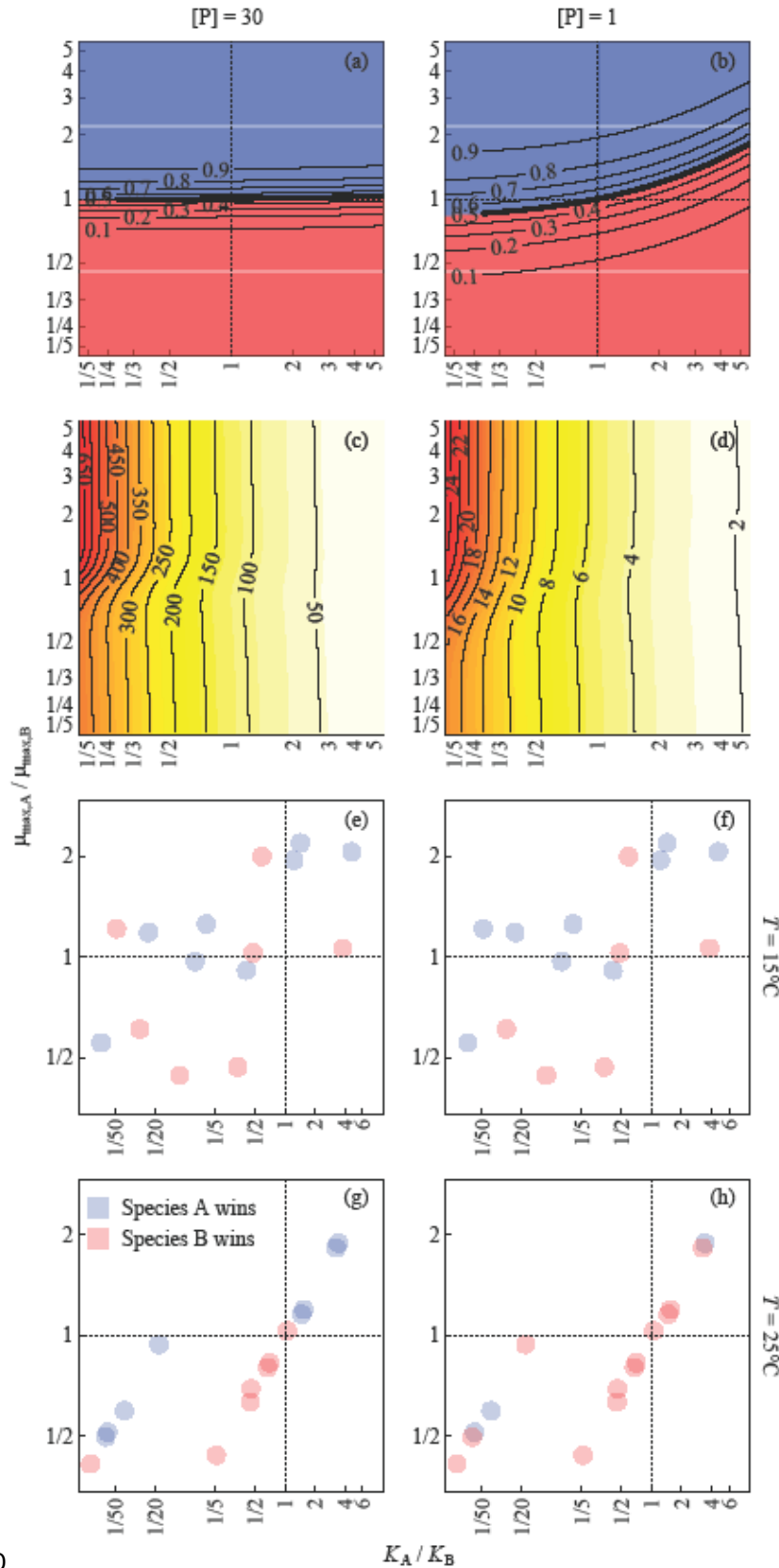
681



682

683**Fig. 2:**

684Competition between species. For each pair of species, the proportion of cells from species A  
685over the total number of cells at day 14. Colours represent the temperature of the trial, blue:  
68615°C, red: 25°C; and fills represent the nutrient conditions of the trial, white: non-saturated  
687nutrient solution (1  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate), green, saturated nutrient solution (30  $\mu\text{mol}\cdot\text{L}^{-1}$  of  
688phosphate). Boxplots represent the values of the 6 replicates per condition. The dotted line  
689represents the situation where 50% of the total number of cells pertain to the species A.



690

691**Fig 3:**

692 Relative importance of mismatches in  $\mu_{\max}$  and  $K_S$  in determining competitive  
 693 outcomes. Panels (a, b) show the proportion of cells belonging to species A after 14 days  
 694 according to the competition model, for a range of mismatches in both traits (see  
 695 Supplementary Section S6 for details). Panels (c, d) show the relative importance of a small

increase in the mismatches of the two traits on competitive outcomes. For example, a value of 10 means that a small increase in the  $\ln$  ratio of  $\mu_{\max}$  has a 10 times greater impact on the competitive outcome than does the same small increase in the  $\ln$  ratio of  $K_s$ . Panels (e-h) show the equivalent experimental results for  $T=15^\circ\text{C}$  (e, f) and  $25^\circ\text{C}$  (g, h), and competitive outcomes (colour of points) refer to the median of six replicates. Panels (a, c, e, g) are for a starting nutrient concentration of  $30\ \mu\text{mol}\cdot\text{L}^{-1}$  and (b, d, e, f) are for a starting concentration of  $1\ \mu\text{mol}\cdot\text{L}^{-1}$ . The legend in (g) applies to all panels except (c, d).

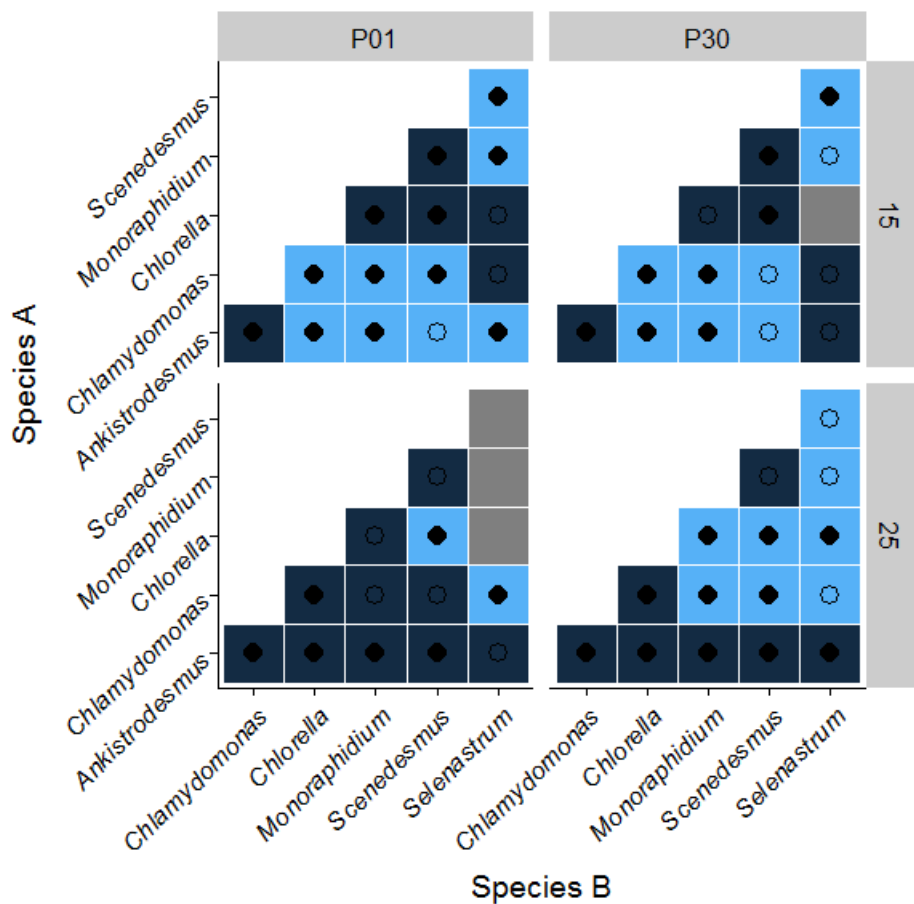


Figure 4: Outcome of the competition and validity of the model predictions for each pair of species depending on the temperature and phosphate level. The color indicates the identity of the winning species (lightblue species A wins, darkblue species B wins), and the shape inside shows whether the model prediction is correct (filled, the model is correct, empty, the model is incorrect).

