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Sulfur and oxygen isotope fractionation during sulfate reduction coupled to anaerobic oxidation of methane is dependent on methane concentration



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ABSTRACT

Isotope signatures of sulfur compounds are key tools for studying sulfur cycling in the modern environment and throughout earth's history. However, for meaningful interpretations, the isotope effects of the processes involved must be known. Sulfate reduction coupled to the anaerobic oxidation of methane (AOM-SR) plays a pivotal role in sedimentary sulfur cycling and is the main process responsible for the consumption of methane in marine sediments — thereby efficiently limiting the escape of this potent greenhouse gas from the seabed to the overlying water column and atmosphere. In contrast to classical dissimilatory sulfate reduction (DSR), where sulfur and oxygen isotope effects have been measured in culture studies and a wide range of isotope effects has been observed, the sulfur and oxygen isotope effects by AOM-SR are unknown. This gap in knowledge severely hampers the interpretation of sulfur cycling in methane-bearing sediments, especially because, unlike DSR which is carried out by a single organism, AOM-SR is presumably catalyzed by consortia of archaea and bacteria that both contribute to the reduction of sulfate to sulfide.

We studied sulfur and oxygen isotope effects by AOM-SR at various aqueous methane concentrations from 1.4 ± 0.6 mM up to 58.8 ± 10.5 mM in continuous incubation at steady state. Changes in the concentration of methane induced strong changes in sulfur isotope enrichment (ε^{34} S) and oxygen isotope exchange between water and sulfate relative to sulfate reduction ($\theta_{\rm O}$), as well as sulfate reduction rates (SRR). Smallest ε^{34} S ($21.9\pm1.9\%$) and $\theta_{\rm O}$ (0.5 ± 0.2) as well as highest SRR were observed for the highest methane concentration, whereas highest ε^{34} S ($67.3\pm26.1\%$) and $\theta_{\rm O}$ (2.5 ± 1.5) and lowest SRR were reached at low methane concentration. Our results show that ε^{34} S, $\theta_{\rm O}$ and SRR during AOM-SR are very sensitive to methane concentration and thus also correlate with energy yield. In sulfate–methane transition zones, AOM-SR is likely to induce very large sulfur isotope fractionation between sulfate and sulfide (i.e. >60%) and will drive the oxygen isotope composition of sulfate towards the sulfate–water oxygen isotope equilibrium value. Sulfur isotope fractionation by AOM-SR at gas seeps, where methane fluxes are high, will be much smaller (i.e. 20 to 40%).

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1. Introduction

1.1. Sulfur and oxygen isotopes as tracers of present and past environmental processes

Sulfur isotopes are essential for the reconstruction of sulfur cycling through earth's history. They have been used to explore

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themes such as Earth's early atmosphere (Farguhar et al., 2000; Ono et al., 2005), the antiquity of sulfate reduction (Philippot et al., 2007; Shen et al., 2009), the evolution of atmospheric oxygen content over the Phanerozoic (Berner et al., 2000), and biogeochemical sulfur cycling over a broad range of geological time periods (e.g. Burdett et al., 1989; Canfield and Teske, 1996; Paytan et al., 1998; Ohkouchi et al., 1999; Kampschulte et al., 2001; Strauss et al., 2001; Paytan et al., 2004; Johnston et al., 2005; Wortmann and Chernyavsky, 2007; for a review see Bottrell and Newton, 2006). In recent years, it has become evident that the oxygen isotope composition of sulfate provides information on oxidative processes within the sulfur cycle that cannot be elucidated by the analysis of sulfur isotopes alone (Turchyn and Schrag, 2004). The study of the oxygen isotope composition of sulfate has turned out to be crucial for the understanding of deep biosphere sulfur cycling (Böttcher et al., 1998, 2000; Wortmann et al., 2007; Riedinger et al., 2010; Antler et al., 2013), oxidative sulfur cycling at the sediment-water interface (Ku et al., 1999; Böttcher and Thamdrup, 2001; Böttcher et al., 2001), the exploration of the formation of phosphatic laminites (Arning et al., 2009) and diagenetic gypsum (Pierre, 1985; Pirlet et al., 2010). Furthermore, the oxygen isotope composition of sulfate has been found to be a useful tool for evaluating sulfur cycling in soils and aquifers contaminated with aromatic hydrocarbons (Knöller et al., 2006, 2008), as well as for fingerprinting of microbial activity in acid mine drainage (e.g. Balci et al., 2007; Brunner et al., 2008; Heidel and Tichomirowa, 2011; Müller et al., 2013).

1.2. Sulfate reduction coupled to the anaerobic oxidation of methane

Sulfate reduction coupled to the anaerobic oxidation of methane (AOM-SR) is considered a process of worldwide relevance, because it is responsible for methane consumption in anoxic marine environments and thereby limits the escape of this potent greenhouse gas from the seabed (Reeburgh, 2007). Availability of dissolved methane regulates AOM-SR, with maximum solubility of methane being dependent mainly on pressure and temperature conditions. Consequently, methane availability varies strongly between different sites where AOM-SR is active, which can be gas seeps, mud volcanoes, hydrate fields or marine sediments rich in organic matter (Knittel and Boetius, 2009), Additionally, at sites where AOM-SR occurs, methane concentrations can vary strongly depending on spatial and temporal variations in methane supply and microbial turnover rates. Allegedly, AOM-SR is catalyzed by syntrophic consortia of methanotrophic archaea (ANME) that generally appear to operate in concert with Desulfosarcina-like (DSS) bacterial partners belonging to the Deltaproteobacteria class (for review, see Knittel and Boetius, 2009) according to the net reaction:

$$CH_4 + SO_4^{2-} \rightarrow HCO_3^- + HS^- + H_2O$$

 $\Delta G^0 = -16.6 \text{ kJ/mol}$ (1)

Recently Milucka et al. (2012) proposed that AOM-SR may be carried out by the ANME and not (only) by their bacterial partners and suggested that the ANME reduce sulfate to zero-valent sulfur, which is disproportionated to sulfate and sulfide by the DSS (Fig. 1b). Immuno-assay studies of the same AOM consortium found no indication that the ANME possess ATP sulfurylase (Sat) or dissimilatory sulfite reductase (Dsr), key enzymes of the classic pathways to activate sulfate to APS and to produce the final product, sulfide (Milucka et al., 2012).

The possibility that ANME use an alternative sulfate reduction pathway to canonical DSR and partner with sulfur disproportionating DSS that likely use the canonical DSR pathway by partially operating it in reverse direction (Frederiksen and Finster, 2003;

Finster et al., 2013; Fig. 1) raises the question if the relationship between sulfur and oxygen isotope fractionation observed for AOM-SR may be similar to the observed relationship for DSR, despite the differences in metabolic pathways.

1.3. Sulfur and oxygen isotope effects by DSR – the links between reversibility, energy yield and sulfate reduction rate

Sulfur isotope enrichment effects (ε^{34} S) reported for DSR cover a wide range from observations of -3% to the theoretical sulfur isotope equilibrium between sulfate and sulfide at +75%, or even higher (for a review, see Sim et al., 2011a). The oxygen isotope effects by DSR are somewhat more subtle, driving the oxygen isotope composition of sulfate towards the isotope equilibrium fractionation value between sulfate and water (Mizutani and Rafter, 1969; Fritz et al., 1989; Brunner et al., 2005) of approximately \sim 28% at 4°C (Zeebe, 2010). There are also cases where no oxygen isotope fractionation was detected (Turchyn et al., 2010).

