A unified dataset of co-located sewage pollution, periphyton, and benthic macroinvertebrate community and food web structure from Lake Baikal (Siberia)

Michael F. Meyer1\*

Ted Ozersky2

Kara H. Woo3

Kirill Shchapov2

Aaron W. E. Galloway4

Julie B. Schram4

Daniel D. Snow5

Maxim A. Timofeyev6

Dmitry Yu. Karnaukhov6

Matthew R. Brousil2

Stephanie E. Hampton2

1. School of the Environment, Washington State University, Pullman, WA, USA

2. Large Lakes Observatory, University of Minnesota - Duluth, Duluth, MN, USA

3. Center for Environmental Research, Education, and Outreach, Washington State University,

Pullman, WA, USA

4. Oregon Institute of Marine Biology, University of Oregon, Charleston, OR, USA

5. School of Natural Resources, University of Nebraska-Lincoln, Lincoln, NE, USA

6. Biological Research Institute, Irkutsk State University, Irkutsk, Irkutsk Oblast, Russia

\*corresponding author: michael.f.meyer@wsu.edu

**Author Contribution Statement**

Conceptualized the project: MFM, SEH, TO

Collected samples in the field: MFM, TO, KHW, SEH

Processed samples: MFM, KS, JBS, DDS, TO, AWEG, SEH

Wrote and Reviewed R scripts: MFM, MRB, KHW

Data management: MFM, MRB

Wrote and edited the manuscript: All authors

Approved the final manuscript: All authors

**Grant sponsor information:**

Funding was provided by the National Science Foundation (NSF-DEB-1136637) to S.E.H., a Fulbright Fellowship to M.F.M., a NSF Graduate Research Fellowship to M.F.M. (NSF-DGE-1347973), and the Russian Ministry of Education and Science Research Project (No. GR 01201461929; 1354-2014/51).

**Key Words:** pharmaceuticals, microplastics, fatty acids, stables isotopes, amphipod, mollusk, diatom, spirogyra

**URL of the Dataset and Metadata with permanent identifier:**

EDI will provide the DOI

**Code URL with permanent identifier:**

EDI will provide the DOI

**Measurement(s):** Chlorophyll a, Fatty Acids, Pharmaceuticals and Personal Care Products, Microplastics, Periphyton community abundance, benthic macroinvertebrate abundance, Stable Isotopes, nitrate, ammonium, total phosphorus

**Technology Type(s): GC/MS, LC/MS, Spectrophotometry**

**Temporal range:** 19 – 23 August 2015

**Frequency or sampling interval:** single snapshot in time

**Spatial scale:** site-based

**Abstract (147 of 150 words)**

Sewage released from lakeside development can introduce nutrients and micropollutants that can restructure aquatic ecosystems. Lake Baikal, the world’s most ancient, biodiverse, and voluminous lake, has been experiencing localized sewage pollution from lakeside settlements. Nearby increases in filamentous algal abundance suggest benthic communities are responding. We surveyed 40-km of Baikal’s southwestern shoreline between 19-23 August 2015 for sewage indicators, including pharmaceuticals, personal care products, and microplastics with co-located periphyton and macroinvertebrate abundance, stable isotopes, and fatty acids. Unique identifiers corresponding to sampling locations are retained throughout all data files to facilitate interoperability among the dataset’s 150+ variables. The data are structured in a tidy format (a tabular arrangement familiar to limnologists) to encourage reuse. For Baikal studies, these data can support continued monitoring and research efforts. For global studies of lakes, these data can help characterize sewage prevalence and ecological consequences of anthropogenic disturbance across spatial scales.

**Background and Motivation**

Globally, sewage pollution is a common and often concentrated source of nitrogen and phosphorus inputs that can reshape aquatic ecosystems. Sewage inputs are often associated with increased primary production (Edmondson 1970; Moore et al. 2003), which can eventually lead to nuisance algal blooms (Hall et al. 1999; Lapointe et al. 2015). Even in instances where sewage pollution is mitigated, restoring systems can be complicated and necessitate system-specific (Jeppesen et al. 2005), long-term mitigation strategies (Hall et al. 1999; Tong et al. 2020). As such, effective sewage monitoring can require merging a suite of chemical, biological, and ecological data to synthesize locations and timing of inputs with associated shifts in ecological communities (Rosenberger et al. 2008; Hampton et al. 2011).

Definitively identifying sewage as the source of excess nutrients in a system can be challenging. Nutrients can originate from multiple sources, such as agriculture (Powers et al. 2016) or melting permafrost (Turetsky et al. 2000), which can obfuscate wastewater signals. Unlike nutrients, sewage-specific micropollutants, such as enhanced δ15N stable isotope signatures (Costanzo et al. 2001; Camilleri and Ozersky 2019), pharmaceuticals and personal care products (PPCPs) (Bendz et al. 2005; Rosi-Marshall and Royer 2012; Meyer et al. 2019) and microplastics (Barnes et al. 2009), can be highly specific to human wastewater. Accordingly, sewage-associated micropollutants have garnered global attention for their usefulness in identifying presence and quantifying magnitude of wastewater inputs. While indicators may accumulate differentially in certain taxa (Gartner et al. 2002; Green 2016; Vendel et al. 2017; Richmond et al. 2018), acutely dangerous concentrations are not common in most aquatic systems (Kolpin et al. 2002; Focazio et al. 2008; Yang et al. 2018). However, chronic exposure to microplastics and PPCPs at even minute concentrations (e.g., µg/L) can still disrupt ecological processes (Richmond et al. 2017). For example, oxazepam can increase feeding rate and decrease sociability of river perch (Brodin et al. 2013), and microplastics cam release dissolved organic carbon, thereby altering microbial communities (Romera-Castillo et al. 2018). The pervasiveness and diversity of sewage-associated micropollutants in tandem with their potency as ecologically disrupting compounds necessitates investigation within and across systems, thereby enabling synthesis of how micropollutants alter ecosystems.

When assessing biological responses to increased nutrient loading, littoral benthic algal, and macroinvertebrate communities often respond most markedly, as their physical proximity to the shoreline puts them in the path of sewage pollution entering the lake (Rosenberger et al. 2008; Hampton et al. 2011). Filamentous algae, for example, can quickly increase abundance near sewage loading (Rosenberger et al. 2008; Hampton et al. 2011). As algal communities change, food webs can also restructure to include different feeding groups, such as detritivores, or with altered nutritional value of primary producers. Among the suite of food quality metrics, availability of essential fatty acids (EFAs) offers a nuanced understanding of food quality as primary producers usually maintain a consistent EFA signatures (Taipale et al. 2013) and consumers acquire EFAs by grazing (Dalsgaard et al. 2003) or trophic upgrading (Sargent and Falk-Petersen 1988; Dalsgaard et al. 2003).

Together, food web structure, community composition, and sewage indicator data can be powerful tools to assess biological impacts of sewage pollution. Despite their utility, these data are not often available for many limnological systems. PPCPs, for example, have historically been less measured in lake environments (Meyer et al. 2019). In instances where data are available, efficiently merging disparate data into a single, analytically-friendly format can be challenging and require relatively complex, computationally intensive workflows (Meyer et al. 2020).

