

Species occurrence as a function of both emergent biological traits
and environmental context

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Abstract

2

Introduction

⁴ All the world's a stage, And all the men and women merely players; They have their
exits and their entrances...

⁶ (Shakespeare, *As You Like It*, Act II, Scene VII)

A regional species pool is the set of species which form communities in a specific region. Local scale processes like resource competition only affect the regional species pool if all communities are affected. The taxonomic and functional composition of a regional species pool changes over time due to speciation, migration, extinction. How do species pools change over time as species are recruited or go extinct? When are specific species ecologies enriched or depleted in the species pool? How does global and regional environmental context affect the set of species ecotypes (e.g. guilds) in a regional species pool? All of these questions fall under a single umbrella of analysis of ecotypic diversity and diversification.

Functional diversity is frequently broken into or thought of as a set of guilds, which are a set of
16 species with similar sets of interactions and interactors (i.e. macroecology) (Bambach, 1977; Brown
and Maurer, 1989; Simberloff and Dayan, 1991; Valentine, 1969; Wilson, 1999). Species within a
18 guild are expected to have more similar macroecological dynamics than species in different guilds.
Building on the concept of guilds and a macroecological niche, Bush et al. (2007) presented a
20 three-dimensional construct, or ecocube, for describing the macroecological role of a marine
invertebrate species by their physical position (i.e. tiering), motility, and trophic role. Unique
22 combinations along the three ecological trait axes indicate which among the possible ecotypes are
observed. This approach has proven quite popular as it attempts to operationalize the guild concept
24 in terms of shared characteristics that are indicative of the type of interactions experience by
species of that macroecology (Bambach et al., 2007; Bush and Bambach, 2011; Bush et al., 2007;

26 Bush and Novack-Gottshall, 2012; Novack-Gottshall, 2007; Villéger et al., 2011), but the overall
utility of this approach is limited due to its condition as just a data type.

28 Previous analysis of mammal diversity and hypotheses as to the processes that have shaped it tend
to be through one or more of the following lenses: diversity of an entire system (e.g. continent)
30 (Alroy, 1996; Alroy et al., 2000; Figueirido et al., 2012; Liow et al., 2008), guild based (Janis et al.,
2004; Janis, 2008; Janis et al., 2000; Janis and Wilhelm, 1993; Jernvall and Fortelius, 2004; Pires
32 et al., 2015), clade based (Cantalapiedra et al., 2017; Fraser et al., 2015; Quental and Marshall,
2013; Silvestro et al., 2015; Slater, 2015), and environment based (Badgley and Finarelli, 2013;
34 Badgley et al., 2017; Blois and Hadly, 2009; Eronen et al., 2015; Fraser et al., 2015; Janis, 1993;
Janis and Wilhelm, 1993). Rarely are more than two of these lenses considered simultaneously, and
36 integration across the resulting diversity of observations and hypotheses tends to be based on
coincidence. One of the goals of this study is to present a framework for simultaneously analyzing a
38 diversity of hypotheses by integrating both species traits and environmental factors into a single
model in order to infer a more holistic multi-level picture of the processes which may have shaped
40 mammal species diversity and diversification.

The principle species trait considered in this study is a species' ecotype, defined here as the unique
42 combination of species dietary category and locomotor category (e.g. arboreal omnivore versus
unguligrade herbivore). These classifications can be considered analogous to guilds or unique
44 ecocube combinations as discussed above (Bambach et al., 2007; Bush and Bambach, 2011; Bush
et al., 2007). Species mass was also included as a species trait, but its inclusion is principally to
46 control for that effect on the other covariates that are the focus of this study.

Translating previous work into hypotheses applicable to this analysis is difficult for a variety of
48 reasons. Taxonomic groupings such as order or family are frequently invoked as an important factor
in many proposed hypotheses for how mammal diversity is structured (Janis, 2008; Janis and
50 Wilhelm, 1993; Pires et al., 2015; Quental and Marshall, 2013; Slater, 2015). Because taxonomic
grouping conflates both species macroecology with shared evolutionary history, there are few clear
52 ways to translate and operationalize these hypotheses in terms of macroecological change viewed

through the lens of species interactions. Hypotheses as to macroecological change viewed through
54 the lens of species interactions. Specifically, this issue arrises when trying to generalize previous
observations from taxonomy-based framework to ecology-based one.

56 There is little convincing evidence of any major or sudden cross-ecotypic or cross-taxonomic
turnover events in history of North American mammal diversity, unlike the Neogene record
58 European mammals (Alroy, 1996, 2009; Alroy et al., 2000; Eronen et al., 2015; Janis, 1993). Instead
of being concentrated in time, turnover has been found to be distributed through time. It is then
60 expected then that, for this analysis, turnover events or periods of rapid diversification or depletion
should not occur simultaneously for all ecotypes.

62 Jernvall and Fortelius (2004) found that for the Neogene of Europe the relative abundance of
mammal guilds was stable over time even in the face of high turnover rates, though they only
64 considered large bodied taxa from a small set of mammal orders. Similar results have been observed
for some taxonomic groups in North America CITATIONS. These results imply that there the
66 types of interactions happening in local communities observed over a region are constant over time
even if the interactors are constantly changing. MORE ABOUT DIVERSITY DEPENDENCE
68 HERE. WHAT DO PEOPLE THINK THE MAMMAL DIVERSITY CURVE REPRESENTS?
CAN BE ANYTHING IF YOU THINK ABOUT IT HARD ENOUGH.

70 The diversity history of ungulate herbivores has been characterized as more recently originating
taxa having longer legs, higher crowned teeth, and a shift from graze-dominated to
72 browse-dominated diets than their earlier originating counterparts (Cantalapiedra et al., 2017;
Fraser et al., 2015; Janis et al., 2004; Janis, 2008; Janis et al., 2000; Janis and Wilhelm, 1993); all of
74 which have all been attributed to some combination of environmental change itself or tectonic
activity driving environmental change (Badgley et al., 2017; Blois and Hadly, 2009; Eronen et al.,
76 2015; Janis, 2008). Additionally, it has been observed that these cursorial ungulate forms arose
prior to cursorial carnivore forms, an observation attributed to the reorganization of plant
78 communities towards the end of the Cenozoic and the latter emergence of “modern” environments
and communities (Janis and Wilhelm, 1993).

80 Within the canid guild of North America (e.g. plantigrade and digitigrade carnivores) there is
evidence that their diversity is self-regulating or somehow limited. Specifically, it has been proposed
82 that different canid clades have replaced each other as the dominate members of that
macroecological role within the species pool (Silvestro et al., 2015). A pattern of generally constant
84 diversity through time is also observed within the canid carnivore subguilds of hypercarnivore,
hypocarnivore, and mesocarnivores identified by Slater (2015) even in the face of constant species
86 turnover is consistent with limited possibility of increased diversity, even though there was no
evidence of diversity-dependence in trait (e.g. body size) evolution (Slater, 2015). THERE IS
88 MORE TO THIS STORY; CHECK THE WORK OF VANVALKENBURG; DANCES WITH
WOLF AVATARS?

90 There is some uncertainty and a lack of consensus as to the effect of species body size on mammal
diversity and aspects of the diversification processes, specifically extinction (Liow et al., 2008, 2009;
92 Smits, 2015; Tomiya, 2013). Species body size is frequently framed as an important biological
descriptor because of how it is correlated with other important and relevant traits such as
94 metabolic rate and home range size CITATIONS. It is also relatively easy to estimate for extinct
species using proxy measures and regression equations, as was done in this study (see below).
96 However, body size is normally considered without reference to other ecological descriptors of the
species (Liow et al., 2008), but see (Smits, 2015); this combined with the high amount of correlation
98 between life history traits and body size limits processed-based inference because the actual causal
mechanisms underlying an observed pattern are obscured or missing.
100 Smits (2015) found that the individual traits which form this study's ecotypes have strong effects
on mammal extinction risk. Omnivorous taxa were found to have, on average, a greater duration
102 than other dietary categories, while arboreal taxa were found to have a shorter duration than other
locomotor categories (Smits, 2015). Two possible scenarios that could yield this pattern were
104 proposed: the extinction risk faced by arboreal is constant and high or the Paleogene and Neogene
represent different regimes and extinction risk increased in the Neogene, thus driving up the
106 Cenozoic average extinction risk. These two possible explanations have clear and testable
predictions with respect to the diversity history of arboreal taxa: 1) the extinction risk arboreal

108 taxa increased in the Neogene compared to the Paleogene, driving the average extinction risk of
arboreal mammals up and leading to the loss of arboreal taxa from the species pool, or 2) if
110 arboreal taxa have just a generally higher extinction risk than other ecotypes but have maintained
a constant diversity for the Cenozoic. By inspecting the inferred diversity histories of the ecotypes,
112 it should be possible to distinguish amongst these hypotheses.

Fundamentally, all species respond differently to climate and environmental change (Blois and
114 Hadly, 2009). Macroecological patterns are emergent patterns due to the similarities among species
in how they respond to a similar “stimulus.”

116 The effect of climate on diversity and the diversification process has been the focus of considerable
research with a slight consensus favoring diversification being more biologically-mediated than
118 climate-mediated (Alroy, 1996; Alroy et al., 2000; Clyde and Gingerich, 1998; Figueirido et al.,
2012). However, differences in temporal and geographic scale seem to underly the contrast between
120 these two perspectives. For example when the mammal fossil record analyzed at small temporal and
geographic scales a correlation between diversity and climate are observable (Clyde and Gingerich,
122 1998). However, when the record is analyzed at the scale of the continent and most of the Cenozoic
there is no correlation with diversity and climate (Alroy et al., 2000). This results, however, does not
124 go against the idea that there may be short periods of correlation and that the correlation between
diversity and climate can change or even reverse direction over time; this type result means that
126 there is no single direction of correlation between diversity and climate (Figueirido et al., 2012).

In the case of a fluctuating correlation between diversity and climate it is hard to make the
128 argument for an actual causal link between the two without modeling the underlying ecological
differences between species; after all, species respond differently based on their individual ecologies
130 CITATION. When analysis is based on diversity or taxonomy alone no mechanisms are possible to
infer. Taxonomy, like body size, stands in for many important species traits to the point that
132 mechanistic or process based inference is impossible. While emergent patterns might correspond to
taxonomic grouping, this itself is an emergent phenomenon. Instead, by framing hypotheses in
134 terms of species traits and their environmental context, these emergent phenomenon can be

observed rather than assumed.

136 The climate history of the Cenozoic is generally characterized by a global cooling trend and the
development of polar ice-caps during the Neogene; there are, of course, a few notable exceptions to
138 this broad characterization (Cramer et al., 2011; Zachos et al., 2008, 2001). The environmental
context of North America for the Cenozoic is additionally characterized by an environmental
140 transition from the closed, partially forested environments of the Paleogene to the savannah and
grasslands environments of the Neogene (Blois and Hadly, 2009; Janis, 1993; Janis et al., 2000;
142 Strömberg, 2005).

A lot of the climate and environmental changes observed for North America have been attributed to
144 tectonic activity or uplift (Badgley and Finarelli, 2013; Blois and Hadly, 2009; Eronen et al., 2015;
Janis, 2008) CITATIONS. Tectonic uplift changes weather patterns (e.g. rain shadow) and
146 mobilizes grit into the environment CITATTIONS. Increased grit in the environment combined
with decreased rain fall is considered the primary reason behind the trend of increased hypsodonty,
148 or high crowned teeth, among herbivore groups over the Cenozoic of both North America and
Europe CITATIONS.

150 The Eocene-Oligocene transition has been observed to be associated with extinction of many
ungulate taxa (Janis, 2008). This boundary also marks the transition from the Paleogene to the
152 Neogene and from herbivores being browsing dominated to grazing dominated CITATION. WHEN
DO ICE CAPS DEVELOP? WHEN DO GRASSES DOMINATE? There is an observed stability in
154 estimates of global temperature from the E/O transition till the end of the Miocene called the
Mid-Miocene climatic optimum (Zachos et al., 2008, 2001). The Mid-Miocene climatic optimum is
156 bookended by periods of temperature decline. We would then expect that, for the Miocene,
turnover and other diversification events would most likely be due to biological interactions or
158 immigration and not biotic-abiotic interactions because of the constancy of the climate, and that
those groups that are driven primarily by environmental factors, the Miocene would be a period of
160 marked by an absence of major changes to diversity or the diversification process.

The environmental factors included in this study are estimates of global temperature and the

¹⁶² changing floral groups present in North America across the Cenozoic CITATIONS. These covariates
¹⁶⁴ were chosen because they provide high level characterizations of the environmental context of the
¹⁶⁶ entire North American regional species pool for most of the Cenozoic. Importantly, the effects of a
species ecotype on diversity are themselves modeled as functions of environmental factors (Fig. 1)
¹⁶⁸ allowing for inference as to how a species ecology can mediate selective pressures do to its
environmental context.

¹⁷⁰ Fourth-corner modeling is an approach to explaining the patterns of either species abundance or
presence/absence as a product of species traits, environmental factors, and the interaction between
¹⁷² traits and environment (Brown et al., 2014; Jamil et al., 2013; Pollock et al., 2012; Warton et al.,
2015); effectively uniting species distribution modeling (SDMs) with trait-based community
¹⁷⁴ assembly (CATS). In modern ecological studies, what is being modeled is species occurrences at
localities distributed across a region (Jamil et al., 2013; Pollock et al., 2012). In this study, what is
¹⁷⁶ being modeled is the pattern of species occurrence over time for most of the Cenozoic in North
America (Fig. 1). By adding an additional dimension (time) to the fourth-corner framework we can
gain better inference of how an instantaneous species pool (i.e. the Modern) is assembled over time.
These two approaches, modern and paleontological, are different views of the same
¹⁷⁸ three-dimensional pattern: species at localities over time. The temporal limitations of modern
ecological studies and difficulties with uneven spatial occurrences of fossils in paleontological studies
¹⁸⁰ means that these approaches are complimentary but reveal different patterns of how species are
distributed in time and space.

¹⁸² All observations, paleontological or modern, are made with uncertainty. With presence/absence
data this uncertainty comes from now knowing if an absence is a “true” absence or just a failure to
¹⁸⁴ observe (Foote, 2001; Foote and Sepkoski, 1999; Lloyd et al., 2011; Royle and Dorazio, 2008; Royle
et al., 2014; Wang and Marshall, 2016). For paleontological data, the incomplete preservation of
¹⁸⁶ whatever species were present into fossil form combined with incomplete sampling of what
organisms were actually fossilized means that the true times of origination or extinction may not be
¹⁸⁸ observed (Foote, 2001; Foote and Sepkoski, 1999; Wang et al., 2016; Wang and Marshall, 2016).

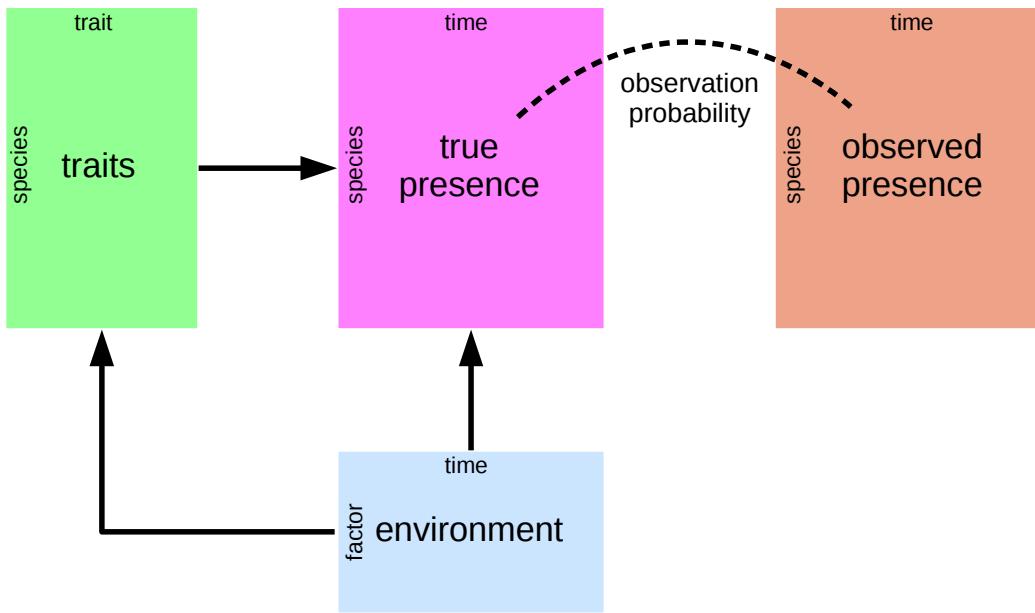


Figure 1: Conceptual diagram of the paleontological fourth corner problem. The observed presence matrix (orange) is the empirical presence/absence pattern for all species for all time points; this matrix is an incomplete observation of the “true” presence/absence pattern (purple). The estimated true presence matrix is modeled as a function of both environmental factors over time (blue) and multiple species traits (green). Additionally, the affect of environmental factors on species traits are also modeled as traits are expected to mediate the effects of a species environmental context. This diagram is based partially on material presented in Brown et al. (2014) and Warton et al. (2015).

