

Finding suitable weather indices for novel fisheries index insurance using machine learning

Working Paper not for Distribution

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Index insurance is a financial tool gaining traction for application in fisheries. It will cover fishers losses under extreme weather events that impact fishery productivity. This is the first assessment to determine the feasibility of such programs and whether suitable indices exist. Catch and revenue data from 74 California fisheries are matched to 20 environmental variables using three prediction models: linear regression, LASSO regression, and random forests. The models are used to calculate the marginal willingness to pay through utility models.

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1 Introduction

Predicting fishery output from weather variables is notoriously difficult. It is widely established that climate and weather affect fishing populations (Lehodey *et al.* 2006), but most stock assessment models use little to no year to year environmental data (Privitera-Johnson and Punt 2020). Variations in environmental conditions are now the leading cause of fishery closures and disaster relief payouts in the United States (Bellquist *et al.* 2021). Disaster declarations are becoming more frequent straining a slow, inequitable system (Holland and Leonard 2020; Jardine *et al.* 2020). Calls for new financial tools to alleviate fisher income shocks have grown (Mumford *et al.* 2009; Sethi 2010).

Index insurance has risen as a prime candidate to protect fishing communities during disasters (Watson *et al.* 2023). Index insurance is a financial product that pays out when an independently verified index, such as rainfall or temperature, falls below a predetermined threshold. The index is chosen to be highly correlated with the asset being insured. However, it is difficult to establish clear, concise weather impacts on fishery productivity. The biological dynamics of the system can lead to lower stock health persisting for years after a negative shock (Hilborn *et al.* 2003). Individual fisheries can cover enormous areas in ocean basins. The expansive spatial coverage of fisheries makes it unclear where and how much specific weather variables impact biological abundance. In addition, if weather impacts have been observed, they are most likely highly non-linear adding further complexity. The greatest impediment to the development of fishery insurance policies is reconciling these challenges to find suitable indices that can predict fishery productivity (Watson *et al.* 2023).

The difficulty in modelling fishery productivity with environmental indices leads to basis risk. Formally, basis risk is the probability that policyholders experience a harmful shock to their income, but the index does not trigger. Basis risk lowers demand for index insurance and remains a significant roadblock in setting up new programs (Binswanger-Mkhize 2012; Clarke 2016; Clement *et al.* 2018).

It is impossible to completely eliminate basis risk. Well designed policies can capture up to 90% of the income variation as shown in Kenyan pasture grazing indices (Jensen *et al.* 2019). Abysmal correlations are prevalent in the Rainfall Index Insurance for Pasture, Rangeland, and Forage (RI-PRF) program where correlations as low as 0.071 exist in California, which leads to 46% additional points of basis risk (Keller and Saitone 2022). The program has a 26% probability of not paying out when damages are suffered in Nebraska and Kansas (Yu *et*

al. 2019). Subsidies covering up to 60% of ranchers paid premiums are needed to stimulate demand in the RI-PRF program (Goodrich *et al.* 2019).

Designing indices with stronger correlations to fishery losses is the most effective way to reduce basis risk (Jensen *et al.* 2019). Agricultural researchers continually seek new methods and data sources to improve the correlation between loss and weather variables. Quantile regressions improve Kazak wheat farmers utility between 0.1-22% over linear models depending on the underlying measure of utility (Conradt *et al.* 2015). Remote sensing variables leveraging the latest satellite data on vegetative cover and rainfall provide better coverage than county wide averages (Dalhaus and Finger 2016; Valverde-Arias *et al.* 2020).

Machine learning has exploded as a new method to define better indices in agriculture index insurance (Feng *et al.* 2019; Cesarini *et al.* 2021; Schmidt *et al.* 2022; Chen *et al.* 2024). Machine learning models excel in index insurance because indemnity contracts only need predictive relationships. Whereas, fishery stock assessments build complex models with biological foundations to accurately inform management of future fish stocks, index insurance can look retroactively at data to uncover relationships and test out of sample predictive quality. Machine learning may be able to capture catch and weather relationships in fisheries that traditional methods in index insurance contract cannot.

The application of machine learning is growing in fisheries as researchers explore data questions beyond formal stock assessments. Ensemble models built through combinations of random forests, boosted trees, and dynamic linear models improved Bristol Bay sockeye salmon forecasts by 15% compared to a standard lagged regression model (Ovando *et al.* 2022). Environmental variables of importance to groundfish populations in Alaska were uncovered using single index varying coefficient models regularized with LASSO (Correia 2021). Random Forests models better predict fish catch in Indonesia than traditional linear models (Rahman *et al.* 2022). The expected non-linear interactions of weather and fishery productivity merit the use of machine learning in fisheries.

Recent expansions in oceanic remote sensing has led to a plethora of new environmental indices that could be used to predict fishery productivity. Fishery data collection continues to improve with better reporting systems with longer and more detailed catch histories. This study aims to leverage these advances to determine how much fishers would be willing to pay for potential realized index insurance contracts. Furthermore, this study will investigate how to effectively design insurance contracts to incentivize purchase by examining what weather indices, fishery loss measures, and prediction models are most effective. Whether the contracts need to be subsidized or are viable in a free market will be measured by comparing the willingness to pay for insurance and payout frequency.

The rest of the paper is structured as follows. Section 2 describes the insurance model tested in this study. Section 3 describes the data collection, transformations, and sources. Fisheries data comes from newly open-access sources provided by the California Department of Fish and Wildlife. Section 4 describes the algorithms used to predict fishery productivity and evaluate

the utility of index insurance. Section 5 demonstrates some preliminary results and highlights some of the challenges present. Future steps are outlined in Section 6.

2 Insurance Model

The seminal work by Clarke (2016) proves the theoretical impacts of basis risk on insurance demand. Policyholders will choose non-zero coverage if the expected claim payment conditional on incurring a loss is higher than the paid premium (Equation 9 of Clarke (2016)). Clarke (2016) proceed to use data from 270 insurance products in India to show that Indian farmers would be willing to pay a premium with a loading factor of up to 1.56 above the actuarially fair rate for the product despite low Pearson correlation coefficients. His model uses explicit measures of crop loss and realized payment data. This presents two challenges in fisheries settings. First, the definition of loss arising from weather variability is not as clear in fisheries as it is in agriculture. Second, there are no existing fishery index insurance products to calculate realized losses and payments to fit Clarke’s ratio.

