

Analyzing mixing systems using a new generation of Bayesian tracer mixing models

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The ongoing evolution of tracer mixing models has resulted in a confusing array of software tools that differ in terms of data inputs, model assumptions, and associated analytic products. Here we introduce MixSIAR, an inclusive, rich, and flexible Bayesian tracer (e.g. stable isotope) mixing model framework implemented as an open-source R package. Using MixSIAR as a foundation, we provide guidance for the implementation of mixing model analyses. We begin by outlining the practical differences between mixture data error structure formulations and relate these error structures to common mixing model study designs in ecology. Because Bayesian mixing models afford the option to specify informative priors on source proportion contributions, we outline methods for establishing prior distributions and discuss the influence of prior specification on model outputs. We also discuss the options available for source data inputs (raw data versus summary statistics) and provide guidance for combining sources. We then describe a key advantage of MixSIAR over previous mixing model software—the ability to include fixed and random effects as covariates explaining variability in mixture proportions and calculate relative support for multiple models via information criteria. We present a case study of *Alligator mississippiensis* diet partitioning to demonstrate the power of this approach. Finally, we conclude with a discussion of limitations to mixing model applications. Through MixSIAR, we have consolidated the disparate array of mixing model tools into a single platform, diversified the set of available parameterizations, and provided developers a platform upon which to continue improving mixing model analyses in the future.

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22 Abstract

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47 Bayesian statistics

48 Introduction

49 Mixing models, or models used to estimate the contribution of different sources to a mixture, are
50 widely used in the natural sciences. Typically, these models require tracer data that characterize
51 the chemical or physical traits of both the sources and mixture – these traits are assumed to
52 predictably transfer from sources to mixtures through a mixing process. In ecology, the majority
53 of mixing model applications use stable isotope signatures as tracers in efforts to assess the
54 contribution of prey (sources) to a consumer (mixture) diet, although other applications include
55 pollutant sourcing, plant water use, carbon sources in soils, etc. (Phillips et al. 2014). However,
56 in recent years, researchers have leveraged other tracers, such as fatty acid signatures to assess
57 predator-prey relationships (Neubauer and Jensen 2015, Galloway et al. 2015). Regardless of the
58 tracers or mixing system considered, all mixing model applications are rooted in the same
59 fundamental mixing equation:

$$60 \quad Y_j = \sum_k p_k \mu_{jk}^s,$$

61 where the mixture tracer value, Y_j , for each of j tracers is equal to the sum of the k source tracer
62 means, μ_{jk}^s , multiplied by their proportional contribution to the mixture, p_k . This basic
63 formulation assumes that (1) all sources contributing to the mixture are known and quantified,
64 (2) tracers are conserved through the mixing process, (3) source mixture and tracer values are
65 fixed (known and invariant), (4) the p_k terms sum to unity, and (5) source tracer values differ.
66 Given a mixing system with multiple tracers such that the number of sources is less than or equal
67 to the number of tracers + 1, the p_k terms in the set of Y_j equations can be solved for analytically,
68 given the unity constraint (Schwarcz 1991, Phillips 2001). In most natural mixing systems an
69 analytical solution to the set of mixing equations is not possible without simplifying the mixing
70 system or the data. In other words, in order to establish a solvable set of equations, researchers

have traditionally reduced the number of sources through aggregation. Additionally, because the analytic solution requires that the source and mixture signatures to be fixed (invariant), researchers used the mean variable tracer data and ignored uncertainty.

More recently, researchers have turned to more sophisticated mixing model formulations that provide probabilistic solutions to the mixing system that are not limited by the ratio of sources to tracers (i.e. under-determined systems), and that integrate the observed variability in source and mixture tracer signatures. The first of such models, IsoSource (Phillips and Gregg 2003), provided distributions of feasible solutions to the mixing system based on a “tolerance” term; IsoSource iteratively identified unique solutions for the p_k terms that resulted in Y_j solutions falling near the true value of the mixture (typically defined by the mean of mixture data), where “near” was arbitrarily defined by the model user through the specification of tolerance. Subsequently, Moore and Semmens (2008) introduced a Bayesian mixing model formulation, MixSIR, that established a formal likelihood framework for estimating source contributions while accounting for variability in the source and mixture tracer data. An updated version of this modeling tool with a slightly different error parameterization, SIAR, continues to be broadly applied in the ecological sciences and beyond (Parnell et al. 2010). Since 2008, Bayesian mixing models have rapidly evolved to account for hierarchical structure (Semmens et al. 2009), uncertainty in source data mean and variance terms (Ward et al. 2010), covariance in tracer values (Hopkins and Ferguson 2012) and covariates within the mixing system (Francis et al. 2011). In short, Bayesian mixing models have developed into a flexible linear modeling framework, summarized by Parnell et al. (2013).

In light of these analytic innovations, we have created an open-source R software package, MixSIAR, that unifies the existing set of mixing model parameterizations into a

customizable tool that can meet the needs of most environmental scientists studying mixing systems. MixSIAR can be run as a graphical user interface (GUI) or script, depending on the user's familiarity with R. Either version can be used to load data files and specify model options; then MixSIAR writes a custom JAGS (Just Another Gibbs Sampler, Plummer 2003) model file, runs the model in JAGS, and produces diagnostics, posterior plots, and summary statistics. As with any sophisticated modeling tool, researchers should take care in establishing situation-specific applications of the tool based on the data in hand and the mixing system targeted for inference. At present, however, guidance on the parameterization and implementation of Bayesian mixing model analyses is lacking in the literature. As a consequence, many researchers are unsure of the correct application and interpretation of existing mixing model tools such as MixSIR (Semmens and Moore 2008) and SIAR (Parnell et al. 2010).

In this paper we introduce and provide guidance on using MixSIAR for the application of Bayesian mixing models. Given early debate in the literature regarding appropriate error parameterizations (Jackson et al. 2009, Semmens, Moore, & Ward 2009), we begin by clarifying the underlying error structures for MixSIAR and provide recommendations for the use of specific error formulations based on the methods of data collection. The integration of prior information is a key advantage of Bayesian approaches to model fitting. However, since Moore and Semmens (2008), few studies have implemented methods for generating prior distributions in mixing model formulations. We therefore provide a set of basic approaches to establishing prior distributions for the proportional contribution terms, and demonstrate how to incorporate informative priors in MixSIAR. Next, we provide guidance for source assignment in the mixing system (e.g. lumping or splitting source groupings). Arguably, the primary advantage of MixSIAR over previous mixing model software is the ability to incorporate covariate data to

explain variability in the mixture proportions via fixed and random effects. As such, we provide guidance on applying covariate data within mixing models and illustrate this using MixSIAR in a case study on American alligator (*Alligator mississippiensis*) diet partitioning. Finally, we discuss limitations of mixing models and issues with under-determined systems. The complete set of MixSIAR equations with additional explanation is attached as Article S1, and the MixSIAR code is available at <https://github.com/brianstock/MixSIAR>.

Understanding MixSIAR error structures for mixture data

In most published results stemming from Bayesian mixing models, little if any detail is reported regarding the assumed error structure of the mixture data. However, assumptions about variability, and the specific parameterizations used to characterize this variability, in the mixing system have been the focus of most of the innovations in mixing model tools in recent years (Parnell et al. 2010, 2013, Ward et al. 2010, Hopkins and Ferguson 2012, Stock and Semmens 2016b). The specific error formulation matters both because it relates to the assumptions regarding how the process of mixing occurs (e.g. how consumers feed on prey populations), and because the estimates of proportional source contributions can be affected (Stock and Semmens 2016b). In this section, we discuss the suite of error parameterizations available in MixSIAR that account for variability in the tracer values of the mixture. Note that this section deals only with “residual” variability in the mixture tracer data after accounting for variability resulting from fixed or random effects (see case study and Article S1 for how these effects interact with the error terms). For simplicity in the equations below, we ignore discrimination factors, concentration dependence and tracer covariance in our notation. Note, however, that MixSIAR accounts for each of these components, should an analyst specify a model appropriate to do so (see Article S1 for complete MixSIAR equations).

Researchers sometimes use “integrated sampling”—pooling many subsamples into one sample that is then analyzed—to characterize the source means while keeping processing time and costs low. Thus, the most basic formulation for mixing models implemented in MixSIAR assumes that the k source means for the j tracers, μ_{jk}^s , are fixed and invariant (but might be observed imperfectly; Fig. 1A). Under this assumption the mixture value for each tracer will also be an invariant weighted (by source proportions, p_k) combination of the source means. Observations of these means, however, are imperfect and thus the i mixture data for tracer j , Y_{ij} , are assumed to follow the distribution,

$$Y_{ij} \sim N\left(\sum_k p_k \mu_{jk}^s, \sigma_j^2\right), \#(1)$$

where σ_j^2 represents residual error variance, or the variability in observations associated with the mixture data points for the j^{th} tracer. This error distribution is appropriate in situations where, for instance, each source and/or mixture data point was generated through the combination of many samples from the source population. For instance, if an analyst were interested in assessing the relative contribution of dissolved organic carbon (DOC) and particulate organic matter (POM) to a filter feeder’s diet, this model formulation would be appropriate since each source isotope signature comes from an integrated sample of the source isotopic signatures (as opposed to isotopic signatures of individual particles).

In contrast, for many mixing models applied to ecological systems, the tracers of individual source items (prey, e.g. individual deer) and mixtures (consumers; e.g. individual wolves) are analyzed separately, and the variability across source tracers is assumed to translate into consumer signature variability—in other words, different wolves eat different deer, and their tracer signatures should differ accordingly. Since the introduction of Bayesian stable isotope

162 mixing models, nearly all published formulations have assumed that each mixture data point i for
 163 tracer j is derived from a normal distribution with the same mean as in Eq. 1, and, importantly, a
 164 variance similarly generated from a weighted combination of source variances, $\omega_{jk}^{s^2}$:

$$165 \quad Y_{ij} \sim N\left(\sum_k p_k \mu_{jk}^s, \sum_k p_k^2 \omega_{jk}^{s^2}\right). \#(2)$$

166 In situations where there is covariance in tracers (typical of stable isotope studies), Eq. 2 can be
 167 modified to account for a weighted average of source covariance matrices (Stock and Semmens
 168 2016b).

169 MixSIAR uses this model formulation only in the special case where the analyst provides
 170 a single mixture value for each of the j tracers considered. This formulation must be used in this
 171 special case because it is not possible to estimate a variance term, σ_j^2 , from a single data point. In
 172 diet partitioning applications, the above formulation assumes that, for a given tracer j , a
 173 consumer i takes a single IID sample from each of k sources and combines these samples in
 174 accordance with the proportional estimates p_k . In other words, each wolf eats exactly one deer,
 175 and thus incorporates the tracer value of only that deer. Because the prey-specific isotopic
 176 signatures will be different for each consumer due to sampling error, the weighted combination
 177 of sampled source isotopic signatures will also vary. We refer to this model of mixture variance
 178 as “process error” because it is derived from an assumption about the mixing process.

