

## Part I

Moving towards the sustainable fisheries framework for Pacific herring: data, models, and alternative assumptions.

## Part II

Stock Assessment and Management Advice for the British Columbia Pacific Herring Stocks:  
2011 Assessment and 2012 Forecasts

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## Abstract

Estimates of herring abundance in British Columbia (B.C.) waters has been based on catch-age data and spawn survey abundance information. These data are typically interpolated using a statistical catch-age framework; however, virtual methods (e.g., VPA) have been used in the past. This assessment also uses a statistical catch age model. This document is broken into two parts: Part I deals with moving the herring assessments towards Canada's sustainable fisheries framework and introduces a new integrated statistical catch-age model for jointly estimating the abundance of Pacific herring and associated reference points to be used in the sustainable fisheries framework. Part II of this document implements this new assessment framework using the data for the five major and two minor regions. Finally, we present catch advice based on decision tables that utilize poor, average, and good age-3 recruitment forecasts.

In Part I of this document we provide a very brief description of the new assessment framework (a full technical description of the model is provided in the Appendix of this document). We then conduct some simulation testing with perfect information to demonstrate that the model is capable of estimating all the parameters. We further explore precision and bias in parameter estimates based on simulation data with both observation and process errors. We then parameterize the new assessment model such that the assumptions of the previous assessment model (Herring Catch Age Model, or HCAM) are mostly met and compare parameter estimates and estimates of spawning stock biomass (using data from 1951:2010). Using data from the Strait of Georgia only, we then compare alternative assumptions about the spawn survey scaling coefficient ( $q$ ), natural mortality and selectivity, and examine how these alternative assumptions influence estimates of key parameters (unfished spawning biomass, steepness, average natural mortality). Relaxing assumptions about  $q$  and natural mortality rates had the largest impacts on estimated parameters. Lastly, we compared estimates of spawning stock biomass from HCAM with the new model for all five major areas to understand the subtle differences between the assumptions in the two models.

In Part II of this document we present updated data from the herring fisheries and surveys in 2011, a brief description of the analytical methods used to construct the decision tables, and present the results of the application of the new assessment model to the 2011 data. New this year is a Bayesian prior for the dive survey spawn index ( $q$ ) and the development of this prior is detailed in the appendix. The expected value of  $q$  was estimated to be 0.587 with a standard deviation of 0.155. To summarize the overall fit to the model, maximum likelihood estimates derived quantities and residuals between observed and predicted variables are used. Retrospective analysis (i.e., the sequential removal of the most recent data) is used as a diagnostic for model misspecification. Catch advice (decision tables) are based on the median values of random samples from the joint posterior distribution and not the maximum likelihood estimates. Visual inspection of the trace plots from the posterior samples and pair plots were used to judge if the samples were taken from a stationary distribution. Historically catch advice was based on cutoff values that were derived from 1996 estimates of the unfished biomass ( $B_0$ , cutoff values are set at  $0.25B_0$ ). This assessment provides updated estimates of  $B_0$  and presents catch advice based on new cutoff values. An alternative decision table, where catch advice is based on old cutoffs, is also presented.

Median estimates of the 2011 spawning stock biomass is as follows: Haida Gwaii (HG) –16,579 t, Prince Rupert District (PRD) – 27,046 t, Central Coast (CC) – 14,666 t, Strait of Georgia (SOG) 125,261 t, West Coast Vancouver Island (WCVI) 14,679 t. Implementation of the current harvest control rule (HCR) advises no fishing in HG under poor recruitment and no fishing in CC under poor and average recruitment. Based on the 20% harvest rate and application of the harvest control rule the estimated maximum available harvest ranges from 4,296 t in HG to 27,690t in SOG (assuming good recruitment). Catch advice for the minor areas is based on a 10% fixed exploitation rate with no cutoffs and ranges from 91 t in Area 27 (assuming poor recruitment) to 614 t in Area 2W (assuming good recruitment).

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## **Part I**

# **Moving towards a sustainable fisheries framework for Pacific herring: data, models and alternative assumptions**

## 1.1 Introduction

There are four major objectives of this paper: (1) to describe in detail an alternative integrated statistical catch-age model ( $i\text{SCA}_M$ ), (2) examine parameter estimation performance using iSCAM, (3) perform a side-by-side comparison of the previous HCAM and iSCAM on the five major herring stocks, and (4) explore alternative assumptions about selectivity, catchability, and natural mortality using iSCAM. The most recent assessment of BC herring stocks was conducted in 2010 using the Herring Catch Age Model (HCAMv2) which is documented in [Cleary and Schweigert \(2010\)](#). Furthermore, a review sponsored by the Herring Research and Conservation Society (HRCS) was conducted June 17-18, 2010 in Nanaimo, BC where an expert panel addressed specific questions about the current implementation of the HCAMv2 model and suggested recommendations for each of the questions. This paper also attempts to address some of the points brought up in the review.

BC herring are currently managed as five major stocks and 2 minor stocks (Figure 1.1). Annual catch advice for each of these areas is based on current estimates of stock status, and a 20% exploitation rate if the stock is above the cutoff level for the five major stocks and a 10% exploitation rate for the two minor stocks. Cutoff levels for the five major stocks are  $0.25B_o$ , and estimates of unfished biomass were established first in 1985 ([Haist et al., 1986](#)). Estimates of  $B_o$  were updated most recently in 1996, and in last years HCAM assessment marginal posterior distributions for  $B_o$  were also presented but were not used to update cutoffs ([Cleary and Schweigert, 2010](#)). These cutoffs are currently thought to be more conservative than the current default Limit Reference Point of  $0.4B_{MSY}$  ([Fisheries and Oceans Canada, 2006](#)). However, estimates of  $B_o$  and MSY based reference points have not been examined for Pacific herring for some time. In this paper we also describe the methods for updating estimates of  $B_o$  and MSY based reference points using the  $i\text{SCA}_M$  model framework. We also compare estimates of MSY based reference points for the Strait of Georgia herring under the previously mentioned alternative assumptions (see point (4) in the previous paragraph).

We do not provide a detailed description of HCAM (v2) in this paper and we refer the reader to [Schweigert et al. \(2009\)](#) and [Cleary and Schweigert \(2010\)](#). We first begin with a description of the input data required and assumptions about the data, followed by a detailed description of the analytical methods and assumptions in iSCAM. We then present the analytical methods and assumptions for exploring alternative hypotheses about selectivity, catchability and natural mortality, followed by a description of the elements that make up the joint posterior distribution (i.e., likelihoods, priors, and penalties). Parameter estimation and quantifying uncertainty is carried out using AD Model Builder ([ADMB Project, 2009](#)). We then explore estimation performance in iSCAM using simulation experiments where the model is used to generate simulated observations with known parameter values, then estimate parameter, and repeat this exercise a number of times to evaluate bias and precision in parameter estimates. Finally, we present forecasts of pre-fishery biomass and available harvest options using the cutoffs (e.g., reproduce Table 5 in [Cleary and Schweigert, 2010](#)) as well as available harvest options based on the Sustainable Fisheries Framework (i.e., [Fisheries and Oceans Canada, 2006](#)) for comparison.

## 1.2 Methods

### 1.2.1 Analytical methods

We present a new stock assessment platform for the assessment of BC Pacific herring. This platform is based on the general statistical catch age model first described by [Fournier and Archibald \(1982\)](#). The software platform used is called  $i\text{SCA}_M$ , which stands for integrated Statistical Catch Age Model. The source code and documentation for  $i\text{SCA}_M$  is freely available from <https://sites.google.com/site/iscamproject/>, or from a subversion repository at <http://code.google.com/p/iscam-project/>. The subversion repository is more likely to be up to date; whereas, the project website has periodic updates with corresponding version numbers. Ideally, the results of this report could easily be repeated just by downloading the necessary software and using the data and control files presented in the appendix of this paper. A complete technical description of  $i\text{SCA}_M$  is provided in Appendix A.1 of this report.

At times this document reads more like a users manual for the  $i\text{SCA}_M$  software and this is intentional as we expect that reviewers may wish to repeat the efforts to verify model results. The details of the analytical

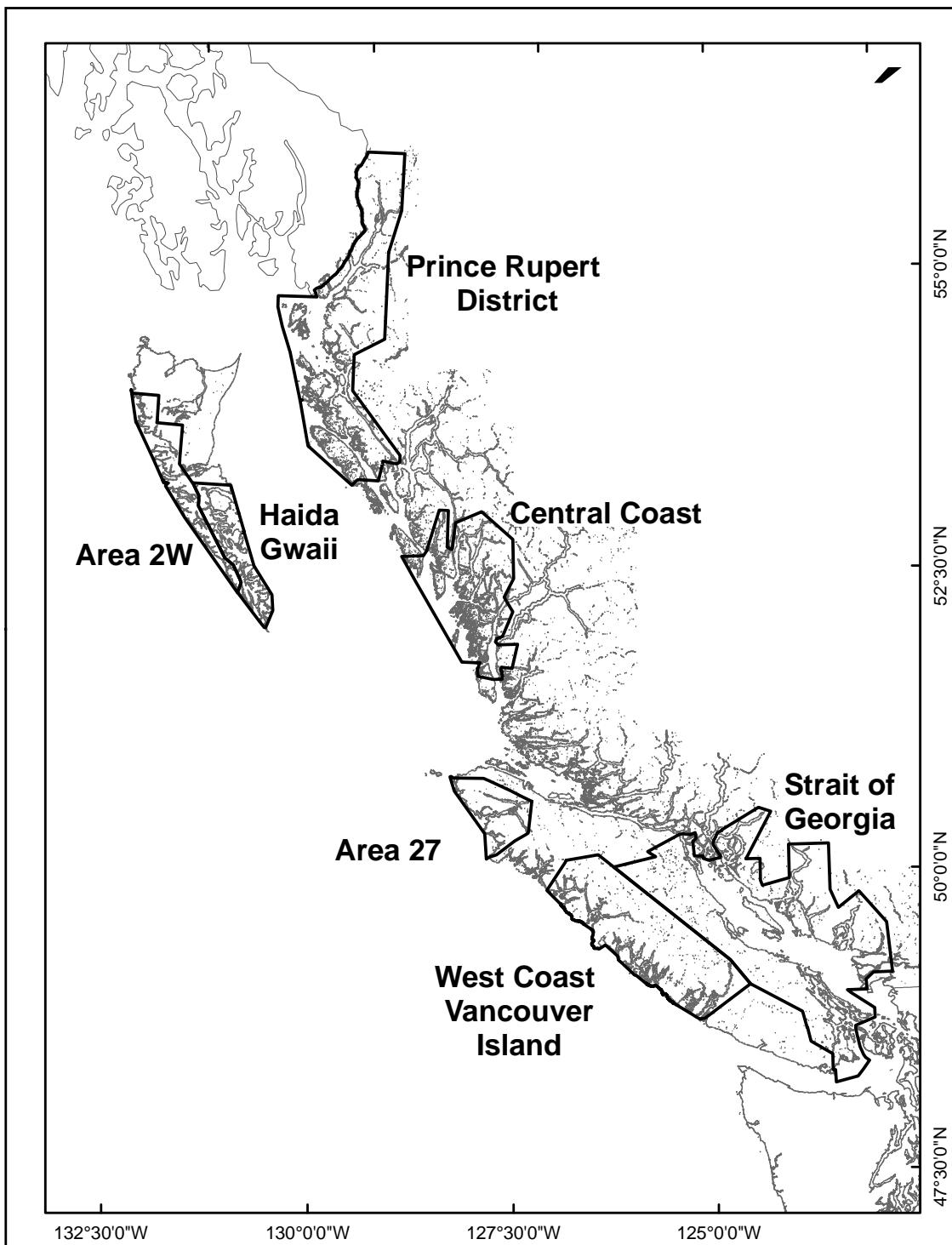


Figure 1.1: B.C. herring major stock areas: Haida Gwaii (HG or QCI 2E), Prince Rupert District (PRD), Central Coast (CC), Strait of Georgia (SOG), West Coast Vancouver Island (WCVI), and minor stock areas: Area 2W and Area 27.

equations are identified in Appendix A.1, and the text herein will often refer to how to implement the scenario in the software itself.

In short, for each stock two input files are required for  $^i\text{SCA}_M$ : (1) a data file that contains the historical catch, survey, life-history and age-composition information, and (2) a control file that specifies initial parameter values, priors, selectivity options, and various other controls that specify options for time-varying natural mortality, type of recruitment model, etc. Each major and minor stock has its own data and control file and these are provided in Appendix B.2 such that these results can be verified by an independent reviewer using the  $^i\text{SCA}_M$  software.

Estimated model parameters includes the initial numbers-at-age, annual age-2 recruits, annual fishing mortality rates for each gear, selectivity parameters, natural mortality rates, parameters that describe the observation error and process error variance, and the unfished age-2 recruits and the steepness of the stock recruitment relationship. The total number of estimated parameters differs for each stock assessment region depending on the number of years of active fishing, assumptions about selectivity, and the number of assumed nodes in natural mortality rates.

### 1.2.2 Sustainable Fisheries Framework reference points

The Sustainable Fisheries Framework provides the basis for ensuring Canadian fisheries are conducted in a manner which support conservation and sustainable use. The framework incorporates existing fisheries management policies along with new and evolving policies. The framework consists of two main elements: conservation and sustainable use policies, and planning and monitoring tools. The conservation and sustainable use policies incorporate the precautionary and ecosystem approaches to fisheries management.

The general framework for developing a harvest strategy that is compliant with the precautionary approach is to divide the stock into three stock status zones; healthy, cautious, and critical zones ([Fisheries and Oceans Canada, 2006](#)). In this work we define these stock status thresholds as  $0.4B_{MSY}$  and  $0.8B_{MSY}$  when crossing from the critical-cautious zone and cautious-healthy zone, respectively. A critical component to this interpretation is the definition of  $B_{MSY}$ . In the case of a single fishing fleet using a fixed gear,  $B_{MSY}$  is normally defined as the spawning stock biomass that would, on average, support the largest surplus yield. In the case of multiple fishing fleets, each using a different gear with different selectivities, the definition of  $B_{MSY}$  is more complex and is a function of allocation of mortality to each gear type. For example, if one gear harvest fish at a much younger age than the other gear, an increase in allocation to the gear that catches younger sexually immature fish will shift  $B_{MSY}$  upwards. Also, changes in selectivity over time (perhaps due to changes in growth) will result in changing  $B_{MSY}$ . This precise definition for  $B_{MSY}$  increases the difficulty in estimating reference points when there are non-stationary parameters in the model (i.e., time-varying natural mortality rates, selectivity, growth in the case of Pacific herring).

Estimates of reference points are based on equilibrium calculations (see Appendix A.1), where the average natural mortality rate over the time period in question is used along with the most recent empirical estimates of weight-at-age and fecundity. In lieu of any formal allocation arrangements for the gear-types that harvest herring, the ratios of average catch over the past 20 years for each gear type, in each SAR, are used for allocation purposes and calculating MSY based reference points.

### 1.2.3 Simulation testing

The purpose of conducting simulation testing is two fold: (1) to demonstrate that the model is capable of estimating model parameters given perfect information, and (2) to examine precision and bias in parameter estimates (and corresponding management quantities) in the presence of observation and process errors. To conduct simulation testing using  $^i\text{SCA}_M$ , the following command line option `-sim 1234` is used, where 1234 is a unique random number seed. There is also a special seed number `-sim 000` that generates data with no error. That is the simulation model is deterministic, the relative abundance data are directly proportional with 0 observation error, and the age-composition data replicates precisely the true vulnerable proportions-at-age. The simulation model is conditioned on the historical catch data specified in the data file, and the true parameter values used to simulate the data are those that are specified in the control file.

## Estimation performance with perfect information

When simulating data with perfect information, there are a couple of things that need to be highlighted here when trying to estimate parameters from data that contain no error. First, the phases for the precision and variance partitioning parameters ( $\vartheta$  and  $\rho$ ) should be set to a negative number. There are no error in the data, therefore, there is no need to estimate the variance terms for the error distributions. Second, in the control file, the initial value for the precision ( $\vartheta$ ) should be set to an extremely large number (e.g., 4999.999, assuming the upper bound is 5000). The reason to set this number large is to minimize the slight bias due to the lognormal bias correction in the stock recruitment relationship (i.e., the  $-0.5\tau^2$  term in [T3.13](#), or [T3.14](#)). The control file used to simulate the fake data is provided in [Appendix B.2.8](#).

## Bias & precision with observation & process errors

To determine bias and precision of parameter estimates when the model is confronted with both observation error and process error, a series of Monte Carlo trials are performed and  $\log_2$  ratios are used to measure the distribution of estimated parameters ( $\hat{\theta}$ ) from the true value( $\theta$ ). The  $\log_2$  ratio

$$\log_2 \left( \frac{\hat{\theta}}{\theta} \right)$$

is zero when  $\hat{\theta} = \theta$ , is 1 when  $\hat{\theta} = 2\theta$ , and is -1 when  $\hat{\theta} = 0.5\theta$ . Box plots are used to examine the distribution of 50 trials where a unique random number seed is used for each trial. For the purposes of the simulation experiments only, we assume that the proportion of the total variance associated with observation error is known ( $\rho = 0.25$ ) and estimate the total variance. The total precision is set to 2.50 which is equivalent to a total standard deviation of 0.4 (i.e.,  $1/2.50=0.4$ ). The control file used for the Monte Carlo procedures is provided in [Appendix B.2.8](#), on page [119](#).

The simulation testing presented here is not very extensive in that only a single scenario based on the catch time series in the Strait of Georgia and an arbitrary parameter set  $\theta$  was used to explore bias and precision. In practicality, additional simulation testing should be conducted preferably using the MLE estimates of  $\hat{\theta}$  as true parameter values to determine if the data and model structure combined are informative about the true parameter values. It is possible to obtain a spurious results based on the choice of  $\theta$  that was used in simulation trials.

### 1.2.4 Comparison of HCAM with $i\text{SCA}_M$

There are a number of different statistical assumptions and structural differences between the previous assessments using HCAM (Herring Catch Age Model) and  $i\text{SCA}_M$ . Here we briefly summarize the differences and similarities between the two approaches, and we first attempt to formulate the  $i\text{SCA}_M$  model to be as similar as possible to the last implementation of HCAM used in [Cleary and Schweigert \(2010\)](#).

The objective function in the HCAM model has four major components to it: 1) the likelihood of the age composition data, 2) the likelihood of the commercial catch data, 3) the likelihood of the spawn data, and 4) the prior densities for estimated model parameters. The following subsections describe how  $i\text{SCA}_M$  was set up to best approximate HCAM.

For the gillnet fishery, HCAM implements a time varying selectivity scheme as a function of the average weight-at-age. A similar implementation was also developed in  $i\text{SCA}_M$ . Alternative options for changes in selectivity will be investigated later in this paper. For the remainder of this paper and especially in the Figures, the definitions for Gear is as follows: Gear 1 = winter purse seine fishery, Gear 2 = seine-roe fishery, Gear 3 = gillnet fishery.

#### Age-composition data

There are two alternative likelihoods specified for the age-composition data in HCAM (see Table 8 in Appendix B in [Cleary and Schweigert, 2010](#)); a multinomial likelihood (T8.1), and a robust normal approximation to the multinomial(T8.2). In  $i\text{SCA}_M$ , a multivariate logistic negative log-likelihood for age-composition

data is used (see equation A.10 above); this likelihood weights age-composition data based on the conditional maximum likelihood estimate of the variance. Whereas, the multinomial likelihoods weights data based on the effective sample size; the effective sample size is normally determined iteratively by examining the distributions of residuals relative to other sources of information that the model is fit to (see [Gavaris and Ianelli, 2002](#), for full details).

In addition,  $iSCAM$  requires a minimum observed proportion in each age class to reduce the influence on the likelihood of observations from extremely weak cohorts that would be difficult to detect due to measurement errors. We assume in years and ages where the observed proportion is less than 2%, the consecutive ages are grouped into a single age-class which reduces the effective number of age-classes (this is some what analogous to a plus group). For example, assume the age-proportions in 1970 of age-2 and age-3 fish are 3.5% and 0.1%, respectively. If the user specifies a minimum proportion of 2% then  $iSCAM$  will treat these observations as 3.6% age 2-3 fish in 1970. Pooling the data this way has been shown to reduce the influence of measurement errors on weak cohorts ([Richards et al., 1997](#)).

### Commercial catch data

In the HCAM assessment, commercial catch was assumed to be known with a high degree of certainty; observation errors were assumed lognormal, and the standard deviation specified in the code is fixed at 0.0707 (variance of 0.005) for all three gear types. To implement these assumptions in  $iSCAM$  we fix the assumed standard deviation for the catches in the last phase to 0.0707 for each of the gear-types for all years. In other words, we assume that errors in estimating the catch by gear-type is constant over the 1951-2011 time period.

### Spawn survey data

For the Strait of Georgia spawn survey, the assumed standard deviations in HCAM were specified at 0.35 and 0.3 for the pre- and post-1988 periods. To carry out the same assumptions in  $iSCAM$  the relative weights for the pre and post 1988 survey data were fixed at 1.0 and 1.1666, respectively. In  $iSCAM$  the total error (or precision=1/(std. dev.)) is estimated and partitioned into components of observation error (spawn survey residuals) and process error (recruitment deviations). To implement the same observation error and process errors in  $iSCAM$  (standard deviations of 0.35 and 0.8, respectively for observation errors and process errors in HCAM) the total precision was fixed at  $\vartheta = 1/1.15$ , and the proportion assigned to observation error was fixed at  $\rho = 0.35/1.15$ . The total variance is calculated as  $(0.35^2 + 0.8^2)^{-1}$ .

### Specification of prior distributions

Starting with the prior density for natural mortality in HCAM, the average natural mortality rate is assumed to be normal with a mean of 0.45, and a standard deviation of 0.2 (see Table 3 in [Cleary and Schweigert, 2010](#)). The average natural mortality rate in  $iSCAM$  is estimated in the log scale; using a normal prior for the  $\ln(M)$  is equivalent to a lognormal prior for  $M$ . A lognormal prior is appropriate for this parameter as natural mortality rates must be positive; however, there is no equivalent analytical transformation to the normal distribution that was used in the HCAM assessment. Here we have specified a normal prior for  $\ln(M)$  with a log mean of  $\ln(0.45) = -0.7985$  and a log standard deviation of 0.4 to approximate the variance specified in the normal distribution used in the previous HCAM assessment. Note that the use of a lognormal prior for natural mortality rates will result in a slight downward bias in the maximum likelihood estimate of natural mortality rates. Based on the prior distribution described above, the mode of this distribution is approximately equivalent to an  $M = 0.38$ .

The base HCAM model also allows for a random walk in natural mortality rate implemented as:

$$M_t = \begin{cases} \psi, & t = t' \\ M_{t-1} \exp(d_t^M), & t > t' \end{cases}$$

where  $d_t^M$  are annual natural mortality deviations that are assumed to be normally distributed with a mean 0 and a standard deviation of 0.10, and  $\psi$  is an estimated initial value for natural mortality. The implemen-

tation of time varying natural mortality is similar in  $iSCA_M$  in that it is a random walk process, but the components of the objective function include a prior for the initial value of  $M$  (as specified in the previous paragraph) and that the first differences between natural mortality deviations are normally distributed. Again, this structure allows natural mortality rates to drift away from central tendency (i.e., a biased random walk) and long-term changes in  $M$  could have profound effects on reference point calculations. HCAM assumed an unbiased random walk and the central tendency of  $d_t$  is 0; whereas,  $iSCA_M$  assumes a biased random walk and the first differences in annual natural mortality deviations were assumed to have a mean 0 and a standard deviation of 0.10.

Annual recruitment deviations in the HCAM implementation were assumed to be normally distributed on a log scale with a mean of zero and a standard deviation of 0.8. To set up an equivalent assumption in  $iSCA_M$ , the total variance ( $\vartheta$ ) and ratio of the total variance ( $\rho$ ) that explains observation error in the spawn survey must be specified *a priori*. In the HCAM model the variance terms for the observation errors and process errors are not estimated and assumed to be known; the standard deviation for recruitment variation was set at 0.8 and the standard deviation for observation errors in the spawn survey was fixed at 0.35 and 0.3 for the pre and post 1988 data, respectively. These variance terms can be estimated within the  $iSCA_M$  model, or treated as fixed constants; however,  $iSCA_M$  estimates the total error and partitions it into observation ( $\sigma$ ) and process error ( $\tau$ ) components. To make the same assumptions about the error terms in  $iSCA_M$  as those that were used in HCAM the following values were used  $\vartheta = 1/1.15 = .8695$ , and  $\rho = 0.3043$ , and the weights assigned to the post 1988 spawn data were set at 1.1666 and 1.0 for the pre 1988 spawn data.

The prior for steepness in HCAM was based on a lognormal distribution with a log mean of 0.67 and a standard deviation of 0.17. For the Beverton-Holt stock recruitment model, steepness must lie in the interval of  $0.2 < h \leq 1.0$ ; a Beta distribution is an appropriate density function for this parameter. In the  $iSCA_M$  implementation, a Beta distribution is used and to approximate the distribution used in the HCAM model, the shape and rate parameters specified are 10.0 and 4.925373, respectively. These values corresponds to a mean of 0.67 and a standard deviation of 0.1178 for the Beta prior.

The last informative prior that is not explicit in the table of priors in the HCAM model is the scaling parameter ( $q$ ) for the spawn survey. The spawn survey data are broken into two separate time series, pre- and post-1988 when the survey switched from a surface estimate to dive surveys for estimating total egg deposition. In the HCAM implementation, a very informative prior for  $q$  in the post 1988 period was used where the mean was fixed at  $q = 1.0$  and not permitted to vary (i.e.,  $\sigma_q = 0$ ). The scaling parameter in the first period was then freely estimated using an uninformative prior. Again, to emulate these assumptions in  $iSCA_M$  a normal prior for  $\ln(q)$  with a mean =0 and a standard deviation of 0.001 was used for the post-1988 data and a uninformative prior for the pre-1988 data.

## 1.3 Results

### 1.3.1 Simulation testing

#### Model validation against perfect information

Given perfect information about trends in relative abundance and age composition information, and a deterministic stock-recruitment relationship,  $iSCA_M$  was able to estimate 216 parameters without any substantial errors. Estimation performance is easily demonstrated by comparing estimates of spawning biomass and fishing mortality rates to the true values that were used to simulate the data. As shown in Figure 1.2 estimates of spawning biomass and fishing mortality rates were exactly the same as the true values. There is no measurable difference between the observed and predicted trends in the relative abundance information (Figs. 1.2cd).

Residuals between the observed and predicted age-composition data (not shown) were also extremely small and are easily summarized by the conditional maximum likelihood estimates of the residual variance  $\hat{\tau}^2$  for the winter seine fishery  $\hat{\tau}^2 = 2.49e - 03$ , the seine roe fishery  $\hat{\tau}^2 = 1.25e - 27$ , and the gillnet fishery  $\hat{\tau}^2 = 1.24e - 27$ .

This perfect fit to the data is used only to judge if the code is syntactically correct and to determine if it

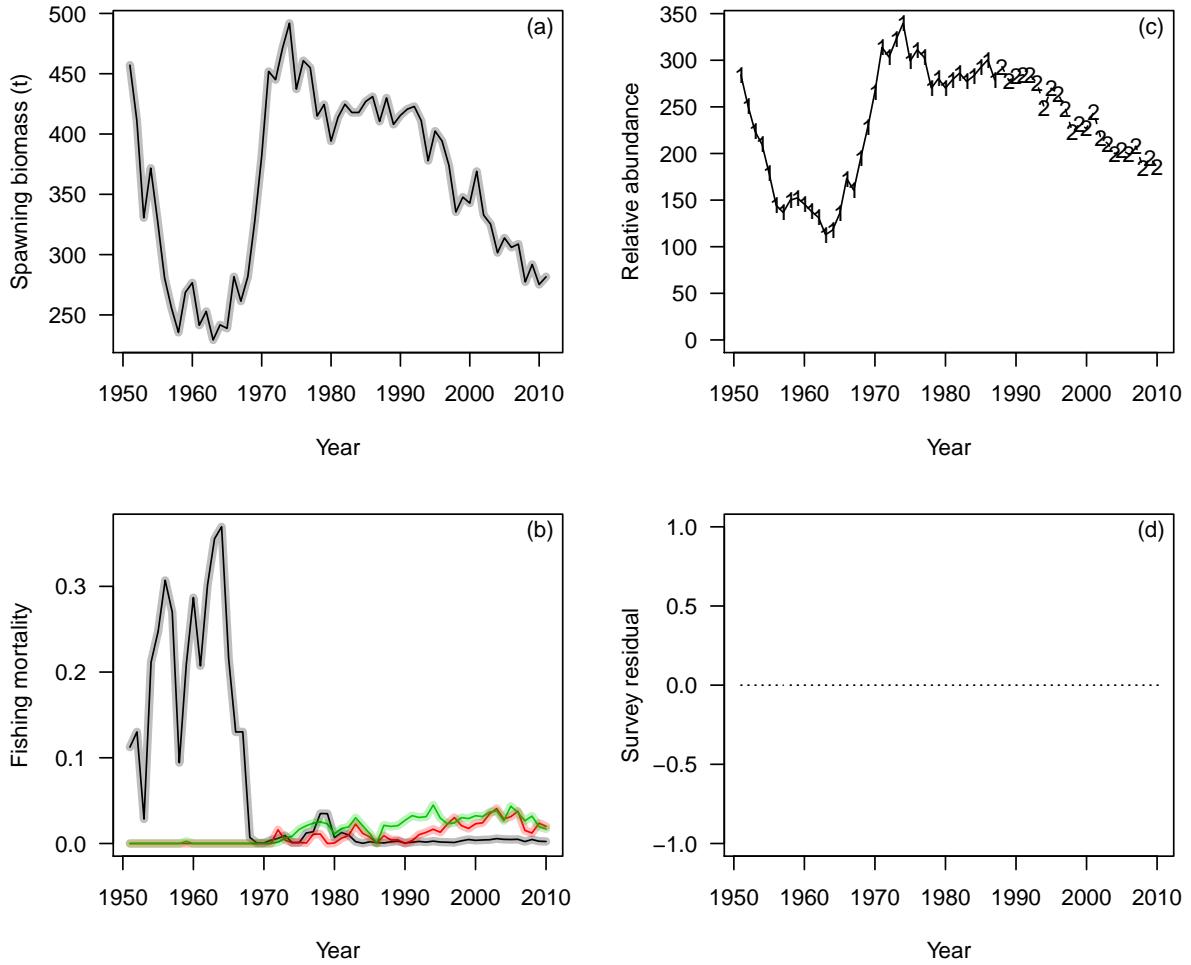


Figure 1.2: True (thin line) and estimated (thick shaded line) spawning biomass (a), fishing mortality rates by gear (b), observed and predicted relative abundance (c), and residuals between observed and predicted relative abundance (d) for the SOG herring simulation with perfect information and a deterministic stock-recruitment relationship.

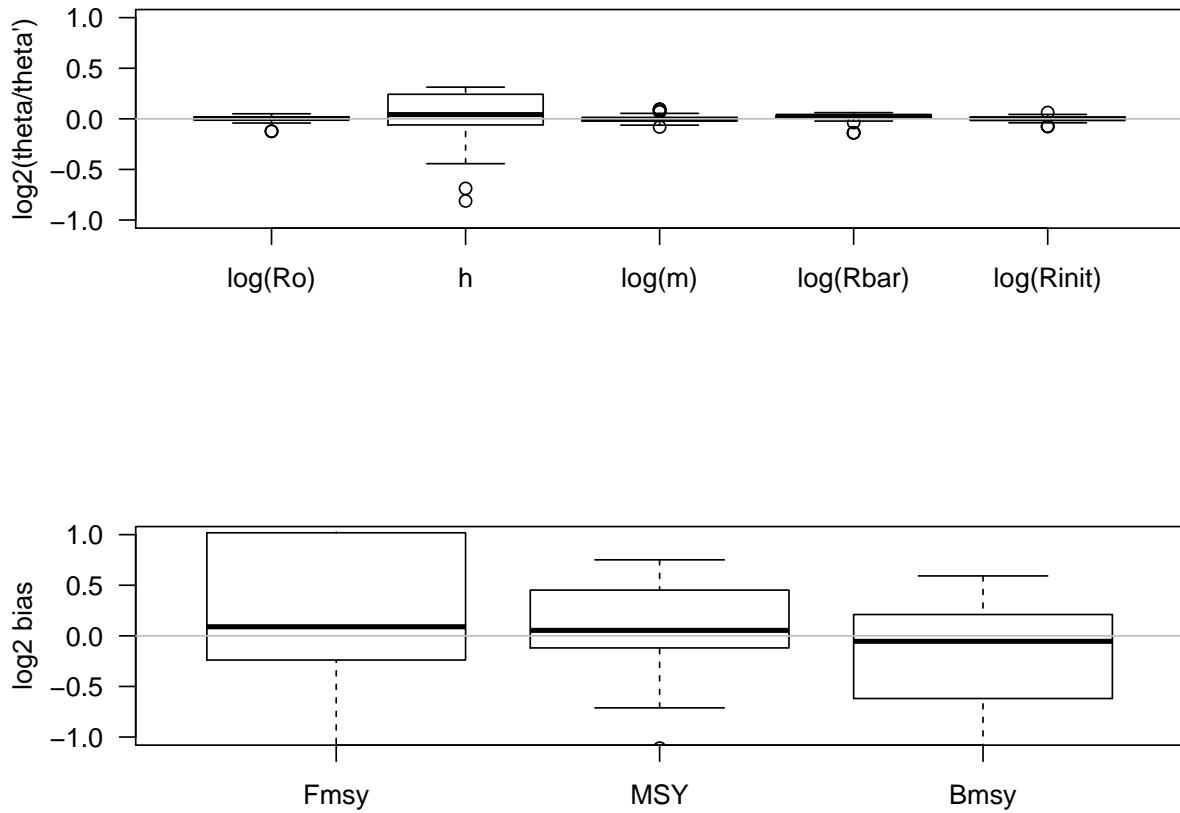


Figure 1.3: Estimates of precision and bias in key model parameters and MSY based reference points for 50 simulated data sets conditioned on the Strait of Georgia herring catch. The  $\log_2$  ratio of estimated (numerator) versus true (denominator) value is plotted; values of 1 and -1 correspond to a twice or half the true value, respectively.

is capable of estimating model parameters exactly given perfect information. In the following section, observation errors and process errors are introduced to determine bias and precision in parameter estimates.

#### Bias & precision with observation & process errors

In the simulation experiments conducted with both observation error and process error included in the simulated data, there was no appreciable bias in the estimates of unfished recruitment ( $\ln(R_o)$ ) and the natural mortality rate ( $\ln(M)$ ), the average recruitment ( $\ln(\bar{R})$ ) and the initial recruitment ( $\ln(R_{\text{init}})$ ), as shown in Figure 1.3. There was, however, a very slight upward bias in the estimate of the steepness parameter for the Beverton-Holt stock recruitment relationship. Also, steepness was estimated with the least amount of precision. The estimability of steepness depends on many factors including the precision of the observations, but also the history of exploitation, the true value of steepness and the natural mortality rate (Conn et al., 2010).

Its not surprising to see a slight upward bias in  $h$  for these simulations because the true value of natural

mortality was set quite high ( $M = 0.45$ ) along with steepness ( $h = 0.8$ ). These high values of  $M$  and  $h$  imply a very productive stock, and the spawning biomass would have to be driven to very low levels in order to generate data that would be informative about the underlying production function. Note that  $M$  and  $h$  are confounded because  $M$  is required to calculate the spawning stock biomass per recruit in unfished conditions. Furthermore, the simulated exploitation history involved a strong depletion signal between the 1950s and 1960's followed by very light exploitation from 1970 onward. On average the assumed parameter values for the simulation and the catch time series generated sufficient contrast to reliably estimate key model parameters without the use of informative priors.

The lower panel of Figure 1.3 shows the apparent precision and bias in the estimates of MSY based reference points. There is a very slight upward bias in the estimate of  $F_{MSY}$  associated with the slight upward bias in steepness. Estimates of  $B_{MSY}$  are also slightly biased in a downward direction, and MSY in a slight upward direction. Overall, MSY is the most precisely estimated and  $F_{MSY}$  is the least precisely estimated management variable.

It should also be noted here that this simulation is somewhat unrealistic in comparison to applying the model to the real data. The simulated conditions have a much lower total variance in comparison to the real assessments, we assume the ratio of observation error to process error is precisely known, and  $M$  and selectivity are constant. Whereas, in the real assessments the total error is estimated and partitioned into observation error and process error using an informative prior,  $M$  and selectivity in the gill net fisheries are not constant and allowed to vary over time. These assumptions should actually be simulation tested in future work to better characterize the nature of the data and model assumptions.

### 1.3.2 Comparison of HCAM with $iSCAM$

Based on the description of the priors and model setup in Section 1.2.4 on page 5, a comparison of the spawning stock biomass between  $iSCAM$  and HCAM were very similar (Figure 1.4a) for the Strait of Georgia stock. Between 1951 and 1969 the absolute difference in spawning biomass is minimal and post 1970 estimates of spawning biomass are slightly higher for the HCAM model. The only real difference between the two models during this period is a difference in the assumptions about the error structure for the age-composition data.

Estimates of spawning depletion are based on the post fishery spawning biomass relative to the estimated unfished spawning biomass (Figure 1.4b). The three coloured zones demarcate the critical zone, cautious zone, and healthy zones with transitions defined by  $0.4B_{MSY}$  and  $0.8B_{MSY}$ , respectively.

Maximum likelihood estimates of key model parameters for both  $iSCAM$  and HCAM are summarized in Table 1.1. The  $iSCAM$  model has a lower estimate of  $B_o$  in comparison to the HCAM model and a higher value of steepness ( $h$ ). These two parameters are usually negatively correlated and it is expected that if  $B_o$  was higher in one model in comparison to the other, then  $h$  would normally be lower to compensate. Information to estimate  $B_o$  and  $h$  come from the apparent stock-recruitment data and the structural form of the stock recruitment relationship. Both models assume that stock-recruitment is in the form of Beverton-Holt, only the prior for steepness differs between the two models.

Table 1.1: A comparison of key parameters from  $iSCAM$  and the HCAM model with the  $iSCAM$  model set up like HCAM.

Parameter	$iSCAM$	HCAM
Unfished spawning biomass ( $B_o$ 1000 t)	108.492	190.817
Steepness ( $h$ )	0.811	0.683
Average natural mortality rate	0.563	0.334
Survey $q$ for period 1	0.985	1.1105

Average natural mortality rate is higher in the  $iSCAM$  model (Table 1.1), and the survey  $q$  for the pre-1988 spawn survey data is nearly identical in the two models (Figure 1.4f). The more contemporary survey data were forced to scale with  $q = 1.0$ . There is some pattern in the residuals for the overall fits to the

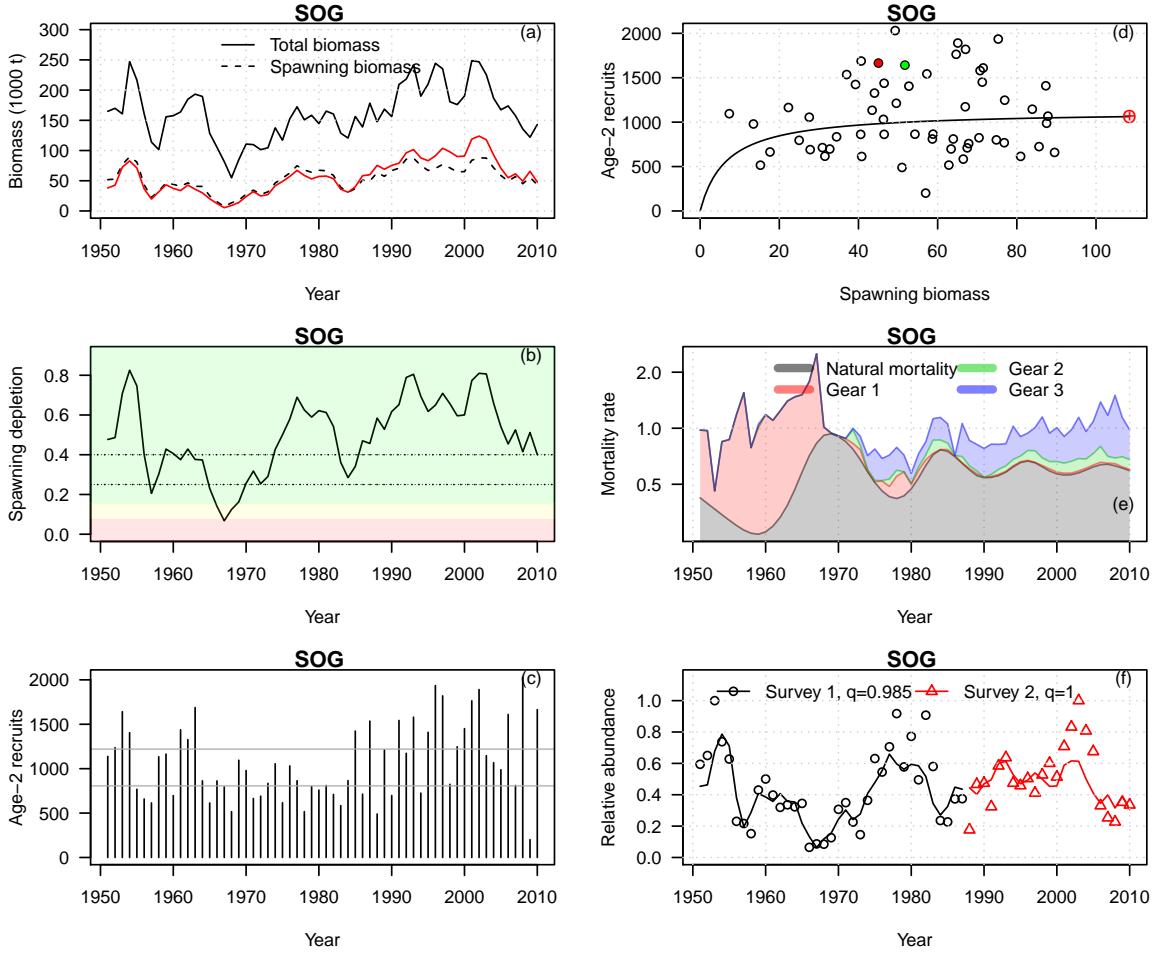


Figure 1.4: Results based on trying to configure the  $i\text{SCAM}$  model as similar as possible to the previous HCAM assessment. Maximum likelihood estimates of pre-fishery biomass (defined as the numbers-at-age times the mean weight-at-age at the start of the year) and post fishery spawning biomass in the Strait of Georgia (a), spawning biomass depletion (b), age-2 recruits (c), stock-recruitment relationship and unfished reference points (d), components of total mortality (log-scale, panel e), and observed (points) and predicted (lines) spawn survey data (f). The red line in panel (a) is the MLE estimate of spawning biomass from HCAM.

survey data (Figure 1.4f), the model fails to predict the large increases in abundance in the late 1970s and early 2000s and the recent sharp decline in the mid 2000s. The assumed standard deviation for the survey errors was 0.35 and 0.3 for the pre and post 1988 survey data, the standard deviation of the residual errors in Figure 1.4f is 0.335 and 0.334, respectively.

Estimates of the components of total mortality for the comparison with the HCAM model are shown in Figure 1.4e. The fishing mortality rates for each gear represent the average fishing mortality rate over all age-classes, and the natural mortality rate is assumed to be age-independent. During the 1950s through to 1968, fishing mortality rates for Pacific herring in the Strait of Georgia were extremely high; this period was almost exclusively a winter purse-seine fishery where fish were taken for fishmeal (the reduction fishery). After the fishery reopened in the early 1970s fishing mortality rates were greatly reduced and targeted the spring spawning aggregations as the market was for herring roe.

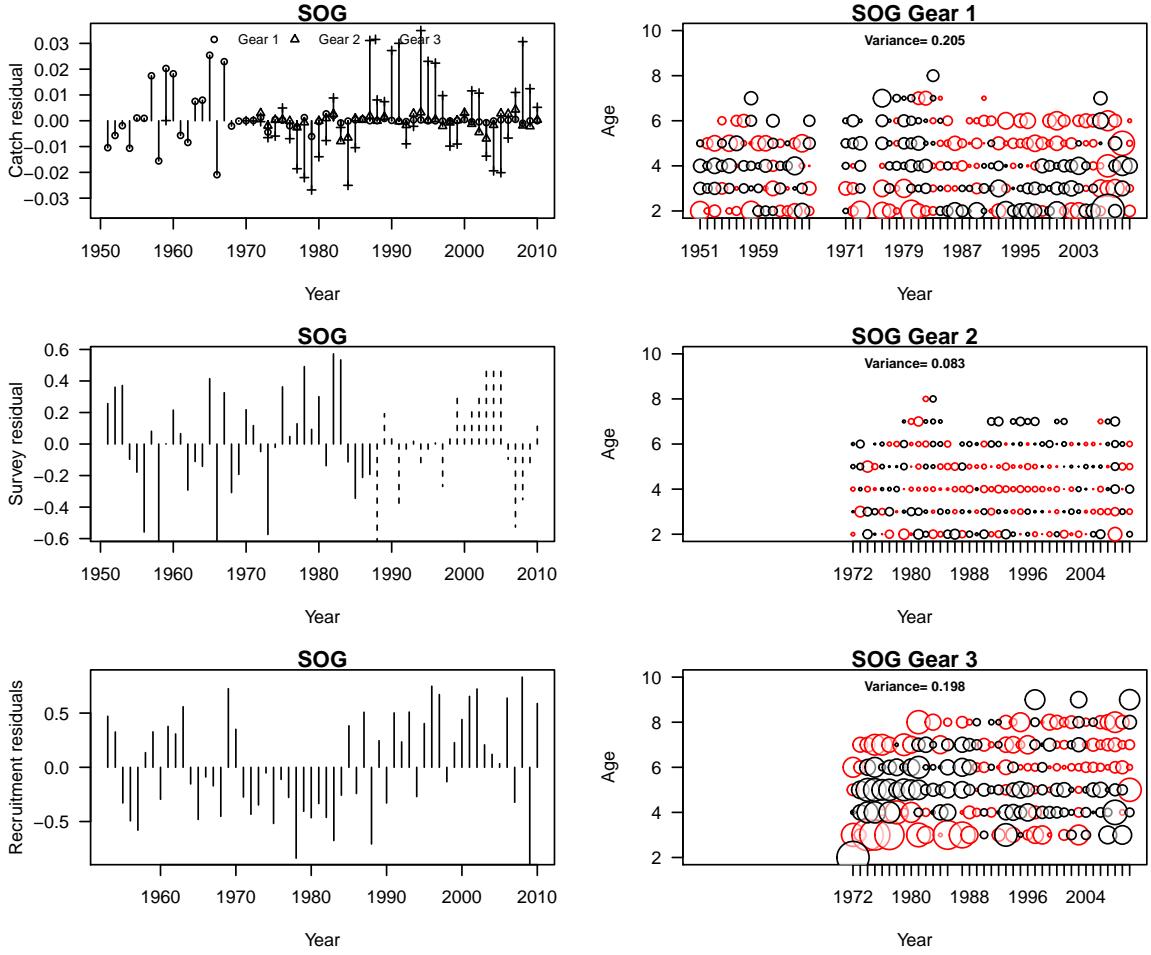


Figure 1.5: Results are based on configuring  $i\text{SCA}_M$  to be similar to HCAM. From top to bottom in the left column: the log residuals between observed and predicted catch for each gear, the residuals between the observed and predicted spawn survey index, and the annual deviations between age-2 recruitment and that predicted by the Beverton-Holt model and the estimated spawning stock biomass. Right column: residual patterns in the age-composition data (observed - predicted, where black is a positive residual) for each of the three commercial gears in the Strait of Georgia.

Estimates of natural mortality are based on a random walk process, initially starting at a value of 0.401 in 1951 and declining to a very low value of 0.217 in 1959, then increasing to a maximum of 0.915 in 1969 (Figure 1.4e). Information to estimate natural mortality rates comes from the age-composition data, and assumptions about selectivity in the fishery. In this comparison, the  $i\text{SCA}_M$  model assumes selectivity is invariant for the purse seine gears and is a function of weight-at-age for the gillnet gear. Much of the residual variation in the age-composition is explained by variation in  $M$  and variation in age-2 recruits (see Figures 1.4c & d for age-2 recruitment and stock recruitment relationship). The HCAM model has a very similar trend in the estimates of natural mortality but the variability is much less than that of the  $i\text{SCA}_M$  assessment. This is almost certainly due to the differences in the assumptions about the error structure in the age-composition data (or relative weights associated with the age-composition).

There is good correspondence between the observed and predicted catch, however, the residual patterns do not appear to be iid for each of the fleets (Fig. 1.5). Recall that the fit to the catch data is largely

determined by an assumed variance for the observation errors in the reported catch ( $\sigma_C^2 = 0.005$ ).

The fit to the spawn survey data in Strait of Georgia is nearly iid for the 1951:1987 time period. There is some pattern in the residuals post 1988 that appears to be in contradiction with other information (i.e., age-composition data and structural assumptions about selectivity) (Fig. 1.5). The pattern in the recruitment residuals for the Strait of Georgia suggest periods of below average-recruitment in the 1970s and early 1980s and above average recruitment starting in the early 1990s.

Fits to the age-composition data for the purse seine-roe fishery were best in comparison to the winter seine and gillnet fisheries. The conditional maximum likelihood estimates of the variance of the age-composition data are 0.205, 0.083, and 0.198 for the winter purse seine, seine-roe, and gillnet fisheries, respectively. The smaller the standard deviation, the better correspondence between the observed and predicted age-composition data. Also the pattern of residuals does indicate some model mis-specification (e.g., the gillnet fishery in the Strait of Georgia).

### 1.3.3 Alternative assumptions about catchability, mortality & selectivity

Here we briefly explore the differences between relaxing the informative prior on catchability for the survey, reducing the number of natural mortality parameters being estimated and exploring alternative selectivity options to try and reduce the residual pattern in the gillnet fishery age-composition data. For the comparisons, we only examine the maximum likelihood fits to the data and the overall objective function value. Table 1.2 presents a few summary statistics for the following sub sections for easier comparison.

Table 1.2: Summary statistics for alternative structural assumptions about the Strait of Georgia herring assessment from 1951 to 2010. Definitions are No. is the number of estimated parameters,  $f$  the objective function value,  $B_o$  unfished spawning biomass,  $h$  steepness,  $M$  is the average natural mortality rate,  $B_{MSY}$  the biomass at maximum sustainable yield,  $q$  survey scalers for pre and post 1988.

Model	No.	$f$	$B_o$	$h$	$M$	$B_{MSY}$	$q_1$	$q_2$
Fixed $q$	279	-1166.62	108.49	0.812	0.56	20.962	0.985	1.0
Prior $q$	279	-1161.83	110.57	0.798	0.59	21.964	0.912	0.912
Fixed $M$	219	-1055.48	128.06	0.687	0.67	25.727	0.688	0.796
Gillnet Selectivity	279	-1196.06	121.65	0.76	0.655	25.27	0.817	0.683

#### Impacts of informative priors on $q$ 's

The implication of relaxing the informative prior on  $q$  for the contemporary spawn survey data was examined by using a less informative prior for the spawn survey  $q$ . The results shown in Fig. 1.6 is a comparison of the HCAM parameterized version of the model as shown in previous sections with a version where the prior on  $\ln(q)$  was assumed normal ( $\mu = 0, \sigma = 0.274$ ) for both the contemporary and surface survey data.

The net result of relaxing this prior on  $q$  is a slight increase in the global scaling (the spawn biomass increases by roughly 12%) in comparison to the fixed  $q = 1$  scenario. There is no appreciable difference in the overall fits to the data (see objective function value in Table 1.2).

#### Implications of variable natural mortality rate $M_t$

In this next scenario we use the less informative prior for  $\ln(q)$  as in the previous section but do not allow natural mortality rates to vary over time (i.e., fixed  $M$ ). The natural mortality rate and  $q$  are confounded so it does not make sense to fix  $q$  and estimate  $M$  because the corresponding estimate of  $M$  would simply be conditional on the assumed value of  $q$ .

In the case where  $M$  is assumed to be time invariant, estimates of average  $M$  over the entire time series does increase slightly as well as the overall scaling of population size (Table 1.2). There are 60 fewer estimated parameters in this case but the overall fit to the data is slightly degraded (Fig 1.7, and see objective function values in Table 1.2).

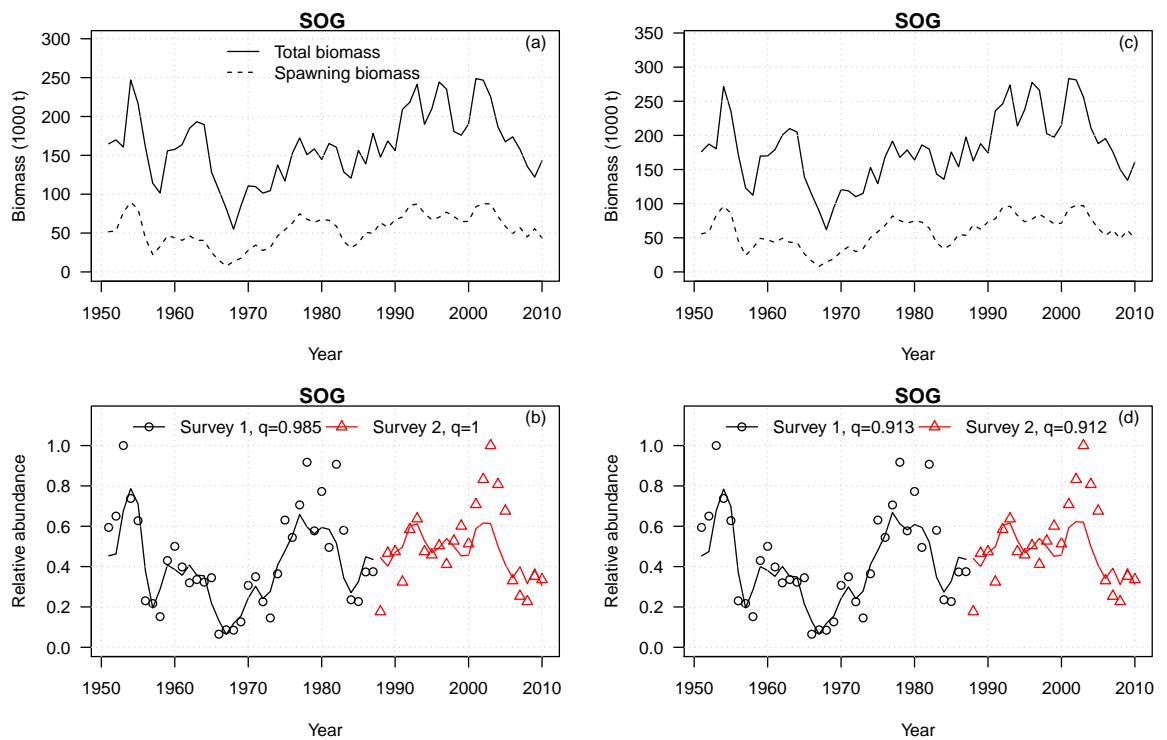


Figure 1.6: A comparison of the estimated biomass and spawning biomass and fits to the survey data when  $q$  is either fixed at 1 for Survey 2 (panels a, b), or estimated using an informative prior with an expected mean of 0 and a log standard deviation of 0.274 (panels c and d).

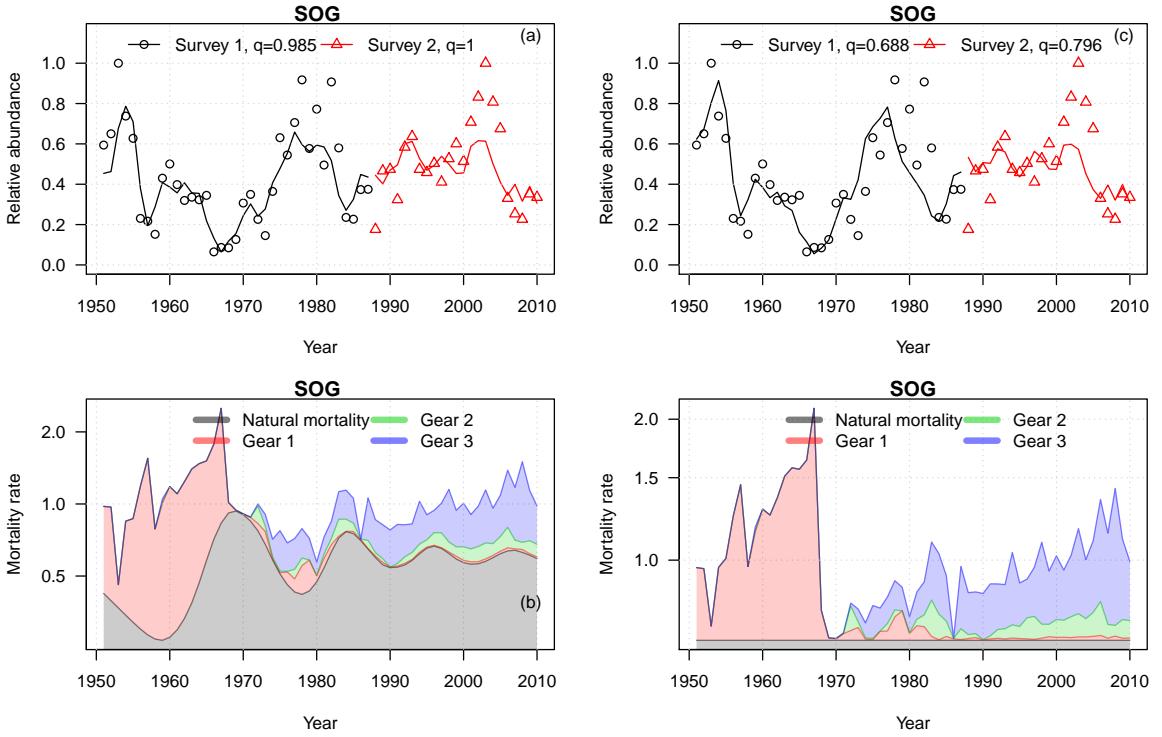


Figure 1.7: A comparison of fits to the survey data and estimated components of average mortality by year when  $M$  is allowed to vary via a random walk process or is estimated and assumed time invariant. Note that the y-axis in panels (b) and (d) are on a log scale and that the estimate of  $M$  is 0.67 in panel (d).

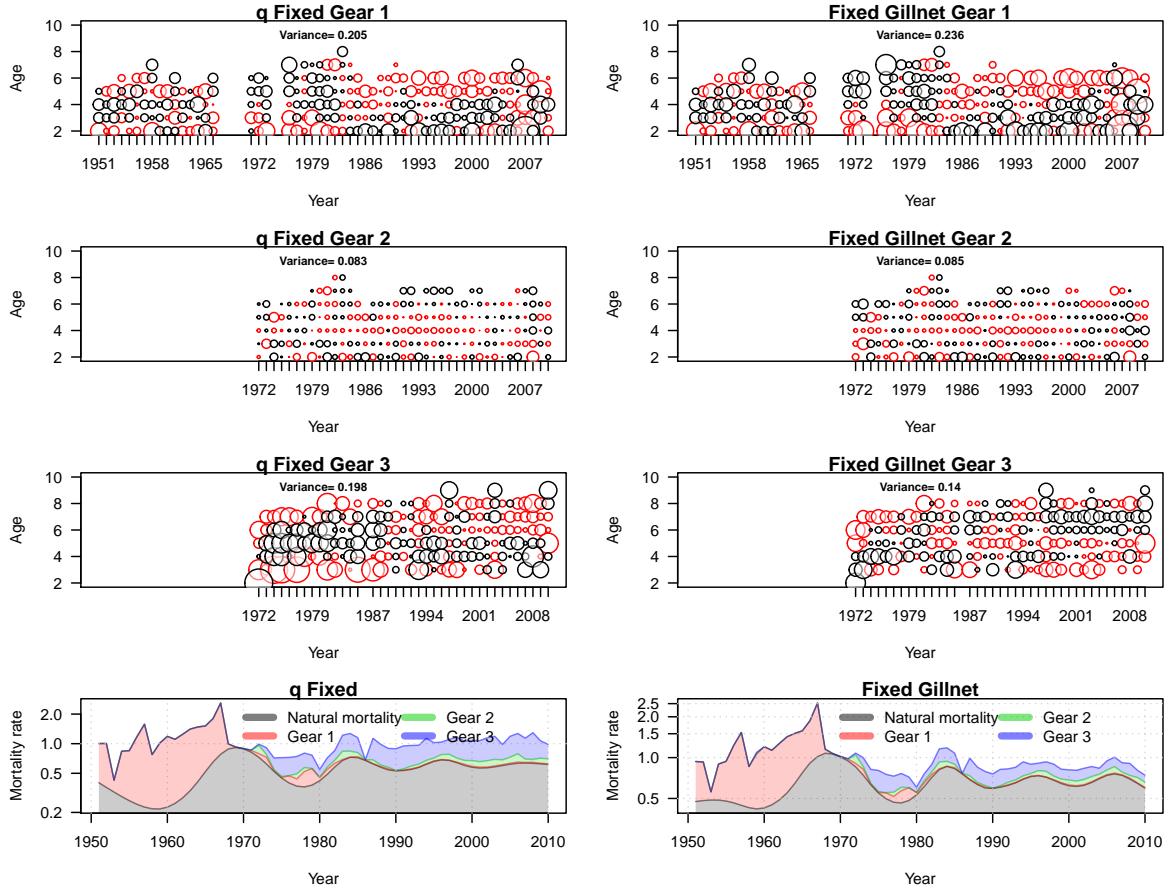


Figure 1.8: Residuals in the age-composition data when gillnet selectivity is a function of mean weight-at-age (left column with  $q$  Fixed in each caption) and or is a logistic function of age and time invariant (right column with Fixed Gillnet in each caption).

#### Implications of variable selectivity in directed fisheries

Finally, we also explored the option of treating the gillnet selectivity as time invariant and estimated two parameters that describe age-specific selectivity using a logistic curve. In this case we also allowed for a random walk process in natural mortality rates so the total number of estimated parameters remains the same.

Better fits to the gillnet fishery age-composition data were obtained with constant selectivity, and the residual pattern also appears to improve under the assumption of constant selectivity (Fig. 1.8). There was also a slight degradation to the age composition data in the winter purse seine fishery (Gear 1), but residual patterns were nearly identical in both the seine fisheries (Fig. 1.8). The same number of model parameters were estimated and there was a slight improvement in the overall objective function with constant selectivity (Table 1.2). Trends in the estimates of natural mortality rates are similar when the gillnet selectivity is assumed constant.

#### 1.3.4 Preliminary assessments for all other areas

For the five major stock assessment regions there was very good correspondence between the estimated spawning stock biomass between the HCAM and new  $i\text{SCA}_M$  models (Fig. 1.9). The control file used for

each of the assessment regions was the same as that used for the Strait of Georgia. No additional changes were required (i.e., each SAR had the same initial starting parameter values etc.). In some areas there are minor differences in the trends and overall scaling (e.g., SOG, PRD), these were also identified as cases that had persistent negative residuals in the HCAM model.

