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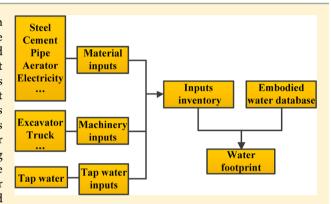


Water Footprint Assessment for Wastewater Treatment: Method, Indicator, and Application

Ling Shao[†] and G. Q. Chen*,[†],[‡]

Supporting Information

ABSTRACT: The water footprint in terms of the sum of both direct and indirect water cost of wastewater treatment is for the first time accounted in this work. On the basis of the hybrid method as a combination of process analysis and input-output analysis, a detailed water footprint accounting procedure is provided to cover the supply chain of a wastewater treatment plant. A set of indices intending to reveal the efficiency as well as renewability of wastewater treatment systems are devised as parallels of corresponding indicators in net energy analysis for energy supply systems. A case study is carried out for the Beijing Space City wastewater treatment plant as a landmark project. The high WROI (water return on investment) and low WIWP (water investment in water purified) indicate a high efficiency and



renewability of the case system, illustrating the fundamental function of wastewater treatment for water reuse. The increasing of the wastewater and sludge treatment rates are revealed in an urgent need to reduce the water footprint of China and to improve the performance of wastewater treatment.

1. INTRODUCTION

Water footprint, also regarded as virtual water or embodied water, is defined as the volume of total freshwater directly and indirectly required to produce a commodity or service. Given the rapidly growing water consumption and water resource scarcity, abundant studies on water footprint accounting have been carried out.²⁻⁸ Wastewater treatment plays a vital role in people's daily lives with regard to purifing the wastewater and discharging ready-for-reuse water to human society. Due to its significant ecological implication in water reuse, the resource uses and environmental emissions of wastewater treatment are quantified in different categories, such as greenhouse gas emission, energy consumption, and land use. 9-11 However, no study on assessing the water footprint as an important indicator of a specific wastewater treatment system has yet been reported. In a few studies, the input-output analysis, as a top-down method compared to the process analysis, has been relied on to calculate the average water footprint of the water industry within a nation or region.^{8,12} These studies are considered meaningful and helpful to guide the policy making on an industry level. Subjected to the aggregation level of statistical data, however, most of the input-output analyses can only give an average sector resolution and usually do not allow for detailed discussions of a specific product, especially industry-atypical products. In the meantime, in most countries, wastewater treatment is not listed as an individual industry in relevant statistics, which prevents us from assessing the water footprint of wastewater treatment based only on input-output analysis.

Under these circumstances, the accounting of the water footprint of a specific wastewater treatment system is performed for the first time in this work. A process-based method supported by the averaged macro-economic data provided by input-output analysis (termed as the hybrid method hereafter) is employed to estimate the water footprint of a specific wastewater treatment system. The hybrid method has many successful applications in assessing environmental impacts of various systems, including the wastewater treatment plant after its proposal in the 1970s. ^{11,13-15} It is employed not only to avoid the systematic truncation errors inherent in process analysis, but also to make the input-output analysis applicable for specific micro systems.

On the basis of the calculation of water footprint, a comprehensive evaluation of purification efficiency and renewability of wastewater treatment necessitates some indicators. A wastewater treatment plant is exactly like the energy supply system with regard to the fact that their common essence is to deliver available resources to human society. As the net energy analysis is a mature and wildly used method to compare the

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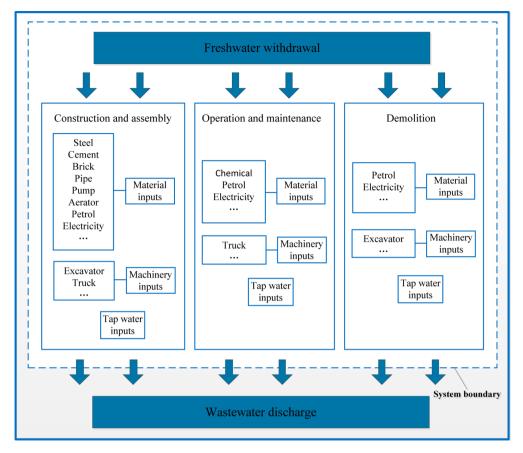


Figure 1. Diagram for water footprint assessment of a wastewater treatment system.

amount of total energy directly and indirectly required by an energy supply system to the energy it has delivered, ¹³ the proposed indices can be transplanted to assess the water purification efficiency of a wastewater treatment system. As a renewable system of water recourse, the renewability assessment of wastewater treatment can also draw lessons from the cases of renewable energy technologies, of which the renewability has been extensively discussed in the past decade. ^{16,17}

This work aims at providing an explicit accounting procedure to calculate the water footprint as an environmental impact of a wastewater treatment system. A set of indices are also devised to reveal the purification efficiency and renewability of a wastewater treatment system. Besides the practical implication in providing clear guidance for designing water saving strategies as well as prioritizing options of wastewater treatment technologies, this study can also contribute to estimation of water footprint of wastewater treatment industry in China. In order to demonstrate the method, a case study has been carried out by systematically calculating the water footprint of a typical conventional wastewater treatment system in Beijing on the basis of detailed inventories and latest embodied water intensity database for the Chinese economy.

