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2	Re-examining chemical conditions of past chamber studies of secondary
3	organic aerosol formation
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12	Abstract
13	The simulation of secondary organic aerosol (SOA) in 3D models generally relies on
14	measurements made in laboratory chamber studies. However, many of the laboratory studies that
15	underpin the aerosol parameterizations used in these models were carried out more than fifteen
16	years ago, before recent developments in our understanding of peroxy radical (RO <sub>2</sub> ) chemistry
17	and its role in aerosol formation. As a result, limitations of past chamber experiments and the
18	incomplete understanding of their chemical conditions (e.g., the initiating oxidants, RO2 fate),
19	may skew SOA representation in models. In this work, literature SOA chamber studies,
20	specifically those referenced by the SOA scheme in the GEOS-Chem global model, are
21	simulated using a modified version of the Master Chemical Mechanism. This enables explicit
22	estimation of typically-unconstrained parameters which affect experimental outcomes, including
23	the relative importance of different oxidants, and the relative loss of RO2 to different
24	unimolecular and bimolecular processes. This work demonstrates that reaction conditions are
25	dynamic, changing within and between experiments, and that many experimental conditions
26	involve more than one oxidant or RO <sub>2</sub> fate, complicating model parameterizations. However, we
27	also find that RO <sub>2</sub> isomerization is important under many of the experimental conditions used,
28	meaning that some RO <sub>2</sub> isomerization processes are to some extent embedded into 3D model
29	SOA estimates, despite no explicit representation of this chemistry.
30	Synopsis
31	We model past secondary organic aerosol chamber studies with explicit, detailed chemistry to
32	explore the chemical conditions of the laboratory experiments that underpin aerosol
33	parameterizations in 3D models.

#### Introduction

- 35 Secondary organic aerosol (SOA), formed from the oxidation and subsequent condensation of
- volatile organic compounds (VOCs), is the major component of organic aerosol in the
- atmosphere<sup>1,2</sup> and therefore substantially impacts the climate and contributes to air pollution.
- However, organic aerosol has long posed a challenge for 3D models, which typically do not
- 39 accurately predict its observed variability,<sup>3,4</sup> likely due to the complexities and uncertainties
- associated with the formation of SOA.
- SOA formation is strongly influenced by key branch points in the chemical oxidation
- mechanism, each of which operate as a function of reaction conditions.<sup>5</sup> Two of the most
- 43 important branch points are the identity of the initiating oxidant and the fate of the resulting
- peroxy radical. VOCs are most commonly oxidized by OH, O<sub>3</sub>, or NO<sub>3</sub>, and these different
- oxidation pathways can affect SOA yield. For example, the reaction of isoprene with OH often
- results in relatively rapid incorporation of oxygen into the molecule, reducing volatility, while
- 47 the reaction of isoprene with O<sub>3</sub> fragments the carbon chain, resulting in two oxidized but more
- volatile products. The oxidation of a VOC typically results in the formation of a peroxy radical
- 49 (RO<sub>2</sub>), a second key branch point that controls the distribution of potential products. Under a
- 50 conventional understanding of RO<sub>2</sub> chemistry, RO<sub>2</sub> species typically react with NO, forming an
- alkoxy radical or organonitrate, or HO<sub>2</sub>, forming an organic peroxide. <sup>5</sup> RO<sub>2</sub> species can also
- react with other RO<sub>2</sub>, forming disproportionation or dimer products, and some RO<sub>2</sub> may react
- with NO<sub>2</sub>, forming peroxyacyl nitrates, or with NO<sub>3</sub>, forming alkoxy radicals.<sup>6</sup> These different
- RO<sub>2</sub> fates can affect SOA yields, for example by favoring fragmentation (yielding higher
- volatility products) or incorporation of additional oxygen atoms (yielding lower volatility
- 56 products).
- However, advances in our understanding of RO<sub>2</sub> chemistry over the last fifteen years have
- 58 brought to light several new RO<sub>2</sub> reaction pathways. In particular, peroxy radical isomerization,
- sometimes referred to as autoxidation, has been recognized as an important pathway through
- which oxygen may sequentially be incorporated into an organic molecule, rapidly forming a low-
- volatility compound that may condense to efficiently contribute to SOA mass.<sup>7,8</sup> RO<sub>2</sub>
- 62 isomerization rates are highly dependent on the structure of the RO<sub>2</sub> species, varying over many
- orders of magnitude, such that for some compounds, it may be a dominant RO<sub>2</sub> fate, while for
- others it may be irrelevant. While additional RO<sub>2</sub> reactions such as the rapid formation of
- ROOR species by  $RO_2 + RO_2^{10}$  or the reaction of  $RO_2$  with  $OH^{11}$  have also been discovered in
- recent years, the importance of these for atmospheric SOA formation is not yet clear.
- The formation of SOA is sufficiently complex such that its explicit modeling on a regional and
- 68 global scale is prohibitive. SOA forms through the condensation of numerous products of VOC
- oxidation, each of which is derived from a different complicated (and incompletely understood)
- sequence of oxidation reactions, and each of which differ in vapor pressure and reactivity. In
- order to make this process computationally tractable, 3D models rely on substantial
- simplifications. In the simplest case, SOA is treated as a fixed yield product of the reaction or

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emission of VOC precursors.<sup>3,12</sup> However, SOA is more realistically represented when
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       absorptive partitioning of gas-phase organic species into the organic aerosol phase is
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       considered, <sup>13</sup> typically through the use of the two-product model <sup>14</sup> or volatility basis set (VBS). <sup>15</sup>
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       In either case, yields of surrogate species of different volatilities are assigned by fitting the
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       laboratory-measured SOA yield as a function of the organic aerosol concentration (c_{OA}), thereby
       linking laboratory results to 3D model output. While SOA yields in a VBS scheme can be
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       defined simply based on the identity of the VOC precursor, they may also be specified based on
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       VOC/oxidant pairs, or NO<sub>X</sub>-conditions ("high NO<sub>X</sub>" and "low NO<sub>X</sub>"), such that aerosol yields in
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       the model can be defined for different chemical environments. Within this framework, aerosol
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       yields may be parameterized such that under intermediate chemical conditions (e.g., competing
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       RO<sub>2</sub> + NO and RO<sub>2</sub> + HO<sub>2</sub> reactions), SOA yields may be linearly interpolated. <sup>16–18</sup>
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       Because SOA parameterizations fundamentally rely on laboratory chamber experiments, they
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       therefore integrate any limitations from these experiments. Many limitations of chamber
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       experiments have been documented previously, including large differences between atmospheric
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       and chamber temperature, humidity, and NO<sub>X</sub> concentrations, <sup>19</sup> as well as the partitioning of
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       semi-volatile and low-volatility species to chamber walls, a process which impacts observed
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       aerosol formation.<sup>20</sup> An important additional limitation is that the chemical conditions used in
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       past chamber experiments are not typically well-characterized.<sup>21</sup> Many laboratory chamber
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       studies that continue to be used as a basis for regional and global model aerosol modules were
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       carried out more than fifteen years ago, before the previously-discussed advances in our
       understanding of RO<sub>2</sub> chemistry (e.g., isomerization). Earlier chamber experiments were not
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       designed to study these unimolecular reaction pathways, and instead focused on separating
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       aerosol yield by reaction with specific oxidants, and/or by "high-NOx" vs "low-NOx" pathways.
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       Furthermore, while assessments of RO<sub>2</sub> and oxidant fate in contemporary chamber experiments
       using mechanistic modeling is increasingly common, it was much less common in older studies.
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       Instead, oxidant and RO<sub>2</sub> fate was often assumed based on the oxidant added or the initial NO<sub>X</sub>
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       concentrations, and any such assumptions are subsequently embedded in aerosol
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       parameterizations. While contemporary model schemes rely on a range of laboratory studies,
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       they often pull from some of the same older experiments; commonly referenced collections of
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       volatility basis set fits include Pye et al., <sup>18</sup> Farina et al., <sup>12</sup> and Zhang et al. <sup>22</sup>
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       Given our evolving understanding of radical chemistry, the chemical conditions as reported in
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       older publications and the chemical conditions actually used in those experiments may not
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       always align. Where an aerosol parameterization relies on assigned RO<sub>2</sub> conditions (e.g., reaction
       with NO vs. reaction with HO<sub>2</sub>), any deviation in the true experimental RO<sub>2</sub> fate from this binary
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       classification (which may arise from the importance of other reactions such as isomerization or
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       RO<sub>2</sub> + RO<sub>2</sub>) may affect the accuracy of the aerosol scheme. Similarly, oxidation processes may
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       differ from an assumption of a single dominant oxidant, as multiple oxidants may be formed
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       (e.g., OH from O<sub>3</sub> + alkene reactions, O<sub>3</sub> or NO<sub>3</sub> from irradiated VOC + NO<sub>3</sub> mixtures). These
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background NO<sub>X</sub> or by the chemistry associated with added compounds (e.g., OH scavengers).

effects may be unexpected, for example where these processes are influenced by trace

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This work seeks to re-examine chamber conditions used in literature SOA experiments, by using 113 explicit box modeling to estimate typically-unconstrained chemical parameters. The specific 114 focus here is on the initiating oxidant and the fate of RO<sub>2</sub> radicals, which to our knowledge have 115 never been systematically assessed in past SOA chamber studies. The set of historical chamber 116 experiments examined is taken from the basis for the "Complex" aerosol scheme<sup>18</sup> from GEOS-117 Chem, a widely used chemical transport model. This scheme describes SOA formation in terms 118 of both oxidant and RO<sub>2</sub> fate, and therefore represents an appropriate study to use as a guide for 119 the selection of experiments. While this set of experiments is not a comprehensive list, it 120 includes chamber studies that form a basis for aerosol schemes in other 3D models, such that the 121 results of this work may be more broadly applicable. Chamber experiments are represented using 122 explicit chemical mechanisms, modified where possible to include appropriate isomerization 123 reactions. Finally, explicitly modeled oxidant and RO<sub>2</sub> fate are interpreted in the context of their 124 use in aerosol parameterizations, giving insight into challenges facing experimental design and 125 126 the creation of accurate parameterizations.

