Climate Ecology as a Driver of Global Breeding Periods in Anurans

1

2 3 Bryan H. Juarez and Lauren A. O'Connell 4 5 Department of Biology; Stanford University; Stanford, California 94305, USA. 6 7 Co-Corresponding Authors Email: bryanhjuarez@gmail.com, loconnel@stanford.edu 8 9 ORCIDs: Juarez: 0000-0002-5474-596X; O'Connell: 0000-0002-2706-4077 10 11 Key Words: amphibian, frogs, reproduction, phenology, environmental change, conservation, 12 microhabitat, body size 13 Running Title: Global climate effects on amphibian reproduction 14 Author contributions: BHJ and LAO conceptualized the project; BHJ gathered the data and 15 performed all analyses; BHJ wrote the manuscript and both authors edited and discussed the 16 text. 17 Data accessibility: All data and code are publicly available on datadryad.org 18 (https://doi.org/10.5061/dryad.547d7wmc2) and included as supplementary material. Code is 19 also found at github.com/bhjuarez/breeding-period in R, Rmarkdown, and HTML formats. 20 Number of words: Abstract: 150; Main text: 4,708. 21 Number of references: 97. 22 Number of figures and tables: 4 Figures, 2 Tables.

Abstract

Climate can have profound effects on reproductive behavior and physiology, especially in ectothermic animals. In amphibians, the evolution of breeding period due to climate has received little attention despite its direct connection to fitness and relevance to conservation. We used phylogenetic comparative methods to analyze a global dataset of frog breeding periods, climate, body size, and microhabitat for 497 species (41 of 54 families). We found support for the hypothesis that breeding periods are longer in the warmer, wetter, aseasonal tropics and shorter in the colder, dryer, seasonal temperate zone. However, this latitudinal reproduction gradient is more complex than expected; the effect of climate depends on body size.

Additionally, breeding periods were similar across microhabitats. Overall, our model describes ~20% of variation in breeding periods and these results are robust to phylogenetic uncertainty. Our results set within an eco-physiology framework have broad implications for more targeted priorities in amphibian conservation.

#### Introduction

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59

60

61

Climate change is increasingly threatening global biodiversity with increasing extinctions as environments are altered (Arneth et al. 2020; Bellard et al. 2012; Thomas et al. 2004; Urban 2015). While climate change is a growing problem in the present, its past effects have molded the behavior and physiology of species still around today. Thus, by studying coevolution between organisms and the environment, we can better understand how climate change will differentially affect species survival. Importantly, evolutionary biology is taking a key role in conservation biology (Zizka et al. 2021), with climate change and anthropogenic impacts at the heart of these studies. As successful reproduction is necessary for survival and fitness, studies of the ecology and evolution of reproductive traits are needed to understand the effects of climate change on species reproductive success. For example, climate change has led to shifting clutch sizes, offspring sizes, and breeding season start times in vertebrates (Bronson 2009; Padilla Perez and Angilletta 2022; Pankhurst and Munday 2011; Telemeco et al. 2013; Todd et al. 2011; Visser et al. 2004). However, limited natural history observations across many species and vertebrate groups has stalled research that would help us better understand how climate impacts various reproductive behaviors. Therefore, ideal model systems for learning how climate affects reproductive behaviors and success include groups with abundant natural history observations and whose reproduction is tied to particular environments. Anurans (frogs and toads) are an ecologically and taxonomically diverse ectotherm group (>7,400 species; AmphibiaWeb 2022) whose reproduction and development is generally dependent on water availability during breeding. Their iconic calling behavior used for reproduction has made them a prime target for natural history studies, making their reproductive

behaviors reasonably well-known. Anurans inhabit a wide array of habitats around the world

63

64

65

66

67

68

69

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

including tropical rainforests, arid deserts, temperate forests, freezing tundras, and more. These habitats are used differently by various species which specialize in aquatic, terrestrial, fossorial (underground), or arboreal (above ground) microhabitats (Moen & Wiens 2017). Studies of amphibian reproduction and climate are dominated by advancements in our understanding of reproductive modes (e.g., Gomez-Mestre et al. 2012) or reproductive timing (Beattie 1985; Benard 2015; Neveu 2009; Reading 2003; Todd et al. 2011; While and Uller 2014). Comparatively little is known about the relationship between climate and breeding period (sensu Wells 1977), the duration of breeding within a population, which is unique from the duration of breeding across populations of a single species. This is despite the stunning contrast of tropical species that breed year-round to explosive-breeding species that limit reproduction to only a few days or weeks of the year (Wells 1977). Common knowledge in herpetology says there exists a latitudinal gradient in breeding periods, where tropical amphibians breed for generally longer periods compared to temperate amphibians (Wells 2007). However, the precise role of climate on breeding periods has yet to be identified across the phylogeny of anurans and further study is necessary given widespread amphibian declines (Campbell Grant et al. 2020; Fisher and Garner 2020). If climate has an effect on breeding periods, we can expect physiology to play a central role since frogs are especially sensitive to the environment as ectotherms and are reliant on water for reproduction. Natural selection has likely led to physiological changes such that populations maximize their reproductive investment leading to longer breeding periods in the context of specific climate exposure or microhabitat use (Huey et al. 2012; Williams et al. 2008). Thus, climate is tied to physiological and reproductive traits to the extent that individuals experience changes in regional climate and atmospheric conditions during the breeding period while in water or on land. These links have been shown previously, where body temperature is correlated with substrate temperature and air temperature in the toad *Bufo canorus* (Cunningham 1963). Both temperature and water are known breeding cues in frogs and warmer

temperatures likely lead to longer breeding periods due to greater food supply, reproductive investment, and egg formation in frogs (Girish and Saidapur 2000; Lardner and Loman 2003; Rastogi *et al.* 1983). Additionally, since reproduction in anurans is generally tied to water, we can expect increased water availability to result in longer breeding periods. These abiotic variables can also interact, as higher temperatures in drier environments may lead to increased rates of water loss. From this perspective, larger-bodied species with lower surface area:volume (SA:V) ratios may benefit from lower rates of water loss relative to smaller species. Alternatively, larger-bodied species may be at a disadvantage in sub-optimal environments due to greater resource demand and limited area for water or heat exchange. Species body size likely influences the effect climate has on breeding period due to trade-offs between reproduction and physiological performance over the course of the breeding period. Knowledge of how climate and body size interact to influence anuran breeding periods is unknown but is likely to allow for more targeted conservation efforts.

The goal of this study is to identify how ecological factors are related to reproduction (breeding periods) in anurans on a global scale. Specifically, we test (H1) whether longer breeding periods are the result of warmer, wetter climates, (H2) whether differences in body size significantly alter the way climate influences breeding periods, (H3) whether species with different non-breeding microhabitat preferences exhibit differences in breeding period, and (H4) whether climate differs significantly between species microhabitat and geographical (temperate and tropical) zones. Here, we provide the first quantitative assessment of the latitudinal gradient of reproduction in anurans, including climatic and possible physiological drivers.

