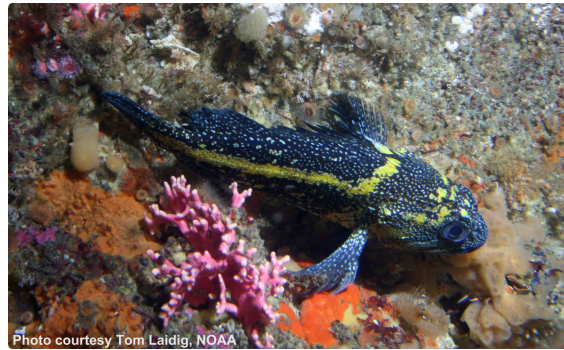# A 2019 catch-only projection from the 2015 assessment of the status of China rockfish (Sebastes nebulosus) along the U.S. West Coast

Maia Sosa Kapur[1]
Owen Hamel[2]

[1]University of Washington, School of Aquatic and Fisheries Sciences, 1122 NE Boat St., Seattle, Washington 98105

[2]Northwest Fisheries Science Center, U.S. Department of Commerce, National Oceanic and Atmospheric Administration, National Marine Fisheries Service, 2725 Montlake Boulevard East, Seattle, Washington 98112

July 2019

A 2019 catch-only projection from the 2015 assessment of the status of China rockfish (Sebastes nebulosus) along the U.S. West Coast

Contents

Executive summary

Stock

This is a catch-only projection of the China rockfish (*Sebastes nebulosus*) resource in U.S. waters off the coast of the California, Oregon, and Washington using actual catch data through 2018 and catch projections for 2019 and 2020, based on models created in 2015 (Dick et al. 2016). China rockfish are modelled with three independent stock assessments to account for spatial variation in exploitation history as well as regional differences in growth and size composition of the catch. The northern area model is defined as Washington state Marine Catch Areas (MCAs) 1-4. The central area model spans from the Oregon-Washington border to 40°10′ N. latitude. The southern area model spans 40°10′ N. latitude to the U.S.-Mexico border. However, very little catch of China rockfish occurs south of Point Conception, California (34°27′ N. latitude).

Catches

China rockfish are most often caught by hook-and-line (both recreational and commercial fisheries) as well as by traps in the commercial live-fish fishery. Although China rockfish were not a major target species, the commercial rockfish fishery along the U.S. Pacific West Coast developed in the late 1800s and early 1990s. Available estimates of China rockfish catch in California begin in the early 1900s, along with small commercial catches in Oregon until recreational landings began to increase in the early 1970s (Figures a-c). Reconstructed recreational landings of China rockfish in the northern assessment begin in 1967. As of 1995, Washington has prohibited commercial nearshore fixed gear in state waters and does not have a historical reconstruction of China rockfish commercial landings. The majority of commercial removals of China rockfish are now landed by live-fish fisheries in California and southern Oregon. The magnitude of total removals over the last 10 years peaked in 2009 (35.52 mt) and has been decreasing since then. In recent years, California has the largest removals of the three states (dominated by the recreational fleet) with smallest removals coming from the Oregon recreational fleet (Table a).

The nearshore live-fish fishery developed in California in the late 1980s and early 1990s and extended into Oregon by the mid-1990s, driven by the market prices for live fish. Northern Oregon (north of Florence) does not contribute significantly to the live-fish fishery (maximum removal of 0.02 mt) as the market for this sector of the fishery is centered in California. Catches from the live-fish fishery in southern Oregon (south of Florence) has composed the majority of the catch in that state since 1999, and peaked in 2002. In California, the landings of live fish begin exceeding the landings of dead fish south of 40°10′ N. latitude in 1998 and north of 40°10′ N. latitude in 1999; and the pattern continues through present.

The historical reconstruction of landings from the recreational fishery for China rockfish in California goes back to 1928, and the fishery began significantly increasing in the late 1940s.

The recreational catches in California are significantly higher than the commercial catches, and have decreased in the last five years (Table a). Recreational catches in California peaked in 1987 at 53.29 mt and have declined to roughly 10-20 mt per year over the last 10 years. The trend is opposite in Oregon, with the magnitude of the commercial landings greater than the recreational landings. The historical landings from the recreational fleet in Oregon start in 1973 at 0.86 mt, peak in 1983 at 6.07 mt and again in 1993 at 6.04 mt. The recreational catches over the last 10 years in Oregon have ranged from 5.60 mt in 2018 to 2.5 mt in 2009. Recreational landings in Washington peaked in 1992 (7.98 mt) and have remained between 2-4 mt from 2005-present.
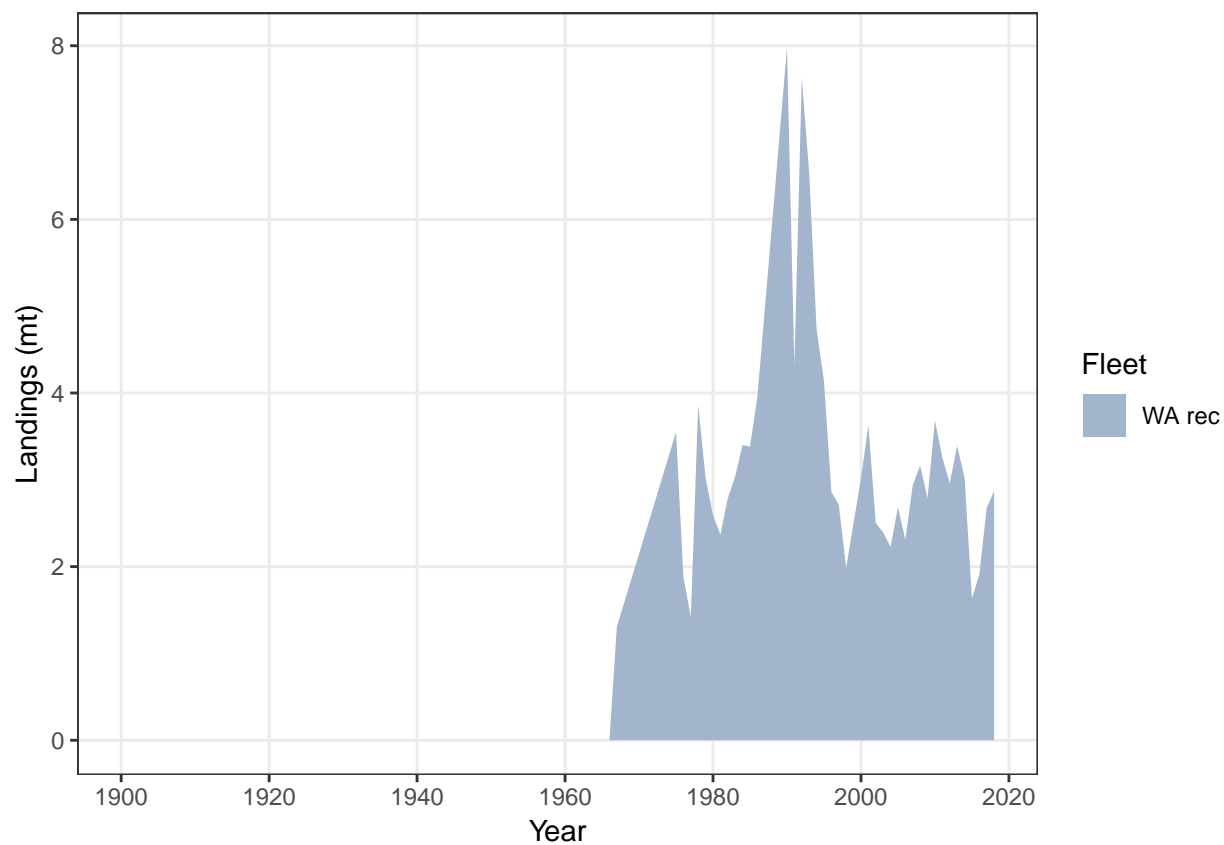
Figure a: China rockfish landings for Washington. Washington does not have a commercial nearshore fishery.
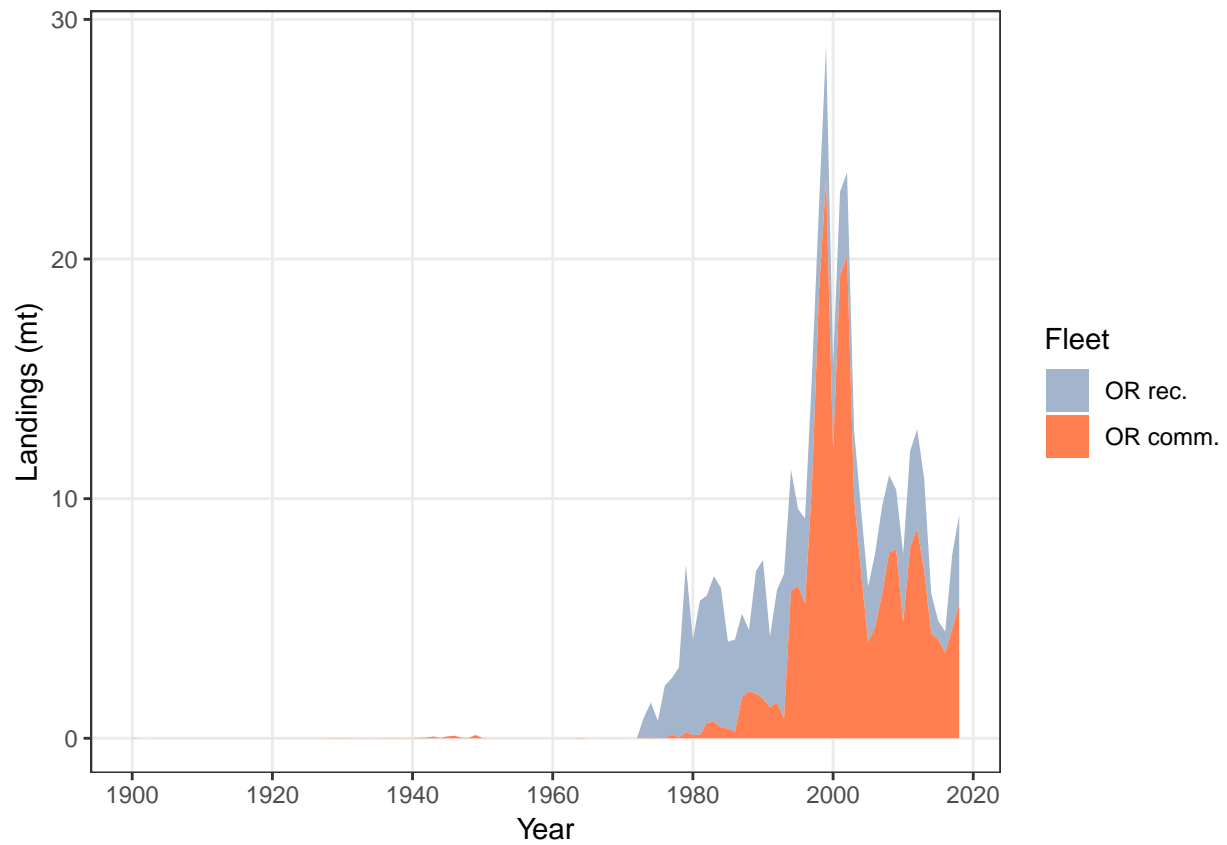
Figure b: Stacked line plot of China rockfish landings history for Oregon by fleet (recreational and commercial).
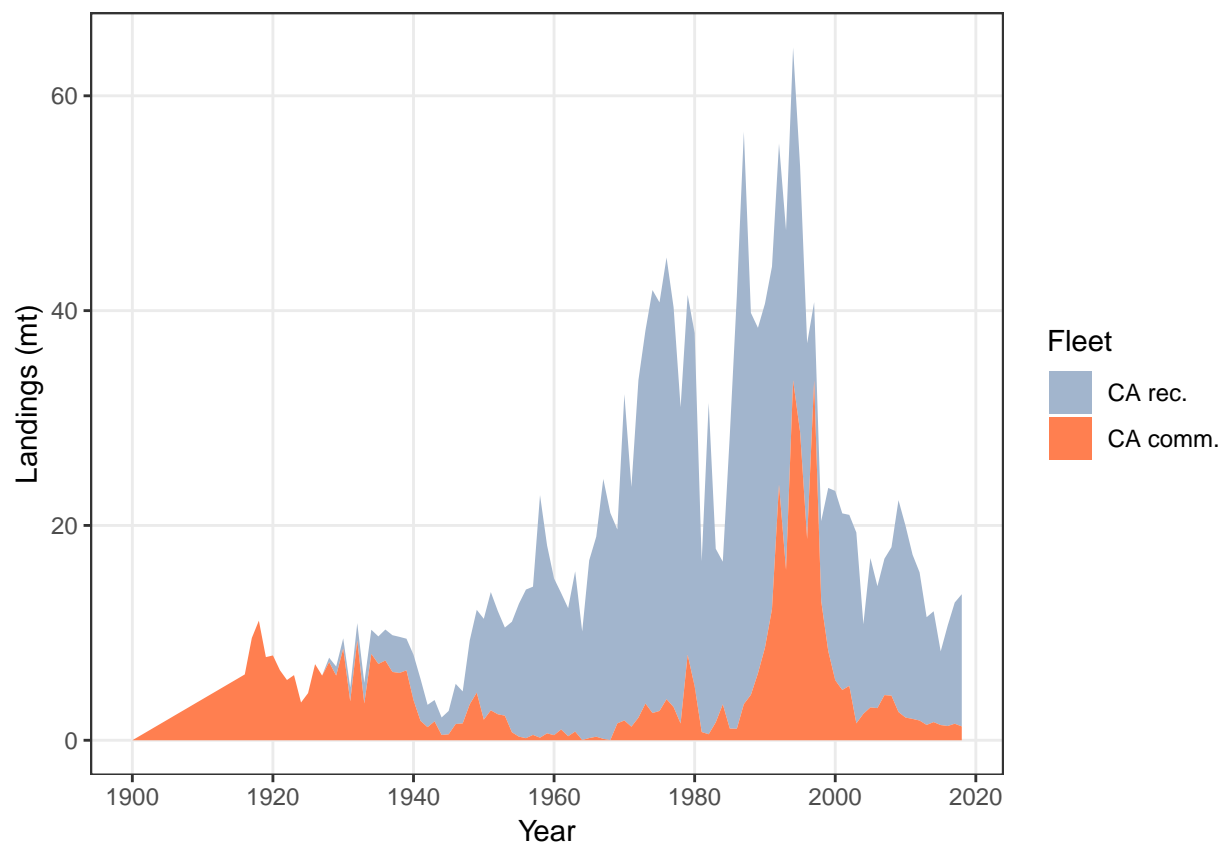
Figure c: Stacked line plot of China rockfish landings history for California by fleet (recreational and commercial).

