Package AquaEnv: an Aquatic modelling Environment in R

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Abstract

AquaEnv is an integrated development toolbox for aquatic chemical model generation focused on (ocean) acidification and CO2 air-water exchange.

- It contains all elements necessary to model the pH, the related CO2 air-water exchange, as well as aquatic acid-base chemistry in general for an arbitrary marine, estuarine or freshwater system. Also chemical batches can be modelled.
- Next to the routines necessary to calculate desired information, AquaEnv also contains a suite of tools to visualize this information.
- Furthermore, **AquaEnv** can not only be used to build dynamic models of aquatic systems, but it can also serve as a simple desktop tool for the experimental aquatic chemist to generate and visualize all possible derived information from a set of measurements with one single easy to use R function.
- Additionally, the sensitivity of the system to variations in the input variables can be visualized.
- AquaEnv also contains a number of example "applications" that make use of the aquatic modelling toolbox that AquaEnv provides:
 - a theoretical titration simulator
 - and a routine to determine total alkalinity ([TA]), the total dissolved inorganic carbon concentration ([\sum CO2]), as well as additionally the electrode standard potential (E₀) and the first dissociation constant of the carbonate system ($K_{CO_2}^*$) from titration data.

Keywords: aquatic modelling, pH, pH scales, dissolved inorganic carbon, total alkalinity, total alkalinity curve fitting, theoretical titration, revelle factor, omega, solubility products, CO₂, ocean acidification, estuaries, carbonate system, seawater, R.

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1 Introduction

AquaEnv is a toolbox for aquatic modelling that serves several purposes

- It provides functions to calculate the stoichiometric equilibrium constants (K*) for key acid base systems in natural seawater, the Henry's constants (K₀), as well as the solubility products (K_{sp}) for calcite and aragonite. This functionality is provided via the functions K_CO2, K_HCO3, K_BOH3, K_W, K_HSO4, K_HF, K_NH4, K_H2S, K_H3PO4, K_H2PO4, K_HPO4, K_SiOH4, K_SiOOH3, KO_CO2, KO_O2, Ksp_aragonite, and Ksp_calcite.
- It is designed to make its use as easy as possible: all the information that can be calculated from the set of parameters know of a system or sample can be obtained by one single function: aquaenv. This function returns a list of class aquaenv that contains next to the input parameters
 - the clorinity, the ionic strength, $[\sum B(OH)_3]$, $[\sum H_2SO_4]$, $[\sum HF]$, $[Cl^-]$, $[Cl^-]$, $[\sum Br]$, $[Na^+]$, $[Mg^{2+}]$, $[Ca^{2+}]$, $[K^+]$, $[Sr^{2+}]$ calculated from salinity as given in DOE (1994) (Please note that if values for $[\sum B(OH)_3]$, $[\sum H_2SO_4]$, $[\sum HF]$ are given as input parameters, these parameters are used and not the ones calculated from salinity.)
 - the hydrostatic pressure calculated from the given depth and the seawater density calculated from temperature and salinity as given by Millero and Poisson (1981)
 - a set of conversion factors to convert between different pH scales (Dickson 1984;
 Zeebe and Wolf-Gladrow 2001) and between mol/kg-H₂O and mol/kg-solution (inferred from Roy, Roy, Vogel, PorterMoore, Pearson, Good, Millero, and Campbell (1993b) and DOE (1994))
 - the Henry's constants for CO₂ (Weiss 1974) and for O₂ (inferred from Weiss 1970) calculated from temperature and salinity as well as the associated saturation concentrations of CO₂ and O₂.
 - the ion product of water (Millero 1995), the stoichiometric equilibrium constants of HSO₄⁻ (Dickson 1990a), HF(Dickson and Riley 1979a), CO₂ (Roy et al. 1993b), HCO₃⁻ (Roy et al. 1993b), B(OH)₃ (Dickson 1990a), NH4⁺(Millero, Yao, and Aicher 1995), H2_S (Millero 1995), H₃PO4(Millero 1995), H₂PO4⁻ (Millero 1995), HPO4²⁻ (Millero 1995), SiOH4 (Millero, Plese, and Fernandez 1988), SiOOH3⁻ (Wischmeyer, Del Amo, Brzezinski, and Wolf-Gladrow 2003), HNO2 (Riordan, Minogue, Healy, O'Driscoll, and Sodeau 2005), HNO3, H2SO4 (Atkins 1996), HS (Atkins 1996) mostly calculated as functions of temperature and salinity and pressure corrected according to Millero (1995).
 - the solubility products of calcite and aragonite (Mucci 1983) as well as the associated Ω 's if a full speciation is calculated (see below)
 - the partial pressure of CO₂ if a full speciation is calculated (see below)
 - if $[\sum CO_2]$ and pH are given [TA] is calculated, if $[\sum CO_2]$ and [TA] are given pH is calculated, if $[\sum CO_2]$ and $[CO_2]$ or pCO₂ are given, pH and [TA] are calculated.
 - if either one of the pairs pH and $[CO_2]$ or pCO₂, pH and [TA], or [TA] and $[CO_2]$ or pCO₂ is given, $[\sum CO_2]$ is calculated