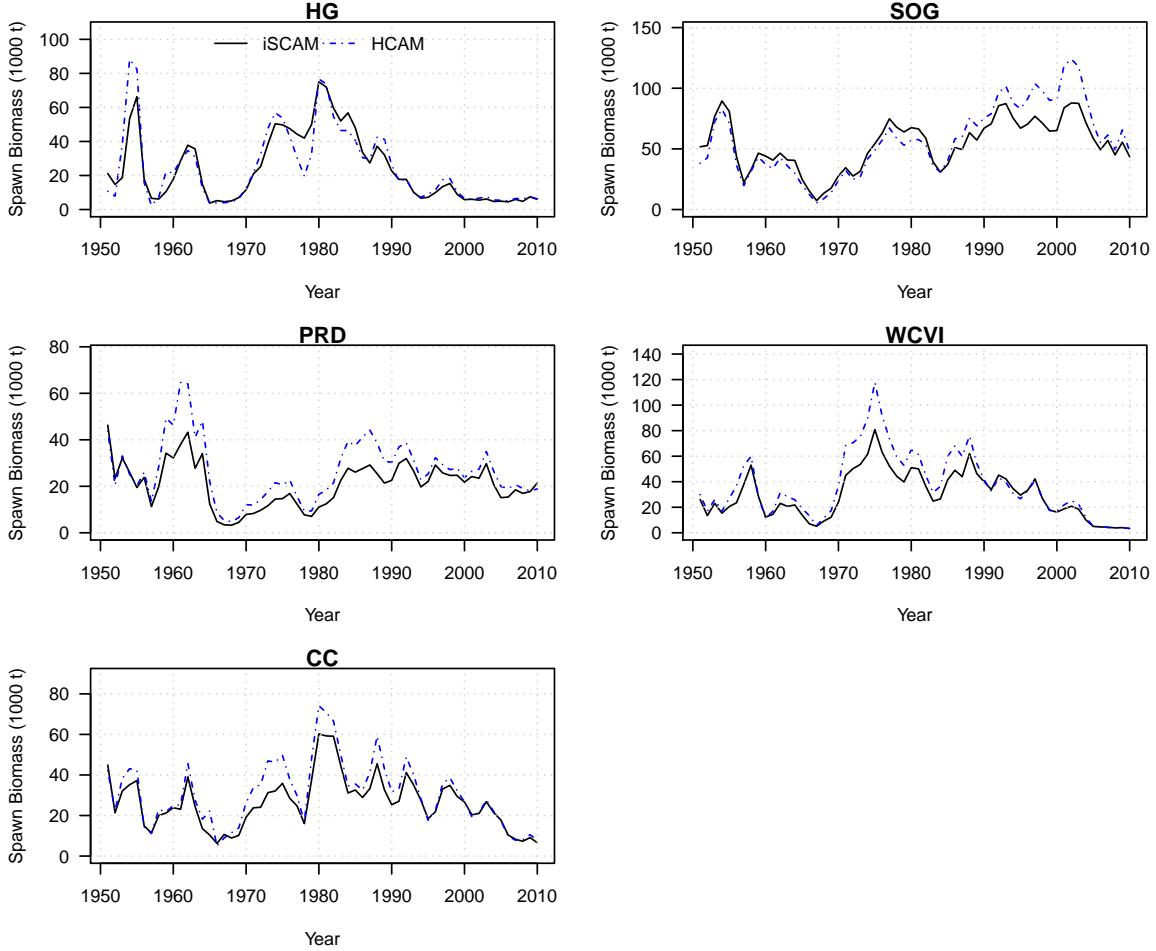


Figure 1.9: A comparison of estimated spawning stock biomass between HCAM and  $i\text{SCA}_M$  for the five major stock assessment regions using data from 1951 to 2010 and setting up  $i\text{SCA}_M$  similar to HCAM.

## 1.4 Discussion

The simulation studies showed a minor upward bias in the estimates of  $F_{MSY}$ , MSY and a slight bias in steepness. Unfortunately, the true parameter values used in the simulation study did not provide significant contrast in the data to generate information over a wide range of spawning depletion. In general, such widely contrasting data are required to resolve parameter confounding (Hilborn and Walters, 1992). The simulated fishing mortality rates were extremely low (i.e., less than 0.3 between 1951 and 1969, and less than 0.1 post 1970.) and the stock was never depressed to a level where, on average, recruitment would be limited by spawner abundance. Future simulation studies should impose a more severe mortality schedule to see if this resolves the slight parameter bias.

There are a few significant differences between  $iSCAM$  and the previous assessment using HCAM, despite efforts to parametrize  $iSCAM$  as close as possible to HCAM. These differences include: different likelihood function for the age-composition data, using the conditional maximum likelihood estimate of  $q$ , estimation of variance, and jointly estimating parameters for the stock recruitment relationship. Here we briefly discuss how these structural differences influence model results.

In our attempts to parametrize the  $iSCAM$  model to reproduce the results of last years HCAM assessment, we fixed the variance and variance partitioning parameters such that the same weights were placed on the spawn observation errors and variation in annual recruitment. The age-composition data, however, were not given the same weight as the multinomial samples sizes that were used in the HCAM model. In this case, the  $iSCAM$  model assumes a homogenous variance in the age-composition data and the weight assigned to each observation is based on the conditional maximum likelihood estimate of the variance (i.e., this is akin to using concentrated likelihoods for the age-composition data). The main affect of this difference in likelihood formulations is that in the  $iSCAM$  model each year is given the same weight regardless of sample size (i.e., small and large sample sizes are given too much or too little weight, respectively, in comparison to the multinomial likelihood used in HCAM). That being said, there is also much debate about what the effective sample size should be when using the multinomial likelihood due to correlation structure in the samples (e.g., Francis, 2011). Another subtle difference between the two approaches is the pooling of samples that are less than 2% of the age-composition into the adjacent year class.

A second major difference between the two modelling platforms is the how  $q$  for the dive survey data is treated. In the previous HCAM model, the dive survey  $q$  was fixed at 1, and was not treated as a latent variable. As a result the residuals between the observed and predicted survey data did not have a mean of 0 due to contradictions between the spawn survey data and the age-composition data in the SOG assessment. The residuals for the spawn survey data post 1988 were nearly all negative. In the  $iSCAM$  model, the conditional maximum likelihood estimate of  $q$  was used to scale the spawn survey data along with a very informative prior for  $q$  was used to force  $q = 1$  for the post 1988 spawn survey. In  $iSCAM$  model the sum of the residuals between the observed and predicted spawn survey data did have a mean equal to zero and there is no apparent bias in scaling parameters ( $q$ 's) associated with contradictions in data sources.

In the comparison between HCAM and  $iSCAM$  models, the variance parameters in the  $iSCAM$  model were fixed such that the same observation error and process error were used in both models. The fixed variance and fixed  $q$  values could also partially explain the persistent negative residual patterns observed in the HCAM model; these data may be weighted less relative to the age-composition data. Moreover, fixing these variance parameters (specifically the ratio of observation to process errors) is also likely to bias estimates of uncertainty in spawning biomass, and other variables. We also explored the possibility of estimating the total variance and the fraction associated with observation error for the SOG model. In this case, reasonable estimates of the variance parameters were obtained and this is only possible because there is information on relative abundance and age-composition in this fishery. We further estimate all the variance parameters in Part II of this document for all major and minor stock areas. Estimates of uncertainty are likely to be less biased now that the variance parameters are jointly estimated.

Lastly, another less subtle difference between the two modelling approaches was the estimation of the stock recruitment parameters. In the HCAM model an improper prior was used for the steepness parameter, where steepness was bound between 0.2-0.99 and a lognormal prior was used. In  $iSCAM$ , we used a Beta distribution as the prior rescaled for values in the range of 0.2-1.0 for the Beverton-Holt model. Estimating the stock-recruitment parameters is critical in establishing reference points and moving towards the Sustainable Fisheries Framework (SFF). If there are no information in the data to reliably estimate steepness,

then the form of the prior distribution will influence derived reference points.

The default reference points suggested in the SFF are based on the concept of Maximum Sustainable Yield (e.g., the suggested LRP is  $0.4B_{MSY}$ , and the USR is  $0.8B_{MSY}$ ). Such a default requires that we can accurately calculate  $B_{MSY}$ . In order to calculate  $B_{MSY}$ , precise estimates of population parameters ( $B_0, h$ ), life-history parameters ( $M, L_\infty, k, t_o, a, b, \hat{a}, \hat{\gamma}$ ), and selectivity parameters ( $\hat{a}, \hat{\gamma}$ ) are required. See parameters defined in Table A-1 for a full explanation of how these are related to  $B_{MSY}$ . Estimating these parameters is one of the major goals of a fisheries stock assessment model, but calculating reference points based on the concept of MSY has a potential problem when there are multiple fleets involved that have very different selectivity curves. That problem relates to allocation of yield to each of the fleets. If each fleet has the same selectivity curve, then each fleet imposes the same age-specific mortality and there will be no impacts on the MSY estimates. If however one or more fleets harvest fish at a much younger (or older) age, then estimates of MSY (and the corresponding removal rate reference point) will shift up or down depending on the ratio of maturity to vulnerability. In general, the later the fish recruit to the fishing gear, the higher the sustainable fishing mortality rate.

Calculating reference points for the SFF is more about determining allocation of herring to the purse seine and gillnet fleets. The gillnet gear tends to catch older herring in comparison to the winter seine and seine roe fisheries. Larger allocations to the seine fisheries come at a tradeoff of reducing potential yield for the gillnet fishery. There is a real optimization problem here that goes just beyond the biology of the species, namely, profitability of each fishery. Furthermore, the calculation of reference points assumes that these data come from a unit-stock and not a collection of mixed sub-stocks that may differ in productivity.

## **Part II**

# **Stock Assessment and Management Advice for the British Columbia Pacific Herring Stocks: 2011 Assessment and 2012 Forecasts**

cutoff

## 2.5 Introduction

The objectives of this section of the report are: (1) present the data used in the 2011 assessment, (2) provide a summary overview of the integrated statistical catch-age model (hereafter,  $i\text{SCA}_M$ ), (3) present the 2011 stock assessment and forecast for 2012, and (4) describe in detail the decision table used to provide advice to fisheries management.

BC herring are currently managed as five major stocks and 2 minor stocks (Figure 1.1). Annual catch advice for each of these areas is based on current estimates of stock status, and a 20% exploitation rate if the post-fishery stock is above the cutoff level for the five major stocks and a 10% exploitation rate for the two minor stocks. Cutoff levels for the five major stocks historically were based on the 1996 estimate of  $0.25B_0$ . These cutoffs are currently thought to be more conservative than the suggested default Limit Reference Point of  $0.4B_{MSY}$  ([Fisheries and Oceans Canada, 2006](#)). For example,  $B_{MSY}$  is normally in the range of 35% of the unfished biomass for many fish stocks; therefore, 40% of  $B_{MSY}$  is roughly 14% of unfished which is significantly lower than the 25% $B_0$  that is currently used for Pacific herring. Alternative cutoffs based on updated estimates of  $B_0$  are also provided in this document.

This years assessment is based on a new model,  $i\text{SCA}_M$ , where alternative assumptions about survey  $q$ , and the form of the error distribution for the age-composition data are the major differences in comparison to the 2010 assessment using HCAM. In addition to the changes in likelihoods, we also present an alternative parametrization of the gillnet selectivity to determine if the residual variation in gillnet age-composition data are better explained by systematic changes in the empirical weight-at-age data or selectivity has been relatively constant and natural mortality rates have varied over time.

In this part of the document, we first describe the five major and two minor Stock Assessment Regions (SARS) that comprise the BC herring stocks. We then present the input data used in this years assessment, briefly describe the analytical methods and diagnostics, describe the recruitment and catch forecasts, and the Harvest Control Rule (HCR) used for generating catch advice. We then present the maximum likelihood estimates of residual patterns and overall fits to the observations, summarize MSY based reference points and maximum likelihood estimates of  $B_0$ . Lastly, we present the results of integrating the joint posterior distribution, diagnostics for ensuring convergence, marginal parameter distributions (with prior distributions overlaid), and catch advice based on the median values of the joint posterior distribution. The last section presents the data, MLE results, marginal distributions and catch advice for the two minor areas (the HCR in the minor area differs from the major areas).

## 2.6 BC Herring Stocks

The geographic boundaries used to delineate the B.C. herring stock assessment regions have remained consistent since 1993. Boundaries and locations of the major stock and minor stock areas are identified in Figure 1.1. The Haida Gwaii (HG) or Queen Charlotte Islands (QCI2E) stock assessment region includes most of Statistical Area 2E, spanning from Cumshewa Inlet in the north to Louscoone Inlet in the south. The Prince Rupert District (PRD) stock assessment region encompasses Statistical Areas 03 to 05. The Central Coast (CC) assessment region separates the major migratory stocks from the minor spawning populations in the mainland inlets. The Central Coast assessment region includes Statistical Area 07 plus Kitasoo Bay in Area 06, Kwakshua Channel in Section 085 and Fitz Hugh Sound in Section 086. The Strait of Georgia (SOG) stock assessment region includes all of Statistical Areas 14 to 19, 28, and 29 (excluding Section 293), Deepwater Bay and Okisollo Channel, both in Section 132, and Section 135. The west coast of Vancouver Island (WCVI) assessment region encompasses Statistical Areas 23 to 25. The minor stocks include all of Area 27 and Area 2W (excluding Louscoone Inlet in Section 006). Current geographic stock boundaries are outlined in [Midgley \(2003\)](#), although note that SOG sections 280 and 291 do not appear as they were added in 2006.

## 2.7 Methods

### 2.7.1 Input data & assumptions

#### Catch data

For each of the statistical areas, the required input data for  $i\text{SCA}_M$  consists of a catch time series for each of the fishing fleets. For the BC herring fishery, the annual total removals has been partitioned into three distinct fishing fleets (or fishing periods, see Figure 2.1). The first fleet is a winter seine fishery that has been in operation since the start of the assessment in 1951, the second is a seine-roe fishery that commenced in 1972 in the Strait of Georgia, and the third fleet is a gillnet fishery that targets females on the spawning grounds. The model is fit to the catch time series information and assumes measurement errors are log-normal, independent and identically distributed. The assumed standard deviation in the catch observation data must be specified in the control file and it is assumed that measurement errors in the catch is the same for all fishing periods. The units of the catch are given in 1000s of metric tons.

In addition to the commercial catch, removals from fisheries independent surveys must also be specified in  $i\text{SCA}_M$ . Two additional fleets are specified to represent the spawn survey, where the spawn survey is broken into two distinct time periods pre-1988 and post-1988, the year when the survey switched from surface surveys to dive surveys. This partitioning of the data is done for two reasons: (1) to allow for different catchability coefficients to be specified for the early and late periods, and to allow for more weight to be placed on the contemporary data due to improved precision in the estimates of egg layers.

#### Relative abundance data

Herring spawn surveys have been conducted throughout the B.C. coast beginning in the 1930s. Prior to 1988, spawn surveys were conducted from the surface either by walking the beach at low tide or using a drag from a skiff to estimate the shoreline length and width of spawn. Egg layers were sampled visually and are used to calculate egg densities following the methods of [Schweigert \(2001\)](#). Beginning in 1988, herring spawn surveys using SCUBA methods were introduced and were implemented coastwide within a couple of years initially being conducted by DFO staff and eventually through contract divers hired through the test fishing program. Prior to the 2006 Larocque ruling, the test fishing program was funded through an allocation of fish by industry. In years since the 2006 Larocque ruling, the availability of resources to conduct dive surveys in all areas has been reduced. For 2011, dive surveys were conducted in all major and minor assessment regions, with the exception of Area 2W where snorkelling and surface survey methods were also used. As in earlier years, a few minor spawning beds outside the main assessment areas were surveyed by SCUBA or surface methods where resources permitted.

The locations of the spawning beds for the five major and two minor stock areas are shown in Figure 2.2. Egg density estimates are used to calculate a fishery-independent index of herring spawning biomass, referred to as the spawn survey index hereafter ([Schweigert, 2001](#)).

The spawn survey is conducted after the fisheries in the area have been completed; therefore, it is assumed that all the mortality for the year has occurred just prior to commencing the spawning survey. The fisheries independent survey estimates egg density and total spawn area, and from this information the total female spawning biomass can be estimated assuming the 200 eggs per gram of female body weight or 100 eggs per gram of mature body weight of both sexes ([Hay, 1985](#); [Hardwick, 1973](#)). The assumed selectivity for the spawn survey is fixed to the maturity schedule for herring and the mean weight-at-age data comes from empirical observations based on biological samples.

#### Biological samples

Biological samples are collected from both commercial catch and from the test fishery program. Commencing in 1975, test fishery charters supplemented biological samples in areas where catch sampling that was not representative of the stock in that area (i.e., fishing solely on spawning aggregations), or in closed areas. Prior to 2006, test fishing charters were funded through an allocation of fish to the test program; the program is now fully funded by DFO. Through a contract with DFO, the Herring Conservation and Research

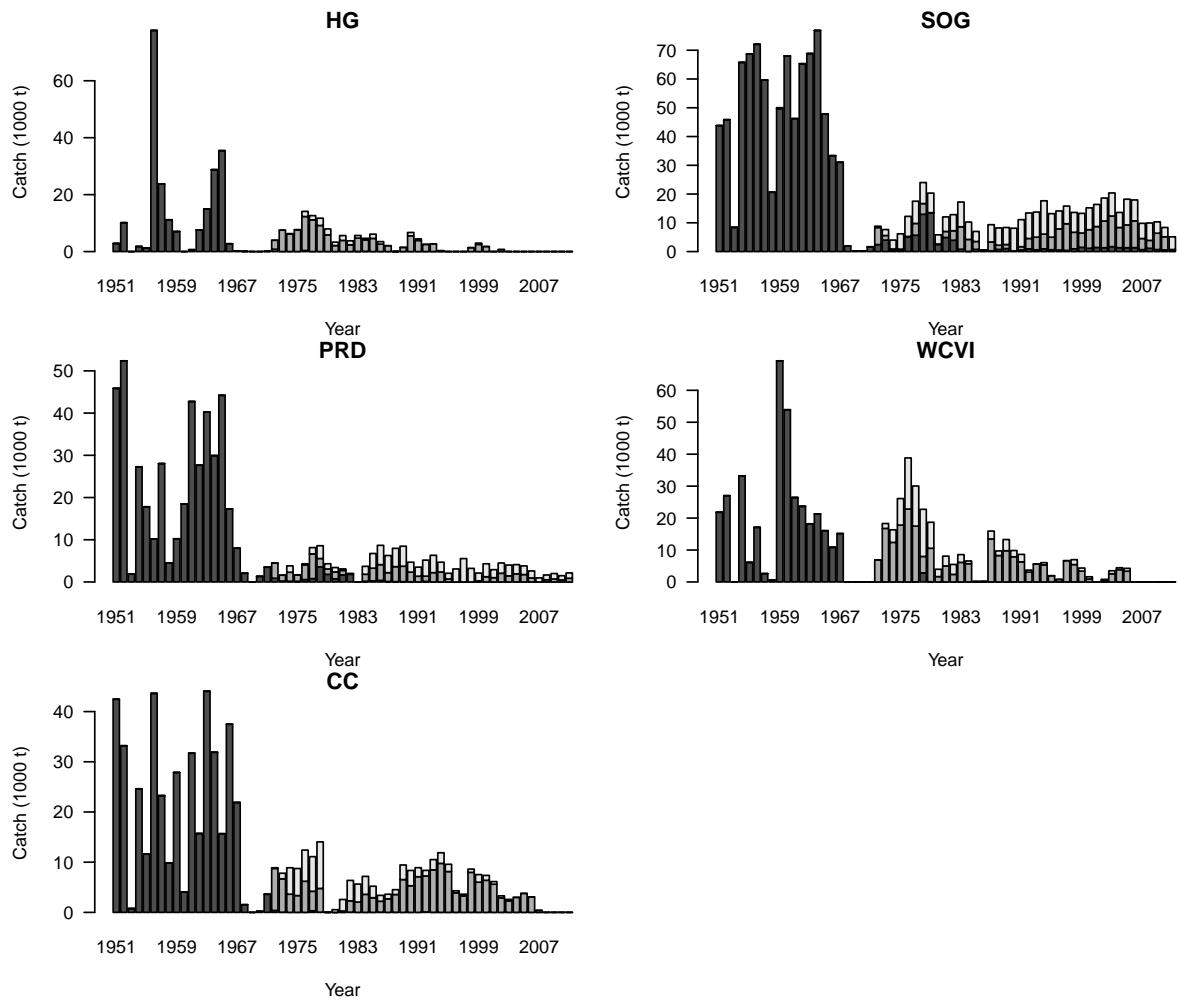


Figure 2.1: Historical catch of herring in the five major stock areas between 1951 and 2011 for the winter purse seine fishery (dark bars), seine-roe fishery (grey bars), and gillnet fishery (light grey bars). Units of catch are in thousands of metric tons.

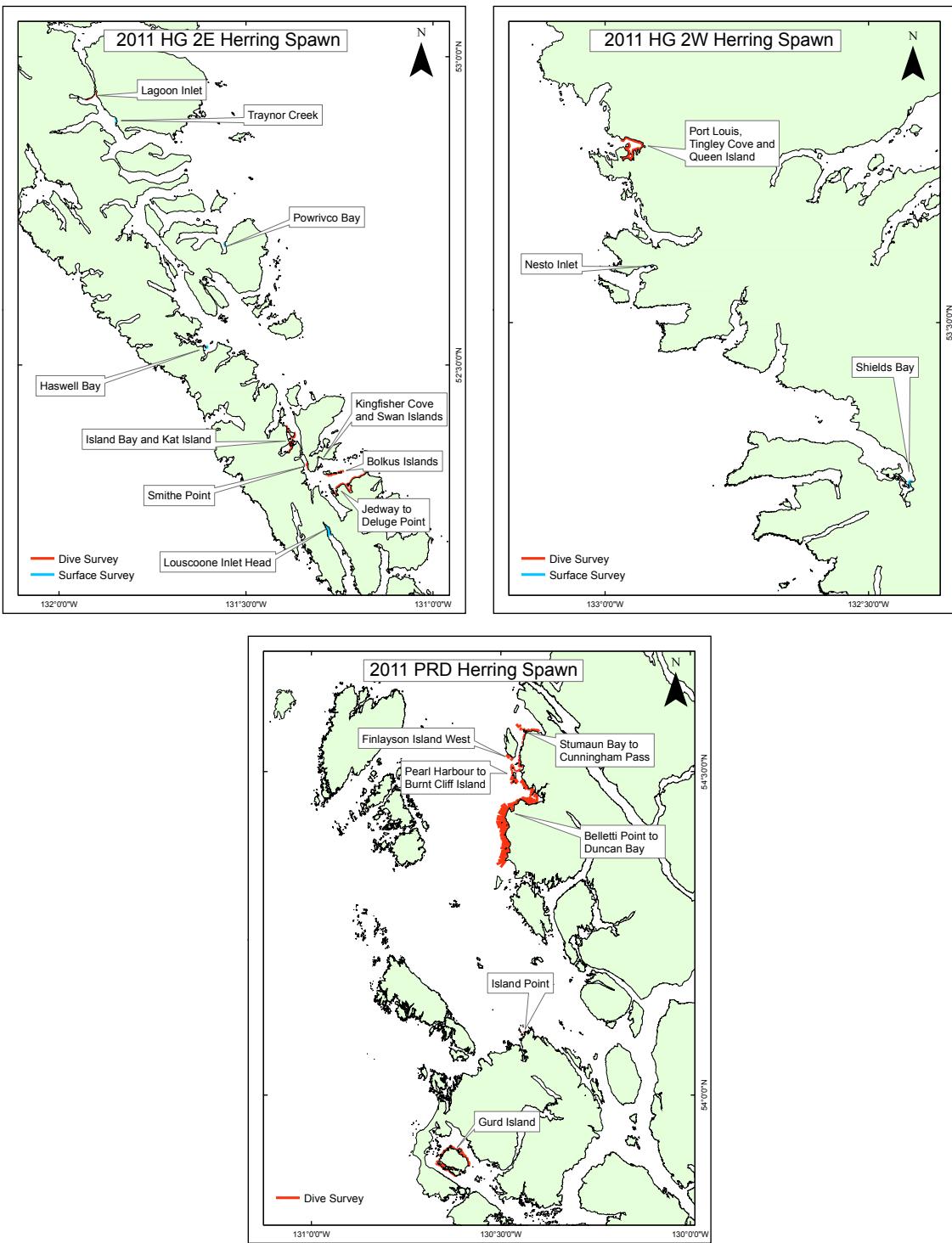


Figure 2.2: Preliminary Spawning activity for Haida Gwaii (top panels) and Prince Rupert District (bottom) in 2011.

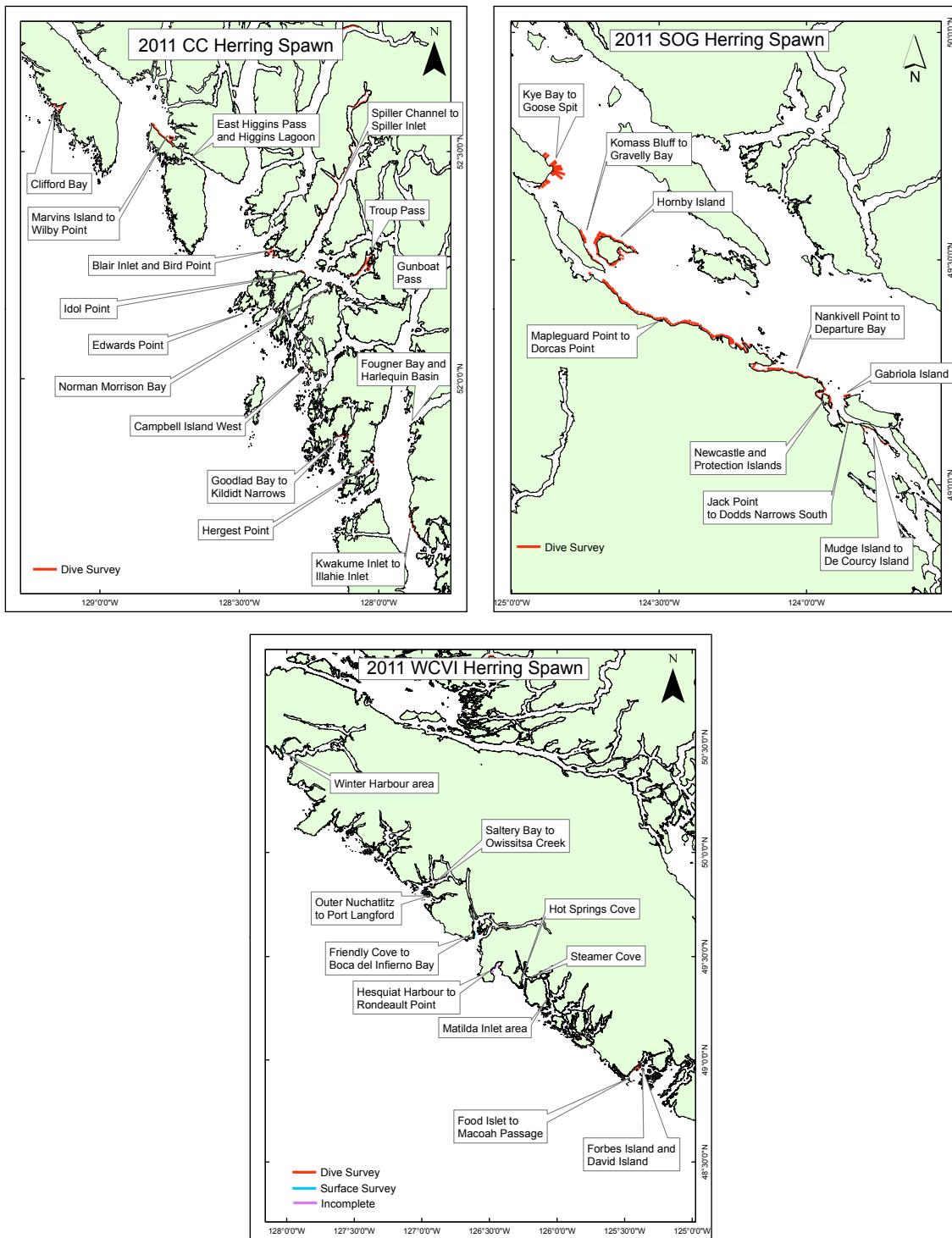


Figure 2.2: Preliminary Spawning activity for Central Coast (top left panel), Strait of Georgia (top right) in 2011 and west coast Vancouver Island (bottom).

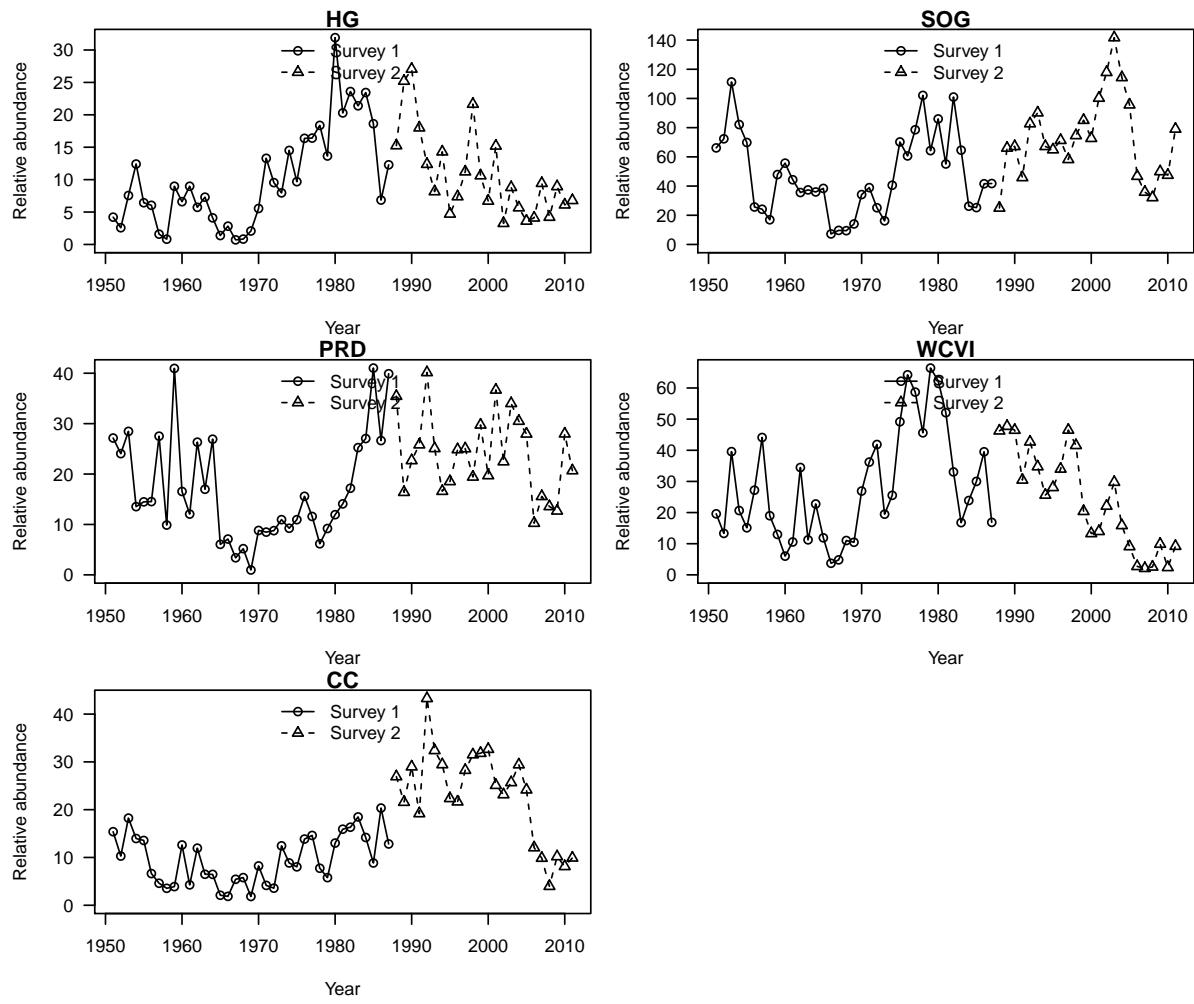


Figure 2.3: Spawn survey index for Strait of Georgia between 1951 and 2011. The units are actual estimates of spawning biomass (1000s tons), but only the trend information is used in the model fitting.

Table 2.1: Summary of biological samples collected and processed from all sources from the 2010/11 herring season.

Stock	Commercial samples				
	Roe fishery	SOK fishery	F&B	Test fishery	Research
HG (QCI 2E)				13	
PRD	29	1		24	
CC				30	
SOG	18		20	60	
WCVI				14	16
Area 2W				10	
Area 27		3			
Other Areas					
Total	57	4	16	151	16

Table 2.2: Summary of biological samples collected and processed from commercial catch and test fishery charters from 2002/03–2010/11.

Fishing season	Commercial fishery samples	Charter and research samples	Total
2002/03	120	287	407
2003/04	79	222	301
2004/05	83	191	274
2005/06	46	164	210
2006/07	114	85	199
2007/08	116	103	219
2008/09	87	136	223
2009/10	78	135	213
2010/11	81	167	248

Society (HCRS) sub-contracts a number of vessels to collect biological samples. Industry also conducts pre-season test sets for roe-quality testing in open areas and supplementary biological samples are provided as part of this program. The following data are collected for all biological samples: fish length, weight, sex, and maturity. Subsequently these sources of data are compiled and used as the information on mean weight-at-age and catch-at-age data that are the essential input data for the stock assessment model.

During the 2010/2011 season a total of 248 biological samples were collected, of which 151 were collected from the test fishery, 57 were collected from the roe fishery, 16 from the food & bait fishery, 4 from Spawn on Kelp (SOK) operations, and 16 from the summer trawl research survey (Table 2.1). Note that the definition of a sample is roughly 100 individual fish. A summary of biological samples collected from commercial and pre-fishery charters from 2002/03–2010/11 is presented in Table 2.2 and the spatial locations of the biosamples are presented in Figure 2.4.

### Age composition data

Ageing data, through the reading of fish scales, are collected from the biological samples taken from the commercial fisheries and test fishery charters. At present, the biological samples from the test fisheries are pooled with the seine-roe fisheries. Future analyses may further disaggregate these data to determine if the test fishery and the seine-roe fishery have very different age-compositions. Age composition data is used to determine proportions-at-age and is an essential source of input data to the herring stock assessment model.

In all of the major SARS, catch-at-age data from the winter seine fishery (top panels of Figures 2.5-2.9)

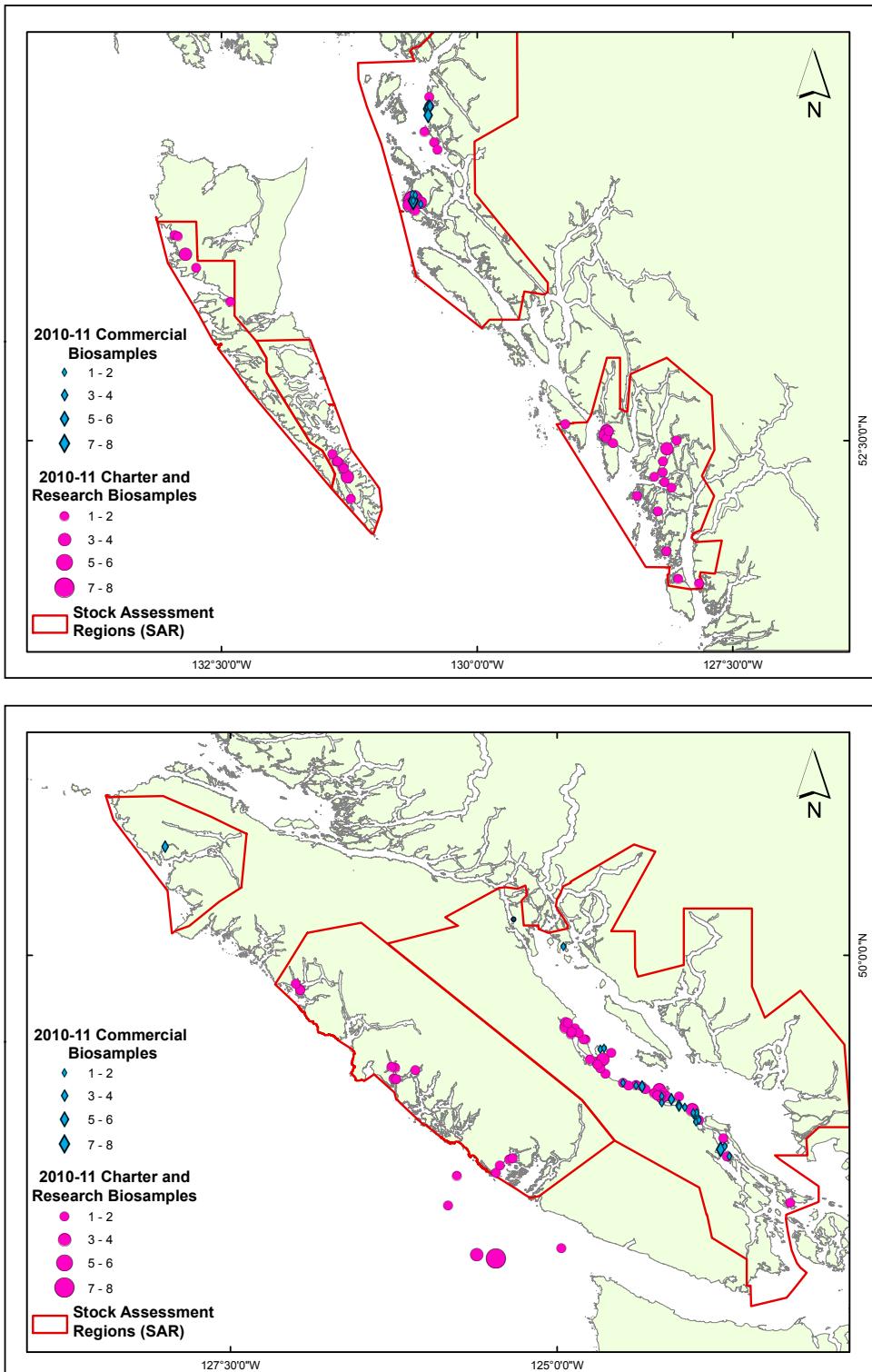


Figure 2.4: Spatial location and sample sizes of 2011 biosamples from commercial and research-charter programs in the north coast (top panel) and south coast (lower panel).

tend to consist of younger fish in comparison to the age composition data from the seine-roe and gillnet fleets post 1970. The shaded polygons in Figures 2.5-2.9 approximates the 95% distribution of ages in the catch. Roughly 90% of the fish landed in the winter seine fishery were younger than age-7, and younger than age-6 in recent years. In both the winter seine and seine-roe fishery age-2 fish are frequently landed; whereas, age-2 fish are rarely landed in the gillnet fishery, and fish do not appear to fully recruit to the gillnet gear until at least 4-5 years of age. The mean age of the catch appears to be increasing between 2008 and 2010 in both the gillnet and winter seine fishery, and there is no obvious trend in the seine roe fishery. There is however a declining trend in the older ages caught in the seine-roe fishery since 2006 (erosion of age-structure).

### Mean weight-at-age data

From the mid-1970s until the present, there has been a measurable decline in weight-at-age for all ages in all major stock areas (Figure 2.10). Samples collected during the 2009/10 fishing year indicate weights-at-age that are among the lowest on record. This declining weight-at-age may be attributed to any number of factors, including: fishing effects (i.e., gear selectivity), environmental effects (changes in ocean productivity), or it may even be attributed to changes in sampling protocols (shorter time frame over which samples are collected). Declining weight-at-age has been observed in all five of the major stocks, and despite area closures over the last 10-years, has continued to occur in the QCI and WCVI stocks. Although the direct cause of this decline is still to be investigated, this trend has been observed in B.C. and U.S. waters, from California to Alaska (Schweigert et al., 2002), and merits further research. The observed mean weight-at-age data appear to have a few errors that need to be investigated as well; for example, see the apparently small age-10 fish in 2001 in Figure 2.10.

Mean weight-at-age data are based on the biological samples taken from the fisheries and test fishery data. The spatial distribution of the biological samples from 2011 are shown in Figure 2.4.

### 2.7.2 Analytical methods

For the 2011 BC herring assessment,  $iSCA_M$  was used to conduct the stock assessment for each of the five major Stock Assessment Regions (SAR) and two minor assessment areas (Area 2W and Area 27). The technical details of this model can be found in Appendix A.1.

### 2.7.3 Retrospective analysis

A retrospective analysis was conducted for each of the major and minor SARs. The retrospective analysis successively removes the last 10-years of data and examines changes in estimates of terminal spawning biomass. The results are then plotted on a single panel to compare how estimates of spawning biomass change as successive years of data are omitted from the analysis.

### 2.7.4 Abundance and recruitment forecasts

The abundance forecast for the upcoming fishing season, also referred to as pre-fishery biomass, is defined as the predicted biomass of age-4 fish and older plus the number of age-3 fish recruiting in year  $T + 1$ . The abundance estimates are based on the median values from the sampled posterior distribution. Age-3 recruits are based on poor, average, and good recruitment scenarios; see next paragraph for definitions of poor, average and good.

The recruitment forecasts are based on the surviving number of age-3 fish at the start of the fishing season times the average weight-at-age 3 in the last 5 years. The definitions of poor, average, and good recruitment are as follows: **Poor** is the average recruitment from the 0-33 percentile, **Average** is the average recruitment from the 33-66 percentile, and **Good** is the average recruitment from the 66-100 percentile. Note that all cohorts from 1951 to 2011 were included in the calculation of recruitment quantiles.

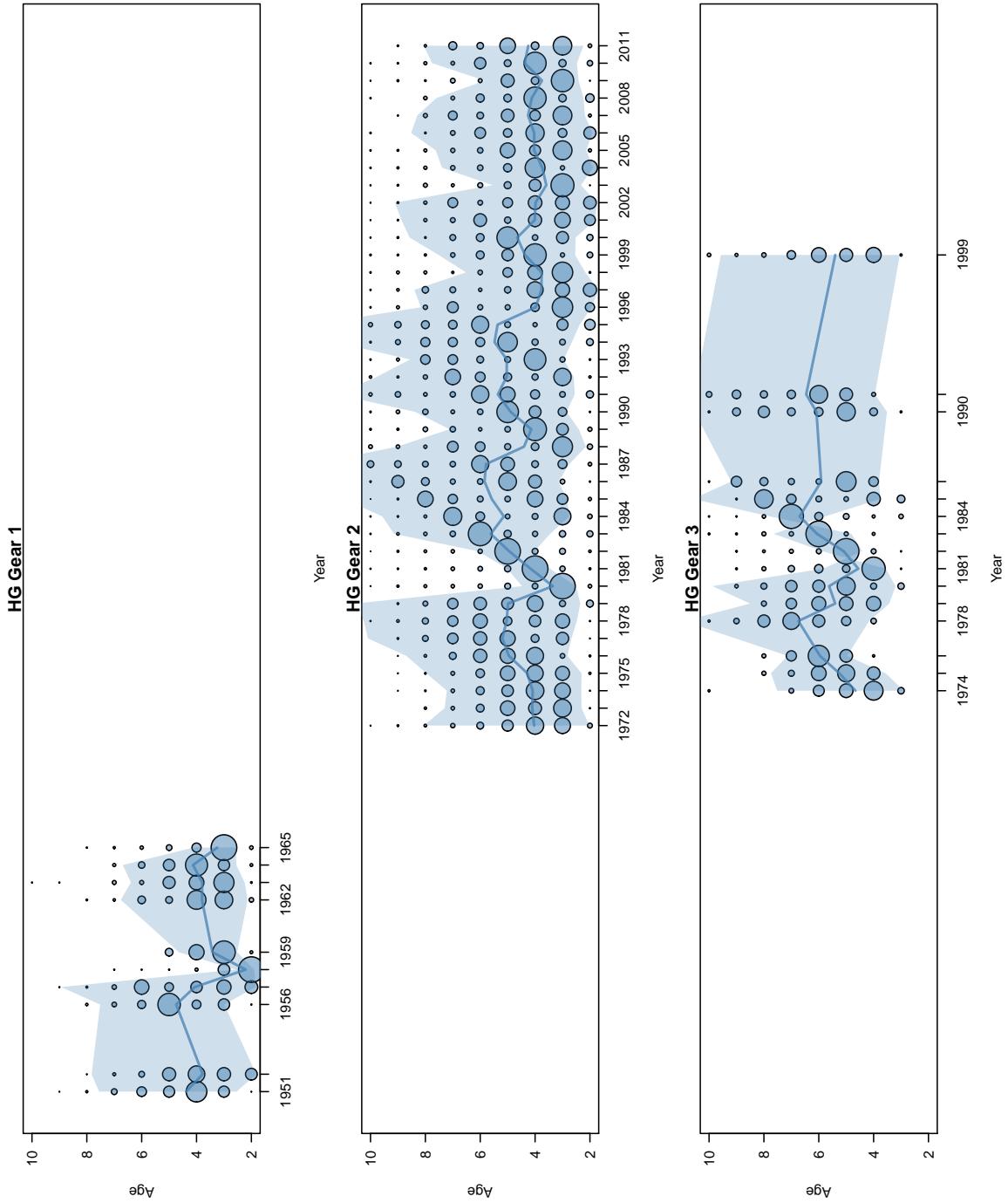


Figure 2.5: Proportions-at-age versus time for the winter purse seine fishery (top), seine roe fishery (middle) and the gillnet fishery (bottom) in Haida Gwaii. The area of the circle reflects the proportion-at-age, each column sums to 1, zeros are not shown, and age 10 is a plus group. Also shown is the mean age of the catch (line) and the approximate 95% distribution of ages (shaded polygon) for each year.

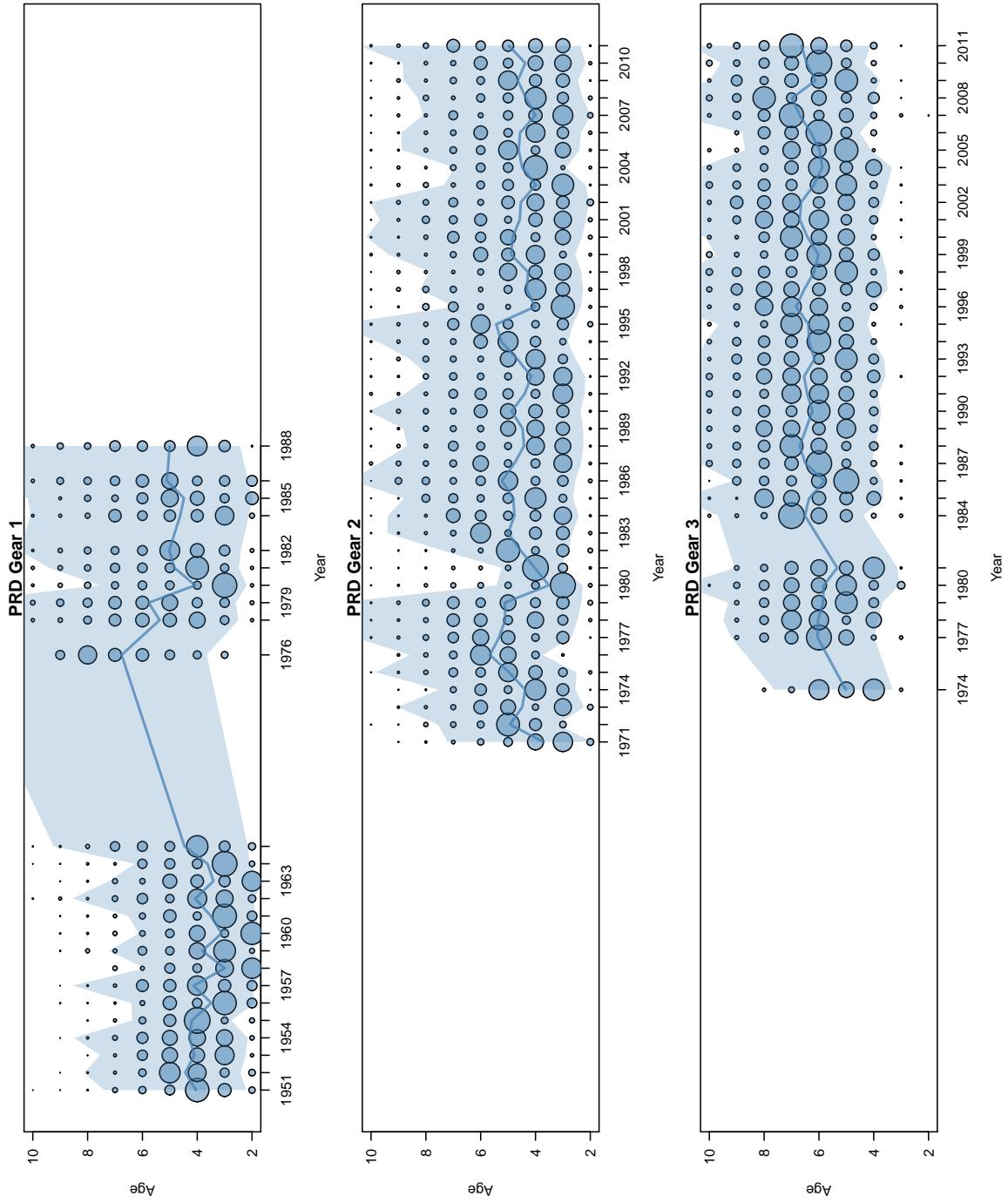


Figure 2.6: Proportions-at-age versus time for the winter purse seine fishery (top), seine roe fishery (middle) and the gillnet fishery (bottom) in Prince Rupert District. The area of the circle reflects the proportion-at-age, each column sums to 1, zeros are not shown, and age 10 is a plus group. Also shown is the mean age of the catch (line) and the approximate 95% distribution of ages (shaded polygon) for each year.

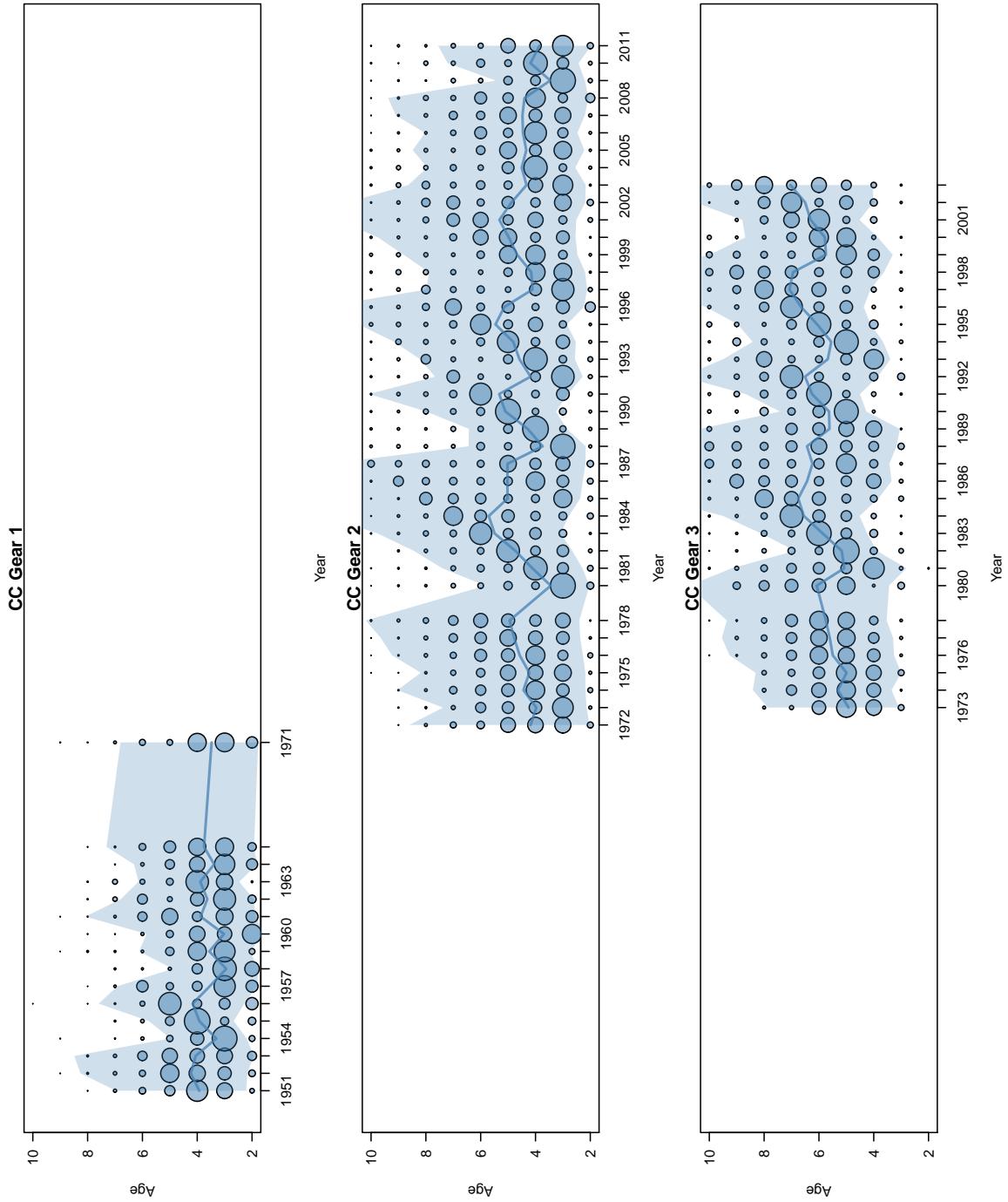


Figure 2.7: Proportions-at-age versus time for the winter purse seine fishery (top), seine roe fishery (middle) and the gillnet fishery (bottom) in the Central Coast region. The area of the circle reflects the proportion-at-age, each column sums to 1, zeros are not shown, and age 10 is a plus group. Also shown is the mean age of the catch (line) and the approximate 95% distribution of ages (shaded polygon) for each year.

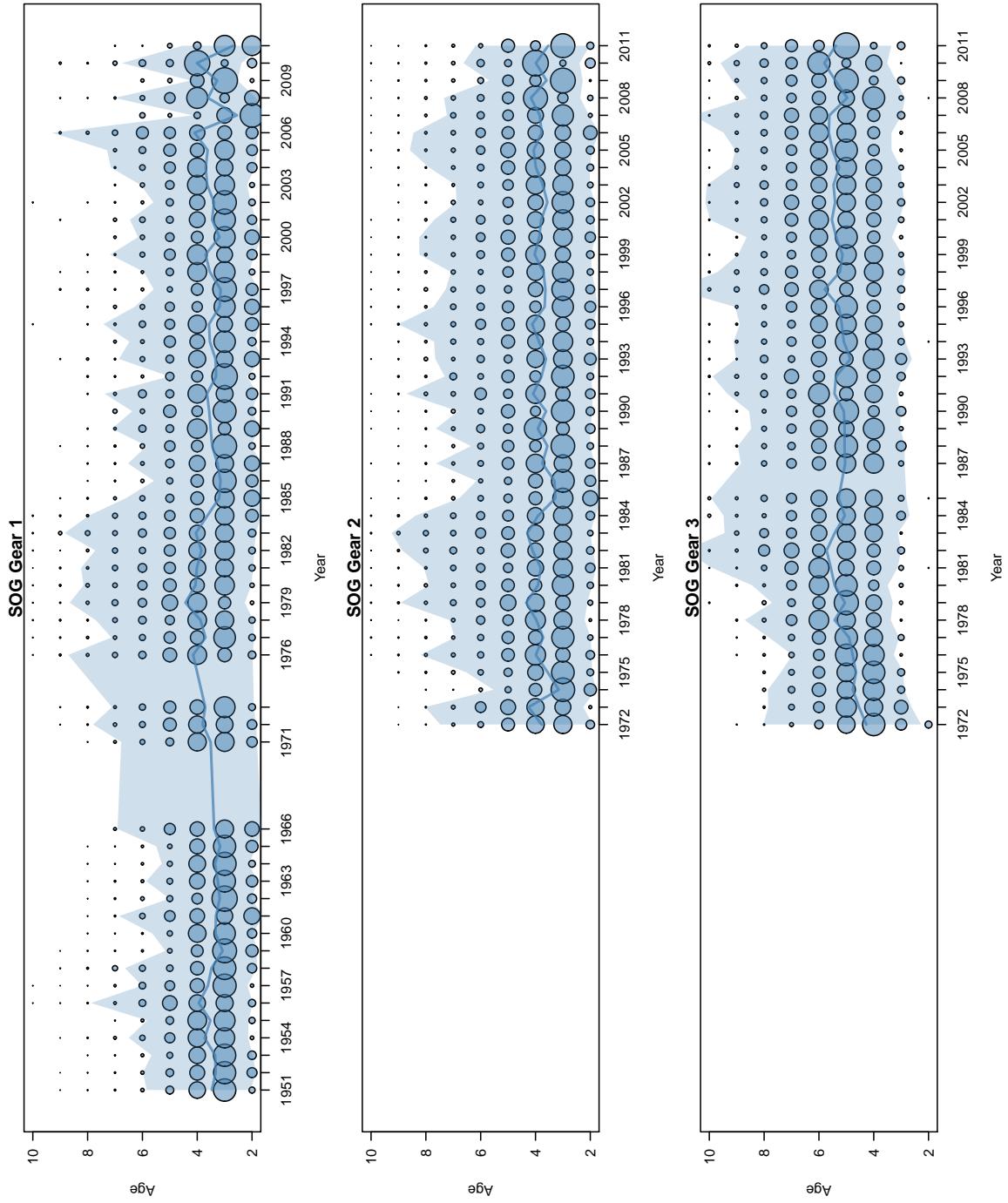


Figure 2.8: Proportions-at-age versus time for the winter purse seine fishery (top), seine roe fishery (middle) and the gillnet fishery (bottom) in the Strait of Georgia. The area of the circle reflects the proportion-at-age, each column sums to 1, zeros are not shown, and age 10 is a plus group. Also shown is the mean age of the catch (line) and the approximate 95% distribution of ages (shaded polygon) for each year.

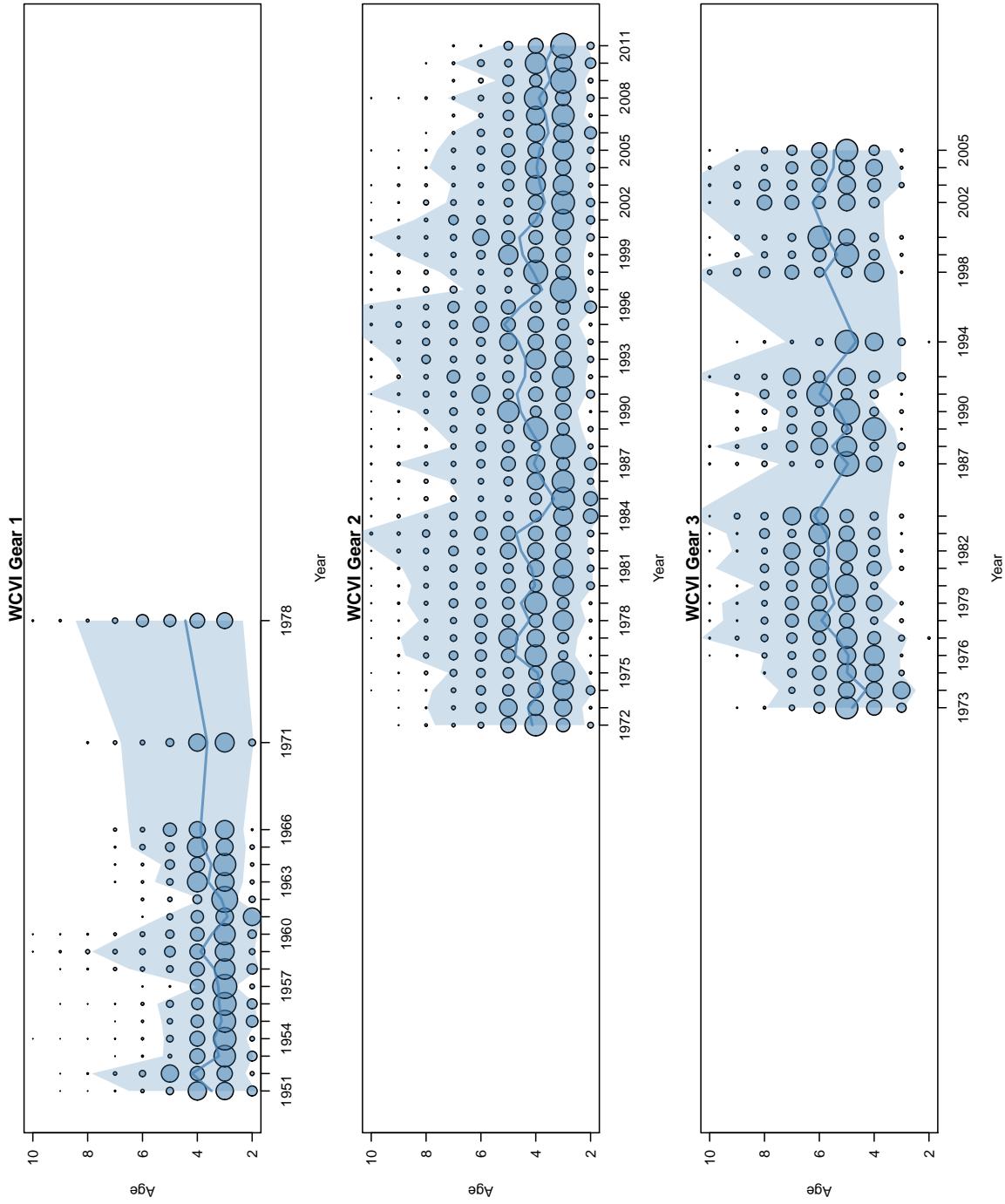


Figure 2.9: Proportions-at-age versus time for the winter purse seine fishery (bottom), seine roe fishery (middle) and the gillnet fishery (bottom) in the West Coast Vancouver Island region. The area of the circle reflects the proportion-at-age, each column sums to 1, zeros are not shown, and age 10 is a plus group. Also shown is the mean age of the catch (line) and the approximate 95% distribution of ages (shaded polygon) for each year.

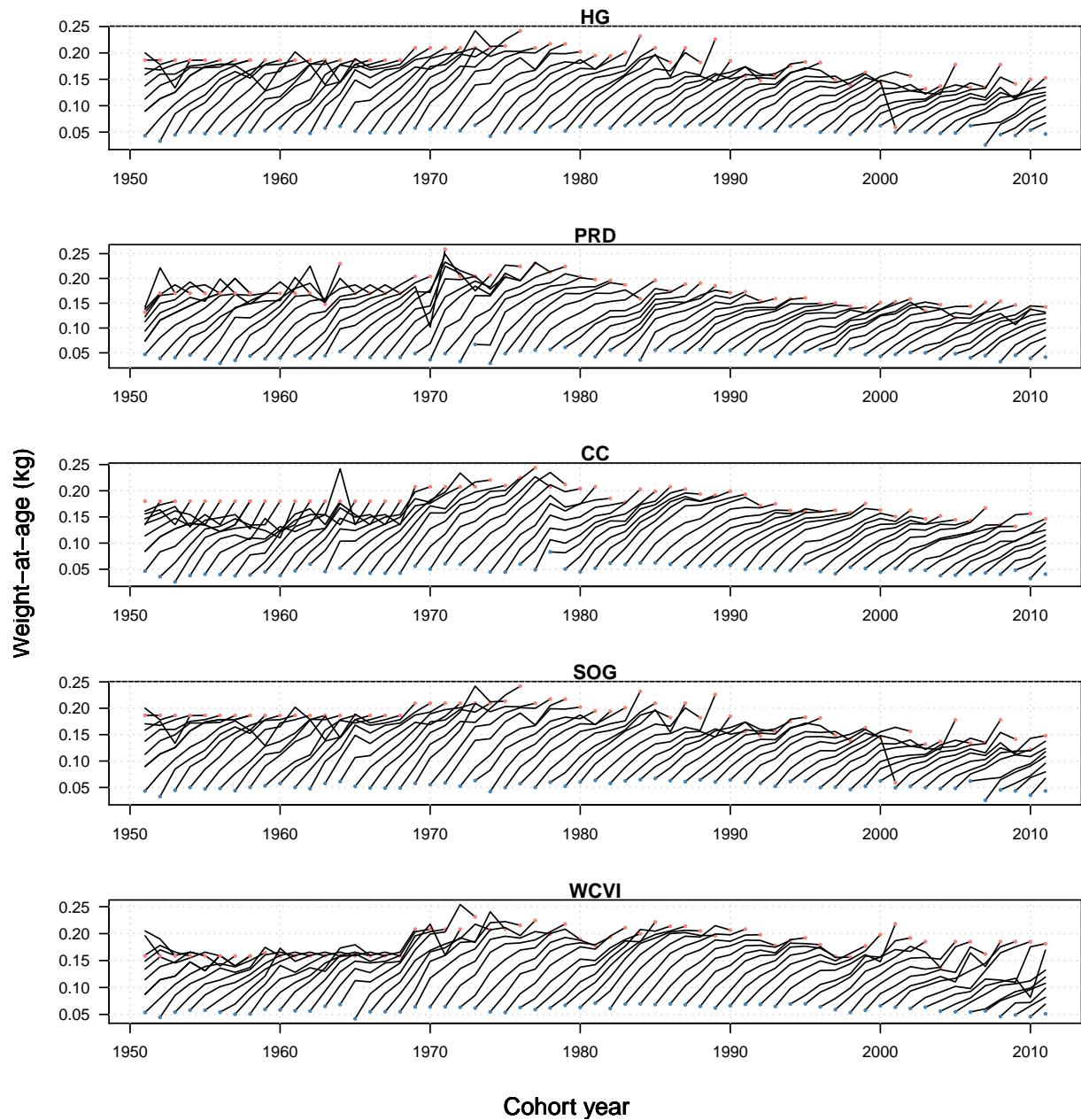


Figure 2.10: Empirical mean weight-at-age data by cohort from 1951 to 2011 for ages 2 to 10 in the five major Stock Assessment Regions.

