

Can constraint closure provide a generalized understanding of community dynamics in ecosystems?

Steven L. Peck^{1*}, Andrew Heiss²

1 Biology Dept, Brigham Young University, Provo, UT, USA

2 Department of Public Management and Policy/Andrew Young School of Policy Studies,
Georgia State University, Atlanta, Georgia, USA

* Corresponding author: steven_peck@byu.edu

1 **Abstract**
2

3 Ecological theorists have generated several yet unresolved disputes that try to untangle the difficulty
4 in understanding the nature of complex ecological communities. In this paper, we combine two
5 recent theoretical approaches that used together suggest a promising way to consider how
6 evolutionary and ecological processes may be used to frame a general theory of community ecology
7 and its functional stability. First, we consider the theoretical proposal by Mark Vellend (2016) to
8 focus on a small set of higher-level evolutionary and ecological processes that act on species within
9 an ecological community. These processes provide a basis for ecological theory similar to the way in
10 which theoretical population genetics has focused on a small set of mathematical descriptions to
11 undergird its theory. Second, we explore ideas that might be applied to ecosystem functioning
12 developed by Alvaro Moreno and Matteo Mossio's (2015) work on how biologically autonomous
13 systems emerge from closure of relevant constraints. To explore the possibility that combining these
14 two ideas may provide a more general theoretical understanding of ecological communities, we have
15 developed a stochastic, agent-based model, with agents representing species, that explores the
16 potential of using evolutionary and ecological processes as a constraint on the flow of species
17 through an ecosystem. We explore how these ideas help illuminate aspects of stability found in many
18 ecological communities. These agent-based modeling results provide in-principle arguments that
19 suggest that constraint closure, using evolutionary and ecological processes, explain general features
20 of ecological communities. In particular, we find that our model suggests a perspective useful in
21 explaining repeated patterns of stability in ecological evenness, species turnover, species richness,
22 and in measures of fitness.
23

1 **Can constraint closure provide a generalized understanding of**
2 **community dynamics in ecosystems?**

3 **Introduction**

4 Ecological theorists have generated several yet unresolved disputes that try to untangle
5 the difficulty in understanding the nature of complex ecological communities, including the
6 question of whether the idea of a community in such systems is coherent. For example, questions
7 about ecological community structure and how it is maintained have been argued about for over
8 a hundred years. These questions are still taking their current principal divisions from camps first
9 staked out by Clements (1916) and Gleason (1926) in the early decades of the 20th century: Are
10 communities assembled non-randomly, conditioned on certain possibilities of co-occurrence and
11 structured by general ecological regularities? Or, alternatively, are they more arbitrarily
12 constructed, brought together by a complex combination of abiotic factors, available species
13 pools, competition, and other ecological processes that suggest a strong element of contingency
14 (Eliot 2007, Götzenberger, et al. 2012)?

15 Currently, there does not seem to be a unified body of theory and practice that explains
16 most features of community structure. The hunt for generality has also been fraught (See Elliott-
17 Graves (2018) for a detailed overview). Theories in the discipline often seem a piecemeal
18 confederation of ideas that carve nature mid-bone, as it were. Often focused on specific natural
19 systems that provide only limited empirical and theoretical support from other, differently
20 structured, ecological communities. It is clear that the theoretical constructions used by
21 ecologists have been instrumental in the development of research programs and have allowed
22 progress in understanding community ecology. For example, Hubbell's neutral theory has been
23 able to give a mechanistic account to assist in understanding statistical regularities in forest tree
24 distributions (Abeyasinghe and Punchi-Manage 2020, Alonso, et al. 2006, Hubbell 2001).

25 However, even basic concepts in the discipline are often viewed differently by different
26 research groups that are often divided into opposing camps. There are even debates about the
27 usefulness of the concept of community ecology. For example, the following questions are still
28 being debated. Ontologically, could ecological communities be considered individuals because
29 they are often maintained through time in the same coherent way that organisms are (Bryant
30 2012, Collier and Cumming 2011, Huneman 2014, Marquet, et al. 2014, Odenbaugh 2010,
31 Odenbaugh 2011, Roughgarden 2009, Sterelny 2006, Sutherland, et al. 2013, Vellend 2016)?
32 Which ecological theories best describe the dynamics of community structure, maintenance,
33 resilience, and development (Chesson 2008, Kirchhoff 2010, Marquet, et al. 2014, Odenbaugh
34 2011, Roughgarden 2009, Sutherland, et al. 2013, Vellend 2016)? Do ecological communities
35 have boundaries (Bryant 2012, Haber and Odenbaugh 2009, Peck 2009). Are they stable; what
36 allows the coexistence of their constituent species (Chesson 2018, Hubbell 2001, Jeltsch 2019,
37 Odenbaugh 2011)? How are communities linked as metapopulations and metacommunities
38 (Chase 2005, Hanski and others 1999, Hanski, et al. 2017, Leibold, et al. 2004)? How are
39 communities assembled? And are they so contingent that each community must be examined as
40 an individual case study (Chase 2003, Rosindell, et al. 2011, Shrader-Frechette and McCoy
41 1994, Simberloff 2004, Vellend, et al. 2014, Weiher, et al. 2011, Wennekes, et al. 2012)? Are
42 ecosystems real, constituent parts of nature, or are they merely constructed as an organizing
43 principle to allow scientific study (Odenbaugh 2007, Sterelny 2001)?

46 These questions are all interrelated, yet the abundant theories used to tame and
47 understand these ecological systems rarely provide a unified or coherent picture among theories.
48 Moreover, in order to make progress in the idea of an ecological community, it seems essential to
49 discover generalities that hold across the numerous possible biotic configurations that comprise
50 these individual community assemblages. It is also important to view this project from multiple
51 perspectives in order to tackle the inherent complexity and provide a more coherent picture
52 among theories to create a more integrated conception of community ecology. There have been
53 many attempts to do so, and the importance of these efforts recognized (de Mazancourt, et al.
54 2013, Evans, et al. 2013, Heleno, et al. 2014, Linquist 2015, Linquist, et al. 2016, Loreau and de
55 Mazancourt 2013, Odenbaugh 2011, Roughgarden 2009, Simberloff 2004, Vellend 2016),
56 though conditioned on some skepticism about the possibility of finding well-supported law-like
57 regularities in community ecology (Lawton 1999, Lockwood 2008, Shrader-Frechette and
58 McCoy 1994, Woodward 2001).

59 In this paper, we combine two recent approaches that used together suggest a promising
60 way to consider how communities are structured, and which may offer the potential for a more
61 generalized theory. In the first approach, we consider a proposal by Vellend (2016) to focus on a
62 small set of high-level processes that provide a basis for ecosystem theory, similar to the way
63 that theoretical population genetics has been fashioned to focus on a small set of processes with
64 relatively simple mathematical formulations. Vellend proposed that we look at ecological
65 communities in light of four evolutionary and ecological processes found in a wide variety of
66 ecosystems: species selection, species-drift, dispersal, and speciation (Vellend 2010, Vellend
67 2016). These community-level processes focusing on species are thought to capture the means
68 whereby most ecosystems are structured.

69 In the second approach to consider how communities are structured, Moreno and Mossio
70 (2015) show how constraints on the flow of matter and energy enable the development of
71 autonomous, self-sustaining, stable structures in chemical and cellular systems. We use their
72 concept to explore the idea that constraints may similarly configure large-scale processes, like
73 the flow of species through an ecological community, to create semi-autonomous, self-
74 sustaining, stable structures in ecosystems (for other attempts to argue this approach, see,
75 (Dussault and Bouchard 2016, Huneman 2014, Montévil and Mossio 2015, Moreno and Mossio
76 2015, Nunes-Neto, et al. 2014)).

77 By combining these two approaches and perspectives, we argue with support from agent-
78 based simulation models, that this might be a productive way to explore general features of
79 stability in ecological communities. We also use the idea that a useful way to think about
80 ecological communities is using trait-based assemblages that focus on trait functional diversity
81 rather than individual species as components of communities (Fontana, et al. 2016, Petchey and
82 Gaston 2002, Zakharova 2019). For example, beak size in Galápagos finches reach a certain size
83 and shape among the different islands, presumably because they use food resources that require
84 particular beak traits. Rather than thinking about species as taxonomic units, species are
85 represented as functional groups based on specific traits. Trait-based functional groups were used
86 by Carturan et al. (2020) to model community dynamics of coral reefs using an agent-based
87 model. We will use a similar approach in our model.

88 Trait-based approaches to species groups has especially useful in creating morpho-
89 species in biological samples taken from a large number of organisms in which species'
90 identifications are difficult, such as in satellite images of forest (Schneider, et al. 2017), in
91 logically hard to sort through samples like dragnet catches of phytoplankton (Salmaso and

92 Padisák 2007, Weithoff and Beisner 2019), other plants (Armbruster, et al. 2002) and animals
93 (Derraik, et al. 2002), and in computer simulations (Marks and Lechowicz 2006). Morpho-
94 species based on functional groups will be used in work described below.

95

96 *Constraint Closure*

97

98 One way to approach the general idea of an ecological community can be drawn from
99 Moreno and Mossio's (2015) work on biologically autonomous or semi-autonomous systems.
100 This paper proposes their work, largely based on cellular and chemically reproducing systems,
101 may be relevant for ecological communities. They identify three dimensions of general
102 biological autonomy: regulation, stability, and historical contingency. Because ecosystems are
103 more permeable in terms of their structure and boundaries, it may be more appropriate to
104 consider them as semi-autonomous, which allows for less well-defined and fuzzy boundaries.

105 The first dimension of biological autonomy, or semi-autonomy (i.e., regulation), of an
106 ecological community, is structured by three things: (a) its abiotic context; (b) the ecological
107 processes that are characteristic of a particular ecosystem, including the boundaries that contain
108 or restrict the flow of energy within the ecological system (e.g., conditions of a lake ecosystem
109 are bound by its shore and inputs through streams), and (c) the processes that allow for the
110 particular conditions to be maintained through internal cycles of biotic components within the
111 system through time (e.g., predator-prey cycles, or the self-maintaining light regime of a tree
112 species growing in a forest).

113 According to Moreno and Mossio's (2015) work, constraint closure is necessary for
114 autonomy or semi-autonomy to emerge. Constraint closure is described as constraints on the
115 flow of matter or energy that form a closed loop and thereby allow for the development of
116 stability. This is a necessary step in the formation of all autonomous systems. Such a
117 configuration of constraints creates a sustainable feedback loop that allows the system to achieve
118 coherence. Constraints emerge from the system processes themselves (e.g., through
119 developmental processes, autocatalysis, or in cellular processes). Fig. 1 illustrates a minimal set
120 of requirements for the occurrence of constraint closure and illustrates how such components
121 modulate the flow of energy and materials within the system; redrawn from Figure 1.3, p. 21,
122 (Moreno and Mossio 2015). Moreno and Mossio (2015, 38) point out that regulation is just a step
123 on the way to autonomy so the ideas of autonomy and regulation have an important distinction.

124 The second dimension of biological autonomy, or semi-autonomy (i.e., stability), is found
125 in interactions conditioned on the larger environment that provides the energy and materials used
126 by the system. In an ecological community, these interactions can take the form of energy flows
127 from sunlight or the movement of organisms in and out of the ecosystem. Such flows might also
128 include forces of disturbance that might restructure the system through things like direct human
129 influence, invasive species, diseases, or abiotic influences such as fire, weather-related events,
130 and other non-biological influences.

131 The third dimension of biological autonomy, or semi-autonomy (i.e., historical
132 contingency), includes the system-specific historical and often contingent aspects of the system.
133 For example, in an ecosystem, it might include the past arrival of new species, ecological
134 processes that have emerged due to the interactions of local configurations of constitutive
135 species, evolutionary dynamics, contingent factors, co-evolutionary adaptations, and other biotic
136 and abiotic influences (Spencer 2020).

137 Given the above, and, as Moreno and Mossio (2015, 105-108) speculate, there are
138 reasons to think that ecological communities might be autonomous or semi-autonomous systems.
139 The energy flow is maintained and channeled within the system by its components. For example,
140 long-term community stability has been shown to be structured by a number of well-studied
141 ecosystem processes, such as the competitive exclusion of species restructuring particular niches,
142 networks of mutualistic relations, and typical ecological interactions like predator-prey relations.

143 Nunes-Neto et al. (2014) make an argument that ecological communities can be seen as
144 autonomous systems because of the functional components of traits that structure most
145 ecosystems. In particular, they look at the function of the components of biodiversity using what
146 is termed the 'organizational approach' (OA) that they use to argue for biological autonomy and
147 stability through constraint closure. Dussault and Bouchard (2016) add a significant
148 consideration to the above account, developing a focus on particular kinds of functions that they
149 call persistence enhancing propensities (PEP). PEPs are a class of functions that allow the
150 currently structured ecosystem to persist stably. The technical details of OA and PEP functions
151 are given in Appendix 3.

