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The Effects of Climate Change on Tradeoffs Among Forest Ecosystem Services

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Abstract

The Effects of Climate Change on
Tradeoffs Among Forest Ecosystem Services

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DRAFT

Forests provide a bounty to humans through ecosystem services such as wildlife habitat, recreation, and water and air purification. Forest managers seek to maximize the provision of ecosystem services and often do so for multiple ecosystem services simultaneously. While many studies predict that climate change will impact forests' ability to provide ecosystem services, no research has addressed the question of how climate change will impact the joint provision of ecosystem services. I address this question here in an attempt to better understand how the relationships between ecosystem services will change with climate. For example, how much additional fire hazard must be assumed in order to maintain an amount of habitat for a particular species. To study this question, I consider the growth of a forested area in the Deschutes National Forest under three climate scenarios of varying intensity. This area provides three competing ecosystem services whose joint provision is assessed under each of the climate scenarios: northern spotted owl habitat, water quality, and resistance to wildfire.

I find that ...

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GLOSSARY

CLIMATE PROJECTION: The IPCC defines a climate projection as a model-derived estimate of future climate. *See* CLIMATE SCENARIO[51].

CLIMATE SCENARIO: The IPCC defines a scenario as a coherent, internally consistent and plausible description of a possible future state of the world. Herein, I use this term interchangeably with CLIMATE PROJECTION, since climate projections often underlie climate scenarios [51].

CLUSTER: Here, a set of contiguous forest stands whose combined area exceeds 200 ha

ECOSYSTEM SERVICE: Benefits that people receive from ecosystems, divided into four categories: supporting, provisioning, regulating and cultural [5]. Examples include food, soil formation, water purification, carbon storage, recreation, and education.

PARETO EFFICIENT: A solution to a multi-objective mathematical program is said to be Pareto efficient if no component of the solution can be improved without compromising at least one other component.

STAND DENSITY INDEX (SDI): Reineke's Stand Density Index is a measure of the stocking of a forest stand. *See* [56].

TRADEOFF: The sacrifice of achievement in one objective in order to achieve more in another.

ACKNOWLEDGMENTS

DRAFT

Thank you to all who contributed to my earning this degree.

DEDICATION

DRAFT

To ma femme and my family

Chapter 1

ASSESSING CHANGES IN TRADEOFFS AMONG ECOSYSTEM SERVICES IN THE DESCHUTES NATIONAL FOREST

1.1 Introduction

Forests play an important role in global ecological, social, and economic processes. They provide ecosystem services such as carbon storage, purification of water and air, wildlife habitat, recreation opportunities, and generate raw materials for goods such as food and lumber [19]. In managed forests, the extent to which forests provide these services depends in part on management practices. Forest management seeks to ensure the sustained provision of these ecosystem services [1].

Like other ecosystems, forests will undergo changes as a result of the changing climate. Researchers anticipate new spatial distributions of tree species [37], increased sediment delivery to streams [32], and increasing disturbance regimes such as wildfires, droughts, and insect infestations [70]. As this transformation occurs, the ability of forests to provide ecosystem services will change. Increased frequency of disturbance regimes will impact the ability of forests to store carbon [7] and provide wildlife habitat [47]. Water supplies that rely on forests' filtration capabilities may be impacted by the rising sediment levels predicted by Goode *et al.* [32].

As a result, forest planners may need to consider the effects of the changing climate. Linder [42] has shown that optimal forest management strategies vary depending on which future climate scenario is assumed. To determine whether and how forest planners need to adapt to climate change, many studies have addressed the effects of climate change on isolated ecosystem services [70][7][47]. However, forests provide these ecosystem services

in concert with one another (see, for example, Tóth and McDill [68]), and little research has been done to understand how climate change impacts the relationships among or the joint provision of ecosystem services. I posit that improvements in this understanding would enable better and more informed forest management decisions.

To improve our understanding of the impacts of climate change on the joint provision of ecosystem services, I propose the use of multi-objective mathematical optimization. Multi-objective optimization has been used in numerous forest management studies involving multiple ecosystem services [59][68][22][8]. The application of multi-objective optimization in these studies reveals the relationships among ecosystem services. When the studies consider competing or “bundled” ecosystem services, multi-objective optimization produces a set of Pareto optimal management alternatives that reveal the trade-offs among the ecosystem services and allow for the quantification of conflict. In such cases, the competition among the ecosystem services prevents the simultaneous maximal provision of all ecosystem services. Instead, trade-offs must be made, where the provision of one or multiple ecosystem services must be reduced in order to increase provision of another. Timber revenues and wildlife habitat provide an example of bundled ecosystem services. In these cases, the optimal management strategy involves finding the best compromise alternatives.

Heterogeneous forcing from climate change on forest ecosystem services will likely also change the trade-off relationships among ecosystem services. This has the potential to disrupt the balance among ecosystem services that define the compromise alternatives from which forest planners choose. In other words, under different climatic conditions, decisions that would once have resulted in a preferred balance among ecosystem services may no longer do so. Without consideration of climate change, forest management plans may restrict the potential for forests to provide ecosystem services most effectively. For instance, would the same amount of old-growth forest have to be converted to fire breaks to ensure a given reduction in fire hazard under various climate projections? Would carbon sequestration cost the same in terms of foregone timber revenues under a changing climate as it does today? To quantify the impacts of climate change on the trade-off relationships among and the joint

provision of forest ecosystem services, I employ multi-objective optimization on a case study of the Drink Area, located in the Deschutes National Forest.

In section 1.2 I define conflict and relationships among ecosystem services and then describe a case study in which these are evaluated. Section 1.3 reports and discusses the results. Finally, I summarize and draw closing remarks in section 1.4.

1.2 Methods

Bundled ecosystem services conflict with one another, and I argue that understanding the change in this conflict over time will enable forest planners to make more informed management decisions. Here, I provide formal definitions of what it means for ecosystem services to conflict, how to measure changes in conflict, and how a change in conflict impacts trade-off relationships.

1.2.1 Terminology

Multi-objective problem and solutions Consider the M -objective optimization problem

Maximize

$$\mathbf{f} = [f_1(\mathbf{x}), f_2(\mathbf{x}), \dots, f_M(\mathbf{x})]^T \quad (1.1)$$

subject to

$$\mathbf{x} \in X \quad (1.2)$$

with *objective functions* $f_i(\mathbf{x}), i \in \{1, \dots, M\}$ and feasible *decision vectors* (or *solutions*) $\mathbf{x} \in \mathbb{R}^n$. n is the number of decision variables in the optimization problem. A set of equality and inequality constraints determine the *feasible decision vector space* X . Each objective function $f_i : \mathbb{R}^n \mapsto \mathbb{R}$ maps decision vectors to scalars in \mathbb{R} . The vector objective function $\mathbf{f} : X \mapsto \mathbb{R}^M$ maps the feasible decision vector space to the *objective space* \mathbb{R}^M .

I define \mathcal{M} as the set of all M objective functions $\mathcal{M} = \{f_1, \dots, f_M\}$. $\mathcal{L} \subset \mathcal{M}$ defines a sub-dimensional objective set that includes only some of the objectives $f_i \in \mathcal{M}$.