Table a: Recent China rockfish landings and discard (mt) by region, used in the catch-only projection. Years 2019 and 2020 are projected values.

Year	North Catch	Central Catch	South Catch	Total
2009	2.79	12.63	21.10	36.52
2010	3.68	8.76	20.44	32.88
2011	3.26	13.30	17.01	33.57
2012	2.96	14.55	15.60	33.11
2013	3.40	12.25	11.29	26.94
2014	3.02	7.04	12.44	22.50
2015	1.63	4.90	13.89	20.42
2016	1.91	4.45	16.37	22.73
2017	1.38	7.67	12.94	21.99
2018	1.47	9.33	13.71	24.51
2019	1.48	7.60	13.32	22.40
2020	1.48	7.58	13.32	22.38

Figure d: Map depicting the boundaries for the three base-case models, Southern model (south of 40°10′ N. latitude), Central model (south of 40°10′ N. latitude to the OR-WA border), and the Northern model (WA state MCAs 1-4).

Data and assessment

China rockfish was assessed as a data moderate stock in 2013 (Cope et al. 2015) using the XDB-SRA modeling framework. The 2015 assessment used as the basis for this update implemented the Stock Synthesis version 3.24u. The model begins in 1900, and assumes the stock was at an unfished equilibrium that year.
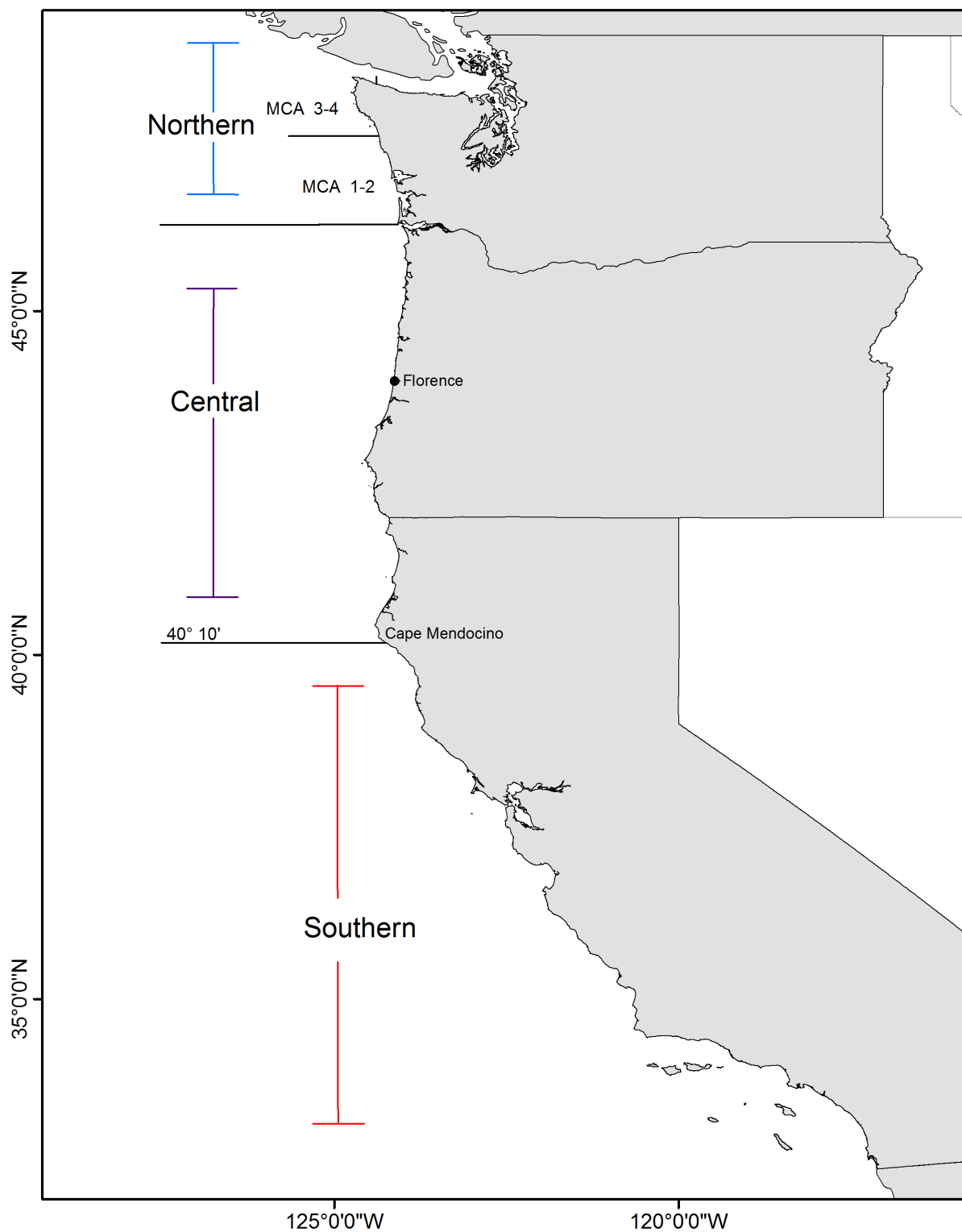
Data within the central and northern models were stratified as follows: the northern model-groups MCAs 1-2 (southern WA) and MCAs 3-4 (northern WA); the central model includes areas north and south of Florence, OR; and the southern model aggreate data south of 40°10′ N. latitude, in part because historical removals from the dominant fisheries (recreational charter and private boat modes) prior to 2004 are not available at a finer spatial scale (Figure d). The data used in the assessments includes commercial and recreational landings, Catch per Unit Effort (CPUE) indices from recreational and commercial fleets, and length and age compositions. Discard data (total discards in mt and size compositions) from the commercial live-fish fishery were modelled south of 40°10′ N. latitude. Where available, age and length compositions for the recreational party/charter (CPFV) and private/rental modes were developed separately.

Stock biomass

Note: the text immediately following reflects outputs from the accepted base case model from 2015. Figures have been updated to reflect the period forecasted in this update. Depletion and stock size estimates during the forecasted period are available in Decision Tables.

Estimated spawning output in the northern area (Washington state) declined between the 1960s and 1990s but has been largely stable during the past two decades (Figure f and Table b). The estimated relative depletion level (spawning output relative to unfished spawning output) of the northern stock in 2015 is 73.4% (~95% asymptotic interval: ± 63.6% - 83.2%) (Figure h).

The central area model for China rockfish estimates that spawning output was just above the biomass target in 2015 (Figure f and Table c). The rate of spawning output decline is estimated to be steepest during the 1980s to 1990s and continued to decline from the early 2000s at a slower rate to an estimated minimum of 39.6% in 2014. The estimated relative depletion level of the central stock in 2015 is 61.5% (~95% asymptotic interval: ± 53.8% - 69.2%) (Figure h).

The assessment for the southern management area suggests that China rockfish were lightly, but steadily exploited since the early 1900s, with more rapid declines in spawning output beginning with development of the recreational fishery in the 1950s (Figure f and Table d). The estimated relative depletion level of the southern stock in 2015 is 29.6% (~95% asymptotic interval: ± 25.0% - 34.3%) (Figure h). Although spawning output in the southern

Table b: Recent trend in beginning of the year biomass and depletion for the northern China rockfish model.

	Year	Spawning Output (billion eggs)	~ 95% confidence interval	Estimated depletion	~ 95% confidence interval
SSB_2010	2010	18.062	(8.96-27.17)	0.739	(0.644-0.834)
SSB_2011	2011	17.993	(8.89-27.1)	0.736	(0.64-0.833)
SSB_2012	2012	17.971	(8.86-27.08)	0.735	(0.638-0.832)
SSB_2013	2013	17.981	(8.87-27.09)	0.736	(0.639-0.833)
SSB_2014	2014	17.944	(8.83-27.06)	0.734	(0.637-0.832)
SSB_2015	2015	17.950	(8.83-27.07)	0.734	(0.637-0.832)
SSB_2016	2016	18.104	(8.98-27.23)	0.741	(0.645-0.836)
SSB_2017	2017	18.225	(9.09-27.36)	0.746	(0.652-0.84)
SSB_2018	2018	18.399	(9.26-27.54)	0.753	(0.661-0.844)
SSB_2019	2019	18.556	(9.41-27.7)	0.759	(0.67-0.848)

area is more depleted than the central and northern areas, it is the only area with an increasing trend over the past 15 years.