179 Recently, Stock and Semmens (2016b) modified the above formulation to include an
 180 additional multiplicative error term for each tracer considered, ξ_j , such that

$$181 \quad Y_{ij} \sim N\left(\sum_k p_k \mu_{jk}^s, \sum_k p_k^2 \omega_{jk}^{s^2} \times \xi_j\right). \#(3)$$

182 The intent of the ξ_j term is to both add biological realism in the mixing equation, and to provide
 183 flexibility on the likelihood error structure such that mixing data not conforming to the mixing

process assumed in the previous likelihood formulation can still be fit appropriately. As before, Eq. 3 can be modified to account for a weighted average of source covariance matrices (see Article S1). This model formulation is appropriate for most ecological mixing model applications (e.g. diet partitioning), with the exception of integrated sampling studies or studies with a single consumer sample, as outlined above. Stock and Semmens (2016b) showed that, compared to existing models (MixSIR, SIAR), Eq. 3 had lower error in p_k point estimates and narrower 95% CI when the true mixture variance is low ($\xi_j < 1$).

When ξ_j is less than 1, the variance in consumer tracer signatures shrinks, presumably due to the biological process of sampling each prey source multiple times from a distribution of tracer values (Fig. 1C). As the number of IID samples a consumer takes from a source population increases, the tracer signature transferred from the source to the consumer will conform more and more closely to the mean source signature. In other words, each wolf eats more than one deer, and thus each wolf incorporates a sample mean of deer tracer values, which becomes closer to the deer tracer mean as the number of deer sampled increases. Thus, ξ_j indicates the amount of food a consumer integrates within a time frame determined by tissue turn-over; the methods for estimating this consumption rate are outlined in Stock and Semmens (2016b). As the value of ξ_j approaches zero, an analyst can assume that the consumers are essentially “feeding at the mean” of the source populations.

Estimates of ξ_j much greater than one indicate that the variability in transfer of tracer signatures from source to consumer is swamping the reduction in consumer variability expected when consumers integrate over multiple samples from prey populations. This could be due to factors such as isotopic routing (Bearhop et al. 2002), or important consumer population structure being absent from the model (e.g. most variability in wolf stable isotope values is

207 explained by random effects of region and pack in Semmens et al. 2009). Alternatively, the
 208 mixing model could be missing a source or underestimating the source variances. In any case,
 209 values of ξ_j much greater than one are an indication that the model mixing system is not
 210 conforming to one or more of the basic assumptions of the mixing model, namely that tracers are
 211 not being consistently conserved through the mixing process, all mixtures are identical and have
 212 the same source proportions (often not the case in biological systems), and/or that the model is
 213 missing at least one source pool.

214 **Constructing informative Bayesian priors**

215 **Priors for compositional data**

216 The analysis of compositional data is not unique to mixing models. Examples of statistical
 217 models for compositional data are widespread in ecology (Jackson 1997), fisheries (Thorson
 218 2014), as well as non-biological fields (Aitchison 1986). The most common choice of prior on
 219 the estimated vector of proportions \mathbf{p} is the Dirichlet distribution; MixSIAR uses this distribution
 220 for estimates of source proportions. The Dirichlet is often referred to as a multivariate extension
 221 of the Beta distribution, and it is important to understand the Beta before transitioning to the
 222 Dirichlet. The Beta distribution has a convenient property that when both its shape parameters
 223 are 1, it is equivalent to a uniform distribution. In other words, if a model tries to estimate the
 224 relative contribution of a 2-component mixture, $p_1 \sim \text{Beta}(1,1)$ is equivalent to p_1
 225 $\sim \text{Uniform}(0,1)$. Because the vector of proportions is constrained so that $\sum_{i=1}^n p_i = 1$, p_2 can
 226 be treated as the derived parameter $p_2 = 1 - p_1$, and thus doesn't require a prior. For the
 227 parameter of interest p_1 , one way to describe the prior distribution is that the $\text{Beta}(1,1)$ prior is

uniform, and an equivalent description is that all possible combinations of p_1 and p_2 are equally likely *a priori*.

For mixtures with more than 2 components, MixSIAR uses the Dirichlet distribution to specify a prior on \mathbf{p} . The hyperparameter of the Dirichlet distribution is a vector $\boldsymbol{\alpha}$, whose length is the same as \mathbf{p} . Like the Beta distribution, the only constraint on the elements of $\boldsymbol{\alpha}$ is that they be positive (they may be discrete or continuous, and the elements of $\boldsymbol{\alpha}$ don't have to be equal). A common choice of hyperparameters for a 3-component mixture is $\boldsymbol{\alpha} = (1,1,1)$, which we refer to as the “uninformative”/generalist prior because 1) while every possible set of proportions has equal probability, the marginal prior likelihood of a given p_k differs across values of p_k , and 2) its mean is $(\frac{1}{3}, \frac{1}{3}, \frac{1}{3})$, corresponding to the assumption of a generalist diet (McCarthy 2007). The first point is illustrated by Figure 2, which shows that the marginal distributions of the proportions are not uniform, instead favoring small values. Part of this confusion can be resolved by examining the joint pairwise distributions of \mathbf{p} (Fig. 2), which illustrates that using a hyperparameter of $\boldsymbol{\alpha} = (1,1,1)$ implies that all combinations of (p_1, p_2, p_3) are equally likely. Thus, this prior is noninformative on the simplex, but is non-uniform with respect to individual p_k parameters. Other choices of a prior may be Jeffreys' prior, $\boldsymbol{\alpha} = (\frac{1}{2}, \frac{1}{2}, \frac{1}{2})$, or the more recently used logit-normal and extensions (Parnell et al. 2013). By default, MixSIAR uses the “uninformative”/generalist prior, where all α_k are set to 1.

246

247 Constructing an informative prior

One of the benefits to conducting mixture models in a Bayesian framework is that information from other data sources can be included via informative prior distributions (Moore

and Semmens 2008, Franco-Trecu et al. 2013). Once an informative prior for the proportional contribution of sources is established, MixSIAR can accept the prior as an input during the model specification process (for details and example, see Stock and Semmens 2016a). For diet studies, these other sources may include fecal or stomach content samples, data from other studies, or expert knowledge. As a simplified example from Moore and Semmens (2008), suppose we wish to construct an informative prior for a 3-source mixing model of 10 rainbow trout diet using sampled stomach contents (30 eggs, 8 fish, 25 invertebrates). The sum of the Dirichlet hyperparameters roughly correspond to prior sample size, so one approach would be to construct a prior with $\alpha = (30, 8, 25)$, where each α_k corresponds to the source k sample size from the stomach contents. A downside of this prior is that a sample size of 63 represents a very informative prior, with much of the parameter space given very little weight (Fig. 3). Keeping the relative contributions the same, the α_k can be rescaled to have the same mean, but different variance. One starting point is to scale the prior to have a total weight equal to the number of sources, K , which is the same weight as the “uninformative”/generalist prior:

$$\alpha_k = \frac{Kn_k}{\sum n_k} \quad (4)$$

The prior constructed from Eq. 4 is shown in Figure 3. Though this rescaling process of Dirichlet hyperparameters may seem arbitrary, it provides a powerful tool for incorporating additional information.

Importantly, choosing a prior—including the “uninformative”/generalist prior—requires explicit consideration of how much weight the prior should have in any analysis. An additional consideration is the turnover time for different types of data. In our example of rainbow trout diet, stomach contents might represent a daily snapshot of prey consumption, whereas stable

isotope and fatty acid signatures likely change on a much longer time scale (e.g. weeks to months). In such cases, we would want to downweight the prior's significance, since a prior constructed from daily information should only be loosely informative on the mixture proportions averaged over weeks to months. Exactly how much to downweight is unclear. However, this challenge lies within the broader issue of how to weight multiple data types, and we follow Francis' (2011) recommendation that users conduct a sensitivity analysis—fit the model using different informative priors (as well as the “uninformative”/generalist prior) and determine how sensitive the primary result is to the choice of prior (as in deVries et al. 2016).

Priors for other model parameters

In addition to specifying prior distributions on proportional contributions, MixSIAR requires priors on variance parameters (Parnell et al. 2013). Because mixing models ultimately are a class of linear models, MixSIAR uses the same weakly informative prior distributions for variances that are widely used in other fields (Gelman et al. 2014). For specific prior formulations associated with residual error, multiplicative error, and variance associated with random effects, we refer the reader to the full set of MixSIAR equations (Article S1). Note, however, that because MixSIAR generates a model file in the JAGS language (Just Another Gibbs Sampler; Plummer, 2003) during each model run, the analyst can access the complete set of prior specifications associated with the model run. Moreover, the model file can be modified and used in a separate model run outside of MixSIAR, should the analyst care to evaluate the sensitivity of model outputs to changes in prior specification.

In some cases, an analyst may wish to incorporate discrete or continuous covariates to explain differences between individual signatures (detailed in the next section; Francis et al.,

2011; Ogle et al., 2014). Ecological examples of these types of covariates may include environmental variables (habitat, temperature) or variables specific to individuals (sex, age, size). Like simple linear regression, including covariates introduces new parameters to be estimated (intercept, slope), but because MixSIAR includes these covariates in transformed compositional space (isometric log ratio, ILR; Aitchison 1986), their prior specification is not straightforward. MixSIAR uses diffuse normal priors in transform space, which are sufficient to establish priors that yield parameter estimates that are essentially informed only by the data (Gelman et al. 2014, McElreath 2016). Analysts who wish to create informative priors in transform space should proceed with caution, because they can have counterintuitive effects when transformed back to proportion space.

Incorporating source data into mixing models

Early versions of Bayesian mixing models treated the estimates of source-specific tracer means and variance as fixed (user specified), and thus only used raw mixture data in calculating the likelihood of source proportions (Moore and Semmens 2008, Parnell et al. 2010). In so doing, the uncertainty in the estimates of source means and variances, typically derived from source isotope data, was ignored. However, Ward et al. (2010) introduced what they termed a “fully Bayesian” model that accounts for estimation uncertainty in source-specific tracer means and variances, and thus treats both the mixture and source information as data within the model framework. More recently Hopkins and Ferguson (2012) incorporated multivariate normality into estimates of source-specific covariance matrices. This multivariate normality accounts for the fact that tracer values often co-vary, particularly for stable isotope studies.

In MixSIAR, the analyst has two options for inputting source data, (1) providing source tracer value summary statistics (mean, variance and sample size), or (2) providing raw tracer data

for each source. In both cases, MixSIAR fits a fully Bayesian model by estimating the “true” source means and variances for each tracer (Ward et al. 2010, Parnell et al. 2013). However, in the case where summary statistics are provided, the tracers are assumed to be independent, since it is not possible to generate estimates of tracer covariance from the summary statistics. Where raw source data are provided, MixSIAR assumes multivariate normality and estimates the variance covariance matrix associated with the tracers for each source (Hopkins and Ferguson 2012). In the event that an analyst wishes to specify fixed (known) means and variances for a particular source-by-tracer combination, we recommend that they provide MixSIAR with summary statistics (mean and variance) with an arbitrarily large sample size (~10,000). In essence, this approach fixes the estimated source means and variances at the values provided.

