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Reaction Kinetics of CO₂ Carbonation with Mg-Rich Minerals

Soonchul Kwon, Maohong Fan, *,*,† Herbert F. M. Dacosta, Armistead G. Russell, and Costas Tsouris

ABSTRACT: Due to their low price, wide availability, and stability of the resulting carbonates, Mg-rich minerals are promising materials for carbonating CO₂. Direct $k = 2.3 \times 10^6 e^{-T}$ (kmol⁶·kg min) carbonation of CO₂ with Mg-rich minerals reported in this research for the first time could be considerably superior to conventional liquid extraction processes from an energy consumption perspective due to its avoidance of the use of a large amount of water with high specific heat capacity and latent heat of vaporization. Kinetic models of

$$k = 2.3 \times 10^6 e^{-\frac{5,193}{T}} \text{ (kmol}^{6.} \text{kg} \cdot \text{min)}$$

$$k_d = 2.5 \times 10^6 e^{\frac{-5,837}{T}}$$
 (L/min)

the reactions of the direct CO₂ carbonation with Mg-rich minerals and within simulated flue gas environments are important to the scale-up of reactor designs. Unfortunately, such models have not been made available thus far. This research was initiated to fill that gap. Magnesium silicate (Mg₂SiO₄), a representative compound in Mg-rich minerals, was used to study CO₂ carbonation reaction kinetics under given simulated flue gas conditions. It was found that the chosen sorbent deactivation model fits well the experimental data collected under given conditions. A reaction order of 1 with respect to CO₂ is obtained from experimental data. The Arrhenius form of CO₂ carbonation with Mg₂SiO₄ is established based on changes in the rate constants of the chosen deactivation model as a function of temperature.

■ INTRODUCTION

The concentration of CO₂ in the atmosphere has been continuously increasing for many years due to the emissions of CO₂ from many sources. 1,2 About 40% of the total emissions of CO₂ in the U.S. are generated by power plants that use fossil fuels including coal, natural gas, and oil as their major energy sources. In the recent years, various technologies have been developed to capture CO₂. The costs associated with the current CO₂. capture technologies are too high to be widely accepted. Without an economically viable CO₂ capture technology, carbon capture and storage (CCS) cannot be justified.¹⁴ CO₂ capture by adsorption does not appear to be thermodynamically limited and, because of the possibility of high CO2 selectivity and low energy demand, 15 adsorption appears to be one of the more promising options among the various technologies proposed for capturing CO2 from power plant flue gases. Consequently, the development of economical and effective adsorbents is desirable. Solid sorbents are an attractive choice because they are easy to handle, are not associated with a risk of releasing solvents in the environment, have low deployment costs, and cause negligible corrosion problems. Further, they cause lower environmental impacts. 16,1

Flue gases from power plants typically contain 8-12% (vol) CO_2 and 8-10% steam (H_2O). Due to the tendency of H_2O molecules to also be adsorbed by the sorbent, many sorbents lose their CO₂ adsorption capacities in the presence of water vapor. 18 Accordingly, sorbents with high adsorption capacities for CO₂ in the presence of water are desirable for their practical applications in power plants. Therefore, it is important to identify solid

sorbents whose CO₂ sorption capacities are not negatively affected due to the presence of water vapor.

Because of their low price, wide availability, and the fact that their CO₂ adsorption capacity and rate increase in the presence of H_2O , magnesium-rich (Mg-rich) minerals are promising candidates for CO_2 separation. Due to the fact that CO_2 is captured as a stable carbonate, mineral-based CO2 removal processes can minimize the risk of CO2 leakage compared to other methods, thereby providing the potential for long-term CO₂ sequestration. Moreover, solid byproducts from the CO₂ mineralization process have many potential applications, including soil remediation.²¹ However, even though their associated reactions are thermodynamically favorable, natural Mg-rich, mineral-based carbonation processes are slow because of the low surface area available. 19,21,22 Therefore, researchers have shown increasing interest in improving the kinetics of mineralbased, Mg-rich CO₂ carbonation methods. One approach is to use an aqueous extraction process under elevated pressure and temperature that can improve the kinetics of reaction between Mg₂SiO₄ and CO₂. However, because it requires either mechanical or chemical pretreatment of Mg-rich minerals, this process is indirect and overly time-consuming.