## Methods

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Historical experiments are chosen based on Pye et al., 18 the basis for the "Complex" aerosol 128 scheme used in GEOS-Chem. While subsequent studies have updated the aerosol module to 129 include aqueous reactive uptake of isoprene oxidation products and some organonitrates 130 separately from the volatility basis set, <sup>23,24</sup> that work is not considered here. The publications 131 covered in this work are shown in Table 1, along with basic experimental details and a brief 132 summary of data availability. Starting experimental conditions (NO<sub>X</sub> levels, VOC levels, etc.) are 133 provided in each publication, and experimental timeseries are scraped from paper figures or 134 provided by authors. All model parameters for each simulation are shown in Tables S.1 and S.2. 135 Each experiment is classified into one of six categories (last column of Table 1) based on the 136 oxidation conditions. While chamber experiments are often grouped simply as "high-NO<sub>X</sub>" or 137 "low-NO<sub>X</sub>," these distinctions do not fully describe the underlying chemistry or RO<sub>2</sub> fate.<sup>25</sup> 138 Experiment categories involving added NO<sub>X</sub> include "HONO" (which use HONO as an oxidant 139 precursor<sup>26,27</sup>), "H<sub>2</sub>O<sub>2</sub>, NO<sub>X</sub>" (which use both H<sub>2</sub>O<sub>2</sub> and added NO<sub>X</sub><sup>26,28</sup>), and "traditional 140 photooxidation" (which use a mixture of VOC, NO<sub>X</sub>, and UV light<sup>29</sup>). Experiments using VOC, 141 NO<sub>X</sub>, UV light, and added O<sub>3</sub><sup>30</sup> are also grouped into this category based on similarities in the 142 observed chemistry. Experiment categories involving no added NO<sub>X</sub> include "H<sub>2</sub>O<sub>2</sub>, low NO<sub>X</sub>," 143 <sup>28,27,26</sup> and "O<sub>3</sub>, low NO<sub>X</sub>". <sup>29,31</sup> In all ozonolysis experiments studied here, an OH scavenger is 144 added to minimize OH oxidation of the precursor. Ozonolysis experiments from Zhang et al.<sup>30</sup> 145 are classified as "O<sub>3</sub>, some NO<sub>X</sub>" since these experiments featured measurable background NO<sub>X</sub> 146 concentrations of 2-15 ppb (sufficiently high to substantially alter observed chemistry). 147

Table 1: Summary of modeled experiments.<sup>a</sup>

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Publication	VOC Precursors	Number of expts. Reference da		Oxidation category	
Griffin et al. <sup>29</sup>	β-pinene β-caryophyllene α-pinene	5	Starting conditions (all expts) + time series (one expt)	O <sub>3</sub> , low NO <sub>X</sub> Traditional photooxidation	
Zhang et al. <sup>30</sup>	limonene	Starting co (all expts series (tw		O <sub>3</sub> , some NO <sub>X</sub> Traditional photooxidation	
Kroll et al. <sup>28</sup>	isoprene	14	Starting conditions (all expts) + max. O <sub>3</sub> concentrations (all expts) + time series (two expts)	H <sub>2</sub> O <sub>2</sub> , low NO <sub>X</sub> H <sub>2</sub> O <sub>2</sub> , NO <sub>X</sub>	
Ng et al. <sup>27</sup>	<i>m</i> -xylene toluene benzene	18	VOC, NO <sub>X</sub> , O <sub>3</sub> time series (all expts)	H <sub>2</sub> O <sub>2</sub> , low NO <sub>X</sub> HONO	
Ng et al. <sup>26</sup>	α-pinene	5	VOC, NO <sub>X</sub> , O <sub>3</sub> time series (all expts)	H <sub>2</sub> O <sub>2</sub> , low NO <sub>X</sub> H <sub>2</sub> O <sub>2</sub> , NO <sub>X</sub> HONO	
Shilling et al. <sup>31</sup>	α-pinene	8	Starting conditions (all expts)	O <sub>3</sub> , low NO <sub>X</sub>	

<sup>a</sup>Model parameters for each experiment are included in Tables S.1 and S.2.

Modeling is carried out using the Framework for 0-D Atmospheric Modeling, version 4.3

(F0AM).<sup>32</sup> Chemical mechanisms are taken from the Master Chemical Mechanism (MCM

v3.3.1). <sup>33,34,34–37</sup> While referenced by Pye et al., <sup>18</sup> α-humulene and naphthalene are not modeled

since they are not represented in the MCM. Similarly, experiments focusing on the reaction of

NO<sub>3</sub> with β-pinene<sup>29</sup> and isoprene<sup>38</sup> are also omitted due to insufficient experimental data to

constrain the model. Only gas-phase chemistry is modeled, with no treatment of partitioning or

particle-phase chemistry.

While some mechanisms in the MCM, most notably for isoprene, include unimolecular RO<sub>2</sub>

reactions, most mechanisms used here do not yet feature unimolecular RO<sub>2</sub> chemistry. While

160 RO<sub>2</sub> isomerization rates are highly dependent on structure<sup>9</sup> and challenging to accurately

measure or calculate, some important isomerization rates have been characterized or estimated

for all of the precursor VOCs studied. Mechanisms are therefore modified to include known or

estimated RO<sub>2</sub> isomerization rate constants for initial RO<sub>2</sub>, defined here as the first RO<sub>2</sub> species

formed immediately after the VOC-oxidant reaction (these species are listed in Table S.3); all

other RO<sub>2</sub> are referred to as "subsequent" RO<sub>2</sub>. We note that because a single generation of

oxidation can include multiple RO<sub>2</sub> species formed in sequence, "initial RO<sub>2</sub>" is not synonymous

with "first-generation RO<sub>2</sub>." Due to the extensive uncertainties and complexities of integrating

unimolecular RO<sub>2</sub> reactions into chemistry that occurs beyond the initial RO<sub>2</sub> species, reactions

for subsequent RO<sub>2</sub> are not modified. To conserve carbon without extensively re-writing each

mechanism, each RO2 isomerization reaction forms an alkyl radical, which reacts with oxygen to

form another RO<sub>2</sub> radical, identical to an RO<sub>2</sub> species from a later point in the mechanism. RO<sub>2</sub>

isomerization rate constants are taken from literature measurements and calculations. 9,39-44

- 173 Mechanism modifications are applied to all precursor mechanisms (except that of isoprene,
- which already features isomerization reactions), and are detailed in Table S.4 and Section S.1.
- Some experiments use additional VOCs such as radical promoters (propene), <sup>29</sup> which facilitate
- additional OH formation through ozonolysis, OH scavengers (2-butanol, n-butanol), <sup>29–31</sup> which
- decrease OH concentrations to maximize O<sub>3</sub> chemistry, or hydrocarbons (*n*-pentane), <sup>30</sup> which are
- used to dilute the precursor VOC for injection. These species are represented by a standard
- mechanism from the MCM and an identical non-interacting tracer mechanism that is run in
- parallel with the main simulation. This allows the subtraction of the influence of these species
- during analysis, such that analysis focuses solely on the chemistry of the VOC of interest.
- All experiments are first modeled based on known parameters. While the publications vary in
- specificity with regard to reported conditions, they generally report initial VOC, O<sub>3</sub>, and NO<sub>X</sub>
- 184 concentrations (See Tables S.1 and S.2 for a summary of all model input parameters). Less well
- quantified starting material concentrations include H<sub>2</sub>O<sub>2</sub> (typically estimated) and OH radical
- sinks (e.g. 2-butanol concentration is "sufficient to consume OH" <sup>29</sup>). Light spectra are taken
- from earlier publications for indoor chambers<sup>45,46</sup> (See Figure S.1). Experiment length is rarely
- stated explicitly and is assumed to be the same as the length of time shown in relevant figures, or
- the length of time in relevant data files. Since the experiments simulated do not quantify vapor
- wall losses, these processes are not represented in the model.
- For each series of experiments, model runs are then tuned to match measured data by adjusting
- free variables that are unmeasured or poorly quantified. Chamber light intensity is adjusted to
- match reported  $j_{NO2}$  values<sup>28</sup> or to fit reported VOC, NO<sub>X</sub>, and O<sub>3</sub> timeseries.<sup>29,30</sup> For the outdoor
- experiments of Griffin et al..<sup>29</sup> the default sunlight intensity based on experimental date and
- location is halved to best match data, likely a result of overhead aerosol or clouds. For
- experiments using HONO as an oxidant precursor, a combination of NO<sub>X</sub> analyzer
- measurements (which measure HONO as "NO<sub>2</sub>"), and fit to measured timeseries of VOC, NO<sub>X</sub>.
- and O<sub>3</sub> are used to constrain HONO concentrations. H<sub>2</sub>O<sub>2</sub> concentrations are unmeasured but
- estimated in several publications. While these are generally taken as the mean value of the
- estimated range, the concentration is loosely constrained by the NO loss rate (which is sensitive
- to H<sub>2</sub>O<sub>2</sub> concentrations) when NO is present in the experiment. For several experiments with no
- 202 added NO<sub>X</sub>, ozone is produced when little is predicted; this is likely due to residual NO<sub>X</sub> species
- off-gassing from the chamber walls. 28 In these cases, ozone timeseries are best matched by
- 204 adding trace NO or NO<sub>2</sub> to the model; in some cases, this is held constant (see Table S.2). Even
- when ozone formation does not constrain  $[NO_X]$ , experimental conditions are unlikely to be truly
- NO<sub>X</sub>-free, so a small amount of NO<sub>2</sub> (0.1 1 ppb), below reported detection limit) is added to the
- model. For variables that are not quantitatively constrained (e.g., a reported  $j_{NO2}$  value), final
- 208 tuning of these input parameters is carried out by eye to achieve a reasonable fit to reported
- VOC,  $NO_X$ , and  $O_3$  timeseries.
- 210 Since some experimental parameters are relatively uncertain, the effect of variability in model
- inputs is explored through sensitivity testing. Light intensity and all nonzero starting

concentrations are varied individually by  $\pm$  30%. These results are first used to assess how

uncertainty in starting conditions impacts the fit between measurement and modeled results. As

described in Results and Discussion, these sensitivities are also used to explore how uncertainty

in starting conditions propagates to measured parameters associated with subsequent chamber

- 216 chemistry.
- 217 Relative VOC reactive loss pathways and RO<sub>2</sub> fates are calculated based on F0AM outputs,
- 218 which include time series of all species in the mechanism and rates over time for each reaction.
- 219 Reactive loss pathways include reaction with OH, O<sub>3</sub>, or NO<sub>3</sub>, as well as direct photolysis for
- 220 photolabile product species such as formaldehyde. We extract reaction rates for each "oxidant"
- (including hy) over time and separate them into reactions with the precursor and reactions with
- other organic products. These time series are interpolated, summed, and normalized to get total
- fractional reactive loss pathways for both precursor and total products. Peroxy radical fate is
- assessed in a similar way. Peroxy radicals are grouped into initial RO2 and subsequent RO2, and
- time-dependent RO<sub>2</sub> reaction rates (RO<sub>2</sub> + HO<sub>2</sub>, RO<sub>2</sub> + NO, RO<sub>2</sub> + NO<sub>2</sub>, RO<sub>2</sub> + RO<sub>2</sub>, RO<sub>2</sub> +
- NO<sub>3</sub>, unimolecular RO<sub>2</sub> reaction) are then extracted from the model output. In order to properly
- account for the relative importance of reversible RO<sub>2</sub> reactions, most notably acyl RO<sub>2</sub> + NO<sub>2</sub>
- reactions which generate peroxyacyl nitrates (PANs), rates are calculated by summing any
- equilibria such that the reported rate is equal to the net flux (using the F0AM function
- ExtractRates with the "sumEq" parameter set to 1). This does not affect calculated rates for other
- RO<sub>2</sub> reactions, which are not reversible in the MCM. As above, grouped RO<sub>2</sub> fate timeseries are
- used to obtain total fractional RO<sub>2</sub> fate for initial and subsequent RO<sub>2</sub>. For all model outputs, the
- 233 influence of radical promoters, OH scavengers, and additional hydrocarbons is removed using
- 234 the output of the tracer mechanism.
- 235 Model outputs are further interpreted with respect to the fraction of RO<sub>2</sub> reacting with NO vs
- HO<sub>2</sub>,  $\beta$ , and the bimolecular peroxy radical lifetime,  $\tau_{bi}$ . As defined in Pye et al. <sup>18</sup>,  $\beta$  enables
- assessment of the competition between reaction with HO<sub>2</sub> and NO (eq. 1), and is used as a
- 238 "mixing parameter" in SOA parameterizations to interpolate aerosol yields measured under
- "high-NO" and "low-NO" conditions. 16–18

$$\beta = \frac{k_{\text{RO}_2 + \text{NO}}[\text{NO}]}{k_{\text{RO}_2 + \text{NO}}[\text{NO}] + k_{\text{RO}_2 + \text{HO}_2}[\text{HO}_2]}$$
(1)