#### Material and Methods

#### Climate data

109

110

111

112

113

114

115

116

117

118

119

120

121

122

123

124

125

126

127

128

129

130

131

132

133

We obtained coordinate data for 497 species representing 41 of 54 anuran families from GBIF.org (GBIF.org 2021a,b,c,d) accessed 3 October 2021. These queries were for presence coordinate data for non-fossil native observations lacking geospatial issues. We cleaned the data by removing coordinates points with >100km uncertainty or outside land boundaries of a high resolution (1:10m) world countries map using the R v4.2.1 (R Core Team 2021) package rworldxtra v1.01 (South 2012). We also removed coordinates incorrectly associated with and within 100m of biodiversity collections outside of the natural range of the species. Additionally, we removed Corythomantis greeningi since its coordinates mapped to a continent far outside its natural range. We restricted data for the invasive Cane Toad (Rhinella marina) to its naturally occurring range from southern United States to South America (AmphibiaWeb 2021). Next, we removed species with less than 3 unique observations from the dataset and retained only unique coordinates for each species to minimize the effects of spatial autocorrelation. Lastly, we removed data outside the spatial 95% Confidence Interval for each species distribution with a minimum of 30 occurrences using the median absolute deviation (mad) method (Huber 1981). This resulted in a final coordinate dataset of 822,953 observations (mean = 1,656 observations per species) after removing species without behavior or body size (see below) or missing from the phylogeny of Jetz and Pyron (2018). Data cleaning was performed using the CoordinateCleaner package v2.0-20 (Zizka et al. 2019). We extracted seven variables from each coordinate using the WorldClim 2.1 (Fick & Hijmans 2017) and envirem (Title & Bemmels 2018) databases at 30 seconds (~1km²) resolution. Our data included variables affecting temperature and water availability and covering all possible breeding period values (0–365 days). The final dataset included annual mean temperature, temperature seasonality, annual precipitation, precipitation seasonality, annual

potential evapotranspiration (PET), PET seasonality, and topographic wetness. Topographic wetness is the land's ability to retain water obtained from estimates of upslope incurrent areas and local slope, which is correlated to vascular plant species richness and soil moisture (see Sørensen *et al.* 2005). We calculated species means for each climate variable to analyze the dataset in a phylogenetic comparative framework.

#### Behavior data

134

135

136

137

138

139

140

141

142

143

144

145

146

147

148

149

150

151

152

153

154

155

We collected species means of breeding period (sensu Wells 1977) for each species in this study. To obtain species means, we collected approximate start and end dates for breeding periods from the primary literature or AmphibiaWeb (2022) and averaged all available breeding period data for individual populations. We then calculated the difference in calendar days between the start and end dates, assuming 365 calendar days. Since some dates were reported as qualitative data (e.g., "May", or "early May"), we ascribed values of either the first or last day of the month when breeding periods were reported as starting or ending on a specific month without a specific date given. When breeding periods were reported as the "early" or "late" parts of a month, we matched this to the end of the 1<sup>st</sup> week and 3<sup>rd</sup> week, respectively. Rarely, breeding periods were reported as specific seasons (e.g., "Summer" or "wet season") and we defined these as the number of days corresponding to each season and country. Populations from species which are explosive breeders are typically associated with breeding periods lasting only a few days to weeks, but never more than a month (Wells 1977). Since we never observed published data reporting explosive breeding over a period longer than 2 weeks, we conservatively assigned a period of 2 weeks to species whose only description of the breeding period was "explosive breeding".

Body size, habitat, and phylogenetic data

We collected maximum male body sizes (snout-vent length) and habitat data for every species in this study from the primary literature and other datasets (AmphibiaWeb 2022; IUCN 2022; Moen and Wiens 2017; Womack and Bell 2020). Species were assigned to microhabitat categories following Moen and Wiens (2017) and included: (1) aquatic, (2) arboreal, (3) burrowing, (4) terrestrial, and (5) torrential. We also included partial categories such as: (6) semi-arboreal, (7) semi-aquatic, and (8) semi-burrowing. Finally, we calculated phylogenetic covariance matrices from the consensus tree of Jetz and Pyron (2018) to incorporate evolutionary history as weights in phylogenetic regression. Additionally, we used 1,000 samples from the pseudoposterior distribution of trees from Jetz and Pyron (2018) to assess models in terms of phylogenetic uncertainty (see below).

#### Statistical analysis and visualization

All analyses were performed in R. We used *ggtree* v3.4.0 (Yu 2020; Yu *et al.* 2018; Yu *et al.* 2017) to plot average species breeding periods and temperate-tropical geographical zones. We analyzed the data using phylogenetic generalized least squares (PGLS; Grafen 1989; Martins & Hansen 1997) implemented through *RRPP* v1.3.0 (Collyer and Adams 2018, 2019). Notably, *RRPP* uses a residual randomized permutation procedure to fit models, evaluate significance, and calculate effect sizes for model terms. We used 9,999 iterations for significance testing in all models. We estimated phylogenetic covariance matrices using *ape* v5.6-2 (Paradis & Schliep 2019).

We scaled each climate variable and performed a principal component analysis (PCA) to avoid collinearity issues in our models. To test our hypotheses, we used phylogenetic multiple regression to examine the relationship of breeding period, all 7 principal components (PCs), and body size. To account for allometry of body size and since the natural logarithm of surface area-

to-volume ratio is expected to scale linearly with the natural logarithm of body size (Castro *et al.* 2021; Klein *et al.* 2016; Santini *et al.* 2018), we used the natural logarithm of body size in our models. Additionally, we fit a phylogenetic ANCOVA model which included interactions between each PC and body size. We also used ANOVA to test whether adding these interactions resulted in a significant improvement in model performance. To aid in interpretation of these models, we used phylogenetic regression to determine the relationship between breeding period and each individual climate variable. We determined the effects of phylogenetic uncertainty in our models by rerunning the ANCOVA model once each for 1,000 trees from the pseudoposterior distribution of Jetz and Pyron (2018), following Baken *et al.* (2021). From these analyses we obtained one distribution of 1,000 Z-scores per model term and visualized the proportion of trees under which terms were not significant. We plotted these distributions using ggridges v0.5.3 (Wilke 2021).

We used phylogenetic ANOVA to test whether breeding periods differ between microhabitat categories, followed by evaluation of pairwise comparisons. We controlled for Type I error rates due to multiple comparisons using the Sidak-like step-down procedure, which is more powerful than the Sidak and Bonferroni methods (Holland & Copenhaver 1987), using mutoss v0.1-12 (MuToss Coding Team et al. 2017). We used a phylogenetic factorial MANCOVA to determine how climate (all 7 PC's) differs among microhabitats and temperate/tropical zones. Finally, we visualized these data by plotting conver hulls for groupings of microhabitat and geographical zone using geomorph v4.0.4 (Adams et al. 2021; Baken et al. 2021b).