To offer a template for harmonizing sewage indicator and biological data, we present a unified data product, which contains disparate data collected from 14 littoral and 3 pelagic sites at Lake Baikal between 19 through 23 August 2015 (Figure 1). Located in Siberia, Lake Baikal is the oldest, most voluminous, and deepest freshwater lake in the world (Hampton et al. 2018). Baikal also has the global distinction of being the most biodiverse lake, with the highest endemism (Moore et al. 2009). The lake is experiencing rapid warming associated with climate change, including notable loss of winter ice cover (Moore et al. 2009), and it exhibits offshore plankton community changes associated with warming (Hampton et al. 2008, 2014, Katz et al. 2015, Izmest’eva et al. 2016). Less is known of the change occurring nearshore in Baikal, where warming is occurring and human activity also directly alters the environment. Nearshore change is particularly important to understand in Baikal, since the majority of the lake’s biodiversity and endemic species occur in the littoral zone (Kozhova and Izmest’eva 1998). While Lake Baikal’s pelagic zone is generally ultra-oligotrophic (Yoshida et al. 2003; O’Donnell et al. 2017), littoral areas abutting lakeside settlements have recently shown distinct signs of eutrophication (Timoshkin et al. 2016; Volkova et al. 2018).

As a means of identifying sewage from small, localized lakeside towns and associated ecological responses, the dataset consists of over 150 variables collected at 14 littoral and 3 pelagic sampling sites. We structured the dataset in a tidy format, where each row is a sample, each column is a variable, and each CSV file is an observable unit (Wickham 2014). Independent CSV files can be merged using unique locational identifiers as relational keys, enabling future researchers to customize analyses around a particular suite of variables. As a result of the dataset’s interoperability, reproducibility, and extensive variable content, it is well poised for future reuse as supporting evidence of sewage pollution in Lake Baikal. Additionally, the data’s flexibility and consistent structure enable it to be merged with similar datasets, so as to synthesize biological responses to sewage across systems and scales.

To our knowledge, no raw data on Lake Baikal macroinvertebrates, periphyton, or nearshore water quality are public in a machine-readable format, for any variable (i.e. abundance, fatty acid content, stable isotopes, nutrient and pollutant concentration), and no georeferenced data on pharmaceuticals and personal care products or microplastics appear to be publicly available for any boreal, subarctic, or arctic lakes or rivers in this region of the world. Thus, the dataset fills a substantial gap for future studies, providing a window into nearshore biotic assemblages and water quality in a unique, ancient ecosystem that holds 20% of the world’s liquid surface water (Moore et al. 2009).

**Data Description**

The final, replicate-level data products are available on the Environmental Data Initiative (EDI) repository, where they can be freely accessed without potential barriers such as paywalls or account registrations. The final data are provided as 11 separate CSV files, each structured in a tabular format and containing a “site” column that can be used to merge tables. The repository also contains a compressed folder of R scripts (scripts.tar.gz), which were used in the main analysis of the dataset (Meyer et al., 202X).

chlorophylla.csv

This file contains chlorophyll a concentrations as well as fluorometric corrections for each littoral and pelagic sampling location.

*site*

Unique alphanumeric identifier for a sampling location.

*replicate*

Replicate number.

*filtered\_volume\_ml*

Lake water volume filtered in milliliters for a given replicate.

*sample\_volume\_ml*

Sample volume filtered for chlorophyll a extraction.

*raw\_fluo*

Raw, uncorrected fluorometric reading for chlorophyll analysis.

*adjusted\_raw*

Corrected fluorometric reading for chlorophyll analysis.

*chl\_conc*

Chlorophyll a concentration in milligrams per liter.

distance\_weighted\_population\_metrics.csv

This file contains human population data for each sampled location. Although the majority of sites do not have adjacent shoreline human developments, we calculated inverse distance weighted (IDW) population for each sampling location. IDW population is a generalized representation of the size of and proximity to a sampling location’s neighboring human settlements. A full description of the methods used to calculate IDW population can be found in the companion manuscript Meyer et al. (20XX).

*site*

Unique alphanumeric identifier for a sampling location.

*distance\_weighted\_population*

Inverse distance weighted population for a given sampling location and estimated as number of people. Because this interpolation process is a function of the size of and proximity to neighboring developed sites, values can contain decimal values.

fatty\_acid.csv

This file contains fatty acid concentrations for various benthic macroinvertebrate genera, periphyton, and endemic *Draparnaldia* spp. collected from the 14 littoral sampling locations.

*site*

Unique alphanumeric identifier for a sampling location.

*Genus*

Genus of the analyzed organism.

*Species*

Species of the analyzed organism. When organism was identified solely to genus, the Species value is NA.

*c12\_0*

Concentration of 12:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*i\_14\_0*

Concentration of i-14:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c14\_0*

Concentration of 14:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c14**\_4w5*

Concentration of 14:4n-5 fatty acid as micrograms of fatty acid per milligram of tissue.

*i\_15\_0*

Concentration of i-15:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*a\_15\_0*

Concentration of a-15:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c15\_0*

Concentration of 15:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c15\_1w7*

Concentration of 15:1ω7 fatty acid as micrograms of fatty acid per milligram of tissue.

*i\_16\_0*

Concentration of i-16:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_0*

Concentration of 16:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_1w9*

Concentration of 16:1ω9 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_1w8*

Concentration of 16:1ω8 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_1w7*

Concentration of 16:1ω7 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_1w6*

Concentration of 16:1ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_1w5*

Concentration of 16:1ω5 fatty acid as micrograms of fatty acid per milligram of tissue.

*i\_17\_0*

Concentration of i-17:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*a\_17\_0*

Concentration of a-17:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c17\_0*

Concentration of 17:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c17\_1n7*

Concentration of 17:1n-7 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_2w7*

Concentration of 16:2ω7 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_2w6*

Concentration of 16:2ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_2w4*

Concentration of 16:2ω4 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_3w6*

Concentration of 16:3ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_3w4*

Concentration of 16:3ω4 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_3w3*

Concentration of 16:3ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_4w3*

Concentration of 16:4ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c16\_4w1*

Concentration of 16:4ω1 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_0*

Concentration of 18:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_1w9*

Concentration of 18:1ω9 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_1w7*

Concentration of 18:1ω7 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_2w6t*

Concentration of 18:2ω6t fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_2w6*

Concentration of 18:2ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_3w6*

Concentration of 18:3ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_3w3*

Concentration of 18:3ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_4w4*

Concentration of 18:4ω4 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_4w3*

Concentration of 18:4ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c18\_5w3*

Concentration of 18:5ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_0*

Concentration of 20:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_1w9*

Concentration of 20:1ω9 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_1w7*

Concentration of 20:1ω7 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_2w5\_11*

Concentration of 20:2-5-11 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_2w5\_13*

Concentration of 20:2-5-13 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_2w6*

Concentration of 20:2ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_3w6*

Concentration of 20:3ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_4w6*

Concentration of 20:4ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_3w3*

Concentration of 20:3ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_4w3*

Concentration of 20:4ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c20\_5w3*

Concentration of 20:5ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_0*

Concentration of 22:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_1w9*

Concentration of 22:1ω9 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_1w7*

Concentration of 22:1ω7 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_2w6*

Concentration of 22:2ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_4w6*

Concentration of 22:4ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_5w6*

Concentration of 22:5ω6 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_3w3*

Concentration of 22:3ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_4w3*

Concentration of 22:4ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_5w3*

Concentration of 22:5ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c22\_6w3*

Concentration of 22:6ω3 fatty acid as micrograms of fatty acid per milligram of tissue.

*c24\_0*

Concentration of 24:0 fatty acid as micrograms of fatty acid per milligram of tissue.

*comments*

Quality flag column. Two samples spilled during fatty acid extraction. These samples are flagged as such. Although concentrations are lower than other samples, proportions between fatty acids are consistent.

invertebrates.csv

This file contains abundance for benthic macroinvertebrates collected at each of the 14 littoral sampling locations. Only amphipod taxa were identified to species. Mollusks and isopods were identified to genus.

*site*

Unique alphanumeric identifier for a sampling location.