- The bimolecular peroxy radical lifetime,  $^{47-49}\tau_{bi}$ , can be used to assess the relative importance of
- unimolecular RO<sub>2</sub> chemistry. Since unimolecular RO<sub>2</sub> reaction rates vary over many orders of
- magnitude,  $\tau_{bi}$  can be used as a threshold to indicate whether a given unimolecular reaction
- would be competitive, relative to the unimolecular lifetime,  $\tau_{uni}$  (the inverse of the unimolecular
- RO<sub>2</sub> reaction rate constant). While prior work has often defined  $\tau_{bi}$  only with respect to
- bimolecular reaction with NO and  $HO_2$ , <sup>49,21</sup> we define  $\tau_{bi}$  with respect to reaction with NO,  $HO_2$ ,
- 247 RO<sub>2</sub>, NO<sub>2</sub>, and NO<sub>3</sub> (eq. 2), since under some chamber conditions, these additional reactive sinks
- 248 are found to contribute to RO<sub>2</sub> loss.

$$\tau_{bi} = \frac{1}{k_{\text{RO}_2 + \text{NO}}[\text{NO}] + k_{\text{RO}_2 + \text{HO}_2}[\text{HO}_2] + k_{\text{RO}_2 + \text{RO}_2}[\text{RO}_2] + k_{\text{RO}_2 + \text{NO}_2}[\text{NO}_2] + k_{\text{RO}_2 + \text{NO}_3}[\text{NO}_3]} (2)$$

The bimolecular lifetime is calculated using rates extracted from the model, and does not take into account substantially faster recent measurements of the RO<sub>2</sub> + RO<sub>2</sub> rate constants.<sup>10</sup>

## **Results and Discussion**

## Model-measurement agreement and example model output

Figure 1A shows an example model-measurement comparison from a "traditional photooxidation" experiment in which β-pinene was oxidized in the presence of NO<sub>X</sub>, propene, and sunlight.<sup>29</sup> The model agrees qualitatively with the measurements, and nearly all values fall within the envelope of timeseries generated by the sensitivity testing (lightly shaded areas). While modeled NO<sub>2</sub> diverges from the measurements late in the experiment, this might not be a model error, and instead could be a result of NOy species (e.g., PANs, organonitrates) reported as NO<sub>2</sub> by commercial NO<sub>X</sub> monitors. For most experiments, measurements similarly qualitatively agree with model estimates (Figures S.2 and S.3).

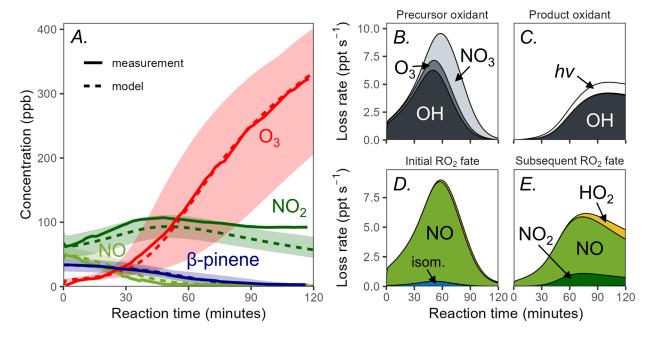


Figure 1: Model-measurement comparison from an example experiment (Panel A), and stacked time series of oxidant and  $RO_2$  loss rates (panels B-E). The experiment shown is a traditional photooxidation experiment in which  $\beta$ -pinene is oxidized in the presence of  $NO_X$  and propene (expt. #1, see Table S.1).<sup>29</sup> Panel A shows both modeled and measured time series for four species, with shaded regions representing the upper and lower bounds based on sensitivity testing. Panels B-E separate precursor from product loss pathways, and separate initial and subsequent  $RO_2$  fate.

However, the tuned model deviates from measurements in several experiments, most notably for  $H_2O_2 + NO_X$  experiments, <sup>26,28</sup> in which measured  $O_3$  is well below the model value. The cause of

- 271 this is unclear, and is not resolved by modulating light intensity or H<sub>2</sub>O<sub>2</sub> concentrations. To
- explore the impact of this difference, modeled O<sub>3</sub> concentrations are fixed to measured values
- and model outputs are recalculated for one experiment for which this occurs. These differ little
- 274 from those derived from the original model run (see Figure S.4).
- Figure 1B-E shows oxidation rates and RO<sub>2</sub> fate over time for the example experiment. The
- precursor β-pinene reacts mostly with OH, but NO<sub>3</sub> becomes an important oxidant as the
- experiment proceeds (Figure 1B). Organic products also react mostly with OH, but photolysis
- increases in relative importance as formaldehyde and other photolabile aldehydes are formed
- 279 (Figure 1C). Similarly, RO<sub>2</sub> fate differs for initial vs subsequent RO<sub>2</sub>. Initial RO<sub>2</sub> reacts mainly
- with NO, with a small contribution from isomerization (Figure 1D). While NO is still the main
- 281 contributor to subsequent RO<sub>2</sub> fate, RO<sub>2</sub> + HO<sub>2</sub> and acyl RO<sub>2</sub> + NO<sub>2</sub> reactions become
- increasingly important later in the experiment (Figure 1E). This demonstrates how oxidant and
- 283 RO<sub>2</sub> fate can evolve substantially over the course of a given experiment.

#### Relative VOC reactive losses

- Integrated VOC reactive loss pathways for a larger selection of experiments are presented in
- Figure 2. This figure shows fractional loss rates to each oxidant for the precursor VOC (first
- column) and the organic products (second column) for a selection of experiments. These
- experiments are chosen to represent a range of precursors and oxidation categories, illustrating
- 289 the variety of oxidant fates seen across the experiments modeled in this study.
- In these experiments, researchers generally sought to achieve chemical conditions in which one
- oxidant was dominant, for example by using an OH scavenger in ozonolysis experiments to
- encourage O<sub>3</sub> chemistry to dominate. This is largely achieved for several experiment types—
- 293 HONO, H<sub>2</sub>O<sub>2</sub> + NO<sub>X</sub>, low-NO<sub>X</sub> H<sub>2</sub>O<sub>2</sub>, and low-NO<sub>X</sub> ozonolysis all feature precursor oxidation
- largely driven by a single oxidant (OH or O<sub>3</sub>). However, this is not the case for traditional
- 295 photooxidation experiments, or the "O<sub>3</sub>, some NO<sub>X</sub>" experiments. These experiments feature a
- 296 mix of oxidants, in which oxidation by NO<sub>3</sub> competes with oxidation by OH and O<sub>3</sub>.
- The distribution of reactive pathways for reaction products (second column of Figure 2) is
- dramatically different from that for precursors. Ozone reactivity is decreased as double bonds are
- consumed, and photolysis becomes an important loss pathway, since many products include
- 300 photolabile functional groups. A majority of this photolysis is of small aldehydes (e.g.,
- formaldehyde, glyoxal), but larger species can also contribute, depending on the VOC precursor.
- While changing reactivity explains many of the differences between fractional loss of the
- precursor and the products, these differences are also affected by temporal shifts in oxidant
- 304 concentrations across an experiment (e.g., increasing O<sub>3</sub> levels).

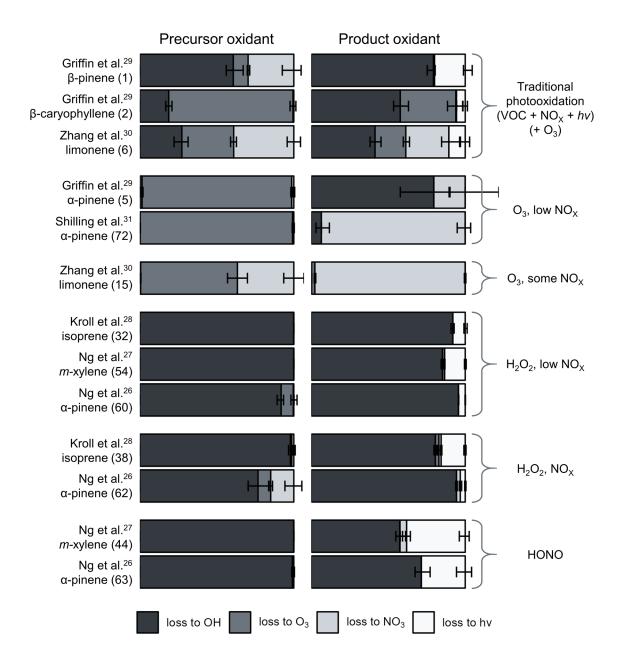


Figure 2: Fractional precursor and product loss by different oxidants. Precursor losses are shown in the left column, and product losses are shown in the right column. Experiments are grouped by experimental conditions, and selected such that each represents a distinct set of starting conditions. Experiment numbers are found in parentheses following the VOC names (see Table S.1). Error bars reflect the range of values observed in response to sensitivity tests and represent the range for each oxidant, not the cumulative range. Error bars do not reflect any uncertainties propagated from errors or uncertainties in the chemical mechanisms.

Figure 2 also shows error bars representing the range of values observed across sensitivity tests. Despite varying model inputs by 30%, the ranges observed here are far lower in most cases, indicating that these results are relatively robust to uncertainties in experimental starting conditions. The main exception is uncertainties driven by trace NO<sub>2</sub> added to low-NO<sub>X</sub> ozonolysis experiments (since none of these experiments are carried out under perfectly NO<sub>X</sub>-

- free conditions). Perturbations to trace NO<sub>2</sub>, e.g., in α-pinene oxidation carried out by Griffin et
- al.,<sup>29</sup> substantially affect the fraction of oxidation products lost to reaction with NO<sub>3</sub>, resulting in
- larger error bars. This effect is similarly observed for Shilling et al.<sup>31</sup> where relatively little total
- product oxidation occurs, leaving later oxidation dominated by NO<sub>3</sub> reactions (since NO<sub>3</sub> is not
- 321 consumed by reaction with the OH scavenger *n*-butanol in the MCM). We note that the
- uncertainty ranges shown here do not include possible uncertainties in the chemical mechanism,
- which may be substantial but are beyond the scope of this work.