# Results

Frog breeding periods follow a latitudinal gradient

We examined phylogenetic patterns of average breeding period (mean = 148 days, range = 1–365 days, std. error = 4.5 days) and geographical zone (temperate/tropical) (Fig. 1) and found species with longer breeding periods tend to live in tropical regions, whereas temperate species breed for shorter periods. However, this pattern has several exceptions, such as temperate species with long breeding periods and tropical species with very short breeding periods. Overall, temperate species breed for an average of 127.8 days (median = 107 days, range = 1–365 days, std. error = 5.87 days) and tropical species breed for an average of 167.3 days (median = 181 days, range = 5–365 days, std. error = 6.52 days). Climate principal components (PCs) 1–3 explained 90% of data variance (Table 1; SI 1). We found variable loadings for PC1 to be similar to known latitudinal climate differences and determined PC1 was highly correlated (*r* = -0.94) with the absolute value of median latitude for each species (SI 1).

## Climate and body size impact breeding periods

As body size can play a large role in thermal physiology and desiccation, we compared model fits between two models with and without interactions between climate and body size (snout-vent length). Indeed, including body size substantially improved model fit (F = 2.40, p = 0.0082; Table 2). Overall, this model explained 18% of variation in breeding periods.

We next describe this model in the context of phylogenetic uncertainty and given the individual effects of each climate variable (see SI 1). We found larger species from the tropics (high scores on PC1) tend to exhibit longer breeding periods relative to smaller species and the magnitude of this effect can be up to 60-90 days. However, this pattern reverses in the temperate zone where relatively smaller species exhibit longer breeding periods (Fig. 2A). For

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240

241

242

243

244

245

246

example. Rana pretiosa (median latitude = 42.60) can be 14mm larger than Rana macrocnemis (median latitude = 45.33) but exhibit a breeding period 48 days shorter (61-13=48 days). This effect was due to latitudinal differences in temperature, temperature seasonality, precipitation, and potential evapotranspiration (PET) seasonality, each showing similar effects when analyzed individually (SI 1 Table 1). Since PET did not have a significant individual effect on breeding periods, it likely does not affect breeding periods although it loads highly onto PC1; we do not interpret PET further for any PC. The effect of PC1 was robust to phylogenetic uncertainty and significant in 89.7% of surveyed trees while its combined effect with body size was supported in 64.2% of trees (Fig. 3). Independent of latitude (PC1), we found a significant impact of PC2 and body size on breeding periods (Fig. 2B). Specifically, we found that relatively larger species breed for longer periods in rainy environments with low topographic wetness (high values for PC2) and smaller species breed for longer periods in environments with little rain but high topographic wetness. Although topographic wetness had no individual effect on breeding periods, it did exhibit a significant impact when paired with body size. This trend showed moderate support across surveyed phylogenies with significance of PC2 and the interaction with size supported in 0.714% and 0.704% of trees, respectively (Fig. 3). Additionally, we found support for a size-independent role of PC3, where species living regions with low precipitation seasonality (PC3) exhibit longer breeding periods (Fig. 2C). This pattern was significant in 87.9% of surveyed trees compared to only 31.4% of trees when taking size into account (Fig. 3). Finally, we also found a size-independent effect of PC7 associated with warmer temperatures, independent of temperature effects of other PC's, where longer breeding periods are associated with warmer temperatures. This pattern was supported in 66.9% of surveyed phylogenetic histories.

## Breeding periods are not different between microhabitats

The microhabitats species occupy can have different effects on breeding period depending on differences in temperature and moisture within temperate or tropical regions. Therefore, we tested whether anuran species with different microhabitat preferences exhibit differences in breeding periods. We initially found support for this relationship, where species from various microhabitats have different breeding periods (F = 4.09, p = 0.0062). However, pairwise differences between microhabitats were not significant after accounting for multiple comparisons (see SI 1 Table 2). In contrast to our expectation, we found large overlap in breeding periods across microhabitat categories (Fig. 4A), including a wide breeding period range associated with burrowing and semi-burrowing species. Semi-arboreal species had the greatest proportion of species with a long breeding period. We next used a factorial MANCOVA used examine climatic differences corresponding to combinations of microhabitat and geographical zone (i.e., their interaction; F = 2.05,  $r^2 = 0.02$ , Z = 2.27, p = 0.0124), but none of these were significant after correcting for multiple comparisons (see SI 1 Table 3). Finally, we found significant climate differences among microhabitats (F = 7.01,  $r^2 = 0.08$ , Z = 3.77, p =0.0001: Fig. 4B) and between temperate and tropical zones (F = 87.46,  $r^2 = 0.14$ , Z = 8.93,  $\rho =$ 0.0001) but we did not find pairwise differences in the former after accounting for geographical zone (see SI 1 Table 4).

# Discussion

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

This study documents the evolution of reproductive behavior associated with climate in one of the largest vertebrate groups. Here we present the first quantitative evidence, to our knowledge, of the latitudinal gradient of breeding period in anurans and its relationship to climate. Our overall findings support longer breeding periods where frogs invest more time into reproduction in warmer and wetter habitats with little seasonality. We also discovered that the

time invested into reproduction, within the context of particular climates, is influenced by body size. We discuss our results in the context of physiology and reproduction and suggest how this new information can guide conservation management priorities of anurans in different climates.

271

272

273

274

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

292

293

294

295

296

Body size played an important role in determining how climate affects breeding periods. This is likely due to its effect on physiological processes such as regulating body temperature and water loss, where larger frogs are better at maintaining body temperatures and are slower to dessicate (Castro et al. 2021; Klein et al. 2016; Santini et al. 2018; Tracy et al. 2010). The size-dependent effect of climate on anuran reproduction is most notable when comparing 1) temperate versus tropical environments, and 2) regions with high rainfall and limited propensity to hold water versus regions with low rainfall and high topographic water retention. These results are consistent with our hypothesis that larger body sizes are associated with greater resource demands and limited area for water and heat exchange, resulting in shorter breeding periods in environments with little rain, cooler temperatures, and low topographic water retention. High temperature seasonality and high potential evapotranspiration seasonality also seem to shorten breeding periods in larger species, possibly due to their combined effect and this may be due to meeting physiological demands in relatively unpredictable environments. Interestingly, environments which shorten breeding periods in larger species seem to be less detrimental to smaller species, possibly due to their lower resource demand and relatively greater area for water and heat exchange. Furthermore, larger-bodied species may be at a disadvantage in cool and dry environments because the amount of energy used for maintaining homeostasis (hydration) may limit the amount of reproductive investment females are able to make, since egg formation is energetically costly for females. This seems likely, given that we know larger frog species exhibit greater reproductive investment in the form of forming larger clutches and larger eggs (Prado & Haddad 2005; Silva et al. 2020). Body size is also correlated with reproductive investment in seasonally breeding endotherms like lagomorphs, whose body size is correlated with neonatal mass, gestation time, and maternal reproductive effort (Swihart

298

299

300

301

302

303

304

305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

1984). Temperature and water preferences and limits can evolve in a correlated fashion (Bonetti and Wiens 2014; Guevara-Molina *et al.* 2020; Köhler *et al.* 2011) and likely involve underexplored molecular and morphological traits associated with the skin and body size in amphibians (Finn *et al.* 2014; Shibata *et al.* 2014; Tracy *et al.* 2010). How temperature and water limits have shaped anuran breeding physiology is unclear and is an area that needs more research.