By contrast, not only might the use of Mg-rich minerals for the direct carbonation of CO2 in flue gases overcome certain shortcomings of an aqueous extraction process, it might also help to identify methods for improved rates of capture. However, to the authors' knowledge, studies on kinetic models

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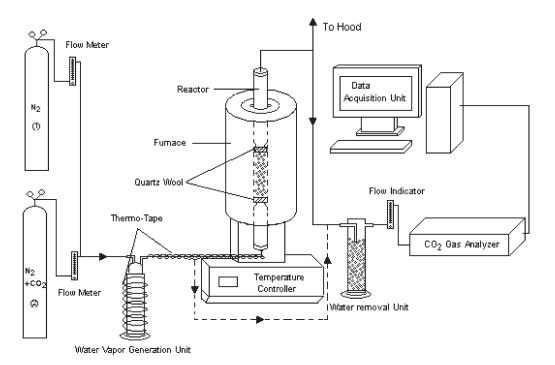


Figure 1. Schematic diagram of experimental setup for Mg₂SiO₄-based CO₂ sorption.

of the reactions between Mg-rich minerals and gaseous CO_2 within a simulated flue gas environment, including rate equation and Arrhenius form, have not yet been reported. Thus, the development of kinetic models for the early application of Mg-rich CO_2 carbonation in coal-fired power plants is desirable.

Pure magnesium silicate (Mg_2SiO_4) was selected in this research to study the kinetics of the reaction between Mg-rich minerals and CO_2 within a simulated flue gas environment. The overall reaction between Mg_2SiO_4 and H_2O is

$$^{1}/_{2}Mg_{2}SiO_{4} + CO_{2} \stackrel{H_{2}O}{\Longleftrightarrow} MgCO_{3} + ^{1}/_{2}SiO_{2}$$
 (R1)

 Mg_2SiO_4 is chosen as a model CO_2 sorbent because it represents major Mg-rich minerals existing in nature (e.g., olivine and serpentine) and because the use of a single compound simplifies the kinetic study.

■ EXPERIMENTAL SECTION

Materials. Mg₂SiO₄ was supplied by Alfa Aesar and used without further treatment. Mg₂SiO₄ is in the form of a white powder with 99% purity, a median particle size of 3.5 μ m, and a bulk density of 3.21 g/cm³. The X-ray diffraction (XRD) patterns of the Mg₂SiO₄ both before and after carbonation were analyzed using an X'PERT model X-ray diffractometer (Philips), with Cu Kα as the radiation resource, and with scans performed in the 2 θ range from 5° to 95° with 0.02°/s.

Experimental Apparatus. The schematic of the experimental setup for CO_2 capture (Figure 1) consists of three parts: a flue gas simulation unit, a CO_2 sorption system, and CO_2 analysis equipment. Gas cylinders containing pure N_2 and a mixture of CO_2 and N_2 with a flow rate of 0.5 L/s were used to conduct all tests. Flow rates of the feed gases were manually controlled with

Matheson Trigas FM-1050 flowmeters. An additional flowmeter was used to determine if the flow of the inlet gas analyzer was normal during all CO_2 sorption operations.

Adsorption of CO₂ with Mg₂SiO₄ was carried out in a quartz tubular reactor with a length of 610 mm and inside diameter of 9 mm. The fixed sorbent bed in the reactor was formed by placing the Mg₂SiO₄ particles between two pieces of quartz wool to serve as bed holders. The quartz reactor was placed inside a TF55030A-1 tube furnace (Thermo Corporation, Asheville, NC), with a UT150 temperature controller (Yokogawa M&C Corp., Newnan, GA) to control the CO₂ sorption temperature. The CO₂ sorption unit was connected to a steam generation unit to introduce water into the simulated dry flue gas stream containing CO2 and N2. In order to avoid steam condensation prior to the flow of the simulated wet flue gas into the sorbent bed, thermotape was used to wrap the pipelines between the vapor injection point and the inlet gas-reactor connection, with temperature controllers employed to control the heating rates of the thermotape. The outlet gas from the tubular reactor was passed through a steam removal unit before entering the ZRE gas analyzer (California Analytical Instruments, Inc.), where the CO_2 concentration was measured. The CO₂ concentration in the simulated dry flue gas stream was also measured before each sorption test started. A data acquisition system was connected directly to the gas analyzer in order to continuously record measured CO₂ concentrations.