Ultimately, the goals of this analysis are to understand when unique ecotypes enriched or
 190 depleted in the North American mammal regional species pool and how changes in ecotypic diversity are related to changes in species' environmental context. In the analyses done here, many
 192 covariates which describe both a species' macroecology and its environmental context are considered. In order to analyze this complex and highly structured data set, I developed a
 194 hierachal Bayesian model combing the forth-corner modeling approach with a model of an observation-occurrence or observation-originatation-extinction process. The complexity and nuance
 196 inherent in questions that are focus of this study, it is possible to consider and test a large number

of possible hypotheses. The hierarchical Bayesian modeling approach used here is appropriate for
198 mitigating complications arising from both this complexity and the plethora of testable hypotheses
(e.g. multiple comparisons, garden of forking paths) CITATIONS.

200 Materials and Methods

Taxon occurrences and species-level information

202 All fossil occurrence information used in this analysis was downloaded from the Paleobiology
Database (PBDB). The initial download restricted all occurrences to all Mammalia observed in
204 North America between the Maastrichtian (XX Mya) and Gelasian (XX Mya) stages. Occurrences
were then further limited to those occurring between 64 and 2 million years ago (Mya). Taxonomic,
206 stratigraphic, and ecological metadata for each occurrence and species was also downloaded. A new
download for a raw, unfiltered PBDB datafile following the same criterion used here is available at
208 <http://goo.gl/2s1geU>. The raw datafiles used as a part of this study, along with all code for
filtering and manipulating this download is available at GITHUB.

210 After download, the raw occurrence data was then sorted, cleaned, and manipulated
programmatically before analysis. Many species taxonomic assignments as present in the raw
212 PBDB data were updated for accuracy and consistency. For example, species classified in the order
Artiodactyla were reclassified as Cetartiodactyla. These re-assignments follow Smits (2015) which
214 were based on taxonomies present in the Encyclopedia of Life WEBSITE and Janis et al. (2008,
1998). All taxa who's life habit was classified as either volant (i.e. Chiroptera) or aquatic (e.g.
216 Cetacea) were excluded from this analysis because of both differences in fossilization potential and
environmental context as well as their lack of direct applicability to the study of terrestrial species
218 pools.

Species ecotype is defined here as the interaction between life habit and diet categories, the goal of
220 which is to classify species based on the manner with which they interact with their environment.
Most mammal species records in the PBDB have life habit and dietary category assignments. In

222 order to simplify interpretation, analysis, and per ecotype sample size these classifications were
 223 coarsened in a similar manner to (Smits, 2015) following Table 1. Additionally for this study, the
 224 life history category was further broken up to better reflect the diversity of mammal locomotor
 225 modes. Ground dwelling species locomotor categories were reassigned based on their ankle posture
 226 associated with their taxonomic group as described in Table 2 (Carrano, 1999). Ecotype categories
 227 with less than 10 total species of that combination were excluded, yielding a total of 18 observed
 228 ecotypes out of a possible 24.

Table 1: Species trait assignments in this study are a coarser version of the information available in the PBDB. Information was coarsened to improve per category sample size and uniformity and followed this table.

This study		PBDB categories
Diet	Carnivore	Carnivore
	Herbivore	Browser, folivore, granivore, grazer, herbivore.
	Insectivore	Insectivore.
	Omnivore	Frugivore, omnivore.
Locomotor	Arboreal	Arboreal.
	Ground dwelling	Fossorial, ground dwelling, semifossorial, saltatorial.
	Scansorial	Scansorial.

Table 2: Ankle posture assignment as based on taxonomy. Assignments are based on (Carrano, 1999). Taxonomic groups are presented alphabetically and without reference for the nestedness of families in orders. As such, do not infer higher-order structure from this table.

Order	Family	Stance
	Ailuridae	plantigrade
	Allomyidae	plantigrade
	Amphicyonidae	plantigrade
	Amphilemuridae	plantigrade
	Anthracotheriidae	digitigrade
	Antilocapridae	unguligrade
	Apheliscidae	plantigrade
	Aplopontidae	plantigrade

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Table 2 – continued from previous page

Order	Family	Stance
	Apternodontidae	scansorial
	Arctocyonidae	unguligrade
	Barbourofelidae	digitigrade
	Barylambdidae	plantigrade
	Bovidae	unguligrade
	Camelidae	unguligrade
	Canidae	digitigrade
	Cervidae	unguligrade
	Cimolodontidae	scansorial
	Coryphodontidae	plantigrade
	Cricetidae	plantigrade
	Cylindrodontidae	plantigrade
	Cyriacotheriidae	plantigrade
	Dichobunidae	unguligrade
Dinocerata		unguligrade
	Dipodidae	digitigrade
	Elephantidae	digitigrade
	Entelodontidae	unguligrade
	Eomyidae	plantigrade
	Erethizontidae	plantigrade
	Erinaceidae	plantigrade
	Esthonychidae	plantigrade
	Eutypomyidae	plantigrade
	Felidae	digitigrade
	Florentiamyidae	plantigrade

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Table 2 – continued from previous page

Order	Family	Stance
	Gelocidae	unguligrade
	Geolabididae	plantigrade
	Glyptodontidae	plantigrade
	Gomphotheriidae	unguligrade
	Hapalodectidae	plantigrade
	Heteromyidae	digitigrade
	Hyaenidae	digitigrade
	Hyaenodontidae	digitigrade
	Hypertragulidae	unguligrade
	Ischyromyidae	plantigrade
	Jimomyidae	plantigrade
Lagomorpha		digitigrade
	Leptictidae	plantigrade
	Leptochoeridae	unguligrade
	Leptomerycidae	unguligrade
	Mammutidae	unguligrade
	Megalonychidae	plantigrade
	Megatheriidae	plantigrade
	Mephitidae	plantigrade
	Merycoidodontidae	digitigrade
Mesonychia		unguligrade
	Mesonychidae	digitigrade
	Micropternodontidae	plantigrade
	Mixodectidae	plantigrade
	Moschidae	unguligrade

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Table 2 – continued from previous page

Order	Family	Stance
	Muridae	plantigrade
	Mustelidae	plantigrade
	Mylagaulidae	fossorial
	Mylodontidae	plantigrade
	Nimravidae	digitigrade
	Nothrotheriidae	plantigrade
Notoungulata		unguligrade
	Oromerycidae	unguligrade
	Oxyaenidae	digitigrade
	Palaeomerycidae	unguligrade
	Palaeoryctidae	plantigrade
	Pampatheriidae	plantigrade
	Pantolambdidae	plantigrade
	Peritychidae	digitigrade
Perissodactyla		unguligrade
	Phenacodontidae	unguligrade
Primates		plantigrade
	Procyonidae	plantigrade
	Proscalopidae	plantigrade
	Protoceratidae	unguligrade
	Reithroparamyidae	plantigrade
	Sciuravidae	plantigrade
	Sciuridae	plantigrade
	Simimyidae	plantigrade
	Soricidae	plantigrade

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Table 2 – continued from previous page

Order	Family	Stance
	Suidae	digitigrade
	Talpidae	fossorial
	Tayassuidae	unguligrade
	Tenrecidae	plantigrade
	Titanoideidae	plantigrade
	Ursidae	plantigrade
	Viverravidae	plantigrade
	Zapodidae	plantigrade

²³⁰ Estimates of species mass used as a covariate in this study were sourced from multiple databases and papers, especially those focusing on similar macroevolutionary or macrecological questions
²³² (Brook and Bowman, 2004; Freudenthal and Martín-Suárez, 2013; McKenna, 2011; Raia et al., 2012;
Smith et al., 2004; Tomiya, 2013); this is similar to what was done in Smits (2015). When species
²³⁴ mass was not available, proxy measures were used and then transformed into estimates of mass. For example, given a measurement of a mammal tooth size, it is possible and routine to estimate its
²³⁶ mass given some regression equation. The PBDB has one or more body part measures for many species. These were used as body size proxies for many species, as was the case in Smits (2015).
²³⁸ Mass was log-transformed and then mean-centered and rescaled by dividing by two-times its standard deviation; this insures that the magnitude of effects for both continuous and discrete
²⁴⁰ covariates are directly comparable (Gelman, 2008; Gelman and Hill, 2007).

All fossil occurrences from 64 to 2 million years long (Mya) were binned into 31 2 million year (My)
²⁴² bins. This temporal length was chosen because it is approximately the resolution of the North American mammal fossil record (Alroy, 1996; Alroy et al., 2000; Marcot, 2014).

Table 3: Regression equations used in this study for estimating body size. Equations are presented with reference to taxonomic grouping, part name, and reference.

Group	Equation	log(Measurement)	Source
General	$\log(m) = 1.827x + 1.81$	lower m1 area	Legendre (1986)
General	$\log(m) = 2.9677x - 5.6712$	mandible length	?
General	$\log(m) = 3.68x - 3.83$	skull length	?
Carnivores	$\log(m) = 2.97x + 1.681$	lower m1 length	?
Insectivores	$\log(m) = 1.628x + 1.726$	lower m1 area	?
Insectivores	$\log(m) = 1.714x + 0.886$	upper M1 area	?
Lagomorph	$\log(m) = 2.671x - 2.671$	lower toothrow area	Tomiya (2013)
Lagomorph	$\log(m) = 4.468x - 3.002$	lower m1 length	Tomiya (2013)
Marsupials	$\log(m) = 3.284x + 1.83$	upper M1 length	?
Marsupials	$\log(m) = 1.733x + 1.571$	upper M1 area	?
Rodentia	$\log(m) = 1.767x + 2.172$	lower m1 area	Legendre (1986)
Ungulates	$\log(m) = 1.516x + 3.757$	lower m1 area	?
Ungulates	$\log(m) = 3.076x + 2.366$	lower m2 length	?
Ungulates	$\log(m) = 1.518x + 2.792$	lower m2 area	?
Ungulates	$\log(m) = 3.113x - 1.374$	lower toothrow length	?

244 Environmental and temporal covariates

The environmental covariates used in this study are collectively referred to as group-level covariates
 246 because they predict the response of a “group” of individual-level observations (i.e. species
 occurrences of an ecotype). Additionally, these covariates are defined for temporal bins and not the
 248 species themselves; as such they predict the parts of each species occurrence history. The
 group-level covariates in this study are two global temperature estimates and the Cenozoic “plant
 250 phases” defined by Graham (2011). Global temperature across most of the Cenozoic was calculated
 from Mg/Ca isotope record from deep sea carbonates (Cramer et al., 2011). Mg/Ca based
 252 temperature estimates are preferable to the frequently used $\delta^{18}\text{O}$ temperature proxy (Alroy et al.,
 2000; Figueirido et al., 2012; Zachos et al., 2008, 2001) because Mg/Ca estimates do not conflate
 254 temperature with ice sheet volume and depth/stratification changes. The former is particularly
 important to this analysis as the current polar ice-caps appeared and grew over the Cenozoic.
 256 These properties make Mg/Ca based temperature estimates preferable for macroevolutionary and
 macroecological studies (Ezard et al., 2016). Two aspects of the Mg/Ca-based temperature curve
 258 were included in this analysis: mean and range. Both were calculated as the mean of all respective

Table 4: Definitions of the start and stop times of the three plant phases used this study as defined by Graham (2011).

Plant phase	Phase number	Start	Stop
Paleocene-Eocene	1	66	50
Eocene-Miocene	2	50	16
Miocene-Pleistocene	3	16	2

estimates for each 2 My temporal bins. Both mean and range were then rescaled by subtracting the
 260 mean and then dividing by twice the standard deviation.

The second set of environmental factors included in this study are the Cenozoic plant phases defined
 262 in Graham (2011). Graham’s plant phases are holistic descriptors of the taxonomic composition of
 12 ecosystem types, which plants are present at a given time, and the relative modernity of those
 264 plant groups with younger phases representing increasingly modern taxa (Graham, 2011). Graham
 (2011) defines four intervals from the Cretaceous to the Pliocene, though only three of these
 266 intervals are included in this analysis. Graham’s plant phases was included as a series of “dummy
 variables” encoding the three phases included in this analysis. This means that the first phase is
 268 synonymous with the intercept and subsequent phases are defined by their differences from the first
 phase. The temporal boundaries of these plant phases are defined in Table 4.

270 Modelling species occurrence

Two different models were used in this study: a pure-presence model and a birth-death model. Both
 272 models at their core are a hidden Markov model where the latent process has an absorbing state
 (Allen, 2011). The difference between these two models is if the probabilities of a species originating
 274 or surviving are considered equal or different (Table 5). Something that is important to realize is
 that while there are only two state “codes” in a presence-absence matrix (i.e. 0/1), there are in fact
 276 three states in a birth-death model: not having originated (yet), extant, and extinct. The last of
 these is the absorbing state, as once a species has gone extinct it cannot re-originate (Allen, 2011);
 278 this is made obvious in the transition matrices as the probability of an extinct species changing
 states is 0 (Table 5). See below for parameter explanations (Tables 7, and 8).

		State at $t + 1$		
		0_{never}	1	$0_{extinct}$
State at t	0_{never}	$1 - \theta$	θ	0
	1	0	θ	$1 - \theta$
	$0_{extinct}$	0	0	1

(a) Pure-presence

		State at $t + 1$		
		0_{never}	1	$0_{extinct}$
State at t	0_{never}	$1 - \phi$	ϕ	0
	1	0	π	$1 - \pi$
	$0_{extinct}$	0	0	1

(b) Birth-death

Table 5: Transition matrices for the pure-presence (5a) and birth-death (5b) models. Both of these models share the core machinery of discrete-time birth-death processes but make distinct assumptions about the equality of originating and surviving (Eq. 2, and 3). Note also that while there are only two state “codes” (0, 1), there are in fact three states: never having originated 0_{never} , present 1, extinct $0_{extinct}$ (Allen, 2011).

280 Data augmentation

All empirical presence/absence observations are potentially incomplete or observed with error. The hidden Markov model at the core of this analysis allows for observed absences to be used meaningfully to estimate the number of unobserved species. Of concern in this analysis is the unknown “true” size of the dataset; how many species could have actually been observed? While many species have been observed, the natural incompleteness of all observations, especially in the case of paleontological data, there are obviously many species which were never sampled (Royle and Dorazio, 2008; Royle et al., 2007).

Let N by the total number of observed species, M be the upper limit of possible species that could have existed given some model of species occurrence, and N^* is the number of all-zero histories added to the presence absence matrix y where $N^* = M - N$. This approach assumes that $\hat{N} \sim \text{Binomial}(M, \psi)$ where \hat{N} is the estimated “true” number of species and ψ is the probability that any augmented species should actually be “present.” Because M is user defined, this approach effectively gives ψ a uniform prior over N to M (Royle and Dorazio, 2008). For this study, $M = \lfloor 1.25 \times N \rfloor$.

Data imputation is the process of estimating missing data for partially observed covariates given the other fully-observed observations and some model (Gelman and Hill, 2007; Rubin, 1996), this is simple in a Bayesian context because data are also parameters (Gelman et al., 2013). Augmented species are fully imputed species and thus have no known mass so a mass estimate must be imputed

Table 6: Observation parameters

Parameter	dimensions	explanation
y	$N \times T$	observed species presence/absence
z	$N \times T$	“true” species presence/absence
p	T	probability of observing a species that is present at time t
m	N	species log mass, rescaled
α_0	1	average log-odds of p
α_1	1	change in average log-odds of p per change mass
r	T	difference from α_0 associated with time t
σ	1	standard deviation of r

for each possible species (Royle and Dorazio, 2012). Assuming that mass values for augmented
300 species are from the same distribution as observed species, the distribution of observed mass values
are estimated as part of the model and new mass values are then generated from this distribution.
302 This approach is an example of imputing covariate information that is missing completely at
random (Gelman and Hill, 2007; Royle and Dorazio, 2012). Because log mass values are rescaled as
304 a part of this study, the body mass distribution is already known ($\mathcal{N}(0, 0.5)$) the body mass of the
augmented species are generated by simple random draws from this distribution. In addition to
306 body mass information, the augmented species need an ecotype classification. Because these species
are completely unknown, they were all classified as “augmented” to indicate their unknown biology.
308 This classification has no biological interpretation.

