

Controls of litter chemistry over early lignin decomposition in beech litter ¹

Lukas Kohl¹, Wolfgang Wanek¹, Katharina Keiblinger^{2,3}, Sonja Leitner^{1,3}, Maria Mooshammer¹, Ieda Hämmerle¹, Lucia Fuchslueger¹, Jörg Schneckner¹, Sophie Zechmeister-Boltenstern^{2,3}, Andreas Richter¹,

1 Department of Chemical Ecology and Ecosystem Research, University of Vienna, Althanstraße 14, A-1090 Vienna, Austria

2 Federal Research and Training Centre for Forests, Natural Hazards and Landscape, Department of Soil Biology, Seckendorff-Gudent-Weg 8, A-1131 Vienna, Austria

3 Current address: Institute for Soil Science, University of Natural Resources and Life Sciences, Peter Jordan-Straße 82, A-1190, Vienna, Austria

*** E-mail: Corresponding author@institute.edu**

Abstract

Lignin is considered the most recalcitrant component of plant litter, which is accumulated during early decomposition and is degraded only in late decomposition stages when its concentration limits litter decomposition rates. A recent study based on the more specific methods challenges this concept, reporting highest lignin decomposition rates during early litter decomposition. Until now, no further studies exploring early lignin decomposition were published, and its potential controls remain unknown.

We follow lignin and carbohydrate decomposition during early litter decay with analytical pyrolysis in a climate-chamber decomposition experiment, focusing on resource control over microbial carbon substrate preferences. Beech litter with different C:N:P stoichiometry but identical initial microbial communities was incubated to identify the control of litter chemistry on the developing microbial community and its decomposition activity.

During the first 6 month fundamental differences in lignin degrading activities were found between sites. Lignin discrimination in litter decomposition ranges from insignificant amounts of lignin decomposed to lignin decomposition at the same rate bulk litter, leading to different niveaus of lignin accumulation. Between 6 and 15 month, no lignin discrimination was found, but different lignin contents aquired earlier reminded.

[results]

Plant litter biomass is dominated by macromolecular compounds. In deciduous foliar litter, lignin and carbohydrate polymers together make up 40-60% of litter dry mass [1], while leachable substances ("DOM") account for only 1.5-6% [2].

Litter decomposition models follow the concept that organic compounds in litter form up to three independent pools of increasing recalcitrance (i.e. soluble compounds, cellulose and hemi-celluloses, and lignin). During decomposition, soluble compounds are most accessible to microbes and are consumed first, followed by carbohydrates (i.e. cellulose). Lignin can be decomposed only by specialized fungi and is not degraded until accumulated to a certain, critical level when it inhibits the degradation of less recalcitrant compounds [3-6]. The pools are usually quantified by gravimetric determination of cellulose, hemi-celluloses and lignin after sequential extractions with selective solvents. These methods were repeatedly criticized as unspecific for lignin determination [7]. When analyzed with alternative methods (NMR, CuO-oxidation, Pyrolysis-GC/MS), extracted lignin fractions contain many other than the proclaimed substances (e.g. [8]).¹

Recent studies based on more specific methods to determine litter lignin content question the assumed intrinsic recalcitrance of lignin. Experiments using isotope labeling used to calculate mean residence times for lignin in soils and litter/soil mixtures in both laboratory and outdoor incubation reported lignin residence times no longer than that of other carbon compounds or bulk soil organic matter [9,10]. It was also shown that several bacterial taxa are able to degrade lignin².

For leaf litter, lignin depletion during early decomposition and lignin decomposition rates decreasing with litter age were recently reported [11]. Based on this results, the authors proposed a new concept for lignin degradation in which fastest lignin degradation occurs during early litter decomposition when labile carbon availability is high. Lignin decomposition during late decomposition is limited by (dissolved organic) carbon availability. However, they do not elaborate stoichiometric constraints of lignin decomposition³. During radical polymerization, significant amounts of cellulose and protein are incorporated into lignin structures [12]. In isolated lignin fractions from fresh beech litter, N contents twice as high as in bulk litter were found [13]. It was therefore argued that, while yielding little C and energy, lignin decomposition makes occluded cell wall protein accessible to decomposers, and lignin decomposition is therefore not driven by C but by N demand ("Nitrogen mining theory") [14].

Nitrogen fertilization experiments with litter and soils indicated that litter N contents are important

¹[lit CuO], lit[Pyr]

²ref.

³ev. nutrient contents instead of stoichiometry?

controls of lignin degradation: N addition increased mass loss rates in low-lignin litter while slowing down³ decomposition in lignin-rich litter [15]. High nitrogen levels were reported to inhibit lignolytic enzymes in forest soils [16]. Moreover, cellulose triggered a stronger priming effect in fertilized than in unfertilized soils indicating that the mineralization of recalcitrant C is controlled by an interaction of labile C and N availability [17].

Addition of N has very different effects on litter decomposition than varying N levels in litter. This is due to the fact that leaf litter N is stored in protein and lignin structures and not directly available to microorganisms, while fertilizer N is added in the form of readily available inorganic N (NH_4^+ , NO_3^-). N-fertilization experiments can simulate increased N-deposition rates but not the effect of litter N on decomposition processes [18].

In this study we analyze samples from climate-chamber incubated beech litter varying in N and P content with pyrolysis-GC/MS (pyr-GC/MS). The experiment was designed to study the effect of resource stoichiometry on microbial decomposition, exclude decomposing fauna and keep climatic conditions constant. We address the following questions:

- (1) Is lignin decomposition delayed until late decomposition stages or are significant amounts of lignin degraded during early litter decomposition?
- (2) Are lignin and carbohydrate degradation rates controlled⁴ by litter N, P, and soluble C content? Do these controls change during decomposition?
- (3) Do high lignin degradation rates correspond to higher fungal activity?

Results

Mass loss, respiration and extractable organic carbon

Litter mass loss was not significant after 2 weeks and 3 month, significant for 2 litter types after 6 month. After 15 month, litter mass loss was significant for all litter types, and strongly correlated to litter N content ($R=0.794$, $p=***$). Detailed results were reported by [19]. After 15 month, between 5 and 12% of the initial dry mass was lost. This is less than reported in litter decomposition studies on other species, but in a similar range as recently reported for beech litter from an in-situ litterbag-study [20] .

Highest respiration rates were measured after 14 days incubation (150-350 $\mu\text{g CO}_2\text{-C d}^{-1}\text{ g}^{-1}\text{ litter-C}$), dropped to rates between between 75 and 100 $\mu\text{g CO}_2\text{-C d}^{-1}\text{ g}^{-1}\text{ litter-C}$ after 97 days. After 181 and

⁴correlated?

375 days, respiration rates for AK and OS further decreased, while SW and KL show a second maximal respiration after 181 days.

Soluble organic carbon content decreased between the first three harvests (14 to 181 days), to strongly increase after 475 days (0.1 to 0.7 mg C g⁻¹ d.w. were found after 14, 97 and 181 days, and increased to amounts between 1.5 and 4 mg/g after 375 days. After 14 and 97 days, the highest C content was found in SW litter followed by AK (data not shown). DOC content was loosely correlated to litter N content after 14 (R=0.69, p<0.001) and 97 days (R = 0.65, p = <0.01), they were strictly correlated after 181 days (R = 0.85, p=<0.001) and 375 days (R=0.90, p=<0.001).

Microbial biomass abundance and stoichiometry

Microbial biomass contents ranged from 0.5 to 6 mg C, 0.05 to 5.5mg N and 0.05 to 3.5 mg P per g litter (d.w.). In KL and OS biomass buildup reaches a plateau after 3 month, AK and SW show further growth reaching a maximum of microbial C and N contents after 6 month (AK also for P). Microbial C:N ratios measured range between 1:6 and 1:18, C:P ratios between 1:8 and 1:35, and N:P ratios between 1:0.5 and 1:3.5. Microbial C:N ratios (Fig. 3).

Litter microbial biomass is homeostatic during the first 6 month (no or marginally negative correlation between microbial stoichiometry and litter stoichiometry) [19], but not after 15 month, when all three ratios show correlations (R 0.53 - 0.64, all p <0.002, H_{C:N}=2.01, H_{C:P}=1.68, H_{N:P}=2.29). Microbial C:N ratios are tightly constrained after 3 (1:14.5 - 1:18.2) and 6 month (1:6.9 - 1:9.0), but significantly different between the two time points. C:P and N:P ratios are less constrained, with the highest variance between litter from different sites after 3 month incubation (Fig. 3).

Potential enzyme activities

Absolute potential enzyme activities were correlated to litter N, respiration and other other decomposition processes (all R >0.8, p <0.001). For all enzymes and at all time points, SW showed the highest and AK the lowest activity. Cellulase was below detecten limit after 14 days, oxidativte enzymes after 15 month. Cellulases activity is highest after 3 month and decreases between 97 and 181 days. Peroxidase and Peroxidase activities reach their maximum after 181 (fig. ??). After between 6 and 15 month, cellulase activity strongly increased. After 475 days, the activity of oxidative enzymes was below the detection limit [data not shown]

The ratio between the potential activities of cellulases and oxidative enzymes was lowest for AK at all time points. Microbial communities in AK litter invest more energy and nitrogen into degrading lignin and

less into degrading carbohydrates than other litter types. (fig. ??)

Pyrolysis-GC/MS and Lignin content

Litter pyrolysis products and different sites are reported in detail elsewhere (Kohl, in preparation). We found only minor changes during pyrograms during decomposition, differences between sites were small but well preserved during decomposition. The high similarity allowed tracing small changes in lignin and carbohydrate abundance during decomposition.

When measured by pyr-GC/MS, lignin derived compounds make up between 29 and 31 %TIC in the initial litter, with an increase of up to 3 %TIC over the first 3 month. Carbohydrate derived pyrolysis products account for 26 to 29 %TIC in initial litter and decrease by up to 2.6 % during litter decomposition. Carbohydrate depletion and lignin accumulation were correlated ($R = 0.47$, $p < 0.01$) in all samples measured. The initial (pyrolysis-) LCI index (applied to excludes influences of changes in the abundance of other pyrolysis products) ranges between 0.517 and 0.533. During decomposition, it increases by up to 8.7% of the initial value, with SW showing the highest and KL the lowest increase. This increase almost completely occurs over the first 6 month, with insignificant changes in both directions between 6 and 15 month incubation. Figure ??⁵ shows changes in the relative abundance of in pyrolysis products versus incubation time and accumulated respiration. Lignin to carbohydrate ratios in a similar range (increasing from 0.565 to 0.588 over 24 month) were reported for in situ oak litter decomposition by [21] using thermochemolysis.⁶

During the first 6 month of litter decomposition, between one and 6% of the initial lignin pool and between 4 and 17% of the initial carbohydrate pool were degraded. Lignin decomposition was highest in AK and KL litter, while KL and SW decomposed the highest part of their carbohydrate pools. Lignin discrimination (compared to carbohydrates) was highest in SW and lowest in AK litter. In AK litter, lignin molecules were 50% more likely to be decomposed than carbohydrates, while in SW litter carbohydrates were 10 times more likely to be decomposed (fig. 4).

Between 6 and 15 month, no further discrimination occurs, lignin and carbohydrate are degraded at the same rates and their content in pyrograms remains constant (fig. ??).

