

Avian Digestive Tract Simulation To Study the Effect of Grit Geochemistry and Food on Pb Shot Bioaccessibility

MONICA MARTINEZ-HARO,^{*,†}
MARK A. TAGGART,[†] ANDY J. GREEN,[‡]
AND RAFAEL MATEO^{*,†}

Instituto de Investigación en Recursos Cinegéticos, IREC (CSIC-UCLM-JCCM), 13071 Ciudad Real, Spain, and Department of Wetland Ecology, Estación Biológica de Doñana—CSIC, Américo Vespucio s/n, 41092 Sevilla, Spain

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Lead shot dissolution was investigated in a dynamic *in vitro* simulated avian gizzard-intestine system. The method allows simulated digestive fluid to pass (at intervals) from a gizzardlike environment to an intestine-based one, and then considers the dissolution of Pb shot (0–3 pellets) in the presence of differing grit geochemistries (siliceous and calcareous) and variable amounts of food (0–4 g of partially milled wheat seed). Dissolved Pb levels in simulated gizzards were consistently higher in the presence of siliceous, than with calcareous, grit. This was also seen in simulated intestines, except when less food was used (0–1 g), when Pb levels in solution were higher in calcareous systems. The Pb concentrations in gizzard and intestine solutions increased directly with the number of Pb shot used. In all treatments Pb levels in intestine liquids were lower than in gizzard liquids. Calcareous grit simulations maintained 2.5–34 times more Ca in solution than those that used siliceous grit. Dietary supplementation with calcareous grit may reduce Pb bioaccessibility of ingested Pb shot in birds by reducing gizzard acidity, by enhancing Pb precipitation (as Pb-carbonate), and by promoting higher dissolved Ca levels in the intestine, which may then compete with Pb for intestinal absorption.

Introduction

Birds commonly ingest and maintain grit in their muscular gizzards to help grind up food (1). However, ingestion can pose an ecotoxicological risk to birds if, for example, Pb shot from hunting is mistakenly consumed (2, 3), or if grit is naturally or anthropogenically enriched with heavy metals (4–6). Shot ingestion and associated Pb poisoning has been shown to be a significant cause of mortality for many avian species, in numerous scenarios globally (7, 8). Shot ingestion is currently estimated to cause mortality of almost one million waterfowl per wintering season in Europe alone, which represents 8.7% of the wintering populations of >11 million birds from 17 species (8). Waterfowl population trends in Europe are also inversely correlated with Pb shot ingestion

prevalences for different species (8), suggesting that Pb poisoning is having effects on population levels at a global scale. In North America, before waterfowl hunting with Pb shot was prohibited, annual mortality from lead poisoning was estimated at ≈ 2 –3% of the total waterfowl population (9, 10). Lead poisoning from ingestion of spent lead ammunition has also been implicated in the decline of several globally endangered species, such as white-headed duck (*Oxyura leucocephala*) (11) and Californian condor (*Gymnogyps californianus*) (12).

While conservationists continue to call for a worldwide ban on the use of Pb shot to protect birds, finding effective methods of mitigating the risks posed by shot already present in the environment is also a priority. Several studies (7, 13, 14) have suggested that where Pb shot pellets are abundant in certain habitats and grit availability is limited, dietary supplementation with clean grit may promote reduced shot ingestion in susceptible species. The amount of grit normally ingested by a bird is related to species (15) whereas grit type (or geochemistry) is limited by local environmental availability (1). In wild birds, two geochemically dissimilar grit types are commonly found in gizzards, i.e., quartz/quartzite based (siliceous grit) and limestone/calcite based (calcareous grit) (1). While facilitating the physical break down of food, such grit also acts as an important source of major and trace elements (1). Likewise, the elements Ca or P and carbonaceous minerals affect both the solubility (16) and absorption of Pb within the digestive tract (17). One competitive metabolic interrelationship between Pb and Ca (for example) is already well-known (18, 19) whereby, with low dietary Ca, the parathyroid gland increases synthesis of intestinal Ca-binding proteins (to increase Ca absorption), but these also have a high affinity for Pb. Hence, birds exposed to Pb (via drinking water or Pb shot) and a low Ca diet develop more severe clinical signs of Pb poisoning, and deposit elevated amounts of Pb in soft tissues (20, 21). As well as competing with Ca, under certain Eh and pH conditions, Pb in solution will also readily form complexes or precipitates with (for example) carbonates and organic compounds (16).

Several *in vitro* methods have been described to investigate the dissolution of Pb shot in simulated avian digestive systems (22), to look at either the effect of grit on the breakdown of plant material (23) or the bioaccessibility of metals from contaminated soil (24, 25), and these can provide practical viable alternatives to *in vivo* methods. However, while current techniques provide valid information, many perhaps lack the degree of complexity that really exists in an avian digestive tract. Here, we develop and use a dynamic *in vitro* simulation and consider interactions among several factors that have not been considered together previously. The proposed novel system allows for variations in grit composition (and/or volume), the amount of food ingested (and/or type), and the number of Pb shot (or potentially, other toxicants). We apply the technique to investigate the dissolution of varying numbers of Pb shot in the presence of variable amounts of food and grit types of differing geochemistry. Further, while the system monitors Pb dissolution over time within a simulated gizzard environment, it also permits the transfer of portions of gizzard solution (at intervals) into differing simulated intestine conditions (where Pb may precipitate out). Since absorption of Pb in solution predominantly occurs in the intestine, and not the gizzard (26–28), we suggest that considering the transfer of fluid from one set of conditions (the gizzard) to another (the intestine) is an important element in an effective simulation of this type.

* Corresponding author phone: +34 926 29 54 50; fax: +34 926 24 54 51; e-mail: rafael.mateo@uclm.es.

[†] Instituto de Investigación en Recursos Cinegéticos.

[‡] Estación Biológica de Doñana—CSIC.