Observation process

310 The type of hidden Markov model used in this study has three characteristic probabilities:
probability p of observing a species given that it is present, probability ϕ of a species surviving from
312 one time to another, and probability π of a species first appearing (Royle and Dorazio, 2008). In
this formulation, the probability of a species going extinct is $1 - \pi$. For the pure-presence model
314 $\phi = \pi$, while for the birth-death model $\phi \neq \pi$.

The probability of observing a species that is present p is modeled as a logistic regression was a
316 time-varying intercept and species mass as a covariate. The effect of species mass on p was assumed
linear and constant over time and given a prior reflecting a possible positive relationship; these

Table 7: Parameters for the model of presence in the pure-presence model

Parameter	dimensions	explanation
z	$N \times T$	“true” species presence/absence
θ	$N \times T - 1$	probability of $z = 1$
a	$T - 1 \times D$	ecotype-varying intercept; mean value of log-odds of θ
m	N	species log mass, rescaled
b_1	1	effect of species mass on log-odds of θ
b_2	1	effect of species mass, squared, on log-odds of θ
U	$T \times D$	matrix of group-level covariates
γ	$U \times D$	matrix of group-level regression coefficients
Σ	$D \times D$	covariance matrix of a
Ω	$D \times D$	correlation matrix of a
τ	D	vector of standard deviations for each ecotype a_d

318 assumptions are reflected in the structure of the model Equation 1, presented here:

$$\begin{aligned}
 y_{i,t} &\sim \text{Bernoulli}(p_{i,t} z_{i,t}) \\
 p_{i,t} &= \text{logit}^{-1}(\alpha_0 + \alpha_1 m_i + r_t) \\
 r_t &\sim \mathcal{N}(0, \sigma).
 \end{aligned} \tag{1}$$

The parameters associated with Equation 1 are described in Table 6.

320 Pure-presence process

For the pure-presence model there is only a single probability dealing with the presence of a species
 322 θ (Table 5a). This probability was modeled as multi-level logistic regression with both species-level
 and group-level covariates (Gelman et al., 2013; Gelman and Hill, 2007). The parameters associated
 324 with pure-presence model are presented in Table 7 and the full sampling statement in Equation 2.

The species-level of the model (Eq. 2) is a logistic regression where the intercept varies by ecotype.
 326 Additionally, species mass was included as a covariate associated with two regression coefficients
 allowing a quadratic relationship with log-odds of occurrence. This assumption is based on the
 328 known distribution of mammal body masses where species with intermediate mass values are more
 common than either small or large bodied species. These assumptions are also reflected in the

330 choice of priors for these regression coefficients.

The values of each ecotype's intercept are themselves modeled as regressions using the group-level
332 covariates associated with environmental context. Each of these regressions has an associated
variance of possible values of each ecotype's intercept (Gelman and Hill, 2007). In addition, the
334 covariances between ecotype intercepts, given this group-level regression, are modeled (Gelman and
Hill, 2007).

336 All parameters not modeled elsewhere were given weakly informative priors (Gelman et al., 2013)
CITATION STAN MANUAL STATISTICAL RETHINKING. Weakly informative means that
338 priors do not necessarily encode actual prior information but instead help regularize or weakly
constrain posterior estimates. These priors have a concentrated probability density around and near
340 zero; this has the effect of tempering our estimates and help prevent overfitting the model to the
data (Gelman et al., 2013) CITATION STAN MANUAL STATISTICAL RETHINKING. The
342 general line of thinking behind this approach is that a result of 0 or “no effect” is more preferable
to a wrong or extremely weak result. The sampling statement, excluding the imputation of body
344 mass associated with the augment species, is as follows

$$\begin{aligned} y_{i,t} &\sim \text{Bernoulli}(p_{i,t} z_{i,t}) & \alpha_0 &\sim \mathcal{N}(0, 1) \\ p_{i,t} &= \text{logit}^{-1}(\alpha_0 + \alpha_1 m_i + r_t) & \alpha_1 &\sim \mathcal{N}(1, 1) \\ r_t &\sim \mathcal{N}(0, \sigma) & \sigma &\sim \mathcal{N}^+(1) \\ z_{i,1} &\sim \text{Bernoulli}(\rho) & b_1 &\sim \mathcal{N}(0, 1) \\ z_{i,t} &\sim \text{Bernoulli}(\theta_{i,t}) & b_2 &\sim \mathcal{N}(-1, 1) \\ \theta_{i,t} &= \text{logit}^{-1}(a_{t,j[i]} + b_1 m_i + b_2 m_i^2) & \gamma &\sim \mathcal{N}(0, 1) \\ a &\sim \text{MVN}(u\gamma, \Sigma) & \tau &\sim \mathcal{N}^+(1) \\ \Sigma &= \text{diag}(\tau)\Omega\text{diag}(\tau) & \Omega &\sim \text{LKJ}(2) \end{aligned} \tag{2}$$

Birth-death process

- ³⁴⁶ In the birth-death version of the model, $\phi \neq \pi$ and so each of these probabilities are modeled
 separately but each is handled in a similar manner to how θ is modeled in the pure-presence model
³⁴⁸ (Eq. 2, Table 5b). The parameters associated with the birth-death presence model are presented in
 Table 8 and the full sampling statement, including observation (Eq. 1), is described in Equation 3:

$$\begin{aligned}
 y_{i,t} &\sim \text{Bernoulli}(p_{i,t} z_{i,t}) & \Sigma^\phi &= \text{diag}(\tau^\phi) \Omega^\phi \text{diag}(\tau^\phi) \\
 p_{i,t} &= \text{logit}^{-1}(\alpha_0 + \alpha_1 m_i + r_t) & \Sigma^\pi &= \text{diag}(\tau^\pi) \Omega^\pi \text{diag}(\tau^\pi) \\
 r_t &\sim \mathcal{N}(0, \sigma) & \rho &\sim U(0, 1) \\
 \alpha_0 &\sim \mathcal{N}(0, 1) & b_1^\phi &\sim \mathcal{N}(0, 1) \\
 \alpha_1 &\sim \mathcal{N}(1, 1) & b_1^\pi &\sim \mathcal{N}(0, 1) \\
 \sigma &\sim \mathcal{N}^+(1) & b_2^\phi &\sim \mathcal{N}(-1, 1) \\
 z_{i,1} &\sim \text{Bernoulli}(\phi_{i,1}) & b_2^\pi &\sim \mathcal{N}(-1, 1) \\
 z_{i,t} &\sim \text{Bernoulli} \left(z_{i,t-1} \pi_{i,t} + \sum_{x=1}^t (1 - z_{i,x}) \phi_{i,t} \right) & \gamma^\phi &\sim \mathcal{N}(0, 1) \\
 \phi_{i,t} &= \text{logit}^{-1}(a_{t,j[i]}^\phi + b_1^\phi m_i + b_2^\phi m_i^2) & \gamma^\pi &\sim \mathcal{N}(0, 1) \\
 \pi_{i,t} &= \text{logit}^{-1}(a_{t,j[i]}^\pi + b_1^\pi m_i + b_2^\pi m_i^2) & \tau^\phi &\sim \mathcal{N}^+(1) \\
 a^\phi &\sim \text{MVN}(U\gamma^\phi, \Sigma^\phi) & \tau^\pi &\sim \mathcal{N}^+(1) \\
 a^\pi &\sim \text{MVN}(U\gamma^\pi, \Sigma^\pi) & \Omega^\phi &\sim \text{LKJ}(2) \\
 && \Omega^\pi &\sim \text{LKJ}(2).
 \end{aligned} \tag{3}$$

- ³⁵⁰ Similar to the pure-presence model, both ϕ and π are modeled as logistic regressions with
 varying-intercept and one covariate associated with two parameters. The possible relationships
³⁵² between mass and both ϕ and π are reflected in the parameterization of the model and choice of
 priors (Eq. 3).
³⁵⁴ The intercepts of ϕ and π both vary by species ecotype and those values are themselves the product
 of group-level regression using environmental factors as covariates (Eq. 3); this is identical to the

Table 8: Parameters for the model of presence in the pure-presence model

Parameter	dimensions	explanation
z	$N \times T$	“true” species presence/absence
ϕ	$N \times T$	probability of $z_{-,t} = 1 z_{-,t-1} = 0$; origination
π	$N \times T - 1$	probability of $z_{-,t} = 1 z_{-,t-1} = 1$; survival
a^ϕ	$T - 1 \times D$	ecotype-varying intercept; mean value of log-odds of θ
a^π	$T - 1 \times D$	ecotype-varying intercept; mean value of log-odds of θ
m	N	species log mass, rescaled
b_1^ϕ	1	effect of species mass on log-odds of ϕ
b_1^π	1	effect of species mass on log-odds of π
b_2^ϕ	1	effect of species mass, squared, on log-odds of ϕ
b_2^π	1	effect of species mass, squared, on log-odds of π
U	$T \times D$	matrix of group-level covariates
γ^ϕ	$U \times D$	matrix of group-level regression coefficients
γ^π	$U \times D$	matrix of group-level regression coefficients
Σ^ϕ	$D \times D$	covariance matrix of a^ϕ
Σ^π	$D \times D$	covariance matrix of a^π
Ω^ϕ	$D \times D$	correlation matrix of a^ϕ
Ω^π	$D \times D$	correlation matrix of a^π
τ^ϕ	D	vector of standard deviations for each ecotype a_d^ϕ
τ^π	D	vector of standard deviations for each ecotype a_d^π

356 pure presence model (Eq. 2).

Posterior inference and model adequacy

358 Computer programs that implement joint posterior inference for the above models (Eqs. 2, 3) were
written in the probabilistic programming language Stan CITATION. The models used here both
360 feature latent discrete parameters in the large matrix z (Tables 6, 7, 8; Eqs. 1, 2, 3). All methods
for posterior inference implemented in Stan are derivative based which causes complications for
362 actually implementing the above models because integers do not have derivatives. Instead of
implementing a latent discrete parameterization, the log posterior probabilities of all possible states
364 of the latent parameters z were calculated and summed (i.e. marginalized).

Species durations at minimum range-through from the FAD to the LAD, but the incompleteness of
366 all observations means that the actual time of origination or extinction is unknown. The
marginalization approach used here means that the probabilities all possible histories for a species

368 are calculated, from the end members of the species having existed for the entire study interval and
 370 the species having only existed between the directly observed FAD and LAD to all possible
 intermediaries (Fig 2) CITATION STAN MANUAL. This process is identical, language-wise, to
 assuming range-through and then estimating the possibility of range extension due to incomplete
 372 sampling.

	Time Bin							
	1	2	3	4	5	6	7	8
Observed	0	0	0	1	0	1	1	0
-----	-----	-----	-----	-----	-----	-----	-----	-----
Certain	?	?	?	1	1	1	1	?
.....
Potential	0	0	0	1	1	1	1	0
Potential	0	0	1	1	1	1	1	0
Potential	1	1	1	1	1	1	1	0
Potential	0	0	0	1	1	1	1	1
Potential	0	0	1	1	1	1	1	1
Potential	0	1	1	1	1	1	1	1
Potential	1	1	1	1	1	1	1	1

Figure 2: Conceptual figure of all possible occurrence histories for an observed species. The first row represents the observed presence/absence pattern for a single species at eight time points. The second row corresponds to the known aspects of the “true” occurrence history of that species. The remaining rows correspond to all possible occurrence histories that are consistent with the observed data. The process of parameter marginalization described in the text

The combined size of the dataset and large number of parameters in both models (Eqs. 2, 3),
 374 specifically the total number of latent parameters that are the matrix z , means that stochastic
 approximate posterior inference is computationally very slow even using NUTS based HMC
 376 implemented. Instead, an approximate Bayesian approach was used: variational inference. A
 recently developed automatic variational inference algorithm called “automatic differentiation
 378 variational inference” (ADVI) is implemented in Stan and was used here CITATION. ADVI
 assumes that the posterior is Gaussian but still yields a true Bayesian posterior; this assumption is

similar to quadratic approximation of the likelihood function commonly used in maximum likelihood based inference CITATION. The principal limitation of assuming the joint posterior is Gaussian is that the true topology of the log-posterior isn't estimated; this is a particular burden for scale parameters which are bound to be positive (e.g. standard deviation).

Of additional concern for posterior inference is the partial identifiability of observation parameters $p_{t=1}$ and $p_{t=T}$ (Royle and Dorazio, 2008). This issue means that the estimates of sampling probabilities at the “edges” of the time series cannot fully be estimated because there are no known “gaps” in species occurrence histories that are guaranteed to be filled. Instead, the values of the first and final columns of the “true” presence-absence matrix z for those observations that do not already have presences in the observed presence-absence matrix y cannot be estimated (Royle and Dorazio, 2008). The hierarchical modeling approach used here helps mitigate this problem by pulling the values of $p_{t=1}$ and $p_{t=T}$ towards the overall mean of p (Gelman et al., 2013), and in fact this approach might be more analytically sound than the more ad-hoc approaches that are occasionally used to overcome this hurdle (Royle and Dorazio, 2008). Additionally, because $p_{t=1}$ and $p_{t=T}$ are only partially identifiable, estimates of occurrence θ and origination ϕ at $t = 1$ and estimates of θ , ϕ and survival pi at $t = T$ may suffer from similar edge effects. Again, the hierarchical modeling approach used here may help correct for this reality by drawing these estimates towards the overall means of those parameters.

After fitting both models (Eqs. 2, 3) using ADVI, model adequacy and quality of fit was assessed using a simple posterior predictive check CITATION CITATION. Because all Bayesian models are inherently generative, simulations of new data sets is “free” CITATION. By simulating 100 theoretical data sets using the observed covariate information the congruence between predictions made by the model and the observed empirical data can be assessed. These datasets are simulated by starting with the observed states of the presence-absence matrix at $t = 1$; from their, the time series roll forward as stochastic processes with covariate information given from the empirical observations. Importantly, this is fundamentally different from observing the posterior estimates of the “true” presence-absence matrix z . The posterior predictive check used in this study is to compare the observed average number of observations per species to a distribution of simulated

⁴⁰⁸ averages; if the empirically observed value sits in the middle of the distribution than the model can
be considered adequate in reproducing the observed number of occurrences per species.

- ⁴¹⁰ The ADVI assumption of a purely Gaussian posterior limits the utility and accuracy of the posterior predictive checks because parameter estimates do not reflect the true posterior
- ⁴¹² distribution and are instead just an approximation CITATIONS. Because of this, posterior predictive estimates are themselves only approximate checks of model adequacy. The posterior
- ⁴¹⁴ predictive check that is used in this study focuses on mean occurrence and not to any scale parameters that might be most affected by the ADVI assumptions.
- ⁴¹⁶ Given parameter estimates, diversity and diversification rates are estimated through posterior predictive simulations. Given the observed presence-absence matrix y , estimates of the true
- ⁴¹⁸ presence-absence matrix z can be simulated and the distribution of possible occurrence histories can be analyzed. This is conceptually similar to marginalization where the probability of each
- ⁴²⁰ possible occurrence history is estimated (Fig. 2).

The posterior distribution of z gives the estimate of standing diversity N_t^{stand} for all time points as

$$N_t^{stand} = \sum_{i=1}^M z_{i,t}. \quad (4)$$

- ⁴²² Given estimates of N^{stand} for all time points, the estimated number of originations O_t are be estimated as

$$O_t = \sum_{i=1}^M z_{i,t} = 1 | z_{i,t-1} = 0 \quad (5)$$

- ⁴²⁴ and number of extinctions E_t estimated as

$$E_t = \sum_{i=1}^M z_{i,t} = 0 | z_{i,t-1} = 1. \quad (6)$$

Per-capita growth D^{rate} , origination O^{rate} and extinction E^{rate} rates are then calculated as

$$\begin{aligned} O_t^{rate} &= \frac{O_t}{N_{t-1}^{stand}} \\ E_t^{rate} &= \frac{E_t}{N_{t-1}^{stand}} \\ D_t^{rate} &= O_t^{rate} - E_t^{rate}. \end{aligned} \tag{7}$$

426 Results

The results of the analyses described above take one of two forms: direct inspection of parameter
428 posterior estimates from both models, and downstream estimates of diversity and diversification
rates based on posterior predictive simulations from the birth-death model; the reason for which is
430 explained below in the comparison of the models' posterior predictive check results.