*replicate*

Replicate for sampling location. While three replicates were collected in the field, some samples were poorly preserved and not enumerated so as to prevent potential errors.

*Acroloxidae*

Mollusk genus

*Asellidae*

Endemic isopod genus

*Baicaliidae*

Mollusc genus, most of which are endemic

*Benedictidae*

Mollusc genus, most of which are endemic

*Brandtia\_latissima*

Endemic amphipod species

*Brandtia\_parasitica\_parasitica*

Endemic amphipod species

*Caddisflies*

General grouping; were not identified to species.

*Cryptoropus\_inflatus*

Endemic amphipod species

*Cryptoropus\_pachytus*

Endemic amphipod species

*Cryptoropus\_rugosus*

Endemic amphipod species

*Eulimnogammarus\_capreolus*

Endemic amphipod species

*Eulimnogammarus\_cruentes*

Endemic amphipod species

*Eulimnogammarus\_cyaneus*

Endemic amphipod species

*Eulimnogammarus\_grandimanus*

Endemic amphipod species

*Eulimnogammarus\_juveniles*

Endemic amphipod genus. Identification kept at genus level so as to prevent misclassification.

*Eulimnogammarus\_maackii*

Endemic amphipod species

*Eulimnogammarus\_marituji*

Endemic amphipod species

*Eulimnogammarus\_verucossus*

Endemic amphipod species

*Eulimnogammarus\_viridis\_viridis*

Endemic amphipod species

*Eulimnogammarus\_vittatus*

Endemic amphipod species

*Flatworms*

Not identified beyond order.

*Leeches*

Not identified beyond order, although 12 endemic species do exist.

*Maackia*

Mollusc genus, most of which are endemic

*Pallasea\_brandtia\_brandtia*

Endemic amphipod species

*Pallasea\_brandtii\_tenera*

Endemic amphipod species

*Pallasea\_cancelloides*

Endemic amphipod species

*Pallasea\_cancellus*

Endemic amphipod species

*Pallasea\_viridis*

Endemic amphipod species

*Planorbidae*

Mollusc genus, most of which are endemic

*Poekilogammarus\_crassimus*

Endemic amphipod species

*Poekilogammarus\_ephippiatus*

Endemic amphipod species

*Poekilogammarus\_juveniles*

Endemic amphipod genus. Identifying to species introduced risk of misclassification.

*Poekilogammarus\_megonychus\_perpolitus*

Endemic amphipod species

*Poekilogammarus\_pictus*

Endemic amphipod species

*Valvatidae*

Mollusk genus, most of which are endemic

metadata.csv

This file contains metadata for each of the pelagic and littoral sampling locations. Missing data are assigned as NA.

*site*

Unique alphanumeric identifier for a sampling location.

*year*

Year sampling occurred.

*month*

Month sampling occurred.

*day*

Day sampling occurred.

*time*

Time sampling occurred as Hours:Minutes.

*lat*

Latitude of sampling location in decimal degrees.

*long*

Longitude of sampling location in decimal degrees.

*site\_description*

Researchers’ description of sampling location at the time of sampling.

*distance\_to\_shore\_m*

Distance from *in situ* sampled location to the shoreline in meters.

*depth\_m*

Depth at *in situ* sampling location in meters.

*air\_temp\_celsius*

Temperature of air at sampling location in Celsius.

*surface\_temp\_celsius*

Temperature of water’s surface at sampling location in Celsius.

*mid\_temp\_celsius*

Temperature of water midway between surface and bottom at sampling location in Celsius.

*bottom\_temp\_celsius*

Temperature of water near sediment at sampling location in Celsius.

*comments*

Notes in the field describing sampling conditions.

*shore\_photo*

Whether or not photos of the shoreline were taken. Photos are available on the project’s Open Science Framework page (DOI).

*substrate\_photo*

Whether or not photos of the substrate were taken.

*sponges*

Whether or not sponges were present at a sampling location.

*brandtia*

Whether or not *Brandtia spp.* (endemic amphipod) was present at a sampling location.

microplastics.csv

This file contains microplastics counts for each of the pelagic and littoral sampling locations.

*site*

Unique alphanumeric identifier for a sampling location.

*replicate*

Replicate for a given sampling location. Replicate values of “C” indicate a control.

*fragments*

Number of microplastic fragments observed.

*fibers*

Number of microplastic fibers observed.

*beads*

Number of microplastic beads observed.

*comments*

Observer comments while counting microplastics

*volume\_filtered\_ml*

Volume in milliliters for a given replicate filtered.

nutrients.csv

This file contains nutrient concentrations for each of the associated sampling locations. Nutrient samples were not filtered prior to analysis, meaning that nitrogen concentrations have the potential to be spurious.

*site*

Unique alphanumeric identifier for a sampling location.

*replicate*

Replicate for a given sampling location.

*nh4\_mg\_dm3*

Ammonium concentration in milligrams of ammonium per cubic decimeter.

*no3\_mg\_dm3*

Nitrate concentration in milligrams of nitrate per cubic decimeter

*tp\_mg\_dm3*

Total phosphorus concentration in milligrams of phosphorus per cubic decimeter.

*tpo43\_mg\_dm3*

Total phosphate concentration as phosphate in milligrams per cubic decimeter.

periphyton.csv

This file contains periphyton abundance data for each of the sampled littoral locations. For poorly preserved samples, counts are listed as NA for each taxonomic grouping, and a note in the “comments” column is provided.

*site*

Unique alphanumeric identifier for a sampling location.

*replicate*

Replicate number for a given sampling site.

*subsamples\_counted*

Number of 10 microliter subsamples counted for a given replicate.

*diatom*

Number of diatom cells counted for a given replicate.

*spirogyra*

Number of *Spirogyra spp.* cells counted for a given replicate.

*spirogyra\_filament*

Number of *Spirogyra spp.* filaments counted for a given replicate.

*ulothrix*

Number of *Ulothrix spp.* cells counted for a given replicate.

*ulothrix\_filament*

Number of *Ulothrix spp.*filaments counted for a given replicate.

*tetrasporales*

Number of *Tetrasporales* spp. cells counted for a given replicate

*pediastrum*

Number of *Pediastrum spp.*cells counted for a given replicate.

*desmidales*

Number of *Desmidales* spp. cells counted for a given replicate.

*comments*

Notes from the observer.

ppcp.csv

This file contains Pharmaceutical and Personal Care Product (PPCP) concentrations for each littoral and pelagic sampling location.

*site*

Unique alphanumeric identifier for a sampling location.

*paraxanthine*

Concentration of paraxanthine, also known as 1,7-dimethylxanthine, in nanograms per liter. Paraxanthine is the main human metabolite of caffeine

*acetaminophen*

Concentration of acetaminophen, also known as paracetamol, in nanograms per liter.

*amphetamine*

Concentration of amphetamine in nanograms per liter.

*caffeine*

Concentration of caffeine in nanograms per liter.

*carbamazepine*

Concentration of carbamazepine in nanograms per liter.

*cimetidine*

Concentration of cimetidine in nanograms per liter.

*cotinine*

Concentration of cotinine, which is the main human metabolite of nicotine, in nanograms per liter.

*diphenhydramine*

Concentration of diphenhydramine in nanograms per liter.

*mda*

Concentration of methylene​dioxy​amphetamine in nanograms per liter.

*mdma*

Concentration of methyl​enedioxy​methamphetamine in nanograms per liter.

*methamphetamine*

Concentration of methamphetamine in nanograms per liter.

*morphine*

Concentration of morphine in nanograms per liter.

*phenazone*

Concentration of phenazone in nanograms per liter.