152

153 *Vellend's Set of Processes as Constraints*

154

155 By combining the above ideas articulated by Moreno and Mossio about autonomy with
156 Vellend's proposal to focus on four high-level processes structuring species dynamics within a
157 community, we can see a possible theoretical space open by combining the two views to consider
158 the ways certain constraints may stabilize ecological communities. It might reasonably be argued
159 that the processes Vellend identifies, species selection, species-drift, dispersal, and speciation,
160 constrain and structure the 'flow of species' through an ecosystem. Each of the four processes
161 that Vellend identified restrict, modify, or enhance the flow of species, and thus act as constraints
162 on the system. Because the constraints that Moreno and Mossio discuss are not limited to the
163 kinds of structures typically discussed in the flow of matter and energy in chemical or cellular
164 systems, it is not unreasonable to consider the potential of constraint closure to organize
165 ecological communities in the way they discuss for chemical systems. It might be objected that
166 this type of analogy in which evolutionary and ecological processes affect species flow through a
167 community might not hold in the same way that systems of cellular or chemical processes
168 maintain stability through constraint closure do. These basic physical systems have more well-
169 understood mechanisms that constrain the flow of energy in material systems that might not hold
170 in ecological systems. Therefore, stretching the notion of constraint closure to something as
171 complex and multivalent as evolutionary and ecological processes that act on individual species
172 and their flow should rightfully be approached with caution. However, we believe that the results
173 described below provide warrant.

174 Consider that in Vellend's four processes, selection constrains the success of a species by
175 affecting its ability to compete in its current abiotic and biotic situation and providing fitness
176 conditions for the species' success or failure in surviving in a particular environment. Species-
177 drift constrains the population genetic structure and what variation is available to a community's
178 individual species, and how stable the presence of a species is in a bounded region, including
179 how it drifts in adjusting or modifying its current niche conditions. Speciation, in addition to the
180 effects of selection, constrains how many new species emerge in an area due to genetic factors,
181 and the interplay between trait differences and the biological fitness of those differences.
182 Likewise, dispersal constrains species flow by influencing a combination of a species' ability to

183 spread beyond its original location, for instance, in the sense that geneticist Sewall Wright
184 proposed in his three-step shifting balance theory, that is, (1) initial establishment, (2) population
185 growth, and (3) species dispersal of individuals to new areas (Pigliucci 2008). This suggests a
186 hypothesis that these processes constrain ecosystem function.

187 Constraints on species flow through the system may set up such communities as
188 autonomous or semi-autonomous systems through constraint closure analogous to the way
189 Moreno and Mossio (2015) envisioned about how autonomous systems are maintained for a
190 period of time. In particular, species and the ecological functional groups they define flow
191 through a community frame as a set of persistence enhancing propensities that influence stability
192 and other aspects of community dynamics. There may be other possible constraints on species
193 flow. For example, things like boundaries over which species cannot cross may affect which
194 species enter a particular community. The spatial patch dynamics of particular landscapes, such
195 as the metapopulation structure, creates a constraint on species flow. Other processes like the
196 frequency of disturbance and mutualistic relationships among species within an ecosystem,
197 which affect the survival probabilities of cooperating species, may also play a role. In particular,
198 mutualist relationships are a major influence on ecosystem dynamics and stability, for example,
199 the mutualistic relationships in mycorrhizal networks (Booth 2004, Heijden, et al. 2015, Mnasri,
200 et al. 2017, Peay 2016, Simard, et al. 2012, Yang, et al. 2014).

202 *A Model of Constraint Closure in Autonomous Ecological Communities*

203
204 To explore the possibility of using constraint closure to understand ecological
205 communities, we developed an agent-based model that explores the potential of understanding
206 the semi-autonomy and stability of ecological communities as an effect of constraints on the flow
207 of species through the ecosystem. We provide in principle arguments that constraint closure may
208 be a way to structure general insights into measures stability in some ecological communities.

210 **Brief Model Summary**

211
212 A complete, detailed model description that follows the ODD (Overview, Design concepts,
213 Details) protocol for describing individual- and agent-based models (Grimm, et al. 2006), as
214 updated by (Grimm, et al. 2020), is provided in the supplementary material in Appendix 1. The
215 reader is encouraged to consult this section to understand the model more fully.

216
217 Briefly, the overall *purpose* of this model is to understand and provide in principle arguments
218 about how evolutionary and ecological dynamics may be viewed as constraints that confer
219 autonomy and stability through constraint closure in ecological communities. Specifically, we
220 address the *questions*: Can combining the insights of Vellend's (2016) work on species
221 evolutionary and ecological processes in community ecology and Moreno and Mossio's (2015)
222 work on constraint closure be used to shed light on common characteristics of ecological
223 community stability and autonomy? The model is designed to provide theoretical insights into
224 ecosystem stability and provide motivation for ecologists to consider examining their particular
225 systems for the kinds of constraints proposed here.

226 To consider our model adequate to its purpose, we hypothesize specifically that *patterns*
227 of standard measures of community dynamics, such as species richness, evenness, and aspects of
228 species fitness, will stabilize in response to evolutionary and ecological dynamics acting as

closed loops of constraints. The fidelity criteria we aim for will be achieved if the model shows that such patterns of stability *emerge* in ways that suggest the included dynamics act as constraints in the way that Moreno and Mossio (2015) propose, that is, that these evolutionary processes are sufficient to confer known *emergent* attributes of autonomy or semi-autonomy, and stability on the ecological community. We make no claim that this model will represent any specific target system but that it will provide insight into general ecological systems. If the results of this model hold up in real-world ecological systems, it will generate testable hypotheses about how particular ecosystems function.

The model includes two kinds of entities, trait-based agents representing functional groups of morpho-species (referred to as ‘species’ in the following) and cells representing niches in “niche space”, not physical space (Fig. 2). Each cell represents, quantified by a single number (dubbed niche score) the set of possible trait-based functional groups or ‘species slots’ that could occupy this niche. This niche is thus not a classical one-dimensional niche but represents a more complex set of environmental configurations that are suitable for certain a suite of species. Each species found in a cell represents a single trait-based functional group that finds the niche (cell) habitable. The number of species that a cell can hold is limited, implying that the cell niche can only hold a given number of species.

Movement of species in niche space is restricted to the eight neighbor cells. Species differ in niche preference, and each species has a preferred niche, also characterized by a single measure called the niche-desirability score. The extinction probability of a species is defined as the absolute difference between the cell’s niche-score and the species niche-desirability score. The state variables of the cells and species are found in Table 1 and Table 2, respectively.

Evolutionary and ecological constraints: Table 3 gives a list and description of the proposed evolutionary and ecological processes thought to constrain the flow of species in an ecosystem and thereby produce ecological measures of stability. Constraints include the processes proposed by Vellend as well as adding two others: mutualism and ecological perturbation. The most important *constraints* of the model, which are repeated at each time step, and executed in the following order: establishing mutualistic networks of species, spatial dispersal to adjacent cells, competition among species and concomitant species extinction in the cell, species-drift, and lastly, the possibility of ecological collapse and perturbation (i.e., species loss irrespective of extinction probability).

Because the theory of constraint closure developed by Moreno and Mossio (2015) about autonomy and stability suggests that at least three constraints are necessary to create closure in the relevant sense, we examine in our analyses the number of constraints. We also examine the influence of each of the processes to assess how each of the constraints we included affects the outcome patterns of the indices of stability emerging in the simulated community. We have no a priori reason for thinking that one of the given constraints might contribute to ecological measures of stability more than another, so we first examine measures of stability against the number of constraints in our analysis.

We also analyze how each of the constraints contributes to the ecological stability measures using statistical random forest analyses to determine the evolutionary and ecological importance of each of the constraints (note: random forests are a statistical technique and do *not* refer to an ecotype consisting of trees). For each of the constraints, random forests were used on the individual outcomes as dependent variables. The constraints as independent variables provide information about a variable’s importance. A good description of the use and evaluation of

random forest analyses can be found in (Strobl, et al. 2007), but briefly, a constraint's importance was determined by calculating the mean standard error (MSE) for each statistical model of the data generated by the agent-based model outcomes. This is accomplished by randomly permuting one of the constraint outcomes and then calculating the percentage change from the unpermuted model (for example, comparing the model's MSE before and after shuffling the values for the variable *competition*). High values of importance indicate that a given constraint is necessary for maintaining model adequacy, while low values indicate that a constraint does not influence model error.

The basic *design concepts* in the model include evolution driving the *adaptation* of species to be more closely aligned with the niches represented within the cell they reside; the collective networks of species forming mutualistic relationships; and the emergence of stability through the process of evolutionary and ecological dynamics thought of as constraints imposing stability through constraint closure. The time step of the model is the average time to species extinction. See Fig. 3 for an illustration of the process flow in the model. Table 4 contains the parameters for the model.

Also important are the *stochastic* elements framed in the following ways: (a) the initial cell structure of the landscape is randomly generated; (b) The proportion of species chosen to form links is a random variable; (c) the species that become extinct in the drift stage are randomly chosen with probability proportional to a species' probability of extinction; drift is defined in terms of the species particular functional group in the cell (d) species with a low probability of extinction are chosen to move randomly to one of their 8-Moore-neighborhood neighbors; (e) The variation in the probability of extinction in newly speciated species is a random variable, and (f) the stochastic perturbation events of species going extinct due to dynamics other than their inherent probability of extinction.

Observations are collected from a Latin hyperspace exploration of the parameters described in Table 4 by taking random samples from the range of their possible parameter values in 2000 replications. The model was run for all 64 possible combinations of constraints, with each constraint being turned on or off for a complete characterization of the state space of the model with regard to the offered evolutionary and ecological constraints considered. The number of constraints included varies as these different permutations on the state space of the model are examined. Because we are looking for constraint closure, which requires at least three constraints, the number of constraints used is relevant for demonstrating the theory outlined above. Each run consisted of 500-time steps for each landscape produced in the combinations listed above.

The measures of stability examined in the model are species evenness, turnover rate, species richness, and fitness landscape as defined by the probability of extinction. Evenness was measured using a modified Pielou's evenness index, modified for algorithmic efficiency. Rather than the mathematical equation, $J' = \frac{H'}{H'_{max}}$ where H' is Shannon's Index, $H' = - \sum_{i=1}^S \frac{p_i}{\ln p_i}$, where p_i is the proportion of the i^{th} species, and S is the maximum number of species, we modify it as follows: Because the number of species per cell is fixed, the algorithm used is computed by the number of unique functional groups in a cell, divided by the number of species in the cell. This has the range $\left[\frac{1}{S}, 1\right]$, where the lower range is found in the case where every species in the cell is functionally unique, and the higher when there is only one species type. Turnover is modeled such that new species arise when there is an opening in relation to the packed niches of each cell. This opening occurs in two places: (1) when there is selection between the linked and unlinked

320 species in which the least fit species is eliminated; and (2) in perturbation events when a large
321 number of species go extinct without regard to their fitness status as might happen in a large
322 natural disaster such as a wildfire.

325 Results and Discussion

326 These simulation results provide evidence that constraint closure may lead to measures of
327 ecological stability through long-term evolutionary and ecosystem processes. It is evident that
328 such processes included in the simulation model do affect the flow of species through the
329 ecosystem in much the same way that Moreno and Mossio (2015) propose that stable
330 autonomous systems are created by constraint closure. This model provides simulation evidence
331 (see Parker (2021) for how simulations can be used as evidence (Parker 2020)) that certain
332 evolutionary and ecological processes can, in principle, act as stabilization conditions and
333 provide semi-autonomy in ecological communities.

334 In Fig. 4a, the effect of the number of constraints on fitness (defined as the extinction
335 probability) can be assessed. We would expect as the average fitness among species across the
336 community landscape of the model begins to evolve over time, that the average extinction
337 probability decreases as the constituent species adapt to the conditions of the local cell in which
338 they are found. As the number of evolutionary and ecological constraints increases, fitness
339 increases, and the variation in fitness decreases. Two noteworthy results on fitness emerge: (1)
340 without any processes present in the model run (The number of constraints = 0), the fitness stays
341 at a baseline value; when all 6 constraints are present in the model, maximum fitness is achieved
342 in the model. This suggests that the species flowing through the ecosystem are channeled by the
343 candidate constraints creating constraint closure in the relevant sense that Moreno and Mossio
344 outlined (Moreno and Mossio 2015).

345 The effect of the number of constraints on ecological evenness, as measured by the
346 modified Pielou's evenness index described above, is given in Fig. 4b. As the number of
347 constraints increases, the maximum mean evenness is achieved when there are three constraints
348 present, which pattern mirrors the required three constraints that Moreno and Mossio (2015)
349 argue are necessary to create constraint closure. Increasing the number of constraints causes the
350 evenness to decrease, although not as low as when there are no evolutionary and ecological
351 constraints present in the simulation model.

352 Fig. 4c evidences that the effect of the number of constraints on the species richness
353 (based on the number of functional traits present in the cell) averaged over all the cells making
354 up the community in the niche-cell. Species richness reaches a minimum when there are three
355 constraints and then increases with added constraints. The species richness has lower numbers of
356 species creating depauperate within-cell diversity as the number of constraints increases. This
357 shows species richness increases with added evolutionary and ecological constraints as these
358 constraints affect species flow. So while individual cells may trend toward being more uniform
359 within the cell, landscape richness maintains variation. For example, from Fig. 4c, it is clear that
360 as the number of species shrinks and becomes < 10 species per cell, the overall community
361 ecology of the entire landscape retains its variation in these functional groups.