Combining sources

No amount of increased sophistication in mixing model methods can overcome the problem of poorly specified mixing systems. If, for instance, an analyst specifies a mixing model with >7 sources contributing to a mixture based on 2 tracers (e.g., $\delta^{13}\text{C}$, $\delta^{15}\text{N}$), it is unlikely the model products will be precise or interpretable. The source data (number of sources and their sample sizes, means, and variances relative to mixture data) have a large influence on the estimated proportions. As such, including several largely extraneous sources with few mixture data points will divert p_k from the truly important sources (as $\sum p_k = 1$). We note, however, that there are ways to constrain the p_k such that models converge—two methods are discussed in sections to follow: informative priors, and including covariates on the p_k as fixed or random effects. Nonetheless, MixSIAR can estimate posterior distributions of source proportions regardless of how under-determined the mixing system is (e.g., many more sources than tracers). This under-

determination, together with the variability in source and mixture isotopic values, often results in quite diffuse probability distributions for many of the proportional contribution estimates, limiting the interpretability of the results (Phillips et al. 2014). Reducing the number of sources by combining several of them together may improve model inference. Either *a priori* or *a posteriori* aggregation (Phillips et al. 2005) may be used with MixSIAR (see “combine_sources” function for *a posteriori* aggregation).

The *a priori* approach typically involves pre-processing the input data by conducting frequentist tests for equality of means of sources and subsequently combining sources without significant differences before running a mixing model (e.g. Ben-David et al. 1997). If tracer data are approximately normally distributed, a Hotelling’s T^2 test can be used to evaluate whether sources are not different from each other, given multivariate data (multiple tracers; Welch and Parsons 1993). If tracers are not normally distributed, a K nearest-neighbor randomization test can be used to assess differences in sources (Rosing et al. 1998). Note that in both cases, a Bonferroni-type correction is typically used when multiple source comparisons are made. Regardless of the test used, if sources appear similar, their data can be aggregated. In general, mixing model outputs will be more interpretable if the sources combined have a logical connection (e.g. same trophic guild, taxon, etc.) so that the aggregated source has some biological meaning, rather than a disparate set of unrelated sources that happen to have similar isotopic values, although this is not an absolute requirement.

Using a frequentist approach (e.g. Hotelling’s T^2 test) to decide on whether sources should be combined *a priori* often presents problems. The amount of data available for each source directly influences the equality of the means tests; the power to reject a null hypothesis of no mean difference between tracer values of sources is thus related to the amount of tracer data,

and is not exclusively a function of the mixing system. Furthermore, in situations when many tracers are available (e.g. fatty acids as tracers; Galloway et al. 2015) there is a high probability that at least some equality of mean tests will fail (reject the null hypothesis) even if the sources are, in reality, identical. Finally, when only the mean, variance and sample size of each source is available (rather than raw data), there is no easy test for equality of the means and methods for aggregating sources are not apparent.

Using the *a posteriori* procedure, the analyst uses the full set of sources to generate posterior distributions of proportional source contributions, and then post-processes the results to combine several sources together. For each posterior draw, the new combined source proportion is simply the sum of the proportions of the original sources. Thus, we obtain a posterior distribution for the new combined source proportion that accounts for correlation between the original source proportions. This new posterior distribution may then be analyzed as before. Importantly, this approach does not require that the isotopic signatures of the combined sources are similar; thus, an analyst is free to combine sources based on functional similarities in the mixing system, regardless of isotopic similarity.

Like the *a priori* approach, combining posteriors from multiple sources as a means of source aggregations is not without issues. One caveat is that each additional source included in the mixing model increases the number of parameters to be estimated, particularly when the model includes random effects. We could easily imagine that a mixing model with 20 sources and random effects may take days to run successfully, and may not converge at all. In models with many more sources than tracers, the source proportions are more likely to be confounded, and therefore highly negatively correlated. In such cases, it is less likely the model will converge. Another potential issue with the *a posteriori* approach is that the combination of multiple diet

proportions estimated with an “uninformative”/generalist Dirichlet prior (each source given equal prior weight) also combines the prior weight for these sources. For instance, given an “uninformative”/generalist Dirichlet prior, the act of aggregating two source posteriors results in a combined source posterior that reflect an aggregated prior with twice the weight of the remaining non-aggregated source priors. As such, the more sources that are combined into an aggregate source group *a posteriori*, the more strongly the prior will be weighted towards increased proportional contributions of this aggregate source to the consumer diet. MixSIAR alerts users to this issue by plotting the aggregated prior when combining sources using the “combine_sources” function (Fig. 4). This is not an issue, however, when the same number of sources are combined into new groupings (e.g. deVries et al. 2016, where six sources were combined into two groups of three). In general, combining sources *a posteriori* can lead to lower variance in diet proportion estimates, particularly when the posteriors for the combined sources show strong negative correlation (Semmens et al. 2013). For most situations, we prefer the *a posteriori* approach to source aggregation, provided the analyst is aware of the cautions mentioned above.

These *a priori* and *a posteriori* approaches to combining sources may be accomplished by simple pre-processing of MixSIAR input data sets and post-processing of MixSIAR output using the “combine_sources” function, respectively. Ward et al. (2011) outlined a Bayesian approach that probabilistically identifies source groupings and generates weighted posterior probabilities associated with various combinations of sources. However, their method requires specialized MCMC sampling, and is computationally impractical for complicated mixing systems. We expect that future refinements to the modeling approach they outlined will yield

409 more robust techniques for treating source combinations as parameters to be estimated, rather
410 than fixed *a priori* or *a posteriori*.

411 **Incorporating covariates via fixed and random effects**

412 In many cases, covariate data (also called explanatory or independent variables) are available for
413 incorporation into a Bayesian mixing model to answer important questions about the mixture
414 (Francis et al. 2011, Ogle et al. 2014). Neglecting to include covariates that are relevant to the
415 mixture proportions can lead to pseudoreplication, since the model assumes all mixtures are
416 independent and identically distributed (Hurlbert 1984). Some examples from diet partitioning
417 applications include:

- 418 1. Consumers (mixtures) are of different sexes and an analyst has interest in whether the
419 dietary proportions differ between sexes (fixed categorical effect).
- 420 2. An analyst has additional numerical measures on the consumers such as weight, length,
421 etc., and would like to see whether the dietary proportions are affected by this value
422 (fixed continuous effect).
- 423 3. An analyst has samples of consumers and/or sources in different regions. It is likely that
424 the consumers' dietary proportions are similar between regions so it makes sense that the
425 estimates should 'borrow strength' between the groups (random effect).

426 In each case it is possible to run a traditional mixing model separately for each sex, region, time
427 point, etc. However, this process can be time-consuming and will often lead to inefficient
428 inference with greater uncertainties in the dietary proportions for three main reasons. First, there
429 will be no direct estimate of the effect size between groups. Second, additional residual error
430 terms will be fit (a residual error term for each level of the fixed/random effect, instead of one
431 error term shared across levels). Third, there is no way to "borrow strength" between groups,

since each set of dietary proportions must be estimated independently. The solution lies in adding the extra information as covariates through the dietary proportions in the mixing model directly. To illustrate the application of fixed and random effects using MixSIAR software we describe a case study on *Alligator mississippiensis* diet partitioning, which executes multiple model formulations and evaluates their relative support using information criteria (Nifong et al. 2015; for data and R code see Data S1).

A common question is how to choose whether to use fixed or random effects. We recognize that the terms “fixed” and “random” effects are unclear (Gelman 2005), and in Gelman’s “constant” versus “varying” terminology, both fixed and random effects in MixSIAR are varying (different for each factor level). Nonetheless, Gelman (2005) recommends using random effects (as defined in MixSIAR, Article S1) when possible, since borrowing strength between groups is a desirable property, and always allows for the model to choose large random effect standard deviations that will yield nearly equivalent estimates to those resulting from fixed effects structure when the analyst has reasonably informative isotopic data. The random effects model draws offsets from a shared distribution, which is appropriate if the factor levels are related, as they often are in biological systems. The random effects model also allows inference on the relative importance of multiple factors through variance partitioning. For example, Semmens et al. (2009) showed that for British Columbia wolves, $\gamma_{Region}^2 > \gamma_{Pack}^2 > \gamma_{Individual}^2$, which means that Region explained most variance in wolf diet, followed by Pack and Individual. However, when the number of groups is small (<5) there can be difficulties in estimating the random effect standard deviations, and fixed effects should always be used when a factor has only two groups.

454 Technical details

455 For covariates to be included, the model must allow for dietary proportions to be specified per
456 individual, e.g. the mixture likelihood must be of a form similar to:

$$457 Y_{ij} \sim N\left(\sum_k p_{ik} \mu_{jk}^s \sum_k p_{ik}^2 \omega_{jk}^2 * \xi_j\right).$$

458 Where p_{ik} is now the dietary proportion for source k on individual i .

459 Regardless of which fixed or random effects are used, MixSIAR establishes a base set of
460 diet proportions \mathbf{p} using a Dirichlet prior that can be modified with prior information. Once
461 specified, these proportions are isometric log-ratio (ILR) transformed into ILR-space parameters,
462 $\boldsymbol{\beta}_0$ (Parnell et al. 2013). This transformation maps a composition in the k -part Aitchison-simplex
463 isometrically to a $k-1$ dimensional Euclidean vector. Each of the $\boldsymbol{\beta}_0$ transformed components are
464 normally distributed and independent of each other and can thus be broached by standard
465 multivariate analysis methods.

466 Once transformed, these $\boldsymbol{\beta}_0$ terms can be modified through the incorporation of
467 covariates, and then subsequently back-transformed into individual-specific vectors of diet
468 proportions \mathbf{p}_i . For instance, for a simple fixed effects structure like that described in example 1
469 above, we have:

$$470 \mathbf{p}_i = \text{inverse.ILR}(\boldsymbol{\beta}_0 + \boldsymbol{\beta}_1 \text{Sex}_i).$$

471 The parameters in the vector $\boldsymbol{\beta}_1$ cumulatively represent the change in dietary proportions for the
472 difference between female and male. In this instance, the categorical fixed effect Sex_i is coded so
473 that male=1 and female=0 (or vice versa).

474 If the covariate is continuous, as in example 2, the structure changes only very slightly:

$$475 \mathbf{p}_i = \text{inverse.ILR}(\boldsymbol{\beta}_0 + \boldsymbol{\beta}_1 \text{Weight}_i).$$

Now the parameters in the vector β_1 represent the change in dietary proportions according to a unit increase in the weight of the consumer.

Covariates are included as random effects in a similar manner. For example 3 given above, we might have:

$$p_i = \text{inverse.ILR}(\beta_0 + \beta_{\text{Region}(i)})$$

where each of the $k-1$ random effect terms in the vector $\beta_{\text{Region}(i)}$, have an extra constraint:

$\beta_{\text{Region}(i),k} \sim N(0, \gamma_{\text{Region}}^2)$. This constraint allows the model to borrow strength between groups.

If γ_{Region}^2 is small, then the groups are similar and the dietary proportions will not change much

between regions. If γ_{Region}^2 is large however, the regions will be very different and this will be

reflected in the dietary proportions. If multiple random effects are included in the model, the

differences between γ^2 terms for each covariate illustrate their relative importance to the

consumer diet (as in Semmens et al. 2009, where $\gamma_{\text{Region}}^2 > \gamma_{\text{Pack}}^2 > \gamma_{\text{Individual}}^2$, indicating that

Region explained more of the diet variability than Pack or Individual).

Since there is no one-to-one relation between the original parts and the transformed variables (i.e. each β_k acts on all p_k terms simultaneously), interpretation of model findings after back-transforming is prudent. MixSIAR therefore provides summary output statistics and preserves posterior draws on the back-transformed proportions for fixed categorical and random effects. In the case of continuous fixed effects (see below), MixSIAR generates a plot of the fitted line in the untransformed proportion space that spans the range of the provided covariate data. For the full set of MixSIAR equations and additional explanation, see Article S1.