Operating Procedures. The CO_2 adsorption capacity of Mg_2SiO_4 under a given condition was determined based on changes in measured CO_2 concentrations after adsorption. Each test run was conducted with fresh Mg_2SiO_4 . First, 0.50 g of Mg_2SiO_4 was lightly packed into the tubular reactor, which was preheated for 10 min to ensure constant operating conditions at the desired reaction temperature. The tubular reactor was then connected to the gas supply unit and the gas analyzer. At the same time, the data acquisition unit was turned on. The

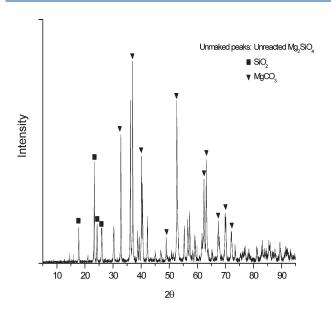


Figure 2. XRD patterns of spent Mg_2SiO_4 [(inlet gas conditions, H_2O_7 , 4.1 mmol/L; CO_2 , 4.1 mmol/L; total gas flow rate, 0.5 L/min); weight of Mg_2SiO_4 , sorption temperature, 150 °C].

composition of effluent gas from the steam removal unit was measured immediately by the ZRE gas analyzer. When the $\rm CO_2$ concentration in the effluent gas stream was within 1% of the initial inlet gas-stream concentration (indicating that the $\rm Mg_2SiO_4$ particle surfaces were saturated with $\rm CO_2$), the flow of simulated flue gas into the reactor was stopped. Each reported data point in this study represents the average value of three tests under the same operating conditions.

■ RESULTS AND DISCUSSION

Determination of the Temperature Range for Kinetic Study. Thermodynamic calculations have shown that the Gibbs free energy changes of (R1) are $-32.95,\,-20.88,\,\mathrm{and}\,-3.88\,\mathrm{kJ/mol}$ at temperatures of 25, 100, and 200 °C, respectively, indicating that Mg₂SiO₄-based CO₂ mineralization should occur in the given temperature range, which is supported by the results (Figures 2 and 3). CO₂ adsorption occurred on the surface of Mg₂SiO₄ at 200 °C, as demonstrated by the peaks of carbonation products (MgCO₃ and SiO₂) of (R1) (Figure 2), a finding consistent with the observations of other researchers.

The increase in the Gibbs free energy changes of (R1) with temperature signifies that higher temperatures could have considerable negative effects on the thermodynamics of CO_2 on Mg_2SiO_4 . On the other hand, the rate of reaction improves at higher temperatures. Therefore, an appropriate temperature range should be chosen for study of the reaction kinetics of (R1). Our preliminary tests showed that the reaction rate of (R1) was quite slow in the temperature range of $0-100\,^{\circ}C$, even though the sorbent could achieve reasonably large CO_2 sorption capacities over very long sorption periods. However, given that the volume of the emitted CO_2 -containing flue gas is so large, slow sorption would require the construction of very bulky adsorber units to extend considerably the time of contact between Mg_2SiO_4 and CO_2 in order to achieve high CO_2 capture efficiencies, a prohibitively expensive approach.

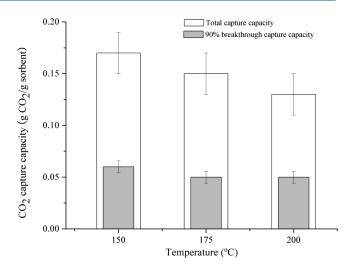


Figure 3. Total and breakthrough CO_2 sorption capacities of Mg_2SiO_4 at different temperatures [(inlet gas conditions, H_2O , 4.1 mmol/L; CO_2 , 4.1 mmol/L; total gas flow rate, 0.5 L/min); weight of Mg_2SiO_4 , 0.5 g].