*sulfachloropyridazine*

Concentration of sulfachloropyridazine in nanograms per liter.

*sulfamethazine*

Concentration of sulfachloropyridazine in nanograms per liter.

*sulfamethoxazole*

Concentration of sulfamethoxazole in nanograms per liter.

*thiabendazole*

Concentration of thiabendazole in nanograms per liter.

*trimethoprim*

Concentration of trimethoprim in nanograms per liter.

*collection\_year*

Year sample was collected in the field.

*collection\_month*

Month sample was collected in the field.

*collection\_day*

Day sample was collected in the field.

*analysis\_year*

Year sample was analyzed.

*analysis\_month*

Month sample was analyzed.

*analysis\_day*

Day sample was analyzed.

stable\_isotopes.csv

This file contains carbon (δ13C) and nitrogen (δ15N) values for various benthic macroinvertebrate genera and periphyton collected from the 14 littoral sampling locations.

*site*

Unique alphanumeric identifier for a sampling location.

*C13*

Carbon (δ13C) stable isotope values in parts per thousand.

*N15*

Nitrogen (δ15N) stable isotope values in parts per thousand.

*Genus*

Genus of the analyzed organism.

*Species*

Species of the analyzed organism. When organism was identified solely to genus, the Species value is NA.

*comments*

Quality flag column where δ13C samples were outside of the range of standards.

total\_lipid.csv

*site*

Unique alphanumeric identifier for a sampling location.

*Genus*

Genus of the analyzed organism.

*Species*

Species of the analyzed organism. When organism was identified solely to genus, the Species value is NA.

*total\_lipid\_mg\_per\_g*

Total amount of lipids in a sample in milligrams of Lipid per gram of tissue.

*deviation*

Samples were weighed three times and deviation in measurement was calculated. All values are reported in milligrams of Lipid per gram of tissue.

*comments*

Quality flag column. Two samples spilled during fatty acid extraction. These samples are flagged as such.

**Methods**

*Inverse distance weighted (IDW) population calculation for each sampling location*

We recognized that sewage indicator concentrations at each sampling location may be related to a sampling location’s spatial position relative to both the size and proximity of neighboring developed sites. Therefore, we created the inverse distance weighted (IDW) population metric to compress, into a single metric, information about human population size, density, and location along the shoreline as well as distance between developed sites and sampling locations.

Our workflow for calculating IDW population required five steps. First, we traced polygons and shorelines from satellite imagery for each developed site in Google Earth. Second, polygon and line geometries were downloaded from Google Earth as a .kml file. Third, the .kml file was imported into the R statistical environment (R Core Team, 2019), where using the sf package (Pebesma, 2018) we calculated shoreline length, polygon area, and centroid location for each developed site. Fourth, we joined point locations of each sampling site with the spatial polygons to calculate the distance from each sampling location to each developed site’s centroid. Fifth, we calculated IDW population for each sampling location, using formula (1).

(1)

where *I* is the IDW population at sampling location *j*, *P* is the population at each of the three developed sites Listvyanka (LI), Bolshie Koty (BK), Bolshoe Goloustnoe (BGO), *A* is the area of a developed site in km2, *L* is the shoreline length at a developed site in km, and *D* is the distance from developed site *j* to each developed site’s centroid in km.

*Nutrients*

Water samples for nutrient analyses were collected in 150 mL glass jars that had been washed with phosphate-free soap and rinsed three times with water from the sampling location. Samples were collected in duplicates and immediately frozen at -20°C until processing at the A.P.Vinogradov Institute of Geochemistry (Siberian Branch of the Russian Academy of Sciences, Irkutsk). Samples were not filtered prior to freezing, meaning that nitrogen and ammonium concentrations may potentially include intracellular nitrogen and overestimate nitrogenous forms in the water column.

For each water sample, nitrate, ammonium, and total phosphorus concentrations were measured. For ammonium (2016a) and nitrate (2017) concentrations, samples were analyzed with a spectrophotometer following the addition of Nessler’s reagent and disulfuric acid respectively. Total phosphorus concentration was measured with a spectrophotometer following the addition of persulfate (2016b).

*Chlorophyll a*

Water samples were collected in 1.5 L plastic bottles from a depth of approximately 0.75 m. Within 12 h of collection, three subsamples (up to 150 mL each) were filtered through 25-mm diameter, 0.2 µm pore size nitrocellulose filters. Filters were then placed in a 35 mm petri dish and frozen in the dark until processing.

Chlorophyll samples were processed in a manner similar to that of Parson (1963) and Lorenzen (1967). Nitrocellulose filters were ground in 90% acetone, in which chlorophyll extraction was allowed to proceed overnight. Samples were then centrifuged for 15-20 minutes. After centrifugation, absorbance of the chlorophyll extract was measured in a spectrophotometer at 630, 645, 665, and 750 nm. Concentrations were calculated using the formula: ; where A is the absorbance value of a particular wavelength, V1 is the volume of the filtered water, and V2 is the volume of extract. Concentrations are reported as mg/L.

*Pharmaceuticals and Personal Care Products (PPCPs)*

Water samples for PPCP analysis were collected in 250 mL amber glass bottles that were rinsed with either methanol or acetone and then three times with sample water prior to collections. Following collection, samples were refrigerated and kept in the dark until solid phase extraction (SPE).

Within 12 h of collection, samples were filtered directly from the amber glass bottle using a single-stream 25-mm GF/F SPE cartridge setup (Waters Corporation, Milford, MA). Lab personnel wore gloves and face masks to minimize contamination. Prior to filtration, SPE cartridges were primed with at least 5 mL of either methanol or acetone and then washed with at least 5 mL of sample water. Rate of SPE occurred at approximately 1 drop per second. Extraction proceeded until water could no longer pass through the SPE cartridge or until all collected water was filtered. Cartridges were stored in Whirlpacks at -20°C until analysis for PPCPs following methods of Lee et al. (2016).

*Microplastics*

At each location, samples were collected in triplicate using 1.5 L clear plastic bottles that were washed thoroughly with sample water before each collection. Samples were collected by hand for each littoral site and with a metal bucket from aboard the ship for pelagic sites.

For processing, each sample was vacuum filtered on to a 47-mm diameter GF/F filter. During filtration, aluminum foil was used to cover the filtration funnel to prevent contamination from airborne microplastic particles. After filtration, filters were dried under vacuum pressure and then stored in 50-mm petri dishes. Following filtration of all three replicates, the filtrate was collected and then re-filtered through a GF/F filter as a control for contamination from the plastic vacuum funnel or potentially airborne microplastics.

Microplastic counting involved visual inspection of the entire GF/F according to Van Cauwenberghe et al. (2015). Visual enumeration was conducted under a stereo microscope with ~100x magnification, and microplastics were classified into one of three categories: fibers, fragments, or beads. For all categories, plastics were defined as observed objects with apparent artificial colors, to avoid enumerating plastics contributed from the sampling bottle itself. Fibers were defined as smooth, long plastics with consistent diameters. Fragments were defined as plastics with irregularly sharp or jagged edges. Beads were defined as spherical plastics. During enumeration, GF/Fs remained in the petri dish to minimize potential for contamination from the air.

*Periphyton abundance*

At each littoral site, we haphazardly selected three rocks representative of local substrate. A plastic stencil was used to define a surface area of each rock from which we scraped a standardized 14.5 cm2 patch of periphyton. Samples were preserved with Lugol’s solution and stored in plastic scintillation vials. Additional periphyton was collected in composite from each site for fatty acid and stable isotope analysis.