Sulfur and oxygen isotope effects during DSR are the result of a sequence of isotope effects intrinsic to enzymatically-catalyzed steps in the sulfate reduction cascade and of the reversibility of this process (Fig. 1a). The extent to which these intrinsic isotope effects are expressed as overall isotope fractionation by DSR depends on how much back flux relative to the forward flux (i.e. reversibility) occurs in single steps (e.g. Rees, 1973; Brüchert, 2004; Brunner and Bernasconi, 2005; Farquhar et al., 2008; Bradley et al., 2011). The effects of increasing reversibility are two-fold. First, the sulfur isotope offset (expressed as ε^{34} S) between substrate sulfate and produced sulfide increases. Second, the oxygen isotope exchange between sulfate and water, mediated by rapid oxygen isotope exchange between sulfur-oxy intermediates in the DSR pathway, relative to the sulfate reduction rate (overall expressed as θ_0 , see Brunner et al., 2012 and references therein) also increases.

The reversibility of DSR has long since been recognized as the main controlling factor for the expression of sulfur and oxygen isotope effects, while the parameters that control the reversibility of the DSR network and its individual steps remain less certain. Nevertheless, it is evident that ultimately, thermodynamics control the reversibility of individual biochemical reactions, and thereby the reversibility of a sequence of such steps. This can best be exemplified for a simple enzymatic reaction from a substrate to a product (A \rightarrow P). The reversibility is given by the relationship

$$\frac{f_{-}}{f_{+}} = \frac{[P]}{K_{e} \cdot [A]} = e^{\Delta G/(R \cdot T)}$$
(2)

where f_- and f_+ denote back and forward flux, K_e the equilibrium constant ($K_e = e^{-\Delta G/(R \cdot T)}$), and ΔG the free energy of the reaction (for a discussion, see Holler et al., 2011a). The net flux ($f_{\rm net}$) is given as

$$f_{\text{net}} = f_{+} - f_{-} = f_{+} \cdot \left(1 - \frac{[P]}{K_{e} \cdot [A]} \right)$$
$$= f_{+} \cdot \left(1 - e^{\Delta G/(R \cdot T)} \right) \tag{3}$$

Eqs. (2) and (3) show that if the free energy of the reaction is small, which is the case when the reaction approaches thermodynamic equilibrium, reversibility becomes large (i.e. close to unity) while $f_{\rm net}$ becomes small (i.e. close to zero). The above equations do not imply that thermodynamics govern kinetics; they merely state that for a certain reaction rate (i.e., f_+), thermodynamics determine the rate in the opposite direction (i.e., f_-).

Intuitively, one also expects that reaction rates control the expression of isotope effects: reaction steps in DSR that proceed rapidly are assumed to deplete internal sulfur-oxy pools and

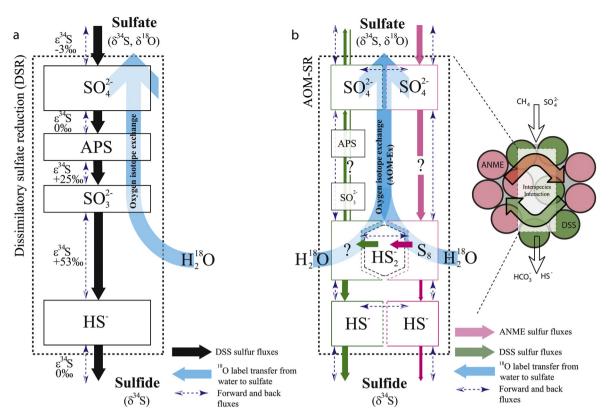


Fig. 1. Scheme of AOM-SR and putative mass and isotope fluxes in DSR. (a) Mass fluxes, sulfur isotope fractionation and oxygen isotope exchange for DSR. In DSR, extracellular sulfate is transported into the cell and subsequently reduced to sulfide via sulfur-oxy intermediates such as adenosine phosphosulfate (APS) and sulfite. The sulfur isotope enrichment fractionation factors $\varepsilon^{34}S_{APS-sulfite}$ and $\varepsilon^{34}S_{sulfite-sulfide}$ are assumed to be +25% and +53%, respectively. The DSR pathway is reversible and the ratio of individual forward and backward fluxes is assumed to be thermodynamically controlled. Oxygen isotope exchange between water and sulfate is mediated via oxygen isotope exchange between water and sulfate and subsequent back flux to sulfate. (b) AOM-SR is purportedly catalyzed by two phylogenetically distinct partner organisms – ANME and DSS – in a syntrophic consortium. Recently, it was proposed that likely both organisms are involved in the reduction of sulfate to sulfate, where the ANME partner reduces sulfate to zero-valent sulfur by an unknown pathway and the DSS partner disproportionates zero-valent sulfur to sulfide and back into sulfate (Milucka et al., 2012), potentially by partially reversing the DSR pathway. Probably, zero-valent sulfur is shuttled from the ANME to the DSS in the form of disulfide (HS $^-_2$).

thereby limit the back-flow of isotopically altered sulfur compounds which leads to a suppression of the intrinsic isotope effects, leading to the conclusion that there should be a link between sulfur and oxygen isotope fractionation and SRR (Böttcher et al., 1998, 1999; Sim et al., 2011b and references therein; Antler et al., 2013; Leavitt et al., 2013). Postulating a relationship between SRR and sulfur and oxygen isotope effects ignores that a back reaction may scale with the forward reaction, keeping the ratio between the fluxes constant, but increasing the observed net reaction. Such a scenario is given when DSR cultures grow, a reason why, typically, ε^{34} S and θ_0 are compared to cell specific sulfate reduction rates (SRR_{cs}; Harrison and Thode, 1958; Chambers et al., 1975; Chambers and Trudinger, 1979; Canfield, 2001; Detmers et al., 2001; Kleikemper et al., 2004: Johnston et al., 2007: Turchyn et al., 2010: Sim et al., 2011b). Even the normalization of SRR to cell numbers may not resolve this issue entirely, because sulfate reducing microorganisms can increase their pool of enzymes, thereby changing SRR_{cs} without affecting ε^{34} S or θ_0 . Nevertheless, a compilation by Sim et al. (2011b) showed that there is correlation between ε^{34} S and SRR_{cs}, which appears to be species-specific. Furthermore, experiments with batch cultures of Desulfovibrio sp. DMSS-1 provided with different electron donors and experiments with continuous cultures of D. sp. DMSS-1 under different concentrations of lactate reveal that low SRR_{cs} and high $arepsilon^{34}$ S coincide with refractory organic substrates and substrate limitation whereas high SRR_{cs} and low $\varepsilon^{34}S$ coincide with easily degradable substrates and at high substrate concentrations. These findings were corroborated by Leavitt et al. (2013) who observed that ε^{33} S and ε^{34} S by the sulfate reducing bacterium Desulfovibrio vulgaris Hildenborough increase when lactate concentrations decrease, leading to the hypothesis that combined ³³S-³⁴S signatures could be a proxy to reconstruct environmental sulfate reduction rates in the geologic past.