497 Case study: *Alligator mississippiensis* diet partitioning

498 This case study highlights the main advantage of MixSIAR over previous mixing model
 499 software—the ability to include fixed and random effects as covariates explaining variability in
 500 mixture proportions and calculate relative support for multiple models via information criteria.
 501 Nifong et al. (2015) analyzed stomach contents and stable isotopes to investigate cross-
 502 ecosystem (freshwater vs. marine) resource use by the American alligator (*Alligator*
 503 *mississippiensis*), and how this varied with ontogeny (total length), sex, and between individuals.
 504 They used 2-source (marine, freshwater), 2-tracer ($\delta^{13}\text{C}$, $\delta^{15}\text{N}$) mixing models and posed three
 505 questions:

506 Q1. What is p_{marine} vs. $p_{\text{freshwater}}$?

507 Q2. How does p_{marine} vary with the covariates Length, Sex, and Individual?

508 Q3. How variable are individuals' diets relative to group-level variability?

509 Nifong et al. (2015) grouped the consumers into eight subpopulations (all combinations of Sex :
 510 Size Class, where Sex \in {male, female} and Size Class \in {small juvenile, large juvenile,
 511 subadult, adult}) and ran separate mixing models for each using SIAR (Parnell et al. 2010). To
 512 calculate p_{marine} estimates for the overall population, they also ran a mixing model with all
 513 consumers. In addition to inadequately addressing Q3 on individual diet variability, this
 514 approach is likely inefficient, as it fits nine residual error terms for each tracer and does not
 515 capitalize on the fact that diets of different-sex and different-sized alligators are probably related.
 516 We propose that a more natural, statistically efficient approach is to fit several models with fixed
 517 and random effects as covariates, and then evaluate the relative support for each model using
 518 information criteria (see “compare_models” function in MixSIAR).

We used MixSIAR to fit eight mixing models with different covariate structures (Table 1, Data S1). Since each model is fit to the same data ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values for each of 181 alligators), we can compare the models using information criteria. Deviance information criterion (DIC) is a commonly-used generalization of Akaike information criterion (AIC) for Bayesian model selection which estimates out-of-sample predictive accuracy using within-sample fits. DIC, however, has several undesirable qualities (e.g. can produce negative estimates of the effective number of parameters, is not defined for singular models, and is not invariant to model parameterization; Vehtari, Gelman, & Gabry 2017). Therefore, MixSIAR implements the widely applicable information criterion (WAIC) and approximate leave-one-out cross-validation (LOO), both of which are more robust to the concerns associated with DIC (Vehtari, Gelman, & Gabry 2017). For a set of candidate models fit to the same mixture data, we can calculate the relative support for each model using LOO and Akaike weights, which are estimates of the probability that each model will make the best predictions on new data (Burnham and Anderson 2002, McElreath 2016).

We found that the models with Length as a continuous fixed effect are heavily preferred over the models that break length into four size classes (combined weight of ‘Length’ and ‘Length + Sex’ = 99%, Table 1). There is little evidence for including sex in addition to length or size class, although it cannot be ruled out (adding sex increases LOO in both cases, but ‘Length + Sex’ still receives 20% weight, Table 1). While the original analysis by Nifong et al. (2015) predicts p_{marine} as a function of subpopulation membership, the ‘Length’ model predicts p_{marine} as a function of length (Fig. 5). Under the ‘Size class : Sex’ model of Nifong et al. (2015), the p_{marine} estimate for adult males is 0.76 (median, 95% CI 0.68-0.84), while the ‘Length’ model estimate of p_{marine} for the largest individual, a 315.5 cm adult male, is 0.96 (median, 95% CI

0.91-0.99). Although Nifong et al. (2015) clearly document an ontogenetic shift in alligator resource use, the data support the conclusion that this shift likely occurs as a continuous function of body size, instead of in discrete stages.

This case study also highlights the interaction between covariates and the multiplicative error term, ξ_j . As covariates are included that increasingly explain the observed variability in alligator isotope values, the estimates of ξ_j shrink (ξ_C decreases from 8.4 to 5.2, ξ_N decreases from 2.2 to 1.0; Table 1). The ξ_N estimate from the ‘Length’ model (1.0) is about what we expect given the assumptions about how predators sample prey. The ξ_C estimate (5.2) is very high, however, indicating that there remains an important process that is unaccounted for in the model. There are several possible explanations (see section on ‘Understanding MixSIAR error structures for mixture data’), with one being that individuals’ diets likely differ based on other processes than sex or length—all models in Table 1 assume that individuals of the same sex, length, and/or size class share the same diet proportions. We can, however, relax this assumption by including Individual as a random effect in addition to Length (or other covariates). Then the diet proportion for the i th individual becomes:

$$p_i = \text{inverse.ILR}(\beta_0 + \beta_1 \text{Length}_i + \beta_{ind}),$$

$$\beta_1 \sim N(0, 1000),$$

$$\beta_{ind} \sim N(0, \sigma_{ind}^2),$$

$$\sigma_{ind}^2 \sim U(0, 20).$$

This ‘Length + Individual’ model allows p_{marine} for individual alligators to vary around the expectation based on Length (Fig. 6).

Like many ecologists, Nifong et al. (2015) were interested in how variable individuals’ diets are, relative to group-level variability (Q3). They calculated the specialization index (ϵ) of

Newsome et al. (2012) for their overall population model, 0.26 ± 0.05 , concluded that alligators are mostly generalists, and “the diet of the majority of individuals is expected to be comprised of similar proportions of freshwater and marine prey.” The proper interpretation, however, is clearer with the best performing model (‘Length’)—the specialization index of an alligator of *average length* is low, but small and large alligators are highly specialized (Fig. 7). Additionally, since the ‘Length + Individual’ model estimates individuals’ diet proportions, we can plot the distribution of ε_{ind} and see directly that most alligators are specialists ($\varepsilon > 0.8$, Fig. 8). Nifong et al. (2015) performed a well-designed study, and their main conclusions are robust—we only reanalyze their data here to highlight advantages of MixSIAR over other mixing model software.

Limitations of Bayesian Mixing Models

Like any statistical model, inference from mixing models is only as good as the data being used. In some situations, data may not be informative – these situations may arise when models are mis-specified, or data are limited. These situations may be difficult to diagnose, because they often require a detailed examination of the likelihood or posterior distributions (which may appear flat with respect to the parameter of interest). Similar situations arise in all statistical models – for example fitting a regression model to a constant response $Y = (3,3,3,...)$ returns an estimate that is a perfect fit to the data, but does not produce standard errors or test-statistics (the response is assumed to be normally distributed, but the variance of $Y = 0$). Several recent papers have illustrated some of these same points with respect to mixing models, and we detail those here.

As a first limitation, Bond and Diamond (2011) illustrated that recently developed mixing models are sensitive to the choice of discrimination factors (systematic changes in the tracer values through the mixing process). This issue arises because the discrimination factors and

estimated source contributions are not completely identifiable. In other words, these parameters are difficult to estimate simultaneously, and one or the other is generally fixed (in food web studies, the discrimination factor is typically specified as fixed *a priori*). At present, MixSIAR does not provide the option to estimate discrimination from user-provided data, although such functionality could easily be added; we anticipate adding this functionality into a future software release.

A second limitation of mixing models is that systems may be under-determined (as discussed in the introduction). Phillips and Gregg (2003) demonstrated several examples of this problem for the 2-tracer scenario, but the issue of underdetermined problems generally arises when the number of sources exceeds the number of tracers plus one. In such instances, posterior estimates of source contributions can be broad and multi-modal, owing to the fact that multiple, often disparate, solutions to the underlying mixing equations exist. Fry (2013) proposed a graphical approach to separate data-supported aspects of solutions from any assumed aspects of solutions method. Essentially, this approach is a *post hoc* means of evaluating model performance, and can easily be applied to the products of any mixing model (including the products of a MixSIAR model run).

A larger issue with underdetermined systems is that in some cases, the choice of Bayesian prior will play a large role. In completely determined systems with reasonable sample sizes and separation of sources, the choice of prior has little impact on results. When systems are underdetermined, however, data may be less informative, and as a result the priors can be relatively influential. Moreover, as the variability within sources increases (the variability around source means), the prior plays an even larger role. Brett (2014) described the interaction between the prior and the shape of the mixing polygon (which arises from the sources and their

variability) as a bias of mixing models. This phenomenon may be better described as weakly informative data, but we agree that approaches like Brett (2014)’s surface area metric may be useful in recognizing *a priori* when these situations may arise. As such, we have incorporated Brett’s surface area metric as a diagnostic output in MixSIAR (“calc_area” function). However, work still needs to be done to generalize this metric to situations with any number of tracers and sources.

Conclusion

Analysts applying modern mixing model software typically must navigate a challenging array of model choices, from source groupings to covariate data, to error parameterization. In the past, those analysts not capable of developing their own models have been faced with the choice between different software packages, each with differing statistical model structures and assumptions. Through the creation of MixSIAR, we have incorporated the disparate suite of mixing model advances into a single tool with the flexibility to meet most analyst’s needs. Because MixSIAR is open source and collaborative, we anticipate that new developments in mixing model methods, from parameterizations to model performance diagnostics, will continue to be incorporated into the functionality of MixSIAR. As such, the software provides a single tool that can meet the diverse needs of the rapidly increasing pool of stable isotope analysts, and affords developers a platform upon which to continue improving and diversifying mixing model analyses.

631 References

- 632 Aitchison, J. 1986. The statistical analysis of compositional data. Chapman and Hall, London.
- 633 Bearhop, S., S. Waldron, S. C. Votier, and R. W. Furness. 2002. Factors that influence
- 634 assimilation rates and fractionation of nitrogen and carbon stable isotopes in avian blood
- 635 and feathers. *Physiological and Biochemical Zoology* 75(5):451-458.
- 636 Ben-David, M., R. W. Flynn, and D. M. Schell. 1997. Annual and seasonal changes in diets of
- 637 martens: evidence from stable isotope analysis. *Oecologia* 111(2):280-291.
- 638 Bond, A.L., and A.W. Diamond. 2011. Recent Bayesian stable-isotope mixing models are highly
- 639 sensitive to variation in discrimination factors. *Ecological Applications* 21(4):1017-1023.
- 640 Brett, M.T. 2014. Resource polygon geometry predicts Bayesian stable isotope mixing model
- 641 bias. *Marine Ecology Progress Series* 514:1-12.
- 642 Burnham, K.P. and D.R. Anderson. 2002. Model selection and multimodel inference: a practical
- 643 information-theoretic approach. Springer Science & Business Media.
- 644 deVries, M.S., B.C. Stock, J.H. Christy, G.R. Goldsmith, and T.E. Dawson. 2016. Specialized
- 645 morphology corresponds to a generalist diet: linking form and function in smashing
- 646 mantis shrimp crustaceans. *Oecologia* 182(2):429-442.
- 647 Francis, R.C. 2011. Data weighting in statistical fisheries stock assessment models. *Canadian*
- 648 *Journal of Fisheries and Aquatic Sciences* 68(6):1124-1138.
- 649 Francis, T. B., D. E. Schindler, G. W. Holtgrieve, E. R. Larson, M. D. Scheuerell, B. X.
- 650 Semmens, and E. J. Ward. 2011. Habitat structure determines resource use by
- 651 zooplankton in temperate lakes. *Ecology Letters* 14(4):364-372.