Dominance and frontiers A solution \mathbf{x}^1 is said to *dominate* another solution \mathbf{x}^2 ($\mathbf{x}^1 \succ \mathbf{x}^2$) if

$$\forall f_i \in \mathcal{M} f_i(\mathbf{x}^1) \geq f_i(\mathbf{x}^2) \text{ and } \exists f_i \in \mathcal{M} : f_i(\mathbf{x}^1) > f_i(\mathbf{x}^2) \quad (1.3)$$

The set of non-dominated solutions to the multi-objective problem (1.1) and (1.2) is referred to as the *Pareto-optimal set* $P = \{\mathbf{x} \in X | \nexists \mathbf{y} \in X : \mathbf{y} \succ \mathbf{x}\}$.

The *Pareto-optimal frontier*, the *efficient frontier* or, simply, the *frontier* Z is the corresponding set of M -dimensional *objective vectors* $\mathbf{z} = [f_1(\mathbf{x}), f_2(\mathbf{x}), \dots, f_M(\mathbf{x})]$. That is,

$$Z = \{\mathbf{z} = [f_1(\mathbf{x}), \dots, f_M(\mathbf{x})] \mid \mathbf{x} \in P\} \quad (1.4)$$

Objective vectors components are referred to in subscripts:

$$\mathbf{z} = [z_1, z_2, \dots, z_M] \quad (1.5)$$

Trade-offs The trade-off between two objective vectors \mathbf{z}^1 and \mathbf{z}^2 is the vector of differences in their objective achievements

$$\tau_{\mathbf{z}^1, \mathbf{z}^2} = [z_1^2 - z_1^1, z_2^2 - z_2^1, \dots, z_M^2 - z_M^1] \quad (1.6)$$

Note that $\mathbf{t}_{\mathbf{a}, \mathbf{b}} = -\mathbf{t}_{\mathbf{b}, \mathbf{a}}$. Because all solutions are assumed to be Pareto efficient, not all components will be of the same sign.

Presence of conflict I determine whether a set of objectives $\mathcal{L} \subset \mathcal{M}$ is *stacked* or *non-conflicting* by checking if $\forall z^1, z^2 \in Z, i \in \mathcal{L}$

$$(z_i^1 \geq z_i^2) \Rightarrow (z_j^1 \geq z_j^2) \quad \forall j \in \mathcal{L}, j \neq i \quad (1.7)$$

This definition is functionally equivalent to that suggested by Zitzler (2009) [9] and Purshouse and Fleming (2003) [54].

If the above holds, then the objectives are stacked. Else, there is conflict among the objectives, and the set \mathcal{M} defines a bundle. Note that the bundle may still be reduced by

identifying subsets of \mathcal{M} that are stacked. If the statement (1.7) holds for some pair of objectives $i, j \in \mathcal{M}$, $j \neq i$, then the objectives are stacked, and one may be removed from the set \mathcal{M} in order to define a smaller bundle.

Measure of conflict Many authors have previously measured conflict (CITEABUNCH-LIKEZITZLERANDPURSHOUSEANDGALLEBERLING). Here, we Take all pairs of lines on the PCoords plot. How many of them have same sign on slope vs how many have diff sign on slope? Zero slope = skip that pair in determining test?

1.2.2 Study system

To quantify the impacts of climate change on the relationships among and the joint provision of bundled forest ecosystem services, I employ multi-objective optimization on a study system in the Deschutes National Forest known as the Drink Planning Area. The Drink Area is a 7056 ha area on the east slopes of the Cascade Mountain Range (see Figure 1.1). Like many forests, the Drink is managed for the simultaneous provision of multiple ecosystem services. The ecosystem services were selected through a process involving stakeholder meetings and discussions regarding compatible ecosystem services. For the current work, a subset of the chosen ecosystem services was selected for study. While the USFS manages for many services simultaneously, many of the services are “stacked” rather than bundled, meaning that the ecosystem services are not in conflict. That is, maximizing the provision of one ecosystem service simultaneously maximizes another. Stacked ecosystem services need not all be studied with multi-objective optimization, since the selection and maximization of one ecosystem service entails the maximization of all in the stack. For that reason, we disregard non-conflicting ecosystem services and select a minimal bundle on which to employ multi-objective optimization. Those that do not conflict can be stacked post-optimization.

The minimal bundle selected for this case study is a set of three ecosystem services. The Forest Service seeks to ensure their sustained provision, and this may require an understanding of how these ecosystem services are impacted, jointly and independently, by climate

change.



Figure 1.1: Overview of the study system, the Drink Planning Area (in purple), consisting of 7056 ha in the Deschutes National Forest.

The first of these ecosystem services is the provision of habitat for the northern spotted owl (NSO) (*Strix occidentalis caurina*). The NSO is a common, if controversial, indicator species in Pacific Northwest forests. Because of the availability of dense old growth forest in the Drink, approximately 43% of the area serves as habitat for the NSO (see Figure 1.2). The USFS is required to protect this species as it is listed as threatened and therefore protected by the Endangered Species Act of 1973 [12].

The second ecosystem service the USFS seeks to provide is protection from high severity wildfire. This protection is achieved by way of silvicultural treatments applied to designated treatment areas (forest stands) across the Drink. The types of treatments assigned are

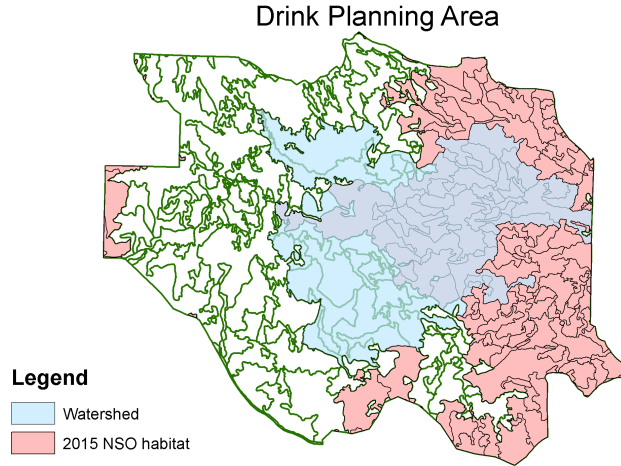


Figure 1.2: Location of the municipal watershed and the suitable NSO habitat in the Drink area at the beginning of the planning horizon (2015). Interior polygons are the 303 management units.

defined in the appendix, §B. The efficacy of these treatments is measured by comparing the fire hazard rating of the stand before and after treatment. The fire hazard rating is described in more detail later and is summarized in Table 1.1. Implementing the silvicultural treatments to reduce the fire hazard rating of the Drink is critical not only because it protects the habitat of the NSO, but also because the Drink Area houses the municipal watershed for the cities of Bend, OR and Sisters, OR (see Figure 1.2). Wildfires pose a threat to these cities' water supply, because wildfires can cause soil water repellency, surface runoff, and debris torrents [36] which would degrade the quality of the watershed. In addition, the Drink has never before undergone fuels treatments, which increases the expected severity of a fire should one occur.