362 The effect of the number of constraints on the difference between turnover in species
363 with and without mutualistic relationships (linked and unlinked species) is provided in Fig. 4d.

366 The turnover rate is the number of new species created at each time step in each cell. The pattern
367 is consonant with networked species alone, creating an increase in turnover rate due to selection.

368 Fig. 5a shows the percent increase in mean standard error (MSE) in a random forest
369 model for landscape fitness when each constraint is removed. When all six constraints are
370 present in the model, the maximum fitness is achieved. As expected, the MSE as constraint
371 *selection* and *competition* are the most important processes for the overall fitness of the
372 landscape.

373 The constraint with the greatest influence on *evenness* (Fig. 5b), was *species-drift*. As
374 expected, as species drift away from parental type, *species-drift* strongly influences evenness
375 measures, and structures the community through its influence on the set of constituent species.
376 The evolutionary processes of *selection* and *competition* are also highly influential in structuring
377 evenness scores, while *dispersal* and *mutualistic networks* are weaker influencers on *evenness*
378 compared with *selection* and *competition*.

379 Similar to *evenness*, *species-drift* was the constraint with the greatest influence on
380 *richness*, Fig. 5c. This makes sense as species drift from parental type, which in turn strongly
381 influences the number of species. The evolutionary influences of *selection* and *competition* were
382 highly influential in structuring richness scores, while *dispersal* and networks of *mutualism* were
383 the least influential for similar reasons to evenness discussed above.

384 The pattern of constraint influence seen in the difference in turnover between species
385 forming mutualistic networks and those not, Fig 5d. This pattern is consonant with networked
386 species alone creating an increase in turnover rate due to *selection*, the second most influential
387 constraint of interest. As *competition* is set up between species without regard to *mutualistic*
388 relationships, it unsurprisingly drops to the lowest constraint of interest in the analysis.

389 Fig. 6, shows the effect of the number of constraints on the variance of landscape fitness.
390 The pattern of mean distribution and the variance associated with the progression as the number
391 of constraints increases is evident in the scatter of the variances.

392 One criticism that might be leveled in the above results is that species richness might
393 achieve within-cell population sizes that are too low to be meaningful. While richness is often
394 low in the simulated landscape with increasing constraints, the mean and median are such that a
395 large proportion of low species diversity in the cell does not change the landscape variance in
396 richness, as illustrated in Fig. 7. This also suggests that while individual cells in the ecosystem
397 may become depauperate, the overall variation in diversity of species in the ecosystem remains
398 remarkably consistent. This might have ecological reality, illustrating that if a set of niches
399 within a community becomes dominated by a particular kind of species (e.g., a patch of forest
400 becoming meadow), the ecological community does not necessarily become simplified to the
401 point where low richness among the functional groups translates into low community richness.
402 As can be seen, there is an almost consistent variance in the simulated community as the number
403 of emergent constraints increases.

404 Further information on how each of the emergent constraints affects fitness singly and in
405 interaction is found in Appendix 2, and in Fig. S1, S2, and S3.

407 Conclusions

408
409 The proposal that evolutionary and ecological processes can act as constraints in the
410 appropriate sense that it creates autonomy (or semi-autonomy) and stability needs more
411 theoretical and practical examination and development. This model is not a knockdown

412 argument for evolutionary constraints acting on the flow of species and thereby structuring
413 ecological stability. Even so, it is a starting point worth further thought and consideration. The
414 work presented here points to the possibility that evolutionary and ecological forces might be
415 explored as a possible source of stability in community ecology.

416 We believe the idea of combining Vellend-like processes with Moreno and Mossio's
417 (2015) approach to autonomous systems has enough warrant to be productively explored by
418 ecologists and other theorists and modelers. They might help formalize the structural arguments
419 argued here to explore whether the analogy holds more rigorously. If these proposed
420 evolutionary and ecological dynamics constrain species in the way we suggest, and to which our
421 model points, then the structures of community ecology, such as stability and long-term
422 persistence, should emerge as constant patterns due to specific evolutionary and ecological
423 processes, suggesting a new research program for community ecology. In addition, such work
424 may discover other general features of these systems that suggests constraint closure allows for
425 their continuance.

426 The emergence of particular pattern formations also may imply empirical work that lends
427 evidentiary support as to whether these evolutionary dynamics form propensity enhancing
428 propensities in the right way to meet the criteria as defined by Dussault and Bouchard (2016)
429 {Dussault, 2016 #29. Further work is, of course, needed to establish these informed speculations,
430 but we argue here that it is warranted and may be highly productive.

431
432
433
434

435 **Figure Legends**

436
437 Fig. 1. Constraint Closure. Constraint closure is described as constraints on the flow of matter or
438 energy that form a closed loop allowing the evolution of stability and autonomy. According to
439 Moreno and Mossio (2015) this is a necessary step in the formation of an autonomous system.
440 Such a configuration of constraints provides a sustainable feedback loop that allows the system
441 to achieve coherence for a period of time, τ . In the original theoretical formulation by Moreno
442 and Mossio these are envisioned to have causal effects on any processes that makes
443 transformations from one condition to another. In the system depicted above, the A_i are
444 processes, and the C_j 's are constraints. The arrows indicate the flow of energy or material
445 effecting causal changes to either the A or C system components indicated. Each vertical row of
446 processes and constraints represent different times. The processes enclosed in the ellipse together
447 form a closed-loop creating “constraint closure,” producing stability, which serves as a
448 foundation for the emergence of autonomy in the system. It is proposed that closed constraints
449 may help stabilize ecological communities in which ecological and evolutionary forces act as
450 constraints on species flowing through ecosystems on relevant time scales. Redrawn from Figure
451 1.3, p. 21, Moreno and Mossio (2015).

452
453 Fig. 2. Distribution of habitat (black) vs nonhabitat (grey) in model landscapes with the given
454 parameter values of *spatial_habitat_patch_distribution_parameter_1* (SHP). See ODD section
455 of Appendix 1 for details. Such random landscapes are generated in the initialization phase of the
456 model. Fig. 2A, SHP = 2 ; Fig. 2B, SHP = 6; Fig. 2C, SHP = 12; Fig. 2D, SHP = 20.

457
458 Fig. 3. Flow of model. See ODD section of Appendix 1 for details. Process Flow within model.
459 The agent-based model used in this paper cycles through Vellend's proposed higher-level
460 ecological and evolutionary processes that operate in ecological communities, with the added
461 ecological processes of mutualism, and periodic, catastrophic species loss. These structures will
462 emerge as constraints to create semi-autonomous structures. The flow chart shows the order of
463 events and possible flow paths taken in the model.

464
465 Fig. 4. The theory of constraint closure developed by Moreno and Mossio (2015) suggests that at
466 least three constraints are necessary to create closure. In this figure, we focus only on the number
467 of constraints and how they affect the given measures of ecological stability. Boxplots show
468 ecological measures across parameter space for different numbers of constraints generated by
469 2000 iterations of the simulation model across all 64 possible permutations of the proposed
470 constraints being present or absent, and with uniform random values of parameters pulled from
471 their possible values in the model iterations. The whiskers in the box plot show the range of
472 variation. The actual data used to construct the boxplots are shown fine gray dots to the right of
473 each box plot. (a) Effect of the number of constraints on landscape fitness. (b) Effect of the
474 number of constraints on ecological evenness. (c) Effect of the number of constraints on the
475 average cell species richness (number of species defined as functional groups) present in the cell-
476 niche. (d) Effect of the number of constraints on the difference between turnover in mutualistic
477 linked and unlinked species. The turnover rate is the number of new species created at each time
478 step in each cell.

480 Fig 5. Percent increase in mean standard error in a random forest model of ecological measures
481 when each constraint is removed (see main text for an explanation of this statistical technique).
482 This analysis gives a measure of the influence of each constraint on the model. (a) For *landscape*
483 *fitness*, the mean square error (MSE) suggests *selection* and *competition* are the most important
484 constraints for the overall fitness of the landscape. This is not surprising given that these
485 constraints are well known to be directly tied to the ideas of fitness. (b) For *evenness* the percent
486 increase in MSE in a random forest model, *Species-drift* was the constraint with the most
487 influence on evenness. (c) For *species richness* the percent increase in MSE in a random forest,
488 similar to *evenness*, *species-drift* was the constraint with the most influence on richness. (d)
489 Difference in turnover rate measures the difference in *turnover* rate between species that can
490 form mutualistic networks and those that cannot. The percent increase in MSE in a random forest
491 model when each constraint is removed is consonant with *networked* species being the constraint
492 with the most influence. *Selection*, is the second most influential constraint of interest.
493

494 Fig. 6. Effect of the number of constraints on the variance of average species fitness, defined as
495 extinction probability, across the landscape and across parameter space. Boxplots show
496 landscape variance for different numbers of constraints generated by 2000 iterations of the
497 simulation model across all 64 possible permutations of the proposed constraints being present or
498 absent, and with uniform random values of parameters pulled from their possible values in the
499 model iterations. The whiskers in the box plot show the range of variation. The actual data used
500 to construct the boxplots are shown fine gray dots to the right of each box plot.

501 Fig. 7. Overall landscape variation in species richness. Recall that richness in this analysis of
502 model results is measured on the functional group of the species rather than the number of
503 species. Because the cells represent packed niches, the number of species it holds is a constant.
504 Boxplots show the landscape-level variance in the difference between the turnover in mutualistic
505 linked and unlinked species. These differences are plotted against different numbers of
506 constraints, generated by 2000 iterations of the simulation model, across all 64 possible
507 permutations of the proposed constraints being present or absent, and with uniform random
508 values of parameters pulled from their possible values in the model iterations. The whiskers in
509 the box plot show the range of variation. The actual data used to construct the boxplots are
510 shown fine gray dots to the right of each box plot.
511

522

Variable name	Variable type and units	Meaning
<i>species_preferred_niche</i>	Real number; range (0, 1)	A single value integrating a species niche requirement. This will be compared with <i>cell_niche_score</i> to assess the probability of a species extinction.
<i>maximum_number_of_links</i>	Integer, range [2, 100]	For species that can form links, the maximum number of links that it can form with other species.
<i>current_number_of_links</i>	Integer, range [0, <i>maximum_number_of_links</i>]	The current number of other species with which a species forms a link.
<i>probability_of_species_extinction</i>	Real number; range (0, 1)	The probability of extinction for a species based on the absolute difference between a species <i>species_preferred_niche</i> and the <i>cell_niche_score</i> on the cell in which it currently resides.
<i>species_niche_function</i>	Integer, range [1, <i>maximum_number_of_links</i>]	This identifies the trait-based functional group within a niche to which the species belongs.

523

524

Table 1: Species state variables

525

526

Variable name	Variable type and units	Meaning
cell location; the code uses NetLogo's built in coordinate variables	Real, static; scale is arbitrary and represents a scale of community size for an average number of possible species	The X and Y coordinates of the cells with origin in bottom left corner.
<i>cell_niche_score</i>	Real, static, range [0, 1), a value of 0 designates an uninhabitable patch	The cell-niche characterization by which species will be compared in assessing their extinction probability on a patch.

527

528 Table 2: Cell state variables

529

530

Model Constraint	Description	Definition
<i>create_species_networks</i>	Establishment and maintenance of mutualistic network of linked species within a cell. If this constraint is turned off mutualistic networks are not established.	This constraint of species flow through the community establishes a set of mutualistic relationships among a subset of species in a given cell. In species competition, and those species that are linked, have an extinction probability equal to the lowest extinction probability among the linked species.
<i>disperse_species</i>	Disperses species among adjacent cells based on probability of extinction among species inhabiting the cell.	In dispersal events the species with the lowest extinction probability is selected to spread from the cell. This constraint sorts the species, by extinction probability. If there is room for another species in a randomly chosen neighbor, the species is instantiated in the chosen cell. Note: it is not dispersal in the sense of an individual moving away from one cell and into another. Rather, a species becomes established in the new cell while the species continues to remain present in the cell of origin.
<i>species_compete</i>	Competition within a cell between a randomly chosen species that has mutualistic relationships with other species in the cell, and another	This constraint compares two species chosen at random, one with mutualistic

	randomly chosen species in the cell that does not have mutualistic relationships within the cell.	associations and the other without. The species with the highest probability of extinction goes extinct. This opens a space in the cell for another species to move into the cell from other processes such as dispersal or speciation.
<i>speciate_without_species_drift</i>	Species spread to a new opening in the cell with attributes derived from the mean value of the parental species' species_niche_function score and variation from that score defined by the parameter variation_in_extinction_probability from which the extinction probability of the new species is derived.	This constraint represents the spread of the species to an opening into an empty niche space within a cell. The species created from the parental species has the same mean value niche score as the parental species but with a given variance specified with a parameter given in column left. Species that can form mutualistic relationships remain able to form such relationships.
<i>speciate_with_species_drift</i>	<i>speciate_with_species_drift</i> , is identical to <i>speciate_without_species_drift</i> except it adds the possibility of species to drift from their trait-based functional group. This includes the possibility of networked species becoming non-networked species and vice-versa.	This constraint focuses the spread species to an opening into an empty niche space within a cell. The species created from the parental species has a potentially different mean value niche score as the parental species and with a given variance specified with the parameter given in

		column left. Species can change from being able to form mutualistic relationships to those who cannot, and vice versa.
<i>cause_catastrophe</i>	<i>cause_catastrophe</i> causes catastrophic ecological collapse at predefined intervals set by, <i>frequency_species_eco_perturbation</i> , and in intensity defined by <i>percent-species-to-wipeout</i>	This constraint causes random species loss without regard to their extinction probability. The species are selected for extinction with probability proportional the parameter in left column of percent species affected. These events happen periodically set with a parameter randomly selected prior to initiation of the model.

