## U.S. ambient air monitoring network has inadequate coverage

- 2 under new PM<sub>2.5</sub> standard
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#### 8 Abstract

- 9 The Clean Air Act (CAA) in the United States relies heavily on regulatory monitoring networks,
- 10 yet monitoring sites are sparsely located, especially among historically disadvantaged
- 11 communities. For ambient fine particulate matter (PM<sub>2.5</sub>), we compare the air quality monitoring
- data with spatially complete concentrations derived from empirical models to quantify the gaps
- of existing U.S. monitoring networks in capturing concentration hotspots and exposure
- disparities. Recently, the U.S. Environmental Protection Agency adopted a more stringent
- annual-average air quality standard for  $PM_{2.5}$  (9  $\mu g/m^3$ ). Here, we demonstrate that 44% of urban
- areas exceeding this new standard encompassing ~ 20 million people would remain
- undetected because of gaps in the current PM<sub>2.5</sub> monitoring network. Crucially, we find that
- 18 "uncaptured" hotspots, which contain 2.8 million people in census tracts that are misclassified as
- in attainment of the new PM<sub>2.5</sub> standard, have substantially higher percentages of minority
- 20 populations (i.e., people of color, disadvantaged communities, and low-income populations)
- 21 compared to the overall US population. To address these gaps, we highlight 10 priority locations
- 22 that could reduce the population in the uncaptured hotspots by 67%. Overall, our findings
- 23 highlight the urgent need to address gaps in the existing monitoring network.
- 24 **Keywords:** PM<sub>2.5</sub>, Clean Air Act, air quality monitoring, environmental justice, NAAQS
- 25 Synopsis: Existing air quality monitoring networks are insufficient to capture concentration
- 26 hotspots, disproportionately impacting minority populations.

## Introduction

	Ambient air pollution causes hundreds of billions of dollars in health damages per year in the
29	United States, driven principally by the health effects of fine particulate matter (PM <sub>2.5</sub> ). These
30	exposures and health burdens disproportionately affect people of color (POC) and low-income
31	populations. <sup>1–3</sup> The US Environmental Protection Agency (EPA), implementing the Clean Air
32	Act (CAA) over the past five decades, has dramatically reduced exposures to criteria air
33	pollutants for hundreds of millions of Americans, yielding enormous health benefits. <sup>4</sup>
34	Nonetheless, we don't all breathe the same air, and major disparities in exposure remain. <sup>2,5–8</sup>
35	The CAA relies on State and Local Air Monitoring Station (SLAMS) networks for determining
36	hotspots and background concentrations, the health and welfare impacts of air pollution, and
37	compliance with the National Ambient Air Quality Standards (NAAQS) (see Supporting
38	Information [SI], Section 1)9. However, due to the high capital and operational cost of
39	monitoring stations, the existing SLAMS network is sparsely located across the US, often
40	missing localized concentration variations 10,11 and causing millions of high-exposure populations
41	to be undetected and unprotected by the monitors. 12-14
12	Moreover, there are disproportionately fewer monitoring sites in communities with higher
43	shares of POC and low-income people. 14-17 While new measurement approaches such as low-
14	cost sensors and mobile monitoring have made denser monitoring networks and high-resolution
45	concentration surfaces feasible, 18-21 such data are still unevenly distributed among those
45 46	concentration surfaces feasible, <sup>18–21</sup> such data are still unevenly distributed among those communities <sup>22–24</sup> and have not been incorporated in the NAAQS nonattainment process.
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46 47	communities <sup>22–24</sup> and have not been incorporated in the NAAQS nonattainment process. On February 7, 2024, EPA revised the annual primary standard for $PM_{2.5}$ , from 12 $\mu g/m^3$ to 9
46 47 48	communities <sup>22–24</sup> and have not been incorporated in the NAAQS nonattainment process. On February 7, 2024, EPA revised the annual primary standard for PM <sub>2.5</sub> , from 12 $\mu$ g/m <sup>3</sup> to 9 $\mu$ g/m <sup>3</sup> . <sup>25</sup> At present, EPA is modifying the PM <sub>2.5</sub> monitoring network design to include an
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- 55 find that the existing SLAMS are inadequate for capturing concentration hotspots and
- disparities. Adding monitors can improve the representation of concentration hotspots, but not
- 57 concentration disparities. This study provides the first quantification of the monitoring gaps
- under the new and future decreasing standards and informs policies for addressing the
- 59 monitoring gaps.