Finally, the Forest Service seeks to provide a watershed with minimal sediment content. While the silvicultural treatments intend to provide long-term protection of the watershed, the implementation of the treatments has the potential to introduce short-term increases in sediment delivery [50]. This is expected to be especially true in the Drink Area, where local Forest Service staff have noted that the watershed is unusually susceptible to spikes in

sediment delivery as a result of foot traffic and activities that occur within the watershed.

The changing climate will likely impact the provision of these ecosystem services and their relationships with one another. The extent of these changes will depend on the severity of the realized climate change. Thus, to understand the potential impacts, multiple climate change scenarios representing a range of severities must be considered. The following section describes the climate scenarios considered in this case study.

1.2.3 Climate Scenarios Considered

In their assessments on the changing climate, the Intergovernmental Panel on Climate Change (IPCC) uses a scenario-based approach, considering many models of future climates from research groups around the world. They make no attempt to predict which of the future climates is most likely or to quantify the probability of realization of any one scenario. This same scenario-based approach is employed here in studying the potential impacts of climate change on trade-off relationships among bundled ecosystem services.

Here, the alternative future climates considered are climate scenarios from the first working group (WG1) of the IPCC’s Fifth Assessment (AR5) [35]. Given the large number of potential future climates considered by the IPCC (see the list of experiments considered in AR5 [24]) combined with the computational complexity involved in the study of each one, I selected a subset of three future climate scenarios for this analysis. Hereafter the scenarios are referred to as “None”, “Ensemble RCP 4.5”, and “Ensemble RCP 8.5”.

The first scenario, “None”, is the assumption of no climate change. While the number of studies incorporating climate change is increasing, this is still the assumption used for many modern studies such as Schroder (2013) [59], from which this study is derived. Because it has served as the basis for many studies and assumes a static climate resembling today’s, the “None” climate scenario serves as a control against which to compare the other two future climate scenarios.

As their names suggest, the second and third scenarios are ensembles. Each ensemble is an assembly of 17 global circulation models (GCMs) used in IPCC AR5. The selection of

component GCMs in the ensembles was performed by the USFS’s Climate-Forest Vegetation Simulator (FVS) [23] team. The list of the 17 scenarios included in the ensemble can be found in Crookston (2016) [15]. Each component GCM has a corresponding climate surface which contains a vector of 35 climate parameters at over 11,000 global locations for three time periods. The climate surfaces for the ensembles were created by averaging the values of all component GCMs for each climate parameter and each time period for each location. The result is a climate surface that, while temporally sparse, is spatially robust. Such a configuration is well-suited for use in the Drink Area given the area’s variance in elevation and slow vegetation growth.

The two ensembles are comprised of the same 17 GCMs, but the assumed representative concentration pathways (RCP) in the component GCMs differ. The RCP indicates the additional radiative forcing in W/m^2 above pre-industrial levels, with higher values of forcing indicative of more severe climate change. The GCMs in the Ensemble RCP 4.5 scenario assume 4.5 W/m^2 of additional radiative forcing, and the GCMs in the Ensemble RCP 8.5 scenario assume 8.5 W/m^2 of additional radiative forcing.

These three chosen scenarios represent a range of predicted climate change severity, from a $0^\circ C$ warming by the year 2100 under the “None” scenario to a $2.6 - 4.8^\circ C$ warming under RCP 8.5 [35]. Comparing the trade-off relationships among the ecosystem services under this range of climate change severities allows for the quantification of the impacts of climate.

1.2.4 *The Multi-objective Optimization Model*

In order to determine trade-off relationships among ecosystem services under each climate scenario, I employ multi-objective spatial optimization. This section describes the multi-objective zero-one mathematical program to optimize the joint provision of ecosystem services in the Drink Area. The model operates over an 80-year planning horizon (2015-2095) in which it assigns silvicultural treatments to forest stands that will be performed in the first or second 20 year period (2015-2035 or 2035-2055). Determining which treatment type to apply to a stand was done *a priori* and is entirely dependent on silvicultural characteristics; the rules

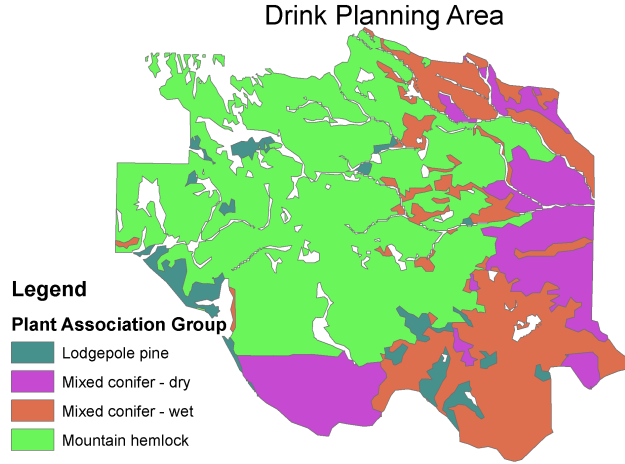


Figure 1.3: Plant association groups in the Drink Planning Area that were selected for potential treatments. Other plant association groups exist in the area but were not considered for treatment.

governing this assignment of treatment type can be found in the appendix, §B.

The model minimizes the fire hazard rating of the Drink at the end of the 80-year planning horizon, maximizes the area of NSO habitat at the end of each planning period, and minimizes the short-term spikes in sediment delivery resulting from the application of silvicultural treatments.

Treatments are assumed to be performed at the midpoint year in the treatment period, years 2025 and 2045 for the first and second periods, respectively. A schematic of the planning horizon including the time of these events is shown in Figure 1.4.

Notation

The following notation is used throughout the model:

Parameters

- $i \in I$: the set of all 303 forest stands comprising the Drink area

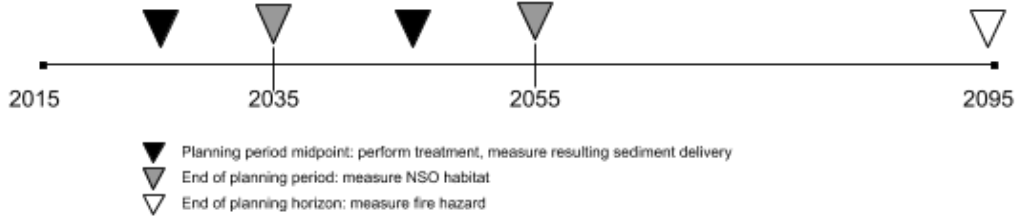


Figure 1.4: The planning horizon used in the analysis spans the 80 year period from 2015 to 2095. Treatments may be performed in the first period, the second period, both, or neither. Treatments are assumed to be performed at the mid-point years of each period (black triangles). Sediment delivery is measured on treatment years. Stands' suitability for NSO habitat is measured at the end of the planning periods (gray triangles), and stands' fire hazard ratings are measured at the end of the planning horizon (white triangle).