1.4. Sulfur and oxygen isotope effects of sulfate reduction coupled to the anaerobic oxidation of methane

So far, direct measurements of the sulfur and oxygen isotope effects of AOM-SR in culture are lacking. Sulfur isotope trends from sulfate, sulfide, and sulfide minerals obtained from marine sediments close to active or past sulfate-methane transition zones (SMTZs) indicate that similar to DSR, AOM-SR produces sulfide that is depleted in ³⁴S relative to sulfate (e.g. Böttcher et al., 2003; Jørgensen et al., 2004; Borowski et al., 2013). Judging from the scarce oxygen isotope data available for natural environments that may have been affected by AOM-SR (i.e. where sulfate concentrations reach depletion; Aharon and Fu, 2000, 2003; Antler et al., 2013) there is no oxygen isotope signature for AOM-SR that is distinct from the isotope signature of DSR.

While the details of the biochemical reactions and couplings of AOM-SR remain unresolved (Milucka et al., 2012; Fig. 1), it is obvious that the overall process has a low energy yield (Eq. (1)), which is shared between the microorganisms that form the consortium. Sulfate reduction, methane oxidation, and each individual enzymatic reaction are expected to operate very close to thermodynamic equilibrium and thus must operate reversibly, a concept recently tested for AOM-SR. Yoshinaga et al. (2014) showed that enzymatic reversibility in AOM at low substrate concentra-

tions strongly influences carbon isotope signatures of methane and dissolved inorganic carbon, Holler et al. (2011a) previously found that cultures of highly enriched AOM consortia converted labeled catabolic products ¹⁴C-bicarbonate and ³⁵S-sulfide to ¹⁴C-methane and ³⁵S-sulfate, respectively, thus indicating full process reversibility. However, there remains the caveat that ³⁵S label transfer from sulfide to sulfate does not unambiguously prove the full reversibility of sulfur transformations carried out by the ANME, as abiotic sulfur isotope exchange between ³⁵S sulfide and zero-valent sulfur followed by sulfur disproportionation by DSS, as lately postulated by Milucka et al. (2012), could be responsible for the formation of ³⁵S sulfate. Nevertheless, under typical thermodynamic constraints for AOM-SR, the reversibility of sulfur transformations should be high and strongly depend on methane concentrations (Eq. (2)). We, therefore, expect to observe a strong dependency of $\varepsilon^{34}S$ and θ_0 of AOM-SR to methane concentrations, with large sulfur isotope fractionation up to the theoretical value of approximately 75% and rapid oxygen isotope exchange between sulfate and water relative to the SSR at low methane concentrations.

2. Materials and methods

2.1. Organisms and medium composition

Biomass slurries were prepared from naturally enriched environmental samples of microbial mats from Black Sea microbial reefs at 44°46′N, 31°59′E (RV Meteor cruise M72/2, 2007, Project MUMM, Max Planck Institute for Marine Microbiology, Bremen). For a microbiological and molecular characterization of the mats see Basen et al. (2011). After recovery the samples were stored under methanotrophic conditions with 0.2 MPa of methane in anoxic artificial seawater (Widdel and Bak, 1992). During storage, AOM-SR activity was monitored by measuring sulfide production. Medium was exchanged on a weekly basis to avoid inhibition of AOM-SR by sulfide accumulation. All manipulations of biomass were carried out under anoxic conditions. Mat pieces were homogenized using a tissue grinder prior to use in the experiments. The incubations were carried out at 20 °C with anoxic artificial seawater medium. Artificial seawater medium with sulfate (~8 mM between 0 and 110 days, ~10 mM between 110 and 129 days) was prepared with ammonium (4 mM), phosphate (1 mM), trace elements, vitamins, bicarbonate (30 mM) and sulfide (0.5 mM) after (Widdel and Bak, 1992). Sulfate concentrations were adjusted to relatively low values as compared to seawater composition to facilitate measurements of changes in the sulfur and oxygen isotope composition of sulfate (δ^{34} S, δ^{18} O, respectively) at low overall turnover rates. However, sulfate concentrations were well within the range of natural concentrations, which were measured to be as low as 6 mM in microbial mat chimneys in the Black Sea (Arnds, 2009). Prior to incubation the medium was spiked with $^{18}\text{O-water}$ (+610% vs. VSMOW). This amendment is necessary to obtain reliable values for θ_0 , but comes at the price that a direct comparison to environmental data is no longer possible (i.e. due to a much higher equilibrium value). The dry weight (DW) was determined at the start of the long-term incubation experiment from an aliquot of 10 mL inoculated medium containing homogenized microbial mat samples which was dried at 80 °C for 24 h after separation from the fluid by centrifugation (10000 rpm). The initial biomass dry weight concentration was 4.6 $g_{DW} L^{-1}$. Due to the strong tendency for biomass aggregation and sedimentation, and the very low dilution rates (D = flux/incubation reactor volume) of 1.7 d⁻¹, the biomass was completely retained within the incubation vessel during long-term incubation.

2.2. High pressure incubations, apparatus and sampling

To carry out the AOM-SR experiments at different methane concentrations we used a two stage high-pressure continuous incubation experimental system (HP-CI, see Fig. 2 for setup scheme). A detailed description of the HP-CI system is given in Deusner et al. (2010). All experiments presented here were carried out in continuous operation at steady state to allow for repeated measurements, particularly at low methane concentrations. The two-stage HP-CI system consists of pressure vessels for gas enrichment and biomass incubation. Biomass incubation was carried out at a constant hydrostatic pressure of 15 MPa to avoid effects of pressure changes on the culture. Liquid samples during continuous operation of the HP-CI system were taken directly into glass vials closed with butyl stoppers from the medium reservoir bottle and downstream of both reaction vessels for gas enrichment and biomass incubation, respectively, at sampling ports (Fig. 2). Samples for sulfide and sulfate (concentration and isotope composition) were collected directly into closed vials that contained zinc acetate solution (5% w/v), which prevents loss of gaseous hydrogen sulfide that might out-gas together with methane if sampling was done into open vials. Samples for methane concentrations were collected into closed vials containing a saturated sodium chloride solution. The methane concentration was measured from the headspace of the sampling vials.

The total duration of high-pressure biomass incubation was 129 days (February 2009-June 2009). During this period the influent concentration of aqueous methane was stepwise increased from 1.4 ± 0.6 mM to 58.8 ± 10.5 mM, with intervals of batch incubation, flushing and stabilization between measurement periods. During stabilization intervals the bioreactor system was first flushed at fourfold increased flow rate (2 mL/min instead of 0.5 mL/min during measurement intervals) and after that operated at the respective incubation conditions for at least 42 h prior to each measurement period. The system was judged to be at steady state when the reactor incubation volume (420 mL) was exchanged at least three times at the respective methane concentration and flow rate. Once steady state was reached, several samples were taken without biomass depressurization and the system was subsequently adjusted to a higher methane concentration. At each methane concentration step, incubation was carried out for 1.8 to 6.2 days, with repeated sampling and analysis of influent and effluent fluids from the incubation vessel.