Experimental Section

Justification for Experimental Design. The avian digestive tract is complex, dynamic, and markedly different from that of mammals. Many birds have powerful muscular gizzards where food is predigested, dissolved, degraded, and physically compressed/ground down [often with the aid of grit (23)]. The gizzard contains enzyme-rich digestive fluid with a basal pH of ≈ 2.6 , containing gastroliths, pepsin, HCl, and sodium chloride (27). Secretion of this fluid is promoted by the presence of food (27), and it occurs continuously during digestion. In chickens, the rate of secretion has been calculated to be $\approx 15.4 \text{ mL h}^{-1}$ (27). As food is digested, digestive liquid containing the finest particulates is allowed to pass into the intestine, where a different intestinal fluid is secreted. These secretions are of a higher pH (5.2–7.2) and also contain bicarbonate ions, bile salts, and pancreatin (25, 27). Undigested food material is ultimately removed as an intestinal pellet (in feces). The duration of the whole digestion process (depending on the ingesta type/size) is between 6–7 h in mallards (*Anas platyrhynchos*; 29).

To simulate avian gizzard digestive juice (Gizz^{DJ}) we prepared a solution of 1 N NaCl (Prolabo) containing 10 g L^{-1} of pepsin (Merck), adjusted to pH 2.0 with concentrated HCl (Panreac; after (22) and (24)). To simulate intestinal digestive juice (Int^{DJ}) we prepared a concentrated solution containing bile extract (3.5%) and porcine pancreatin (0.35%; Sigma) which was diluted upon experimental addition to effective concentrations of 0.35% and 0.035% (after (25)), respectively (see below). We used two grit types, siliceous (Grit^{sil}) with 99% SiO₂ and 0.6% CaO and calcareous (Grit^{cal}) with 94% CaO and 4% SiO₂ [XRF analytical methods and information is given in Supporting Information (SI)]. Grit^{sil} was purchased from a pet shop (aquarium fine gravel), and Grit^{cal} from a floor tile factory (Sanluqueña de Pavimentos, S.A.). Before undertaking experiments, grit was thoroughly washed with Milli-Q water. The grit size used was 1–3 mm, as this is most commonly found in gizzards of waterfowl species that are more frequently affected by Pb poisoning (15, 30). For all experiments we used 2 g of grit, a typical amount previously reported in hunted mallards (30). For food, we used briefly milled wheat seed with a particle size ranging from fine flour to whole seed. We created one batch of this material which was subdivided into portions and then used for all experiments. Pb shot pellets used were size no. 6, with an average weight of 109 mg ($\pm 8 \text{ mg}$). All digestive solutions used in experiments were heated to 42 °C before use [average mallard body temperature (31)], and all chemicals used were analytical grade or better. High purity nitric acid (69% Analytical grade, Panreac) and hydrogen peroxide (30% v/v Suprapur, Merk) were used for the total digestions mentioned.

Simulated Gizzard-Intestine. Gizzard simulations were carried out in triplicate in 50 mL polypropylene centrifuge tubes. Initially, 12 mL of Gizz^{DJ} was added to the tube, then 2 g of grit (Grit^{sil} or Grit^{cal}). Lead dissolution was investigated using 1 Pb shot and a variable amount of food material (0, 1, 2, or 4 g). We undertook a control experiment free of Pb with 2 g of grit and 4 g of food, and another experiment using 4 g of food but 3 Pb shot. A fourth replicate of all treatments was used solely to monitor pH change during each experiment; this replicate was undertaken first.

The 50 mL tubes containing each simulation [12 mL of Gizz^{DJ}, 2 g grit (2 types), 0–4 g food, and 0–3 Pb shot pellets] were incubated for a total of almost 3 h at 42 °C, and constantly shaken on their sides at high speed on a flat bed revolving shaker (at 350 rpm). Shaking was used to partially simulate the physical action of the muscular gizzard, i.e., permitting aggressive contact between grit, food, and Pb shot. To simulate the flow of gizzard solution (Gizz^{soln}) containing suspended material into the intestine, and the associated

excretion of fresh Gizz^{DJ} that would normally occur in an avian gizzard, after 6 min of shaking, tubes were opened, 3 mL of existing Gizz^{soln} was removed, and 3 mL of fresh Gizz^{DJ} was added. Tubes were then closed and placed back on the heated shaker for another 6 min. In each experiment Gizz^{DJ} was added 10 times and the total run time (including the steps required to add or remove liquid, purge headspace, etc.) was approximately 180 min. Before starting each simulation, and after each addition of fresh Gizz^{DJ}, tube headspace was purged with N₂ to maintain a low oxygen environment within the tubes.

Each 3 mL of Gizz^{soln} removed from each treatment were immediately divided into two 1.5 mL aliquots. The first was immediately centrifuged at 9500 g for 10 min, and 1 mL of supernatant was then digested in 0.5 mL of concentrated HNO₃ and 0.5 mL of 30% H₂O₂ for 1 h, at 90 °C, in 15 mL polypropylene centrifuge tubes in a water bath. This was diluted to 10 mL with Milli-Q water, and stored at 4 °C until analysis. The second 1.5 mL aliquot was immediately taken forward for further incubation under simulated intestine conditions.

The second 1.5 mL aliquot (in a 1.5 mL microtube) of Gizz^{soln} was adjusted to pH 6.2 using a saturated solution of NaHCO₃ [Merck; after (25)]. The amount required to reach this pH (20 μL for Grit^{sil}, 0–10 μL for Grit^{cal}) was predetermined from the control replicate used to monitor pH change during each experiment. After vortex mixing, 150 μL was discarded, and then 150 μL of Int^{DJ} (10 times concentrate) was added. The tube headspace was then purged under a stream of N₂, and solutions were incubated for 3 h at 42 °C in an end over end shaker at slow speed (to simulate slow intestinal passage). Finally, intestinal solutions were centrifuged at 9500 g for 10 min, and 1 mL of supernatant was acid digested in a 15 mL polypropylene centrifuge tube (as above). Remaining solid material (effectively the feces) was oven-dried at 40 °C for 48 h, and then digested in quartz tubes. A 1 mL portion of concentrated HNO₃ was added to the sample, tubes were left at room temperature overnight, then heated for 1 h at 90 °C, a further 1 mL of 30% H₂O₂ was added, and the tubes were heated for another 2 h at 90 °C. Finally, digests were diluted to 20 mL with Milli-Q water and stored at 4 °C until analysis. In order to attain enough solid material, all three replicate intestinal pellets were combined for total digestion.