- $r \in R$: the set of treatment schedule prescriptions:

$$r = \begin{cases} 1 & \text{treatment applied in the first period (2015-2035)} \\ 2 & \text{treatment applied in the second period (2035-2055)} \\ 3 & \text{treatment applied in both periods} \\ 0 & \text{no treatment applied in either period} \end{cases}$$

- $F_{i,r}$: the area-weighted fire hazard rating of stand i at the end of the planning horizon if prescribed to treatment schedule r
- $I_{\omega,t}$: the set of stands that can qualify as NSO habitat at the end of planning period t
- a_i : the area of stand i
- e : the discount factor applied to NSO habitat that is less than 200 ha in size
- $j \in R_{i,t}$: the set of treatment schedules such that stand i qualifies as NSO habitat in planning period t
- $s_{i,t}$: the contribution in tons of sediment delivered from performing fuel treatments on stand i in planning period t

- $c \in C$: the set of all clusters of stands whose combined area exceeds 200 hectares
- $i \in D_c$: the set of all stands that comprise cluster c
- $c \in C_i$: the set of all clusters that contain stand i
- A : the maximum area in hectares that may be treated in either planning period
- ℓ, u : the lower and upper bounds, respectively, on the relative fluctuation in the area treated in periods 1 and 2

Decision Variables

$$x_{i,r} = \begin{cases} 1 & \text{if stand } i \text{ is prescribed to treatment schedule } r \\ 0 & \text{otherwise} \end{cases}$$

Indicator Variables

- $q_{c,t} = 1$ if all stands in cluster c qualify as NSO habitat in planning period t and $q_{c,t} = 0$ otherwise
- $p_{i,t} = 1$ if in planning period t stand i is part of a cluster c such that $q_{c,t} = 1$; $p_{i,t} = 0$ otherwise

Accounting Variables

- S_t : the contribution in tons of sediment delivered from performing fuel treatments in planning period t
- O_t : the amount of NSO habitat in hectares at the end of planning period t
- H_t : the area in hectares treated in planning period t

Parameterization

The model was parameterized as follows:

- $F_{i,r}$: the metric for fire hazard rating used in this analysis originated in the work by Schroder *et al.* [59]. This metric was developed for the Drink area. It uses fire characteristics from Anderson’s fuel models [4] to assign a fire hazard rating. I expanded the rating system to include fuel models not present in Schroder *et al.* See Table 1.1.

The stands’ fuels and vegetation characteristics to determine the fire hazard rating were generated using the US Forest Service’s Climate-Forest Vegetation Simulator (FVS). Input vegetation data to Climate-FVS came from the 2012 GNN structure map (<http://lemma.forestry.oregonstate.edu/data/structure-maps>) from Oregon State University’s Landscape Ecology, Modeling, Mapping & Analysis (LEMMA) group. Plots from the LEMMA database were mapped to the stands in the Drink area in order to produce tree and stand lists. These lists were used with Climate-FVS to simulate the stands’ vegetation and fuels characteristics forward for the duration of the planning horizon under each climate scenario. Input climate data for Climate-FVS was obtained through the Climate-FVS climate data server [16].

- $I_{\omega,t}$: the set of stands that qualify as NSO habitat at the end of a planning period t are those that meet the following three criteria, as specified by the USFS:
 1. elevation less than 1830 m
 2. the presence of trees with DBH no less than 76 cm
 3. canopy closure of at least 60%

The elevation requirement was checked using a digital elevation model from the US Department of Agriculture’s GeoSpatial Data Gateway; canopy closure and large tree requirements were determined using the simulated vegetation characteristics output from Climate-FVS.

To account for the NSO's large habitat requirements, stands must also be members of a cluster exceeding 200 ha in size, the entirety of which meets the aforementioned NSO habitat criteria. Stands not part of such a cluster have their contributions to owl habitat discounted by a factor of e .

- e : the discount factor for sub-200 ha NSO habitat was set to $e = 0.5$ following the convention used in Schroder *et al.* [59].
- $j \in R_{i,t}$: each stand-treatment schedule combination is evaluated at the end of each planning period to determine its suitability as NSO habitat. Treatment schedules for which stand i meets the NSO habitat criteria at the end of treatment period t become members of the set $R_{i,t}$.
- $s_{i,t}$: the contributions of sediment delivery from treatment of stand i in period t were determined using the Watershed Erosion Prediction Project (WEPP) online GIS tool [29]. This tool takes as input soil textures, treatment types, duration of simulation, and custom climate data. Soil texture data for the Drink area was obtained from the USDA's Soil Survey Geographic (SSURGO) database, treatment types are those specified in §B, and the years of simulation correspond to the treatment years in the model's planning horizon (2015-2095). The custom climate data are those described above for use with Climate-FVS, obtained through the Climate-FVS data server.
- C, D_c, C_i : the formulation of clusters was performed *a priori* according to the algorithm used in Rebain and McDill (2003) [55]. The enumerated clusters can then be used to immediately determine cluster members D_c and cluster-stand relationships C_i .
- A : the maximum area that may be treated in either planning period was defined to be 6000 acres, or approximately 2428 ha

- ℓ , u : the relative fluctuation in the area treated in periods 1 and 2 was defined to be 20%. That is, the lower bound $\ell = 0.8$, and the upper bound $u = 1.2$.

Fuel Model	Fire Hazard Rating	Group	Flame length (m)	Rate of spread (m/hr)	Total fuel load (tons/ha)
4*	5	Shrub	5.79	1508.76	32.12
5	4	Shrub	1.22	362.10	8.65
8	1	Timber	0.30	32.19	12.36
9*	2	Timber	0.79	150.88	8.65
10	2	Timber	1.46	158.92	29.65
11*	2	Logging Slash	1.07	120.7	28.42
12	4	Logging Slash	2.44	261.52	85.50
13	5	Logging Slash	3.20	271.58	143.57

Table 1.1: Fire hazard rating system used here, originally employed by Schroder *et al.* [59].

* denotes fuel models not present in Schroder *et al.*

The fuel model column refers to the Anderson fuel model ratings [4].

Formulation

The formulation of the model is as follows:

Minimize

$$\sum_{i \in I} \sum_{r \in R} F_{i,r} x_{i,r} \quad (1.8)$$

$$\max\{S_1, S_2\} \quad (1.9)$$

Maximize

$$\min\{O_1, O_2\} \quad (1.10)$$

Subject to:

$$\sum_{i \in I_{\omega,t}} \left(a_i p_{i,t} + e a_i \left(\sum_{j \in R_{i,t}} x_{i,j} - p_{i,t} \right) \right) = O_t \quad \forall t \in \{1, 2\} \quad (1.11)$$

$$\sum_{i \in I} \sum_{r \in 1,3} s_{i,1} x_{i,r} = S_1 \quad (1.12)$$

$$\sum_{i \in I} \sum_{r \in 2,3} s_{i,2} x_{i,r} = S_2 \quad (1.13)$$

$$\sum_{i \in D_c} \sum_{j \in R_{i,t}} x_{i,j} - |c| q_{c,t} \geq 0 \quad \forall t \in \{1, 2\}, c \in C \quad (1.14)$$

$$\sum_{c \in C_i} q_{c,t} - p_{i,t} \geq 0 \quad \forall t \in \{1, 2\}, i \in I_{\omega,t} \quad (1.15)$$

$$\sum_{r \in R} x_{i,r} = 1 \quad \forall i \in I \quad (1.16)$$

$$\sum_{i \in I} \sum_{r \in 1,3} a_i x_{i,r} = H_1 \quad (1.17)$$

$$\sum_{i \in I} \sum_{r \in 2,3} a_i x_{i,r} = H_2 \quad (1.18)$$

$$H_t \leq A \quad \forall t \in \{1, 2\} \quad (1.19)$$

$$\ell H_1 - H_2 \leq 0 \quad (1.20)$$

$$-u H_1 + H_2 \leq 0 \quad (1.21)$$

$$x_{i,r}, p_i, q_c \in \{0, 1\} \quad \forall i \in I, r \in R, c \in C \quad (1.22)$$

Equations (1.8)-(1.10) are the objective functions: equation (1.8) minimizes the cumulative fire hazard rating of the Drink area at the end of the 80-year planning horizon, equation (1.9) minimizes the maximum peak in sediment delivery for the two planning periods, and equation (1.10) maximizes the minimum NSO habitat available at the end of the planning periods. Equation set (1.11) defines the amount of NSO habitat available at the end of the planning horizons. Note that if stand i does not belong to a cluster of NSO habitat exceeding 200 hectares, then its area contribution to total NSO habitat is discounted by a factor of e . Equations (1.12) and (1.13) define the sediment delivered in planning periods one and two,

respectively.