432 Comparing parameter estimates from the pure-presence and birth-death models

Comparison of the posterior predictive results from the pure-presence and birth-death models
434 reveals a striking difference in performance of either model to predict the structure of the
underlying data (Fig. 3). The simulated datasets generated from the birth-death model are clearly
436 able to better reproduce the observed average number of occurrence than the pure-birth model
which greatly underestimates the observed average number of occurrences. This result means that
438 inferences based on the birth-death model are more likely to be representative of the underlying
data than inferences based on the pure-presence model. Further inspection of the posterior
440 parameter estimates from both models gives further insight into the reasons for this difference in
posterior predictive results (Gelman et al., 2013).

442 Occurrence probabilities estimated from the pure-presence model (Fig. 4) are broadly similar to the
estimates of origination probability from the birth-death model (Fig. 5) but not the survival
444 probability estimates (Fig. 6). This result supports the idea that changes to the North American

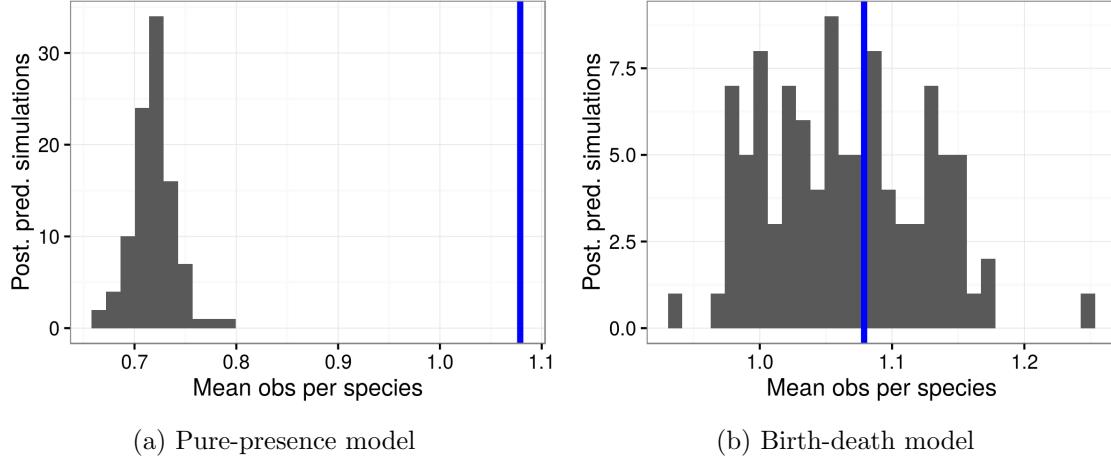


Figure 3: Comparison of the average observed number of occurrences per species (blue line) to the average number of occurrences from 100 posterior predictive datasets using the posterior estimates from the pure-presence and birth-death models.

regional species pool is more likely due to changes in origination than extinction, a result that is

446 returned to later in the discussion of per-capita diversification, origination, and extinction rates.

For most ecotypes, both estimated occurrence probabilities from the pure-presence model (Fig. 4)

448 and origination probabilities estimated from the birth-death model (Fig. 5) increase with time. This

makes sense given that, over time, all species that have at least one observed occurrence must have

450 had that occurrence by the last time point, so our certainty in a species occurring must increase

with time. Importantly, there are potential issues surrounding the partial identifiability of the

452 observation parameters p which may contribute to edge effects in estimates of occurrence,

origination, and extinction (Royle and Dorazio, 2008). Notably, ecotypes with arboreal components

454 do not appear to follow a similar pattern; instead, occurrence and origination probabilities appear

relatively flat for most of the Cenozoic; this is most likely caused by those species of those ecotypes

456 no longer originating or originating very rarely.

The dramatic differences in the estimates origination and survival probabilities are indicative of

458 how differently these processes affect the diversification process and may also be responsible for the

better posterior predictive performance of the birth-death model over the pure-presence model (Fig.

460 3a, and 3b). While the estimates at all points along both time series have high variance, what is

striking is how mean origination probability changes over time while most ecotype survival
 462 probabilities have relatively stable means for the entire Cenozoic (Fig. 5, and 6).

For most ecotypes, the estimates of origination probabilities are with less uncertainty than similar
 464 estimates of survival probabilities (Fig. 5, and 6). In logistic regression, high uncertainty in the
 estimates of the underlying log-odds of occurrence, origination, or survival tends to be indicative of
 466 extreme rarity or complete absence of the specific ecotype; the latter is called complete separation
 which occurs when there is no uncertainty in the effect of a covariate on presence/absence, the
 468 effect of which has been mitigated by the hierarchical modeling strategy used here (Gelman et al.,
 2013; Gelman and Hill, 2007) CITATION Statistical Rethinking.

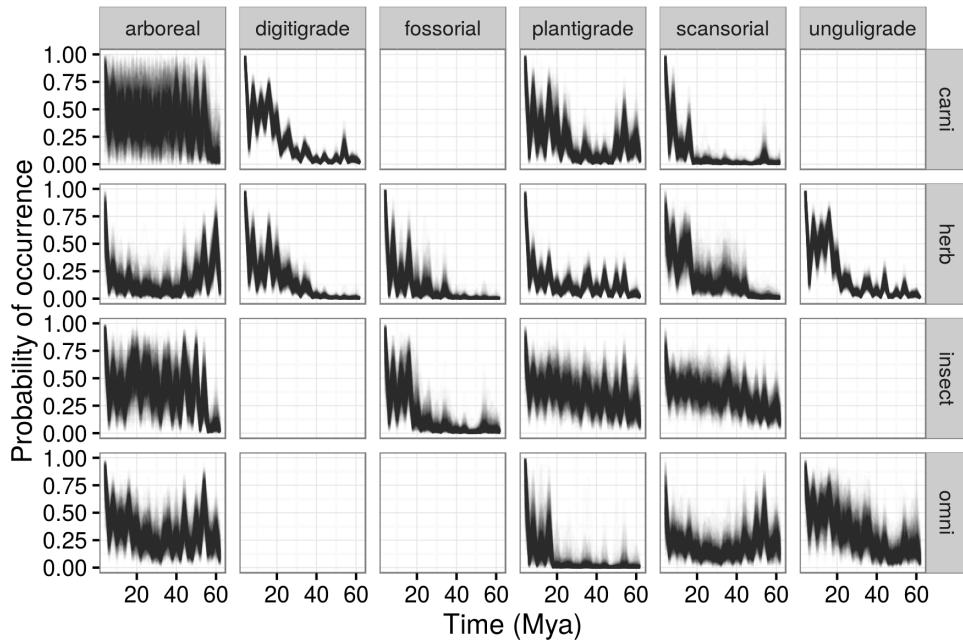


Figure 4: Probability of a mammal ecotype occurring over time as estimated from the pure-presence model. Each panel depicts 100 random samples from the model's posterior. The columns are by locomotor category and rows by dietary category; their intersections are the observed and analyzed ecotypes. Panels with no lines are ecotypes not observed in the dataset.

470 The pure-presence and birth-death models also differ in the estimated effect of mass on the
 probability of sampling a species that is present (Fig. 7). For the pure-presence model, mass is
 472 estimated to not have a strong effect on the probability of sampling a species that is presence (Fig.
 7a). Contrastingly, for the birth-death model mass is found to have a negative relationship with

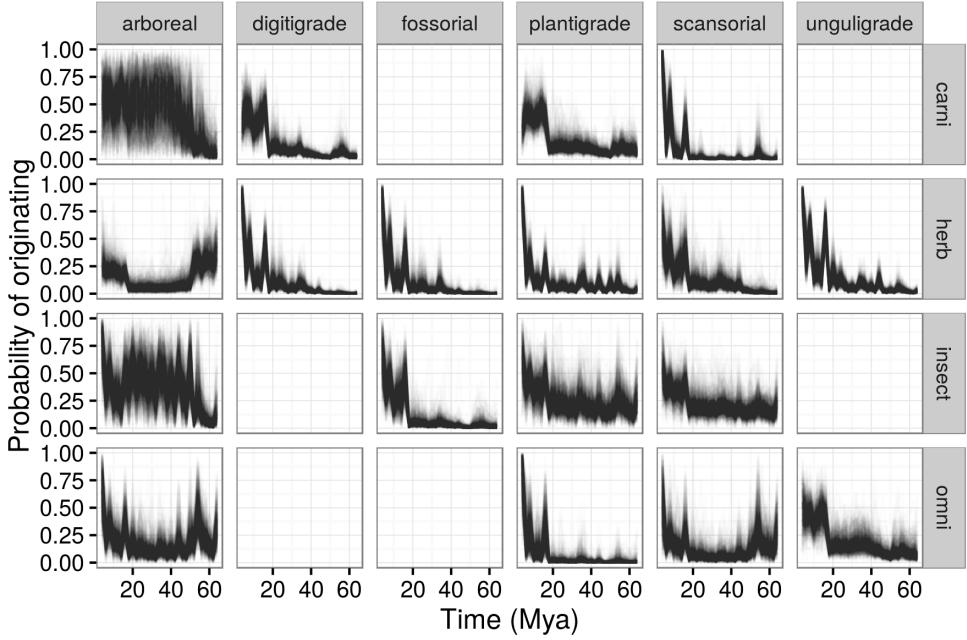


Figure 5: Probability of a mammal ecotype origination probabilities at each time point as estimated from the birth-death model. Each panel depicts 100 random samples from the model’s posterior. The columns are by locomotor category and rows by dietary category; their intersections are the observed and analyzed ecotypes. Panels with no lines are ecotypes not observed in the dataset.

- 474 observation such that larger species are less likely to be observed if present than smaller species
 (Fig. 7b).
- 476 The result from the birth-death model may be considered unexpected given that it is generally
 assumed that larger mammals are more likely to have been collected than smaller mammals
- 478 CITATION. However, collection is not preservation; similarities in preservation rate indicate
 similarities in how gap-filled species records are. What this result means is that the record of large
 480 bodied species is expected on average to have more gaps in sampling and a less consistent record
 from time point to time point than smaller bodied species. Additionally, as this is presence/absence
 482 data higher preservation and collection in terms of individual specimens at a location or a single
 temporal horizon does not necessarily translate to high preservation over multiple time points.
- 484 The average sampling probabilities for both the pure-presence model and birth-death model are
 both at the point where (rescaled log) mass equals 0; visual comparison indicates that, on average,

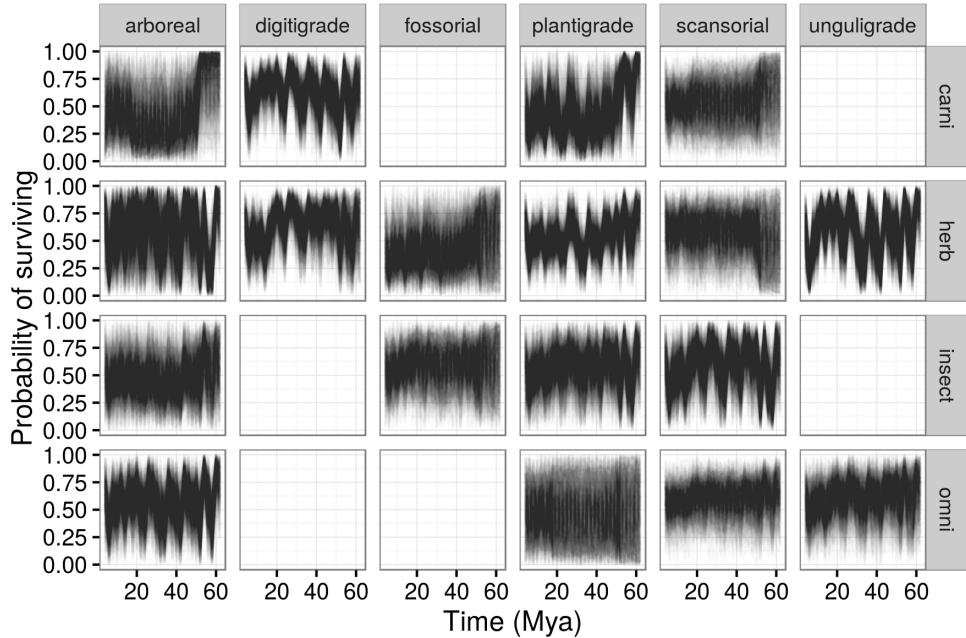


Figure 6: Probability of a mammal ecotype survival probabilities at each time point as estimated from the birth-death model. Each panel depicts 100 random samples from the model’s posterior. The columns are by locomotor category and rows by dietary category; their intersections are the observed and analyzed ecotypes. Panels with no lines are ecotypes not observed in the dataset.

486 sampling probability has greater posterior estimate in the pure-presence model than the birth-death
 model (Fig. 7). The probability that one estimate is different from the other, however, are not
 488 directly calculable as they come from different models; what this tells us is how adding more
 information to the model (i.e. replacing occurrence with origination and extinction) changes
 490 parameter estimates in the model.

The effect of species mass on probability of occurrence as estimated from the pure-presence (Fig. 8)
 492 are most similar to the estimated effect of species mass on probability of origination for the
 birth-death model (Fig. 9). The striking pattern observable in both sets of estimates is the higher
 494 probability of occurrence for species with body sizes closer to the mean than either extremes. This
 result is consistent with the canonically normal distribution of mammal body sizes (Smith et al.,
 496 2004); it is then expected that the most likely to occur species would be those from the middle of
 the distribution, and that species originating will on average be of average mass, especially
 498 considering species shared common ancestry (Felsenstein, 1985). Note that all variation in estimates

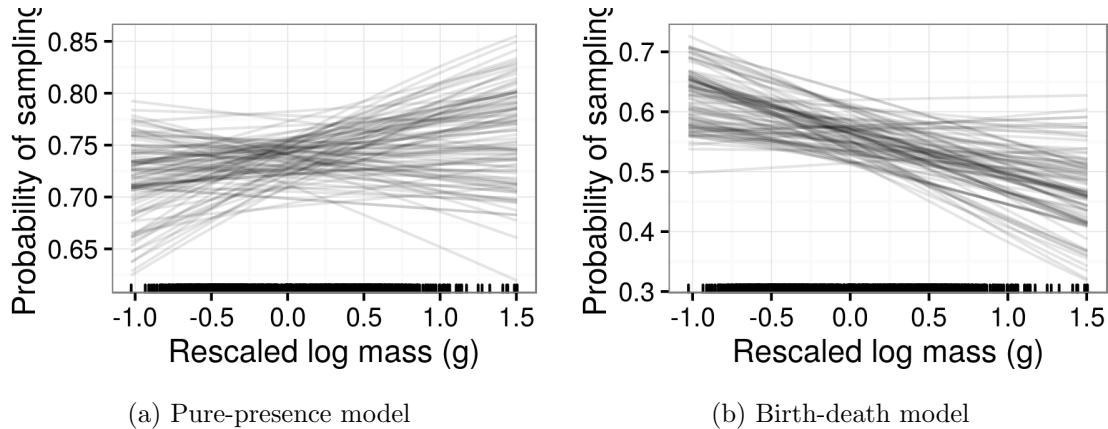


Figure 7: Estimates of the effect of species mass on probability of sampling a present species (p). Mass has been log-transformed, centered, and rescaled; this means that a mass of 0 corresponds to the mean of log-mass of all observed species and that mass is in standard deviation units. Estimates are from both the pure-presence and birth-death models.

between ecotypes (Fig. 9) is due to differences in ecotype-specific survival probability and the associated effects of plant phase; the effect of mass was considered constant for all ecotypes.

In contrast, the effect of species mass on probability of survival as estimated from the birth-death model (Fig. 10) is consistent with previous findings that there is little effect of mass on extinction for North American mammals for the Cenozoic (Smits, 2015; Tomiya, 2013). Note that all variation between ecotypes depicted in Figure 10 is due to differences in ecotype-specific survival probability and the associated effects of plant phase; the effect of mass was considered constant for all ecotypes (Eqs. 2, 3).

Similarities in parameters estimates between ecotypes may be due to similar response to environmental factors (Fig. 11, 12, and 13). As with previous comparisons between posterior estimates from the pure-presence and birth-death models, the effects of the group-level covariates in the pure-presence model (Fig. 11) are more similar to those estimates of the group-level effects on origination (Fig. 12) as opposed to survival (Fig. 13). As demonstrated in the comparisons of the effect of mass on occurrence from the pure-presence model (Fig. 8) with the effect of mass on origination and survival from the birth-death model (Fig. 9, and 10), there is considerable variation in the effect of plant phases on ecotype-specific estimates.