## 711Tables

712

713**Table 1:** Proportion of competitive outcomes correctly predicted by mismatches in individual  
 714photosynthetic capacity (rP158), in  $\mu$  (the growth rate from the Buchanan model at each  
 715temperature and nutrient concentration combination),  $\mu_{max}$  (maximum growth rate at saturating  
 716nutrient concentrations from the Monod model),  $K_S$  (where a lower  $K_S$  is assumed to be  
 717beneficial), and both  $\mu_{max}$  and  $K_S$  (using the competition model), for all competitions, and by  
 718subsets (by temperature, by nutrient concentration, and by species where only competitions involving  
 719a specific species are included). Numbers in brackets show the proportion of 10,000 random runs with  
 720greater predictive power (see Supplementary Information Section S6). The experimental competition  
 721data uses the LDA discrimination method, and we here used the median value of the proportion of  
 722cells of a competitor across the six replicates. Monod model parameters (  $\mu_{max}$  and  $K_S$  ) are the  
 723parameter estimates from the mixed effects model described in Methods.

724

Subset	rP158	$\mu$	$\mu_{max}$	$K_S$	Model	N
<i>Full dataset</i>	0.70 (0.001)	0.62 (0.026)	0.60 (0.061)	0.40 (0.683)	0.65 (0.044)	60
<i>By temperature</i>						
$T=15^\circ\text{C}$	0.80 (0.000)	0.67 (0.021)	0.57 (0.175)	0.43 (0.489)	0.67 (0.085)	30
$T=25^\circ\text{C}$	0.60 (0.114)	0.57 (0.186)	0.63 (0.068)	0.30 (0.666)	0.63 (0.126)	30
<i>By nutrient concentration</i>						
$P=1\ \mu\text{mol}\cdot\text{L}^{-1}$	0.70 (0.009)	0.57 (0.160)	0.60 (0.062)	0.43 (0.525)	0.67 (0.032)	30
$P=30\ \mu\text{mol}\cdot\text{L}^{-1}$	0.70 (0.016)	0.67 (0.035)	0.60 (0.078)	0.37 (0.745)	0.63 (0.089)	30
<i>By species</i>						
<i>Ankistrodesmus</i>	0.90 (0.000)	0.45 (0.510)	0.85 (0.005)	0.35 (0.598)	0.80 (0.039)	20
<i>Chlamydomonas</i>	0.75 (0.006)	0.65 (0.046)	0.60 (0.114)	0.30 (0.769)	0.70 (0.046)	20
<i>Chlorella</i>	0.70 (0.059)	0.85 (0.003)	0.70 (0.087)	0.30 (0.698)	0.75 (0.087)	20
<i>Monoraphidium</i>	0.60 (0.077)	0.65 (0.027)	0.50 (0.241)	0.60 (0.061)	0.65 (0.068)	20
<i>Scenedesmus</i>	0.80 (0.000)	0.70 (0.004)	0.60 (0.059)	0.40 (0.553)	0.60 (0.125)	20
<i>Selenastrum</i>	0.45 (0.520)	0.40 (0.689)	0.35 (0.731)	0.45 (0.390)	0.40 (0.753)	20

725

## 726Table2:

727Link between competitive reversals and reversals in traits due to temperature. For each pair  
 728of species and each phosphate level, the sum of reversals observed in traits (  $\mu_{max}$  and  
 729  $K_S$  ) or competitive outcomes between 15 and 25°C. Out of the 18 competitive reversals  
 730observed between temperature levels, 8 were linked to reversals in  $\mu_{max}$  alone, 6 were  
 731linked to reversals in both traits  $\mu_{max}$  and  $K_S$  while 4 were not linked to any kind of  
 732reversals between nutrient physiology traits. For a more detailed description of the  
 733competitive outcomes, see Supplementary material S6

734

	Reversal in $\mu_{max}$		Reversal in $K_S$		Reversal in $\mu_{max}$ and $K_S$	
Reversal in	Yes (N = 20)	No (N = 10)	Yes (N = 10)	No (N = 20)	Yes (N = 10)	No (N = 20)