- 652 Franco-Trecu, V., M. Drago, F. G. Riet-Sapriz, A. Parnell, R. Frau, and P. Inchausti. 2013. Bias
653 in diet determination: incorporating traditional methods in Bayesian mixing models.
654 PLoS ONE 8(11):e80019.
- 655 Fry, B. 2013. Alternative approaches for solving underdetermined isotope mixing problems.
656 Marine Ecology Progress Series 472:1–13.
- 657 Galloway, A.W.E., M.T. Brett, G.W. Holtgrieve, E.J. Ward, A.P. Ballantyne, C.W. Burns, M.J.
658 Kainz, D.C. Müller-Navarra, J. Persson, J.L. Ravet, U. Strandberg, S.J. Taipale, and G.
659 Alhgren. 2015. A fatty acid based Bayesian approach for inferring diet in aquatic
660 consumers. PLoS ONE 10(6):e0129723.
- 661 Gelman, A. 2005. Analysis of variance—why it is more important than ever. The Annals of
662 Statistics 33(1):1-53.
- 663 Gelman, A., J.B. Carlin, H.S. Stern, and D.B. Rubin. 2014. Bayesian data analysis (2nd ed.).
664 Boca Raton, FL: CRC Press.
- 665 Hurlbert, S.H. 1984. Pseudoreplication and the design of ecological field experiments.
666 Ecological Monographs 54(2):187-211.
- 667 Hopkins, J.B., III, and J.M. Ferguson. 2012. Estimating the diets of animals using stable isotopes
668 and a comprehensive Bayesian mixing model. PLoS ONE 7(1):e28478.
- 669 Jackson, A.L., R. Inger, S. Bearhop, and A. Parnell. 2009. Erroneous behaviour of MixSIR, a
670 recently published Bayesian isotope mixing model: a discussion of Moore & Semmens
671 (2008). Ecology Letters, 12:E1-E5.
- 672 Jackson, D.A. 1997. Compositional data in community ecology: the paradigm or peril of
673 proportions? Ecology 78(3):929-940.
- 674 McCarthy, M.A. 2007. Bayesian methods for ecology. Cambridge University Press, London.

- 675 McElreath, R. 2016. Statistical rethinking: a Bayesian course with examples in R and Stan. CRC
676 Press.
- 677 Moore, J.W., and B.X. Semmens. 2008. Incorporating uncertainty and prior information into
678 stable isotope mixing models. *Ecology Letters* 11(5):470-480.
- 679 Neubauer, P., and O.P. Jensen. 2015. Bayesian estimation of predator diet composition from
680 fatty acids and stable isotopes. *PeerJ* 3:e920.
- 681 Newsome, S.D., J.D. Yeakel, P.V. Wheatley, and M.T. Tinker. 2012. Tools for quantifying
682 isotopic niche space and dietary variation at the individual and population level. *Journal*
683 *of Mammalogy* 93(2):329-341.
- 684 Ogle, K., C. Tucker, and J.M. Cable. 2014. Beyond simple linear mixing models: process-based
685 isotope partitioning of ecological processes. *Ecological Applications* 24(1):181-195.
- 686 Parnell, A.C., R. Inger, S. Bearhop, and A.L. Jackson. 2010. Source partitioning using stable
687 isotopes: coping with too much variation. *PLoS ONE* 5(3):e9672.
- 688 Parnell, A.C., D.L. Phillips, S. Bearhop, B.X. Semmens, E.J. Ward, J.W. Moore, A.L. Jackson, J.
689 Grey, D.J. Kelly, and R. Inger. 2013. Bayesian stable isotope mixing models.
690 *Environmetrics* 24(6):387-399.
- 691 Phillips, D.L. 2001. Mixing models in analyses of diet using multiple stable isotopes: a critique.
692 *Oecologia* 127(2):166-170.
- 693 Phillips, D.L., and J.W. Gregg. 2003. Source partitioning using stable isotopes: coping with too
694 many sources. *Oecologia* 136(2):261-269.
- 695 Phillips, D.L., R. Inger, S. Bearhop, A.L. Jackson, J.W. Moore, A.C. Parnell, B.X. Semmens,
696 and E.J. Ward. 2014. Best practices for use of stable isotope mixing models in food-web
697 studies. *Canadian Journal of Zoology* 92(10):823-835.

- 698 Phillips, D.L., S.D. Newsome, and J.W. Gregg. 2005. Combining sources in stable isotope
699 mixing models: alternative methods. *Oecologia* 144(4):520-527.
- 700 Plummer, M. 2003. JAGS: A program for analysis of Bayesian graphical models using Gibbs
701 sampling. Proceedings of the 3rd international workshop on distributed statistical
702 computing. Vienna, Austria.
- 703 Rosing, M.N., M. Ben-David, and R.P. Barry. 1998. Analysis of stable isotope data: a K nearest-
704 neighbors randomization test. *The Journal of Wildlife Management* 62(1):380-388.
- 705 Schwarcz, H.P. 1991. Some theoretical aspects of isotope paleodiet studies. *Journal of*
706 *Archaeological Science* 18(3):261-275.
- 707 Semmens, B.X., J.W. Moore, J.W., and E.J. Ward. 2009. Improving Bayesian isotope mixing
708 models: a response to Jackson et al. (2009). *Ecology Letters*, 12:E6-E8.
- 709 Semmens, B.X., E.J. Ward, J.W. Moore, and C.T. Darimont. 2009. Quantifying inter- and intra-
710 population niche variability using hierarchical Bayesian stable isotope mixing models.
711 *PLoS ONE* 4(7):e6187.
- 712 Semmens, B.X., E.J. Ward, A.C. Parnell, D.L. Phillips, S. Bearhop, R. Inger, A. Jackson, and J.
713 W. Moore. 2013. Statistical basis and outputs of stable isotope mixing models: Comment
714 on Fry (2013). *Marine Ecology Progress Series* 490:285-289.
- 715 Stock, B.C. and B.X. Semmens. 2016a. MixSIAR GUI user manual. Version 3.1.
716 <https://github.com/brianstock/MixSIAR>. doi:10.5281/zenodo.56159.
- 717 Stock, B.C., and B.X. Semmens. 2016b. Unifying error structures in commonly used biotracer
718 mixing models. *Ecology* 97(10):2562-2569.
- 719 Thorson, J.T. 2014. Standardizing compositional data for stock assessment. *ICES Journal of*
720 *Marine Science: Journal du Conseil* 71(5):1117-1128.

- 721 Van Der Boogaart, K.G., and R. Tolosana-Delgado. 2006. Compositional data analysis with 'R'
722 and the package 'compositions.' Geological Society, London, Special Publications
723 264:119-127.
- 724 Vehtari, A., A. Gelman, and J. Gabry. 2017. Practical Bayesian model evaluation using leave-
725 one-out cross-validation and WAIC. *Statistics and Computing* 27(5):1413-1432.
- 726 Ward, E.J., B.X. Semmens, D.L. Phillips, J.W. Moore, and N. Bouwes. 2011. A quantitative
727 approach to combine sources in stable isotope mixing models. *Ecosphere* 2(2):1-11.
- 728 Ward, E.J., B.X. Semmens, and D.E. Schindler. 2010. Including source uncertainty and prior
729 information in the analysis of stable isotope mixing models. *Environmental Science &*
730 *Technology* 44(12):4645-4650.
- 731 Welch, D.W., and T.R. Parsons. 1993. $\delta^{13}\text{C}$ - $\delta^{15}\text{N}$ values as indicators of trophic position and
732 competitive overlap for Pacific salmon (*Oncorhynchus spp.*). *Fisheries Oceanography*
733 2(1):11-23.

Table 1(on next page)

Comparison of mixing models fit using MixSIAR on the alligator diet partitioning data from Nifong et al. (2015).

dLOOic is the difference in LOOic between each model and the model with lowest LOOic. The 'Length' model had the lowest LOOic and received 79% of the Akaike weight, indicating a 79% probability it is the best model. The 'Length + Sex' model cannot be ruled out (20% weight). Note that as variability in the mixture data is better explained by covariates, the estimates of ξ_j decrease.

Model	LOOic	SE(LOOic)	dLOOic	SE(dLOOic)	Weight	ξ_C	ξ_N
Length	820.8	31.4	0	--	0.789	5.3	1.0
Length + Sex	823.6	31.4	2.8	2.1	0.195	5.2	1.0
Size class	829.5	31.6	8.7	11.7	0.010	5.4	1.1
Size class + Sex	831.4	31.5	10.6	12.1	0.004	5.3	1.1
Size class : Sex	832.9	29.8	12.1	13.6	0.002	4.9	1.1
Habitat	890.7	28.7	69.9	43.4	0	6.4	1.5
Sex	973.8	17.7	153.0	30.1	0	8.4	2.2
--	977.0	16.7	156.2	31.5	0	8.4	2.2

Table 1. Comparison of mixing models fit using MixSIAR on the alligator diet partitioning data from Nifong et al. (2015). dLOOic is the difference in LOOic between each model and the model with lowest LOOic. The 'Length' model had the lowest LOOic and received 79% of the Akaike weight, indicating a 79% probability it is the best model. The 'Length + Sex' model cannot be ruled out (20% weight). Note that as variability in the mixture data is better explained by covariates, the estimates of ξ_j decrease.

Figure 1(on next page)

Representation of the 3 different methods MixSIAR uses for modeling variability in mixture data, assuming a two source (k), 1 tracer (j) scenario

A) In the "residual error only" formulation, the means of each source (upper black dots; typically estimated within the model based on source data) are additively combined, after weighting based on estimated proportional source contributions, in order to generate the expected mean value of the mixture signatures (Eq. 1). Actual mixture measurements deviate from this mean due to residual error, σ_j^2 . B) Given a single mixture data point, MixSIAR assumes this mixture value is drawn from a normal distribution defined by the same mean, with the variance generated by a weighted combination of source variances (Eq. 2). C) In the "multiplicative error" formulation (Eq. 3), the model assumes the mixture data are generated from the process as in (B), but the variance of this distribution is modified by a multiplicative term, ξ_j , that allows the distribution to shrink (as would be expected if consumers are sampling multiple times from each source pool) or expand (as would be expected if the model is missing a non-negligible source, or processes such as isotopic routing introduce significant additional variability into the mixing system).

