

Manuscript Draft

What is long-distance dispersal? And a taxonomy
of dispersal events

Supplementary Material

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Methods

Species and Study Site. The study species, *Prunus mahaleb* (L.) (Rosaceae), is a shrub or small tree that produces fleshy fruits that are consumed by frugivores, who disperse seeds after regurgitating or defecating them. This species is frequently visited during July to mid-August by small- and medium-sized birds and carnivorous mammals that include fruits in their diets during late summer to winter (Jordano & Schupp, 2000). *P. mahaleb* occurs in a patchy distribution at the regional scale, with relatively isolated populations consisting of dozens to hundreds of trees. Our study population included a total of 196 adult reproductive trees distributed over an area of 26 ha in patches of variable density. Other populations within 20 km exist as scattered patches of 10–150 trees, with some containing $\geq 1,000$ trees. The nearest population is 1.5 km away. Additional information on the study population is reported by Jordano *et al.* (2007) and Garcia *et al.* (2009).

Sampling dispersed seeds. To estimate the relative contribution of each dispersal vector to the different categories of dispersal events defined in Table 1, we first collected dispersed seeds, following different sampling schemes according to the functional group of dispersal vector. We used this grouping of frugivores giving the difficulties of resolving the identification of scats, pellets and regurgitated seeds down to species level just based on visual cues. We differentiated four major frugivore types: large carnivorous mammals (such as foxes, badgers, and stone martens); two species of medium- and large-sized frugivorous birds, mistle thrushes (*T. viscivorus*), and carrion crows (*C. corone*); and a pool of small-sized frugivorous birds, including warblers, redstarts, and robins (Jordano *et al.*, 2007).

Seeds were collected in 1997–1999 and 2003–2005. The sampling schemes are described in detail elsewhere (Jordano *et al.*, 2007; Garcia *et al.*, 2009) and include a combination of seed traps and direct sampling of mammal feces along fixed transects. We haphazardly collected 130 samples of mammal feces during the *P. mahaleb* fruit ripening period and recorded their location relative to potential source trees. Overall, we genotyped 167 seeds from 20 fecal samples. Most samples

31 were from red fox (*Vulpes vulpes*) and stone marten (*Martes foina*); some (10
32 samples) were from badger (*Meles meles*) (Jordano *et al.*, 2007).

33 In addition we sampled directly the pellets of large corvids (*Corvus corone*) and
34 from *Turdus viscivorus*, the latter by direct sampling beneath pine trees and scats
35 from seed traps (see Jordano *et al.*, 2007, for details). Finally, a seed sample di-
36 rectly from seed traps included seeds dispersed by small- and medium-sized passer-
37 ines, such as *Phoenicurus ochruros*, *Turdus merula*, *Erithacus rubecula*, *Sylvia com-*
38 *munis*, *Sylvia atricapilla*, etc. (Jordano *et al.*, 2007). The total seed sample thus
39 consisted of seed endocarps collected from the seed traps (mostly small passerines)
40 ($n=465$), mammal scats ($n=167$), and *C. corone* pellets ($n=23$).

41 **Seed genotyping.** We used material described in Jordano *et al.* (2007), and
42 genotyping methods described in detail in previous work (Godoy & Jordano, 2001;
43 Garcia *et al.*, 2007, 2009). Briefly, we used a set of 10 polymorphic microsatellite
44 markers (simple DNA sequence repeats) (Godoy & Jordano, 2001) to obtain the
45 multilocus genotypes of both of the adult trees (candidate source trees from the
46 study population) and the sample of seed endocarps. Given that all adult trees
47 in the population had a distinct multilocus genotype, an unambiguous assignment
48 of each seed to its source tree could be made. When a full match between the
49 endocarp genotype and any of the adult-tree genotypes in the population was not
50 possible, we assumed that the seed came from another population. To assess the
51 effect of genotyping errors, we reexamined the exclusion of genotypes due to a single
52 locus mismatch, two loci mismatches, etc. At the analysis level, any exclusion of
53 identity between a seed and a potential mother tree based on mismatches of only
54 one or two loci was rechecked. We used GIMLET software (Valière, 2002) to find
55 the matching adult multilocus genotype for each endocarp with eight or more
56 loci successfully typed. Because each seed belonged to one of the four groups of
57 dispersers, we could thus derive the relative contribution of each frugivore group
58 to different classes of seed dispersal events and to seed immigration.

59 **Contribution of dispersal vectors to types of dispersal events.** We con-
60 sidered each dispersed seed as an independent replicate, because each represented

61 a dispersal event from the perspective of plant population genetics, i.e., an in-
62 dependent "arrival" event resulting from the dispersal process mediated by the
63 frugivore.