# 324 *RO*<sub>2</sub> *fate*

- Figure 3 shows modeled RO<sub>2</sub> fate for the same set of experiments as shown in Figure 2. This
- figure shows total fractional RO<sub>2</sub> fate separated between initial and subsequent RO<sub>2</sub>, grouped
- 327 again by experiment type.
- Most OH experiments (with the exception of the traditional photooxidation experiments) feature
- 329 conditions designed to favor one RO<sub>2</sub> fate. For experiments in which researchers sought to force
- all RO<sub>2</sub> through reaction with NO (e.g. "HONO" and "H<sub>2</sub>O<sub>2</sub>, NO<sub>X</sub>" experiments), this is often
- achieved for initial RO<sub>2</sub>. Likewise, reactions with high concentrations of H<sub>2</sub>O<sub>2</sub> but no added NO<sub>X</sub>
- typically involve reaction with HO<sub>2</sub> as a majority of initial RO<sub>2</sub> fate (e.g. Kroll et al. low-NO<sub>X</sub>
- oxidation of isoprene $^{28}$ ).
- However, numerous experiments feature mixed RO<sub>2</sub> fates, particularly with regard to the
- influence of RO<sub>2</sub> isomerization. RO<sub>2</sub> isomerization represents an important initial RO<sub>2</sub> reaction
- channel, particularly for the "low-NO<sub>X</sub> ozonolysis" experiments, <sup>29,31</sup> but also for the "low-NO<sub>X</sub>
- H<sub>2</sub>O<sub>2</sub>" experiments<sup>26,28</sup> and even for some experiments in which NO<sub>X</sub> was added.<sup>27,29,30</sup> In many
- cases, the isomerization rates are fast enough to dominate the chemistry of a given RO<sub>2</sub>,
- outcompeting bimolecular reaction, but limited by the fact that only some RO<sub>2</sub> can undergo
- isomerization. In the extreme example of this, ring closure reactions for aromatic species are
- considered fast enough  $(\tau_{uni} < \sim 10^{-4} \text{ s})^{42}$  to outcompete bimolecular reaction under all conditions
- (for this reason, the only bimolecular reactions for initial  $RO_2$  from m-xylene (Fig. 3) stem from
- the minor OH reaction channel). For RO<sub>2</sub> isomerization reactions that involve H-atom
- abstraction and hence have slower isomerization rates (e.g.,  $\tau_{uni} = 0.25 \text{ s}^{-1}$  for one  $\alpha$ -pinene-
- derived RO<sub>2</sub><sup>44</sup>), reaction conditions can substantially affect the degree of isomerization. For
- example, the importance of isomerization in the oxidation of  $\alpha$ -pinene by Ng et al.<sup>26</sup> shifts
- considerably between experiments with and without added NO<sub>X</sub> (Figure 3). While the
- 348 concentrations of HO<sub>2</sub> and NO required to largely shut off the isomerization pathway vary with
- the isomerization rate, Figure 3 suggests that HO<sub>2</sub> concentrations for H<sub>2</sub>O<sub>2</sub> experiments are
- typically not high enough to fully suppress unimolecular chemistry. This agrees with previous
- work showing that isomerization (assuming a rate constant of  $\geq 0.1 \text{ s}^{-1}$ ) is only partially
- suppressed even when [H<sub>2</sub>O<sub>2</sub>] is as high as 10 ppm.<sup>50</sup> In contrast, reaction with NO can fully
- suppress these pathways, but only when large excesses of NO are added (e.g., hundreds of ppb
- NO in the oxidation of α-pinene with HONO<sup>26</sup>).

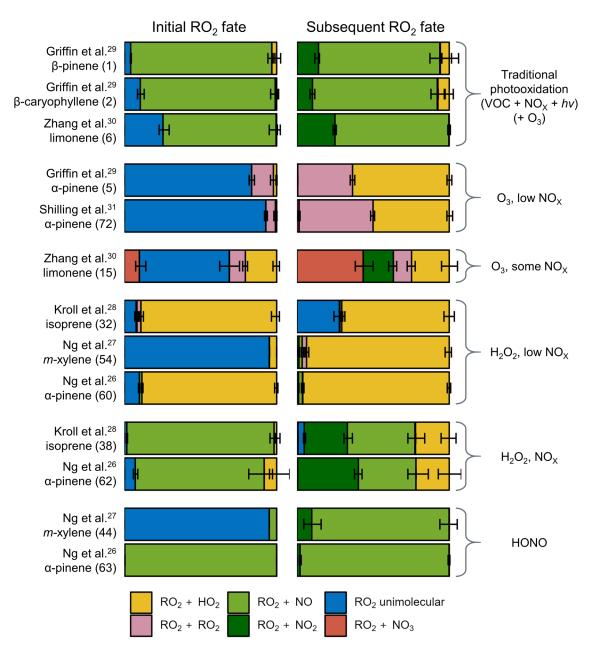


Figure 3: Fractional RO<sub>2</sub> loss rates, grouped by initial RO<sub>2</sub> (left column) and subsequent RO<sub>2</sub> (right column) for selected experiments (same as shown in Figure 2). As in Figure 2, error bars reflect the range of values observed in response to sensitivity tests and represent the range for each type of reaction, not the cumulative range; they do not reflect any uncertainties propagated from errors in the chemical mechanisms.

Subsequent RO<sub>2</sub> chemistry can be dramatically different than that of the initial RO<sub>2</sub>. While this is impacted by a lack of isomerization reactions in the reaction mechanisms (see Methods), bimolecular pathways also shift substantially, due to changing reactivity and changing reactant concentrations. One clear change is the increase in reaction with NO<sub>2</sub>, driven by increased NO<sub>2</sub>/NO ratios (Figure 1) and formation of acyl peroxy radicals, leading to increased RO<sub>2</sub> + NO<sub>2</sub> reactions forming PAN. RO<sub>2</sub> + RO<sub>2</sub> reactions notably also increase for "low-NO<sub>X</sub> ozonolysis

- experiments," where RO<sub>2</sub> concentrations are enhanced by the products of OH reaction with
- added OH scavengers, and NO and HO<sub>2</sub> concentrations remain relatively low. Finally, in the case
- of isoprene, for which the MCM includes several unimolecular reactions of subsequent RO<sub>2</sub>,
- isomerization increases in relative importance, in part to due to the gradual increases in  $\tau_{bi}$  that
- occur throughout some experiments. While much of this subsequent isomerization is of RO<sub>2</sub>
- species formed directly from the initial RO<sub>2</sub> isomerization, it also includes substantial
- 372 contributions from the oxidation of first-generation reaction products (e.g., OH + isoprene
- epoxydiol and OH + methacrolein).
- Like in Figure 2, the range of RO<sub>2</sub> fates simulated with perturbed starting conditions (error bars
- in Figure 3) is smaller than the range of perturbations, suggesting these values are relatively
- robust to error in initial conditions. Imperfections in the chemical mechanisms, perhaps most
- notably the lack of later-generation RO<sub>2</sub> isomerization reactions likely contribute a larger (but
- unknown) error to RO<sub>2</sub> fate.

### Comparison with SOA parameterizations

- 380 Chemical conditions as modeled here can be compared to how these experiments are used in the
- 381 GEOS-Chem aerosol parameterization. As mentioned above, the "Complex" aerosol scheme in
- GEOS-Chem specifies aerosol yield parameters as a function of  $\beta$ , the fraction of RO<sub>2</sub> reacting
- with NO instead of HO<sub>2</sub>. Table 2 presents the classification of experiments used for GEOS-
- 384 Chem<sup>18</sup> alongside modeled oxidant and RO<sub>2</sub> chemistry parameters.
- In most cases, RO<sub>2</sub> fates calculated here and those assumed in the parameterization generally
- agree. This is particularly evident where conditions are pushed to very high levels of NO or HO<sub>2</sub>
- and  $\beta$  falls near 1 or 0.<sup>27,28</sup> However, several differences stand out. Sesquiterpene oxidation
- experiments from Griffin et al.<sup>29</sup> feature a mean  $\beta$  value of 0.93, indicating NO as a dominant
- RO<sub>2</sub> fate, but are used for parameterizing SOA at low- $\beta$  (RO<sub>2</sub> + HO<sub>2</sub>) conditions. While  $\beta$  values
- from Zhang et al.<sup>30</sup> (limonene) and Shilling et al.<sup>31</sup> ("monoterpenes") experiments fall near
- expected extremes,  $\beta$  may not appropriately characterize these experiments due to the importance
- of other RO<sub>2</sub> pathways. A large fraction of the overall bimolecular RO<sub>2</sub> fate (Table 2) does not
- involve NO or HO<sub>2</sub> in these sets of experiments, indicating that these experiments may not serve
- as an ideal basis for SOA parameterizations relying on β.

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GEOS-Chem				This work			
Parent hydrocarbon	Citation	Assumed RO <sub>2</sub> fate	Assumed oxidant	Expt. #	Mean β (1σ)	Mean bimolecular RO <sub>2</sub> fate (HO <sub>2</sub> % / NO % / other %)	Mean precursor oxidant fraction (OH % / O <sub>3</sub> % / NO <sub>3</sub> %)
Limonene	30	NO	OH, O <sub>3</sub>	16 – 27	0.88 (0.14)	2 / 23 / 75	1 / 27 / 73
Monoterpenes <sup>b</sup>	с	NO	OH, O <sub>3</sub>	-	-	-	-
Sesquiterpenes <sup>b</sup>	d	NO	OH, O <sub>3</sub>	-	-	-	-
Limonene	30	HO <sub>2</sub>	OH, O <sub>3</sub>	8 – 15	0 (0)	13 / 0 / 87	2 / 74 / 24
Monoterpenes <sup>b</sup>	31	HO <sub>2</sub>	OH, O <sub>3</sub>	65 - 72	0 (0)	25 / 0 / 75	0 / 97 / 3
Sesquiterpenes <sup>b</sup>	29	HO <sub>2</sub>	OH, O <sub>3</sub>	2-4	0.93 (0.04)	6 / 86 / 8	19 / 81 / 0
Isoprene	28	-	ОН	28 - 35	0 (0)	98 / 0 / 1	100 / 0 / 0
Benzene	27	NO	ОН	50	1 (-)	0 / 95 / 4	100 / 0 / 0
Toluene	27	NO	ОН	46 - 49	1 (0)	0 / 99 / 1	100 / 0 / 0
Xylene	27	NO	ОН	42 - 45	1 (0)	0 / 97 / 3	100 / 0 / 0
Benzene	27	HO <sub>2</sub>	ОН	59	0.04 (-)	92 / 4 / 4	100 / 0 / 0
Toluene	27	HO <sub>2</sub>	ОН	55 – 58	0.01(0)	95 / 1 / 3	100 / 0 / 0
Xylene	27	HO <sub>2</sub>	OH	51 – 54	0.02 (0.01)	95 / 2 / 3	100 / 0 / 0

<sup>&</sup>lt;sup>a</sup> This table is based on Table 1 from Pye et al., <sup>18</sup> omitting rows that do not correspond to model runs in this study. Values correspond to modeled oxidant and RO<sub>2</sub> chemistry from experiments used to generate the VBS parameterization. Note that the total mean bimolecular RO<sub>2</sub> fate does not include isomerization reactions. Standard deviations reflect the variation within the set of experiments.