<b>competition</b>						
Yes (N = 18)	14	4	6	12	6	12
No (N = 12)	6	6	4	8	4	8

735

736



737 **Supplementary Information**

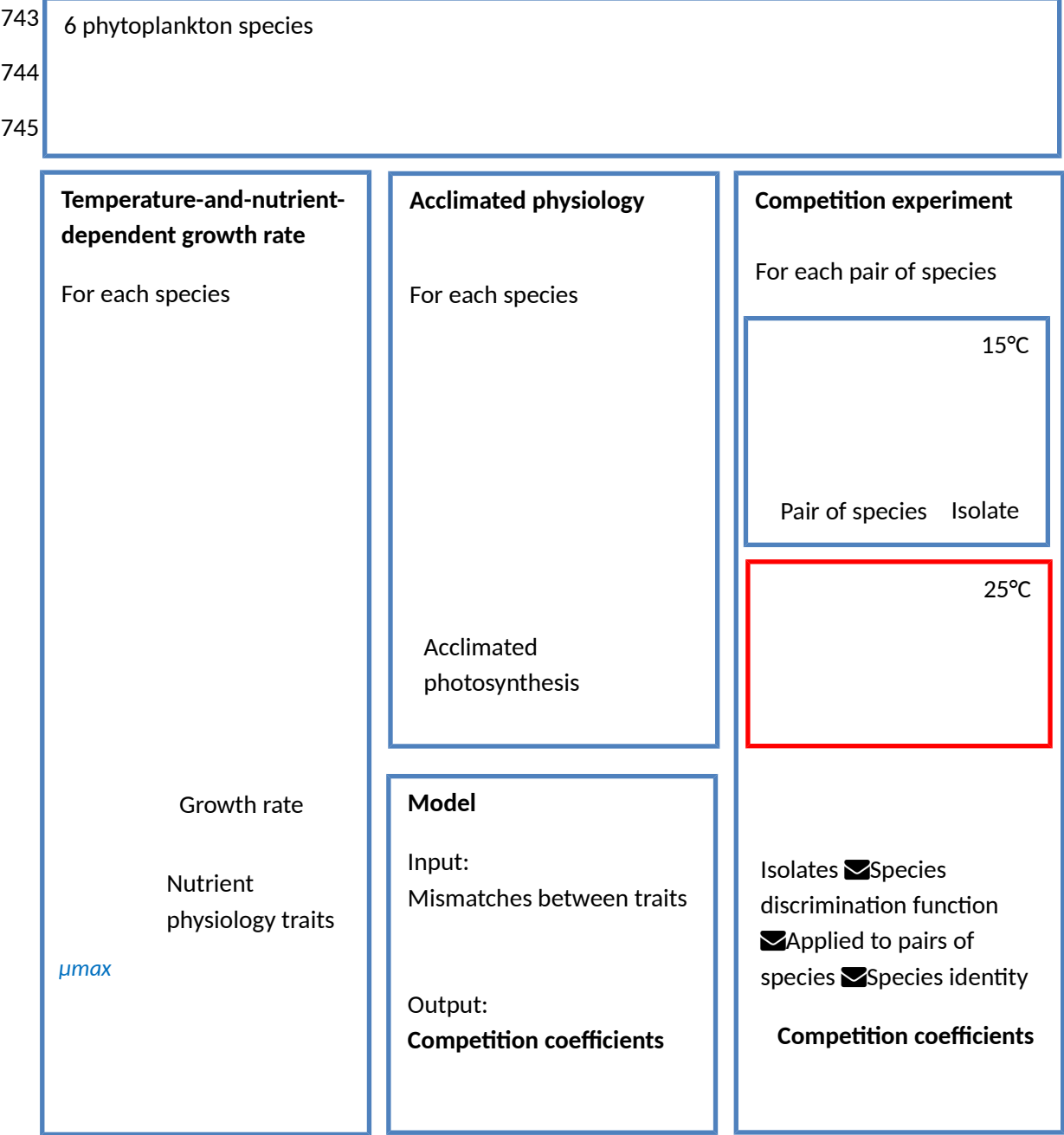
738

739 **S1: Experimental design**

740

741 **Figure S1A: Flow chart of the experimental design**

742



**746Table S1A:**

747Detailed information about the 6 species

Species name	Class	Order	Strain	Origin
<i>Ankistrodesmus nannoselene</i> Skuja (1948)	Chlorophyceae	Sphaeropleales	CCAP 202/6A	Siggeforsajon, Sweden
<i>Chlamydomonas moewusii</i> Gerlof (1940)	Chlorophyceae	Chlamydomonadales	CCAP 11/5A	Freshwater
<i>Chlorella sorokiniana</i> Shihira & Krauss (1965)	Trebouxiophyceae	Chlorellales	CCAP 211/8K	Austin, Texas, USA
<i>Monoraphidium minutum</i> (Nägeli) Komarkova-Legnerova (1969)	Chlorophyceae	Sphaeropleales	CCAP 278/3	Texas, USA
<i>Scenedesmus obliquus</i> (Turpin) Kützing (1833)	Chlorophyceae	Sphaeropleales	CCAP 276/3B	Lund, Sweden
<i>Selenastrum capricornutum</i> Printz (1913)	Chlorophyceae	Sphaeropleales	CCAP 278/4	Akershus, Norway

748

749

**750Table S2A:**

751Phosphate concentration levels for each solution in  $\mu\text{mol}\cdot\text{L}^{-1}$  and  $\mu\text{g}\cdot\text{L}^{-1}$ . We created 13  
752solutions of different phosphate concentrations ranging from 0.01  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate to  
75350  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate by mixing different amounts of COMBO medium without  
754potassium phosphate dibasic (P- COMBO) and normal COMBO medium (P+ COMBO) in  
75540 mL nuncnons. We used a modified version of the standard COMBO medium without  
756animal trace solution in which we increased the fraction of carbonate by adding 10 mL of a  
757stock solution of 55.8 g.  $\text{L}^{-1}$  of sodium bicarbonate to maintain a DIC of more than 6.6  
758 $\text{mmol}\cdot\text{L}^{-1}$  in order to prevent carbon limitation, which allowed a C:N:P ratio of 132:20:1 in  
759the P+ COMBO solution, above the Redfield ratio of 106:16:1.

760

Phosphate concentration ( $\mu\text{mol}\cdot\text{L}^{-1}$ )	50	40	30	20	10	8	6	4	2	1	0.5	0.1	0.01
Phosphate concentration ( $\mu\text{g}\cdot\text{L}^{-1}$ )	4750	3800	2850	1900	950	760	570	380	190	95	47.5	9.5	0.95
Amount of P+ COMBO (mL)	40	32	24	16	8	6.4	4.8	3.2	1.6	0.8	0.4	0.08	0.008
Amount of P- COMBO (mL)	0	8	16	24	32	33.6	35.2	36.8	38.4	39.2	39.6	40	40

761

## 762S2: Discrimination between species in the competition experiment

763

### 764Table S2A:

765A: Proportion of correct assignments for each discrimination method (LDA: linear  
766discriminant analysis, Random Forest analysis, RPART: recursive partitioning and regression  
767tree) summarised by phosphate and nutrient conditions for all pair of species. B: Proportion  
768of correct assignments for each discrimination method summarised by pair of species for all  
769nutrient and thermal conditions. C: Proportion of correct assignments for each discrimination  
770method summarised by species for all nutrient and thermal conditions

771A

Temperatur e	Nutrien t	LDA	Random forest	RPART
15	1	0.79	0.68	0.64
15	30	0.85	0.8	0.76
25	1	0.7	0.69	0.68
25	30	0.64	0.66	0.62
<b>Mean</b>		<b>0.75</b>	<b>0.71</b>	<b>0.68</b>

772

773B

Species	LDA	Randomforest	RPART
<i>Ankistrodesmus</i>	0.91	0.86	0.72
<i>Chlamydomonas</i>	0.93	0.93	0.81
<i>Chlorella</i>	0.85	0.86	0.67
<i>Monoraphidium</i>	0.84	0.78	0.65
<i>Scenedesmus</i>	0.83	0.77	0.61
<i>Selenastrum</i>	0.70	0.68	0.48
<b>Mean</b>	<b>0.84</b>	<b>0.81</b>	<b>0.66</b>

774

775

776C

<b>Pair of species</b>	<b>LDA</b>	<b>Randomfores t</b>	<b>RPAR T</b>
<i>Ankistrodesmus-Chlamydomonas</i>	1	1	0.94
<i>Ankistrodesmus-Chlorella</i>	0.91	0.88	0.73
<i>Ankistrodesmus-Monoraphidium</i>	0.87	0.74	0.71
<i>Ankistrodesmus-Scenedesmus</i>	0.95	0.93	0.71
<i>Ankistrodesmus-Selenastrum</i>	0.82	0.73	0.52
<i>Chlamydomonas-Chlorella</i>	0.96	0.96	0.79
<i>Chlamydomonas-Monoraphidium</i>	0.96	0.97	0.86
<i>Chlamydomonas-Scenedesmus</i>	0.94	0.92	0.74
<i>Chlamydomonas-Selenastrum</i>	0.78	0.8	0.74
<i>Chlorella-Monoraphidium</i>	0.83	0.85	0.7
<i>Chlorella-Scenedesmus</i>	0.86	0.84	0.65
<i>Chlorella-Selenastrum</i>	0.67	0.76	0.48
<i>Monoraphidium-Scenedesmus</i>	0.88	0.69	0.63
<i>Monoraphidium-Selenastrum</i>	0.68	0.67	0.33
<i>Scenedesmus-Selenastrum</i>	0.53	0.46	0.34
<b>Mean</b>	<b>0.84</b>	<b>0.81</b>	<b>0.66</b>