2. METHODS AND MATERIALS

2.1. The Hybrid Method. The process analysis as a bottom-up method has been relied on to trace individual environmental impact of a wastewater treatment plant through its lifecycle by a plenty of studies.^{9,10} Although it can provide detailed process information, it is time-consuming and has suffered from truncation error. The input—output analysis is a

top-down method compared to the process analysis. It is relatively fast and deemed to be complete, as it can model all transaction activities within an economy. However, due to the high level of aggregation in industry or commodity classifications of the input-output table, the results of inputoutput analysis is often too rough for concrete assessments of some particular technologies or products in an economic sector. Therefore, the hybrid method intending to combine the strengths and reduce the weaknesses of both methods has been advocated. So far, there have been several alternative models of the hybrid method, e.g., tiered hybrid analysis, input-output based hybrid analysis and integrated hybrid analysis. 18,19 The simplest model of tiered hybrid analysis contributed by Bullard et al. 13 is adopted in this study. The approach is simple, quick, and does not require any additional data. 18 Through greater or lesser modification of the input-output table, other hybrid methods are complex and suffer from high data requirements,²⁰ which are far from realizable in water footprint accounting, especially for the assessment of wastewater treatment in China.

To evaluate the virtual water of an individual wastewater treatment system reflected by some process model as a microsystem in the macro-economy, the indirect fluxes originated outside the process boundary can be well traced by averaged sectoral intensities provided by proper input—output analysis of the economy. A wastewater treatment plant shares a tiny part of the whole wastewater treatment industry, which would avoid the possible double counting of the tiered hybrid method. The simplest model of the tiered hybrid analysis assumes that the product of interest is well approximated by its industry sector. Therefore, the water footprint of an input can be easily obtained by simply

Table 1. Indices of an Energy Supply System and a Wastewater Treatment System

element	index	definition	explanation
efficiency	NE (net energy)	$NE = E_{gained} - E_{required}$	E_{gained} : the energy gained
	EROI (energy return on investment)	$EROI = E_{gained}/E_{required}$	E_{required} : the energy required to get that energy
	NW (net water)	$NW = W_{purified} - W_{required}$	W_{purified} : the water purified
	WROI (water return on investment)	WROI = $W_{\text{purified}}/W_{\text{required}}$	W_{required} : the water required to purify that water
renewability	NEIED (nonrenewable energy investment in energy delivered)	$NEIED = NRE/E_{gained}$	NRE: the nonrenewable energy required to get that energy
			E_{gained} : the energy gained
	WIWP (water investment in water purified)	WIWP = $W_{\text{required}}/W_{\text{purified}}$	W_{required} : the water required to purify that water
			W_{purified} : the water purified

multiplying the monetary cost to the water footprint intensity of the sector.

2.2. Diagram for Water Footprint Assessment of a Wastewater Treatment System. The freshwater withdrawal during the whole lifetime of the wastewater treatment system is set as the accounting goal of this study. Besides the usual functional unit as 1 m³ of treated wastewater, the water footprints of 1 kg removed pollutant (BOD and COD) and of 10 000 Chinese Yuan (CNY) output value are also calculated. Water footprint assessment for a wastewater treatment system is diagrammed in Figure 1. As shown in the figure, the inputs of wastewater treatment are classified into three categories as material inputs, machinery inputs and tap water inputs. Material inputs are directly provided by other industries and consumed by wastewater treatment. The water footprint of material inputs are the virtual water content of the products. The machinery refers to the involved equipment that can be used in other place. For instance, the rooter and truck used to dig and deliver soil in the construction stage are machineries. But the pump and aerator used to draw and aerate wastewater during the operation stage are materials. By assuming that the virtual water embedded in the machinery is uniformly distributed along its lifecycle, the water footprint of machinery inputs can be estimated as the water footprint of the product times its work time and divided by its total lifetime.²¹ The lifetime here indeed refers to the total work time. As for the fuel consumption of machinery, they are classified as materials.

Tap water is indeed a kind of material input. It is listed as a distinctive category to avoid possible misunderstanding by simply regarding the volume of tap water as its water footprint in water footprint accounting. Tap water is the typical and major product of the water supply industry, which needs a lot of inputs from other industries in its production just like the secondary energy. Therefore, the water footprint of tap water should be the virtual water embodied in its lifetime, not its direct water volume. The data source of input inventory of the wastewater treatment system should be reliable and exhaustive. First-hand project data provided by the builder are preferred.