Periphyton taxonomic identification and enumeration was performed by subsampling 10 μL aliquots from each preserved sample. Cells, filaments, and colonies were counted for each taxonomic group until at least 300 cells were identified. For all 10 μL aliquots, the entire subsample was counted, even if 300 cells were counted prior to completing the aliquot. Taxa were classified into broad categories consistent with Baikal algal taxonomy (Izhboldina, 2007), using coarse groupings to capture general patterns in relative algal abundance. As a result, algal groups consisted of diatoms, *Ulothrix*, *Spirogyra*,and the green algal Order Tetrasporales.

*Benthic macroinvertebrate abundance*

Three kick-net samples were collected for assessment of benthic community composition and abundance. Using a D-net, we collected macroinvertebrates by flipping over 1-3 rocks, and then sweeping five times in a left-to-right motion across approximately 1 m. After the series of sweeps, the catch was rinsed into a plastic bucket. For each replicate, bucket contents were concentrated using a 64 μm mesh and placed in glass jars with 40% ethanol (vodka; the only preservative available to us at the time) for preservation and refrigerated at 4°C aboard the research vessel. The 40% ethanol preservative was replaced with ~80% ethanol upon return to the lab within 24 to 48 hours, and samples were stored at ~4°C.

Invertebrate taxonomic identification and enumeration were performed under a stereo microscope. All invertebrates were identified to species with the exception of juveniles (Taakhteev, 2015 for amphipods; Sitnikova, 2012 for molluscs; Table 2). All samples contained oligochaetes and polychaetes, but due to poor preservation, these taxa were not counted. Six samples of the 42 collected were not well-preserved and were excluded from further analyses, in order to reduce errors in identification. KD-1 and LI-1 were the only sites with 1 sample counted. BK-2 and KD-2 each had two samples counted.

*Stable Isotope Analysis*

Measurements of δ15N and δ13C were performed on an elemental analyzer-isotope ratio mass spectrometer (EA-IRMS; Finnigan DELTAplus XP, Thermo Scientific) at the Large Lakes Observatory, University of Minnesota Duluth.

*Fatty Acid Analysis*

Following freeze-drying, samples were transferred to 10 mL glass centrifuge vials, and 2 mL of 100% chloroform was added to each under nitrogen gas. Samples were allowed to sit in chloroform overnight at -80°C. Fatty acid extractions generally involved three phases: (1) 100% chloroform extraction, (2) chloroform-methanol extraction, and (3) fatty acid methylation. Fatty acid extraction methods were adapted from Schram et al. (2018).

After overnight chloroform extraction, samples underwent a chloroform-methanol extraction three times. To each sample, we added 1 mL cooled 100% methanol, 1 mL chloroform:methanol solution (2:1), and 0.8 mL 0.9% NaCl solution. Samples were inverted three times and sonicated on ice for 10 minutes. Next, samples were vortexed for 1 minute, and centrifuged for 5 minutes (3,000 rpm) at 4°C. Using a double pipette technique, the lower organic layer was removed and kept under nitrogen. After the third extraction, samples were evaporated under nitrogen flow, and resuspended in 1.5 mL chloroform and stored at -20°C overnight.

Once resuspended in chloroform, 1 mL of chloroform extract was transferred to a glass centrifuge tube with a glass syringe as well as an internal standard of 4 μL of 19-carbon fatty acid. Samples were then evaporated under nitrogen, and then 1 mL of toluene and 2 mL of 1% sulfuric acid-methanol was added. The vial was closed under nitrogen gas and then incubated in 50°C water bath for 16 hours. After incubation, samples were removed from the bath, allowed to reach room temperature and stored on ice. Next, we performed a potassium carbonate-hexane extraction twice. To each sample, we added 2 mL of 2% potassium bicarbonate and 5 mL of 100% hexane, inverting the capped vial so as to mix the solution. Samples were centrifuged for 3 minutes (1,500 rpm) at 4°C. The upper hexane layer was then removed and placed in a vial to evaporate under nitrogen flow. Once almost evaporated, 1 mL of 100% hexane was added and stored in a glass amber autosampler vial for GC/MS quantification. GC/MS quantification was performed with a Shimadzu QP2020 GC/MS following Schram et al. (2018).

*A commitment to FAIR and TRUST principles*

Throughout the dataset’s development, we strove to incorporate both FAIR (Findable, Accessible, Interoperable, and Reproducible) and TRUST (Transparency, Responsibility, User Focus, Sustainability, and Technology) principles where applicable.

With respect to FAIR principles (Wilkinson et al. 2016), the data are openly accessible in a standardized, replicate-level format on the EDI portal. The 11 CSV files contained within the dataset are entirely interoperable using the “site” column, enabling all variables to efficiently merged together with little data wrangling involved. Finally, all analytical and some data wrangling scripts are available on the EDI portal in a compressed format, such that future users can reproduce data manipulation and analyses described in Meyer et al. (202X).

With respect to TRUST principles (Lin et al. 2020), we strove to document additional metadata and data-cleaning practices in a public Open Science Framework (OSF) repository (DOI). These steps are not necessarily critical to the core EDI dataset, but provide increased transparency for future users wishing recreate the dataset de novo. All “raw” data are provided in the OSF portal, including an initial cleaning script (00\_disaggregated\_data\_cleaning.R) to remove incorrect spellings, erroneous data values, and inconsistent column names. This repository also includes photographs of both field notes as well as photographs of shoreline and substrate from sampling locations. To empower and expedite future reuse, all directories are accompanied with documentation that details directory contents, and all associated scripts are documented and annotated. While many of the files are redundant from the EDI repository, the OSF repository is meant to supplement the EDI repository, so as to enable sustainable, user-focused transparency of how data were collected and cleaned from their raw formats.

**Technical Validation**

The dataset had three main validation procedures: taxonomic, analytical, and reproducible.

For taxonomic validation, all phylogenetic groupings were based off most recent identification keys. Amphipods were identified according to Taakhteev & Didorenko (2015). Mollusks were identified according to Sitnikova (2012). Algal taxa were identified according to Izhboldina (2007). For consistency, all taxa were identified by one person (Michael F. Meyer), who was trained by experts in Baikal algal and macroinvertebrate taxonomy.

For analytical validation, internal standards were used for all mass-spectroscopy analyses. PPCP analyses involved an internal nicotinamide standard to assess yield. Stable isotope values were calibrated against certified reference materials including L-glutamic acid (NIST SRM 8574), low organic soil and sorghum flour (standards B-2153 and B-2159 from Elemental Micro-analysis Ltd., Okehampton, UK) and in-house standards (acetanilide and caffeine). Replicate analyses of external standards showed a mean standard deviation of 0.06 ‰ and 0.09 ‰, for δ13C and δ15N, respectively. Finally, fatty acid estimations used an internal 19:0 standard to assess oxidation of fatty acids during extraction, methylation, and quantification.

For data reproducibility, we systematically reviewed all R code to format and aggregate source data. An initial cleaning script (00\_disaggregated\_data\_cleaning.R) removed incorrect spellings, erroneous data values, and inconsistent column names from raw data. This step created the standardized .csv files detailed above, which are available on the EDI repository. Raw data files are available on the project’s Open Science Framework page (DOI) but are not included in the EDI repository to prevent confusion or incorrect usage. Data hosted on EDI are at the replicate-level but can be aggregated to the sampling-site-level using script 01\_data\_cleaning.R. In addition to aggregation scripts, six R scripts used for analyses in Meyer et al. (202X) are also available on the EDI repository within the compressed entity “scripts.tar.gz”. All R code for data aggregation was written by one person (Michael F. Meyer) and then independently reviewed by two others (Matthew R. Brousil and Kara H. Woo) to confirm that code performed as intended, was well documented, and annotations were complete.