A good sorbent should, therefore, have not only a high total sorption capacity but also a high breakthrough capacity. High breakthrough capacity is determined by inherent characteristics of the sorbent, including molecular structure, pore structure, surface defect or active site, BET (Brunauer, Emmett and Teller) surface area, and sorption operation conditions. Since it is directly related to the reaction kinetics of (R1), temperature is expected to be one of the more important operational factors affecting the breakthrough capacity of Mg₂SiO₄. The total and breakthrough CO₂ sorption capacities of Mg₂SiO₄ at 150, 175, and 200 °C (Figure 3) show that, compared to other solid sorbents, Mg_2SiO_4 has good CO_2 sorption performance at the three temperatures. While the total sorption capacity of Mg₂SiO₄ decreases with an increase in temperature, the breakthrough capacity does not change considerably (Figure 3). Our tests showed that the breakthrough sorption capacity achieved at 150 °C, when CO₂ separation efficiencies were close to 100%, was lower than that obtained at 200 °C. Further, the ratio of breakthrough sorption capacity to the total sorption capacity increased with temperature under the given experimental conditions (Figure 3) and was higher than those of most of the reported sorbents.^{31–36} Because its performance in the 150-200 °C range suggests that Mg₂SiO₄ could react with CO₂ reasonably fast and with a considerable CO₂ sorption capacity, a 100-200 °C temperature range was chosen to study the kinetics of reaction between Mg₂SiO₄ and CO₂.

Sorption Kinetics. Determination of Reaction Order and Rate Constants. Park³⁷ successfully used a deactivation model to describe the decrease in activity of Na₂CO₃ during reaction of NaOH with CO₂. Because the heterogeneous CO₂ adsorption process investigated in this research is very similar to that of Park's,³⁷ his deactivation model is used to establish the rate equation of (R1).

According to the deactivation model, Mg₂CO₃ forms gradually during the carbonation process and covers the surface of Mg₂SiO₄, thus reducing the activity or carbonation rate of the sorbent. Assuming that the pseudo-steady-state hypothesis is applicable within a constant water vapor concentration environment, the deactivation equation of the Mg₂SiO₄-based CO₂ carbonation process in a packed bed reactor shown in Figure 1

can be expressed as³⁷

$$a = \frac{C_{\text{CO}_2}}{C_{\text{CO}_2,0}}$$

$$= \exp \left[\frac{\left[1 - \exp \left(\frac{kC_{\text{H}_2\text{O}} W_{\text{Mg}_2\text{SiO}_4}}{Q_{\text{g}}} \right) (1 - \exp(-k_{\text{d}}t)) \right]}{1 - \exp(-k_{\text{d}}t)} \exp(-k_{\text{d}}t) \right]$$
(E1)

or

 C_{CO_2}

$$= C_{\text{CO}_2,0} \exp \left[\frac{\left[1 - \exp\left(\frac{kC_{\text{H}_2\text{O}}W_{\text{Mg}_2\text{SiO}_4}}{Q_g}\right) (1 - \exp(-k_d t))\right]}{1 - \exp(-k_d t)} \exp(-k_d t) \right]$$
(E2)

where $C_{\rm CO_2,0}$ and $C_{\rm CO_2}$ are the inlet and outlet concentrations $({\rm kmol/m^3})$ of ${\rm CO_2}$ in the gas stream, respectively, at any sorption time (t) and k is the initial sorption rate constant $[{\rm m^6/(kmol\cdot kg\cdot min)}]$ in the following ${\rm CO_2}$ isothermal conservation equation

$$-Q_{g}\frac{dC_{CO_{2}}}{dw_{Mg_{2}SiO_{4}}}-kC_{H_{2}O}C_{CO_{2}}^{n(CO_{2})}\alpha_{Mg_{2}SiO_{4}}^{m(Mg_{2}SiO_{4})}=0$$
 (E3)

Here, $C_{\rm H_2O}$ is the concentration of water vapor (kmol/m³), $W_{\rm Mg_2SiO_4}$ is the weight of Mg₂SiO₄ (kg) in the reactor, $Q_{\rm g}$ is the volumetric flow rate of the inlet gas mixture (L/min), $w_{\rm Mg_2SiO_4}$ is consumed Mg₂SiO₄, $n_{\rm CO_2}$ is the reaction order with respect to CO₂, $m_{\rm Mg_2SiO_4}$ is the power of $\alpha_{\rm Mg_2SiO_4}$ (the activity of Mg₂SiO₄ defined as $(W_{\rm Mg_2SiO_4} - w_{\rm Mg_2SiO_4})/W_{\rm Mg_2SiO_4}$, varying from 0 to 1), and $k_{\rm d}$ is the deactivation rate constant [m³/(kmol·min)] when $n_{\rm CO_2}$ and $m_{\rm Mg_2SiO_4}$ are 1 in the following sorbent deactivation equation

$$-\frac{d\alpha_{Mg_2SiO_4}}{dt} = k_d C_{CO_2}^{n(CO_2)} \alpha_{Mg_2SiO_4}^{m(Mg_2SiO_4)}$$
(E4)