Quality Control/Assurance and Analytical Procedure.

Certified reference material (bush, branches and leaves, NCS DC 73349) was digested using the procedures noted above (for the intestinal pellet) to provide quality control data (Table S1, Supporting Information). Mean Pb and Ca recoveries were 101% ($n = 6$) and 102% ($n = 6$), respectively.

The triplicate gizzard and intestine liquid digests, and the combined digest for the intestine pellet, were analyzed using graphite furnace atomic absorption spectroscopy [GF-AAS; AAnalyst800 with autosampler AS800 (Perkin-Elmer)] for Pb, using 50 μg of NH₄H₂PO₄ and 3 μg of Mg(NO₃)₂ as matrix modifiers for each atomization. One of the triplicate gizzard and intestine liquid digests was also analyzed using flame atomic emission spectroscopy [AES; AAnalyst800 with AS90 plus autosampler (Perkin-Elmer)] for Ca.

Statistical Analyses. Lead levels in solution in the gizzard (Pb-GIZZ^{soln}), the intestine (Pb-INT^{soln}) liquid, and the intestine solid (Pb-INT^{solid}) were log-transformed to attain a normal distribution and analyzed using generalized linear models (GLMs), to examine which factor variations caused significantly different concentration levels. The factors examined were grit type, number of Pb shot, amount of food (g), and time point. Posthoc Tukey tests were then used to examine differences found in both grit systems, due to the amount of food used, or time point. Differences between treatments in pH and Ca levels in the different compartments were tested using Mann–Whitney and Kruskal–Wallis tests

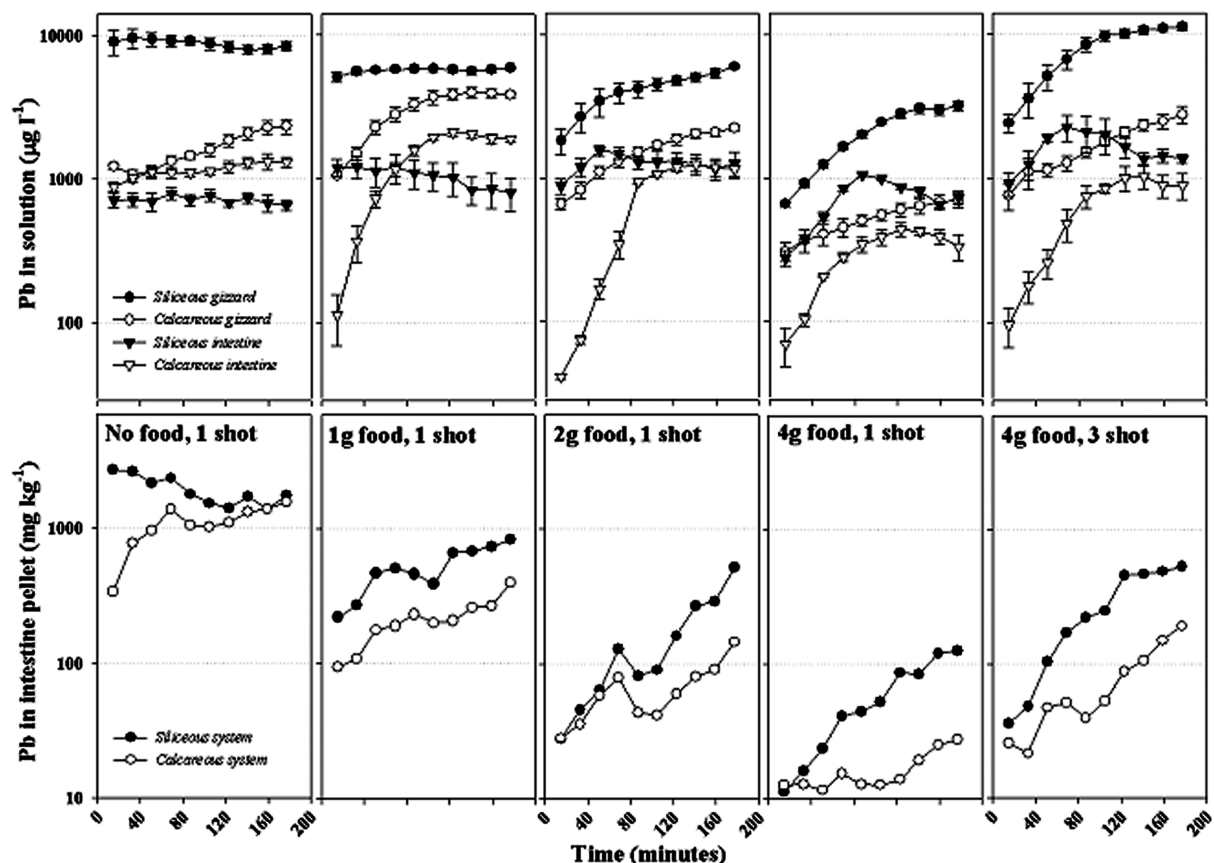


FIGURE 1. Mean (\pm SE based on three replicates) over time of Pb concentrations in solution (above) and in the composite intestinal pellet (below), for experiments using 0–4 g of food, 1 or 3 Pb shot, and 2 g of calcareous or siliceous grit, 124×84 mm (150×150 DPI).

since we observed a lack of homogeneity of variances when using GLMs. Correlations between Pb-GIZZ^{soln} and Pb-INT^{soln} levels as well as between Pb-INT^{soln} and Pb-INT^{solid} levels were tested with Pearson correlations for each treatment. Differences between Pb-GIZZ^{soln} and Pb-INT^{soln} were tested with a paired student *t*-test. The level of statistical significance used was $p < 0.05$, and all tests were performed using SPSS 15.0.