2.3. Measurement of sulfate, sulfide and methane concentrations

Precipitated zinc sulfide was re-suspended and the total sulfide concentration (here abbreviated as [HS $^-$]) was determined colorimetrically according to the methylene-blue reaction in a miniaturized assay (Aeckersberg et al., 1991). The sulfate concentration ([SO_4^2]) was measured from the sample supernatant with ion chromatography with a SYKAM Solvent Delivery System coupled to a Millipore Waters Conductivity Detector Model 430 and a Waters IC PAK column or a Metrohm Advanced Compact IC 761 equipped with an electrical conductivity detector and a Metrosep A Supp 5–50 column. Methane concentrations were determined with a GC 14B gas chromatograph (Shimadzu) equipped with a Supel-Q Plot column (30 m \times 0.53 mm; Supelco) and a flame ionization detector. The carrier gas was N2 at a flow rate of 3 mL min $^{-1}$. The column temperature was 110 °C.

2.4. Isotope analysis

Sulfate was precipitated with an acidified barium chloride (BaCl₂) solution as barium sulfate (BaSO₄). The BaSO₄ precipitate

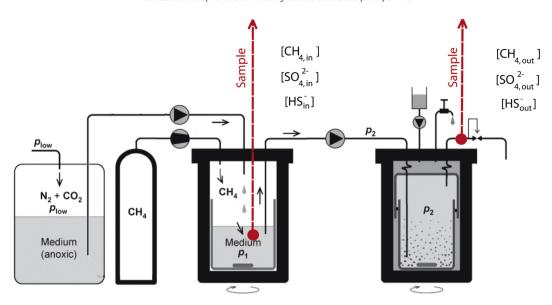


Fig. 2. High pressure flow-through incubation system. Two-stage experimental system for high pressure continuous incubation experiments (Deusner et al., 2010). Anoxic medium is delivered at a constant rate with a high pressure pump into the gas saturation vessel which is pressurized with methane to a desired pressure (p_1). Continuously stirred medium is rapidly enriched with methane according to the pressure-temperature conditions. The methane-enriched medium is delivered with a second pump into the incubation vessel. The pressure (p_2) inside the incubation vessel is regulated with a backpressure valve downstream of the vessel. To prevent effects from corrosion of the high-pressure components, both pressure vessels were equipped with internal PEEK containers. Samples of influent and effluent sulfate, sulfide and methane were taken upstream and downstream of the incubation vessel. Sampling points are shown in red. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

was recovered by centrifugation, where the supernatant was decanted. Samples were two times re-suspended with de-ionized water, followed by centrifugation. Subsequently, the BaSO₄ was re-dissolved with a strong chelator (DTPA, Bao, 2006) to liberate water from the medium (enriched in ¹⁸0) that is potentially trapped in the BaSO₄ precipitate. The BaSO₄ was re-precipitated by addition of acidified barium chloride (BaCl2) solution, washed with de-ionized water and dried at 50°C for at least 24 h. Zinc sulfide was converted into silver sulfide (Ag₂S) with a silver nitrate (AgNO₃) solution. The precipitate was rinsed with de-ionized water and washed with 1 M NH₄OH solution to remove any silver chloride. After repeating the rinse step with de-ionized water the precipitate was dried at 50 °C overnight. Sulfur and oxygen isotope measurements of solid samples were performed with a gas source stable isotope ratio mass spectrometer (GS-IRMS) model DeltaPLUS (Finnigan; Thermo Fisher Scientific, Waltham, MA, USA) at the Max Planck Institute for Marine Microbiology in Bremen, Germany. The sulfur isotopic composition is reported with respect to the Vienna Cañon Diablo Troilite (VCDT) standard. The sulfur isotope measurements were calibrated with reference materials NBS 127 $(\delta^{34}S = 20.3\%, \text{ Hut, } 1987) \text{ and IAEA-SO-6 } (\delta^{34}S = -34.1\%). \text{ The}$ true value for NBS 127 may be higher, i.e. 21.1% (Halas and Szaran, 2001), which would mean that our values would need to be slightly corrected ($\delta^{34}S_{corr} = \delta^{34}S_{reported} \cdot 1.01 + 0.3$), which shifts our values maximally by 0.5%, and does not affect the outcome of our study. The oxygen isotopic composition of sulfate and water is reported with respect to Vienna Standard Mean Ocean Water (VSMOW). The oxygen isotope measurements were calibrated with NBS 127 ($\delta^{18}O = 8.6\%$; Boschetti and Iacumin, 2005) and IAEA-SO-5 ($\delta^{18}O = 12.0\%$; Boschetti and Iacumin, 2005) and IAEA-SO-6 ($\delta^{18}O = -11.3\%$; Halas et al., 2007). The standard deviation of the measurements was less than 0.2% for $\delta^{34} S$ values and less than 0.3% for $\delta^{18}O$ values. The oxygen isotope composition of the experiment medium was determined via CO₂ equilibration and subsequent isotope analysis of CO₂ by GS-IRMS (Delta V, Finnigan; Thermo Fisher Scientific, Waltham, MA, USA) at the Geological Institute, ETH Zurich, Switzerland. The system was calibrated with the international reference waters SMOW, GISP and SLAP and the oxygen isotope ratios are reported in the δ -notation relative to VS-

MOW. The reproducibility of the measurements based on repeated measurements of standards is less than 0.1%.

2.5. Calculation of SRR, ε^{34} S and θ_0

Sulfate reduction rates and isotope fractionation values were calculated from mass and isotope balances, respectively. For calculations, steady state conditions were assumed. Sulfate reduction rates were quantified by either calculating sulfide production or sulfate consumption rates based on the mass conservation Eqs. (4) and (5), respectively.

$$SRR = D \cdot ([HS_{out}^{-}] - [HS_{in}^{-}])$$
(4)

$$SRR = D \cdot \left(\left\lceil SO_{4 \text{ in}}^{2-} \right\rceil - \left\lceil SO_{4 \text{ out}}^{2-} \right\rceil \right) \tag{5}$$

Terms in brackets are concentration values for sulfate and sulfide in the inflow (in) and outflow (out) of the incubation vessel, respectively, whereas D is the hydraulic dilution rate (dimension: t^{-1}). Consequently, in continuous incubation at steady state, Eq. (6) should hold:

$$[HS_{out}^{-}] - [HS_{in}^{-}] = [SO_{4 in}^{2-}] - [SO_{4 out}^{2-}]$$
(6)

Under steady state conditions, the sulfur isotope composition of effluent sulfate and sulfide only depends on SRR (i.e. consumption of sulfate and production of sulfide within the reactor) and on the sulfur isotope enrichment factor by AOM-SR, i.e. ε^{34} S. The isotope enrichment factor, ε^{34} S, is related to the isotope fractionation factor α^{34} S, which is commonly used in studies on sulfur isotope fractionation, according to Eq. (7).

$$\varepsilon^{34} S = 1000\% \cdot (1 - \alpha^{34} S) \tag{7}$$

The value of $\varepsilon^{34}\mathrm{S}$ was calculated according to Eq. (8) and Eq. (9) (for derivation see Supplementary Information). Data presented in this report consider averaged values calculated based on sulfate and sulfide mass balances from multiple measurements at steady state.