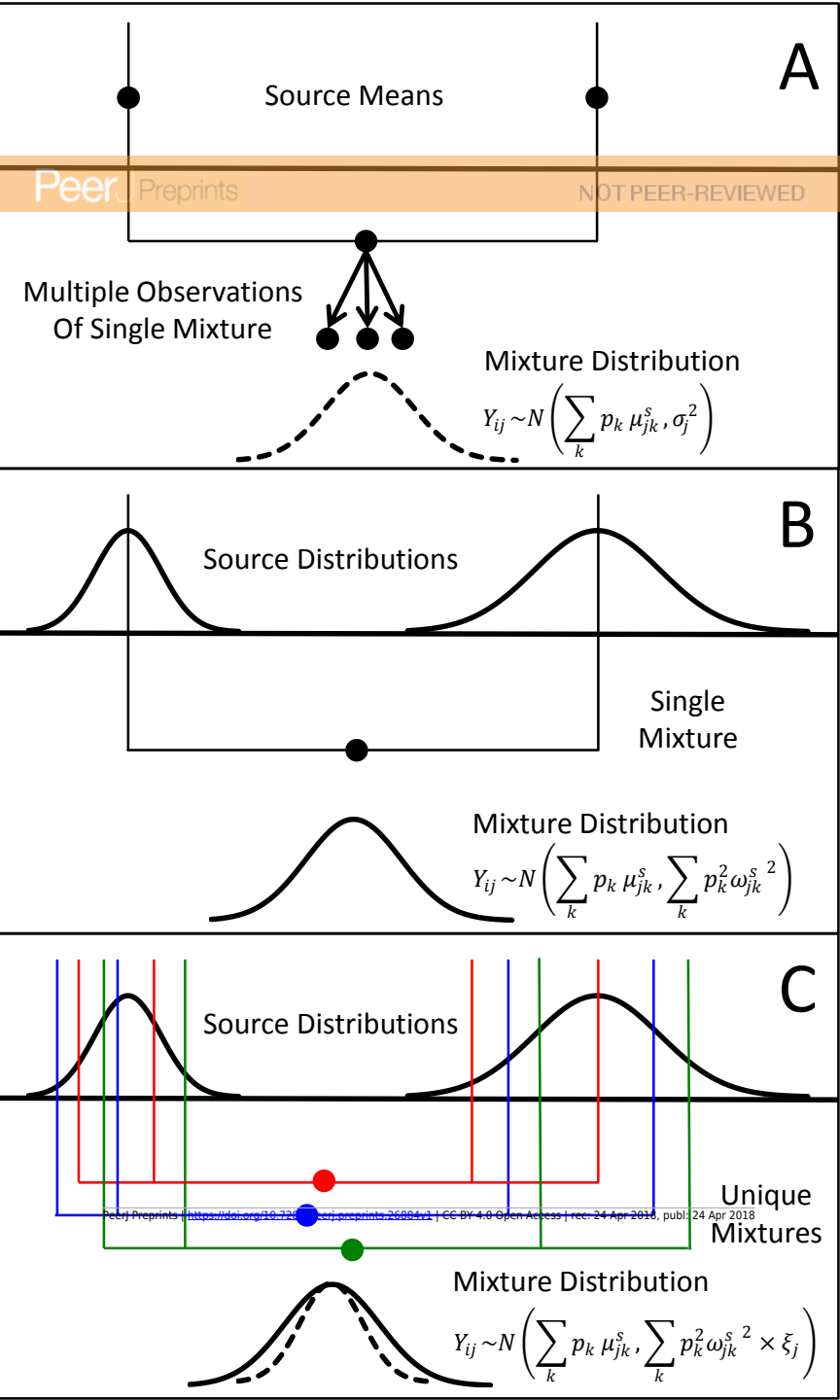
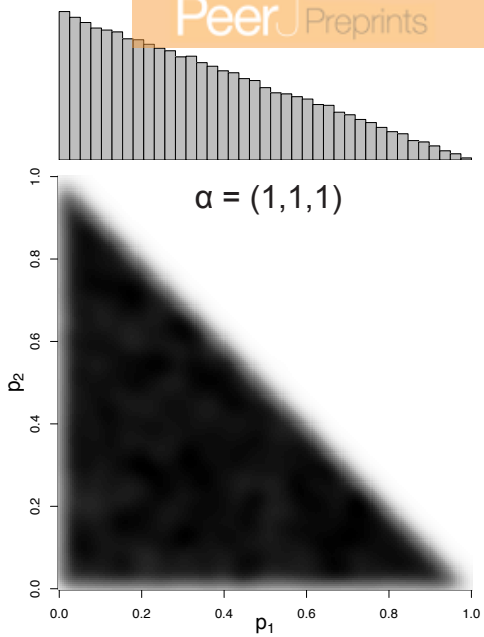


Figure 2(on next page)

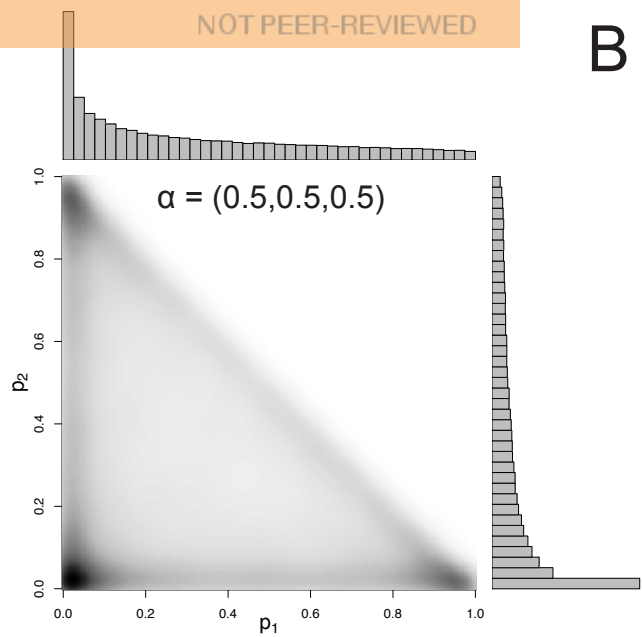
Examples of joint and marginal distributions of p_1 and p_2 for a 3-component Dirichlet distribution, across 4 sets of hyperparameters.

(A) $\alpha = 1$, (B) $\alpha = 0.5$, (C) $\alpha = 10$, and (D) $\alpha = 100$. All simulations were done with the 'rdirichlet' function in the 'compositions' library in R (Van Der Boogaart and Tolosana-Delgado 2006).

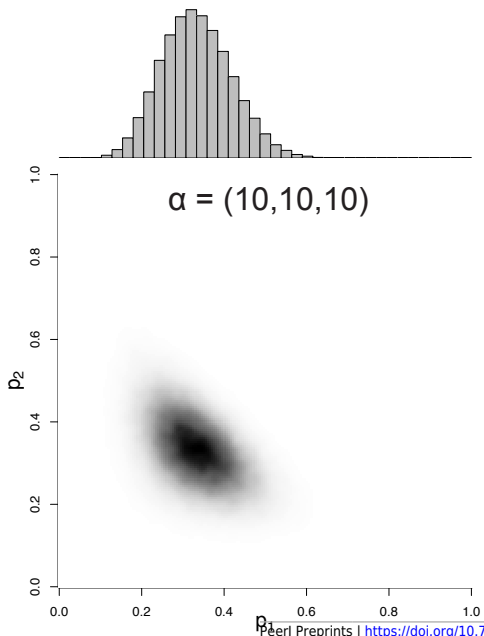
A



B



C



D

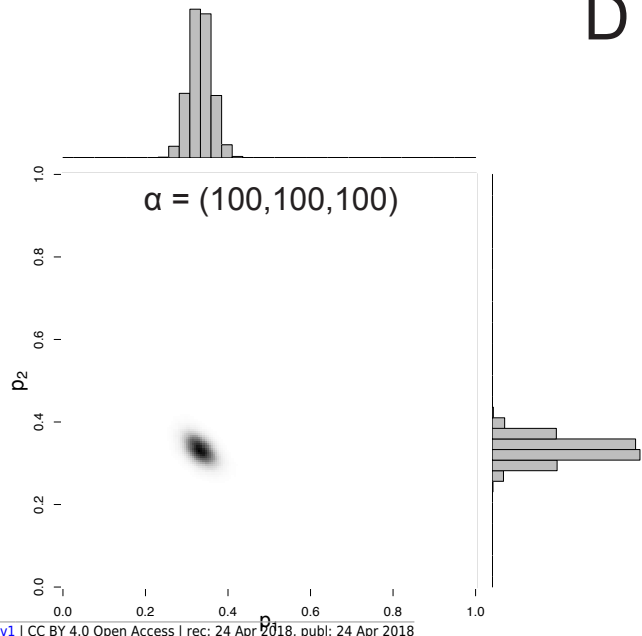


Figure 3

Illustration of alternative priors for a mixing model of rainbow trout (consumers/mixture) diet comprised of 3 sources: eggs, fish, and invertebrates

(Left) The "uninformative"/generalist Dirichlet prior MixSIAR uses by default, $\alpha = (1, 1, 1)$.
(Middle) A strongly informative prior with $\alpha = (30, 8, 25)$, where each α_k corresponds to the sample size of source k from stomach contents. (Right) A moderately informative prior with the same mean, but each α_k rescaled such that $\sum \alpha_k = 3$, the number of sources. Note that both informative priors have the same mean but differ in their "informativeness".

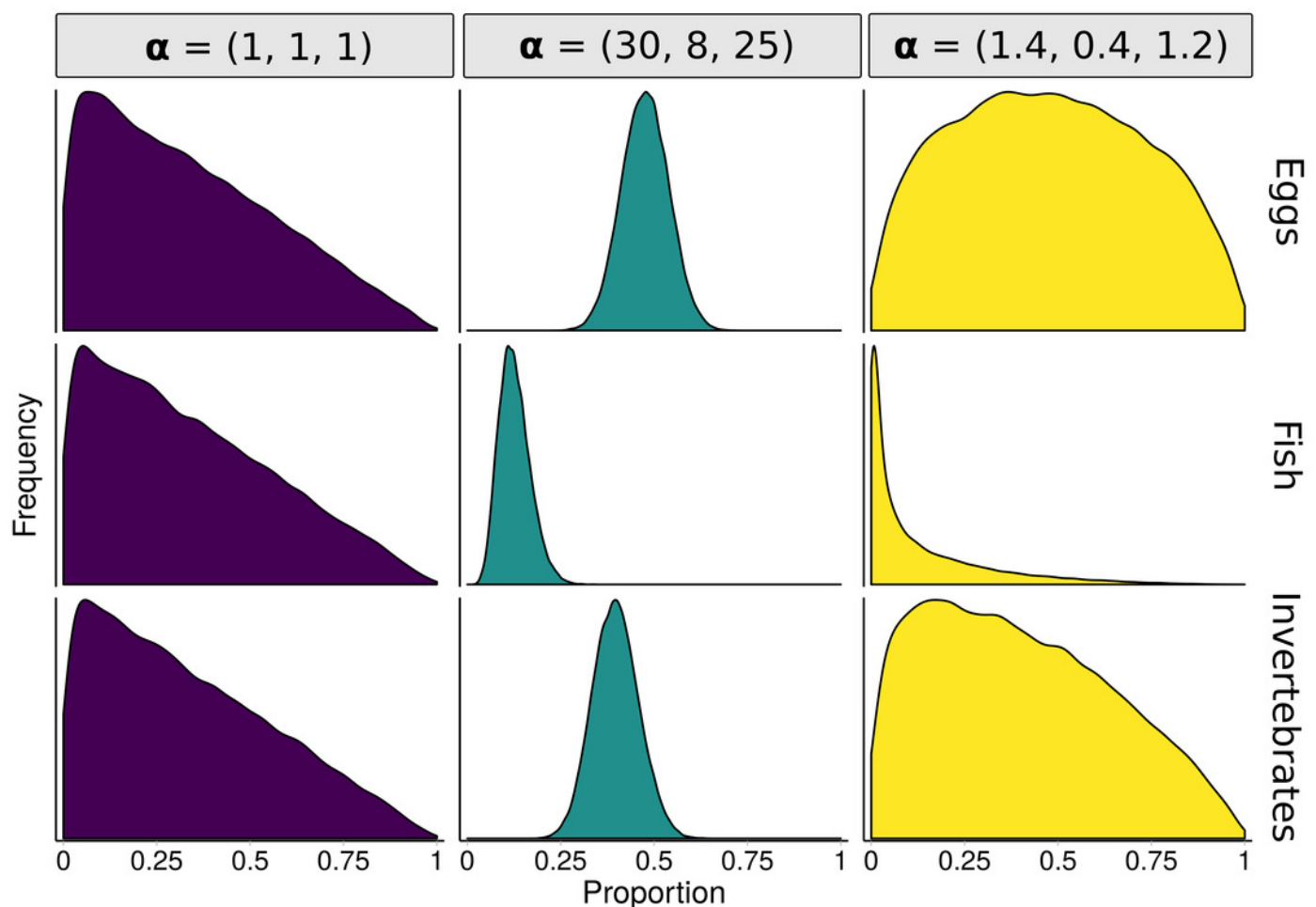
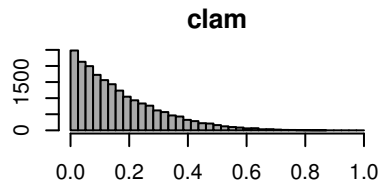
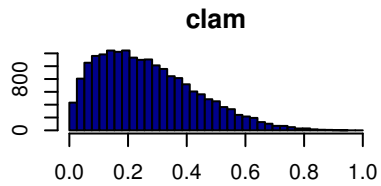
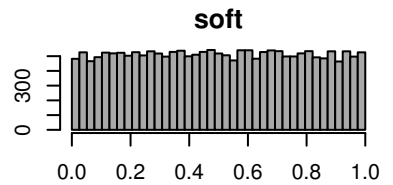
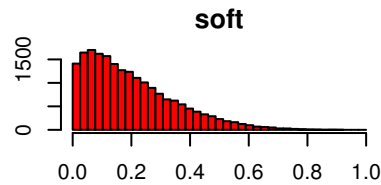
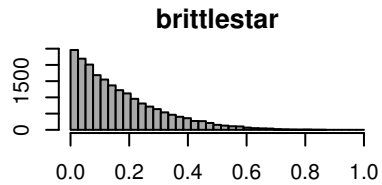
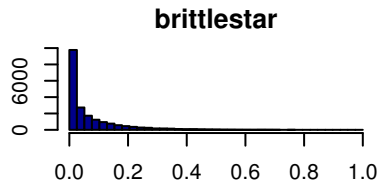
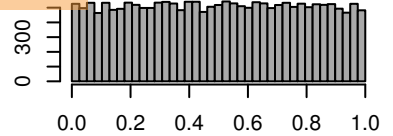
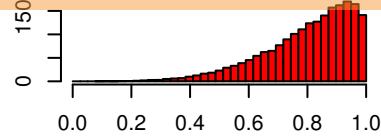
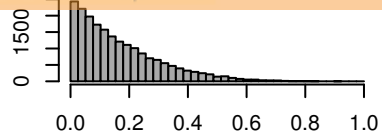
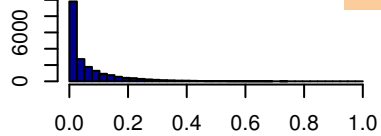