Results

Irrespective of treatment, (i.e., 1 or 3 shot, and 0, 1, 2, or 4 g of food), Pb-GIZZ^{soln} levels were consistently higher in simulations with Grit^{sil} than Grit^{cal} (GLM $p < 0.001$; Figure 1, Table 1). For Pb-INT^{soln} levels, this was also the case with 2 or 4 g of food and 1 or 3 Pb shot, but higher Pb-INT^{soln} concentrations were recorded in Grit^{cal} simulations without food and, at the temporal end of the simulation, with 1 g of food and 1 Pb shot (Figure 1; Figure S1, Supporting Information). For Pb-INT^{solid} in simulations with Grit^{sil}, Pb levels were always higher than in those with Grit^{cal} (GLM $p < 0.001$, Figure 1).

For experiments using 4 g of food, where grit type, number of shot (1 or 3), and time were factors, all factors were significant for Pb-GIZZ^{soln}, Pb-INT^{soln}, and Pb-INT^{solid} levels. As might be expected, Pb-GIZZ^{soln} concentrations increased with the number of Pb shot used (GLM $p < 0.001$, Figure 1), as did Pb-INT^{soln} and Pb-INT^{solid} levels (GLM both $p < 0.001$). In solution, levels were 3.7 (GIZZ^{sil}), 3.3 (GIZZ^{cal}), 2.3 (INT^{sil}), and 2.1 (INT^{cal}) times higher in systems with 3 shot rather than 1. In INT^{solid}, levels were 4.6 and 4.7 times higher in Grit^{sil} and Grit^{cal} systems, respectively (Table 1). Without Pb shot, Pb concentrations in solution were below our limits of detection for all treatments.

In treatments with 1 Pb shot, the amount of food used in each simulation affected Pb-GIZZ^{soln} and Pb-INT^{soln} levels (GLM both $p < 0.001$, Table 1). In gizzards with Grit^{sil}, the Pb-GIZZ^{soln} concentrations increased as food content decreased (Tukey all at $p < 0.001$). With Grit^{cal}, Pb-GIZZ^{soln} levels were also lowest with 4 g of food, but were highest (higher than with no food and 2 g) with 1 g of food ($p < 0.001$). In terms of Pb-INT^{soln} levels, with Grit^{sil}, concentrations were similar with 4 g and no food, and higher with 1 g and then 2 g of food (Tukey $p < 0.001$). Using Grit^{cal}, increased amounts of food (1–4 g) caused lower Pb-INT^{soln} levels (Tukey $p < 0.001$), and again, at the temporal end of the 1 g food simulation, levels were higher than when using no food (Figure 1). With time, Pb-GIZZ^{soln} concentrations increased, especially in simulations where food was included (GLM $p < 0.001$). Pb-INT^{soln} levels varied with time in all treatments (GLM $p < 0.001$), with maximum concentrations during the middle and at the end of simulations with Grit^{sil} and Grit^{cal} respectively (Figure 1).

In all experiments, for both grit types, Pb-INT^{soln} concentrations were lower than Pb-GIZZ^{soln} levels (paired Student *t* test $p < 0.001$; Figure 1, Table 1). Pb-INT^{soln} and Pb-GIZZ^{soln} levels were also correlated in all Grit^{cal} simulations (all at $r > 0.79$, $p < 0.001$), but only with 1 Pb shot + 4 g food when using Grit^{sil} (Table S2, Supporting Information). Pb-INT^{soln} and Pb-INT^{solid} levels were also correlated in Grit^{cal} simulations with no food and 1 g of food ($r > 0.738$, $p < 0.05$), and also with 4 g of food and 3 Pb shot ($r = 0.672$, $p = 0.05$). However, with Grit^{sil}, levels were negatively correlated when using 1 g of food ($r = -0.883$, $p = 0.001$).

GIZZ^{soln} pH was consistently lower in simulations with Grit^{sil} than with Grit^{cal} (Mann–Whitney $p < 0.001$). With Grit^{sil}, in the gizzard phase of experiments, the increased presence

TABLE 1. Mean (\pm SE) Concentrations in Solution of Pb (in $\mu\text{g L}^{-1}$) and Ca (mg L^{-1}) in Gizzard and Intestine Liquids and in the Intestine Pellet (mg kg^{-1}) with Ranges in Parentheses and pH Range Recorded During Each Experiment

	<i>n</i>	without food 1 shot	1 g food 1 shot	2 g food 1 shot	4 g food 1 shot	4 g food 3 shot
Gizzard Liquid-Siliceous						
Pb	10 ^a	8733 \pm 273 ^{A1} (5490–1075)	5660 \pm 74 ^{B1} (4295–6241)	4179 \pm 257 ^{C1} (1411–6107)	2099 \pm 170 ^{D1} (608–3595)	7856 \pm 604 ¹ (1737–12173)
Ca	10	7.8 \pm 1.1 (4.5–17.1)	28.3 \pm 3.3 (15.6–41.4)	45.2 \pm 5.5 (22.2–74.7)	64.1 \pm 4.9 (44.7–92.4)	77.8 \pm 13.2 (45.2–161.7)
pH	10	1.9–2.0	2.5–3.1	2.9–3.8	3.4–4.5	3.4–4.6
Gizzard Liquid-Calcareous						
Pb	10 ^a	1631 \pm 93 ^{B2} (937–2814)	3017 \pm 203 ^{A2} (1016–4588)	1537 \pm 101 ^{B2} (577–2428)	524 \pm 30 ^{C2} (219–884)	1727 \pm 132 ² (444–3402)
Ca	10	261.5 \pm 3.0 (250.0–285.8)	231.8 \pm 1.1 (227.3–237.5)	232.6 \pm 9.0 (206.7–282.9)	217.9 \pm 3.9 (202.0–243.8)	196.7 \pm 1.1 (188.1–199.4)
pH	10	6.1–6.4	5.4–5.4	5.2–5.4	5.4–5.5	5.2–5.3
Intestine Liquid-Siliceous						
Pb	10 ^a	715 \pm 21 ^{C2} (498–914)	1038 \pm 66 ^{B1} (523–1754)	1270 \pm 57 ^{A1} (736–1774)	715 \pm 46 ^{C1} (209–1112)	1625 \pm 116 ¹ (704–3292)
Ca	10	7.0 \pm 0.3 (5.9–8.3)	25.4 \pm 2.4 (16.1–36.7)	41.9 \pm 4.8 (22.5–65.8)	58.8 \pm 5.6 (39.7–91.8)	64.7 \pm 4.9 (50.6–97.4)
pH	10	6.4–6.7	6.4–6.6	6.4–6.5	6.4–6.7	6.3–6.5
Intestine Liquid-Calcareous						
Pb	10 ^a	1148 \pm 36 ^{A1} (763–1645)	1383 \pm 134 ^{A1} (26–2343)	742 \pm 94 ^{B2} (39–1471)	300 \pm 25 ^{C2} (48–517)	643 \pm 72 ² (50–1378)
Ca	10	231.8 \pm 2.4 (222.4–244.1)	196.3 \pm 5.8 (173.5–243.3)	174.8 \pm 6.9 (147.7–214.1)	163.5 \pm 3.4 (143.4–175.6)	158.8 \pm 3.7 (136.7–172.4)
pH	10	6.1–6.4	6.6–7.0	6.3–6.8	6.3–6.9	6.3–6.6
Intestine Pellet-Siliceous						
Pb	10	2001 \pm 172 ^{A1} (1392–2761)	521 \pm 63 ^{B1} (220–825)	168 \pm 48 ^{C1} (28–521)	60 \pm 13 ^{D1} (11–126)	277 \pm 61 ¹ (36–533)
Ca	10	402 \pm 22 (295–520)	356 \pm 22 (246–443)	376 \pm 35 (231–582)	476 \pm 16 (422–559)	356 \pm 40 (221–581)
Intestine Pellet-Calcareous						
Pb	10	1034 \pm 101 ^{A2} (338–1395)	213 \pm 27 ^{B2} (94–397)	66 \pm 11 ^{C2} (28–145)	16 \pm 2 ^{D2} (7–25)	77 \pm 18 ² (22–191)
Ca	10	11805 \pm 524 (10000–14538)	4621 \pm 229 (3632–6242)	3156 \pm 184 (2408–4300)	1998 \pm 104 (1428–2499)	2306 \pm 172 (1567–3361)