#### **Materials and Methods**

- 61 Air pollution data and attainment status definition
- 62 The U.S. EPA uses ambient measurements from SLAMS to determine whether a specific
- 63 geographical area is in attainment of the NAAQS. Attainment here is assessed for Core-Based
- 64 Statistical Areas (CBSAs), which each correspond to one or more adjoining counties that
- encompass a large urban area or population nucleus. We employ CBSAs because they are
- usually used for determining area-wide air quality levels and planning new monitors (see SI
- 67 Section S1). There are 894 CBSAs in the contiguous U.S., home to 320 million people: 379
- metropolitan statistical areas (MSAs; population  $\geq$  50,000) and 515 micropolitan statistical areas
- 69 (μSAs; population 10,000-49,999).
- 70 To investigate whether SLAMS are potentially missing areas of elevated PM<sub>2.5</sub> in excess of the
- NAAQS, we employ a spatially complete dataset of census-tract level PM<sub>2.5</sub> estimates for the
- 72 contiguous U.S. from the empirical model of the Center for Air, Climate and Energy Solutions
- 73 (CACES, www.caces.us), <sup>28,29</sup> building on partial least squares regressions with universal kriging
- framework. For the model years we consider here (2017-2019), the predictions have high-
- 75 fidelity to out-of-sample validation measurements (R<sup>2</sup>: 0.77-0.83; standardized root mean square
- error: 14%-16%).<sup>29</sup> We compute three-year averaged tract-level concentrations from the annual
- 77 model predictions to match EPA's design values (three-year averaged measurements)<sup>30</sup> and
- further reduce the influence of model uncertainties and extreme events (see SI Section S2).
- Next, we obtain the design values and geographical coordinates of the 2017-2019 active PM<sub>2.5</sub>
- 80 monitoring sites (n = 988) from the EPA's Air Quality System and match them with CACES
- 81 predictions (Figure S1). To further validate the empirical model, we check if model predictions
- 82 correctly classify monitoring sites exceeding 9 μg/m<sup>3</sup> NAAQS (Figure S2). The model's low

- bias makes our conclusions slightly conservative in identifying exceeding tracts. As sensitivity
- tests, we separately employ years 2017, 2018, and 2019 from CACES, and an alternative dataset
- of remotely-sensed  $0.01^{\circ} \times 0.01^{\circ}$  resolution (~1.1 km) PM<sub>2.5</sub> predictions (see SI Section S2).<sup>31</sup>
- For each CBSA, we compare PM<sub>2.5</sub> model predictions at monitoring sites with PM<sub>2.5</sub>
- 87 distributions for all census tracts. EPA determines a CBSA as "nonattainment" if any SLAMS
- 88 monitors' design values exceed the NAAQS. We adapt this by defining nonattainment as having
- 89 three or more tracts within a CBSA exceeding the standard, allowing us to focus on areas with
- 90 elevated concentrations that likely affect thousands of people, rather than small-location
- 91 hotspots. As sensitivity tests, we employ alternative nonattainment definitions (see SI Section
- 92 2). Finally, we classify nonattainment CBSAs by whether the PM<sub>2.5</sub> estimates at monitoring
- 93 locations exceed the NAAQS (see Table S1). CBSAs are considered to be "captured" if they are
- orrectly identified as nonattainment by monitoring locations; and "uncaptured" if they are
- 95 misclassified as in attainment by monitoring locations, but have other unmonitored hotspots
- 96 exceeding the NAAQS. Uncaptured CBSAs are of special concern here. As another sensitivity
- 97 test, we evaluate nonattainment at the county level (see SI Section S2).
- 98 Demographic data and exposure disparities
- 99 We consider three demographic groupings: (1) race-ethnicity, (2) disadvantaged community
- 100 (DAC) status, and (3) median household income, all by Census Tract for 2020. The five racial-
- ethnic groups based on US Census data are: non-Hispanic White (58%; "White"), Latino or
- Hispanic (19%; "Hispanic"), non-Hispanic Black or African American (12%; "Black"), non-
- Hispanic Asian and Pacific Islander (5%; "Asian"), and American Indian, another race, or
- multiracial (3%; "Other"). All except non-Hispanic White are grouped as People of Color
- 105 (POC).
- Second, DACs are defined by combining six publicly available national screening tools from the
- 107 federal government (see SI Section S3; Table S2). We identify a census tract as DAC if it
- surpasses the specified thresholds by three or more tools (~25% of the total US population;
- Figures S3-S6). The reasons for combining six tools are to avoid the ineffectiveness or
- uncertainty in any single tool<sup>32</sup> and to highlight locations of highest concern or federal funding.

Third, median household income is from the 2020 American Community Survey. We classify income into tertiles: high (> \$76,164), middle (\$51,168 - \$76,164), and low (<\$51,168).

We calculate  $PM_{2.5}$  exposure disparities by race-ethnicity, DAC status, and household income, respectively, as the population-weighted average concentration for POC, DAC, and low-income populations, minus the overall population-average concentration. Disparities are calculated for all census tracts and tracts near monitors (n = 4,360; defined here as centroids within 1-km buffer of a monitoring site).

#### **Results and Discussion**

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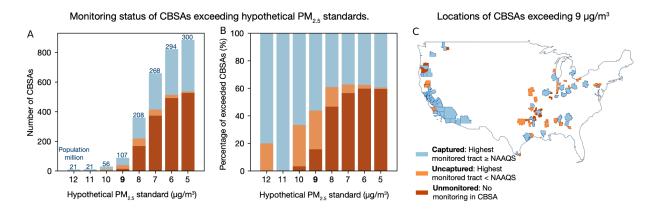