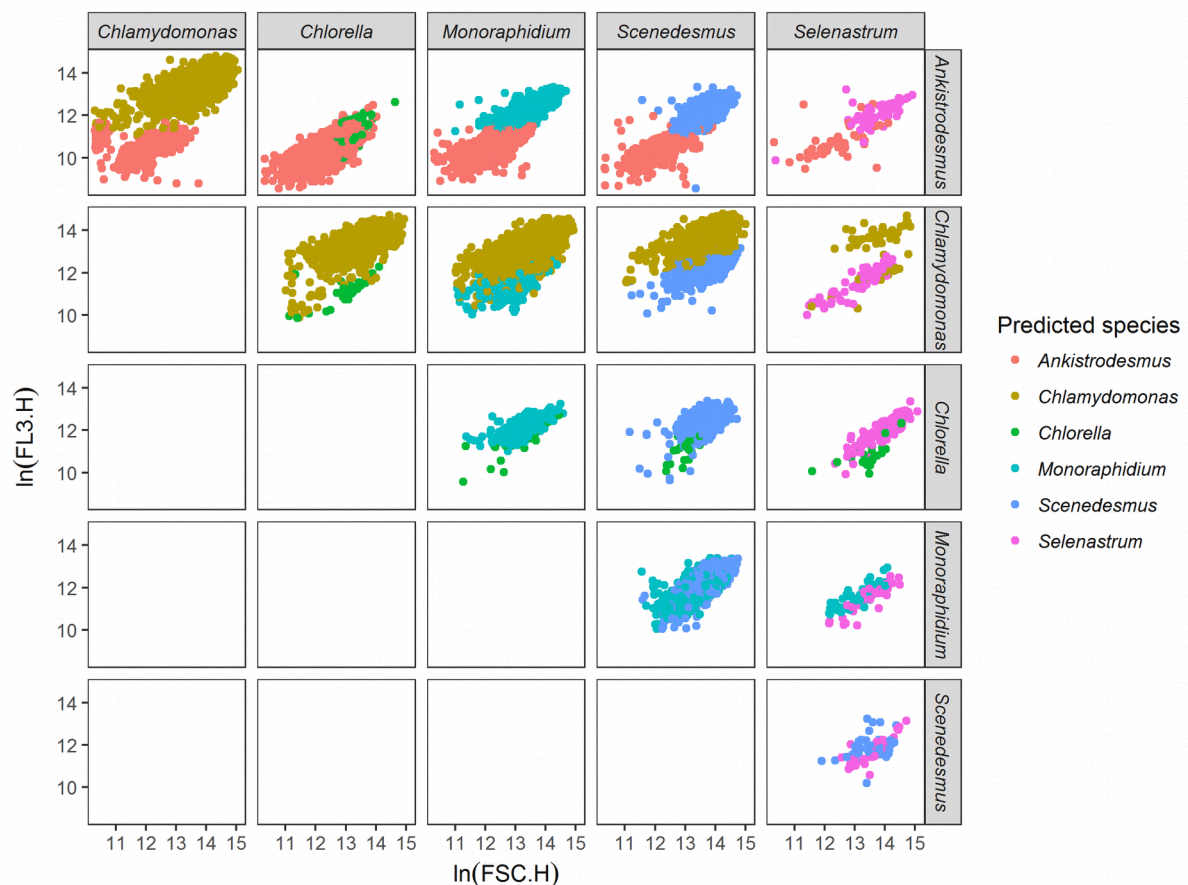
777

778

779

780

781



782

### 783Figure S2A:

784 Example of discrimination between species among pairs of species, here for species grown at  
 78515°C in saturating nutrient conditions after 14 days of experiment. Each dot represents a cell,  
 786here mapped on FSC.H (size proxy) and FL3.H (chlorophyll a proxy) characteristics from the  
 787flow cytometer. Colours represent the species predicted by the discrimination algorithm. The  
 788discrimination algorithm is a linear discriminant analysis trained with flow cytometer data  
 789(FSC.H, FSC.A, SSC.H, SSC.A, FL1.H, FL1.A, FL2.H, FL2.A, FL3.H, FL3.A, FL4.H, and  
 790FL4.A) from the species grown in isolates at the same temperature and nutrient conditions.  
 791For example, *Chlamydomonas* outcompetes *Chlorella* in these nutrient and temperature  
 792conditions.

793

### 794S3: Temperature dependency of the estimates from the Monod model

795

796

#### 797Table S3A:

798Results from the GAM models investigating  $\mu_{\max}$  as a function of temperature for each species.

799See Fig 1 for the representation of the GAM model.

800

Species	edf	F	p-value	R <sup>2</sup>
<i>Ankistrodesmus</i>	2	6.74	0.011*	0.45
<i>Chlamydomonas</i>	2	3.42	0.066.	0.26
<i>Chlorella</i>	2	54.02	>0.001***	0.88
<i>Monoraphidium</i>	2	63.67	>0.001***	0.90
<i>Scenedesmus</i>	2	0.41	0.674	-0.09
<i>Selenastrum</i>	2	5.82	0.017*	0.41

801

#### 802Table S3B:

803Results from the GAM models investigating  $K_s$  as a function of temperature for each species. See

804Fig 1 for the representation of the GAM model.

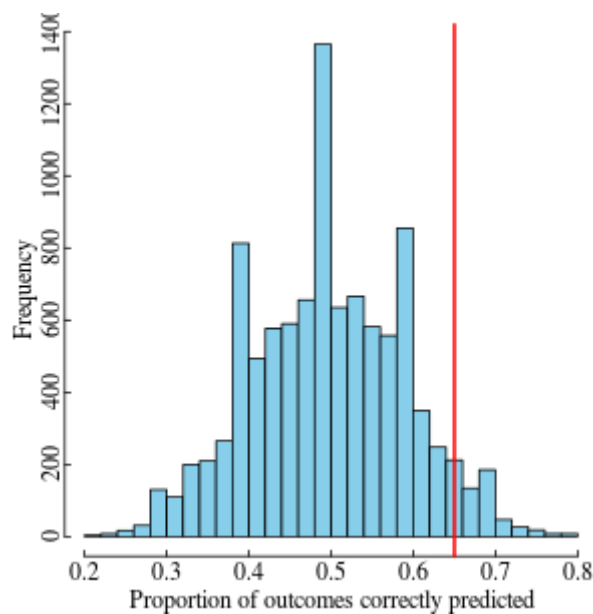
Species	edf	F	p-value	R <sup>2</sup>
<i>Ankistrodesmus</i>	2	6.29	0.013*	0.43
<i>Chlamydomonas</i>	2	4.53	0.034*	0.34
<i>Chlorella</i>	2	6.37	0.013*	0.43
<i>Monoraphidium</i>	2	2.17	0.157	0.14
<i>Scenedesmus</i>	2	1.32	0.302	0.04
<i>Selenastrum</i>	2	7.92	0.006**	0.50

805

806

## 807S5: Significance of competitive outcomes predicted by the model

808To quantify the significance of the model's ability to predict competitive outcomes, we ran the  
809competition model 10,000 times, sampling the values of  $\mu_{\max}$  and  $K_S$  independently with  
810replacement from the pool of available values. When assessing model performance for a particular  
811subset, for example, for competitions at  $T=15$  °C,  $\mu_{\max}$  and  $K_S$  were sampled  
812independently from all values at  $T=15$  °C only. The analysis produced 10,000 values of  
813proportion of competitive outcomes correctly predicted, for the 10,000 random parameter  
814combinations. Figure S1 shows an example distribution, for the full dataset. The proportion of runs  
815that correctly predicted a greater number of competitive outcomes than the model with the real  
816values of  $\mu_{\max}$  and  $K_S$  is then given as the P value in Table 1. Therefore,  $P=0.05$  means that  
817500 out of 10,000 random parameter combinations correctly predicted a greater proportion of  
818competitive outcomes.



819

### 820Figure S5A:

821Histogram of the proportion of competitive outcomes correctly predicted for the 10,000  
822random combinations of  $\mu_{\max}$  and  $K_S$ . The red line indicates the performance of the  
823model with the real values of  $\mu_{\max}$  and  $K_S$ . Here, there were 427 random parameter  
824combinations that correctly predicted a greater number of competitive outcomes ( $P = 0.04$  in  
825Table 1).