Ecology, in addition to physiology, plays an important role in determining breeding periods. Our results suggest seasonal environments present the challenge of overcoming unreliable cues for reproduction, similar to findings at shorter ecological scales (Todd et al. 2011). Here, we found size-dependent patterns, themselves independent of latitude, where breeding periods in smaller species decline in rainier environments with lower topographic wetness. These qualities are characteristic of high-altitude rainforests, or cloud forests, such as those found in Central and South America, which hold much of the world's anuran diversity (Cortés et al. 2008; Duellman 1988; Giaretta et al. 1999). One explanation for this pattern is that cloud forests exhibit relatively rapid local drainage of water resources necessary for prolonged amphibian reproduction. Another possible explanation is that areas with low topographic wetness and correspondingly low vascular plant species richness (Sørensen et al. 2005) presents a challenge for prolonged breeding periods in smaller species by limiting protection from canopy cover or reproductive opportunities associated with leaf- or phytotelma-breeding species (Donnelly and Guyer 1994; Poelman et al. 2013; Schulte et al. 2010). Predation risk may also differ between small and large species, where larger species are less reliant on canopy cover due to their large size reducing predation pressure (Nakazawa et al. 2013), resulting in longer breeding periods. Future research investigating the role of plant diversity and predator-prey relationships in anurans seems likely to explain this variability in amphibian reproduction.

323

324

325

326

327

328

329

330

331

332

333

334

335

336

337

338

339

340

341

342

343

344

345

346

347

Microhabitat use across and within species undoubtedly plays a critical role in how animals experience broader regional climates and together they define climate exposure (Huey et al. 2012; Williams et al. 2008). While ambient and body temperatures are connected (Cunningham 1963), water stream and air temperatures are also correlated to atmospheric temperatures in both warmer and colder climatic zones (Erickson and Stefan 1996; Mohseni and Stefan 1999; Pilgrim et al. 1995). However, support for our hypothesis of differences in breeding periods between species occupying different microhabitats was lacking. Instead, we found a minority of burrowing species exhibit explosive breeding (extremely short breeding periods) and an overabundance of semi-arboreal species which breed for longer periods. The latter is likely related to radiations of small semi-arboreal frogs in the tropics of Central and South America, one of the most diverse regions of the world. Perhaps unsurprisingly, we did not find discernible climatic differences between microhabitats. This could reflect a low correlation between regional climate and ambient conditions within microhabitats (e.g., a burrow). However, such a strong role of microhabitat buffering seems unlikely if we expect, for example, microhabitats within a temperate desert to resemble each other more closely than microhabitats within a tropical rainforest. Instead, the lack of breeding period differences among microhabitats most likely reflects the ecological breadth of frog communities and global convergent evolution of microhabitat use across anurans (Moen et al. 2015). Further study of the relationships between microhabitat, reproductive traits, and climate would likely result in additional model improvements and better predictions of breeding period. For example, previous work found decreased fecundity in anurans was associated with warmer winters within species (Benard 2015; Neveu 2009; Reading 2003) and warm, wet climates across terrestrial egg-laying species (Gomez-Mestre et al. 2012). Notably, species in seasonal environments may have evolved larger clutches in response to shorter breeding periods (Sheridan 2009). Future research will continue to disentangle the relationships between behaviors, the environment, and further investigate the role of life history in determining the effect of climate on reproductive ecology.

Unfortunately, we still know very little about the life histories of many frogs. A sustained push for deeper understanding of life history diversity is necessary for identifying causal mechanisms related to eco-evolutionary patterns.

348

349

350

351

352

353

354

355

356

357

358

359

360

361

362

363

364

365

366

367

368

369

370

371

372

Adaptation to changing climates is a key part of amphibian evolutionary history and should be centered in future amphibian conservation efforts. Previous research has shown conclusive evidence of adaptation to changing climates over ecological timescales (<40 yrs.: Beattie 1985; Benard 2015; Freidenburg and Skelly 2004; Neveu 2009; Reading 2003; Skelly and Freidenburg 2008; Skelly et al. 2007; Todd et al. 2011). This work, along with previous studies on breeding start date (Beattie 1985; Ulloa et al. 2019; While and Uller 2014), identify both water and temperature as important environmental cues for the start of the breeding period. Associated neuroendocrine systems, thermal preferences, and critical thermal limits seem to show flexibility in responding to changing climate, but this might change as the effects of climate change become more severe. Huey et al. (2012) provide a model used to assess extinction vulnerability over time, which can be affected by behavioral thermoregulation, acclimation, and adaptation. Unfortunately, most anuran species live in warm, dry environments and these same species exhibit narrow climatic niche breadths (Bonetti & Wiens 2014) which can restrict dispersal (Chejanovski & Wiens 2014). The current study shows how species with restricted breeding periods living in the cold, dry, and seasonal environments may be less tolerant of climate change. Within the temperate zone, larger frogs with greater resource demands may be particularly vulnerable, while smaller frogs are most vulnerable in the tropics. Independent of latitude differences, we also found larger species may be vulnerable in areas receiving little rain. Without physiological adaptation to climate change, frogs might mistime their reproduction and be faced with either extinction, or potentially speciation through temporal isolation of reproduction. This might be one of the possible mechanisms linking diversification rates and rates of climatic niche change in anurans (Moen & Wiens 2017).

374

375

376

377

378

379

380

381

382

383

384

385

386

387

388

389

390

391

392

393

394

395

396

397

Amphibians are facing unprecedented threats due to climate change from the lens of eco-physiology. Notably, climate change seems to amplify the current amphibian chytrid fungus epidemic due to size-related osmoregulatory costs (Wu et al. 2018). Given the several factors contributing to growing extinction risks faced by frogs and toads, an attempt to maintain reproductive fitness by maximizing breeding duration seems critical. Overall, our research suggests a regional-based approach toward conserving species' flexibility in the timing and period of breeding. There are multiple factors impacting amphibian breeding periods but regional climate has a noticeable average effect on breeding periods worldwide. Extinction risks due to the mistiming of reproduction or shortened breeding periods might be highest in smaller tropical species which are common in tropical lowlands. Species living in tropical lowlands along with those living in cloud forests have both been identified as being under threat of extinction (von May et al. 2019; Meza-Parral and Pineda 2015; Murrieta-Galindo et al. 2013; Ponce-Reyes et al. 2013). Further research is needed to understand the neuroendocrine cues used for timing reproduction in frogs, thermal preferences, and the role of molecular or morphological adaptations related to water and heat stress. Without this research, future temperature instability will increase extinction risk without either dispersal to suitable habitats or the ability to acclimate and adapt to their changing environment (Huey et al. 2012). Perhaps the existence of tropical explosive breeders shows capacity for adaptation and persistence in the face of climate change. For example, species faced with limited breeding periods may also adapt to produce more eggs (Sheridan 2009) and habitat shifts associated with thermal preference and choosing egg-laying sites may help mitigate negative thermal climatic effects on tadpoles (Pintanel et al. 2022; Todd et al. 2011). Amphibians have much to benefit from dedicated conservation efforts to protect large and diverse landscapes allowing dispersal, protecting plant diversity, and preserving a thermally-diverse range of habitat options. Similar efforts have been successful in partially restoring amphibian diversity in cloud forests (Díaz-García et al. 2020). Future research might seek to combine environmental, behavioral, and physiological data to assess extinction risk and habitat suitability (Greenberg and Palen 2021; Griffis-Kyle 2016; Huey 1991).