Figure 1. Core-based statistical areas (CBSAs) exceed a hypothetical PM<sub>2,5</sub> standard, classified by monitoring status. Here, we consider only those CBSAs with three or more census tracts that have modeled PM<sub>2.5</sub> exceeding a range of hypothetical PM<sub>2.5</sub> standards, which we thereby consider to be in nonattainment. We classify the (a) number and (b) percentage of CBSAs exceeding the PM<sub>2.5</sub> standard into three distinct groups. In blue, we present "captured" CBSAs. These CBSAs are correctly identified as exceeding the standard, by virtue of having monitors located in tracts that exceed the standard. In orange, we present "uncaptured CBSAs," which would be misclassified as in attainment based on present monitoring locations. In these uncaptured CBSAs, the highest monitored tract does not exceed the standard, despite other unmonitored hotspot tracts exceeding the standard. Finally in red, we show CBSAs that exceed a given standard value and have no monitors at all. There are no red or orange bars in (a) and (b) when the standard is set at 11 µg/m³, as the highest concentrations in all nonattainment CBSAs fall between 11 and 12 µg/m<sup>3</sup>. Therefore, with a standard of 11 µg/m³, there are no "uncaptured" or "unmonitored" CBSAs (see Figure S8 for details). In (c), we illustrate the geographic distribution of CBSAs for the new PM<sub>2.5</sub> NAAQS of 9  $\mu g/m^3$ .

The median number of PM<sub>2.5</sub> monitoring stations in an MSA is 1 (μSA: 0) (population-weighted median: 5 [MSA], 1 [μSA]; Figure S7). On average, there is one site per 250,000 people. For NAAQS attainment status, the results reveal that 89 CBSAs (total population: 107 million) exceed the new PM<sub>2.5</sub> standard (9 μg/m³) (Figure 1a). Among the nonattainment CBSAs, 44% (n=39; 20 million people) are not captured by monitoring (Figures 1b and S8), because the CBSA has either no monitoring stations or the existing locations fail to capture the concentration hotspots (see Figure S9 for case studies). Most uncaptured nonattainment CBSAs are in the Midwest and South (Figure 1c). The estimations of monitoring gaps are robust considering model errors, using alternative nonattainment definitions, separate years, alternative concentration data, and at the county level (see SI Section S2; Figures S10-S13; Table S3). Under future decreasing standards (e.g., to the World Health Organization guideline, 5 μg/m³), ~60% of nonattainment CBSAs would not be captured by existing monitors (Figure 1b).

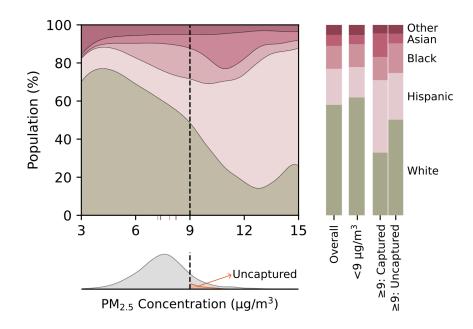
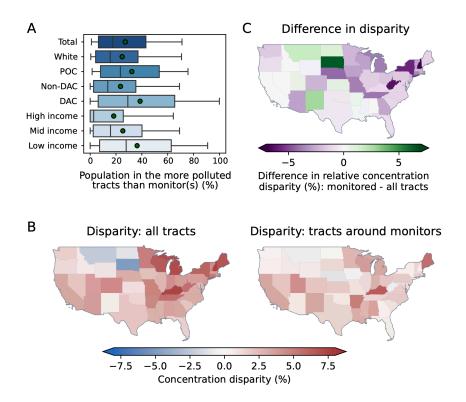


Figure 2. Tract-level racial-ethnic composition under different PM<sub>2.5</sub> exposure levels. Instead of considering the whole CBSAs (Figure 1), this figure considers the census tracts themselves. If one high-exposure census tract ( $\geq 9~\mu g/m^3$ ) is in the captured CBSAs (blue colors in Figure 1 c), then the tract is defined as captured; otherwise, the high-exposure census tract is defined as uncaptured. Left panel: tract-level racial-ethnic composition (White, Hispanic, Black, Asian, or Other; upper row) and concentration distribution (population-weighted; bottom row)

across the concentration range (3-15 µg/m<sup>3</sup>). The new standard (9 µg/m<sup>3</sup>) is represented as 157 158 black dashed lines. The uncaptured high-exposure tracts (≥ 9 µg/m³) are represented by the 159 orange shadow (bottom-left panel). Right panel: racial-ethnic composition for (i) overall census 160 tracts; (ii) census tracts with concentrations < 9 µg/m³; (iii) census tracts with concentrations ≥ 9 161 µg/m<sup>3</sup> and located in nonattainment CBSAs that are captured by monitors (blue color in Figure 162 1c); (iv) census tracts with concentrations  $\geq 9 \mu g/m^3$  and not in the captured nonattainment 163 CBSAs. There are three reasons for non capturing: the census tracts are in nonattainment 164 CBSAs that are uncaptured by monitors (orange and red colors in Figure 1c); the CBSAs where 165 the census tracts are located don't have three or more census tracts exceeding the standard; or 166 the census tracts are rural tracts (not within any CBSAs). The latter two reasons include high-167 exposure tracts that are not located in the blue-, orange-, or red-colored CBSAs in Figure 1c. 168 169 Considering only the census tracts exceeding 9 µg/m<sup>3</sup> PM<sub>2.5</sub> (i.e., only the tracts themselves, 170 rather than the whole CBSAs; "hotspot" tracts), 44 million people (14% of the U.S. population) 171 live in exceeding tracts, of which most (41 million) live in tracts captured by monitors, and the rest (2.8 million) live in tracts not captured by monitors (Figures 2 and S14). The average 172 173 concentration in the captured hotspots (10.2 µg/m<sup>3</sup>) is higher than the uncaptured hotspots (9.2 174 μg/m<sup>3</sup>). Crucially, both captured and uncaptured hotspots contain higher percentages of POC 175 (68% and 50%, respectively) compared to the overall population (42%) (Figure 2). Those 176 hotspots also contain higher percentages of DAC (42% and 41%) and low-income populations 177 (28% and 39%) than the overall population (25% [DAC]; 28% [low-income]; Figures S15-S16). 178 Minority population percentages in the uncaptured hotspots are higher than the state average in 179 most states (Figure S17). This suggests that the existing monitors are insufficient to identify 180 concentration hotspots, disproportionately impacting minority populations. According to the 181 Code of Federal Regulations, 9 regulatory monitors primarily focus on area-wide air quality, not 182 concentration hotspots. However, the EPA is planning new monitors in at-risk communities, 183 including minority communities, to capture source impacts (see SI Section S1). Our results 184 indicate that new monitors are essential for detecting hotspots in those communities. 185 We also examine whether monitoring locations represent exposure hotspots, average exposure 186 levels, and disparities by demographic group. On average, 23% of the overall population lives in 187 census tracts with higher concentrations than the highest monitored concentrations in the