Table c: Recent trend in beginning of the year biomass and depletion for the central (north of 40°10′ N. latitude to the OR-WA border) China rockfish model.

	Year	Spawning Output (billion eggs)	~ 95% confidence interval	Estimated depletion	~ 95% confidence interval
SSB_2010	2010	40.125	(27.05-53.2)	0.616	(0.541-0.692)
SSB_2011	2011	40.380	(27.29-53.47)	0.620	(0.545-0.695)
SSB_2012	2012	40.112	(27.01-53.21)	0.616	(0.54-0.692)
SSB_2013	2013	39.706	(26.6-52.82)	0.610	(0.533-0.687)
SSB_2014	2014	39.573	(26.45-52.7)	0.608	(0.53-0.686)
SSB_2015	2015	40.033	(26.88-53.19)	0.615	(0.538-0.692)
SSB_2016	2016	40.721	(27.54-53.9)	0.625	(0.55-0.701)
SSB_2017	2017	41.441	(28.23-54.65)	0.637	(0.563-0.71)
SSB_2018	2018	41.778	(28.55-55.01)	0.642	(0.569-0.714)
SSB_2019	2019	41.909	(28.66-55.16)	0.644	(0.571-0.716)

Table d: Recent trend in beginning of the year spawning output and depletion for the southern (south of 40°10′ N. latitude) China rockfish model.

	Year	Spawning Output (billion eggs)	~ 95% confidence interval	Estimated depletion	~ 95% confidence interval
SSB_2010	2010	16.361	(10.75-21.97)	0.246	(0.186-0.306)
SSB_2011	2011	16.444	(10.73-22.16)	0.247	(0.186-0.309)
SSB_2012	2012	16.758	(10.91-22.6)	0.252	(0.189-0.315)
SSB_2013	2013	17.168	(11.18-23.15)	0.258	(0.193-0.323)
SSB_2014	2014	17.899	(11.73-24.07)	0.269	(0.203-0.336)
SSB_2015	2015	18.565	(12.23-24.9)	0.279	(0.211-0.347)
SSB_2016	2016	19.110	(12.62-25.6)	0.287	(0.219-0.356)
SSB_2017	2017	19.462	(12.86-26.06)	0.293	(0.223-0.362)
SSB_2018	2018	20.082	(13.35-26.82)	0.302	(0.231-0.373)
SSB_2019	2019	20.656	(13.8-27.51)	0.311	(0.239-0.382)
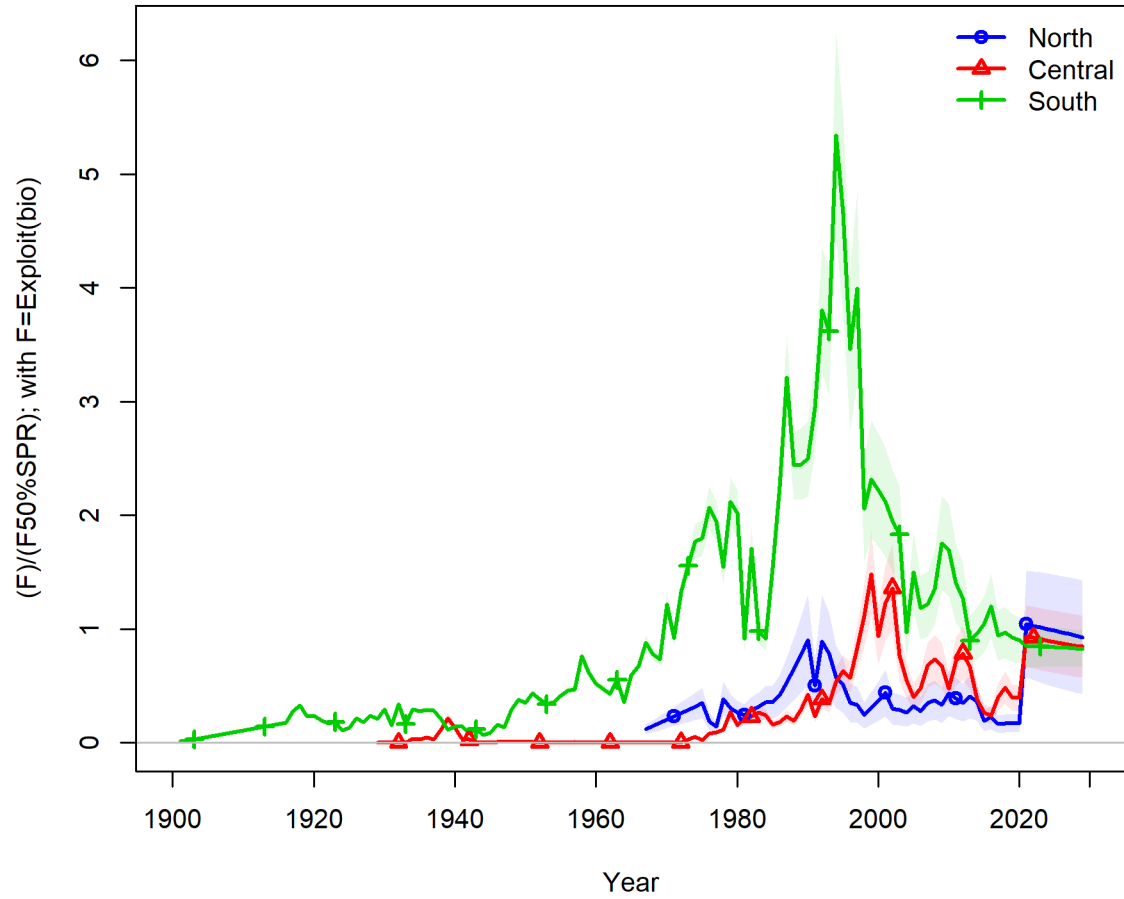
Figure e: Time series of estimated summary harvest rate (points and line: median; shaded areas: 95% credibility intervals) for the three models of China rockfish (North=Washington state, Central = 40°10′ N. latitude to the OR/WA border, and South = south of 40°10′ N. latitude).

11
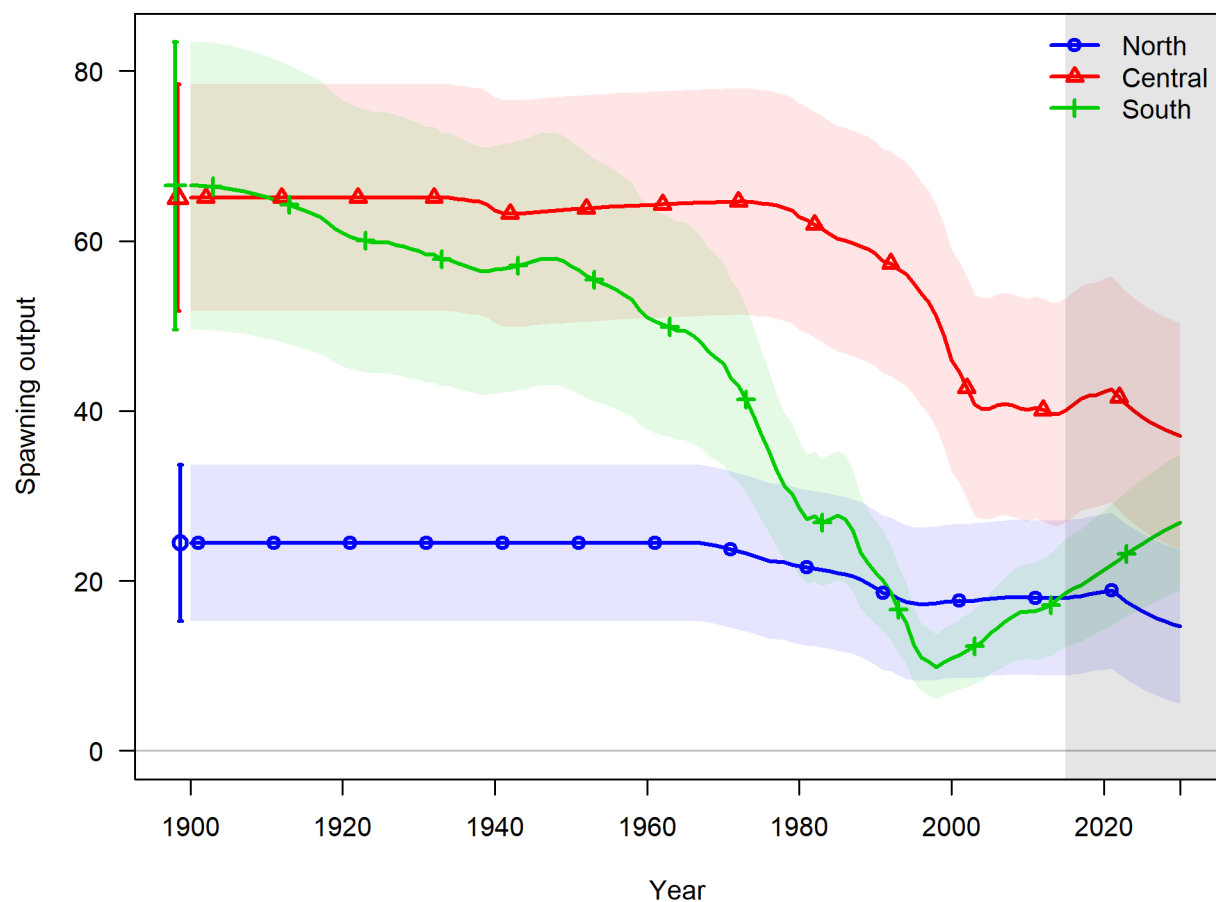
Figure f: Time series of spawning output trajectory (circles and line: median; shaded area: 95% credibility intervals) for the three models of China rockfish (North=Washington state, Central = 40°10′ N. latitude to the OR/WA border, and South = south of 40°10′ N. latitude).

Recruitment

Length and age composition data for China rockfish contain insufficient information to reliably resolve year-class strength. Therefore, all three base models assume that recruitment follows a deterministic Beverton-Holt stock-recruitment relationship, so trends in recruitment reflect trends in estimated spawning output. Given the assumed value of steepness and estimates of current stock status, estimated recruitment has remained fairly constant in the central and northern models, while the estimated biomass in the southern area has declined enough to impact spawning output (Figure g, Tables e, f and g).

Table e: Recent recruitment for the northern model (Washington state MCAs 1-4).

	Year	Estimated Recruitment (1,000s)	~ 95% confidence interval
Recr_2010	2010	33.31	(21.35 - 45.26)
Recr_2011	2011	33.30	(21.34 - 45.25)
Recr_2012	2012	33.29	(21.33 - 45.25)
Recr_2013	2013	33.29	(21.33 - 45.25)
Recr_2014	2014	33.29	(21.33 - 45.25)
Recr_2015	2015	33.29	(21.33 - 45.25)
Recr_2016	2016	33.31	(21.36 - 45.27)
Recr_2017	2017	33.34	(21.39 - 45.29)
Recr_2018	2018	33.37	(21.42 - 45.31)
Recr_2019	2019	33.39	(21.45 - 45.33)

Table f: Recent recruitment for the central model ($40°10'$ N. latitude to the OR/WA border).