Inequality set (1.14) controls the value of the cluster variables $q_{c,t}$ indicating clusters meeting NSO habitat criteria in each of the planning periods. Inequality set (1.15) controls the value of the $p_{i,t}$ variables indicating stands' inclusion in NSO habitat clusters.

The set of equalities (1.16) enforces the logical constraint that each stand must be prescribed to exactly one treatment schedule. Equations (1.17) and (1.18) are accounting constraints for the total area treated in each planning period, and inequalities (1.19) ensure that this area does not exceed the predefined per-period maximum. Inequalities (1.20) and (1.21) bound the fluctuation in treated area between the planning periods. Finally, constraint (1.22) defines the decision and indicator variables as binary.

1.2.5 Model Solution and Comparing Efficient Frontiers

I developed an implementation of Tóth's Alpha-Delta algorithm [65] to solve the models utilizing the IBM ILOG CPLEX optimization engine. For a problem with N objectives, the Alpha-Delta algorithm finds the optimal set of solutions by iteratively slicing the N -dimensional objective space with a tilted $N - 1$ dimensional plane. The algorithm was implemented using an alpha parameter of $\alpha = .01$ and delta parameters of $\delta_{Hab} = 1$ ha and $\delta_{Sed} = 2$ tons for the NSO habitat and sediment delivery objectives, respectively.

The solution to a bounded and non-degenerate multi-objective optimization problem with N objectives is a set of objective vectors (also called "solutions") $\mathbf{z} \in Z$ where $\mathbf{z} = [z^1, \dots, z^N]$. The set of solutions Z is referred to as the Pareto-optimal frontier or efficient frontier or, simply, frontier. The solutions comprising an efficient frontier have the special relationship such that no component of a solution \mathbf{z}^i can be improved upon without one of the other components \mathbf{z}^j ($j \neq i$) degrading. This quality is known as Pareto efficiency. For example, this relationship in the current problem means that further reducing the value of fire hazard in a solution would result in either additional sediment delivery, a reduction of NSO habitat, or both.

Thus the efficient frontier provides information on the trade-off relationship that exists

between ecosystem services. Parameterizing and solving the above model for each of the climate scenarios generates three frontiers: Z_{None} , $Z_{4.5}$, and $Z_{8.5}$ for the None, Ensemble RCP 4.5, and Ensemble RCP 8.5 scenarios, respectively. Since climate data alone differentiates the models and their resulting frontiers, comparing the frontiers reveals the impacts of climate on the trade-off relationships among the ecosystem services. However, no standardized procedure exists to compare frontiers.

One applicable metric to compare frontiers is the volume of the N -dimensional objective space bounded by the frontier, known as the hypervolume indicator. Together with Sándor Tóth, I devised an algorithm to compute the value of the hypervolume indicator any general N -dimensional frontier. The algorithm proceeds by sorting the solutions according to one objective, then iteratively adding them to the frontier, each time computing the additional volume enclosed by the solution. Details of the algorithm may be found in the appendix, §A.

We developed this algorithm independently and later discovered that researchers in the field of Evolutionary Multi-objective Optimization (EMO) have developed similar algorithms to compute the hypervolume indicator. In the present study, the hypervolume indicator is used as a measure of conflict among ecosystem services. Comparing hypervolume indicators across frontiers offers a quantification of the impact of climate change on trade-off relationships among ecosystem services. In EMO, the hypervolume indicator is used to assess the performance of stochastic multi-objective optimization algorithms. Hence, while the metric is the same, the algorithm to compute it and its application are entirely unique in this study.

Upon realization of the use of the hypervolume indicator in EMO, I discovered additional frontier comparison metrics used in this field and have adopted them for use here. These include the following indicators: the additive binary epsilon, binary hypervolume, unary distance, additive unary epsilon, and unary spacing. Information on these indicators can be found in the appendix, §C.1.

In addition to frontier-level comparisons that measure conflict between all ecosystem services simultaneously, it is also worthwhile to consider how climate change may impact

the relationship between two specific ecosystem services within the frontier. Here, I use two methods to determine this: 1) the hypervolume indicator of the nondominated frontier points in a 2D projection, and 2) the Pearson correlation coefficient between objectives. Details on these methods may be found in the appendix, §C.2.

1.3 Results and Discussion

DRAFT

The frontiers for each climate scenario can be found in Figure ...

1.4 Conclusion

DRAFT

I find that climate change has positive impacts on the tradeoff structure between managed ecosystem services in the Drink Area ...

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Appendix A

COMPUTING A FRONTIER'S HYPERVOLUME INDICATOR

Given a set of Pareto optimal solutions \mathcal{P} to a multi-objective mathematical programming model with a set of objectives O of cardinality $N := |O|$, this algorithm computes the volume V of the objective space bounded by the Pareto frontier defined by the solutions $x \in \mathcal{P}$. The objectives are assumed to be normalized so that the objective space is the N -dimensional unit hypercube with the origin and the point $\vec{1}$ defining the nadir objective vector and the ideal objective vector, respectively. That is, all objectives are assumed to be maximized with bounds $[0, 1]$.

The algorithm projects the objective space into $N - 1$ dimensions by eliminating the dimension associated with an (arbitrarily-chosen) objective $p \in O$. The set of objectives is $\bar{O} := O \setminus \{p\}$. It is assumed that $x \in \mathcal{P}$ are sorted in descending order according to p . The algorithm proceeds by sequentially adding solutions to the $(N - 1)$ -dimensional space, and calculating the contribution to the frontier volume as a product of the volume contribution in $N - 1$ dimensions and its achievement in objective p .

Let \bar{V}_x be the $(N - 1)$ -dimensional volume contribution of solution x and x_p be the achievement of solution x in objective p . Further, let F be the set of non-dominated solutions in $N - 1$ dimensions. I compute the N -dimensional volume of the frontier V as follows.