CBSAs. However, for POC, DAC, and low-income populations, the numbers are 32%, 39% and 36%, respectively, indicating that monitoring is less representative of the upper bounds of the population-concentration distribution for these minority groups (Figure 3a). Comparing the concentration disparities for all census tracts and tracts around monitors, monitored locations underestimate state-level disparities in most states (Figures 3b-3c, S18-S19). For example, the national racial-ethnic relative disparity of PM<sub>2.5</sub> concentration is 6.1% for all tracts; the relative disparity for tracts around monitors is only 4.3% (a 30% underestimation).



**Figure 3.** Representativeness of monitoring locations for exposure hotspots and exposure disparities by demographics. (A) Percentages of populations in each CBSA that are exposed to concentrations higher than the maximum concentrations in the monitored tracts. Populations are grouped by race-ethnicity, DAC status, and income levels. The box-and-whisker represents the 10th, 25th, 50th, 75th, and 90th percentiles, and the green circle represents the population-weighted mean. (B) State-level racial-ethnic concentration (relative) disparities in PM<sub>2.5</sub> for all census tracts and census tracts around (within 1-km circular buffer) monitoring sites. (C) The difference in the two disparities, calculated as disparities for all census tracts minus disparities for tracts around monitors. The purple colors represent that the monitoring locations underestimate racial-ethnic disparities; the green colors represent that monitoring locations overestimate racial-ethnic disparities.

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Lastly, we examine approaches for addressing these monitoring gaps and disparities (see SI Section S4), consistent with recent federal and state legislation supporting enhanced monitoring for DACs. <sup>33,34</sup> Here, we present an approach for prioritizing new monitor locations, following a simple scheme that identifies optimal census tracts for monitoring based on the size of the additional population in census tracts that would be newly captured (i.e., correctly reclassified as nonattainment) through the addition of a marginal monitoring site (see SI for full details). Our results imply that adding only 10 new monitor locations could reduce the population in the uncaptured hotspots by 67% (from 2.8 million to 0.9 million; Figure 4). This approach would reduce the percentage of POC population in uncaptured hotspots by 20% (from 50% to 40%; Figure S20), but would provide less benefit to DAC and low-income populations (see Figures S20-S22 for other approaches, which might better target those subpopulations). Nonetheless, although adding a small number of targeted monitor locations could sharply reduce the number of people "uncaptured" by the existing monitoring network, it would not meaningfully improve the ability of the SLAMS to characterize nationwide PM<sub>2.5</sub> disparities (Figure S23). To accurately evaluate exposure disparities, other methods or tools (instead of regulatory monitoring), with much finer spatial resolution, are likely needed.

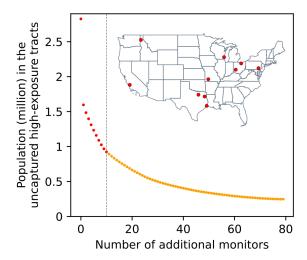


Figure 4. Number of remaining population residing in high concentration census tracts that are not captured by monitoring (total = 2.8 million people). By selecting the first 10

CBSAs with the highest number of people residing in uncaptured census tracts (10 red locations), and adding one additional appropriately-sited monitor to each CBSA, the population remaining in uncaptured hotspots would be reduced by 67% to 0.9 million people. The addition of these monitors would result in each of these 10 CBSAs (total population = 13 million) being classified as in non-attainment of the new PM2.5 NAAQS based on the 2017-2019 design value. Note that after all hotspots in the CBSAs are captured, there remains a non-urban high-exposure population of ~ 0.2 million people that is located outside of the CBSAs.