- **2.3. Procedure of Water Footprint Assessment for a Wastewater Treatment System.** By employing the hybrid method as an integration of process analysis and input—output analysis, the framework of water footprint assessment for wastewater treatment can be established as follows.
- (1) Itemize all of the materials, machinery, and tap water inputs required in the construction, operation, and demolition stages to form the inputs inventory of the wastewater treatment system. As the embodied water intensity, i.e., the water footprint or the virtual water content of each input product provided by input-output analysis is in the unit of m³/10 000 CNY, the monetary

- cost (C_j) of the *j*-th input should be listed in the inventory.
- (2) Choose an appropriate embodied water intensity database for all inputs. To avoid the truncation error of process-based analysis when tracing each process linked to a specific product in the network of the economy, the database obtained from input—output analysis is necessary here to support a comprehensive and accurate research. The economy input—output table as the fundamental data source of input—output analysis is updated every fixed time period, five years in China for instance. So it is essential to choose a database whose publishing year is in accordance with the year when the wastewater treatment system was built.
- (3) Identify the corresponding production sector for each item of the inventory with reference to statistical methods for economy input—output table provided by the national or regional statistics. The corresponding embodied water intensity (I_j) of each input can be promptly derived from the database. For example, pumps are produced by sector 67 (manufacture of pump, valve, and similar machinery) in the Chinese economy 2007, so the embodied water intensity of pumps is determined as the intensity of sector 67 listed in the database. This step may involve some extra efforts. For example, some inputs from the same industry and having the same embodied water intensity may be combined to simplify the account.
- (4) Multiply each item's cost by its embodied water intensity to give the water footprint (WF_j), then the total water footprint (WF) of a wastewater treatment system is readily obtained as follows:

WF
$$\equiv \sum_{j=1}^{n} WF_j = \sum_{j=1}^{n} (C_j \times I_j)$$

2.4. Water Footprint Indices for Assessing Purification Efficiency and Renewability of a Wastewater Treatment **System.** As mentioned in the Introduction, the water footprint as the virtual water cost of a wastewater treatment system is analogous to the energy cost of an energy supply system and the indices developed in energy analysis can be transplanted here. Among all of the indices provided by net energy analysis to reveal the conversion coefficients of energy supply systems, the NE (net energy) denoted as the energy gained and EROI (energy return on investment), denoted as the ratio of the energy extracted or delivered by an energy supply system to the energy consumed directly and indirectly in its supply chain, are regarded as the most basic and representative indices since the 1970s.^{22,23} The indices of NW (net water) as the net water purified and WROI (water return on investment) as the ration of the water purified by a water treatment system to the water

Table 2. Water Footprint Account of the BSCWWTP

			embodied water intensities $(m^3/10\ 000\ CNY)$			water footprint (m³)		
item	sector code	sector contents	agricultural water	industrial water	total	agricultural water	industrial water	total
pump	67	manufacture of pump, valve and similar machinery	1.81×10^{1}	5.91×10^{1}	7.72×10^{1}	2.47×10^{2}	8.08×10^{2}	1.06×10^{3}
steel griller	63	manufacture of metal products	2.13×10^{1}	8.56×10^{1}	1.07×10^{2}	1.92×10^{2}	7.71×10^{2}	9.64×10^{2}
equipment	72	manufacture of other special purpose machinery	2.25×10^{1}	7.83×10^{1}	1.01×10^{2}	2.77×10^3	9.65×10^{3}	1.24×10^4
automatic control system	78	manufacture of equipments for power transmission and distribution and control	2.21×10^{1}	6.98×10^{1}	9.19×10^{1}	7.49×10^2	2.37×10^3	3.11×10^3
pipe and accessory	49	manufacture of plastic	2.89×10^{1}	8.95×10^{1}	1.18×10^{2}	1.23×10^{3}	3.82×10^{3}	5.04×10^{3}
civil work	95	construction	2.93×10^{1}	6.50×10^{1}	9.44×10^{1}	6.74×10^{3}	1.49×10^{4}	2.17×10^{4}
tap water	94	production and distribution of water	1.46×10^{1}	1.95×10^{4}	1.96×10^{4}	1.26×10^{1}	1.68×10^{4}	1.69×10^{4}
electricity	92	production and supply of electric power and heat power	1.41×10^{1}	9.31×10^{1}	1.07×10^2	1.24×10^4	8.16×10^4	9.37×10^4
chemical	44	manufacture of special chemical products	5.28×10^{1}	1.14×10^{2}	1.67×10^{2}	2.90×10^{3}	6.26×10^{3}	9.16×10^{3}
total						2.72×10^{4}	1.37×10^{5}	1.64×10^{5}
per m ³ wastewater						5.17×10^{-4}	2.61×10^{-3}	3.12×10^{-3}
per kg BOD ₅						2.20×10^{-3}	1.11×10^{-2}	1.33×10^{-2}
per kg COD						1.72×10^{-3}	8.68×10^{-3}	1.04×10^{-2}
per 10 ⁴ CNY						6.47×10^{0}	3.26×10^{1}	3.90×10^{1}

required, i.e., water footprints of the system are therefore suggested here in this work (see Table 1).

Renewable energy sources, such as biofuel, solar power, and wind power, have been developed to reduce the fossil fuel cost and thus mitigate the greenhouse gas emission since the end of the last century. While energy efficiency indicated by EROI remains meaningful, it does not suffice to evaluate the renewability of renewable energies. Chen et al.²⁴ suggested an indicator of NEIED (nonrenewable energy investment in energy delivered) to address how much nonrenewable energy instead of inclusive energy is consumed to produce a presumed renewable energy. Wastewater treatment is devised as a water reuse technology for renewing polluted water resources. Renewability is among the most fundamental properties of a wastewater treatment system. The index as WIWP (water investment in water purified) is therefore utilized to assess the renewability of a wastewater treatment system (see Table 1). From the definition, it can be known that a low WIWP is preferred.