$$\begin{split} \varepsilon^{34} S &\approx \delta^{34} S_{SO_{4 \text{ out}}^{2-}} - \delta^{34} S_{HS_{\text{out}}^{-}} \\ &+ \frac{[HS_{\text{in}}^{-}]}{[HS_{\text{out}}^{-}] - [HS_{\text{in}}^{-}]} \cdot \left(\delta^{34} S_{HS_{\text{in}}^{-}} - \delta^{34} S_{HS_{\text{out}}^{-}} \right) \end{split} \tag{8}$$

$$\varepsilon^{34} S \approx \frac{[SO_{4 \text{ in}}^{2-}]}{[SO_{4 \text{ in}}^{2-}] - [SO_{4 \text{ out}}^{2-}]} \cdot \left(\delta^{34} S_{SO_{4 \text{ out}}^{2-}} - \delta^{34} S_{SO_{4 \text{ in}}^{2-}}\right)$$
(9)

The bulk apparent oxygen isotope exchange between sulfate and water (AOM-Ex) mediated by AOM-SR is calculated according to Eq. (10).

$$\begin{split} \text{AOM-Ex} &= D \cdot \left[\text{SO}_{4 \, \text{in}}^{2-} \right] \\ &\cdot \frac{\delta^{18} \text{O}_{\text{SO}_{4 \, \text{out}}^{2-}} - \delta^{18} \text{O}_{\text{SO}_{4 \, \text{in}}^{2-}}}{\delta^{18} \text{O}_{\text{water}} + \varepsilon^{18} \text{O}_{\text{H}_2\text{O} \Leftrightarrow \text{SO}_4^{2-}} - \delta^{18} \text{O}_{\text{SO}_{4 \, \text{out}}^{2-}}}, \end{split} \tag{10}$$

where $\varepsilon^{18}O_{\text{H}_2\text{O}\Leftrightarrow\text{SO}_4^2}$ corresponds to the oxygen isotope equilibrium fractionation between sulfate and water at 20 °C (\sim 24‰, Zeebe, 2010). The oxygen isotope exchange parameter θ_0 is defined according to Eq. (11) (Brunner et al., 2012).

$$\theta_{\rm O} = \frac{\rm AOM\text{-}Ex}{\rm SRR} \tag{11}$$

2.6. Calculation of Gibbs free energies

Gibbs free energies of reaction (ΔG_r) were calculated based on energies of formation (Thauer et al., 1977) for overall AOM-SR (Eq. (1)). The following activity coefficients for ionic species were considered: SO_4^{2-} : 0.2, HCO_3^{-} : 0.7, HS^{-} : 0.6. The activity coefficient for aqueous CH₄ was calculated to be approximately 1.2. The effect of increased hydrostatic pressure (15 MPa) was considered in thermodynamic calculations.

2.7. Calculation of SRR_{bm}

Biomass specific SRR (SRR_{bm}, dimension mmol/g_{DW}/d) were calculated based on the initial biomass dry weight concentration of 4.6 g_{DW}/L. This procedure neglects biomass growth effects. Growth of organisms performing AOM-SR is generally slow (e.g. Nauhaus et al., 2007) and expected to be of minor importance for the observed isotope effects. Black Sea biomass performing AOM-SR is mainly constituted by large aggregates consisting of active biomass, inactive or non-AOM-SR cells, as well as extracellular polymeric substances and carbonate precipitates (Michaelis et al., 2002). Since biomass growth could not be measured directly in a robust way and biomass sampling from the high-pressure experimental system would have caused significant process disturbance and biomass loss, uncertainties were accounted for by showing a possible range of growth related changes of $\mbox{SRR}_{\mbox{\scriptsize bm}},$ as highlighted in Figs. 4 and 5, respectively. This range is calculated based on previous estimates for growth of biomass performing AOM-SR. Ambient to low pressure incubations found doubling times of 1.1-1.5 (Girguis et al., 2005), 3.8 (Meulepas et al., 2009) and 7 months (Nauhaus et al., 2007). Similar growth rates were also found for thermophilic cultures with a doubling time of 2.3 months (Holler et al., 2011b) and in high-pressure bioreactor studies (2.5 months, Zhang et al., 2011).

3. Results

This study reports $\varepsilon^{34} S$ and θ_0 for a naturally enriched biomass actively catalyzing AOM-SR during a continuous incubation experiment where the concentration of aqueous methane was stepwise increased over 92 days from 1.4 ± 0.6 mM to 58.8 ± 10.5 mM, with intervals of stabilization between measurement periods (Fig. 3a).

3.1. Sulfate and sulfide concentration

During the course of our experiment, the offset in concentration of sulfate or sulfide between the inflow and outflow increases with increasing methane concentration (Fig. 3). Sulfate removal exceeded sulfide production by a factor of 1.9 relative to the expected 1:1 ratio (Eq. (3)) at highest SRR. Theoretically, sulfide and sulfate concentration balances are equally well suited as a measure for SRR under controlled incubation conditions (Eq. (6)), and both sulfide production and sulfate removal are considered in calculation of SRR and SRR_{bm} throughout this study. As a consequence, the variance in SRR_{bm} is large (Table 1, Figs. 4 and 5). We have not found any evidence that sulfide is lost or oxidized during fluid sampling or analysis. The samples were instantaneously fixed by the formation of ZnS to prevent sulfide oxidation during sampling. However, some effluent samples contained measurable amounts of zero-valent sulfur (up to 18%, on average 2.2% of measured sulfide). This amount of dissolved zero-valent sulfur alone cannot explain the deviations between sulfate and sulfide mass balance. Considering the low solubility of zero-valent sulfur and polysulfides (Kamyshny et al., 2004, 2006, 2007), it is likely that solid or biomass associated sulfur accumulated during the experiments, which may account for the discrepancy in the sulfur mass balance but was not analyzed in our experiments.

3.2. Isotope composition of sulfate and sulfide

The δ^{34} S values of influent and effluent sulfate and sulfide, as well as δ^{18} O values of influent and effluent sulfate are shown in Fig. 3. Our isotope data set shows two unexpected features. First, the δ^{34} S values of effluent sulfide at low methane concentrations are associated with a large variance. We attribute this large range to the mixing of the small amount of sulfide that was actually produced by AOM-SR at these low methane concentrations with sulfide from the medium, which was used as a reducing agent. Under these conditions, the scatter in the data is amplified by slight changes in the flow rate of medium through the reactor. At higher rates of AOM-SR (higher methane concentrations), the variance becomes small because the relatively small amount of sulfide from the medium does not affect the sulfur isotope signature in the sulfide of the outflow to a measurable degree. Secondly, in the last experimental step, at a methane concentration of approximately 60 mM, the δ^{34} S value and to a lesser extent the δ^{18} O value of sulfate from the inflow was high. This shift in the isotope composition does not have an effect for the subsequent calculation of ε^{34} S and θ_0 because the δ^{34} S and values δ^{18} O for effluent sulfate mirrored this change. Contamination of the medium reservoir with AOM-SR active biomass that would have altered the isotope composition of sulfate in the inflow is highly unlikely because of the extremely low growth rates of AOM consortia. Likely, the change in $\delta^{34}S$ and $\delta^{18}O$ of influent sulfate was caused by the use of a different stock sulfate when the medium for the experimental phase between 40 and 120 days was prepared.