Our study comprehensively quantifies gaps and disparities in the existing regulatory monitoring

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### Implications for future policy

networks, revealing the following key points. First, the existing SLAMS regulatory monitoring network fails to capture 44% of nonattainment CBSA under the new PM<sub>2.5</sub> NAAQS, providing inadequate protection to tens of millions of highly-exposed people. These uncaptured populations are higher than previously documented under the old PM<sub>2.5</sub> standards, <sup>12–14</sup> highlighting the urgent need for additional monitors to implement the new standard effectively. Second, existing monitoring networks have disproportionately less coverage among the highexposure minority populations. Those populations are already more vulnerable and sensitive to the health effects from air pollution. 35,36 Our findings indicate that adding a small number of additional monitors can noticeably reduce the number of unmonitored exceeding locations; that step will benefit the overall population and help reduce injustices via State Implementation Plans. Third, the monitoring stations underestimate exposure disparities. Unfortunately, adding a moderate number of monitors would be ineffective at addressing this gap (Figure S23). Indeed, since empirical models may underestimate hotspot concentrations, <sup>2,37</sup> the true underestimation in disparities by the monitoring networks is likely to be even greater than is estimated here. Our results imply that other technologies and tools with higher spatial resolution, such as mobile monitoring, <sup>37–40</sup> well-calibrated low-cost or portable sensors, <sup>22–24,41–44</sup> and satellite-based models, <sup>45–49</sup> could aid in representing exposure hotspots and disparities. Those tools may also be useful for nonattainment designation. Thus, an important open question is whether new

256	data/tools need to be incorporated into the CAA policies (e.g., for identifying at-risk
257	communities, planning new monitors, and determining attainment/nonattainment status).
258	Our study informs the implementation of the new PM <sub>2.5</sub> NAAQS, in terms of regulatory
259	monitoring. Our findings reveal that as the "umbrella" to protect the US population, the existing
260	PM <sub>2.5</sub> SLAMS network has consequential monitoring gaps. Effective and straightforward
261	solutions exist (i.e., adding a small number of monitors) to address the monitoring gaps
262	identified here. Doing so would protect the overall population, but would not substantially
263	change the underestimation of disparities by the monitoring network. Our results use 2017-2019
264	data, while EPA's nonattainment designations will rely on post-November 2024 measurements.
265	However, air quality trends have been broadly steady since 2016 (Figure S24), suggesting our
266	findings offer insight into near-future attainment status, though actual concentrations may differ.
267	Previous research indicated that simply tightening NAAQS standards without targeting specific
268	locations will not address disparities. <sup>8,32</sup> Therefore, improvement in monitoring networks,
269	incorporating other high-resolution tools, and more effective location-based strategies are all
270	urgently needed, in addition to stricter NAAQS standards, to address exposure disparities. Future
271	studies could further investigate state-level solutions for reducing pollution levels, eliminating
272	disparities, and designing monitoring networks to support both goals. Our methodologies for
273	investigating monitoring gaps may apply to other pollutants (e.g., nitrogen dioxide).
274	Acknowledgment
275	This publication was developed with funding from Google.org (Project TF2203-106429).
276	Supporting information
277	Additional methodological details, sensitivity analyses, and supporting information tables and
278	figures.
279	References
280	(1) Clark, L. P.; Millet Dylan B.; Marshall Julian D. Changes in Transportation-Related Air

Pollution Exposures by Race-Ethnicity and Socioeconomic Status: Outdoor Nitrogen

Dioxide in the United States in 2000 and 2010. Environ. Health Perspect. 2017, 125 (9),

11

097012. https://doi.org/10.1289/EHP959.