## 2.7.5 Harvest control rule

Catch advice is based on the application of the harvest control rule (HCR). A formal HCR has been used to provide management advice for the major BC herring stocks since 1986 ([Stocker, 1993](#)). The herring HCR has three components:

1. Reference points
2. Harvest rate
3. Decision rules

These three components are consistent with the DFO harvest strategy that is compliant with the precautionary approach ([Fisheries and Oceans Canada, 2006](#)). In this strategy, there are two reference points: (1) the limit reference point (LRP) which is a minimum stock size where fishing activity is ceased if the stock falls below the LRP into the critical zone, and (2) the upper stock reference (USR) that defines the boundary between the cautious zone and healthy zones.

### Reference Points

The harvest control rule that is currently used to provide catch advice for the five major BC herring stocks is a hybrid between a fixed escapement policy and a fixed exploitation rate policy. For each of the major stocks, the reference point is defined as a cutoff level (or escapement target) and is set at 25% of the unfished spawning stock biomass. The cutoff is intended to maintain a minimum spawning stock biomass of 25% of the estimated unfished biomass. Simulation studies in the past ([Haist et al., 1986; Hall et al., 1988](#)) suggest that 25% of the unfished spawning biomass leaves a sufficient spawning reserve to ensure long-term sustainability of the resource.

At present, there are no formal definitions for LRP and USR for the five major herring stocks. The cutoff values for each of the stocks are thought to be more conservative than the default LRP of  $0.4B_{MSY}$ . For example, surplus production in most fish stocks is usually maximized when the stock is depleted in a range of 30%-45% of its unfished state. If we assume that herring production was maximized at a depletion level of 45% or  $B_{MSY}=0.45B_o$ , then the default LRP for herring would be equal to 18% of the unfished biomass (i.e. 40% of  $B_{MSY}/B_o$ ). This document also presents the maximum likelihood estimates of spawning biomass depletion, and these results are overlaid on coloured panels that define the default  $0.4B_{MSY}$  and  $0.8B_{MSY}$  LPR and USR, respectively (see Figure 2.20).

Critical to the HCR is the estimate of unfished spawning biomass ( $B_o$ ). The cutoff levels were last revised in 1996 ([Schweigert et al., 1996](#)), and these same values have been used to provide catch advice ever since. In this assessment, we provide updated estimates of  $B_o$  and the associated cutoff values based on  $0.25B_o$ .

In the case of the minor stock areas, the harvest control rule consists of a fixed exploitation rate and there are now cutoff values associated with these stock assessment regions.

### Harvest rate

The Pacific Science Advice Review Committee (PSARC) has reviewed the biological basis for target exploitation rate, considering both the priority of assuring conservation of the resource and allowing sustainable harvesting opportunities (Schweigert and Ware 1995). The review concluded that 20% is an appropriate exploitation rate for those major stock areas that are well above cutoff levels of 25% of the estimated unfished biomass.. The recommended 20% harvest rate is based on an analysis of stock dynamics which indicates this level will stabilize both catch and spawning biomass while foregoing minimum yield over the long term ([Hall et al., 1988; Zheng et al., 1993](#)).

In the case of minor stock areas, data-limitations present a challenge in providing reliable estimates of unfished biomass, required for the calculation of stock-specific cutoffs. Consequently, the PSARC recommended harvest rate of 10% is applied to the currently estimated biomass for the following year for these areas.

## Decision rules

For the major stock areas, the harvest control rule combines both constant exploitation rate and constant escapement policies, allowing for smaller fisheries in areas where the 20% harvest rate would bring the escapement down to levels below the cutoff. The rule operates as follows:

- If the forecast is less than the cutoff: the area is closed to all commercial harvest.
- If the forecast run ( $B_{t+1}$ ) is greater than the cutoff: A commercial harvest is permitted and the harvest rate is based on the following rules:
  - If  $0.8B_{t+1} > \text{Cutoff}$ , then harvest rate  $u = 20\%$ .
  - If  $0.8B_{t+1} < \text{Cutoff}$ , then harvest rate  $u = \frac{B_{t+1} - \text{Cutoff}}{B_{t+1}}$

In the case of the minor stock areas, the decision to allow for a commercial harvest has been at the discretion of Fisheries Management. In years where a commercial harvest is permitted, a harvest rate of 10% is applied to the estimated biomass for the area.

## 2.8 Results

The results section is broken down into three major subsections, Maximum likelihood fits to the data, marginal posterior distributions, and stock forecasts and catch advice based on samples from the joint posterior distribution.

### 2.8.1 Maximum likelihood fits to the data

Although the maximum likelihood estimates are not explicitly used for constructing the catch advice, we do present the MLE estimates of the residual patterns and fits to the data for comparisons.

#### Catch residuals

Residuals between the observed and predicted catch are largely determined by the user specified standard deviation in each of the control files. In this assessment, the assumed variance for all regions (including minor regions) was set at 0.005, which corresponds to a standard deviation of approximately 0.0707. Overall the residuals for each fishery in each stock assessment region are unremarkable (Fig. 2.11), with exception of a major outlier in the Haida Gwaii in the mid 1950s. In 1956, the reported catch in Haida Gwaii was extremely large ( $> 60,000$  mt) and the model has a difficult time explaining this large catch. In order to explain this large catch in a single year, a large biomass in the region is required.

#### Fits to the spawn survey data

The residuals between the observed and predicted spawn survey index (on a log scale) are shown in Figure 2.12. Recall that the spawn survey data are treated as two independent time series where data between 1951–1987 were based on surface estimates of spawn deposition and data post 1988 are based on diver surveys of spawn deposition. More weight was assigned to the contemporary data. Also, you might be tempted to compare the estimated values of  $q$  in Figure 2.12 with those estimated in Part I of this document (e.g., Table 1.2 on page 13). The results in Table 1.2 are based on data from 1951 to 2010 (i.e., omit the 2011 data) and use a different parameterization of the selectivity function for the gillnet fishery (type 7 as opposed to type 8, see Appendix A.1.5 on page 97).

For most areas, there is little pattern in the residuals between the observed and predicted survey data (Fig 2.12). For the HG, PRD and CC regions, there is very good correspondence between the observed and predicted survey data post 1988. In the SOG, there is a period of positive residuals between 1999 and 2005 where the predicted spawn biomass fails to increase as much as indicated by the survey. Similarly 3–4 year trends also exist in the WCVI spawn survey data after the year 2000.

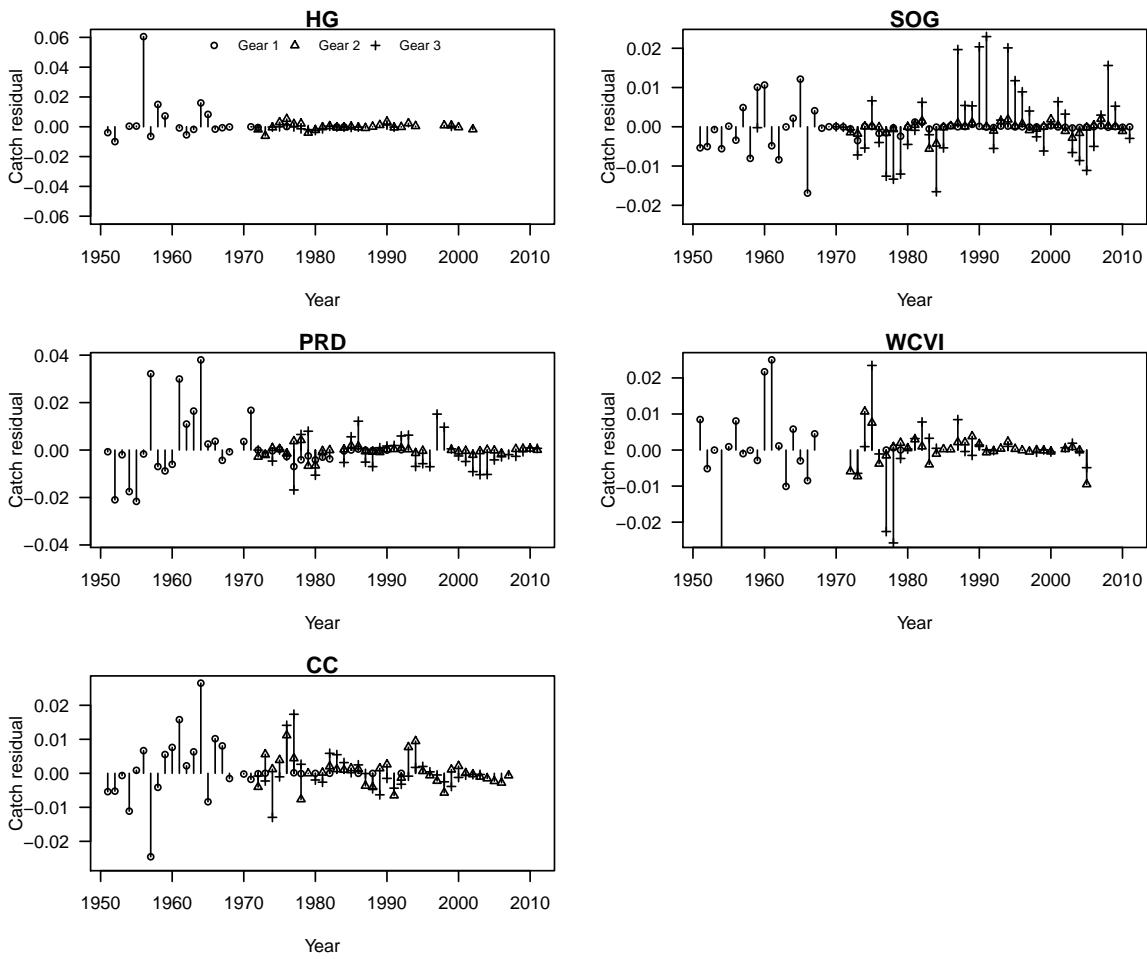


Figure 2.11: Residual for the log difference between observed and predicted catch for the five major SARs for each gear type (Gear 1 = winter seine fishery, Gear 2 = seine-roe fishery, Gear 3 = gillnet fishery).

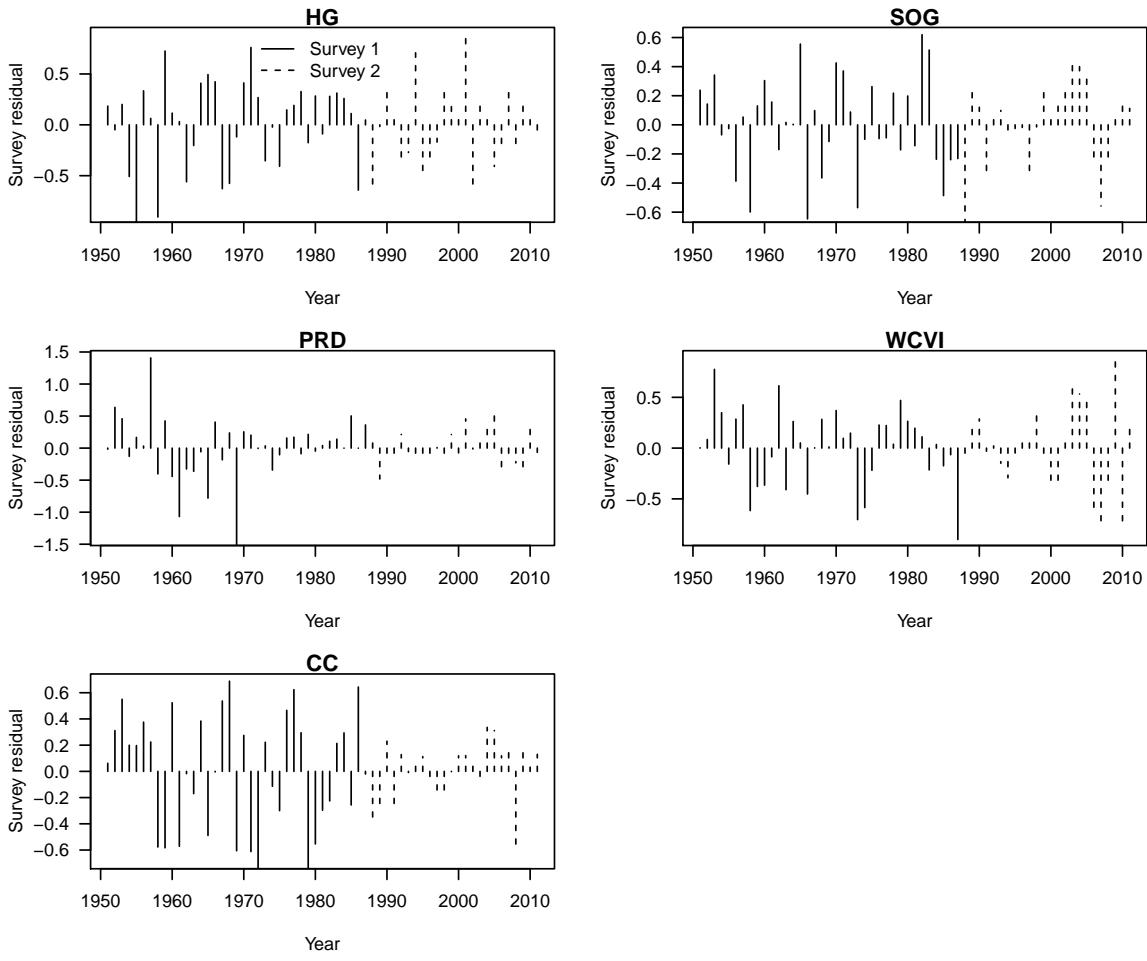


Figure 2.12: Residual patterns for the log difference between observed and predicted spawn survey abundance for the five major SARs. Spawn survey data based on surface estimates are shown as solid lines and data based on diver surveys is shown as dashed lines.

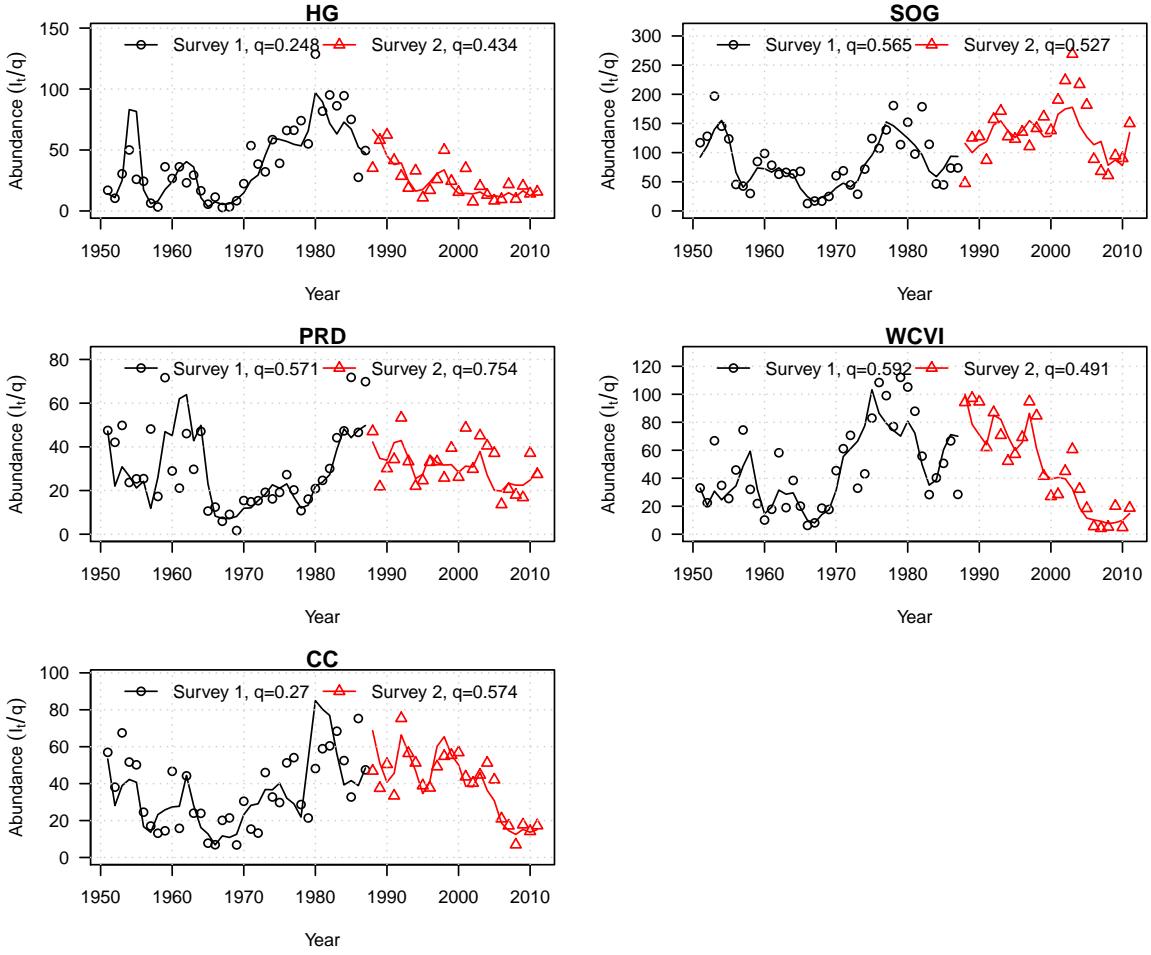


Figure 2.13: Observed (points) and predicted (lines) spawn survey abundance data scaled by the MLE estimate of  $q$  for each of the five major SARs. In each panel, the corresponding scaler ( $q$ ) is presented for each of the surveys.

In comparison to the previous assessment for Pacific herring using the HCAM model, estimates of the catchability coefficient are very different (HCAM assumed  $q=1$  for post 1988 data). In each of the five major assessment regions (and the two minor regions) a less informative prior for the catchability coefficient was used (see Appendix C.3). Maximum Likelihood Estimates (MLE) of the catchability coefficients are presented for each region in Fig. 2.13 along with the observed and predicted trends in the spawn index. Estimates of  $q$  in both time periods are less than 1.0 for all regions. The interpretation of  $q = 1$  is that the spawn survey data is an absolute measure of spawn abundance,  $q < 1$  implies that the survey under-estimates the spawn abundance and  $q > 1$  implies an over-estimate. For example, in the HG region the MLE values for  $q$  are 0.248 and 0.434 for the pre- and post-1988 data, respectively. This could be interpreted as the spawn survey, on average, sees 24.8% and 43.4% of the deposited spawn each year. This interpretation however is conditional on the specification of mature biomass in the stock assessment model and the methods used to extrapolate egg density to spawning biomass. Values of  $q < 1$  could also be interpreted as the fraction of eggs remaining at the time the spawn survey was conducted (i.e.,  $(1-q)$  of the eggs survived predation, storms, etc.)

### Age composition residuals

The assumed error distribution for the age-composition data has changed in this assessment from a multinomial distribution implemented in HCAM to a multivariate-logistic distribution. In the former implementation the age-composition data were weighted by the annual samples sizes in each region for each age and year. In the  $iSCAM$  implementation the age-composition data for all years is given the same weight (i.e., we assume the observation errors is homogenous) based on the conditional maximum likelihood estimate of the variance (see Appendix A.1 for full details). We further pool age-proportions that are less than 2% into the adjacent younger year class to reduce the influence of small outliers and weak cohorts.

In HG the MLE estimates of the variance for each gear is 0.102, 0.106 and 0.306, for the winter seine, seine-roe and gillnet fleets, respectively (Fig. 2.14). In general there is fairly good agreement between the observed and predicted age-composition data in this region, with poorer fits to the gillnet age-composition data. There is no persistent pattern in the residuals.

For the PRD region, the fits to the age-composition data are slightly poorer, with MLE estimates of the variance ranging from 0.164 to 0.269 for the gillnet and winter seine fleets (Fig. 2.15). There is no remarkable pattern in the winter seine fishery, the seine-roe fishery tends to have positive residuals for age-3 and age 7+ fish, and negative residuals for ages 5-6 fish. Residuals in the gillnet fishery are mostly negative for age-4 fish post 1988. The gillnet gear tends to catch older fish than both seine gears.

For the Central Coast (CC) region, there is also good correspondence between the observed and predicted age-composition data, with MLE estimates of the variance ranging from 0.135 to 0.201 (Fig. 2.16). There is no striking temporal pattern in the residuals for any of the fishing fleets. There is a tendency to overestimate the proportion-at-age 4 in the seine-roe fishery.

For the Strait of Georgia, there is also very good correspondence between the observed and predicted age-composition data for all three gears (Fig 2.17). The MLE estimates of the variance range from 0.089 to 0.263 for the seine-roe and winter seine fleets, respectively. In the gillnet fleet there has been a tendency to under-estimate the proportions-at-age 6-7 between the 1996 to 2011. Recall that selectivity for the gillnet fishery can be influenced by the empirical weight-at-age data, which has been trending to small fish in recent years. In this case, the age-composition data do not suggest that changes in mean weight-at-age has influenced the selectivity patterns (see results for selectivities).

In the case of WCVI, there is good correspondence between the observed and predicted age composition data for the seine fisheries and less so for the gillnet fishery (Fig 2.18). The MLE estimates of the variance range from 0.092 to 0.237 for the seine-roe and gillnet fisheries, respectively. Residual patterns in the seine fisheries and gillnet fisheries are unremarkable. The size of the residuals are fairly homogenous over time for all gears.

### 2.8.2 Biomass estimates & reference points

Maximum likelihood estimates of total biomass (age 2+) and the spawning stock biomass for each of the five major assessment regions in summarized in Figure 2.19. Estimates of spawning stock depletion ( $B_t/B_0$ ) for the five major regions is summarized in Figure 2.20 along with estimates of the sustainable fisheries framework reference points. With the exceptions of CC and WCVI, estimates of spawning stock depletion in 2011 are all currently at or above 40% of their estimated unfished state. In the CC and WCVI, spawning stock depletion is estimated to be 25% and 25% of their unfished state, respectively (Fig 2.20).

Maximum likelihood estimates of spawning stock biomass in 2011 were as follows: HG – 16,723 tonnes, PRD – 27,288 tonnes, CC – 14,624 tonnes, SOG – 129,070 tonnes, and WCVI – 14,909 tonnes (Table 2.3). These estimates are considerably higher in comparison to last years HCAM estimates; the difference largely owes to the substantial change in spawn survey scaling coefficient ( $q$ ).

In addition to the current estimates of spawning biomass, Table 2.3 also summarizes estimates of reference points and the total number of estimated parameters for each of the five major stock assessment regions. Each region contained data from 1951 to 2011, and the number of estimated parameters ranges from 159 in HG to 235 in SOG. The difference in the number of estimated parameters owes to the difference in the number of years of catch data for each region.

Estimates of unfished spawning biomass for each region is as follows: HG – 40,684 tonnes, PRD – 68,761 tonnes, CC – 59,365 tonnes, SOG – 135,523 tonnes, and WCVI – 57,462 tonnes. Applying the same cutoff

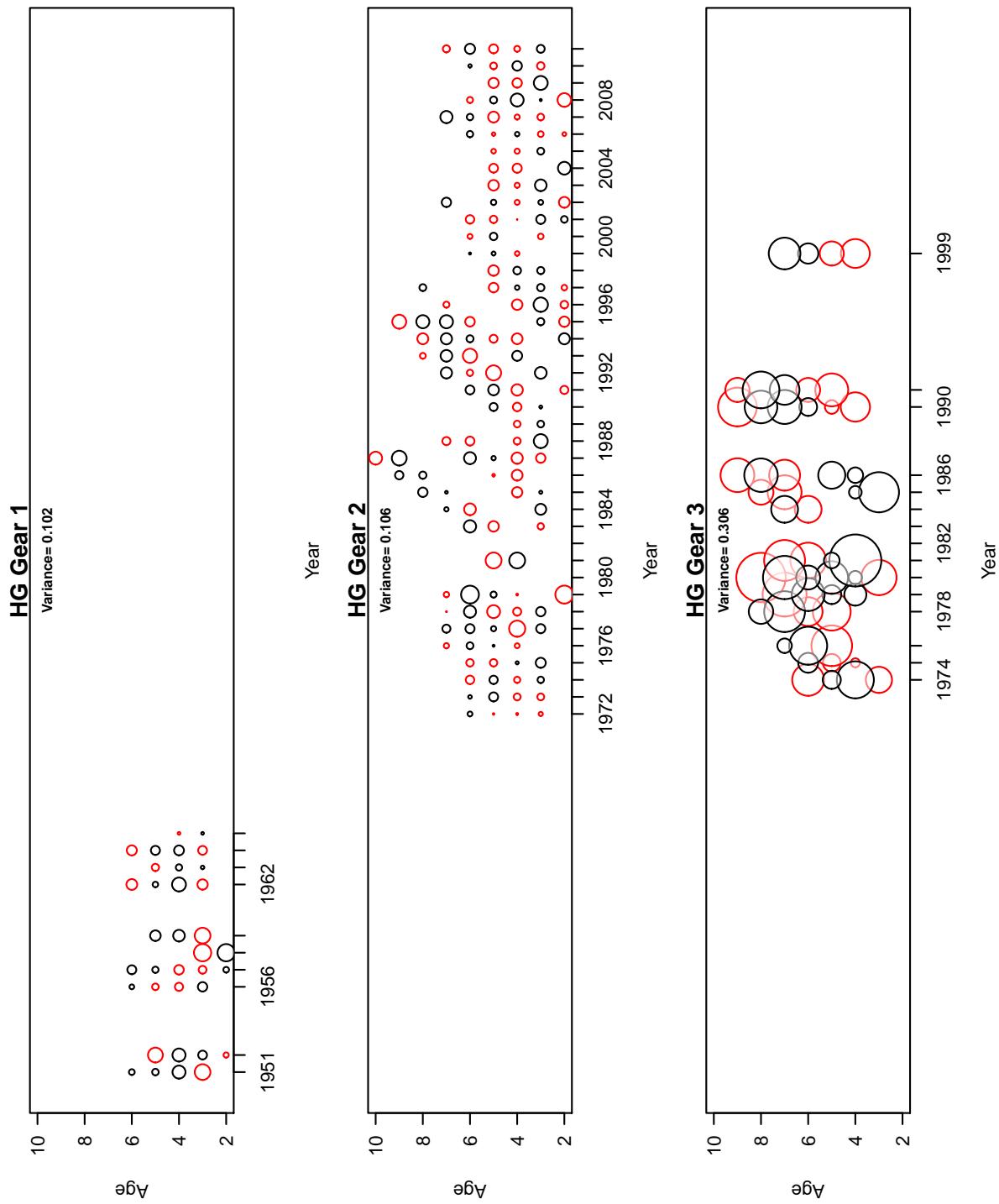


Figure 2.14: Residual difference between the observed and predicted proportions-at-age for HG for each of the three gear types (Gear 1 = winter seine, Gear 2 = seine-roe, Gear 3 = gillnet). The area of each circle is proportional to the residua, black is positive, and red is negative. The corresponding MLE estimates of the residual variance is displayed in each panel.

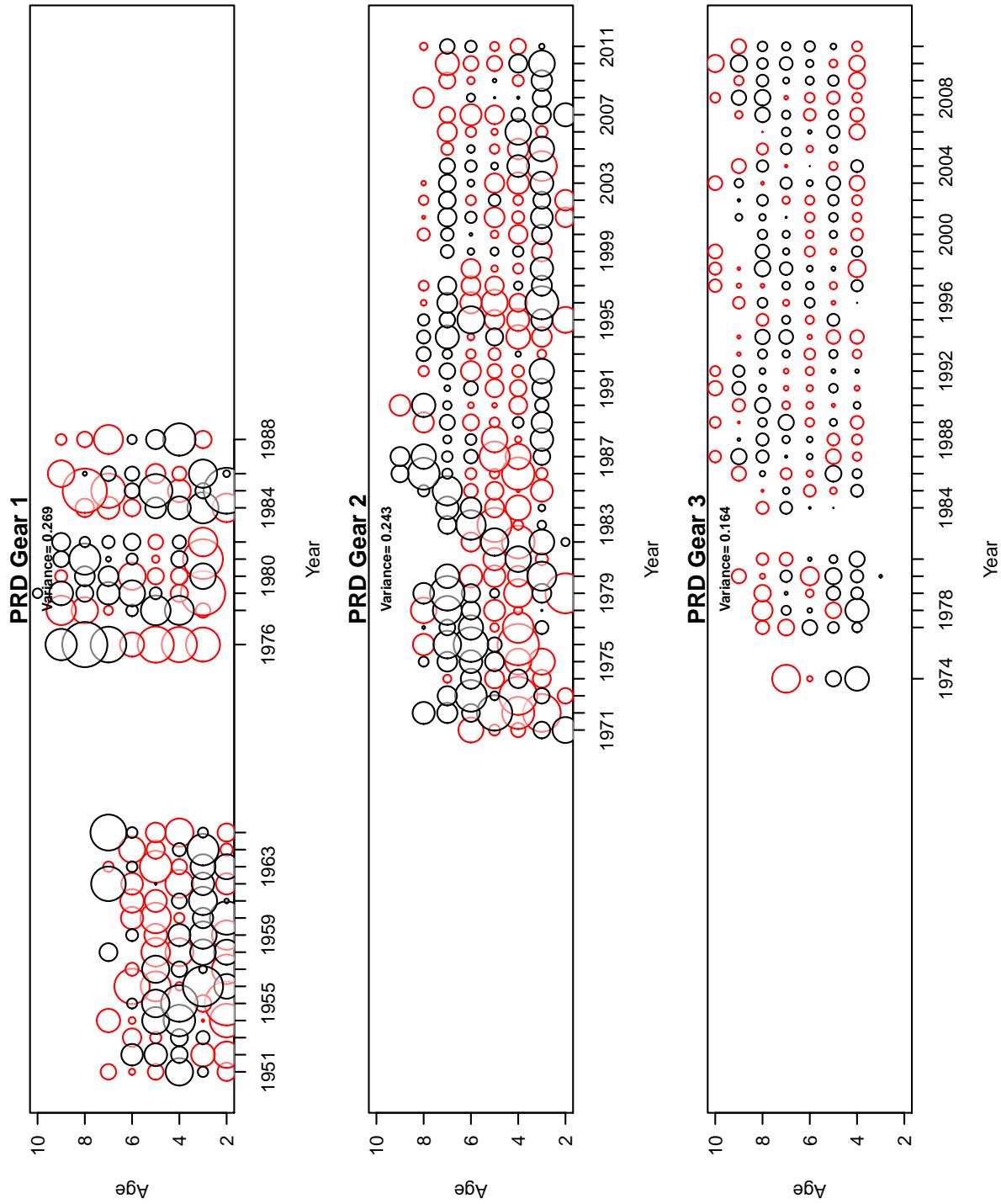


Figure 2.15: Residual difference between the observed and predicted proportions-at-age for PRD for each of the three gear types (Gear 1 = winter seine, Gear 2 = seine-roe, Gear 3 = gillnet). The area of each circle is proportional to the residua, black is positive, and red is negative. The corresponding MLE estimates of the residual variance is displayed in each panel.

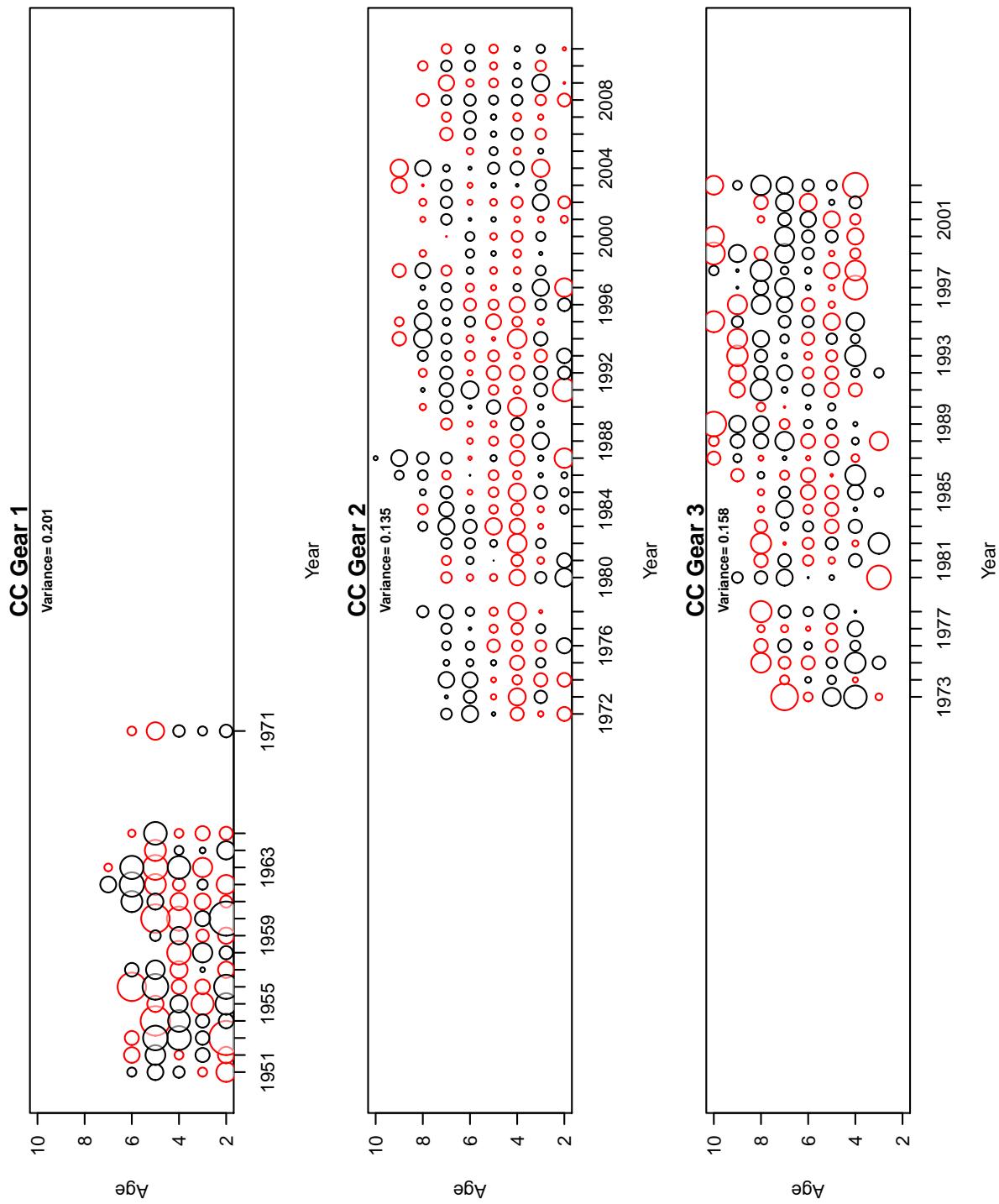


Figure 2.16: Residual difference between the observed and predicted proportions-at-age for CC for each of the three gear types (Gear 1 = winter seine, Gear 2 = seine-roe, Gear 3 = gillnet). The area of each circle is proportional to the residua, black is positive, and red is negative. The corresponding MLE estimates of the residual variance is displayed in each panel.

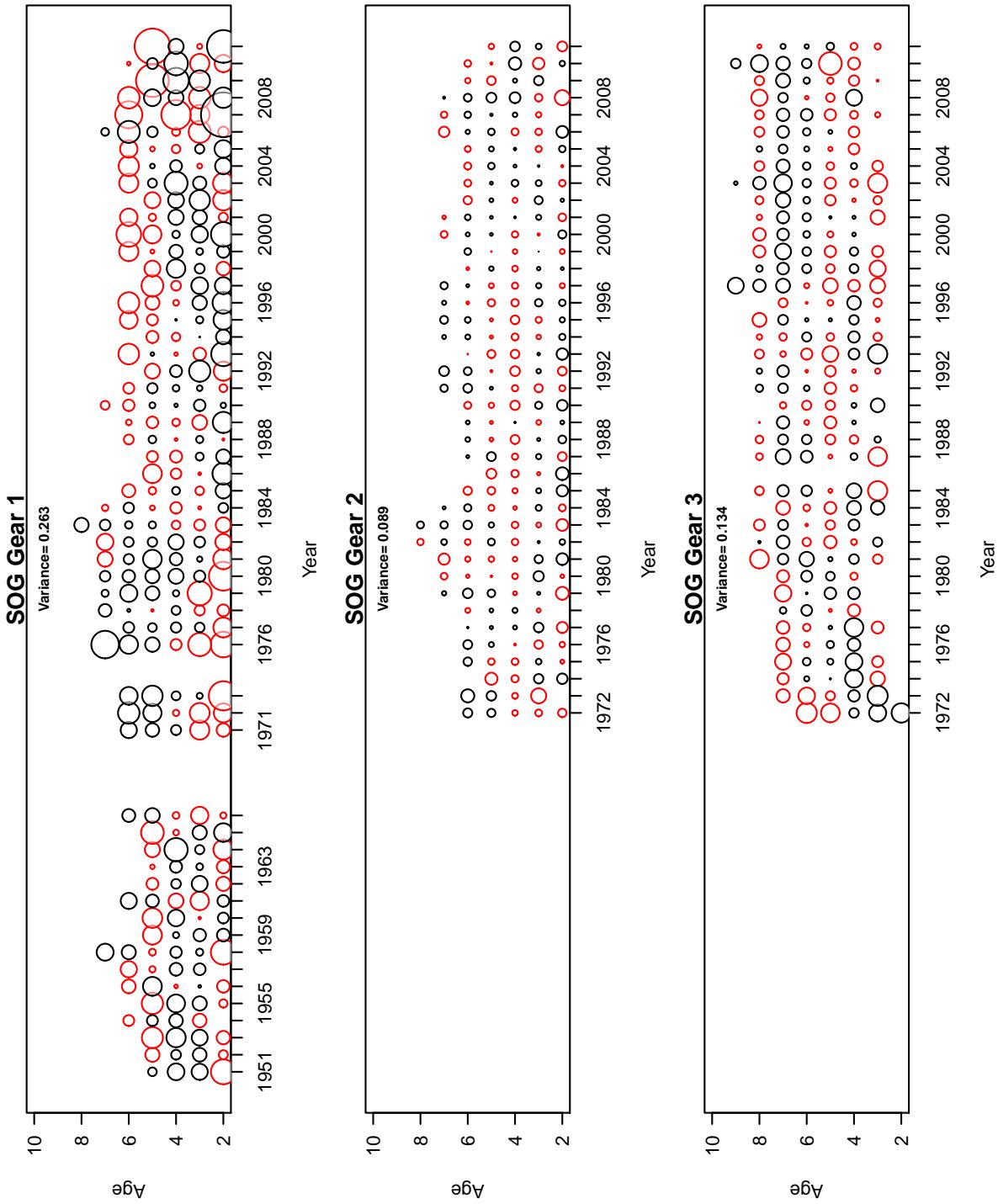


Figure 2.17: Residual difference between the observed and predicted proportions-at-age for SOG for each of the three gear types (Gear 1 = winter seine, Gear 2 = seine-roe, Gear 3 = gillnet). The area of each circle is proportional to the residua, black is positive, and red is negative. The corresponding MLE estimates of the residual variance is displayed in each panel.

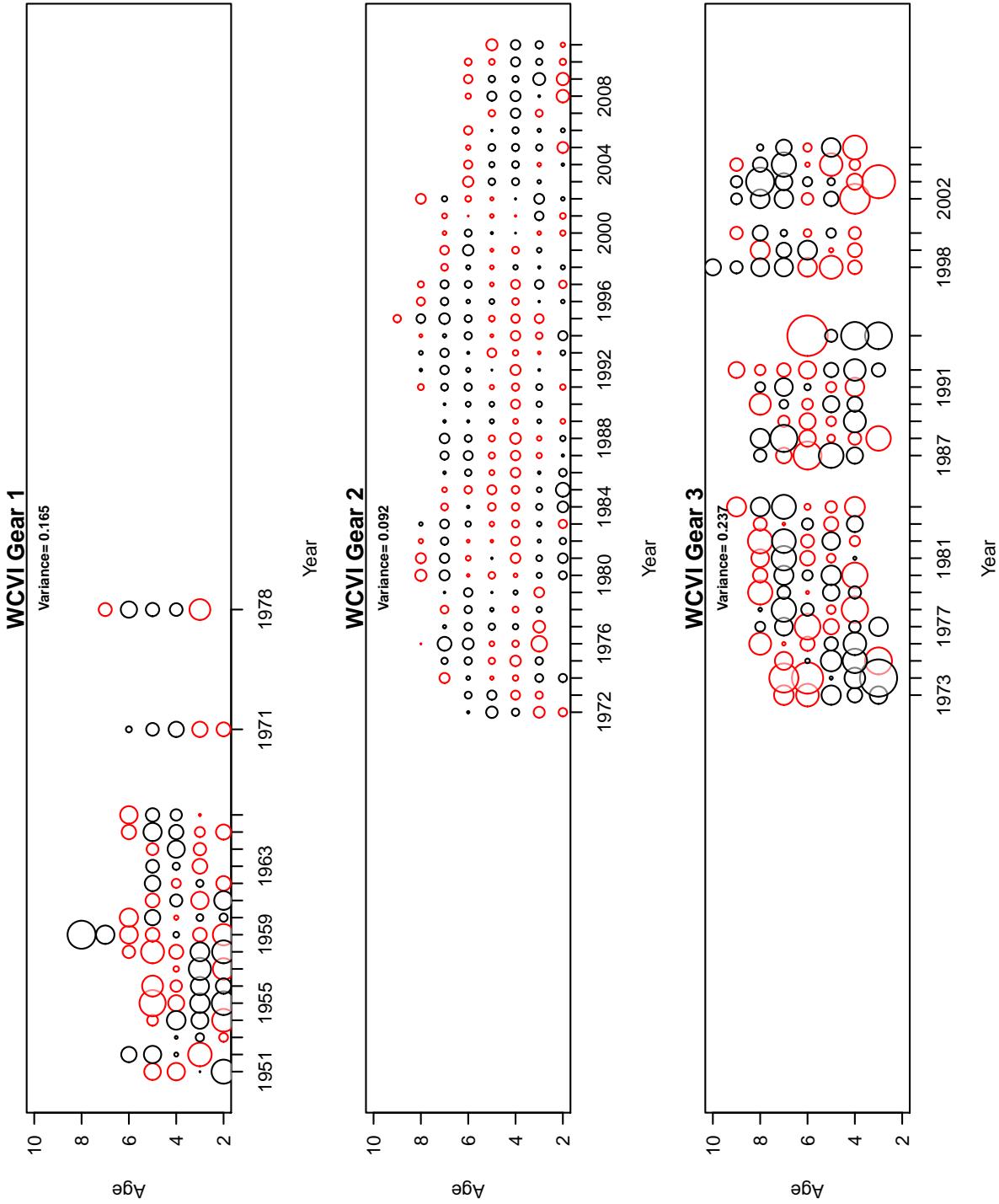


Figure 2.18: Residual difference between the observed and predicted proportions-at-age for WCVI for each of the three gear types (Gear 1 = winter seine, Gear 2 = seine-roe, Gear 3 = gillnet). The area of each circle is proportional to the residua, black is positive, and red is negative. The corresponding MLE estimates of the residual variance is displayed in each panel.

Table 2.3: Summary of maximum likelihood estimates for each of the five major stock areas. No. is the total number of estimated parameters,  $F_{MSY}$  the average instantaneous fishing rate to achieve the maximum sustainable yield (MSY),  $B_0$  is the unfished spawning biomass,  $B_{MSY}$  is the spawning biomass that achieves maximum sustainable yield,  $B_t$  is the spawning biomass at the end of the 2011 fishing season, and  $B_t/B_0$  is the spawning depletion level at the end of the 2011 fishing season.

Stock	HG	PRD	CC	SOG	WCVI
No.	159	206	190	235	174
$F_{MSY}$	2.36	0.54	1.31	1.4	0.98
MSY	8,761	6,669	9,104	27,442	10,260
$B_0$	40,684	68,761	59,365	135,523	57,462
$0.25B_0$	10,171	17,190	14,841	33,881	14,366
$B_{MSY}$	8,708	18,600	11,514	28,211	11,281
$0.8B_{MSY}$	6,966	14,880	9,211	22,568	9,025
$0.4B_{MSY}$	3,483	7,440	4,605	11,284	4,512
$B_t$	16,723	27,288	14,624	129,070	14,909
$B_t/B_0$	0.41	0.4	0.25	0.95	0.26

rule used in previous assessments (25% of  $B_0$ ), results in a substantial change in the cutoff levels for PRD, CC, SOG, and WCVI. The previous cutoffs cutoff level for HG was estimated at 10,700 tonnes, and in this assessment there is a minor downward revision to 10,171 tonnes. In the case of PRD, the previous cutoff was 12,100 tonnes and in this assessment is now 17,190 tonnes. For the CC, the previous cutoff was 17,600 tonnes and now 14,841 tonnes. For the SOG, the previous cutoff was 21,200 tonnes, and in this assessment it has been revised upwards to 33,881 tonnes. Lastly, for the WCVI the cutoff has decreased from 18,800 tonnes to 14,366 tonnes. Note however, that these revised  $B_0$ 's and cutoffs are Maximum Likelihood Estimates (MLE) and not median values from the joint posterior distribution (see section 2.8.10)

### 2.8.3 Estimates of mortality

The most recent HCAM assessment model allowed for annual estimates of  $M_t$  where natural mortality was modelled as a random walk process. The same random walk model has been adopted in this *iSCAM* implementation; however, a reduced number of parameters (12 nodes instead of 60 annual deviations) was estimated and interpolated using a bicubic spline. The number of estimated nodes does have minor influences on the various trends in natural mortality; we came to arrive at estimating 12 nodes by ensuring the estimated trends were very similar to trends in  $M$  when estimating 60 annual natural mortality rate deviations (NB. the use of formal model selection criterion should be used to determine the optimal number of nodes).

For all of the five major stock assessment regions, estimates of natural mortality rates have trended upwards since the 1950s (Figure 2.21). Trends in estimates of natural mortality are also consistent with the trends in natural mortality from last years HCAM model (see Figure 18 in Cleary and Schweigert, 2010). There was no relationship between trends in natural mortality and trends in the empirical mean weight-at-age data. Information about natural mortality comes from the age-composition data; therefore, changes in natural mortality rates over time is not necessarily linked with changes in growth rates, and vice versa. In the mid to late 1970s, estimates of natural mortality rates were very low during a time when most of the stocks were recovering from the earlier reduction fishery. In the last decade, estimates of natural mortality rates for herring have been at an all time high, and in all locations there is indication that natural mortality rates may be starting to decline. Estimates of  $M_t$  in the most recent years, however, are highly suspect because there are incomplete cohorts to infer estimates of total mortality rates and  $M_t$  is also confounded with selectivity.

Estimates of fishing mortality rates in each of the regions, between 1951 and 1970 were very high due to the reduction fishery by the winter purse seine (Gear 1). After the fishery re-opened in the early 1970s fishing mortality rates have been greatly reduced and periodic since the early 1990s due to the implementation

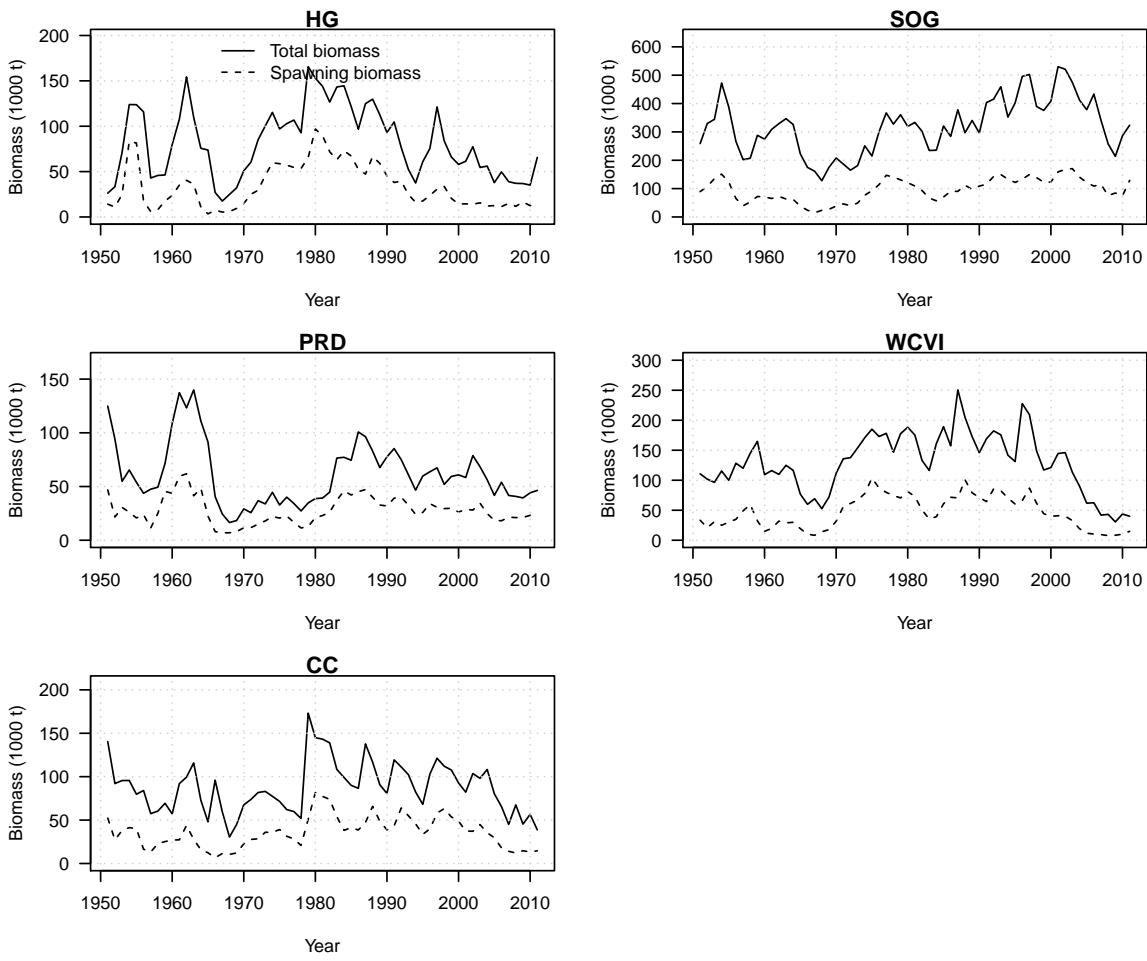


Figure 2.19: Estimates of total biomass at the start of the year (numbers times empirical weight-at-age) and spawning stock biomass (post fishery) for the five major SARs.

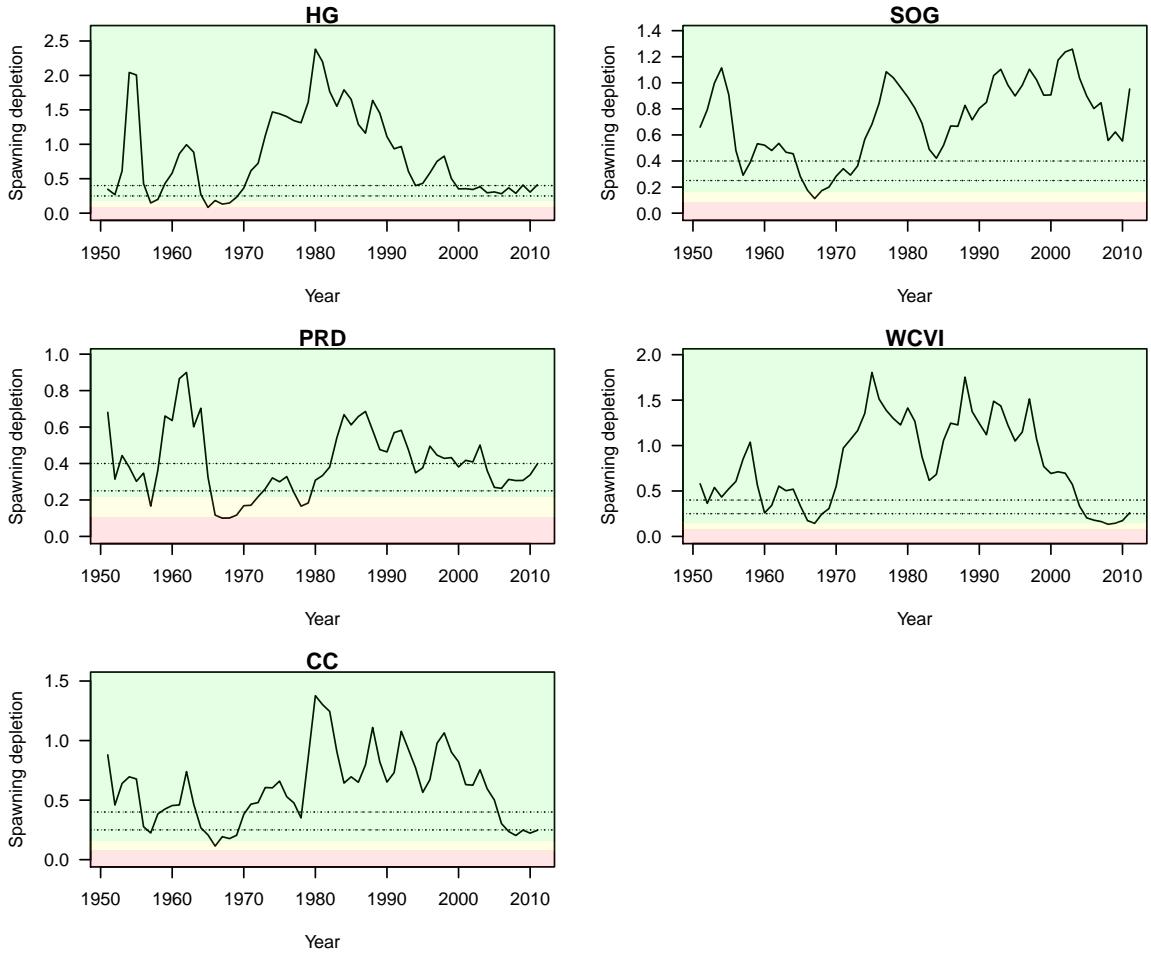


Figure 2.20: Estimates of spawning biomass depletion ( $B_t/B_0$ ) for each of the five major stock areas. Horizontal dotted lines represent 25% and 40% depletion levels, and the shaded regions demarcate reference points based on  $<40\%$   $B_{MSY}/B_0$ (critical zone) and  $40\text{--}80\%$   $B_{MSY}/B_0$ (cautious zone) and  $>80\%$   $B_{MSY}/B_0$ (healthy zone). Note that in calculating the  $B_{MSY}$  reference points, the average catch ratios over the last 20 years was used to partition fishing mortality to each of the gears.

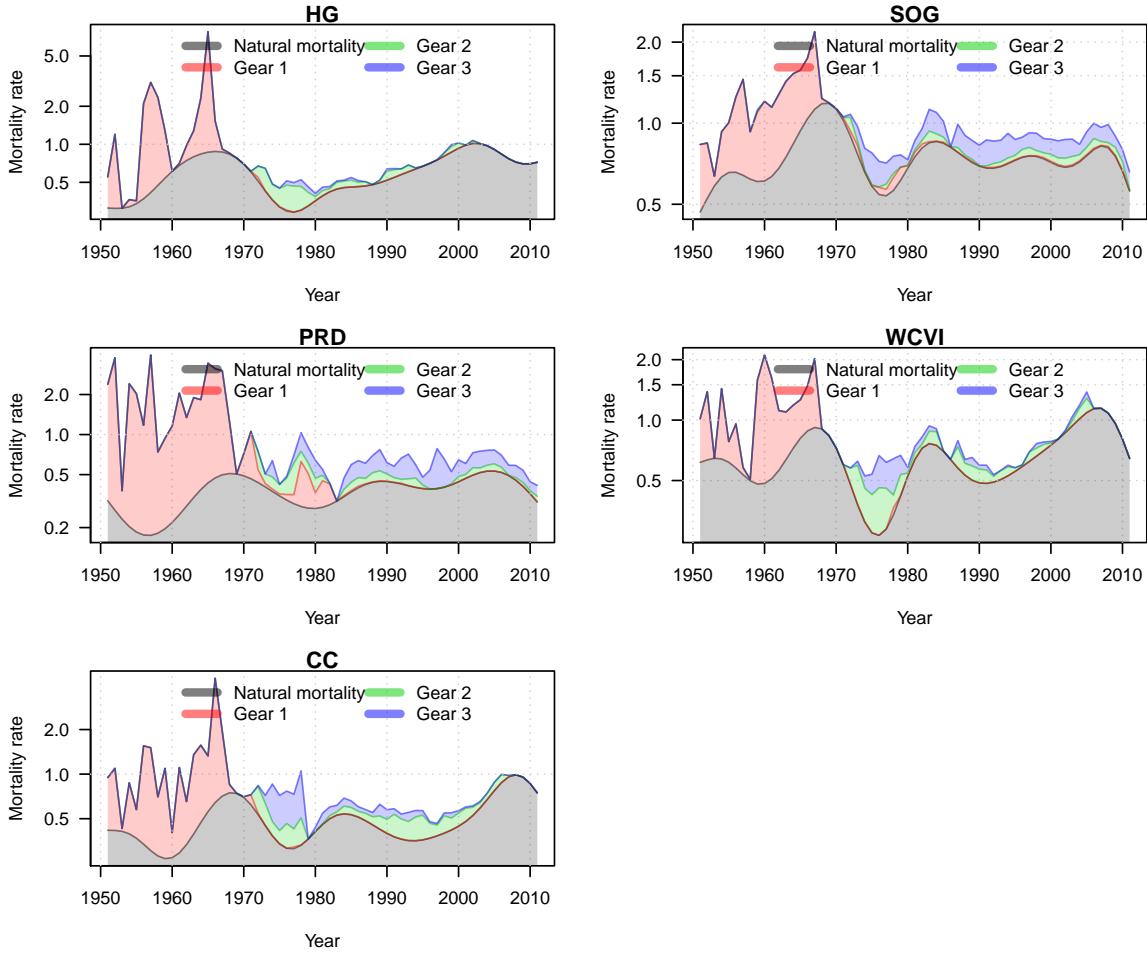


Figure 2.21: Maximum likelihood estimates of the components of average total mortality for each of the five major stock assessment regions. Note that the y-axis is plotted on a log scale, natural mortality (grey) is age-independent, fishing mortality is age-specific and the average fishing mortality rate over all age-classes is plotted here.

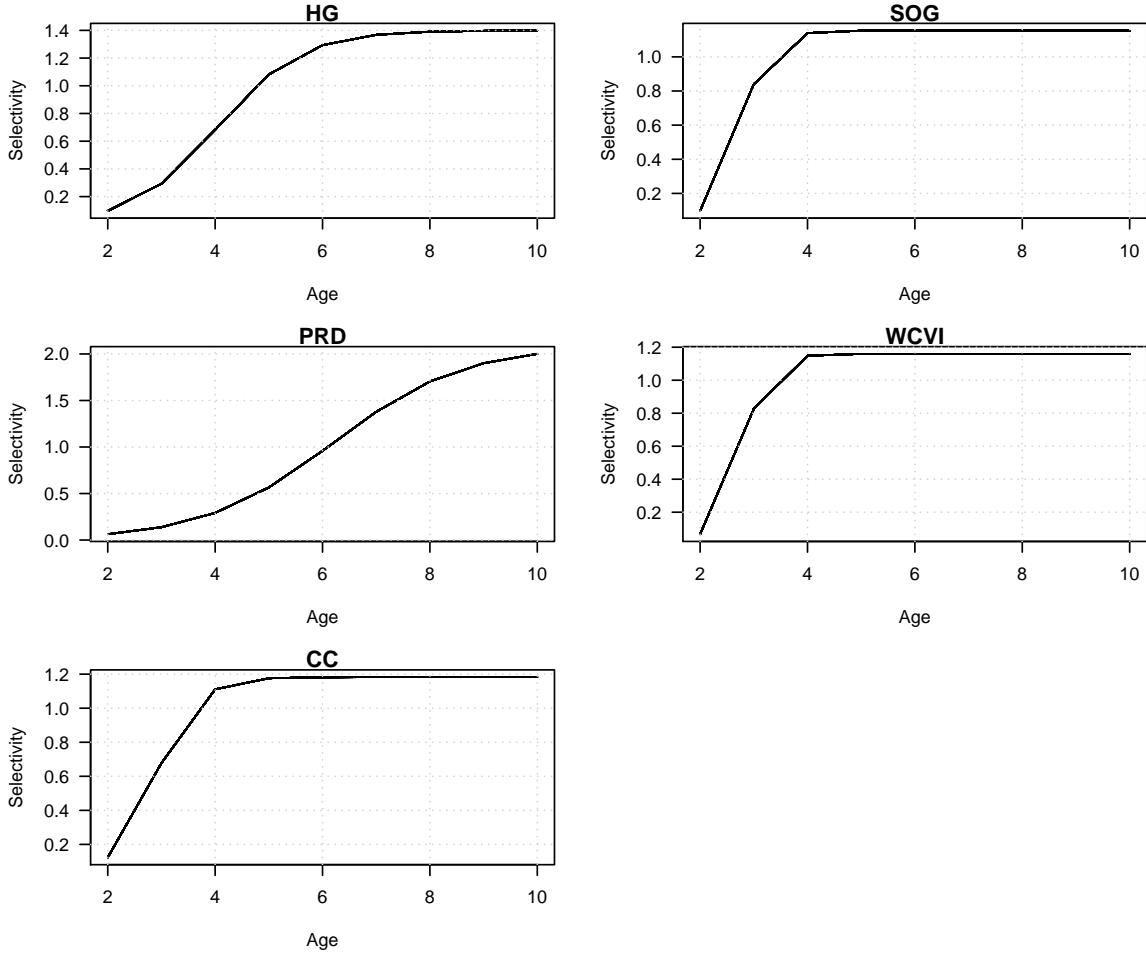


Figure 2.22: Maximum likelihood estimates of age-specific selectivity coefficients for the winter seine fishery for each of the major stock areas.

of a harvest control rule with target escapements (cutoffs). Of notable exceptions are the fishing mortality rates for the gillnet fishery in PRD and SOG have been substantially higher than other regions and consistently open each and every year (Fig. 2.21). Note that fishing mortality rates for each gear in Fig. 2.21 reflect the average fishing mortality over all age-classes and are not comparable among gears due to differences in gear selectivity. Fishing mortality rates for the gillnet fishery tend to be higher than the seine-roe fishery because recruitment to the gillnet gear is much older in comparison to the seine-roe gear.

#### 2.8.4 Selectivity

Maximum likelihood estimates of selectivity for the winter seine fishery, seine-roe fishery and the gillnet fishery for each of the five major SARs are shown in Figures 2.22, 2.23, and 2.24, respectively. Selectivities for the seine fisheries were assumed time-invariant, and selectivity for the gillnet fishery varies over time due to changes in the mean weight-at-age data.

For the winter seine fishery, age-specific selectivity coefficients were somewhat variable among the assessment regions (Fig. 2.22). The age at which herring were fully recruited to the gear was roughly age 5 for CC, SOG, and WCVI. Age at full recruitment for HG and PRD was much older, 9- and 10-years, respectively.