The choices for $n_{\rm CO_2}=1$ and $m_{\rm Mg_2SiO_4}=1$ are based on Park's studies³⁷ on the $\rm Na_2CO_3$ reaction system, which is very similar to the one studied in this research. An iterative method was used to determine k and $k_{\rm d}$ at a given temperature. The k and $k_{\rm d}$ ultimately chosen should make the $C_{\rm CO_2}$ vs t relationship predicted by (E2) under given experimental conditions match the $C_{\rm CO_2}$ vs t profiles recorded by the $\rm CO_2$ analyzer, as shown in Figure 4, for the results obtained at a given temperature under different flow rates. The differences of the two $C_{\rm CO_2}$ vs t relationships, shown in Figure 4 to become larger after a period of reaction time will be explained in one of the following paragraphs.

We chose to use the sorption profile obtained at 200 °C (Figure 4) to test the deactivation model (E2), since the slope resolution of the sorption profile at 200 °C is higher than that at other temperatures due to the increases of k and $k_{\rm d}$ with temperature. On the basis of the assumption that $n_{\rm CO_2}=1$ and $m_{\rm Mg,SiO_4}=1$ at 200 °C, and under other given reaction

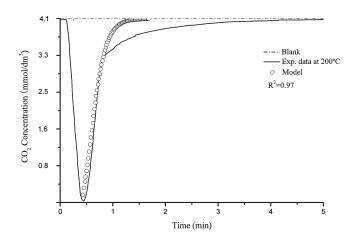


Figure 4. Mg₂SiO₄ based CO₂ sorption profile at 200 °C [(inlet gas conditions, H₂O, 4.1 mmol/L; CO₂, 4.1 mmol/L; total gas flow rate, 0.5 L/min); weight of Mg₂SiO₄, 0.5 g].

conditions, (E4) can be written as

$$-\frac{d\alpha_{Mg_2SiO_4}}{dt} = k_d C_{CO_2} \alpha_{Mg_2SiO_4}$$
 (E5)

If the rate-limiting reaction in the overall reaction mechanism of (R1) does not change within the range of 100-200 °C, $m_{\rm Mg_2SiO_4}$ and $n_{\rm CO_2}$ should not vary either, and their values can therefore be obtained with the sorption data collected at any temperature within the range. However, $m_{\rm Mg_2SiO_4}=1$ and $n_{\rm CO_2}=1$ are assumed values and were tested only under limited test conditions. Thus, their applicability within broader ranges of test conditions should be further supported with theoretical derivations, as discussed below.

Assuming (R1) proceeds through the following elementary steps in the presence of water during the CO₂ sorption process

$$CO_2 + H_2O \Leftrightarrow 2H^+ + CO_3^{2-}$$
 (R2)

$$^{1}/_{2}Mg_{2}SiO_{4} + 2H^{+} \rightarrow Mg^{2+} + H_{2}O + ^{1}/_{2}SiO_{2}$$
 (R3)

$$Mg^{2+} + CO_3^{2-} \rightarrow MgCO_3$$
 (R4)

(R2) is a reversible reaction that can reach its equilibrium state quickly, leading to

$$k_{(R2), forward} C_{CO_2} C_{H_2O} - k_{(R2), reverse} C_{net, H^+}^2 C_{net, CO_3^{2-}} = 0$$
(E6)

or

$$C_{\text{net, H}^{+}}^{2} = \frac{k_{(R2), \text{forward}}}{k_{(R2), \text{reverse}}} \frac{C_{\text{CO}_{2}} C_{\text{H}_{2}\text{O}}}{C_{\text{net, CO}_{3}^{2^{-}}}}$$
(E7)

where $k_{(R2),forward}$ and $k_{(R2),reverse}$ represent the forward and reverse reaction rate constants of (R2), respectively, and C_{CO_2} and C_{H_2O} correspond to the concentrations of CO_2 and C_{H_2O} respectively. Since C_{H_2O} is actually neither consumed nor generated when both (R2) and (R3) are considered, C_{H_2O} can be treated as a constant. C_{L_2O} is simply an indispensible C_{L_2O} and the concentration of C_{L_2O} is the sum of the concentrations of the C_{L_2O} and the forward reaction of (R2) and the C_{L_2O} and the reverse reaction of (R2).