Agriculture has long historical records on crop losses. Farmers plant set amounts of crop that equate to a maximum potential yield. Weather impacts the harvest leading to actual yield. The difference between maximum potential yield and actual yield is yearly loss. Farmers report these numbers to authorities to provide county/area level information (Tack and Ubilava 2015). There is no equivalent framing in fisheries. When fishers deploy their gear, they are not guaranteed a catch. Does the discrepancy arise from the fisher’s skill, the weather, or the fish population? The question of identifying loss in fisheries was a leading reason for the denial of a salmon fishery insurance in Alaska in the early 2000s (Herrmann *et al.* 2004).

I explore multiple definitions of loss to uncover the most viable options moving forward in fisheries. Each measure of loss revolves around defining a productivity measure of a fishery π , such as catch or revenue. Setting the threshold for loss for each of the productivity measures is the next challenge. Average historical harvest is the most common choice in agriculture. I will use this as the baseline threshold for each π . The biological dynamics of fisheries and residual impacts of weather variables suggest fisheries may need time dependent thresholds. A moving average over the last 5 years of π could capture more accurate measures of loss. Finally, biological thresholds could be used as limits to define loss (Deng *et al.* 2008). For example, fisheries that operate below maximum sustainable yield could be considered to have incurred a loss.

To tackle the second challenge, I blend the model of Clarke (2016) with the data driven utility models of Conradt *et al.* (2015) and Kenduiywo *et al.* (2021). Utility measures are used to evaluate the effectiveness of index insurance policies while providing equatable measures of welfare improvement across models and settings. Kenduiywo *et al.* (2021) define a ratio on the relative improvement of utility for policyholders between a theoretical “perfect” insurance contract with no basis risk and a contract with observed basis risk. Conradt *et al.* (2015) compare the relative improvement in utility using an out of sample prediction approach

for consistent comparison between linear and quantile regression models and different utility models.

I will determine the marginal willingness to pay for insurance for a given contract in a fishery by finding the premium loading factor that equates the utility of having insurance with the utility of not having insurance.

$$\mathbb{E}[U_{ni}(\pi)] = \mathbb{E}[U_i(\pi, I(\omega), \rho)] \quad (1)$$

The expected utility of not having insurance, $\mathbb{E}[U_{ni}]$, is the average utility over all years in the sample for any variable of interest π . The expected utility of having insurance, $\mathbb{E}[U_i]$, includes the contract payout schedule based on models built with weather variables, $I(\omega)$. The premium ρ is calculated as the expected value of the payout function times the premium loading factor. The premium loading factor will be varied to equate both states of the world to represent the marginal willingness to pay for insurance. Essentially, it reflects the highest premium a policyholder would be willing to pay for a contract with a certain level of basis risk.

Insurance contracts are specified by calculating payout functions ($I(\omega)$) based on independently measured weather variables. Contracts need productivity losses and thresholds to effectively compensate policyholders. Payouts will be issued when models predict deviations from defined thresholds. The three prediction models ($k \in \{\text{LR}, \text{LA}, \text{RF}\}$) are a linear regression (LR), a LASSO regression (LA), and a random forest (RF). Example thresholds include deviations from long run average (Equation 2), a j year moving average (Equation 3), or a biological threshold (Equation 4).

$$I(\omega, l, c) = \max(0, (\bar{\pi} \cdot c - \hat{\pi}_t^k(\omega)) \cdot l) \quad (2)$$

$$I(\omega) = \max(0, (\frac{1}{j} \sum_{i=n-j+1}^n \pi_t \cdot c - \hat{\pi}_t^k(\omega)) \cdot l) \quad (3)$$

$$I(\omega) = \max(0, (\pi_t(b_t) \cdot c - \hat{\pi}_t^k(b_t, \omega)) \cdot l) \quad (4)$$

Where k is the prediction model, l is the level of scale, c is the coverage, $\hat{\pi}_t^k(w)$ is the predicted fishing variable from ω weather variables. Scale is the amount of protection in unit loss a policyholder chooses to protect protect, and coverage is the deviation from the index that initiates a payout. For example, when $c = 1$, payouts are distributed anytime the index falls below the long run average in Equation 2. Lower values of c require larger disasters to trigger payouts. Both variables are often chosen by policyholders when purchasing insurance. Coverage is usually constrained as set choices e.g. $c \in \{0.7, 0.85, 1\}$, whereas scale is a continuous choice. The premium ρ is calculated as the expected value of the payout function times

the premium loading factor m (Equation 5). The premium loading factor is the variable that will be adjusted to equate utility states. Values above 1 indicate high willingness to pay for insurance. Values below 1 imply subsidies will be needed to stimulate demand.

$$\rho(\omega) = \mathbb{E}[I(\omega, l, c)]m \quad (5)$$

We use log utility measures as our base case for constant relative risk aversion, and use exponential utility as robustness checks to validate results. Fishing variables may vary extensively from fishery to fishery. We normalize utility by dividing all measured payouts and fishing variables by the maximum observed value in each fishery. Expected utility for a given fishery is the average utility over all years in the sample for any variable of interest π . Fishers choose insurance scale l to maximize expected utility over the time period for an offered insurance contract built on the models. Two coverage levels are provided exogenous to fishers, $c \in \{0.7, 1\}$ to test whether fishers are better off with more frequent payouts or disaster coverage.

$$\mathbb{E}[U_i] = \max_{l_t} \frac{1}{n} \sum_t^T u(\pi_t + I(\omega, l_t, c) - \rho(w)) \quad (6)$$

3 Data

This study attempts to cover breadth, not depth in possible indices. Each fishery has unique ecological characteristics that interact with environmental variables in different and non-linear ways. By studying a wide collection of fisheries and environmental variables we can uncover the potential feasibility of index insurance for fisheries holistically, and then further refine measures with ecologically sound models in future applications.

3.1 Fishery Data

Landings revenue, and participation data comes from the West Coast Fish data package (Free *et al.* 2022). It is a reconstruction of California Department of Fish and Wildlife catch data combined with PacFin receipts for Washington and Oregon. The last three years of data are updated from the CDFW Marine Fisheries Data Explorer (MFDE). Names are matched to each species within the West Coast Fish data package.