826

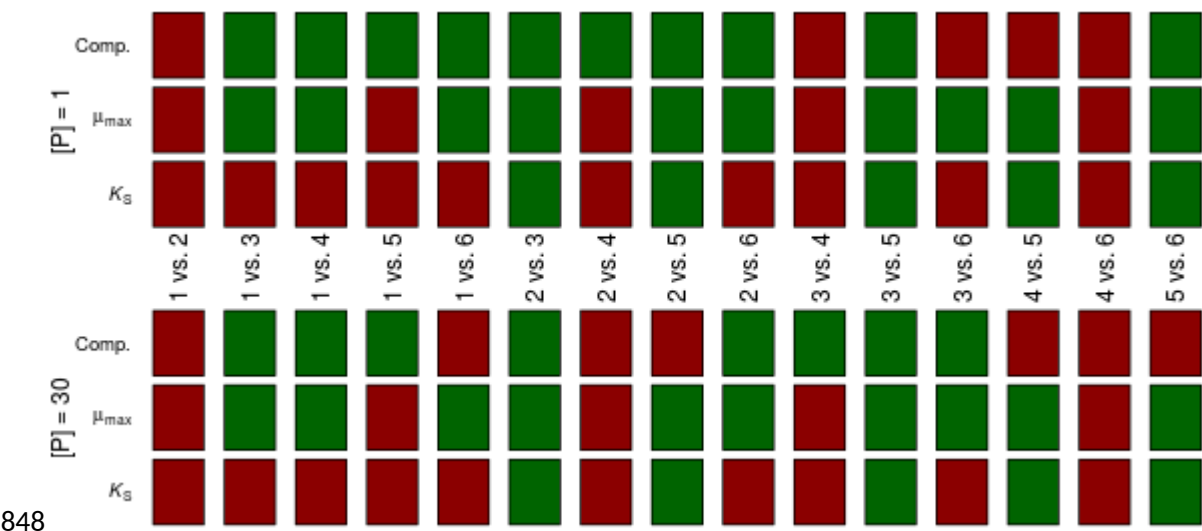


## 827 **S6. Simulations to compare the relative effects of mismatches**

828 We use the competition model to assess the relative importance of mismatches in  $\mu_{\max}$  and  
 829  $K_S$  for determining the competitive outcome (Figure 3 in the main text). Mismatches are  
 830 here defined as the ratios in the two traits between the two competitors. Ratios allow for  
 831 direct comparison of the relative importance of mismatches in  $\mu_{\max}$  and  $K_S$  despite the  
 832 different units. We assumed one competitor to have values of  $\mu_{\max}$  and  $K_S$  close to the  
 833 respective median values across all species and treatments (  $\mu_{\max}=1$  and  $K_S=0.15$   
 834 respectively), while the second competitor's  $\mu_{\max}$  and  $K_S$  values were chosen such that  
 835 the ln ratio parameter space was evenly sampled. Results were insensitive to the choice of  
 836 values for the fixed competitor. The reason to use ln ratios is to ensure that a ratio and its  
 837 inverse are equidistant from a ratio of one. For all combinations of mismatches in traits we  
 838 ran the competition model and extracted the proportion of total cells belonging to competitor  
 839 A at day 14, for different starting nutrient concentrations (Figure 3a, b in the main text). As in  
 840 the experiments, both species had a starting population density of 100 cells·mL<sup>-1</sup>.

841 To compare the relative importance of mismatches in the two traits directly, we quantified by  
 842 how much the competitive outcome changed due to a small increase in the ln ratio of  $\mu_{\max}$   
 843 and due to the same small increase in the ln ratio of  $K_S$ , and took their ratio. For example,  
 844 a value of 10 means that a small increase in the mismatch in  $\mu_{\max}$  had a 10 times greater  
 845 effect on the competitive outcome than did the same small increase in the mismatch in  $K_S$   
 846 (Figure 3c, d in the main text).

847



849 **Figure S6A:**

850 Reversals of competitive outcomes and traits due to temperature. Red boxes mean no reversal,  
 851 while green boxes mean a reversal was observed between 15°C and 25°C. The different columns are  
 852 for different competitions (e.g., species 1 vs. species 2). 'Comp.' refers to the experimentally  
 853 observed competitive outcome (using the LDA discrimination method), while  $\mu_{\max}$  and  $K_S$   
 854 refer to species' traits that define the Monod curve (estimated using the mixed effects model).  
 855 Results are shown for the two nutrient concentrations separately. For example, for species 1 vs.  
 856 species 2, no reversals were observed in competitive outcomes or traits, while for species 1 vs.  
 857 species 3, a reversal was observed in both the competitive outcome and  $\mu_{\max}$  (at both nutrient

858 concentrations), but not in  $K_S$ . There were 18 reversals in competitive outcomes, of which 14  
859 coincided with reversals in mismatches in  $\mu_{max}$ , and six with reversals in mismatches in  $K_S$ .  
860 Numbers represent the identity of the species: 1 = *Ankistrodesmus*, 2 = *Chlamydomonas*, 3 =  
861 *Chlorella*, 4 = *Monoraphidium*, 5 = *Scenedesmus* and 6 = *Selenastrum*.

862

## 863 **S7: Robustness of the results to different statistical methods**

864

865 Estimates for  $K_S$  and  $\mu_{max}$  for the Monod model were obtained from a non-linear mixed  
866 model approach with the 'nlme' function in R, and were then used in models investigating  
867 competitive outcome throughout the manuscript. To test the robustness of the model to the  
868 method of determination of  $K_S$  and  $\mu_{max}$ , we also fitted the Monod model to each species  
869 and temperature level using the 'nlsLM' function in the 'minpack.lm' package. Parameter  
870 estimation was achieved by running 1000 different random combination of starting parameters  
871 picked from a uniform distribution and returning the parameter set that returned the lowest AIC  
872 score. The two modelling approaches gave concordant results (Fig S7A). Thus in a second time,  
873 we used fits from this latter approach to feed in the later competition modelling (Table S7A).  
874 The results were extremely similar, with a slightly higher variance on the effect of  
875  $K_S$ , which did not affect the predictive power of the model overall.

876 A second source of uncertainty was due to the method of discrimination between species. We  
877 used three different methods of discrimination, a linear discriminant analysis, a randomforest  
878 analysis and a recursive partitioning and regression tree (rpart, see Supplementary Material  
879 S2). Because the linear discriminant analysis was found to have the best predictive power  
880 overall (Table S2A), we used this method throughout the manuscript. However, we tested  
881 whether our results were robust to the method of species discrimination by comparing results  
882 from the competition model to predictions using the randomforest analysis and the rpart  
883 discrimination method, first with the mixed effect parameters from the Monod model (Table  
884 S7B and S7D), and second with the Monod parameters estimated using nonlinear least squares  
885 (Table S7C and S7E). The results were similar, with a lower predictive power of each variable  
886 and of the model due to the lower discrimination power of the two methods, but no significant  
887 discrepancies between species and temperature and nutrient conditions.

888 We also redid the figures comparing the relative effects of the mismatches (Fig S6A) with the  
889 Monod parameters estimated using nonlinear least squares or mixed effect model and the three  
890 discrimination analysis (LDA, rpart and randomforest, Figs S7B-S7F). The results were generally  
891 congruent.

892

893

894

895

**896Table S7A:**

897Same as Table 1 in the main text, using the LDA discrimination method for the competition data, and  
 898Monod parameters estimated using nonlinear least squares.

899

Subset	rP158	$\mu$	$\mu_{max}$	$K_S$	Model	N
<i>Full dataset</i>	0.70 (0.001)	0.62 (0.025)	0.60 (0.059)	0.37 (0.790)	0.65 (0.041)	60
<i>By temperature</i>						
$T=15\text{ }^{\circ}\text{C}$	0.80 (0.000)	0.67 (0.024)	0.57 (0.172)	0.43 (0.498)	0.67 (0.082)	30
$T=25\text{ }^{\circ}\text{C}$	0.60 (0.115)	0.57 (0.183)	0.63 (0.066)	0.30 (0.827)	0.63 (0.125)	30
<i>By nutrient concentration</i>						
$P=1\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.70 (0.008)	0.57 (0.160)	0.60 (0.060)	0.40 (0.654)	0.67 (0.035)	30
$P=30\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.70 (0.018)	0.67 (0.035)	0.60 (0.083)	0.33 (0.830)	0.63 (0.089)	30
<i>By species</i>						
<i>Ankistrodesmus</i>	0.90 (0.000)	0.45 (0.502)	0.85 (0.003)	0.35 (0.594)	0.80 (0.037)	20
<i>Chlamydomonas</i>	0.75 (0.004)	0.65 (0.050)	0.60 (0.112)	0.30 (0.768)	0.70 (0.047)	20
<i>Chlorella</i>	0.70 (0.056)	0.85 (0.004)	0.70 (0.087)	0.30 (0.695)	0.75 (0.085)	20
<i>Monoraphidium</i>	0.60 (0.084)	0.65 (0.026)	0.50 (0.237)	0.50 (0.239)	0.65 (0.064)	20
<i>Scenedesmus</i>	0.80 (0.000)	0.70 (0.004)	0.60 (0.062)	0.30 (0.833)	0.60 (0.128)	20
<i>Selenastrum</i>	0.45 (0.512)	0.40 (0.699)	0.35 (0.737)	0.45 (0.389)	0.40 (0.757)	20

900

901

**902Table S7B:**

903Same as Table 1 in the main text, using the rpart discrimination method for the competition data, and  
 904Monod parameters estimated using the mixed effects model.