**Data Use and Recommendations for Reuse**

Recognizing the potential for continued low-level, sewage pollution at Lake Baikal (Timoshkin et al. 2016, 2018; Volkova et al. 2018) and lakes worldwide (Yang et al. 2018; Meyer et al. 2019), the final dataset can be applied to a suite of research questions pertaining to ecological responses to human disturbance. We highlight two main areas for immediate application.

First, the final data products can be harmonized with other littoral sampling efforts throughout Lake Baikal, so as to enhance spatial coverage and data diversity. Since 2010, Lake Baikal has experienced increasing filamentous algal abundance, especially near larger lakeside developments (Kravtsova et al. 2014; Timoshkin et al. 2016, 2018; Volkova et al. 2018). Recent benthic algal surveys throughout Baikal’s entirety have suggested that cosmopolitan filamentous algae, such as *Spirogyra spp.*, tend to be more abundant near larger lakeside developments (Timoshkin et al. 2016; Volkova et al. 2018). For example, Listvyanka is a small town located at the beginning of the Angara River, Baikal’s only surface outflow. While Listvyanka’s permanent population is approximately 2,000 persons, the town is a growing tourism hub, which hosts over 1.2 million tourists per year (Interfax-Tourism 2018). Surveys conducted near Listvyanka have suggested increased *Spirogyra spp.* abundance is associated with wastewater release (Timoshkin et al. 2016). Although wastewater inputs are likely low and are diluted to negligible concentrations offshore (Meyer et al., 202X), combining monitoring efforts across spatial and temporal scales are necessary to evaluate the spatial and temporal extent of wastewater entering Baikal. As such, our data could complement previous, current, and future monitoring efforts, where observations may be missing.

Second, the final data products are useful to expanding freshwater PPCP, microplastic, and associated biological responses across large spatial scales. Recent syntheses of the PPCP literature have reported that studies involving lakes are less abundant relative to those focused on lotic systems (Meyer et al. 2019). Likewise, microplastic studies have noted that freshwater environments are less represented in the literature relative to marine ecosystems (Horton et al. 2017). For both PPCPs and microplastics, toxic responses to even minute concentrations can be uncertain and differ between ecosystem types (e.g., Rosi-Marshall et al. 2013 for lotic and Shaw et al. 2015 for lentic). As a result of PPCPs and microplastics garnering increasing attention worldwide, sampling of PPCPs and microplastics with co-located biological data across multiple spatial and temporal scales would be necessary to synthesize biotic responses to micropollutants across systems. Although our data constitute a limited number of such data that exist globally, our final data products are highly structured and flexible for merging with similar datasets. Additionally, our dataset’s sequential harmonization workflow could be adopted by similar monitoring efforts, thereby facilitating data interoperability. Through integration with similar monitoring efforts, our dataset can contribute to global synthesis of emerging contaminant consequences, especially in a region of the world that is often not easily accessible to many researchers.

**References**

Barnes, D. K. A., F. Galgani, R. C. Thompson, and M. Barlaz. 2009. Accumulation and fragmentation of plastic debris in global environments. Philos Trans R Soc Lond B Biol Sci **364**: 1985–1998. doi:10.1098/rstb.2008.0205

Bendz, D., N. A. Paxéus, T. R. Ginn, and F. J. Loge. 2005. Occurrence and fate of pharmaceutically active compounds in the environment, a case study: Höje River in Sweden. Journal of Hazardous Materials **122**: 195–204. doi:10.1016/j.jhazmat.2005.03.012

Brodin, T., J. Fick, M. Jonsson, and J. Klaminder. 2013. Dilute Concentrations of a Psychiatric Drug Alter Behavior of Fish from Natural Populations. Science **339**: 814–815. doi:10.1126/science.1226850

Camilleri, A. C., and T. Ozersky. 2019. Large variation in periphyton δ13C and δ15N values in the upper Great Lakes: Correlates and implications. Journal of Great Lakes Research **45**: 986–990. doi:10.1016/j.jglr.2019.06.003

Costanzo, S. D., M. J. O’Donohue, W. C. Dennison, N. R. Loneragan, and M. Thomas. 2001. A New Approach for Detecting and Mapping Sewage Impacts. Marine Pollution Bulletin **42**: 149–156. doi:10.1016/S0025-326X(00)00125-9

Dalsgaard, J., M. St. John, G. Kattner, D. Müller-Navarra, and W. Hagen. 2003. Fatty acid trophic markers in the pelagic marine environment, p. 225–340. *In* Advances in Marine Biology. Elsevier.

Edmondson, W. T. 1970. Phosphorus, Nitrogen, and Algae in Lake Washington after Diversion of Sewage. Science **169**: 690–691.

Focazio, M. J., D. W. Kolpin, K. K. Barnes, E. T. Furlong, M. T. Meyer, S. D. Zaugg, L. B. Barber, and M. E. Thurman. 2008. A national reconnaissance for pharmaceuticals and other organic wastewater contaminants in the United States - II) Untreated drinking water sources. SCIENCE OF THE TOTAL ENVIRONMENT **402**: 201–216. doi:10.1016/j.scitotenv.2008.02.021

Gartner, A., P. Lavery, and A. J. Smit. 2002. Use of delta N-15 signatures of different functional forms of macroalgae and filter-feeders to reveal temporal and spatial patterns in sewage dispersal. Mar. Ecol.-Prog. Ser. **235**: 63–73. doi:10.3354/meps235063

Green, D. S. 2016. Effects of microplastics on European flat oysters, Ostrea edulis and their associated benthic communities. Environmental Pollution **216**: 95–103. doi:10.1016/j.envpol.2016.05.043

Hall, R. I., P. R. Leavitt, R. Quinlan, A. S. Dixit, and J. P. Smol. 1999. Effects of agriculture, urbanization, and climate on water quality in the northern Great Plains. Limnology and Oceanography **44**: 739–756. doi:10.4319/lo.1999.44.3\_part\_2.0739

Hampton, S. E., S. C. Fradkin, P. R. Leavitt, and E. E. Rosenberger. 2011. Disproportionate importance of nearshore habitat for the food web of a deep oligotrophic lake. Marine and Freshwater Research **62**: 350. doi:10.1071/MF10229

Hampton, S. E., S. McGowan, T. Ozersky, and others. 2018. Recent ecological change in ancient lakes. Limnology and Oceanography **63**: 2277–2304. doi:10.1002/lno.10938

Horton, A. A., A. Walton, D. J. Spurgeon, E. Lahive, and C. Svendsen. 2017. Microplastics in freshwater and terrestrial environments: Evaluating the current understanding to identify the knowledge gaps and future research priorities. Science of The Total Environment **586**: 127–141. doi:10.1016/j.scitotenv.2017.01.190

Interfax-Tourism. 2018. Байкал с января по август 2018 года посетили 1,2 миллиона туристов (1.2 million tourists vistied Baikal from January through August 2018). Interfax-Tourism, October 25

Jeppesen, E., M. Søndergaard, J. P. Jensen, and others. 2005. Lake responses to reduced nutrient loading – an analysis of contemporary long-term data from 35 case studies. Freshwater Biology **50**: 1747–1771. doi:10.1111/j.1365-2427.2005.01415.x

Kolpin, D. W., E. T. Furlong, M. T. Meyer, E. M. Thurman, S. D. Zaugg, L. B. Barber, and H. T. Buxton. 2002. Pharmaceuticals, Hormones, and Other Organic Wastewater Contaminants in U.S. Streams, 1999−2000: A National Reconnaissance. Environmental Science & Technology **36**: 1202–1211. doi:10.1021/es011055j

Kozhova, O. M., and L. R. Izmest’eva. 1998. Lake Baikal: Evolution and Biodiversity, Backhuys Publishers.