We select California fisheries with a minimum of 30 years of consecutive, non-confidential catch records at both the state and port-complex level. Unclassified catch records are dropped i.e. “Other Sharks” and similar categories. Fisheries with an average revenue from 2010-2019 greater than \$100,000 at the state and \$75,000 at port-complex level are analyzed. Twenty four fisheries at the state level and 50 fisheries at the port complex level meet these criteria.

These fisheries contain the most economically important fisheries in California and their mean values are shown in Table A1 and at the port complex in Table A2.

Fisheries have complex spatial dynamics. Agriculture has clear, quantifiable impacts of weather in grids that are well suited for index insurance. Drought on a single farm directly leads to crop loss for that farm. Whether there is sufficient spatial coverage to identify impacts down to an individual farm remains a challenge in agriculture (Dalhaus and Finger 2016; Leppert *et al.* 2021; Stigler and Lobell 2024). Fish and fishers can move thousand of miles in a given year, thus more consideration must be given to the location of weather impacts in fisheries. We spatially refine catch histories using the California CDFW fishing blocks records from the MFDE Data Explorer. Summarized catch histories of all landed fish within each block provide an average representation of effort for a given fishery. Spatial catch history is measured at both the state and port-complex level. The spatial location refines the location of environmental variables. Local weather is more likely to affect fishery productivity and catch than observations thousands of miles away.

3.1.1 Uncollected Data

While the fishery data from CDFW is the most comprehensive available, it is not without its limitations. There are other data needs that may be relevant for the analysis that I have not collected yet. All of these data sources could add to the uniqueness of fishery index insurance.

Catch per unit effort data may be valuable as it might be able to indicate loss more effectively than harvest measures. Aggregate harvest measures are sensitive to the total amount of fishers in a given year. The variation in fishers can lead to large swings in catch data that are not related to weather. Catch per unit effort data is often used as a direct calculation of the underlying biomass, which will be affected by the weather. Individual catch histories with records days fished would be the most ideal data source, but I would need approval through the CDFW to access.

Management covariates may be needed to indicate the relative influence of changes in regulation or weather to productivity. Quotas would be an excellent way to account for limitations on catch, but not all fisheries in the data set are managed through quotas. Instead other regulations could account for the weather independent effects. For example, Dungeness Crabs are managed through trap and bot limits that have changed over the years in addition to delays in season openings. A time series of management covariates would allow the models to separate institutional influences from weather influences. PacFin possesses current quota data and the stock assessments used in their determination.

Biomass estimates could be the most direct estimate of fishery productivity. Stock assessments are the most common way to estimate biomass, but they are not available for all fisheries. The Pacific Fisheries Management Council (PFMC) provides stock assessments for many of the

fisheries in the data set, but not all. Additionally, some measures of biomass are built off of CPUE so I would need to avoid double counting those effects in the model.

3.2 Environmental Data

Fisheries are highly sensitive to marine heatwaves and water temperature. Sea surface temperature is a natural variable to first consider in fisheries index insurance. Sea surface temperature data comes from the NOAA DHW data set that provides 5-km resolution of monthly temperature from 1985 to 2023. The 5-km grids are averaged within the nearest California fishing block to provide an annual time series of temperature for each fishery. Temperature is lagged from 1 to 3 years prior to account for residual impacts that carry over due to fishery biological dynamics.

Upwelling provides vital nutrients to stimulate primary productivity. The coast of California is a highly productive ecosystem due to its patterns of upwelling (Chelton *et al.* 1982; Huyer 1983). We capture upwelling through monthly observations of Coastal Upwelling Transport Index (CUTI) and Biological Effective Upwelling Transport Index (BEUTI). Both indices create measures of vertical movement in the mixed layer at 1 degree latitude intervals extending 75 km along the entire US West Coast (Jacox *et al.* 2018). The closest layer to the surface was used in this analysis as the correlation between surface index values and deeper index values are high. CUTI examines the physical measures of wind, Ekman transport, and cross-shore geostrophic transport to indicate the strength of upwelling in a given month. BEUTI adds nitrate concentration in its calculation to capture more biological effects of upwelling. Fishing blocks are matched to the nearest 1 degree latitude interval to provide a monthly time series of upwelling for each fishery. Seasonal strengths of upwelling are captured by averaging CUTI and BEUTI within each quarter of the year. Spring upwelling in early March and April are especially important to a wide array of fish species. Yearly average and amplitude values (the difference between minimum observed upwelling and maximum) are also calculated. These indices are the most temporally limited datasets in this analysis, only extending from 1988 to 2023.

The Habitat Compression Index measures the area extent of water below average temperatures thresholds along the US West Coast (Schroeder *et al.* 2022). Habitat compression is a measure of the spatial extent of cold water habitats that are important for fish species. The index is broken down into four distinct oceanographic regions ranging from 3.5 degrees to 5.5 degrees latitude in size with coverage out to 150 km offshore. We use the cumulative habitat compression index that sums the index value in each month to provide a yearly time series of habitat compression for each fishery. The cumulative index showed stronger correlations with biological productivity measures than monthly measures (Schroeder *et al.* 2022)

The final environmental variables are the Pacific Decadal Oscillation (PDO) and the El Nino Southern Oscillation (ENSO). Both indices are well known to affect marine ecosystems and

fisheries. Both indices are averaged over a given year. PDO data is taken from the PDO ERSST V5, and ENSO data is taken from the multivariate ENSO Index Version 2 (MEI.v2).

Summary statistics for the environmental data are presented in Table A3. In total, 74 fisheries with 35 years of catch data and 20 weather variables are spatially matched with annual coverage from 1988 to 2023.

4 Methods

We use three models to predict yearly fishing revenue and landings at state and port-complex levels. Linear models are used as the base model given its ubiquitous use in index insurance policies. We compare utility improvements with the adoption of more robust LASSO regression and random forest models.

In all class of models, the final utility maximization choice of coverage leverage is found through a box constrained quasi-Newton Method using the `optim` function in R. Choices are constrained to be non-negative. Premium schedules are found by the model output below the trigger values in Equation 2 and then averaged over the total fishery data. A root finder then determines what m will equate the expected utility of having insurance with the expected utility of not having insurance.

4.1 Linear Models

Perfect regression coefficients mimic the optimal choice of scale in index insurance contracts (Mahul 1999). Combined with the ease of implementation, linear models on single weather indices are the most common design choice for index insurance policies. They offer a basic starting place to consider the viability of fishery index insurance.