	Year	Estimated Recruitment (1,000s)	~ 95% confidence interval
Recr_2010	2010	68.17	(54.47 - 81.87)
Recr_2011	2011	68.22	(54.52 - 81.91)
Recr_2012	2012	68.17	(54.46 - 81.87)
Recr_2013	2013	68.09	(54.36 - 81.81)
Recr_2014	2014	68.06	(54.32 - 81.8)
Recr_2015	2015	68.15	(54.43 - 81.87)
Recr_2016	2016	68.28	(54.58 - 81.98)
Recr_2017	2017	68.41	(54.74 - 82.09)
Recr_2018	2018	68.47	(54.81 - 82.14)
Recr_2019	2019	68.50	(54.84 - 82.16)

Table g: Recent recruitment for the southern model (south of 40°10′ N. latitude).

	Year	Estimated Recruitment (1,000s)	~ 95% confidence interval
Recr_2010	2010	126.27	(109.96 - 142.57)
Recr_2011	2011	126.42	(109.97 - 142.87)
Recr_2012	2012	126.99	(110.52 - 143.46)
Recr_2013	2013	127.71	(111.29 - 144.13)
Recr_2014	2014	128.94	(112.72 - 145.15)
Recr_2015	2015	129.99	(113.95 - 146.03)
Recr_2016	2016	130.80	(114.9 - 146.7)
Recr_2017	2017	131.31	(115.48 - 147.14)
Recr_2018	2018	132.18	(116.5 - 147.85)
Recr_2019	2019	132.94	(117.4 - 148.47)
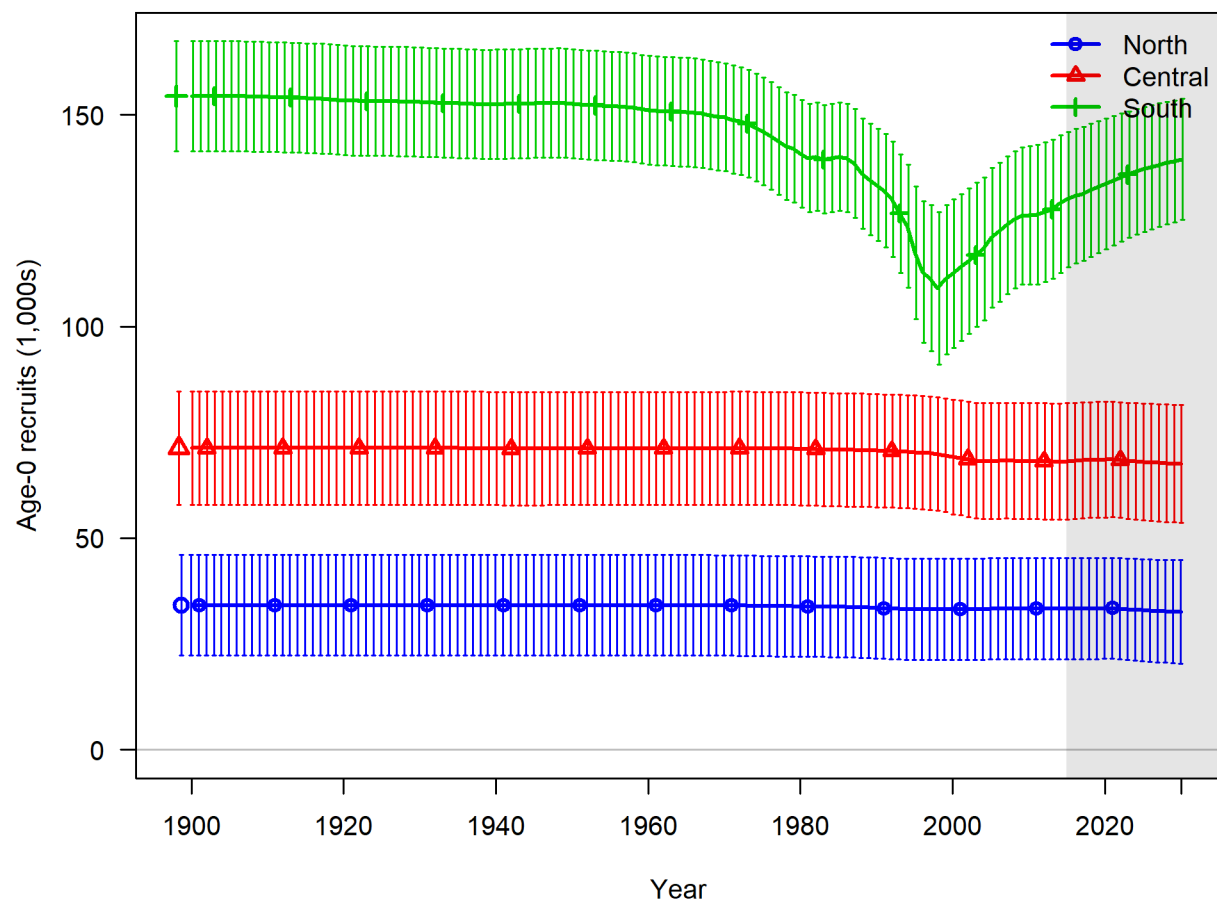
Figure g: Time series of estimated China rockfish recruitments for the three base-case models with 95% confidence or credibility intervals.

Exploitation status

Harvest rates estimated by the northern area model for Washington have never exceeded management target levels (Table h and Figure i). Model results for the central area suggest that harvest rates have briefly exceeded the current proxy MSY value around 2000, but has remained below the management target since then (Table i and Figure i). Historical harvest rates for China rockfish rose steadily in the southern management area until the mid-1990s and exceeded the target SPR harvest rate for several decades, and is just below the target harvest rate as of 2013 (Table j and Figure i). A summary of China rockfish exploitation histories for the northern, central, and southern areas is provided as Figure j.
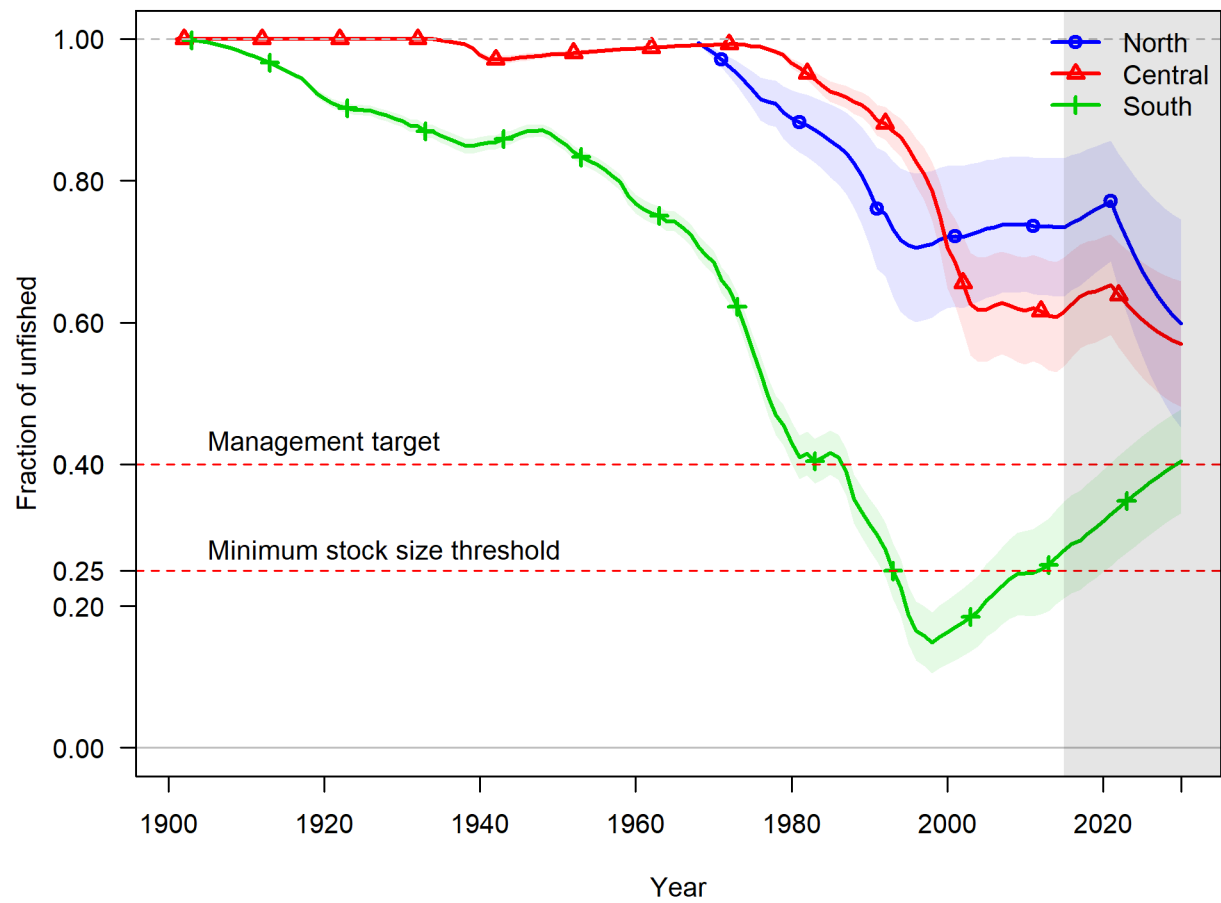
Figure h: Estimated relative depletion with approximate 95% asymptotic confidence intervals (dashed lines) for the three base case assessment models.

Table h: Recent trend in spawning potential ratio and exploitation for the northern China rockfish model (Washington state MCAs 1-4). Fishing intensity is (1-SPR) divided by 50% (the SPR target) and exploitation is F divided by F_{SPR}.

	Year	Fishing intensity	~ 95% confidence interval	Exploitation rate	~ 95% confidence interval
SPRratio_2010	2010	0.559416	(0.36-0.76)	0.437591	(0.24-0.64)
SPRratio_2011	2011	0.512533	(0.32-0.7)	0.388894	(0.21-0.57)
SPRratio_2012	2012	0.477292	(0.3-0.66)	0.353453	(0.19-0.52)
SPRratio_2013	2013	0.528937	(0.34-0.72)	0.405772	(0.22-0.59)
SPRratio_2014	2014	0.484801	(0.3-0.67)	0.361013	(0.19-0.53)
SPRratio_2015	2015	0.295276	(0.17-0.42)	0.195266	(0.1-0.29)
SPRratio_2016	2016	0.334050	(0.2-0.47)	0.226385	(0.12-0.33)
SPRratio_2017	2017	0.251859	(0.14-0.36)	0.162729	(0.09-0.24)
SPRratio_2018	2018	0.263758	(0.15-0.38)	0.172013	(0.09-0.25)
SPRratio_2019	2019	0.262875	(0.15-0.37)	0.171648	(0.09-0.25)

Table i: Recent trend in spawning potential ratio and exploitation for the central China rockfish model (40°10′ N. latitude to the OR/WA border). Fishing intensity is (1-SPR) divided by 50% (the SPR target) and exploitation is F divided by F_{SPR}.