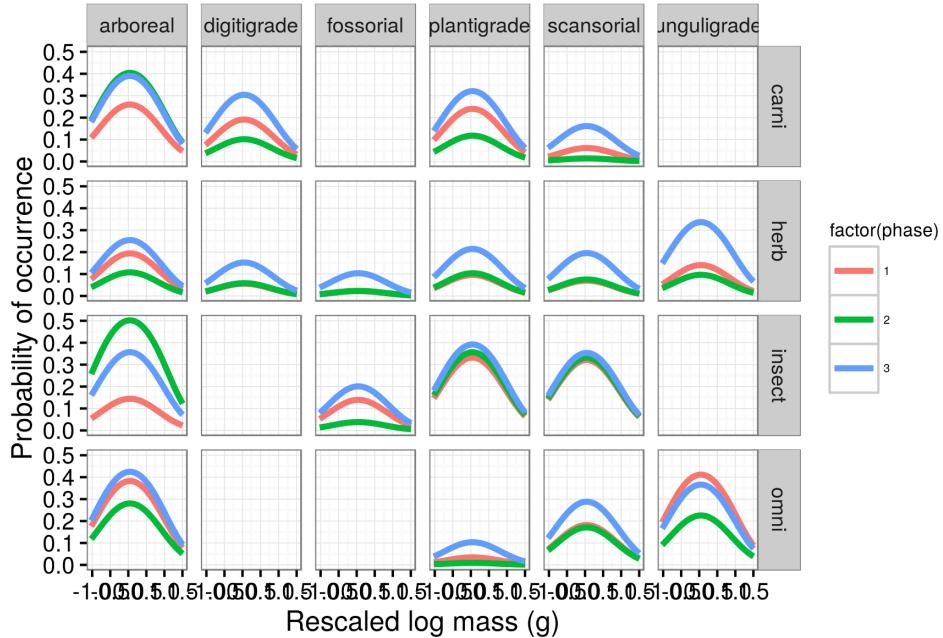


Figure 8: Mean estimate of the effect of species mass on the probability of a species occurrence for each of the three plant phases. The effect of mass is considered constant over time and that the only aspect of the model that changes with plant phase is the intercept of the relationship between mass and occurrence. The three plant phases are indicated by the color of the line. Mass has been log-transformed, centered, and rescaled; this means that a mass of 0 corresponds to the mean of log-mass of all observed species and that mass is in standard deviation units. Only the mean estimates of the effects of both mass and plant phase are plotted for clarity; these estimates are obviously made with uncertainty.

An association between plant phase and differences in the log-odds of occurrence (Fig. 11),

516 origination (Fig. 12), or extinction (Fig. 13) is interpreted to mean that if the set of possible
 517 mammal-plant interactions was either favorable (positive association) or adverse (negative
 518 association) to those ecotypes. In the case of species origination, for example, favorable conditions
 519 for an ecotype may be indicative of an increasing number of possible and available mammal-plant
 520 interactions (e.g. ecological opportunity); while adverse conditions may translate to a decreasing set
 521 of interactions or loss of appropriate environmental context. Note that favorable versus adverse
 522 condition of a plant phase is definitionally relative to the other two plant phases.

Plant phases are associated with large differences in log-odds for occurrence and origination
 524 probabilities (Tab. 9), 10), though there is little evidence of plant phase being an important

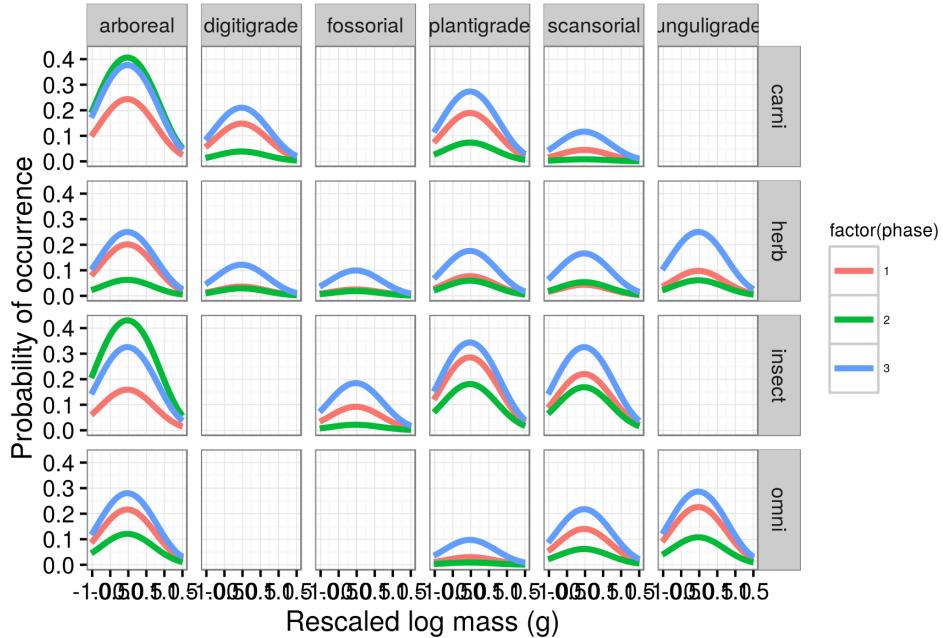


Figure 9: Mean estimate of the effect of species mass on the probability of a species originating for each of the three plant phases. The effect of mass is considered constant over time and that the only aspect of the model that changes with plant phase is the intercept of the relationship between mass and origination. The three plant phases are indicated by the color of the line. Mass has been log-transformed, centered, and rescaled; this means that a mass of 0 corresponds to the mean of log-mass of all observed species and that mass is in standard deviation units. Only the mean estimates of the effects of both mass and plant phase are plotted for clarity; these estimates are obviously made with uncertainty.

distinguishing factor in species survival as only a few ecotypes demonstrate strong affinities with
 526 some plant phases (Tab. 11). As with previous comparisons between parameter estimates associated
 with species occurrence and species origination, parameters associated with probability of newly
 528 originating appear as a more “tempered” version of those associated with probability occurrence.

The almost universal pattern of the effect of plant phase on ecotype occurrence or origination is
 530 that the during first and last plant phases ecotypes have a greater log-odds of
 occurrence/origination than the second plant phase (Fig. 4, 5). The three exceptions to this pattern
 532 are fossorial herbivores, scansorial herbivores, and arboreal insectivores.

The difference between the third plant phase and the other two plant phases, for all ecotypes except
 534 arboreal carnivores, is obvious upon inspection the occurrence and origination time series as there

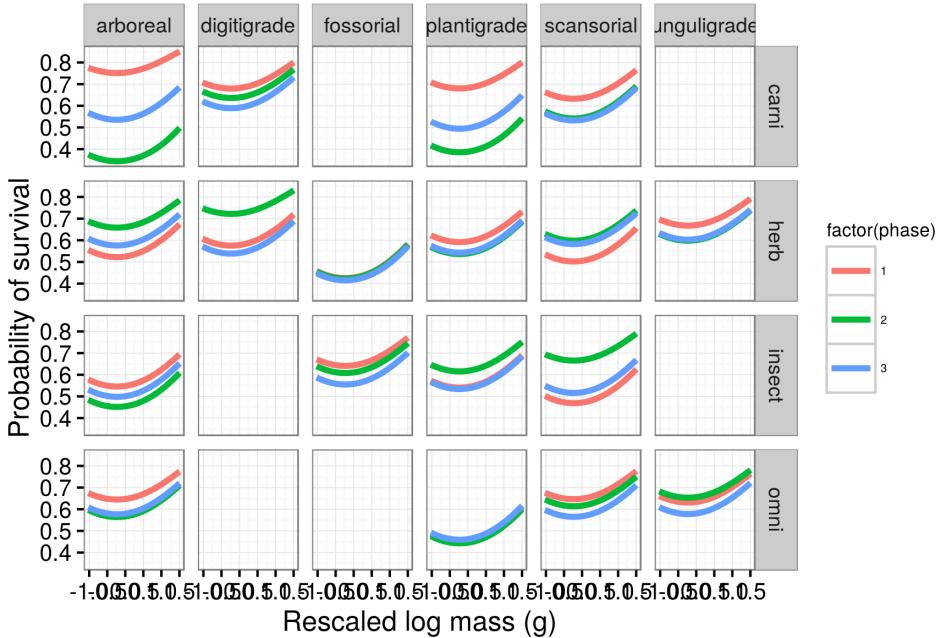


Figure 10: Mean estimate of the effect of species mass on the probability of a species survival for each of the three plant phases. The effect of mass is considered constant over time and that the only aspect of the model that changes with plant phase is the intercept of the relationship between mass and survival. The three plant phases are indicated by the color of the line. Mass has been log-transformed, centered, and rescaled; this means that a mass of 0 corresponds to the mean of log-mass of all observed species and that mass is in standard deviation units. Only the mean estimates of the effects of both mass and plant plant are plotted for clarity; these estimates are obviously made with uncertainty.

is large up-tick in probability of occurring or originating towards the modern (Fig. 4, 5). The

536 differences in mean probability of occurring or originating attributable to the plant phases are
observable as shifts along the time series correponding to the phase barriers (Table 4). For example,
538 scansorial herbivore occurrence and origination probabilities demonstrate clear shifts at 50 Mya and
16 Mya (Fig. 4, 5).

540 Both aspects of global temperature analyzed here are estimated to have strong effects on species
occurrence and origination for most mammal ecotypes (Tables 12, 13). Similarity, temperature is
542 only expected to have a strong effect on species extinction for very few ecotypes (Tab. 14). For the
occurrence and origination probabilities of many ecotypes, both temperature covariates have
544 negative estimates which means that as temperature decreases, occurrence or origination are

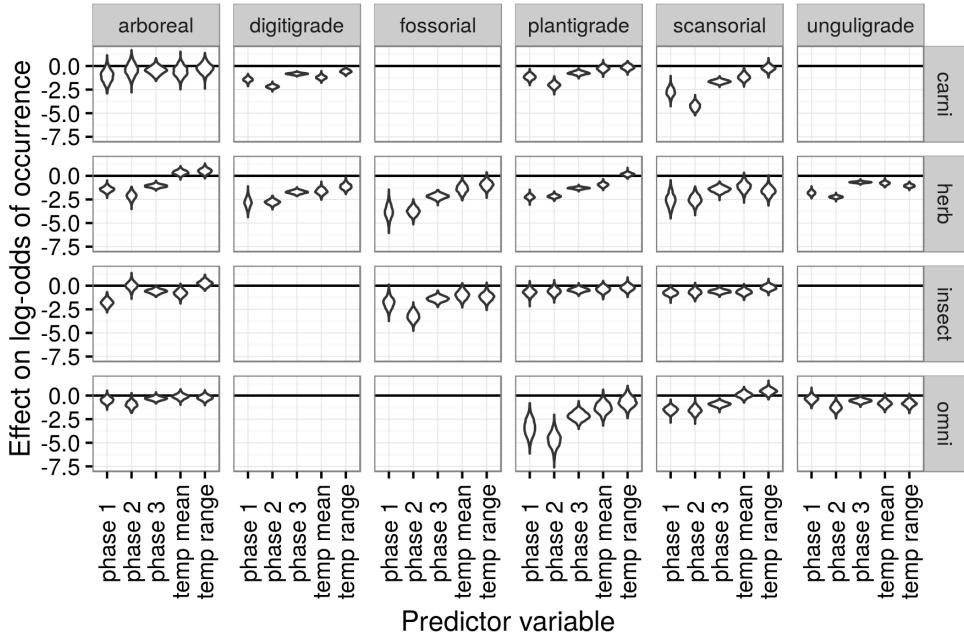


Figure 11: Estimated effects of the group-level covariates describing environmental context on log-odds of species occurrence. These estimates are from the pure-presence model.

expected to increase. The only strongly positive estimate (e.g. temperature decrease, origination
546 decrease) is for the effect of temperature range on arboreal herbivores. Contrastingly, the only
strong ecotype associations for either of the temperature covariates are with plantigrade carnivores,
548 plantigrade herbivores, and to a less certain extent arboreal herbivores and scansorial insectivores
(Tab. 14). The effects of the temperature covariates on these ecotypes are all estimated to be
550 positive (e.g. temperature range increase, increase in survival).

Analysis of diversity

552 All of the analyses of diversification and macroevolutionary rates has been done using only the
birth-death model because of the model's better posterior predictive check performance (Fig. 3).
554 The general pattern of the estimated North American total mammal diversity for the Cenozoic is
“stable” in that mean standing diversity does not fluctuate wildly and rapidly over the Cenozoic
556 (Fig. 14a). In broad strokes, the first 15 or so million years of the Cenozoic are characterized by a

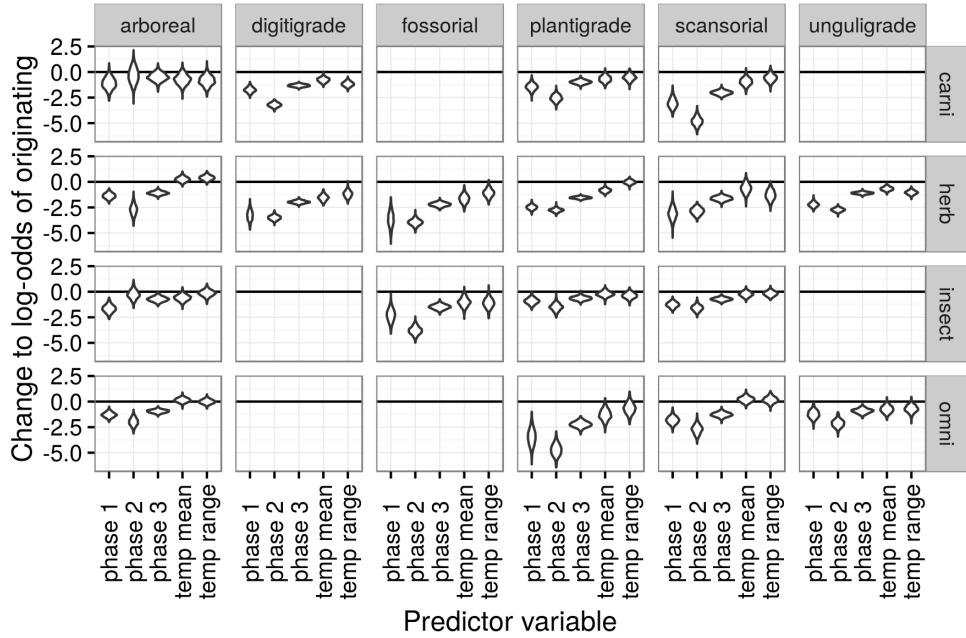


Figure 12: Estimated effects of the group-level covariates describing environmental context on log-odds of species origination. These estimates are from the birth-death model.

gradual decline in standing diversity until approximately 45-50 million years ago (early-middle

558 Eocene). Following this decline, standing diversity is broadly constant from 45 to 18 Mya (early
Miocene). After this, there is a rapid spike in diversity followed by a slight decline in diversity up to
560 the Modern. This characterization of the estimated diversity history is knowingly broad strokes and
diversity time series is not without variation and vagaries.