Yearly aggregated fishing variables are regressed on each environmental variable individually. Weather variables are spatially matched to the location of catch. We perform a 10 fold cross validation method to determine the best individual weather variable based on root mean square error (RMSE). To preserve the time series element of the data, we used a rolling split to partition the training and testing data. For example, the first fold contains the first 70% of data as training (1988-2011), and the last 30% as testing (2013-2023). The final date of the training set is extended in each fold until the year 2020 to create 10 folds. Models with the lowest average RMSE are selected and trained on the full set before being passed to the utility optimization procedure in Equation 6.

4.2 LASSO Regression

Least Absolute Shrinkage and Selection Operator (LASSO) regression is a popular regularization technique to assist model selection. It attempts to minimize the residual sum of squared errors through Ordinary Least Squares (OLS), but adds a penalty constraint on the absolute sum of selected coefficient values (Equation 7).

$$\hat{\beta}^{lasso} = \arg \min_{\beta} \left\{ \sum_{i=1}^n (y_i - \beta_0 - \sum_{j=1}^p \omega_{ij} \beta_j)^2 + \lambda \sum_{j=1}^p |\beta_j| \right\} \quad (7)$$

Where, y_i is our fishing variable, β the regression coefficients, n , the number of observations, p the number of predictors, and ω the total collection of weather variables. The λ is the penalty term that controls the amount of shrinkage. Models are trained using the `glmnet` package in R. The LASSO regression model is trained on 200 bootstrapped samples of the training data. The optimal λ is selected through a grid search method that selects the minimum RMSE. This choice is to ensure the most parsimonious model that still captures the most important weather variables. LASSO is particularly well suited for this research design as the absolute value of the penalty term shrinks coefficients to zero. Overfitting is a concern with so few observations in the initial training set; the shrinkage towards zero will help minimize this bias by reducing the parameter space.

4.3 Random Forests

While LASSO offers us the ability to simultaneously explore a wide collection of weather variables including lagged effects, it remains linear in its predictions. Random Forests are tree-based ensemble models that capture non-linear interactions through recursive partitioning. They are less sensitive to over fitting through the aggregation of many trees.

We tune two hyperparameters to create the best performing random forest for each fishery: The number of variables to consider at each split and the minimum number of observations in a leaf node. We use a grid search method to find the best hyperparameters based on RMSE through the year based cross validation method presented in the linear models. The final model is trained on the full dataset and passed to the utility optimization procedure in Equation 6.

4.4 Weather variables of importance

Machine learning algorithms are inherently “black boxes” that sacrifice interpretability for predictive accuracy. Fishers will be less likely to purchase complicated products that do not correspond to their experiences. Extracting the relative contribution of weather variables will assist translating products to fishers. Additionally, it can help ground-truth the chosen variables with previous biological modelling.

The cross-validation in the linear models provides a simple weather variable comparison. We calculate the frequency a given weather variable is chosen as the best performing linear model.

We use `vip` package in R to extract importance measures for both the LASSO and random forests ¹. Feature extraction will occur for each fishery product, and the importance of each will be normalized then aggregated in order to compare all features.

5 Preliminary Results

Before running the analysis on all models, we examine only the difference between linear models and random forests for the state wide fishery data. Preliminary results indicate there exists a fundamental difference between the training and testing datasets that is causing the random forest models to overfit. The linear models generally outperform the random forests in the testing data. The LASSO models are performing similarly to the linear models. The random forests are capturing the training data well, but are not generalizing to the testing data. The random forests are overfitting to the training data.

The lack of management covariates is probably the primary cause of the discrepancy. Random forests are robust to overfitting, but if there are fundamental differences in the testing and training sets, or omitted variables, random forests can lead to poor out of sample performance. Figure 2 provides a clear demonstration of how management could be affecting the models. Cabezón had little management and became overfished in the early 2000s. The Pacific Fisheries Management Council implemented a quota system in the mid 2000s to help the stock recover that remains in place today. The peak of catch in the late 1990s corresponded to the strong 1997 El Niño event with high sea surface temperatures. The random forests attempt to capture similar warm water events in the 2015-2016 blob and drastically overestimate the predicted catch. The random forests could not properly capture the management effects of the quota.

The performance of the utility model is biased by the underlying predicted model performance, but also demonstrate how average strike levels may be problematic.

6 Future Steps

- **Variable importance**

Once the performance of the models is refined with better data or alternative testing-training splits, the trained models will indicate which weather variables are the most important to predicting catch or revenue. For example, the linear models that had the best performance

¹Nathan note: I need to read more exactly how this package will extract between permutations or variance measures

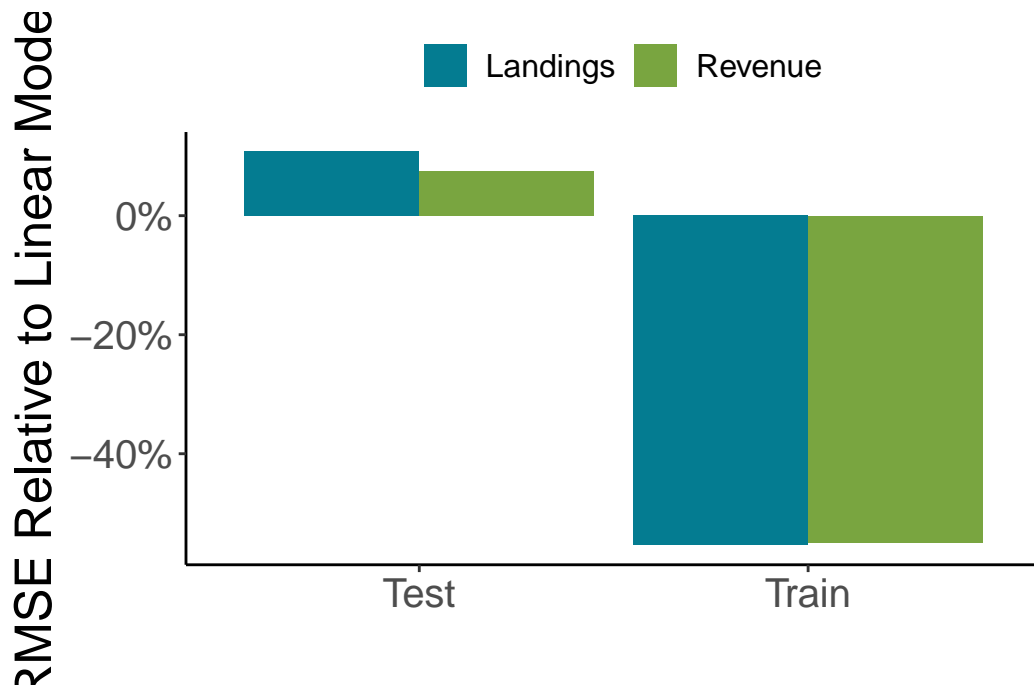


Figure 1: Root Mean Square Error performance is slightly better for linear models than random forests in the testing set. Random forests are overfitting both landings (blue) and revenue (green) in the training set.