532

533 Table 3: Evolutionary and ecological processes thought to constrain the flow of species through an
 534 ecosystem.

535

Parameter name	Parameter type and units	Parameter meaning
<i>spatial_habitat_patch_distribution_parameter_1</i>	Integer, range [1, 20]	Sets the spatial distribution and patch structure of model, including density, and permeability of cells at landscape level.
<i>maximum_time_to_run</i>	Integer, range [1, 100,000]	Number of time steps to run model.
<i>percent_species_loss_eco_perturbation</i>	Real random variable, range [0, 100]	The number of species lost in a cell due to ecology perturbation without regard to extinction probability.
<i>frequency_species_eco_perturbation</i>	Integer random variable, range [0, 10,000]	Time steps between ecological perturbations.
<i>maximum_number_species_per_patch</i>	Integer, range [5, 100]	The number of species occupying the packed niche space of the cell.
<i>number_linked_species_per_patch</i>	Integer, range [2, 100]	The number of linked species possible.
<i>maximum_intial_proportion_links</i>	Real, range [0.01, 0.99]	The initial proportion of linked species, out of total maximum linked species, present in the patch.
<i>variation_in_extinction_probability</i>	Real random variable, range [0, 0.25]	When a new species is spawned from an old, the variation in extinction probability from parental species.
<i>species_drift_variation_speciation</i>	Real random variable, range [0, 0.1]	The variation between parental species and new speciation in functional group of parental species.

538 Table 4: Global parameters initialized at the beginning of the model (all random variables are drawn
 539 from a uniform distribution).

542 **Acknowledgements**

543

544 We want to thank the anonymous reviewers and the editors for helping shepherd this
545 manuscript forward. We would especially like to thank Volker Grimm for his patience and
546 insight in making this a much better paper than it would have been without his help.

547

548

549

550

- 551 **Bibliography**
- 552
- 553 Abeysinghe, W. and Punchi-Manage, R. 2020. Beyond neutrality: adding habitat filtering to neutral
554 models. - *Ecological Processes* 9: 1-9.
- 555 Alonso, D., et al. 2006. The merits of neutral theory. - *Trends in ecology & evolution* 21: 451-457.
- 556 Armbruster, P., et al. 2002. Factors influencing community structure in a South American tank
557 bromeliad fauna. - *Oikos* 96: 225-234.
- 558 Booth, M. G. 2004. Mycorrhizal networks mediate overstorey-understorey competition in a
559 temperate forest. - *Ecology Letters* 7: 538-546.
- 560 Bryant, R. 2012. What If ecological communities are not wholes? - *The Environment: Philosophy,
561 Science, and Ethics*: 37-56.
- 562 Carturan, B. S., et al. 2020. Combining agent-based, trait-based and demographic approaches to
563 model coral-community dynamics. - *Elife* 9: e55993.
- 564 Chase, J. M. 2003. Community assembly: when should history matter? - *Oecologia* 136: 489-498.
- 565 Chase, J. M. 2005. Towards a really unified theory for metacommunities. - *Functional Ecology* 19:
566 182-186.
- 567 Chesson, P. 2008. Quantifying and testing species coexistence mechanisms. - *Unity in diversity:
568 reflections on ecology after the legacy of Ramon Margalef*: 119-164.
- 569 Chesson, P. 2018. Updates on mechanisms of maintenance of species diversity. - *Journal of ecology*
570 106: 1773-1794.
- 571 Clements, F. E. 1916. Plant Succession: An Analysis of the Development of Vegetation. - Carnegie
572 Institution of Washington.
- 573 Collier, J. and Cumming, G. 2011. A Dynamical Approach to Ecosystem Identity. *Philosophy of
574 Ecology*. North-Holland, pp. 201-218.
- 575 de Mazancourt, C., et al. 2013. Predicting ecosystem stability from community composition and
576 biodiversity. - *Ecology Letters* 16: 617-625.
- 577 Derraik, J. G. B., et al. 2002. Arthropod morphospecies versus taxonomic species: a case study with
578 Araneae, Coleoptera, and Lepidoptera. - *Conservation Biology* 16: 1015-1023.
- 579 Dussault, A. C. and Bouchard, F. 2016. A persistence enhancing propensity account of ecological
580 function to explain ecosystem evolution. - *Synthese*: 1-31.
- 581 Eliot, C. 2007. Method and metaphysics in Clements's and Gleason's ecological explanations. -
582 *Studies in History and Philosophy of Science Part C: Studies in History and Philosophy of Biological
583 and Biomedical Sciences* 38: 85-109.
- 584 Elliott-Graves, A. 2018. Generality and causal interdependence in ecology. - *Philosophy of Science*
585 85: 1102-1114.
- 586 Evans, M. R., et al. 2013. Do simple models lead to generality in ecology? - *Trends in Ecology &
587 Evolution* 28: 578-583.
- 588 Fontana, S., et al. 2016. Individual-level trait diversity concepts and indices to comprehensively
589 describe community change in multidimensional trait space. - *Functional Ecology* 30: 808-818.
- 590 Gleason, H. A. 1926. The individualistic concept of the plant association. - *Bulletin of the Torrey
591 Botanical Club* 53: 7-26.
- 592 Götzenberger, L., et al. 2012. Ecological assembly rules in plant communities---approaches, patterns
593 and prospects. - *Biological Reviews* 87: 111-127.
- 594 Grimm, V., et al. 2006. A standard protocol for describing individual-based and agent-based models.
595 - *Ecological modelling* 198: 115-126.

- 596 Grimm, V., et al. 2020. The ODD protocol for describing agent-based and other simulation models:
597 A second update to improve clarity, replication, and structural realism. - *Journal of Artificial
598 Societies and Social Simulation* 23.
- 599 Haber, M. H. and Odenbaugh, J. 2009. The edges and boundaries of biological objects. - *Biological
600 Theory* 4: 219-224.
- 601 Hanski, I. and others. 1999. Metapopulation ecology. - Oxford University Press.
- 602 Hanski, I., et al. 2017. Ecological and genetic basis of metapopulation persistence of the Glanville
603 fritillary butterfly in fragmented landscapes. - *Nature Communications* 8: 1-11.
- 604 Heijden, M. G. A., et al. 2015. Mycorrhizal ecology and evolution: the past, the present, and the
605 future. - *New Phytologist* 205: 1406-1423.
- 606 Heleno, R., et al. 2014. Ecological networks: delving into the architecture of biodiversity. - *Biology
607 Letters* 10.
- 608 Hubbell, S. P. 2001. The unified neutral theory of biodiversity and biogeography (MPB-32). -
609 Princeton University Press.
- 610 Huneman, P. 2014. Individuality as a theoretical scheme. I. Formal and material concepts of
611 individuality. - *Biological Theory* 9: 361-373.
- 612 Huneman, P. 2014. Individuality as a Theoretical Scheme. II. About the Weak Individuality of
613 Organisms and Ecosystems. - *Biological Theory* 9: 374-381.
- 614 Jeltsch, F., et al. 2019. Give chance a chance: from coexistence to coviability in biodiversity theory. -
615 *Ecosphere*, 10: e02700.
- 616 Kirchhoff, T., et al. 2010. The onesidedness and cultural bias of the resilience approach. – *Gaia-
617 Ecological Perspectives for Science and Society* 19: 25-32.
- 618 Lawton, J. H. 1999. Are there general laws in ecology? - *Oikos* 84: 177-192.
- 619 Leibold, M. A., et al. 2004. The metacommunity concept: a framework for multi-scale community
620 ecology. - *Ecology letters* 7: 601-613.
- 621 Linquist, S. 2015. Against Lawton's contingency thesis; or, why the reported demise of community
622 ecology is greatly exaggerated. - *Philosophy of Science* 82: 1104-1115.
- 623 Linquist, S., et al. 2016. Yes! There are resilient generalizations (or ``laws'') in ecology. - *The
624 Quarterly Review of Biology* 91: 119-131.
- 625 Lockwood, D. 2008. When logic fails ecology. - *The Quarterly Review of Biology* 83: 57-64.
- 626 Loreau, M. and de Mazancourt, C. 2013. Biodiversity and ecosystem stability: a synthesis of
627 underlying mechanisms. - *Ecology Letters* 16: 106-115.
- 628 Marks, C. O. and Lechowicz, M. J. 2006. Alternative designs and the evolution of functional
629 diversity. - *The American Naturalist* 167: 55-66.
- 630 Marquet, P. A., et al. 2014. On theory in ecology. - *BioScience* 64: 701-710.
- 631 Mnasri, M., et al. 2017. Comparison of arbuscular mycorrhizal fungal effects on the heavy metal
632 uptake of a host and a non-host plant species in contact with extraradical mycelial network. -
633 *Chemosphere* 171: 476-484.
- 634 Montévil, M. and Mossio, M. 2015. Biological organisation as closure of constraints. - *Journal of
635 Theoretical Biology* 372: 179-191.
- 636 Moreno, A. and Mossio, M. 2015. Biological Autonomy: A Philosophical and Theoretical Enquiry. -
637 Springer Netherlands.
- 638 Nunes-Neto, N., et al. 2014. Function in ecology: an organizational approach. - *Biology &
639 Philosophy* 29: 123-141.
- 640 Odenbaugh, J. 2007. Seeing the forest *and* the trees: realism about communities and
641 ecosystems. - *Philosophy of Science* 74: 628-641.

- 642 Odenbaugh, J. 2010. On the very idea of an ecosystem. - In: Hazlett, A. (ed.) New Waves in
643 Metaphysics. Palgrave Macmillan UK, pp. 240-258.
- 644 Odenbaugh, J. 2011. Complex ecological systems. - Philosophy of complex systems.
- 645 Odenbaugh, J. 2011. A general unifying theory of ecology. - A general theory of ecology: 51-64.
- 646 Parker, W. S. 2020. Evidence and Knowledge from Computer Simulation. - Erkenntnis: 1-18.
- 647 Peay, K. G. 2016. The Mutualistic Niche: Mycorrhizal Symbiosis and Community Dynamics. -
648 Annual Review of Ecology, Evolution, and Systematics 47.
- 649 Peck, S. L. 2009. Whose boundary? An individual species perspectival approach to borders. -
650 Biological Theory 4: 274-279.
- 651 Petchey, O. L. and Gaston, K. J. 2002. Functional diversity (FD), species richness and community
652 composition. - Ecology letters 5: 402-411.
- 653 Pigliucci, M. 2008. Sewall Wright's adaptive landscapes: 1932 vs. 1988. - Biology & Philosophy 23:
654 591-603.
- 655 Rosindell, J., et al. 2011. The unified neutral theory of biodiversity and biogeography at age ten. -
656 Trends in Ecology & Evolution 26: 340--348.
- 657 Roughgarden, J. 2009. Is there a general theory of community ecology? - Biology & Philosophy 24:
658 521-529.
- 659 Salmaso, N. and Padisák, J. 2007. Morpho-functional groups and phytoplankton development in
660 two deep lakes (Lake Garda, Italy and Lake Stechlin, Germany). - Hydrobiologia 578: 97-112.
- 661 Schneider, F. D., et al. 2017. Mapping functional diversity from remotely sensed morphological and
662 physiological forest traits. - Nature communications 8: 1-12.
- 663 Shrader-Frechette, K. and McCoy, E. D. 1994. Applied ecology and the logic of case studies. -
664 Philosophy of Science 61: 228-249.
- 665 Simard, S. W., et al. 2012. Mycorrhizal networks: mechanisms, ecology and modelling. - Fungal
666 Biology Reviews 26: 39-60.
- 667 Simberloff, D. 2004. Community Ecology: Is It Time to Move On? (An American Society of
668 Naturalists Presidential Address). - The American Naturalist 163: 787-799.
- 669 Spencer, H. G. 2020. Beyond Equilibria: The Neglected Role of History in Ecology and Evolution. -
670 The Quarterly Review of Biology 95: 311-321.
- 671 Sterelny, K. 2001. The reality of ecological assemblages: a palaeo-ecological puzzle. - Biology and
672 Philosophy 16: 437-461.
- 673 Sterelny, K. 2006. Local ecological communities. - Philosophy of Science 73: 215-231.
- 674 Strobl, C., et al. 2007. Bias in random forest variable importance measures: Illustrations, sources and
675 a solution. - BMC Bioinformatics 8: 25.
- 676 Sutherland, W. J., et al. 2013. Identification of 100 fundamental ecological questions. - Journal of
677 Ecology 101: 58-67.
- 678 Vellend, M. 2010. Conceptual synthesis in community ecology. - The Quarterly Review of Biology
679 85: 183-206.
- 680 Vellend, M. 2016. The Theory of Ecological Communities (MPB-57). - Princeton University Press.
- 681 Vellend, M., et al. 2014. Assessing the relative importance of neutral stochasticity in ecological
682 communities. - Oikos 123: 1420-1430.
- 683 Weiher, E., et al. 2011. Advances, challenges and a developing synthesis of ecological community
684 assembly theory. - Philosophical Transactions of the Royal Society B: Biological Sciences 366: 2403.
- 685 Weithoff, G. and Beisner, B. E. 2019. Measures and approaches in trait-based phytoplankton
686 community ecology—from freshwater to marine ecosystems. - Frontiers in Marine Science 6: 40.
- 687 Wennekes, P. L., et al. 2012. The neutral---niche debate: a philosophical perspective. - Acta
688 Biotheoretica 60: 257-271.