Figure 4(on next page)

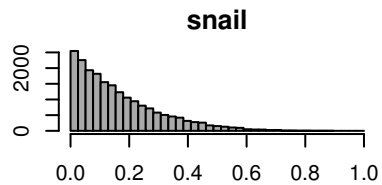
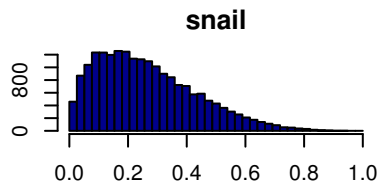
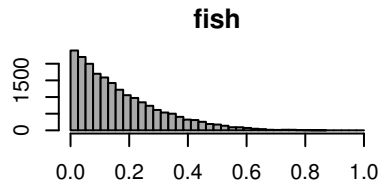
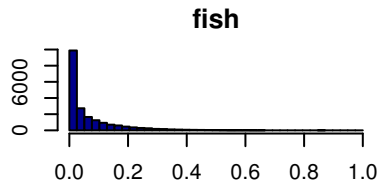
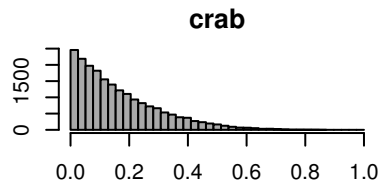
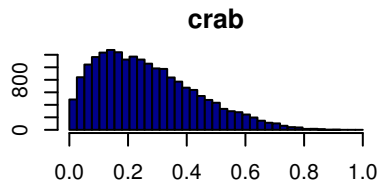
Effect of aggregating sources *a posteriori* on priors in mixing models, produced by the "combine_sources" function in MixSIAR as a warning to the user.

Columns from left to right: the original, unaggregated prior on 6 sources from the mantis shrimp example (dark blue); the "uninformative"/generalist prior on 6 sources (grey); the prior resulting from aggregating the 6-source prior in dark blue into 2 sources (hard-shelled = clam + crab + snail, soft-bodied = alphworm + brittlestar + fish, red); and the prior resulting from aggregating the 6-source "uninformative"/generalist prior into the same 2 sources (grey).



New prior
(4.8,1.2)

"Uninformative" prior
(1,1)



Original prior
(0.4,0.4,1.6,1.6,0.4,1.6)

"Uninformative" prior
(1,1,1,1,1,1)

Figure 5

Posterior distributions for alligator diet proportions as a function of length from the best performing model, 'Length'.

Small/young alligators depend upon freshwater prey and shift to a marine-based diet as they increase in size. Lines depict posterior medians, and shading displays the 90% credible intervals. The 'Length' model estimate of p_{marine} (blue curve) for the largest individual, a 315.5 cm adult male, is 0.96 (median, 95% CI 0.91-0.99). Estimates of p_{marine} for the smallest (37.7 cm) and median-sized (116.9 cm) alligators are 0.09 (0.04-0.15) and 0.32 (0.24-0.39), respectively.

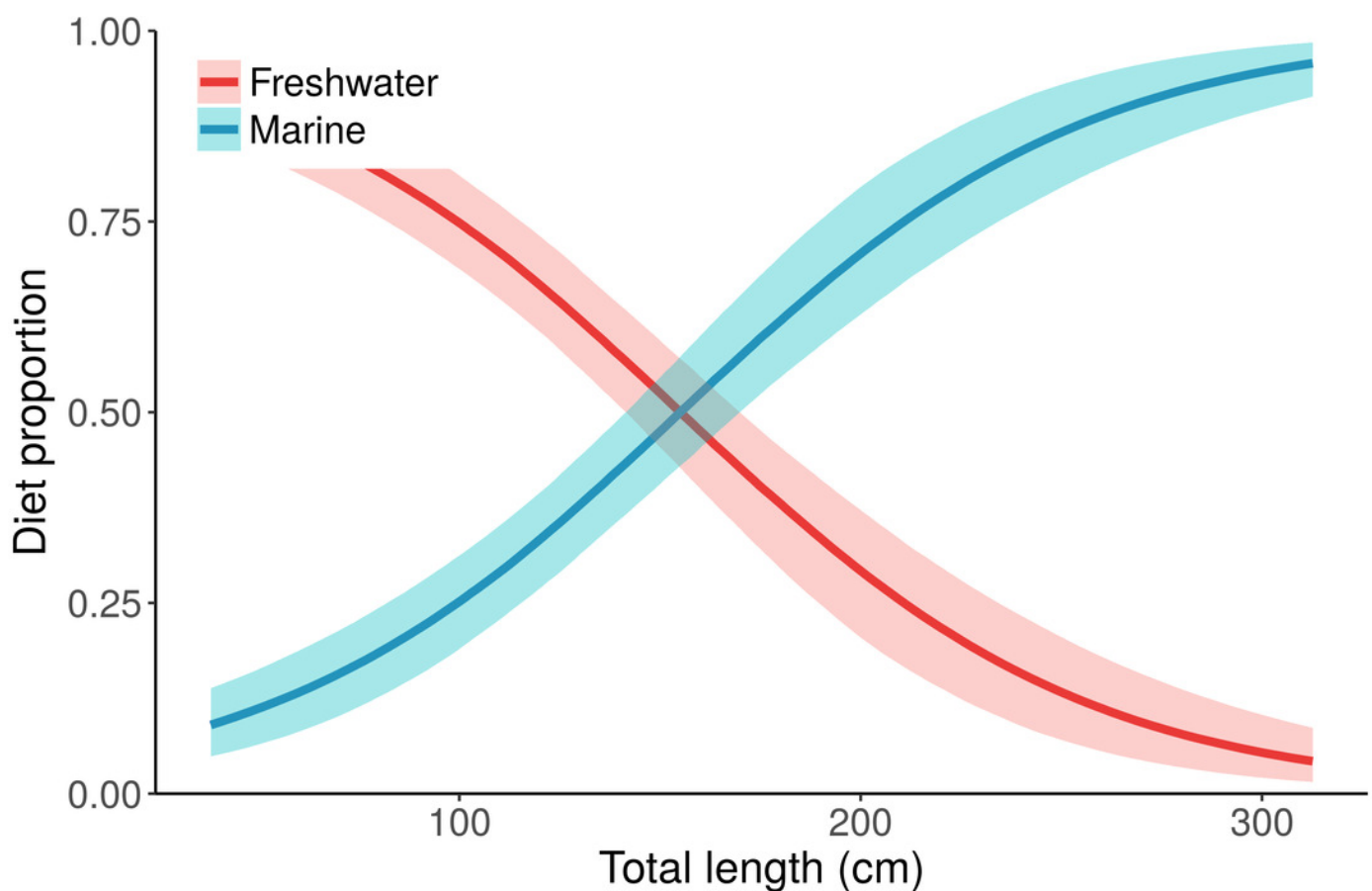


Figure 6

Posterior distributions for the marine proportion, p_{marine} , of alligator diet as a function of length from the 'Length + Individual' model.

Whereas the 'Length' model estimates one diet for all alligators of a given length, the 'Length + Individual' model allows p_{marine} for individual alligators to vary around the expectation based on Length. For most alligators around 100 cm total length, the p_{marine} is very low, but for some it is above 80%. Likewise, the model estimates that most large (> 200 cm) alligators' diets are dominated (> 95%) by marine prey, but p_{marine} for three large individuals is less than 10%.

Dark blue line and points indicate posterior medians, light lines and shading show 90% credible intervals.