Cabezon

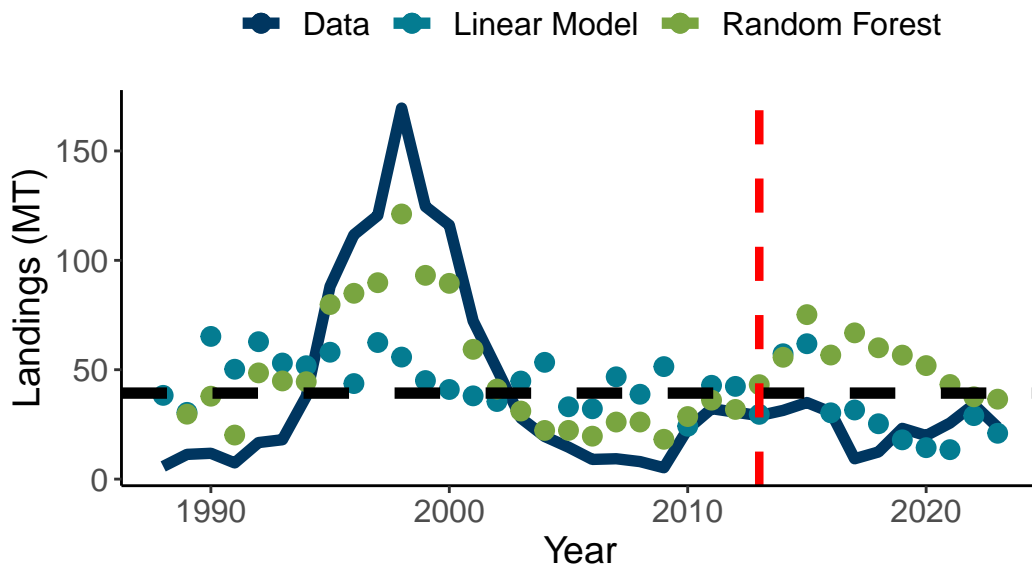


Figure 2: Measured Cabezon landings in metric tons from 1988-2023 (blue line). Predictions from random forestes (green points) perform well in the training sample compared to the linear model predictions (blue points). However, in the testing period post 2013 (dashed vertical red line) there is a distinct misalignment in predictions.

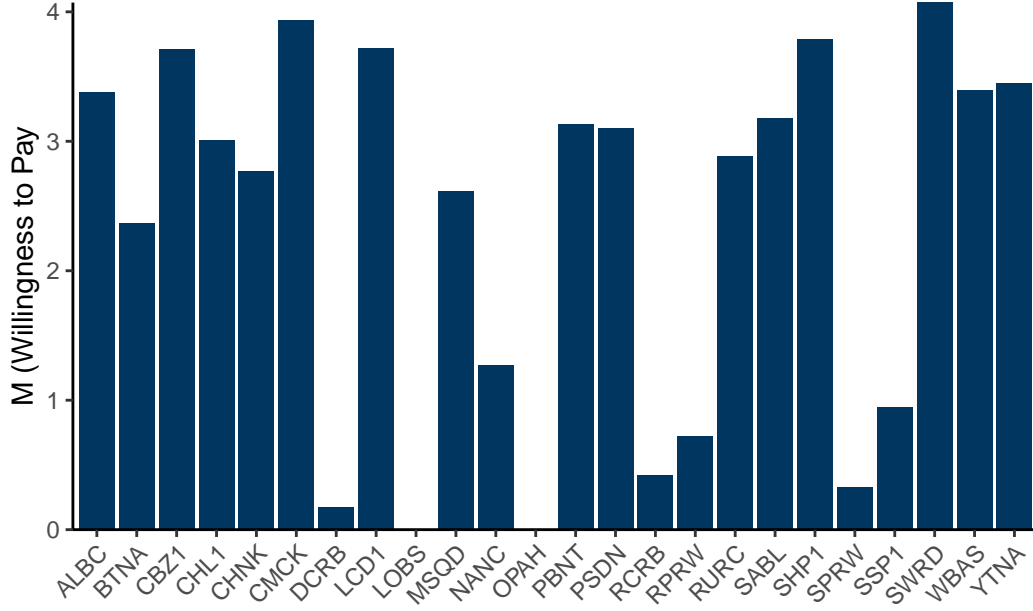


Figure 3: The marginal willingness to pay for insurance based on a linear model. Fishery species code on the x-axis.

in the cross validation show upwelling indices have the strongest relationship with catch. Analyzing the variable importance will help ground interpretation of the models for both fishers and fishery scientists. Describing black-box machine learning models will be an uptake barrier that can be alleviated by breaking down the model into more interpretable terms.

- **Use management measurements**

The performance of the models is concerning without accounting for fundamental differences in the training and testing sets. Once I have collected some indicator of management the prediction models will need to be retrained for each productivity measure. Perhaps even a simple dummy variable will help the models perform better without the explicit measures of management.

- **Trigger on biomass**

The weather data collected connects to biological productivity. Biomass could be a more direct measure of the underlying fishery health without the tenuous connections between weather and catch. The amount of fish is also an input so it will correlate tightly with catch. Estimates of biomass may not exist for all fisheries. I can trim down the analysis to just those that have stock abundance estimates over time. Catch per unit effort does serve as a proxy for biomass, so using CPUE for π may suffice.