- 689 Woodward, J. 2001. Law and explanation in biology: invariance is the kind of stability that matters. -
690 Philosophy of Science 68: 1-20.
- 691 Yang, G., et al. 2014. The interaction between arbuscular mycorrhizal fungi and soil phosphorus
692 availability influences plant community productivity and ecosystem stability. - Journal of Ecology
693 102: 1072-1082.
- 694 Zakharova, L., , et al. 2019. Trait-based modelling in ecology: A review of two decades of research. -
695 - Ecological Modelling 407: 108703.

Appendix 1: Supplementary Information S1: ODD

This model description follows the ODD (Overview, Design concepts, Details) protocol for describing individual- and agent-based models (Grimm, et al. 2006), as updated by (Grimm, et al. 2020). A copy of the peer-reviewed model code can be found at Steve L. Peck (2020, December 06). “Evolution of Ecological Communities: Testing Constraint Closure” (Version 1.0.0). CoMSES Computational Model Library, and can be retrieved from <https://www.comses.net/codebases/c9d55182-4444-4cc1-9a6f-dcf9c4ca942a/releases/1.0.0/>.

1. Purpose and patterns

The following theoretical model has been designed to provide in principle arguments for how constraint closure of evolutionary dynamics may provide ecosystem stability in the sense described by Moreno and Mossio (Moreno and Mossio 2015). We explore this using a set of Mark Vellend's proposed processes for conceptually framing evolutionary changes in a set of species comprising the niche structure of ecological communities. In particular, the model includes Vellend's suggested processes, selection, species-drift, speciation, and dispersal (Vellend 2016), as well as ecological perturbation and mutualism. We hypothesize that these will allow several patterns of standard measures of ecological stability to emerge as a result of constraint closure. We argue that evolutionary dynamics acting on the flow of species through an ecosystem can be considered constraints, in which closed loops providing constraint closure emerge. In this model, we suggest that these processes act in such a way that the semi-autonomy of the ecological community emerges, creating conditions of stability. Specifically, we expect the model to show increases in stability measures as the number of constraints increases, especially as the number of constraints moves beyond the three proposed necessary to achieve constraint closure. The model is written in NetLogo 6.0.2 an Agent-Based Programming Environment, (Wilensky 1999).

2. Entities, state variables, and scales

Following Vellend's work, the model comprises two types of entities, species (rather than individual organisms), and grid cells that do not represent a physical metric space but a space of fixed ecological niches. Species compete within grid cells, send lines of descendants to adjacent grid cells, and can go extinct. The entities in the model include species representing the inhabitants of an ecological community. Grid cells represent spatially adjacent ecological niches, and a global environment representing a landscape comprised of inhabitable and uninhabitable locations.

2.1 State variables: species (agent)

Rationale: Species are the only agent in the model. The state variables for species are given in Table 1. Each species has a state variable called *species_preferred_niche_score*, which represents a one-dimensional measure of the species' resource needs. This variable corresponds to the cell state variable, *cell_niche_score*, which in turn is a measure of what needs that habitat

will provide. A species' extinction probability will be defined as the absolute difference between these two variables.

Each of the species in the model belongs to a functional group located in each cell defined by the species state variable *species_niche_function*. This variable defines a trait-based functional group that the species plays in the community (e.g. in a real ecosystem, these might be such ecological functions as predator, herbaceous plant, etc., in the model the functional group is defined as an integer value that characterizes a set of traits with a single value), and different species can be in the same functional group. This trait-based functional group can change by wandering randomly through species-drift. Cells may be completely homogeneous or contain the entire gamut of possible ecological functional groups that a species can take on in the model. Because the number of species in a cell is fixed, is variable serves as a surrogate for differences in species in ecological evenness measures.

A subset of the species in the model may form mutualistic relationships with other species that can affect the extinction risk of the species involved.

Species can send lines of descent into adjacent cells in which the niche is not completely packed with species, and there is room among the niches for another species. Species are in competition with others in a given cell regardless of linked or unlinked status. They can also undergo catastrophic ecological collapse, species selection, species-drift, and other processes discussed in (Vellend 2016).

2.2. State variables: cells

Rationale: All of the state variables of the cells are given in Table 2. The unsuitable habitat cells are places where species cannot exist and form a barrier or impediment to migration. Each habitat cell represents a packed set of niches that can hold a given number of species representing *maximum_number_species_per_patch* ecological functional trait groups in the niches of the cell. Each cell contains an environmental variable, called *cell_niche_score*, which a habitat suitability index is chosen from a uniform random distribution between 0 and 1.

2.3 Spatial and temporal scales

Rationale: The space represented in the cell is non-Euclidean and space is only defined in terms of adjacency, i.e., cells may be near or far from each other, but there is no measure of distance x between cells. Each cell represents a fixed number of possible species viewed as functional groups. Time is specifically modeled, and each time step represents the mean time any species is likely to exist in a community before cell extinction or species evolution. The model is structured on a 16x16 (although much larger grid sizes, this captured the basic dynamics and was used for the exploration of the parameter space) spatial toroidal grid composed of cells that form clusters of suitable and unsuitable areas. The procedure for setting up the topological aspects of the grid is based on O'Sullivan and Perry (O'Sullivan and Perry 2013).

3. Process overview and scheduling

After establishing the world's initialization (details below), the following processes take place in the following order.

3.1 Module: `create_species_networks`

. **Brief Description:** This creates the linkages that will serve as mutualistic relationships among species. This creates formal links among subsets of species within a cell which will influence the fitness relationships of those species able to form such links.

3.2 Module: `disperse_species`. **Brief Description:** Species dispersal copies the species with the lowest extinction probability and moves the new instantiation to an adjacent cell if there is room for a new species. In the new cell, new extinction probabilities are calculated and join the network if it is a species able to form mutualistic relationships. In this way, species send descendants into adjacent cells in which there is room to hold another species. They can also undergo catastrophic ecological collapse, species selection, species-drift, and other processes discussed in (Vellend 2016).

3.3 Module: `Species_compete`. **Brief Description:** Competition among networked and non-networked species. A random species is chosen from networked and non-networked species, respectively. The species with the lower extinction probability goes extinct (note: in the competition between the species chosen, there is no advantage to being networked), and the winner speciates. Either (4a) or (4b) is executed depending on the initialization choices.

3.4.1 Module: `speciate_without_species_drift`. **Brief Description:** Speciation without drift New species are created from the parental species. No possibility of species-drift.

3.4.2 `speciate_with_species_drift`. **Brief Description:** Speciation with drift New species are created from the parental species. Possibility of species-drift.

3.5 Module: `cause_catastrophe`. **Brief Description:** Periodic ecological perturbation Periodic ecological perturbation at given intervals that causes the local extinction of a proportion of the species in a cell without regard to extinction probabilities.

3.6 Module: `select_with_drift_species_w_lowest_extinction_probability`. **Brief description:** Extinction of species In each cell (randomly examined), all of the species' *probability_of_species_extinction* is calculated as the absolute value of the difference between the *species_preferred_niche_score* and the cell's *cell_niche_score*. An undirected network is established between species in a subset of species chosen at initialization. The extinction probability of a linked species differs from unlinked species in that its demise in various places in the model is calculated as the lowest extinction probability of all the linked species considered as a collective. In species that are networked, the probability of extinction is the lowest probability of extinction among the set of all of the networked species. The species with the highest extinction probability goes extinct.

4. Design Concepts

4.1 Basic Principles

This agent-based theoretical model explores whether basic evolutionary processes found in ecological communities can form constraint closure in a way that creates ecological community stability. By using the ideas of constraint closure, we hope to provide a theoretical proof of concept for some of the patterns found in such communities and offer plausible interpretations of such patterns. The following is a generalized model. We are not trying to target a specific ecosystem found in a field study reported in the scientific literature but rather to keep an eye toward the kind of interactions found in the broad class of community systems. To put the following in terms of the model mapping schema provided by Weisberg (2013) (Weisberg 2013) the model below is a generalized, agent-based computational model, using a minimalist idealization of key processes thought to potentially function in ecological communities, i.e., evolutionary dynamics such as selection, species-drift, speciation, and dispersal. In addition, including elements common to ecological communities such as mutualism, and in addition, perturbations that indiscriminately affect species in an ecosystem (e.g., fires, hurricanes, and habitat destruction by human activity). The model considers the kind of constraints thought to capture the dynamics that confer biological autonomy and stability. These two ideas should allow community stability, as recognized by signature patterns to emerge. The fidelity criteria we aim for will be achieved if such patterns are present and suggest that these processes are sufficient to confer known emergent attributes to these model systems.

4.2 Emergence

We expect emergence to occur in two ways. In the first, we expect the evolutionary processes encoded in the model that affect species flow through the cells to emerge in the model as constraints. Second, we expect ecological stability to emerge as certain combinations of constraints form a closed loop of interaction and create constraint closure among the ecological processes described above.

4.3 Collectives

Mutualism among species in the model is explicitly included as one of the ecological processes considered as a possible constraint on species flow. Some of the species have the ability to form networks that lower the extinction probability of all species in a collective. The collective in effect has a lower extinction probability than those not in the collective.

4.4 Adaptation

Species that form networks compete with those that do not. Species are replaced each time step with instantiations of new species with the same mean niche score as the parental species within some variance determined by the difference between the cell's resource score and the species resource needs score. At each time step, species with the highest extinction probability go extinct, creating room for species with lower extinction probabilities to replace them through dispersal, speciation, or competition.

4.5 Sensing

The observer can sense if a cell in the adjacent eight cells has room for another species.

4.6 Interaction

Species within a cell compete as individual species.

4.7 Stochasticity

Stochasticity enters into the model in several places: (a) the initial cell structure of the landscape is randomly generated; (b) the proportion of species chosen to form links is a random variable; (c) the species that go extinct in the drift stage are randomly chosen with probability proportional to fitness; species with the lowest probability of extinction are chosen to move, and then move randomly to one of their 8-Moore-neighborhood neighbors; (d) the variation in fitness around newly formed species is a random variable.

4.8 Observation

Testing: The module BehaviorSpace in NetLogo 6.0.2 was used to test the overall behavior of the model, in a Latin hypersphere run of the model with random non-constraint independent variables. The following analyses were used to reconstruct the effects of the constraints on the model landscape. Much larger grid sizes were tested to find the minimal size needed for running a complete iteration set of the binary switches turning on and off the constraint *submodels* and for three values of the possible number of links per species, and two values of each of the parameter values given types given in Table 3. leading to 256 possible combinations that compromise the permutation set used to explore the constraint space, with 500 timesteps for each sample from the model hyperspace. A uniform random distribution of the variables found in Table 4, were drawn from the 2000 replications of each of the possible constraint permutation sets for each of the runs. These were sampled with replacement over the multidimensional parameter space. As White et al. (2014) note (White, et al. 2014), these results should not be interpreted as a test of statistical significance of parameter differences as the likelihood values, and statistical tests, are easily influenced by the number of iterations used in the model testing. Therefore, p-values were provided only as a heuristic.

4.9 The following variables were used to assess the patterns found in the model:

4.9.1 Evenness: Evenness was measured using a modified Pielou's evenness index. It was modified for algorithmic efficiency. Rather than the mathematical equation, $J' = \frac{H'}{H'_{max}}$ where H' is Shannon's Index, $H' = - \sum_{i=1}^S \frac{p_i}{\ln p_i}$, where p_i is the proportion of the i^{th} species, and S is the maximum number of species, we modify it as follows: Because the number of species per cell is fixed, the algorithm used is computed by the number of unique functional morpho-species in a cell, divided by the number of species in the cell. This has the range $\left[\frac{1}{S}, 1 \right]$, where the lower range is found in the case where every species in the cell is functionally unique, and the higher when there is only one species type.