64 Once the maternal source tree of each individual seed was identified (or its prove-
65 nance from outside the study population determined) we assessed the dispersal
66 distance and grouped the seeds separately as coming from trees located within or
67 outside the population. In addition, for seeds originating from local trees we de-
68 termined whether dispersal distances were ≥ 45 m to sort out LDD_{loc} dispersal
69 events from SDD_{loc} events. All the events involving immigrant seeds were con-
70 sidered LDD_{ss} by definition, given that the neighborhood size was very reduced
71 (radius = 0.045 km) relative to the geographic limits of the study population (max-
72 imum length for a within-population dispersal event: 1220 m) (Garcia *et al.*, 2009).
73 Along this reasoning, LDD_{neigh} events were considered non-existent in this partic-
74 ular case study given that neighborhood size area was smaller than the population
75 area.

Table S1. Summary of neighborhood area sizes and estimated neighborhood radius for tree species with different combinations of dispersal modes. Data from Nason *et al.* (1998); Garcia *et al.* (2005, 2007) and present study.

Species	Pollinator	Seed disperser	Density (ha^{-1})	Breeding unit (km^2)	Radius (km)
<i>Ficus dugandii</i>	Fig wasp	Vertebrates	0.004	631.7	14.2
<i>Ficus obtusifolia</i>	Fig wasp	Vertebrates	0.072	105.9	5.8
<i>Prunus mahaleb</i>	Bees, flies	Vertebrates	0.003	0.87	0.042
<i>Frangula alnus</i>	Bees, flies	Vertebrates	0.0004	0.45	0.013
<i>Astrocaryum mexicanum</i>	Beetle	Vertebrates	1364.0	0.011	0.06
<i>Calophyllum longifolium</i>	Bees	Vertebrates	0.28	1.241	0.629
<i>Platypodium elegans</i>	Bees	Wind	0.78	0.866	0.525
<i>Cedrus atlantica</i>	Wind	Wind	61.7	0.151	0.22
<i>Fraxinus americana</i>	Wind	Wind	24.7	0.008	0.05
<i>Pseudotsuga menziesii</i>	Wind	Wind	25.0	0.078	0.158

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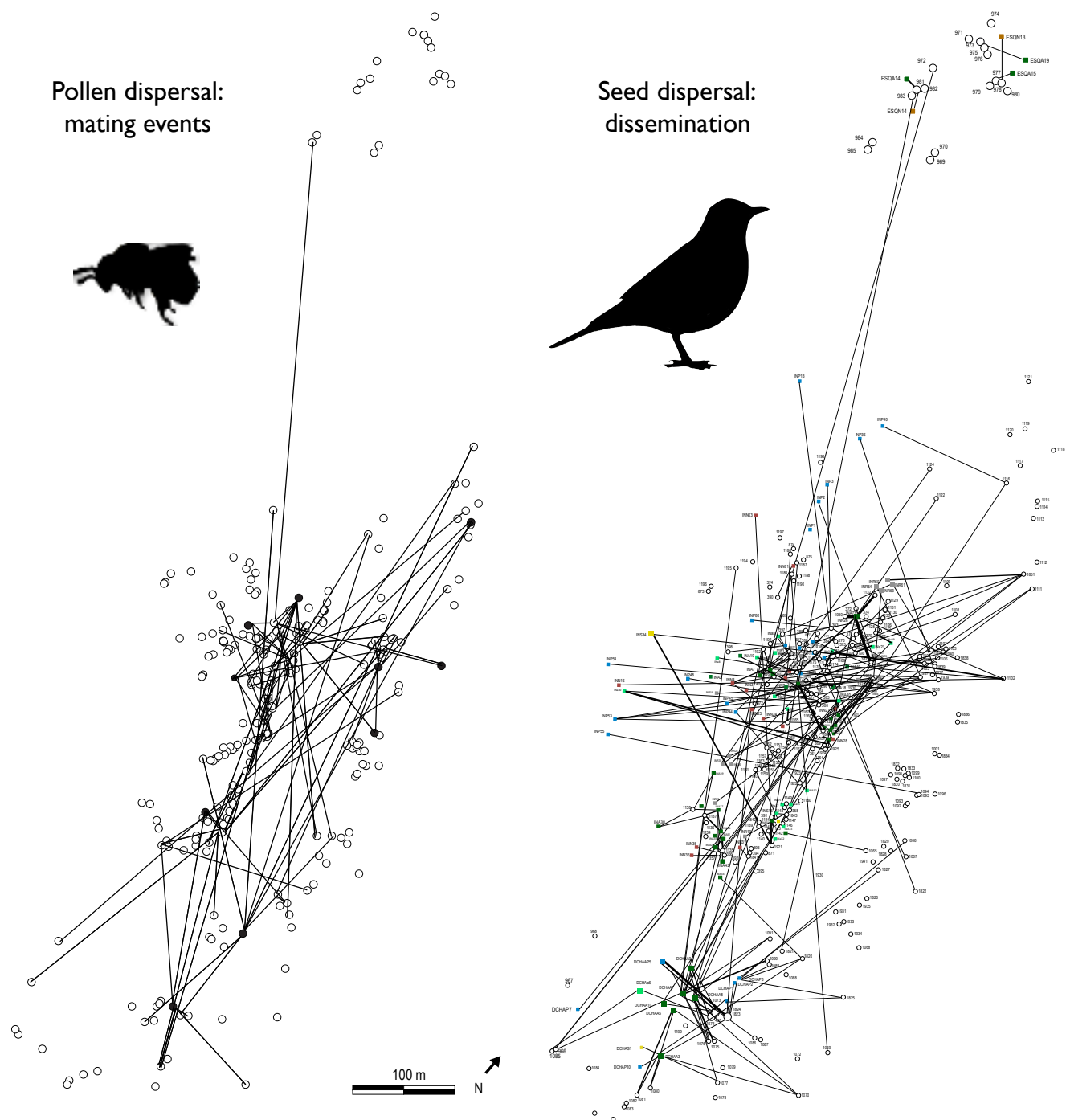