	Year	Fishing intensity	~ 95% confidence interval	Exploitation rate	~ 95% confidence interval
SPRratio_2010	2010	0.613719	(0.48-0.75)	0.474023	(0.33-0.61)
SPRratio_2011	2011	0.802993	(0.65-0.96)	0.715904	(0.5-0.93)
SPRratio_2012	2012	0.850236	(0.69-1.01)	0.787748	(0.55-1.02)
SPRratio_2013	2013	0.770326	(0.62-0.93)	0.668997	(0.47-0.87)
SPRratio_2014	2014	0.530370	(0.4-0.66)	0.385534	(0.27-0.5)
SPRratio_2015	2015	0.399404	(0.29-0.5)	0.265925	(0.19-0.35)
SPRratio_2016	2016	0.363132	(0.27-0.46)	0.237870	(0.17-0.31)
SPRratio_2017	2017	0.543302	(0.42-0.67)	0.403889	(0.29-0.52)
SPRratio_2018	2018	0.620584	(0.49-0.76)	0.488070	(0.35-0.63)
SPRratio_2019	2019	0.536404	(0.41-0.66)	0.396857	(0.28-0.51)

Table j: Recent trend in spawning potential ratio and exploitation for the southern China rockfish model 40°10′ N. latitude). Fishing intensity is (1-SPR) divided by 50% (the SPR target) and exploitation is F divided by F_{SPR}.

	Year	Fishing intensity	~ 95% confidence interval	Exploitation rate	~ 95% confidence interval
SPRratio_2010	2010	1.337550	(1.2-1.48)	1.696970	(1.29-2.1)
SPRratio_2011	2011	1.246290	(1.1-1.4)	1.405290	(1.06-1.75)
SPRratio_2012	2012	1.199550	(1.05-1.35)	1.269020	(0.96-1.58)
SPRratio_2013	2013	1.018540	(0.86-1.18)	0.900269	(0.68-1.12)
SPRratio_2014	2014	1.044190	(0.89-1.2)	0.960927	(0.73-1.19)
SPRratio_2015	2015	1.117330	(0.97-1.27)	1.042890	(0.8-1.29)
SPRratio_2016	2016	1.185510	(1.04-1.33)	1.204450	(0.92-1.49)
SPRratio_2017	2017	1.014270	(0.86-1.17)	0.940792	(0.72-1.16)
SPRratio_2018	2018	1.029420	(0.88-1.18)	0.971817	(0.75-1.2)
SPRratio_2019	2019	0.999773	(0.85-1.15)	0.924289	(0.71-1.13)
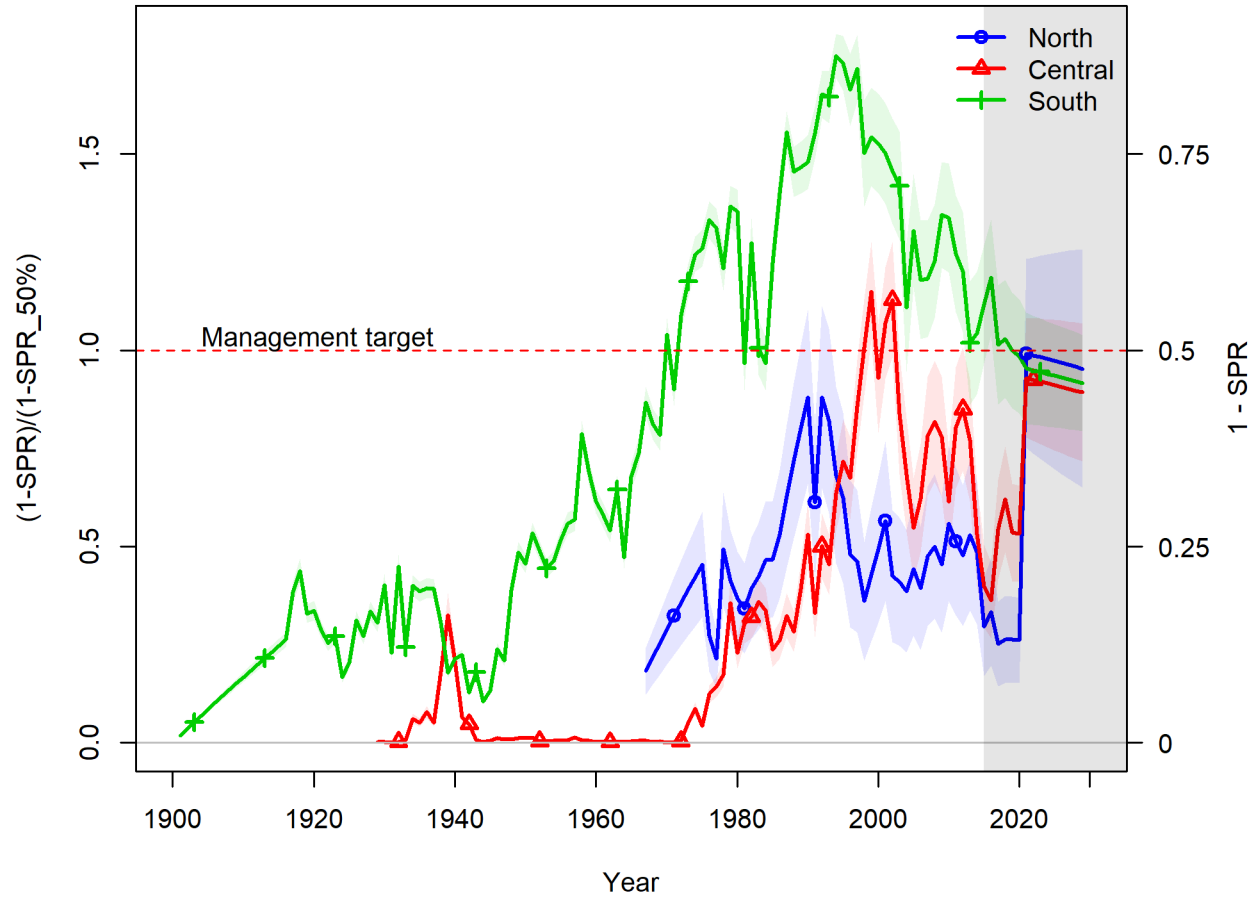
Figure i: Estimated spawning potential ratio (SPR) for the northern, central, and southern base-case models. One minus SPR is plotted so that higher exploitation rates occur on the upper portion of the y-axis. The management target is plotted as a red horizontal line and values above this reflect harvests in excess of the overfishing proxy based on the $SPR_{50\%}$ harvest rate. The last year in the base model is 2014; grey shading indicates forecast period.
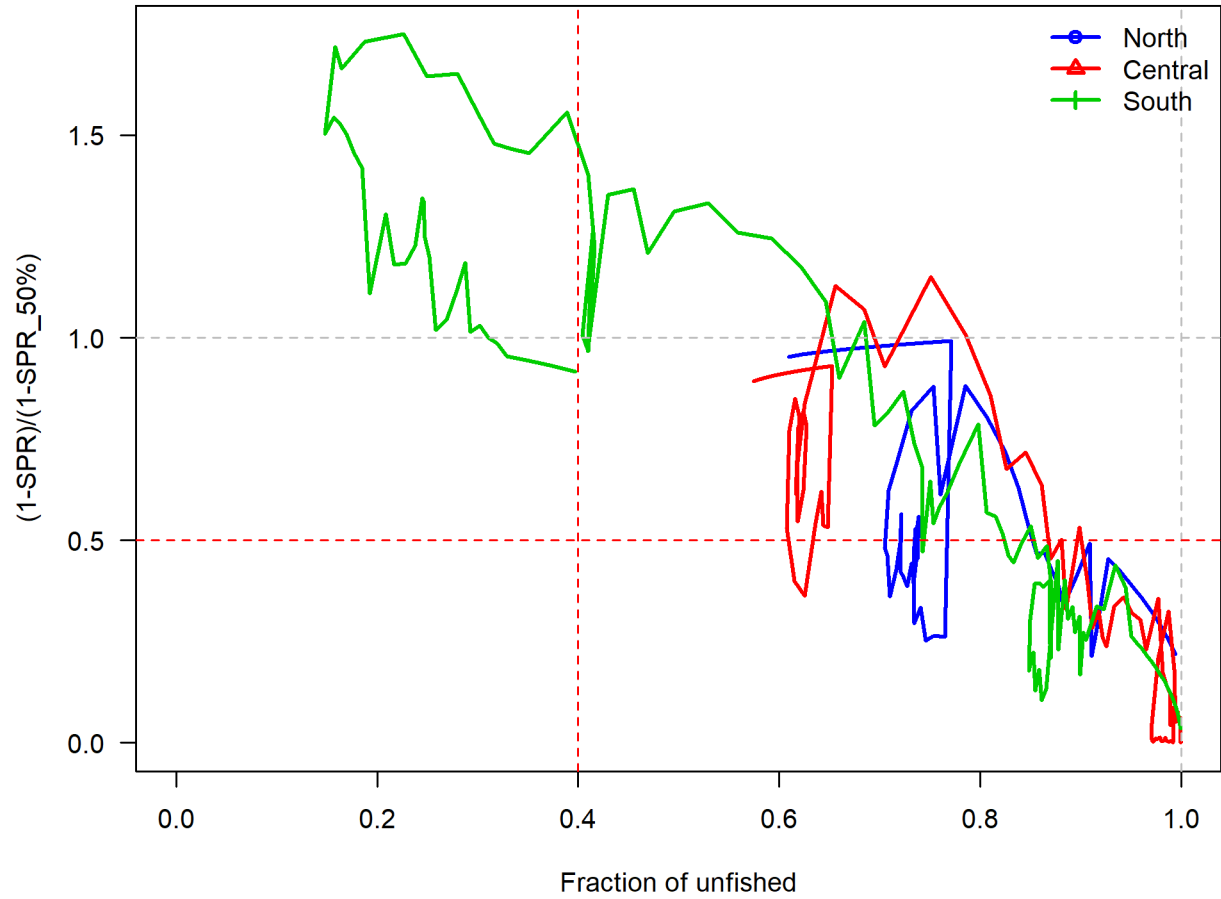
Figure j: Phase plot of estimated relative (1-SPR) vs. relative spawning biomass for the southern, central, and northern base case models. The relative (1-SPR) is (1-SPR) divided by 50% (the SPR target). Relative depletion is the annual spawning biomass divided by the unfished spawning biomass.

Ecosystem considerations

In this assessment, ecosystem considerations were not explicitly included in the analysis. This is primarily due to a lack of relevant data and results of analyses (conducted elsewhere) that could contribute ecosystem-related quantitative information for the assessment.

Recently available habitat information was used to select the data used in the onboard observer indices.

Reference points

Note: The text below and Tables k - l reflect the accepted base case models from the 2015 assessment (Dick et al. 2016).

The management line for China rockfish is at $40°10'$ N. latitude, with differing management guidelines north and south. From 2005-2010, the Nearshore Rockfish Complexes north and south of $40°10'$ N. latitude were managed by a total catch Optimum Yield (OY). As of the Pacific Fishery Management Council (PFMC) 2011-12 management cycle, China rockfish has a component OFL and ABC within the northern and southern Nearshore Rockfish Complexes, based on the work by Dick and MacCall (2010).

This stock assessment estimates that China rockfish in the north are above the biomass target. The spawning output of the stock declined between the 1960s and 1990s but has largely been stable during the past few decades. The estimated relative depletion level in 2015 is 73.4% (~95% asymptotic interval: $\pm$ 63.7% - 83.2%, corresponding to an unfished spawning output of 24.4 billion eggs (~95% asymptotic interval: 15.2 - 33.7 billion eggs) of spawning output in the base model (Table k). Unfished age 5+ biomass was estimated to be 240.8 mt in the base case model. The target spawning output based on the biomass target $(SB_{40\%})$ is 9.8 billion eggs, which gives a catch of 6.3 mt. Equilibrium yield at the proxy F_{MSY} harvest rate corresponding to $SPR_{50\%}$ is 5.8 mt.