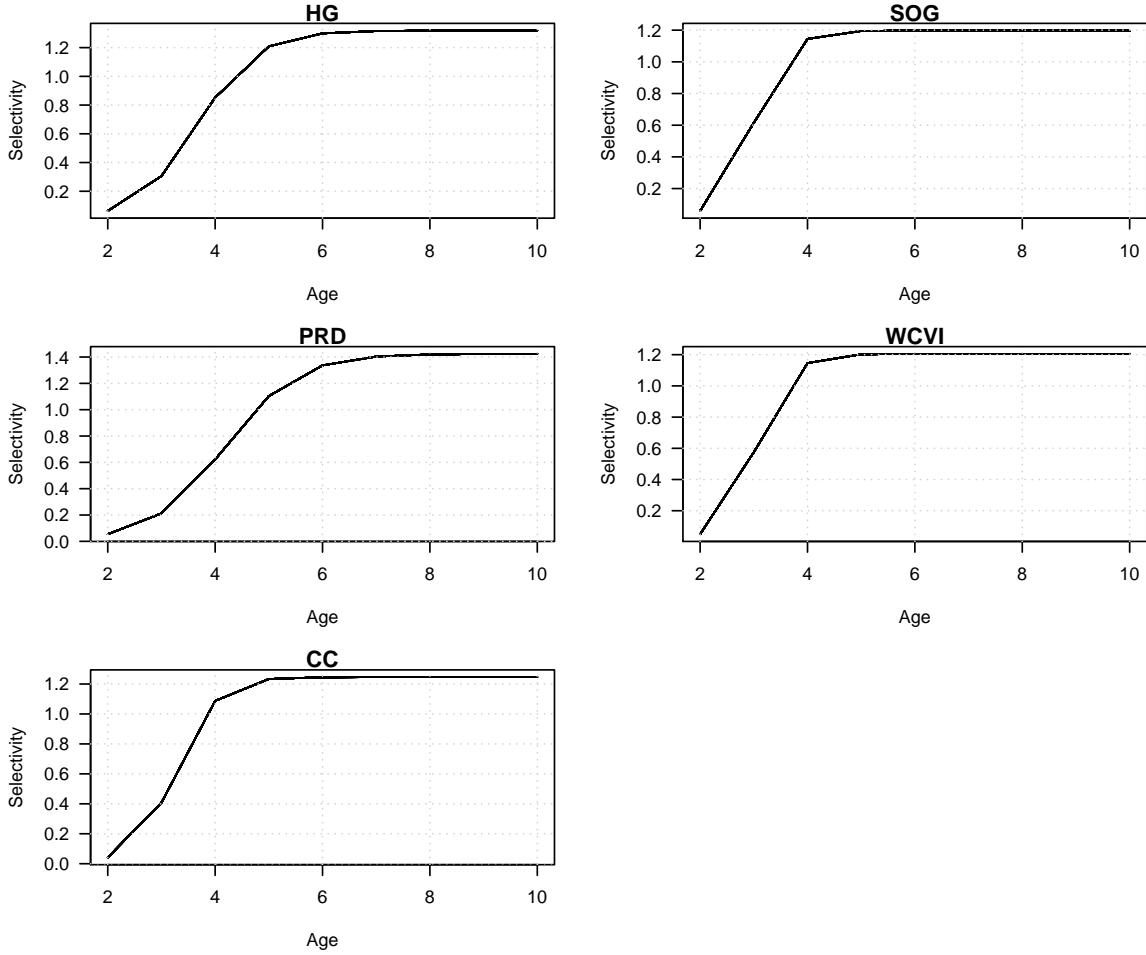


Figure 2.23: Maximum likelihood estimates of age-specific selectivity coefficients for the seine-roe fishery for each of the major stock areas.

For the seine-roe fishery, maximum likelihood estimates of selectivity were much more consistent among regions than the winter seine fishery (Fig. 2.23). Age at full recruitment to this gear type was roughly 5-6 years, and roughly the age at 50% vulnerability was roughly 3-4 years, with a tendency to recruit to the fishery at a younger age in the southern regions.

In the case of the gillnet fishery, selectivity was allowed to vary over time according to variation in the empirical weight-at-age data (Fig. 2.24). Recall that selectivity for the gillnet fishery was modelled as a logistic function of age with the addition of age-specific deviations where selectivity can increase if the weight-at-age is above average for that year. This selectivity function consists of three latent variables: two that describe the age-at-50% vulnerability and standard deviation in vulnerability-at-age, and a third parameter that describes the influence of variation in weight-at-age on departures from the logistic selectivity function ( $\lambda^{(a)}$ ). Maximum likelihood estimates for these parameters for the gillnet fishery are presented in Table 2.4. With the exception of PRD and WCVI, estimates of  $\lambda^{(a)}$  are negative and close to 0 implying no affect of variation in weight-at-age on selectivity or a slight negative effect (i.e., vulnerability to the gear declines for fish that are larger than the average weight). In the case of PRD and WCVI, the variation in weight-at-age explains approximately 4.1% and 8.7% of the residual variation in the age-composition data (Table 2.4).

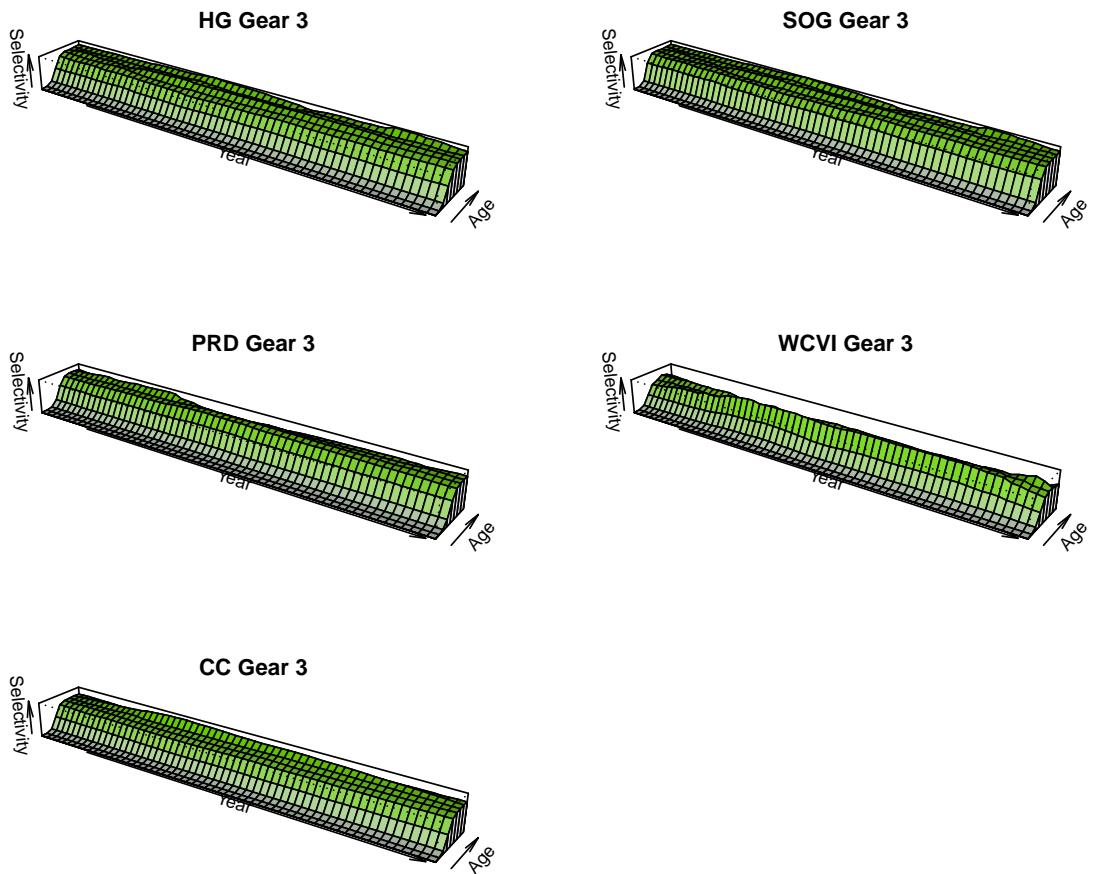


Figure 2.24: Estimates of selectivity for the gillnet fleet for each of the five major stock assessment regions. In this case selectivity is a logistic function of the empirical weight-at-age data; due to declining growth there is a tendency for selectivity to shift to older ages.

Table 2.4: Maximum likelihood estimates of gillnet selectivity parameters, where  $\mu_a$  is the age-at-50% vulnerability,  $\sigma_a$  is the standard deviation in selectivity, and  $\lambda^{(a)}$  is the coefficient that describes the influence of growth on selectivity ( $\lambda^{(a)}=0$  implies no effect,  $\lambda^{(a)} > 0$  implies a positive effect).

Stock	$\ln(\mu_a)$	$\ln(\sigma_a)$	$\lambda^{(a)}$	$\sigma_{\lambda^{(a)}}$
HG	1.598	-0.68125	-0.030581	0.045
PRD	1.727	-0.66217	0.040988	0.108
CC	1.604	-0.8050	-0.019404	0.028
SOG	1.540	-0.9797	-0.02835	0.068
WCVI	1.608	-0.6647	0.08677	0.014

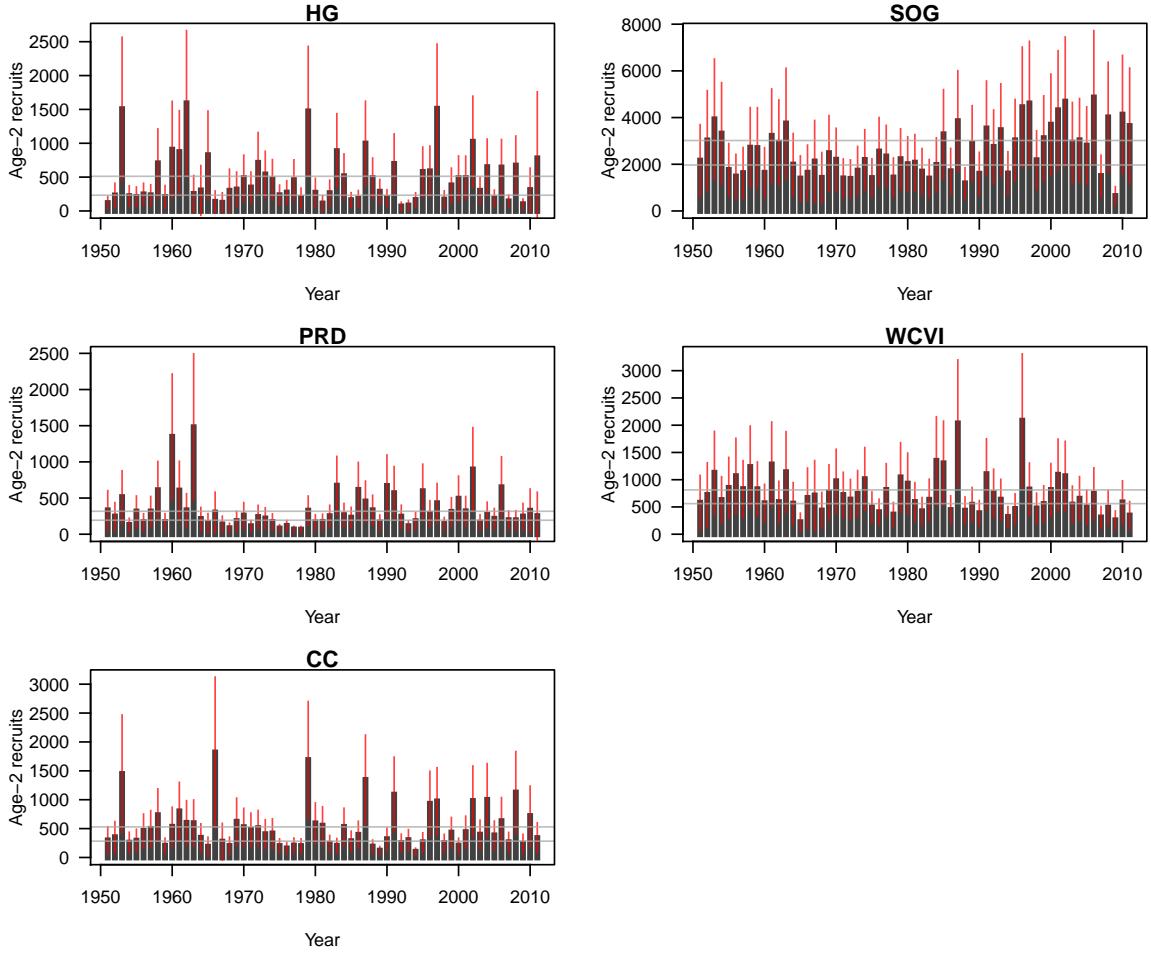


Figure 2.25: Maximum likelihood estimates of age-2 recruits for each of the five major stock areas and the asymptotic estimates of the 95% confidence interval. The horizontal divisions demarcate the 0.33 and 0.66 quantiles that define poor, average, and good recruitment.

### 2.8.5 Recruitment and stock-recruitment relationships

Recruitment to each stock is defined as the number of age-2 fish entering the population at the beginning of each year (i.e., May 1). Age-2 recruitment is estimated as a free parameter within  $i\text{SCAM}$ , subject to the constraint that annual estimates vary around a Beverton-Holt stock recruitment relationship with an estimated unknown standard deviation. Maximum likelihood estimates of age-2 recruits are shown in Figure 2.25 along with horizontal lines that demarcate the 0.33 and 0.66 quantiles that was traditionally used to categorize recruitment as poor, average, and good in previous assessments.

Estimates of age-2 recruits for 2010 and 2011 were average and good in HG, average in PRD, good and average in CC, good in SOG, and average and poor in the WCVI region. Note however, that estimates of age-2 recruits are highly uncertain in the most recent years of a stock assessment because these age-classes are only partially recruited to the fishing gears.

The underlying stock-recruitment relationship is key for determining reference points for this stock. Maximum likelihood estimates of the age-2 recruits versus spawning biomass, along with the corresponding Beverton-Holt stock recruitment model are shown in Figure 2.26. The Beverton-Holt stock recruitment

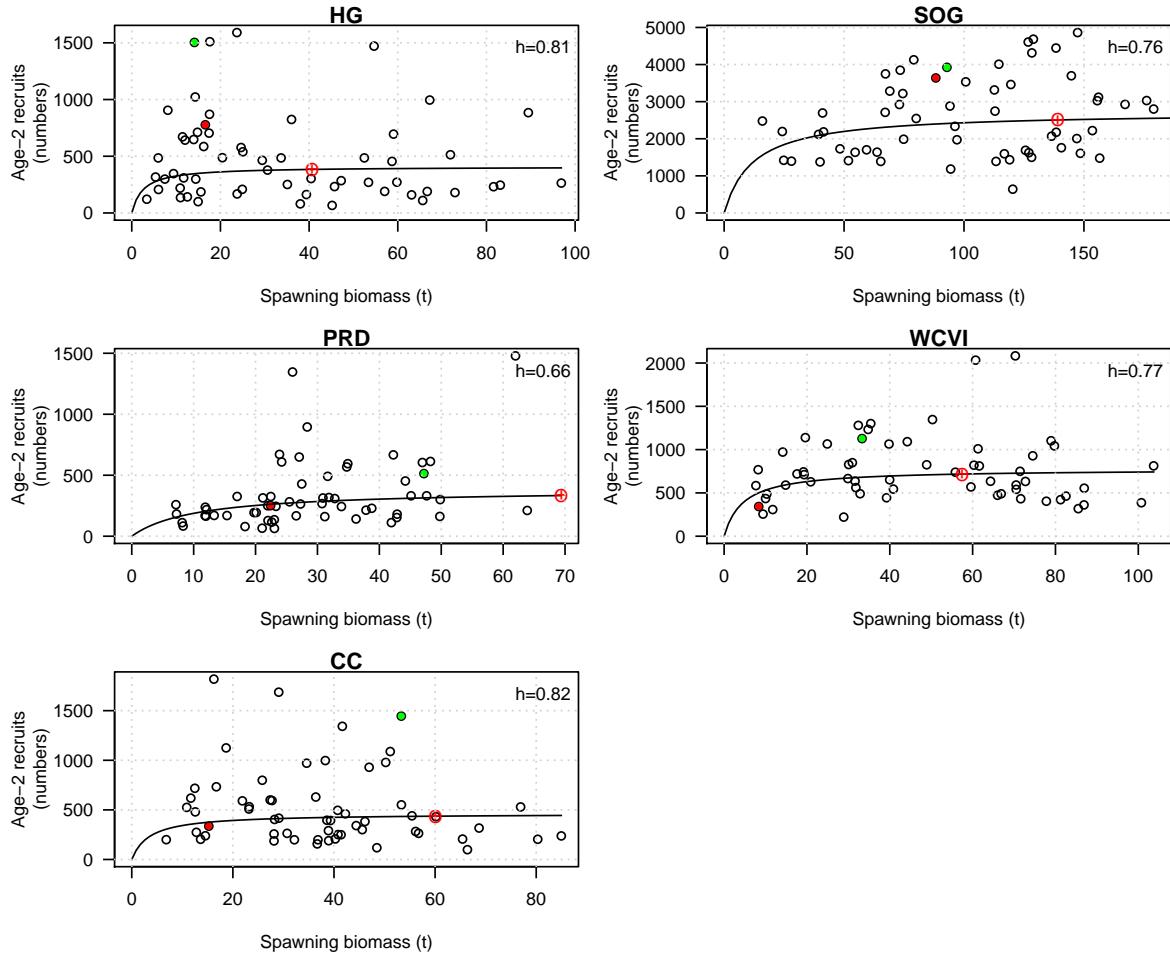


Figure 2.26: Maximum likelihood estimates of age-2 recruits versus estimated spawning stock biomass in each of the five major assessment regions. The green and red circles indicate the start (recruits in 1952) and end (recruits in 2011) of the series, the circle plus (red) corresponds to the maximum likelihood estimate of unfished spawning biomass ( $B_0$ ) and unfished age-2 recruitment  $R_0$ , the line is the Beverton-Holt stock recruitment model fitted to these data.

model was jointly fitted to these data by estimating the steepness of the stock recruitment relationship ( $h$ ) and the unfished age-2 recruits ( $R_0$ ). The unfished spawning biomass was determined by using the average fecundity and average natural mortality rates (from 1951-2011) to calculate the average spawning biomass per recruit. Alternative stock-recruitment models (e.g., Ricker model) were not explored to determine if they provided a better fit.

Between 1951 and 2011, four of the five major stock areas have fluctuated above the estimate of unfished spawning biomass; the exception is the PRD area. In HG, age-2 recruitment has been remarkably stable over a very wide range of spawner abundance. This is also the case for CC and WCVI. In PRD and SOG, variation in recruitment appears to be lower at low spawning abundance and the average recruitment rate tends to drop. Maximum likelihood estimates for steepness for these five stocks are as follows: HG – 0.81, PRD – 0.66, CC – 0.82, SOG – 0.76, WCVI – 0.77.

The log residual differences between the estimated age-2 recruits and that predicted by the estimated spawning-biomass and Beverton-Holt model for each of the major stock areas is shown in Figure 2.27. There

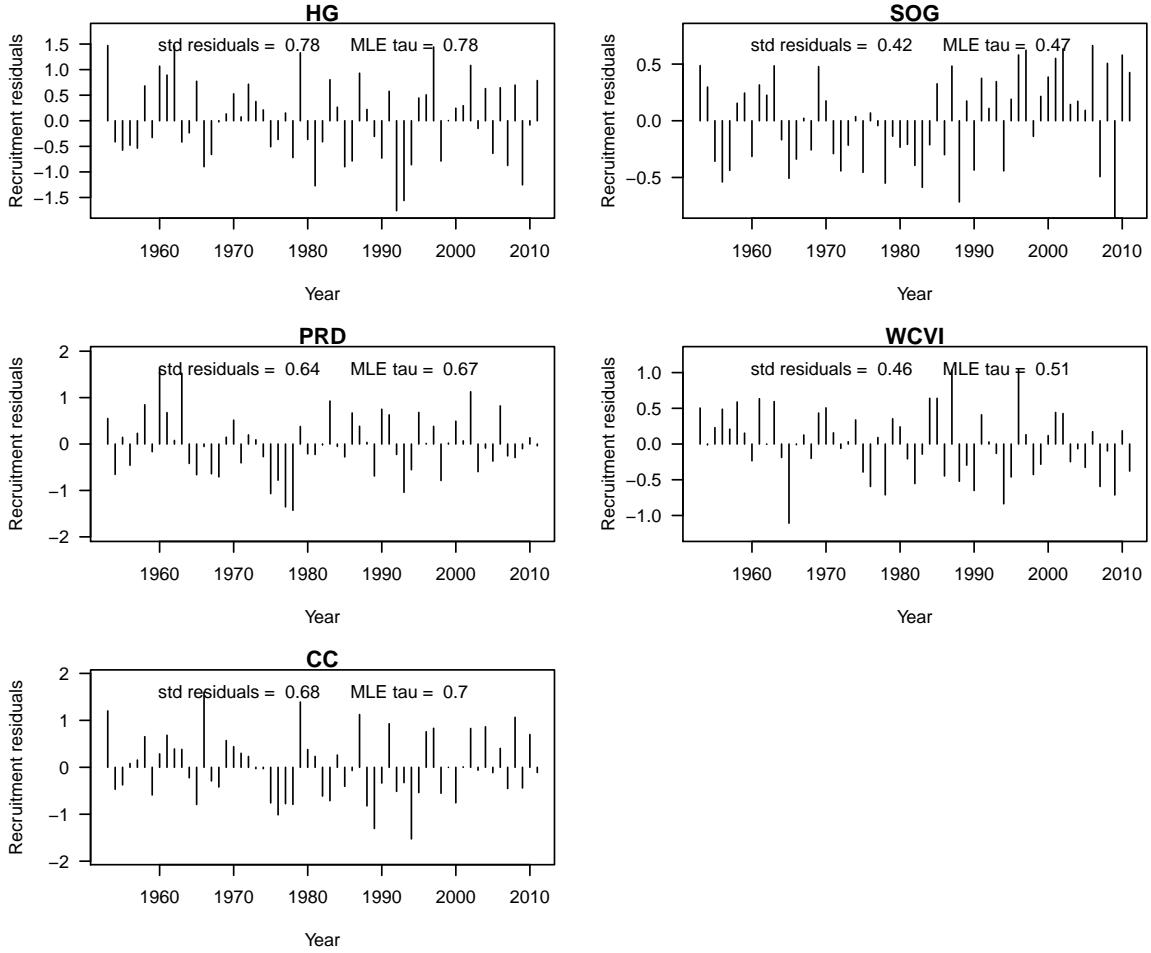


Figure 2.27: Log residual differences between estimated age-2 recruits and the recruitment predicted by the Beverton-Holt model and estimated spawning stock biomass. The standard deviations of the residuals along with the MLE estimate of the process error standard deviations are displayed at the top of each panel.

is no strong autocorrelation in recruitment, except perhaps 5-8 year periods of poor and good recruitment in SOG. There is good correspondence between the standard deviations of the residuals and the estimated standard deviation of the process error variance ( $\tau$ ).

### 2.8.6 Retrospective analysis

Four of the five major regions contained little to no retrospective bias in the estimates of spawning stock biomass when fitting the data back to 2001 (10 years, Figure 2.28). The PRD region does show a strong retrospective bias; as each year of data is removed estimates of the terminal spawning biomass that year increase. This pattern of declining estimates of biomass as data are added is persistent for all of the years in which the retrospective estimates were examined which implies that estimates of biomass in this region are positively biased. The direction of bias in PRD warrants extra caution as uncertainty in biomass estimates are also likely biased upwards.

In SOG, there was no retrospective bias; however, there is some retrospective error that occurs when data from 2008 and onwards are removed from the analysis. Information in the data prior to 2008 suggest

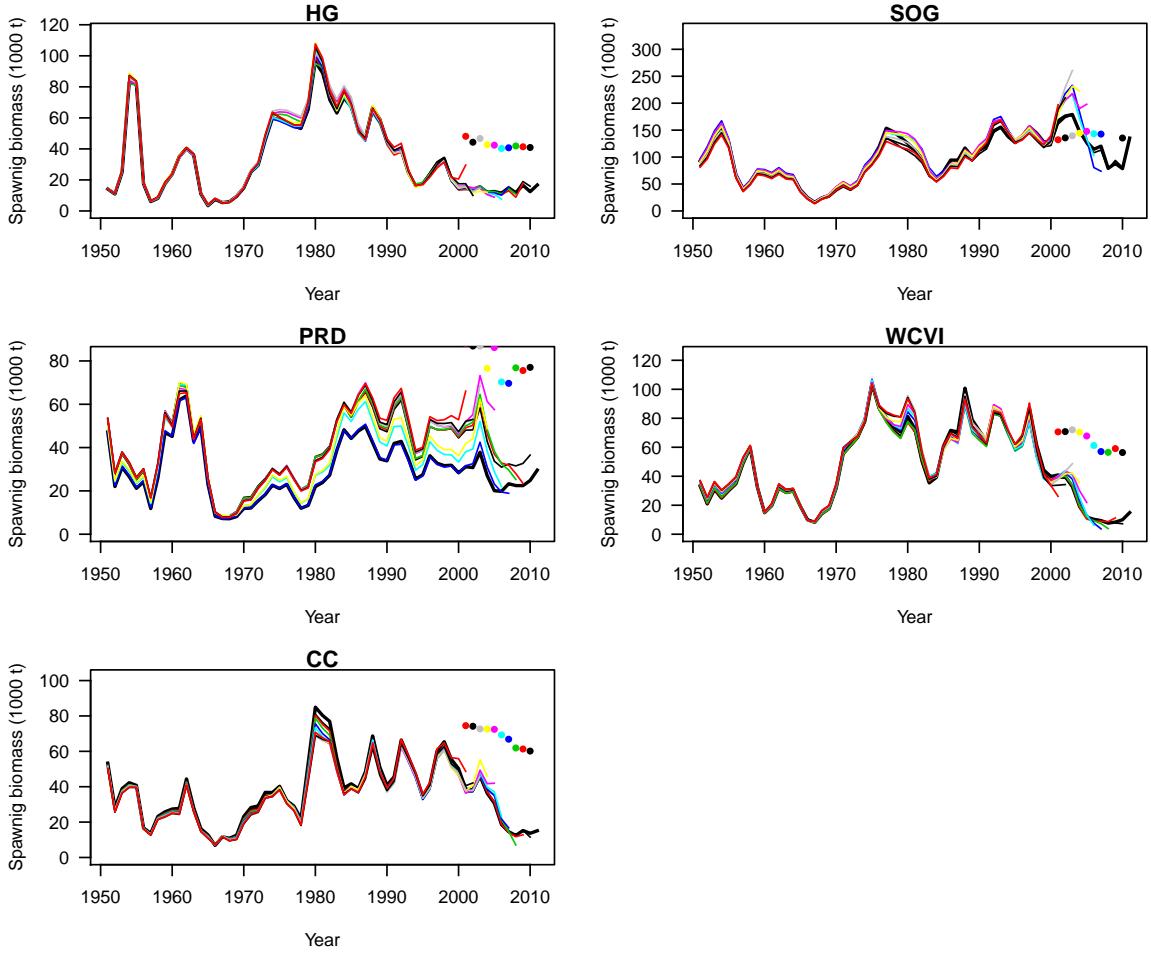


Figure 2.28: Retrospective estimates of spawning stock biomass and estimates of unfished biomass ( $B_o$ , shown as circles) for each of the five major stock assessment areas. The model was sequentially fitted to the full data set, then from 1951:2010, 1951:2009, ... 1951:2001.

large increases in spawning abundance in the early 2000s. This increase is revised downwards when the 2008 data are included in the assessment. It is not clear what components of the data suggest this downward revision.

Key to the harvest control rule is the estimate of unfished spawning biomass in each of the SARs. The points plotted in Figure 2.28 show how estimates of  $B_o$  change as more data accumulates over time. If estimates of  $B_o$  increase over time, then cutoff levels would also increase, and vice versa. In all regions with the exception of PRD, retrospective estimates of  $B_o$  have been relatively consistent and trending downwards slightly as more and more data have accumulated over time. Retrospective estimates of  $B_o$  in PRD are much more variable in comparison to other areas. A more conservative policy would be to set cutoff levels at the upper end of the  $B_o$  estimates for this region.

### 2.8.7 Marginal posterior distributions

Marginal posterior distributions for estimated model parameters were constructed using AD Model Builders built in Metropolis-Hastings algorithm (Gelman, 2004). For each of the major and minor assessment areas,

Table 2.5: Estimates of 2011 spawning biomass,  $B_0$ , and depletion based on 2000 systematic samples from the joint posterior distribution drawn from a chain of length 1,000,000.

Stock	$SB_{2011}$			$B_0$			$SB_{2011}/B_0$		
	Median	2.5%	97.5%	Median	2.5%	97.5%	Median	2.5%	97.5%
HG	<b>16.58</b>	7.70	33.63	<b>41.74</b>	30.05	61.51	<b>0.39</b>	0.19	0.75
PRD	<b>27.05</b>	14.45	50.59	<b>78.56</b>	54.15	150.18	<b>0.34</b>	0.15	0.68
CC	<b>14.67</b>	7.28	27.28	<b>62.40</b>	48.47	85.06	<b>0.23</b>	0.12	0.41
SOG	<b>125.14</b>	70.43	217.95	<b>140.05</b>	110.47	184.24	<b>0.89</b>	0.53	1.45
WCVI	<b>14.68</b>	6.99	27.63	<b>59.58</b>	46.84	78.53	<b>0.24</b>	0.12	0.43

a systematic sample of 2,000 points from a chain of length 1,000,000 and is intended to represent a random sample from the joint posterior distribution. These samples were then used to construct marginal distributions for derived quantities (e.g.,  $B_0$ ). All areas with the exception of the SOG used the inverse Hessian matrix as the jumping distribution. In the case of SOG, the hessian matrix had to be re-scaled (using the `-mcmult 2.0` option in ADMB) in order to invert the Hessian matrix.

### Diagnostic trace plots

No formal statistical tests were carried out to determine if the samples from the joint posterior distribution were taken from a converged distribution. Visual inspection was used to determine overall convergence and the trace plots for each of the five major regions are shown in Figures 2.29–2.33.

### 2.8.8 Parameter confounding

To examine the level of confounding among the estimated parameters, 200 randomly selected points from the joint posterior distribution for the seven leading parameters were plotted against each other in a pairs plot (e.g., Fig. 2.34). Only 200 points were plotted to reduce the file size. Among the seven leading estimated parameters ( $R_0, h, M, \bar{R}, \ddot{R}, \rho, \vartheta$ ) there was very little confounding (Figures 2.34 – 2.38).

There is, however, some strong confounding between the estimated parameters and a few of the derived reference points. In all major areas, there was a strong positive correlation between steepness ( $h$ ) and  $F_{MSY}$ ; similarly there is a strong positive correlation between  $B_0$  and  $R_0$ . Among the reference points alone, there is a negative correlation between  $B_{MSY}$  and  $F_{MSY}$ , and a positive correlation between  $F_{MSY}$  and  $MSY$ . This level of confounding among the derived variables is not cause for concern from a parameter estimation standpoint; it does, however, highlight the tradeoffs that must be made from a decision makers perspective.

### 2.8.9 Marginal posterior distributions

Marginal posterior distributions and along with the prior densities for the seven leading parameters are shown in Figure 2.39. In all cases, the steepness parameter, followed by the instantaneous natural mortality rate appears to be the most influenced by the prior density. Uniform prior distributions were assumed for the scaling parameters ( $R_0, \bar{R}$ , and  $\ddot{R}$ ). There were good posterior updates for the total variance and variance portioning parameters ( $\vartheta, \rho$ ).

Median estimates of the 2011 spawning stock biomass and the 95% credible interval for each of the five major assessment regions are summarized in Table 2.5. Current estimates of depletion and the corresponding uncertainty is based on the ratio of the 2011 spawning biomass divided by the corresponding estimate of  $B_0$ . Depletion levels less than 0.25 would be considered to be below the cutoff level.

### 2.8.10 Forecast and catch advice based on the joint posterior distribution

Catch advice has historically been provided in the form of a decision table based on median values of the joint posterior distribution. The decision table contains columns specifying the 2011 SSB the age 4+

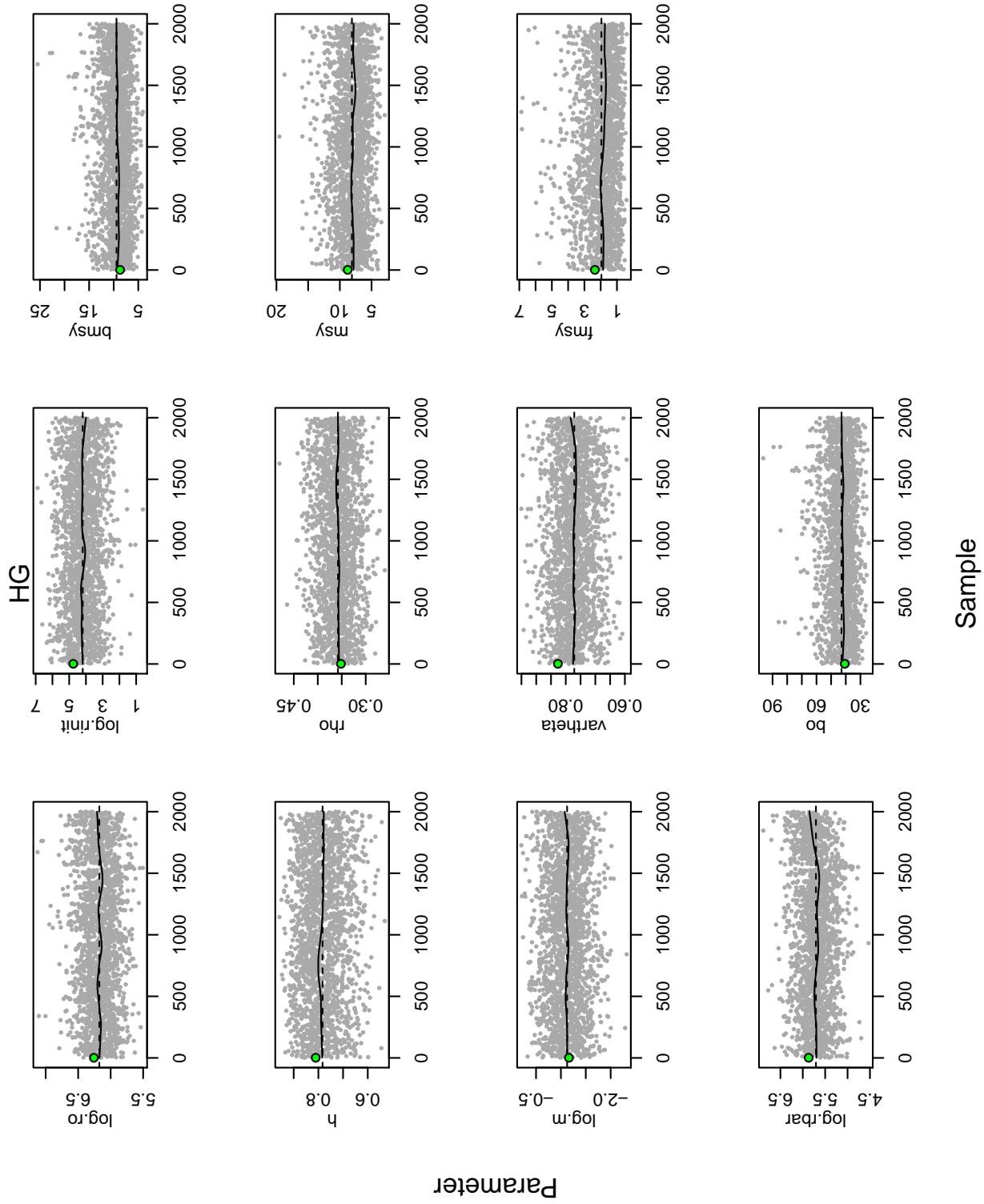


Figure 2.29: A systematic sample of 2,000 from an MCMC chain of length 1,000,000 of leading parameters and derived variables used in reference point calculations for HG. Green circle corresponds to the MLE estimates and the solid line is a lowess smooth fit to the data ( $f=1/4$ ), and the dashed line is the mean of the distribution.

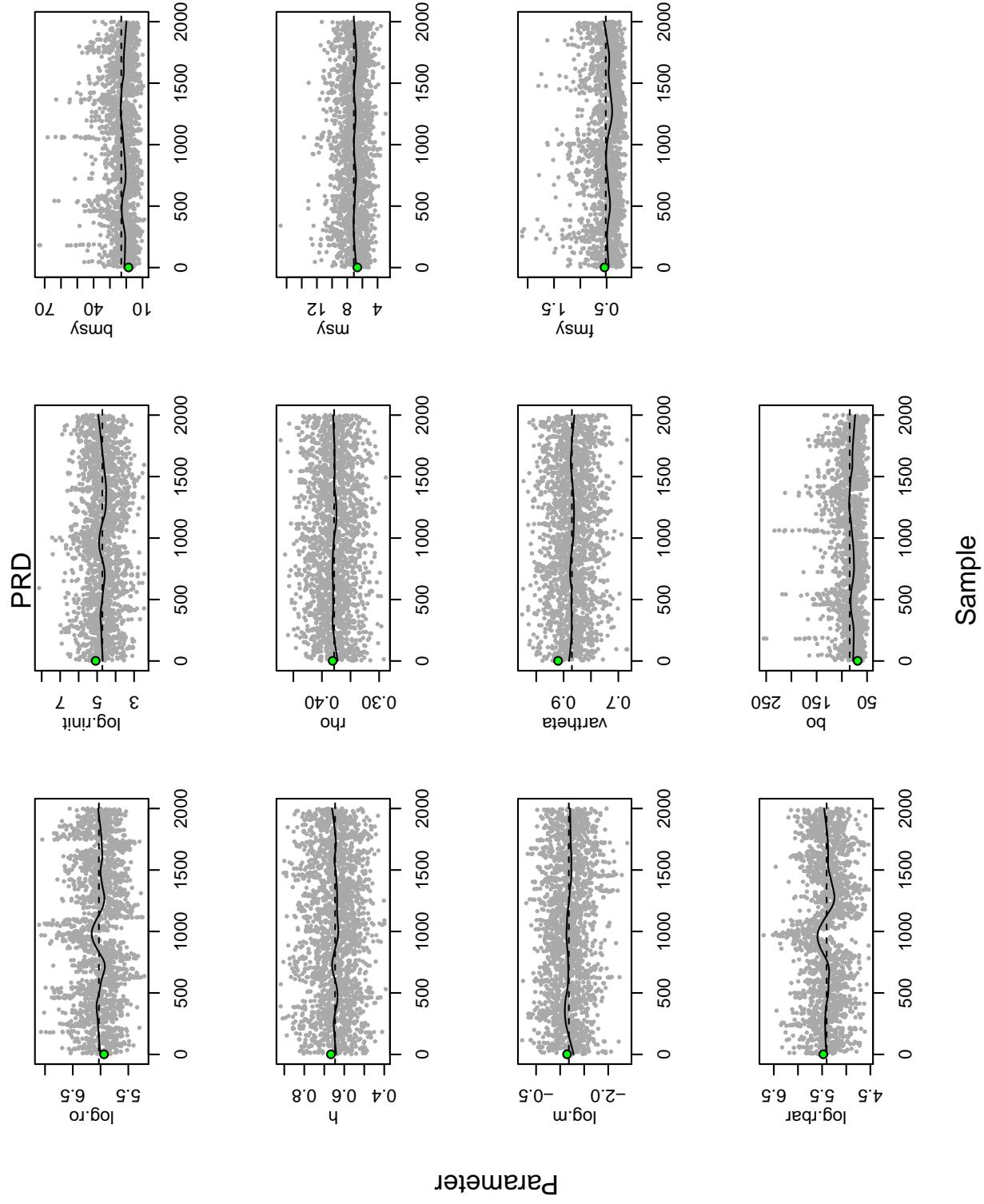


Figure 2.30: A systematic sample of 2,000 from an MCMC chain of length 1,000,000 of leading parameters and derived variables used in reference point calculations for PRD. Green circle corresponds to the MLE estimates and the solid line is a lowess smooth fit to the data ( $f=1/4$ ), and the dashed line is the mean of the distribution.

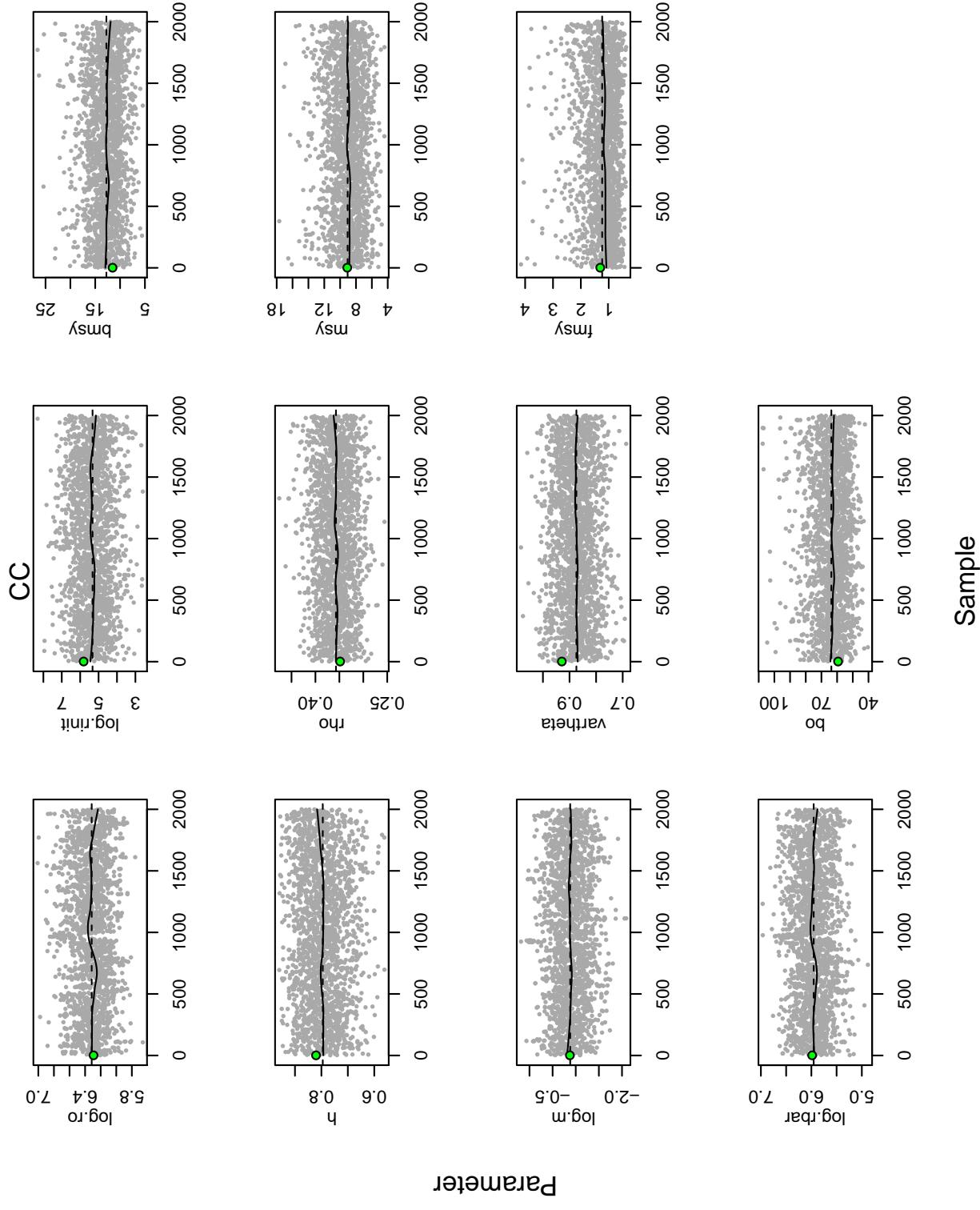


Figure 2.31: A systematic sample of 2,000 from an MCMC chain of length 1,000,000 of leading parameters and derived variables used in reference point calculations for CC. Green circle corresponds to the MLE estimates and the solid line is a lowess smooth fit to the data ( $f=1/4$ ), and the dashed line is the mean of the distribution.

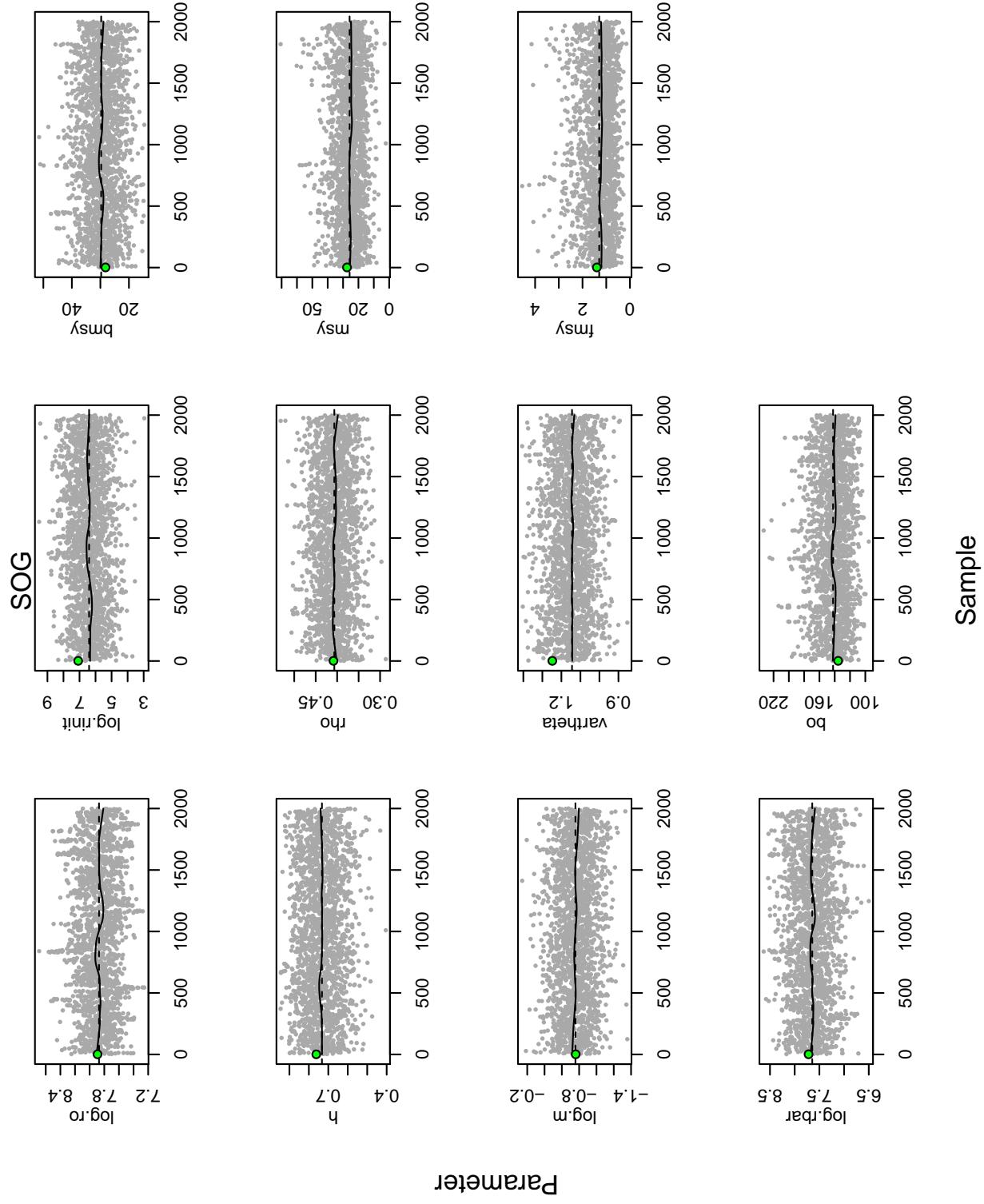


Figure 2.32: A systematic sample of 2,000 from an MCMC chain of length 1,000,000 of leading parameters and derived variables used in reference point calculations for SOG. Green circle corresponds to the MLE estimates and the solid line is a lowess smooth fit to the data ( $f=1/4$ ), and the dashed line is the mean of the distribution.

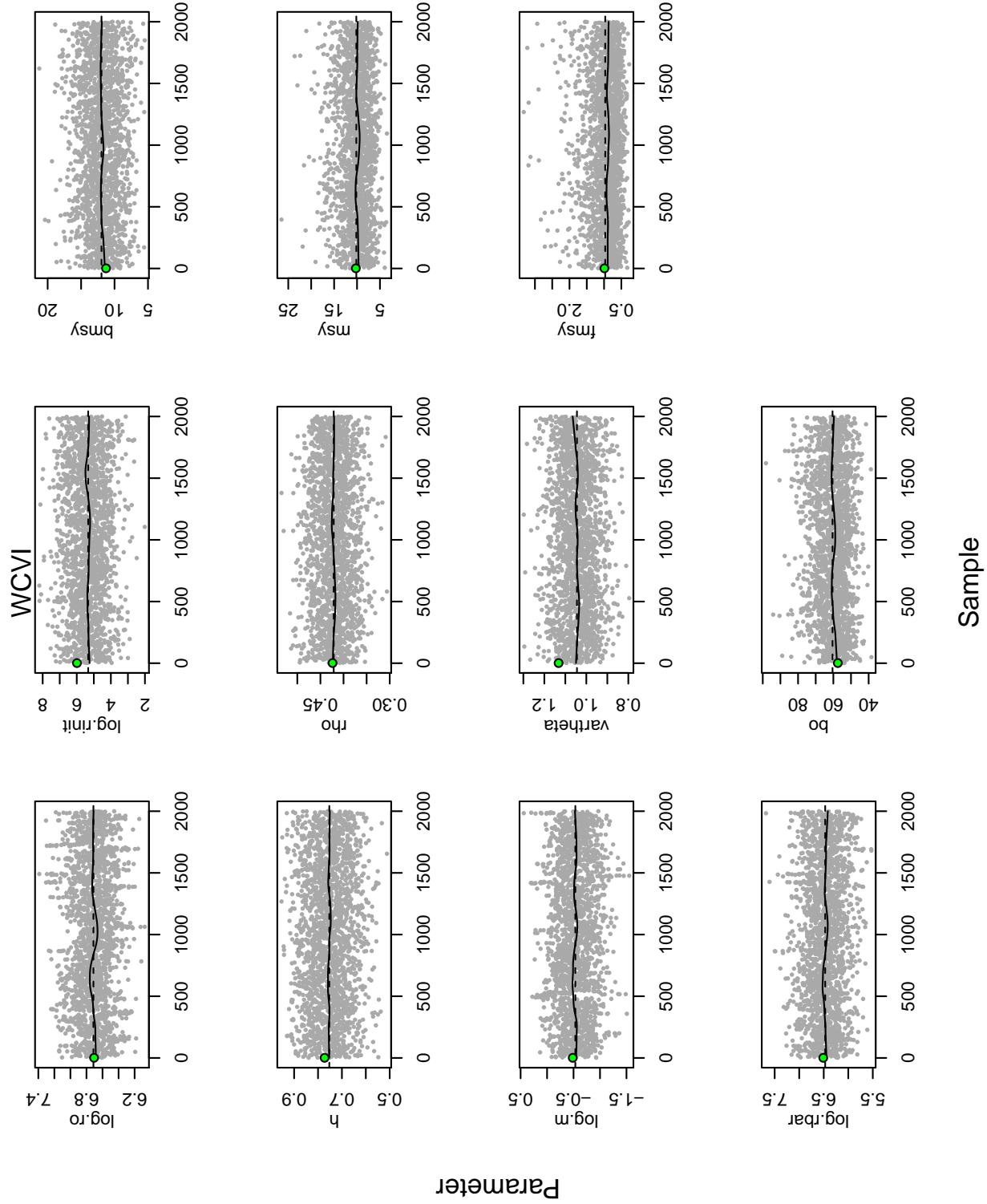


Figure 2.33: A systematic sample of 2,000 from an MCMC chain of length 1,000,000 of leading parameters and derived variables used in reference point calculations for WCVI. Green circle corresponds to the MLE estimates and the solid line is a lowess smooth fit to the data ( $f=1/4$ ), and the dashed line is the mean of the distribution.

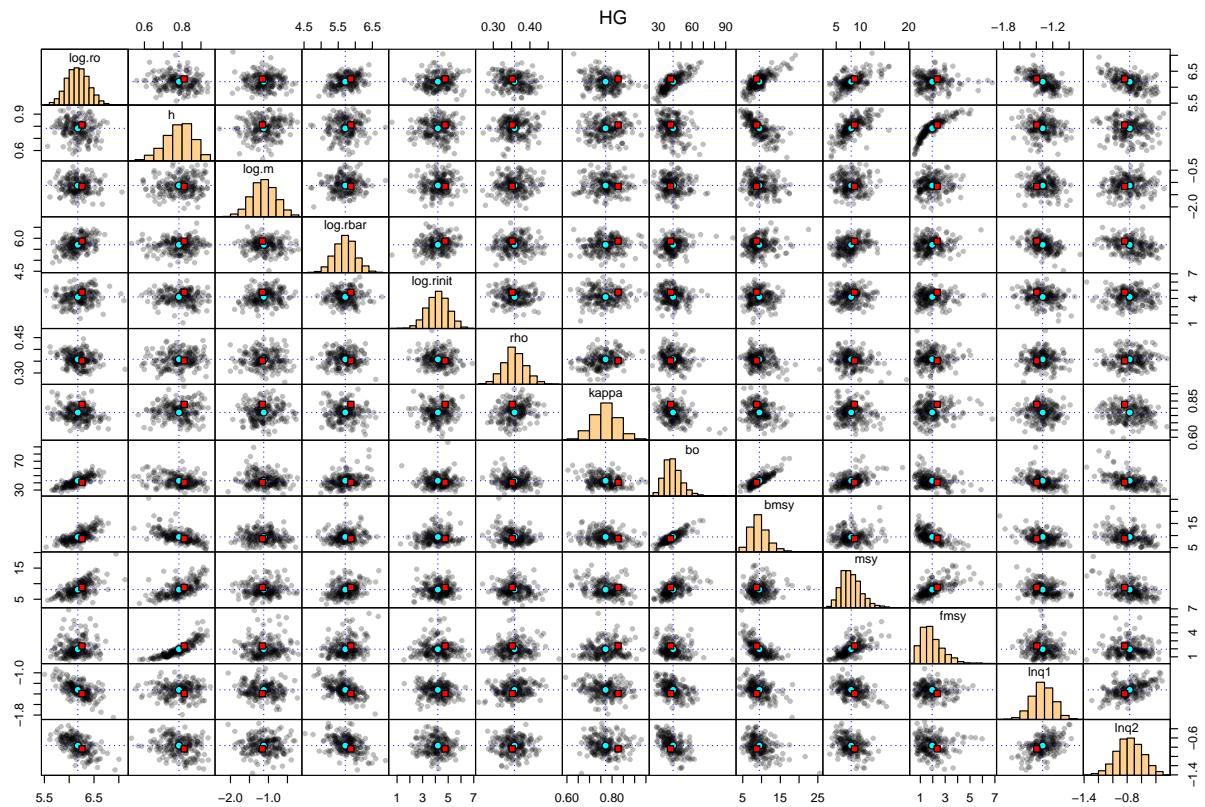


Figure 2.34: Pairs plot and marginal distributions for leading parameters in HG region (200 random samples). The dotted lines correspond to the means of the distributions, circle is the mean, and the red square is the mode of the distribution.

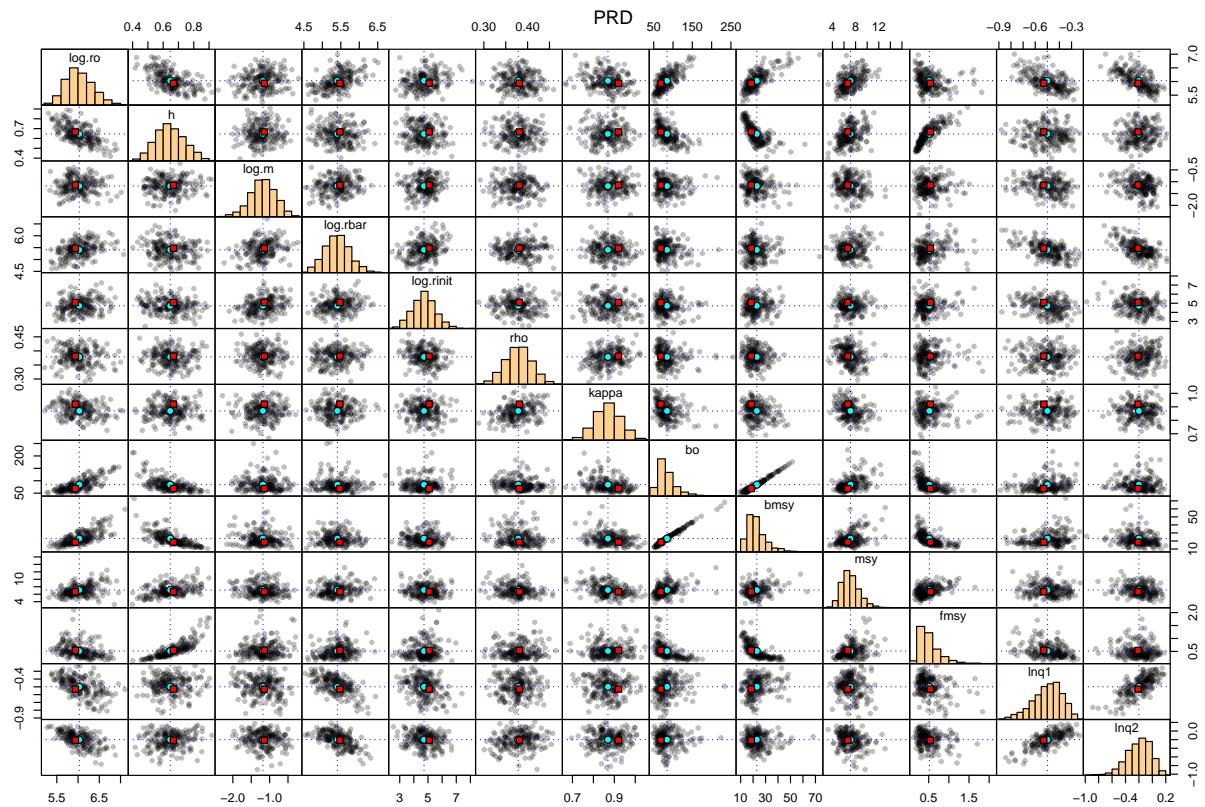


Figure 2.35: Pairs plot and marginal distributions for leading parameters in PRD region (200 random samples). The dotted lines correspond to the means of the distributions, circle is the mean, and the red square is the mode of the distribution.

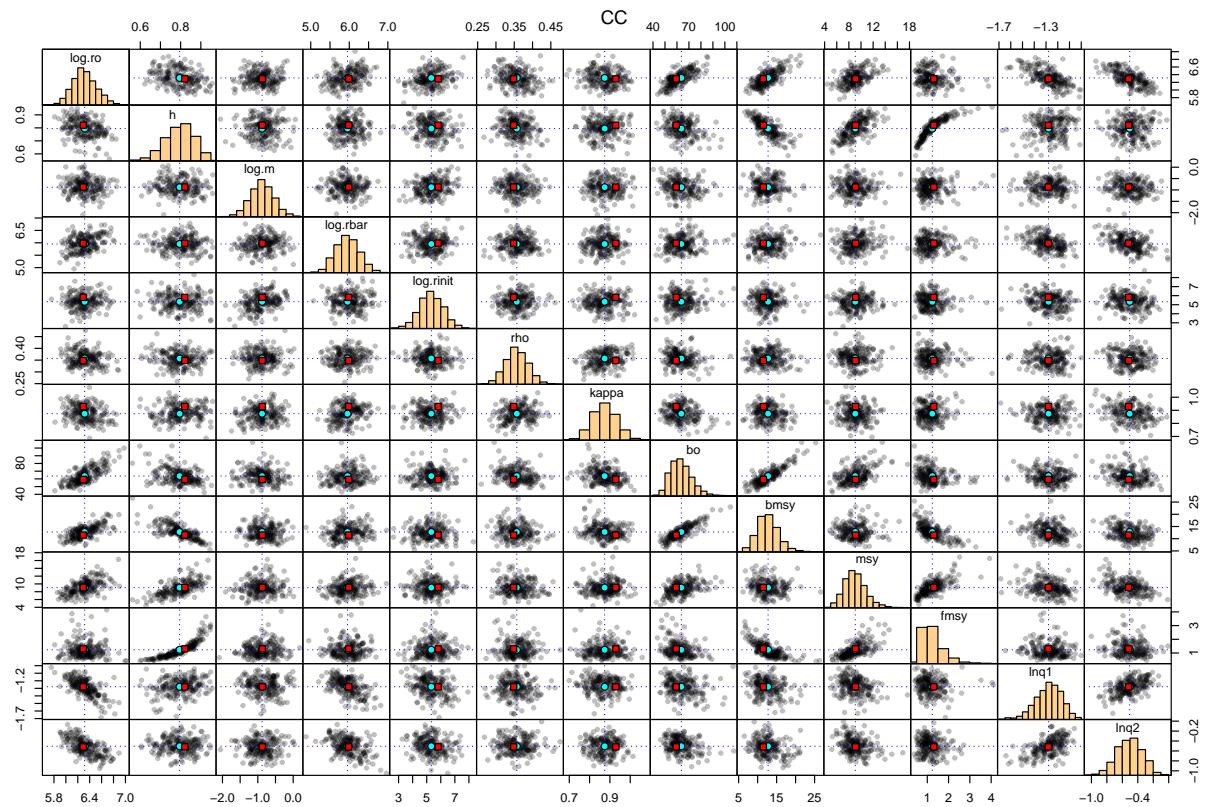


Figure 2.36: Pairs plot and marginal distributions for leading parameters in CC region (200 random samples). The dotted lines correspond to the means of the distributions, circle is the mean, and the red square is the mode of the distribution.

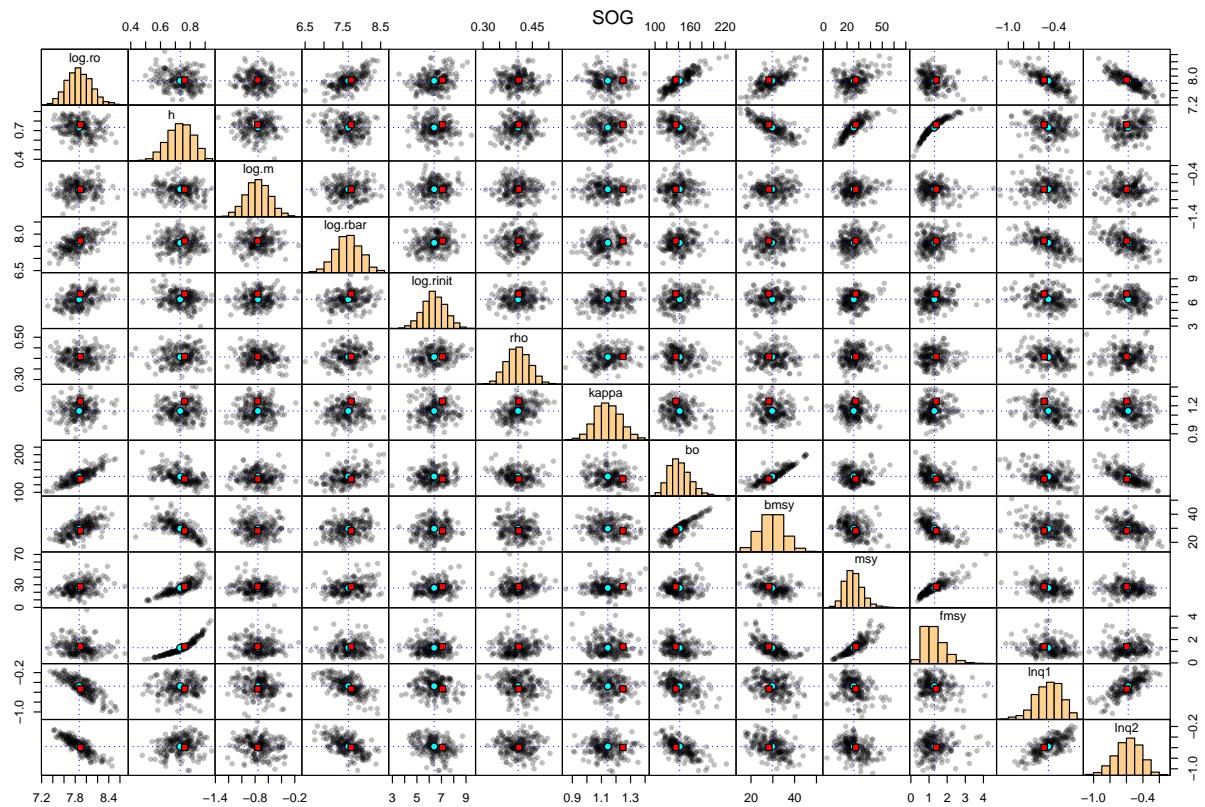


Figure 2.37: Pairs plot and marginal distributions for leading parameters in SOG region (200 random samples). The dotted lines correspond to the means of the distributions, circle is the mean, and the red square is the mode of the distribution.