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- (2) Liu, J.; Clark, L. P.; Bechle, M. J.; Hajat, A.; Kim, S.-Y.; Robinson, A. L.; Sheppard, L.;
   Szpiro, A. A.; Marshall, J. D. Disparities in Air Pollution Exposure in the United States by
   Race/Ethnicity and Income, 1990–2010. *Environ. Health Perspect.* 2021, 129 (12), 127005.
   https://doi.org/10.1289/EHP8584.
- (3) Geldsetzer, P.; Fridljand, D.; Kiang, M. V.; Bendavid, E.; Heft-Neal, S.; Burke, M.; Thieme,
   A. H.; Benmarhnia, T. Disparities in Air Pollution Attributable Mortality in the US
   Population by Race/Ethnicity and Sociodemographic Factors. *Nat. Med.* 2024, 1–9.
   https://doi.org/10.1038/s41591-024-03117-0.
- 292 (4) Currie, J.; Voorheis, J.; Walker, R. What Caused Racial Disparities in Particulate Exposure 293 to Fall? New Evidence from the Clean Air Act and Satellite-Based Measures of Air Quality. 294 *Am. Econ. Rev.* **2023**, *113* (1), 71–97. https://doi.org/10.1257/aer.20191957.
- (5) Tessum, C. W.; Paolella, D. A.; Chambliss, S. E.; Apte, J. S.; Hill, J. D.; Marshall, J. D.
   PM<sub>2.5</sub> Polluters Disproportionately and Systemically Affect People of Color in the United
   States. Sci. Adv. 2021, 7 (18), eabf4491. https://doi.org/10.1126/sciadv.abf4491.
- 298 (6) Colmer, J.; Hardman, I.; Shimshack, J.; Voorheis, J. Disparities in PM<sub>2.5</sub> Air Pollution in the United States. *Science* **2020**, *369* (6503), 575–578. https://doi.org/10.1126/science.aaz9353.
- 300 (7) Jbaily, A.; Zhou, X.; Liu, J.; Lee, T.-H.; Kamareddine, L.; Verguet, S.; Dominici, F. Air Pollution Exposure Disparities across US Population and Income Groups. *Nature* **2022**, *601* (7892), 228–233. https://doi.org/10.1038/s41586-021-04190-y.
- (8) Wang, Y.; Apte, J. S.; Hill, J. D.; Ivey, C. E.; Patterson, R. F.; Robinson, A. L.; Tessum, C.
   W.; Marshall, J. D. Location-Specific Strategies for Eliminating US National Racial-Ethnic
   PM<sub>2.5</sub> Exposure Inequality. *Proc. Natl. Acad. Sci.* 2022, 119 (44), e2205548119.
   https://doi.org/10.1073/pnas.2205548119.
- (9) Kelp, M. M.; Lin, S.; Kutz, J. N.; Mickley, L. J. A New Approach for Determining Optimal
   Placement of PM<sub>2.5</sub> Air Quality Sensors: Case Study for the Contiguous United States.
   Environ. Res. Lett. 2022, 17 (3), 034034. https://doi.org/10.1088/1748-9326/ac548f.
- 310 (10) 40 CFR Part 58 -- Ambient Air Quality Surveillance. https://www.ecfr.gov/current/title-311 40/part-58 (accessed 2024-09-08).
- (11) Di, Q.; Amini, H.; Shi, L.; Kloog, I.; Silvern, R.; Kelly, J.; Sabath, M. B.; Choirat, C.;
   Koutrakis, P.; Lyapustin, A.; Wang, Y.; Mickley, L. J.; Schwartz, J. An Ensemble-Based
   Model of PM<sub>2.5</sub> Concentration across the Contiguous United States with High
   Spatiotemporal Resolution. *Environ. Int.* 2019, 130, 104909.
   https://doi.org/10.1016/j.envint.2019.104909.
- Fowlie, M.; Rubin, E.; Walker, R. Bringing Satellite-Based Air Quality Estimates Down to Earth. *AEA Pap. Proc.* **2019**, *109*, 283–288. https://doi.org/10.1257/pandp.20191064.
- 319 (13) Sullivan, D.; Krupnick, A. *Using Satellite Data to Fill the Gaps in the US Air Pollution*320 *Monitoring Network*; RFF Working Paper Series 18–21; Resources for the Future, 2018.
  321 https://econpapers.repec.org/paper/rffdpaper/dp-18-21.htm (accessed 2024-06-11).
- (14) Dobkin, F.; Kerr, G. Demographic Disparities in United States Clean Air Act
   PM<sub>2.5</sub> Attainment Counties: Assessing Population Living in Nonattainment Conditions. *J. Environ. Stud. Sci.* 2024. https://doi.org/10.1007/s13412-024-00933-1.
- 325 (15) Grainger, C.; Schreiber, A. Discrimination in Ambient Air Pollution Monitoring? *AEA* 326 *Pap. Proc.* **2019**, *109*, 277–282. https://doi.org/10.1257/pandp.20191063.
- (16) Pedde, M.; Adar, S. D. Representativeness of the US EPA PM Monitoring Site Locations
   to the US Population: Implications for Air Pollution Prediction Modeling. *J. Expo. Sci. Environ. Epidemiol.* 2024, 1–6. https://doi.org/10.1038/s41370-024-00644-3.

- Kelp, M. M.; Fargiano, T. C.; Lin, S.; Liu, T.; Turner, J. R.; Kutz, J. N.; Mickley, L. J.
   Data-Driven Placement of PM<sub>2.5</sub> Air Quality Sensors in the United States: An Approach to
   Target Urban Environmental Injustice. *GeoHealth* 2023, 7 (9), e2023GH000834.
   https://doi.org/10.1029/2023GH000834.
- Apte, J. S.; Messier, K. P.; Gani, S.; Brauer, M.; Kirchstetter, T. W.; Lunden, M. M.;
  Marshall, J. D.; Portier, C. J.; Vermeulen, R. C. H.; Hamburg, S. P. High-Resolution Air
  Pollution Mapping with Google Street View Cars: Exploiting Big Data. *Environ. Sci. Technol.* 2017, 51 (12), 6999–7008. https://doi.org/10.1021/acs.est.7b00891.
- 338 (19) Barkjohn, K. K.; Gantt, B.; Clements, A. L. Development and Application of a United 339 States-Wide Correction for PM<sub>2.5</sub> Data Collected with the PurpleAir Sensor. *Atmospheric* 340 *Meas. Tech.* **2021**, *14* (6), 4617–4637. https://doi.org/10.5194/amt-14-4617-2021.
- (20) Considine, E. M.; Braun, D.; Kamareddine, L.; Nethery, R. C.; deSouza, P. Investigating
   Use of Low-Cost Sensors to Increase Accuracy and Equity of Real-Time Air Quality
   Information. *Environ. Sci. Technol.* 2023, 57 (3), 1391–1402.
   https://doi.org/10.1021/acs.est.2c06626.
- 345 (21) Apte, J. S.; Manchanda, C. High Resolution Air Pollution Mapping. *Science* **2024**, *In Press*.