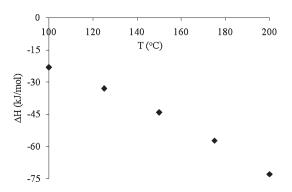


Figure 5. Effect of temperature on enthalpy change of CO₂ hydrolysis.

 $C_{\rm net,CO_3^{2-}}$ is the net concentration of ${\rm CO_3^{2-}}$, which is the sum of the concentrations of the ${\rm CO_3^{2-}}$ generated from the forward reaction of (R2) and the ${\rm CO_3^{2-}}$ consumed by (R4) and the reverse reaction of (R2). According to the pseudo-steady-state theory, 38 $C_{\rm net,CO_3}^{2-}$ can be treated as a constant. Furthermore, $C_{\rm net,CO_3}^{2-}$ should be low for two reasons. First, since exothermal characteristics of the reaction (Figure 5) result in ${\rm CO_2}$ having low solubility in water in the tested temperature range (particularly at 200 °C, a relatively high temperature for sorption), the forward reaction rate of (R2) is slow. Therefore, (R2) or the interaction between H₂O and ${\rm CO_2}$ is the rate-controlling step of the overall ${\rm CO_2}$ sorption process. Second, given that high pressure is not used, (R3) and (R4) should be much quicker than the forward reaction of (R2), leading to quick consumption of (R2)'s slowly generated ${\rm CO_3}^{2-}$. Thus, $C_{\rm net,CO_3}^{2-}$ can be considered to be a small constant.

On the basis of (R3), a decrease of $\alpha_{Mg_2SiO_4}$ with time can be expressed as

$$-\frac{d\alpha_{\text{Mg}_2\text{SiO}_4}}{dt} = k_{(R3)}C_{\text{net},H^+}^2\alpha_{\text{Mg}_2\text{SiO}_4}^{0.5}$$

$$= k_{(R3)}\frac{k_{(R2),\text{forward}}}{k_{(R2),\text{reverse}}} \frac{C_{\text{CO}_2}C_{H_2O}}{C_{\text{net},\text{CO}_3}^{2-}}\alpha_{\text{Mg}_2\text{SiO}_4}^{0.5}$$

$$= \frac{k_{(R2),\text{forward}}k_{(R3)}}{k_{(R2),\text{reverse}}} \frac{C_{H_2O}}{C_{\text{net},\text{CO}_3}^{2-}}C_{\text{CO}_2}\alpha_{\text{Mg}_2\text{SiO}_4}^{0.5}$$

$$= k'C_{\text{CO}_2}\alpha_{\text{Mg}_2\text{SiO}_4}^{0.5}$$
(E8)

where

$$k' = \frac{k_{(\text{R2}), \text{forward}} k_{(\text{R3})}}{k_{(\text{R2}), \text{reverse}}} \frac{C_{\text{H}_2\text{O}}}{C_{\text{net, CO}_3}^{2-}}$$

is a constant. Comparing (E8) to (E5), we find that the reaction orders with respect to CO₂, i.e., $n_{\text{CO}_2,\text{experiment}}$ in (E5) and $n_{\text{CO}_2,\text{theory}}$ in (E8), obtained with experimental results and derived with a postulated reaction mechanism, respectively, are exactly the same. However, $m_{\text{Mg}_2\text{SiO}_4,\text{experiment}} = 1$ in (E5), is larger than $m_{\text{Mg}_2\text{SiO}_4} = 0.5$ in (E8). In the meantime, $m_{\text{Mg}_2\text{SiO}_4,\text{experiment}} = 1$ still makes (E2) fit the experimental data well at the early stage of sorption. This is because when $\alpha_{\text{Mg}_2\text{SiO}_4}$ is within 0.85-1, which is the range used in Figure 4 for comparing $C_{\text{CO}_2,\text{measured}}$ and $C_{\text{CO}_2,\text{predicted with (E2)}}$

$$\frac{\alpha_{Mg_2SiO_4}^{0.5} - \alpha_{Mg_2SiO_4}}{\alpha_{Mg_2SiO_4}^{0.5}} \times 100\% \Bigg|_{maximum}$$