Similarly, Table 2 illustrates that most experiments feature precursor oxidation by OH or  $O_3$ , consistent with assumptions in the aerosol parameterization. The most notable exception to this however is the oxidation of limonene in high-NO<sub>X</sub> experiments, <sup>30</sup> where ~73% of the precursor is oxidized by NO<sub>3</sub>. These experiments may be better suited to a parameterization of SOA formation following reaction with NO<sub>3</sub>, since the formation of organonitrate species may impact

formation following reaction with NO<sub>3</sub>, since the formation of organonitrate species may impact aerosol yields<sup>52</sup> relative to reaction only with OH and O<sub>3</sub>.

Perhaps the most important difference between the SOA parameterization<sup>18</sup> and the box

modeling results (Figure 3) is that  $\beta$ -based parameterizations do not explicitly capture

isomerization. The potential importance of RO2 isomerization can be described in terms of the

RO<sub>2</sub> bimolecular lifetime,  $\tau_{bi}$ .<sup>47,49</sup> This metric enables comparison of RO<sub>2</sub> conditions in chamber

experiments to atmospheric conditions,<sup>21</sup> in addition to facilitating the interpretation of trends in

<sup>&</sup>lt;sup>b</sup> Monoterpene parameterization is based on  $\alpha$ -pinene. Sesquiterpene parameterization is based on  $\beta$ -caryophyllene and  $\alpha$ -humulene, but only  $\beta$ -caryophyllene model data is used here since  $\alpha$ -humulene is not in the MCM.

<sup>&</sup>lt;sup>c</sup> In Pye et al., <sup>18</sup> corresponding low-NO<sub>X</sub> monoterpene data<sup>31</sup> is scaled by a factor of 0.5 based on Ng et al. <sup>26</sup> and Pathak et al. <sup>51</sup> There are therefore no direct model runs that correspond to this parameterization.

<sup>&</sup>lt;sup>d</sup> In Pye et al., <sup>18</sup> corresponding "low-NO<sub>X</sub>" sesquiterpene data<sup>29</sup> is scaled by a factor of 2 based on Ng et al. <sup>26</sup> There are therefore no direct model runs that correspond to this parameterization.

RO<sub>2</sub> isomerization. Figure 4 shows the bimolecular lifetime for each experiment considered in this study, in comparison with the global distribution of bimolecular lifetimes (weighted by the isoprene oxidation rate) taken from GEOS-Chem (version 13.4.0, from hourly output for January and July 2016).<sup>21</sup>

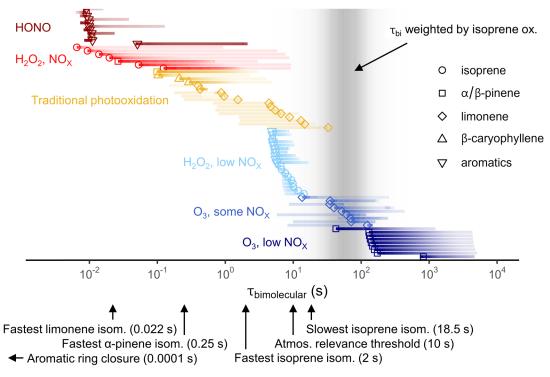


Figure 4:  $\tau_{bi}$  in chamber experiments and in the atmosphere. The shaded gray region represents the atmospheric distribution of  $RO_2$  bimolecular lifetimes taken from GEOS-Chem, weighted by total isoprene reaction rate. The shading is proportional to the frequency of a given atmospheric  $\tau_{bi}$ . Modeled chamber experiments are represented by points with tails. Each point represents the  $\tau_{bi}$  at the start of the experiment while the tail shows the evolution of  $\tau_{bi}$  over the course of the experiment. The tail transparency is scaled to the percentage of precursor remaining. Point shape and color correspond to precursor identity and experiment category (Table 1), respectively. Lifetimes corresponding to a range of relevant isomerization rates included in the box model are annotated at the bottom of the plot. Values of  $\tau_{uni}$  are included in parentheses.

Figure 4 demonstrates that experimental  $\tau_{bi}$  varies over many orders of magnitude. Given this wide range, only a limited number of experiments fall within the global atmospheric lifetime distribution of RO<sub>2</sub> radicals weighted by isoprene oxidation (centered at ~ 60 s and shown in gray; the distribution is similar when weighted by monoterpene oxidation<sup>21</sup>), demonstrating that laboratory conditions usually differ from those found in the atmosphere. However, despite the large overall spread, experiments can clearly be grouped by their oxidation conditions. Experiments featuring high levels of added NO<sub>X</sub> fall around  $\tau_{bi} < \sim 1$  s, H<sub>2</sub>O<sub>2</sub> experiments fall near  $\tau_{bi} = 10$  s, ozonolysis with some NO<sub>X</sub> fall around  $\tau_{bi} = 10 - 100$  s, and low-NO<sub>X</sub> ozonolysis falls around  $\tau_{bi} > \sim 100$  s. While many "traditional photooxidation" experiments feature  $\tau_{bi}$  on the order of 1 s, some experiments<sup>30</sup> feature longer  $\tau_{bi}$  despite high NO<sub>X</sub> levels, due to suppression of NO by O<sub>3</sub>. The RO<sub>2</sub> bimolecular lifetime also varies within experiments (the "tails" in Fig. 4), often increasing by more than an order of magnitude due to decreasing HO<sub>X</sub> and NO<sub>X</sub> levels.

- While for very long or very short  $\tau_{bi}$ , this change over time may not substantially affect observed
- chemistry, shifts across intermediate  $\tau_{bi}$  may affect the importance of unimolecular reactions as
- the experiment proceeds.
- Comparing the RO<sub>2</sub> bimolecular lifetime to the unimolecular lifetime of a given isomerization
- reaction indicates whether that reaction will occur. The unimolecular lifetimes corresponding to
- a selection of isomerization rates used in the model are annotated in Figure 4 (slower
- isomerization reactions are not shown, as only the relatively fast ones are included in the
- mechanisms). While these vary by several orders of magnitude, all are shorter than the mean of
- the atmospheric  $\tau_{bi}$  distribution, indicating that these reactions will play an important role under
- most atmospheric conditions. This is consistent with an "atmospherically relevant" threshold of
- 452  $0.1 \text{ s}^{-1}$  ( $\tau_{\text{uni}} = 10 \text{ s}$ ), which has been used to represent a lower bound for the rate at which
- unimolecular chemistry can proceed as a substantial fraction of RO<sub>2</sub> fate. The majority of low-
- NO<sub>X</sub> experiments involve  $\tau_{bi}$  near or greater than this threshold, indicating that these chamber
- experiments included the effects isomerization reactions, even well before the importance of RO<sub>2</sub>
- isomerization reactions was appreciated. In contrast, higher  $NO_X$  experiments feature a shorter  $\tau_{bi}$
- that prevents all but the fastest reactions from occurring.
- The wide variability of both experimental  $\tau_{bi}$  and isomerization rates complicate the relationship
- between experimental measurements and model SOA parameterizations. Where an isomerization
- channel is extremely fast (e.g., ring closure reactions in aromatic species) so as to outcompete
- bimolecular reactions under all conditions, the exclusion of isomerization in the parameterization
- does not matter, since this reaction will always occur. In contrast, many "atmospherically-
- relevant" isomerization reactions ( $\tau_{uni} \le 10 \text{ s}$ ) occur only under low-NO<sub>X</sub> H<sub>2</sub>O<sub>2</sub> and ozonolysis
- 464 conditions, thereby playing a role in the gas-phase chemistry (and ultimately aerosol formation)
- in these experiments. While the importance of these reactions in such experiments may not
- perfectly match that of atmospheric conditions, 50 their occurrence in chamber experiments means
- that rapid isomerization processes are to some extent built into the global model aerosol
- parameterization for RO<sub>2</sub> + HO<sub>2</sub> conditions. However, chamber experiments are unlikely to
- perfectly capture the importance of isomerization under most atmospheric conditions. Moreover,
- under high-NO<sub>X</sub> conditions, isomerization is largely quenched, such that little to no
- isomerization is built into the aerosol parameterization for the high-NO<sub>X</sub> conditions.

## Changing chemical conditions across yield curves

- As mentioned above, aerosol yield curves (aerosol yield vs. mass concentration of organic
- aerosol) can be used to derive a volatility distribution that is useful for representing SOA in 3D
- 475 models. An underlying assumption for an aerosol yield curve is that the only modified variable is
- 476 the mass of organic aerosol into which the gas-phase species partition. However, since the curve
- 477 is typically made up of measurements across a series of experiments (which typically involve
- different starting precursor concentrations), variability in chemical conditions across these
- experiments may affect the resulting aerosol yield curve.

Figure 5 illustrates how chemical conditions can vary across a series of aerosol yield experiments. In this series of experiments, Zhang et al.<sup>30</sup> oxidize limonene under dark ozonolysis conditions, with low but nonzero NO<sub>X</sub>. While these experiments were designed to favor the reaction of limonene with ozone, the NO<sub>X</sub> present is sufficient to make reaction with NO<sub>3</sub> an important contributor to limonene oxidation. As shown in Figure 5a, this is most pronounced for experiments with low initial precursor concentrations and hence low organic aerosol concentrations ( $c_{OA}$ ); this is due to variation of the NO<sub>3</sub>/O<sub>3</sub> among experiments. Further, at high precursor concentrations (and high  $c_{OA}$ ), OH generated by the ozonolysis of limonene starts to become an important oxidant, exceeding the capacity of the OH scavenger (5 ppm 2-butanol) to keep [OH] low.

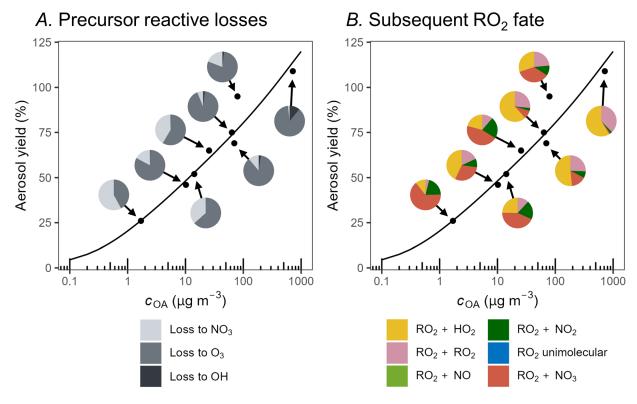


Figure 5: Modeled chemical conditions as a function of measured organic aerosol mass and organic aerosol yield for an example set of experiments. Panel A shows the distribution of precursor reactive loss (integrated over the entire chamber run) for experiments featuring the oxidation of limonene  $(O_3, some\ NO_X)$  from Zhang et al.<sup>30</sup> (expt. 8 – 15). Each experiment is depicted as a pie chart, plotted across the aerosol mass vs aerosol yield space. Panel B shows subsequent  $RO_2$  fate for the same series of experiments plotted in the same way.