This stock assessment estimates that central area China rockfish are just above the biomass target. The rate of spawning output decline is estimated to be steepest during the 1980s to 1990s and has continued to decline since the 1990s at a slower rate. The estimated relative depletion level in 2015 is 61.5% (~95% asymptotic interval: $\pm$ 53.8% - 69.2%), corresponding to an unfished spawning output of 65.1 billion eggs (~95% asymptotic interval: 51.8 - 78.4 billion eggs) of spawning output in the base model (Table l). Unfished age 5+ biomass was estimated to be 591.5 mt in the base case model. The target spawning output based on the biomass target $(SB_{40\%})$ is 26 billion eggs, which gives a catch of 15.7 mt. Equilibrium yield at the proxy F_{MSY} harvest rate corresponding to $SPR_{50\%}$ is 14.5 mt.

This stock assessment estimates that China rockfish south of $40°10'$ N. latitude are below the biomass target, but above the minimum stock size threshold, and have been increasing over the last 15 years. The estimated relative depletion level in 2015 is 27.9% (~95% asymptotic interval: $\pm$ 21.2% - 34.7%), corresponding to an unfished spawning output of 66.5 billion eggs

(~95% asymptotic interval: 49.6 - 83.4 billion eggs) of spawning output in the base model (Table m). Unfished age 5+ biomass was estimated to be 768.6 mt in the base case model. The target spawning output based on the biomass target ($SB_{40\%}$) is 26.6 billion eggs, which gives a catch of 21.1 mt. Equilibrium yield at the proxy F_{MSY} harvest rate corresponding to $SPR_{50\%}$ is 19.5 mt.

Table k: Summary of reference points and management quantities for the northern (Washington state MCAs 1-4) base case model.

	Quantity	Estimate	~95% Confidence Interval
SSB_Unfished	Unfished spawning output (billions of eggs)	24.4	(15.2-33.7)
TotBio_Unfished	Unfished age 5+ biomass (mt)	240.8	(153-328.7)
Recr_Unfished	Unfished recruitment (R0, thousands)	34.2	(22.3-46)
SPB_2015	Spawning output (2015, billions of eggs)	17.9	(8.8-27.1)
Depletion_2015	Depletion (2015)	0.7344	(0.6369-0.8319)
Refpt_sB	**Reference points based on SB_B 40%**		
SSB_Btgt	Proxy spawning output B40%	9.8	(6.1-13.5)
SPR_Btgt	SPR resulting in B40% (SPR_B40%)	0.444	(0.444-0.444)
Fstd_Btgt	Exploitation rate resulting in B40%	0.0551	(0.0522-0.058)
TotYield_Btgt	Yield with SPR_B40% at B_40% (mt)	6.3	(4-8.5)
Refpt_SPR	**Reference points based on SPR proxy for MSY**		
SSB_SPRtgt	Spawning output	11.3	(7-15.5)
SPR_proxy	SPR_{proxy}	0.5	
Fstd_SPRtgt	Exploitation rate corresponding to SPR proxy	0.0458	(0.0435-0.0482)
TotYield_SPRtgt	Yield with SPR proxy at SB SPR (mt)	5.8	(3.7-7.9)
Refpts_MSY	**Reference points based on estimated MSY values**		
SSB_MSY	Spawning output at MSY (SBMSY)	5.6	(3.5-7.8)
SPR_MSY	SPR MSY	0.2875	(0.2823-0.2927)
Fstd_MSY	Exploitation rate at MSY	0.0924	(0.0863-0.0985)
TotYield_MSY	MSY (mt)	7	(4.5-9.4)

Table l: Summary of reference points and management quantities for the central ($40°10'$ N. latitude to the OR/WA border) base case model.

	Quantity	Estimate	~95% Confidence Interval
SSB_Unfished	Unfished spawning output (billions of eggs)	65.1	(51.8-78.4)
TotBio_Unfished	Unfished age 5+ biomass (mt)	591.5	(473.7-709.3)
Recr_Unfished	Unfished recruitment (R0, thousands)	71.3	(57.9-84.6)
SPB_2015	Spawning output (2015, billions of eggs)	40	(26.9-53.2)
Depletion_2015	Depletion (2015)	0.6149	(0.5381-0.6918)
Refpt_sB	**Reference points based on SB_B 40%**		
SSB_Btgt	Proxy spawning output B40%	26	(20.7-31.4)
SPR_Btgt	SPR resulting in B40% (SPR_B40%)	0.444	(0.444-0.444)
Fstd_Btgt	Exploitation rate resulting in B40%	0.0584	(0.0567-0.0602)
TotYield_Btgt	Yield with SPR_B40% at B_40% (mt)	15.7	(12.6-18.7)
Refpt_SPR	**Reference points based on SPR proxy for MSY**		
SSB_SPRtgt	Spawning output	30	(23.8-36.1)
SPR_proxy	SPR_{proxy}	0.5	
Fstd_SPRtgt	Exploitation rate corresponding to SPR proxy	0.0484	(0.0469-0.0498)
TotYield_SPRtgt	Yield with SPR proxy at SB SPR (mt)	14.5	(11.7-17.3)
Refpts_MSY	**Reference points based on estimated MSY values**		
SSB_MSY	Spawning output at MSY (SBMSY)	15.4	(12.2-18.6)
SPR_MSY	SPR MSY	0.2925	(0.29-0.295)
Fstd_MSY	Exploitation rate at MSY	0.098	(0.094-0.1019)
TotYield_MSY	MSY (mt)	17.3	(14-20.7)

Table m: Summary of reference points and management quantities for the southern (south of 40°10′ N. latitude) base case model.

	Quantity	Estimate	~95% Confidence Interval
SSB_Unfished	Unfished spawning output (billions of eggs)	66.5	(49.6-83.4)
TotBio_Unfished	Unfished age 5+ biomass (mt)	768.6	(660.1-877)
Recr_Unfished	Unfished recruitment (R0, thousands)	154.5	(141.5-167.4)
SPB_2015	Spawning output (2015, billions of eggs)	18.6	(12.2-24.9)
Depletion_2015	Depletion (2015)	0.2791	(0.2113-0.3469)
Refpt_sB	**Reference points based on SB_B 40%**		
SSB_Btgt	Proxy spawning output B40%	26.6	(19.8-33.4)
SPR_Btgt	SPR resulting in B40% (SPR_B40%)	0.444	(0.444-0.444)
Fstd_Btgt	Exploitation rate resulting in B40%	0.057	(0.0491-0.065)
TotYield_Btgt	Yield with SPR_B40% at B_40% (mt)	21.1	(19.9-22.3)
Refpt_SPR	**Reference points based on SPR proxy for MSY**		
SSB_SPRtgt	Spawning output	30.6	(22.8-38.4)
SPR_proxy	SPR_{proxy}	0.5	
Fstd_SPRtgt	Exploitation rate corresponding to SPR proxy	0.0476	(0.041-0.0541)
TotYield_SPRtgt	Yield with SPR proxy at SB SPR (mt)	19.5	(18.4-20.6)
Refpts_MSY	**Reference points based on estimated MSY values**		
SSB_MSY	Spawning output at MSY (SBMSY)	15.5	(11.2-19.9)
SPR_MSY	SPR MSY	0.2898	(0.2832-0.2965)
Fstd_MSY	Exploitation rate at MSY	0.0938	(0.0784-0.1092)
TotYield_MSY	MSY (mt)	23.4	(22.1-24.8)

Management performance

Note: The text below and Table n reflect the accepted base case models from the 2015 assessment (Dick et al. 2016). Actual catch has been input for years 2015 and 2016.}

China rockfish is managed in the northern and southern Nearshore Rockfish Complex (split at 40°10′ N. latitude. Since the 2015 management cycle, China rockfish has had a contribution OFL and ACL within each the northern and southern Nearshore Rockfish Complexes (Table n). The estimated catch of China rockfish north of 40°10′ N. latitude of Nearshore Rockfish Complex has been above both the China rockfish contribution to the northern Nearshore Rockfish Complex OFL and ACL in all years (2011-2016). The estimated catch of China rockfish south of 40°10′ N. latitude of Nearshore Rockfish Complex has been below the China rockfish contribution to the northern Nearshore Rockfish Complex OFL and ACL in all years (2011-2016). A summary of these values as well as other base case summary results can be found in Table r.

Table n: Recent trend in total catch and commercial landings (mt) relative to the management guidelines. Estimated total catch reflect the commercial landings plus the model estimated discarded biomass.

Year	Management guideline	Nearshore rockfish north	China contrib. north	Estimated catch north	Nearshore rockfish south	China contrib. south	Estimated catch south
2005	ABC			10.1			16.7
	Total Catch OY	122			615		
2006	ABC			11.3			13.6
	Total Catch OY	122			615		
2007	ABC			15.8			14.2
	Total Catch OY	142			564		
2008	ABC			16.9			16
	Total Catch OY	142			564		
2009	ABC			15.4			21
	Total Catch OY	155			650		
2010	ABC			12.4			19.3
	Total Catch OY	155			650		
2011	ABC	116	11.7	16.6	1156	19.8	16.2
	Total Catch OY	99	9.8		1001	16.5	
2012	ABC	116	11.7	17.5	1145	19.8	14.1
	Total Catch OY	99	9.8		990	16.5	
2013	ABC	110	9.8	15.6	1164	16.6	10.4
	Total Catch OY	94	8.2		1005	13.8	
2014	ABC	110	9.8	10.1	1160	16.6	11.8
	Total Catch OY	94	8.2		1001	13.8	
2015	ABC	88	7.2	6.53	1313	55.2	8.28
	Total Catch OY	69	6.6		1114	50.4	
2016	ABC	88	7.4	6.36	1288	52.7	10.69
	Total Catch OY	69	6.8		1006	50.4	

Unresolved problems and major uncertainties

As in most/all stock assessments, the appropriate value for stock-recruit steepness remains a major uncertainty for China rockfish. In this assessment a prior value was available from a meta-analysis, allowing bracketing of the uncertainty. Exploration of the southern model during the STAR panel meeting established that the range of uncertainty in current and projected biomass status provided by this bracketing was very similar to the range due to natural mortality, and that natural mortality alone would be used to bracket uncertainty in model results for management advice.

While the northern and the southern area models are able to estimate a plausible value of natural mortality with an apparently good level of precision, this was not possible with the central area model.

The fishery-dependent abundance indices used in the assessment are relatively noisy. There is no fishery-independent index. The assessments assume that trends in CPUE indices are representative of population trends.

Assessment results for the central and the northern area models are dependent on the method used for weighting the conditional age-at-length data. This is an area of active research and there is a lack of consensus on an agreed approach. A workshop is planned for later this year that might provide guidance. For this assessment, the Panel recommended use of harmonic mean method, because it is a well-understood and frequently applied method that provided intermediate results compared to other alternatives.

The current term of reference for stock assessment require development of a single decision table with states of nature ranging along the dominant axis of uncertainty. This presumes

that uncertainty is consequential only for a single variable or estimated quantity, such as natural mortality, steepness, or ending biomass. This approach may fail to capture important elements of uncertainty that should be communicated to the Council and its advisory bodies. Additional flexibility in the development of decision tables is needed.