4.9.2 Turnover rate: Turnover is modeled such that new species arise when there is an opening in relation to the packed niches of each cell. This opening occurs in two places: (1) when there is selection between the linked and unlinked species in which the least fit species is eliminated; and (2) in perturbation events when a large number of species go extinct without regard to their fitness status as might happen in a large natural disaster such as a wildfire.

Functional species groups: Because the number of species in a cell is fixed, the richness used in the model defines the number of functional types present. Therefore, species of a unique type, rather than species, as such. For example, if the cells have only one kind of species, the richness as defined here would be 1.

4.9.3 Landscape fitness: This is the mean extinction probability of a species averaged over all inhabitable cells. It is expected that this will increase over time as the species better aligns with the cell niche score.

5. Initialization

5.1 Cells

Cells are set up as being a random grid of suitable habitat, in which species can exist, or unsuitable habitat, which serves as an uninhabitable matrix and a barrier to migration or movement, Fig. 2. A landscape parameter, *spatial_habitat_patch_distribution_parameter_1*, is used to determine the landscape of cell features such as permeability, how the habitat cells are clustered, and other macroscale features of the landscape cell distribution. Each cell is given a niche attribute, *cell_niche_score*, with which species will be compared on their, *species_preferred_niche_score*. Their closeness to this value will determine their extinction probability in the cell. Cells are initialized to have the maximum number of species they can hold, representing packed niches.

5.2 Species

Species are initialized on a habitat cell with a random location on the matrix of cells, and a niche score, *species_preferred_niche_score*, that will be compared with the niche score of the habitat cell, *cell_niche_score*, on which they land is used to calculate the extinction probability. They are also assigned a functional trait group, *species_trait_group*, a score that will be used to determine its fit to the ecosystem defined in a cell. The functional group is used to separate the species. These functional types are not meant to represent actual species but are designed to mirror the idea of trait-based functional groups. Networks of linked species are established.

5.3 Model Initialization Parameters

For the model runs, several parameters are set that will be used throughout the simulation detailed in the *submodels* found below. A complete list of these is given in Table 3.

6. Submodels

6.1 *create_species_networks*

create_species_networks works on individual species. First, it looks to see whether there are any species present in the cell. Then, it establishes directed links with those species with a species number less than the *number_linked_species_per_patch* variable. The species first receives a species number sequentially at first. Therefore, this forms links only for those whose species number is less than the number that can form links. Cells can be composed of completely

linkable species or non-linkable species depending on the evolutionary dynamics that play out elsewhere in the model.

Algorithm for submodule

Uses link set: mutualisms

1. For each species in the cell, if any are present;
2. form a set of all species that are eligible to form an undirected linked network
3. calculate the number of links to create as *maximum_number_of_links* * *maximum_initial_proportion_links*
4. create the given number of links from step #3 among the identified eligible species

6.2 create_little_network

Operates identically as described above, but used for a single species newly created or migrating from another cell.

6.3 select_species_with_lowest_prob_extinction,

select_species_with_highest_prob_extinction, first sets the probability of extinction as the given species to the lowest probability of extinction of those species it has a link with (so that being in a network increases the fitness advantage of the entire network). It then causes the extinction of the species with the lowest fitness.

Algorithm for submodule

1. For all linked species in a cell
2. Change the probability of extinction to the lowest probability among those species with which it is linked.
3. Sort by probability of extinction from high to low.
4. Cause the extinction of the species with the highest probability of extinction.

6.4 disperse_species

disperse_species, sorts the species by extinction probability within a cell, and checks the existence conditions that appropriate species are present in the cell. It then takes the species with the lowest extinction probability, hatches a new species and then moves it to an adjacent habitable cell with room for a new species, lets the species establish mutualisms with their new neighbors, and recalculate its extinction probability. The "else" segment moves species without links to the new cell and calculates its new fitness in the new cell. It hatches a new species first, then looks for room, then moves there, or it goes extinct.

Algorithm for submodule

Within a cell

1. Find the species with the lowest extinction probability
2. Create an new instance of that species
3. Randomly search the adjacent cells to see if there is room for another species
4. If so, move species to that cell
 - (a) If it is a species able to form a network establish network in a new cell
 - (b) Establish its extinction probability in the new cell
5. Else if there is nowhere to go, cause the extinction of new instantiation of species

6.5 species_compete

species_compete picks one of the linked species and one of the unlinked species and compares the two. The winner of the contest speciates, while the loser dies. This function also allows the possibility of species-drift.

Algorithm for submodule

Within a cell

1. Pick one species randomly from the linked species and one from the unlinked species.
2. Pick a competitive winner by comparing the two species' probability of extinction and choosing the one with the highest extinction probability.
3. Speciate with or without species-drift (set by initialization parameter).
4. Cause cell extinction in the non-winning species.

6.6 Speciate_without_species_drift

speciate_without_species_drift is an algorithm designed to create a speciation event such that the new species has the same attributes as the parental species, but with some variation. It takes a species and creates a new species based on the parental species mean extinction probability and adds variation, *variation_in_extinction_probability*, which is set at initiation. For those species with links, it creates links with *create-little-network*.

Algorithm for submodule

1. For each species targeted for speciation—
2. Add or subtract random variation from its extinction probability based on a random variate from the variable, *variation_in_extinction_probability*.
3. Check it is not < 1 or > 0
4. If it is, reflect back the amount remaining above or below the boundary
5. Create a new instantiation of that species
6. Set its extinction probability to the calculated value

6.7 speciate_with_species_drift

speciate_with_species_drift is identical to *speciate_without_species_drift*, except that it adds the possibility of species to drift from their functional group. This includes the possibility of networked species becoming non-networked species and vice-versa.

Algorithm for submodule

1. The algorithm of this submodule is identical to steps 1-6 of the algorithm in *speciate_without_species_drift*.
2. With equal probability increase or decrease the *species_trait_group* by one.
3. Check it is not over or under the allowable boundaries
4. If it goes outside of the boundaries, reflect back the increase or decrease.

6.8 cause_catastrophe

cause_catastrophe causes catastrophic ecological collapse at predefined intervals defined by, *frequency_species_eco_perturbation*, and in intensity defined by *percent-species-to-wipeout*.

1. At time interval $\text{mod}(\text{frequency_species_eco_perturbation}) = 0$
2. In all cells
3. Loop through all species
4. Cause a species to go extinct with probability *percent-species-to-wipeout*

Appendix 2: Supplementary Information S2: Influence of evolutionary and ecological processes that emerge act constraints on model outcomes

Fig. S1. *Average community fitness across parameter space for all possible combinations of constraints.* See the ODD in supplemental information S1 for more details. This figure provides a closer look at the processes analyzed in the main body of the article and described in Fig. 5 of the paper. Fitness is defined in the model as the extinction probability ($\text{prob}(\text{extinction})$). Lower values of extinction probability imply higher fitness. This figure provides specific information on how each of the evolutionary and ecological processes individually and in interaction, with all other represented constraints, affect the simulated landscape representing a community of species. The figure shows the effect of constraints on average community fitness across all combinations of the model's parameter space. As the number of constraints increases, the overall pattern of the most influential evolutionary processes in interaction with the others can be examined. For example, when $\text{number constraints} = 0 \Rightarrow \text{prob}(\text{extinction}) > .3$ which serves as a baseline fitness with which to compare how adding constraints and their interaction affect average fitness levels of species in the simulated community. When the model is run with $\text{number of constraints} = 1$ the evolutionary processes representing *selection* and *competition* increase fitness more than other processes ($\text{Prob}(\text{extinction}) < 0.1$) for both processes. This is not unexpected as these are evolutionary forces known to influence fitness strongly. Of special note is when there are three constraints present, 16 of the 20 possible interactions show a high fitness, with the *extinction probability* < 1.3 implying that when $\text{number constraints} \geq 3$ there is a pattern of higher average landscape fitness in the digital ecology of the simulated community than when there are fewer constraints.

Fig. S2 *Average ecological evenness across parameter space for all possible combinations of constraints.* See the ODD in supplemental information S1 for more details on how ecological evenness is defined. This figure provides a closer look at the processes analyzed in the main body of the article and described in Fig. 6 of the paper. This figure provides specific information on how each of the evolutionary and ecological processes individually and in interaction with all other represented constraints affect the simulated landscape representing a community of interacting species. The figure shows the effect of constraints on average community evenness across all combinations of the model's parameter space. As the number of constraints increases, the overall pattern of the most influential evolutionary processes in interaction with the others can be examined. For example, when $\text{number constraints} = 0 \Rightarrow \text{average ecology evenness} > .7$ which serves as a baseline evenness with which to compare how adding constraints and their interaction affect average fitness levels of species in the simulated community. When the model is run with $\text{number of constraints} = 1$ the evolutionary processes representing *selection* and *competition* increase evenness more than other processes ($\text{average ecology evenness} < 0.2$) and for the process *ecological catastrophe* ($\text{average ecology evenness} < 0.3$). Of special note is when there are two constraints present, 7 out of 15 interactions average ecological evenness < 0.2 and for three interactions, 8 of the 20 possible interactions show ($\text{average ecology evenness} < 0.2$), with the average ecological evenness increasing for higher parameter values. This suggests greater diversity but not so high that only unique individual species occur in each cell and away from very low diversity being present in a cell for higher values of ecological evenness. Recall from the ODD found in that

This has the range, $\left[\frac{1}{S}, 1\right]$, where S is the number of unique functional group in which the lower range is found in the case where every species in the cell is functionally unique, and the higher when there is only one species type. This shows that the model itself because of the structure of the simulated community special components does not allow the theoretical maximum to occur, which should be taken into account when evaluating the meaning of the evenness achieved for the various constraint numbers and interactions.

Fig. S3. Average ecological richness as defined by number of functional groups present in a cell across parameter space for all possible combinations of constraints. See the ODD in supplemental information S1 for more details on how richness is defined and calculated. This figure provides a closer look at the processes analyzed in the main body of the article and described in Fig. 7 of the paper. It provides specific information on how each of the evolutionary and ecological processes individually and in interaction with all other represented constraints affect the simulated landscape representing a community of interacting species. The plots show the effect of constraints on the average of the cell species' richness across all combinations of the model's parameter space. The is at a maximum when there are no constraints when $number\ constraints = 0 \Rightarrow average\ richness > 15$ with which to compare how adding constraints and their interaction affect average fitness levels of species in the simulated community. When the model is run with $number\ of\ constraints = 1$ the evolutionary processes representing *selection* and *competition* restrict the richness severely (while noting from Fig. 8, in the main article the number of constraints does not impact the overall community variance in richness so depauperate individual cells do not imply a less rich community just less within cell richness). the average richness of these two processes ($average\ richness < 5$) and for the process ecological catastrophe which also reduced cell richness ($average\ richness < 10$). Of special note is when there are two constraints present, 7 out of 15 interactions average richness < 5 and for three interactions, 8 of the 20 possible interactions show ($average\ richness < 5$), with the average richness increasing for some of the constraints including those with the two most relevant to species richness reduction. However, the propensity for species to rebound to greater richness even in the presence of *selection* and *competition* suggests that even after three constraints, a greater number of interactions are increasing in richness. This increase in richness continues as constraints are added despite evolutionary and ecological forces that reduce richness in individual cells.

Appendix 3: Organizational approach (OA) to ecological communities and persistence enhancing functions (PEP)

Mossio, Saborido, and Moreno (Mossio, et al. 2009) formally define biological function as, "A trait T has a function in the organization O of a system S if and only if:

- C1 : T contributes to the maintenance of the organization O of S;
- C2 : T is produced and maintained under some constraints exerted by O;
- C3 : S is organizationally differentiated." p. 828

This is in contrast to the Etiological Approach (EA), which uses evolutionary history to define function as doing that which it was selected to do in an adaptationist sense. OA then considers only what the function does in the current ecosystem, not as it is embedded in its historical or evolutionary framing. This, drawing on the work of Moreno and Mossio (Moreno and Mossio 2015) on constraint closure, leads to an ecological definition of function, by Nunes et al.. An ecological function is a precise (differentiated) effect of a given constraining action on the flow of matter and energy (process) performed by a given item of biodiversity in an ecosystem closure of constraints " (Nunes-Neto, et al. 2014).(p. 131)

This provides a handle for seeing ecological communities as autonomous systems, as maintained by constraint closure. The constraints are modulations induced by the components of biodiversity acting on the flow of matter and energy through the system, such that they feed back into the maintenance of the ecosystem. It is important to point out that the claim of its being an autonomous system is not the same as the claim that they are individuals in the biological sense, although they do seem to exhibit weak individuality (Huneman 2011).