Peroxy radical chemistry may also shift as a function of precursor VOC concentration (and therefore  $c_{OA}$ ). While initial peroxy radical fate is relatively stable across these experiments (see Figure S.5), Figure 5B shows that subsequent RO<sub>2</sub> fate can also shift across a series of experiments. At low precursor concentrations (low  $c_{OA}$ ), reactions with NO<sub>3</sub> and NO<sub>2</sub> are relatively important, while RO<sub>2</sub> + RO<sub>2</sub> reactions become more important at higher precursor concentrations (and higher  $c_{OA}$ ). While our modeled subsequent chemistry does not include

- isomerization reactions that can redistribute RO<sub>2</sub> fate, variations in bimolecular chemistry, due
- for example to differing total RO<sub>2</sub> concentrations, are likely to persist.
- While Fig 5 (Zhang et al.<sup>30</sup>) provides an example of how chemical conditions may shift across an
- aerosol yield curve, conditions can be more uniform for other experiment types, as shown in
- Figures S.7 to S.12. For experiments which clearly favor a single oxidant or RO<sub>2</sub> fate (e.g., low-
- NO<sub>X</sub> H<sub>2</sub>O<sub>2</sub> experiments), changes in VOC concentration or NO<sub>X</sub> background have a smaller
- 508 impact on modeled chemical conditions. Nonetheless some shifts, such as changes in the
- prevalence of  $RO_2 + RO_2$  reactions, can appear even under these more extreme conditions
- 510 (Figure S.8 and S.10).

## **Implications**

- This work demonstrates that chemical conditions in an environmental chamber often represent a
- 513 complex mix of VOC reactive loss pathways and RO<sub>2</sub> fates, making it challenging to assign a
- single pathway to a given experiment. Conditions often vary over the course of an experiment,
- and between relatively similar experiments. These changes occur in both the VOC loss pathways
- and the peroxy radical fate, and are caused by multiple factors. VOC loss pathways can shift as a
- result of changes not only in oxidant concentration, but also in reactivity, as ozone reactivity and
- capacity for photolysis changes as the VOC mixture evolves. Peroxy radical fates similarly shift
- as an experiment proceeds, with NO<sub>2</sub> driving later-generation PAN formation, greater total RO<sub>2</sub>
- concentrations increasing  $RO_2 + RO_2$  reactions, and decreasing NO and  $HO_2$  leading to increases
- in the importance of unimolecular reactions (due to increases in  $\tau_{bi}$ ).
- Patterns in relative VOC loss pathways and RO<sub>2</sub> fate arise that are correlated with the varied sets
- of starting oxidation conditions. While chamber experiments using HONO, H<sub>2</sub>O<sub>2</sub> (with or
- without NO<sub>X</sub>), and ozone (without NO<sub>X</sub>) all feature oxidation dominated by a single oxidant,
- other types of experiments (traditional photooxidation and "O<sub>3</sub>, some NO<sub>X</sub>") involve a mixture
- of oxidants, due to the formation of substantial OH, O<sub>3</sub>, and NO<sub>3</sub> concentrations. RO<sub>2</sub> conditions
- can similarly be grouped by experiment type, with comparable bimolecular lifetimes and
- fractional RO<sub>2</sub> fate between like experiments; few experimental conditions (low-NO<sub>X</sub> H<sub>2</sub>O<sub>2</sub>,
- HONO) consistently favor a single RO<sub>2</sub> fate.
- The dynamic chemistry observed here may be relevant for the calculation of aerosol yield
- curves, which represent the basis for aerosol parameterizations. As shown in this work, chemical
- conditions, including both oxidant and peroxy radical fate, can sometimes vary across a series of
- 533 SOA experiments from which an aerosol yield curve is derived. Shifts in chemical conditions
- which vary as a function of precursor concentration (and hence  $c_{OA}$ ) may skew measured yields,
- resulting in fitted volatility distributions that do not represent a single chemical condition. The
- extent to which these changes may affect measured aerosol formation is unclear, as these
- changes are often subtle, and the specific impact of shifting oxidant or RO<sub>2</sub> fate on SOA
- formation is often not well understood.

- When considering the application of SOA chamber studies to 3D modeling, we find that
- experimental oxidant and RO<sub>2</sub> conditions for a given experiment do not always match how it is
- used in a 3D model aerosol parameterization. Most notably, the GEOS-Chem SOA
- parameterization bases some low-NO<sub>X</sub> SOA formation on experiments in which NO<sub>X</sub>
- concentrations are high enough to influence observed chemistry. With more experimental data
- now available, reevaluating and constraining aerosol parameterizations may represent a valuable
- avenue for future research. Further, since the atmosphere can involve mixed chemical fates,
- future work should additionally consider whether the basis for SOA parameterizations should
- expand beyond experiments targeting single oxidants or RO<sub>2</sub> fates.<sup>21</sup>
- Finally, this work highlights the importance of considering RO<sub>2</sub> fates beyond NO and HO<sub>2</sub> for
- regional and global SOA parameterizations. Despite the limitations of previous SOA
- experiments, RO<sub>2</sub> isomerization reactions, whose importance for SOA formation was not yet
- appreciated when these experiments were carried out, proceed under many of the experimental
- 552 conditions used. Ozonolysis or H<sub>2</sub>O<sub>2</sub> experiments with no added NOx feature relatively long
- bimolecular lifetimes ( $\tau_{bi} > \sim 10 \text{ s}$ ), allowing for reasonably rapid isomerization ( $\tau_{uni} < \sim 10 \text{ s}$ ) to
- occur. Because of this, HO<sub>2</sub>-pathway aerosol yields in 3D models implicitly include the effects
- of some RO<sub>2</sub> isomerization on aerosol yields. Still, chamber experiments rarely fall into the same
- range of bimolecular lifetimes that occur in the atmosphere, indicating that laboratory peroxy
- radical conditions generally do not fully match those in the real environment.<sup>21</sup> Ultimately, it is
- not clear whether this approach to yield parameterizations is sufficient to characterize aerosol
- formation in the real atmosphere, which may involve slower isomerization processes that do not
- occur in a lab setting, or isomerization processes that may at times compete with reaction with
- 561 NO.
- This study seeks to explore unmeasured chamber conditions used in key historical SOA
- experiments, and in doing so, identifies patterns that may be useful both for designing laboratory
- experiments and for implementing parameterizations of laboratory chemistry into models. Future
- experimental work should consider the role of changing chemical fate with respect to first- and
- later-generation compounds, and the influence of changing starting conditions across sets of
- aerosol yield experiments (Figure 5). Alternative experimental designs, such as the use of steady-
- state continuous flow chamber experiments, may help mitigate the challenges associated with
- conditions that change over the course of an experiment. Future SOA model schemes must at
- least consider the uncertainties in experimental chemical conditions, if not explore the
- development of wholly new frameworks to translate experimental yield measurements to the
- 572 modeling sphere.

#### Data availability

- Data for this article, including all figure data, Matlab code, and model input conditions are
- available on the Kroll Group publication website at http://krollgroup.mit.edu/publications.html.

#### **Author contributions** 576

- 577 MBG performed all box modeling, analyzed the data, and wrote the paper. HSK assisted in
- framing the project, provided GEOS-Chem output, and shared code relevant to RO<sub>2</sub> fate analysis. 578
- CLH aided in project design and manuscript preparation. JHK aided in project conceptualization, 579
- project design, data interpretation, and manuscript preparation. 580

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#### References

- Zhang, O.; Jimenez, J. L.; Canagaratna, M. R.; Allan, J. D.; Coe, H.; Ulbrich, I.; Alfarra, M. 594 R.; Takami, A.; Middlebrook, A. M.; Sun, Y. L.; Dzepina, K.; Dunlea, E.; Docherty, K.; 595 DeCarlo, P. F.; Salcedo, D.; Onasch, T.; Jayne, J. T.; Miyoshi, T.; Shimono, A.; Hatakeyama, 596 S.; Takegawa, N.; Kondo, Y.; Schneider, J.; Drewnick, F.; Borrmann, S.; Weimer, S.; 597 Demerjian, K.; Williams, P.; Bower, K.; Bahreini, R.; Cottrell, L.; Griffin, R. J.; Rautiainen, 598 J.; Sun, J. Y.; Zhang, Y. M.; Worsnop, D. R. Ubiquity and Dominance of Oxygenated Species 599 in Organic Aerosols in Anthropogenically-Influenced Northern Hemisphere Midlatitudes. 600 Geophys. Res. Lett. 2007, 34 (13). https://doi.org/10.1029/2007GL029979. 601
- (2) Jimenez, J. L.; Canagaratna, M. R.; Donahue, N. M.; Prevot, A. S. H.; Zhang, Q.; Kroll, J. 602 H.; DeCarlo, P. F.; Allan, J. D.; Coe, H.; Ng, N. L.; Aiken, A. C.; Docherty, K. S.; Ulbrich, 603 I. M.; Grieshop, A. P.; Robinson, A. L.; Duplissy, J.; Smith, J. D.; Wilson, K. R.; Lanz, V. 604 A.; Hueglin, C.; Sun, Y. L.; Tian, J.; Laaksonen, A.; Raatikainen, T.; Rautiainen, J.; 605 Vaattovaara, P.; Ehn, M.; Kulmala, M.; Tomlinson, J. M.; Collins, D. R.; Cubison, M. J.; E.; 606 Dunlea, J.; Huffman, J. A.; Onasch, T. B.; Alfarra, M. R.; Williams, P. I.; Bower, K.; Kondo, 607 Y.; Schneider, J.; Drewnick, F.; Borrmann, S.; Weimer, S.; Demerjian, K.; Salcedo, D.; 608 609 Cottrell, L.; Griffin, R.; Takami, A.; Miyoshi, T.; Hatakeyama, S.; Shimono, A.; Sun, J. Y.; Zhang, Y. M.; Dzepina, K.; Kimmel, J. R.; Sueper, D.; Jayne, J. T.; Herndon, S. C.; Trimborn, 610 A. M.; Williams, L. R.; Wood, E. C.; Middlebrook, A. M.; Kolb, C. E.; Baltensperger, U.; 611 Worsnop, D. R. Evolution of Organic Aerosols in the Atmosphere. Science 2009, 326 (5959), 612
- 1525–1529. https://doi.org/10.1126/science.1180353. 613