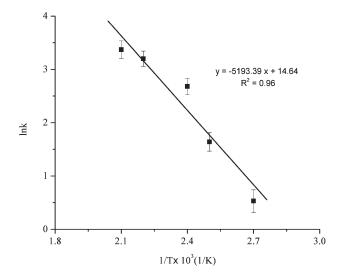


Figure 6. Determination of parameters of Arrhenius form of initial CO₂ sorption [H₂O, 4.1 mmol/dm³; CO₂, 4.1 mmol/dm³; total gas flow rate, 0.5 L/min; weight of Mg₂SiO₄, 0.5 g; sorption temperature, 100–200 °C].

is less than 8%. Clearly, when $\alpha_{\mathrm{Mg_2SiO_4}}$ continues to decrease, the difference between the assumed $m_{\mathrm{Mg_2SiO_4}}$ (1) and the $m_{\mathrm{Mg_2SiO_4}}$ (0.5) that was postulated with the reaction mechanism leads to an increase in $(C_{\mathrm{CO_2,predicted\,with\,(E2)}} - C_{\mathrm{CO_2,measured}})$, as shown in the later $\mathrm{CO_2}$ mineralization stage in Figure 4. Even though an analytical solution or explicit expression of (E1) or (E2) cannot be derived from (E3) and (E4), and numerical methods must be used, it is better to use $m_{\mathrm{Mg_2SiO_4}} = 0.5$ in (E3) and (E4) for predicting the $\mathrm{Mg_2SiO_4}$ deactivation rate.

Arrhenius Form. The relationship between rate constants (k or k_d) and reaction temperature (T) can be correlated using the Arrhenius form³⁸

$$k = Ae^{-E/RT} (E9)$$

$$k_{\rm d} = A_{\rm d} e^{-E_{\rm d}/RT} \tag{E10}$$

where A and $A_{\rm d}$ are pre-exponential factors treated as constants in the studied temperature range, E and $E_{\rm d}$ are the activation energy values corresponding to k and $k_{\rm d}$, and R is the ideal gas constant. The pre-exponential factors would also be impacted by reactor conditions (e.g., geometry, loading), and these were kept constant between experiments.

The values of k and $k_{\rm d}$ at 100, 125, 150, and 175 °C were obtained using the same method cited previously for the k and $k_{\rm d}$ values at 200 °C. The ln k vs T^{-1} and ln $k_{\rm d}$ vs T^{-1} plots of (R1) are shown in Figure 6 and Figure 7, respectively. Regression found an E value of 43.2 \pm 3.7 kJ/mol and an A value of (2.3 \pm 0.6) \times 10⁶ (kmol⁶·kg·min), respectively. In addition, regression found an $E_{\rm d}$ value of 48.5 \pm 2.4 kJ/mol and an $A_{\rm d}$ value of (2.5 \pm 0.8) \times 10⁵ (L/min), respectively. Thus, the Arrhenius forms of CO₂ carbonation with Mg₂SiO₄ within an H₂O environment for the geometry of our experiments are

$$k = 2.3 \times 10^6 e^{5193/T} \text{ (kmol}^6 \cdot \text{kg} \cdot \text{min)}$$
 (E11)

$$k_{\rm d} = (2.5 \times 10^6) {\rm e}^{-5837/T} \, ({\rm L/min})$$
 (E12)

The reaction order and activation energy of (R1) should not be affected by the flow rate of the simulated flue gas mixture

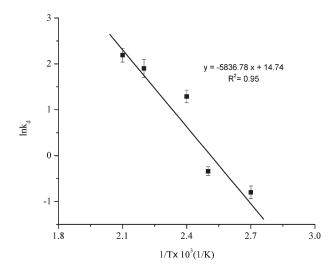


Figure 7. Determination of parameters of Arrhenius form of Mg_2SiO_4 deactivation [H_2O , 4.1 mmol/dm³; CO_2 , 4.1 mmol/dm³; total gas flow rate, 0.5 L/min; weight of Mg_2SiO_4 , 0.5 g; sorption temperature, $100-200\ ^{\circ}C$].