#### Acknowledgements

We thank the O'Connell and Hadly labs at Stanford University for providing valuable feedback on previous versions of this manuscript. This material is based upon work supported by the NSF Postdoctoral Research Fellowships in Biology Program under Grant No. 2109850 (to BHJ). BHJ and LAO are supported by the New York Stem Cell Foundation. LAO is a New York Stem Cell – Robertson Investigator. We are indebted to the following museums for making their data available online: the Department of Herpetology, California Academy of Sciences (CAS); the Museum of Southwestern Biology (MSB); the Museum of Vertebrate Zoology, University of California, Berkeley (MVZ); and the University of Washington Burke Museum (UWBM).

#### **Tables**

**Table 1. PCA summary and loadings of climate variables.** PCA based on means for 497 species. Cumulative proportion of variance is given in parentheses next to PC heading. Loadings with a magnitude >=0.30 are in bold. Asterisks indicate variables with a significant individual effect on breeding periods and carets indicate variables with a significant interaction with body size (see SI).

Variable	PC1 (52)	PC2 (76)	PC3 (90)	PC4 (94)	PC5 (97)	PC6 (99)	PC7 (100)
Annual Mean Temperature*^	0.47	-0.28	0.08	-0.02	0.24	0.15	0.78
Temperature Seasonality*^	-0.47	-0.14	-0.09	0.44	0.25	0.70	0.04
Annual Precipitation*^	0.38	0.38	0.13	0.75	0.27	-0.21	-0.14
Precipitation Seasonality*^	0.15	-0.26	-0.90	0.23	-0.16	-0.15	0.00

Annual Potential Evapotranspiration (PET)	0.42	-0.4	0.00	-0.23	0.47	0.21	-0.59
PET Seasonality*^	-0.45	-0.28	0.01	0.02	0.59	-0.60	0.11
Topographic Wetness^	0.01	-0.67	0.40	0.37	-0.47	-0.15	-0.10

Table 2. Model summary of the effects of climate and body size on breeding periods in 497 frog species. Climate shown as PC's. Body size is snout-vent length (SVL) in mm. Breeding periods are species means. Df = degrees of freedom, SS = sum of squares, MS = mean square, Rsq = coefficient of determination, F = F-statistic, Z = standardized effect size (Z-score), and P = p-value. Rows with significant terms (p < 0.05) are in bold.

Term	Df	ss	MS	Rsq	F	Z	P
PC1	1	6279.32	6279.32	0.05	26.34	3.69	1e-04
PC2	1	2443.75	2443.75	0.02	10.25	2.59	0.0017
PC3	1	6756.84	6756.84	0.05	28.34	3.83	1e-04
PC4	1	22.69	22.69	0	0.10	-0.71	0.7519
PC5	1	734.96	734.96	0.01	3.08	1.39	0.0810
PC6	1	601.68	601.68	0	2.52	1.22	0.1152
PC7	1	1643.48	1643.48	0.01	6.89	2.12	0.0096
SVL	1	1058.17	1058.17	0.01	4.44	1.71	0.0389
PC1:SVL	1	1555.92	1555.92	0.01	6.53	2.07	0.0115

PC2:SVL	1	2228.56	2228.56	0.02	9.35	2.49	0.0020
PC3:SVL	1	831.79	831.79	0.01	3.49	1.51	0.0599
PC4:SVL	1	0.12	0.12	0	0	-2.05	0.9812
PC5:SVL	1	30.6	30.6	0	0.13	-0.62	0.7241
PC6:SVL	1	57.72	57.72	0	0.24	-0.32	0.6277
PC7:SVL	1	103.47	103.47	0	0.43	0.02	0.5090
Residuals	481	114677.84	238.42	0.82	NA	NA	NA
Total	496	139026.91	NA	NA	NA	NA	NA

Figure Legends

**Figure 1. Phylogenetic differences in average breeding period and geographic region across 497 species of frogs. A.** Phylogeny of 497 species with breeding period and region mapped at the tips. Node labels identify the 4 highest-sampled families in our dataset. Inner ring shows tropical versus temperate regions. The tropical zone is shown in red and the temperate zone is shown in blue. Outer ring shows the breeding period. Species with shorter breeding periods are shown in lighter colors while species with longer breeding periods are shown in darker colors. **B.** Boxplot showing differences in breeding period between temperate (N = 237) and tropical (N = 260) species. Temperate species breed for an average of 127.8 days (median = 107 days, range = 1–365 days, std. error = 5.87 days) and tropical species breed for an average of 167.3 days (median = 181 days, range = 5–365 days, std. error = 6.52 days).

Combined, anurans breed for a mean of 148 days (range = 1–365 days, std. error = 4.5 days).

**Figure 2. Climate and body size (snout-vent length) influence breeding periods in 497 frog species.** All relationships are significant at p < 0.0115. Regression lines for A,B represent effects at 10th and 90th percentiles of body size (snout-vent length; 24 and 80 mm). **A.** High scores on PC1 represent warm, rainy areas with low temperature and annual potential evapotranspiration seasonality and are

444

445

446

447

448

449

450

451

452

453 454

455

456

457

458

459

460 461

462

463

464

465

466

467

correlated with latitude. Larger species have longer breeding periods in the tropics; smaller species have longer breeding periods in the temperate zone. Combined effects of PC1 and body size yield Rsq = 0.07. **B.** High PC2 scores are regions with relatively higher annual precipitation but low topographic wetness. Larger species have longer breeding periods in rainy climates with low topographic wetness. Smaller species have longer breeding periods in less rainy climates with high topographic wetness. The combined effect of PC2 and body size yields Rsq = 0.05 C. High PC3 scores are regions with relatively lower precipitation seasonality but higher topographic wetness. Regression line shown in red. Longer breeding periods are common in areas with lower precipitation seasonality and higher topographic wetness. Rsq = 0.05. **D.** Density plot showing the distribution of raw (untransformed) body sizes of species used in this study. Figure 3. Effect of phylogenetic uncertainty on effect sizes (z-scores) relating climate and body size to breeding periods. Z-scores for each model term calculated from 1,000 trees of the pseudoposterior distribution of Jetz and Pyron (2018). PC's are climate variables; SVL is snout-vent length. Vertical line denotes the Z = 1.645 significance cutoff for empirically generated Z-scores. Values to the right of distributions indicate percent of significant Z-scores per distribution. Red highlights portions of the density within the 10th percentile. Figure 4. Breeding period and climate differences between microhabitats across 497 frog species. A. Average breeding periods do not differ between microhabitats: most ranges overlap, semi-arboreal species stand out in having the greatest proportion of species with longer breeding periods. B. Scatterplot of first two PCA axes corresponding to climate variables. Polygon outlines encase species in microhabitats corresponding in color. PC1 is correlated with absolute latitude (r = -0.94). Temperate species denoted by circles and encased in solid lines. Tropical species denoted by triangles and encased in dashed lines.