Figure A.1: Algorithm to compute the unary hypervolume indicator of a Pareto frontier.

```

1:  $V \leftarrow 0$ 
2:  $\bar{V} \leftarrow 0$ 
3:  $F \leftarrow \emptyset$ 
4: for all  $x \in \mathcal{P}$  do
5:    $\bar{V}_x \leftarrow \prod_{o \in \bar{O}} x_o - \bar{V}$ 
6:   for all  $f \in F$  do
7:     if  $f_o < x_o \forall o \in \bar{O}$  then
8:        $F \leftarrow F \setminus \{f\}$ 
9:     end if
10:  end for
11:  for all  $o \in \bar{O}$  do
12:     $F_{x,o} := \{f \in F : f_o > x_o\}$ 
13:    Sort  $f \in F_{x,o}$  in ascending order by their  $o$ th component,  $f_o$ 
14:     $v_i \leftarrow x_o$ 
15:    for all  $f \in F_{x,o}$  do
16:       $v_t \leftarrow f_o$ 
17:       $\delta_o := v_t - v_i$ 
18:       $\bar{V}_x \leftarrow \bar{V}_x + \delta_o \prod_{\sigma \in \bar{O} \setminus \{o\}} f_\sigma$ 
19:       $v_i \leftarrow v_t$ 
20:    end for
21:  end for
22:   $F \leftarrow F \cup \{x\}$ 
23:   $\bar{V} \leftarrow \bar{V} + \bar{V}_x$ 
24:   $V \leftarrow V + x_p \bar{V}_x$ 
25: end for

```

Appendix B

TREATMENT SPECIFICATIONS FOR THE DRINK AREA

Vegetation conditions were assessed at the midpoint of each planning period. If a set of conditions as listed in Table B.1 were met, then the corresponding treatment was applied. Otherwise, no action was taken. Table adapted from Schroder [59].

Table B.1: Rules governing treatment assignments.

SDI ¹	CBD ²	TPH _{<18} ³	Fuel model ⁴	BA _{MHD+WF,>46} ⁵	Treatment
Lodgepole pine (LPD) plant association					
< 87	N/A	N/A	N/A	N/A	Prescribed burn
≥ 87	> 0.037	> 49	≥ 10	N/A	Thin, pileburn slash and fuels ⁶
			< 10	N/A	Thin, pileburn slash
Mixed conifer wet (MCW) or mountain hemlock (MHD) plant associations					
< 87	N/A	N/A	N/A	N/A	Prescribed burn

¹Stand Density Index, calculated in metric units (trees per ha).

²Crown bulk density (kg/m^3)

³Number of trees per hectare whose diameter at breast height (DBH) is less than 18 cm

⁴According to the Anderson rating system[4]

⁵Basal area in m^2 of all mountain hemlock (MHD) and white fir (WF) trees with DBH > 46cm.

⁶Pileburning slash involves removal of thinned trees only, while pileburning slash and fuels also involves removal of materials that were on the ground before thinning (Wall, Powers, 2012; personal communication)

≥ 87	> 0.037	> 49	$= 10$	> 7.5	Thin, pileburn slash and fuels, prescribed burn
				≤ 7.5	Thin, pileburn slash and fuels
			> 10	N/A	Thin, pileburn slash and fuels
	≤ 0.037	≤ 49	< 10	N/A	Thin, pileburn slash
			$= 10$	≥ 7.5	Prescribed burn
			$= 10$	≥ 7.5	Prescribed burn
	N/A	N/A	$\in \{6, 8, 9, 10\}$	N/A	Prescribed burn ⁷
Mixed conifer dry (MCD) plant association					
< 87	N/A	N/A	N/A	N/A	Prescribed burn
≥ 87	> 0.037	> 49	$\in \{10, 11\}$	N/A	Thin, pileburn slash and fuels, prescribed burn
			≥ 12	N/A	Thin, pileburn slash and fuels
			< 10	N/A	Thin, pileburn slash
	≤ 0.037	≤ 49	$\in \{10, 11\}$	N/A	Prescribed burn
			$\in \{10, 11\}$	N/A	Prescribed burn
			$\in \{10, 11\}$	N/A	Prescribed burn
	N/A	N/A	$\in \{6, 8, 9, 10\}$	N/A	Prescribed burn ⁷

⁷Only if prescribed burn was assigned in period 1 (applies to period 2 treatment assignments only)

Appendix C

INTER- AND INTRA-FRONTIER COMPARISON METRICS

This chapter describes in more detail the metrics used to compare the frontiers generated by solving the multi-objective model described in §1.2.4. The metrics are broken up into two groups. *Inter-frontier comparison metrics* are those metrics that quantify some feature of a frontier. That is, the metric has only a single value per frontier. These metrics may be used either alone or with other metrics to make comparisons at the frontier level. *Intra-frontier comparison metrics* are those metrics that quantify some feature within a frontier. There may be multiple values of this metric per frontier, depending on the metric and the number of objectives. These metrics may be used either alone or with other metrics to make comparisons at the objective (or ecosystem service) level.

C.1 Inter-Frontier Comparison Metrics

Researchers in the field of EMO develop algorithms to generate a set of non-dominated solutions that best represents the true Pareto-optimal frontier [20]. To test their algorithms, they solve a benchmark multi-objective optimization problem and compare their resulting frontiers to the known Pareto front for that problem [40]. There is no assurance of optimality of the solutions derived using these algorithms, so they require a means of comparing the resulting frontiers to determine if one algorithm produces a “better” non-dominated frontier than another. Zitzler et al. provide a review of comparison methods [75]. These methods aim to quantify certain traits about a frontier that can be used to measure their success in approximation of the true frontier.

When necessary, the normalization of the objective space is such that all objectives are maximized, and each frontier is contained within the unit hypercube. That is, each

objective is bounded between 0 and 1, yielding a frontier bounded by $[0, 1]^N$. Defining the nadir solution \mathbf{z}_{nad} of a frontier of points $z \in Z$ as the objective vector with components

$$\mathbf{z}_{\text{nad}}^i = \inf_z \{z^i\} \quad \forall 1 \leq i \leq N \quad (\text{C.1})$$

and the ideal solution as the objective vector with components

$$\mathbf{z}_{\text{ideal}}^i = \sup_z \{z^i\} \quad \forall 1 \leq i \leq N \quad (\text{C.2})$$

then under my normalization, the nadir solution is the origin and the ideal solution is the N -dimensional vector of ones $\mathbf{1}_N$.

The definitions of dominance terms used here are in Table C.1.

Relation	Solutions		Frontiers	
Strictly dominates	$\mathbf{z}_1 \succ \succ \mathbf{z}_2$	\mathbf{z}_1 is better than \mathbf{z}_2 in all objectives	$Z_1 \succ \succ Z_2$	Every solution in Z_2 is strictly dominated by at least one solution in Z_1
Dominates	$\mathbf{z}_1 \succ \mathbf{z}_2$	\mathbf{z}_1 is better than \mathbf{z}_2 in at least one objective and is not worse in any objective	$Z_1 \succ Z_2$	Every solution in Z_2 is dominated by at least one solution in Z_1
Better			$Z_1 \triangleright Z_2$	Every solution in Z_2 is weakly dominated by at least one solution in Z_1 and $Z_1 \neq Z_2$
Weakly dominates	$\mathbf{z}_1 \succeq \mathbf{z}_2$	\mathbf{z}_1 is at least as good as \mathbf{z}_2 in all objectives	$Z_1 \succeq Z_2$	Every solution in Z_2 is weakly dominated by at least one solution in Z_1
Incomparable	$\mathbf{z}_1 \mathbf{z}_2$	Neither \mathbf{z}_1 nor \mathbf{z}_2 weakly dominates the other	$Z_1 Z_2$	Neither Z_1 nor Z_2 weakly dominates the other

Table C.1: Definitions of dominance relationships between solutions and between frontiers, reproduced from Zitzler *et al.* [75].