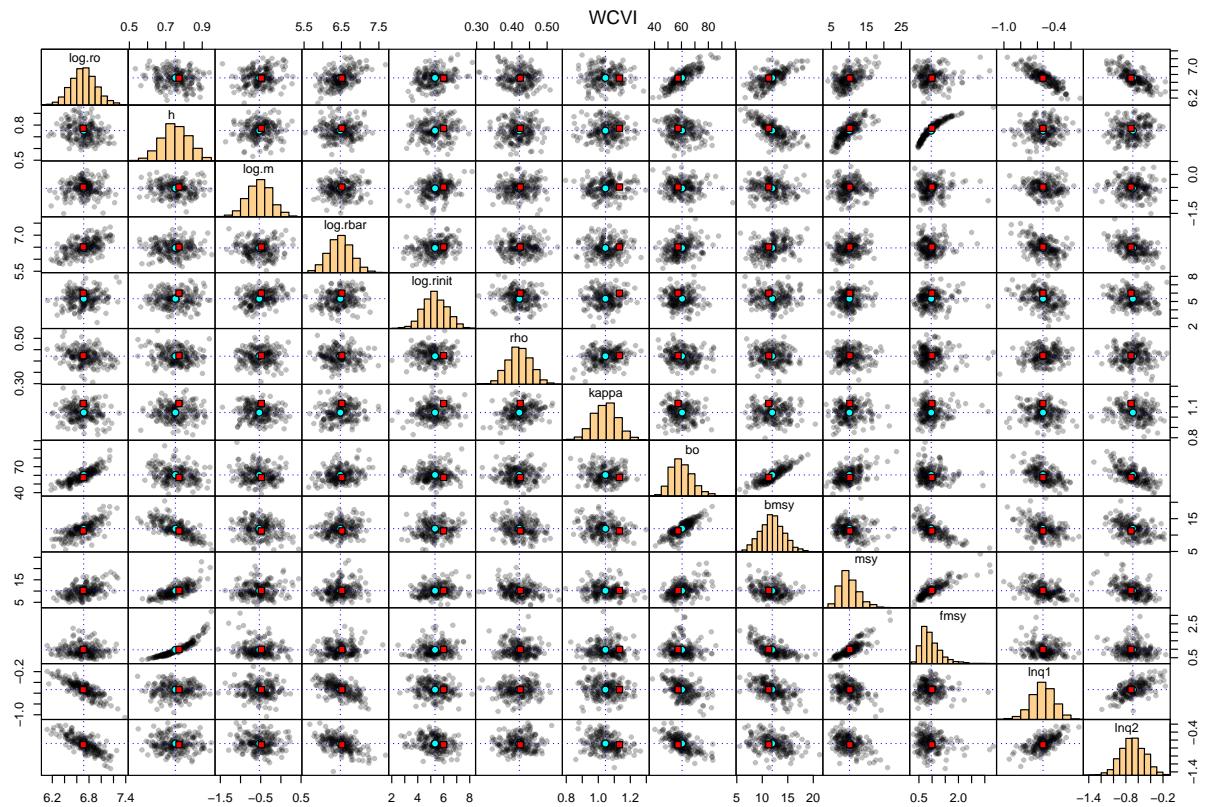


Figure 2.38: Pairs plot and marginal distributions for leading parameters in WCVI region (200 random samples). The dotted lines correspond to the means of the distributions, circle is the mean, and the red square is the mode of the distribution.

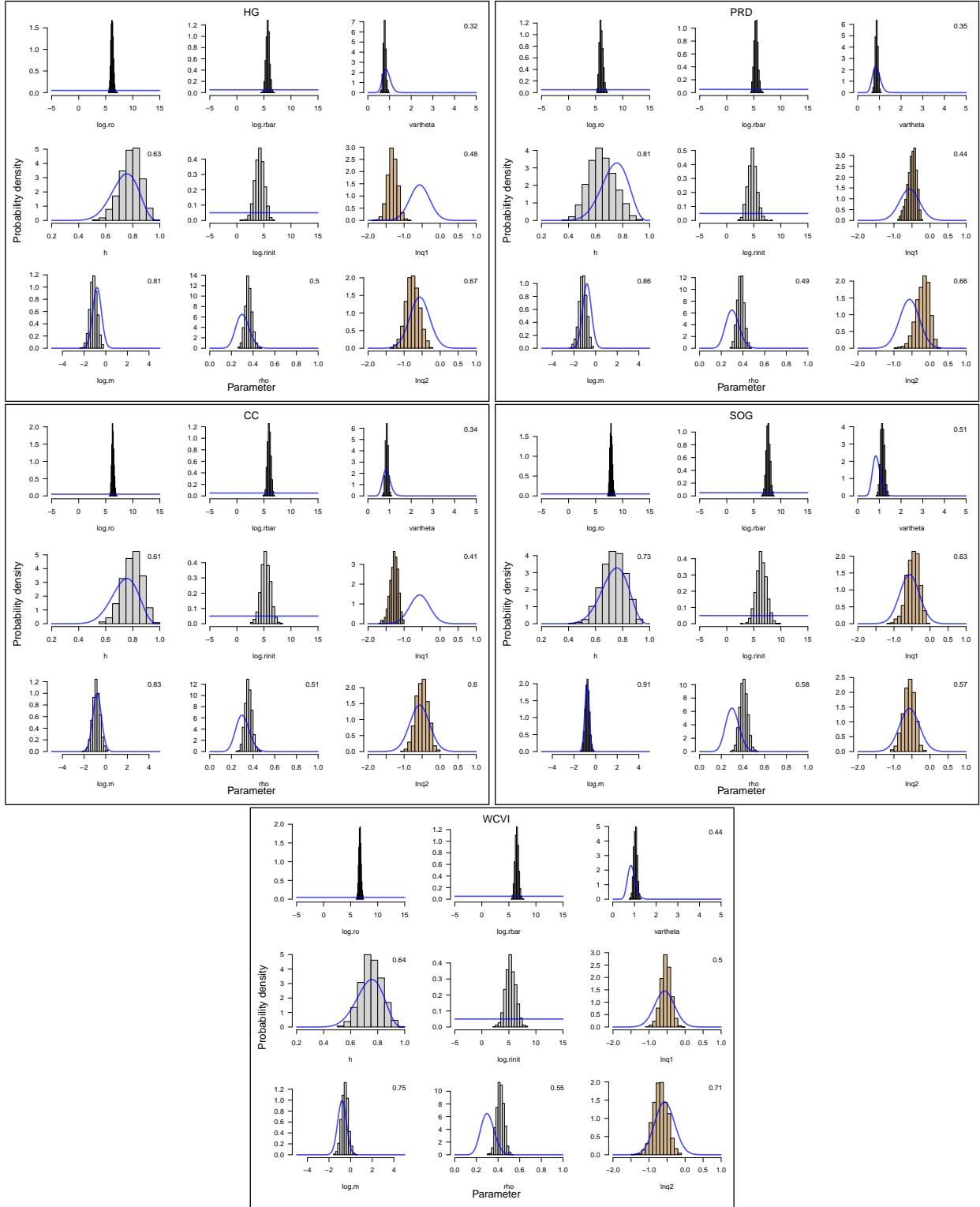


Figure 2.39: Marginal posterior densities (histograms) and prior densities (lines) for the seven leading parameters and spawn survey scaler ( $\ln(q)$ , tan colour) for each of the five major assessment regions. Note the x-axis range corresponds to the lower and upper bounds of each model parameter, and the number in the top right corner is the ratio of posterior standard deviation to prior standard deviation (a ratio close to one implies no information in the data to estimate the parameter).

Table 2.6: Estimated spawning stock biomass, age-4+ biomass and pre-fishery biomass for poor average and good recruitment, old cutoffs, and available harvest based on median values from the joint posterior distribution.

Stock	SSB	4+ Biomass	Pre-fishery forecast biomass			Cutoff	Available harvest		
			Poor	Average	Good		Poor	Average	Good
HG	16,579	7,089	9,618	12,892	21,478	10,700	0	2,192	4,296
PRD	27,046	20,593	24,150	27,492	37,286	12,100	4,830	5,498	7,457
CC	14,666	7,809	11,357	14,709	22,883	17,600	0	0	4,577
SOG	125,261	72,937	94,703	112,856	138,448	21,200	18,941	22,571	27,690
WCVI	14,679	8,267	15,321	20,906	31,130	18,800	0	2,106	6,226

Table 2.7: Estimated spawning stock biomass, age-4+ biomass and pre-fishery biomass for poor average and good recruitment, new cutoffs (based on median value of  $0.25B_0$  estimated within the  $iSCA_M$  model), and available harvest based on the median values from the joint posterior distribution.

Stock	SSB	4+ Biomass	Pre-fishery forecast biomass			Cutoff	Available harvest		
			Poor	Average	Good		Poor	Average	Good
HG	16,579	7,089	9,618	12,892	21,478	10,436	0	2,456	4,296
PRD	27,046	20,593	24,150	27,492	37,286	19,641	4,510	5,498	7,457
CC	14,666	7,809	11,357	14,709	22,883	15,600	0	0	4,577
SOG	125,261	72,937	94,703	112,856	138,448	35,013	18,941	22,571	27,690
WCVI	14,679	8,267	15,321	20,906	31,130	14,894	427	4,181	6,226

total biomass, estimates of age-3 recruit biomass for poor, average, and good recruitment, cutoff levels, and the available harvest under poor, average, and good recruitment scenarios. Moving towards DFO's sustainable fisheries framework (SFF) is a necessary next step. The SFF calls for the establishment of three reference levels: (1) the limit reference point (LRP) that defines the transition between the critical zone and the cautious zone, (2) an Upper Stock Reference (USR) that defines the transition between the cautious zone and the healthy zone, and (3) the Removal Reference (RR) which is the maximum removal rate when the stock is in the healthy zone ([Shelton and Sinclair, 2008](#)). Suggested default levels for these reference levels are  $0.4B_{MSY}$  and  $0.8B_{MSY}$  for the LRP and USR, respectively and  $F_{MSY}$  for the RR. At present these suggested reference points have not undergone simulation testing. Moreover, the definition of MSY based reference points for this fishery require an *a priori* allocation of the total catch to each of the different gears that harvest herring. Finally, there is also a great deal of uncertainty associated with estimates of  $B_{MSY}$  for each of the SARs; probabilities of exceeding these reference points must also be established *a priori* before we can move this fishery towards the sustainable fisheries framework.

Cutoff levels for the BC herring stocks is defined as  $0.25B_0$ . The historical cutoff levels have not been updated for over 10 years now (1996/1997). Recent stock assessments (since 2001) for BC herring have assumed  $q = 1$  for the spawn survey data. In this assessment we have relaxed this assumption and as a consequence estimates of herring biomass have increased substantially. Due to significant changes in population scaling it would not make sense to continue to use the previous cutoff levels as this may lead to policies that would result in overfishing or under utilization of the resource. We therefore present catch advice based on both the old cutoffs and the new cutoffs in Tables 2.6-2.7 for comparison. Under the new model and the median estimates of unfished biomass, cutoff levels decreased in HG, CC and the WCVI, and increased in PRD and the SOG. In the case of SOG, cutoff values increased from 21,200 t to 35,013 t. In PRD, cutoff values increased from 12,100 t to 19,641 t. In both cases the catch advice based on the current HCR does not differ for the average and good recruitment scenarios as the projected 3+ biomass is well above these cutoff values.

In addition to Tables 2.6 and 2.7 we also provide an additional risk-based decision table that attempts to integrate over all of the uncertainty in the model (Tables 2.10-2.12). This decision table is also represented

graphically in Figure 2.40. Figure 2.40 should be interpreted as follows: the probability of the spawning stock biomass falling below the cutoff level is determined by drawing a vertical line that intersect the cumulative probability curve and reading off the corresponding probability level. The reverse of this process was used to construct Tables 2.10-2.12.

Table 2.8: Decision table for HG where the risk level represents the probability of exceeding the quantities specified in the headers of each column. Three performance measures are considered: the probability of the spawning stock biomass in 2013 falling below the cutoff level, the probability of the spawning stock biomass in 2013 declining from 2012, and the probability of 2012 exploitation rate exceeding the 20% level that is used in the harvest control rule. To use this table, first determine the appropriate level of risk (e.g. 0.25 or 25% chance), then choose the appropriate management quantity (e.g. spawning biomass falling below the cutoff), and then read off the recommended catch (e.g., 3,060 tonnes).

Risk level	$P(SB_{2013} < \text{Cutoff})$	$P(SB_{2013} < SB_{2012})$	$P(U_{2012} < 0.2)$
0.05	0	0	2,228
0.1	0	0	3,552
0.15	583	0	4,372
0.2	1,940	0	4,989
0.25	3,060	0	5,498
0.3	4,039	0	5,944
0.35	4,928	0	6,348
0.4	5,760	0	6,727
0.45	6,558	0	7,090
0.5	7,340	0	7,445
0.55	8,121	0	7,801
0.6	8,919	907	8,164
0.65	9,751	2,547	8,542
0.7	10,640	4,299	8,947
0.75	11,619	6,229	9,392
0.8	12,740	8,439	9,902
0.85	14,096	11,113	10,519
0.9	15,898	14,666	11,338
0.95	18,809	20,404	12,662

Table 2.9: Decision table for PRD where the risk level represents the probability of exceeding the quantities specified in the headers of each column. Three performance measures are considered: the probability of the spawning stock biomass in 2013 falling below the cutoff level, the probability of the spawning stock biomass in 2013 declining from 2012, and the probability of 2012 exploitation rate exceeding the 20% level that is used in the harvest control rule. To use this table, first determine the appropriate level of risk (e.g. 0.25 or 25% chance), then choose the appropriate management quantity (e.g. spawning biomass falling below the cutoff), and then read off the recommended catch (e.g., 1,788 tonnes).

Risk level	$P(SB_{2013} < \text{Cutoff})$	$P(SB_{2013} < SB_{2012})$	$P(U_{2012} < 0.2)$
0.05	0	0	3,743
0.1	0	0	4,914
0.15	0	0	5,639
0.2	655	182	6,185
0.25	1,788	787	6,636
0.3	2,777	1,315	7,030
0.35	3,675	1,795	7,388
0.4	4,516	2,244	7,722
0.45	5,322	2,675	8,043
0.5	6,112	3,097	8,358
0.55	6,902	3,518	8,673
0.6	7,708	3,949	8,994
0.65	8,549	4,398	9,328
0.7	9,447	4,878	9,686
0.75	10,437	5,406	10,080
0.8	11,569	6,011	10,531
0.85	12,940	6,743	11,077
0.9	14,761	7,716	11,802
0.95	17,702	9,287	12,973

Table 2.10: Decision table for CC where the risk level represents the probability of exceeding the quantities specified in the headers of each column. Three performance measures are considered: the probability of the spawning stock biomass in 2013 falling below the cutoff level, the probability of the spawning stock biomass in 2013 declining from 2012, and the probability of 2012 exploitation rate exceeding the 20% level that is used in the harvest control rule. To use this table, first determine the appropriate level of risk (e.g. 0.25 or 25% chance), then choose the appropriate management quantity (e.g. spawning biomass falling below the cutoff), and then read off the recommended catch (e.g., 402 tonnes).

Risk level	$P(SB_{2013}) < \text{Cutoff}$	$P(SB_{2013} < SB_{2012})$	$P(U_{2012} < 0.2)$
0.05	0	0	2,450
0.1	0	1,521	3,195
0.15	0	4,063	3,657
0.2	0	5,977	4,005
0.25	402	7,557	4,292
0.3	994	8,938	4,543
0.35	1,530	10,192	4,770
0.4	2,033	11,366	4,984
0.45	2,515	12,491	5,188
0.5	2,987	13,593	5,388
0.55	3,459	14,696	5,589
0.6	3,940	15,821	5,793
0.65	4,443	16,995	6,006
0.7	4,980	18,249	6,234
0.75	5,571	19,629	6,485
0.8	6,247	21,210	6,772
0.85	7,067	23,124	7,119
0.9	8,155	25,665	7,581
0.95	9,913	29,771	8,327

Table 2.11: Decision table for SOG where the risk level represents the probability of exceeding the quantities specified in the headers of each column. Three performance measures are considered: the probability of the spawning stock biomass in 2013 falling below the cutoff level, the probability of the spawning stock biomass in 2013 declining from 2012, and the probability of 2012 exploitation rate exceeding the 20% level that is used in the harvest control rule. To use this table, first determine the appropriate level of risk (e.g. 0.25 or 25% chance), then choose the appropriate management quantity (e.g. spawning biomass falling below the cutoff), and then read off the recommended catch (e.g., 51,565 tonnes).

Risk level	$P(SB_{2013} < \text{Cutoff})$	$P(SB_{2013} < SB_{2012})$	$P(U_{2012} < 0.2)$
0.05	32,082	0	32,080
0.1	39,969	0	37,840
0.15	44,852	0	41,406
0.2	48,528	0	44,091
0.25	51,565	0	46,308
0.3	54,218	0	48,246
0.35	56,627	0	50,005
0.4	58,882	0	51,651
0.45	61,043	0	53,230
0.5	63,161	0	54,777
0.55	65,280	0	56,324
0.6	67,441	0	57,902
0.65	69,696	0	59,549
0.7	72,105	0	61,308
0.75	74,758	0	63,245
0.8	77,794	0	65,463
0.85	81,471	0	68,148
0.9	86,354	0	71,714
0.95	94,241	8,031	77,474

Table 2.12: Decision table for WCVI where the risk level represents the probability of exceeding the quantities specified in the headers of each column. Three performance measures are considered: the probability of the spawning stock biomass in 2013 falling below the cutoff level, the probability of the spawning stock biomass in 2013 declining from 2012, and the probability of 2012 exploitation rate exceeding the 20% level that is used in the harvest control rule. To use this table, first determine the appropriate level of risk (e.g. 0.25 or 25% chance), then choose the appropriate management quantity (e.g. spawning biomass falling below the cutoff), and then read off the recommended catch (e.g., 184 tonnes).

Risk level	$P(SB_{2013} < \text{Cutoff})$	$P(SB_{2013} < SB_{2012})$	$P(U_{2012} < 0.2)$
0.05	0	907	2,269
0.1	0	6,355	3,125
0.15	0	9,728	3,655
0.2	0	12,267	4,054
0.25	184	14,365	4,384
0.3	969	16,197	4,671
0.35	1,682	17,861	4,933
0.4	2,349	19,418	5,178
0.45	2,989	20,912	5,412
0.5	3,616	22,375	5,642
0.55	4,243	23,838	5,872
0.6	4,882	25,331	6,106
0.65	5,549	26,888	6,351
0.7	6,262	28,552	6,613
0.75	7,047	30,385	6,900
0.8	7,946	32,482	7,230
0.85	9,034	35,022	7,629
0.9	10,478	38,395	8,159
0.95	12,812	43,842	9,015

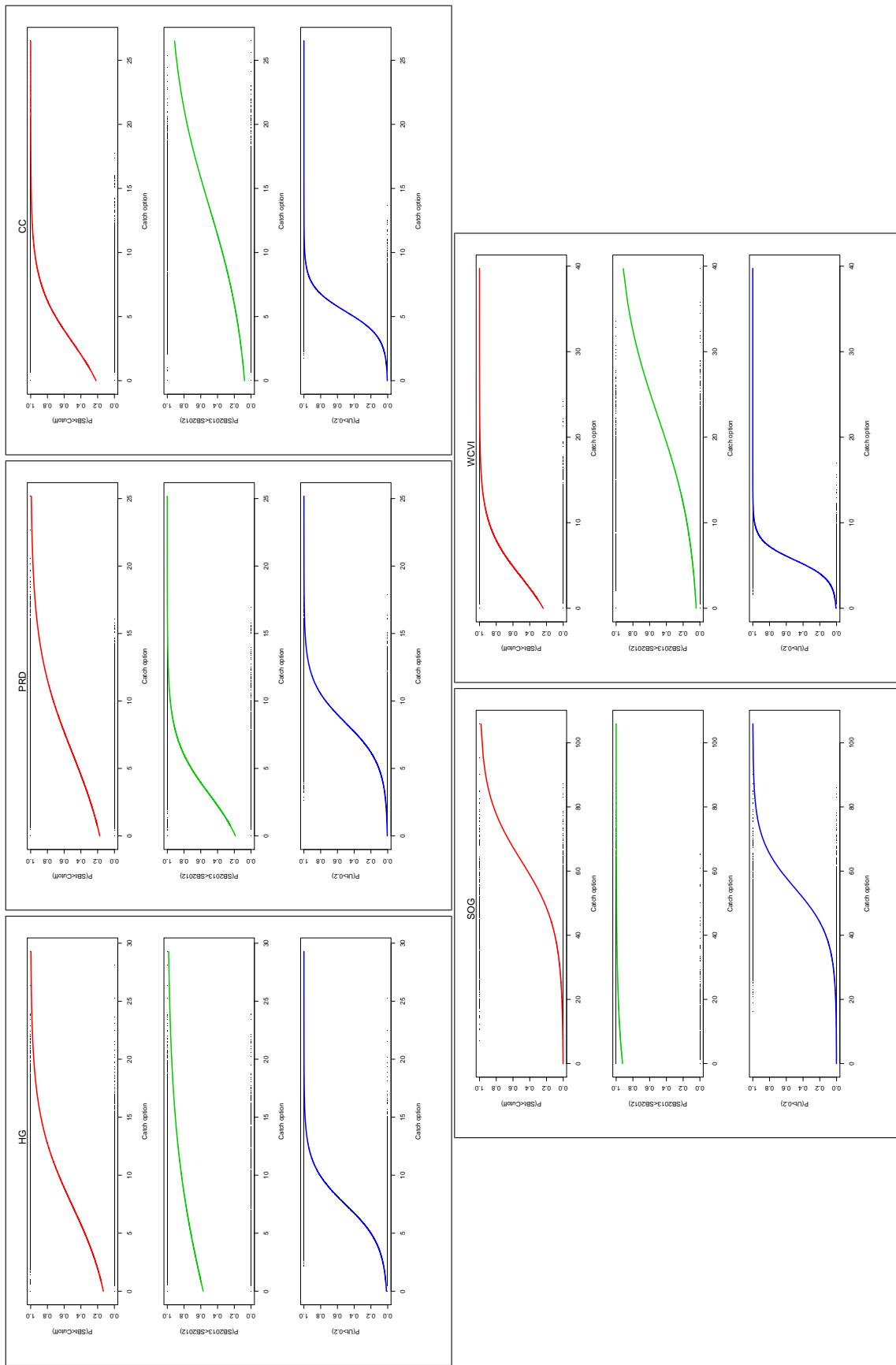


Figure 2.40: Top Panels: probability of the spawning stock biomass in 2013 falling below the cutoff level versus the 2012 catch option. Middle Panels: probability of the spawning stock in 2012 being less than the spawning stock biomass in 2012 versus the 2012 catch option. Bottom Panels: probability of the 2012 harvest rate (catch/3+ biomass) being greater than the target harvest rate of 0.2.

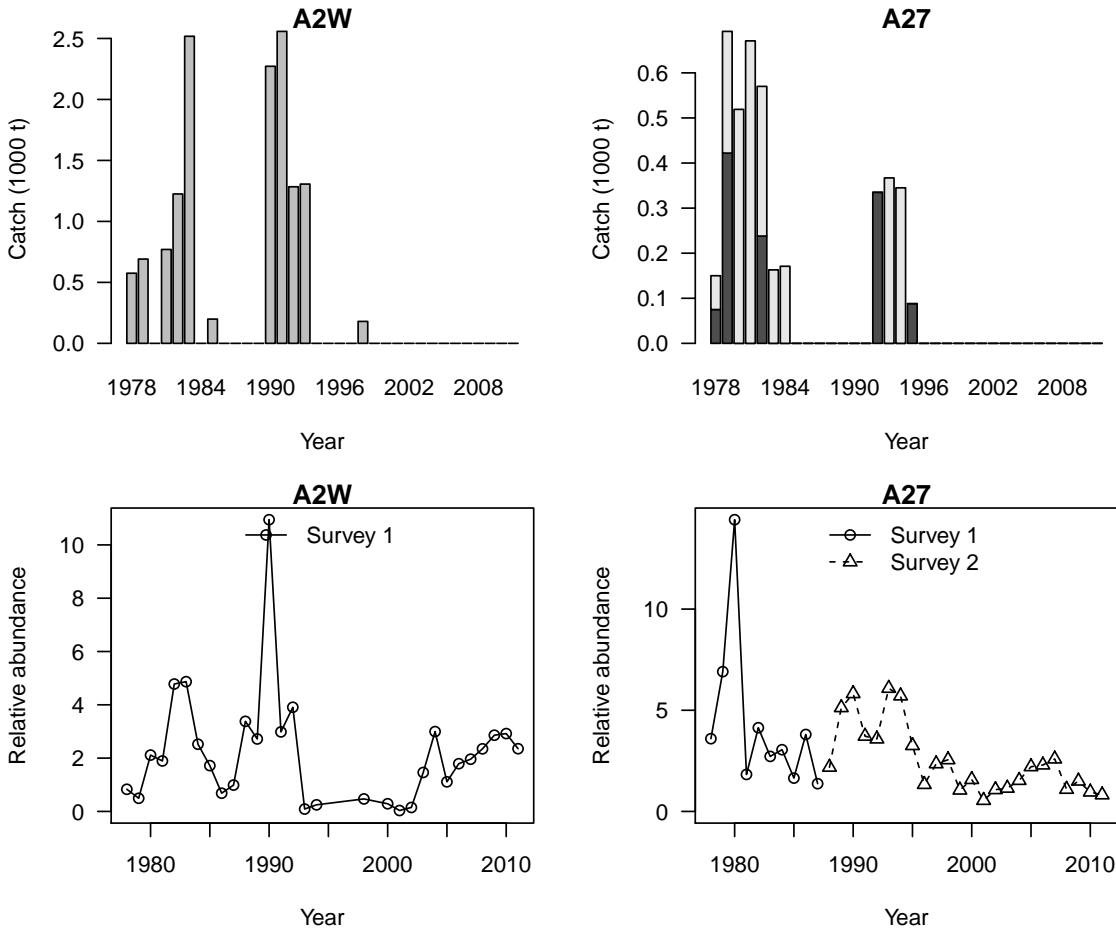


Figure 2.41: Catch and survey data for minor stock Areas 2W and Area 27.

## 2.9 Stock assessments for minor stock areas

Abundance estimates for the minor stock areas, Area 2W and Area 27 were also obtained using the  $i\text{SCA}_M$  model. For these minor areas, there were some minor differences in the treatment of the data and model assumptions. Also, the gillnet selectivity in area 27 was not allowed to vary over time due to the sparse amount of information (and presumably biological samples) available to reliably estimate minor changes in selectivity for this fishery. Selectivity for area 27 gillnet fishery was assumed to be a logistic function of age and invariant over time.

The input data (Catch and relative abundance) for the minor areas is shown in Figure 2.41. As in the previous assessments of Area 2W, the spawn survey data is treated as a single continuous series from 1978 to 2011. Area 27 however, the time series is split into two series between 1978-1987 and 1988-2011. The age-composition data used in fitting the model is shown in Figure 2.42.

### 2.9.1 Maximum likelihood estimates of biomass

Spawning biomass in 2011 for Area 2W and Area 27 was estimated at 4,671 tonnes and 928 tonnes, respectively (Table 2.13). The time series of total biomass and spawning biomass for these two areas is presented

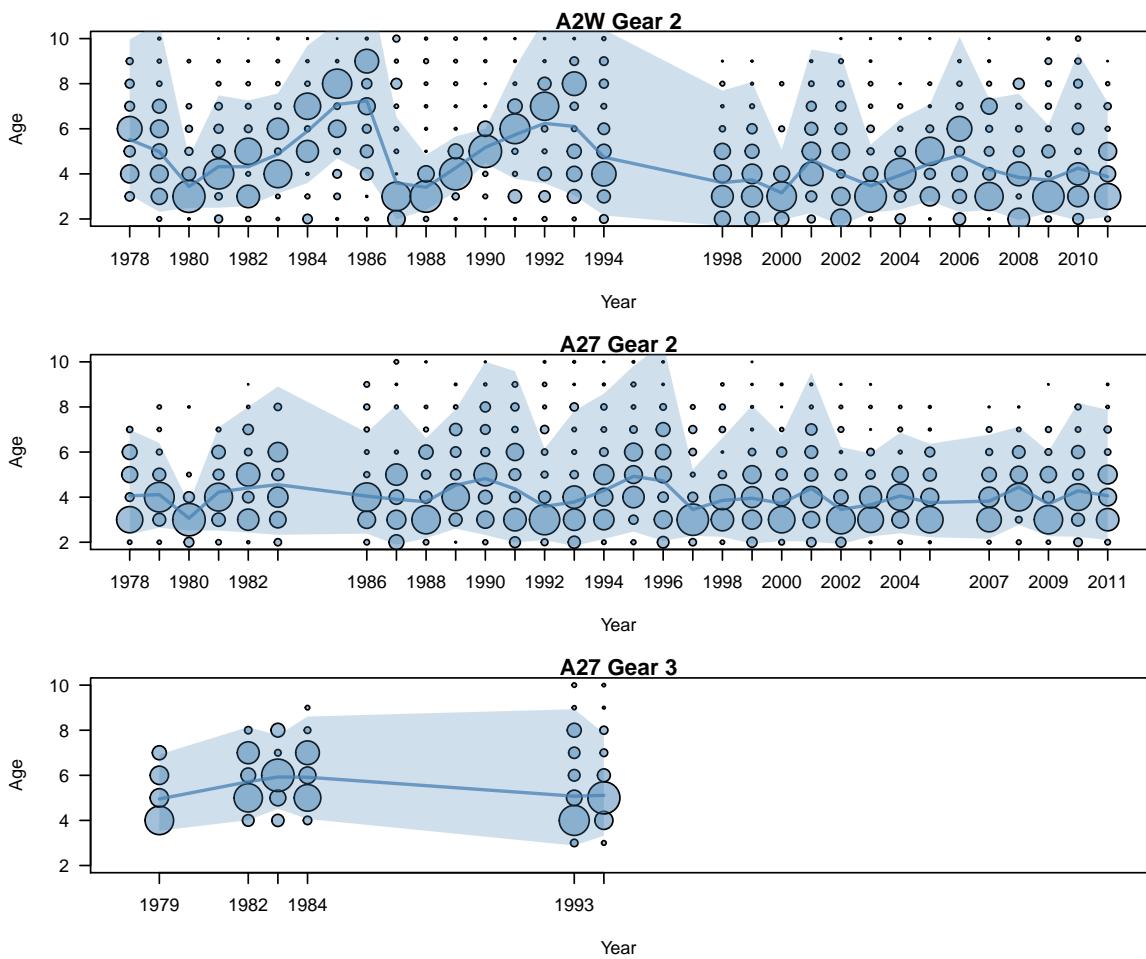


Figure 2.42: Age composition data for Area 2W and Area 27 for the seine-roe fishery (Gear 2) and the gillnet fishery (Gear 3).

Table 2.13: Summary of maximum likelihood estimates for the two minor stock areas. No. is the total number of estimated parameters,  $F_{MSY}$  the average instantaneous fishing rate to achieve the maximum sustainable yield (MSY),  $B_0$  is the unfished spawning biomass,  $B_{MSY}$  is the spawning biomass that achieves maximum sustainable yield,  $B_t$  is the spawning biomass at the end of the 2011 fishing season, and  $B_t/B_0$  is the spawning depletion level at the end of the 2011 fishing season.

Stock	A2W	A27
No.	74	79
$F_{MSY}$	0.34	1.9
MSY	265	304
$B_0$	2,915	2,084
$0.25B_0$	729	521
$B_{MSY}$	705	447
$0.8B_{MSY}$	564	358
$0.4B_{MSY}$	282	179
$B_t$	4,671	924
$B_t/B_0$	1.6	0.44

in Figure 2.43

### 2.9.2 Estimates of recruitment and reference points

Maximum likelihood estimates of age-2 recruitment, stock-recruitment relationships and residuals in the stock-recruitment model is shown in Figure 2.44. Estimates of age-2 recruits in area 2W have been poor-to-average for much of the time-series. There have been 4 periods of above average recruitment for area 2W (late 1970s, mid 1980s, 2002, and 2008-2010). Recruitment in area 27 has been much more consistent by comparison. Estimates of unfished spawning biomass for areas 2W and 27 are 2,915 tonnes and 2,112 tonnes.

### 2.9.3 Retrospective analysis

There is almost no retrospective bias for the estimates of spawning stock biomass in area 27 using data between 1951:2001 and 1951:2011 (Fig. 2.45). In Area 2W, there is a slight retrospective bias in the estimates of spawning stock biomass. As the more recent data are fit in the model estimates of spawning biomass in the mid 2000s are revised downwards. In other words, there is a positive bias in estimates of spawning biomass in area 2W.

### 2.9.4 Marginal posterior distributions and trace plots

Information for the catch advice for the two minor areas is based on the median values of the joint posterior distribution. Therefore, it is important to show posterior samples to ensure proper convergence and the marginal posterior distributions for the leading parameter estimates and derived variables that are of management interest.

The trace plots for the two minor areas are summarized in Figure 2.46, and the marginal distributions for the leading parameters is shown in Figure 2.47. Again, no formal convergence statistics were examined to determine if MCMC chain converged to a stable distribution. Visual inspection of the trace plots appear to have a homogenous distribution over the course of the 2000 samples. In both of the statistical areas, the posterior updates did occur (Figure 2.47). The marginal posterior for steepness in area 27 does appear to be influenced considerably by the assumed prior distribution.

### 2.9.5 Catch advice

Catch advice for the minor areas differs from that of the major areas in that there are not cutoffs for these two areas and the reference exploitation rate is reduced from 20% to 10%. The same decision table format is

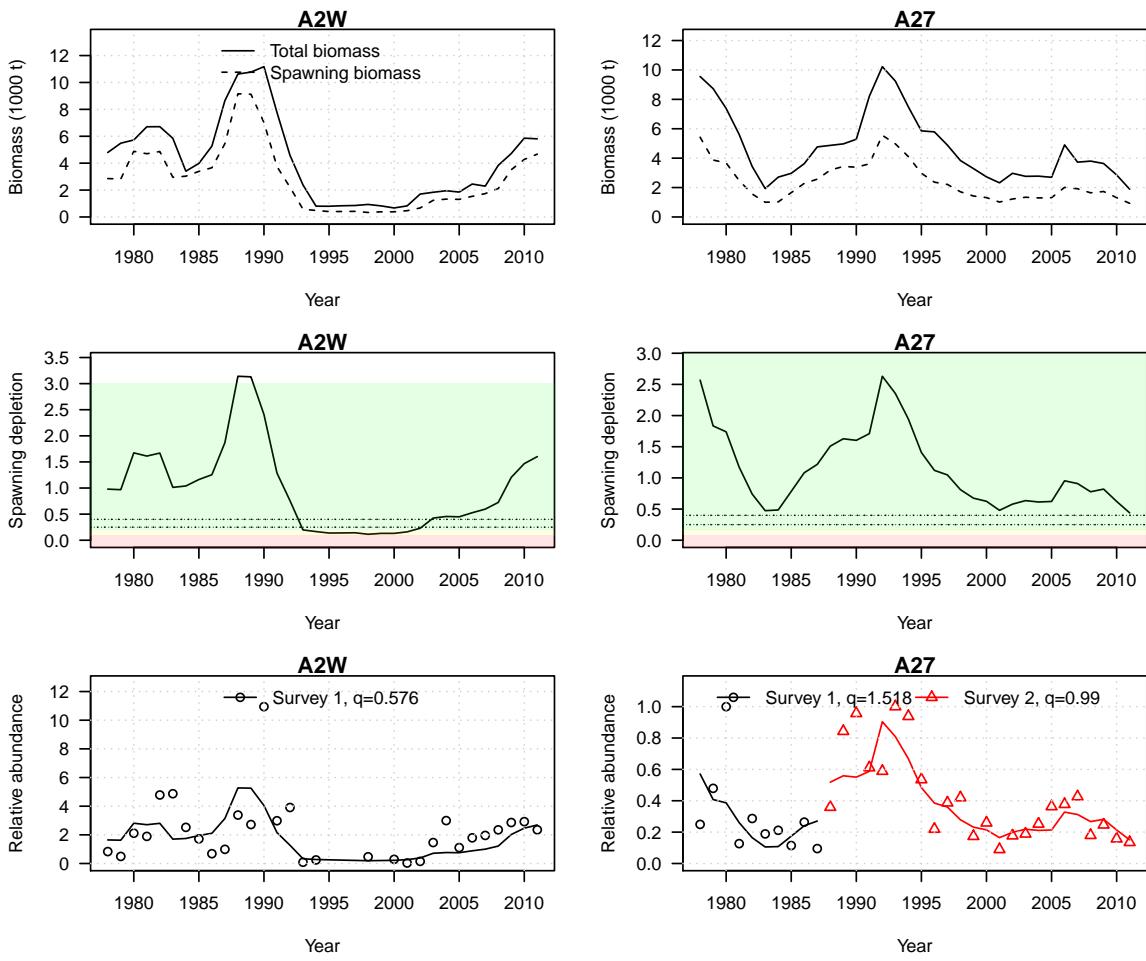


Figure 2.43: Maximum likelihood estimates of total biomass, spawning biomass, spawning depletion and fits to the spawn survey data for the two minor stock areas.

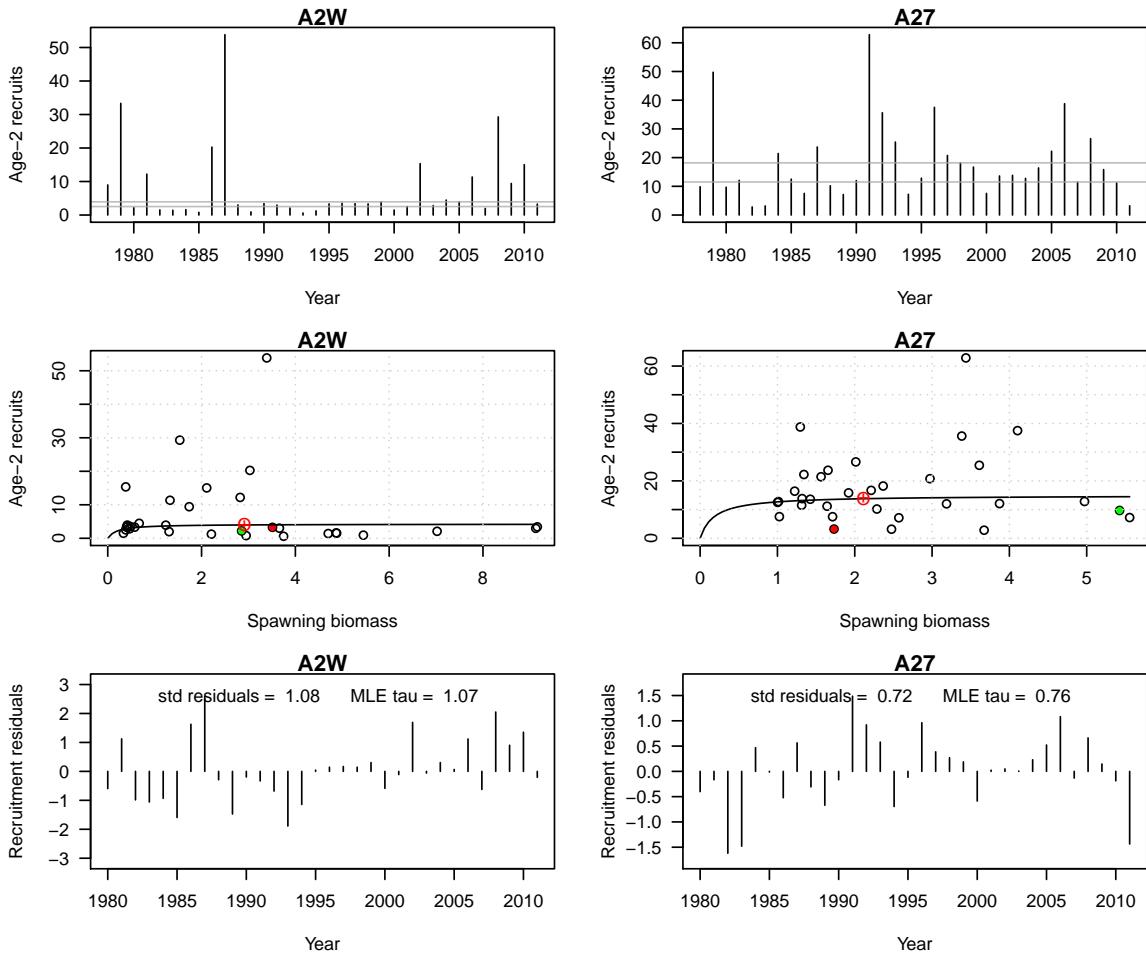


Figure 2.44: Maximum likelihood estimates of age-2 recruits, spawner-recruit relationships with the fitted Beverton Holt model and unfished reference points ( $B_o, R_o$ ), and the residuals between the estimated age-2 recruits and that predicted by the Beverton-Holt model.

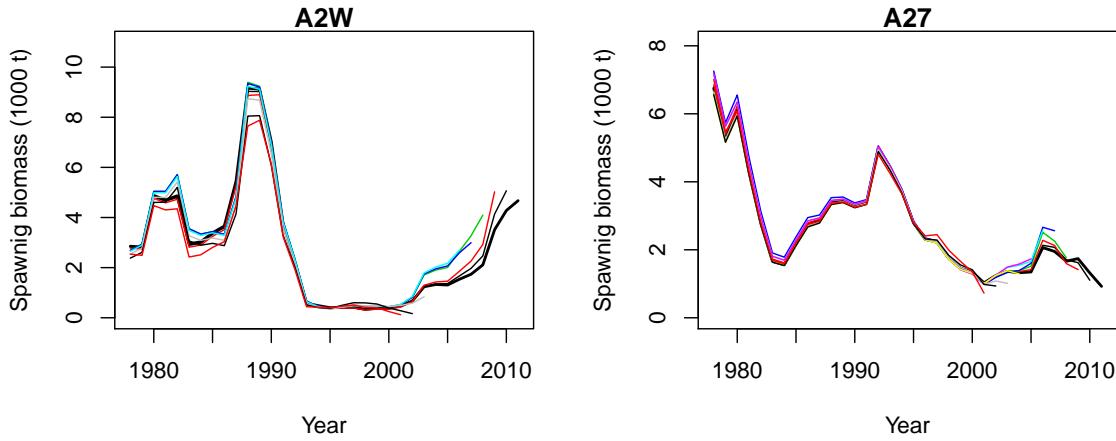


Figure 2.45: Retrospective estimates of spawning stock biomass for each of the minor stock assessment areas. The model was sequentially fitted to the full data set, then from 1951:2010, 1951:2009, ... 1951:2001.

Table 2.14: Estimated spawning stock biomass, age-4+ biomass and pre-fishery biomass for poor average and good recruitment, cutoffs, and available harvest.

Stock	Pre-fishery forecast biomass						Available harvest		
	SSB	4+ Biomass	Poor	Average	Good	Cutoff	Poor	Average	Good
A2W	5,448	5,204	5,294	5,398	6,141	809	529	540	614
A27	1,077	692	909	1,123	1,736	602	91	112	174

provided with catch advice based on poor, average, and good recruitment. Catch advice for the two minor areas is summarized in Table ??.

## 2.10 Outstanding Issues

The catch advice provided this year is based on the old  $0.25B_0$  rule for establishing Cutoffs for each SAR. Also, in moving towards a Sustainable Fisheries Framework and perhaps adopting the suggested MSY-based reference points, presents technical issue with regard to setting these reference points when population parameters are changing over time. In this assessment, we have used the average weight-at-age to calculate  $B_0$ . Then in an inconsistent manner, we subsequently use the average weight-at-age age (and fecundity-at-age) over the last 5 years to determine  $B_{MSY}$  and  $F_{MSY}$  reference points. This outstanding issue should be examined more carefully before moving towards the SFF.

There was a strong retrospective bias for the PRD region. At this time, the sources of this bias are unknown, but likely due to two or more sources of data that contradict each-other or model misspecification. The current bias problem suggest that biomass has typically been over-estimated in recent years. The source of this bias should be investigated more closely.

There are a number of changes implemented in this  $^iSCA_M$  modelling framework that have not been formally evaluated using statistical criterion. Model selection criterion such as Analysis of Deviance (DIC) should be used when adopting new formulations. A number of alternative hypotheses (e.g., changes in selectivity, natural mortality rates) should be formally evaluated using model selection criterion.

An informative prior for the spawn survey has been used in this assessment. The marginal posterior

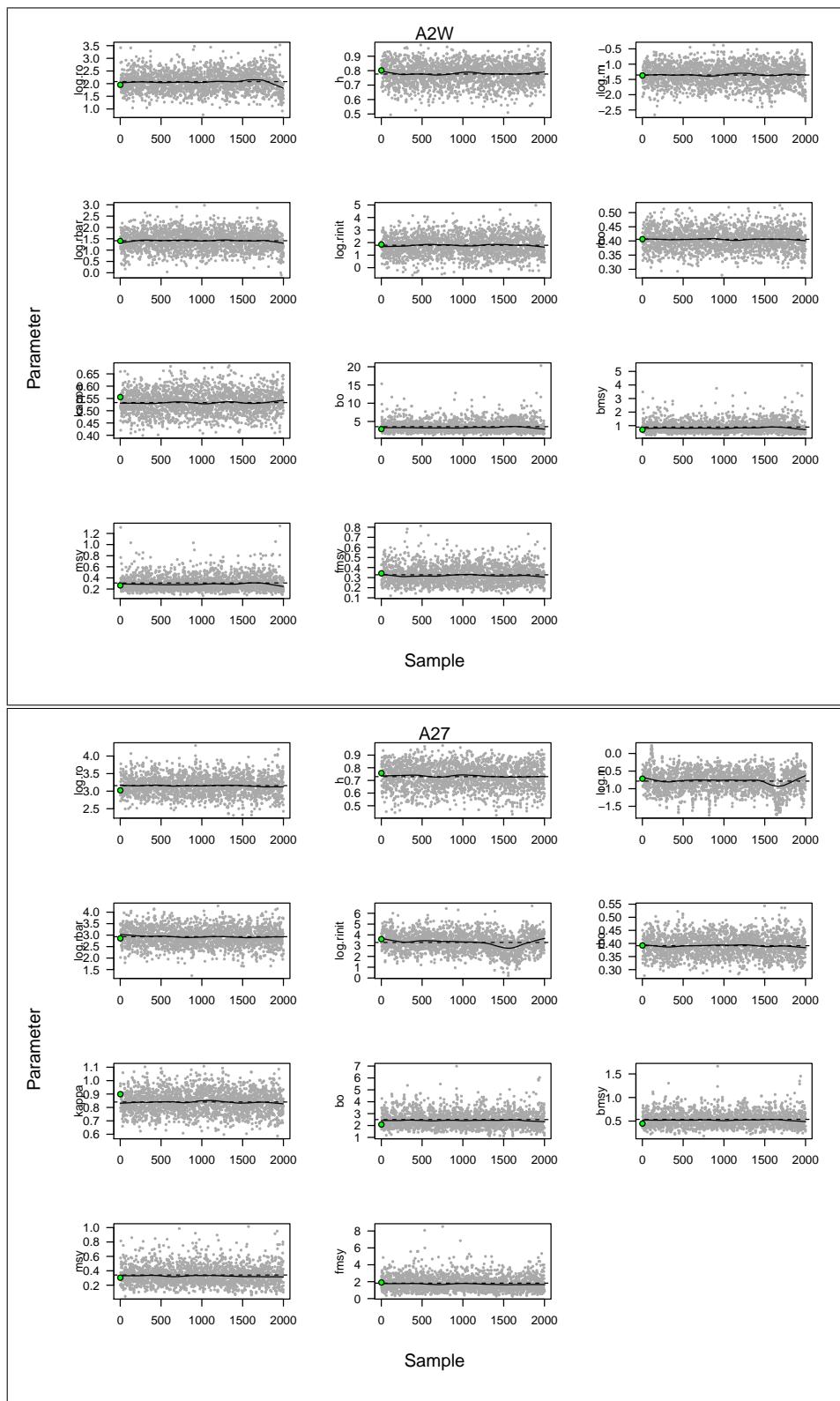


Figure 2.46: A systematic sample of 2000 points from a chain of length 1,000,000 from the joint posterior distribution for areas 2W and area 27.

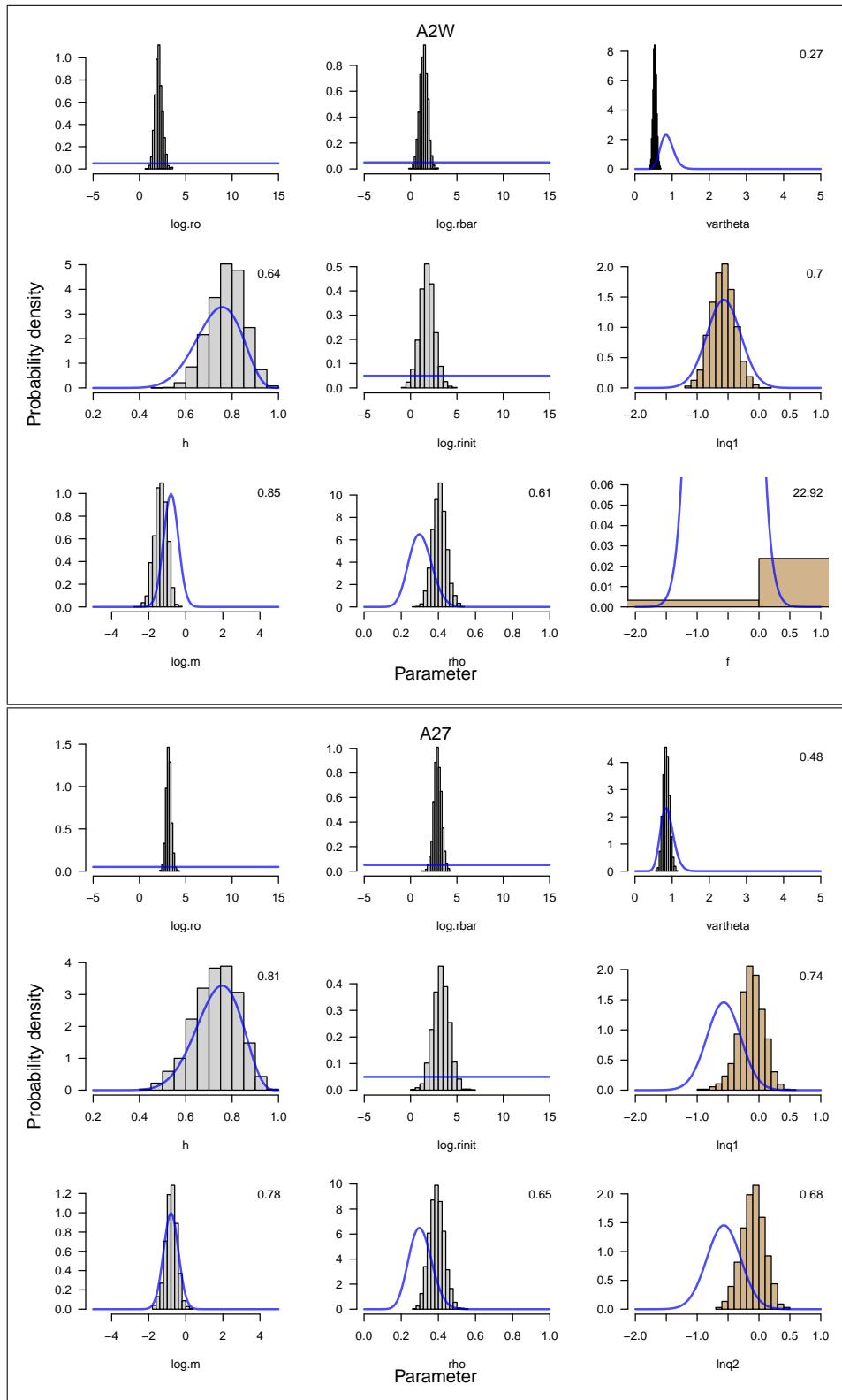


Figure 2.47: Marginal distributions (bars) and priors (lines) for the leading parameters based on a systematic sample of 2000 points from a chain of length 1,000,000 from the joint posterior distribution for areas 2W and area 27. Note the x-axis range corresponds to the lower and upper bounds of each model parameter, and the number in the top right corner is the ratio of posterior standard deviation to prior standard deviation (a ratio close to one implies no information in the data to estimate the parameter).

distributions for  $q$  along with the prior distribution have been plotted and indicate that there is some information in the data to inform estimates of  $q$ . However, there are interesting geographic patterns in the estimates of  $q$ : areas to the north (HG, PRD, CC) the dive survey  $q$  is higher than the surface survey  $q$ , whereas in the south the dive survey  $q$  is less than the surface survey  $q$ . Also, in Part I of this document there was an indication that the change from selectivity based on weight-at-age to fixed, or using weight-at-age as a covariate, had a significant impact on the estimates of  $q$  in the SOG. Further work should also examine the other SARs to determine if a change in selectivity for the gillnet fishery also implies large changes in  $q$ .

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The authors would like to thank Charles Fort and Kristen Daniel for their continued efforts in error checking, reviewing and updating the catch, spawn survey and biological sampling databases. We would also like to acknowledge Howard Stiff for providing programming support for the MS Access database used to summarize the assessment data time series.

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We also extend our thanks to Jim Ianelli for ideas and help with the cubic splines for selectivity and time varying natural mortality rates. Also, our huge appreciation to the members of the ADMB Foundation for continued development and support of AD Model Builder.

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# **Part III**

# **Appendices**

## A.1 Technical description of $i\text{SCA}_M$

### A.1.1 Analytic methods

The section contains the documentation in mathematical form of the underlying age structured model, and its steady state version that is used to calculate MSY-based reference points, the observation models used in predicting observations, and the components of the objective function that formulate the statistical criterion that is used to estimate model parameters. All of the model equations are laid out in tables and are intended to represent the order of operations, or pseudocode, in which to implement the model.  $i\text{SCA}_M$  was implemented in AD Model Builder version 10.1 (ADMB Project, 2009). This appendix also describes some of the optional features in  $i\text{SCA}_M$  for estimating nonparametric selectivities.

It should be noted here that MSY-based reference points assume steady-state conditions, and the model structure that is implemented for the BC herring stocks is non-stationary due to time-varying changes in natural mortality rates ( $M_t$ ) and selectivity. Estimates of MSY are conditional on the estimates of  $M$ , selectivity and mean weight-at-age; all of which change over time in the herring assessments. In the calculations of reference points, we use the average natural mortality between 1951-2010 and estimated selectivities and the empirical weight-at-age data in 2011.

### A.1.2 Equilibrium considerations

Steady-state conditions are presented in Table A-1, in here we assume the parameter vector  $\Theta$  in (T1.1) is unknown (with the exception of  $F_e$ ) and would eventually be estimated by fitting  $i\text{SCA}_M$  to time series data. The definition of  $F_e$  is the steady-state fishing mortality rate, and the value of  $F_e$  that maximizes equilibrium yield corresponds to  $F_{\text{MSY}}$  (see section A.1.3). For a given set of growth parameters (or if available empirical weight-at-age data) and maturity-at-age parameters defined by (T1.2), growth is assumed to follow the von Bertalanffy model (T1.3), mean weight-at-age is given by the allometric relationship in (T1.4), and the age-specific vulnerability is given by a logistic function (T1.5). Note, however, there are alternative selectivity functions implemented in  $i\text{SCA}_M$ , the logistic function used here is simply for demonstration purposes. Mean fecundity-at-age is assumed to be proportional to the mean weight-at-age of mature fish, where maturity at age is specified by the parameters  $\dot{\alpha}$  and  $\dot{\gamma}$  for the logistic function.

Survivorship for unfished and fished populations is defined by (T1.7) and (T1.8), respectively. It is assumed that all individuals ages  $A$  and older (i.e., the plus group) have the same total mortality rate. The incidence functions refer to the life-time or per-recruit quantities such as spawning biomass per recruit ( $\phi_E$ ) or vulnerable biomass per recruit ( $\phi_b$ ). Note that upper and lower case subscripts denote unfished and fished conditions, respectively. Spawning biomass per recruit is given by (T1.9), the vulnerable biomass per recruit is given by (T1.10) and the per recruit yield to the fishery is given by (T1.11). Unfished recruitment is given by (T1.12) and the steady-state equilibrium recruitment for a given fishing mortality rate  $F_e$  is given by (T1.13). Note that in (T1.13) we assume that recruitment follows a Beverton-Holt model of the form:

$$R_e = \frac{s_o R_e \phi_e}{1 + \beta R_e \phi_e}$$

where

$$\begin{aligned} s_o &= \kappa / \phi_E, \\ \beta &= \frac{(\kappa - 1)}{R_o \phi_E}, \end{aligned}$$

which simplifies to (T1.13). The equilibrium yield for a given fishing mortality rate is (T1.14). These steady-state conditions are critical for determining various reference points such as  $F_{\text{MSY}}$  and  $B_{\text{MSY}}$ . The description of calculating steady-state yield for a given value of  $F_e$  in Table A-1 is written assuming that only one fishing fleet exists. The actual calculations are slightly more complicated for the BC herring fishery, as there are three distinct fishing fleets that each have different selectivities. The actual selectivities and calculations of survivorship involve a matrix of age-specific fishing mortalities, where each row of this matrix corresponds to one fishing fleet. In this case  $F_e$  is the total fishing mortality rate for fully selected fish summed over all fleets. In order to calculate the fleet specific fishing mortality rate, a fixed allocation of the total yield must be specified *a priori*.

Table A-1: Steady-state age-structured model assuming unequal vulnerability-at-age, age-specific natural mortality, age-specific fecundity and Beverton-Holt type recruitment. Note that  $M$  is the average natural mortality rate between 1951-2011.

Parameters	
$\Theta = (B_o, \kappa, M, \hat{a}, \dot{\gamma}, F_e)$	(T1.1)
$B_o > 0; \kappa > 1; M > 0; F_e \geq 0$	
$\Phi = (l_\infty, k, t_o, a, b, \dot{a}, \dot{\gamma})$	(T1.2)
Age-schedule information	
$l_a = l_\infty(1 - \exp(-k(a - t_o)))$	(T1.3)
$w_a = a(l_a)^b$	(T1.4)
$v_a = (1 + \exp(-(\hat{a} - a)/\gamma))^{-1}$	(T1.5)
$f_a = w_a(1 + \exp(-(\dot{a} - a)/\dot{\gamma}))^{-1}$	(T1.6)
Survivorship	
$\iota_a = \begin{cases} 1, & a = 1 \\ \iota_{a-1}e^{-M}, & a > 1 \\ \iota_{a-1}/(1 - e^{-M}), & a = A \end{cases}$	(T1.7)
$\hat{\iota}_a = \begin{cases} 1, & a = 1 \\ \hat{\iota}_{a-1}e^{-M - F_e v_{a-1}}, & a > 1 \\ \hat{\iota}_{a-1}e^{-M - F_e v_{a-1}}/(1 - e^{-M - F_e v_a}), & a = A \end{cases}$	(T1.8)
Incidence functions	
$\phi_E = \sum_{a=1}^{\infty} \iota_a f_a, \quad \phi_e = \sum_{a=1}^{\infty} \hat{\iota}_a f_a$	(T1.9)
$\phi_B = \sum_{a=1}^{\infty} \iota_a w_a v_a, \quad \phi_b = \sum_{a=1}^{\infty} \hat{\iota}_a w_a v_a$	(T1.10)
$\phi_q = \sum_{a=1}^{\infty} \frac{\hat{\iota}_a w_a v_a}{M + F_e v_a} \left(1 - e^{(-M - F_e v_a)}\right)$	(T1.11)
Steady-state conditions	
$R_o = B_o/\phi_B$	(T1.12)
$R_e = R_o \frac{\kappa - \phi_E/\phi_e}{\kappa - 1}$	(T1.13)
$C_e = F_e R_e \phi_q$	(T1.14)

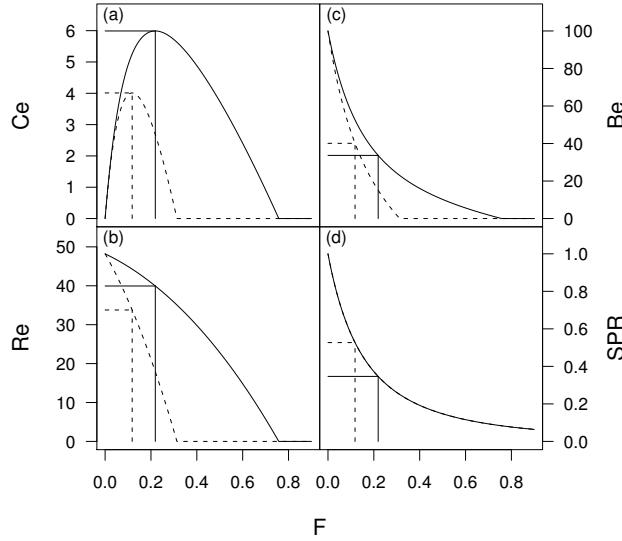


Figure A.48: Equilibrium yield (a), recruits (b), biomass (c) and spawner per recruit ( $\phi_e/\phi_E$ ) (d) versus instantaneous fishing mortality  $F_e$  for two different values of the recruitment compensation ratio ( $\kappa = 12$  solid lines,  $\kappa = 4$  dashed lines). Vertical lines in each panel correspond to  $F_{MSY}$  and horizontal lines correspond to various reference points that would achieve MSY.

### A.1.3 MSY based reference points

$^i\text{SCA}_M$  calculates MSY-based reference points by finding the value of  $F_e$  that results in the zero derivative of the steady-state catch equation (T1.14). This is accomplished numerically using a Newton-Raphson method where an initial guess for  $F_{MSY}$  is set equal to  $1.5M$ , then use (A.1) to iteratively find  $F_{MSY}$ . Note that the partial derivatives in (A.1) can be found in Table A-2.

$$F_{e+1} = F_e - \frac{\frac{\partial C_e}{\partial F_e}}{\frac{\partial^2 C_e}{\partial F_e}} \quad (\text{A.1})$$

where

$$\begin{aligned} \frac{\partial C_e}{\partial F_e} &= R_e \phi_q + F_e \phi_q \frac{\partial R_e}{\partial F_e} + F_e R_e \frac{\partial \phi_q}{\partial F_e} \\ \frac{\partial^2 C_e}{\partial F_e} &= \phi_q \frac{\partial R_e}{\partial F_e} + R_e \frac{\partial \phi_q}{\partial F_e} \end{aligned}$$

The algorithm usually converges in less than 10 iterations depending on how close the initial guess of  $F_{MSY}$  is to the true value. A maximum of 20 iterations are allowed in  $^i\text{SCA}_M$ , however, if  $\frac{\partial C_e}{\partial F_e} < 10^{-5}$  the algorithm stops. Note also, that this is only performed on data type variables and not differentiable variables within AD Model Builder.

Given an estimate of  $F_{MSY}$ , other reference points such as MSY are calculated use the equations in Table A-1 where each of the expressions is evaluated at  $F_{MSY}$ . A graphical representation of MSY based reference points for two alternative values of the recruitment compensation parameter  $\kappa$  is show in Figure A.48.

There are some additional technical details about calculating MSY based reference points when considering multiple fishing gears with different selectivities. The maximum sustainable yield summed over all fishing gears is a function of the selectivities of each gear type and what fraction of the total catch is

Table A-2: Partial derivatives, based on components in Table A-1, required for the numerical calculation of  $F_{MSY}$  using (A.1).