905

Subset	rP158	$\mu$	$\mu_{max}$	$K_S$	Model	N
<i>Full dataset</i>	0.65 (0.010)	0.63 (0.015)	0.52 (0.216)	0.47 (0.405)	0.62 (0.065)	60
<i>By temperature</i>						
$T=15\text{ }^{\circ}\text{C}$	0.77 (0.000)	0.63 (0.051)	0.53 (0.255)	0.47 (0.413)	0.63 (0.139)	30
$T=25\text{ }^{\circ}\text{C}$	0.53 (0.240)	0.63 (0.058)	0.50 (0.305)	0.47 (0.399)	0.60 (0.145)	30
<i>By nutrient concentration</i>						
$P=1\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.63 (0.049)	0.63 (0.047)	0.47 (0.360)	0.53 (0.172)	0.63 (0.065)	30
$P=30\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.67 (0.032)	0.63 (0.069)	0.57 (0.138)	0.40 (0.638)	0.60 (0.145)	30
<i>By species</i>						
<i>Ankistrodesmus</i>	0.85 (0.000)	0.50 (0.366)	0.80 (0.006)	0.40 (0.523)	0.75 (0.049)	20
<i>Chlamydomonas</i>	0.70 (0.023)	0.70 (0.024)	0.55 (0.195)	0.35 (0.649)	0.65 (0.092)	20
<i>Chlorella</i>	0.70 (0.045)	0.85 (0.001)	0.60 (0.197)	0.35 (0.589)	0.70 (0.100)	20
<i>Monoraphidium</i>	0.55 (0.179)	0.60 (0.082)	0.45 (0.385)	0.65 (0.019)	0.60 (0.123)	20

<i>Scenedesmus</i>	0.75 (0.003)	0.75 (0.004)	0.45 (0.411)	0.55 (0.182)	0.55 (0.291)	20
<i>Selenastrum</i>	0.35 (0.770)	0.40 (0.609)	0.25 (0.855)	0.50 (0.262)	0.45 (0.487)	20

906

907**Table S7C :**

908Same as Table 1 in the main text, using the rpart discrimination method for the competition data, and  
 909Monod parameters estimated using nonlinear least squares.

910

Subset	rP158	$\mu$	$\mu_{max}$	$K_S$	Model	N
<i>Full dataset</i>	0.65 (0.009)	0.63 (0.015)	0.52 (0.221)	0.43 (0.532)	0.62 (0.071)	60
<i>By temperature</i>						
$T=15\text{ }^{\circ}\text{C}$	0.77 (0.001)	0.63 (0.045)	0.53 (0.254)	0.47 (0.411)	0.63 (0.134)	30
$T=25\text{ }^{\circ}\text{C}$	0.53 (0.245)	0.63 (0.058)	0.50 (0.307)	0.40 (0.581)	0.60 (0.161)	30
<i>By nutrient concentration</i>						
$P=1\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.63 (0.043)	0.63 (0.044)	0.47 (0.367)	0.50 (0.250)	0.63 (0.065)	30
$P=30\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.67 (0.036)	0.63 (0.068)	0.57 (0.139)	0.37 (0.740)	0.60 (0.143)	30
<i>By species</i>						
<i>Ankistrodesmus</i>	0.85 (0.000)	0.50 (0.367)	0.80 (0.005)	0.40 (0.526)	0.75 (0.047)	20
<i>Chlamydomonas</i>	0.70 (0.023)	0.70 (0.025)	0.55 (0.198)	0.35 (0.657)	0.65 (0.093)	20
<i>Chlorella</i>	0.70 (0.042)	0.85 (0.001)	0.60 (0.192)	0.35 (0.585)	0.70 (0.103)	20
<i>Monoraphidium</i>	0.55 (0.182)	0.60 (0.082)	0.45 (0.389)	0.55 (0.111)	0.60 (0.122)	20
<i>Scenedesmus</i>	0.75 (0.004)	0.75 (0.004)	0.45 (0.412)	0.45 (0.413)	0.55 (0.296)	20
<i>Selenastrum</i>	0.35 (0.771)	0.40 (0.608)	0.25 (0.855)	0.50 (0.263)	0.45 (0.484)	20

911

912

913

914**Table S7D:**

915Same as Table 1 in the main text, using the random forests discrimination method for the competition  
 916data, and Monod parameters estimated using the mixed effects model.

917

Subset	rP158	$\mu$	$\mu_{max}$	$K_S$	Model	N
<i>Full dataset</i>	0.70 (0.001)	0.62 (0.032)	0.57 (0.118)	0.43 (0.559)	0.68 (0.021)	60
<i>By temperature</i>						
$T=15\text{ }^{\circ}\text{C}$	0.80 (0.000)	0.67 (0.023)	0.57 (0.172)	0.43 (0.489)	0.67 (0.084)	30
$T=25\text{ }^{\circ}\text{C}$	0.60 (0.130)	0.57 (0.199)	0.57 (0.196)	0.43 (0.502)	0.70 (0.058)	30
<i>By nutrient concentration</i>						
$P=1\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.67 (0.039)	0.60 (0.119)	0.50 (0.312)	0.53 (0.221)	0.70 (0.027)	30
$P=30\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.73 (0.003)	0.63 (0.058)	0.63 (0.032)	0.33 (0.844)	0.67 (0.038)	30
<i>By species</i>						
<i>Ankistrodesmus</i>	0.90 (0.000)	0.45 (0.512)	0.85 (0.005)	0.35 (0.602)	0.80 (0.041)	20
<i>Chlamydomonas</i>	0.80 (0.000)	0.60 (0.100)	0.65 (0.035)	0.25 (0.891)	0.75 (0.011)	20
<i>Chlorella</i>	0.70 (0.066)	0.85 (0.005)	0.60 (0.203)	0.40 (0.516)	0.75 (0.087)	20
<i>Monoraphidium</i>	0.60 (0.088)	0.65 (0.028)	0.50 (0.231)	0.60 (0.062)	0.65 (0.062)	20
<i>Scenedesmus</i>	0.80 (0.001)	0.80 (0.001)	0.50 (0.297)	0.50 (0.291)	0.60 (0.190)	20

<i>Selenastrum</i>	0.40 (0.696)	0.35 (0.843)	0.30 (0.817)	0.50 (0.261)	0.55 (0.280)	20
--------------------	--------------	--------------	--------------	--------------	--------------	----

918

919

920**Table S7E:**

921Same as Table 1 in the main text, using the random forests discrimination method for the competition  
922data, and Monod parameters estimated using nonlinear least squares.