Dussault and Bouchard(Dussault and Bouchard 2016)add a significant consideration to the above account, developing a focus on particular kinds of functions that they call persistence enhancing propensities (PEP). PEPs are a class of functions that allow the currently structured ecosystem to persist. The focus is not on these functions in an adaptive sense, such that their historical grounding has been necessarily worked out. Rather, it is future-directed such that it attends to the extant functions that contributed to current stability--defined as the ability to return to equilibrium after a disturbance. They define a PEP function as

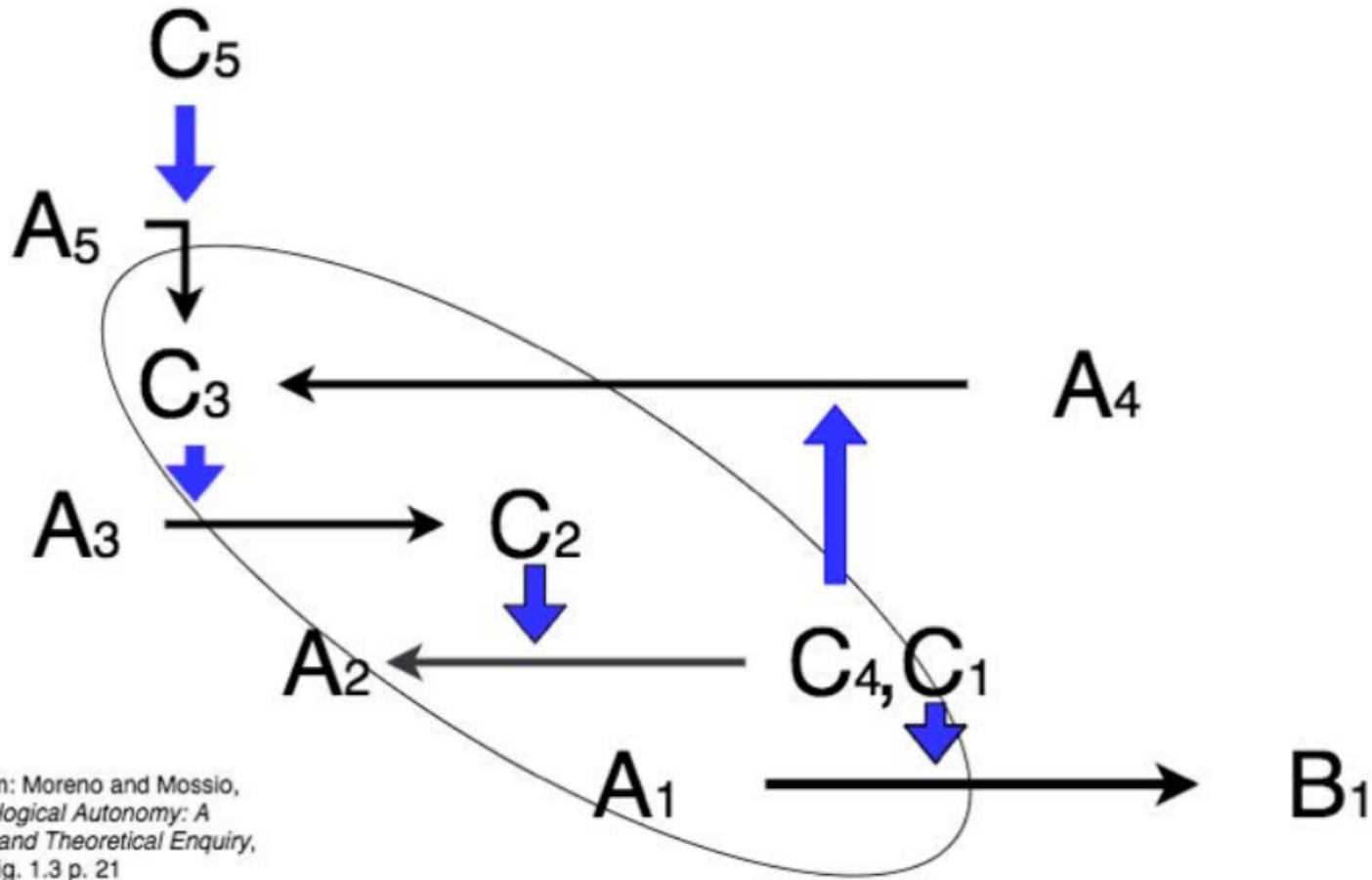
"The PEP function of x in an ecosystem E is to F if, and only if, x is capable of doing F and x's capacity to F contributes to E's propensity to persist." (Dussault and Bouchard 2016) p. 8)

PEPs then allow us to explore the question of what features provide for the ability of this ecosystem to survive disturbances to equilibrium. It pulls away from questions about how particular assemblages come to be constructed and looks for general characteristics of ecosystem functioning and construction that are based on the functional group that species play in the current community.

These two proposals (OA and PEP) offer possible connections with the evolutionary constraints we develop below regarding ecological communities and that influence stability and structure, which allow them to be considered autonomous systems.

- Dussault, A. C. and Bouchard, F. 2016. A persistence enhancing propensity account of ecological function to explain ecosystem evolution. - *Synthese*: 1-31.
- Grimm, V., et al. 2006. A standard protocol for describing individual-based and agent-based models. - *Ecological modelling* 198: 115-126.
- Grimm, V., et al. 2020. The ODD protocol for describing agent-based and other simulation models: A second update to improve clarity, replication, and structural realism. - *Journal of Artificial Societies and Social Simulation* 23.
- Huneman, P. 2011. About the conceptual foundations of ecological engineering: Stability, individuality and values. - *Procedia Environmental Sciences* 9: 72-82.
- Moreno, A. and Mossio, M. 2015. Biological Autonomy: A Philosophical and Theoretical Enquiry. - Springer Netherlands.
- Mossio, M., et al. 2009. An organizational account of biological functions. - *The British Journal for the Philosophy of Science* 60: 813-841.
- Nunes-Neto, N., et al. 2014. Function in ecology: an organizational approach. - *Biology & Philosophy* 29: 123-141.
- O'Sullivan, D. and Perry, G. L. W. 2013. Spatial simulation: exploring pattern and process. - John Wiley.
- Vellend, M. 2016. The Theory of Ecological Communities (MPB-57). - Princeton University Press.
- Weisberg, M. 2013. Simulation and Similarity: Using Models to Understand the World. - OUP USA.
- White, J. W., et al. 2014. Ecologists should not use statistical significance tests to interpret simulation model results. - *Oikos* 123: 385--388.
- Wilensky, U. 1999. Netlogo.

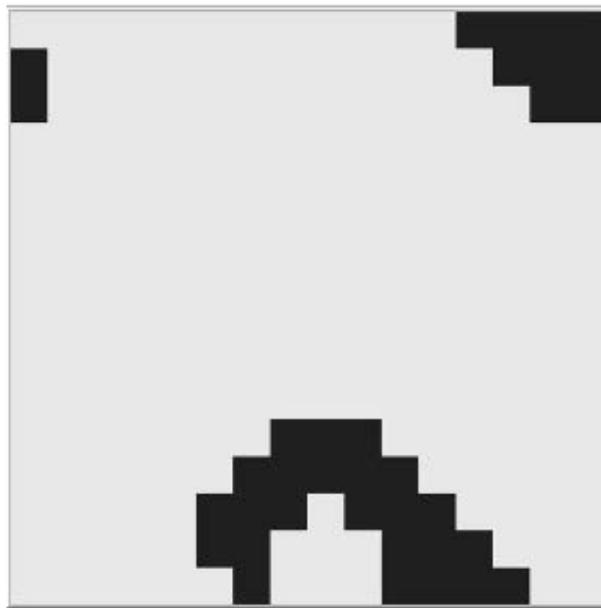
Fig. 1



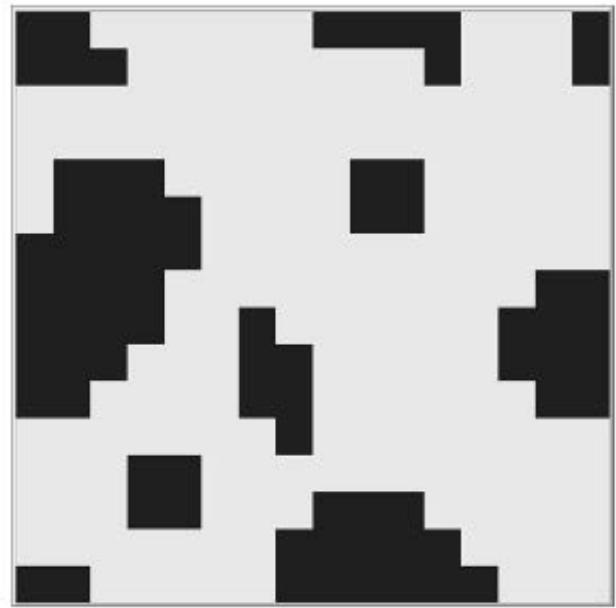
Adapted from: Moreno and Mossio,
2015, *Biological Autonomy: A
Philosophical and Theoretical Enquiry*,
Fig. 1.3 p. 21

Fig. 2

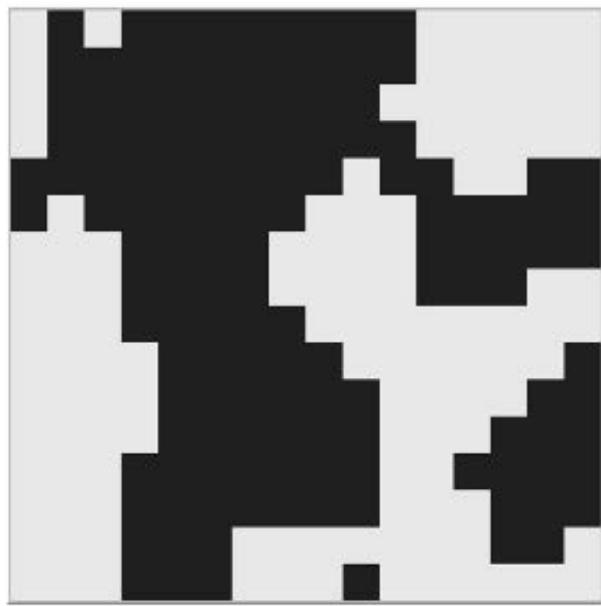
A



B



C



D

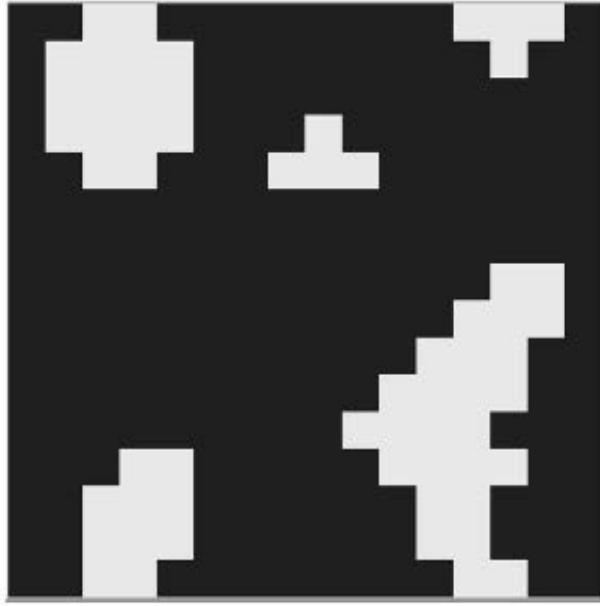
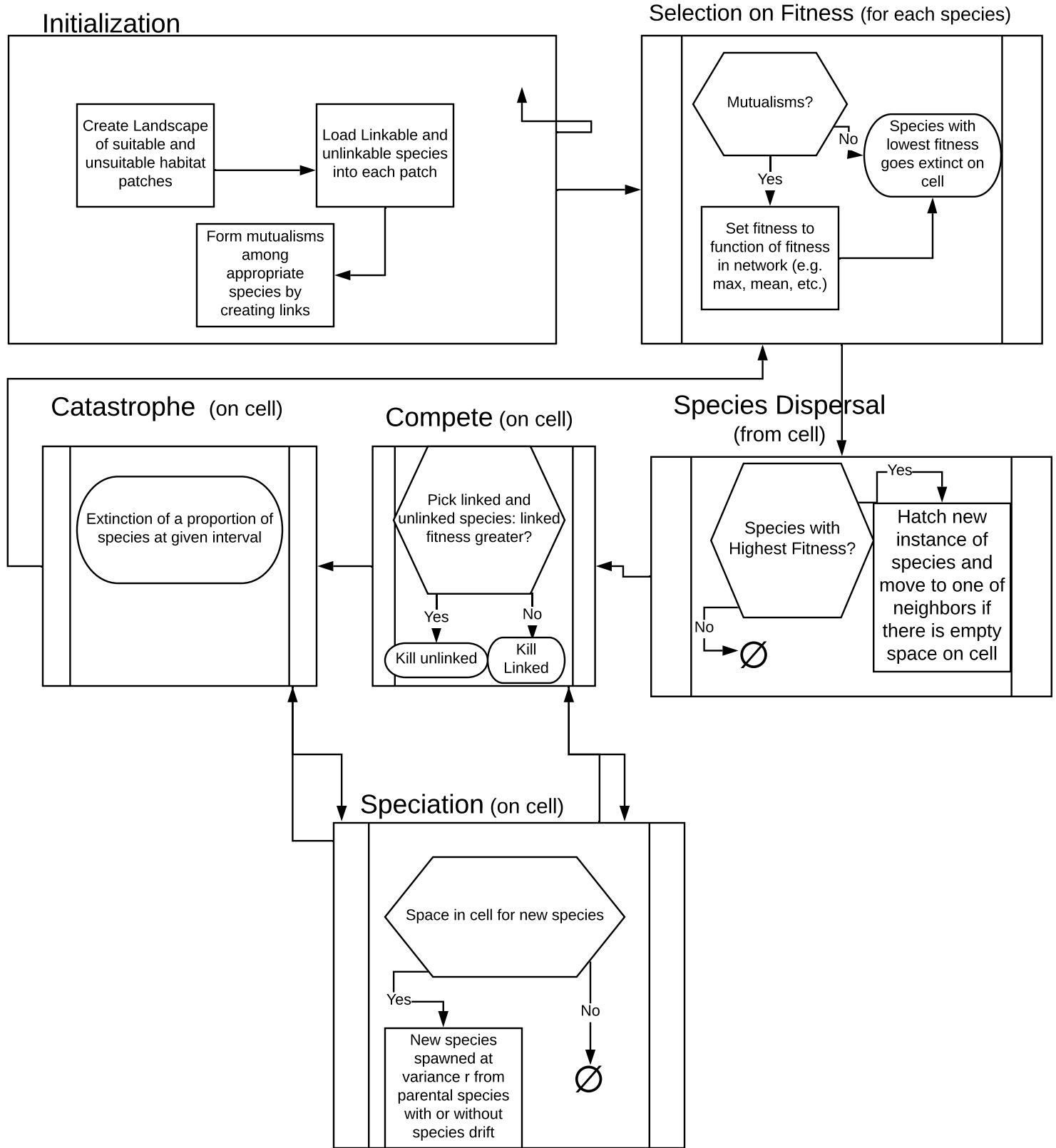