⁵check fig.

⁶I converted the L:C ratio stated by Snajdr to $L/(L+C)$. This demonstrates a surprising coherence between quite different analytical methods, different peaks analyzed.

Fungi:Bacteria ratio

Metaproteome fungi:bacteria ratios were highest (1:5 - 1:12) after 14 days and decrease during decomposition (1:1.7 - 1:3 after 475 days). Differences between litter from different sites decrease during decomposition. Fungi are most dominant in SW, bacteria in AK. The fungi to bacteria ratio is negatively correlated to lignin decomposition during the first 6 month 11.

Correlations between lignin and carbohydrate decomposition and litter chemistry, microbial community and decomposition processes

Table 2 provides linear regressions found between lignin and carbohydrate degradation, litter chemistry, microbial biomass and decomposition processes after 6 month incubation including data presented by [19] and [22]. We found The lignin to cellulose degradation ratio was correlated to phenoloxidase to cellulase and peroxidase to cellulase enzymatic activity ratios ($R=0.729$ and $R=0.863$, $p=?$). Lignin accumulation and carbohydrate depletion were found to increase with enzymatic activities measured (including lignolytic enzymes) N, and P gross depolymerization rates but not with glucan depolymerization.

While carbohydrate degradation and depletion was correlated litter N content, C:N ratio and C:N imbalances. lignin degradation and accumulation were correlated to litter P, litter C:P and N:P ratios, C:P and N:P imbalances and extractable organic C and PO_4 . High lignin accumulation and carbohydrate depletion were also connected to wide C:N, C:P and N:P ratios.

Discussion

Our experimental approach allowed us to single out the effects of litter quality on the microbial decomposer community as well as decomposition processes, while excluding effects of fauna, climate and the initial microbial community. By exploiting intra-specific differences in beech litter stoichiometry, we were able to minimize differences in the macromolecular composition of initial litter (lignin and cellulose content), while exploring the effect of litter nutrient contents on lignin and carbohydrate decomposition. Therefore, we can attribute different rates of carbohydrate and lignin decomposition to the intrinsic qualities of different litter types.

Contradicting the traditional concepts of litter decomposition, our results demonstrate that relevant amounts of lignin are degraded during the the first 6 months of incubation. Lignin decomposition rates during this early stage depended on litter quality (N & P) and ranged from non-significant amounts decomposed

140 to degradation rates similar to bulk carbon mineralization rates (i.e. no discrimination against lignin). We
141 can therefore confirm that early lignin decomposition rates are by far underestimated, as recently proposed
142 Klotzbücher et. al. [11], based on a complementary analytic approach. Unlike them, we found no decreases in
143 lignin contents and constant or increasing lignin degradation rates during early decomposition. Additionally,
144 we found a change in the controls of lignin discrimination after this initial period. While the preference for
145 carbohydrate to lignin decomposition was controlled by litter chemistry over the first 6 month, all components
146 of litter were degraded at similar rates thereafter.

147 The differences we found in early lignin accumulation do not result from high lignin contents, as is
148 suggested by traditional decomposition models. Differences in initial lignin contents were marginal (below
149 10%), and lignin contents of sites with high initial lignin decomposition rates were not higher than that of
150 sites with low rates. Also, these differences are not caused by a lack of Mn or Fe, as was suggested for
151 late lignin decomposition [3]. While Mn and Fe contents strongly varied between litter collected at different
152 sites, but both Mn and Fe contents are lowest in the litter with the highest lignin decomposition (AK, see
153 tab.1). Low contents of these Elements would explain inhibited, not enhance lignin decomposition. Also,
154 soluble carbon was not limiting lignin decomposition since we found it's highest amounts in litter from the
155 two different sites who show the highest and the lowest lignin degradation.

156 We found strong evidences that C:N:P stoichiometry exerts a key control over the extent of lignin de-
157 composition during initial decomposition. Carbohydrate decomposition was correlated positively to litter N
158 contents and negatively to litter C:N ratios, as were a majority of decomposition processes (mass loss, respi-
159 ration, potential extracellular enzymatic activities). In contrast, relative decomposition rates of Lignin were
160 positively correlated with litter C:P ratios and negatively with a number of P pools analyzed. Correlation
161 was highest when lignin decomposition was compared to resource/consumer C:P ratios.

162 Cultivation studies showed that lignin decomposition in fungi is triggered by nitrogen starvation, and
163 that lignin does not provide sufficient energy to maintain the decomposer's metabolism without the use
164 of another carbon sources⁷. Lignin decomposition was found in wild-type *A. thaliana* litter, but not in
165 a low-cellulose mutant during 12 month incubation in a boreal forest [18]. In the N- and P- co-limited
166 situation during early litter decomposition, in which lignin is degraded to access additional nutrients or to
167 use a C surplus by decomposing an less C efficient but nutrient enriched substrate. However, a stimulation
168 of lignin decomposition by a high P imbalance or a delay of lignin decomposition under high P availability,
169 as indicated by the high correlation to P pools we found, was not reported yet.

⁷citation

For the decomposition of litter lignin and carbohydrates, microbial decomposers rely on the production and excretion of hydrolytic and oxidative extracellular enzymes. While the absolute amounts, in which these enzymes were produced, was controlled by N availability, the ratio in which they were produced corresponded to differences in cellulose and lignin degradation (fig.??). While the mode of this regulation is unknown, we find a corresponding shift in the microbial community composition. Fungi colonize litter faster than bacteria and therefore dominate early litter decomposition, but the bacteria/fungi ratio decrease over the entire incubation period. Talbot et. al. [18] suggested that lignin decomposition might be interpreted as a k-strategy used by microbes to be able to colonize more lignin-rich and nutrient-poor substrates, while r-strategists grow fast based on labile carbon. Low nutrient availability might favor this strategy, as the high P- demands of a fast growing microbial community can not be met. Indeed we found that lignin degradation is higher in litter, where the C:P and N:P are highest, i.e. high P demand limits growth.⁸

While we found different levels of lignin degradation during the first 6 month, lignin contents remained constant in all litter types between 6 and 15 month. This indicates that lignin was decomposed at the same rate as other litter compounds, but differences in lignin contents acquired during the first 6 month remain in place. The controls which lead to differences in the extend of lignin discrimination over the first six month are no longer predominant between 6 and 15 month. Lignin loss in this decomposition stage was positively correlated to litter N content, as was litter mass loss and respiration. The higher fungi/bacteria ratio was higher than during initial decomposition, and differences between sites became (especially after 15 months incubation). Fungi, which grew fast in N and P rich litter become out-competed by a bacteria-dominated lignin degrading community.

The change in decomposition dynamics corresponds to change in soluble carbon. While during the first 3 month, extractable carbon contents were not or to a lesser extend correlated to litter N, soluble carbon is strictly correlated to litter N and actual respiration after 6 and 15 month⁹. KltzbÄijcher et. al. [11] suggest a change in decomposition dynamics after 100 to 200 days of incubation, after which lignin decomposition rates decrease due to lack of labile carbon. They also report a correlation between respiration rates and extractable carbon after this change. The authors interpret this correlation as carbon limitation to respiration, and suggest that lignin decomposition is inhibited under such a limitation. We can confirm the correlation between extractable carbon and respiration after 181 days, but not the inhibition of lignin decomposition. Also, we found that both respiration and the production of soluble carbon are both controlled by litter N content. The process of degrading macromolecular compounds into soluble molecules is conducted by

⁸naja..

⁹stats

200 extracellular enzymes and is therefore N intensive that the mineralization of labile carbon, depolymerization⁹
201 is the point in the decomposition process where a N limitation is most likely to become effective.

202 Another notable change occurs in the homeostasis of the microbial community. While is was strictly
203 homeostatic during the first 6 month, substrate stoichiometry had a minor, but significant influence on mi-
204 crobial stoichiometry after 15 month. Together, those changes indicate that the microbial community is able
205 to compensate for differences in substrate quality (on the expense of community growth and overall decom-
206 position speed) and can select preferred compounds during the first 6 month. However, this compensation
207 is limited and imbalances can not be upheld at the same intensity after the first 200 days.

208 Conclusions

209 Our results further question the concept that lignin decomposition is inhibited until late decomposition.
210 While traditional litter decomposition concepts locate lignin decomposition only during late decomposition,
211 we find substancial amounts of lignin decomposed over the first six month. The extend to which lignin
212 is decomposed was controlled by litter chemistry over the first 6 month. However, we did not find lignin
213 decomposition rates controlled by litter quality between 6 and 15 month. Also, soluble carbon contents were
214 not restrictive to lignin decomposition.

215 While carbohydrate decomposition was stimulated by high N contents, early lignin decomposition rates
216 were highly correlated C:P ratio and resource:consumer imbalances. High lignin contents accumulated during
217 this stages remained in place during later decomposition. Litter with high lignin decomposition rates had
218 higher bacterial abundance.

219 For further studies, this raises the question, to which extend late decomposition is influenced by this
220 early, stoichiometry-controlled accumulation of recalcitrant compounds.

221 Material and methods

222 Litter decomposition experiment

223 A detailed description of our litter decomposition experiment was published in [23]. Briefly, beech litter was
224 collected at four different sites in Austria (Achenkirch (AK), Klausenleopoldsdorf(KL), Ossiach(OS), and
225 Schottenwald(SW); refered to as litter types) in October 2008. Litter was cut to pieces of approximately

0.25cm², homogenized, sterilized twice by gamma¹⁰ radiation (35 kGy, 7 days between irradiations) and¹⁰
inoculated (1.5% w/w) with a mixture of litter and soil to assure that all litter types share the same initial
microbial community. From each type, four samples of litter were taken after inoculation and stored dried
at room temperature. Samples of 60g litter (fresh weight) were incubated at 15 °C and 60% water content
in mesocosms for a duration between 2 weeks to 15 month. For each litter type 5 replicas were removed and
analyzed after 14, 97, 181 and 475 days.

Litter chemistry as analyzed 14 days after incubation is listed in table 1. C:N ratios between 1:41 and
1:58 and C:P ratios between 1:700 and 1:1300 were found, N:P ratios ranged between 1:15 and 1:30. No
significant changes occurred during litter incubation except a slight decrease of the C:N ratio (1:41.8 to
1:37.4) found in the most active litter type (SW) after 15 month. Fe content were more than twice as high
for OS (approx. 450 ppm) than for other litter types (approx. 200 ppm). Litter Mn also was highly variable
between litter types, ranging between 170 and 2130 ppm. Changes of micro-nutrient concentrations during
litter incubation were significant, but in all cases <15% of the initial concentration.

Bulk litter, extractable, and microbial biomass nutrient content

To calculate litter mass loss, litter dry mass content was measurement in 5 g litter (fresh weight) after 48 h
at 80 °C. Dried litter was ball-milled for further chemical analysis. Litter C and N content were determined
using an elemental analyzer (Leco CN2000, Leco Corp., St. Joseph, MI, USA). Litter phosphorus content
was measured with ICP-AES (Vista-Pro, Varian, Darmstadt, Germany) after acid digestion [24]).