562 When viewed through the lens of diversification rate, some of the structure behind the estimated
diversity history begins to take shape (Fig. 14b). For most of the Cenozoic, the diversification rate

564 hovers around zero, punctuated by both positive and negative spikes. The largest spike in
diversification rate is at 18 Mya, which is early Oligocene (Fig. 14b). Other notable increases in
566 diversification rate occur 56, 46, 38, and 6 Mya (Table 15), though the last of these may be due
edge effects surrounding the partial-identifiability of $p_{t=T}$. Notable decreases in diversification rate
568 occur 60, 54, 50, 44, 34, 20, 16, 12, and 8 Mya (Table 15), meaning that diversification rate has
more major decreases than increases. Given that diversification rate more closely resembles
570 origination rate than extinction rate (Fig. 14b, 14c, 14d), these decreases in diversification rate may

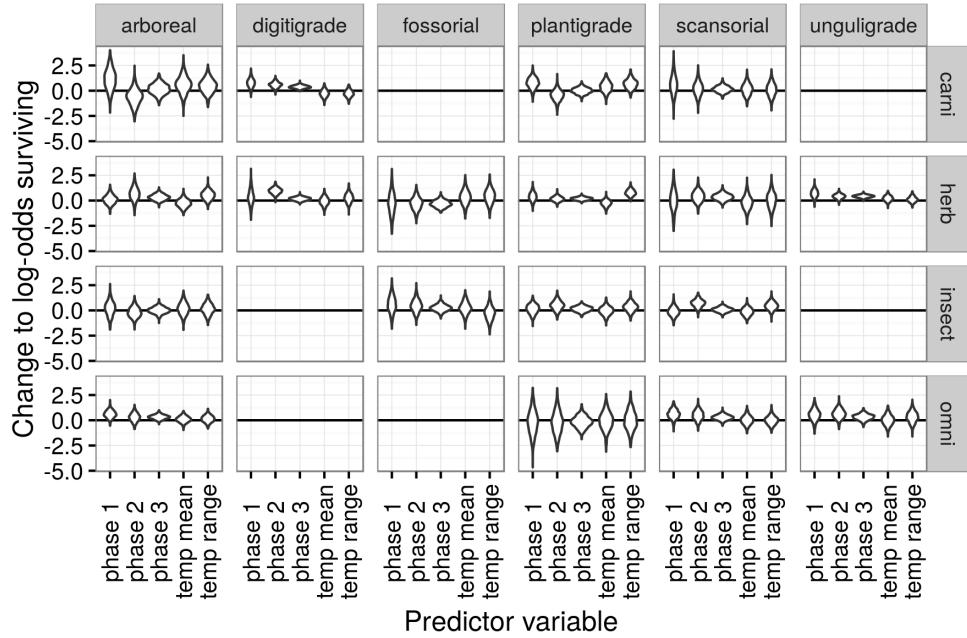


Figure 13: Estimated effects of the group-level covariates describing environmental context on log-odds of species survival. These estimates are from the birth-death model.

be indicative of “depletions” (failure to replace extinct taxa) rather than pulses of extinction.

- 572 The comparison between per capita origination and extinction rate estimates reveals how diversification rate is formed (Fig. 14c, 14d). As expected given previous inspection of the ecotype
- 574 specific estimates of origination and survival probabilities from the birth-death model, diversification rate seems most driven by changes in origination rate as opposed to extinction rate.
- 576 Extinction rate, on the other hand, demonstrates an almost saw-toothed pattern around a constant mean (Fig. 14d).
- 578 Diversity partitioned by ecotype reveals a lot of the complexity behind the pattern of mammal diversity for the Cenozoic (Fig. 15).
- 580 Arboreal ecotypes obtain peak diversity early in the Cenozoic and then decline for the rest of the time series, becoming increasingly rare or absent as diversity approaches the Modern (Fig. 15).
- 582 Arboreal herbivores and omnivores obtain peak diversity at the beginning of the Cenozoic then go into decline while remaining a small part of the species pool, while arboreal carnivores and

Table 9: Posterior probability of the differences in the log-odds of an ecotype originating based on plant phase. These probabilities are calculated as $P(\text{Phase 1} > \text{2}) = (\gamma_{\text{phase1}} - \gamma_{\text{phase1}} + \gamma_{\text{phase2}})/100$ and similarly for the other comparisons. These estimates are from the pure-presence model.

	P(Phase 1 > Phase 2)	P(Phase 2 > Phase 3)	P(Phase 1 > Phase 3)
arboreal carnivore	0.460	0.776	0.866
digitigrade carnivore	1.000	0.000	1.000
plantigrade carnivore	1.000	0.040	1.000
scansorial carnivore	1.000	0.001	1.000
arboreal herbivore	1.000	0.540	1.000
digitigrade herbivore	1.000	0.995	1.000
fossorial herbivore	1.000	0.920	1.000
plantigrade herbivore	1.000	0.998	1.000
scansorial herbivore	0.999	0.754	1.000
unguligrade herbivore	1.000	0.000	1.000
arboreal insectivore	0.028	1.000	0.999
fossorial insectivore	1.000	0.161	1.000
plantigrade insectivore	0.706	0.774	0.985
scansorial insectivore	0.630	0.937	1.000
arboreal omnivore	0.981	0.165	0.944
plantigrade omnivore	1.000	0.325	1.000
scansorial omnivore	0.987	0.746	1.000
unguligrade omnivore	0.990	0.344	0.997

584 insectivores obtain peak diversity 52-50 Mya and then quickly decline and become extremely rare or entirely absent from the species pool.

586 The diversity of both digitigrade and unguligrade herbivores increase over the Cenozoic (Fig. 15). In contrast, plantigrade herbivore diversity does not have a single, broad-strokes pattern; instead, 588 diversity increases, decreases, and may have then increased till the Modern. Contrastingly, fossorial and scansorial herbivores demonstrate a much flatter history of diversity, with a slight increase in 590 diversity that over time is more pronounced among fossorial taxa than scansorial taxa.

Digitigrade carnivores have a multi-modal diversity history, with peaks 54-52 and 12-10 Mya 592 (Fig.15). Between these two peaks digitigrade carnivore diversity dips below average diversity following the first peak and then grows slowly until the second peak. Plantigrade carnivores obtain 594 peak diversity in the early Cenozoic and then maintain a relatively stable diversity until another peak at the end of the Cenozoic.

Table 10: Posterior probability of the differences in the log-odds of an ecotype originating based on plant phase. These probabilities are calculated as $P(\text{Phase 1} > 2) = (\gamma_{\text{phase1}} - \gamma_{\text{phase1}} + \gamma_{\text{phase2}})/100$ and similarly for the other comparisons. These estimates are from the birth-death model.

	P(Phase 1 > Phase 2)	P(Phase 2 > Phase 3)	P(Phase 1 > Phase 3)
arboreal carnivore	0.460	0.776	0.866
digitigrade carnivore	1.000	0.000	1.000
plantigrade carnivore	1.000	0.040	1.000
scansorial carnivore	1.000	0.001	1.000
arboreal herbivore	1.000	0.540	1.000
digitigrade herbivore	1.000	0.995	1.000
fossorial herbivore	1.000	0.920	1.000
plantigrade herbivore	1.000	0.998	1.000
scansorial herbivore	0.999	0.754	1.000
unguligrade herbivore	1.000	0.000	1.000
arboreal insectivore	0.028	1.000	0.999
fossorial insectivore	1.000	0.161	1.000
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arboreal omnivore	0.981	0.165	0.944
plantigrade omnivore	1.000	0.325	1.000
scansorial omnivore	0.987	0.746	1.000
unguligrade omnivore	0.990	0.344	0.997

- 596 There are some broad similarities in diversity histories of insectivorous and omnivorous taxa. The
 597 diversity histories of arboreal, plantigrade, and scansorial insectivorous taxa all demonstrate a
 598 decreasing pattern with time, while fossorial insectivores have a flat diversity history with a peak
 599 approximately 10 Mya (Fig. 15). Arboreal and scansorial omnivores decrease in diversity from their
 600 initial peaks early in the Cenozoic, and plantigrade omnivores have a generally flat diversity history
 601 with a sudden peak in diversity late in the Cenozoic (Fig. 15). Unguligrade omnivores also
 602 demonstrate a possible decrease in diversity over the Cenozoic, but not as clearly as arboreal and
 603 scansorial omnivores.
- 604 Many of the estimated ecotype specific diversity histories share a similar increases in diversity to
 605 one degree or another at the late Cenozoic 16-14 Mya (Fig. 15); these increases are either sustained
 606 or temporary: digitigrade carnivores, plantigrade carnivores, scansorial carnivores, unguligrade
 607 herbivores, fossorial insectivores, and plantigrade omnivores.
- 608 When ecotype diversity is decomposed into the number of origination events per time bin (Fig. 16)

Table 11: Posterior probability of the differences in the log-odds of an ecotype surviving based on plant phase. These probabilities are calculated as $P(\text{Phase 1} > 2) = (\gamma_{\text{phase1}} - \gamma_{\text{phase1}} + \gamma_{\text{phase2}})/100$ and similarly for the other comparisons. These estimates are from the birth-death model.

	P(Phase 1 > Phase 2)	P(Phase 2 > Phase 3)	P(Phase 1 > Phase 3)
arboreal carnivore	0.904	0.121	0.382
digitigrade carnivore	0.181	0.248	0.004
plantigrade carnivore	0.857	0.195	0.519
scansorial carnivore	0.477	0.438	0.310
arboreal herbivore	0.278	0.510	0.140
digitigrade herbivore	0.001	0.978	0.175
fossorial herbivore	0.480	0.723	0.816
plantigrade herbivore	0.558	0.192	0.111
scansorial herbivore	0.444	0.286	0.133
unguligrade herbivore	0.548	0.022	0.002
arboreal insectivore	0.691	0.359	0.492
fossorial insectivore	0.334	0.488	0.221
plantigrade insectivore	0.189	0.677	0.308
scansorial insectivore	0.017	0.918	0.375
arboreal omnivore	0.549	0.196	0.074
plantigrade omnivore	0.528	0.537	0.618
scansorial omnivore	0.326	0.442	0.125
unguligrade omnivore	0.191	0.487	0.145

and the number of extinction events per time bin (Fig. 17) the estimates are clearly similar; there

610 are no obvious major cross-ecotype origination or extinction events, and there is no evidence of a
 sudden turnover as expected peaks in originations proceed peaks in peaks in the number of
 612 extinctions. Also, it is clear that the sustained increases in digitigrade and unguligrade herbivore
 diversity observed above (Fig. 15) is driven by an increase in the average number of originations as
 614 with a relatively constant number of extinctions over time (Fig. 16, 17).

Discussion

616 Both the composition of a species pool and its environmental context changes over time, though not
 necessarily at the same rate or concurrently. Local communities, who's species are drawn from the
 618 regional species pool, have “roles” in their communities defined by their interactions with a host of
 biotic and abiotic interactors (i.e. a species’ niche). For higher level ecological characterizations like

Table 12: Posterior probability the the effects of the two temperature covariates on the the log-odds of an ecotype occurring are greater than 0. What is estimated is the probability that these estimates are greater than 0; high or low probabilities indicate the “strength” of the covariate in that direction (positive and negative, respectively). These estimates are from the pure-presence model.

	$P(\gamma_{temp\ mean} > 0)$	$P(\gamma_{temp\ range} > 0)$
arboreal carnivore	0.169	0.317
digitigrade carnivore	0.000	0.000
plantigrade carnivore	0.168	0.304
scansorial carnivore	0.000	0.206
arboreal herbivore	0.943	0.969
digitigrade herbivore	0.000	0.000
fossorial herbivore	0.001	0.022
plantigrade herbivore	0.000	0.832
scansorial herbivore	0.009	0.003
unguligrade herbivore	0.000	0.000
arboreal insectivore	0.006	0.783
fossorial insectivore	0.016	0.003
plantigrade insectivore	0.127	0.260
scansorial insectivore	0.009	0.238
arboreal omnivore	0.337	0.191
plantigrade omnivore	0.012	0.120
scansorial omnivore	0.597	0.935
unguligrade omnivore	0.002	0.002

620 ecotypes and guilds, these roles are broad and not defined by specific interactions but by the genre
of interactions species within that grouping participate in. The diversity of species within an
622 ecotype or guild can be stable over millions of years despite constant species turnover (Jernvall and
Fortelius, 2004; Slater, 2015) CITATIONS. This implies that the size and scope of the role of an
624 ecotype or guild in local communities, and the regional species pool as a whole, is preserved even as
the individual interactors change. This also implies the structure of regional species pools can be
626 constant over time despite a constantly changing set of “players.”

Comparison of the posterior predictive check results for the pure-presence and birth-death models
628 supports the conclusion that regional species pool dynamics cannot simply be described by a single
occurrence probability and is instead the result of the interplay between origination and extinction.
630 Additionally, changes to the ecotypic composition and diversification rate for the North American
regional species pool are driven primarily by variation in origination rates. These aspects of how

Table 13: Posterior probability that the effects of the two temperature covariates on the log-odds of an ecotype origination are greater than 0. What is estimated is the probability that these estimates are greater than 0; high or low probabilities indicate the “strength” of the covariate in that direction (positive and negative, respectively). These estimates are from the birth-death model.

	$P(\gamma_{temp\ mean} > 0)$	$P(\gamma_{temp\ range} > 0)$
arboreal carnivore	0.086	0.045
digitigrade carnivore	0.001	0.000
plantigrade carnivore	0.013	0.054
scansorial carnivore	0.007	0.062
arboreal herbivore	0.853	0.957
digitigrade herbivore	0.000	0.001
fossorial herbivore	0.000	0.002
plantigrade herbivore	0.000	0.428
scansorial herbivore	0.106	0.003
unguligrade herbivore	0.000	0.000
arboreal insectivore	0.028	0.314
fossorial insectivore	0.010	0.006
plantigrade insectivore	0.188	0.090
scansorial insectivore	0.182	0.224
arboreal omnivore	0.749	0.482
plantigrade omnivore	0.007	0.117
scansorial omnivore	0.765	0.699
unguligrade omnivore	0.016	0.023

632 regional species pool diversity is shaped is not directly observable in studies of the Modern where
 time scales are short and macroevolutionary dynamics are inferable solely from phylogeny (Fritz
 634 et al., 2013).

Extinction rate for the entire regional species pool through time is highly variable and demonstrates
 636 a saw-toothed pattern around an apparently constant mean. While a constant mean extinction rate
 is consistent with previous observation (Alroy, 1996; Alroy et al., 2000), the degree to which
 638 extinction rate is actually variable may not have been equally appreciated. What is most consistent
 with previous observations (Alroy, 1996; Alroy et al., 2000), however, is that diversity seems to be
 640 most structured by origination than extinction.

Comparison of the ecotype specific diversity histories supports the conclusion that there was no
 642 large scale or sudden simultaneous reorganization of all local communities; instead these results
 support a more gradual and idiosyncratic shifts in relative ecotypic diversity over time (Fig. 15).

Table 14: Posterior probability that the effects of the two temperature covariates on the log-odds of an ecotype survival are greater than 0. What is estimated is the probability that these estimates are greater than 0; high or low probabilities indicate the “strength” of the covariate in that direction (positive and negative, respectively). These estimates are from the birth-death model.

	$P(\gamma_{temp\ mean} > 0)$	$P(\gamma_{temp\ range} > 0)$
arboreal carnivore	0.777	0.745
digitigrade carnivore	0.236	0.211
plantigrade carnivore	0.763	0.929
scansorial carnivore	0.596	0.554
arboreal herbivore	0.261	0.878
digitigrade herbivore	0.438	0.720
fossorial herbivore	0.676	0.731
plantigrade herbivore	0.215	0.997
scansorial herbivore	0.377	0.535
unguligrade herbivore	0.768	0.655
arboreal insectivore	0.614	0.610
fossorial insectivore	0.673	0.337
plantigrade insectivore	0.470	0.787
scansorial insectivore	0.364	0.879
arboreal omnivore	0.620	0.645
plantigrade omnivore	0.476	0.484
scansorial omnivore	0.514	0.494
unguligrade omnivore	0.513	0.729

- 644 The closest examples to a sudden increase or decrease of a specific ecotype is the jump in standing
 diversity of scansorial carnivorans and, to a lesser extent, fossorial insectivores at 16 Mya (i.e. the
 646 start of the third plant phase). This result may, however, not reflect the dynamics of individual
 local communities as this is an analysis of the entire North American mammal regional species pool.
 648 Arboreal taxa disappear from the regional species pool over the Cenozoic, with massive
 disappearance by the Paleogene-Neogene transition ~22 Mya. This is consistent with one of the two
 650 possible patterns presented here and in Smits (2015) that would result in arboreal taxa having a
 greater extinction risk than other ecotypes: the Paleogene and Neogene were different selective
 652 regimes and while the earliest Cenozoic may have been neutral wrt arboreal taxa, they disappeared
 quickly over the Cenozoic which may account for their higher extinction risk. In addition to all
 654 arboreal taxa, the diversity of plantigrade and scansorial insectivores decreases with time (Fig. 15).
 Digitigrade carnivorans have a relatively stable diversity history through the Cenozoic and can be

656 characterized as varying around a constant mean diversity. This ecotype has a large amount of
overlap with the carnivore guild which has been the focus of much research CITATIONS. This
658 result is consistent with some form of “control” on the ecotype, such as environmental stability,
diversity-dependence, or similar Silvestro et al. (2015); Slater (2015).