^a Mean of results for 30 digestions; 3 replicates were analyzed for each of 10 time points. Superscript capital letters illustrate whether means within the same row differ significantly or not (Tukey test $p < 0.05$), i.e., whether amount of food influences Pb in solution when 1 Pb shot was present.

of food tended to cause less acidic conditions, with the most acidic conditions (pH 1.9–2.0) observed without food (Kruskal–Wallis $p < 0.001$, Table 1). In contrast, with Grit^{cal}, food caused a reduction in pH (to around 5.2–5.5) in comparison to systems without food, where the pH was 6.1–6.4.

Ca-GIZZ^{soln}, Ca-INT^{soln}, and Ca-INT^{solid} levels were all higher in simulations using Grit^{cal} than with Grit^{sil} (Mann–Whitney all at $p < 0.001$; Table 1, Figure S2, Supporting Information). Ca-GIZZ^{soln} and Ca-INT^{soln} levels decreased with increasing food in Grit^{cal} simulations (Kruskal–Wallis both $p < 0.001$), but increased in Grit^{sil} systems (Kruskal–Wallis both $p < 0.001$).

Discussion

Grit Geochemistry and Associated pH Modification. Results show that Pb shot dissolution and concentrations within the digestive tract (gizzard and intestine) are affected significantly by coingested grit geochemistry. With Grit^{sil} in the gizzard, Pb-GIZZ^{soln} concentrations were consistently higher than when Grit^{cal} was present. This difference is probably due to geochemical interactions and/or the linked effect of grit geochemistry on prevailing gizzard liquid pH, since systems with Grit^{sil} maintained a more acidic environment than did those with Grit^{cal}. In both cases pH values were within the active range for duck pepsin (28). In previous *in vitro* digestion

models, prevailing pH has also been described as one of the most important limiting factors for Pb solubility (25, 32, 33). Gastric fluid is the most acidic digestive liquid in an avian digestive tract, and tends to solubilize the greatest fraction of various metals (25, 32, 34). Ultimately, our results are intrinsically linked to Pb solubility under differing Eh/pH conditions (35). At ambient temperature/pressure, in a Pb²⁺ + H₂O + CO₂ system (with $1 \times 10^{-6} \text{ mol L}^{-1} \text{ Pb}^{2+}$), under the Eh conditions likely in a digestive tract (mildly reducing), Pb²⁺ tends to remain in solution up to pH 5.9. Above this, and up to pH 7.7, PbCO₃ will tend to precipitate. Pb-INT^{soln} levels were also lower than Pb-GIZZ^{soln} levels in all treatments, and this indicates that Pb taken into solution in the gizzard subsequently precipitated out (with, for example, carbonates) within the intestine phase. In Grit^{sil} systems, 66–92% of Pb in solution at the gizzard stage did not remain so during the intestine phase, while in Grit^{cal} systems 30–63% reductions were recorded. Hence, while higher levels of Pb in solution were maintained in the gizzards with Grit^{sil} than with Grit^{cal}, the decreases within the intestine were also proportionally greater. Mean levels in solution within the intestine were not actually very dissimilar for the two grit types; Pb-INT^{soln} levels were 0.6–2.5 times higher in Grit^{sil} systems, while Pb-GIZZ^{soln} levels were 1.9–5.4 times higher than in Grit^{cal} systems (Table 1). Lower Pb concentrations in intestine phases, than in gizzard phases, have also been reported in previous *in vitro*

studies (25, 32, 33), and in some, Pb concentrations were below or near detection limits (25). The decreases observed here (66–92%) in Pb-INT^{soln} levels in Grit^{sil} systems were presumably not driven by carbonates derived from grit dissolution, as is likely to be the case where Grit^{cal} was present. Instead, carbonate derived from food material, from the NaHCO₃ added to adjust the intestine pH, or other potential coprecipitants (such as phosphates) and organic ligands (also from the food) may have acted to draw Pb levels in solution down (36, 37).