- if sufficient information is given and the flag speciation=TRUE is set, a full speciation of $[\sum CO2]$, $[\sum NH4]$, $[\sum H_2S]$, $[\sum HNO3]$, $[\sum HNO2]$, $[\sum H_3PO4]$, $[\sum Si(OH)_4]$, $[\sum B(OH)_3]$, $[\sum H_2SO_4]$, $[\sum HF]$, as well as water itself is calculated
- if the flag revelle=TRUE is set, the revelle factor (Zeebe and Wolf-Gladrow 2001) is calculated. item if the flag revelle=TRUE is set, all necessary quantities for the explicit "direct substitution approach" (DSA) to pH modelling as given in Hofmann, Meysman, Soetaert, and Middelburg (2008b) are calculated. These are the buffer factor (the partial derivative of [TA] with respect to $[H^+]$) and the partial derivatives of [TA] with respect to the other total quantities. Furthermore, the partial derivatives of [TA] with respect to changes in the equilibrium constants (K^*) , multiplied with the partial derivatives of the equilibrium constants with respect to their variables needed for the DSA with time variable equilibrium constants as described in Hofmann, Meysman, Soetaert, and Middelburg (2008a) are calculated. Finally, the ionization fractions as defined by Stumm and Morgan (1996) and used in Hofmann, Middelburg, Soetaert, Wolf-Gladrow, and Meysman (2008c) are calculated for the full speciation.
- Input for aquaenv has to be supplied in standard SI units, the free proton pH scale and in molinity¹ (mol/kg-solution). Conversion of input parameters to this necessary units and pH scale can be done with the generic function convert.
- The information created with aquaenv is also supplied in standard SI units and in molinity. All elements of an object of class aquaenv of a certain unit or pH scale can be converted into other units or pH scales with the function convert as well.
- One can use input vectors of temperature T, salinity S or depth d for aquaenv to obtain vectors of all calculated information as function of the input vector. This can be visualized in a large variety of ways using the plot function specially defined for objects pf class aquaenv.
- Objects of class aquaenv can be used in dynamic models to define the state of the system in each timestep of the numerical integration (done e.g. with **deSolve**). with the function aquaenv and the flag from.data.frame=TRUE it is possible to convert output of those dynamic models into objects pf class aquaenv which allows the user to use the whole suite of visualisation tools that is provided by the function plot in AquaEnv.
- As mentioned above Hofmann et al. (2008b), Hofmann et al. (2008a), and Hofmann et al. (2008c) describe methods for an "explicit" pH modelling which allows for the quantification of the influences of kinetically modelled processes on the pH. Objects pf class aquaenv provide all needed quantities (partial derivatives of [TA], ionization fractions, etc.) to employ both of those methods in dynamic models. Furthermore, AquaEnv provides the functionality to cumulatively plot the obtained influences on the pH.
- As an example of how to use the aquatic chemical toolbox that is provided by AquaEnv, two applications are provided

¹Note that it is not sufficient to give a gravimetric concentration in mol/kg since there is a substancial difference between mol/kg-H₂O (molality) and mol/kg-solution (molinity).

- The function titration: creates theoretical titrations which can be used e.g. to create bjerrum plots with the function plot.aquaenv in AquaEnv.
- The function TAfit: a routine based on a method in DOE (1994) that makes use of that theoretical titration function and allows for determining total alkalinity ([TA]), the total dissolved inorganic carbon concentration ([\sum CO2]), as well as additionally the electrode standard potential (E₀) and the first dissociation constant of the carbonate system ($K_{CO_2}^*$) using the Levenberg-Marquart algorithm (least squares optimization procedure) as provided in the R package **minpack.lm**.

2 The elements of an object of class aquaenv

The function aquaenv, the central function of **AquaEnv**, returns an object of class aquaenv. This object is a list of different elements which can be accesses with the \$\\$ character or with the [[]] operator

```
> test <- aquaenv(10, 35)
> test$Tc

[1] 10
attr(,"unit")
[1] "deg C"

> test[["Tc"]]

[1] 10
attr(,"unit")
[1] "deg C"
```