- Tsigaridis, K.; Daskalakis, N.; Kanakidou, M.; Adams, P. J.; Artaxo, P.; Bahadur, R.; 614 Balkanski, Y.; Bauer, S. E.; Bellouin, N.; Benedetti, A.; Bergman, T.; Berntsen, T. K.; 615 Beukes, J. P.; Bian, H.; Carslaw, K. S.; Chin, M.; Curci, G.; Diehl, T.; Easter, R. C.; Ghan, 616 S. J.; Gong, S. L.; Hodzic, A.; Hoyle, C. R.; Iversen, T.; Jathar, S.; Jimenez, J. L.; Kaiser, J. 617 W.; Kirkevåg, A.; Koch, D.; Kokkola, H.; Lee, Y. H.; Lin, G.; Liu, X.; Luo, G.; Ma, X.; 618 Mann, G. W.; Mihalopoulos, N.; Morcrette, J.-J.; Müller, J.-F.; Myhre, G.; Myriokefalitakis, 619 S.; Ng, N. L.; O'Donnell, D.; Penner, J. E.; Pozzoli, L.; Pringle, K. J.; Russell, L. M.; Schulz, 620 M.; Sciare, J.; Seland, Ø.; Shindell, D. T.; Sillman, S.; Skeie, R. B.; Spracklen, D.; Stavrakou, 621 T.; Steenrod, S. D.; Takemura, T.; Tiitta, P.; Tilmes, S.; Tost, H.; van Noije, T.; van Zyl, P. 622 G.; von Salzen, K.; Yu, F.; Wang, Z.; Zaveri, R. A.; Zhang, H.; Zhang, K.; Zhang, Q.; Zhang, 623 X. The AeroCom Evaluation and Intercomparison of Organic Aerosol in Global Models. 624 Atmospheric Chem. Phys. 2014, 14 (19), 10845-10895. https://doi.org/10.5194/acp-14-625 10845-2014. 626
- 627 (4) Pai, S. J.; Heald, C. L.; Pierce, J. R.; Farina, S. C.; Marais, E. A.; Jimenez, J. L.; Campuzano-628 Jost, P.; Nault, B. A.; Middlebrook, A. M.; Coe, H.; Shilling, J. E.; Bahreini, R.; Dingle, J. 629 H.; Vu, K. An Evaluation of Global Organic Aerosol Schemes Using Airborne Observations. 630 Atmospheric Chem. Phys. 2020, 20 (5), 2637–2665. https://doi.org/10.5194/acp-20-2637-631 2020.
- 632 (5) Kroll, J. H.; Seinfeld, J. H. Chemistry of Secondary Organic Aerosol: Formation and Evolution of Low-Volatility Organics in the Atmosphere. *Atmos. Environ.* **2008**, *42* (16), 3593–3624. https://doi.org/10.1016/j.atmosenv.2008.01.003.
- Vaughan, S.; Canosa-Mas, C. E.; Pfrang, C.; Shallcross, D. E.; Watson, L.; Wayne, R. P.
   Kinetic Studies of Reactions of the Nitrate Radical (NO3) with Peroxy Radicals (RO2): An
   Indirect Source of OH at Night? *Phys. Chem. Chem. Phys.* 2006, 8 (32), 3749.
   https://doi.org/10.1039/b605569a.
- 639 (7) Crounse, J. D.; Nielsen, L. B.; Jørgensen, S.; Kjaergaard, H. G.; Wennberg, P. O. 640 Autoxidation of Organic Compounds in the Atmosphere. *J. Phys. Chem. Lett.* **2013**, *4* (20), 641 3513–3520. https://doi.org/10.1021/jz4019207.
- 642 (8) Ehn, M.; Thornton, J. A.; Kleist, E.; Sipilä, M.; Junninen, H.; Pullinen, I.; Springer, M.; Rubach, F.; Tillmann, R.; Lee, B.; Lopez-Hilfiker, F.; Andres, S.; Acir, I.-H.; Rissanen, M.; Jokinen, T.; Schobesberger, S.; Kangasluoma, J.; Kontkanen, J.; Nieminen, T.; Kurtén, T.; Nielsen, L. B.; Jørgensen, S.; Kjaergaard, H. G.; Canagaratna, M.; Maso, M. D.; Berndt, T.; Petäjä, T.; Wahner, A.; Kerminen, V.-M.; Kulmala, M.; Worsnop, D. R.; Wildt, J.; Mentel, T. F. A Large Source of Low-Volatility Secondary Organic Aerosol. *Nature* 2014, 506 (7489), 476–479. https://doi.org/10.1038/nature13032.
- (9) Vereecken, L.; Nozière, B. H Migration in Peroxy Radicals under Atmospheric Conditions.
   Atmospheric Chem. Phys. 2020, 20 (12), 7429–7458. https://doi.org/10.5194/acp-20-7429-2020.
- (10) Berndt, T.; Scholz, W.; Mentler, B.; Fischer, L.; Herrmann, H.; Kulmala, M.; Hansel, A. 652 Accretion Product Formation from Self- and Cross-Reactions of RO2 Radicals in the 653 Atmosphere. Angew. Chem. Int. 2018. (14),3820-3824. Ed.57 654 https://doi.org/10.1002/anie.201710989. 655

- 656 (11) Berndt, T.; Chen, J.; Kj, E. R.; Hoffmann, E. H.; Herrmann, H.; Crounse, J. D.; Wennberg, P. O.; Kjaergaard, H. G. Hydrotrioxide (ROOOH) Formation in the Atmosphere. **2022**, 5.
- 658 (12) Farina, S. C.; Adams, P. J.; Pandis, S. N. Modeling Global Secondary Organic Aerosol 659 Formation and Processing with the Volatility Basis Set: Implications for Anthropogenic 660 Secondary Organic Aerosol. *J. Geophys. Res. Atmospheres* **2010**, *115* (D9). 661 https://doi.org/10.1029/2009JD013046.
- 662 (13) Pankow, J. F. An Absorption Model of the Gas/Aerosol Partitioning Involved in the Formation of Secondary Organic Aerosol. *Atmos. Environ.* **1994**, *28* (2), 189–193. https://doi.org/10.1016/1352-2310(94)90094-9.
- (14) Odum, J. R.; Hoffmann, T.; Bowman, F.; Collins, D.; Flagan, R. C.; Seinfeld, J. H.
   Gas/Particle Partitioning and Secondary Organic Aerosol Yields. *Environ. Sci. Technol.* 1996, 30 (8), 2580–2585. https://doi.org/10.1021/es950943+.
- (15) Donahue, N. M.; Robinson, A. L.; Stanier, C. O.; Pandis, S. N. Coupled Partitioning,
   Dilution, and Chemical Aging of Semivolatile Organics. *Environ. Sci. Technol.* 2006, 40 (8),
   2635–2643. https://doi.org/10.1021/es052297c.
- (16) Presto, A. A.; Donahue, N. M. Investigation of α-Pinene + Ozone Secondary Organic Aerosol
   Formation at Low Total Aerosol Mass. *Environ. Sci. Technol.* 2006, 40 (11), 3536–3543.
   https://doi.org/10.1021/es052203z.
- (17) Henze, D. K.; Seinfeld, J. H.; Ng, N. L.; Kroll, J. H.; Fu, T.-M.; Jacob, D. J.; Heald, C. L.
   Global Modeling of Secondary Organic Aerosol Formation from Aromatic Hydrocarbons:
   High- vs. Low-Yield Pathways. *Atmospheric Chem. Phys.* 2008, 8 (9), 2405–2420.
   https://doi.org/10.5194/acp-8-2405-2008.
- 678 (18) Pye, H. O. T.; Chan, A. W. H.; Barkley, M. P.; Seinfeld, J. H. Global Modeling of Organic 679 Aerosol: The Importance of Reactive Nitrogen (NO<sub>X</sub> and NO<sub>3</sub>). *Atmospheric Chem. Phys.* 680 **2010**, 10 (22), 11261–11276. https://doi.org/10.5194/acp-10-11261-2010.
- 681 (19) Porter, W. C.; Jimenez, J. L.; Barsanti, K. C. Quantifying Atmospheric Parameter Ranges for 682 Ambient Secondary Organic Aerosol Formation. *ACS Earth Space Chem.* **2021**, *5* (9), 2380– 683 2397. https://doi.org/10.1021/acsearthspacechem.1c00090.
- (20) Krechmer, J. E.; Day, D. A.; Jimenez, J. L. Always Lost but Never Forgotten: Gas-Phase
   Wall Losses Are Important in All Teflon Environmental Chambers. *Environ. Sci. Technol.* 2020, 54 (20), 12890–12897. https://doi.org/10.1021/acs.est.0c03381.
- 687 (21) Kenagy, H. S.; Heald, C. L.; Tahsini, N.; Goss, M. B.; Kroll, J. H. Can We Achieve 688 Atmospheric Chemical Environments in the Laboratory? An Integrated Model-Measurement 689 Approach to Chamber SOA Studies. *Sci. Adv.* **2024**, *10* (37), eado1482. 690 https://doi.org/10.1126/sciadv.ado1482.
- (22) Zhang, X.; Cappa, C. D.; Jathar, S. H.; McVay, R. C.; Ensberg, J. J.; Kleeman, M. J.; Seinfeld, 691 692 J. H. Influence of Vapor Wall Loss in Laboratory Chambers on Yields of Secondary Organic Natl. 2014. 5802-5807. 693 Aerosol. Proc. Acad. Sci. 111 (16),https://doi.org/10.1073/pnas.1404727111. 694
- 695 (23) Fisher, J. A.; Jacob, D. J.; Travis, K. R.; Kim, P. S.; Marais, E. A.; Chan Miller, C.; Yu, K.; 696 Zhu, L.; Yantosca, R. M.; Sulprizio, M. P.; Mao, J.; Wennberg, P. O.; Crounse, J. D.; Teng,