Kravtsova, L. S., L. A. Izhboldina, I. V. Khanaev, and others. 2014. Nearshore benthic blooms of filamentous green algae in Lake Baikal. Journal of Great Lakes Research **40**: 441–448. doi:10.1016/j.jglr.2014.02.019

Lapointe, B. E., L. W. Herren, D. D. Debortoli, and M. A. Vogel. 2015. Evidence of sewage-driven eutrophication and harmful algal blooms in Florida’s Indian River Lagoon. Harmful Algae **43**: 82–102. doi:10.1016/j.hal.2015.01.004

Lin, D., J. Crabtree, I. Dillo, and others. 2020. The TRUST Principles for digital repositories. Scientific Data **7**: 144. doi:10.1038/s41597-020-0486-7

Meyer, M. F., S. G. Labou, A. N. Cramer, M. R. Brousil, and B. T. Luff. 2020. The global lake area, climate, and population dataset. Scientific Data **7**: 174. doi:10.1038/s41597-020-0517-4

Meyer, M. F., S. M. Powers, and S. E. Hampton. 2019. An Evidence Synthesis of Pharmaceuticals and Personal Care Products (PPCPs) in the Environment: Imbalances among Compounds, Sewage Treatment Techniques, and Ecosystem Types. Environ. Sci. Technol. **53**: 12961–12973. doi:10.1021/acs.est.9b02966

Moore, J. W., D. E. Schindler, M. D. Scheuerell, D. Smith, and J. Frodge. 2003. Lake eutrophication at the urban fringe, Seattle region, USA. AMBIO: A Journal of the Human Environment **32**: 13–18.

O’Donnell, D. R., P. Wilburn, E. A. Silow, L. Y. Yampolsky, and E. Litchman. 2017. Nitrogen and phosphorus colimitation of phytoplankton in Lake Baikal: Insights from a spatial survey and nutrient enrichment experiments. Limnology and Oceanography **62**: 1383–1392. doi:10.1002/lno.10505

Powers, S. M., T. W. Bruulsema, T. P. Burt, and others. 2016. Long-term accumulation and transport of anthropogenic phosphorus in three river basins. Nature Geoscience **9**: 353–356. doi:10.1038/ngeo2693

Richmond, E. K., M. R. Grace, J. J. Kelly, A. J. Reisinger, E. J. Rosi, and D. M. Walters. 2017. Pharmaceuticals and personal care products (PPCPs) are ecological disrupting compounds (EcoDC). Elem Sci Anth **5**: 66. doi:10.1525/elementa.252

Richmond, E. K., E. J. Rosi, D. M. Walters, J. Fick, S. K. Hamilton, T. Brodin, A. Sundelin, and M. R. Grace. 2018. A diverse suite of pharmaceuticals contaminates stream and riparian food webs. Nature Communications **9**: 4491. doi:10.1038/s41467-018-06822-w

Romera-Castillo, C., M. Pinto, T. M. Langer, X. A. Álvarez-Salgado, and G. J. Herndl. 2018. Dissolved organic carbon leaching from plastics stimulates microbial activity in the ocean. Nat Commun **9**: 1–7. doi:10.1038/s41467-018-03798-5

Rosenberger, E. E., S. E. Hampton, S. C. Fradkin, and B. P. Kennedy. 2008. Effects of shoreline development on the nearshore environment in large deep oligotrophic lakes. Freshwater Biology **53**: 1673–1691. doi:10.1111/j.1365-2427.2008.01990.x

Rosi-Marshall, E. J., D. W. Kincaid, H. A. Bechtold, T. V. Royer, M. Rojas, and J. J. Kelly. 2013. Pharmaceuticals suppress algal growth and microbial respiration and alter bacterial communities in stream biofilms. Ecological Applications **23**: 583–593. doi:10.1890/12-0491.1

Rosi-Marshall, E. J., and T. V. Royer. 2012. Pharmaceutical Compounds and Ecosystem Function: An Emerging Research Challenge for Aquatic Ecologists. Ecosystems **15**: 867–880. doi:10.1007/s10021-012-9553-z

Sargent, J. R., and S. Falk-Petersen. 1988. The lipid biochemistry of calanoid copepods. Hydrobiologia **167–168**: 101–114. doi:10.1007/BF00026297

Shaw, L., C. Phung, and M. Grace. 2015. Pharmaceuticals and personal care products alter growth and function in lentic biofilms. Environmental Chemistry **12**: 301. doi:10.1071/EN14141

Taipale, S., U. Strandberg, E. Peltomaa, A. W. E. Galloway, A. Ojala, and M. T. Brett. 2013. Fatty acid composition as biomarkers of freshwater microalgae: analysis of 37 strains of microalgae in 22 genera and in seven classes. Aquatic Microbial Ecology **71**: 165–178. doi:10.3354/ame01671

Timoshkin, O. A., M. V. Moore, N. N. Kulikova, and others. 2018. Groundwater contamination by sewage causes benthic algal outbreaks in the littoral zone of Lake Baikal (East Siberia). Journal of Great Lakes Research. doi:10.1016/j.jglr.2018.01.008

Timoshkin, O. A., D. P. Samsonov, M. Yamamuro, and others. 2016. Rapid ecological change in the coastal zone of Lake Baikal (East Siberia): Is the site of the world’s greatest freshwater biodiversity in danger? Journal of Great Lakes Research **42**: 487–497. doi:10.1016/j.jglr.2016.02.011

Tong, Y., M. Wang, J. Peñuelas, and others. 2020. Improvement in municipal wastewater treatment alters lake nitrogen to phosphorus ratios in populated regions. Proc Natl Acad Sci USA **117**: 11566–11572. doi:10.1073/pnas.1920759117

Turetsky, M. R., R. K. Wieder, C. J. Williams, and D. H. Vitt. 2000. Organic matter accumulation, peat chemistry, and permafrost melting in peatlands of boreal Alberta. Écoscience **7**: 115–122. doi:10.1080/11956860.2000.11682608

Vendel, A. L., F. Bessa, V. E. N. Alves, A. L. A. Amorim, J. Patrício, and A. R. T. Palma. 2017. Widespread microplastic ingestion by fish assemblages in tropical estuaries subjected to anthropogenic pressures. Marine Pollution Bulletin **117**: 448–455. doi:10.1016/j.marpolbul.2017.01.081

Volkova, E. A., N. A. Bondarenko, and O. A. Timoshkin. 2018. Morphotaxonomy, distribution and abundance of *Spirogyra* (Zygnematophyceae, Charophyta) in Lake Baikal, East Siberia. Phycologia **57**: 298–308. doi:10.2216/17-69.1

Wickham, H. 2014. Tidy Data. Journal of Statistical Software **59**: 1–23. doi:10.18637/jss.v059.i10

Wilkinson, M. D., M. Dumontier, Ij. J. Aalbersberg, and others. 2016. The FAIR Guiding Principles for scientific data management and stewardship. Sci Data **3**. doi:10.1038/sdata.2016.18

Yang, Y., W. Song, H. Lin, W. Wang, L. Du, and W. Xing. 2018. Antibiotics and antibiotic resistance genes in global lakes: A review and meta-analysis. Environment International **116**: 60–73. doi:10.1016/j.envint.2018.04.011

Yoshida, T., T. Sekino, M. Genkai-Kato, and others. 2003. Seasonal dynamics of primary production in the pelagic zone of southern Lake Baikal. Limnology **4**: 53–62. doi:10.1007/s10201-002-0089-3

2016a. Methods for determination of nitrogen-containing matters (with corrections) (Методы определения азотсодержащих веществ (с Поправками)).