Mortality & Survival	
$Z_a = M + F_e v_a$	(T2.1)
$S_a = 1 - e^{-Z_a}$	(T2.2)
Partial for survivorship	
$\frac{\partial \hat{v}_a}{\partial F_e} = \begin{cases} 0, & a = 1 \\ e^{-Z_{a-1}} \left( \frac{\partial \hat{v}_{a-1}}{\partial F_e} - \hat{v}_{a-1} v_{a-1} \right), & 1 < a < A \\ \frac{\partial \hat{v}_{a-1}}{\partial F_e} - \frac{\hat{v}_{a-1} e^{-Z_{a-1}} v_a e^{-Z_a}}{(1 - e^{-Z_a})^2}, & a = A \end{cases}$	(T2.3)
Partials for incidence functions	
$\frac{\partial \phi_e}{\partial F_e} = \sum_{a=1}^{\infty} f_a \frac{\partial \hat{v}_a}{\partial F_e}$	(T2.4)
$\frac{\partial \phi_q}{\partial F_e} = \sum_{a=1}^{\infty} \frac{w_a v_a S_a}{Z_a} \frac{\partial \hat{v}_a}{\partial F_e} + \frac{\hat{v}_a w_a v_a^2}{Z_a} \left( e^{-Z_a} - \frac{S_a}{Z_a} \right)$	(T2.5)
Partial for recruitment	
$\frac{\partial R_e}{\partial F_e} = \frac{R_o}{\kappa - 1} \frac{\phi_E}{\phi_e^2} \frac{\partial \phi_e}{\partial F_e}$	(T2.6)

allocated to each gear. In the Pacific herring fishery, there are three distinct fleets that all have different selectivities; the purse-seine gears tend to catch smaller younger fish, while the gill net fishery tends to target larger mature females. The optimum fishing mortality rate for each gear that would maximize the yield depends on what the other gears are removing; this in itself is another optimization problem that fisheries management must contend with. For the purposes of this assessment,  ${}^i\text{SCA}_M$  requires an allocation of the total catch (summed across gear type) to each gear before it proceeds with calculating reference points.

For this herring assessment, the average catch over the past 20 years used to determine the allocation scheme for each of the stock assessment regions. For the Strait of Georgia this corresponds to 6.9% for the winter seine fishery, 41.4% for the seine roe fishery, and 51.8% for the gill net fishery. We further assume that 100% of the total mortality takes place prior to spawning, and the start of each biological year is the month of April.

#### A.1.4 Dynamic age-structured model

The estimated parameter vector in  ${}^i\text{SCA}_M$  is defined in (T3.1), where  $R_0$ ,  $\kappa$  and  $M$  are the leading unknown population parameters that define the overall population scale in the form of unfished recruitment and productivity in the form of recruitment compensation and natural mortality. The total variance  $\vartheta^2$  and the proportion of the total variance that is associated with observation errors  $\rho$  are also estimated, then the variance is partitioned into observation errors ( $\sigma^2$ ) and process errors ( $\tau^2$ ) using (T3.2).

The unobserved state variables (T3.3) include the numbers-at-age year year  $t$  ( $N_{t,a}$ ), the spawning stock biomass ( $B_t$ ) and the total age-specific total mortality rate ( $Z_{t,a}$ ).

The initial numbers-at-age in the first year (T3.4) and the annual recruits (T3.5) are treated as estimated parameters and used to initialize the numbers-at-age matrix. Age-specific selectivity for gear type  $k$  is a function of the selectivity parameters  $\gamma_k$  (T3.6), and the annual fishing mortality for each gear  $k$  in year  $t$  ( $F_{k,t}$ ). The vector of log fishing mortality rate parameters  $F_{k,t}$  is a bounded vector with a minimum value

Table A-3: Statistical catch-age model using the Baranov catch equation, where  $R_0$  and  $\kappa$  are the leading parameters that define population scale and productivity, respectively.

Estimated parameters	
$\Theta = \left( R_0, \kappa, M, \bar{R}, \ddot{R}, \rho, \vartheta, \vec{\gamma}_k, F_{k,t}, \{\ddot{\omega}_a\}_{a=\dot{a}+1}^{a=A}, \{\omega_t\}_{t=1}^{t=T}, \{\varphi_t\}_{t=2}^T \right)$	(T3.1)
$\sigma = \rho/\vartheta, \quad \tau = (1 - \rho)/\vartheta$	(T3.2)
Unobserved states	
$N_{t,a}, B_t, Z_{t,a}$	(T3.3)
Initial states ( $t = \dot{t}$ )	
$N_{t,a} = \ddot{R} e^{\ddot{\omega}_a} \exp(-M_t)^{(a-\dot{a})}; \quad t = \dot{t}; \dot{a} \leq a \leq A$	(T3.4)
$N_{t,a} = \bar{R} e^{\omega_t}; \quad \dot{t} \leq t \leq T; a = \dot{a}$	(T3.5)
$v_{k,a} = f(\vec{\gamma}_k)$	(T3.6)
$M_t = M_{t-1} \exp(\varphi_t), \quad t > 1, \varphi_t \sim N(0, \sigma_M)$	(T3.7)
$F_{k,t} = \exp(F_{k,t})$	(T3.8)
State dynamics ( $t > \dot{t}$ )	
$B_t = \sum_a N_{t,a} f_a$	(T3.9)
$Z_{t,a} = M_t + \sum_k F_{k,t} v_{k,t,a}$	(T3.10)
$\hat{C}_{k,t} = \sum_a \frac{N_{t,a} w_a F_{k,t} v_{k,t,a} (1 - e^{-Z_{t,a}})}{Z_{t,a}} e^{\eta_t}$	(T3.11)
$N_{t,a} = \begin{cases} N_{t-1,a-1} \exp(-Z_{t-1,a-1}) & a > \dot{a} \\ N_{t-1,a} \exp(-Z_{t-1,a}) & a = A \end{cases}$	(T3.12)
Recruitment models	
$R_t = \frac{s_o B_{t-k}}{1 + \beta B_{t-k}} e^{\delta_t - 0.5\tau^2} \quad \text{Beverton-Holt}$	(T3.13)
$R_t = s_o B_{t-k} e^{-\beta B_{t-k} + \delta_t - 0.5\tau^2} \quad \text{Ricker}$	(T3.14)

of -30 and an upper bound of 3.0. In arithmetic space this corresponds to a minimum value of 9.36e-14 and a maximum value of 20.01 for annual fishing mortality rates. In years where there are 0 reported catches for a given fleet, no corresponding fishing mortality rate parameter is estimated and the implicit assumption is there was no fishery in that year.

There is an option to treat natural mortality as a random walk process (T3.7), where the natural mortality rate in the first year is the estimated leading parameter (T3.1) and in subsequent years the mortality rate deviates from the previous year based on the estimated deviation parameter  $\varphi_t$ . If the mortality deviation parameters are not estimated, then  $M$  is assumed to be time invariant.

State variables in each year are updated using equations T3.9–T3.12, where the spawning biomass is the product of the numbers-at-age and the mature biomass-at-age (T3.9). The total mortality rate is given by (T3.10), and the total catch (in weight) for each gear is given by (T3.11) assuming that both natural and fishing mortality occur simultaneously throughout the year. The numbers-at-age are propagated over time using (T3.12), where members of the plus group (age  $A$ ) are all assumed to have the same total mortality rate.

Table A-4: An incomplete list of symbols, constants and description for variables used in  $i\text{SCA}_M$ .

Symbol	Constant value	Description
<u>Indexes</u>		
a		index for age
t		index for year
k		index for gear
<u>Model dimensions</u>		
$\acute{a}, A$	2, 10	youngest and oldest age class ( $A$ is a plus group)
$\acute{t}, T$	1951, 2010	first and last year of catch data
K	5	Number of gears including survey gears
<u>Observations (data)</u>		
$C_{k,t}$		catch in weight by gear $k$ in year $t$
$I_{k,t}$		relative abundance index for gear $k$ in year $t$
$p_{k,t,a}$		observed proportion-at-age $a$ in year $t$ for gear $k$
<u>Estimated parameters</u>		
$R_o$		Age- $\acute{a}$ recruits in unfished conditions
$\kappa$		recruitment compensation
M		instantaneous natural mortality rate
$\bar{R}$		average age- $\acute{a}$ recruitment from year $\acute{t}$ to $T$
$\ddot{R}$		average age- $\acute{a}$ recruitment in year $\acute{t} - 1$
$\rho$		fraction of the total variance associated with observation error
$\vartheta$		total precision (inverse of variance) of the total error
$\vec{\gamma}_k$		vector of selectivity parameters for gear $k$
$F_{k,t}$		logarithm of the instantaneous fishing mortality for gear $k$ in year $t$
$\ddot{\omega}_a$		age- $\acute{a}$ deviates from $\ddot{R}$ for year $\acute{t}$
$\omega_t$		age- $\acute{a}$ deviates from $\ddot{R}$ for years $\acute{t}$ to $T$
$\varphi_t$		logarithm of annual change in natural mortality rate
<u>Standard deviations</u>		
$\sigma_M$	0.1	standard deviation in random walk for natural mortality
$\sigma$		standard deviation for observation errors in survey index
$\tau$		standard deviation in process errors (recruitment deviations)
$\sigma_C$	0.0707	standard deviation in observed catch by gear
<u>Residuals</u>		
$\delta_t$		annual recruitment residual
$\eta_t$		residual error in predicted catch

Recruitment to age  $k$  can follow either a Beverton-Holt model (T3.13) or a Ricker model (T3.14) where the maximum juvenile survival rate ( $s_o$ ) in either case is defined by  $s_o = \kappa/\phi_E$ . For the Beverton-Holt model,  $\beta$  is derived by solving (T3.13) for  $\beta$  conditional on estimates of  $\kappa$  and  $R_o$ :

$$\beta = \frac{\kappa - 1}{R_o \phi_E},$$

and for the Ricker model this is given by:

$$\beta = \frac{\ln(\kappa)}{R_o \phi_E}$$

### A.1.5 Options for selectivity

At present, there are eight alternative age-specific selectivity options in  $i\text{SCA}_M$ . The simplest of the selectivity options is a simple logistic function with two parameters where it is assumed that selectivity is time-invariant. The more complex selectivity options assume that selectivity may vary over time a may have as many as (A-1)-T parameters. For time-varying selectivity, cubic and bicubic splines are used to reduce the number of estimated parameters. The last two options consider how selectivity may vary over time based on changes in mean weight-at-age. Prior to parameter estimation,  $i\text{SCA}_M$  will determine the exact number of selectivity parameters that need to be estimated based on which selectivity option was chosen for each gear type. It is not necessary for all gear types to have the same selectivity option. For example it is possible to have a simple two parameter selectivity curve for say a survey gear, and a much more complicated selectivity option for a commercial fishery.

**Logistic selectivity** The logistic selectivity option is a two parameter model of the form

$$v_a = \frac{1}{1 + \exp(-(a - \mu_a)/\sigma_a)}$$

where  $\mu_a$  and  $\sigma_a$  are the two estimated parameters representing the age-at-50% vulnerability and the standard deviation, respectively.

**Age-specific selectivity coefficients** The second option also assumes that selectivity is time-invariant and estimates at total of  $A-1$  selectivity coefficients, where the plus group age-class is assumed to have the same selectivity as the previous age-class. For example, if the ages in the model range from 1 to 15 years, then a total of 14 selectivity parameters are estimated, and age-15+ animals will have the same selectivity as age-14 animals.

When estimating age-specific selectivity coefficients, there are two additional penalties that are added to the objective function that control how much curvature there is and limit how much dome-shaped can occur. To penalize the curvature, the square of the second differences of the vulnerabilities-at-age are added to the objective function:

$$\lambda_k^{(1)} \sum_{a=2}^{A-1} (v_{k,a} - 2v_{k,a-1} + v_{k,a-2})^2 \quad (\text{A.2})$$

The dome-shaped term penalty as:

$$\begin{cases} \lambda_k^{(2)} \sum_{a=1}^{A-1} (v_{k,a} - v_{k,a+1})^2 & (\text{if } v_{k,a+1} < v_{k,a}) \\ 0 & (\text{if } v_{k,a+1} \geq v_{k,a}) \end{cases} \quad (\text{A.3})$$

For this selectivity option the user must specify the relative weights  $(\lambda_k^{(1)}, \lambda_k^{(2)})$  to add to these two penalties.

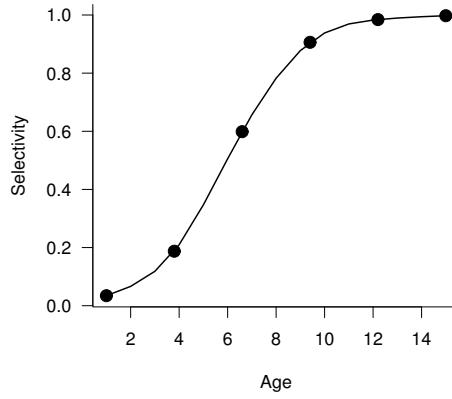


Figure A.49: Example of a natural cubic spline interpolation for 15-selectivity coefficients based on estimating 6 nodes (true selectivity was based on a logistic function). In <sup>i</sup>SCAM the user specifies the number of nodes (e.g., 6 circles) to estimate; then the 15 age-specific selectivity coefficients are interpolated using a natural cubic spline.

**Cubic spline interpolation** The third option also assumes time-invariant selectivity and estimates a selectivity coefficients for a series age-nodes (or spline points) and uses a natural cubic spline to interpolate between these nodes (Figure A.49). Given  $n + 1$  distinct knots  $x_i$ , selectivity can be interpolated in the intervals defined by

$$S(x) = \begin{cases} S_0(x) & x \in [x_0, x_1] \\ S_1(x) & x \in [x_1, x_2] \\ \dots \\ S_{n-1}(x) & x \in [x_{n-1}, x_n] \end{cases}$$

where  $S''(x_0) = S''(x_n) = 0$  is the condition that defines a natural cubic spline.

The same penalty functions for curvature and dome-shaped selectivity are also invoked for the cubic spline interpolation of selectivity.

**Time-varying selectivity with cubic spline interpolation** A fourth option allows for cubic spline interpolation for age-specific selectivity in each year. This option adds a considerable number of estimated parameters but the most extreme flexibility. For example, given 40 years of data and estimated 5 age nodes, this amounts 200 (40 years times 5 ages) estimated selectivity parameters. Note that the only constraints at this time are the dome-shaped penalty and the curvature penalty; there is no constraint implemented for say a random walk (first difference) in age-specific selectivity. As such this option should only be used in cases where age-composition data is available for every year of the assessment.

**Bicubic spline to interpolate over time and ages** The fifth option allows for a two-dimensional interpolation using a bicubic spline (Figure A.50). In this case the user must specify the number of age and year nodes. Again the same curvature and dome shaped constraints are implemented. It is not necessary to have age-composition data each and every year as in the previous case, as the bicubic spline will interpolate between years. However, it is not advisable to extrapolate selectivity back in time or forward in time where there are no age-composition data unless some additional constraint, such as a random-walk in age-specific selectivity coefficients is implemented (as of January 3, 2012, this has not been implemented).

**Selectivity as a logistic function of weight-at-age** The seventh option for selectivity is to parameterize a logistic function in terms of the weight-at-age in year  $t$  ( $w_{a,t}$ ). In this case changes in weight-at-age over

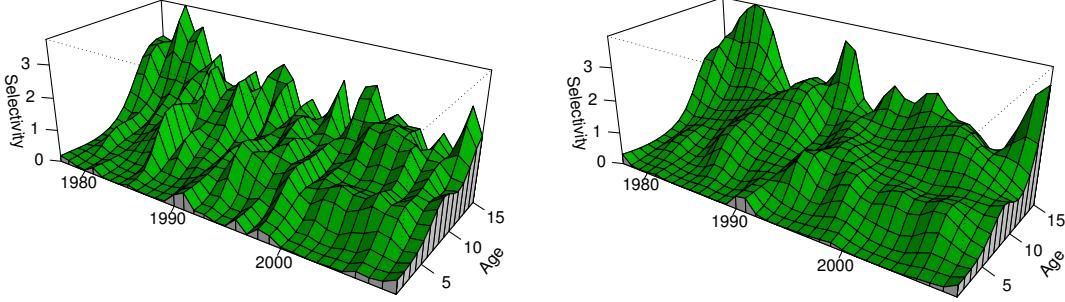


Figure A.50: Example of a time-varying cubic spline (left) and bicubic spline (right) interpolation for selectivity based on data from the Pacific hake. The panel on the left contains 165 estimated selectivity parameters and the bicubic interpolation estimates 85 selectivity parameters, or 5 age nodes and 17 year nodes. There are 495 actual nodes (selectivity parameters) being interpolated.

time allow for changes in selectivity. Such a weight-based function may be appropriate for size selective gears such as gill nets.

$$v_{a,t} = \frac{1}{1 + \exp(-(w_{a,t} - \mu_a)/\sigma_a)}$$

**Using weight as a covariate** The eighth option for selectivity is to use a logistic function based on age, but allow selectivity to vary based on deviations in the mean weight-at-age over time. In this case:

$$v_{a,t} = \frac{1}{1 + \exp(-(a - \mu_a)/\sigma_a)} \exp(\lambda^{(a)} \delta_{a,t})$$

where  $\lambda^{(a)}$  is a latent variable that describes the residual variation in the age-composition data that is due to changes in selectivity, and  $\delta_{a,t}$  is a standardized ( $\mu = 0, \sigma = 1$ ) annual age-specific deviation in mean weight-at-age. In this case, estimates of  $\lambda^{(a)} = 0$  imply that variation in the empirical weight-at-age data explain none of the residual variation in the age-composition data. Values of  $\lambda^{(a)} \neq 0$  imply a positive or negative affect of variation in growth on selectivity.

### A.1.6 Options for natural mortality

There is an option in  $^i\text{SCA}_M$  to estimate a time series of annual changes in natural mortality rates ( $\varphi_t$ ). If not estimated, natural mortality  $M$  is assumed to be invariant over time and age. If, however,  $M$  is thought to vary over time, then  $^i\text{SCA}_M$  models natural mortality as a random walk process (T3.7). In such cases where  $M$  is allowed to freely vary over time, the user must specify two additional components in the control file. First, the phase in which the vector of deviations  $\varphi_t$  is estimated must be specified (use a -ve phase to turn off the estimation), and the user must also specify a standard deviation in the rate of change  $\sigma_M$ . If estimated, then an additional component is added to the objective function to constrain the first differences in the deviation parameters. This first difference constraint only limits how quickly  $M$  may increase or

decrease over time and does not penalize deviations from an underlying mean. Thus it is possible for  $M$  to drift (increase or decrease) away from some central tendency. This drifting can have profound effects on reference point calculations as it also allows for non-stationarity in the underlying production function.

### A.1.7 Residuals, likelihoods & objective function value components

There are 3 major components to the overall objective function that are minimized. These components consist of the likelihood of the data, prior distributions and penalty functions that are invoked to regularize the solution during intermediate phases of the non-linear parameter estimation. This section discusses each of these in turn, starting first with the residuals between observed and predicted states followed by the negative loglikelihood that is minimized for the catch data, relative abundance data, age-composition, and stock-recruitment relationships.

### A.1.8 Catch data

It is assumed that the measurement errors in the non-zero catch observations are log-normally distributed, and the residuals is given by:

$$\eta_{k,t} = \ln(C_{k,t}) - \ln(\hat{C}_{k,t}), \quad (\text{A.4})$$

The residuals are assumed to be normally distributed with a user specified standard deviation  $\sigma_C$ . At present, it is assumed that observed catches for each gear  $k$  is assumed to have the same standard deviation. To aid in parameter estimation, two separate standard deviations are specified in the control file: the first is the assumed standard deviation used in the first, second, to N-1 phases, and the second is the assumed standard deviation in the last phase. The negative loglikelihood (ignoring the scaling constant) for the catch data is given by:

$$\ell_C = \sum_k \left[ T_k \ln(\sigma_C) + \frac{\sum_{t \in \hat{C}_{k,t} \neq 0} (\eta_{k,t})^2}{2\sigma_C^2} \right], \quad (\text{A.5})$$

where  $T_k$  is the total number of non-zero catch observations for gear type  $k$ .

### A.1.9 Relative abundance data

The relative abundance data are assumed to be proportional to biomass that is vulnerable to the sampling gear:

$$V_{k,t} = \sum_a N_{t,a} e^{-\lambda_{k,t} Z_{t,a}} v_{k,a} w_{a,t}, \quad (\text{A.6})$$

where  $v_{k,a}$  is the age-specific selectivity of gear  $k$ , and  $w_a$  is the mean-weight-at-age. A user specified fraction of the total mortality  $\lambda_{k,t}$  adjusts the numbers-at-age to correct for survey timing. In the case of Pacific herring spawn surveys, the vulnerability is fixed to the assumed maturity ogive and the empirical weight-at-age data are used to construct the predicted relative abundance. Also, it was assumed that all the mortality (post-fishing) had occurred during the time the survey took place (i.e.,  $\lambda_{k,t} = 1$ ). The residuals between the observed and predicted relative abundance index is given by:

$$\epsilon_{k,t} = \ln(I_{k,t}) - \ln(q_k) - \ln(V_{k,t}), \quad (\text{A.7})$$

where  $I_{k,t}$  is the observed relative abundance index,  $q_k$  is the catchability coefficient for index  $k$ , and  $V_{k,t}$  is the predicted vulnerable biomass at the time of sampling. The catchability coefficient  $q_k$  is evaluated at its conditional maximum likelihood estimate:

$$q_k = \frac{1}{N_k} \sum_{t \in I_{k,t}} \ln(I_{k,t}) - \ln(V_{k,t}),$$

where  $N_k$  is the number of relative abundance observations for index  $k$  (see [Walters and Ludwig, 1994](#), for more information). The negative loglikelihood for relative abundance data is given by:

$$\ell_I = \sum_k \sum_{t \in I_{k,t}} \ln(\sigma_{k,t}) + \frac{\epsilon_{k,t}^2}{2\sigma_{k,t}^2} \quad (\text{A.8})$$

where

$$\sigma_{k,t} = \frac{\rho\vartheta}{\omega_{k,t}},$$

where  $\rho\vartheta$  is the proportion of the total error that is associated with observation errors, and  $\omega_{k,t}$  is a user specified relative weight for observation  $t$  from gear  $k$ . The  $\omega_{k,t}$  terms allow each observation to be weighted relative to the total error  $\rho\vartheta$ ; for example, to omit a particular observation, set  $\omega_{k,t} = 0$ , or to give 2 times the weight, then set  $\omega_{k,t} = 2.0$ . To assume all observations have the same variance then simply set  $\omega_{k,t} = 1$ . Note that if  $\omega_{k,t} = 0$  then equation (A.8) is undefined; therefore,  $i\text{SCAM}$  adds a small constant to  $\omega_{k,t}$  (1.e-10, which is equivalent to assuming an extremely large variance) to ensure the likelihood can be evaluated.

In the case of the Pacific herring assessment, the spawn survey data post-1988 were assumed to be twice as precise as the pre-dive survey data (1951-1987). To implement this, weights for the 1951-1987 data were set equal to  $\omega_{k,t} = 1.0$  and the contemporary data was assigned  $\omega_{k,t} = 2.0$ . The standard deviation in the observation errors is conditional on estimated values of  $\rho$  and  $\varphi^2$ .

### A.1.10 Age composition data

Sampling theory suggest that age composition data are derived from a multinomial distribution ([Fournier and Archibald, 1982](#)); however,  $i\text{SCAM}$  assumes that age-proportions are obtained from a multivariate logistic distribution ([Schnute and Richards, 1995](#); [Richards et al., 1997](#)). The main reason  $i\text{SCAM}$  departs from the traditional multinomial model has to do with how the age-composition data are weighted in the objective function. First, the multinomial distribution requires the specification of an effective sample size; this may be done arbitrarily or through iterative re-weighting ([McAllister and Ianelli, 1997](#); [Gavaris and Ianelli, 2002](#)), and in the case of multiple and potentially conflicting age-proportions this procedure may fail to converge properly. The assumed effective sample size can have a large impact on the overall model results.

A nice feature of the multivariate logistic distribution is that the age-proportion data can be weighted based on the conditional maximum likelihood estimate of the variance in the age-proportions. Therefore, the contribution of the age-composition data to the overall objective function is “self-weighting” and is conditional on other components in the model.

Ignoring the subscript for gear type for clarity, the observed and predicted proportions-at-age must satisfy the constraint

$$\sum_{a=1}^A p_{t,a} = 1$$

for each year. The multivariate logistic residuals between the observed  $(p_{t,a})$  and predicted proportions  $(\widehat{p}_{t,a})$  is given by:

$$\eta_{t,a} = \ln(p_{t,a}) - \ln(\widehat{p}_{t,a}) - \frac{1}{A} \sum_{a=1}^A [\ln(p_{t,a}) - \ln(\widehat{p}_{t,a})]. \quad (\text{A.9})$$

The conditional maximum likelihood estimate of the variance is given by

$$\widehat{\tau}^2 = \frac{1}{(A-1)T} \sum_{t=1}^T \sum_{a=1}^A \eta_{t,a}^2,$$

and the negative loglikelihood evaluated at the conditional maximum likelihood estimate of the variance is given by:

$$\ell_A = (A-1)T \ln(\widehat{\tau}^2). \quad (\text{A.10})$$

In short, the multivariate logistic likelihood for age-composition data is just the log of the residual variance weighted by the number observations over years and ages.

There is also a technical detail in (A.9), where observed and predicted proportions-at-age must be greater than 0. It is not uncommon in catch-age data sets to observe 0 proportions for older, or young, age classes or weak year classes. In  $i\text{SCA}_M$  the same approach described by Richards et al. (1997) is adopted where the definition of age-classes is altered to require that  $p_{t,a} \geq \hat{p}$  for every age in each year, where  $\hat{p}$  is the minimum percentage specified by the user (e.g.,  $\hat{p} = 0.02$  corresponds to 2%). This is accomplished by grouping consecutive ages, where  $p_{t,a} < \hat{p}$ , into a single age-class and reducing the effective number of age-classes in the variance calculation ( $\hat{\tau}^2$ ) by the number of groups created. The minimum proportion (including 0) is set by the user and can influence the results, especially in cases where there is sparse aging information. In the case of  $\hat{p} = 0$ , the pooling of the adjacent age-class still occurs, this ensures that (A.9) is defined.

In the Strait of Georgia herring example, we set the minimum proportion to 2% to reduce the influence of the large numbers of 0 proportions in the purse-seine fleets, especially prior to 1970 during the reduction fishery.

### A.1.11 Stock-recruitment

There are two alternative stock-recruitment models available in  $i\text{SCA}_M$ : the Beverton-Holt model and the Ricker model. Annual recruitment and the initial age-composition are treated as latent variables in  $i\text{SCA}_M$ , and residuals between estimated recruits and the deterministic stock-recruitment models are used to estimate unfished spawning stock biomass and recruitment compensation. The residuals between the estimated and predicted recruits is given by

$$\delta_t = \ln(\bar{R}e^{w_t}) - \ln(f(B_{t-\acute{a}})) \quad (\text{A.11})$$

where  $f(B_{t-k})$  is given by either (T3.13) or (T3.14), and  $\acute{a}$  is the age at recruitment. Note that a bias correction term for the lognormal process errors is included in (T3.13) and (T3.14).

The negative log likelihood for the recruitment deviations is given by the normal density (ignoring the scaling constant):

$$\ell_\delta = n \ln(\tau) + \frac{\sum_{t=1+k}^T \delta_t^2}{2\tau^2} \quad (\text{A.12})$$

Equations (A.11) and (A.12) are key for estimating unfished spawning stock biomass and recruitment compensation via the recruitment models. The relationship between  $(s_o, \beta)$  and  $(B_o, \kappa)$  is defined as:

$$s_o = \kappa/\phi_E \quad (\text{A.13})$$

$$\beta = \begin{cases} \frac{\kappa-1}{B_o} & \text{Beverton-Holt} \\ \frac{\ln(\kappa)}{B_o} & \text{Ricker} \end{cases} \quad (\text{A.14})$$

where  $s_o$  is the maximum juvenile survival rate,  $\beta$  is the density effect on recruitment, and  $B_o$  is the unfished spawning stock biomass. Unfished steady-state spawning stock biomass per recruit is given by  $\phi_E$ , which is the sum of products between age-specific survivorship and relative fecundity. In cases where the natural mortality rate is allowed to vary over time, the calculation of  $\phi_E$ , and the corresponding unfished spawning stock biomass ( $B_o$ ) is based on the average natural mortality rate over the entire time period. This subtle calculation has implications for reference point calculations in cases where there are increasing or decreasing trends in natural mortality rates over time; as estimates of natural mortality rates trend upwards, estimates of  $B_o$  decrease.

For the Strait of Georgia Pacific herring example, only the Beverton-Holt recruitment model was considered. The description of the Ricker model is included here for the sake of completely documenting the features in the  $i\text{SCA}_M$  platform.

### A.1.12 Parameter Estimation and Uncertainty

Parameter estimation and quantifying uncertainty was carried out using the tools available in AD Model Builder ([ADMB Project, 2009](http://admb-project.org/)). AD Model Builder (ADMB) is a software for creating computer programs to estimate the parameters and associated probability distributions for nonlinear statistical models. The software is freely available from <http://admb-project.org/>. This software was used to develop  $i\text{SCAM}$ , and the source code and documentation for  $i\text{SCAM}$  is freely available from <https://sites.google.com/site/iscampoint/>, or from a subversion repository at <http://code.google.com/p/iscam-project/>.

Suffice it to say that there is a lot more going on in the  $i\text{SCAM}$  software than just minimizing the sum of the four negative loglikelihood functions defined in the previous section. There are actually five distinct components that make up the objective function that ADMB is minimizing:

$$f = \text{negative loglikelihoods} + \text{constraints} + \text{priors for parameters} + \text{survey priors} + \text{convergence penalties}.$$

The purpose of this section is to completely document all of the components that make up the objective function. Such transparency is absolutely necessary to better understand estimation performance, as well as, to ensure the results are repeatable.

### A.1.13 Negative loglikelihoods

The negative loglikelihoods pertain specifically elements that deal with the data and variance partitioning and have already been described in detail in section A.1.7. There are four specific elements that make up the vector of negative loglikelihoods:

$$\vec{\ell} = \ell_C, \ell_I, \ell_A, \ell_\delta. \quad (\text{A.15})$$

To reiterate, these are the likelihood of the catch data  $\ell_C$ , likelihood of the survey data  $\ell_I$ , the likelihood of the age-composition data  $\ell_A$  and the likelihood of the stock-recruitment residuals  $\ell_\delta$ . Each of these elements are expressed in negative log-space, and ADMB attempts to estimate model parameters by minimizing the sum of these elements.

### A.1.14 Constraints

There are two specific constraints that are described here: 1) parameter bounds, and 2) constraints to ensure that a parameter vector sums to 0. In  $i\text{SCAM}$  the user must specify the lower and upper bounds for the leading parameters defined in the control file ( $\ln(R_o)$ ,  $h$ ,  $\ln(M)$ ,  $\ln(\bar{R})$ ,  $\rho$ ,  $\vartheta$ ). All estimated selectivity parameters  $\vec{\gamma}_k$  are estimated in log space and have a minimum and maximum values of -5.0 and 5.0, respectively. These values are hard-wired into the code, but should be sufficiently large/small enough to capture a wide range of selectivities. Estimated fishing mortality rates are also constrained (in log space) to have a minimum value of -30, and a maximum value of 3.0. Log annual recruitment deviations are also constrained to have minimum and maximum values of -15.0 and 15.0 and there is an additional constraint to ensure the vector of deviations sums to 0. This is necessary in order to be able to estimate the average recruitment  $\bar{R}$ . Finally, the annual log deviations in natural mortality rates are constrained to lie between -2.0 and 2.0.

An array of selectivity parameters (i.e., `init_bounded_matrix_vector`) is estimated within  $i\text{SCAM}$  where each matrix corresponds to a specific gear type, and the number of rows and columns of each depends on the type of selectivity function assumed for the gear and if that selectivity changes over time. In cases where the nodes of a spline are estimated these nodes also have an additional constraint to sum to 0. This is effectively implemented by adding to the objective function:

$$1000 \left( \frac{1}{N_{\vec{\lambda}_k}} \sum \vec{\lambda}_k \right)^2.$$

This additional constraint is necessary to ensure the model remains separable and the annual fishing mortality rates are less confounded with selectivity parameters.

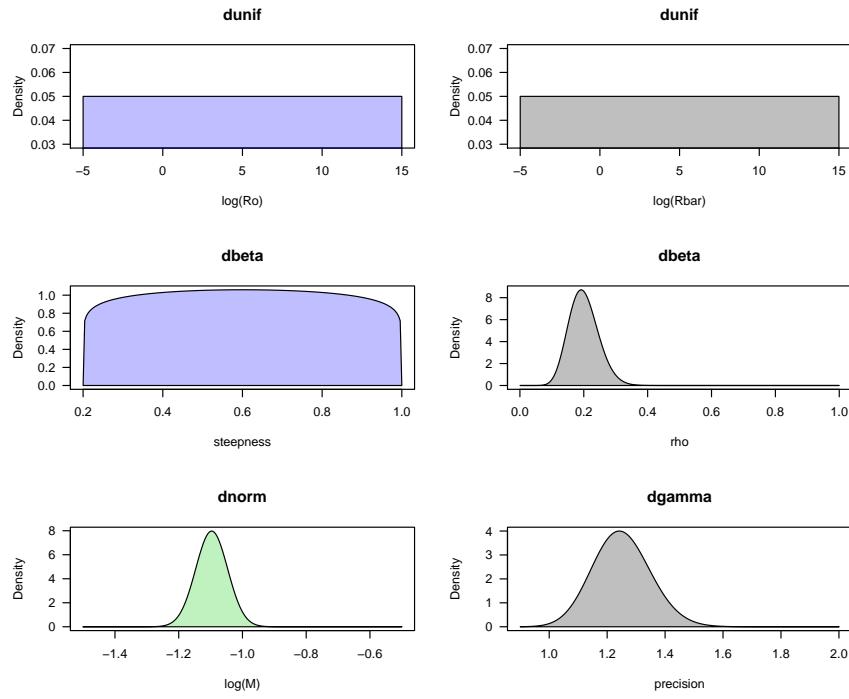


Figure A.51: Prior distributions used for  $\ln(R_o)$ ,  $h$ ,  $\ln(M)$ ,  $\ln(\bar{R})$ ,  $\rho$ ,  $\vartheta$  in the herring assessment models.

### A.1.15 Priors for parameters

Each of the six leading parameters specified in the control file ( $\ln(R_o)$ ,  $h$ ,  $\ln(M)$ ,  $\ln(\bar{R})$ ,  $\rho$ ,  $\vartheta$ ) are declared as bounded parameters and in addition the user can also specify an informative prior distribution for each of these parameters. Five distinct prior distributions can be implemented: uniform, normal, lognormal, beta and a gamma distribution. For the Strait of Georgia herring, a bounded uniform prior was specified for the log of unfished recruitment  $U(-5.0, 15)$ , a vague beta prior was assumed for steepness  $Beta(1.01, 1.01)$ , a normal prior was specified for the log of natural mortality rate  $N(-1.0966, 0.05)$ , a bounded uniform prior for the log of average recruitment  $U(-5.0, 15.0)$ , a beta prior for the variance partitioning parameter  $\rho$   $Beta(15, 60)$ , and a gamma prior for the precision parameter  $\vartheta$ ,  $Gamma(156.25, 125.0)$ . These prior distributions based on the parameter specified above are shown in Figure A.51.

In addition to the priors specified for the six leading parameter, there are several other informative distributions that are invoked for the non-parametric selectivity parameters. In cases where age-specific selectivity coefficients are estimated, or nodes of a spline function are estimated, two additional penalties are added to the objective function to control how smooth the selectivity changes (A.2) and how much dome-shape is allowed in the nonparametric selectivities (A.3).

### A.1.16 Survey priors

The scaling parameter  $q$  for each of the surveys is not treated as an unknown parameter within the code; rather, the maximum likelihood estimate for  $q$  conditional on all other parameters is used to scale the predicted spawning biomass to the observed spawn survey index. In the case of Pacific herring, the relationship between fecundity and mature female biomass is relatively invariant at about 200 eggs per gram (Hay, 1985; Hardwick, 1973). This relationship has been used to convert total egg deposition from the spawn survey to total female spawning biomass, and assuming all spawning was accounted for, then a reasonable estimate for  $q$  should be 1.0.

In the Strait of Georgia herring assessment, we specified an informative normal prior on  $\ln(q)$  with a mean of 0, and a standard deviation of 0.05 for the contemporary data. For the pre-1988 spawn survey

data, we explored three alternative priors including a non-informative prior, and a normal prior with a mean 0 and standard deviations of 0.05, or 0.1. The informative prior for the contemporary data implies a 95% confidence interval of 0.82 to 1.22 for  $q$ .

### A.1.17 Convergence penalties

For the Strait of Georgia herring assessment, there are well over 200 estimated parameters, the exact number depends on the model configuration. Needless to say, non-linear parameter estimation is often very sensitive to the initial starting conditions, and the end results may differ depending on the initial values of the model parameters or even the phase at which parameters are included into the estimation problem. There is no guarantee that the algorithm will converge to the global minimum every time. AD Model Builder is unique in that the estimation process can be conducted in a series of phases where more and more parameters are ‘freed up’ as the model progress through each phase. Furthermore, the actual objective function can change between phases such that during the initial phases large penalties can be used to, as Dave Fournier would say, “regularize the solution”. For example, in the initial phases of parameter estimation  $i\text{SCA}_M$  uses fairly steep quadratic penalties for the annual recruitment deviations and average fishing mortality rates to initially aid in finding reasonable values of the average recruitment, natural mortality and selectivity parameters. In the final phase, these quadratic penalties are relaxed.

In the case of the annual recruitment deviations, the quadratic penalty term is:

$$100 \sum_{t=1-A}^T \omega_t^2,$$

which is approximately a normal density with a standard deviation equal to 0.07. In the last phase this constraint is relaxed with a large standard deviation of 5.0.

A similar penalty (a normal distribution for the log mean fishing rate) is also invoked for the mean fishing mortality rate, but in this case the user specifies the mean fishing mortality rate and the standard deviations in the initial phases and the last phase. Normally, a rather small standard deviation is used in the initial phases (e.g., 0.01) and this is then relaxed to a much larger value (e.g., 5.0) in the last phase. These standard deviations are specified by the user in the control file.

## B.2 Data and Control files

### B.2.1 Haida Gwaii

```

#NB The data herein were taken from qc12010_final.dat for the HCAM model.
## -----
## ____Model Dimensions_____
1951 #first year of data
2011 #last year of data
2 #age of youngest age class
10 #age of plus group
5 #number of gears (ngear)
## flags for fishery (1) or survey (0) in ngears
#1 1 0 0
0.0686819 0.4137555 0.5175626 0.0000000 0.0000000
## -----
## _____
#Age-schedule and population parameters
#natural mortality rate (m)
0.334
#growth parameters (linf,k,to) (from fishbase)
27.0, 0.48, 0
#length-weight allometry (a,b)
4.5e-6, 3.1270
#maturity at age (am=log(3)/k) & gm=std for logistic
2.055, 0.05
## -----
#Time series data
#Observed catch (1951-2011, 1000s metric t)
#Year P1 P2 P3 S1 S2
1951 2.847 0.000 0.000 0 0
1952 10.147 0.000 0.000 0 0
1953 0.000 0.000 0.000 0 0
1954 1.786 0.000 0.000 0 0
1955 1.234 0.000 0.000 0 0
1956 77.681 0.000 0.000 0 0
1957 23.711 0.000 0.000 0 0
1958 11.168 0.000 0.000 0 0
1959 7.027 0.000 0.000 0 0
1960 0.000 0.000 0.000 0 0
1961 0.653 0.000 0.000 0 0
1962 7.632 0.000 0.000 0 0
1963 14.980 0.000 0.000 0 0
1964 28.777 0.000 0.000 0 0
1965 35.448 0.000 0.000 0 0
1966 2.746 0.000 0.000 0 0
1967 0.213 0.000 0.000 0 0
1968 0.080 0.000 0.000 0 0
1969 0.000 0.000 0.000 0 0
1970 0.000 0.000 0.000 0 0
1971 0.102 0.000 0.000 0 0
1972 0.849 3.124 0.000 0 0
1973 0.000 7.520 0.000 0 0
1974 0.000 6.191 0.127 0 0
1975 0.000 7.619 0.105 0 0
1976 0.374 11.939 1.802 0 0
1977 0.000 11.146 1.489 0 0
1978 0.000 9.172 2.553 0 0
1979 0.000 5.867 2.086 0 0
1980 0.000 2.106 1.210 0 0
1981 0.000 3.926 1.705 0 0
1982 0.000 2.371 1.407 0 0
1983 0.067 4.661 0.929 0 0
1984 0.096 4.016 0.535 0 0
1985 0.000 4.616 1.493 0 0
1986 0.000 2.613 0.890 0 0
1987 0.000 2.061 0.000 0 0
1988 0.000 0.032 0.000 0 0
1989 0.000 1.461 0.000 0 0
1990 0.000 5.542 1.170 0 0
1991 0.000 3.899 0.543 0 0
1992 0.000 2.524 0.000 0 0
1993 0.000 2.699 0.000 0 0
1994 0.000 0.299 0.000 0 0
1995 0.000 0.000 0.000 0 0
1996 0.000 0.000 0.000 0 0
1997 0.000 0.000 0.000 0 0
1998 0.000 1.372 0.000 0 0
1999 0.000 2.500 0.473 0 0
2000 0.000 1.764 0.000 0 0
2001 0.000 0.000 0.000 0 0
2002 0.000 0.706 0.000 0 0
2003 0.000 0.000 0.000 0 0
2004 0.000 0.000 0.000 0 0
2005 0.000 0.000 0.000 0 0
2006 0.000 0.000 0.000 0 0
2007 0.000 0.000 0.000 0 0
2008 0.000 0.000 0.000 0 0
2009 0.000 0.000 0.000 0 0
2010 0.000 0.000 0.000 0 0
2011 0.000 0.000 0.000 0 0
#Relative Abundance index from fisheries independent survey (it) 1970-2008
#nit
#it
#2
#nit_nobs
37 24
#Survey type
## 1 = survey is proportional to vulnerable numbers
## 2 = survey is proportional to vulnerable biomass
## 3 = survey is proportional to spawning biomass (e.g., herring spawn survey)
3 3
#it gear wt survey timing

```

```

1990 2 11 487 918 3033 199 93 193 86 14
1991 2 227 140 361 972 1303 125 61 135 51
1992 2 23 1243 159 270 402 992 77 19 27
1993 2 12 128 2240 165 225 448 436 43 9
1994 2 75 52 61 590 129 133 132 39 5
1995 2 68 75 11 21 178 46 38 25 11
1996 2 103 515 89 31 32 149 23 11 2
1997 2 372 430 549 86 25 73 88 14 3
1998 2 10 1470 758 315 73 18 33 30 7
1999 2 108 58 1610 433 204 64 16 10 9
2000 2 107 398 84 1270 171 97 9 10 3
2001 2 175 363 256 58 240 35 16 3 1
2002 2 602 750 706 369 86 371 42 13 3
2003 2 2 1685 453 159 80 28 52 10 3
2004 2 248 20 428 74 34 22 12 5 2
2005 2 17 606 205 374 51 31 16 6 3
2006 2 136 72 305 67 108 20 3 0 2
2007 2 6 247 78 114 32 56 12 1 0
2008 2 86 68 583 70 79 17 15 0 2
2009 2 1 645 76 222 20 29 4 5 1
2010 2 39 70 644 62 170 18 13 3 2
2011 2 21 522 90 371 65 100 9 4 0
1974 3 0 9 76 40 26 5 0 0 1
1975 3 0 0 9 16 12 2 1 0 0
1976 3 0 0 1 29 81 19 3 0 0
1978 3 0 0 6 17 29 56 29 6 1
1979 3 0 0 48 44 46 26 6 0 0
1980 3 0 29 27 229 104 93 27 9 0
1981 3 0 2 583 61 77 44 19 4 0
1982 3 0 1 16 425 16 11 5 2 0
1983 3 0 0 7 14 532 16 14 3 3
1984 3 0 11 5 18 35 313 7 1 1
1985 3 0 20 59 7 11 30 113 1 0
1986 3 0 0 41 172 13 17 29 49 1
1990 3 0 2 32 174 39 33 68 33 2
1991 3 0 0 8 79 153 34 25 36 16
1999 3 0 4 185 137 175 60 16 8 11
#n_wt_obs
61
#Mean weight-at-age in kilograms (interpolated: from HCAM.rep)
#>$yr V1 V2 V3 V4 V5 V6 V7 V8 V9
1951 0.0417 0.0897 0.1126 0.1376 0.1586 0.1706 0.2001 0.1865 0.1865
1952 0.0432 0.0896 0.1133 0.1385 0.1599 0.1756 0.1677 0.1780 0.1865
1953 0.0329 0.0763 0.0958 0.1255 0.1498 0.1599 0.1335 0.1784 0.1865
1954 0.0447 0.0847 0.1038 0.1278 0.1571 0.1721 0.1751 0.1860 0.1865
1955 0.0499 0.0893 0.1043 0.1262 0.1448 0.1738 0.1720 0.1784 0.1865
1956 0.0471 0.0845 0.1060 0.1205 0.1410 0.1587 0.1785 0.1733 0.1865
1957 0.0481 0.0828 0.1084 0.1371 0.1506 0.1640 0.2100 0.1784 0.1865
1958 0.0438 0.0763 0.1083 0.1421 0.1538 0.1647 0.1762 0.1888 0.1865
1959 0.0502 0.0836 0.0993 0.1314 0.1598 0.1677 0.1523 0.1780 0.1865
1960 0.0532 0.0902 0.1128 0.1279 0.1597 0.1290 0.1770 0.1784 0.1865
1961 0.0576 0.0786 0.1111 0.1170 0.1372 0.1650 0.1720 0.1784 0.1865
1962 0.0501 0.0880 0.0993 0.1366 0.1449 0.1530 0.2020 0.1784 0.1865
1963 0.0477 0.0832 0.1049 0.1116 0.1494 0.1778 0.1790 0.1784 0.1865
1964 0.0576 0.0993 0.1159 0.1378 0.1586 0.1850 0.1400 0.1784 0.1865
1965 0.0611 0.1051 0.1224 0.1451 0.1452 0.1770 0.1440 0.1784 0.1865
1966 0.0520 0.1067 0.1487 0.1681 0.1783 0.1888 0.1720 0.1784 0.1865
1967 0.0484 0.0880 0.1108 0.1332 0.1532 0.1677 0.1720 0.1784 0.1865
1968 0.0488 0.0880 0.1104 0.1332 0.1530 0.1677 0.1720 0.1784 0.1865
1969 0.0488 0.0880 0.1104 0.1332 0.1530 0.1677 0.1720 0.1784 0.1865
1970 0.0578 0.0888 0.1175 0.1399 0.1608 0.1733 0.1871 0.1927 0.2096
1971 0.0553 0.1056 0.1319 0.1537 0.1742 0.1895 0.1690 0.1927 0.2096
1972 0.0586 0.0897 0.1290 0.1478 0.1671 0.1778 0.1973 0.1850 0.2096
1973 0.0525 0.0993 0.1274 0.1588 0.1749 0.1928 0.1977 0.2010 0.2096
1974 0.0632 0.0877 0.1392 0.1653 0.2093 0.1733 0.2420 0.1927 0.2096
1975 0.0421 0.0828 0.1115 0.1423 0.1698 0.1934 0.1998 0.2122 0.2096
1976 0.0498 0.0849 0.1234 0.1455 0.1743 0.1914 0.2033 0.2248 0.2133
1977 0.0571 0.0888 0.1173 0.1389 0.1616 0.1910 0.2010 0.2027 0.2418
1978 0.0501 0.0853 0.1099 0.1321 0.1513 0.1672 0.1683 0.1999 0.2095
1979 0.0597 0.0848 0.1178 0.1405 0.1604 0.1754 0.1990 0.2051 0.2174
1980 0.0522 0.0780 0.1083 0.1352 0.1588 0.1718 0.1831 0.1984 0.2172
1981 0.0609 0.0857 0.1083 0.1348 0.1566 0.1706 0.1873 0.1802 0.2022
1982 0.0635 0.0945 0.1153 0.1292 0.1532 0.1618 0.1698 0.1688 0.1950
1983 0.0576 0.0898 0.1174 0.1349 0.1431 0.1580 0.1707 0.1851 0.1941
1984 0.0624 0.0881 0.1146 0.1345 0.1564 0.1703 0.1869 0.1926 0.2320
1986 0.0670 0.0895 0.1108 0.1324 0.1495 0.1700 0.1971 0.1954 0.2096
1987 0.0628 0.0874 0.1056 0.1226 0.1366 0.1522 0.1660 0.1580 0.1823
1988 0.0609 0.0900 0.1138 0.1302 0.1423 0.1547 0.1642 0.2006 0.2096
1989 0.0644 0.0841 0.1068 0.1273 0.1398 0.1474 0.1561 0.1580 0.1820
1990 0.0604 0.0851 0.1068 0.1283 0.1465 0.1572 0.1600 0.1453 0.2260
1991 0.0642 0.0894 0.1101 0.1280 0.1426 0.1553 0.1632 0.1508 0.1850
1992 0.0598 0.0905 0.1118 0.1319 0.1494 0.1592 0.1738 0.1739 0.1563
1993 0.0579 0.0922 0.1118 0.1286 0.1411 0.1529 0.1565 0.1600 0.1475
1994 0.0522 0.0809 0.1047 0.1207 0.1350 0.1405 0.1512 0.1609 0.1563
1995 0.0612 0.0856 0.1106 0.1314 0.1446 0.1617 0.1631 0.1791 0.1753
1996 0.0619 0.0829 0.1061 0.1264 0.1457 0.1557 0.1719 0.1694 0.1830
1997 0.0497 0.0826 0.1017 0.1197 0.1368 0.1463 0.1541 0.1670 0.1817
1998 0.0504 0.0725 0.0939 0.1086 0.1194 0.1331 0.1438 0.1559 0.1485
1999 0.0460 0.0795 0.0991 0.1134 0.1255 0.1341 0.1434 0.1508 0.1390
2000 0.0523 0.0724 0.0948 0.1110 0.1291 0.1387 0.1530 0.1599 0.1630
2001 0.0622 0.0853 0.0990 0.1197 0.1332 0.1483 0.1552 0.1454 0.1440
2002 0.0494 0.0793 0.0959 0.1076 0.1284 0.1324 0.1413 0.1640 0.0590
2003 0.0518 0.0772 0.0930 0.1050 0.1113 0.1282 0.1401 0.1280 0.1563
2004 0.0498 0.0729 0.0888 0.0993 0.1091 0.1125 0.1241 0.1219 0.1320
2005 0.0476 0.0741 0.0910 0.1064 0.1175 0.1257 0.1305 0.1215 0.1373
2006 0.0483 0.0714 0.0882 0.1023 0.1113 0.1208 0.1279 0.1394 0.1780
2007 0.0620 0.0753 0.0831 0.0994 0.1154 0.1230 0.1301 0.1432 0.1340
2008 0.0257 0.0658 0.0868 0.0936 0.1030 0.1105 0.1155 0.1331 0.1350
2009 0.0453 0.0644 0.0868 0.1033 0.1156 0.1251 0.1347 0.1542 0.1780
2010 0.0435 0.0580 0.0790 0.0850 0.1118 0.1189 0.1158 0.1119 0.1413
2011 0.0536 0.0697 0.0803 0.0985 0.1116 0.1206 0.1287 0.1288 0.1500
#eof
999
## -----
## SOG HERRING CONTROLS
## CONTROLS FOR ESTIMATED PARAMETERS
## Prior descriptions:
## -o uniform (0,0)
## -1 normal (p1=mu,p2=sig)
## -2 lognormal (p1=log(mu),p2=sig)
## -3 beta (p1=alpha,p2=beta)
## -4 gamma(p1=alpha,p2=sig)
## -----
## npar ival lb ub phz prior p1 p2 parameter name
## 7 15 4 0 -5.0 15 #log_ro
## 0.67 0.2 1.0 4 3 10.0 4.925373 #steepness
## -0.7985077 -5.0 5.0 3 1 -0.7985077 0.4 #log_m
## 7.40 -5.0 15 1 0 -5.0 15 #log_avgrec
## 7.20 -5.0 15 1 0 -5.0 15 #log_recinit
## 0.3043478 0.001 0.999 3 3 17.08696 39.0559 #rho
## 0.8695652 0.01 5.0 3 4 25.0 28.75 #kappa (precision)
## -----
## SELECTIVITY PARAMETERS
## OPTIONS FOR SELECTIVITY:
## 1) logistic selectivity parameters
## 2) selectivity coefficients
## 3) a constant cubic spline with age-nodes
## 4) a time varying cubic spline with age-nodes
## 5) a time varying bicubic spline with age & year nodes.
## 6) fixed logistic (set isel_type=6, and estimation phase to -1)
## 7) logistic function of body weight.
## sign=0.05 0.10 0.20 0.30 0.40 0.50
## wt=200. 50.0 22.2 12.5 5.56 3.12 2.00
## Gear 1:3 fishery: Gear 4-5 survey
## isel_type
## 1 1 8 6 6
## Age at 50% selectivity (logistic)
## 3.0 3.0 0.6 2.055 2.055
## STD at 50% selectivity (logistic)
## 0.25 0.25 0.15 0.05 0.05
## No. of age nodes for each gear (0 to ignore).
## 5 5 5 0 0
## No. of year nodes for each gear (0 to ignore).
## 12 3 10 0 0
## Estimation phase
## 2 2 2 -1 -1
## Penalty weight for 2nd differences w1/(2*sig^2)
## 125. 125. 12.5 12.5 12.5
## Penalty weight for dome-shaped selectivity 1/(2*sig^2)
## 50.0 50.0 200.0 200.0 200.0
## -----
## Priors for Survey q
## nits #number of surveys
## 2
## priors 0=uniform density 1=normal density
## 1 1
## prior log(mean)
## -0.569 -0.569
## prior sd
## 0.274 0.274
## -----
## OTHER MISCELLANEOUS CONTROLS
## 0 ## 1 verbose ADMB output (0=off, 1=on)
## 1 ## 2 recruitment model (1=beverton-holt, 2=ricker)
## 0.100 ## 3 std in observed catches in first phase.
## 0.0707 ## 4 std in observed catches in last phase.
## 0 ## 5 Assume unfished in first year (0=FALSE, 1=TRUE)
## 0.05 ## 6 Minimum proportion to consider in age-proportions for dmvlogistic
## 0.20 ## 7 Mean fishing mortality for regularizing the estimates of Ft
## 0.01 ## 8 std in mean fishing mortality in first phase
## 2.00 ## 9 std in mean fishing mortality in last phase
## 3 ## 10 phase for estimating _m_deviations (use -1 to turn off mdevs)
## 0.1 ## 11 std in deviations for natural mortality
## 12 ## 12 number of estimated nodes for deviations in natural mortality
## 1.00 ## 13 fraction of total mortality that takes place prior to spawning
## 1 ## 14 switch for age-composition likelihood (1=dmvlogistic, 2=dmultinom)
## -----
## eofc
999

```

## B.2.2 Prince Rupert District

```

#NB The data herein were taken from qci2010_final.dat for the HCAM model.
## -----
## ____Model Dimensions_____
1951 #first year of data
2011 #last year of data
2 #age of youngest age class
10 #age of plus group
5 #number of gears (ngear)
## flags for fishery (1) or survey (0) in ngears
#1 1 0 0
0.0686819 0.4137555 0.5175626 0.0000000 0.0000000
## -----
## _____
##age-schedule and population parameters
#natural mortality rate (m)
0.334
#growth parameters (linf,k,to) (from fishbase)
27.0, 0.48, 0
#length-weight allometry (a,b)
4.5e-6, 3.1270
#maturity at age (am=log(3)/k) & gm=std for logistic
2.055, 0.05
## -----
#time series data
#observed catch (1951-2010, 1000s metric t)
#Year P1 P2 P3 S1 S2
1951 45.865 0.000 0.000 0 0
1952 52.379 0.000 0.000 0 0
1953 1.865 0.000 0.000 0 0
1954 27.277 0.000 0.000 0 0
1955 17.806 0.000 0.000 0 0
1956 10.182 0.000 0.000 0 0
1957 28.035 0.000 0.000 0 0
1958 4.523 0.000 0.000 0 0
1959 10.224 0.000 0.000 0 0
1960 18.476 0.000 0.000 0 0
1961 42.746 0.000 0.000 0 0
1962 27.660 0.000 0.000 0 0
1963 40.228 0.000 0.000 0 0
1964 29.930 0.000 0.000 0 0
1965 44.211 0.000 0.000 0 0
1966 17.295 0.000 0.000 0 0
1967 7.998 0.000 0.000 0 0
1968 2.068 0.000 0.000 0 0
1969 0.000 0.000 0.000 0 0
1970 1.330 0.000 0.000 0 0
1971 3.500 0.000 0.000 0 0
1972 0.877 3.613 0.004 0 0
1973 0.218 1.388 0.000 0 0
1974 0.182 2.122 1.515 0 0
1975 0.155 1.536 0.011 0 0
1976 0.564 3.466 0.276 0 0
1977 0.792 5.856 1.494 0 0
1978 3.582 1.974 3.031 0 0
1979 1.810 1.271 1.236 0 0
1980 0.738 1.641 1.046 0 0
1981 1.682 1.051 0.356 0 0
1982 1.815 0.170 0.000 0 0
1983 0.000 0.000 0.000 0 0
1984 0.173 1.653 1.880 0 0
1985 0.253 3.018 3.476 0 0
1986 0.375 3.732 4.573 0 0
1987 0.122 2.077 4.071 0 0
1988 0.079 3.550 4.340 0 0
1989 0.071 3.657 4.745 0 0
1990 0.043 2.285 2.361 0 0
1991 0.000 1.366 2.143 0 0
1992 0.142 1.238 3.797 0 0
1993 0.000 2.208 4.112 0 0
1994 0.000 2.363 2.324 0 0
1995 0.000 0.706 1.355 0 0
1996 0.000 0.000 3.086 0 0
1997 0.000 0.000 5.541 0 0
1998 0.000 0.000 3.217 0 0
1999 0.000 0.266 1.859 0 0
2000 0.000 1.239 3.076 0 0
2001 0.000 1.012 1.906 0 0
2002 0.000 2.061 2.432 0 0
2003 0.000 1.451 2.562 0 0
2004 0.000 1.919 2.192 0 0
2005 0.000 1.750 2.050 0 0
2006 0.000 0.957 1.661 0 0
2007 0.000 0.000 0.969 0 0
2008 0.000 0.513 1.148 0 0
2009 0.000 0.713 1.286 0 0
2010 0.000 0.475 1.010 0 0
2011 0.000 0.883 1.264 0 0
#Relative Abundance index from fisheries independent survey (it) 1970-2008
#nit
#61
2
#nit_nobs
37 24
#survey type
## 1 = survey is proportional to vulnerable numbers
## 2 = survey is proportional to vulnerable biomass
## 3 = survey is proportional to spawning biomass (e.g., herring spawn survey)
3 3
#it gear wt survey timing
1951 27.149 4 1 1
1952 24.047 4 1 1
1953 28.468 4 1 1
1954 13.535 4 1 1
1955 14.482 4 1 1
1956 14.533 4 1 1
1957 27.518 4 1 1
1958 9.882 4 1 1
1959 40.961 4 1 1
1960 16.545 4 1 1
1961 12.059 4 1 1
1962 26.329 4 1 1
1963 16.981 4 1 1
1964 26.919 4 1 1
1965 6.055 4 1 1
1966 7.105 4 1 1
1967 3.386 4 1 1
1968 5.197 4 1 1
1969 0.965 4 1 1
1970 8.814 4 1 1
1971 8.480 4 1 1
1972 8.774 4 1 1
1973 10.959 4 1 1
1974 9.244 4 1 1
1975 10.949 4 1 1
1976 15.587 4 1 1
1977 11.589 4 1 1
1978 6.164 4 1 1
1979 9.195 4 1 1
1980 11.937 4 1 1
1981 14.087 4 1 1
1982 17.186 4 1 1
1983 25.247 4 1 1
1984 27.041 4 1 1
1985 41.028 4 1 1
1986 26.638 4 1 1
1987 39.905 4 1 1
1988 35.444 5 1.1666 1
1989 16.379 5 1.1666 1
1990 22.679 5 1.1666 1
1991 25.811 5 1.1666 1
1992 40.145 5 1.1666 1
1993 25.071 5 1.1666 1
1994 16.589 5 1.1666 1
1995 18.516 5 1.1666 1
1996 24.854 5 1.1666 1
1997 25.037 5 1.1666 1
1998 19.420 5 1.1666 1
1999 29.745 5 1.1666 1
2000 19.694 5 1.1666 1
2001 36.684 5 1.1666 1
2002 22.449 5 1.1666 1
2003 34.007 5 1.1666 1
2004 30.493 5 1.1666 1
2005 27.956 5 1.1666 1
2006 10.251 5 1.1666 1
2007 15.562 5 1.1666 1
2008 13.553 5 1.1666 1
2009 12.684 5 1.1666 1
2010 27.979 5 1.1666 1
2011 20.673 5 1.1666 1
#Age composition data by year, gear (ages 2-15+)
#na_gears
3
#na_nobs
25 41 34
#a_sage
2 2 2
#a_page
10 10 10
#yr gear V2 V3 V4 V5 V6 V7 V8 V9 V10 #Number aged.
1951 1 203 852 2739 486 263 124 12 2 1
1952 1 282 522 1994 2679 364 61 18 2 0
1953 1 17 541 327 361 158 14 1 0 0
1954 1 56 753 772 638 351 69 16 1 0
1955 1 31 55 795 177 59 12 2 0 0
1956 1 169 978 160 319 43 9 3 2 0
1957 1 401 610 1482 597 558 45 12 1 0
1958 1 339 256 64 82 13 17 0 0 0
1959 1 54 973 539 144 157 34 35 3 0
1960 1 1903 252 972 286 119 71 16 7 0
1961 1 400 2348 276 649 155 54 16 5 0
1962 1 30 153 190 38 58 17 3 5 1
1963 1 1326 434 550 673 100 89 13 2 0
1964 1 109 2174 339 371 300 23 20 4 1
1965 1 184 412 1603 336 350 312 58 14 4
1976 1 0 8 11 16 27 28 57 14 0
1978 1 11 80 265 188 191 145 64 22 12
1979 1 22 125 140 348 234 216 113 75 31
1980 1 13 708 76 90 90 62 48 23 9
1981 1 69 459 3366 586 601 562 266 145 68
1982 1 37 288 485 1005 246 161 124 52 16
1984 1 18 237 103 91 64 107 18 7 5
1985 1 92 48 110 165 69 24 23 5 0
1986 1 198 182 155 451 261 152 95 82 20
1988 1 1 53 159 46 43 46 17 16 4
1971 2 39 309 211 64 34 11 4 1 0
1972 2 0 38 128 460 42 27 17 1 1
1973 2 37 336 47 262 219 31 12 6 0
1974 2 1 113 336 47 104 28 2 1 0
1975 2 41 298 695 1362 355 306 78 20 4
1976 2 0 6 49 226 357 52 17 6 0
1977 2 3 327 125 406 561 240 72 20 6
1978 2 10 100 269 79 163 150 23 7 2
1979 2 27 181 113 290 103 166 53 14 3
1980 2 57 2507 239 164 129 104 45 17 5
1981 2 36 494 3840 170 79 68 20 10 6
1982 2 42 290 114 1024 44 21 6 3 0
1983 2 62 954 813 241 2253 171 52 27 9
1984 2 17 1138 436 314 444 721 31 9 4
1985 2 18 330 2288 528 268 432 329 8 4
1986 2 99 778 534 2616 611 294 401 313 3
1987 2 42 1904 490 327 1423 281 165 136 59
1988 2 19 1303 1638 251 351 485 82 61 10
1989 2 22 784 1307 1001 178 162 129 23 8