- A. P.; Nguyen, T. B.; St. Clair, J. M.; Cohen, R. C.; Romer, P.; Nault, B. A.; Wooldridge, P. J.; Jimenez, J. L.; Campuzano-Jost, P.; Day, D. A.; Hu, W.; Shepson, P. B.; Xiong, F.; Blake, D. R.; Goldstein, A. H.; Misztal, P. K.; Hanisco, T. F.; Wolfe, G. M.; Ryerson, T. B.; Wisthaler, A.; Mikoviny, T. Organic Nitrate Chemistry and Its Implications for Nitrogen Budgets in an Isoprene- and Monoterpene-Rich Atmosphere: Constraints from Aircraft (SEAC<sup>4</sup>RS) and Ground-Based (SOAS) Observations in the Southeast US. *Atmospheric Chem. Phys.* **2016**, *16* (9), 5969–5991. https://doi.org/10.5194/acp-16-5969-2016.
- (24) Marais, E. A.; Jacob, D. J.; Jimenez, J. L.; Campuzano-Jost, P.; Day, D. A.; Hu, W.;
  Krechmer, J.; Zhu, L.; Kim, P. S.; Miller, C. C.; Fisher, J. A.; Travis, K.; Yu, K.; Hanisco, T.
  F.; Wolfe, G. M.; Arkinson, H. L.; Pye, H. O. T.; Froyd, K. D.; Liao, J.; McNeill, V. F.
  Aqueous-Phase Mechanism for Secondary Organic Aerosol Formation from Isoprene:
  Application to the Southeast United States and Co-Benefit of SO<sub>2</sub> Emission Controls.
  Atmospheric Chem. Phys. 2016, 16 (3), 1603–1618. https://doi.org/10.5194/acp-16-1603-2016.
- 711 (25) Wennberg, P. O. Let's Abandon the "High NO<sub>X</sub>" and "Low NO<sub>X</sub>" Terminology. *ACS EST Air* **2024**, *1* (1), 3–4. https://doi.org/10.1021/acsestair.3c00055.
- (26) Ng, N. L.; Chhabra, P. S.; Chan, A. W. H.; Surratt, J. D.; Kroll, J. H.; Kwan, A. J.; McCabe,
   D. C.; Wennberg, P. O.; Sorooshian, A.; Murphy, S. M.; Dalleska, N. F.; Flagan, R. C.;
   Seinfeld, J. H. Effect of NOx Level on Secondary Organic Aerosol (SOA) Formation from
   the Photooxidation of Terpenes. *Atmos Chem Phys* 2007, 16.
- 717 (27) Ng, N. L.; Kroll, J. H.; Chan, A. W. H.; Chhabra, P. S.; Flagan, R. C.; Seinfeld, J. H. Secondary Organic Aerosol Formation from *m*-Xylene, Toluene, and Benzene. *Atmospheric Chem. Phys.* **2007**, *7* (14), 3909–3922. https://doi.org/10.5194/acp-7-3909-2007.
- 720 (28) Kroll, J. H.; Ng, N. L.; Murphy, S. M.; Flagan, R. C.; Seinfeld, J. H. Secondary Organic 721 Aerosol Formation from Isoprene Photooxidation. *Environ. Sci. Technol.* **2006**, *40* (6), 1869– 722 1877. https://doi.org/10.1021/es0524301.
- 723 (29) Griffin, R. J.; Cocker, D. R.; Flagan, R. C.; Seinfeld, J. H. Organic Aerosol Formation from 724 the Oxidation of Biogenic Hydrocarbons. *J. Geophys. Res. Atmospheres* **1999**, *104* (D3), 725 3555–3567. https://doi.org/10.1029/1998JD100049.
- 726 (30) Zhang, J.; Huff Hartz, K. E.; Pandis, S. N.; Donahue, N. M. Secondary Organic Aerosol Formation from Limonene Ozonolysis: Homogeneous and Heterogeneous Influences as a 727 728 Function of  $NO_X$ . J. Phvs. Chem. 2006. 110 (38),11053-11063. https://doi.org/10.1021/jp062836f. 729
- (31) Shilling, J. E.; Chen, Q.; King, S. M.; Rosenoern, T.; Kroll, J. H.; Worsnop, D. R.; McKinney,
   K. A.; Martin, S. T. Particle Mass Yield in Secondary Organic Aerosol Formed by the Dark
   Ozonolysis of α-Pinene. *Atmos Chem Phys* 2008, 16.
- 733 (32) Wolfe, G. M.; Marvin, M. R.; Roberts, S. J.; Travis, K. R.; Liao, J. The Framework for 0-D Atmospheric Modeling (F0AM) v3.1. *Geosci. Model Dev.* **2016**, *9* (9), 3309–3319. https://doi.org/10.5194/gmd-9-3309-2016.
- 736 (33) Jenkin, M. E.; Saunders, S. M.; Pilling, M. J. The Tropospheric Degradation of Volatile 737 Organic Compounds: A Protocol for Mechanism Development. *Atmos. Environ.* **1997**, *31* 738 (1), 81–104. https://doi.org/10.1016/S1352-2310(96)00105-7.

- 739 (34) Jenkin, M. E.; Saunders, S. M.; Wagner, V.; Pilling, M. J. Protocol for the Development of 740 the Master Chemical Mechanism, MCM v3 (Part B): Tropospheric Degradation of Aromatic 741 Volatile Organic Compounds. *Atmospheric Chem. Phys.* **2003**, *3* (1), 181–193. 742 https://doi.org/10.5194/acp-3-181-2003.
- (35) Jenkin, M. E.; Wyche, K. P.; Evans, C. J.; Carr, T.; Monks, P. S.; Alfarra, M. R.; Barley, M.
   H.; McFiggans, G. B.; Young, J. C.; Rickard, A. R. Development and Chamber Evaluation of the MCM v3.2 Degradation Scheme for β-Caryophyllene. *Atmospheric Chem. Phys.* 2012, 12 (11), 5275–5308. https://doi.org/10.5194/acp-12-5275-2012.
- 747 (36) Jenkin, M. E.; Young, J. C.; Rickard, A. R. The MCM v3.3.1 Degradation Scheme for 748 Isoprene. *Atmospheric Chem. Phys. Katlenburg-Lindau* **2015**, *15* (20), 11433–11459. 749 http://dx.doi.org/10.5194/acp-15-11433-2015.
- 750 (37) Bloss, C.; Wagner, V.; Jenkin, M. E.; Volkamer, R.; Bloss, W. J.; Lee, J. D.; Heard, D. E.; Wirtz, K.; Martin-Reviejo, M.; Rea, G.; Wenger, J. C.; Pilling, M. J. Development of a Detailed Chemical Mechanism (MCMv3.1) for the Atmospheric Oxidation of Aromatic Hydrocarbons. *Atmospheric Chem. Phys.* **2005**, *5* (3), 641–664. https://doi.org/10.5194/acp-5-641-2005.
- 755 (38) Ng, N. L.; Kwan, A. J.; Surratt, J. D.; Chan, A. W. H.; Chhabra, P. S.; Sorooshian, A.; Pye, H. O. T.; Crounse, J. D.; Wennberg, P. O.; Flagan, R. C.; Seinfeld, J. H. Secondary Organic Aerosol (SOA) Formation from Reaction of Isoprene with Nitrate Radicals (NO<sub>3</sub>). Atmospheric Chem. Phys. 2008, 8 (14), 4117–4140. https://doi.org/10.5194/acp-8-4117-2008.
- (39) Iyer, S.; Rissanen, M. P.; Valiev, R.; Barua, S.; Krechmer, J. E.; Thornton, J.; Ehn, M.;
   Kurtén, T. Molecular Mechanism for Rapid Autoxidation in α-Pinene Ozonolysis. *Nat. Commun.* 2021, *12* (1), 878. https://doi.org/10.1038/s41467-021-21172-w.
- (40) Møller, K. H.; Otkjær, R. V.; Chen, J.; Kjaergaard, H. G. Double Bonds Are Key to Fast
   Unimolecular Reactivity in First-Generation Monoterpene Hydroxy Peroxy Radicals. J.
   Phys. Chem. A 2020, 124 (14), 2885–2896. https://doi.org/10.1021/acs.jpca.0c01079.
- (41) Pang, J. Y. S.; Novelli, A.; Kaminski, M.; Acir, I.-H.; Bohn, B.; Carlsson, P. T. M.; Cho, C.;
   Dorn, H.-P.; Hofzumahaus, A.; Li, X.; Lutz, A.; Nehr, S.; Reimer, D.; Rohrer, F.; Tillmann,
   R.; Wegener, R.; Kiendler-Scharr, A.; Wahner, A.; Fuchs, H. Investigation of the Limonene
   Photooxidation by OH at Different NO Concentrations in the Atmospheric Simulation
   Chamber SAPHIR (Simulation of Atmospheric PHotochemistry In a Large Reaction
   Chamber). Atmospheric Chem. Phys. 2022, 22 (13), 8497–8527. https://doi.org/10.5194/acp 22-8497-2022.
- 773 (42) Vereecken, L. Reaction Mechanisms for the Atmospheric Oxidation of Monocyclic Aromatic 774 Compounds. In *Advances in Atmospheric Chemistry*; Advances in Atmospheric Chemistry; 775 WORLD SCIENTIFIC, 2018; Vol. Volume 2, pp 377–527. 776 https://doi.org/10.1142/9789813271838\_0006.
- 777 (43) Vereecken, L.; Vu, G.; Wahner, A.; Kiendler-Scharr, A.; Nguyen, H. M. T. A Structure 778 Activity Relationship for Ring Closure Reactions in Unsaturated Alkylperoxy Radicals. 779 *Phys. Chem. Chem. Phys.* **2021**, *23* (31), 16564–16576. 780 https://doi.org/10.1039/D1CP02758A.

- (44) Xu, L.; Møller, K. H.; Crounse, J. D.; Otkjær, R. V.; Kjaergaard, H. G.; Wennberg, P. O.
   Unimolecular Reactions of Peroxy Radicals Formed in the Oxidation of α-Pinene and β-Pinene by Hydroxyl Radicals. J. Phys. Chem. A 2019, 123 (8), 1661–1674.
   https://doi.org/10.1021/acs.jpca.8b11726.
- 785 (45) Cocker, D. R.; Flagan, R. C.; Seinfeld, J. H. State-of-the-Art Chamber Facility for Studying 786 Atmospheric Aerosol Chemistry. *Environ. Sci. Technol.* **2001**, *35* (12), 2594–2601. 787 https://doi.org/10.1021/es0019169.
- 788 (46) Presto, A. A.; Huff Hartz, K. E.; Donahue, N. M. Secondary Organic Aerosol Production from Terpene Ozonolysis. 1. Effect of UV Radiation. *Environ. Sci. Technol.* **2005**, *39* (18), 7036–7045. https://doi.org/10.1021/es050174m.
- (47) da Silva, G.; Graham, C.; Wang, Z.-F. Unimolecular β-Hydroxyperoxy Radical
   Decomposition with OH Recycling in the Photochemical Oxidation of Isoprene. *Environ.* Sci. Technol. 2010, 44 (1), 250–256. https://doi.org/10.1021/es900924d.
- 794 (48) Crounse, J. D.; Paulot, F.; Kjaergaard, H. G.; Wennberg, P. O. Peroxy Radical Isomerization 795 in the Oxidation of Isoprene. *Phys. Chem. Chem. Phys.* **2011**, *13* (30), 13607–13613. 796 https://doi.org/10.1039/C1CP21330J.
- 797 (49) Teng, A. P.; Crounse, J. D.; Wennberg, P. O. Isoprene Peroxy Radical Dynamics. *J. Am. Chem. Soc.* **2017**, *139* (15), 5367–5377. https://doi.org/10.1021/jacs.6b12838.
- (50) D'Ambro, E. L.; Møller, K. H.; Lopez-Hilfiker, F. D.; Schobesberger, S.; Liu, J.; Shilling, J. 799 E.; Lee, B. H.; Kjaergaard, H. G.; Thornton, J. A. Isomerization of Second-Generation 800 Isoprene Peroxy Radicals: Epoxide Formation and Implications for Secondary Organic 801 Aerosol Yields. Environ. Sci. Technol. 2017, 51 (9),4978-4987. 802 https://doi.org/10.1021/acs.est.7b00460. 803
- 804 (51) Pathak, R. K.; Stanier, C. O.; Donahue, N. M.; Pandis, S. N. Ozonolysis of α-Pinene at Atmospherically Relevant Concentrations: Temperature Dependence of Aerosol Mass 805 Fractions (Yields). Geophys. Res. **Atmospheres** 2007, 806  $J_{\cdot}$ 112 (D3). https://doi.org/10.1029/2006JD007436. 807
- (52) Fry, J. L.; Draper, D. C.; Barsanti, K. C.; Smith, J. N.; Ortega, J.; Winkler, P. M.; Lawler, M.
  J.; Brown, S. S.; Edwards, P. M.; Cohen, R. C.; Lee, L. Secondary Organic Aerosol
  Formation and Organic Nitrate Yield from NO3 Oxidation of Biogenic Hydrocarbons.
  Environ. Sci. Technol. 2014, 48 (20), 11944–11953. https://doi.org/10.1021/es502204x.