2.5. Case Description and Data Source. The rapid economic growth and accelerating urbanization of China have a profound impact on the availability of China's natural resource, especially the water resource. The water scarcity tension is extremely prominent in most cities in Northern China, due to a series of factors such as unbalanced spatial distribution, large population, and heavy water pollution. Therefore wastewater treatment becomes one of the most promising technologies. Projected in the early 1990s, Beijing Space City (BSC) is located in Haidian District, northwest of Beijing, which is far from the downtown of Beijing and thus is out of the municipal wastewater pipe network. With so many research institutes and accompanied life centers entering in, the wastewater becomes a critical problem for creating a graceful environment of BSC. Against this background, a wastewater treatment plant (WWTP) was designed and constructed here by the Environment Design and Research Center of the Chinese People's Liberation Army in 1998 to deal with the wastewater of BSC.²⁵

The case system lies in the north (1.4 km) of BSC and has a designed daily treatment capacity of 7200 m³. It employs a Cyclic Activated Sludge System (CASS) to treat the wastewater.²⁶ A brief introduction of the CASS and a detailed description of the case system can be found in the Support Information (SI), Section 1.

It is recommended that a database adapted to the place and year where and when the wastewater treatment plant was built should be the priority. Our case system is located in China and was built in 1998. However, in that year, the Chinese Government did not publish an economy input-output table. Among the very few input-output tables available, the table of Chinese economy in 2007 is the latest and most detailed economy input-output table published by National Bureau of Statistics of China. A previous work contributed by Chen and Chen¹² has already calculated the water footprint intensity database for the Chinese economy in 2007 based on it (SI, Section 2). It is adopted in this study as the supporting data. Only water use instead of water consumption data are available in China Statistical Yearbook, and the water usage has been classified as agricultural water, industrial water, domestic water and municipal ecological protection water.²⁷ Chen and Chen¹² have simplified the assortment by categorizing the water resources as agricultural water (with reference to agricultural water only) and industrial water (with reference to industrial water, domestic water, and municipal ecological protection water) in their work. The Producer Price Indices of Industrial Products (PPI) of China is referred to convert the construction cost of construction year-based into 2007-based (SI, Section $3).^{28}$

3. RESULTS AND DISCUSSION

3.1. The Water Footprint of the Case System. Only the water footprints of construction and operation stages are accounted for, with that of the demolition stage ignored. The designed lifetime of BSCWWTP is set at 20 years, in accordance with the usual practice of China. The inputs

inventory accompanied with their corresponding input—output sectors is listed in Table S4 of the SI. As mentioned in Section 2, the simplest model of tiered hybrid analysis which assumes that the product of interest is well approximated by its corresponding industry sector is employed in this study. It can be seen that each input is the typical product of its production industry. Therefore, the method is reasonable and reliable and will not bring much inaccuracy to our results. In earlier life cycle environmental impacts studies of a wastewater treatment system, their inputs were traced to a raw material base. ¹⁰ It may cause errors to the final results since the assembly processes of purchased products also involve extra impacts. For example, the water footprint intensities of steel and pump are quite different.

The calculations of the water footprint of BSCWWTP are shown in Table 2, with some inputs from the same production sector having the same intensity merged for convenience.

The water footprint of BSCWWTP is estimated as 1.64×10^5 m³ freshwater, with intensities of 3.12×10^{-3} m³ freshwater/m³ wastewater, 1.33×10^{-2} m³ freshwater/kg BOD₅, 1.04×10^{-2} m³ freshwater/kg COD, and 3.90×10^1 m³/10 000 CNY. The components of the water footprint of BSCWWTP are shown in Figure 2. The electricity accounting

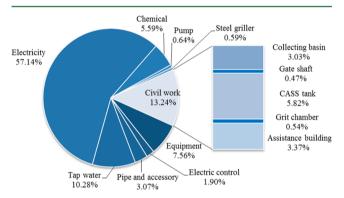


Figure 2. The components of the water footprint.

for more than one-half is the largest source of water footprint. Energy saving practice has been promoted all over the world for decades. According to the results, energy saving policy should be further enhanced for wastewater treatment as a high energy-consuming industry, due to its equal importance for water footprint reduction.

As the second largest water footprint source, tap water accounts for about 10% of the total water footprint. The direct water consumption, i.e., the volume of tap water, is less than 3% of the total water footprint of the case system, which strongly demonstrates the importance and essentiality of the water footprint account. The real water content (4365 $\rm m^3$) of tap water is only 25% of its virtual water content (1.69 \times 10⁴ $\rm m^3$). The civil work consisting of the construction of basins and tanks shares 13.24% of the total water footprint, followed by the equipment with a total contribution of 7.56%.