Decision Tables

The forecasts of stock abundance and yield were developed using the final base models. The total catches in 2019 and 2020 are projections. The exploitation rate for 2021 and beyond is based upon an SPR harvest rate of 50%. The average of 2019-2020 catch by fleet was used to distribute catches in forecasted years. The forecasted projections of the OFL for each model are presented in Table r. For the 'constant' catch scenario, catches were input as the 2019-2020 average for that region. The upper catch stream is 50% above the regional forecasted catch in 2021.

Uncertainty in the forecasts is based upon the three states of nature agreed upon at the STAR panel and are based on a low value of M, 0.05, and a high value, 0.09.

Current medium-term forecasts based on the alternative states of nature project that the northern stock, under the current control rule as applied to the base model, will decline towards the target stock size Table o yet remain above 58%. The current control rule under the low state of nature results in a stock decline into the precautionary zone, while the high state of nature maintains the stock at near unfished levels. The upper catch stream combined with a low state of nature resulted in the stock declining to an extremely low level (5%) by 2030, but remaining at or above the target level in the base and high states of nature. Under constant catches, all states of nature remained at or above 54% depletion.

Current medium-term forecasts based on the alternative states of nature for the central model project that the stock, under the current control rule as applied to the base model, will remain above the target stock size (Table p). The current control rule under the low state of nature results in a stock in the precautionary zone, while the high state of nature maintains the stock above 83% depletion from 2021-2030. Removing the high level catches (upper catch stream) under the low M states of nature results in the population going to very low levels during the projection period, and nearing 40% depletion for the last four years even under a base M state of nature. However, removing 2019-2020 average catches results in the stock remaining above the current target stock size under all states of nature.

Assuming that catches beginning in 2021 follow the default ACL harvest control rule, projections of expected China spawning output from the southern base model suggest the stock will be at roughly 33% of unfished spawning output in 2021, and increase to 40% by 2030 (Table q). The stock is expected to remain at or below the target stock size (40% of unfished spawning output) in the base model and "low M" states of nature through 2030 for any of the tested catch scenarios, and to only exceed target size for all years in the "high M" scenario with either default or constant catches, assuming stationarity in the stock-recruitment assumptions.

Table o: Summary of 10-year projections beginning in 2021 for alternate states of nature based on an axis of uncertainty for the northern model. Columns range over low, mid, and high states of nature, and rows range over different assumptions of catch levels.

| | Year | Catch | *States of nature | | | | | |
| | | | Low M 0.05 | | Base M 0.07 | | High M 0.09 | |
			Spawning Output	Depletion	Spawning Output	Depletion	Spawning Output	Depletion
Constant (2019-2020 Average)	2021	1.48	10.73	0.57	18.85	0.77	59.94	0.94
	2022	1.48	10.86	0.58	18.98	0.78	60.07	0.94
	2023	1.48	10.98	0.59	19.11	0.78	60.18	0.94
	2024	1.48	11.10	0.59	19.23	0.79	60.29	0.94
	2025	1.48	11.21	0.60	19.35	0.79	60.39	0.94
	2026	1.48	11.32	0.61	19.45	0.80	60.48	0.94
	2027	1.48	11.42	0.61	19.55	0.80	60.56	0.95
	2028	1.48	11.52	0.62	19.64	0.80	60.64	0.95
	2029	1.48	11.62	0.62	19.73	0.81	60.71	0.95
	2030	1.48	11.71	0.63	19.81	0.81	60.78	0.95
40-10 Rule	2021	9.09	10.73	0.57	18.85	0.77	59.94	0.94
	2022	8.68	10.02	0.54	18.16	0.74	59.27	0.93
	2023	8.32	9.35	0.50	17.54	0.72	58.66	0.92
	2024	7.97	8.74	0.47	16.97	0.69	58.12	0.91
	2025	7.66	8.17	0.44	16.45	0.67	57.64	0.90
	2026	7.38	7.64	0.41	15.99	0.65	57.22	0.89
	2027	7.13	7.17	0.38	15.58	0.64	56.86	0.89
	2028	6.90	6.73	0.36	15.22	0.62	56.55	0.88
	2029	6.68	6.34	0.34	14.90	0.61	56.28	0.88
	2030	6.50	5.98	0.32	14.62	0.60	56.06	0.88
Upper Stream	2021	13.64	10.73	0.57	18.85	0.77	59.94	0.94
	2022	13.64	9.52	0.51	17.68	0.72	58.79	0.92
	2023	13.64	8.32	0.45	16.53	0.68	57.69	0.90
	2024	13.64	7.13	0.38	15.42	0.63	56.63	0.88
	2025	13.64	5.98	0.32	14.36	0.59	55.64	0.87
	2026	13.64	4.86	0.26	13.33	0.55	54.69	0.85
	2027	13.64	3.78	0.20	12.35	0.51	53.81	0.84
	2028	13.64	2.76	0.15	11.43	0.47	52.98	0.83
	2029	13.64	1.80	0.10	10.55	0.43	52.21	0.82
	2030	13.64	0.94	0.05	9.71	0.40	51.49	0.80

Table p: Summary of 10-year projections beginning in 2021 for alternate states of nature based on an axis of uncertainty for the central model. Columns range over low, mid, and high states of nature, and rows range over different assumptions of catch levels.

| | Year | Catch | *States of nature | | | | | |
| | | | Low M 0.05 | | Base M 0.07 | | High M 0.09 | |
			Spawning Output	Depletion	Spawning Output	Depletion	Spawning Output	Depletion
Constant (2019-2020 Average)	2021	7.59	20.66	0.42	42.52	0.65	110.87	0.86
	2022	7.59	20.79	0.42	42.79	0.66	111.19	0.86
	2023	7.59	20.92	0.42	43.05	0.66	111.49	0.87
	2024	7.59	21.03	0.43	43.30	0.67	111.76	0.87
	2025	7.59	21.14	0.43	43.53	0.67	112.02	0.87
	2026	7.59	21.25	0.43	43.76	0.67	112.26	0.87
	2027	7.59	21.36	0.43	43.97	0.68	112.49	0.87
	2028	7.59	21.46	0.43	44.17	0.68	112.70	0.88
	2029	7.59	21.56	0.44	44.36	0.68	112.89	0.88
	2030	7.59	21.66	0.44	44.55	0.68	113.07	0.88
40-10 Rule	2021	18.16	20.66	0.42	42.52	0.65	110.87	0.86
	2022	17.58	19.62	0.40	41.59	0.64	109.99	0.86
	2023	17.06	18.65	0.38	40.75	0.63	109.20	0.85
	2024	16.56	17.75	0.36	40.00	0.61	108.52	0.84
	2025	16.13	16.92	0.34	39.33	0.60	107.92	0.84
	2026	15.73	16.17	0.33	38.74	0.60	107.42	0.84
	2027	15.35	15.48	0.31	38.22	0.59	106.99	0.83
	2028	15.03	14.87	0.30	37.78	0.58	106.64	0.83
	2029	14.73	14.31	0.29	37.40	0.57	106.35	0.83
	2030	14.44	13.80	0.28	37.08	0.57	106.12	0.83
Upper Stream	2021	27.24	20.66	0.42	42.52	0.65	110.87	0.86
	2022	27.24	18.62	0.38	40.56	0.62	108.96	0.85
	2023	27.24	16.59	0.34	38.66	0.59	107.12	0.83
	2024	27.24	14.61	0.30	36.82	0.57	105.38	0.82
	2025	27.24	12.68	0.26	35.05	0.54	103.73	0.81
	2026	27.24	10.82	0.22	33.37	0.51	102.19	0.79
	2027	27.24	9.03	0.18	31.77	0.49	100.75	0.78
	2028	27.24	7.33	0.15	30.25	0.46	99.41	0.77
	2029	27.24	5.72	0.12	28.81	0.44	98.16	0.76
	2030	27.24	4.23	0.09	27.45	0.42	97.00	0.75

Table q: Summary of 10-year projections beginning in 2021 for alternate states of nature based on an axis of uncertainty for the southern model. Columns range over low, mid, and high states of nature, and rows range over different assumptions of catch levels.

| | | | *States of nature | | | | | |
| | | | Low M 0.05 | | Base M 0.07 | | High M 0.09 | |
	Year	Catch	Spawning Output	Depletion	Spawning Output	Depletion	Spawning Output	Depletion
	2021	13.32	15.56	0.22	21.88	0.33	25.38	0.43
	2022	13.32	15.97	0.23	22.49	0.34	26.00	0.44
	2023	13.32	16.39	0.24	23.09	0.35	26.61	0.45
	2024	13.32	16.82	0.24	23.69	0.36	27.20	0.46
	2025	13.32	17.24	0.25	24.29	0.37	27.77	0.47
Constant (2019-2020 Average)	2026	13.32	17.67	0.25	24.87	0.37	28.33	0.48
	2027	13.32	18.10	0.26	25.45	0.38	28.87	0.49
	2028	13.32	18.53	0.27	26.02	0.39	29.39	0.50
	2029	13.32	18.96	0.27	26.58	0.40	29.89	0.51
	2030	13.32	19.38	0.28	27.13	0.41	30.38	0.52
	2021	12.83	15.56	0.22	21.88	0.33	25.38	0.43
	2022	13.06	16.01	0.23	22.53	0.34	26.04	0.44
	2023	13.30	16.45	0.24	23.15	0.35	26.66	0.46
	2024	13.50	16.87	0.24	23.75	0.36	27.25	0.47
	2025	13.69	17.28	0.25	24.33	0.37	27.81	0.47
40-10 Rule	2026	13.89	17.68	0.25	24.88	0.37	28.34	0.48
	2027	14.06	18.07	0.26	25.42	0.38	28.83	0.49
	2028	14.21	18.45	0.27	25.93	0.39	29.29	0.50
	2029	14.33	18.81	0.27	26.42	0.40	29.73	0.51
	2030	14.54	19.17	0.28	26.89	0.40	30.14	0.51
	2021	19.21	15.56	0.22	21.88	0.33	25.38	0.43
	2022	19.21	15.58	0.22	22.05	0.33	25.56	0.44
	2023	19.21	15.61	0.22	22.21	0.33	25.73	0.44
	2024	19.21	15.63	0.23	22.37	0.34	25.90	0.44
	2025	19.21	15.66	0.23	22.54	0.34	26.06	0.45
Upper Stream	2026	19.21	15.70	0.23	22.71	0.34	26.23	0.45
	2027	19.21	15.75	0.23	22.88	0.34	26.40	0.45
	2028	19.21	15.80	0.23	23.06	0.35	26.56	0.45
	2029	19.21	15.86	0.23	23.24	0.35	26.73	0.46
	2030	19.21	15.93	0.23	23.43	0.35	26.90	0.46

Table r: China rockfish base case results summary.