3.3. ε^{34} S, AOM-Ex and θ_{0}

The calculated ε^{34} S, AOM-Ex and θ_{0} are shown in Figs. 4 and 5 as a function of methane concentration (Figs. 4a, 5a), biomass specific rate SRR_{bm} (Figs. 4b, 5b) and free energy yield ΔG_{r} (Figs. 4c, 5c). The large variance associated with the calculated maximum values for ε^{34} S and θ_{0} are due to deviations in influent sulfide concentrations and low turnover rates at low methane concentrations, and in the case of ε^{34} S also due to scatter in the δ^{34} S values of effluent sulfide. Changes in the concentrations of aqueous

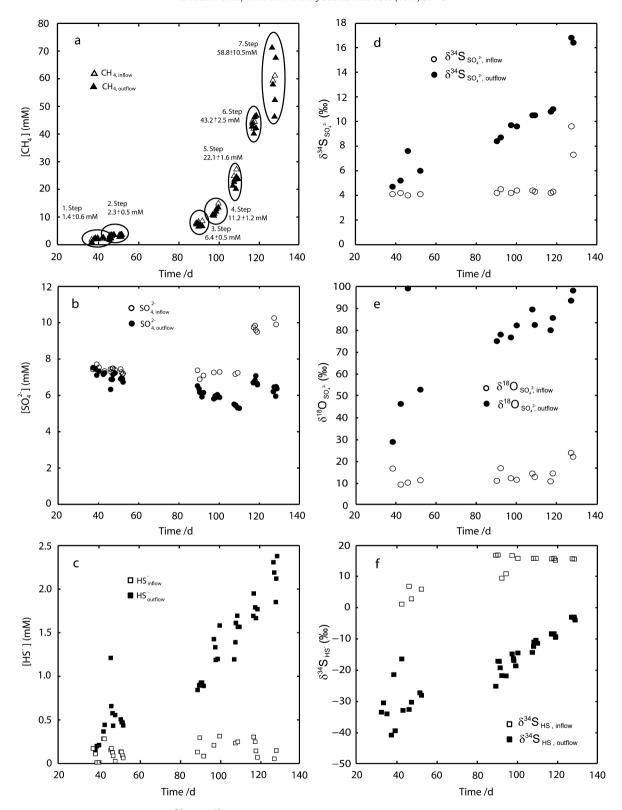


Fig. 3. Methane, sulfate and sulfide concentrations, δ^{34} S and δ^{18} O values of influent and effluent sulfate and sulfide during continuous incubation steps. Influent and effluent concentrations of (a) methane (b) sulfate, and (c) sulfide during seven incubation steps spanning the experimental period. Data are grouped representing single incubation steps at different methane concentrations (a, black ellipsoids). Data indicate that sulfate is consumed (b) and sulfide produced (c). Sulfide was present in the influent because sodium sulfide was used as reducing agent. Sulfur and oxygen isotope compositions of sulfate (d and e, respectively) show that effluent sulfate is enriched in the heavy isotopes ³⁴S and ¹⁸O relative to influent sulfate. Effluent sulfide is depleted in ³⁴S relative to influent sulfide (f).

methane during AOM-SR coincide with strong changes in sulfate reduction rates, but also in AOM-Ex, ε^{34} S and θ_0 . We observe the highest SSR_{bm} and AOM-Ex at the highest methane concentrations, where ε^{34} S and θ_0 are lowest (21.9 \pm 1.9‰ and 0.5 \pm 0.2, re-

spectively). The maximum values for ε^{34} S of $67.3 \pm 26.1\%$ and θ_0 of 2.5 ± 1.5 were observed at the lowest methane concentration, which coincide with the lowest sulfate reduction rates and AOM-Ex (Figs. 4, 5).

 Table 1

 Sulfate removal and sulfide production at different methane concentrations. $SRR_{sulfate}$ and $SRR_{sulfate}$ are volumetric rates calculated from differences in influent and effluent sulfate or sulfide concentrations, respectively (values are shown as arithmetic means \pm SD, see also Table S1 in Supplementary Information for concentration values).

Step	Days	[CH ₄] (mM)	SRR _{sulfate} (mmol L ⁻¹ d ⁻¹)	SRR _{sulfide} (mmol L ⁻¹ d ⁻¹)	SRR _{sulfate} /SRR _{sulfide} (/)
1	37-43	1.4 ± 0.6	0.2 ± 0.3	0.2 ± 0.2	1.0
2	46-52	2.3 ± 0.5	0.8 ± 0.3	0.7 ± 0.1	1.1
3	89-92	6.4 ± 0.5	1.6 ± 0.4	1.3 ± 0.1	1.2
4	97-100	11.2 ± 1.2	2.3 ± 0.2	1.8 ± 0.3	1.3
5	107-110	22.1 ± 1.6	3.1 ± 0.2	2.2 ± 0.3	1.4
6	117-119	43.2 ± 2.5	5.0 ± 0.3	2.7 ± 0.2	1.9
7	127-129	58.8 ± 10.5	6.5 ± 0.4	3.5 ± 0.4	1.9

4. Discussion

4.1. Relationship between $\varepsilon^{34} {\rm S}, \, \theta_{\rm O}, {\rm SRR}_{\rm bm},$ energy yield and methane concentration

Small ε^{34} S and θ_0 coincide with high methane concentrations and relatively high energy yield, large ε^{34} S and θ_0 with low methane concentrations and low energy yield (Figs. 4a, c; 5a, c). However, $\varepsilon^{34} {\rm S}$ and $\theta_{\rm O}$ also correlate with SRR_{bm} (Figs. 4b; 5b), and conversely, SRR_{bm} correlate with methane concentrations and the calculated energy yield. This interrelation between all parameters is not unexpected for two reasons. First, ε^{34} S and θ_0 are expected to be a measure of the reversibility of AOM-SR, which should depend on the energy yield (for a simple enzymatically catalyzed reaction, see Eq. (2)). Second, the net sulfate consumption rate, SRR_{hm}, depends on the rate of individual enzymatically catalyzed reactions, which should depend on substrate and product concentrations as well (Eq. (3)). In principle, an organism could employ an increased number of enzymes to counteract low SRR_{bm} under low energy yield, however, this appears unlikely as maintaining a higher number of enzymes comes at additional energy cost.

4.2. Relationship between $\varepsilon^{34}{\rm S}$ and $\theta_{\rm O}$ for AOM-SR and comparison to DSR and sulfur disproportionation

We observe a strong correlation between $\varepsilon^{34}\mathrm{S}$ and θ_0 in AOM-SR, resonating the predictions – and to some extent observations – that such a correlation exists for the DSR pathway (Mangalo et al., 2007, 2008; Einsiedl, 2008; Turchyn et al., 2010; Brunner et al., 2012). The $\varepsilon^{34}\mathrm{S}$ – θ_0 relationship for AOM-SR from our experiment covers almost the entire range of previously observed $\varepsilon^{34}\mathrm{S}$ – θ_0 relationships for DSR, and even expands it to higher $\varepsilon^{34}\mathrm{S}$ values (Fig. 6). Notably, our observed $\varepsilon^{34}\mathrm{S}$ – θ_0 relationship for AOM-SR falls within the theoretical $\varepsilon^{34}\mathrm{S}$ – θ_0 field for DSR predicted by Brunner et al. (2012).