The contrast between the two categories of water footprint as agricultural and industrial water footprint during the construction and the operation stages for the case system is shown in Figure 3. Due to the data availability, only the electricity and chemicals are concerned for the operation stage while other inputs such as the maintenance of the plant are ignored. Even so, the water footprint in the operation stage is larger than that in the construction stage. The result is in good accordance with numerous former studies showing that the environmental

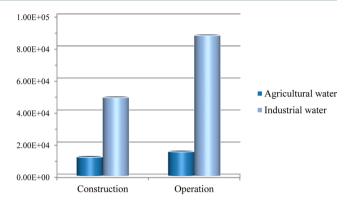


Figure 3. The water footprint of construction and operation stages of BSCWWTP.

impacts of the operation stage are much greater than that of the construction stage for wastewater treatment systems. 11,29,30

The difference between the operation stage and construction stage is mainly from the industrial water. It can be seen that the wastewater treatment depends more on industrial products than on agricultural products. Almost nothing in the inputs inventory of the case system is directly from the agricultural industry, but the percentage of agricultural water as 16.57% is still considerable. It clearly shows the connection between primary and secondary industries. In contrast to abundant studies on estimating the water footprint of agricultural products, the water footprint of industrial products has seldom been given attention. It should be enhanced in the future since the close interactions of different industries are associated with water footprint exchanges.

3.2. The Efficiency and Renewability of the Case **System.** The NW and WROI of BSCWWTP are calculated as 5.24×10^7 m³ in its whole lifetime and 3.20×10^2 m³ purified water/m³ invested water. The WROI in this study is much larger than the EROI of energy supply systems (usual less than 100), indicating a high efficiency of wastewater treatment. However, the EROI for the energy system is becoming smaller day by day due to the increasing difficulty in fossil energy extraction. For example, the running average EROI for the discovery and production of U.S. domestic oil has dropped from greater than 100 in the 1930s to about 30 in the 1970s and to 11-18 nowadays.³¹ The freshwater withdrawal will become more and more difficult in the future. Also, the base pollutant concentration of freshwater would increase as a result of the continuous pollution caused by human activity. Therefore, the WROI of wastewater treatment would be expected to decrease in the future too.

Water is deemed as a renewable resource and has a wide range of renewable spans. It can either be renewed through natural processes or purified by manmade wastewater treatment. But a wastewater treatment system has its own water footprint. Only when its water footprint is less than its purified wastewater can a wastewater treatment system be regarded as a renewable water system. Moreover, the renewable degrees are different among different wastewater treatment technologies, which can be assessed by the indicator of WIWP. For renewable energy systems, NEIED < 1 means that more energy is produced than nonrenewable energy invested, while NEIED > 1 means that more nonrenewable energy is consumed than energy produced. The same meaning has been transplanted to the index of WIWP. The WIWP of the case system is calculated as 3.12×10^{-03} m³ freshwater/m³

purified water, which is much less than 1 and indicates a high renewability of the case system. This finding coincides with the fact that wastewater treatment is widely devised as an effective water reuse technology.

3.3. Comparison between This Study and Some Other Water Footprint Accounting Studies. Hoekstra and Chapagain (as well as their fellows) have made significant contributions by systematically classifying the water footprint as blue water footprint, green water footprint, and gray water footprint. This perspective of classification emphasizing the scarcity of water has contributed significantly by providing clear guidance for prioritizing options of water footprint reduction. In earlier water footprint studies, only used water resources were considered as blue water and green water. They refer to surface or groundwater and freshwater brought by rain, respectively. Later, this concept was extended by quantifying the impacts of pollution from the quality aspect. The gray water was then proposed and defined as the volume of freshwater required to assimilate the already generated pollutants to meet the existing ambient water quality standards.³² Such definition was derived from the concept of ecological footprint, which transforms the environmental emission to the land area needed for uptake of the emission. This is not to mean that people are going to dilute the pollutants, but to provide an indicator to measure the pollution.1

Due to the data availability, only the blue water footprint is accounted for in the case system. Meanwhile, the implication of wastewater treatment in eliminating the gray water is discussed. There are two calculation methods for the water footprint. In most bottom-up process analysis based water footprint studies, especially the water footprint accounting of agricultural products, only consumptive water use, e.g., evaporation is concerned.^{32,33} It can either be estimated by field measurement or some empirical estimation methods. In contrast to this, in most top-down based studies, especially the water footprint accounting of industrial products, the total water use, e.g., the irrigation or withdrawn water, is concerned. 4,8 It is due to the fact that available statistics of countries and regions only provide total water withdrawals. The databases that contain typical data on consumptive water use for various types of manufacturing processes hardly exist. Some studies have made strong assumptions to relate the water withdrawal to consumptive water. For example, in Mekonnen and Hoekstra's work to calculate the water footprint of agricultural and industrial commodities of each country, it is assumed that 5% of the water withdrawn for industrial purposes is actual consumption (blue water footprint) and that the remaining fraction is return flow; for the domestic water supply sector a consumptive portion of 10% is assumed.³³ The presented study only concerns the water withdrawal, and the result (3.90×10^{1}) m³/10 000 CNY) is more than three times larger than that of aforementioned study concerning the consumptive water (9.82 m³ /1000 U.S. \$, average blue water footprint of China during 1996-2005). It reveals the gap between consumptive water and withdrawal water studies. Attention to this issue should be paid in future water footprint accounting studies.