| Year	Predicted OFL (mt)	ABC Catch (mt)	ACL Catch (mt)	Age 5+ Biomass (mt)	Spawning Biomass (mt)	Depletion	Region
2019	**10.65**	**1.48**	**1.48**	**187.62**	**18.56**	**0.76**	**North**
2020	**10.74**	**1.48**	**1.48**	**188.83**	**18.71**	**0.77**	**North**
2021	**10.82**	**9.09**	**9.09**	**189.99**	**18.85**	**0.77**	**North**
2022	**10.43**	**8.68**	**8.68**	**184.03**	**18.16**	**0.74**	**North**
2023	**10.07**	**8.32**	**8.32**	**178.68**	**17.54**	**0.72**	**North**
2024	**9.75**	**7.97**	**7.97**	**173.88**	**16.97**	**0.69**	**North**
2025	**9.45**	**7.66**	**7.66**	**169.62**	**16.45**	**0.67**	**North**
2026	**9.19**	**7.38**	**7.38**	**165.86**	**15.99**	**0.65**	**North**
2027	**8.96**	**7.13**	**7.13**	**162.53**	**15.58**	**0.64**	**North**
2028	**8.75**	**6.90**	**6.90**	**159.59**	**15.22**	**0.62**	**North**
2029	**8.57**	**6.68**	**6.68**	**157.01**	**14.90**	**0.61**	**North**
2030	**8.41**	**6.50**	**6.50**	**154.76**	**14.62**	**0.60**	**North**
2019	21.24	7.60	7.60	396.06	41.91	0.64	Central
2020	21.41	7.58	7.58	398.46	42.22	0.65	Central
2021	21.57	18.16	18.16	400.76	42.52	0.65	Central
2022	21.08	17.58	17.58	393.23	41.59	0.64	Central
2023	20.64	17.06	17.06	386.51	40.75	0.63	Central
2024	20.25	16.56	16.56	380.56	40.00	0.61	Central
2025	19.89	16.13	16.13	375.33	39.33	0.60	Central
2026	19.58	15.73	15.73	370.78	38.74	0.60	Central
2027	19.30	15.35	15.35	366.81	38.22	0.59	Central
2028	19.06	15.03	15.03	363.40	37.78	0.58	Central
2029	18.86	14.73	14.73	360.44	37.40	0.57	Central
2030	18.69	14.44	14.44	357.91	37.08	0.57	Central
2019	**14.39**	**13.32**	**13.32**	**303.18**	**20.66**	**0.31**	**South**
2020	**14.81**	**13.32**	**13.32**	**310.07**	**21.27**	**0.32**	**South**
2021	**15.23**	**12.81**	**12.83**	**316.91**	**21.88**	**0.33**	**South**
2022	**15.68**	**13.06**	**13.06**	**323.97**	**22.53**	**0.34**	**South**
2023	**16.12**	**13.32**	**13.30**	**330.74**	**23.15**	**0.35**	**South**
2024	**16.54**	**13.52**	**13.50**	**337.19**	**23.75**	**0.36**	**South**
2025	**16.94**	**13.72**	**13.69**	**343.38**	**24.33**	**0.37**	**South**
2026	**17.33**	**13.91**	**13.89**	**349.29**	**24.88**	**0.37**	**South**
2027	**17.70**	**14.08**	**14.06**	**354.95**	**25.42**	**0.38**	**South**
2028	**18.05**	**14.23**	**14.21**	**360.38**	**25.93**	**0.39**	**South**
2029	**18.40**	**14.35**	**14.33**	**365.58**	**26.42**	**0.40**	**South**
2030	**18.73**	**14.47**	**14.54**	**370.59**	**26.89**	**0.40**	**South**
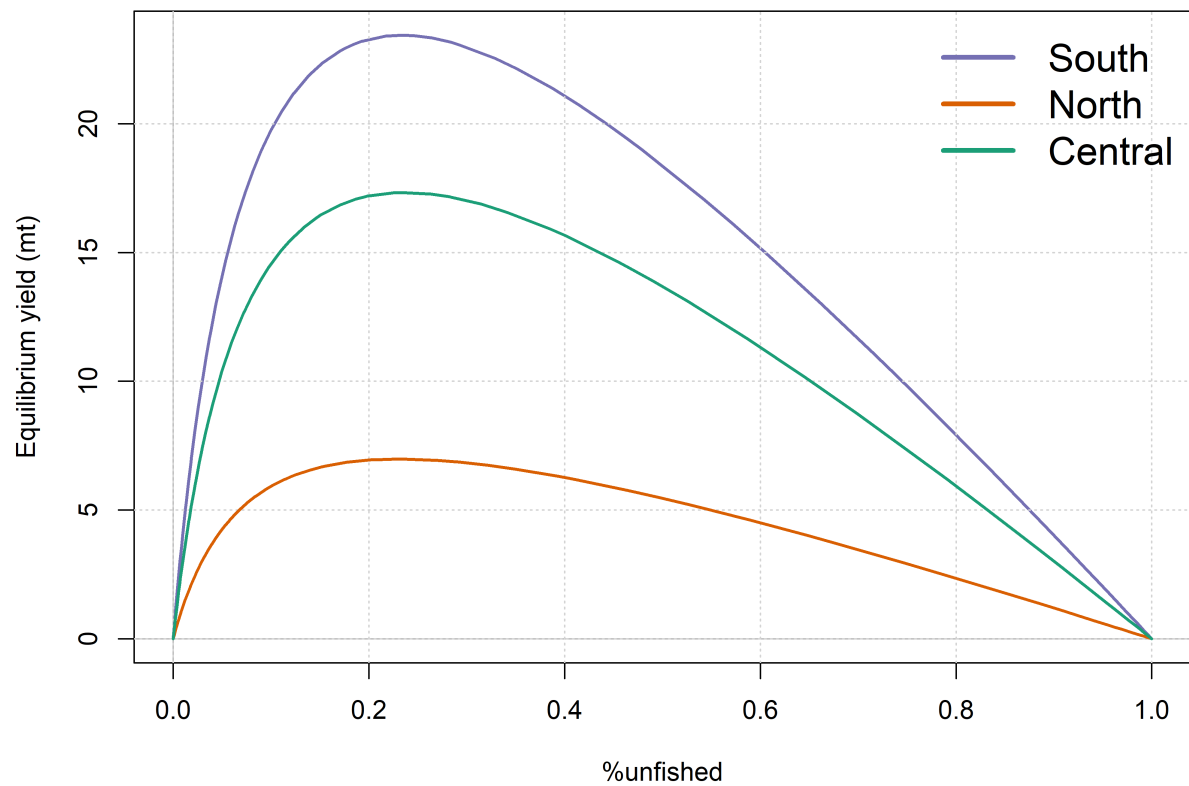
Figure k: Equilibrium yield curve for the base case models. Values are based on the 2014 fishery selectivity and with steepness fixed at 0.773.

Research and data needs

We recommend the following research be conducted before the next assessment:

1. The number of hours fished in Washington should be recorded for each dockside sample (vessel) so that future CPUE can be measured as angler hours rather than just number of anglers per trip. This will allow for a more accurate calculation of effort.

2. The number of hours fished in Oregon should be recorded for each dockside sample (vessel), instead of the start and end times of the entire trip. This will allow for a more accurate calculation of effort.

3. Compare the habitat-based methods used to subset data for the onboard observer indices to Stephens-MacCall and other filtering methods.

4. Explore the sensitivity of Stephens-MacCall when the target species is "rare" or not commonly encountered in the data samples.

5. A standardized fishery independent survey sampling nearshore rockfish in all three states would provide a more reliable index of abundance than the indices developed from catch rates in recreational and commercial fisheries. However, information value of such surveys would depend on the consistency in methods over time and space and would require many years of sampling before an informative index could be obtained.

6. A coastwide evaluation of genetic structure of China rockfish is a research priority. Genetic samples should be collected at sites spaced regularly along the coast throughout the range of the species to estimate genetic differences at multiple spatial scales (i.e., isolation by distance).

7. Difficulties were encountered when attempting to reconstruct historical recreational catches at smaller spatial scales, and in distinguishing between landings from the private and charter vessels. Improved methods are needed to allocate reconstructed recreational catches to sub-state regions within each fishing mode.

8. There was insufficient time during the STAR Panel review to fully review the abundance indices used in the China rockfish assessments. Consideration should be given to scheduling a data workshop prior to STAR Panel review for review of assessment input data and standardization procedures for indices, potentially for all species scheduled for assessment. The nearshore data workshop, held earlier this year, was a step in this direction, but that meeting did not deal with the modeling part of index development.

9. The Marine Recreational Fisheries Statistics Survey (MRFSS) index in Oregon was excluded from the assessment model because it was learned that multiple intercept interviews were done for a single trip. Evaluate whether database manipulations or some other approach can resolve this issue and allow these data to be used in the assessment.

10. Many of the indices used in the China rockfish assessment model used the Stephens-MacCall (2004) approach to subset the CPUE data. Research is need to evaluate the performance of the method when there are changes in management restrictions and in relative abundance of different species. Examination of the characteristics of trips retained/removed should be a routine part of index standardization, such as an evaluation of whether there are time trends in the proportion of discarded trips.

11. Fishery-dependent CPUE indices are likely to be the only trend information for many nearshore species for the foreseeable future. Indices from a multi-species hook-and-line fishery may be influenced by regulatory changes, such as bag limits, and by interactions with other species (e.g., black rockfish) due to hook competition. It may be possible to address many of these concerns if a multi-species approach is used to develop the indices, allowing potential interactions and common forcing to be evaluated.

12. Consider the development of a fishery-independent survey for nearshore stocks. As the current base model structure has no direct fishery-independent measure of stock trends, any work to commence collection of such a measure for nearshore rockfish, or use of existing data to derive such an index would greatly assist with this assessment.

13. Basic life history research may help to resolve assessment uncertainties regarding appropriate values for natural mortality and steepness.

14. Examine length composition data of discarded fish from recreational onboard observer programs in California and Oregon. Consider modeling discarded catch using selectivity and retention functions in Stock Synthesis rather than combining retained and discarded catch and assuming they have identical size compositions. Another option would be to model discarded recreational catch as a separate fleet, similar to the way commercial discards were treated in the southern model.

15. Ageing data were influential in the China rockfish stock assessments. Collection and ageing of China rockfish otoliths should continue. Samples from younger fish not typically selected by the fishery are needed to better define the growth curve.

16. Consider evaluating depletion estimators of abundance using within season CPUE indices. This approach would require information on total removals on a reef-by-reef basis.

17. The extensive use of habitat information in index development is a strength of the China rockfish assessment. Consideration should be given to how to further incorporate habitat data into the assessment of nearshore species. The most immediate need seems to be to increase the resolution of habitat maps for waters off Oregon and Washington, and standardization of habitat data format among states.

18. Although all the current models for China rockfish estimated implausibly large recruitment deviations when allowed to do so, particularly early in the modeled time period,

further exploration of available options in stock synthesis could produce acceptable results. In addition, this work may provide guidance on any additional options that could be added to stock synthesis to better handle this situation. For example, assuming different levels autocorrelation in the stock-recruit relationship for data-moderate stocks may help curb the tendency to estimate extreme recruitment with sparse datasets.

19. Research is needed on data-weighting methods in stock assessments. In particular, a standard approach for conditional age-at-length data is needed. The Center for the Advancement of Population Assessment Methodology (CAPAM) data weighting workshop, scheduled for later this year, should make important progress on this research need.

Cope, J., Dick, E., Maccall, A., Monk, M., Soper, B., and Wetzel, C. 2015. Data-moderate stock assessments for brown, China, copper, sharpchin, stripetail, and yellowtail rockfishes and English and rex soles in 2013. Pacific Fishery Management Council, Portland, OR.

Dick, E.J., and MacCall, A.D. 2010. Estimates of sustainable yield for 50 data-poor stocks in the Pacific Coast Groundfish Fishery Management Plan. NOAA Technical Memorandum NMFS NOAA-TM-NMFS-SWFSC-460.

Dick, E.J., Monk, M., Taylor, I., Haltuch, M., Tsou, T.-S., and Mirick, P. 2016. Status of China rockfish off the U.S. Pacific Coast in 2015. (April 2016): 1–515. Available from http://www.pcouncil.org/wp-content/uploads/2016/05/2015{_}China{_}assessment{_}Final.pdf.