Figure S1. Dispersal events for pollen (left) and seeds (right) traced for *Prunus mahaleb* trees (white dots). All the adult, reproductive, trees in the population are mapped. Lines indicate mating events of pollen dispersal among trees (left) or seed dissemination events from source fruiting trees to seed traps (squares; right). Line thickness is proportional to the number of events recorded.

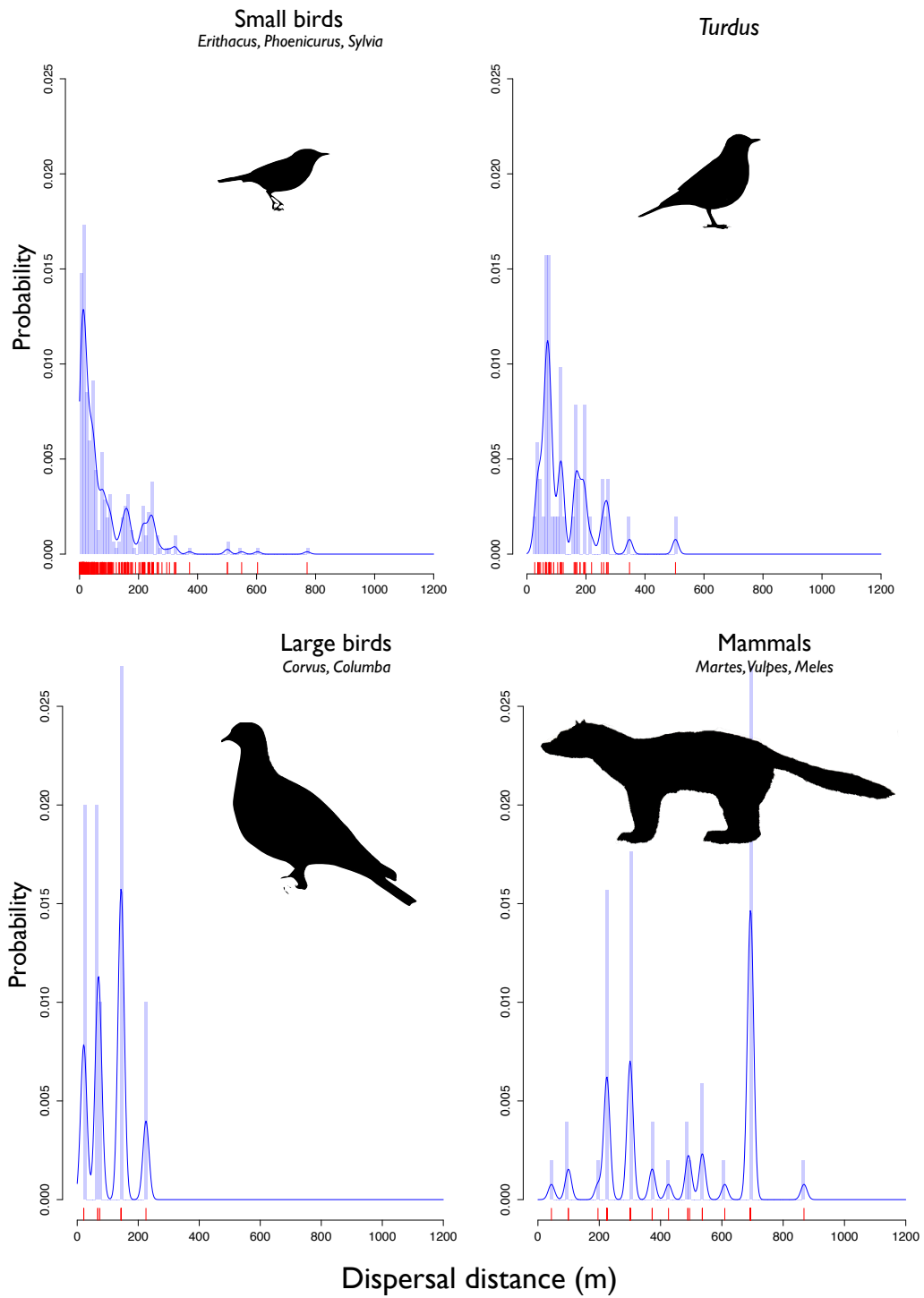


Figure S2. Differential contributions of functional groups of frugivores to the short- (SDD_{loc}) and long-distance (LDD_{loc}) local seed dispersal events for *Prunus mahaleb*.