```

```

1990 2 33 920 1143 1431 1040 203 168 109 13
1991 2 39 1979 391 519 649 391 68 36 39
1992 2 15 1699 1587 251 228 287 146 26 17
1993 2 5 432 1783 1216 162 177 175 63 4
1994 2 44 325 885 3248 1487 276 248 96 31
1995 2 140 673 297 495 1898 692 107 56 25
1996 2 29 1763 241 76 115 316 140 10 5
1997 2 35 615 1447 216 68 133 128 50 5
1998 2 4 702 465 768 94 30 23 27 2
1999 2 17 95 706 350 425 76 18 15 13
2000 2 77 1111 381 1132 498 646 89 20 10
2001 2 79 1430 875 235 702 315 260 39 4
2002 2 240 867 1553 871 187 442 167 82 10
2003 2 16 2387 538 605 313 92 131 47 20
2004 2 23 50 1700 273 238 98 19 28 2
2005 2 21 856 268 1297 279 166 59 13 10
2006 2 29 327 887 176 460 78 32 9 2
2007 2 27 355 161 78 22 72 9 7 1
2008 2 69 578 2062 448 310 65 138 29 9
2009 2 11 847 703 1723 286 197 45 59 2
2010 2 41 1095 888 377 676 108 54 10 12
2011 2 15 1082 1055 680 494 893 160 62 27
1974 3 0 1 41 22 36 3 1 0 0
1977 3 0 3 6 56 152 41 19 4 0
1978 3 0 0 31 9 49 50 10 2 0
1979 3 0 0 24 120 54 66 22 5 0
1980 3 0 13 20 92 52 51 22 9 1
1981 3 0 1 105 44 62 39 12 5 0
1984 3 0 5 10 65 108 290 17 6 4
1985 3 0 2 90 82 87 120 164 2 3
1986 3 0 5 54 713 249 115 101 80 1
1987 3 0 10 55 125 1041 293 163 108 60
1988 3 0 3 46 51 153 318 83 36 13
1989 3 0 0 22 145 65 112 104 16 11
1990 3 0 0 34 116 231 56 63 33 8
1991 3 0 0 39 171 288 287 61 40 28
1992 3 0 3 123 85 219 236 183 37 31
1993 3 0 0 62 302 71 138 99 61 7
1994 3 0 0 24 160 434 110 101 54 15
1995 3 0 1 10 144 295 334 35 16 10
1996 3 0 4 21 29 132 167 135 16 6
1997 3 0 1 123 73 88 128 130 70 15
1998 3 0 7 33 466 222 107 122 76 40
1999 3 0 0 78 119 357 97 33 14 21
2000 3 0 1 17 187 166 342 76 9 10
2001 3 0 3 58 97 337 215 266 55 9
2002 3 0 1 87 243 117 285 161 145 16
2003 3 0 3 40 323 226 92 107 46 32
2004 3 0 1 244 151 412 172 55 53 17
2005 3 0 0 6 350 136 195 44 10 7
2006 3 0 0 17 58 332 90 72 8 0
2007 3 1 11 40 208 108 630 150 65 14
2008 3 0 1 126 102 224 108 519 77 36
2009 3 0 1 20 406 187 144 53 92 9
2010 3 0 0 19 72 492 145 78 31 26
#2011 3 49 138 282 601 108 45 7 11 1
2011 3 0 2 49 138 282 601 108 45 20
#n_lt_obs
61
#Mean weight-at-age in kilograms (interpolated from HCAM.rep)
#A$yr V1 V2 V3 V4 V5 V6 V7 V8 V9
1951 0.0383 0.0737 0.0940 0.1128 0.1229 0.1316 0.1420 0.1380 0.1320
1952 0.0467 0.0838 0.1158 0.1308 0.1496 0.1657 0.1672 0.2215 0.1700
1953 0.0388 0.0803 0.1108 0.1307 0.1433 0.1572 0.1690 0.1871 0.1700
1954 0.0405 0.0724 0.1057 0.1308 0.1453 0.1617 0.1828 0.1930 0.1700
1955 0.0458 0.0808 0.1013 0.1215 0.1429 0.1593 0.1540 0.1871 0.1700
1956 0.0373 0.0764 0.0944 0.1136 0.1405 0.1518 0.1700 0.1990 0.1700
1957 0.0287 0.0752 0.1041 0.1225 0.1368 0.1654 0.1887 0.2000 0.1700
1958 0.0348 0.0758 0.1216 0.1414 0.1393 0.1679 0.1690 0.1871 0.1700
1959 0.0436 0.0857 0.1022 0.1205 0.1494 0.1547 0.1659 0.1480 0.1700
1960 0.0386 0.0674 0.1038 0.1211 0.1398 0.1545 0.1686 0.1700 0.1700
1961 0.0395 0.0740 0.1074 0.1306 0.1443 0.1594 0.1554 0.1930 0.1700
1962 0.0439 0.0791 0.1088 0.1388 0.1641 0.1797 0.1807 0.2020 0.1700
1963 0.0393 0.0662 0.1042 0.1281 0.1545 0.1664 0.1902 0.2250 0.1700
1964 0.0442 0.0724 0.0932 0.1224 0.1333 0.1590 0.1520 0.1560 0.1480
1965 0.0527 0.0983 0.1143 0.1400 0.1566 0.1637 0.1797 0.2001 0.2300
1966 0.0408 0.0774 0.1054 0.1270 0.1442 0.1599 0.1690 0.1871 0.1700
1967 0.0409 0.0774 0.1054 0.1270 0.1442 0.1599 0.1690 0.1871 0.1700
1968 0.0408 0.0774 0.1054 0.1270 0.1442 0.1599 0.1690 0.1871 0.1700
1969 0.0408 0.0774 0.1054 0.1270 0.1442 0.1599 0.1690 0.1871 0.1700
1970 0.0485 0.0831 0.1134 0.1382 0.1571 0.1734 0.1836 0.1913 0.2042
1971 0.0362 0.0688 0.1103 0.1317 0.1443 0.1766 0.1720 0.1020 0.2042
1972 0.0485 0.1004 0.1370 0.1628 0.1991 0.2247 0.2331 0.2490 0.2590
1973 0.0326 0.0626 0.1172 0.1641 0.1791 0.1976 0.2103 0.2158 0.2042
1974 0.0670 0.0863 0.1212 0.1655 0.1845 0.1947 0.2040 0.1750 0.2042
1975 0.0288 0.0658 0.1138 0.1375 0.1648 0.1673 0.1827 0.1786 0.2063
1976 0.0485 0.0887 0.1325 0.1583 0.1734 0.2029 0.2106 0.2270 0.2042
1977 0.0540 0.0862 0.1178 0.1510 0.1688 0.1838 0.1955 0.1956 0.2242
1978 0.0554 0.0984 0.1249 0.1491 0.1701 0.1856 0.2022 0.2323 0.2304
1979 0.0566 0.0966 0.1289 0.1483 0.1661 0.1844 0.1911 0.2136 0.2129
1980 0.0609 0.0807 0.1165 0.1471 0.1697 0.1792 0.1879 0.2094 0.2241
1981 0.0452 0.0811 0.0999 0.1306 0.1551 0.1692 0.1827 0.2014 0.2028
1982 0.0421 0.0775 0.1084 0.1162 0.1483 0.1703 0.1778 0.1833 0.1978
1983 0.0568 0.0795 0.1042 0.1222 0.1343 0.1536 0.1697 0.1914 0.1961
1984 0.0457 0.0752 0.0897 0.1112 0.1237 0.1351 0.1569 0.1769 0.1870
1985 0.0356 0.0787 0.0975 0.1095 0.1219 0.1336 0.1490 0.1766 0.1590
1986 0.0555 0.0919 0.1185 0.1369 0.1472 0.1578 0.1685 0.1795 0.1963
1987 0.0549 0.0835 0.1070 0.1283 0.1418 0.1526 0.1596 0.1719 0.1746
1988 0.0507 0.0742 0.0967 0.1165 0.1351 0.1514 0.1519 0.1642 0.1887
1989 0.0565 0.0751 0.0964 0.1156 0.1361 0.1471 0.1659 0.1596 0.1906
1990 0.0504 0.0890 0.1078 0.1222 0.1379 0.1524 0.1658 0.1763 0.1850
1991 0.0548 0.0763 0.1058 0.1205 0.1298 0.1413 0.1483 0.1601 0.1711
1992 0.0470 0.0764 0.0934 0.1201 0.1334 0.1405 0.1483 0.1671 0.1732
1993 0.0538 0.0767 0.0964 0.1093 0.1268 0.1367 0.1417 0.1514 0.1543
1994 0.0425 0.0717 0.0935 0.1061 0.1160 0.1340 0.1375 0.1413 0.1594
1995 0.0480 0.0741 0.0920 0.1123 0.1207 0.1309 0.1487 0.1577 0.1589
1996 0.0524 0.0723 0.0947 0.1111 0.1290 0.1343 0.1429 0.1495 0.1608
1997 0.0565 0.0677 0.0839 0.1039 0.1190 0.1300 0.1377 0.1454 0.1496
1998 0.0448 0.0670 0.0799 0.0924 0.1018 0.1204 0.1303 0.1462 0.1505
1999 0.0579 0.0791 0.0956 0.1040 0.1155 0.1180 0.1363 0.1394 0.1443
2000 0.0465 0.0691 0.0852 0.1042 0.1099 0.1184 0.1303 0.1308 0.1416
2001 0.0424 0.0674 0.0917 0.1051 0.1245 0.1264 0.1368 0.1376 0.1513
2002 0.0466 0.0660 0.0847 0.1047 0.1176 0.1287 0.1334 0.1476 0.1532
2003 0.0505 0.0701 0.0853 0.1096 0.1262 0.1400 0.1457 0.1521 0.1588
2004 0.0500 0.0649 0.0865 0.1003 0.1148 0.1312 0.1432 0.1524 0.1340
2005 0.0382 0.0644 0.0709 0.1000 0.1059 0.1194 0.1376 0.1392 0.1470
2006 0.0479 0.0631 0.0796 0.0900 0.1097 0.1213 0.1309 0.1433 0.1205
2007 0.0399 0.0581 0.0702 0.0902 0.1069 0.1103 0.1196 0.1274 0.1440
2008 0.0442 0.0580 0.0816 0.0947 0.1078 0.1175 0.1316 0.1323 0.1516
2009 0.0322 0.0725 0.0823 0.1017 0.1134 0.1198 0.1292 0.1374 0.1540
2010 0.0449 0.0661 0.0869 0.0985 0.1121 0.1180 0.1266 0.1068 0.1464
2011 0.0388 0.0687 0.0819 0.1020 0.1113 0.1256 0.1379 0.1447 0.1383
##eofc
999

## -----
## SOG HERRING CONTROLS
## -----
## Prior descriptions:
## -0 uniform (0,0)
## -1 normal (p1=mu,p2=sig)
## -2 lognormal (p1=log(mu),p2=sig)
## -3 beta (p1=alpha,p2=beta)
## -4 gamma(p1=alpha,p2=beta)
## -----
## npar
## ival lb ub phz prior p1 p2 parameter name ##
## 7
## 1 7.60 -5.0 15 4 0 -5.0 15 #log_ro
## 2 0.67 0.2 1.0 4 3 10.0 4.925373 #steepness
## 3 -0.7985077 -5.0 5.0 3 1 -0.7985077 0.4 #log.m
## 4 7.40 -5.0 15 1 0 -5.0 15 #log_avgrec
## 5 7.20 -5.0 15 1 0 -5.0 15 #log_recinit
## 6 0.3043478 0.001 0.999 3 3 17.08696 39.0559 #rbo
## 7 0.8695652 0.01 5.0 3 4 25.0 28.75 #kappa (precision)
## -----
## SELECTIVITY PARAMETERS
## -----
## OPTIONS FOR SELECTIVITY:
## 1) logistic selectivity parameters
## 2) selectivity coefficients
## 3) a constant cubic spline with age-nodes
## 4) a time varying cubic spline with age-nodes
## 5) a time varying bicubic spline with age & year nodes.
## 6) fixed logistic (set isel_type=6, and estimation phase to -1)
## 7) logistic function of body weight.
## 8) sign 0.05 0.10 0.20 0.30 0.40 0.50
## wt = 200. 50.0 22.2 12.5 5.56 3.12 2.00
## Gear 1:3 fishery: Gear 4-5 survey
## isel_type
## 1 1 8 6 6
## Age at 50% selectivity (logistic)
## 2 5.0 0.6 2.055 2.055
## STD at 50% selectivity (logistic)
## 0.25 0.25 0.15 0.05 0.05
## No. of age nodes for each gear (0 to ignore).
## 5 5 5 0 0
## No. of year nodes for each gear (0 to ignore).
## 12 3 10 0 0
## Estimation phase
## 2 2 2 -1 -1
## Penalty weight for 2nd differences w1=1/(2*sig^2)
## 125. 125. 12.5 12.5 12.5
## Penalty weight for dome-shaped selectivity 1=1/(2*sig^2)
## 50.0 50.0 200.0 200.0 200.0
## -----
## Priors for Survey q
## -----
## nits number of surveys
## 2
## priors 0=uniform density 1=normal density
## 1 1
## prior log(mean)
## -0.569 -0.569
## prior sd
## 0.274 0.274
## -----
## OTHER MISCELLANEOUS CONTROLS
## 0 ## 1 verbose ADMB output (0=off, 1=on)
## 1 ## 2 recruitment model (1=beverton-holt, 2=ricker)
## 0.100 ## 3 std in observed catches in first phase.
## 0.0707 ## 4 std in observed catches in last phase.
## 0 ## 5 Assume unfished in first year (0=FALSE, 1=TRUE)
## 0.02 ## 6 Minimum proportion to consider in age-proportions for dmvlogistic
## 0.20 ## 7 Mean fishing mortality for regularizing the estimates of Pt
## 0.01 ## 8 std in mean fishing mortality in first phase
## 2.00 ## 9 std in mean fishing mortality in last phase
## 3 ## 10 phase for estimating _m_deviations (use -1 to turn off mdevs)
## 0.1 ## 11 std in deviations for natural mortality
## 12 ## 12 number of estimated nodes for deviations in natural mortality
## 1.00 ## 13 fraction of total mortality that takes place prior to spawning
## 1 ## 14 switch for age-composition likelihood (1=dmvlogistic, 2=dmultinom)
## -----
## eofc
999

```

### B.2.3 Central Coast

```

#NB The data herein were taken from qci2010_final.dat for the HCAM model.
## -----
## Model Dimensions
1951 #first year of data
2011 #last year of data
2 #age of youngest age class
10 #age of plus group
5 #number of gears (ngear)
## flags for fishery (1) or survey (0) in ngears
#1 1 0 0
0.0686819 0.4137555 0.5175626 0.0000000 0.0000000
## -----
## -----
##age-schedule and population parameters
#natural mortality rate (m)
0.334
#growth parameters (linf,k,to) (from fishbase)
27.0, 0.48, 0
#length-weight allometry (a,b)
4.5e-6, 3.1270
#maturity at age (am=log(3)/k) & gm=std for logistic
2.055, 0.05
## -----
#Time series data
#Observed catch (1951-2010, 1000s metric t)
#Year P1 P2 P3 S1 S2
1951 42.458 0.000 0.000 0 0
1952 33.195 0.000 0.000 0 0
1953 0.768 0.000 0.000 0 0
1954 24.616 0.000 0.000 0 0
1955 11.594 0.000 0.000 0 0
1956 43.627 0.000 0.000 0 0
1957 23.261 0.000 0.000 0 0
1958 9.849 0.000 0.000 0 0
1959 27.870 0.000 0.000 0 0
1960 4.037 0.000 0.000 0 0
1961 31.704 0.000 0.000 0 0
1962 15.709 0.000 0.000 0 0
1963 44.054 0.000 0.000 0 0
1964 31.895 0.000 0.000 0 0
1965 15.670 0.000 0.000 0 0
1966 37.482 0.000 0.000 0 0
1967 21.890 0.000 0.000 0 0
1968 1.528 0.000 0.000 0 0
1969 0.009 0.000 0.000 0 0
1970 0.209 0.000 0.000 0 0
1971 3.614 0.000 0.000 0 0
1972 0.388 8.367 0.137 0 0
1973 0.035 6.653 1.112 0 0
1974 0.000 3.621 5.267 0 0
1975 0.000 3.343 5.395 0 0
1976 0.000 6.198 6.213 0 0
1977 0.322 3.881 6.904 0 0
1978 0.048 4.723 9.277 0 0
1979 0.000 0.005 0.000 0 0
1980 0.010 0.000 0.528 0 0
1981 0.000 0.269 2.304 0 0
1982 0.041 2.258 4.071 0 0
1983 0.000 2.061 3.579 0 0
1984 0.000 3.588 3.582 0 0
1985 0.000 2.915 2.294 0 0
1986 0.038 2.173 1.176 0 0
1987 0.000 2.695 0.920 0 0
1988 0.028 3.529 0.970 0 0
1989 0.000 6.531 2.911 0 0
1990 0.000 5.305 3.046 0 0
1991 0.000 7.097 1.806 0 0
1992 0.084 7.163 1.111 0 0
1993 0.000 8.478 2.038 0 0
1994 0.000 9.757 2.122 0 0
1995 0.000 8.131 1.451 0 0
1996 0.000 3.897 0.402 0 0
1997 0.000 3.276 0.344 0 0
1998 0.000 7.976 0.646 0 0
1999 0.000 6.013 1.511 0 0
2000 0.000 6.394 0.972 0 0
2001 0.000 5.613 0.517 0 0
2002 0.000 2.894 0.399 0 0
2003 0.000 2.299 0.289 0 0
2004 0.000 2.988 0.000 0 0
2005 0.000 3.778 0.000 0 0
2006 0.000 3.072 0.000 0 0
2007 0.000 0.398 0.000 0 0
2008 0.000 0.000 0.000 0 0
2009 0.000 0.000 0.000 0 0
2010 0.000 0.000 0.000 0 0
2011 0.000 0.000 0.000 0 0
#Relative Abundance index from fisheries independent survey (it) 1970-2008
#it
#61
2
#init_nobs
37 24
#survey type
## 1 = survey is proportional to vulnerable numbers
## 2 = survey is proportional to vulnerable biomass
## 3 = survey is proportional to spawning biomass (e.g., herring spawn survey)
3 3
#1yr it gear wt survey timing
1951 15.390 4 1 1
1952 10.295 4 1 1
1953 18.237 4 1 1
1954 13.967 4 1 1
1955 13.564 4 1 1
1956 6.626 4 1 1
1957 4.607 4 1 1
1958 3.549 4 1 1
1959 3.904 4 1 1
1960 12.615 4 1 1
1961 4.265 4 1 1
1962 11.948 4 1 1
1963 6.485 4 1 1
1964 6.464 4 1 1
1965 2.097 4 1 1
1966 1.863 4 1 1
1967 5.434 4 1 1
1968 5.790 4 1 1
1969 1.837 4 1 1
1970 8.230 4 1 1
1971 4.156 4 1 1
1972 3.572 4 1 1
1973 12.434 4 1 1
1974 8.852 4 1 1
1975 8.037 4 1 1
1976 13.849 4 1 1
1977 14.613 4 1 1
1978 7.747 4 1 1
1979 5.779 4 1 1
1980 13.012 4 1 1
1981 15.919 4 1 1
1982 16.333 4 1 1
1983 18.482 4 1 1
1984 14.185 4 1 1
1985 8.850 4 1 1
1986 20.342 4 1 1
1987 12.827 4 1 1
1988 26.916 5 1.1666 1
1989 21.561 5 1.1666 1
1990 28.980 5 1.1666 1
1991 19.183 5 1.1666 1
1992 43.274 5 1.1666 1
1993 32.392 5 1.1666 1
1994 29.432 5 1.1666 1
1995 22.348 5 1.1666 1
1996 21.646 5 1.1666 1
1997 28.255 5 1.1666 1
1998 31.503 5 1.1666 1
1999 31.813 5 1.1666 1
2000 32.652 5 1.1666 1
2001 25.109 5 1.1666 1
2002 23.147 5 1.1666 1
2003 25.679 5 1.1666 1
2004 29.407 5 1.1666 1
2005 24.158 5 1.1666 1
2006 12.051 5 1.1666 1
2007 9.857 5 1.1666 1
2008 3.971 5 1.1666 1
2009 10.183 5 1.1666 1
2010 8.121 5 1.1666 1
2011 9.906 5 1.1666 1
#Age composition data by year, gear (ages 2-15+)
#na_gears
3
#na_nobs
16 39 30
#a_sage
2 2 2
#a_page
10 10 10
#yr gear V2 V3 V4 V5 V6 V7 V8 V9 V10 #Number aged.
1951 1 129 1518 2693 638 269 66 3 0 0
1952 1 267 1034 1551 1966 232 79 23 2 0
1953 1 274 822 702 779 297 39 13 0 0
1954 1 126 2192 640 147 41 5 0 2 0
1955 1 156 181 1749 213 36 9 0 0 0
1956 1 852 683 458 2862 145 17 2 0 1
1957 1 758 2357 501 284 688 33 1 0 0
1958 1 880 2272 439 45 20 17 0 0 0
1959 1 189 2430 1831 401 40 22 21 1 0
1960 1 545 303 364 77 16 1 1 0 0
1961 1 450 902 301 826 276 22 3 1 0
1962 1 43 303 117 16 64 13 1 0 0
1963 1 4 324 604 58 30 30 2 0 0
1964 1 164 549 320 118 17 1 0 0 0
1965 1 143 637 591 277 95 6 1 0 0
1971 1 137 377 351 36 40 9 1 1 0
1972 2 75 530 494 448 120 76 19 1 0
1973 2 15 615 228 198 150 25 6 2 0
1974 2 44 278 602 287 187 100 14 3 0
1975 2 103 2932 2269 2477 764 283 60 6 2
1976 2 163 637 2234 1132 912 246 80 13 1
1977 2 18 492 587 818 408 221 51 10 1
1978 2 3 355 212 278 322 151 49 15 5
1980 2 186 2305 214 266 121 76 21 13 1
1981 2 197 755 3325 405 314 145 35 19 6
1982 2 59 548 376 2112 182 160 51 17 0
1983 2 29 381 840 589 3109 274 169 40 11
1984 2 274 460 637 1143 1016 2563 142 52 6
1985 2 149 2052 410 457 698 638 987 24 7
1986 2 240 972 2378 516 384 404 367 697 25
1987 2 256 1161 744 1626 289 230 294 235 275
1988 2 59 3526 606 326 370 87 76 78 44
1989 2 72 260 4300 517 202 158 42 45 36
1990 2 66 383 346 4973 511 258 202 51 28
1991 2 144 1337 480 440 3947 453 166 105 23
1992 2 146 4241 828 199 250 1362 155 44 34

```

```

1993 2 252 586 5608 848 177 225 916 98 28
1994 2 85 1538 620 3888 549 148 199 257 22
1995 2 74 581 2250 894 4604 609 192 220 155
1996 2 667 1114 323 926 388 1698 328 83 43
1997 2 146 3892 1161 249 422 274 583 106 27
1998 2 34 2393 2793 553 155 202 198 192 41
1999 2 39 440 2141 1709 326 81 106 97 55
2000 2 16 865 490 1572 1186 263 53 41 26
2001 2 112 340 1194 517 1173 831 181 38 15
2002 2 269 1851 579 971 338 1124 475 78 13
2003 2 22 2144 1138 365 400 183 317 120 24
2004 2 37 225 2085 542 112 147 75 70 17
2005 2 42 2311 1037 2101 566 125 112 60 30
2006 2 53 702 3246 585 967 199 44 31 3
2007 2 32 700 444 739 190 185 37 10 1
2008 2 144 146 659 184 246 44 43 8 1
2009 2 60 2059 308 238 67 63 8 10 2
2010 2 41 387 1597 133 189 51 52 2 6
2011 2 125 1359 426 671 86 64 17 17 3
1973 3 0 4 28 43 21 2 1 0 0
1974 3 0 2 106 184 116 58 8 0 0
1975 3 0 16 99 171 59 21 9 0 0
1976 3 0 10 230 364 431 144 37 5 1
1977 3 0 5 59 161 143 61 18 6 0
1978 3 0 7 74 277 345 149 30 3 1
1980 3 0 6 1 39 24 21 13 7 0
1981 3 4 22 722 194 213 153 74 27 7
1982 3 0 31 75 944 84 71 28 8 1
1983 3 0 9 124 224 1177 87 67 11 4
1984 3 0 3 34 141 190 655 51 12 6
1985 3 0 43 84 137 303 349 558 18 13
1986 3 0 18 248 126 101 166 134 219 6
1987 3 0 8 76 440 115 77 97 80 84
1988 3 0 23 56 80 144 72 37 51 46
1989 3 0 2 180 159 107 91 33 23 16
1990 3 0 0 8 529 133 50 62 9 12
1991 3 0 3 13 34 377 51 38 19 5
1992 3 0 66 87 61 101 659 98 35 10
1993 3 0 2 342 112 44 45 211 17 8
1994 3 0 30 94 1287 237 69 83 135 12
1995 3 0 3 112 101 823 135 23 29 37
1996 3 0 2 8 102 65 306 59 12 7
1997 3 0 7 15 32 117 99 197 37 7
1998 3 0 5 149 142 90 183 164 217 62
1999 3 0 1 123 382 151 51 44 46 35
2000 3 0 3 14 277 285 71 11 6 14
2001 3 0 0 39 46 422 225 57 9 0
2002 3 0 3 30 105 38 237 83 7 1
2003 3 0 4 33 103 238 104 306 114 20
#n_wt_obs
61
#Mean weight-at-age in kilograms (interpolated: from HCAM.rep)
#$yr V1 V2 V3 V4 V5 V6 V7 V8 V9
1951 0.0480 0.0842 0.1138 0.1371 0.1465 0.1559 0.1610 0.1352 0.1800
1952 0.0466 0.0869 0.1121 0.1311 0.1476 0.1578 0.1636 0.1730 0.1800
1953 0.0358 0.0833 0.1078 0.1273 0.1474 0.1616 0.1699 0.1352 0.1800
1954 0.0268 0.0626 0.0937 0.1167 0.1378 0.1330 0.1545 0.1310 0.1800
1955 0.0386 0.0725 0.0965 0.1204 0.1428 0.1347 0.1545 0.1352 0.1800
1956 0.0406 0.0834 0.1107 0.1269 0.1434 0.1581 0.1220 0.1352 0.1800
1957 0.0400 0.0821 0.1082 0.1222 0.1322 0.1487 0.1730 0.1352 0.1800
1958 0.0374 0.0725 0.0956 0.1157 0.1282 0.1465 0.1545 0.1352 0.1800
1959 0.0391 0.0774 0.0938 0.1098 0.1069 0.1316 0.1345 0.1270 0.1800
1960 0.0447 0.0639 0.0810 0.1034 0.1078 0.1470 0.1240 0.1352 0.1800
1961 0.0379 0.0771 0.0985 0.1206 0.1246 0.1317 0.1223 0.1100 0.1800
1962 0.0470 0.0808 0.1074 0.1314 0.1432 0.1554 0.1220 0.1352 0.1800
1963 0.0589 0.0825 0.1038 0.1236 0.1425 0.1543 0.1655 0.1352 0.1800
1964 0.0457 0.0858 0.1083 0.1271 0.1282 0.1540 0.1545 0.1352 0.1800
1965 0.0525 0.1037 0.1271 0.1474 0.1675 0.1757 0.2420 0.1352 0.1800
1966 0.0426 0.0799 0.1039 0.1240 0.1365 0.1497 0.1545 0.1352 0.1800
1967 0.0426 0.0799 0.1059 0.1240 0.1365 0.1497 0.1545 0.1352 0.1800
1968 0.0426 0.0799 0.1039 0.1240 0.1365 0.1497 0.1545 0.1352 0.1800
1969 0.0426 0.0799 0.1038 0.1240 0.1365 0.1497 0.1545 0.1352 0.1800
1970 0.0558 0.0897 0.1165 0.1386 0.1573 0.1713 0.1858 0.1978 0.2077
1971 0.0502 0.0934 0.1275 0.1576 0.1712 0.1819 0.1760 0.1780 0.2077
1972 0.0602 0.0949 0.1172 0.1418 0.1576 0.1654 0.1973 0.1930 0.2077
1973 0.0594 0.0986 0.1312 0.1560 0.1730 0.1834 0.1972 0.2340 0.2077
1974 0.0492 0.0869 0.1209 0.1416 0.1658 0.1788 0.1974 0.2167 0.2077
1975 0.0448 0.0836 0.1187 0.1444 0.1662 0.1862 0.1989 0.2037 0.2205
1976 0.0444 0.0814 0.1081 0.1356 0.1553 0.1754 0.1913 0.2004 0.2100
1977 0.0598 0.0891 0.1164 0.1382 0.1660 0.1839 0.1987 0.2219 0.2250
1978 0.0498 0.0856 0.1142 0.1338 0.1612 0.1864 0.2160 0.2269 0.2442
1979 0.0832 0.1065 0.1286 0.1534 0.1913 0.2116 0.2350 0.2077
1980 0.0501 0.0815 0.0998 0.1244 0.1440 0.1614 0.1688 0.1981 0.2120
1981 0.0452 0.0749 0.0990 0.1156 0.1338 0.1500 0.1755 0.1747 0.2040
1982 0.0521 0.0880 0.1094 0.1300 0.1392 0.1519 0.1683 0.1824 0.2077
1983 0.0614 0.0907 0.1107 0.1292 0.1424 0.1493 0.1567 0.1727 0.1855
1984 0.0509 0.0900 0.1082 0.1223 0.1352 0.1419 0.1556 0.1761 0.1782
1985 0.0620 0.0951 0.1234 0.1399 0.1502 0.1563 0.1734 0.1753 0.2029
1986 0.0623 0.0989 0.1268 0.1422 0.1551 0.1667 0.1734 0.1802 0.1992
1987 0.0594 0.0918 0.1221 0.1491 0.1683 0.1790 0.1845 0.1957 0.2077
1988 0.0541 0.0842 0.1143 0.1390 0.1711 0.1844 0.1889 0.1957 0.2031
1989 0.0563 0.0830 0.1032 0.1301 0.1465 0.1728 0.1797 0.1804 0.1936
1990 0.0572 0.0843 0.1065 0.1262 0.1476 0.1685 0.1787 0.1877 0.1916
1991 0.0577 0.0846 0.1062 0.1288 0.1447 0.1650 0.1784 0.1868 0.1992
1992 0.0501 0.0860 0.1055 0.1236 0.1362 0.1511 0.1683 0.1874 0.1932
1993 0.0517 0.0858 0.1052 0.1201 0.1334 0.1404 0.1533 0.1674 0.1716
1994 0.0476 0.0831 0.1063 0.1225 0.1334 0.1481 0.1581 0.1630 0.1750
1995 0.0478 0.0796 0.1057 0.1233 0.1348 0.1438 0.1522 0.1561 0.1624
1996 0.0607 0.0781 0.1025 0.1258 0.1397 0.1476 0.1581 0.1655 0.1617
1997 0.0456 0.0761 0.0888 0.1054 0.1320 0.1432 0.1493 0.1596 0.1607
1998 0.0415 0.0716 0.0874 0.1006 0.1173 0.1404 0.1462 0.1540 0.1629
1999 0.0538 0.0681 0.0902 0.1054 0.1138 0.1292 0.1478 0.1545 0.1582
2000 0.0514 0.0771 0.0881 0.1127 0.1268 0.1377 0.1450 0.1660 0.1756
2001 0.0445 0.0734 0.0974 0.1061 0.1255 0.1358 0.1468 0.1569 0.1631
2002 0.0512 0.0670 0.0878 0.1084 0.1187 0.1295 0.1366 0.1414 0.1507
2003 0.0478 0.0773 0.0885 0.1111 0.1264 0.1371 0.1433 0.1508 0.1629
2004 0.0481 0.0694 0.0912 0.1096 0.1125 0.1241 0.1362 0.1365 0.1461
2005 0.0378 0.0675 0.0752 0.1063 0.1091 0.1258 0.1360 0.1404 0.1521
2006 0.0388 0.0605 0.0786 0.0919 0.1106 0.1147 0.1281 0.1347 0.1447
2007 0.0410 0.0651 0.0750 0.0955 0.1025 0.1181 0.1202 0.1397 0.1430
2008 0.0430 0.0606 0.0757 0.0873 0.1034 0.1153 0.1260 0.1344 0.1670
2009 0.0405 0.0664 0.0727 0.0921 0.1035 0.1219 0.1346 0.1315 0.1350
2010 0.0481 0.0667 0.0841 0.0928 0.1046 0.1034 0.1228 0.1540 0.1320
2011 0.0325 0.0601 0.0724 0.0912 0.1000 0.1177 0.1185 0.1342 0.1563
##eofc
999
## -----
## SOG HERRING CONTROLS
## -----
## CONTROLS FOR ESTIMATED PARAMETERS
## -----
## Prior descriptions:
## -0 uniform (0,0)
## -1 normal (p1=mu,p2=sig)
## -2 lognormal (p1=log(mu),p2=sig)
## -3 beta (p1=alpha,p2=beta)
## -4 gamma(p1=alpha,p2=beta)
## -----
## npar
## ival lb ub phz prior p1 p2 parameter name
## -----
## 7
## 6.60 -5.0 15 4 0 -5.0 15 #log_ro
## 0.67 0.2 1.0 4 3 10.0 4.925373 #steepness
## -0.7985077 -5.0 5.0 3 1 -0.7985077 0.4 #log_m
## 7.40 -5.0 15 1 0 -5.0 15 #log_avgrec
## 7.20 -5.0 15 1 0 -5.0 15 #log_recinit
## 0.3043478 0.001 0.999 3 3 17.08696 39.0559 #rho
## 0.8695652 0.01 5.0 3 4 25.0 28.75 #kappa (precision)
## -----
## 1
## 5.0 5 5 0 0
## No. of age nodes for each gear (0 to ignore).
## 12 3 10 0 0
## Gear 1:3 fishery: Gear 4-5 survey
## isel_type
## 1 1 8 6 6
## Age at 50% selectivity (logistic)
## 2.0 3.0 0.6 2.055 2.055
## STD at 50% selectivity (logistic)
## 0.25 0.25 0.15 0.05 0.05
## No. of age nodes for each gear (0 to ignore).
## 5 5 5 0 0
## No. of year nodes for each gear (0 to ignore).
## 12 3 10 0 0
## Estimation phase
## 2 2 2 -1 -1
## Penalty weight for 2nd differences w1/(2*sig^2)
## 125. 125. 12.5 12.5
## Penalty weight for dome-shaped selectivity 1/(2*sig^2)
## 50.0 50.0 200.0 200.0 200.0
## -----
## nits #number of surveys
## 2
## priors 0=uniform density 1=normal density
## 1 1
## prior log(mean)
## -0.569 -0.569
## prior sd
## 0.274 0.274
## -----
## OTHER MISCELLANEOUS CONTROLS
## 0 ## 1 verbose ADMB output (0=off, 1=on)
## 1 ## 2 recruitment model (1=beverton-holt, 2=ricker)
## 0.100 ## 3 std in observed catches in first phase.
## 0.0707 ## 4 std in observed catches in last phase.
## 0 ## 5 Assume unfished in first year (0=FALSE, 1=TRUE)
## 0.02 ## 6 Minimum proportion to consider in age-proportions for dmvlogistic
## 0.20 ## 7 Mean fishing mortality for regularizing the estimates of Ft
## 0.01 ## 8 std in mean fishing mortality in first phase
## 2.00 ## 9 std in mean fishing mortality in last phase
## 3 ## 10 phase for estimating _m_deviations (use -1 to turn off mdevs)
## 0.1 ## 11 std in deviations for natural mortality
## 12 ## 12 number of estimated nodes for deviations in natural mortality
## 1.00 ## 13 fraction of total mortality that takes place prior to spawning
## 1 ## 14 switch for age-composition likelihood (1=dmvlogistic, 2=dmultinom)
## -----
## eofc
999

```

## B.2.4 Strait of Georgia

```

#NB The data herein were taken from sog2010_final.dat for the HCAM model.
## -----
## Model Dimensions-----
1951 #first year of data
2011 #last year of data
2 #age of youngest age class
10 #age of plus group
5 #number of gears (ngear)
## flags for fishery (1) or survey (0) in ngears
#1 1 0 0
0.0686819 0.4137555 0.5175626 0.0000000 0.0000000
## -----
## -----
##age-schedule and population parameters
#natural mortality rate (m)
0.334
#growth parameters (linf,k,to) (from fishbase)
27.0, 0.48, 0
#length-weight allometry (a,b)
4.5e-6, 3.1270
#maturity at age (am=log(3)/k) & gm=std for logistic
2.055, 0.05
## -----
###
#Time series data
#Observed catch (1951-2010, 1000s metric t)
#yr p1 p2 p3 survey
1951 43.798 0.000 0.000 0 0
1952 45.849 0.000 0.000 0 0
1953 8.412 0.000 0.000 0 0
1954 65.767 0.000 0.000 0 0
1955 68.641 0.000 0.000 0 0
1956 72.062 0.000 0.000 0 0
1957 59.608 0.000 0.000 0 0
1958 20.628 0.000 0.000 0 0
1959 49.644 0.000 0.381 0 0
1960 68.037 0.000 0.000 0 0
1961 46.215 0.000 0.000 0 0
1962 65.303 0.000 0.000 0 0
1963 68.847 0.000 0.000 0 0
1964 76.881 0.000 0.000 0 0
1965 47.819 0.000 0.000 0 0
1966 33.333 0.000 0.000 0 0
1967 31.043 0.000 0.000 0 0
1968 1.891 0.000 0.000 0 0
1969 0.194 0.000 0.000 0 0
1970 0.221 0.000 0.022 0 0
1971 1.610 0.000 0.084 0 0
1972 2.434 5.921 0.456 0 0
1973 3.980 1.604 2.064 0 0
1974 0.479 0.425 3.095 0 0
1975 0.378 0.469 5.331 0 0
1976 5.061 0.202 6.975 0 0
1977 5.676 4.098 7.736 0 0
1978 12.963 3.723 7.316 0 0
1979 13.513 0.000 6.825 0 0
1980 2.470 0.169 3.180 0 0
1981 4.904 2.081 5.067 0 0
1982 3.937 3.312 5.583 0 0
1983 0.824 7.780 8.613 0 0
1984 0.087 4.126 6.039 0 0
1985 0.772 2.726 3.495 0 0
1986 0.432 0.162 0.000 0 0
1987 0.244 3.111 5.998 0 0
1988 0.756 1.471 5.968 0 0
1989 1.033 1.417 5.919 0 0
1990 0.233 0.000 7.886 0 0
1991 0.562 1.131 9.410 0 0
1992 0.939 3.610 8.870 0 0
1993 0.617 4.391 8.733 0 0
1994 0.942 5.134 11.572 0 0
1995 0.641 4.359 8.190 0 0
1996 0.541 7.338 6.233 0 0
1997 0.402 9.274 6.148 0 0
1998 0.954 5.755 6.895 0 0
1999 1.471 4.976 6.837 0 0
2000 1.156 6.455 7.593 0 0
2001 1.424 7.274 7.682 0 0
2002 1.328 9.299 7.986 0 0
2003 1.696 10.670 8.010 0 0
2004 1.356 7.019 5.226 0 0
2005 1.332 7.928 8.954 0 0
2006 1.371 9.308 7.277 0 0
2007 0.672 3.865 5.285 0 0
2008 1.139 2.752 6.046 0 0
2009 0.709 5.685 3.937 0 0
2010 0.595 4.540 3.244 0 0
2011 0.713 0.000 4.415 0 0
#Relative Abundance index from fisheries independent survey (it) 1970-2008
#nit
#61
2
#init_nobs
37 24
#survey type
## 1 = survey is proportional to vulnerable numbers
## 2 = survey is proportional to vulnerable biomass
## 3 = survey is proportional to spawning biomass (e.g., herring spawn survey)
3 3
#yr it gear wt survey timing
1951 66.143 4 1 1
1952 72.376 4 1 1
1953 111.307 4 1 1
1954 82.141 4 1 1
1955 69.854 4 1 1
1956 25.667 4 1 1
1957 24.126 4 1 1
1958 16.911 4 1 1
1959 47.864 4 1 1
1960 55.709 4 1 1
1961 44.326 4 1 1
1962 35.574 4 1 1
1963 37.381 4 1 1
1964 35.954 4 1 1
1965 38.390 4 1 1
1966 7.211 4 1 1
1967 9.647 4 1 1
1968 9.442 4 1 1
1969 14.039 4 1 1
1970 34.163 4 1 1
1971 38.921 4 1 1
1972 25.139 4 1 1
1973 16.191 4 1 1
1974 40.571 4 1 1
1975 70.211 4 1 1
1976 60.642 4 1 1
1977 78.562 4 1 1
1978 102.115 4 1 1
1979 64.266 4 1 1
1980 85.991 4 1 1
1981 55.121 4 1 1
1982 100.987 4 1 1
1983 64.575 4 1 1
1984 26.227 4 1 1
1985 25.247 4 1 1
1986 41.575 4 1 1
1987 41.737 4 1 1
1988 24.976 5 1.1666 1
1989 66.052 5 1.1666 1
1990 67.152 5 1.1666 1
1991 45.830 5 1.1666 1
1992 82.714 5 1.1666 1
1993 90.198 5 1.1666 1
1994 67.144 5 1.1666 1
1995 64.899 5 1.1666 1
1996 71.326 5 1.1666 1
1997 58.232 5 1.1666 1
1998 74.616 5 1.1666 1
1999 85.095 5 1.1666 1
2000 72.688 5 1.1666 1
2001 100.248 5 1.1666 1
2002 117.864 5 1.1666 1
2003 141.651 5 1.1666 1
2004 114.352 5 1.1666 1
2005 95.643 5 1.1666 1
2006 46.752 5 1.1666 1
2007 35.865 5 1.1666 1
2008 32.103 5 1.1666 1
2009 49.909 5 1.1666 1
2010 47.480 5 1.1666 1
2011 79.070 5 1.1666 1
#Age composition data by year, gear (ages 2-15+)
#na_gears
3
#na_nobs
55 40 39
#a_sage
2 2 2
#a_page
10 10 10
## 
#yr gear V2 V3 V4 V5 V6 V7 V8 V9 V10 #Number aged.
1951 1 326 4413 2371 556 110 27 8 2 0
1952 1 1008 4997 2262 608 119 24 6 1 0
1953 1 900 6177 3380 447 119 18 6 0 0
1954 1 200 6011 4845 1520 432 124 27 3 0
1955 1 233 2562 2065 355 57 6 0 0 0
1956 1 703 3798 3502 2801 627 106 28 6 3
1957 1 98 4344 1639 782 374 44 5 1 2
1958 1 564 3278 1171 278 244 177 25 4 0
1959 1 1533 5878 1460 223 48 21 11 2 0
1960 1 398 2049 1391 127 15 4 3 0 0
1961 1 1425 1395 1153 607 166 16 2 0 0
1962 1 270 2019 363 116 45 10 1 0 0
1963 1 361 1290 660 94 24 9 2 0 0
1964 1 190 2173 1157 110 26 4 3 0 0
1965 1 500 1703 775 73 22 3 1 0 0
1966 1 168 256 166 100 16 7 0 0 0
1971 1 175 568 516 101 40 13 1 0 0
1972 1 245 740 770 448 96 33 6 1 0
1973 1 58 520 253 176 48 10 4 1 0
1976 1 248 770 1463 788 221 112 37 17 2
1977 1 88 947 383 274 76 26 12 4 1
1978 1 110 919 958 236 165 62 9 8 3
1979 1 58 665 1545 1093 323 159 47 7 3
1980 1 84 1627 874 716 363 93 57 15 3
1981 1 340 2634 2479 1091 636 242 54 21 1
1982 1 223 1824 1378 625 201 131 36 2 4
1983 1 392 2093 1603 833 429 146 149 77 10
1984 1 630 1398 798 443 203 103 22 10 6
1985 1 741 1063 491 193 79 26 9 2 0
1986 1 230 832 245 60 18 4 1 0 0
1987 1 295 406 328 93 18 4 2 0 0
1988 1 77 1051 248 202 37 9 2 1 0
1989 1 331 219 532 85 73 10 2 0 0

```

```

1990 1 62 544 118 164 21 19 0 0 0
1991 1 101 223 265 51 50 7 1 0 0
1992 1 37 645 125 69 8 3 3 0 0
1993 1 252 302 316 67 26 3 6 1 0
1994 1 42 287 134 81 9 5 0 0 0
1995 1 222 311 409 123 54 8 2 0 1
1996 1 419 821 199 157 41 18 3 0 0
1997 1 52 201 50 10 3 1 2 1 0
1998 1 60 758 626 99 13 7 1 1 0
1999 1 151 361 564 176 36 5 2 0 0
2000 1 250 536 169 90 24 8 0 0 0
2001 1 77 414 223 44 28 10 0 1 0
2002 1 46 570 243 55 9 3 1 0 1
2003 1 12 189 181 36 9 1 0 0 0
2004 1 57 156 203 71 18 2 0 0 0
2005 1 22 72 36 28 9 2 0 0 0
2006 1 29 39 47 26 29 6 2 1 0
2007 1 39 17 3 1 2 0 0 0 0
2008 1 231 94 458 118 24 5 1 1 0
2009 1 4 172 49 5 3 0 0 0 0
2010 1 34 19 249 20 18 7 1 2 0
2011 1 335 371 50 19 1 1 0 0 0
2012 2 564 2514 2354 1282 260 77 10 1 0
1973 2 51 1306 1510 1157 588 77 11 1 0
1974 2 144 533 155 37 13 1 0 0 0
1975 2 288 3117 1506 417 180 85 25 10 0
1976 2 183 505 1002 395 97 41 23 6 2
1977 2 100 1717 675 506 133 37 19 6 5
1978 2 30 1253 1545 423 277 53 11 1 2
1979 2 92 765 1121 898 270 126 38 15 5
1980 2 350 3800 1344 1341 694 174 93 22 7
1981 2 1230 4902 3605 1200 1002 398 87 31 6
1982 2 337 1852 1334 1124 254 273 125 29 2
1983 2 434 4122 3745 2285 1428 411 385 161 28
1984 2 604 2784 2099 936 522 244 82 35 10
1985 2 2024 3592 1519 628 268 104 46 5 1
1986 2 889 3799 1477 409 128 53 9 5 0
1987 2 781 2623 2945 1201 276 86 29 10 4
1988 2 301 3848 924 935 250 63 13 4 0
1989 2 651 1177 3491 610 460 104 20 2 1
1990 2 452 3337 652 1159 182 105 23 4 1
1991 2 542 1173 2123 476 775 116 75 10 1
1992 2 257 2762 691 843 176 260 30 16 0
1993 2 832 2096 1737 379 326 76 84 8 2
1994 2 279 2518 1594 1120 238 168 42 9 0
1995 2 580 1251 2048 1005 627 155 62 20 6
1996 2 1059 3964 1160 1192 481 280 57 19 6
1997 2 618 3806 1671 433 464 184 107 9 3
1998 2 383 4176 2784 1049 228 171 64 17 2
1999 2 268 1054 1716 792 274 75 28 7 2
2000 2 859 2759 1385 1591 627 149 21 15 1
2001 2 458 2981 1939 603 559 181 45 7 2
2002 2 490 3042 1535 673 140 116 22 4 0
2003 2 330 3994 3368 1099 322 81 35 9 0
2004 2 251 1353 1982 972 237 74 14 9 1
2005 2 353 1420 1468 1183 387 98 32 11 4
2006 2 968 1257 1134 754 448 103 33 8 1
2007 2 107 2951 1666 749 346 188 42 10 1
2008 2 160 582 3191 717 259 105 41 8 1
2009 2 20 3164 665 606 180 60 25 10 1
2010 2 583 228 3843 297 311 73 26 8 4
#2011 2 1006 3193 665 1028 109 56 21 3 1
2011 2 337 3808 642 1126 115 61 21 3 1
1972 3 46 118 468 286 68 15 2 1 0
1973 3 0 39 68 84 25 7 1 0 0
1974 3 0 45 390 283 158 39 9 0 0
1975 3 0 8 76 53 21 5 1 0 0
1976 3 0 5 322 342 89 22 5 1 0
1977 3 0 56 480 779 270 62 9 2 0
1978 3 0 2 110 165 195 59 8 2 0
1979 3 0 6 121 286 72 29 8 0 1
1980 3 0 4 26 117 90 23 2 1 0
1981 3 1 25 207 262 426 183 32 3 1
1982 3 0 27 94 180 89 123 69 5 2
1983 3 0 2 113 120 96 38 30 7 0
1984 3 0 54 229 234 144 71 12 5 7
1985 3 1 34 286 356 259 101 41 9 9
1987 3 0 48 684 642 317 163 50 11 5
1988 3 0 82 132 426 179 47 20 3 2
1889 3 0 13 331 181 213 64 18 3 0
1990 3 0 115 160 771 167 133 20 4 1
1991 3 0 14 306 187 436 79 51 13 1
1992 3 0 74 174 510 137 221 31 17 5
1993 3 0 104 363 154 196 37 49 2 2
1994 3 1 45 300 537 183 95 30 8 2
1995 3 0 21 243 341 242 52 22 4 2
1996 3 0 21 86 247 119 56 10 4 0
1997 3 0 30 113 104 202 108 54 16 6
1998 3 0 45 450 438 185 191 57 26 5
1999 3 0 18 245 307 176 56 28 5 1
2000 3 0 13 170 530 330 107 25 4 0
2001 3 0 31 190 263 345 154 34 7 2
2002 3 0 45 206 285 149 178 45 5 2
2003 3 0 30 283 439 305 137 83 28 3
2004 3 0 25 278 451 276 116 25 13 1
2005 3 0 5 91 352 207 80 28 9 1
2006 3 0 6 108 265 268 124 30 8 0
2007 3 0 140 384 760 744 512 134 36 6
2008 3 1 32 841 458 309 153 54 17 1
2009 3 0 42 63 466 166 99 29 11 1
2010 3 0 1 222 67 428 114 60 23 7
2011 3 0 103 77 1170 205 260 60 18 8
##_wt_obs
61
Mean weight-at-age in kilograms (interpolated: from HCAM.rep)
##$yr V1 V2 V3 V4 V5 V6 V7 V8 V9
1951 0.0417 0.0897 0.1126 0.1376 0.1586 0.1706 0.2001 0.1865 0.1865
1952 0.0432 0.0896 0.1133 0.1385 0.1599 0.1756 0.1677 0.1780 0.1865
1953 0.0329 0.0763 0.0958 0.1255 0.1498 0.1599 0.1335 0.1784 0.1865
1954 0.0447 0.0847 0.1036 0.1278 0.1571 0.1720 0.1751 0.1860 0.1865
1955 0.0499 0.0894 0.1043 0.1261 0.1448 0.1738 0.1720 0.1784 0.1865
1956 0.0471 0.0841 0.1060 0.1205 0.1410 0.1587 0.1785 0.1733 0.1865
1957 0.0481 0.0826 0.1089 0.1371 0.1506 0.1640 0.2100 0.1784 0.1865
1958 0.0438 0.0763 0.1081 0.1421 0.1535 0.1647 0.1762 0.1885 0.1865
1959 0.0502 0.0836 0.0993 0.1314 0.1596 0.1677 0.1523 0.1580 0.1865
1960 0.0530 0.0902 0.1126 0.1279 0.1597 0.1290 0.1770 0.1784 0.1865
1961 0.0576 0.0784 0.1111 0.1170 0.1372 0.1650 0.1720 0.1784 0.1865
1962 0.0501 0.0880 0.0993 0.1360 0.1449 0.1530 0.2020 0.1784 0.1865
1963 0.0477 0.0832 0.1049 0.1116 0.1494 0.1778 0.1790 0.1784 0.1865
1964 0.0576 0.0993 0.1159 0.1378 0.1586 0.1850 0.1400 0.1784 0.1865
1965 0.0611 0.1051 0.1224 0.1451 0.1452 0.1770 0.1440 0.1784 0.1865
1966 0.0520 0.1067 0.1487 0.1681 0.1783 0.1888 0.1720 0.1784 0.1865
1967 0.0488 0.0880 0.1104 0.1332 0.1530 0.1677 0.1720 0.1784 0.1865
1968 0.0488 0.0880 0.1104 0.1332 0.1530 0.1677 0.1720 0.1784 0.1865
1969 0.0488 0.0880 0.1104 0.1332 0.1530 0.1677 0.1720 0.1784 0.1865
1970 0.0578 0.0880 0.1175 0.1392 0.1608 0.1733 0.1871 0.1927 0.2096
1971 0.0553 0.1056 0.1319 0.1537 0.1742 0.1898 0.1690 0.1927 0.2096
1972 0.0586 0.0897 0.1290 0.1478 0.1671 0.1778 0.1973 0.1850 0.2096
1973 0.0525 0.0993 0.1274 0.1588 0.1749 0.1928 0.1977 0.2010 0.2096
1974 0.0630 0.0877 0.1392 0.1653 0.2093 0.1733 0.2420 0.1927 0.2096
1975 0.0421 0.0828 0.1115 0.1423 0.1694 0.1934 0.1995 0.2122 0.2096
1976 0.0498 0.0849 0.1234 0.1455 0.1743 0.1914 0.2033 0.2248 0.2133
1977 0.0571 0.0884 0.1173 0.1388 0.1616 0.1914 0.2010 0.2027 0.2418
1978 0.0501 0.0853 0.1099 0.1321 0.1531 0.1672 0.1683 0.1998 0.2095
1979 0.0597 0.0848 0.1178 0.1403 0.1604 0.1754 0.1990 0.2051 0.2174
1980 0.0522 0.0780 0.1083 0.1352 0.1666 0.1703 0.1869 0.1928 0.2320
1981 0.0602 0.0857 0.1085 0.1348 0.1561 0.1703 0.1873 0.1802 0.2022
1982 0.0635 0.0945 0.1153 0.1292 0.1531 0.1618 0.1696 0.1688 0.1950
1983 0.0576 0.0898 0.1174 0.1349 0.1431 0.1580 0.1707 0.1851 0.1941
1984 0.0624 0.0881 0.1146 0.1388 0.1569 0.1590 0.1666 0.1762 0.2010
1985 0.0647 0.0867 0.1141 0.1340 0.1564 0.1703 0.1869 0.1928 0.2320
1986 0.0670 0.0894 0.1108 0.1328 0.1495 0.1703 0.1971 0.1954 0.2096
1987 0.0628 0.0874 0.1056 0.1228 0.1366 0.1522 0.1660 0.1550 0.1823
1988 0.0609 0.0900 0.1138 0.1302 0.1423 0.1547 0.1642 0.2006 0.2096
1989 0.0644 0.0841 0.1068 0.1273 0.1390 0.1474 0.1561 0.1580 0.1820
1990 0.0604 0.0851 0.1060 0.1283 0.1468 0.1572 0.1609 0.1453 0.2260
1991 0.0642 0.0894 0.1101 0.1280 0.1426 0.1553 0.1634 0.1508 0.1850
1992 0.0598 0.0902 0.1108 0.1310 0.1494 0.1592 0.1738 0.1739 0.1563
1993 0.0579 0.0922 0.1118 0.1286 0.1411 0.1528 0.1565 0.1600 0.1475
1994 0.0522 0.0803 0.1049 0.1207 0.1350 0.1405 0.1512 0.1601 0.1563
1995 0.0612 0.0856 0.1106 0.1314 0.1446 0.1617 0.1631 0.1791 0.1753
1996 0.0619 0.0829 0.1061 0.1264 0.1457 0.1557 0.1719 0.1694 0.1830
1997 0.0497 0.0826 0.1017 0.1197 0.1368 0.1463 0.1541 0.1670 0.1817
1998 0.0504 0.0725 0.0939 0.1086 0.1194 0.1331 0.1438 0.1559 0.1485
1999 0.0460 0.0798 0.0991 0.1134 0.1255 0.1341 0.1434 0.1508 0.1390
2000 0.0523 0.0724 0.0948 0.1116 0.1291 0.1387 0.1530 0.1598 0.1630
2001 0.0622 0.0854 0.0990 0.1197 0.1332 0.1483 0.1552 0.1454 0.1440
2002 0.0494 0.0793 0.0959 0.1074 0.1254 0.1324 0.1413 0.1640 0.0590
2003 0.0518 0.0772 0.0930 0.1050 0.1113 0.1282 0.1401 0.1280 0.1563
2004 0.0498 0.0729 0.0888 0.0993 0.1091 0.1125 0.1241 0.1219 0.1320
2005 0.0476 0.0741 0.0910 0.1064 0.1175 0.1257 0.1305 0.1215 0.1373
2006 0.0483 0.0714 0.0842 0.1023 0.1113 0.1208 0.1279 0.1394 0.1780
2007 0.0620 0.0753 0.0831 0.0994 0.1154 0.1238 0.1301 0.1432 0.1340
2008 0.0257 0.0654 0.0858 0.0938 0.1030 0.1108 0.1155 0.1331 0.1350
2009 0.0453 0.0644 0.0688 0.1033 0.1156 0.1251 0.1347 0.1542 0.1780
2010 0.0435 0.0580 0.0790 0.0850 0.1118 0.1189 0.1158 0.1199 0.1413
2011 0.0353 0.0686 0.0721 0.0906 0.0948 0.1083 0.1194 0.1430 0.1220
#eof
999
#####
## SOG HERRING CONTROLS
## CONTROLS FOR ESTIMATED PARAMETERS
## Prior descriptions:
## -0 uniform (0,0)
## -1 normal (p1=mu,p2=sig)
## -2 lognormal (p1=log(mu),p2=sig)
## -3 beta (p1=alpha,p2=beta)
## -4 gamma(p1=alpha,p2=beta)
#####
## # npar
## ival lb ub phz prior p1 p2 parameter name
## 7.20 -5.0 15 4 0 -5.0 15 #log_ro
## 0.67 0.2 1.0 4 3 10.0 4.925373 #steepness
## -0.7985077 -5.0 5.0 3 1 -0.7985077 0.2 #log_m
## 6.80 -5.0 15 1 0 -5.0 15 #log_avgrec
## 5.80 -5.0 15 1 0 -5.0 15 #log_recinit
## 0.3043478 0.001 0.999 3 3 17.08696 39.0559 #rho
## 0.8695652 0.01 5.0 3 4 25.0 28.75 #kappa (precision)
#####
## SELECTIVITY PARAMETERS
## OPTIONS FOR SELECTIVITY:
## 1) logistic selectivity parameters
## 2) selectivity coefficients
## 3) a constant cubic spline with age-nodes
## 4) a time varying cubic spline with age-nodes
## 5) time varying bicubic spline with age & year nodes.
## 6) fixed logistic (set isel_type=6, and estimation phase to -1)
## 7) logistic function of body weight.
## sig=0.05 0.10 0.15 0.20 0.30 0.40 0.50
## wt=200. 50.0 22.2 12.5 5.56 3.12 2.00
## Gear 1:3 fishery: Gear 4-5 survey
## isel_type
## 1 1 8 6 6
## Age at 50% selectivity (logistic)
## 1.5 2.0 0.6 2.055 2.055
## STD at 50% selectivity (logistic)
## 0.25 0.25 0.15 0.05 0.05
## No. of age nodes for each gear (0 to ignore).