475

476

477

478

479

480

481

482

483

484

485

486

487

488

489

490

491

492

493

494

495

496

References Adams, D.C., Collyer, M.L., Kaliontzopoulou, A. & Baken E.K. (2021). Geomorph: Software for geometric morphometric analyses. R package version 4.0.2. https://cran.rproject.org/package=geomorph. AmphibiaWeb. (2022). Amphibian Species By the Numbers. Available at: https://amphibiaweb.org/amphibian/speciesnums.html. University of California, Berkeley, CA, USA. Last accessed 12 July 2022. AmphibiaWeb. (2021). Rhinella marina: Marine Toad. Available at: https://amphibiaweb.org/species/229. University of California, Berkeley, CA, USA. Last accessed 12 July 2022. Arneth, A., Shin, Y.-J., Leadley, P., Rondinini, C., Bukvareva, E., Kolb, M., et al. (2020), Post-2020 biodiversity targets need to embrace climate change. Proc. Natl. Acad. Sci. U. S. A., 117, 30882-30891. Baken, E.K., Mellenthin, L.E. & Adams, D.C. (2021a). Is salamander arboreality limited by broadscale climatic conditions? PLoS ONE, 16, e0255393. https://doi.org/10.1371/journal.pone.0255393. Baken, E. K., Collyer, M.L., Kaliontzopoulou, A. & Adams, D.C. (2021b). geomorph v4.0 and qmShiny: enhanced analytics and a new graphical interface for a comprehensive morphometric experience. Methods Ecol. Evol., 12, 2355-2363. Beattie, R.C. (1985). The date of spawning in populations of the Common frog (Rana temporaria) from different altitudes in northern England. J. Zool., 205, 137–154. Bellard, C., Bertelsmeier, C., Leadley, P., Thuiller, W. & Courchamp, F. (2012). Impacts of climate change on the future of biodiversity. *Ecol. Lett.*, 15, 365–377.

498

499

500

501

502

503

504

505

506

507

508

509

510

511

512

513

514

515

516

517

518

519

520

521

522

Benard, M.F. (2015). Warmer winters reduce frog fecundity and shift breeding phenology, which consequently alters larval development and metamorphic timing. Glob. Chang. Biol., 21, 1058-1065. Bonetti, M.F. & Wiens, J.J. (2014). Evolution of climatic niche specialization: a phylogenetic analysis in amphibians. *Proc. Biol. Sci.*, 281, 20133229. http://dx.doi.org/10.1098/rspb.2013.3229. Bronson, F.H. (2009). Climate change and seasonal reproduction in mammals. Philos. Trans. R. Soc. Lond. B Biol. Sci., 364, 3331–3340. Campbell Grant, E.H., Miller, D.A.W. & Muths, E. (2020). A synthesis of evidence of drivers of amphibian declines. Herpetologica, 76, 101–107. Castro, K.M.S.A., Amado, T.F., Olalla-Tárraga, M.Á., Gouveia, S.F., Navas, C.A. & Martinez, P.A. (2021). Water constraints drive allometric patterns in the body shape of tree frogs. Sci. Rep., 11, 1218. Chejanovski, Z.A. & Wiens, J.J. (2014). Climatic niche breadth and species richness in temperate treefrogs. J. Biogeogr., 41, 1936–1946. Collyer, M.L. & Adams, D.C. (2018). RRPP: An R package for fitting linear models to high-dimensional data using residual randomization. Methods Ecol. Evol., 9, 1772–1779. Collyer, M.L. & Adams, D.C. (2019). RRPP: Linear Model Evaluation with Randomized Residuals in a Permutation Procedure. https://CRAN.R-project.org/package=RRPP. Cortés, A.M., Ramírez-Pinilla, M.P., Suárez, H.A. & Tovar, E. (2008). Edge effects on richness, abundance and diversity of frogs in Andean cloud forest fragments. S. Am. J. Herpetol., 3, 213-222. Cunningham, J.D. (1963). Additional observations on the ecology of the Yosemite Toad, Bufo canorus. Herpetologica, 19, 56-61. Díaz-García, J.M., López-Barrera, F., Toledo-Aceves, T., Andresen, E. & Pineda, E. (2020). Does forest restoration assist the recovery of threatened species? A study of cloud forest

523 amphibian communities. Biol. Conserv., 242, 108400. Donnelly, M.A. & Guyer, C. (1994). Patterns of reproduction and habitat use in an assemblage 524 525 of Neotropical hylid frogs. *Oecologia*, 98, 291–302. 526 Duellman, W.E. (1988). Patterns of species diversity in anuran amphibians in the American 527 tropics. Ann. Mo. Bot. Gard., 75, 79–104. 528 Erickson, T.R. & Stefan, H.G. (1996). Correlations of Oklahoma stream temperatures with air 529 temperatures. St. Anthony Falls Laboratory. Available at: 530 https://hdl.handle.net/11299/109509. 531 Fick, S.E. & Hijmans, R.J. (2017). WorldClim 2: new 1-km spatial resolution climate surfaces for 532 global land areas. Int. J. Climatol., 37, 4302-4315. 533 Finn, R.N., Chauvigné, F., Hlidberg, J.B., Cutler, C.P. & Cerdà, J. (2014). The lineage-specific 534 evolution of aquaporin gene clusters facilitated tetrapod terrestrial adaptation. PLoS One, 535 9, e113686. 536 Fisher, M.C. & Garner, T.W.J. (2020). Chytrid fungi and global amphibian declines. Nat. Rev. 537 Microbiol., 18, 332–343. 538 Freidenburg, L.K. & Skelly, D.K. (2004). Microgeographical variation in thermal preference by an 539 amphibian. Ecol. Lett., 7, 369–373. 540 GBIF.org. (2021a). Data from: GBIF Occurrence Download. Available at: 541 https://doi.org/10.15468/dl.354f8k. Accessed: 03 October 2021. 542 GBIF.org. (2021b). Data from: GBIF Occurrence Download. Available at: 543 https://doi.org/10.15468/dl.pdjgte. Accessed: 03 October 2021. 544 GBIF.org. (2021c). Data from: GBIF Occurrence Download. Available at: 545 https://doi.org/10.15468/dl.nbku3v. Accessed: 03 October 2021. 546 GBIF.org. (2021d). Data from: GBIF Occurrence Download. Available at: 547 https://doi.org/10.15468/dl.uejpg6. Accessed: 03 October 2021. 548 Giaretta, A.A., Facure, K.G., Sawaya, R.J., Meyer, J.H.D.M. & Chemin, N. (1999). Diversity and