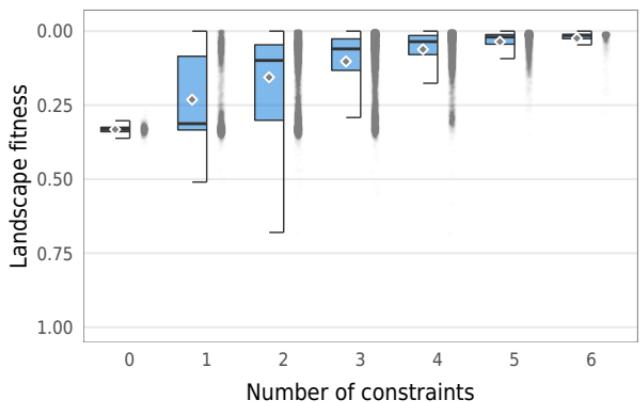
Fig. 3

Model Flow



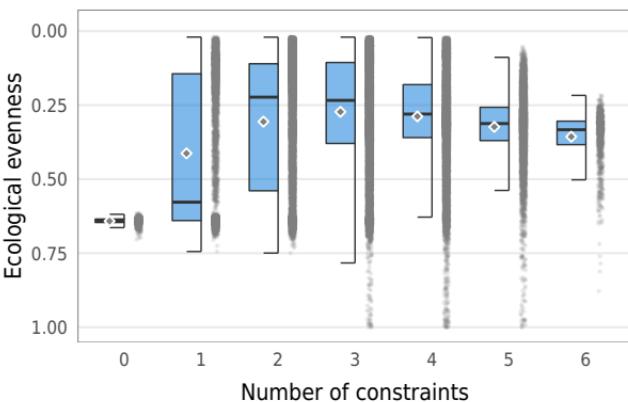
4A: Landscape fitness by number of constraints

Values near zero indicate high fitness



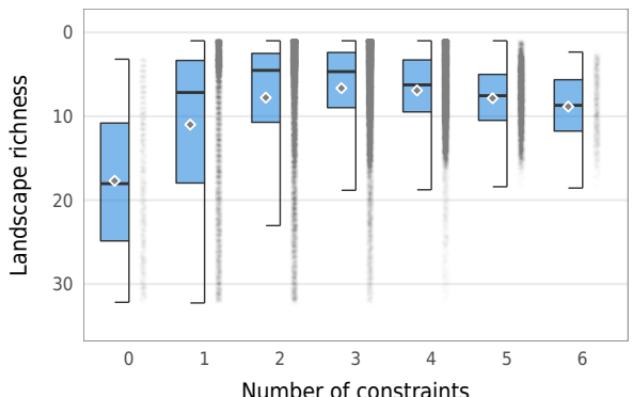
4B: Ecological evenness by number of constraints

Values near zero indicate high evenness

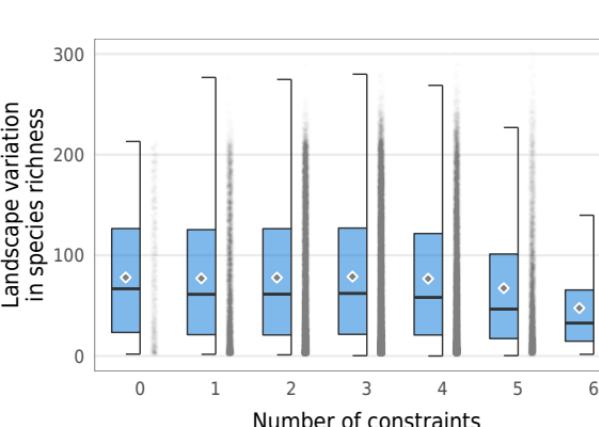


4C: Species richness by number of constraints

Values near zero indicate low species richness



4D: Species turnover by number of constraints

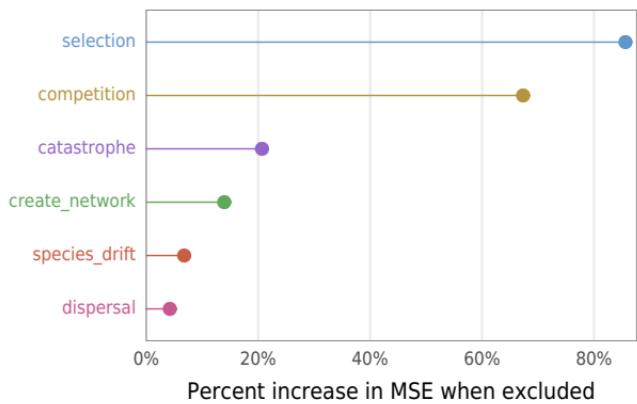


■ Latin hypercube samples

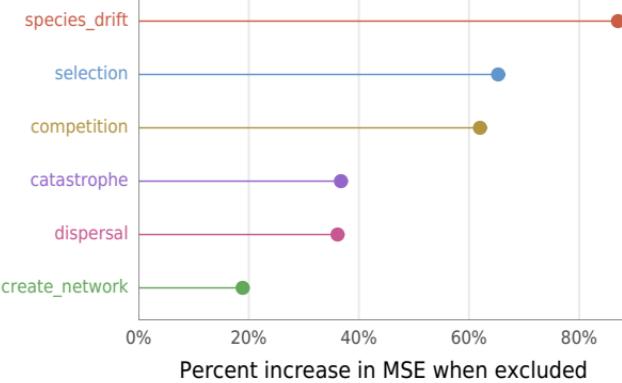
◆ = mean value

5A Constraint importance by landscape fitness

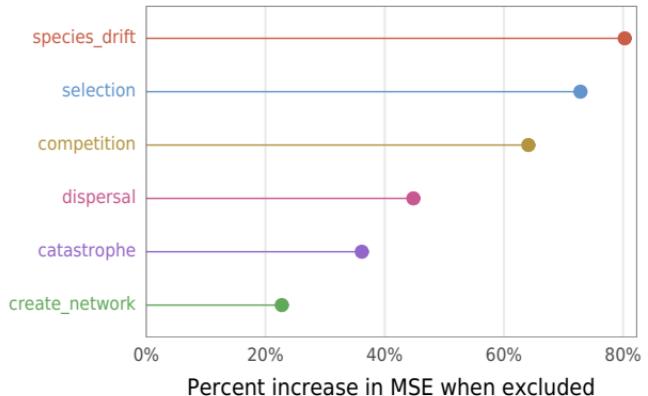
Higher values indicate greater variable importance



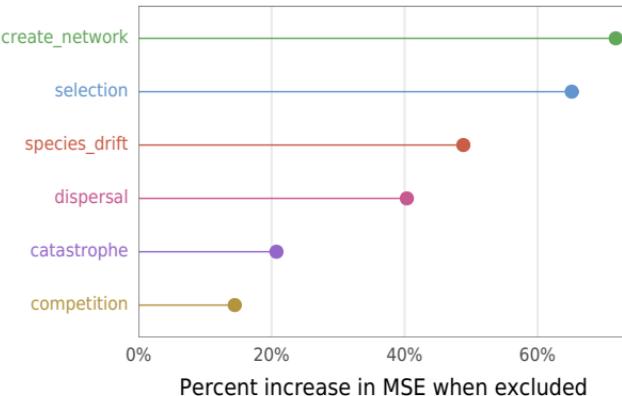
5B Constraint importance for ecological evenness



5C Constraint importance for species richness



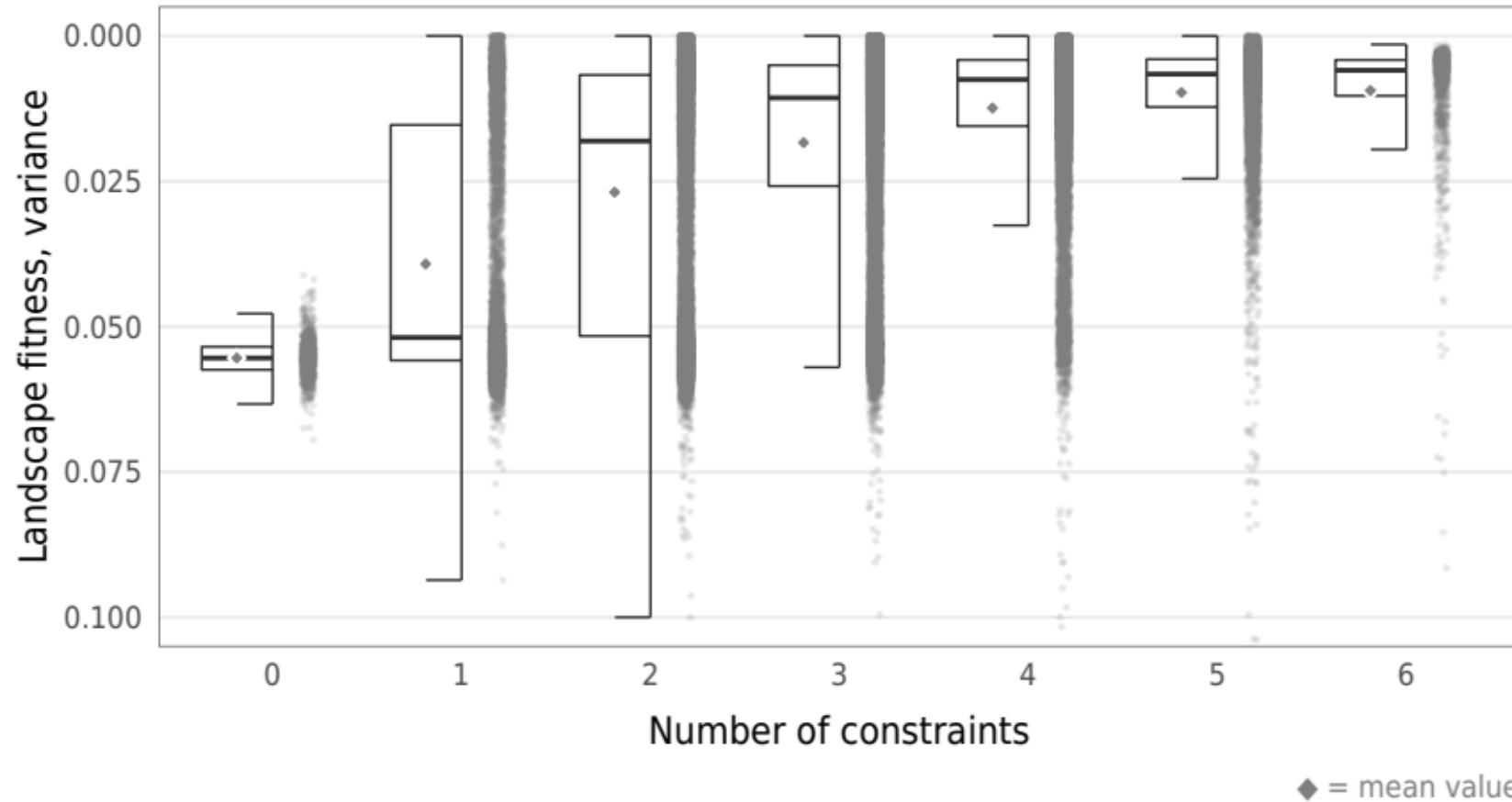
5D Constraint importance for species turnover



—●— Latin hypercube samples

6: Variance in landscape fitness by number of constraints

Latin hypercube samples



7: Variance in species richness by number of constraints

Turnover in linked species – turnover in unlinked species