```

```

5      5      5      0      0
## No. of year nodes for each gear (0 to ignore).
12     3      10     0      0

## Estimation phase
2      2      2      -1      -1
## Penalty weight for 2nd differences w=1/(2*sig^2)
125.   125.   12.5   12.5   12.5
## Penalty weight for dome-shaped selectivity 1=1/(2*sig^2)
50.0   50.0   200.0  200.0  200.0
## -----
## ----- Priors for Survey q -----
## -----
## nits #number of surveys
2
## priors 0=uniform density      1=normal density
1      1
## prior log(mean)
-0.569 -0.569
## prior sd
0.274  0.274

## eofc
999
## ----- OTHER MISCELLANEOUS CONTROLS -----
## 0      ## 1 verbose ADMB output (0=off, 1=on)
1      ## 2 recruitment model (1=beverton-holt, 2=ricker)
0.100  ## 3 std in observed catches in first phase.
0.0707 ## 4 std in observed catches in last phase.
0      ## 5 Assume unfished in first year (0=FALSE, 1=TRUE)
0.02   ## 6 Minimum proportion to consider in age-proportions for dmvlogistic
0.20   ## 7 Mean fishing mortality for regularizing the estimates of Ft
0.05   ## 8 std in mean fishing mortality in first phase
2.00   ## 9 std in mean fishing mortality in last phase
3      ## 10 phase for estimating m_deviations (use -1 to turn off mdevs)
0.1    ## 11 std in deviations for natural mortality
12    ## 12 number of estimated nodes for deviations in natural mortality
0.99   ## 13 defraction of total mortality that takes place prior to spawning
1      ## 14 switch for age-composition likelihood (1=dmvlogistic, 2=dmultinom)
## -----
## 
```

### B.2.5 West Coast of Vancouver Island

```

##NB The data herein were taken from sog2010_final.dat for the HCAM model.
## -----
## -----Model Dimensions-----
1951 #first year of data
2011 #last year of data
2  #age of youngest age class
10 #age of plus group
5 #number of gears (ngear)
## flags for fishery (1) or survey (0) in ngears
1 1 0 0
## -----
## -----
## -----
##Age-schedule and population parameters
#natural mortality rate (m)
0.334
#growth parameters (linf,k,to) (from fishbase)
27.0, 0.48,
#length-weight allometry (a,b)
4.5e-6, 3.1270
#maturity at age (am=log(3)/k) & gm=std for logistic
2.055, 0.05
## -----
## -----
##Time series data
#Observed catch (1951-2010, metric t)
#yr p1 p2 p3 survey
1951 21.821 0.000 0.000 0 0
1952 27.008 0.000 0.000 0 0
1953 0.020 0.000 0.000 0 0
1954 33.209 0.000 0.000 0 0
1955 6.123 0.000 0.000 0 0
1956 17.098 0.000 0.000 0 0
1957 2.612 0.000 0.000 0 0
1958 0.556 0.000 0.000 0 0
1959 69.223 0.000 0.000 0 0
1960 53.911 0.000 0.000 0 0
1961 26.435 0.000 0.000 0 0
1962 23.684 0.000 0.000 0 0
1963 18.206 0.000 0.000 0 0
1964 21.266 0.000 0.000 0 0
1965 16.046 0.000 0.000 0 0
1966 10.843 0.000 0.000 0 0
1967 15.145 0.000 0.000 0 0
1968 0.000 0.000 0.000 0 0
1969 0.000 0.000 0.000 0 0
1970 0.000 0.000 0.000 0 0
1971 0.000 0.000 0.000 0 0
1972 0.000 6.894 0.000 0 0
1973 0.000 16.766 1.537 0 0
1974 0.000 12.394 3.940 0 0
1975 0.000 17.799 8.309 0 0
1976 0.000 22.820 16.005 0 0
1977 0.029 17.458 12.556 0 0
1978 2.839 5.151 14.755 0 0
1979 0.084 10.472 8.138 0 0
1980 0.000 1.682 2.300 0 0
1981 0.000 5.008 3.079 0 0
1982 0.000 2.370 3.115 0 0
1983 0.000 6.141 2.434 0 0
1984 0.000 5.718 0.858 0 0
1985 0.000 0.177 0.000 0 0
1986 0.000 0.203 0.000 0 0
1987 0.000 13.463 2.471 0 0
1988 0.000 8.276 1.448 0 0
1989 0.000 9.774 3.515 0 0
1990 0.000 7.890 1.950 0 0
1991 0.000 6.299 2.336 0 0
1992 0.000 3.086 0.627 0 0
1993 0.000 5.612 0.000 0 0
1994 0.000 5.332 0.706 0 0
1995 0.000 1.947 0.000 0 0
1996 0.000 0.790 0.000 0 0
1997 0.000 6.656 0.000 0 0
1998 0.000 5.450 1.534 0 0
1999 0.000 3.405 0.968 0 0
2000 0.000 0.926 0.700 0 0
2001 0.000 0.000 0.000 0 0
2002 0.000 0.433 0.388 0 0
2003 0.000 2.571 0.945 0 0
2004 0.000 3.861 0.593 0 0
2005 0.000 3.373 0.896 0 0
2006 0.000 0.000 0.000 0 0
2007 0.000 0.000 0.000 0 0
2008 0.000 0.000 0.000 0 0
2009 0.000 0.000 0.000 0 0
2010 0.000 0.000 0.000 0 0
2011 0.000 0.000 0.000 0 0
#
#Relative Abundance index from fisheries independent survey (it) 1970-2008
#nit
2
#nit_nobs
37 24
#survey type
## 1 = survey is proportional to vulnerable numbers
## 2 = survey is proportional to vulnerable biomass
## 3 = survey is proportional to spawning biomass (e.g., herring spawn survey)
3 3
#iyr it gear wt survey timing
1951 19.597 4 1 1
1952 13.310 4 1 1
1953 39.571 4 1 1
1954 20.648 4 1 1
1955 15.112 4 1 1
1956 27.183 4 1 1
1957 44.114 4 1 1
1958 18.986 4 1 1
1959 12.979 4 1 1
1960 6.015 4 1 1
1961 10.556 4 1 1
1962 34.470 4 1 1
1963 11.245 4 1 1
1964 22.761 4 1 1
1965 11.891 4 1 1
1966 3.722 4 1 1
1967 4.813 4 1 1
1968 11.029 4 1 1
1969 10.465 4 1 1
1970 26.912 4 1 1
1971 36.206 4 1 1
1972 41.857 4 1 1
1973 19.481 4 1 1
1974 25.540 4 1 1
1975 49.149 4 1 1
1976 64.222 4 1 1
1977 58.679 4 1 1
1978 45.607 4 1 1
1979 66.397 4 1 1
1980 62.308 4 1 1
1981 52.063 4 1 1
1982 33.047 4 1 1
1983 16.771 4 1 1
1984 23.872 4 1 1
1985 30.010 4 1 1
1986 39.514 4 1 1
1987 16.858 4 1 1
1988 46.242 5 1 1
1989 47.718 5 1 1
1990 46.464 5 1 1
1991 30.456 5 1 1
1992 42.687 5 1 1
1993 34.728 5 1 1
1994 25.625 5 1 1
1995 28.057 5 1 1
1996 33.986 5 1 1
1997 46.490 5 1 1
1998 41.556 5 1 1
1999 20.390 5 1 1
2000 13.267 5 1 1
2001 13.955 5 1 1
2002 22.086 5 1 1

```

```

2003 29.750      5  1       1
2004 15.844      5  1       1
2005 9.075       5  1       1
2006 2.705       5  1       1
2007 2.089       5  1       1
2008 2.548       5  1       1
2009 9.876       5  1       1
2010 2.373       5  1       1
2011 9.196       5  1       1
#
#Age composition data by year, gear (ages 2-15+)
#na_gears
3
#na_nobs
18 40 26
#a_sage
2 2 2
#a_page
10 10
#
#yr gear    V2   V3   V4   V5   V6   V7   V8   V9   V10  #Number aged.
1951  1   508 1519 1666 272   58  12  1   1   0
1952  1   97 1431 1224 1809 241  72  16  2   0
1953  1   465 2220 1086 65   19  2   0   0   0
1954  1   163 3852 1681 338   42  9   5  1   1
1955  1   418 1471 484   86   16  1   0   0   0
1956  1   575 2990 743   282   52  7   2   2   0
1957  1   16  423 146   2     1   0   0   0   0
1958  1   193 770 376   81   34  20  5   1   0
1959  1   148 1607 993 519   140  88  74  21  3
1960  1   254 1561 662 246   80  27 10  4   2
1961  1   226 224 113 26     1   0   0   0   0
1962  1   56  957 112 28     10   0   0   0   0
1963  1   37  804 907 96     14   4   0   0   0
1964  1   16  677 284 118   9   3   0   0   0
1965  1   18  269 372 75     27   3   0   0   0
1966  1   1  101 78 53     6   3   0   0   0
1971  1   25  193 160 34     12   7   2   0   0
1978  1   0  180 156 105   106  20  6  4   2
1972  2   50  279 716 356   51  18 11  1   0
1973  2   18  776 620 817   276  40  7   2   0
1974  2  433 2324 1290 731   476 120 12  2   1
1975  2   60 5405 1983 1140   804 498 128 17  1
1976  2   19  818 4332 1828 1196 746 251 40  0
1977  2   35  830 2077 2487   828 294 112 18  3
1978  2   63 3164 1407 1177 1294 287 87 14  5
1979  2   30  513 1848 525   398 293 59 19  4
1980  2   232 1640 582 747   238 198 82 13  2
1981  2   232 1688 1423 499   603 261 103 29  7
1982  2   156 1144 1309 1244   260 454 130 65  5
1983  2   135 719 696 699   562 142 172 34 26
1984  2   669 1146 418 282   309 182 33 33  5
1985  2   613 1606 426 111   82  98 51 4  6
1986  2   157 2094 1233 344   130  93 73 24  3
1987  2   783 863 1709 1053   351 123 71 52 13
1988  2   193 4587 584 1110   738 209 55 33 15
1989  2   155 960 3693 450   534 271 44 10  3
1990  2   33 1856 849 3233   307 406 125 16  4
1991  2   482 1655 1543 780 2420 220 251 48  2
1992  2   104 2960 662 827   362 1029 129 75 12
1993  2   214 1528 2255 380   416 226 423 51 27
1994  2   183 1407 1485 1875   492 311 330 98 15
1995  2   44 667 1304 1087 1335 348 226 166 31
1996  2   1079 1505 909 1457   945 968 204 85 42
1997  2   145 4109 487 266   402 294 198 32 20
1998  2   123 1224 3239 347   155 176 104 60 11
1999  2   65 960 1044 1641   325 112 63 31 14
2000  2   284 1170 984 894 1374 196 67 35 9
2001  2   165 1074 475 197   178 222 31 5  3
2002  2   348 2658 1136 371   140 157 131 15 1
2003  2   96 2191 2042 705   135 62 42 33 4
2004  2   390 1295 2431 1002   283 64 21 11 0
2005  2   157 1655 939 680   237 71 12 2 3
2006  2   174 430 387 91   62  9  1  0  0
2007  2   7  303 211 66   11  4  0  0  0
2008  2   54 255 559 119   32  8  6  1  1
2009  2   44 1202 283 230   41 10  0  0  0
2010  2   211 577 839 105   88 14  2  0  0
2011  2   43 523 189 64   3  3  0  0  0
1973  3   0  49 131 286   68 17  4  1  0
1974  3   0  46 43 43   16  6  0  0  0
1975  3   0  7  78 88 48   19 1  0  0  0
1976  3   0  8 495 408 179   73 31 4  1
1977  3   2  12 58 144 56   37 17  5  1
1978  3   0  6 23 90 207   74 20 1  1
1979  3   0  5 118 136 113 88 15 2  1
1980  3   0  0 19 188 80 43 26 1  1
1981  3   0  5 59 42 102 53 20 0  0
1982  3   0  4 69 254 67 158 19 2  1
1983  3   0  2 81 136 256 37 56 2  1
1984  3   0  10 40 107 194 190 32 20  1
1987  3   0  10 135 340 30 12 16 5  2
1988  3   0  25 33 192 133 60 14 6  2
1989  3   0  1 208 42 85 36 6  4  0
1990  3   0  6 35 307 37 46 11 3  0
1991  3   0  1 21 39 198 21 25 2  0
1992  3   0  35 75 171 77 166 16 14 2
1994  3   1  35 199 340 33 7  4  1  0
1998  3   0  5 344 99 87 181 111 51 19
1999  3   0  9 113 612 193 58 38 18 2
2000  3   0  8 47 169 330 39 16 14 1
2002  3   0  0 55 154 82 110 120 12 2
2003  3   0  13 87 159 99 49 64 25 3
2004  3   0  5 179 154 158 92 24 14 5
2005  3   0  4 54 249 119 53 19 1  1
#n_wt_obs
61

```

```

    1.5      2.0 0.6 2.05 2.05
## STD at 50% selectivity (logistic)
  0.75      0.5 0.15 0.05 0.05
## No. of age nodes for each gear (0 to ignore).
  5          5 5 0 0
## No. of year nodes for each gear (0 to ignore).
 12         3 10 0 0
## Estimation phase
  2          2 2 -2 -2
## Penalty weight for 2nd differences w=1/(2*sig^2)
 125.0     12.5 12.5 12.5 12.5
## Penalty weight for dome-shaped selectivity 1=1/(2*sig^2)
  3.125    200.0 200.0 200.0 200.0
## -----
## ----- Priors for Survey q -----
## -----
## nits #number of surveys
2
## priors 0=uniform density 1=normal density
1 1
## prior log(mean)
-0.569 -0.569
## prior sd

```

## B.2.6 Area 2W

```

#NB The data herein were taken from a2w2010_final.dat for the HCAM model.
## -----
## -----Model Dimensions-----
1978      #first year of data
2011      #last year of data
2          #age of youngest age class
10         #age of plus group
4          #number of gears (ngear)
## flags for fishery (1) or survey (0) in ngears
#1 1 0
0.100 0.405 0.495 0.0
## -----
## Age-schedule and population parameters
#natural mortality rate (m)
0.334
#growth parameters (linf,k,to) (from fishbase)
27.0, 0.48, 0
#length-weight allometry (a,b)
4.5e-6, 3.1270
#maturity at age (am=log(3)/k) & gm=std for logistic
2.055, 0.05
## -----
#Time series data
#Observed catch (1951-2010, 1000s metric t)
#Year P1 P2 P3 S1
1978 0 0.575 0 0
1979 0 0.691 0 0
1980 0 0.000 0 0
1981 0 0.770 0 0
1982 0 1.225 0 0
1983 0 2.518 0 0
1984 0 0.000 0 0
1985 0 0.199 0 0
1986 0 0.000 0 0
1987 0 0.000 0 0
1988 0 0.000 0 0
1989 0 0.000 0 0
1990 0 2.272 0 0
1991 0 2.558 0 0
1992 0 1.284 0 0
1993 0 1.306 0 0
1994 0 0.000 0 0
1995 0 0.000 0 0
1996 0 0.000 0 0
1997 0 0.000 0 0
1998 0 0.180 0 0
1999 0 0.000 0 0
2000 0 0.000 0 0
2001 0 0.000 0 0
2002 0 0.000 0 0
2003 0 0.000 0 0
2004 0 0.000 0 0
2005 0 0.000 0 0
2006 0 0.000 0 0
2007 0 0.000 0 0
2008 0 0.000 0 0
2009 0 0.000 0 0
2010 0 0.000 0 0
2011 0 0.000 0 0
##Relative Abundance index from fisheries independent survey (it) 1970-2008
#nit
#61
1
#init_nobs
30
#survey type
## 1 = survey is proportional to vulnerable numbers
## 2 = survey is proportional to vulnerable biomass
## 3 = survey is proportional to spawning biomass (e.g., herring spawn survey)
3
#yrs    it gear wt    survey timing
1978   0.832 4 1 1
1979   0.494 4 1 1

```

```

          0.274  0.274
## -----
## ----- OTHER MISCELLANEOUS CONTROLS -----
## 0 ## verbose ADMB output (0=off, 1=on)
## 1 ## recruitment model (1=beverton-holt, 2=ricker)
0.100 ## std in observed catches in first phase.
0.0707 ## std in observed catches in last phase.
0 ## Assume unfished in first year (0=FALSE, 1=TRUE)
0.02 ## Minimum proportion to consider in age-proportions for dmvglogistic
0.20 ## Mean fishing mortality for regularizing the estimates of Ft
0.01 ## std in mean fishing mortality in first phase
5.00 ## std in mean fishing mortality in last phase
3 ## phase for estimating m_deviations (use -1 to turn off mdevs)
0.1 ## std in deviations for natural mortality
12 ## number of estimated nodes for deviations in natural mortality
0.99 ## fraction of total mortality that takes place prior to spawning
1      ## 14 switch for age-composition likelihood (1=dmvglogistic, 2=dmultinom)
## -----
## eofc
999

```

```

1979 0.0528 0.1004 0.1373 0.1557 0.1809 0.1903 0.2110 0.2030 0.1870
1980 0.0600 0.0955 0.1197 0.1686 0.1987 0.2200 0.2124 0.2465 0.2174
1981 0.0686 0.0934 0.1256 0.1572 0.1822 0.1901 0.2018 0.1900 0.2255
1982 0.0657 0.1132 0.1233 0.1559 0.1814 0.1889 0.2148 0.2100 0.2220
1983 0.0758 0.1076 0.1415 0.1579 0.1783 0.1951 0.2024 0.1962 0.2136
1984 0.0728 0.1073 0.1307 0.1563 0.1886 0.1851 0.1841 0.1867 0.2140
1985 0.0858 0.1178 0.1531 0.1791 0.2042 0.2102 0.2188 0.2191 0.2260
1986 0.0800 0.1160 0.1490 0.1623 0.1843 0.2116 0.2270 0.2315 0.2070
1987 0.0628 0.1030 0.1270 0.1702 0.2018 0.1864 0.2227 0.1958 0.2245
1988 0.0707 0.1005 0.1429 0.1583 0.1818 0.2066 0.2209 0.2391 0.2370
1989 0.0626 0.1011 0.1317 0.1579 0.1809 0.1912 0.2034 0.2161 0.2163
1990 0.0588 0.0942 0.1414 0.1638 0.1868 0.1917 0.2304 0.2070 0.2344
1991 0.0622 0.0959 0.1278 0.1676 0.1756 0.1892 0.2002 0.2123 0.2049
1992 0.0698 0.1073 0.1342 0.1460 0.1777 0.1965 0.2094 0.2066 0.2168
1993 0.0685 0.1045 0.1283 0.1460 0.1691 0.1771 0.1892 0.1979 0.1922
1994 0.0748 0.1147 0.1393 0.1514 0.1738 0.1530 0.1999 0.1985 0.1960
1995 0.0655 0.0959 0.1209 0.1482 0.1725 0.1767 0.1877 0.1996 0.2081
1996 0.0655 0.0959 0.1209 0.1482 0.1725 0.1767 0.1877 0.1996 0.2081
1997 0.0658 0.0959 0.1209 0.1482 0.1725 0.1767 0.1877 0.1996 0.2081
1998 0.0703 0.1053 0.1334 0.1677 0.1758 0.1714 0.1973 0.1940 0.2081
1999 0.0713 0.1069 0.1209 0.1476 0.1682 0.1658 0.1340 0.1870 0.2081
2000 0.0693 0.0830 0.0881 0.1482 0.2045 0.1767 0.1110 0.1996 0.2081
2001 0.0688 0.1038 0.1478 0.1717 0.1769 0.1794 0.2155 0.2081
2002 0.0655 0.1075 0.1254 0.1747 0.1969 0.2039 0.2041 0.2041 0.2405
2003 0.0742 0.1030 0.1144 0.1337 0.1828 0.1994 0.1965 0.1922 0.1917
2004 0.0621 0.0952 0.1298 0.1437 0.1622 0.1992 0.2460 0.1996 0.2235
2005 0.0598 0.0842 0.1093 0.1392 0.1558 0.1477 0.1738 0.1996 0.1900
2006 0.0594 0.0773 0.1037 0.1374 0.1698 0.1844 0.2094 0.2113 0.2081
2007 0.0800 0.0822 0.0879 0.1174 0.1406 0.1585 0.1550 0.1753 0.2000
2008 0.0555 0.0753 0.1102 0.1286 0.1558 0.1448 0.1643 0.1996 0.2081
2009 0.0560 0.0877 0.1015 0.1393 0.1563 0.1613 0.1922 0.1903 0.1923
2010 0.0559 0.0923 0.1227 0.1347 0.1684 0.1691 0.1722 0.1852 0.2062
2011 0.0564 0.0941 0.1172 0.1411 0.1277 0.1554 0.1608 0.1570 0.2000
#eof
999

## -----
##          AREA 2W HERRING CONTROLS
##          CONTROLS FOR ESTIMATED PARAMETERS
## -----
## Prior descriptions:
##   -0 uniform (0,0)
##   -1 normal (p1=mu,p2=sig)
##   -2 lognormal (p1=log(mu),p2=sig)
##   -3 beta (p1=alpha,p2=beta)
##   -4 gamma(p1=alpha,p2=beta)
## -----
7 ## npar
## ival lb ub phz prior p1 p2 parameter name
## -----
2.60 -5.0 15 4 0 -5.0 15 #log_ro
0.67 0.2 1.0 4 3 10.0 4.925373 #steepness
-0.7985077 -5.0 5.0 3 1 -0.7985077 0.4 #log_m
2.40 -5.0 15 1 0 -5.0 15 #log_avgrec
2.20 -5.0 15 1 0 -5.0 15 #log_recinit
0.3043478 0.001 0.999 3 3 17.08696 39.0559 #rho
0.8695652 0.01 5.0 3 4 25.0 28.75 #kappa (precision)
## -----
## -----
##          SELECTIVITY PARAMETERS
## -----
## OPTIONS FOR SELECTIVITY:

```

## B.2.7 Area 27

```

#NB The data herein were taken from a2w2010_final.dat for the HCAM model.
## -----
##          Model Dimensions
1978 #first year of data
2011 #last year of data
2  #age of youngest age class
10 #age of plus group
5 #number of gears (ngear)
## flags for fishery (1) or survey (0) in ngears
1 1 0 0
0.100 0.405 0.495 0.0 0.0
## -----
## Age-schedule and population parameters
#natural mortality rate (m)
0.334
#growth parameters (linf,k,to) (from fishbase)
27.0, 0.48, 0
#length-weight allometry (a,b)
4.5e-6, 3.1270
#maturity at age (am=log(3)/k) & gm=std for logistic
2.055, 0.05
## -----
#time series data
#Observed catch (1951-2010, 1000s metric t)
#Year P1 P2 P3 S1 S2
1978 0 0.075 0.075 0 0
1979 0 0.422 0.270 0 0
1980 0 0.000 0.519 0 0
1981 0 0.000 0.671 0 0
1982 0 0.238 0.332 0 0
1983 0 0.000 0.163 0 0
1984 0 0.000 0.171 0 0
1985 0 0.000 0.000 0 0
1986 0 0.000 0.000 0 0
1987 0 0.000 0.000 0 0
1988 0 0.000 0.000 0 0
1989 0 0.000 0.000 0 0
1990 0 0.000 0.000 0 0
1991 0 0.000 0.000 0 0
1992 0 0.335 0.000 0 0
1993 0 0.000 0.367 0 0
1994 0 0.000 0.345 0 0
1995 0 0.088 0.000 0 0
1996 0 0.000 0.000 0 0
1997 0 0.000 0.000 0 0
1998 0 0.000 0.000 0 0
1999 0 0.000 0.000 0 0
2000 0 0.000 0.000 0 0
2001 0 0.000 0.000 0 0
2002 0 0.000 0.000 0 0
2003 0 0.000 0.000 0 0
2004 0 0.000 0.000 0 0
2005 0 0.000 0.000 0 0
2006 0 0.000 0.000 0 0
2007 0 0.000 0.000 0 0
2008 0 0.000 0.000 0 0
2009 0 0.000 0.000 0 0
2010 0 0.000 0.000 0 0
2011 0 0.000 0.000 0 0
#Relative Abundance index from fisheries independent survey (it) 1970-2008
#nit
#61
2
#nit_nobs
10 24
#survey type
## 1 = survey is proportional to vulnerable numbers
## 2 = survey is proportional to vulnerable biomass
## 3 = survey is proportional to spawning biomass (e.g., herring spawn survey)
3 3
#iyr it gear wt survey timing

```

```

1978 3.595 4 1 1
1979 6.909 4 1 1
1980 14.419 4 1 1
1981 1.828 4 1 1
1982 4.137 4 1 1
1983 2.720 4 1 1
1984 3.051 4 1 1
1985 1.648 4 1 1
1986 3.803 4 1 1
1987 1.372 4 1 1
1988 2.184 5 1.1666 1
1989 5.128 5 1.1666 1
1990 5.821 5 1.1666 1
1991 3.717 5 1.1666 1
1992 3.581 5 1.1666 1
1993 6.084 5 1.1666 1
1994 5.707 5 1.1666 1
1995 3.254 5 1.1666 1
1996 1.333 5 1.1666 1
1997 2.354 5 1.1666 1
1998 2.553 5 1.1666 1
1999 1.054 5 1.1666 1
2000 1.576 5 1.1666 1
2001 0.544 5 1.1666 1
2002 1.075 5 1.1666 1
2003 1.147 5 1.1666 1
2004 1.532 5 1.1666 1
2005 2.209 5 1.1666 1
2006 2.295 5 1.1666 1
2007 2.592 5 1.1666 1
2008 1.093 5 1.1666 1
2009 1.498 5 1.1666 1
2010 0.957 5 1.1666 1
2011 0.818 5 1.1666 1
#age composition data by year, gear (ages 2-15+)
#na_gears
2
#na_nobs
31 6
#_sage
2 2
#_page
10 10
#yr gear V2 V3 V4 V5 V6 V7 V8 V9 V10 #Number aged.
1978 2 1 38 4 14 12 2 0 0 0
1979 2 1 10 55 10 2 1 0 0 0
1980 2 20 229 25 4 0 0 1 0 0
1981 2 15 99 435 63 98 11 0 0 0
1982 2 7 370 105 439 43 84 8 1 0
1983 2 4 21 32 11 29 0 4 0 0
1986 2 6 64 172 7 4 5 7 6 0
1987 2 48 78 45 100 3 0 3 1 4
1988 2 8 232 41 23 57 6 3 0 1
1989 2 1 59 268 38 39 53 6 2 0
1990 2 17 210 132 367 54 66 72 6 2
1991 2 33 145 33 38 83 10 18 8 0
1992 2 49 1004 158 48 41 71 14 18 6
1993 2 72 228 248 32 10 9 32 2 3
1994 2 14 300 232 292 52 20 27 5 2
1995 2 24 91 504 348 352 59 19 23 6
1996 2 58 140 45 119 101 84 18 2 2
1997 2 23 441 42 9 23 27 9 0 0
1998 2 4 112 140 14 1 8 7 2 0
1999 2 59 213 257 189 31 4 4 2 1
2000 2 15 355 158 63 49 8 1 3 0
2001 2 13 41 70 25 24 19 2 1 0
2002 2 35 293 73 47 3 11 4 1 0
2003 2 3 295 214 36 23 1 4 1 0
2004 2 5 83 209 76 4 6 3 0 0
2005 2 1 97 43 23 13 1 1 0 0
2007 2 5 209 140 72 16 10 1 0 0
2008 2 6 12 218 80 44 5 1 0 0
2009 2 9 448 73 143 23 18 0 1 0
2010 2 15 35 154 25 36 6 7 0 0
2011 2 6 105 64 74 8 10 2 1 0
1979 3 0 0 29 12 12 7 0 0 0
1982 3 0 0 5 30 8 18 2 0 0
1983 3 0 0 7 12 50 2 9 0 0
1984 3 0 0 18 182 72 144 11 5 0
1993 3 0 17 276 73 41 39 60 5 6
1994 3 0 6 91 287 46 16 18 2 3
#n_wt_obs
34
#Mean weight-at-age in kilograms (interpolated: from HCAM.rep)
#&yr V1 V2 V3 V4 V5 V6 V7 V8 V9
1978 0.0550 0.0778 0.1035 0.1311 0.1539 0.1520 0.1768 0.2015 0.2289
1979 0.0350 0.0827 0.1028 0.1254 0.1360 0.1510 0.1780 0.2015 0.2289
1980 0.0679 0.0833 0.0977 0.1218 0.1559 0.1657 0.1610 0.2015 0.2289
1981 0.0643 0.0928 0.1115 0.1288 0.1379 0.1498 0.1768 0.2015 0.2289
1982 0.0561 0.0929 0.1098 0.1263 0.1358 0.1469 0.1661 0.1450 0.2289
1983 0.0508 0.0884 0.1060 0.1141 0.1275 0.1657 0.1365 0.2015 0.2289
1984 0.0585 0.0932 0.1173 0.1357 0.1559 0.1657 0.1768 0.2015 0.2289
1985 0.0585 0.0932 0.1173 0.1357 0.1559 0.1657 0.1768 0.2015 0.2289
1986 0.0678 0.1139 0.1375 0.1560 0.1988 0.2048 0.1987 0.2267 0.2289
1987 0.0671 0.1072 0.1508 0.1645 0.1827 0.1657 0.2113 0.2330 0.1958
1988 0.0626 0.1002 0.1368 0.1537 0.1782 0.1897 0.1865 0.2015 0.2620
1989 0.0430 0.1039 0.1376 0.1770 0.1994 0.2127 0.1977 0.2495 0.2034
1990 0.0628 0.1014 0.1380 0.1705 0.1969 0.2162 0.2264 0.2417 0.2345
1991 0.0649 0.0944 0.1192 0.1528 0.1738 0.2005 0.2063 0.2045 0.2034
1992 0.0605 0.1020 0.1328 0.1538 0.1821 0.2035 0.2212 0.2344 0.2518
1993 0.0583 0.0894 0.1190 0.1277 0.1747 0.1848 0.1971 0.1860 0.2130
1994 0.0699 0.0948 0.1112 0.1361 0.1547 0.1680 0.1857 0.1882 0.1830
1995 0.0601 0.0998 0.1167 0.1309 0.1514 0.1682 0.1748 0.2007 0.1853
1996 0.0526 0.0907 0.1126 0.1344 0.1425 0.1643 0.1688 0.1910 0.1835
1997 0.0476 0.0820 0.1092 0.1328 0.1336 0.1494 0.1579 0.1808 0.2034
1998 0.0435 0.0751 0.0972 0.0993 0.1240 0.1328 0.1530 0.1490 0.2034
1999 0.0487 0.0720 0.0891 0.1064 0.1054 0.1390 0.1243 0.1750 0.1730
2000 0.0531 0.0798 0.0888 0.1130 0.1342 0.1356 0.1500 0.1343 0.2034
2001 0.0509 0.0743 0.0911 0.1024 0.1108 0.1142 0.1135 0.1210 0.2034
2002 0.0845 0.0917 0.0989 0.1230 0.0963 0.1192 0.1450 0.1650 0.2034
2003 0.0567 0.0998 0.1070 0.1148 0.1328 0.1499 0.1628 0.1490 0.2034
2004 0.0546 0.0816 0.1013 0.1052 0.1288 0.1277 0.1163 0.1808 0.2034
2005 0.0340 0.0684 0.0775 0.1084 0.1340 0.1300 0.1540 0.1808 0.2034
2006 0.0544 0.0845 0.1026 0.1217 0.1369 0.1505 0.1594 0.1808 0.2034
2007 0.0556 0.0675 0.0744 0.0903 0.1003 0.1171 0.1290 0.1808 0.2034
2008 0.0472 0.0660 0.0791 0.0882 0.0960 0.1114 0.1060 0.1808 0.2034
2009 0.0446 0.0735 0.0729 0.1005 0.1131 0.1141 0.1594 0.1540 0.2034
2010 0.0507 0.0677 0.0821 0.0882 0.0919 0.1037 0.0999 0.1808 0.2034
2011 0.0452 0.0642 0.0741 0.0916 0.1000 0.1020 0.1235 0.0580 0.2034
##eofc
999
## -----
## AREA 2W HERRING CONTROLS
## CONTROLS FOR ESTIMATED PARAMETERS
## Prior descriptions:
## -0 uniform (0,0)
## -1 normal (p1=mu,p2=sig)
## -2 lognormal (p1=log(mu),p2=sig)
## -3 beta (p1=alpha,p2=beta)
## -4 gamma(p1=alpha,p2=beta)
## -----
## npar
## ival lb ub phz prior p1 p2 parameter name
## 3.60 -5.0 15 4 0 -5.0 15 #log_ro
## 0.67 0.2 1.0 4 3 10.0 4.925373 #steepness
## -0.7985077 -5.0 5.0 3 1 -0.7985077 0.4 #log_m
## 2.40 -5.0 15 1 0 -5.0 15 #log_avgrec
## 2.20 -5.0 15 1 0 -5.0 15 #log_recinit
## 0.3043478 0.001 0.999 3 3 17.08696 39.0559 #rho
## 0.8695652 0.01 5.0 3 4 25.0 28.75 #kappa (precision)
## -----
## SELECTIVITY PARAMETERS
## OPTIONS FOR SELECTIVITY:
## 1) logistic selectivity parameters
## 2) selectivity coefficients
## 3) a constant cubic spline with age-nodes
## 4) a time varying cubic spline with age-nodes
## 5) time varying bicubic spline with age & year nodes.
## 6) fixed logistic (set isel_type=6, and estimation phase to -1)
## 7) logistic weight of body weight.
## sig=0.05 0.10 0.15 0.20 0.30 0.40 0.50
## wt=.200. 50.0 22.2 12.5 5.56 3.12 2.00
## Gear 1:3 fishery: Gear 4-5 survey
## isel_type
## 1 1 1 6 6
## Age at 50% selectivity (logistic)
## 2.0 3.0 3.6 2.055 2.055
## STD at 50% selectivity (logistic)
## 0.25 0.25 0.25 0.05 0.05
## No. of age nodes for each gear (0 to ignore).
## 5 5 5 0 0
## No. of year nodes for each gear (0 to ignore).
## 12 3 10 0 0
## Estimation phase
## -2 2 2 -1 -1
## Penalty weight for 2nd differences w1/(2*sig^2)
## 125. 50. 12.5 12.5 12.5
## Penalty weight for dome-shaped selectivity 1=(2*sig^2)
## 50.0 50.0 200.0 200.0 200.0
## -----
## Priors for Survey q
## nits #number of surveys
## priors 0=uniform density 1=normal density
## 1 1
## prior log(mean)
## -0.569 -0.569
## prior sd
## 0.274 0.274
## -----
## OTHER MISCELLANEOUS CONTROLS
## 1 verbose ADMB output (0=off, 1=on)
## 2 recruitment model (1=beverton-holt, 2=ricker)
## 3 std in observed catches in first phase.
## 4 std in observed catches in last phase.
## 5 Assume unfished in first year (0=FALSE, 1=TRUE)
## 6 Minimum proportion to consider in age-proportions for dmvlogistic
## 0.20 7 Mean fishing mortality for regularizing the estimates of Ft
## 0.01 8 std in mean fishing mortality in first phase
## 2.00 9 std in mean fishing mortality in last phase
## 3 10 phase for estimating _m_deviations (use -1 to turn off mdevs)
## 0.1 11 std in deviations for natural mortality
## 12 12 number of estimated nodes for deviations in natural mortality
## 1.00 13 fraction of total mortality that takes place prior to spawning
## 1 14 switch for age-composition likelihood (1=dmvlogistic, 2=dmultinom)
## -----

```

## B.2.8 Control file for the simulation studies

The following control file was used in the simulation study where the data were generated with absolutely no error based on catch data from the Strait of Georgia. The purpose of this simulation study was to demonstrate that <sup>i</sup>SCA<sub>M</sub> is capable of estimating the true parameter values given perfect information.

```
## ----- SOG HERRING CONTROLS -----
## ----- CONTROLS FOR ESTIMATED PARAMETERS -----
## Prior descriptions:
##   -0 uniform (0,0)
##   -1 normal (p1=mu,p2=sig)
##   -2 lognormal (p1=log(mu),p2=sig)
##   -3 beta (p1=alpha,p2=beta)
##   -4 gamma(p1=alpha,p2=beta)
## -----
7 ## npar
## ival lb ub phz prior p1 p2 parameter name ##
## -----
7.60 -5.0 15 4 0 -5.0 15 #log_ro/fmsy
0.80 0.2 1.0 4 0 1.1 1.1 #steepness/fmsy
-0.7985077 -5.0 0.0 -3 0 -0.7985 0.2 #log.m
7.60 -5.0 15 3 0 -5.0 15 #log_avgrec
7.60 -5.0 15 3 0 -5.0 15 #log_recinit
0.05 0.001 0.999 -3 3 1.01 1.01 #rho
4999 0.01 5000 -3 4 1.01 1.01 #vartheta
## -----
## -----
## ----- SELECTIVITY PARAMETERS -----
## OPTIONS FOR SELECTIVITY:
##   1) logistic selectivity parameters
##   2) selectivity coefficients
##   3) a constant cubic spline with age-nodes
##   4) a time varying cubic spline with age-nodes
##   5) a time varying bicubic spline with age & year nodes.
##   6) fixed logistic (set isel_type=1, and estimation phase to -1)
## Gear 1:3 fishery: Gear 4-5 survey
## isel_type
1 1 1 1 1
## Age at 50% selectivity (logistic)
1.5 2.5 3.5 2.05 2.05
## STD at 50% selectivity (logistic)
0.5 0.5 0.5 0.05 0.05
## No. of age nodes for each gear (0 to ignore).
5 5 5 0 0
## No. of year nodes for each gear (0 to ignore).
12 3 10 0 0
## Estimation phase
-2 -2 -2 -2 -2
## Penalty weight for 2nd differences w=1/(2*sig^2)
12.5 12.5 12.5 12.5
## Penalty weight for dome-shaped selectivity 1=1/(2*sig^2)
3.125 200.0 200.0 200.0 200.0
## -----
## -----
## ----- Priors for Survey q -----
## nits #number of surveys
2
## priors 0=uniform density 1=normal density
0 0
## prior log(mean)
0 0
## prior sd
1 1
## -----
## ----- OTHER MISCELLANEOUS CONTROLS -----
0 ## 1 verbose ADMB output (0=off, 1=on)
1 ## 2 recruitment model (1=beverton-holt, 2=ricker)
0.05 ## 3 std in observed catches in first phase.
0.01 ## 4 std in observed catches in last phase.
0 ## 5 Assume unfished in first year (0=FALSE, 1=TRUE)
0.01 ## 6 Minimum proportion to consider in age-proportions for dmvlogistic
0.05 ## 7 Mean fishing mortality for regularizing the estimates of Ft
0.01 ## 8 std in mean fishing mortality in first phase
5.00 ## 9 std in mean fishing mortality in last phase
-3 ## 10 phase for estimating m_deviations (use -1 to turn off mdevs)
0.01 ## 11 std in deviations for natural mortality
12 ## 12 number of estimated nodes for deviations in natural mortality
1.00 ## 13 fraction of total mortality that takes place prior to spawning
1 ## 14 switch for age-composition likelihood (1=dmvlogistic,2=dmultinom)
## -----
## eofc
999
```

The following control file was used in the simulation study where data were generated with observation error and process errors.

```
## ----- SOG HERRING CONTROLS for Monte Carlo trials -----
## ----- CONTROLS FOR ESTIMATED PARAMETERS -----
## Prior descriptions:
##   -0 uniform (0,0)
##   -1 normal (p1=mu,p2=sig)
##   -2 lognormal (p1=log(mu),p2=sig)
##   -3 beta (p1=alpha,p2=beta)
##   -4 gamma(p1=alpha,p2=beta)
## -----
7 ## npar
## ival lb ub phz prior p1 p2 parameter name ##
## -----
7.60 -5.0 15 4 0 -5.0 15 #log_ro/fmsy
0.80 0.2 1.0 4 3 1.1 1.1 #steepness/fmsy
-0.7985 -5.0 0.0 3 1 -0.7985 0.2 #log.m
7.40 -5.0 15 1 0 -5.0 15 #log_avgrec
7.40 -5.0 15 1 0 -5.0 15 #log_initrec
0.25 0.001 0.999 -3 3 1.01 1.01 #rho
2.50 0.01 15 -3 4 1.01 1.01 #vartheta
## -----
## -----
## ----- SELECTIVITY PARAMETERS -----
## OPTIONS FOR SELECTIVITY:
##   1) logistic selectivity parameters
##   2) selectivity coefficients
##   3) constant cubic spline with age-nodes
##   4) time varying cubic spline with age-nodes
##   5) a time varying bicubic spline with age & year nodes.
##   6) fixed logistic (set isel_type=1, and estimation phase to -1)
## Gear 1:3 fishery: Gear 4-5 survey
## isel_type
1 1 1 1 1
## Age at 50% selectivity (logistic)
1.5 2.5 3.5 2.05 2.05
## STD at 50% selectivity (logistic)
0.5 0.5 0.5 0.05 0.05
## No. of age nodes for each gear (0 to ignore).
5 5 5 0 0
## No. of year nodes for each gear (0 to ignore).
12 3 10 0 0
## Estimation phase
2 2 2 -2 -2
## Penalty weight for 2nd differences w=1/(2*sig^2)
12.5 12.5 12.5 12.5 12.5
## Penalty weight for dome-shaped selectivity 1=1/(2*sig^2)
3.125 200.0 200.0 200.0 200.0
## -----
## -----
## ----- Priors for Survey q -----
## nits #number of surveys
2
## priors 0=uniform density 1=normal density
0 0
## prior log(mean)
0 0
## prior sd
1 1
## -----
## ----- OTHER MISCELLANEOUS CONTROLS -----
0 ## 1 verbose ADMB output (0=off, 1=on)
1 ## 2 recruitment model (1=beverton-holt, 2=ricker)
0.05 ## 3 std in observed catches in first phase.
0.01 ## 4 std in observed catches in last phase.
0 ## 5 Assume unfished in first year (0=FALSE, 1=TRUE)
0.01 ## 6 Minimum proportion to consider in age-proportions for dmvlogistic
0.05 ## 7 Mean fishing mortality for regularizing the estimates of Ft
0.01 ## 8 std in mean fishing mortality in first phase
5.00 ## 9 std in mean fishing mortality in last phase
-3 ## 10 phase for estimating m_deviations (use -1 to turn off mdevs)
0.01 ## 11 std in deviations for natural mortality
12 ## 12 number of estimated nodes for deviations in natural mortality
1.00 ## 13 fraction of total mortality that takes place prior to spawning
1 ## 14 switch for age-composition likelihood (1=dmvlogistic,2=dmultinom)
## -----
## eofc
999
```

## C.3 Bayesian prior for the dive survey spawn index proportionality constant $q$

### C.3.1 The process

A Bayesian prior for the herring dive survey spawn index proportionality constant ( $q$ ) is developed using a process that combines expert knowledge (in some cases best guesses) and data associated with factors influencing the prior. The process, used to develop acoustic and trawl survey priors for New Zealand fisheries stock assessments (e.g., [Cordue, 2006](#)), is comprised of the following steps:

1. List all factors affecting  $q$ .
2. For each factor, determine the statistical distribution that best describes the uncertainty associated with that factor. Where available, the distribution is based on data (not data that will be used in the assessment); otherwise it is based on expert knowledge.
3. The prior distribution for  $q$  is estimated by integrating across the distributions for each factor. This can be approximated by generating joint random samples from the distributions.
4. Finally, a parametric model is fit to the resulting distribution of replicate random samples to approximate the  $q$  prior.

### C.3.2 Factors affecting $q$ and their distributions

The factors that contribute to the spawn index  $q$  prior include: the proportion of the total spawn that is surveyed; the amount of egg loss that occurs prior to the spawn survey; bias in the estimate of mean egg density; and drift in spawn survey observations over time. The distributions for each of these factors should reflect uncertainty in their average affect over years, not capture inter-annual variation in them.

#### Proportion of total spawn surveyed

The proportion of the total spawn that is surveyed has a natural upper bound of 1, though its central tendency is not known. Reasons for not surveying herring spawns include non-detection (early or late in season or very deep and not observed) or lack of resources to conduct the survey. For the latter case the spawns will be reported, and this occurrence is rare. The proportion of the total spawn that is not-detected is likely higher in more remote locations and when spawning abundance is low. With limited information, we assume a uniform distribution on 0.9–1.0 for the average proportion of total spawn surveyed.

#### Egg loss prior to survey

This factor accounts for egg loss due to predation (seabirds, invertebrates, marine mammals) and translocation between the time of egg deposition and the spawn surveys. The amount of egg loss prior to spawn surveys is determined by the daily egg loss rate and the number of days between a spawn and the subsequent survey.

The herring egg loss literature, recently summarized by ([Hay et al., 2011](#), their Appendix 7), is used to estimate a distribution for daily egg loss rates. All studies conducted on the west coast of North America that estimated total egg loss over the incubation period (or daily egg loss rates) were considered for inclusion in the egg loss rate distribution. Those criteria substantially reduce the available literature (Table C-3). Egg loss estimates from the selected studies were standardized to instantaneous (daily) rates (Table C-4). A normal distribution for the daily egg loss rate, based on the mean and standard deviation of the selected estimates, is assumed.

The second component of the egg loss distribution is the average number of days between the spawn event and the surveys. Information to inform the distribution of this factor was available in the B.C. herring spawn survey database. Only dive survey records were selected, and numerous error checks imposed to remove erroneous data (Table C-4). For each spawn record, the number of days between the spawn event

and subsequent survey was estimated as the difference between the mid-spawn date and the mid-survey date. The mean time between a spawn deposition event and the subsequent survey ranges from 6.4 to 9.2 days across the stock assessment regions (Table C-6). A normal distribution for the average time between egg deposition and surveys is assumed, based on the mean and standard deviation of the mean values for the stock assessment regions (mean=7.7; standard deviation =1.13).

### Bias in mean egg density

The equation predicting egg density from dive survey observations was calculated from field studies conducted through much of the B.C. coast in the mid 1980s. These studies included diver observations of egg layers and percent cover by vegetation classes and subsequent laboratory egg counts of the observed quadrats. While the egg density prediction equation is unbiased, the unexplained residual error is large and the error in the mean egg densities predicted at the stock assessment region/year level were often greater than expected based on the assumption of unbiased iid observations. To allow for potential bias in predicted mean egg density at the stock assessment region level, we assume a normal distribution for this factor with mean 1 and standard deviation 0.2.

### Drift in dive survey observations

The studies to calibrate field observations of herring spawns to egg density estimates were primarily conducted during the mid 1980s by research divers. Since then, Fisheries Officers and subsequently research divers have conducted the coast wide herring spawn surveys. While there is considerable effort to ensure standardization of the surveys, it is possible that there has been drift in how observations are made. There is no direct information on how survey observations may have changed over time, however Hay (in press) suggests that if drift has occurred its direction is to observations that result in lower density estimates (i.e. there has been an increase in trace observations.) For now, we do not include this factor in calculating a prior distribution for the spawn survey  $q$ .

Table C-1 summarizes the factors affecting the  $q$  prior and their assumed distributions.

Table C-1: Factors affecting the  $q$  prior and their assumed distributions.

Factor affecting the $q$ prior	Distribution	Parameters of distribution
Proportion of total spawn surveyed ( $p_i$ )	Uniform	0.9-1.0
Egg loss prior to survey:		
Instantaneous daily egg loss rate ( $Z_i$ )	Normal	Mean 0.0642 Std. dev 0.0187
Days between spawn deposition and survey ( $d_i$ )	Normal	Mean 7.7 Std. dev. 1.13
Bias in mean egg density ( $b_i$ )	Normal	Mean 1 Std. dev. 0.2

### C.3.3 Simulating the dive survey spawn index $q$

Monte-Carlo simulations were conducted, randomly sampling from each factors distribution. The factors will operate independently so covariance structure does not need to be considered. For each of 10,000 replicates (i), a random draw was made for each factor to generate a point in the joint distribution for the  $q$  prior ( $\tilde{q}_i$ ):

$$\tilde{q}_i = p_i b_i \exp(-d_i Z_i)$$

For the  $i\text{SCA}_M$  herring stock assessments, a lognormal prior for the spawn index  $q$  is assumed (i.e.,  $\ln(q)$  is assumed normally distributed). The distribution of the simulated  $\tilde{q}_i$  is reasonably approximated by a lognormal distribution (Figure C.1). Means and standard deviations for the simulated and the natural log of the  $\tilde{q}_i$  are presented in Table C-2

Table C-2: Estimated means and standard deviations for the simulated q prior and natural log of the q prior.

	$\tilde{q}_i$	$\ln(\tilde{q}_i)$
Mean	0.587	-0.569
Std	0.155	0.274

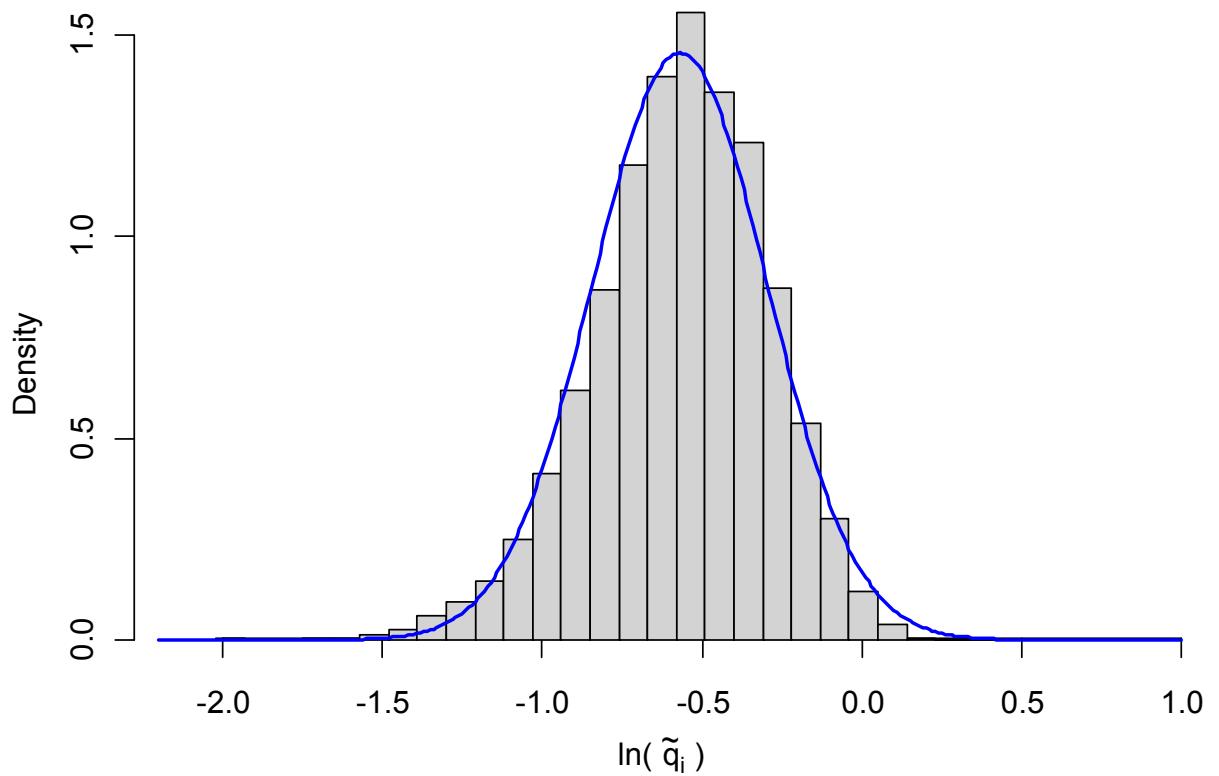


Figure C.1: Distribution of the log of the simulated spawn index q estimates, overlaid with a normal distribution based on the mean and standard deviation of the simulated values.

Table C-3: Summary of west coast North America herring egg loss literature.

Study and summary of pertinent egg loss estimates:	Rationale for inclusion/exclusion from prior estimation:
<a href="#">Bishop and Green (2001)</a> Estimated 31% of herring egg deposition was consumed by 5 species of birds (1994, Prince William Sound), based on a bioenergetics model.	Not included because study estimated only bird predation effect.
<a href="#">Haegele and Schweigert (1991)</a> Estimated 58% herring egg loss over 14 day incubation period (Lambert Channel 1989). Bird and invertebrate predation accounted for 7.1% egg loss; the remainder from physical removal and translocation could not be directly estimated.	Included because comprehensive B.C. study.
<a href="#">Haegele and Schweigert (1989)</a> Estimated 19.5% egg loss from predation based on predator counts and consumption rates (birds and invertebrates). From egg counts, total egg loss estimated at 68.8% over a 14 day incubation period (their Table 3 egg loss equations). Modelled changes in observations of egg layers over the incubation period: egg layers on sea grasses= 2.17 0.07 (day); egg layers on filamentous algae= 3.47 0.13 (day). Equations result in 45% and 52% decrease in egg layers over 14 days incubation period, respectively.	Included because comprehensive B.C. study.
<a href="#">Outram (1958)</a> Seabird predator exclusion study, West coast Vancouver Island (1951-1953). Overall, estimated 39% egg loss due to seabirds over the incubation period. Total egg loss over incubation period ranged from 56% to 99% (based on change in egg biomass for the control plots). Study was restricted to eelgrass beds.	Not included because study restricted to eel grass beds
<a href="#">Palsson (1984)</a> Estimated daily egg loss rates from 16.9% to 51.8% (positively correlated with egg density just after spawning). Large predators account for 20% to 50% of daily egg loss. Initial egg densities were very low, ranging from 400 to 80,000 eggs/m <sup>2</sup> across 9 study sites. Paulsons thesis cites additional egg loss literature that is not included here because the studies generally focussed on a limited range of habitat types.	Not included because study egg densities were much lower than densities generally seen in B.C. spawns.
<a href="#">Rooper et al. (1999)</a> Surveys conducted 1991, 1992, 1994 and 1995. Depth is factor that best accounts for egg loss rates (higher egg loss in shallower waters). Mean daily egg loss rates (Z) were (their Table 2): <ul style="list-style-type: none"><li>• 1990 0.076</li><li>• 1991 0.042</li><li>• 1994 0.096</li><li>• 1995 0.096</li></ul>	Comprehensive study in Prince William Sound. Estimates for 1990 and 1991 only are included because population had crashed by 1994 and abundance was low.
Note that population was much lower in 1994 and 1995.	

Table C-4: Estimates of the instantaneous daily egg loss rate ( $Z$ ) from herring egg loss studies conducted in the Pacific Northwest. The  $Z$  estimates for the Haegele and Schweigert (1989, 1991) studies were calculated from their reported egg loss rates over the study period.

Publication	Study Location	Year	$Z$
Haegele & Schweigert (1991)	SoG	1989	0.056
Haegele & Schweigert (1989)	WCVI	1988	0.083
Rooper et al. (1999)	PWS	1991	0.076
Rooper et al. (1999)	PWS	1992	0.042
		Mean	0.0642
		Std	0.0187

Table C-5: Criteria for selecting herring spawn survey data records for estimating days between spawning and surveys. The "number of records" is the records retained after each successive selection criterion.

Selection criterion	Number of Records
Total dive survey records (1985-2010)	3457
Spawn and survey start/end dates completed	3188
End spawn date $\leq$ start spawn date	
End spawn date - start spawn date $< 20$	
Survey days $\leq 14\ 3130$	3130
End spawn date-end survey date $\leq 2$	
End survey date- end spawn date $< 20$	3074

Table C-6: Average number of days between spawn deposition and spawn survey by stock assessment region and year.

Year	A2W	HG	PRD	CC	SoG	WCVI	A27
1985					5.2	6.6	8.8
1986			6.0	5.5	5.8	9.2	2.2
1987					10.6		
1988		7.4	11.2	8.3	11.0	8.2	
1989		8.2	10.8	5.8	14.5	8.6	10.7
1990	8.9	12.2	8.3	7.9	12.8	7.6	7.8
1991	10.5	5.9	7.5	11.0	12.4	9.2	2.3
1992	3.6	0.1	6.6	10.0	8.1	9.4	5.7
1993	12.2	9.3	4.9	9.1	13.1	7.6	16.8
1994	4.8	12.2	10.8	8.7	12.8	6.2	
1995		10.3	6.1	7.9	9.3	6.4	11.1
1996		7.9	5.7		10.2	4.5	7.9
1997		8.6	10.3	6.1	9.4	7.5	2.8
1998	10.5	13.3	5.5	12.4	10.4	6.4	4.2
1999		10.0	8.1	6.4	8.0	4.9	
2000	6.5	10.8	10.8	7.4	8.8	6.1	6.9
2001	9.0	7.4	8.3	6.8	8.0	6.9	3.8
2002	6.5	9.0	5.7	5.6	9.9	5.5	2.3
2003		7.5	6.2	5.0	9.5	7.0	7.3
2004	8.6	14.0	8.2	4.1	9.5	7.1	8.2
2005	6.3	9.4	8.8	9.2	5.8	5.1	3.3
2006			7.6	6.5	11.7	3.1	5.8
2007		6.0	10.8	0.7	9.3	3.4	
2008		7.4	7.2	4.3	9.5	3.8	2.0
2009	5.3	6.0	9.6	5.3	4.2	5.5	5.6
2010					7.1	5.2	
Mean	7.8	8.7	8.3	6.8	9.2	6.6	6.4

## **D.4 Landings and Survey data**

The following tables present the herring catch by gear and year for each Stock Assessment Region and the spawn survey data. Note that the units are in 1000's of tonnes.

Table D-1: Observed catch by gear type and year for each stock.

Stock	Year	HG		PRD			CC			SOG			WCVI		
		Gear 1	Gear 2	Gear 1	Gear 2	Gear 3	Gear 1	Gear 2	Gear 3	Gear 1	Gear 2	Gear 3	Gear 1	Gear 2	Gear 3
1951	2.860	46.000	42.700	33.300	0.769		44.200	46.100	8.420	21.600	27.200	0.020			
1952	10.200	53.200	1.870	27.400	25.000	11.600	66.300	68.600	72.100	58.900	20.900	34.300	6.110		
1953	1.780												16.900	2.620	
1954	1.230												0.556		
1955	18.000	10.100	43.000	23.800	9.870	27.500	48.900	48.900	46.000	67.000	67.000	69.200			
1956	72.600	26.600	4.550	10.300	18.500	4.000	31.000	15.700	43.500	65.900	68.500	52.100	52.100		
1957	23.700												25.500	23.700	
1958	10.900												18.400		
1959	6.970												21.100		
1960													16.100		
1961	0.654	41.200	27.300	39.400	28.900	44.100	17.100	8.030	21.700	15.800	37.000	34.000	10.900		
1962	7.680												15.000		
1963	15.000														
1964	28.200														
1965	34.900														
1966	2.750														
1967	0.213														
1968	0.080														
1969															
1970	1.330														
1971	0.102	3.450	0.878	3.620	0.004	0.388	8.360	0.137	2.430	5.900	0.456	6.940			
1972	0.849												16.900	1.540	
1973													12.200	3.930	
1974	0.127	0.182	2.120	1.530		3.610	5.320	0.479	0.425	3.110			8.180		
1975	0.105	0.155	1.540	0.011		3.330	5.410	0.378	0.469	5.310			22.800	16.100	
1976	0.374	1.800	0.566	3.470	0.277	6.100	6.130	5.070	0.202	7.020			0.029	17.400	
1977	1.490	0.797	5.820	1.520	0.320	3.850	6.790	5.690	4.110	7.860			12.800		
1978	2.560	3.600	1.960	3.090	0.046	4.760	9.280	13.000	3.730	7.450			5.140	15.200	
1979	2.090	1.810	1.270	1.240		0.005				13.600			0.084	10.400	
1980	1.210	0.740	1.640	1.060	0.010					2.470			1.680	2.300	
1981	1.710	1.690	1.050	0.357		0.269	2.300	4.890	2.080	5.100			4.980	3.070	
1982	1.410	1.820	0.170	0.041		2.250	4.050	3.930	3.930	3.300			3.530	2.370	
1983	0.067	0.929				0.041	2.170	1.170	0.432	0.162			0.177		
1984	0.096	0.535	0.173	1.650	1.890		2.710	0.924	0.244	3.110	5.870		0.203		
1985	1.490	0.253	3.010	3.470		0.038	3.550	0.978	0.756	1.470	5.960		13.400	2.450	
1986	0.890	0.375	3.730	4.540			6.510	2.950	1.030	1.410	5.880		8.250	1.450	
1987		0.122	2.080	4.090		0.028	5.350			0.233	7.690		9.710	3.520	
1988		0.079	3.560	4.430			6.270	3.050	0.270	1.130	9.130		7.860	1.960	
1989		0.071	3.660	4.820			7.130	1.820	0.562	1.120	0.939	3.610	6.300	2.340	
1990	1.170	0.043	2.280	2.370			8.120	2.030	0.617	4.380	8.730	8.920	3.090	0.627	
1991	0.543	1.360	2.140	3.780	0.088		7.120	1.120	0.942	5.120	5.120	5.120	5.610		
1992	0.142	1.240	3.780	4.100			8.340	2.030	0.617	4.380	8.730	8.730	5.310	0.706	
1993		2.210	4.100				2.360	2.350	9.610	2.110					
1994															

Table D-1: (continued)

Year	Gear 1	Gear 2	Gear 1	Gear 2	Gear 3	Gear 1	Gear 2	Gear 3	Gear 1	Gear 2	Gear 3	Gear 1	Gear 2	Gear 3
1995					0.706	1.370			8.150	1.450	0.641	4.360	8.030	1.950
1996					3.100	5.420			3.910	0.403	0.541	7.340	6.120	0.790
1997					3.170	5.420			3.290	0.345	0.402	9.300	6.110	6.670
1998					0.256	1.850			8.060	0.650	0.954	5.760	6.970	5.450
1999					1.240	3.100			6.000	1.530	1.470	4.970	6.900	3.410
2000					1.010	1.920			6.360	0.977	1.160	6.430	7.580	0.927
2001					2.060	2.470			5.620	0.519	1.420	7.280	7.590	0.702
2002					1.450	2.590			2.900	0.400	1.330	9.330	7.900	0.433
2003					1.920	2.210			2.300	0.289	1.700	10.700	8.100	2.570
2004					1.750	2.070			3.000		1.360	7.030	5.320	3.850
2005					0.958	1.670			3.790		1.330	7.910	9.150	0.594
2006					0.512	0.968			3.080		1.370	9.290	7.350	0.389
2007					0.713	1.280			0.398		0.672	3.850	5.240	0.950
2008					0.475	1.000					1.140	2.760	5.850	0.594
2009					0.884	1.260					0.709	5.700	3.860	3.420
2010											0.595	4.570	3.260	0.901
2011											0.713		4.530	

Table D-2: Abundance for each survey by year for each stock.

Year	HG		PRD		CC		SOG		Survey 2		Survey 1		Year		Survey 2		Survey 1		Year		Survey 2		Survey 1				
	Survey 1	Year	Survey 2	Year	Survey 1	Year	Survey 2	Year	Survey 1	Year	Survey 2	Year	Survey 1	Year	Survey 2	Year	Survey 1	Year	Survey 2	Year	Survey 1	Year	Survey 2	Year	Survey 1	Year	
1951	4.213	1988	15.245	1988	27.149	1988	35.444	1988	15.390	1988	26.916	1988	66.143	1988	24.976	1988	19.597	1988	19.51	1988	19.597	1988	19.51	1988	46.212	1988	
1952	2.578	1989	25.201	1989	24.047	1989	16.379	1989	10.295	1989	21.561	1989	1052	1989	72.376	1989	66.052	1989	13.310	1989	13.310	1989	13.310	1989	47.718	1989	
1953	7.355	1990	27.058	1990	28.468	1990	22.679	1990	1953	1990	28.980	1990	1952	1990	67.152	1990	1953	1990	39.571	1990	45.830	1990	1954	1990	46.464	1990	
1954	12.408	1991	17.998	1954	13.535	1991	25.811	1954	1954	1991	19.183	1954	13.967	1991	82.141	1991	45.830	1991	15.112	1991	15.112	1991	15.112	1991	30.456	1991	
1955	6.437	1992	12.376	1992	14.482	1992	40.145	1955	1955	1992	13.564	1955	43.274	1992	1955	1992	69.854	1992	25.667	1992	82.714	1992	1955	1992	42.687	1992	
1956	6.042	1993	8.152	1956	14.533	1993	19.93	1956	1956	1992	6.626	1956	32.392	1993	1956	1993	90.198	1993	1956	1993	27.183	1993	1956	1993	34.728	1993	
1957	1.392	1994	14.293	1957	27.518	1994	16.589	1957	1957	1994	4.607	1957	29.432	1994	1957	1994	67.144	1994	1958	1994	16.911	1994	1958	1994	25.625	1994	
1958	0.815	1995	4.701	1958	9.882	1995	18.516	1958	1958	1995	3.549	1958	22.348	1995	1958	1995	64.899	1995	1958	1995	18.986	1995	1958	1995	28.057	1995	
1959	8.981	1996	7.377	1959	40.961	1996	24.854	1959	1959	1996	3.904	1959	21.646	1996	1959	1996	47.864	1996	1959	1996	12.979	1996	1959	1996	33.986	1996	
1960	6.599	1997	11.215	1960	16.545	1997	25.037	1960	1960	1997	28.255	1960	12.615	1997	1960	1997	58.232	1997	1960	1997	6.015	1997	1960	1997	46.490	1997	
1961	8.981	1998	21.649	1961	12.059	1998	19.420	1961	1961	1998	4.265	1961	31.503	1998	1961	1998	44.326	1998	1961	1998	10.556	1998	1961	1998	41.556	1998	
1962	5.730	1999	10.610	1962	16.329	1999	29.745	1962	1962	1998	11.948	1962	31.813	1999	1962	1999	35.574	1999	1962	1999	85.095	1999	1962	1999	20.390	1999	
1963	7.297	2000	6.698	1963	16.981	2000	19.694	1963	1963	1998	6.485	1963	32.652	1998	1963	1998	72.688	1998	1963	1998	11.245	1998	1963	1998	13.267	1998	
1964	15.195	2001	1964	1964	26.919	2001	36.684	1964	1964	1998	6.464	1964	25.109	1998	1964	1998	100.248	1998	1964	1998	22.761	1998	1964	1998	13.955	1998	
1965	1.378	2002	3.257	1965	6.055	2002	22.449	1965	1965	1998	2.097	1965	200.2	1998	1965	1998	38.390	1998	1965	1998	117.864	1998	1965	1998	22.086	1998	
1966	2.824	2003	8.801	1966	7.105	2003	34.007	1966	1966	1998	1.863	1966	200.3	1998	1966	1998	7.211	1998	1966	1998	3.722	1998	1966	1998	29.750	1998	
1967	0.710	2004	5.668	1967	3.386	2004	30.493	1967	1967	1998	5.434	1967	29.407	1998	1967	1998	9.647	1998	1967	1998	14.167	1998	1967	1998	4.813	1998	
1968	0.883	2005	3.614	1968	5.197	2005	27.356	1968	1968	1998	5.790	1968	200.5	1998	1968	1998	9.442	1998	1968	1998	11.029	1998	1968	1998	9.075	1998	
1969	2.075	2006	4.097	1969	9.436	2006	15.251	1969	1969	1998	1.837	1969	12.051	1998	1969	1998	46.752	1998	1969	1998	10.465	1998	1969	1998	2.705	1998	
1970	5.052	2007	8.814	1970	8.480	2007	15.562	1970	1970	1998	8.230	1970	9.857	1998	1970	1998	35.865	1998	1970	1998	34.163	1998	1970	1998	2.089	1998	
1971	13.291	2008	4.213	1971	17.72	2009	8.935	1972	1972	1998	13.553	1972	4.156	2008	1971	1998	38.921	1998	1971	1998	32.103	1998	1971	1998	2.548	1998	
1972	9.542	2009	8.935	1972	8.774	2009	12.684	1972	1972	1998	3.572	1972	200.9	1998	1972	1998	25.139	1998	1972	1998	47.909	1998	1972	1998	9.876	1998	
1973	7.960	2010	6.091	1973	10.059	2010	26.988	1973	1973	1998	8.075	1973	12.434	2010	1973	1998	16.191	2010	1973	1998	47.480	1998	1973	1998	2.373	1998	
1974	14.510	2011	1974	1974	9.244	2011	8.852	1974	1974	1998	8.852	1974	8.037	2011	1974	1998	40.571	1998	1974	1998	25.540	1998	1974	1998	15.844	1998	
1975	9.086	2012	16.374	1975	10.949	2012	15.587	1975	1975	1998	13.849	1975	19.77	200.5	1998	1975	1998	19.75	1998	1975	1998	19.75	1998	49.149	1998		
1976	16.408	2013	18.371	1976	11.589	2013	18.247	1976	1976	1998	14.819	1976	14.819	200.5	1998	1976	1998	19.76	1998	1976	1998	64.222	1998	1976	1998	17.777	1998
1977	18.371	2014	17.779	1977	6.164	2014	17.747	1977	1977	1998	5.747	1977	17.747	1998	1977	1998	102.115	1998	1977	1998	19.778	1998	1977	1998	45.607	1998	
1978	13.649	2015	11.195	1978	11.937	2015	11.937	1978	1978	1998	13.012	1978	15.919	1998	1978	1998	85.991	1998	1978	1998	19.778	1998	1978	1998	66.397	1998	
1979	31.904	2016	14.087	1979	14.087	2016	14.087	1979	1979	1998	15.919	1979	16.919	1998	1979	1998	102.115	1998	1979	1998	19.779	1998	1979	1998	19.779	1998	
1980	20.294	2017	19.81	1980	17.186	2017	17.186	1980	1980	1998	16.333	1980	16.333	1998	1980	1998	100.987	1998	1980	1998	19.882	1998	1980	1998	33.047	1998	
1981	23.593	2018	19.82	1981	19.82	2018	19.82	1981	1981	1998	18.482	1981	18.482	1998	1981	1998	64.575	1998	1981	1998	16.771	1998	1981	1998	19.882	1998	
1982	21.391	2019	19.83	1982	19.83	2019	19.83	1982	1982	1998	14.185	1982	14.185	1998	1982	1998	19.84	1998	1982	1998	16.771	1998	1982	1998	23.872	1998	
1983	23.439	2020	19.84	1983	19.84	2019	19.84	1983	1983	1998	8.850	1983	8.850	1998	1983	1998	19.85	1998	1983	1998	16.771	1998	1983	1998	19.882	1998	
1984	18.925	2021	18.855	1984	18.855	2019	18.855	1984	1984	1998	20.342	1984	20.342	1998	1984	1998	41.575	1998	1984	1998	19.86	1998	1984	1998	30.010	1998	
1985	18.925	2022	18.867	1985	18.867	2019	18.867	1985	1985	1998	12.827	1985	12.827	1998	1985	1998	41.737	1998	1985	1998	19.87	1998	1985	1998	16.855	1998	