550

551

552

553

554

555

556

557

558

559

560

561

562

563

564

565

566

567

568

569

570

571

572

573

574

abundance of litter frogs in a Montane forest of southeastern Brazil: Seasonal and altitudinal changes. Biotropica, 31, 669-674. Girish, S. & Saidapur, S.K. (2000). Interrelationship between food availability, fat body, and ovarian cycles in the frog, Rana tigrina, with a discussion on the role of fat body in anuran reproduction. J. Exp. Zool., 286, 487-493. Gomez-Mestre, I., Pyron, R.A. & Wiens, J.J. (2012). Phylogenetic analyses reveal unexpected patterns in the evolution of reproductive modes in frogs. Evolution, 66, 3687-3700. Grafen, A. (1989). The phylogenetic regression. Philos. Trans. R. Soc. Lond. B Biol. Sci., 326, 119-157. Greenberg, D.A. & Palen, W.J. (2021). Hydrothermal physiology and climate vulnerability in amphibians. Proc. Biol. Sci., 288, 20202273. Griffis-Kyle. (2016). Physiology and ecology to inform climate adaptation strategies for desert amphibians. Herpetol. Conserv. Biol., 11, 563-582. Guevara-Molina, E.C., Gomes, F.R. & Camacho, A. (2020). Effects of dehydration on thermoregulatory behavior and thermal tolerance limits of Rana catesbeiana (Shaw, 1802). J. Therm. Biol., 93, 102721. Holland, B.S. & Copenhaver, M.D. (1987). An improved sequentially rejective bonferroni test procedure. Biometrics, 43, 417–423. Huber, P.J. (1981). Robust statistics. In: International encyclopedia of statistical science. { ed. Lovric, M. J. Springer, Berlin, Heidelberg, pp. 1248–1251. Huey, R.B. (1991). Physiological consequences of habitat selection. Am. Nat., 137, S91–S115. Huey, R.B., Kearney, M.R., Krockenberger, A., Holtum, J.A.M., Jess, M. & Williams, S.E. (2012). Predicting organismal vulnerability to climate warming: roles of behaviour, physiology and adaptation. Philos. Trans. R. Soc. Lond. B Biol. Sci., 367, 1665–1679. IUCN. (2022). Data from: The IUCN Red List of Threatened Species, version 2021-3. Available at: https://www.iucnredlist.org.

576

577

578

579

580

581

582

583

584

585

586

587

588

589

590

591

592

593

594

595

596

597

598

599

600

Jetz, W. & Pyron, R.A. (2018). The interplay of past diversification and evolutionary isolation with present imperilment across the amphibian tree of life. Nat Ecol Evol, 2, 850-858. Klein, W., Dabés, L., Bonfim, V.M.G., Magrini, L. & Napoli, M.F. (2016). Allometric relationships between cutaneous surface area and body mass in anuran amphibians. Zoologischer Anzeiger - A Journal of Comparative Zoology, 263, 45-54. Köhler, A., Sadowska, J., Olszewska, J., Trzeciak, P., Berger-Tal, O. & Tracy, C.R. (2011). Staying warm or moist? Operative temperature and thermal preferences of common frogs (Rana temporaria), and effects on locomotion. Herpetol. J., 21, 17–26. Lardner, B. & Loman, J. (2003). Growth or reproduction? Resource allocation by female frogs Rana temporaria. Oecologia, 137, 541–546. Martins, E.P. & Hansen, T.F. (1997). Phylogenies and the comparative method: A general approach to incorporating phylogenetic information into the analysis of interspecific data. Am. Nat., 149, 646–667. von May, R., Catenazzi, A., Santa-Cruz, R., Gutierrez, A.S., Moritz, C. & Rabosky, D.L. (2019). Thermal physiological traits in tropical lowland amphibians: Vulnerability to climate warming and cooling. PLoS One. 14, e0219759. Meza-Parral, Y. & Pineda, E. (2015). Amphibian diversity and threatened species in a severely transformed neotropical region in Mexico. *PLoS One*, 10, e0121652. Moen, D.S., Morlon, H. & Wiens, J.J. (2015). Testing convergence versus history: convergence dominates phenotypic evolution for over 150 million years in frogs. Syst. Biol., 65, 145-160. Moen, D.S. & Wiens, J.J. (2017). Microhabitat and climatic niche change explain patterns of diversification among frog families. Am. Nat., 190, 29–44. Mohseni, O. & Stefan, H.G. (1999). Stream temperature/air temperature relationship: a physical interpretation. J. Hydrol., 218, 128–141. Murrieta-Galindo, R., López-Barrera, F., González-Romero, A. & Parra-Olea, G. (2013). Matrix

602

603

604

605

606

607

608

609

610

611

612

613

614

615

616

617

618

619

620

621

622

623

624

625

626

and habitat quality in a montane cloud-forest landscape: amphibians in coffee plantations in central Veracruz, Mexico. Wildl. Res., 40, 25-35 MuToss Coding Team, Blanchard, G., Dickhaus, T., Hack, N., Konietschke, F., Rohmeyer, K., et al. (2017). mutoss: Unified Multiple Testing Procedures. R package version 0.1-12. https://CRAN.R-project.org/package=mutoss. Nakazawa, T., Ohba, S.-Y. & Ushio, M. (2013). Predator-prey body size relationships when predators can consume prey larger than themselves. Biol. Lett., 9, 20121193. Neveu, A. (2009). Incidence of climate on common frog breeding: Long-term and short-term changes. Acta Oecol., 35, 671-678. Padilla Perez, D.J. & Angilletta, M.J., Jr. (2022). Macroclimatic and maternal effects on the evolution of reproductive traits in lizards. Ecol. Evol., 12, e8885. Pankhurst, N.W. & Munday, P.L. (2011). Effects of climate change on fish reproduction and early life history stages. Mar. Freshwater Res., 62, 1015–1026. Paradis, E. & Schliep, K. (2019). ape 5.0: an environment for modern phylogenetics and evolutionary analyses in R. *Bioinformatics*, 35, 526–528. Pilgrim, J.M., Fang, X., Stefan, H.G. (1995). Correlations of Minnesota stream water temperatures with air Temperatures. St. Anthony Falls Laboratory. Available at: https://hdl.handle.net/11299/109299. Pintanel, P., Tejedo, M., Merino-Viteri, A., Almeida-Reinoso, F., Salinas-Ivanenko, S., López-Rosero, A.C., et al. (2022). Elevational and local climate variability predicts thermal breadth of mountain tropical tadpoles. *Ecography*, 2022, e05906. Poelman, E.H., van Wijngaarden, R.P.A. & Raaijmakers, C.E. (2013). Amazon poison frogs (Ranitomeya amazonica) use different phytotelm characteristics to determine their suitability for egg and tadpole deposition. Evol. Ecol., 27, 661–674. Ponce-Reyes, R., Nicholson, E., Baxter, P.W.J., Fuller, R.A. & Possingham, H. (2013). Extinction risk in cloud forest fragments under climate change and habitat loss. *Divers.* 