It is interesting to discuss the observed ε^{34} S and θ_0 for AOM-SR in light of the finding that sulfur cycling by AOM-SR may consist of two interlinked processes - sulfate reduction to zero-valent sulfur by the ANME and sulfur disproportionation by the DSS (Milucka et al., 2012). The biochemical mechanistics and potential sulfur and oxygen isotope effects of the hypothesized sulfate reduction to zero-valent sulfur by the ANME are currently unknown, but some information about the mechanics of sulfur disproportionation utilized by the DSS partner and the associated isotope effects is available. It appears that sulfur disproportionation makes use of a partially reversed DSR pathway (Frederiksen and Finster, 2003; Finster et al., 2013) which implies that if this pathway is operated in a highly reversible manner, sulfur and oxygen isotope effects for sulfur disproportionation might be similar to those for DSR. Similar to the sulfur isotope effects by DSR, $\varepsilon^{34}\mathrm{S}$ for the disproportionation of elemental sulfur to sulfate and sulfide depend on experimental conditions (i.e. type of provided sulfide scavenger) and culture-types, varying between \sim 0 and \sim 35\% (Böttcher et al.,

2005). If DSS in consortia responsible for AOM-SR operate more reversibly than these classical sulfur-disproportionating bacteria, even larger $arepsilon^{34} \mathrm{S}$ should occur. Although it cannot be excluded that DSS also contribute to overall sulfate reduction during AOM-SR via the canonical DSR pathway, it could be speculated that the similarity between the ε^{34} S- θ_0 relationship for AOM-SR and DSR is mainly caused by the similarity of the biochemical mechanics of DSR and sulfur disproportionation, the latter overprinting potential sulfur and oxygen isotope effects related to sulfate reduction by ANME. Consequently, the overlap between $\varepsilon^{34}S-\theta_0$ relationships for AOM-SR and DSR does not imply that the processes are mechanistically identical. Rather, the implication is that the underlying principles of reversibility at low energy yields apply to both processes and that it may be difficult to disentangle DSR from AOM-SR based on $\varepsilon^{34}S-\theta_0$ relationships. Although theoretically the sulfur isotope effects for sulfate reduction by ANME and sulfur disproportionation by DSS could be staggered in analogy to what has been proposed for canonical DSR combined with sulfur disproportionation (Canfield and Thamdrup, 1994), the thermodynamic constraints on the overall AOM-SR process may be the reason why our observed maximum sulfur isotope fractionation did not exceed the maximum for the sulfur isotope fractionation between sulfate and sulfide of approximately 75%. Studies targeting the multiple S isotope signature (32 S, 33 S, 34 S, 36 S) of AOM-SR have not been carried out yet, but may provide additional information that helps elucidating the mechanistic similarities and dissimilarities between DSR and AOM-SR.

4.3. Range of $\varepsilon^{34}{\rm S}$ and $\theta_{\rm O}$ for AOM-SR and application to natural systems

In our experiment on AOM-SR, we observed sulfur isotope fractionation as low as $21.9 \pm 1.9\%$, but also as high as $67.3 \pm 26.1\%$, thereby approaching the sulfur isotope equilibrium fractionation of approximately 75\%. Sulfur isotope fractionations exceeding 60\% were also observed for a pure DSR culture grown on glucose by Sim et al. (2011a), who pointed out that such large fractionations are likely much more relevant for natural environments than the multitude of significantly lower $\varepsilon^{34} S$ found for various cultures of sulfate reducers grown in the laboratory on easily degradable substrates. This raises the question whether the values of $\varepsilon^{34} S$ and θ_0 found in our steady-state experiments with a naturally enriched biomass performing AOM-SR are representative for the natural environment. We believe this is the case, as our experimental approach bridges pure culture studies and field measurements. From the strong dependency of $\varepsilon^{34} S$ and θ_{0} of AOM-SR on methane concentrations we conclude that there is a large variety of sulfur and oxygen isotope signatures for AOM-SR in natural systems, which depends on the methane availability (Fig. 6). We expect large ε^{34} S and θ_0 of AOM-SR where methane concentrations are low. A typical system where this occurs is in sediments where burial of organic matter is elevated to a degree that pore-water sulfate becomes depleted and methanogenesis produces methane

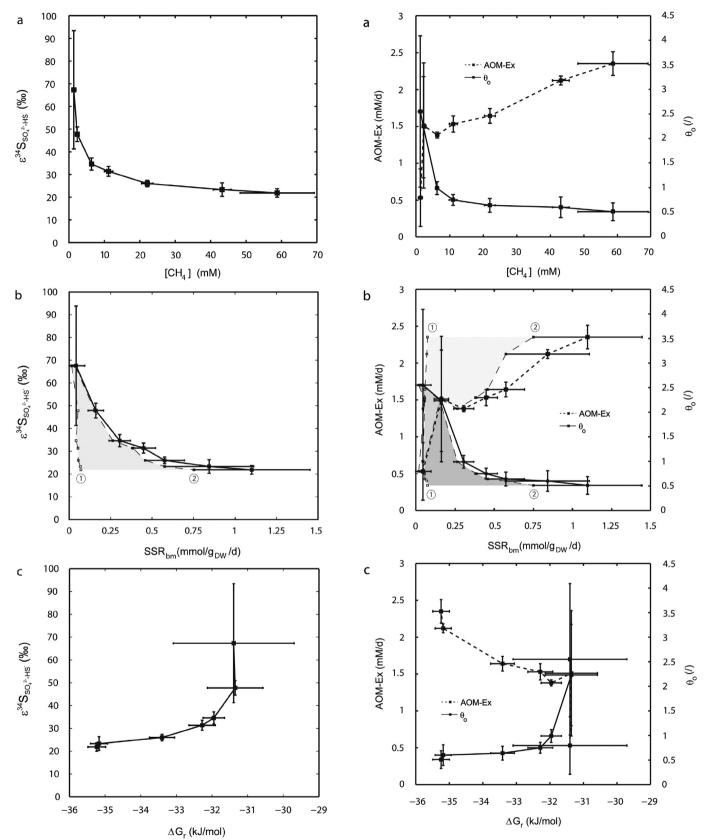


Fig. 4. Sulfur isotope fractionation (ε^{34} S) as a function of methane concentration, biomass specific rate, and Gibbs free energy yield. S-isotope fractionation values (ε^{34} S) are shown as a function of (a) methane concentration, (b) biomass specific rate (SRR_{bm}), and (c) Gibbs free energy yield (ΔG_r). Growth related uncertainties of SRR_{bm} are indicated by grey areas that span data regions assuming organism doubling times between 1 month (dashed line 1, Girguis et al., 2005), and 7 months (dashed line 2, Nauhaus et al., 2007). Data are plotted as arithmetic means \pm SD.