To determine soluble C, N, and P contents, 1.8g litter (fresh weight) were extracted with 50 ml 0.5M
K₂SO₄. Samples were shaken on a reciprocal shaker with the extractant for 30 minutes, filtered with ash-free
filters and frozen at -20 °C until analysis. To quantify microbial biomass C, N and P pools, sample were
extracted under the same conditions after chloroform fumigation. Microbial biomass was determined as the
difference between fumigated and non-fumigated extractions [25]. C and N concentration in extracts were
determined with a TOC/TN analyzer (TOC-VCPH and TNM, Shimadzu), Phosphorous was determined
photometrically.¹¹

Substrate to consumer stoichiometric imbalances $X:Y_{inbal}$ were calculated as

$$X : Y_{inbal} = \frac{X : Y_{litter}}{X : Y_{microbial}} \quad (1)$$

¹⁰greek gamma here

¹¹lit!!

where X and Y stand for one of the elements C, N, or P.

Microbial Respiration

Respiration was monitored weekly during the entire incubation in mesocosms removed after 6 month and on the last incubation day for all mesocosms using an infrared gas analyzer (IRGA, EGM4 with SRC1, PPSystems, USA). CO₂ concentration was measured over 70 seconds and increase per second was calculated based on initial dry mass. Measurements of ambient air were performed before and after each measurement to assess possible leaks or base-line drifts IRGA. Accumulated respiration after 6 month was calculated assuming linear transition between measurements, accumulated respiration after 15 month was estimated from respiration rates after 181 and 475 days.

Enzyme activities

Measurements of potential exo-enzyme activities for cellulases, peroxidases and phenoloxidase were described by [22]. Activities were determined with a series of micro-plate assays based on the hydrolysis of 4-methyl- β -D-cellobioside (cellulase) and L-3,4-dihydroxyphenylalanin (oxidative enzymes). Products of enzyme catalyzed reactions were detected photometrically (oxidative enzymes) or fluouometrically (cellulase) [26–28].

0.1 Metaproteome analysis

[..]

Pyrolysis-GC/MS

Pyrolysis-GC/MS was performed with a Pyroprobe 5250 pyrolysis system (CDS Analytical) coupled to a Thermo Trace gas chromatograph and a DSQ II MS detector (both Thermo Scientific) equipped with a carbowax column (Supelcowax 10, Sigma-Aldrich).

Litter analyzed was sampled immediately after inoculation and after 3, 6, and 15 month incubation. 2-300 μ g dried and finely ball-milled litter were heated to 600°C for 10 seconds in helium atmosphere. GC oven temperature was constant at 50 °C for 2 minutes, followed by an increase of 7°C/min to a final temperature of 260 °C, which was held for 15 minutes. The MS detector was set for electron ionization at 70 EV cycling between m/z 20 and 300.¹²

¹²maybe cite other paper for method?

Peaks were assignment was based on NiSt 05 MS library after comparison with reference material mea¹²
sured. 128 peaks were identified and selected for integration due to their high abundance or diagnostic
value, including 28 lignin and 45 carbohydrate derived substances. For each peak between one and four
dominant mass fragments selected for high abundance and specificity were integrated and converted to TIC
peak areas by a multiplication with a MS response coefficient [29, 30]. For principal component analysis,
unconverted areas were used. Peak areas are stated as % of the sum of all integrated peaks of a sample.

Relative peak areas are different from weight%, but allow tracing of accumulation/depletion of substance
classes during decomposition [29].

We use the terms "accumulation" and depletion to refer to changes in litter composition and "degradation"
to refer to the amount of lignin and carbohydrates decomposed.

A lignin to carbohydrates index was calculated to measure the ratio between these two substance classes
without influences of changes in the abundance of other compounds

$$LCI = \frac{Lignin}{Lignin + Carbohydrates} \quad (2)$$

Accounting for carbon loss, we estimate % lignin and cellulose degraded during decomposition according
to equation 3, where TIC_{init} and TIC_{act} stand for initial and actual %TIC area of lignin or cellulose pyrolysis
products, C_{init} for the initial amount of C and R_{acc} for the accumulated CO₂-C respired by a mesocosm.

$$\%_{loss} = 100 \cdot \frac{TIC_{init} - TIC_{act}}{TIC_{init}} \cdot \frac{(1 - R_{acc})}{C_{init}} \quad (3)$$

We provide % of initial lignin and carbohydrate pools decomposed, % decomposed per % litter carbon
mineralized and the ratio between the twodecomposition rates.

Statistical analysis

All statistical analyses were performed with the software and statistical computing environment R using the
package "vegan" [31]. If not mentioned otherwise, results were considered significant when $p < 0.05$. Due to
the frequent of variance inhomogenities Welch anova and paired Welch's t-tests with Bonferroni corrected p
limits were used. Principal component analysis was performed using vegan function "rda" scaling variables.
All correlations refer to Pearson correlations. We calculated correlations between depletion and degradation
rates found in this study with litter chemistry parametres and process data reported by [19] and [22].

References

1. Berg, B & McClaugherty C (2008) Plant Litter. Decomposition, Humus Formation, Carbon Sequestration. Berlin: Springer.
2. Don A, Kalbitz K (2005) Amounts and degradability of dissolved organic carbon from foliar litter at different decomposition stages. *Soil Biology and Biochemistry* 37: 2171–2179.
3. Berg B, Staaf H (1980) Decomposition rate and chemical changes of Scots pine needle litter. II. Influence of chemical composition. *Ecological Bulletins* : 373–390.
4. Coûteaux MM, Bottner P, Berg B (1995) Litter decomposition, climate and litter quality. *Trends in ecology & evolution* (Personal edition) 10: 63–66.
5. Moorhead DL, Sinsabaugh RL (2006) A theoretical model of litter decay and microbial interaction. *Ecological Monographs* 76: 151–174.
6. Adair EC, Parton WJ, Del Grosso SJ, Silver WL, Harmon ME, et al. (2008) Simple three-pool model accurately describes patterns of long-term litter decomposition in diverse climates. *Global Change Biology* : 2636–2660.
7. Hatfield RD, Romualdo SF (2005) Can Lignin Be Accurately Measured? *Crop Science* 45: 832–839.
8. Preston CM, Trofymow JA, Sayer BG, Niu J (1997) ¹³C nuclear magnetic resonance spectroscopy with cross-polarization and magicâ€šangle spinning investigation of the proximate analysis fractions used to assess litter quality in decomposition studies. *Canadian Journal of Botany* 75: 1601–1613.
9. Thevenot M, Dignac MF, Rumpel C (2010) Fate of lignins in soils: A review. *Soil Biology and Biochemistry* 42: 1200–1211.
10. Bol R, Poirier N, Balesdent J, Gleixner G (2009) Molecular turnover time of soil organic matter in particle - size fractions of an arable soil. *Rapid Communications in Mass Spectrometry* 23: 2551–2558.
11. Klotzbücher T, Kaiser K, Guggenberger G, Gatzek C, Kalbitz K (2011) A new conceptual model for the fate of lignin in decomposing plant litter. *America* 92: 1052–1062.

- 326 12. Achyuthan KE, Achyuthan AM, Adams PD, Dirk SM, Harper JC, et al. (2010) Supramolecular¹⁴
327 self-assembled chaos: polyphenolic lignin's barrier to cost-effective lignocellulosic biofuels. *Molecules*
328 (Basel, Switzerland) 15: 8641–88.
- 329 13. Dyckmans J, Flessa H, Brinkmann K, Mai C, Polle A (2002) Carbon and nitrogen dynamics in acid de-
330 tergent fibre lignins of beech (*Fagus sylvatica* L.) during the growth phase. *Plant, Cell & Environment*
331 25: 469–478.
- 332 14. Craine JM, Morrow C, Fierer N (2007) Microbial nitrogen limitation increases decomposition. *Ecology*
333 88: 2105–13.
- 334 15. Knorr M, Frey S, Curtis P (2005) Nitrogen addition and litter decomposition : A meta-analysis.
335 *Ecology* 86: 3252–3257.
- 336 16. Sinsabaugh RL (2010) Phenol oxidase, peroxidase and organic matter dynamics of soil. *Soil Biology*
337 *and Biochemistry* 42: 391–404.
- 338 17. Fontaine S, Henault C, Aamor a, Bdioui N, Bloor J, et al. (2011) Fungi mediate long term sequestration
339 of carbon and nitrogen in soil through their priming effect. *Soil Biology and Biochemistry* 43: 86–96.
- 340 18. Talbot JM, Treseder KK Interactions between lignin, cellulose, and nitrogen drive litter chemistry -
341 decay relationships. *Ecology* (in press) .
- 342 19. Mooshammer M, , the Crew (2011) Marias paper. *ecosystems* .
- 343 20. Kalbitz K, Kaiser K, Bargholz J, Dardenne P (2006) Lignin degradation controls the production of
344 dissolved organic matter in decomposing foliar litter. *European Journal of Soil Science* 57: 504–516.
- 345 21. Snajdr J, Cajthaml T, Valášková V, Merhautová V, Petránková M, et al. (2011) Transformation
346 of *Quercus petraea* litter: successive changes in litter chemistry are reflected in differential enzyme
347 activity and changes in the microbial community composition. *FEMS microbiology ecology* 75: 291–
348 303.
- 349 22. Leitner S, Wanek W, Wild B, Haemmerle I, Kohl L, et al. (2011) Linking resource quality to decom-
350 position processes: Influence iof litter chemistry and stoichiometry on glucan depolymerization during
351 decomposition of beech (*Fagus silvatica* L.) litter. *Soil Biology and Biochemistry* .

- 352 23. Wanek W, Mooshammer M, Blöchl A, Hanreich A, Keiblinger K, et al. (2010) Determination of gross¹⁵
353 rates of amino acid production and immobilization in decomposing leaf litter by a novel N-15 isotope
354 pool dilution technique. *Soil Biology and Biochemistry* 42: 1293–1302.
- 355 24. Henschler G (1988) *Analysen im biologischen Material*. Weinheim: VCH Verlagsgesellschaft mbH.
- 356 25. Schinner (1996) chloroform fumigation method (dummy) .
- 357 26. Marx M (2001) A microplate fluorimetric assay for the study of enzyme diversity in soils. *Soil Biology*
358 *and Biochemistry* 33: 1633–1640.
- 359 27. Sinsabaugh RL (1999) Characterizing soil microbial communities. *Standard Soil Methods for Long-*
360 *Term Ecological Research* 2: 318–348.
- 361 28. Kaiser C, Koranda M, Kitzler B, Fuchslueger L, Schnecker J, et al. (2010) Belowground carbon alloca-
362 tion by trees drives seasonal patterns of extracellular enzyme activities by altering microbial community
363 composition in a beech forest soil. *New Phytologist* 187: 843–858.
- 364 29. Schellekens J, Buurman P, Pontevedra-Pombal X (2009) Selecting parameters for the environmental
365 interpretation of peat molecular chemistry - A pyrolysis-GC/MS study. *Organic Geochemistry* 40:
366 678–691.
- 367 30. Kuder T, Kruege MA (1998) Preservation of biomolecules in sub-fossil plants from raised peat bogs -
368 a potential paleoenvironmental proxy. *Organic Geochemistry* 29: 1355–1368.
- 369 31. Oksanen J, Blanchet FG, Kindt R, Legendre P, O'Hara R, et al. (2011). *vegan: Community Ecology*
370 *Package*. R packge version 1.17-9. URL <http://cran.r-project.org/package=vegan>.