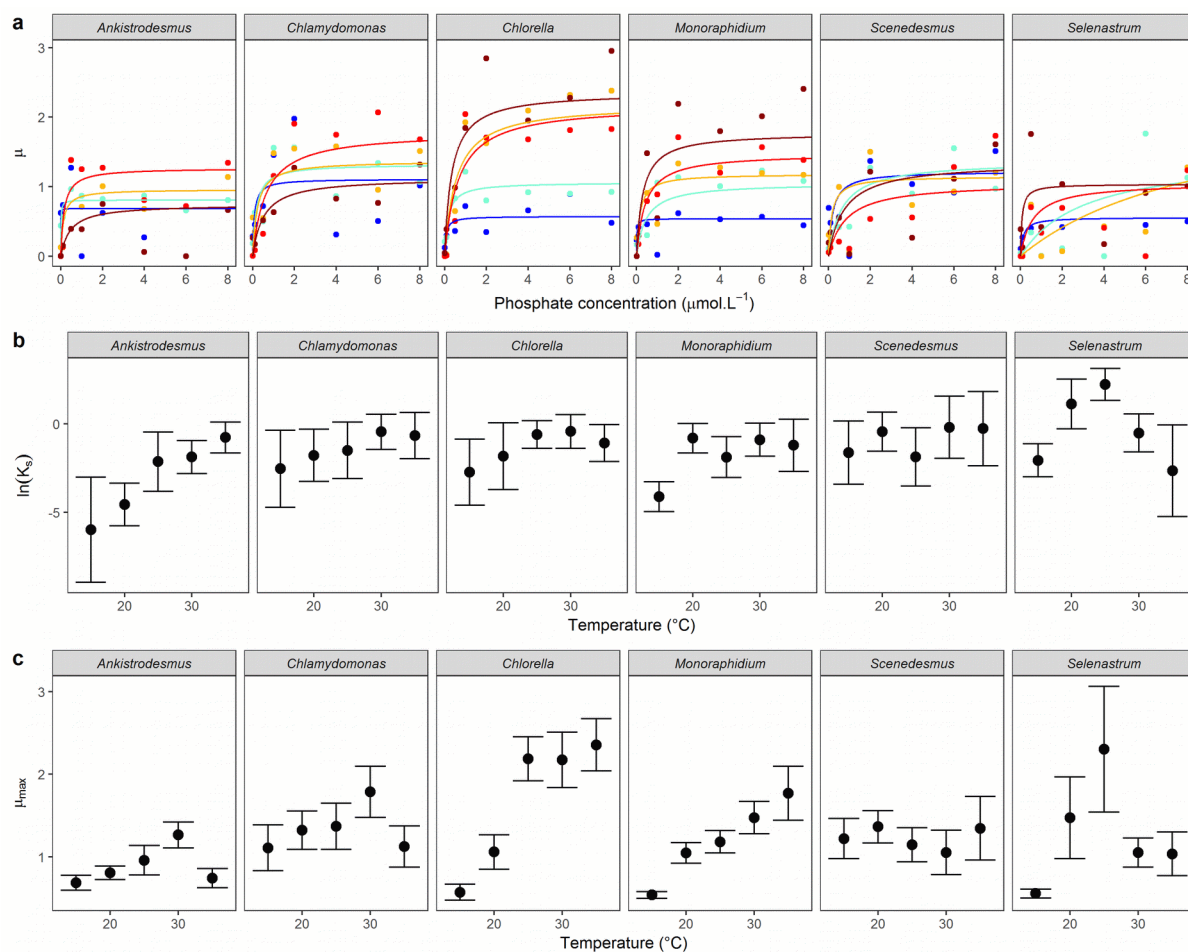
923

<b>Subset</b>	<b>rP158</b>	<b><math>\mu</math></b>	<b><math>\mu_{max}</math></b>	<b><math>K_s</math></b>	<b>Model</b>	<b>N</b>
<i>Full dataset</i>	0.70 (0.002)	0.62 (0.034)	0.57 (0.122)	0.40 (0.674)	0.68 (0.020)	60
<i>By temperature</i>						
$T=15\text{ }^{\circ}\text{C}$	0.80 (0.000)	0.67 (0.025)	0.57 (0.179)	0.43 (0.483)	0.67 (0.087)	30
$T=25\text{ }^{\circ}\text{C}$	0.60 (0.128)	0.57 (0.193)	0.57 (0.202)	0.37 (0.659)	0.70 (0.063)	30
<i>By nutrient concentration</i>						
$P=1\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.67 (0.038)	0.60 (0.118)	0.50 (0.314)	0.50 (0.319)	0.70 (0.028)	30
$P=30\text{ }\mu\text{mol}\cdot\text{L}^{-1}$	0.73 (0.004)	0.63 (0.057)	0.63 (0.035)	0.30 (0.908)	0.67 (0.040)	30
<i>By species</i>						
<i>Ankistrodesmus</i>	0.90 (0.000)	0.45 (0.502)	0.85 (0.003)	0.35 (0.603)	0.80 (0.037)	20
<i>Chlamydomonas</i>	0.80 (0.000)	0.60 (0.094)	0.65 (0.040)	0.25 (0.889)	0.75 (0.010)	20
<i>Chlorella</i>	0.70 (0.062)	0.85 (0.002)	0.60 (0.196)	0.40 (0.523)	0.75 (0.085)	20
<i>Monoraphidium</i>	0.60 (0.076)	0.65 (0.027)	0.50 (0.237)	0.50 (0.242)	0.65 (0.062)	20
<i>Scenedesmus</i>	0.80 (0.001)	0.80 (0.000)	0.50 (0.295)	0.40 (0.535)	0.60 (0.195)	20
<i>Selenastrum</i>	0.40 (0.701)	0.35 (0.840)	0.30 (0.817)	0.50 (0.266)	0.55 (0.279)	20

924

925





926

# 927**Fig S7A:**

928 **(a)** Mean Monod curves for each species growth rate estimated using nonlinear least squares.  
 929 Growth rate  $\mu$  as a function of phosphate concentration in the medium ( $\mu\text{mol}\cdot\text{L}^{-1}$ ) and  
 930 temperature (from blue: 15°C to dark red: 35°C). Points represent the mean of the 3  
 931 replicates. Note that the phosphate concentration levels in the experiment go from 0.01 to 50  
 932  $\mu\text{mol}\cdot\text{L}^{-1}$  but the x-axis was cut at 8  $\mu\text{mol}\cdot\text{L}^{-1}$  for clarity. **(b)** Half-saturation coefficient  $K_s$   
 933 (mean  $\pm$  95%CI) **(c)** Maximum growth rate  $\mu_{\max}$  (mean  $\pm$  95%CI).

934

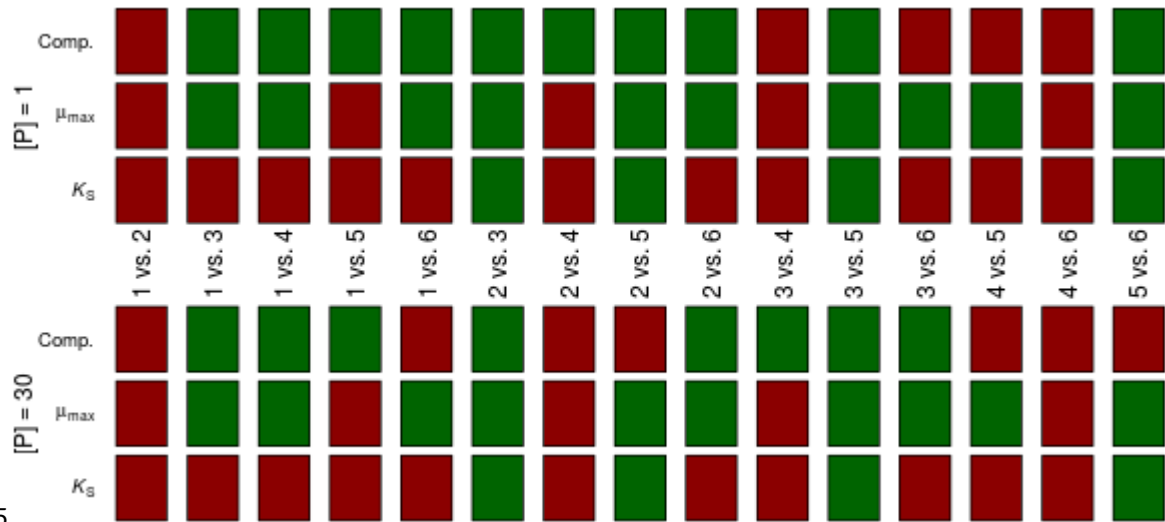


Figure S7B: Same as Figure SX, using the LDA discrimination method in the competition data, and nonlinear least squares parameter estimates of the Monod model for the traits. There were 18 reversals in competitive outcomes, of which 14 coincided with reversals in mismatches in  $\mu_{max}$ , and six with reversals in mismatches in  $K_S$ .