C.1.1 Additive binary epsilon indicator $I_{\epsilon+2}$

Given two frontiers, Z_1 and Z_2 , the additive binary epsilon indicator is defined as [75]

$$I_{\epsilon+2}(Z_1, Z_2) = \inf_{\epsilon \in \mathbb{R}} \{ \forall \mathbf{z}_2 \in Z_2 \exists \mathbf{z}_1 \in Z_1 : \mathbf{z}_1 \succeq_{\epsilon+} \mathbf{z}_2 \} \quad (\text{C.3})$$

where $\succeq_{\epsilon+}$ is the additive ϵ -dominance relationship:

$$\mathbf{z}_1 \succeq_{\epsilon+} \mathbf{z}_2 \iff \epsilon + \mathbf{z}_1^i \geq \mathbf{z}_2^i \quad \forall 1 \leq i \leq N \quad (\text{C.4})$$

Intuitively, ϵ is the minimum amount by which a frontier Z_1 must be translated such that every solution $\mathbf{z}_2 \in Z_2$ is “covered”. See Figure C.1. Positive values of $I_{\epsilon+2}(Z_1, Z_2)$ indicate the presence of points $\mathbf{z}_2 \in Z_2$ that are not dominated by Z_1 . Negative values of $I_{\epsilon+2}(Z_1, Z_2)$ indicate that Z_1 strictly dominates Z_2 ($Z_1 \succ \succ Z_2$).

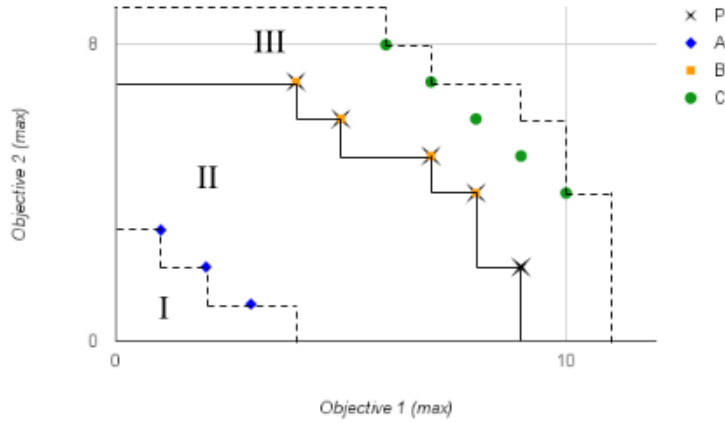


Figure C.1: Depiction of the additive binary epsilon indicator $I_{\epsilon+2}$ and the additive epsilon dominance relationship $\succeq_{\epsilon+}$. In the figure,

$$I_{\epsilon+2}(P, A) = -4 < 0 \quad I_{\epsilon+2}(P, B) = 0 \quad I_{\epsilon+2}(P, C) = 2 > 0$$

Region III is ϵ_+ -dominated for $\epsilon = 2$; region II is ϵ_+ -dominated for $\epsilon = 0$; region I is ϵ_+ -dominated for $\epsilon = -4$. Note that region II also encompasses region I, and region III encompasses region II.

C.1.2 Additive unary epsilon indicator I_{ϵ_+}

I define the unary epsilon indicator as

$$I_{\epsilon_+}(Z) = I_{\epsilon_+2}(Z, \mathbf{z}_{\text{ideal}}) \quad (\text{C.5})$$

That is, the additive unary epsilon indicator is identical to the additive binary epsilon indicator where the second frontier consists of a single point: the ideal solution for the first frontier.

This differs from the unary epsilon indicator traditionally used in EMO [75]. In EMO, the frontier is compared against a reference nondominated set. However, because the frontiers in the present study have guaranteed optimality, there is no reference set against which to compare them.

C.1.3 Unary hypervolume indicator I_{H1} and binary hypervolume indicator I_{H2}

For every solution \mathbf{z}_i in a frontier Z define the hyperrectangle r_i whose diagonal corners are the origin and the objective vector $\mathbf{z}_i = \langle z^1, \dots, z^N \rangle$ (see Figure C.2). Then the unary hypervolume indicator of the frontier Z is the N -dimensional volume of the union of all of the hyperrectangles corresponding to the solutions in Z :

$$I_{H1}(Z) = \text{vol} \left(\bigcup_{i=1}^{|Z|} r_i \right) \quad (\text{C.6})$$

Then define the binary hypervolume indicator of two frontiers Z_1 and Z_2 as [74]

$$I_{H2}(Z_1, Z_2) = I_{H1}(Z_1 + Z_2) - I_{H1}(Z_2) \quad (\text{C.7})$$

where $I_{H1}(Z_1 + Z_2)$ is the unary hypervolume indicator of the frontier consisting of the nondominated points in $Z = \{z \in Z_1 \cup Z_2\}$. See Figure C.3. The binary hypervolume indicator provides the volume of frontier Z_1 that is not contained within frontier Z_2 . Larger values of I_{H1} correspond to frontiers occupying larger amounts of the objective space. In a

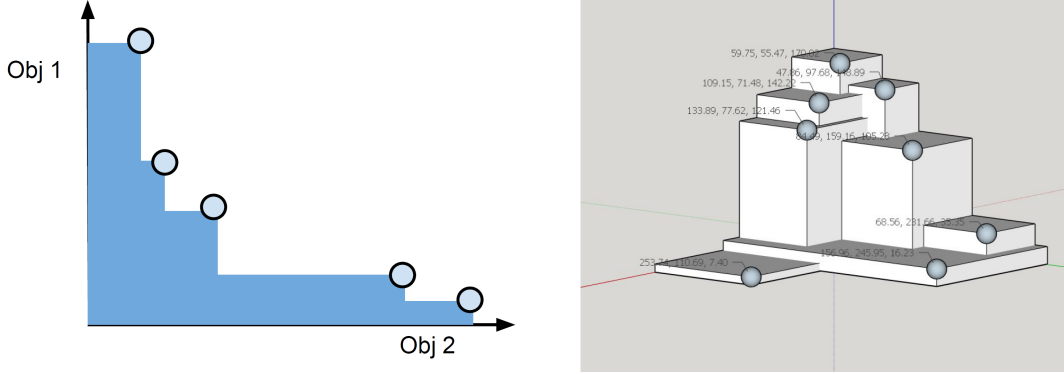


Figure C.2: Depiction of the hypervolumes of frontiers with two objectives (left) and three objectives (right).

normalized objective space, $I_{H2}(Z_1, Z_2) > I_{H2}(Z_2, Z_1)$ indicates areas of less conflict between objectives in Z_1 than in Z_2 .

I developed a custom algorithm to solve for the hypervolume indicators. The details of the algorithm may be found in §A.