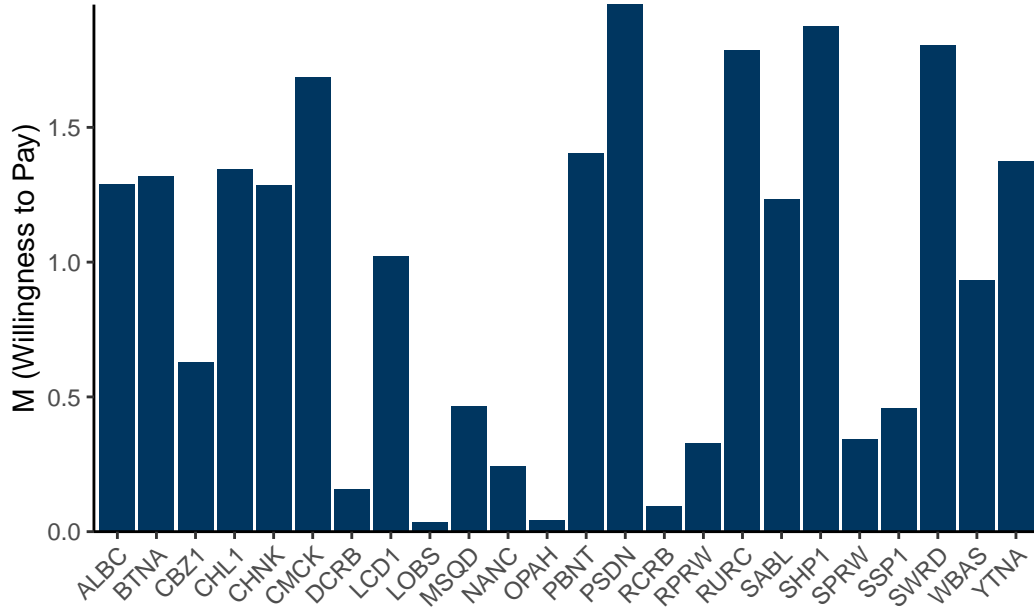


Figure 4: The marginal willingness to pay for insurance based on a random forest model. Fishery species code on the x-axis.

- **Compare between different contracts**

So far, I've only analyzed the utility with contracts based on long run average. AS described above, these may be problematic. The next step needs to use different trigger measures. However, without resolving the difference in the training and testing sets, it may be moot if the underlying models perform poorly.

- **Analyze payouts and insurance company profits**

The high demand present for some linear models stems from consistent payouts in recent years. Insurance companies may be forced to operate at significant losses if they have to make reoccurring payments. A robustness check would be to verify the premium an insurance company would need to charge to make a profit and compare to the willingness to pay estimated for the fishers.

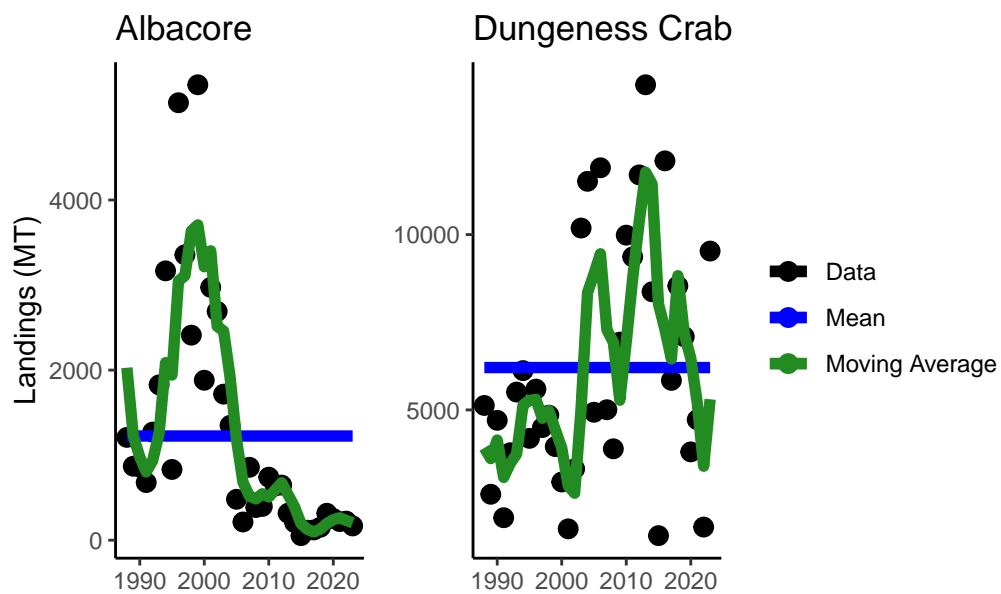


Figure 5: A moving average (green) may be able to more closely match the underlying biological dynamics of the fishery than the historical average (blue). The moving average of landings in metric tons for Albacore and Dungeness Crab.

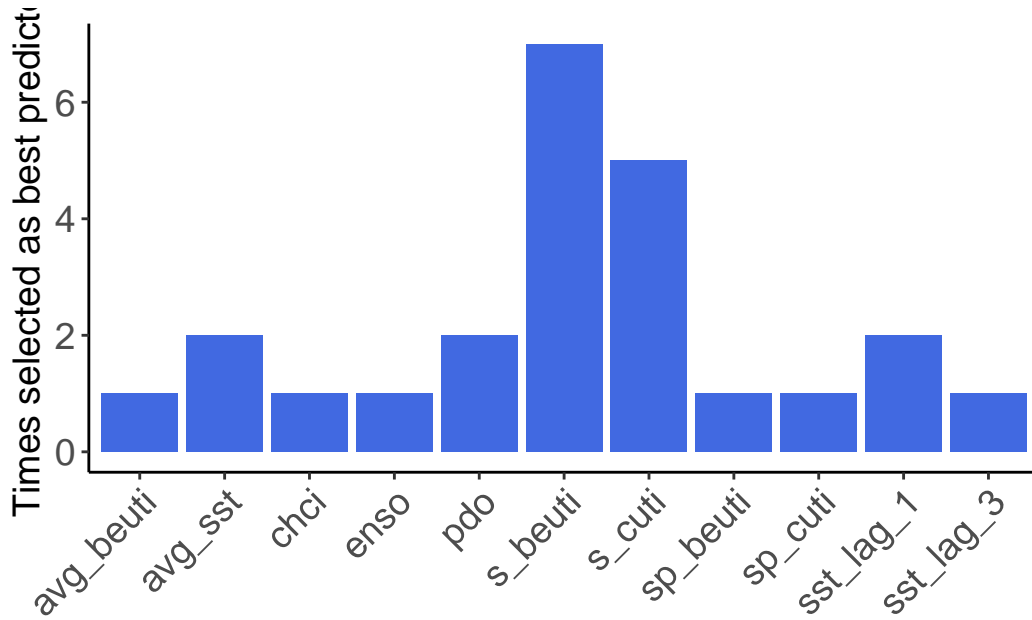


Figure 6: The number of times a weather variable was selected as the best predictor in the linear models.