348

349

354

355

356

357

358

359

360

361

362

363

- (22) deSouza, P.; Kinney, P. L. On the Distribution of Low-Cost PM<sub>2.5</sub> Sensors in the US: Demographic and Air Quality Associations. *J. Expo. Sci. Environ. Epidemiol.* **2021**, *31* (3), 514–524. https://doi.org/10.1038/s41370-021-00328-2.
- (23) Lu, T.; Liu, Y.; Garcia, A.; Wang, M.; Li, Y.; Bravo-villasenor, G.; Campos, K.; Xu, J.;
   Han, B. Leveraging Citizen Science and Low-Cost Sensors to Characterize Air Pollution
   Exposure of Disadvantaged Communities in Southern California. *Int. J. Environ. Res.* Public. Health 2022, 19 (14), 8777. https://doi.org/10.3390/ijerph19148777.
  - (24) Sun, Y.; Mousavi, A.; Masri, S.; Wu, J. Socioeconomic Disparities of Low-Cost Air Quality Sensors in California, 2017–2020. *Am. J. Public Health* **2022**, *112* (3), 434–442. https://doi.org/10.2105/AJPH.2021.306603.
  - (25) US EPA. Reconsideration of the National Ambient Air Quality Standards for Particulate Matter. Federal Register. https://www.federalregister.gov/documents/2024/03/06/2024-02637/reconsideration-of-the-national-ambient-air-quality-standards-for-particulate-matter (accessed 2024-07-08).
  - (26) US EPA. *Biden-Harris Administration Announces \$53 Million for 132 Community Air Pollution Monitoring Projects Across the Nation*. https://www.epa.gov/newsreleases/biden-harris-administration-announces-53-million-132-community-air-pollution (accessed 2023-11-20).
- US EPA. EPA Announces an Additional \$50 Million Under the American Rescue Plan to
   Enhance Air Pollution Monitoring. https://www.epa.gov/newsreleases/epa-announces additional-50-million-under-american-rescue-plan-enhance-air-pollution (accessed 2023-11 20).
- (287) Kim, S.-Y.; Bechle, M.; Hankey, S.; Sheppard, L.; Szpiro, A. A.; Marshall, J. D.
   Concentrations of Criteria Pollutants in the Contiguous U.S., 1979 2015: Role of
   Prediction Model Parsimony in Integrated Empirical Geographic Regression. *PLOS ONE* 2020, 15 (2), e0228535. https://doi.org/10.1371/journal.pone.0228535.
- Lu, T.; Kim, S.-Y.; Marshall, J. D. High-Resolution Geospatial Database: National
   Criteria-Air-Pollutant Concentrations in the Contiguous U.S., 2016-2020. *ChemRxiv* 2024.
   https://10.26434/chemrxiv-2024-zg5f1.

- 376 (30) US EPA, O. *Air Quality Design Values*. https://www.epa.gov/air-trends/air-quality-design-values (accessed 2024-06-11).
- van Donkelaar, A.; Hammer, M. S.; Bindle, L.; Brauer, M.; Brook, J. R.; Garay, M. J.;
  Hsu, N. C.; Kalashnikova, O. V.; Kahn, R. A.; Lee, C.; Levy, R. C.; Lyapustin, A.; Sayer, A.
  M.; Martin, R. V. Monthly Global Estimates of Fine Particulate Matter and Their
  Uncertainty. *Environ. Sci. Technol.* 2021, 55 (22), 15287–15300.
  https://doi.org/10.1021/acs.est.1c05309.
- Wang, Y.; Apte, J. S.; Hill, J. D.; Ivey, C. E.; Johnson, D.; Min, E.; Morello-Frosch, R.;
   Patterson, R.; Robinson, A. L.; Tessum, C. W.; Marshall, J. D. Air Quality Policy Should
   Quantify Effects on Disparities. *Science* 2023, 381 (6655), 272–274.
   https://doi.org/10.1126/science.adg9931.
- 387 (33) Bill Status AB-617 Nonvehicular air pollution: criteria air pollutants and toxic air contaminants.
  389 https://leginfo.legislature.ca.gov/faces/billStatusClient.xhtml?bill\_id=201720180AB617
  390 (accessed 2024-07-18).
- 391 (34) Rep. Yarmuth, J. A. [D-K.-3. *Text H.R.5376 117th Congress (2021-2022): Inflation Reduction Act of 2022*. https://www.congress.gov/bill/117th-congress/house-bill/5376/text (accessed 2024-07-18).
- Josey Kevin P.; Delaney Scott W.; Wu Xiao; Nethery Rachel C.; DeSouza Priyanka;
   Braun Danielle; Dominici Francesca. Air Pollution and Mortality at the Intersection of Race and Social Class. *N. Engl. J. Med.* 2023, 388 (15), 1396–1404.
   https://doi.org/10.1056/NEJMsa2300523.
  - (36) Wang, Y.; Kloog, I.; Coull, B. A.; Kosheleva, A.; Zanobetti, A.; Schwartz, J. D. Estimating Causal Effects of Long-Term PM<sub>2.5</sub> Exposure on Mortality in New Jersey. *Environ. Health Perspect.* **2016**, *124* (8), 1182–1188. https://doi.org/10.1289/ehp.1409671.
  - (37) Chambliss, S. E.; Pinon, C. P. R.; Messier, K. P.; LaFranchi, B.; Upperman, C. R.; Lunden, M. M.; Robinson, A. L.; Marshall, J. D.; Apte, J. S. Local- and Regional-Scale Racial and Ethnic Disparities in Air Pollution Determined by Long-Term Mobile Monitoring. *Proc. Natl. Acad. Sci.* **2021**, *118* (37). https://doi.org/10.1073/pnas.2109249118.
  - (38) Manchanda, C.; Harley, R.; Marshall, J.; Turner, A.; Apte, J. Integrating Mobile and Fixed-Site Black Carbon Measurements to Bridge Spatiotemporal Gaps in Urban Air Quality. ChemRxiv December 25, 2023. https://doi.org/10.26434/chemrxiv-2023-d4q7n.
- 408 (39) Wang, A.; Testi, I.; Paul, S.; Mora, S.; Walker, E.; Nyhan, M.; Duarte, F.; Santi, P.; Ratti,
  409 C. Big Mobility Data Reveals Hyperlocal Air Pollution Exposure Disparities in the Bronx,
  410 New York City. November 29, 2023. https://doi.org/10.21203/rs.3.rs-3595378/v1.
- (40) Wen, Y.; Zhang, S.; Wang, Y.; Yang, J.; He, L.; Wu, Y.; Hao, J. Dynamic Traffic Data in
   Machine-Learning Air Quality Mapping Improves Environmental Justice Assessment.
   Environ. Sci. Technol. 2024, 58 (7), 3118–3128. https://doi.org/10.1021/acs.est.3c07545.
- (41) Lu, Y.; Giuliano, G.; Habre, R. Estimating Hourly PM<sub>2.5</sub> Concentrations at the
   Neighborhood Scale Using a Low-Cost Air Sensor Network: A Los Angeles Case Study.
   Environ. Res. 2021, 195, 110653. https://doi.org/10.1016/j.envres.2020.110653.
- 417 (42) Park, Y. M.; Sousan, S.; Streuber, D.; Zhao, K. GeoAir—A Novel Portable, GPS-418 Enabled, Low-Cost Air-Pollution Sensor: Design Strategies to Facilitate Citizen Science 419 Research and Geospatial Assessments of Personal Exposure. *Sensors* **2021**, *21* (11), 3761.
- 420 https://doi.org/10.3390/s21113761.