Table 1. Effect of Gas Flow Rate on Reaction Order and Activation Energy of the Mg₂SiO₄-Based CO₂ Sorption [(inlet gas condition, H₂O, 4.1 mmol/L; CO₂, 4.1 mmol/L; total gas flow rate, 0.5 L/min); weight of Mg₂SiO₄, 0.5 g; sorption temperature, 200 °C]

| flow rate (L/min) | initial CO_2 sorption E (kJ/mol) | Mg_2SiO_4 deactivation E_d (kJ/mol) |
|-------------------|--------------------------------------|-----------------------------------------|
| 0.25 | 42.4 ± 5.1 | 47.9 ± 4.2 |
| 0.5 | 43.2 ± 3.7 | 48.5 ± 2.4 |
| 1 | 43.5 ± 4.6 | 47.2 ± 5.7 |

under the given CO_2 mineralization conditions, an assumption verified by conducting CO_2 sorption tests with the same initial CO_2 concentration but with gas flow rates of 0.25 and 1.0 L/min. The derived values of E and E_d of (R1) for the three flow rates (0.25, 0.5, and 1.0 L/min) were consistent (see Table 1).

Application Considerations. Reactor Choice. Figure 3 shows that the $\rm CO_2$ breakthrough and total sorption capacities achieved with $\rm Mg_2SiO_4$ under the given test conditions were in the ranges of $\rm 50-60$ and $\rm 125-175$ g of $\rm CO_2/kg$ of $\rm Mg_2SiO_4$, respectively, which are lower than the theoretical sorption capacity of pure $\rm Mg_2SiO_4$, $\rm 630$ g of $\rm CO_2/kg$ of $\rm Mg_2SiO_4$. Therefore, appropriate measures should be taken to enhance the $\rm CO_2$ sorption capacities of $\rm Mg_2SiO_4$ within given reaction conditions, since otherwise a large percentage of $\rm Mg_2SiO_4$ would be wasted.

One of the conventional methods to increase the chemisorption capacities of sorbents is to improve reaction kinetics, which could theoretically be realized by three approaches for the heterogeneous reaction, i.e., elevation of the reaction temperature, reduction of the particle size of Mg₂SiO₄ or Mg-rich minerals, and an increase in gas pressure.

Since (R1)'s activation energy value ($43.2 \pm 3.7 \text{ kJ/mol}$) is relatively large, raising the reaction temperature should have a relatively large effect on the reaction rate of the carbonation process. However, since the Gibbs free energy change of (R1) increases with an increase in temperature, any increase in

temperature is unfavorable to (R1). Indeed, the reaction becomes unfavorable above 200 $^{\circ}$ C.

The median particle size of Mg_2SiO_4 used in this research was 3.5 μ m, which is already small for fixed-bed reactors. Further reductions in Mg_2SiO_4 particle size can create higher Mg_2SiO_4 surface areas, thereby improving CO_2 breakthrough sorption capacities within a given time period. However, this can also lead to a considerable increase in back pressure, undesirable in fixed-bed reactors because, according to Ergun's equation³⁹

$$\Delta P/L = 150 \frac{(1-\varepsilon)^2}{\varepsilon^3} \frac{\mu}{d_{\rm p}^2} u + 1.75 \frac{1-\varepsilon}{\varepsilon^3} \frac{\rho}{d_{\rm p}} u^2 \qquad (E13)$$

where ε is the interparticle porosity of the CO₂ sorption bed, μ is the flue gas viscosity (kg/ms), u is the interstitial velocity (m/s) of the gas stream, ρ is the density (kg/m³) of flue gas, and $\Delta P/L$ is the pressure drop (Pa/m) of gas in the fixed bed. As shown, $\Delta P/L$ is significantly affected by the particle size of Mg₂SiO₄ ($d_{\rm p}$). Accordingly, a fluidized-bed reactor might be used to overcome the stricter particle-size constraints of a fixed bed. The pressure of flue gas should be elevated when fluidized-bed reactors are used. However, use of fluidized-bed reactors does not necessarily entail the increase in the overall CO₂ capture expense due to their advantages over fixed bed reactors.

Other Compounds in Mg-Rich Minerals. Mg-rich minerals in nature often contain other metals in the form of silicates such as Fe₂SiO₄. However, because their concentrations in the minerals are very low, these metals should not have any significant effect on (R1).

 SO_x/NO_x . As with any other CO_2 separation technologies, the effects of SO_x/NO_x in flue gas on the studied CO_2 carbonation process should be considered as well. Acids resulting from the hydrolysis of SO_x/NO_x are much stronger than those from CO_2 and thus can react easily with Mg_2SiO_4 . However, because their concentrations in flue gas are only at parts per million levels (far lower than the percentage levels of CO_2), and Mg-rich minerals are not intended to be used as regenerated sorbents for CO_2 in the first place, the effects of SO_x/NO_x may be negligible.

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