Hoekstra et al.¹ stated that wastewater treatment can bring gray water footprint down. However, the process of wastewater treatment has its own water footprint. The determination of the water footprint of wastewater treatment should be the premise to quantify its real contribution in eliminating the gray water footprint. By assuming that the natural concentration is zero and the maximum acceptable concentration in the receiving

water body is the same as the concentration of effluent in this study, the gray water footprint eliminated by the case wastewater treatment system is calculated as $8.76\times10^8~\text{m}^3~(\text{BOD}_5)$ and $3.68\times10^8~\text{m}^3~(\text{COD})$ with reference to the gray water footprint theory (the equation can be found in SI, Section 4). Our case system does have a significant contribution in eliminating the gray water footprint (1.64 \times $10^5~\text{m}^3$ vs 8.76×10^8 and $3.68\times10^8~\text{m}^3)$, which verifies the statement. Although the ignored green and gray water footprint of the case system may slightly diminish the contribution, the observed advantage is so outstanding that the superiority will not change.

3.4. Uncertainty Analysis. Wastewater treatment can be considered as the typical product or service provided by wastewater treatment industry, therefore the monetary intensity result in this study can be compared with those of other products of different industries. Chen and Chen¹² have obtained the water footprint intensities for 135 industries, i.e., 135 products or services of Chinese economy in 2007. It appeared that the water footprint intensity of the case system is very small compared with them (ranking the least fourth), which is only larger than that of scrap and waste, real estate and banking, security, and other financial activities industries. Meanwhile, it is much lower than that of the water related industry (1.96 \times 10⁴ m³/10 000 CNY). It confirms the necessity of this study to calculate the water footprint of a specific wastewater treatment system and the high efficiency of the case system on one hand. On the other hand, the reason may be attributed to the fact that the case system in this study is a municipal wastewater treatment plant, whose water footprint is deemed to be lower than that of the treatment plants dealing with high-concentration or specific-industrial wastewater. Moreover, the water related industry comprises both the water production and wastewater treatment industries, its high intensity may come from the high water footprint intensity of the water production industry.

The absence of subsequent sludge treatments, the wastewater pipe outside the plant, and some other supporting systems have also contributed to that. The supporting systems, such as the gardening system, are not discussed here as it highly depends on individual devices, of which no reliable data are available. Here, we will estimate the water footprints of pre-wastewater pipes and subsequent sludge treatment outside the plant based on the results of some academic references and the average statistical data of Beijing.

The detailed calculation information is in SI, section 5. The water footprint of pre-wastewater pipes and subsequent sludge treatment are estimated as 6.30×10^3 m³ and 2.23×10^4 m³ freshwater, which are 3.84% and 13.58% of the previous water footprint of the case system. Sludge treatment plays a very important role in wastewater treatment. Not only because of its considerable share of environmental impact, but also the improper treatment of sludge may cause serious secondary pollution, such as heavy metal pollution. If properly cared for, however, sludge can turn into resource, and the reuse benefit may even offset part of the water footprint of wastewater treatment. In China, the sludge treatment is often ignored in the design and operation of wastewater treatment, which has reduced the environmental benefit of wastewater treatment to a discount. It is reported that only about 1/4 of disposal sludge in China is completely harmless to the environment in 2010, while the rest can be more or less considered as pollution sources.³⁴ Therefore, the sludge treatment should be given special

attention, and needs to be practically and effectively included in the environmental assessment of wastewater treatment in the future.

3.5. Water Footprint of Wastewater Treatment of China. CASS is one of the most popular wastewater treatment technologies in China.³⁵ The treatment capacity of the case system is medium among those of wastewater treatment systems in China.³⁶ Therefore, the presented results can be used to estimate the water footprint of wastewater treatment in China. as a first effort.

In China, it is reported that 5.20×10^6 and 1.27×10^7 t COD in municipal and industrial wastewater has been removed by wastewater treatment in 2007.^{34,37} If all wastewater were treated by the same technology as the CASS in this study, then the water footprint of wastewater treatment would be estimated as 2.79×10^8 m³ freshwater, accounting for 0.05% of China's freshwater withdrawal (5.82 \times 10¹¹ m³ freshwater) in 2007 (with pre-wastewater pipes and subsequent sludge treatment involved). For Beijing, they are 7.29×10^6 m³ freshwater and 0.21%, respectively. Although there might be possible underestimates, for instance, part of the inputs of the case system were ignored and the treatment of high-concentration or specific-industrial wastewater may have a much higher water footprint, the water footprint of wastewater treatment is reasonably shown to be insignificant. The proportion for Beijing is about four times as high as the national average. This is partly caused by the much higher municipal wastewater treatment rate of Beijing (about 80%) than that of China (49%) in 2007.³⁸ The Chinese government has been struggling to improve the wastewater treatment rate in recent years. The industrial waste treatment rate has maintained up to 90% these years. The municipal wastewater treatment rates in cities and towns of China have been increased from 56% and 14% in 2006 to 78% and 60% in 2010, and the rates of 85% and 70% have been set as goals to be achieved in 2015.34

It is also estimated that the 4.18×10^{11} m³ gray water footprint has been eliminated by municipal and industrial wastewater treatment of China in 2007. It is a little larger than the final gray water footprint of the national consumption of China $(3.23 \times 10^{11} \text{ m}^3/\text{year} \text{ in the period } 1996-2005).^{33}$ Even so, China is the country with the largest gray water footprint within its borders, which is 26% of the global gray water footprint.³⁹ While the main reason lies in the fact that China has the largest population, part of the reason can be attributed to the relatively low treatment rate of China. Moreover, agricultural wastewater is an important component of wastewater, but it has not been included in the environmental supervision of China yet. It is reported that over than 90% of agricultural wastewater is discharged without any treatment. 40 If proper care is taken, then the increasing of wastewater treatment rates and application of wastewater treatment in agriculture would further greatly eliminate the gray water footprint of China.