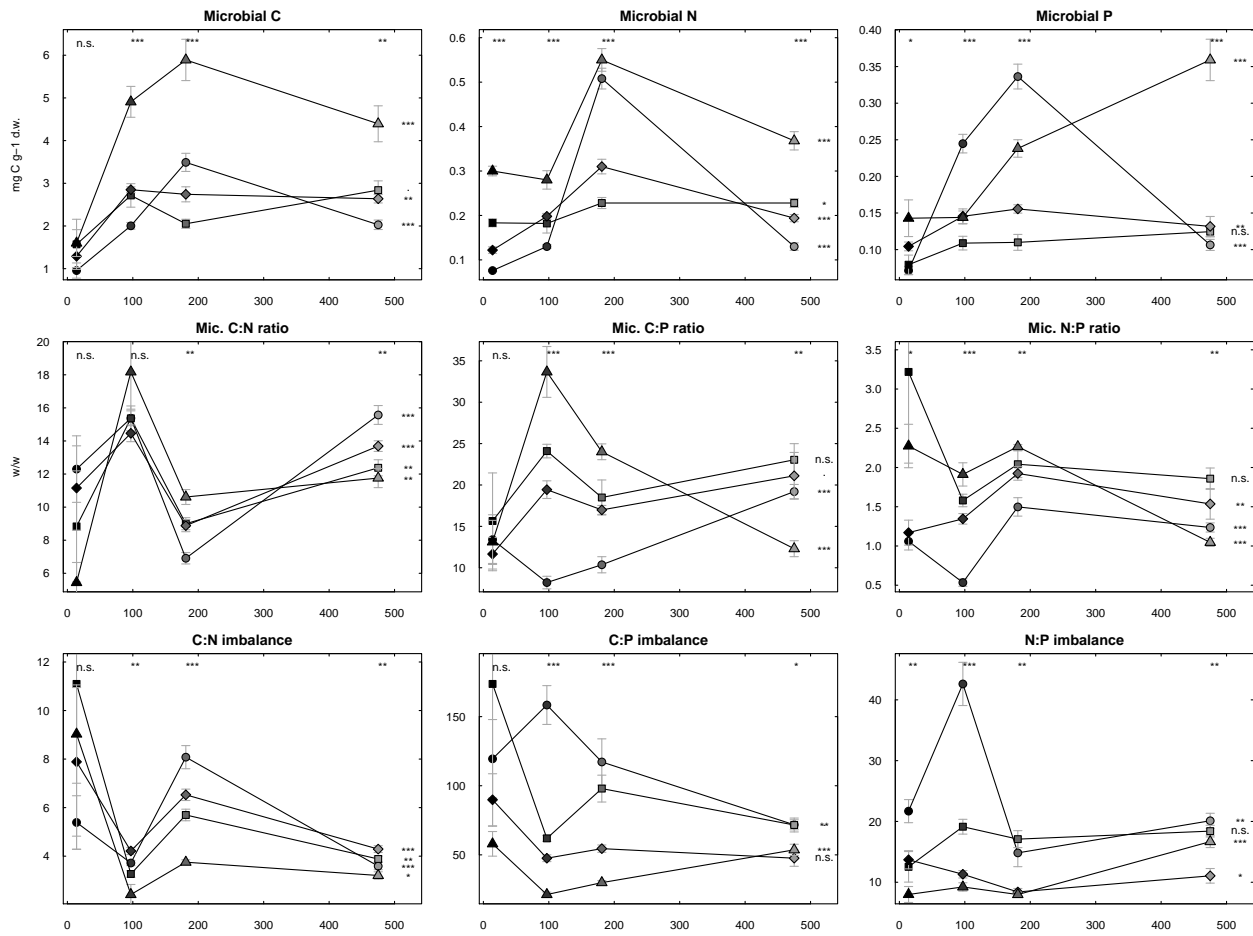


Figure 1. Microbial biomass, microbial stoichiometry and resource:consumer stoichiometric imbalance. Error bars indicate standard errors ($n=5$).

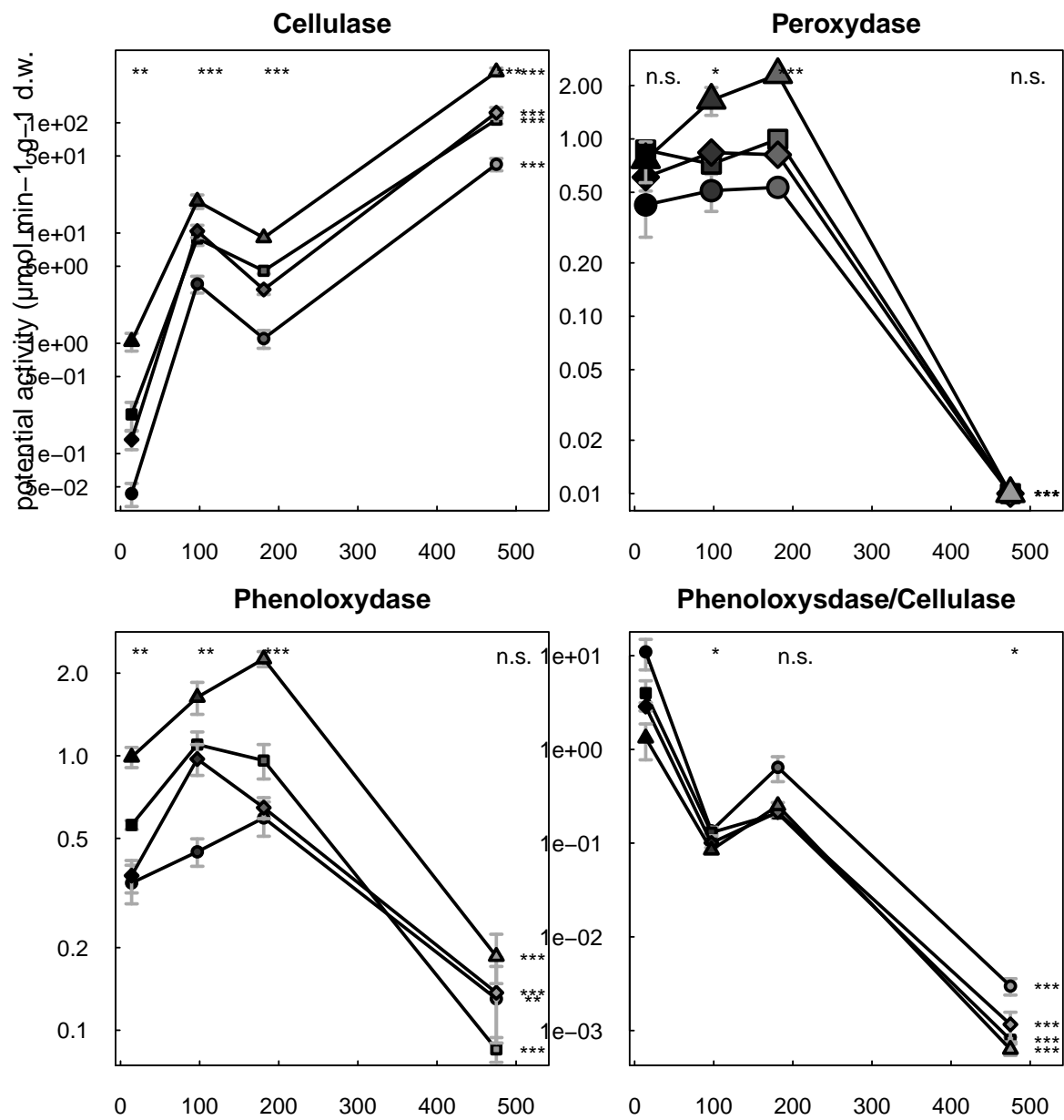


Figure 2. Potential eco-enzyme activities [caption]

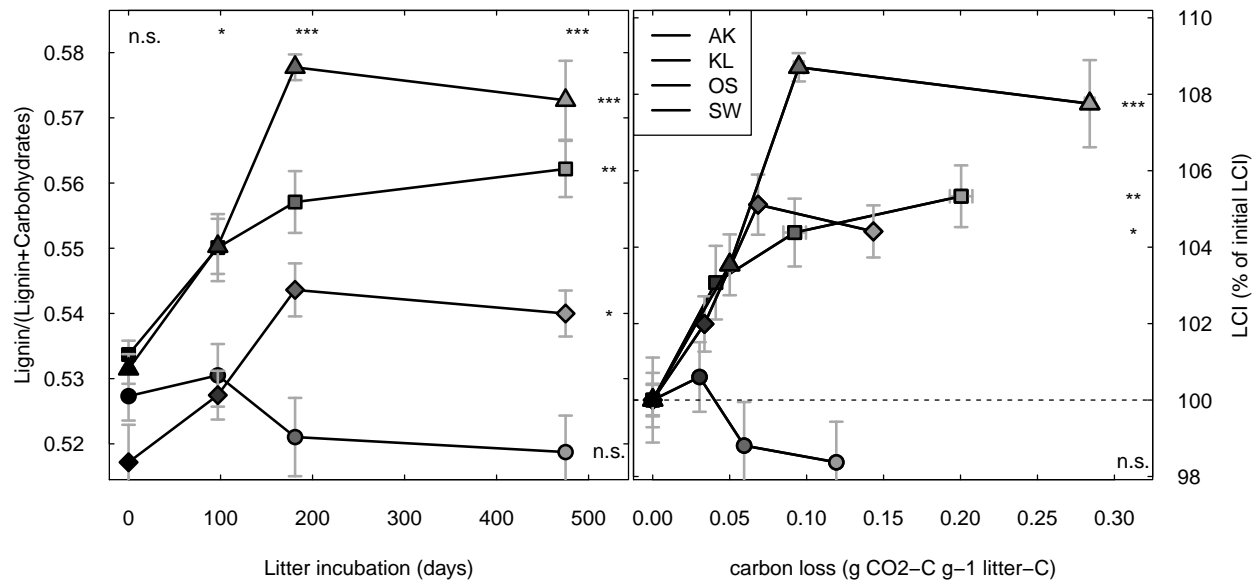


Figure 3. Development of the LCI (lignin/(lignin+carbohydrates)). Errorbars indicate standard errors (n=4-5). The dashed line indicates a constant ratio between lignin and carbohydrates (i.e. no preferential decomposition of carbohydrates).