399

400

401

402

403

404

405

406

407

421 (43) Do, K.; Yu, H.; Velasquez, J.; Grell-Brisk, M.; Smith, H.; Ivey, C. E. A Data-Driven

- Approach for Characterizing Community Scale Air Pollution Exposure Disparities in Inland Southern California. *J. Aerosol Sci.* **2021**, *152*, 105704. https://doi.org/10.1016/j.jaerosci.2020.105704.
  - (44) Bi, J.; Wildani, A.; Chang, H. H.; Liu, Y. Incorporating Low-Cost Sensor Measurements into High-Resolution PM<sub>2.5</sub> Modeling at a Large Spatial Scale. *Environ. Sci. Technol.* **2020**, 54 (4), 2152–2162. https://doi.org/10.1021/acs.est.9b06046.
  - (45) Demetillo, M. A. G.; Harkins, C.; McDonald, B. C.; Chodrow, P. S.; Sun, K.; Pusede, S. E. Space-Based Observational Constraints on NO2 Air Pollution Inequality From Diesel Traffic in Major US Cities. *Geophys. Res. Lett.* **2021**, *48* (17), e2021GL094333. https://doi.org/10.1029/2021GL094333.
  - (46) Kerr, G. H.; Goldberg, D. L.; Harris, M. H.; Henderson, B. H.; Hystad, P.; Roy, A.; Anenberg, S. C. Ethnoracial Disparities in Nitrogen Dioxide Pollution in the United States: Comparing Data Sets from Satellites, Models, and Monitors. *Environ. Sci. Technol.* **2023**. https://doi.org/10.1021/acs.est.3c03999.
  - (47) Lunderberg, D. M.; Liang, Y.; Singer, B. C.; Apte, J. S.; Nazaroff, W. W.; Goldstein, A. H. Assessing Residential PM<sub>2.5</sub> Concentrations and Infiltration Factors with High Spatiotemporal Resolution Using Crowdsourced Sensors. *Proc. Natl. Acad. Sci.* **2023**, *120* (50), e2308832120. https://doi.org/10.1073/pnas.2308832120.
  - (48) Mullen, C.; Flores, A.; Grineski, S.; Collins, T. Exploring the Distributional Environmental Justice Implications of an Air Quality Monitoring Network in Los Angeles County. *Environ. Res.* **2022**, *206*, 112612. https://doi.org/10.1016/j.envres.2021.112612.
  - (49) Li, J.; Carlson, B. E.; Lacis, A. A. How Well Do Satellite AOD Observations Represent the Spatial and Temporal Variability of PM<sub>2.5</sub> Concentration for the United States? *Atmos. Environ.* **2015**, *102*, 260–273. https://doi.org/10.1016/j.atmosenv.2014.12.010.

### 448 TOC Art

# EPA monitoring misses many regions exceeding new PM<sub>2.5</sub> standard