Furthermore, the reuse of reclaimed water has also been promoted by the government. The reclaimed water means the end product of wastewater treatment that meets water quality requirements for biodegradable materials, suspended matter, and pathogens. It can be routed directly to a recycled water system for agriculture and sundry industry uses such as landscaping irrigation or to recharge groundwater aquifers. The reuse rate of reclaimed municipal wastewater is only 10% in 2010 for China, which is much lower than that of the developed countries all over the world.³⁴ Israel is the world's most efficient

recycled water user which purifies and reuses almost 70% of its wastewater each year for agriculture. In Israel, only 11.25% of the total footprint is gray water footprint. As for China, it is as high as 23.63%. By improving the reuse rate of reclaimed water, the gray water footprint can be directly reduced by getting rid of the wastewater discharge. Moreover, the reuse of reclaimed wastewater can also reduce the freshwater withdrawal, which would further reduce the blue water footprint.

ASSOCIATED CONTENT

S Supporting Information

Additional materials as noted in the text. This information is available free of charge via the Internet at http://pubs.acs.org.

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Notes

The authors declare no competing financial interest.

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REFERENCES

- (1) Hoekstra, A. Y.; Chapagain, A. K.; Aldaya, M. M.; Mekonnen, M. M. The Water Footprint Assessment Manual: Setting the Global Standard; Earthscan: Washington, DC, 2011.
- (2) Berger, M.; Warsen, J.; Krinke, S.; Bach, V.; Finkbeiner, M. Water footprint of european cars: Potential impacts of water consumption along automobile life cycles. *Environ. Sci. Technol.* **2012**, *46* (7), 4091–4099.
- (3) Chapagain, A. K.; Hoekstra, A. Y. The water footprint of coffee and tea consumption in the Netherlands. *Ecol. Econ.* **2007**, *64* (1), 109–118.
- (4) Chen, Z. M.; Chen, G. Q. Virtual water accounting for the globalized world economy: National water footprint and international virtual water trade. *Ecol. Indic.* **2012**, *28*, 142–149.
- (5) Chen, Z. M.; Chen, G. Q.; Xia, X. H.; Xu, S. Y. Global network of embodied water flow by systems input-output simulation. *Front. Earth Sci.* **2012**, *6* (3), 331–344.
- (6) Jeswani, H. K.; Azapagic, A. Water footprint: Methodologies and a case study for assessing the impacts of water use. *J. Clean Prod.* **2011**, 19 (12), 1288–1299.
- (7) Stoessel, F.; Juraske, R.; Pfister, S.; Hellweg, S. Life cycle inventory and carbon and water foodprint of fruits and vegetables: Application to a Swiss retailer. *Environ. Sci. Technol.* **2012**, *46* (6), 3253–3262.
- (8) Zhao, X.; Yang, H.; Yang, Z.; Chen, B.; Qin, Y. Applying the input-output method to account for water footprint and virtual water trade in the Haihe River Basin in China. *Environ. Sci. Technol.* **2010**, 44 (23), 9150–9156.
- (9) Dixon, A.; Simon, M.; Burkitt, T. Assessing the environmental impact of two options for small-scale wastewater treatment: Comparing a reedbed and an aerated biological filter using a life cycle approach. *Ecol. Eng.* **2003**, *20* (4), 297–308.
- (10) Emmerson, R. H. C.; Morse, G. K.; Lester, J. N.; Edge, D. R. The life-cycle analysis of small-scale sewage-treatment processes. *Water Environ. J.* **1995**, 9 (3), 317–325.