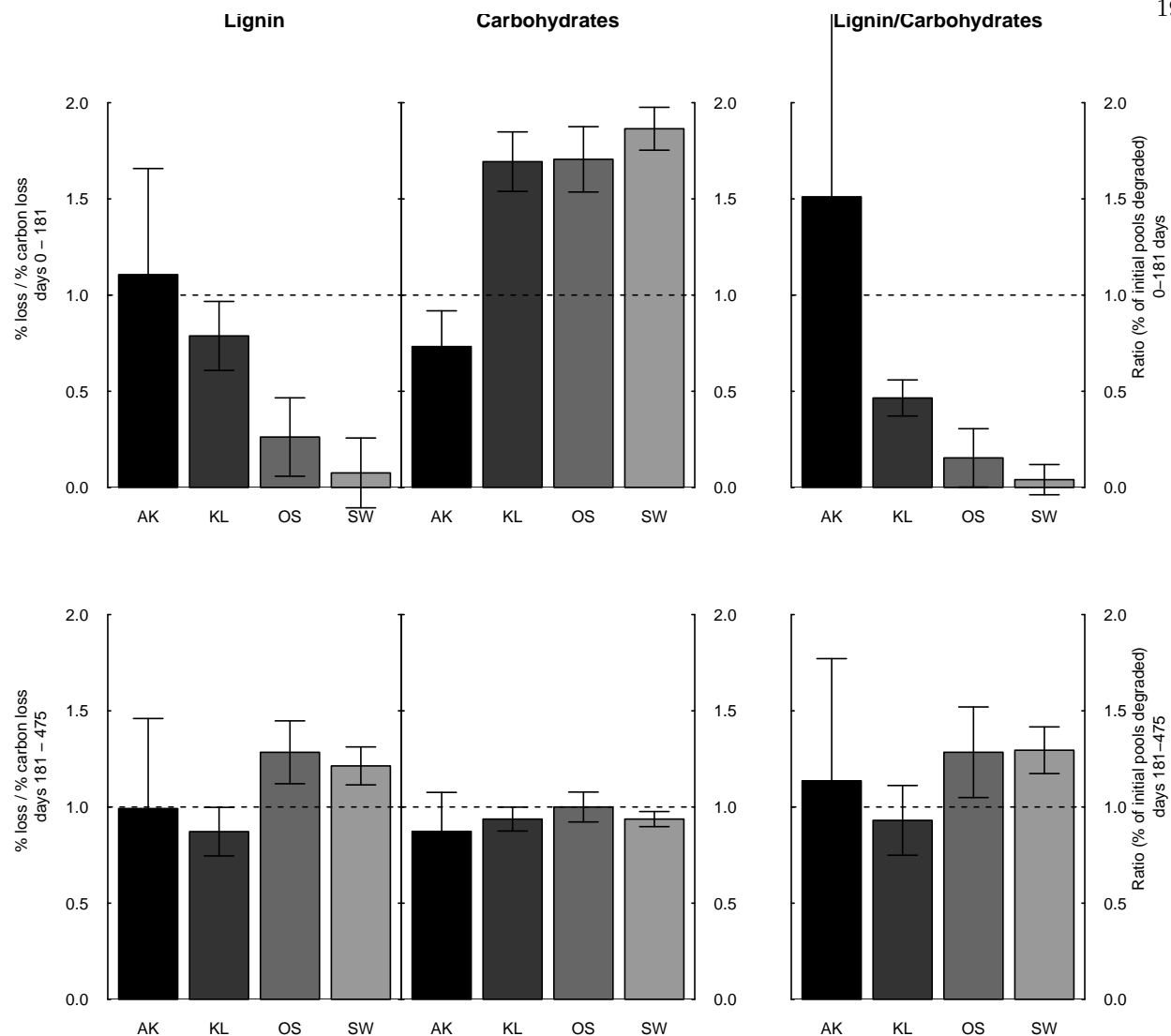


Figure 4. Carbon loss corrected amounts of lignin and carbohydrates degraded. Carbon loss was calculated based on accumulated respiration. Error bars indicate standard errors (n=4-5). The dashed line marks no discrimination between lignin, carbohydrates and bulk carbon loss.

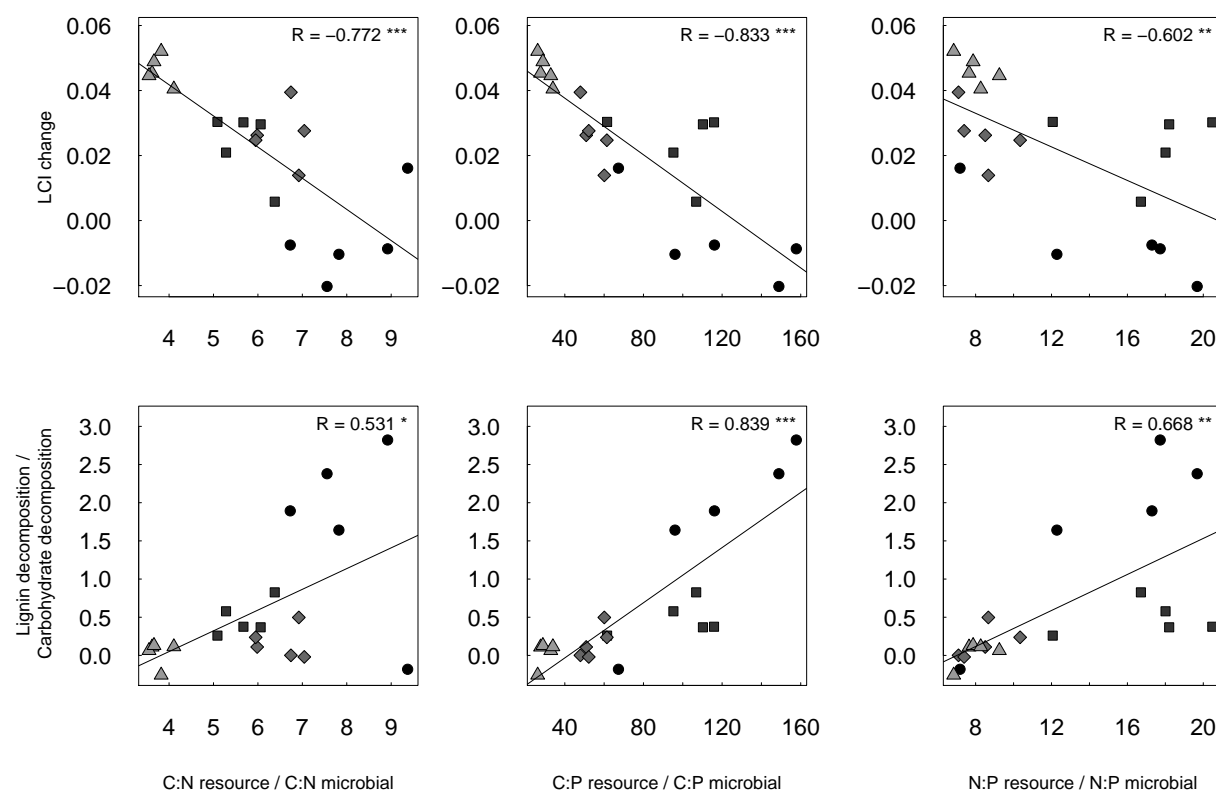


Figure 5. Correlations between Lignin accumulation during the first 6 month of litter incubation and stoichiometric resource:consumer imbalances