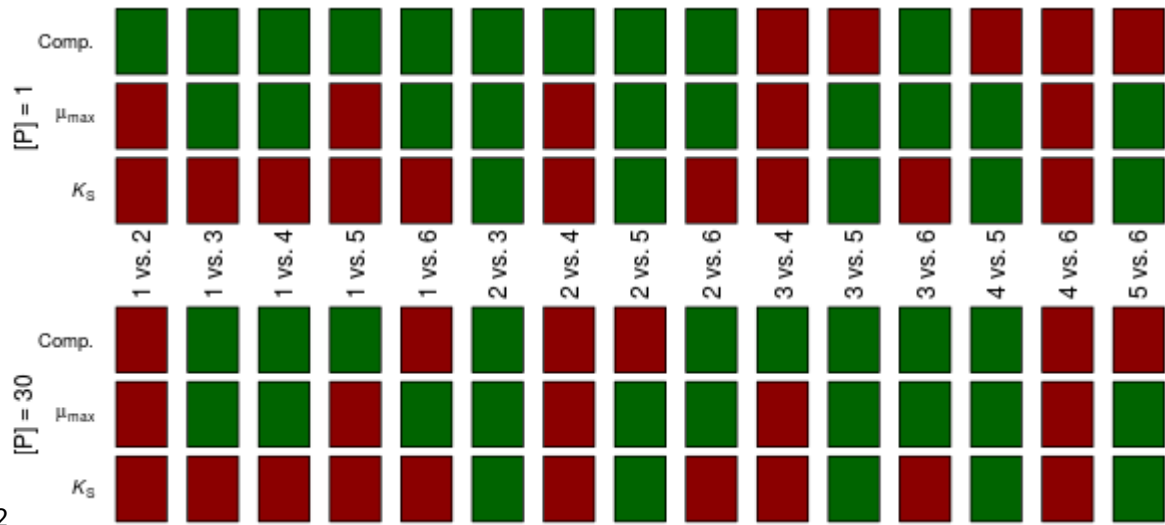


Figure S7C: Same as Figure SX, using the rpart discrimination method in the competition data, and parameter estimates from the mixed effects model of the Monod model for the traits. There were 19 reversals in competitive outcomes, of which 14 coincided with reversals in mismatches in  $\mu_{max}$ , and five with reversals in mismatches in  $K_S$ .

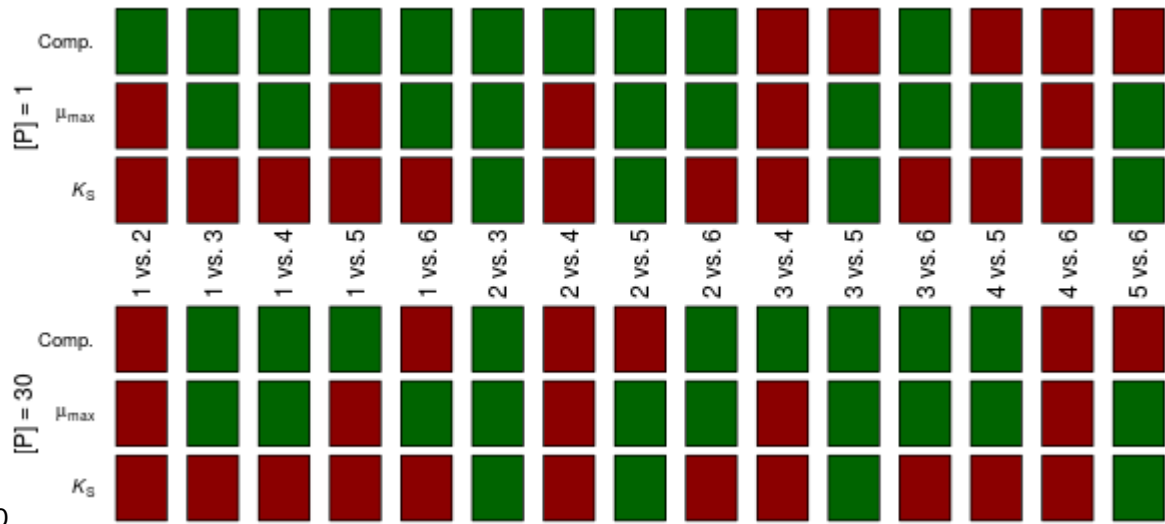


Figure S7D: Same as Figure SX, using the rpart discrimination method in the competition data, and parameter estimates from nonlinear least squares fit of the Monod model for the traits. There were 19 reversals in competitive outcomes, of which 14 coincided with reversals in mismatches in  $\mu_{max}$ , and four with reversals in mismatches in  $K_S$ .

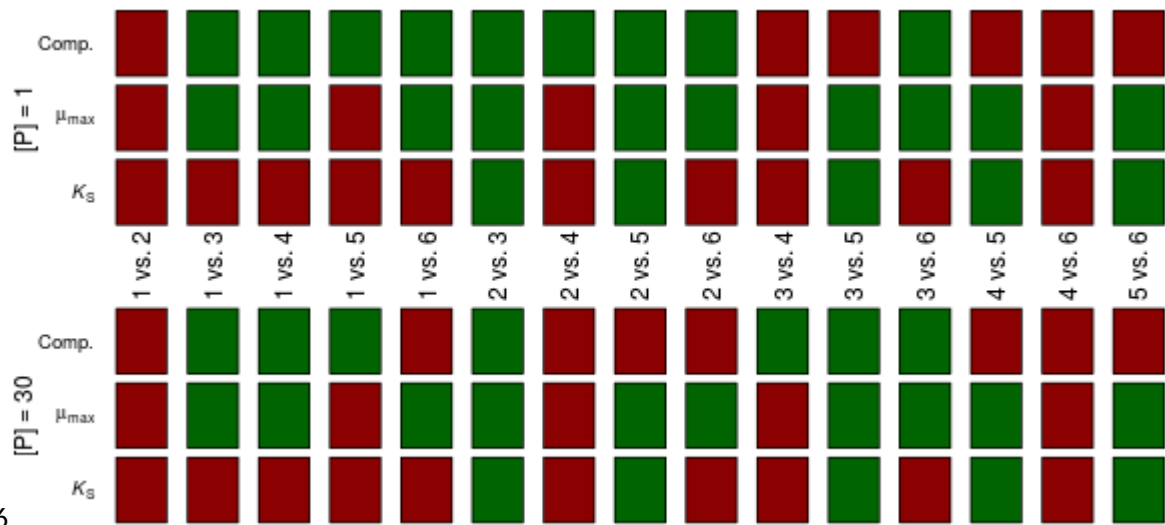
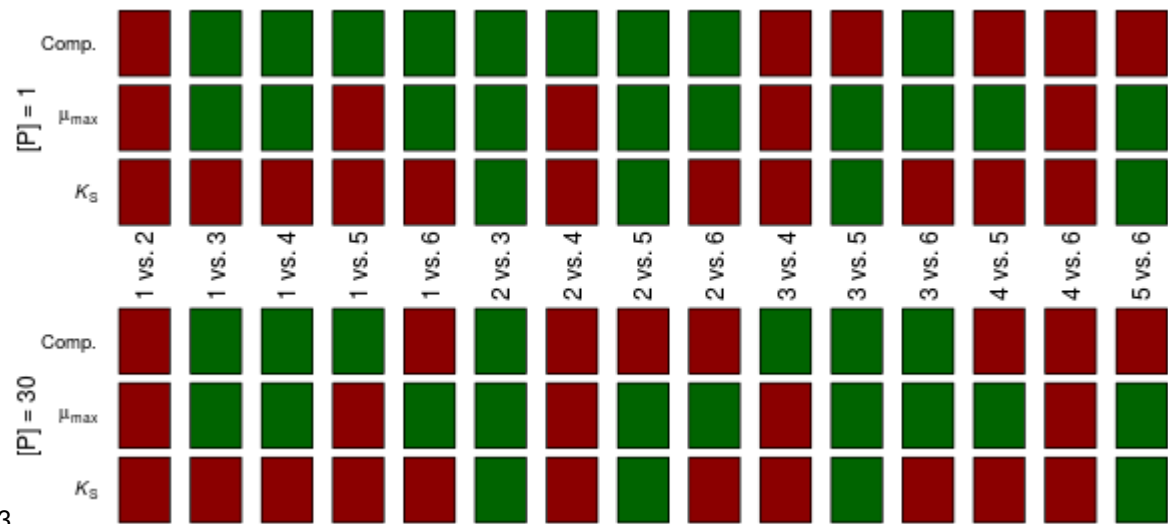


Figure S7E: Same as Figure SX, using the random forests discrimination method in the competition data, and parameter estimates from the mixed effects model of the Monod model for the traits. There were 16 reversals in competitive outcomes, of which 12 coincided with reversals in mismatches in  $\mu_{max}$ , and four with reversals in mismatches in  $K_S$ .



963

964Figure S7F: Same as Figure SX, using the random forests discrimination method in the  
 965competition data, and parameter estimates from nonlinear least squares fit of the Monod  
 966model for the traits. There were 16 reversals in competitive outcomes, of which 12 coincided  
 967with reversals in mismatches in  $\mu_{\max}$ , and four with reversals in mismatches in  $K_s$ .

968

969