A Appendix

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Table A1: Summary statistics of catch from 1988-2023 for California fisheries.

Species	Landings (mt)		Revenue (USD)		MT per Fisher		Number of Fishers	
	Mean	SD	Mean	SD	Mean	SD	Mean	SD
Albacore	2055.4	2646.5	\$3,574,984	4051538.3	5.3	3.1	305.5	307.7
Bluefin tuna	760.5	1131.1	\$1,118,896	1137192.5	9.3	12.1	76.3	46.1
California spiny lobster	314.6	69.6	\$7,752,729	5402314.5	1.7	0.6	202.0	39.8
Chinook salmon	1576.1	1291.7	\$10,383,616	7643140.9	1.5	1.1	1146.6	1022.2
Chub mackerel	11304.3	10208.0	\$1,944,342	1709450.2	80.2	59.5	133.9	78.6
Dungeness crab	5712.8	3334.7	\$29,115,960	23070113.4	11.6	7.9	534.5	121.7
Market squid	51363.6	34634.6	\$27,517,541	24057582.4	462.1	289.5	109.4	23.4
Nom. Cabezon	35.3	38.7	\$306,669	321874.4	0.2	0.1	199.0	115.1
Nom. Calif halibut	374.6	140.9	\$2,605,362	873979.0	0.9	0.3	427.1	87.8
Nom. California sheephead	46.7	38.8	\$342,986	231782.3	0.4	0.2	134.1	88.5
Nom. Lingcod	359.0	421.0	\$388,778	276453.1	0.4	0.3	675.4	374.0
Nom. Shortspine thornyhead	56.4	70.0	\$432,850	465662.0	0.5	0.6	116.2	35.6
Northern anchovy	7696.1	10157.9	\$860,871	643585.4	230.8	246.9	32.7	9.8
Opah	102.4	100.0	\$198,182	257248.5	2.5	3.6	71.0	46.8
Pacific bonito	1090.7	1694.6	\$544,835	886379.0	8.7	10.3	108.2	168.6
Pacific sardine	20665.4	22757.5	\$2,574,205	2614838.8	309.1	325.9	53.3	21.0
Red sea urchin	7947.2	6067.5	\$11,443,321	7684762.0	32.7	12.7	231.1	139.5
Ridgeback prawn	172.8	146.5	\$634,305	473336.8	7.7	5.7	26.1	12.6
Rock crab	620.2	146.3	\$1,750,541	597289.3	3.6	1.2	180.7	47.4
Sablefish	2753.2	1823.2	\$5,759,976	2768043.7	10.4	6.9	269.6	70.3
Spotted prawn	153.1	76.1	\$3,206,726	1989792.4	3.7	2.3	52.6	31.1
Swordfish	808.3	554.7	\$5,671,198	3401866.3	6.7	3.2	129.3	84.3
White seabass	109.2	76.4	\$635,876	429179.6	0.8	0.7	150.3	45.1
Yellowfin tuna	8853.7	17403.5	\$10,230,193	20959459.4	67.4	89.0	74.0	73.9

Table A2: Summary statistics of catch from 1988-2023 for California fisheries split between species and port complex