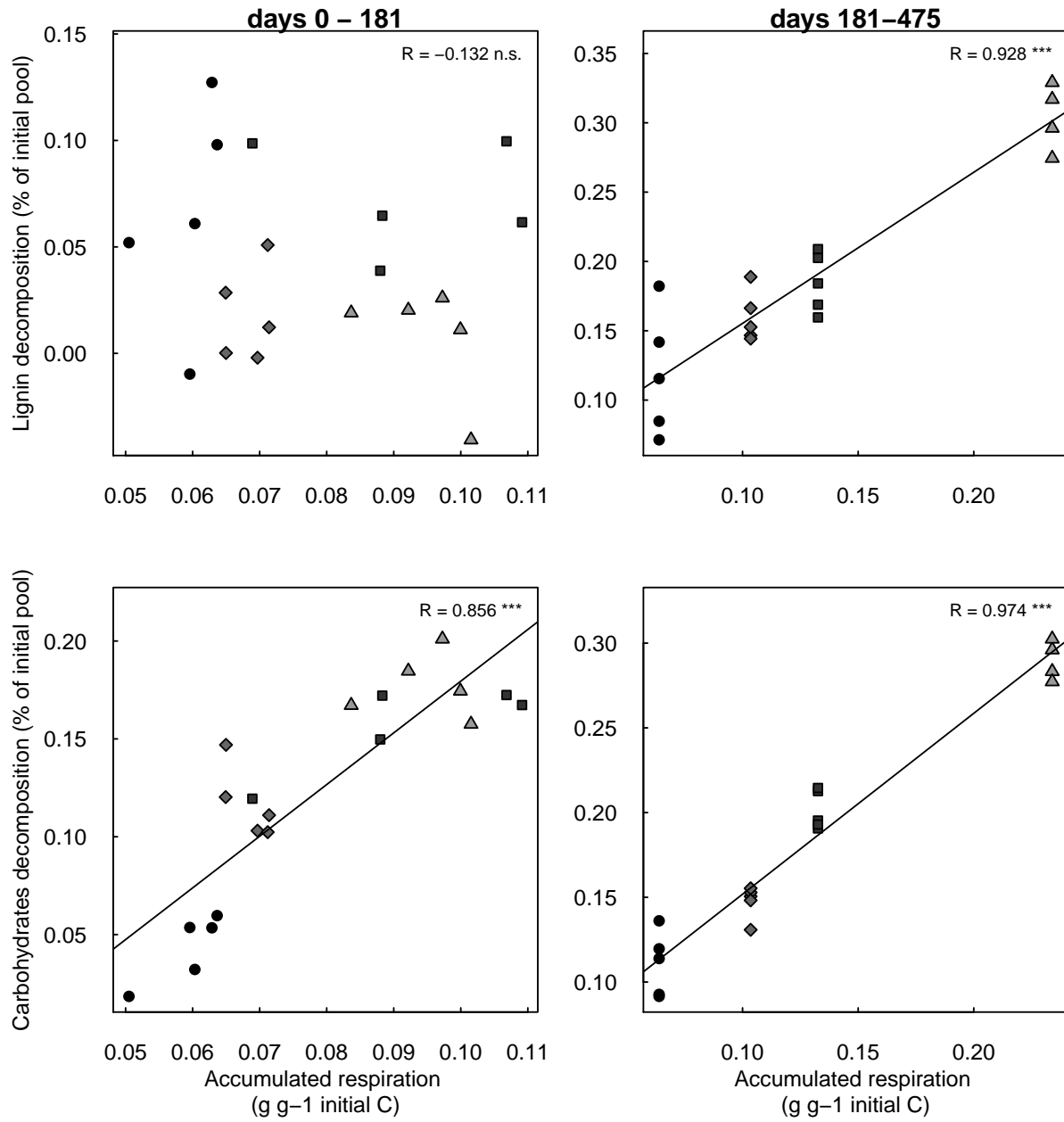


Figure 6. caption

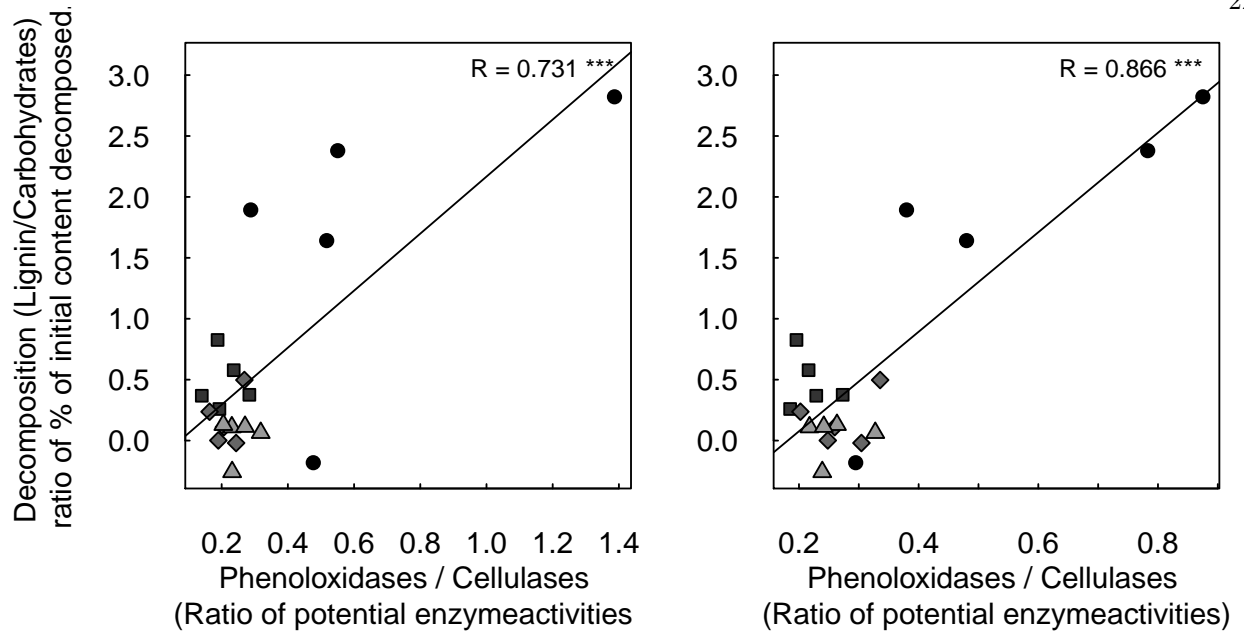


Figure 7. caption

Table 1. Litter stoichiometry and mineral elemental contents measured after 14 days incubation. Standard errors are stated in brackets (n=5). C extr stands for extractable carbon.

	AK	(SE)	KL	(SE)	OS	(SE)	SW	(SE)	p value
C (% d.w.)	50.86	(0.39)	49.41	(0.53)	48.15	(0.39)	48.90	(0.34)	0.002
C extr (mg g ⁻¹)	0.46	(0.03)	0.14	(0.01)	0.21	(0.01)	0.64	(0.03)	0.002
N (% d.w.)	0.878	(0.012)	0.938	(0.012)	0.806	(0.013)	1.172	(0.016)	<0.001
P (% d.w.)	0.040	(0.000)	0.030	(0.000)	0.052	(0.002)	0.070	(0.000)	<0.001
C:N (w/w)	57.86	(0.57)	52.60	(0.49)	59.97	(0.72)	41.78	(0.76)	<0.001
C:P (w/w)	1282	(21)	1548	(25)	905	(15)	699	(9)	<0.001
N:P (w/w)	22.17	(0.47)	29.45	(0.60)	15.10	(0.29)	16.75	(0.39)	<0.001
K (mg g ⁻¹)	0.26	(0.00)	0.54	(0.00)	0.21	(0.00)	0.55	(0.00)	<0.001
Ca (mg g ⁻¹)	1.33	(0.01)	1.26	(0.01)	1.63	(0.01)	1.23	(0.01)	<0.001
Mg (mg g ⁻¹)	0.27	(0.00)	0.14	(0.00)	0.20	(0.00)	0.15	(0.00)	<0.001
Fe (ppm)	210	(2)	208	(4)	453	(12)	192	(4)	<0.001
Mn (ppm)	172	(2)	1430	(10)	776	(9)	2137	(51)	<0.001
Zn (ppm)	30.8	(0.4)	33.0	(0.3)	36.0	(1.0)	42.4	(0.7)	<0.001

Table 2. Correlation (R) between Lignin and Carbohydrate degradation with litter chemistry, microbial community and decomposition processes. Significant ($p < 0.05$) correlations are printed bold. Data taken from [19, 22]. Differences in litter chemistry were calculated between 0 and 181 days, process rates were measured after 181 days.

	L acc	Ch acc	LCI diff	L dec	C dec	L resp	C resp	L/C dec	Per/Cell	Phen/Cell
Massloss	0.291	-0.15	0.245	-0.328	0.106	-0.201	0.125	-0.081	0.048	0.0534
Actual respiration	0.333	-0.723	0.606	-0.0822	0.771	-0.195	0.594	-0.368	-0.268	-0.362
Accumulated Respiration	0.494	-0.704	0.688	-0.132	0.856	-0.332	0.557	-0.525	-0.506	-0.534
Cellulase activity	0.657	-0.76	0.803	-0.431	0.801	-0.497	0.664	-0.589	-0.436	-0.539
Protease activity	0.186	-0.296	0.264	-0.132	0.274	-0.157	0.301	-0.27	-0.26	-0.18
Chitinase activity	0.409	-0.749	0.663	-0.17	0.795	-0.312	0.677	-0.559	-0.49	-0.607
Phosphatase activity	0.549	-0.813	0.776	-0.302	0.851	-0.407	0.702	-0.556	-0.418	-0.522
Phenoxidase activity	0.632	-0.669	0.737	-0.415	0.719	-0.449	0.552	-0.484	-0.305	-0.356
Peroxidase activity	0.599	-0.588	0.677	-0.412	0.639	-0.438	0.47	-0.435	-0.173	-0.302
N mineralization	0.466	-0.664	0.65	-0.167	0.739	-0.299	0.527	-0.387	-0.282	-0.367
Nitrification	0.587	-0.707	0.732	-0.38	0.74	-0.432	0.621	-0.499	-0.369	-0.45
P mineralization	0.665	-0.55	0.684	-0.544	0.596	-0.576	0.414	-0.478	-0.212	-0.255
C litter	-0.545	0.506	-0.578	0.604	-0.368	0.643	-0.618	0.698	0.525	0.581
extractable C	0.609	-0.766	0.782	-0.37	0.814	-0.446	0.658	-0.54	-0.392	-0.484
N litter	0.354	-0.517	0.503	-0.14	0.587	-0.187	0.366	-0.203	-0.119	-0.159
P litter	0.682	-0.222	0.517	-0.747	0.175	-0.68	0.188	-0.491	-0.0728	-0.16
C:N litter	-0.405	0.586	-0.57	0.175	-0.654	0.234	-0.44	0.273	0.195	0.242
C:P litter	-0.636	0.174	-0.453	0.754	-0.0823	0.649	-0.176	0.418	0.049	0.0805
N:P litter	-0.512	-0.0287	-0.264	0.714	0.147	0.577	-0.0202	0.316	-0.0316	-0.0192
C:N mic	0.666	-0.758	0.799	-0.43	0.798	-0.515	0.678	-0.609	-0.584	-0.596
C:P mic	0.692	-0.787	0.834	-0.476	0.814	-0.562	0.726	-0.672	-0.564	-0.648
N:P mic	0.582	-0.729	0.74	-0.415	0.729	-0.508	0.715	-0.67	-0.545	-0.671
C:N imbalance	-0.56	0.81	-0.772	0.288	-0.859	0.391	-0.71	0.531	0.564	0.56
C:P imbalance	-0.817	0.663	-0.833	0.757	-0.61	0.799	-0.668	0.839	0.575	0.67
N:P imbalance	-0.724	0.351	-0.602	0.81	-0.253	0.764	-0.397	0.668	0.301	0.41

Table 3. Correlation (R) between Lignin and Carbohydrate degradation with litter chemistry, microbial community and decomposition processes. Significant ($p < 0.05$) correlations are printed bold. Data taken from [19, 22]. Differences in litter chemistry were calculated between 181 and 475 days, process rates were measured after 475 days.

	L acc	Ch acc	LCI diff	L dec	C dec	L resp	C resp	L/C dec	Per/Cell	Phen/Cell
Massloss	-0.0455	-0.264	0.0665	0.623	0.71	0.505	0.496	-0.118	-0.444	0.403
Actual respiration	-0.374	-0.22	-0.213	0.86	0.83	0.837	0.809	0.0279	-0.403	0.29
Accumulated Respiration	-0.165	-0.29	-0.0113	0.909	0.981	0.753	0.825	-0.119	-0.608	0.486
Cellulase activity	-0.317	-0.307	-0.137	0.861	0.863	0.805	0.91	-0.00551	-0.575	0.414
Protease activity	-0.229	-0.271	-0.086	0.455	0.447	0.434	0.645	-0.0269	-0.456	0.381
Phosphatase activity	0.0425	-0.0182	0.0685	0.334	0.39	0.259	0.487	-0.0904	-0.152	0.0167
Chitinase activity	-0.221	-0.228	-0.0874	0.695	0.7	0.578	0.78	0.0348	-0.58	0.395
Phenoloxidase activity	0.34	-0.436	0.435	-0.196	0.0177	-0.338	-0.121	-0.456	-0.483	0.692
Peroxidase activity	-0.274	0.452	-0.385	0.126	-0.067	0.261	0.0631	0.397	0.546	-0.708
N mineralization	0.175	0.195	0.0757	0.0631	0.111	-0.0805	-0.142	-0.145	0.0624	0.0892
Nitrification	-0.289	0.23	-0.321	0.645	0.573	0.574	0.407	0.164	-0.105	-0.0234
P mineralization	-0.164	0.0616	-0.137	0.475	0.461	0.516	0.402	-0.0877	0.0433	-0.0273
C litter	0.33	0.231	0.176	-0.329	-0.269	-0.358	-0.654	-0.0539	0.501	-0.348
extractable C	-0.205	-0.188	-0.0882	0.884	0.912	0.727	0.774	-0.0383	-0.538	0.409
N litter	-0.17	-0.166	-0.0672	0.854	0.896	0.722	0.644	-0.0751	-0.431	0.349
P litter	-0.4	-0.369	-0.181	0.727	0.701	0.786	0.883	-0.00155	-0.464	0.325
C:N litter	0.124	0.196	0.018	-0.846	-0.912	-0.683	-0.643	0.113	0.49	-0.404
C:P litter	0.508	0.277	0.313	-0.572	-0.463	-0.721	-0.765	-0.144	0.283	-0.162
N:P litter	0.477	0.189	0.325	-0.233	-0.0883	-0.466	-0.5	-0.205	0.048	0.0338
C:N mic	0.216	0.186	0.095	-0.723	-0.745	-0.568	-0.693	0.136	0.57	-0.513
C:P mic	0.395	0.0762	0.312	-0.559	-0.453	-0.599	-0.45	-0.122	0.233	-0.223
N:P mic	0.333	0.0142	0.288	-0.288	-0.169	-0.409	-0.207	-0.174	-0.00191	-0.00931
C:N imbalance	-0.0522	0.084	-0.0756	-0.348	-0.412	-0.311	-0.132	0.00942	0.0273	0.0196
C:P imbalance	0.0913	0.335	-0.0757	-0.114	-0.16	-0.218	-0.499	0.0773	0.16	-0.0317
N:P imbalance	0.0576	0.293	-0.0865	0.0497	0.0088	-0.0352	-0.392	0.128	0.16	-0.0803