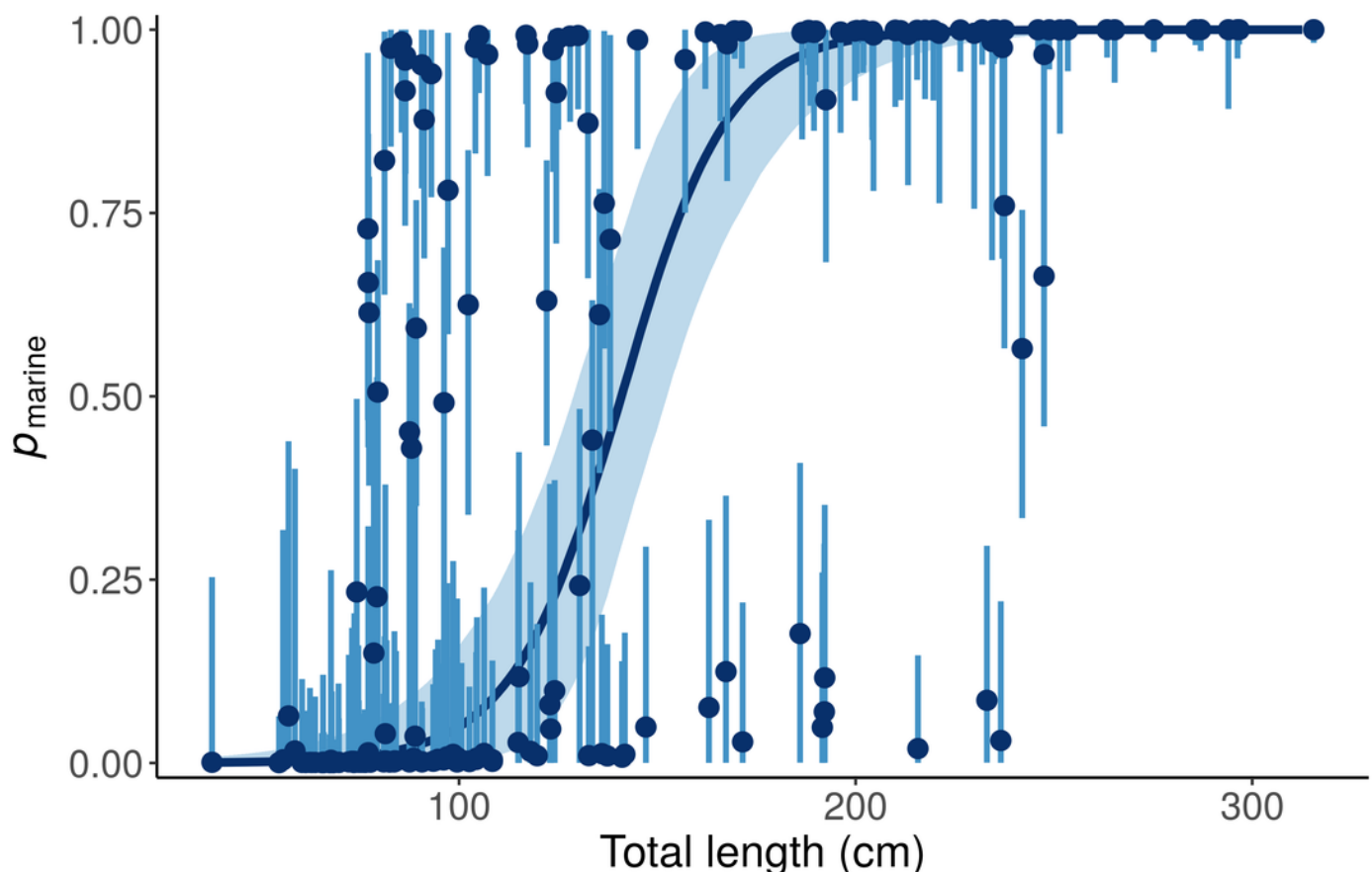


Figure 7

Posterior distribution of the specialization index (ε) as a function of length from the 'Length' model.

Small and large alligators are highly specialized (on freshwater and marine prey, respectively), whereas average-length alligators have low specialization index (i.e. are consuming both freshwater and marine prey). Specialization index is calculated using Eq. 5 in Newsome et al. (2012) from individual MCMC draws of $p_{\text{freshwater}}$ and p_{marine} as a function of length. The line depicts the posterior median and shading displays the 95% credible interval.

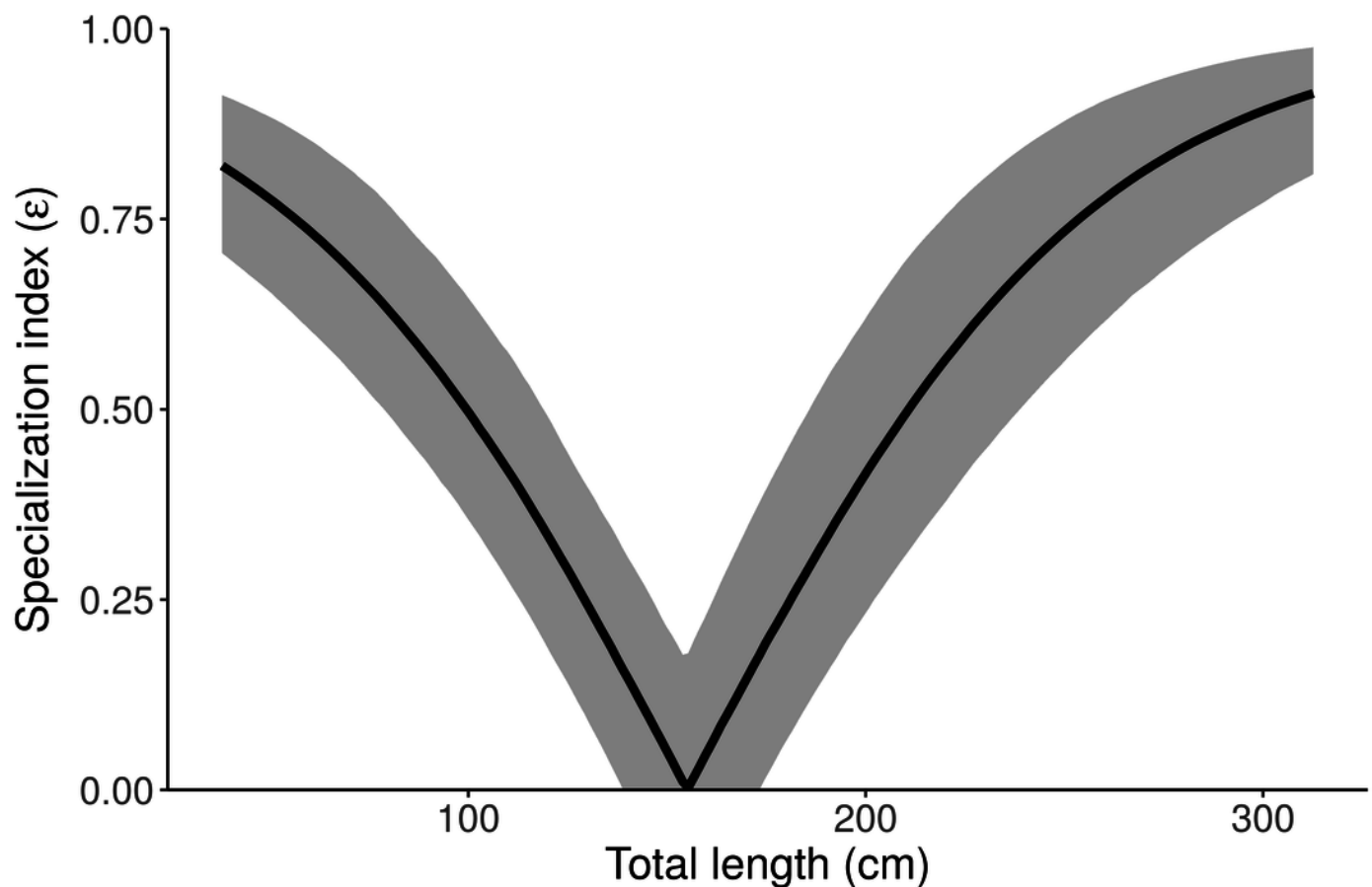
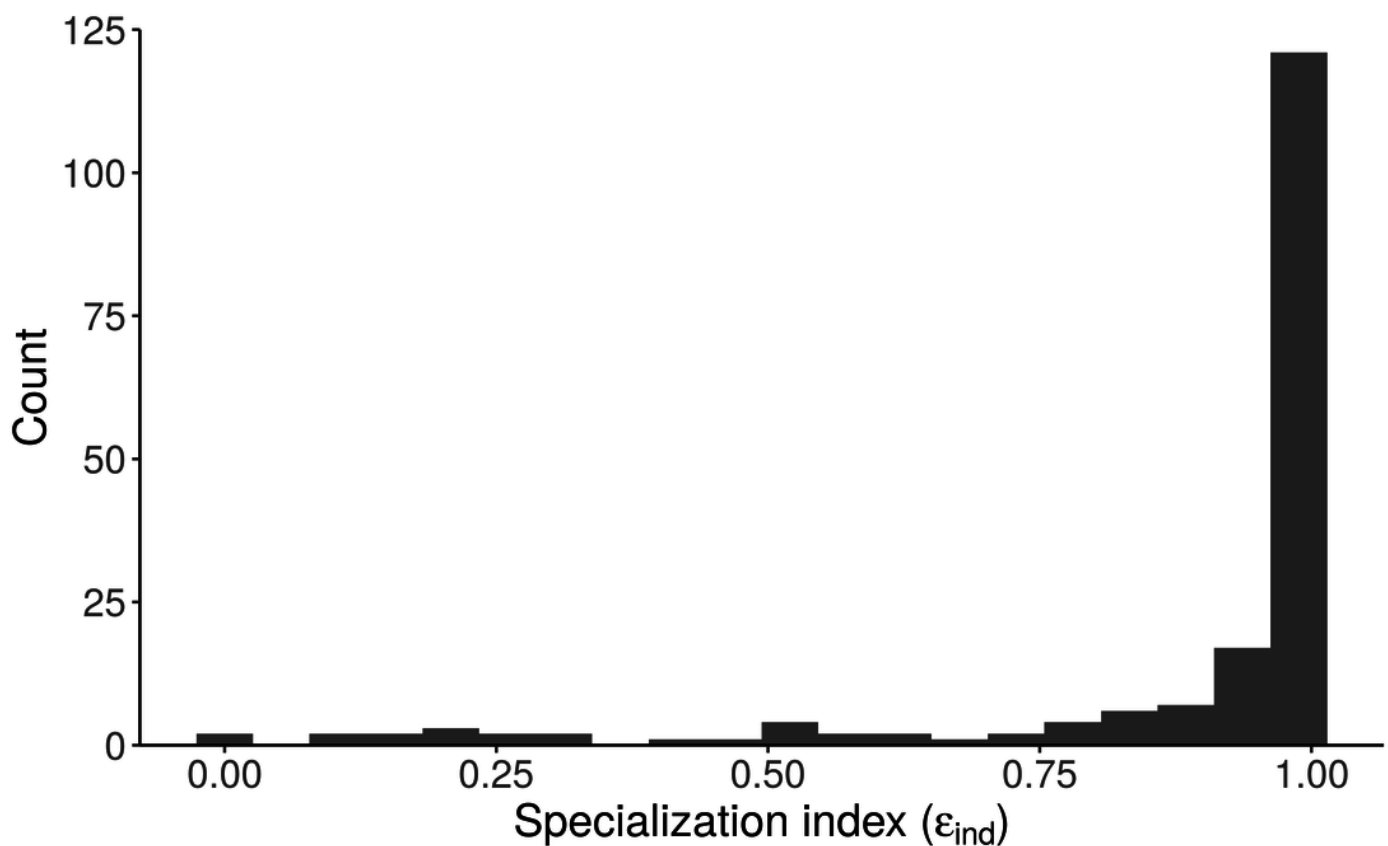


Figure 8

Distribution of the specialization index calculated for each individual (ϵ_{ind} , $n = 181$) from the 'Length + Individual' model estimates of individuals' diet proportions (posterior median of p_{ind}).

The model estimates that most alligators sampled by Nifong et al. (2015) are specialists ($\epsilon > 0.8$).



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