628

629

630

631

632

633

634

635

636

637

638

639

640

641

642

643

644

645

646

647

648

649

650

651

652

Distrib., 19, 518-529. Prado, C.P.A. & Haddad, C.F.B. (2005). Size-fecundity relationships and reproductive investment in female frogs in the Pantanal, South-Western Brazil. Herpetol. J., 15, 181-189. R Core Team. (2022). R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. Available at: https://www.R-project.org/. Rastogi, R.K., Izzo-Vitiello, I., Meglio, M., Matteo, L., Franzese, R., Costanzo, M.G., et al. (1983). Ovarian activity and reproduction in the frog, Rana esculenta. J. Zool., 200, 233– 247. Reading, C.J. (2003). The effects of variation in climatic temperature (1980-2001) on breeding activity and tadpole stage duration in the common toad, Bufo bufo. Sci. Total Environ., 310, 231-236. Santini, L., Benítez-López, A., Ficetola, G.F. & Huijbregts, M.A.J. (2018). Length-mass allometries in amphibians. Integr. Zool., 13, 36–45. Schulte, Rödder, Schulte & Lötters. (2010). Preference and competition for breeding plants in coexisting Ranitomeya species (Dendrobatidae): does height play a role. Salamandra, 186, 180-184. Sheridan, J.A. (2009). Reproductive variation corresponding to breeding season length in three tropical frog species. J. Trop. Ecol., 25, 583-592. Shibata, Y., Sano, T., Tsuchiya, N., Okada, R., Mochida, H., Tanaka, S., et al. (2014). Gene expression and localization of two types of AQP5 in Xenopus tropicalis under hydration and dehydration. Am. J. Physiol. Regul. Integr. Comp. Physiol., 307, R44-56. Silva, N.R., Berneck, B.V.M., da Silva, H.R., Haddad, C.F.B., Zamudio, K.R., Mott, T., et al. (2020). Egg-laying site, fecundity and degree of sexual size dimorphism in frogs. Biol. J. Linn. Soc. Lond., 131, 600-610. Skelly, D.K. & Freidenburg, L.K. (2008). Effects of beaver on the thermal biology of an

653 amphibian. Ecol. Lett., 3, 483-486. Skelly, D.K., Joseph, L.N., Possingham, H.P., Freidenburg, L.K., Farrugia, T.J., Kinnison, M.T., 654 655 et al. (2007). Evolutionary responses to climate change. Conserv. Biol., 21, 1353–1355. 656 Sørensen, R., Zinko, U. & Seibert, J. (2005). On the calculation of the topographic wetness 657 index: evaluation of different methods based on field observations. Hydrol. Earth Syst. Sci. 658 Discuss., 2, 1807-1834. 659 South, A. (2012). rworldxtra: Country boundaries at high resolution. R package version 1.01. 660 https://CRAN.R-project.org/package=rworldxtra. 661 Swihart, R.K. (1984). Body size, breeding season length, and life history tactics of lagomorphs. 662 Oikos, 43, 282-290. 663 Telemeco, R.S., Abbott, K.C. & Janzen, F.J. (2013). Modeling the effects of climate change-664 induced shifts in reproductive phenology on temperature-dependent traits. Am. Nat., 181, 665 637–648. 666 Thomas, C.D., Cameron, A., Green, R.E., Bakkenes, M., Beaumont, L.J., Collingham, Y.C., et 667 al. (2004). Extinction risk from climate change. Nature, 427, 145–148. 668 Title, P.O. & Bemmels, J.B. (2018). ENVIREM: an expanded set of bioclimatic and topographic 669 variables increases flexibility and improves performance of ecological niche modeling. 670 Ecography, 41, 291–307. 671 Todd, B.D., Scott, D.E., Pechmann, J.H.K. & Gibbons, J.W. (2011). Climate change correlates 672 with rapid delays and advancements in reproductive timing in an amphibian community. 673 Proc. Biol. Sci., 278, 2191-2197. 674 Tracy, C.R., Christian, K.A. & Tracy, C.R. (2010). Not just small, wet, and cold: effects of body 675 size and skin resistance on thermoregulation and arboreality of frogs. Ecology, 91, 1477-676 1484. Ulloa, J.S., Aubin, T., Llusia, D., Courtois, É.A., Fouquet, A., Gaucher, P., et al. (2019). 677 678 Explosive breeding in tropical anurans: environmental triggers, community composition and 679 acoustic structure. BMC Ecol., 19, 28. 680 Urban, M.C. (2015). Climate change. Accelerating extinction risk from climate change. Science, 681 348, 571–573. 682 Visser, M.E., Both, C. & Lambrechts, M.M. (2004). Global climate change leads to mistimed 683 avian reproduction. In: Advances in Ecological Research. Academic Press, 89–110. 684 Wells, K.D. (1977). The social behaviour of anuran amphibians. Anim. Behav., 25, 666–693. 685 Wells, K.D. (2007). The Ecology and Behavior of Amphibians. University of Chicago Press. 686 While, G.M. & Uller, T. (2014). Quo vadis amphibia? Global warming and breeding phenology in 687 frogs, toads and salamanders. *Ecography*, 37, 921-929. 688 Wilke, C.O. (2021). ggridges: Ridgeline Plots in 'ggplot2'. R package version 0.5.3. 689 https://CRAN.R-project.org/package=ggridges. 690 Williams, S.E., Shoo, L.P., Isaac, J.L., Hoffmann, A.A. & Langham, G. (2008). Towards an 691 integrated framework for assessing the vulnerability of species to climate change. PLoS 692 Biol., 6, 2621-2626. 693 Womack, M.C. & Bell, R.C. (2020). Two-hundred million years of anuran body-size evolution in 694 relation to geography, ecology and life history. J. Evol. Biol., 33, 1417–1432. 695 Wu, N.C., Cramp, R.L. & Franklin, C.E. (2018). Body size influences energetic and 696 osmoregulatory costs in frogs infected with Batrachochytrium dendrobatidis. Sci. Rep., 8, 697 3739. 698 Yu, G. (2020). Using agtree to visualize data on tree-like structures. Curr. Protoc. 699 Bioinformatics, 69, e96. doi: 10.1002/cpbi.96. 700 Yu, G., Lam, T.T.-Y., Zhu, H. & Guan, Y. (2018). Two methods for mapping and visualizing 701 associated data on phylogeny using ggtree. Mol. Biol. Evol., 35, 3041-3043. doi: 702 10.1093/molbev/msv194. 703 Yu, G., Smith, D., Zhu, H., Guan, Y. & Lam, T.T-Y. (2017). ggtree: an R package for 704 visualization and annotation of phylogenetic trees with their covariates and other

associated data. *Methods Ecol. Evol.*, 8, 28–36. doi:10.1111/2041-210X.12628.
Zizka, A., Barratt, C.D., Ritter, C.D., Joerger-Hickfang, T. & Zizka, V.M.A. (2021). Existing approaches and future directions to link macroecology, macroevolution and conservation prioritization. *Ecography*, 2022, e05557.
Zizka A., Silvestro, D., Andermann, T., Azevedo, J., Duarte Ritter, C., Edler, D. *et al.* (2019).
CoordinateCleaner: standardized cleaning of occurrence records from biological collection databases. *Methods Ecol. Evol.*, 10, 744–751. R package version 2.0-20, Available at: https://github.com/ropensci/CoordinateCleaner. doi: 10.1111/2041-210X.13152.