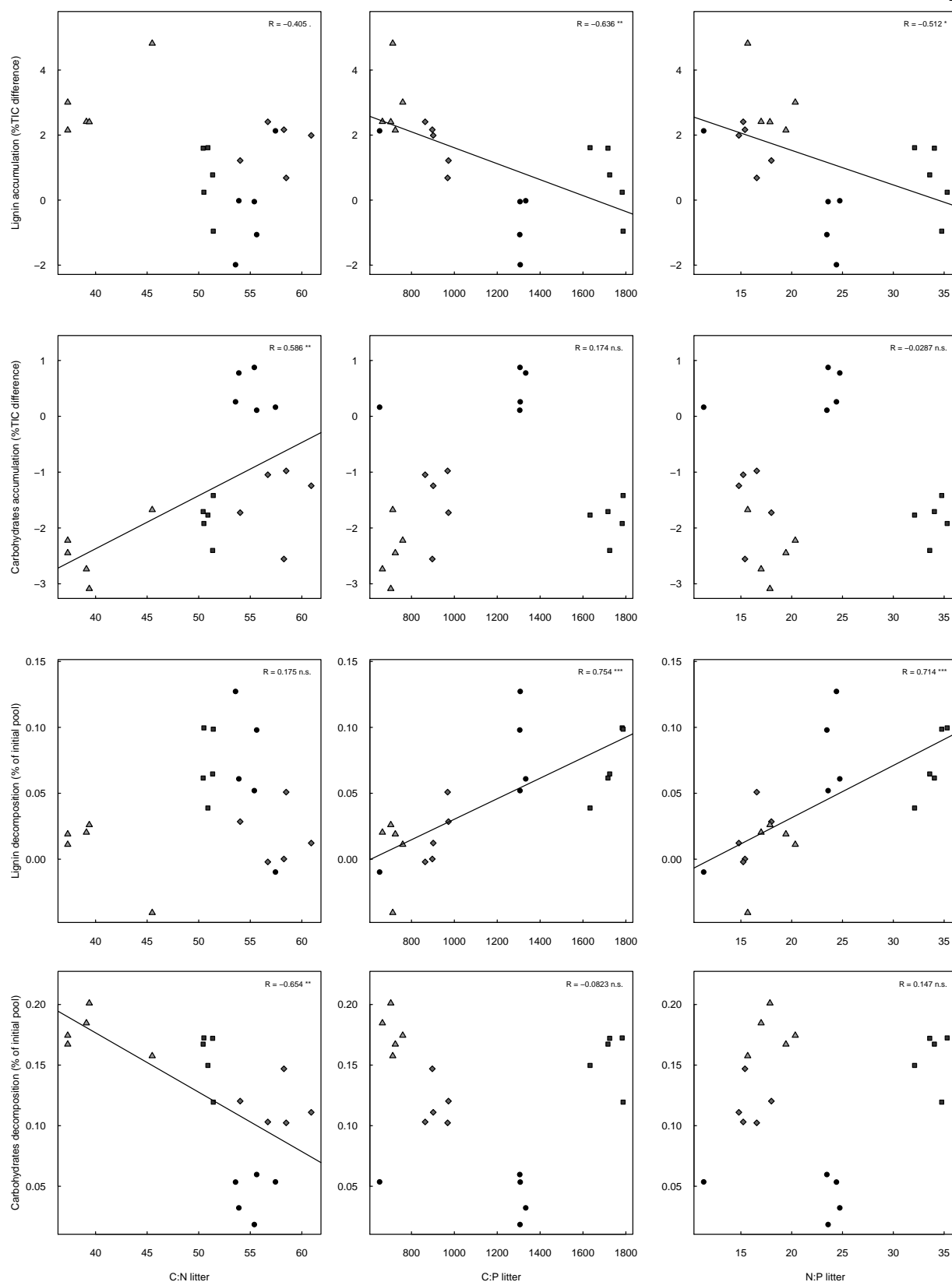


Figure 8. Correlations between Lignin and Carbohydrates accumulation and decomposition during the first 6 months of litter incubation and litter C:N:P ratios

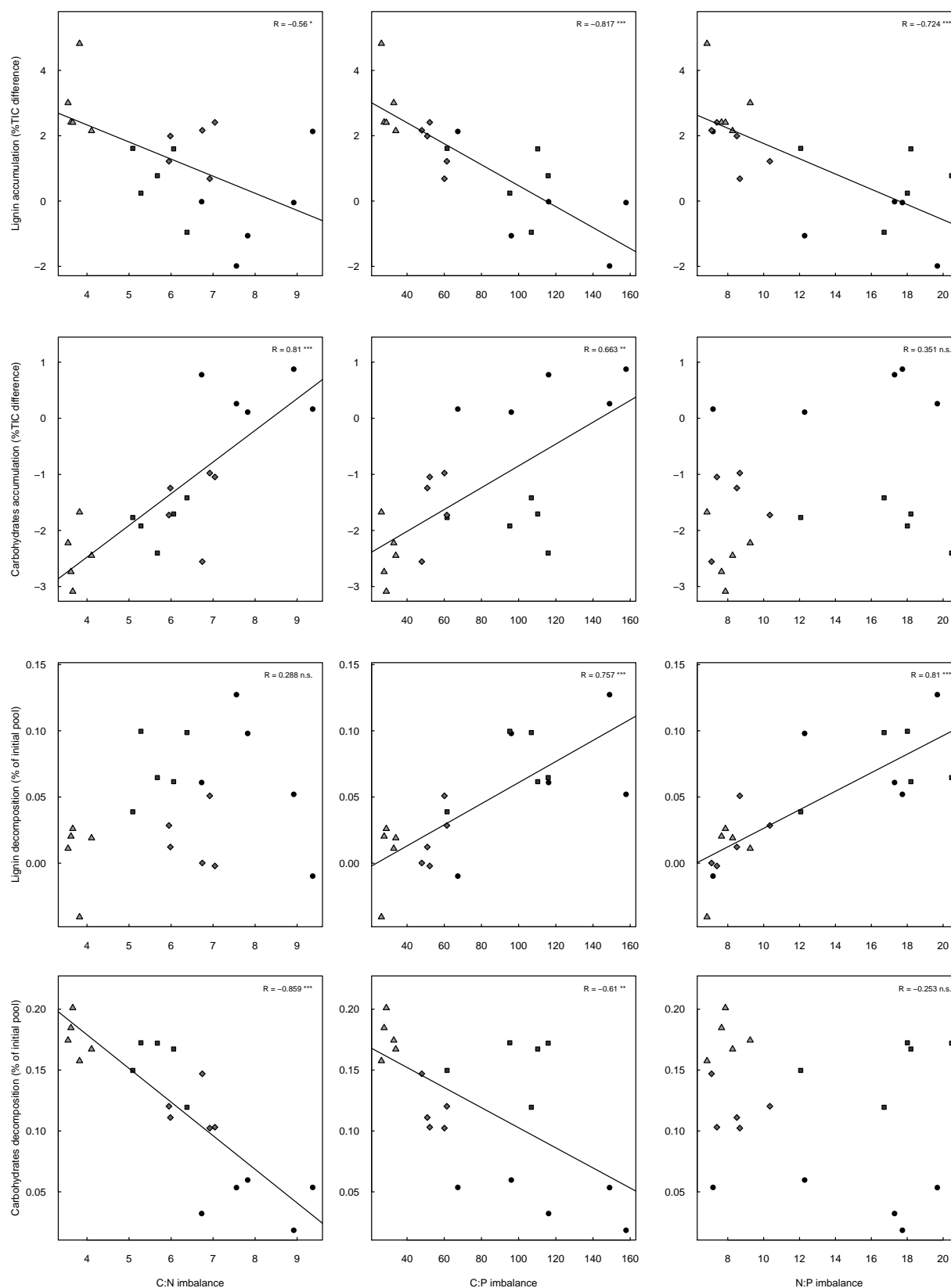


Figure 9. Correlations between Lignin and Carbohydrates accumulation and decomposition during the first 6 month of litter incubation and stoichiometric resource:consumer imbalances

Table 4. Correlations between C25/27/29 alkanes and alkenes, 14:0, 16:0 and 18:0 fatty acids and phytol. Differences between 0 and 181 days.

	alkanacc	alkenacc	faacc	phytolacc	alkandeg	alkendeg	fadeg	phytoldeg	alkanresp	alkenresp	faresp	phytolresp
Massloss	0.17	0.49	-0.0933	0.226	-0.15	-0.348	-0.00722	-0.462	-0.189	-0.227	-0.0146	-0.291
Actual respiration	0.752	0.428	0.701	0.792	-0.357	0.163	-0.656	-0.867	0.119	0.349	-0.767	-0.863
Accumulated Respiration	0.703	0.0679	0.703	0.781	-0.221	0.545	-0.507	-0.714	0.431	0.673	-0.792	-0.87
Cellulase activity	0.665	0.548	0.673	0.894	-0.154	0.112	-0.531	-0.905	0.308	0.332	-0.719	-0.903
Protease activity	0.027	0.0541	0.139	0.304	0.245	0.178	0.09	-0.172	0.344	0.3	-0.0935	-0.21
Phosphatase activity	0.673	0.169	0.756	0.778	-0.229	0.405	-0.574	-0.691	0.231	0.574	-0.75	-0.774
Chitinase activity	0.744	0.519	0.748	0.916	-0.234	0.157	-0.617	-0.931	0.27	0.382	-0.75	-0.93
Peroxidase activity	0.526	0.601	0.574	0.838	-0.157	0.0178	-0.482	-0.911	0.286	0.238	-0.653	-0.87
N mineralization	0.535	0.614	0.478	0.79	-0.0898	-0.0474	-0.625	-0.917	0.282	0.155	-0.551	-0.843
Nitrification	0.724	0.453	0.662	0.828	-0.327	0.132	-0.426	-0.941	0.197	0.298	-0.74	-0.923
P mineralization	0.654	0.487	0.551	0.836	-0.223	0.121	-0.43	-0.912	0.19	0.319	-0.615	-0.876
C litter	-0.0337	0.695	0.368	0.652	-0.052	-0.174	-0.317	-0.762	0.353	0.0259	-0.437	-0.712
extractable C	0.715	-0.192	-0.284	-0.369	-0.244	-0.0544	-0.0979	0.0362	-0.155	0.0115	0.0999	0.0999
N litter	0.688	0.496	0.688	0.906	-0.225	0.17	-0.562	-0.944	0.263	0.374	-0.741	-0.937
P litter	0.0781	0.65	0.076	0.7	-0.376	-0.13	-0.107	-0.429	0.0454	0.0547	-0.645	-0.829
C:N litter	-0.728	-0.636	-0.562	-0.759	0.133	-0.494	-0.107	-0.92	0.197	-0.334	-0.0955	-0.294
C:P litter	0.054	0.0372	0.0372	0.219	0.38	0.0768	0.598	0.943	-0.0723	-0.113	0.689	0.87
N:P litter	0.305	0.24	0.0377	-0.219	-0.216	0.57	-0.0572	0.273	-0.185	0.418	-0.0817	0.135
C:N mic	0.535	-0.561	0.24	0.0377	-0.348	0.584	-0.263	-0.0351	-0.156	0.49	-0.324	-0.139
C:P mic	0.557	0.398	0.62	0.826	0.00763	0.226	-0.345	-0.704	0.439	0.48	-0.609	-0.782
N:P mic	0.486	0.28	0.647	0.864	-0.00386	0.246	-0.38	-0.762	0.458	0.469	-0.624	-0.821
C:N imbalance	-0.695	-0.486	0.613	0.773	0.0068	0.288	-0.358	-0.661	0.396	0.487	-0.554	-0.717
C:P imbalance	-0.289	-0.576	-0.91	-0.91	0.167	-0.193	0.507	0.87	-0.367	-0.42	0.738	0.915
N:P imbalance	0.0124	-0.566	-0.155	-0.684	-0.196	0.0083	0.213	0.566	-0.487	-0.164	0.375	0.874
					-0.31	0.303	-0.00124	0.271	-0.354	0.128	0.0428	0.216