In Grit^{cal} systems we noted consistent correlations between Pb-INT^{soln} and Pb-GIZZ^{soln} levels, which may suggest that Pb dissolution here is controlled predominantly within the gizzard phase, with simple proportional precipitation occurring in the intestine. This may indicate that, in Grit^{cal} systems, grit solubility, the subsequent generation of free carbonate, and associated pH increases were exerting key (somewhat linear) controls over Pb levels in solution (38, 39). In the Grit^{sil} systems, we noted fewer correlations between Pb-INT^{soln} and Pb-GIZZ^{soln} levels, indicating that Pb dissolution and subsequent precipitation/adsorption in this system is more complex (40). In Grit^{sil} systems, high Pb-GIZZ^{soln} levels were recorded, and a high proportion of this precipitated or became absorbed from solution in the intestine. However, simple proportional reductions were often not occurring. Hence, for Grit^{sil} systems at least, it may be difficult to extrapolate *in vitro* data based solely on gizzard simulations, to toxicologically relevant intestine levels. It seems important in this case to couple and include the intestine phase since precipitation/adsorption reactions occur primarily within the intestine, as does Pb absorption into the bloodstream (26). Simple gizzard based models that do not consider the passage of fluid to very differing intestine conditions (or include grit, if appropriate) may ultimately seriously overestimate potential toxicological impacts. While considering worst case scenarios may be useful, more realistic simulations should also be designed.

Effects Caused by Differing Amounts of Food. As noted previously (27, 33, 34, 41), pH and Pb dissolution were not only affected by grit geochemistry, but also by the presence of food in the gizzard in all treatments. In Grit^{sil} systems, increased amounts of food caused pH-GIZZ^{soln} values to rise, and Pb-GIZZ^{soln} levels to fall. Again, more alkaline conditions and increased levels of food derived carbonate, phosphate, and organic ligands are likely to have promoted these reductions (34, 41). Elsewhere, such decreases have been linked to Pb precipitation with organic ligands such as phytic acid salts (38). In Grit^{cal} systems, food had the opposite effect, whereby its presence within the system caused decreased pH, i.e., 5.2–5.5 in systems with food, but 6.1–6.4 in those without. Thus, with 1 g of food, mean Pb-GIZZ^{soln} levels were actually higher than if no food had been added. This may be due to several factors. First (and most importantly), with food, the pH was more acidic and below 5.9 (the level below which Pb²⁺ tends to remain in solution), yet without food, it was above this level and dissolution rates are likely to have been lower, while precipitation with grit derived carbonate was more likely to occur. Second, food itself may have enhanced physical abrasion and predissolution oxidation of the Pb shot surface, and/or acted to bind hydroxide (OH⁻) from solution (resulting in a lowered pH). Both processes would act to allow higher Pb-GIZZ^{soln} levels in systems with 1 g of food, versus those without food. Ultimately, simulations with increasing amounts of food (from 1–4 g) did however have progressively lower Pb levels in solution (as observed in the Grit^{sil} system). In addition to the mechanisms already noted above, in this case higher Ca levels may also have enhanced the sorption of Pb to organic ligands such as phytic acid (42).

Including food in simulations such as this is complicated. However, we believe that, in doing so, more realistic results are generated. Many previous studies regarding *in vitro* gastrointestinal simulations use “fasted” (without food) conditions, since there is bound to be a degree of variation in determined Pb bioaccessibility, depending on the type and amount of food added (25, 32). However, diets high in calcium and protein, and low in fiber, can reduce Pb accumulation in tissues and resultant toxicity (7, 20, 28, 43). While fasted system models may be simpler, and provide valid data, in reality birds such as waterfowl can survive for long periods with ingested Pb shot in their gizzards (7), and it is critical that interactions with food (and ingested grit) are considered, if realistic toxicological simulations are to be developed.

Dietary Grit Supplementation. Irrespective of treatment or phase, Ca concentrations in Grit^{cal} systems were always higher than in corresponding Grit^{sil} experiments. Birds with a low Ca diet who are exposed to Pb (via drinking water, or as Pb shot) have been shown to develop more severe clinical signs of Pb poisoning, and deposit higher amounts of Pb in soft tissues, than birds who are given Ca supplements (20, 21, 44). For example, mallards with a daily Ca intake between 5 and 13 mg showed greater signs of Pb toxicity than those with intakes of 430–840 mg (20), and in non-breeding passerines exposed to Pb, greater accumulation of Pb in tissues occurred when Ca was reduced in the diet [from 3% to 0.3% (21)]. According to our results, Ca-GIZZ^{soln} and Ca-INT^{soln} levels in Grit^{cal} systems were 2.4–33.5 times higher than in Grit^{sil} systems. Although not surprising, these results indicate that, by providing supplementary Grit^{cal} to birds, especially in areas where Pb shot may be prevalent and/or Ca intake may be low (i.e., areas where underlying geology is low in Ca), the effects and bioaccessibility of Pb (from shot) may be reduced. Calcite supplementation would have several effects, i.e., to increase prevailing digestive pH in the gizzard (and hence reduce Pb dissolution), to increase carbonate levels (and promote Pb precipitation within the intestine), and, by increasing levels of Ca, to allow intestinal absorption competition (19). Our data indicate that, in certain scenarios, with low levels of food, Pb-INT^{soln} levels may be higher in systems with Grit^{cal} than with Grit^{sil} (Figure S1, Supporting Information). However, if this data set is taken as a whole, it would be incorrect to suggest that Grit^{sil} would perhaps be more beneficial than Grit^{cal} supplementation, simply because calcite is likely to play not just one, but multiple important geochemical, and potentially beneficial roles.