2016b. Methods for determination of phosphorus-containing matters (with corrections) (Методы определения фосфорсодержащих веществ).

2017. Nitrate concentration in waters: Photometric methods with Giress reagent following stabilization in a cadmium reducer (Массовая концентрация нитратного азота в водах: Методика измерений фотометрическим методом с реактивом Грисса после восстановления в камиевом редукторе).

**Acknowledgements**

We would like to thank the faculty, students, staff, and mariners of the Irkutsk State University’s Biological Research Institute Biostation for their expert field, taxonomic, and laboratory support; Marianne Moore and Bart De Stasio for helpful advice; the researchers and students of the Siberian Branch of the Russian Academy of Sciences Limnological Institute for expert taxonomic and logistical assistance; Oleg A. Timoshkin, Tatiana Ya. Sitnikova, and Irina V. Mekhanikova for offering insights throughout the development of this project. Funding was provided by the National Science Foundation (NSF-DEB-1136637) to S.E.H., a Fulbright Fellowship to M.F.M., a NSF Graduate Research Fellowship to M.F.M. (NSF-DGE-1347973), and the Russian Ministry of Education and Science Research Project (No. GR 01201461929; 1354-2014/51). This work serves as one chapter of M.F.M.’s doctoral dissertation in Environmental and Natural Resource Sciences at Washington State University.

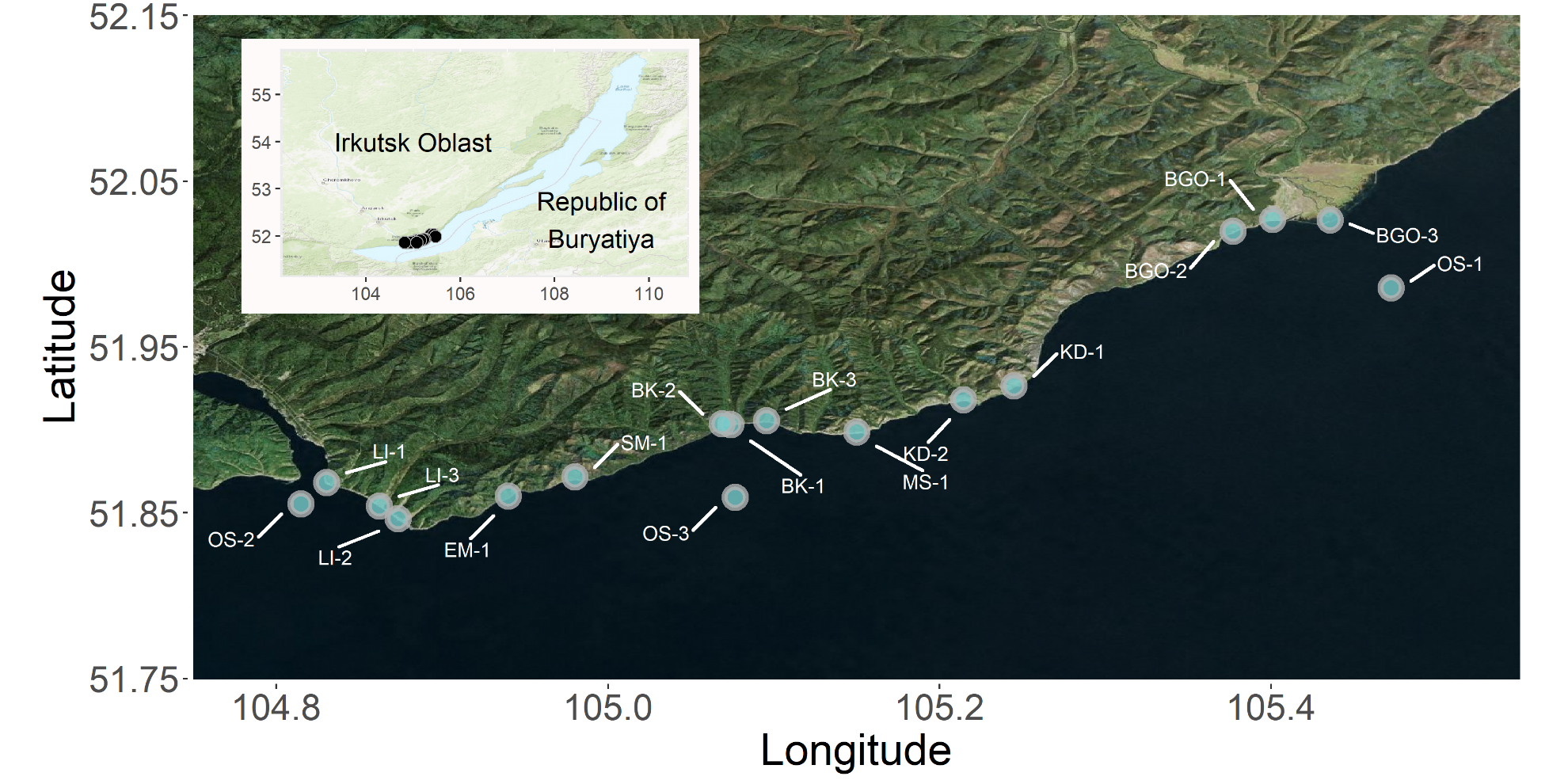


Figure 1: Map of all sampling locations with sites labeled with unique alphanumeric code. The entire transect included three developed sites (i.e., Listvyanka (LI), Bolshie Koty (BK), Bolshoe Goloustnoe (BGO)). Three offshore sites (OS) were also sampled to compare pelagic sewage signals to those in the littoral. Sites without adjacent lakeside development included Emelyanikha Bay (EM), Maloe Kadilnoe (KD), Mys Soboliny (MS), Sredny Mys (SM). Littoral sampling locations were all 8.9-20.75 m from shore and at a depth approximately of 0.75 m, whereas pelagic sites were approximately 2-5 km from shore and ranged in depth from 900 to 1300 m.

|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| **Site** | **Latitude** | **Longitude** | **Depth (m)** | **Distance to shore (m)** |
| BK-1 | 51.90316 | 105.074 | 0.7 | 10 |
| BK-2 | 51.90365 | 105.069 | 0.9 | 17.5 |
| BK-3 | 51.90536 | 105.0957 | 0.8 | 10 |
| BGO-1 | 52.02693 | 105.401 | 0.9 | 18 |
| BGO-2 | 52.0197 | 105.3771 | 1.1 | 14 |
| BGO-3 | 52.02649 | 105.4358 | 0.7 | 21 |
| OS-1 | 51.98559 | 105.4724 | 900 | NA |
| KD-1 | 51.92646 | 105.245 | 0.8 | 20.75 |
| KD-2 | 51.91807 | 105.2146 | 0.9 | 14.5 |
| MS-1 | 51.89863 | 105.1502 | 0.6 | 10.5 |
| SM-1 | 51.87152 | 104.9801 | 0.9 | 11.5 |
| LI-1 | 51.86825 | 104.8304 | 0.6 | 8.9 |
| LI-2 | 51.84626 | 104.8736 | 0.8 | 9.4 |
| LI-3 | 51.85407 | 104.8622 | 0.7 | 9.25 |
| EM-1 | 51.86005 | 104.94 | 0.7 | 15.5 |
| OS-2 | 51.8553 | 104.8148 | 1300 | NA |
| OS-3 | 51.85911 | 105.0769 | 1400 | 5000 |

Table 1: Locational information for each of the 17 sampling stations. “OS” refers to pelagic locations (i.e., “Offshore”), whereas other site abbreviations refer to littoral sampling locations.