Table 5. Correlations between C25/27/29 alkanes and alkenes, 14:0, 16:0 and 18:0 fatty acids and phytol. Differences between 181 and 475 days.

	alkenacc	alkenacc	faacc	phytolacc	alkandeg	alkendeg	fadeg	phytoldeg	alkanresp	alkenresp	faresp	phytolresp
Massloss	-0.634	-0.289	0.0683	0.709	0.448	0.557	0.344	-0.478	0.152	0.525	-0.535	-0.731
Actual respiration	-0.471	-0.301	0.432	0.676	0.313	0.636	0.0176	-0.253	0.157	0.464	-0.326	-0.584
Accumulated Respiration	-0.795	-0.35	0.324	0.835	0.53	0.764	0.242	-0.402	0.241	0.77	-0.63	-0.829
Cellulase activity	-0.596	-0.244	0.662	0.784	0.356	0.613	-0.0828	-0.288	0.317	0.5	-0.274	-0.547
Protease activity	-0.38	-0.00818	0.659	0.439	0.12	0.196	-0.394	-0.0506	0.264	0.194	-0.011	-0.0894
Phosphatase activity	-0.451	0.0332	0.461	0.323	0.25	0.194	-0.224	-0.000321	0.28	0.255	-0.178	-0.192
Chitinase activity	-0.633	-0.104	0.589	0.611	0.3	0.44	-0.189	-0.186	0.244	0.536	-0.353	-0.4
Peroxidase activity	-0.0256	0.134	-0.345	-0.134	-0.0858	-0.13	0.379	0.137	-0.184	0.0737	-0.0468	-0.0142
N mineralization	0.0627	-0.106	0.343	0.127	0.028	0.0833	-0.401	-0.172	0.119	-0.106	0.0258	0.0264
Nitrification	-0.0379	0.0275	-0.156	0.236	-0.0981	0.0926	0.259	-0.395	-0.397	0.283	-0.569	-0.421
P mineralization	-0.496	-0.451	0.253	0.663	0.417	0.628	0.112	-0.424	0.179	0.535	-0.57	-0.657
C litter	-0.303	-0.466	0.206	0.473	0.333	0.543	0.0642	-0.249	0.235	0.211	-0.182	-0.447
C litter	0.0444	-0.336	-0.825	-0.0325	0.223	0.145	0.685	-0.412	0.0238	0.0415	-0.212	-0.3
extractable C	-0.733	-0.294	0.398	0.876	0.452	0.701	0.14	-0.5	0.174	0.735	-0.666	-0.809
N litter	-0.76	-0.181	0.116	0.836	0.618	0.862	0.415	-0.0722	0.28	0.764	-0.652	-0.909
C:N litter	-0.354	-0.548	0.597	0.575	0.254	0.448	-0.246	0.345	0.17	0.0983	-0.261	0.361
C:P litter	0.0939	0.0836	-0.0977	-0.818	-0.614	-0.843	-0.448	0.494	-0.248	-0.804	0.701	0.922
N:P litter	-0.241	-0.0878	-0.682	-0.422	-0.0492	-0.249	0.48	-0.000689	-0.251	0.0934	-0.341	-0.0041
C:N mic	0.549	0.241	-0.627	-0.0957	0.187	0.0732	0.654	-0.191	-0.202	0.445	-0.659	-0.377
C:P mic	0.206	0.413	-0.218	-0.611	-0.18	-0.547	0.04	0.245	0.0205	-0.687	0.569	0.55
N:P mic	-0.0201	0.326	-0.054	-0.503	-0.241	-0.297	-0.0471	0.26	-0.217	-0.175	0.0327	0.355
C:N imbalance	0.429	0.393	0.363	-0.273	-0.175	-0.297	-0.028	0.156	-0.253	0.13	-0.243	0.124
C:P imbalance	-0.0373	-0.357	-0.544	-0.443	-0.586	-0.515	-0.584	0.403	-0.335	-0.297	0.27	0.611
N:P imbalance	-0.2	-0.486	-0.619	0.137	0.13	0.208	0.535	-0.327	-0.143	0.262	-0.427	-0.324
				0.199	0.37	0.391	0.687	-0.468	0.045	0.324	-0.457	-0.522

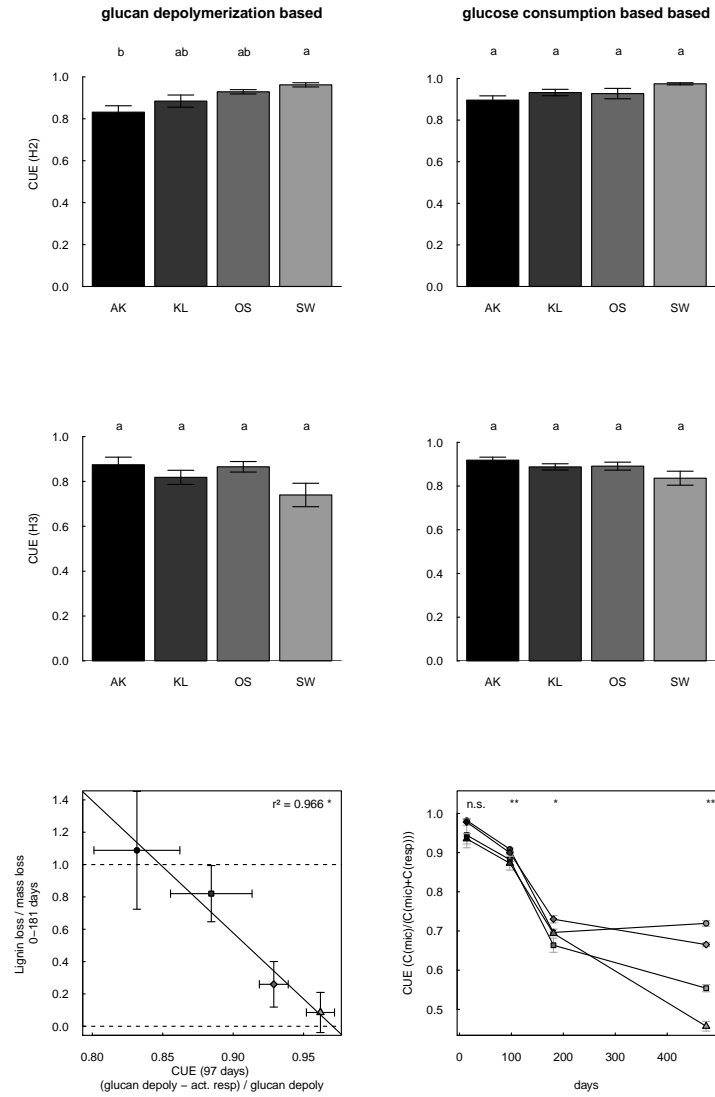


Figure 10. caption

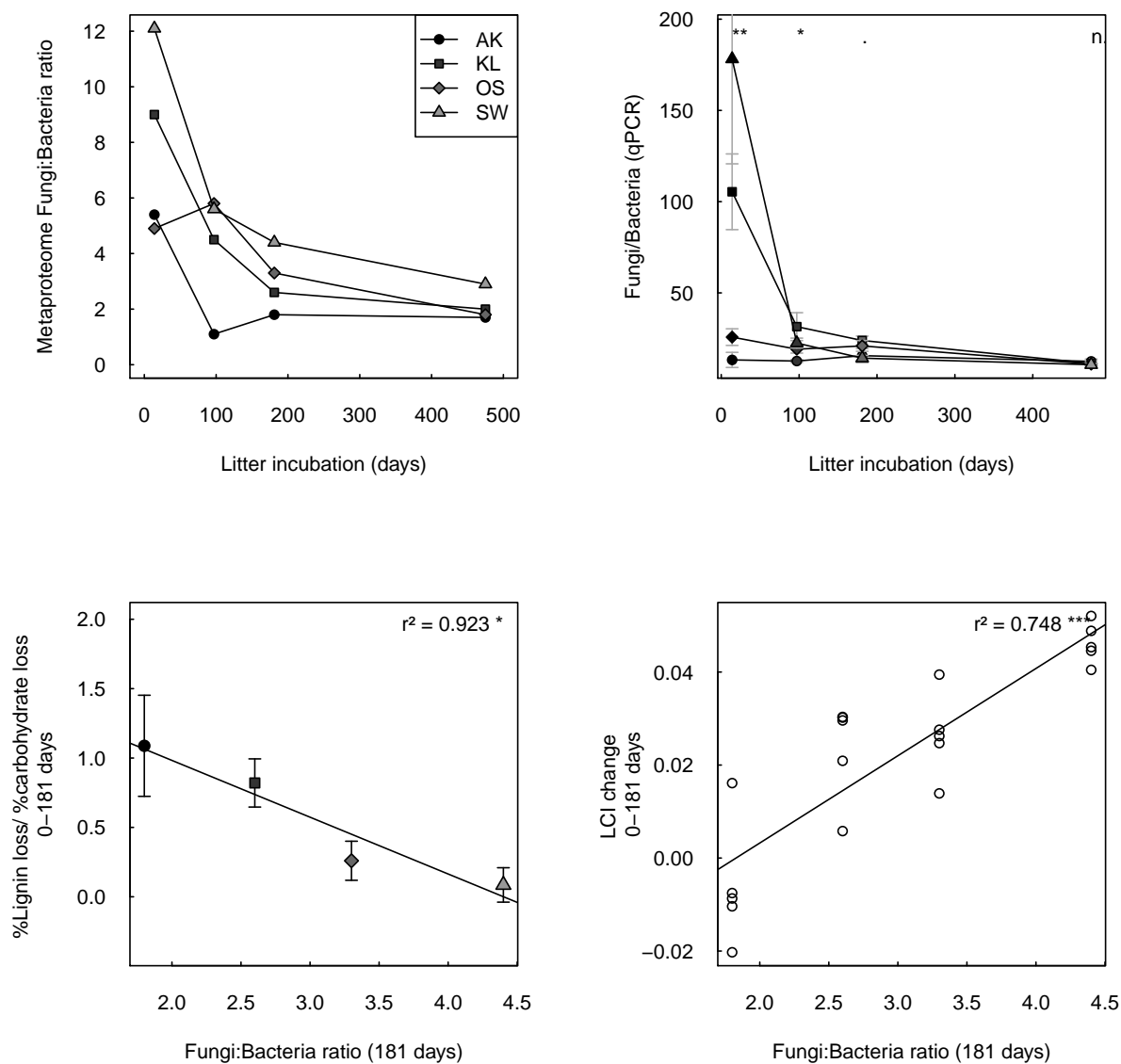


Figure 11. Bold the first sentence. Rest of figure 2 caption. Caption should be left justified, as specified by the options to the caption package.