Although the concept of using grit supplementation to reduce Pb shot ingestion in birds (7, 14, 44), or to reduce lead uptake/accumulation in tissues, has been proposed previously, there are remarkably few experimental studies which have considered this a viable environmental management tool, and there is almost no large scale field trial data available. Experimental work with ducks in captivity given Pb shot and grit have shown that grit rich in calcium may reduce mortality (7). However, Grit^{cal} also has a shorter half-life within the gizzard (than Grit^{sil}), since it dissolves more rapidly (2, 3). As such, birds tend to need to consume more Grit^{cal} per unit time, and hence the risk of Pb shot ingestion may increase. At the same time, higher grit consumption rates/turnover may increase avoidance of shot from the gizzard, but it might also increase Pb shot erosion, and therefore exposure/mortality (7). Further, if grit supplements are given using, for example, a “corn grain + grit” mix, ducks tend to consume twice as much grit than those given “commercial duck pellets + grit”, and again, this may result (in the field) in a greater probability of Pb shot ingestion (2, 3). The only field scale experimental work we are aware of which has studied grit supplementation to reduce the probability of Pb shot ingestion in waterfowl was completed in the Camargue

(France). Here, reduced ingestion was reported 1 year after supplementation, especially for *Aythya fuligula* (from 40% to 8%) (13). Besides grit supplementation, several other alternative environmental management practices have been proposed to reduce Pb shot ingestion in wetlands with high pellet densities, i.e., land cultivation, water level management, and food supplementation (7). However, grit supplementation could be quite easily and cheaply applied near wetlands or feeding grounds at key times of year, and this would ultimately have a very low environmental impact.

Concluding, this dynamic *in vitro* simulated avian gizzard-intestine system has been applied here to study Pb dissolution from Pb shot. However, the system could also be used to study the dissolution of a much wider range of contaminants (As, Cd, Ni, Zn) from a wider set of matrixes (soil, biota). Also, by incorporating different “study specific/relevant” criteria, such as grit type/geochemistry, grit size, food type, or amounts of either, more realistic indications of dissolution rates, intestinal precipitation, and bioaccessibility may be gained for specific avian species. The proposed technique attempts to consider waterfowl physiology in its design, and may therefore provide a more complete type of “sequential extraction” than has previously been suggested, which mimics the avian digestive process more accurately. Ultimately, similar systems could also be created which would more precisely account for the physical/mechanical action of the gizzard upon ingested materials (as suggested in (23)). This physical/mechanical action is difficult to simulate in the laboratory, but is perhaps the most lacking consideration in this and many other previous studies (24, 25).

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Supporting Information Available

Raw data and graphical representations for Pb, Ca, and pH, associated statistical results, XRF methods and results, and quality control/assurance data. This material is available free of charge via the Internet at <http://pubs.acs.org>.