660 Both digitigrade and unguligrade herbivores increase in diversity over the Cenozoic. The increase of
these cursorial forms is consistent with the gradual opening up of the North American landscape
662 (Blois and Hadly, 2009; Graham, 2011; Strömberg, 2005). These herbivore increase in diversity over
the Cenozoic which may be indicative of a long-term shift in the interactors associated with those
664 ecotypes leading to increased contribution to the regional species pool. This result may be
comparable to the increasing percentage of hypsodont (high-crowned teeth) mammals in the
666 Neogene of Europe being due to an enrichment of hyposodont taxa and not a depletion of
non-hypsodont taxa. Smaller scale increases in fossorial herbivore species, and a lesser extent
668 plantigrade herbivores, suggests that the increase of interactors may be associated mostly with the
herbivore dietary category with locomotor category tempering that relationship.

670 An association between plant phase and differences in ecotype occurrence or origination-extinction
probabilities is interpreted to mean that an ecotype enrichment or depletion is due to to
672 associations between that ecotype and whatever plants are dominate at that time and are thus a
contributing factor to the constancy of an ecotype, or the lack there of. Plant phase clearly
674 structures the occurrence and origination probability time series (Fig. 4, 5). These differences in
occurrence or origination translate opaquely to the estimates of diversity and diversification rate;
676 the largest spike in both diversity, diversification rate, and origination rate all correspond to the
onset of the last plant phase (Fig. 14). The clearest example of the diversity of an ecotype
678 increasing at this particular transition is in scansorial carnivores (Fig. 15); similar shifts in other
ecotypes are much more subtle, as was previously noted for fossorial insectivores. Interestingly, all
680 of the ecotypes with sudden changes to diversity at this transition increase in diversity, even if only
temporarily. My interpretation of this result is that, because plant phase associations are only with
682 occurrence or origination probabilities and not survival, these ecotypes were well suited for the
newly available mammal-plant interactions due to the increased modernization of their floral

684 context (Graham, 2011).

There are fewer, less obvious shifts in diversity surrounding the transition from the first to second,
686 with the following ecotypes having apparent shifts in diversity at 50 My: digitigrade carnivores
(down), plantigrade carnivores (down), plantigrade herbivores (up), arboreal omnivores (down), and
688 scansorial omnivores (down). Because plant phase has been found to structure
occurrence/origination (Fig. 4, 5), but not survival (Fig. 6, my interpretation of these results is that
690 new species were not entering the system because there were fewer available mammal-plant
interactions available for those ecotypes. Instead, these ecotypes were poorly suited for the newly
692 available mammal-plant interactions brought upon by the changing environmental context
(Graham, 2011).

694 The estimated effects of temperature on occurrence and origination-extinction probabilities are
similar to those of the plant phases. The occurrence and origination probabilities of many mammal
696 ecotypes have strong relationships with the two temperature covariates (Tables 12, 13). In most
cases, there is a negative association between temperature and probability of occurring or first
698 originating; this means that if temperature decreases, we would then expect the probability of
occurring or first originating would increase. Contrastingly, only temperature range are estimated
700 to be good predictors of survival in four mammal ecotypes and only marginally for two of those
(Table 14). Additionally, in all four of these cases are expected to have positive relationships,
702 meaning that if temperature decreases it is expected that species survival will also decrease.

What is the comparative size of the effects of plant phase and temperature are approximately equal
704 in importance in the sense that they have similar effect sizes on the ecotypes. The focus in previous
research on temperature and major climatic or geological events without other measures of
706 environmental context may have been a mistake and perhaps led to increasing confusion in
discussions of how the “environment” affects mammal diversity and diversification. The
708 environment or climate is not just global or regional temperature, it is the set of all possible biotic
and abiotic interactions that can be experienced by a member of the species pool. By including
710 more descriptors of species’ environmental context a more complete “picture” of the diversification

process is inferred.

712 The effect of species mass on either occurrence or origination and extinction was not allowed to
vary by ecotype or environmental context even though it is not known if this is the case or not
714 CITATION. The primary reason for this modeling choice was that this study focuses on ecotypic
based differences in either occurrence, or origination and extinction. Allowing the effect of body size
716 to vary by ecotype, time, and environmental factors would increase the overall complexity of the
model, something that I felt was not necessary because the overall scope of the study. Instead, body
718 size was included in order to control for its possible underlying effects CITATION. A control means
that if there is variation due to body mass, having a term to “absorb” that effect is better than
720 ignoring it which may affect other parameter estimates. Additionally, the effect of body size was
allowed to have a second-order polynomial form and no higher order polynomials were considered;
722 this was done because it is hard to conceive of a more complex third- or higher-order relationship
between body size and the other parameters. Finally, parameteric forms of nonlinearity have not
724 previously been considered, so the simple act of estimating a potential second-order relationship is
an opportunity to test more complex hypotheses of the relationship between body size and both
726 macroevolutionary and macroecological processes.

The only covariate allowed to affect sampling probability was mass and only as a linear predictor.

728 Other covariates, such as the environmental factors considered here, could have affected the
underlying preservation process that limits sampling probability; their exclusion as covariates of
730 sampling/observation was the product of a few key decisions: model complexity, model
interpretability, the scope of this study, and a lack of good hypotheses related to these covariates to
732 warrant their inclusion.

The potential effects of common ancestry (i.e. phylogeny) on origination and extinction are not
734 directly considered in this analysis. While a birth-death process approximates the
speciation-extinction process underlying the phylogeny (Silvestro et al., 2014) this is not same as
736 considering how the similarity between closely related species may affect the estimates of the effects
of species traits to environmental factors on both origination and extinction (Harnik et al., 2014;

⁷³⁸ Smits, 2015). One of the principle barriers to the inclusion of the effect of phylogeny in either the
pure-presence or birth-death models is computational; with well over 1000 tips, the calculation of
⁷⁴⁰ the scale parameter defining the phylogenetic effect would be very slow and further increase the
already slow computation time necessary for both the marginalization of the discrete occurrence
⁷⁴² histories and data augmentation already included in both models.

HERE GOES A CONCLUSION

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⁷⁵⁰ References

- Allen, L. J. S. 2011. An introduction to stochastic processes with applications to biology. 2nd ed.
⁷⁵² Chapman and Hall/CRC, Boca Raton, FL.
- Alroy, J. 1996. Constant extinction, constrained diversification, and uncoordinated stasis in North
⁷⁵⁴ American mammals. *Palaeogeography, Palaeoclimatology, Palaeoecology* 127:285–311.
- . 2009. Speciation and extinction in the fossil record of North American mammals. Pages
⁷⁵⁶ 302–323 *in* R. K. Butlin, J. R. Bridle, and D. Schlüter, eds. *Speciation and patterns of diversity*.
Cambridge University Press, Cambridge.
- ⁷⁵⁸ Alroy, J., P. L. Koch, and J. C. Zachos. 2000. Global climate change and North American
mammalian evolution. *Paleobiology* 26:259–288.

- 760 Badgley, C., and J. A. Finarelli. 2013. Diversity dynamics of mammals in relation to tectonic and
climatic history: comparison of three Neogene records from North America. *Paleobiology*
762 39:373–399.
- 764 Badgley, C., T. M. Smiley, R. Terry, E. B. Davis, L. R. G. Desantis, D. L. Fox, S. S. B. Hopkins,
T. Jezkova, M. D. Matocq, N. Matzke, J. L. McGuire, A. Mulch, B. R. Riddle, V. L. Roth, J. X.
Samuels, C. A. E. Strömberg, and B. J. Yanites. 2017. Biodiversity and Topographic Complexity:
766 Modern and Geohistorical Perspectives. *Trends in Ecology & Evolution* pages 1–16.
- Bambach, R. K. 1977. Species richness in marine benthic habitats through the Phanerozoic.
768 *Paleobiology* 3:152–167.
- Bambach, R. K., A. M. Bush, and D. H. Erwin. 2007. Autecology and the filling of ecospace: Key
770 metazoan radiations. *Palaeontology* 50:1–22.
- Blois, J. L., and E. A. Hadly. 2009. Mammalian Response to Cenozoic Climatic Change. Annual
772 Review of Earth and Planetary Sciences 37:181–208.
- Brook, B. W., and D. M. J. S. Bowman. 2004. The uncertain blitzkrieg of Pleistocene megafauna.
774 *Journal of Biogeography* 31:517–523.
- Brown, A. M., D. I. Warton, N. R. Andrew, M. Binns, G. Cassis, and H. Gibb. 2014. The
776 fourth-corner solution - using predictive models to understand how species traits interact with
the environment. *Methods in Ecology and Evolution* 5:344–352.
- 778 Brown, J. H., and B. A. Maurer. 1989. Macroecology: the division of food and space among species
on continents. *Science* 243:1145–1150.
- 780 Bush, A. M., and R. K. Bambach. 2011. Paleoecologic Megatrends in Marine Metazoa, vol. 39.
- Bush, A. M., R. K. Bambach, and G. M. Daley. 2007. Changes in theoretical ecospace utilization in
782 marine fossil assemblages between the mid-Paleozoic and late Cenozoic. *Paleobiology* 33:76–97.
- Bush, A. M., and P. M. Novack-Gottshall. 2012. Modelling the ecological-functional diversification
784 of marine Metazoa on geological time scales. *Biology Letters* 8:151–155.

- Cantalapiedra, J. L., J. L. Prado, and M. T. Alberdi. 2017. Decoupled ecomorphological evolution
786 and diversification in Neogene-Quaternary horses. *Science* 355:627–630.
- Carrano, M. T. 1999. What, if anything, is a cursor? Categories versus continua for determining
788 locomotor habit in mammals and dinosaurs. *Journal of Zoology* 247:29–42.
- Clyde, W. C., and P. D. Gingerich. 1998. Mammalian community response to the latest Paleocene
790 thermal maximum: an isotaphonomic study in the northern Bighorn Basin, Wyoming. *Geology*
26:1011–1014.
- 792 Cramer, B. S., K. Miller, P. Barrett, and J. Wright. 2011. Late Cretaceous-Neogene trends in deep
ocean temperature and continental ice volume: Reconciling records of benthic foraminiferal
794 geochemistry ($\delta^{18}\text{O}$ and Mg/Ca) with sea level history. *Journal of Geophysical Research: Oceans*
116:1–23.
- 796 Eronen, J. T., C. M. Janis, C. P. Chamberlain, and A. Mulch. 2015. Mountain uplift explains
differences in Palaeogene patterns of mammalian evolution and extinction between North
798 America and Europe. *Proceedings of the Royal Society B: Biological Sciences* 282:20150136.
- Ezard, T. H. G., A. Purvis, and H. Morlon. 2016. Environmental changes define ecological limits to
800 species richness and reveal the mode of macroevolutionary competition. *Ecology Letters*
19:899–906.
- 802 Felsenstein, J. 1985. Phylogenies and the comparative method. *The American Naturalist* 125:1–15.
- Figueirido, B., C. M. Janis, J. A. Pérez-Claros, M. De Renzi, and P. Palmqvist. 2012. Cenozoic
804 climate change influences mammalian evolutionary dynamics. *Proceedings of the National
Academy of Sciences* 109:722–727.
- 806 Foote, M. 2001. Inferring temporal patterns of preservation, origination, and extinction from
taxonomic survivorship analysis. *Paleobiology* 27:602–630.
- 808 Foote, M., and J. J. Sepkoski. 1999. Absolute measures of the completeness of the fossil record.
Nature 398:415–7.

- 810 Fraser, D., R. Gorelick, and N. Rybczynski. 2015. Macroevolution and climate change influence
phylogenetic community assembly of North American hoofed mammals. *Biological Journal of the*
812 *Linnean Society* 114:485–494.
- Freudenthal, M., and E. Martín-Suárez. 2013. Estimating body mass of fossil rodents. *Scripta*
814 *Geologica* 145:1–130.
- Fritz, S. A., J. Schnitzler, J. T. Eronen, C. Hof, K. Böhning-Gaese, and C. H. Graham. 2013.
816 Diversity in time and space: wanted dead and alive. *Trends in Ecology & Evolution* 28:509–16.
- Gelman, A. 2008. Scaling regression inputs by dividing by two standard deviations. *Statistics in*
818 *Medicine* pages 2865–2873.
- Gelman, A., J. B. Carlin, H. S. Stern, D. B. Dunson, A. Vehtari, and D. B. Rubin. 2013. *Bayesian*
820 *data analysis*. 3rd ed. Chapman and Hall, Boca Raton, FL.
- Gelman, A., and J. Hill. 2007. *Data Analysis using Regression and Multilevel/Hierarchical Models*.
822 Cambridge University Press, New York, NY.
- Graham, A. 2011. *A natural history of the New World: the ecology and evolution of plants in the*
824 *Americas*. University of Chicago Press, Chicago.
- Harnik, P. G., P. C. Fitzgerald, J. L. Payne, and S. J. Carlson. 2014. Phylogenetic signal in
826 extinction selectivity in Devonian terebratulide brachiopods. *Paleobiology* 40:675–692.
- Jamil, T., W. A. Ozinga, M. Kleyer, and C. J. F. Ter Braak. 2013. Selecting traits that explain
828 species-environment relationships: A generalized linear mixed model approach. *Journal of*
Vegetation Science 24:988–1000.
- Janis, C., J. Damuth, and J. M. Theodor. 2004. The species richness of Miocene browsers, and
830 implications for habitat type and primary productivity in the North American grassland biome.
832 *Palaeogeography, Palaeoclimatology, Palaeoecology* 207:371–398.
- Janis, C. M. 1993. Tertiary mammal evolution in the context of changing climates, vegetation, and
834 tectonic events. *Annual Review of Ecology and Systematics* 24:467–500.

- . 2008. An evolutionary history of browsing and grazing ungulates. Pages 21–45 in I. J. Gordon and H. H. T. Prins, eds. *The Ecology of Browsing and Grazing*. Springer-Verlag.
- Janis, C. M., J. Damuth, and J. M. Theodor. 2000. Miocene ungulates and terrestrial primary productivity: where have all the browsers gone? *Proceedings of the National Academy of Sciences* 97:7899–904.
- Janis, C. M., G. F. Gunnell, and M. D. Uhen. 2008. Evolution of Tertiary mammals of North America. Vol. 2. Small mammals, xenarthrans, and marine mammals. Cambridge University Press, Cambridge.
- Janis, C. M., K. M. Scott, and L. L. Jacobs. 1998. Evolution of Tertiary mammals of North America. Vol. 1. Terrestrial carnivores, ungulates, and ungulate-like mammals. Cambridge University Press, Cambridge.
- Janis, C. M., and P. B. Wilhelm. 1993. Were there mammalian pursuit predators in the tertiary? Dances with wolf avatars. *Journal of Mammalian Evolution* 1:103–125.
- Jernvall, J., and M. Fortelius. 2004. Maintenance of trophic structure in fossil mammal communities: site occupancy and taxon resilience. *The American Naturalist* 164:614–624.
- Legendre, S. 1986. Analysis of mammalian communities from the Late Eocene and Oligocene of Southern France. *Paleovertebrata* 16:191–212.
- Liow, L. H., M. Fortelius, E. Bingham, K. Lintulaakso, H. Mannila, L. Flynn, and N. C. Stenseth. 2008. Higher origination and extinction rates in larger mammals. *Proceedings of the National Academy of Sciences* 105:6097–6102.
- Liow, L. H., M. Fortelius, K. Lintulaakso, H. Mannila, and N. C. Stenseth. 2009. Lower Extinction Risk in Sleep or Hide Mammals. *The American Naturalist* 173:264–272.
- Lloyd, G. T., J. R. Young, and A. B. Smith. 2011. Taxonomic Structure of the Fossil Record is Shaped by Sampling Bias. *Systematic Biology* 61:80–89.
- Marcot, J. D. 2014. The fossil record and macroevolutionary history of North American ungulate

- 860 ungulate mammals: standardizing variation in intensity and geography of sampling. *Paleobiology*
40:237–254.
- 862 McKenna, R. T. 2011. Potential for Speciation in Mammals Following Vast , Late Miocene Volcanic
Interruptions in the Pacific Northwest. Masters. Portland State University.
- 864 Novack-Gottshall, P. M. 2007. Using a theoretical ecospace to quantify the ecological diversity of
Paleozoic and modern marine biotas Using a theoretical ecospace to quantify the ecological
866 diversity of Paleozoic and modern marine biotas. *Paleobiology* 33:273–294.
- 868 Pires, M. M., D. Silvestro, and T. B. Quental. 2015. Continental faunal exchange and the
asymmetrical radiation of carnivores. *Proceedings of the Royal Society B: Biological Sciences*
282:20151952.
- 870 Pollock, L. J., W. K. Morris, and P. A. Vesk. 2012. The role of functional traits in species
distributions revealed through a hierarchical model. *Ecography* 35:716–725.
- 872 Quental, T. B., and C. R. Marshall. 2013. How the Red Queen Drives Terrestrial Mammals to
Extinction. *Science* 341:290–292.
- 874 Raia, P., F. Carotenuto, F. Passaro, D. Fulgione, and M. Fortelius. 2012. Ecological specialization
in fossil mammals explains Cope’s rule. *The American Naturalist* 179:328–37.
- 876 Royle, J. A., and R. M. Dorazio. 2008. Hierarchical modeling and inference in ecology: the analysis
of data from populations, metapopulations and communities. Elsevier, London.
- 878 ———. 2012. Parameter-expanded data augmentation for Bayesian analysis of capture-recapture
models. *Journal of Ornithology* 152:521–537.
- 880 Royle, J. A., R. M. Dorazio, and W. a. Link. 2007. Analysis of Multinomial Models With Unknown
Index Using Data Augmentation. *Journal of Computational and Graphical Statistics* 16:67–85.
- 882 Royle, J. A., J. D. Nichols, M. Kéry, E. Ranta, and M. Kery. 2014. detection is of species when
Modelling occurrence and abundance imperfect 110:353–359.