- (11) Shao, L.; Wu, Z.; Zeng, L.; Chen, Z. M.; Zhou, Y.; Chen, G. Q. Embodied energy assessment for ecological wastewater treatment by a constructed wetland. *Ecol. Model.* **2012**, 252 (10), 63–71.
- (12) Chen, G. Q.; Chen, Z. M. Carbon emissions and resources use by Chinese economy 2007: A 135-sector inventory and input-output embodiment. *Commun. Nonlinear Sci. Numer. Simul.* **2010**, *15* (11), 3647–3732.
- (13) Bullard, C. W.; Penner, P. S.; Pilati, D. A. Net energy analysis: Handbook for combining process and input-output analysis. *Resour. Energy* **1978**, *1* (3), 267–313.
- (14) Chen, G. Q.; Shao, L.; Chen, Z. M.; Li, Z.; Zhang, B.; Chen, H.; Wu, Z. Low-carbon assessment for ecological wastewater treatment by a constructed wetland in Beijing. *Ecol. Eng.* **2011**, 37 (4), 622–628.
- (15) Lenzen, M. A guide for compiling inventories in hybrid life-cycle assessments: some Australian results. *J. Clean Prod.* **2002**, *10* (6), 545–572.
- (16) Malça, J.; Freire, F. Renewability and life-cycle energy efficiency of bioethanol and bio-ethyl tertiary butyl ether (bioETBE): Assessing the implications of allocation. *Energy* **2006**, *31* (15), 3362–3380.
- (17) Baral, A.; Bakshi, B. R.; Smith, R. L. Assessing resource intensity and renewability of cellulosic ethanol technologies using Eco-LCA. *Environ. Sci. Technol.* **2012**, *46* (4), 2436–2444.
- (18) Joshi, S. Product environmental life-cycle assessment using input-output techniques. *J. Ind. Ecol.* **1999**, 3 (2–3), 95–120.
- (19) Suh, S.; Huppes, G. Methods for life cycle inventory of a product. J. Clean Prod. 2005, 13 (7), 687–697.
- (20) Stromman, A. H.; Peters, G. P.; Hertwich, E. G. Approaches to correct for double counting in tiered hybrid life cycle inventories. *J. Clean Prod.* **2009**, *17* (2), 248–254.
- (21) Chen, G. Q.; Chen, H.; Chen, Z. M.; Zhang, B.; Shao, L.; Guo, S.; Zhou, S. Y.; Jiang, M. M. Low-carbon building assessment and multi-scale input-output analysis. *Commun. Nonlinear Sci. Numer. Simul.* **2011**, *16* (1), 583–595.
- (22) Cleveland, C. J.; Costanza, R.; Hall, C. A. S.; Kaufmann, R. Energy and the US economy: A biophysical perspective. *Science* **1984**, 225 (4665), 890–897.
- (23) Murphy, D. J.; Hall, C. A. S. Year in review—EROI or energy return on (energy) invested. *Ann. Ny. Acad. Sci.* **2010**, *1185* (1), 102–118.
- (24) Chen, G. Q.; Yang, Q.; Zhao, Y. H. Renewability of wind power in China: A case study of nonrenewable energy cost and greenhouse gas emission by a plant in Guangxi. *Renew. Sust. Energ. Rev.* **2011**, *15* (5), 2322–2329.
- (25) Wang, S. Z.; Zhang, T.; Hou, R. Q.; Zhang, Z. R. CASS wastewater treatment process for Beijing Space Flight Center. *Water Wastewater* 1999, 25 (8), 10–12 (In Chinese)..
- (26) Zhang, T. Intermittent Activated Sludge Wastewater Treatment Technology and Project Cases; Chemical Industry Press: Beijing, 2002, (In Chinese).
- (27) CSY. China Statistical Yearbook (2008). Beijing: China Statistical Publishing House, 2008. (In Chinese).
- (28) CSY. China Statistical Yearbook (1997–2007); China Statistical Publishing House: Beijing, 1997–2007, (In Chinese).
- (29) Lundin, M.; Bengtsson, M.; Molander, S. Life cycle assessment of wastewater systems: Influence of system boundaries and scale on calculated environmental loads. *Environ. Sci. Technol.* **2000**, 34 (1), 180–186.
- (30) Zhang, Q. H.; Wang, X. C.; Xiong, J. Q.; Chen, R.; Cao, B. Application of life cycle assessment for an evaluation of wastewater treatment and reuse project—Case study of Xi'an, China. *Bioresour. Technol.* **2010**, *101* (5), 1421–1425.
- (31) Hall, C. A. S.; Powers, R.; Schoenberg, W. Peak Oil EROI, Investments and the Economy in an Uncertain Future Biofuels, Solar and Wind As Renewable Energy Systems; Pimentel, D., Ed.; Springer: Netherlands, 2008; pp 109–132.
- (32) Chapagain, A. K.; Hoekstra, A. Y. The blue, green and grey water footprint of rice from production and consumption perspectives. *Ecol. Econ.* **2011**, 70 (4), 749–758.

- (33) Mekonnen, M. M.; Hoekstra, A. Y. National Water Footprint Accounts: The Green, Blue and Grey Water Footprint of Production and Consumption; Value of water research report series no. 50; 2011.
- (34) Report on Wastewater Discharge and Treatment of China in 2006–2010. Ministry of Housing and Urban-Rural Development of China: Beijing, 2011, (In Chinese).
- (35) Yu, Z. B.; Wang, D. Q.; Zhang, X. H. The reviewing and expectation of the development on urban sewage dispose technique. *J. Guangxi Normal Univ. (Nat. Sci.)* **2004**, 22 (2), 81–87 (In Chinese).
- (36) Ministry of Environmental Protection of China. *The Inventory of Municipal Wastewater Treatment Plants in China*, 2012. http://www.zhb.gov.cn/gkml/hbb/bgg/201204/t20120424_226655.htm. (In Chinese).
- (37) Ministry of Environmental Protection of China. *China Environment Yearbook* 2008; China Environment Yearbook Press: Beijing, 2008, (In Chinese).
- (38) Ministry of Environmental Protection of China. Annual Report of Chinese Environment Statistic in 2007, Beijing, 2008. (In Chinese).
- (39) Hoekstra, A. Y.; Mekonnen, M. M. The water footprint of humanity. Proc. Natl. Acad. Sci. U.S.A. 2012, 109 (9), 3232–3237.
- (40) Shen, F. J. Application status and development trend of agricultural wastewater treatment technology in China. *Agric. Eng. Technol.* (New Energy Ind.) 2011, 1, 16–19 (In Chinese).