C.1.4 Unary distance indicator I_d

The unary distance indicator measures the average distance from the frontier to the ideal solution:

$$I_d = \frac{\sum_{\mathbf{z} \in Z} \|\mathbf{z}_{\text{ideal}} - \mathbf{z}\|}{N} \quad (\text{C.8})$$

Smaller values of I_d correspond to frontiers that are closer to the ideal solution, which may imply less conflict between objectives. This metric is analogous to the unary distance indicator more commonly used in EMO [18]. Where the metric used here measures the distance to the ideal solution, the traditional metric measures the distance to a reference Pareto frontier.

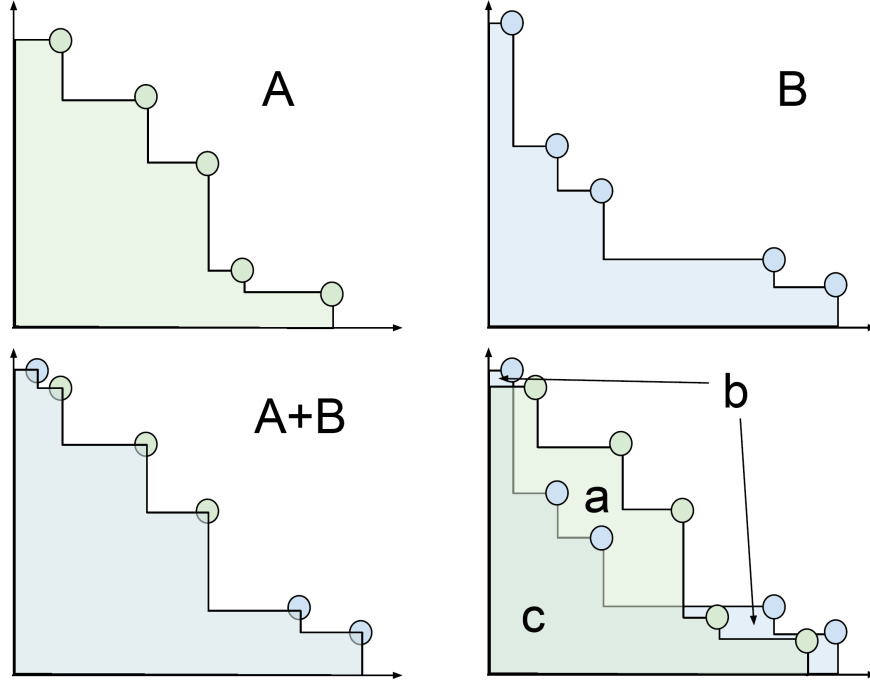


Figure C.3: Depiction of the binary hypervolume indicator. The individual frontiers are shown in the top row: frontier A (left) and frontier B (right). The merged frontier $A + B$ is shown in bottom left - note the absence of points that were dominated when combined. Following the naming of regions as shown in the bottom right figure, the binary hypervolume indicator is equal to

$$I_{H2}(A, B) = (\text{area}_a + \text{area}_b + \text{area}_c) - (\text{area}_b + \text{area}_c) = \text{area}_a$$

C.1.5 Unary Spacing Indicator I_s

The unary spacing indicator, or Schott's spacing metric [58], computes the standard deviation of the distance between points in the frontier:

$$I_s = \sqrt{\frac{1}{N-1} \sum_{\mathbf{z} \in Z} (d_z - \bar{d})^2} \quad (\text{C.9})$$

where

$$d_z = \min_{\mathbf{y} \in Z, \mathbf{y} \neq \mathbf{z}} \|\mathbf{z} - \mathbf{y}\| \quad (\text{C.10})$$

and \bar{d} is the average over all d_z . In EMO, the spacing indicator provides a measure of an algorithm's ability to search the frontier space uniformly. Here, the spacing metric provides a measure of the flexibility afforded to the decision maker under each climate scenario, since smaller values of I_s imply a higher density of solutions and greater flexibility.

C.2 Intra-Frontier Comparison Metrics

While the above methods provide frontier-level metrics of conflict and tradeoffs, there are two methods employed here to also determine the degree of conflict within a single frontier. The first is an approach used in many-objective optimization, and the second is a variant of the unary hypervolume indicator.

C.2.1 Pearson correlation coefficients

Given the increased difficulty in solving many-objective optimization problems [39], researchers in this field seek to reduce the number of objectives considered in the model. To determine which objectives most strongly influence the shape of the frontier, they compute the correlation between each pair of objectives [21]. Objective pairs with strong negative correlation conflict with one another. The Pearson correlation coefficients are computed per

$$\rho_{X,Y} = \frac{\text{cov}(X,Y)}{\sigma(X)\sigma(Y)} \quad (\text{C.11})$$

where, for objectives x and y , X and Y are

$$X = \{\mathbf{z}_1^x, \mathbf{z}_2^x, \dots, \mathbf{z}_{|Z|}^x\} \quad (\text{C.12})$$

$$Y = \{\mathbf{z}_1^y, \mathbf{z}_2^y, \dots, \mathbf{z}_{|Z|}^y\} \quad (\text{C.13})$$

C.2.2 Area of 2D frontier projection A_{xy}

The second intra-frontier comparison metric uses the unary hypervolume indicator described in §C.1.3. Given a frontier with objective vectors in N dimensions, take two objectives x and y , and project the N -dimensional frontier to the two-dimensional xy -plane. Remove

solutions dominated in this projection, and compute the hypervolume indicator (which, in two-dimensions, is simply the area). See Figure C.4. Larger values of A_{xy} imply less conflict between objectives x and y .

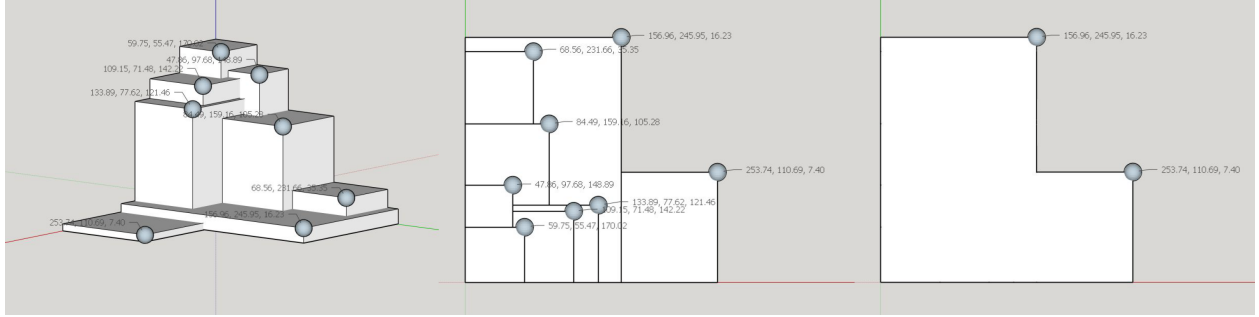


Figure C.4: Comparing conflict between objectives based on the area bounded by two-dimensional frontier projection. Left is the original frontier; middle shows the 2D projection of the frontier; right shows the projected frontier with all dominated solutions removed. Assuming both objectives are maximized, the larger the area bounded by the cross-sectional area, the less conflict between the objectives.