Fig. 5. Bulk apparent oxygen isotope exchange between sulfate and water mediated by AOM-SR and θ_0 . The two parameters AOM-Ex and θ_0 for AOM-SR are shown as a function of (a) methane concentration, (b) biomass specific rate (SRR_{bm}), and (c) Gibbs free energy yield (ΔG_r). Note, that SRR is used in the calculation of θ_0 (Eq. (11)). Growth related uncertainties of SRR_{bm} are indicated by grey areas that span data regions assuming organism doubling times between 1 month (dashed lines 1, Girguis et al., 2005), and 7 months (dashed lines 2, Nauhaus et al., 2007). Data are plotted as arithmetic means \pm SD.

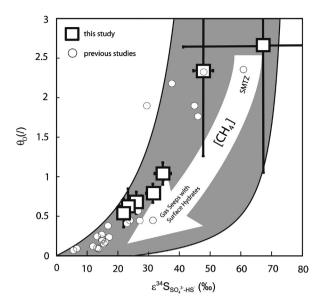


Fig. 6. Comparison of ε^{34} S- θ_0 relationships of AOM-SR to ε^{34} S- θ_0 relationships of DSR and to methane concentrations in selected marine environments. Relationship between ε^{34} S and θ_0 for AOM-SR (open squares) and DSR (open circles). Our results fall within the range of ε^{34} S- θ_0 relationships predicted for DSR based on a sulfur and oxygen isotope fractionation model (grey area; Brunner et al., 2012). The white arrow indicates the trend from high ε^{34} S and θ_0 at low methane concentrations which are typical for SMTZs in marine sediments to low ε^{34} S and θ_0 at high methane concentrations which are typical for gas seeps with surface hydrates. Data are plotted as arithmetic means \pm SD. Additional information on maximum methane concentrations in marine environments as a function of pressure and temperature is provided in Supplementary Fig. S1.

from substrates no longer available to DSR. Such systems are typically diffusion controlled and result in the development of a SMTZ where AOM-SR maintains both low methane and sulfate concentrations (e.g. Iversen and Jørgensen, 1985). Typically, minimum methane concentrations in such systems are even below the lowest methane concentrations in our experiment $(1.4 \pm 0.6 \text{ mM}, \text{Fig. 3a})$, where we observe $\varepsilon^{34}{\rm S}$ to be as high as $67.3\pm26.1\%$. Indeed, large sulfur isotope offsets between sulfate and sulfide have been found for such systems, for example in the Black Sea (Jørgensen et al., 2004) and in the Cariaco Basin (Werne et al., 2003) with 59 to 77% and 53 to 65%, respectively. However, numerous sites with high methane advection rates are also known, such as mud volcano and methane seep systems. Sediments at methane seep systems are often characterized by two-phase gas-water fluid transport and elevated concentrations of aqueous methane in equilibrium with free gas or gas hydrates (see Supplementary Fig. S1 for equilibrium concentrations) which coincide with substantial sulfate concentrations (Haeckel et al., 2004; Heeschen et al., 2007, 2011). In these environments small $\varepsilon^{34} S$ and θ_{0} of AOM-SR would be expected. Mud volcano and seep systems are not only characterized by higher methane fluxes (e.g. Lichtschlag et al., 2010, and references therein), but also by rapid changes in fluxes (e.g. de Beer et al., 2006, and references therein). Considering that our enrichments of biomass performing AOM-SR quickly adapted to the changed methane concentrations we would also expect a higher variability of expressed $\varepsilon^{34}S$ and θ_0 . Special environments to consider may be locations where sulfate-rich water comes in contact with heterogeneously distributed gas hydrates (Boetius et al., 2000; Boetius and Suess, 2004) or where gas hydrates dissociate due to natural or anthropogenic changes in bottom water temperatures and increased amounts of methane are released (Bernd et al., 2014; Biastoch et al., 2011). In such a case, locally and temporarily elevated methane concentrations could increase AOM-SR and corresponding small sulfur isotope offsets between sulfate and sulfide, as well as small $\theta_{\rm O}$ should be expected. As a consequence

of the apparent metabolic capability to adapt to changing methane concentrations, signals from transient events with high methane availability might be imprinted in the S-isotope record and help to reconstruct past methane release events.

5. Conclusions

With naturally enriched, heterogeneous biomass that performs AOM-SR under controlled incubation conditions, we observe a strong dependency of ε^{34} S and θ_0 on methane (substrate) concentration. This finding corroborates the hypothesis that weakly exergonic processes must be highly reversible - to a point where back flux of product to the substrate pool becomes possible (Holler et al., 2011a) - and that the extent of the reversibility, and as such the expression of isotope effects, is substrate concentration dependent. We have shown that naturally enriched active biomass performing AOM-SR is able to quickly adapt to changing substrate concentrations even at methane concentrations strongly exceeding concentrations in the natural habitat, implying the metabolic pathways that constitute AOM-SR allow for efficient use of available substrate by rapid adjustment of specific turnover rates. The presence of dissolved and extracellular zero-valent sulfur and the observation that sulfate consumption exceeds sulfide production in our experiment are consistent with the finding that zero-valent sulfur is likely an intermediate during AOM-SR and that sulfate consumption may be carried out by a different organism (i.e. ANME) than sulfide production (i.e. DSS; Milucka et al., 2012). This also raises the question if under our experimental conditions the growth and/or metabolic rate of ANME was slightly faster than the growth rate of their bacterial partners, leading to the accumulation of elemental sulfur in the reactor.

Under environmental conditions, AOM-SR is likely to induce a wide range of sulfur and oxygen isotope fractionations, depending on ambient methane concentrations. In sedimentary SMTZs where supply with methane is generally low, ε^{34} S resulting from AOM-SR are likely to exceed +50% and even approach the theoretical sulfur isotope equilibrium of +75\%, and to drive oxygen isotope enrichment of sulfate toward the sulfate-water oxygen isotope equilibrium value, which explains large sulfur isotope offsets between sulfate and sulfide from biologically active sulfate-methane transition zones. At methane gas seeps and locations where sulfate-rich water comes in contact with gas hydrates, $\varepsilon^{34}S$ and θ_0 are expected to be much smaller, and also more variable. Factors other than methane concentrations might have an impact on the sulfur and oxygen isotope effects by AOM-SR in the environment and to some extent on the results of our study. For example, increased steady state sulfide concentrations could affect isotope fractionation - either by affecting the overall energy yield of the reaction or by product inhibition and toxic effects on AOM-SR. As AOM-SR proceeds over a wide range of sulfide activities (e.g. Nauhaus et al., 2002) we assume that our enrichment cultures did not suffer from inhibitory effects due to sulfide stress. Because the detailed mechanistic coupling between methane oxidation and sulfate reduction during AOM-SR remains unknown, we believe that it would be valuable to conduct incubation experiments with changing concentrations of substrates (sulfate, methane) and products (sulfide, bicarbonate) in order to assess potential effects on the observed ε^{33} S, ε^{34} S and θ_0 , due to changes in the overall energy yield, and, in the case of sulfate, also due to potential suppression of isotope effects at very low sulfate concentrations (Habicht et al., 2002). This approach could open an avenue to the reconstruction of the concentration of substrates and products of AOM-SR from the rock record, where pyrite minerals and carbonate-associated sulfate from diagenetic carbonates preserve such isotope signatures.

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Appendix A. Supplementary material

Supplementary material related to this article can be found online at http://dx.doi.org/10.1016/j.epsl.2014.04.047.

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