Species	Port	Landings (mt)		Revenue (USD)		MT per Fisher		Number of Fishers	
		Mean	SD	Mean	SD	Mean	SD	Mean	SD
Albacore	Eureka	355.0	377.4	\$672,866	634894.5	5.4	3.4	62.2	58.9
Albacore	San Francisco	175.8	368.5	\$322,734	639778.0	2.6	2.2	50.1	70.0
Bluefin tuna	San Diego	25.2	38.5	\$163,167	334643.0	0.8	1.7	29.2	20.9
Nom. Calif halibut	Los Angeles	47.2	36.2	\$303,840	172626.8	0.7	0.3	64.0	25.9
Nom. Calif halibut	San Diego	17.9	12.3	\$117,380	51099.8	0.6	0.3	27.5	10.3
Nom. Calif halibut	San Francisco	140.4	64.0	\$1,053,595	721430.4	1.3	0.6	109.7	28.0
Nom. Calif halibut	Santa Barbara	97.9	44.4	\$705,370	187398.0	1.1	0.3	87.1	22.3
Chinook salmon	Bodega Bay	329.6	309.3	\$2,258,504	1989960.2	1.0	0.6	357.0	263.0
Chinook salmon	Eureka	103.6	178.4	\$649,797	1000836.3	0.4	0.4	227.5	363.7
Chinook salmon	Fort Bragg	343.9	404.8	\$2,320,248	2367832.3	1.1	1.4	357.0	263.0
Chinook salmon	Monterey	313.7	272.7	\$1,896,626	1161527.4	0.9	0.6	354.0	201.8
Chinook salmon	Morro Bay	72.9	81.4	\$493,105	515982.2	0.6	0.5	107.3	72.5
Chinook salmon	San Francisco	493.7	348.0	\$3,339,479	2075261.9	1.2	0.8	434.7	265.1
Chub mackerel	Los Angeles	10320.4	9344.1	\$1,782,814	1561212.8	150.3	110.6	60.5	27.4
Dungeness crab	Bodega Bay	572.9	554.1	\$3,390,721	3510592.1	6.7	6.2	94.1	30.3
Dungeness crab	Eureka	3571.6	2232.7	\$16,072,727	12672636.9	13.9	10.7	286.2	106.0
Dungeness crab	Fort Bragg	247.3	191.4	\$1,376,466	1437305.5	5.6	3.9	42.5	8.0
Dungeness crab	Monterey	91.7	104.4	\$692,846	868788.1	3.0	2.8	27.8	7.7
Dungeness crab	San Francisco	1174.3	1276.2	\$7,090,167	8035652.0	7.2	6.3	144.1	39.2
California spiny lobster	Los Angeles	98.3	21.3	\$2,326,853	1471851.3	1.3	0.5	78.0	16.5
California spiny lobster	San Diego	99.0	22.5	\$2,179,454	1244460.4	1.5	0.6	72.0	21.4
California spiny lobster	Santa Barbara	116.8	49.4	\$3,239,108	2893952.5	2.0	0.9	61.2	11.8
Market squid	Los Angeles	16986.5	14774.8	\$8,750,561	8927983.2	278.8	175.1	52.6	21.5
Market squid	Santa Barbara	24534.0	20112.5	\$12,790,364	13303266.4	412.4	269.4	53.5	21.0
Opah	San Diego	58.6	66.2	\$121,629	182250.9	2.5	3.6	36.4	20.5
Rock crab	Los Angeles	101.5	105.4	\$259,159	196512.5	2.3	1.7	41.2	13.6
Rock crab	Morro Bay	75.9	65.7	\$183,343	124726.4	3.5	2.4	22.0	15.0
Rock crab	San Diego	67.1	35.0	\$163,988	79313.9	2.3	0.9	30.1	13.0
Rock crab	Santa Barbara	344.2	173.1	\$1,036,433	641375.7	5.3	3.0	71.0	16.4
Ridgeback prawn	Santa Barbara	155.5	139.5	\$562,758	453595.2	7.9	5.6	19.7	7.6
Red sea urchin	Los Angeles	1229.8	1020.2	\$1,890,768	1362341.7	18.0	4.8	63.6	41.9
Red sea urchin	Santa Barbara	3971.6	2451.7	\$6,081,764	3987202.8	32.0	14.5	130.3	83.3
Sablefish	Bodega Bay	90.7	90.1	\$186,193	166543.1	4.1	2.7	17.9	9.1
Sablefish	Eureka	886.0	607.5	\$1,634,560	691917.3	12.9	5.4	68.2	30.9
Sablefish	Fort Bragg	574.5	368.0	\$1,221,487	633518.2	13.5	13.6	45.7	20.8
Sablefish	Los Angeles	227.5	552.3	\$321,915	461035.2	13.6	50.7	18.1	9.5
Sablefish	Monterey	316.0	204.2	\$642,984	415849.4	7.2	6.1	48.7	21.7
Sablefish	Morro Bay	229.8	185.6	\$679,713	939824.4	8.1	5.4	31.5	16.4
Sablefish	San Diego	25.9	22.7	\$137,336	144319.5	2.1	2.3	15.1	6.5
Sablefish	San Francisco	336.2	320.6	\$559,985	268977.2	6.8	4.6	45.0	19.5
Sablefish	Santa Barbara	63.3	74.2	\$352,529	476423.6	3.0	2.9	17.3	6.7
Nom. California sheephead	San Diego	13.9	11.3	\$119,436	84159.3	0.7	0.5	23.4	13.9
Nom. California sheephead	Santa Barbara	20.9	22.7	\$134,335	125858.2	0.4	0.4	48.1	30.8
Spotted prawn	Santa Barbara	53.4	29.2	\$1,116,708	873801.1	3.5	2.4	18.7	9.5
Swordfish	Los Angeles	315.9	360.1	\$2,072,740	1972406.2	5.8	5.2	60.9	49.1
Swordfish	San Diego	178.4	131.0	\$1,416,146	905587.3	3.0	1.3	57.3	31.6
Swordfish	Santa Barbara	66.1	91.5	\$466,637	585529.3	2.3	1.9	25.8	26.4
White seabass	Los Angeles	31.5	30.9	\$146,996	111988.1	1.2	1.0	31.9	16.6
White seabass	San Diego	13.1	26.4	\$73,176	87486.8	0.6	0.7	19.8	9.4
White seabass	Santa Barbara	54.3	43.1	\$331,544	269526.5	1.1	1.0	51.4	14.1

Table A3: Summary statistics of environmental variables from 1988-2023 for California fisheries.

Weather Index	Mean	SD	Temporal Resolution	Spatial Resolution	Source
CUTI Amp	1.3	0.6	Monthly	1 degree latitude	Jacox et al., 2018
CUTI Avg	0.5	0.3	Monthly	1 degree latitude	Jacox et al., 2018
CUTI Fall	0.2	0.3	Monthly	1 degree latitude	Jacox et al., 2018
CUTI Summer	0.6	0.3	Monthly	1 degree latitude	Jacox et al., 2018
CUTI Spring	0.7	0.4	Monthly	1 degree latitude	Jacox et al., 2018
CUTI Winter	0.2	0.3	Monthly	1 degree latitude	Jacox et al., 2018
BEUTI Amp	15.8	10.6	Monthly	1 degree latitude	Jacox et al., 2018
BEUTI Avg	4.1	3.9	Monthly	1 degree latitude	Jacox et al., 2018
BEUTI Fall	1.0	2.0	Monthly	1 degree latitude	Jacox et al., 2018
BEUTI Summer	4.2	4.7	Monthly	1 degree latitude	Jacox et al., 2018
BEUTI Spring	9.1	7.9	Monthly	1 degree latitude	Jacox et al., 2018
BEUTI Winter	1.9	4.2	Monthly	1 degree latitude	Jacox et al., 2018
Cummulative Habitat Compression Index	4.8	2.3	Yearly	1 degree latitude	Integrated Ecosystem Assessment
Average Sea Surface Temperature	14.2	2.0	Monthly	5x5 km	NOAA Coral Bleaching Degree Heating Week
Sea Surface Temperature Lag 1 Year	14.2	2.0	Monthly	5x5 km	NOAA Coral Bleaching Degree Heating Week
Sea Surface Temperature Lag 2 Years	14.2	2.0	Monthly	5x5 km	NOAA Coral Bleaching Degree Heating Week
Sea Surface Temperature Lag 3 Years	14.2	2.0	Monthly	5x5 km	NOAA Coral Bleaching Degree Heating Week
Sea Surface Temperature Lag 4 Years	14.2	2.0	Monthly	5x5 km	NOAA Coral Bleaching Degree Heating Week
ENSO	-0.1	0.7	Monthly	Regional	MEL.v2
Pacific Decadal Oscillation	-0.3	1.0	Monthly	Regional	PDO ERSST V5

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