- 884 Rubin, D. B. 1996. Multiple imputation after 18+ years. *Journal of the American Statistical
Assocaition* 91:473–489.
- 886 Silvestro, D., A. Antonelli, N. Salamin, and T. B. Quental. 2015. The role of clade competition in
the diversification of North American canids. *Proceedings of the National Academy of Sciences of
the United States of America* 112:8684–9.
- 890 Silvestro, D., J. Schnitzler, L. H. Liow, A. Antonelli, and N. Salamin. 2014. Bayesian estimation of
speciation and extinction from incomplete fossil occurrence data. *Systematic biology* 63:349–67.
- 892 Simberloff, D., and T. Dayan. 1991. The Guild Concept and the Structure of Ecological
Communities. *Annual Review of Ecology and Systematics* 22:115–143.
- 894 Slater, G. J. 2015. Iterative adaptive radiations of fossil canids show no evidence for
diversity-dependent trait evolution. *Proceedings of the National Academy of Sciences*
112:4897–4902.
- 896 Smith, F. A., J. Brown, J. Haskell, and S. Lyons. 2004. Similarity of mammalian body size across
the taxonomic hierarchy and across space and time. *The American Naturalist* 163:672–691.
- 898 Smits, P. D. 2015. Expected time-invariant effects of biological traits on mammal species duration.
Proceedings of the National Academy of Sciences 112:13015–13020.
- 900 Strömberg, C. A. E. 2005. Decoupled taxonomic radiation and ecological expansion of open-habitat
grasses in the Cenozoic of North America. *Proceedings of the National Academy of Sciences of
the United States of America* 102:11980–4.
- 902 Tomiya, S. 2013. Body Size and Extinction Risk in Terrestrial Mammals Above the Species Level.
The American Naturalist 182:196–214.
- 904 Valentine, J. W. 1969. Patterns of taxonomic and ecological structure of the shelf benthos during
Phanerozoic time. *Paleontology* 12:684–709.
- 906 Villéger, S., P. M. Novack-Gottshall, and D. Mouillot. 2011. The multidimensionality of the niche

- 908 reveals functional diversity changes in benthic marine biotas across geological time. *Ecology*
letters 14:561–8.
- 910 Wang, S. C., P. J. Everson, H. J. Zhou, D. Park, and D. J. Chudzicki. 2016. Adaptive credible
intervals on stratigraphic ranges when recovery potential is unknown. *Paleobiology* 42:240–256.
- 912 Wang, S. C., and C. R. Marshall. 2016. Estimating times of extinction in the fossil record. *Biology*
Letters 12:20150989.
- 914 Warton, D. I., B. Shipley, and T. Hastie. 2015. CATS regression - a model-based approach to
studying trait-based community assembly. *Methods in Ecology and Evolution* 6:389–398.
- 916 Wilson, J. B. 1999. Guilds, functional types and ecological groups. *Oikos* 86:507–522.
- Zachos, J. C., G. R. Dickens, and R. E. Zeebe. 2008. An early Cenozoic perspective on greenhouse
918 warming and carbon-cycle dynamics. *Nature* 451:279–283.
- Zachos, J. C., M. Pagani, L. Sloan, E. Thomas, and K. Billups. 2001. Trends, rhythms, and
920 aberrations in global climate 65 Ma to present. *Science* 292:686–693.

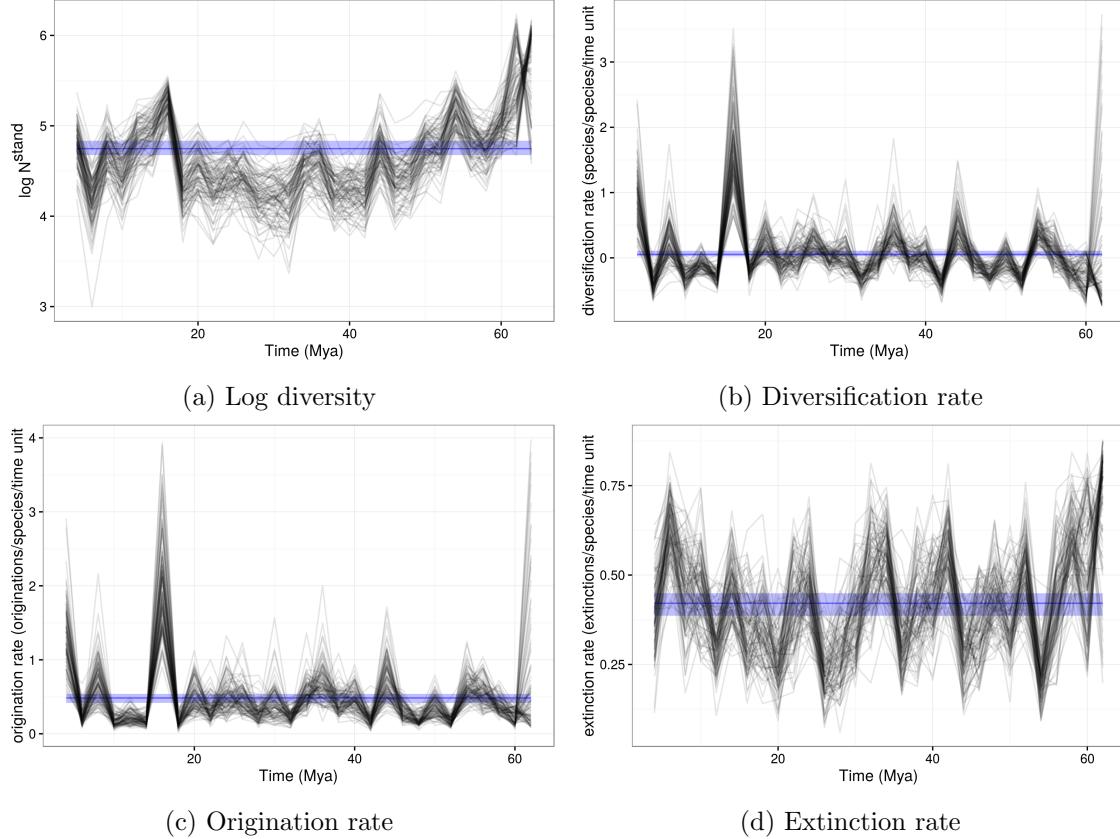


Figure 14: Posterior estimates of the time series of Cenozoic North American mammal diversity and its characteristic macroevolutionary rates; all estimates are from the birth-death model and 100 posterior draws are plotted to indicate the uncertainty in these estimates. The blue horizontal strip corresponds to the 80% credible interval of estimated mean standing diversity, diversification rate, origination rate, and extinction rate respectively; the median estimate is also indicated. What is also plotted is the The dramatic differences between diversity estimates at the first and second time points and the penultimate and last time points in this series are caused by well known edge effects in discrete-time birth-death models caused by $p_{-,t=1}$ and $p_{-,t=T}$ being partially unidentifiable (Royle and Dorazio, 2008); the hierarchical modeling strategy used here helps mitigate these effects but they are still present (Gelman et al., 2013; Royle and Dorazio, 2008). Diversification rate is in units of species gained per species present per time unit (2 My), origination rate is in units of species originating per species present per time unit, and extinction rate is in units of species becoming extinct per species present per time unit.

Table 15: Posterior probabilities of diversity N_t^{stand} or diversification rate D_t^{rate} being greater than average standing diversity \bar{N}^{stand} or average diversification rate \bar{D}^{rate} for the whole Cenozoic. The “Time” column corresponds to the top of each of the temporal bins. Diversification rate can not be estimated for the last time point because it is unknown how many more species originated or went extinct following this temporal bin. The estimates are from the birth-death model.

Time (Mya)	$P(N_t^{stand} > \bar{N}^{stand})$	$P(D_t^{rate} > \bar{D}^{rate})$
64.00	0.99	0.18
62.00	0.93	0.15
60.00	0.93	0.04
58.00	0.53	0.59
56.00	0.72	0.99
54.00	0.99	0.00
52.00	0.59	0.45
50.00	0.57	0.01
48.00	0.05	0.27
46.00	0.04	0.92
44.00	0.53	0.00
42.00	0.01	0.44
40.00	0.00	0.37
38.00	0.01	0.94
36.00	0.23	0.46
34.00	0.22	0.01
32.00	0.00	0.31
30.00	0.00	0.33
28.00	0.00	0.83
26.00	0.03	0.32
24.00	0.02	0.25
22.00	0.01	0.89
20.00	0.15	0.02
18.00	0.02	1.00
16.00	1.00	0.00
14.00	0.83	0.11
12.00	0.67	0.01
10.00	0.11	0.79
8.00	0.40	0.02
6.00	0.00	0.98
4.00	0.59	

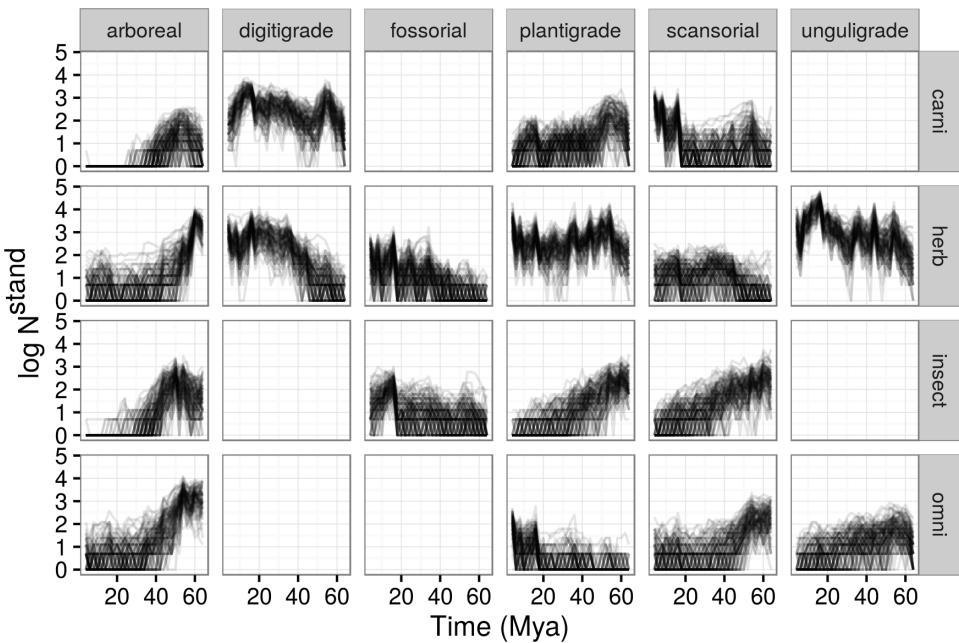


Figure 15: Posterior of standing log-diversity of North American mammals by ecotype for the Cenozoic as estimated from the birth-death model; 100 posterior draws are plotted to indicate the uncertainty in these estimates and what is technically plotted is log of diversity plus 1.

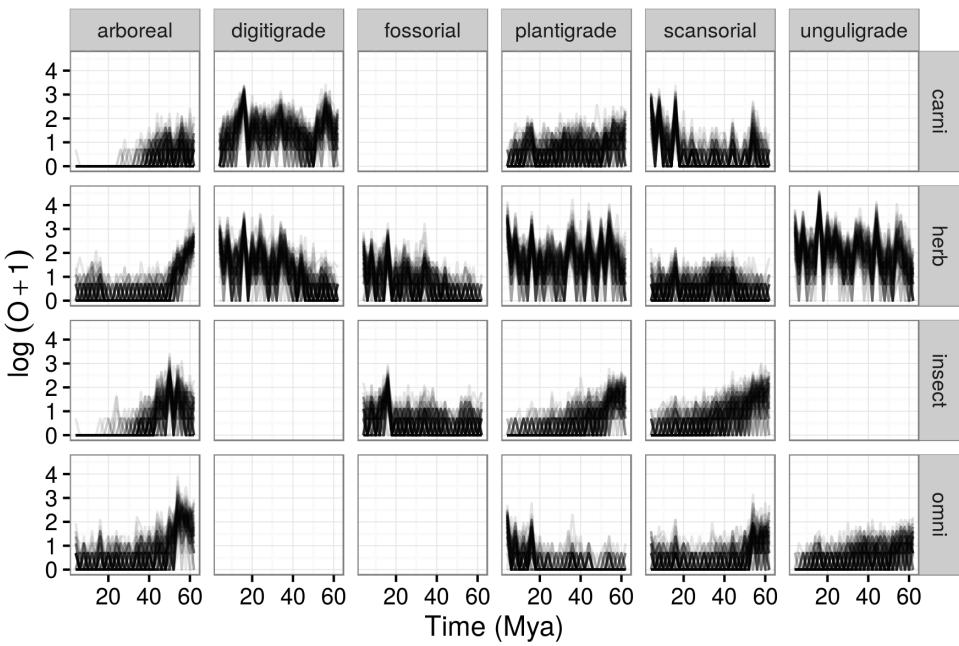


Figure 16: Posterior estimates of the number of origination events from one temporal bin to another, plotted at the bin they originate from. 100 posterior draws are plotted to indicate the uncertainty in these estimates. Also, what is plotted is log of the number of originations plus 1.

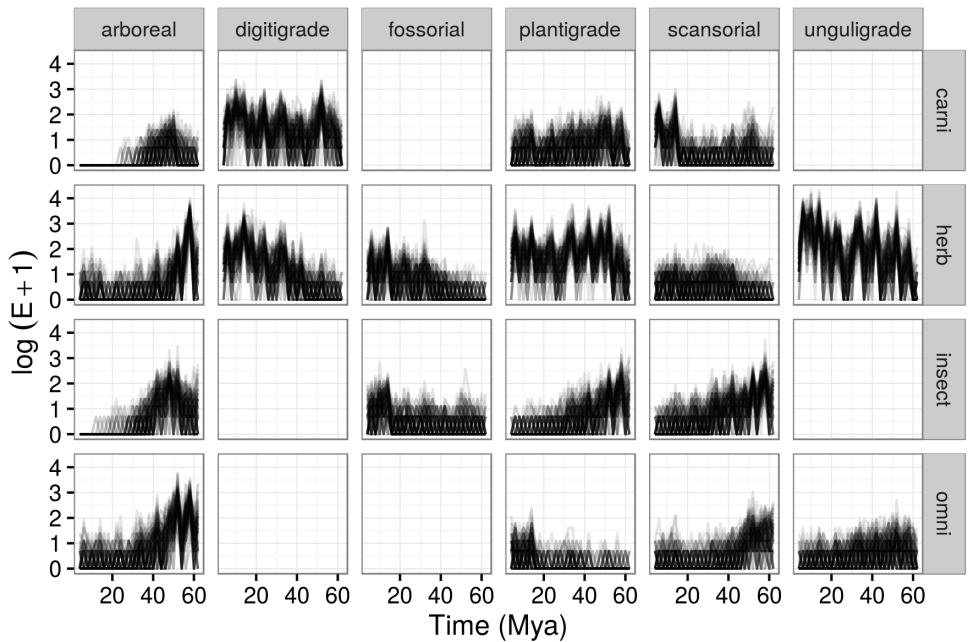


Figure 17: Posterior estimates of the number of extinction events from one temporal bin to another, plotted at the bin they go extinct from. 100 posterior draws are plotted to indicate the uncertainty in these estimates. Also, what is plotted is \log of the number of extinctions plus 1.