Literature Cited

- Gionfriddo, J. P.; Best, L. B. Grit use by birds: A review. *Curr. Ornithol.* **1999**, *15*, 89–148.
- Trost, R. E. Dynamics of grit selection and retention in captive mallards. *J. Wildl. Manage.* **1981**, *45*, 64–73.
- Mateo, R.; Guitart, R. The effects of grit supplementation and feed type on steel-shot ingestion in mallards. *Prev. Vet. Med.* **2000**, *44*, 221–229.
- Bendell-Young, L. I.; Bendell, J. F. Grit ingestion as a source of metal exposure in the spruce grouse, *Dendragapus canadensis*. *Environ. Pollut.* **1999**, *106*, 405–412.
- King, J. R.; Bendell-Young, L. I. Toxicological significance of grit replacement times for juvenile mallards. *J. Wildl. Manage.* **2000**, *64*, 858–862.
- Taggart, M. A.; Mateo, R.; Charnock, J. M.; Bahrami, F.; Green, A. J.; Meharg, A. A. Arsenic rich iron plaque on macrophyte roots - an ecotoxicological risk. *Environ. Pollut.* **2009**, *157*, 946–954.
- Sanderson, G. C.; Bellrose, F. C. A review of the problem of lead poisoning in waterfowl. *Illinois Nat. Hist. Surv. Special Publication* **1986**, *4*, 1–34.
- Mateo, R. Lead poisoning in wild birds in Europe and the regulations adopted by different countries In *Ingestion of Lead from Spent Ammunition: Implications for Wildlife and Humans*; Watson, R. T., Fuller, M., Pokras, M., Hunt, W. G., Eds.; The Peregrine Fund: Boise, ID, 2009; pp 71–98.
- Bellrose, F. C. Lead poisoning as a mortality factor in waterfowl populations. *Illinois Nat. Hist. Surv. Bull.* **1959**, *27*, 235–288.
- Friend, M. *Field guide to wildlife diseases*; U.S Department of the Interior, Fish and Wildlife Service: Washington, DC, 1987.
- Taggart, M. A.; Green, A. J.; Mateo, R.; Svanberg, F.; Hillström, L.; Meharg, A. A. Metal levels in the bones and livers of globally threatened marbled teal and white-headed duck from El Hondo, Spain. *Ecotoxicol. Environ. Safe.* **2009**, *72*, 1–9.
- Cade, T. J. Exposure of California condors to lead from spent ammunition. *J. Wildl. Manage.* **2007**, *71*, 2125–2133.
- Hofmann, L. Le saturnisme fleau de la sauvagine en Camargue. *Terre Vie* **1960**, *107*, 120–131.
- Thomas, V. G.; Scheuhammer, A. M.; Bond, D. E. Bone lead levels and lead isotope ratios in red grouse from Scottish and Yorkshire moors. *Sci. Total Environ.* **2009**, *407*, 3494–3502.
- Figuerola, J.; Mateo, R.; Green, A. J.; Mondain-Monval, J. Y.; Lefranc, H.; Mentaberre, G. Grit selection in waterfowl and how it determines exposure to ingested lead shot in Mediterranean wetlands. *Environ. Conserv.* **2005**, *32*, 226–234.
- Mann, A. W.; Deutscher, R. L. Solution geochemistry of lead and zinc in water containing carbonate, sulphate and chloride ions. *Chem. Geol.* **1980**, *29*, 293–311.
- Blake, K. C. H.; Mann, M. Effect of calcium and phosphorus on the gastrointestinal absorption of ^{203}Pb in man. *Environ. Res.* **1983**, *30*, 188–194.
- Mykkanen, H. M.; Wasserman, R. H. Gastrointestinal absorption of lead (^{203}Pb) in chicks: Influence of lead, calcium, and age. *J. Nutr.* **1981**, *111*, 1757–1765.
- Fullmer, C. S.; Edelstein, S.; Wasserman, R. H. Lead-binding properties of intestinal calcium-binding proteins. *J. Biol. Chem.* **1985**, *260*, 6816–6819.
- Carlson, B. L.; Nielsen, S. W. Influence of dietary calcium on lead poisoning in mallard ducks (*Anas platyrhynchos*). *Am. J. Vet. Res.* **1985**, *46*, 276–282.
- Scheuhammer, A. M. Influence of reduced dietary calcium on the accumulation and effects of lead, cadmium, and aluminum in birds. *Environ. Pollut.* **1996**, *94*, 337–343.
- Kimball, W. H.; Munir, Z. A. Corrosion of lead shot in a simulated waterfowl gizzard. *J. Wildl. Manage.* **1971**, *35*, 360–365.
- Moore, S. J. The comparative functional gizzard morphology of several species of birds. *Aust. J. Zool.* **1998**, *46*, 359–368.
- Levengood, J. M.; Skowron, L. A. Use of a simulated gizzard to measure bioavailability of metals and other elements to waterfowl. *Ecotoxicology* **2002**, *10*, 299–304.
- Furman, O.; Strawn, D. G.; Heinz, G. H.; Williams, B. Risk assessment test for lead bioaccessibility to waterfowl in mine-impacted soils. *J. Environ. Qual.* **2006**, *35*, 450–458.
- Pain, D. J. Lead in the environment. In *Handbook of Ecotoxicology*; Hoffman D. J., Rattner, B. A., Burton, G. A., Cairns, J., Eds.; Lewis Publishers: Boca Raton, FL, 1995; pp 356–391.
- Denbow, D. Gastrointestinal anatomy and physiology In *Sturkie's Avian Physiology*; Whittow G., Ed.; Academic Press, University of Hawaii at Manoa: Honolulu, 2000; pp 299–325.
- Clemens, E. T.; Krook, L.; Aronson, A. L.; Stevens, C. E. Pathogenesis of lead shot poisoning in the mallard duck. *Cornell Vet.* **1975**, *65*, 248–285.
- Clark, R. G.; Gentle, G. C. Estimates of grain passage time in captive mallards. *Can. J. Zool.* **1990**, *68*, 2275–2279.
- Mateo, R.; Guitart, R.; Green, A. J. Determinants of lead shot, rice, and grit ingestion in ducks and coots. *J. Wildl. Manage.* **2000**, *64*, 939–947.
- Dawson, W. R.; Whittow, G. C. Regulation of body temperature In *Sturkie's Avian Physiology*; Whittow G., Ed.; Academic Press, University of Hawaii at Manoa: Honolulu, 2000; pp 343–390.
- Ruby, M. V.; Davis, A.; Schoof, R.; Eberle, S.; Sellstone, C. M. Estimation of lead and arsenic bioavailability using a physiologically based extraction test. *Environ. Sci. Technol.* **1996**, *30*, 422–430.
- Oomen, A. G.; Hack, A.; Minekus, M.; Zeijdner, E.; Cornelis, C.; Schoeters, G.; Verstraete, W.; Van De Wiele, T.; Wragg, J.; Rompelberg, C. J. M.; Sips, A. J. A. M.; Van Wijnen, J. H. Comparison of five *in vitro* digestion models to study the bioaccessibility of soil contaminants. *Environ. Sci. Technol.* **2002**, *36*, 3326–3334.
- Schroder, J. L.; Basta, N. T.; Casteel, S. W.; Evans, T. J.; Payton, M. E.; Si, J. Validation of the *In Vitro* Gastrointestinal (IVG) methods to estimate relative bioavailable lead in contaminated soils. *J. Environ. Qual.* **2004**, *33*, 513–521.
- Cao, X.; Ma, L. Q.; Chen, M.; Hardison Jr, D. W.; Harris, W. G. Weathering of lead bullets and their environmental effects at outdoor shooting ranges. *J. Environ. Qual.* **2003**, *32*, 526–534.
- Gerritse, R. G.; Vriesema, R.; Dalenberg, J. W.; De Roos, H. P. Effect of sewage sludge on trace element mobility in soils. *J. Environ. Qual.* **1982**, *11*, 359–364.
- Bullock, J. I.; Duffin, P. A.; Nolan, K. B.; Smith, T. K. Effect of phytate on the *in-vitro* solubility of Al^{3+} , Ca^{2+} , Hg^{2+} and Pb^{2+}

- as a function of pH at 37°C. *J. Sci. Food Agric.* **1995**, 67, 507–509.
- (38) Schock, M. R. Response of lead solubility to dissolved carbonate in drinking water. *J. Am. Water Works Assoc.* **1980**, 72, 695–704.
- (39) Godelitsas, A.; Astilleros, J. M.; Hallam, K.; Harissopoulos, S.; Putnis, A. Interaction of calcium carbonates with lead in aqueous solutions. *Environ. Sci. Technol.* **2003**, 37, 3351–3360.
- (40) Bickmore, B. R.; Wheeler, J. C.; Bates, B.; Nagy, K. L.; Eggett, D. L. Reaction pathways for quartz dissolution determined by statistical and graphical analysis of macroscopic experimental data. *Geochim. Cosmochim. Acta* **2008**, 72, 4521–4536.
- (41) Ruby, M. V.; Davis, A.; Link, T. E.; Schoof, R.; Chaney, R. L.; Freeman, G. B.; Bergstrom, P. Development of an *in vitro* screening test to evaluate the *in vivo* bioaccessibility of ingested mine-waste lead. *Environ. Sci. Technol.* **1993**, 27, 2870–2877.
- (42) Wise, A.; Gilbert, D. J. Binding of cadmium and lead to the calcium-phytate complex *in vitro*. *Toxicol. Lett.* **1981**, 9, 45–50.
- (43) Douglas-Stroebel, E.; Hoffman, D. J.; Brewer, G. L.; Sileo, L. Effects of lead-contaminated sediment and nutrition on mallard duckling brain growth and biochemistry. *Environ. Pollut.* **2004**, 131, 215–222.
- (44) Dauwe, T.; Snoeijs, T.; Bervoets, L.; Blust, R.; Eens, M. Calcium availability influences lead accumulation in a passerine bird. *Anim. Biol.* **2006**, 56, 289–298.

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