Maximally, i.e., if the enough input data is supplied to define the pH of the system and the flags speciation, dsa, and revelle are TRUE while the flag skeleton is FALSE, an object of class aquaenv contains the following elements

| element | unit | explanation |
|---------------|-------------------------|--|
| Tc | °C | temperature |
| Tk | K | absolute temperature |
| S | "psu" (no unit) | salinity |
| Cl | % 0 | chlorinity |
| I | mol/kg-H ₂ O | ionic strength |
| d | m | depth |
| hydroP | bar | hydrostatic pressure |
| density | kg/m^3 | (seawater) density |
| SumCO2 | mol/kg-soln | $[\sum CO_2]$, total dissolved inorganic carbon concentration |
| SumNH4 | mol/kg-soln | $[\sum NH_4^+]$, total ammonium concentration |
| SumH2S | mol/kg-soln | $[\sum H_2S]$, total sulfide concentration |
| SumHNO3 | mol/kg-soln | $\sum HNO_3$, total nitrate concentration |

| SumHNO2 | mol/kg-soln | $[\sum HNO_2]$, total nitrite concentration |
|-------------|--|---|
| SumH3PO4 | mol/kg-soln | $[\sum H_3PO_4]$, total phosphate concentration |
| SumSiOH4 | mol/kg-soln | $[\sum Si(OH)_4]$, total silicate concentration |
| SumBOH3 | mol/kg-soln | $[\sum B(OH)_3]$, total borates concentration |
| SumH2SO4 | mol/kg-soln | $[\sum H_2SO_4]$, total sulfate concentration |
| SumHF | mol/kg-soln | $\sum HF$, total fluoride concentration |
| SumBr | mol/kg-soln | $\sum_{i=1}^{n} HBr_{i}$, total bromide concentration |
| ClConc | mol/kg-soln | [Cl ⁻], chloride concentration |
| Na | mol/kg-soln | [Na ⁺], sodium concentration |
| Mg | mol/kg-soln | [Mg ²⁺], magnesium concentration |
| Ca | mol/kg-soln | [Ca ²⁺], calcium concentration |
| K | mol/kg-soln | [K ⁺], potassium concentration |
| Sr | mol/kg-soln | [Sr ²⁺], strontium concentration |
| molal2molin | (mol/kg-soln)/(mol/kg-H2O) | concentration conversion factor: from molality to |
| motat2mom | (mor/ kg som)/ (mor/ kg 1120) | molinity |
| free2tot | - | pH conversion factor: free scale to total scale |
| free2sws | - | pH conversion factor: free scale to sawater scale |
| tot2free | - | pH conversion factor: total scale to free scale |
| tot2sws | - | pH conversion factor: total scale to seawater scale |
| sws2free | - | pH conversion factor: seawater scale to free scale |
| sws2tot | - | pH conversion factor: seawater scale to total scale |
| K0_CO2 | mol/(kg-soln*atm) | Henry's constant for CO ₂ |
| K0_O2 | mol/(kg-soln*atm) | Henry's constant for O ₂ |
| CO2_sat | mol/kg-soln | ${\rm CO_2}$ saturation concentration at an atmospheric partial pressure/fugacity of Fugacity\$CO2 |
| O2_sat | mol/kg-soln | O ₂ saturation concentration at an atmospheric partial pressure/fugacity of Fugacity\$O2 |
| $K_{-}W$ | (mol/kg-soln) ² , free pH scale | stoichiometric equilibrium ion product of |
| | | $H_2O: K_W^* = [H^+][OH-]$ |
| K_HSO4 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | | $K_{HSO_4^-}^* = [H^+][SO_4^{2-}]/[HSO_4^-]$ |
| K_HF | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | mor/ ng som, neo pri seme | $K_{HF}^* = [H^+][F^-]/[HF]$ |
| K_CO2 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | | $K_{CO_2}^* = [H^+][HCO_3^-]/[CO_2]$ |
| K_HCO3 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| 11211000 | mor, ng bom, nee pir beare | $K_{HCO_3^-}^* = [H^+][CO_3^2]/[HCO_3^-]$ |
| K_BOH3 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| n_bons | mor/ng both, free pir both | $K_{B(OH)_3}^* = [H^+][B(OH)_4^-]/[B(OH)_3]$ |
| K_NH4 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | mor, ng som, nee pri seere | $K_{NH_{4}^{+}}^{*} = [H^{+}][NH_{3}]/[NH_{4}^{+}]$ |
| K_H2S | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | | $K_{H_2S}^* = [H^+][HS^-]/[H_2S]$ |
| K_H3PO4 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | | $K_{H_3PO_4}^* = [H^+][H_2PO_4^-]/[H_3PO_4]$ |
| K_H2PO4 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| ' | ' | |

| | | $K_{H_2PO_4^-}^* = [H^+][HPO_4^{2-}]/[H_2PO_4^-]$ |
|---------------|--------------------------------|---|
| K_HPO4 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| K_IIF 04 | mor/kg-som, free pri scare | _ |
| | | $K_{HPO_4^{2-}}^* = [H^+][PO_4^{3-}]/[HPO_4^{2-}]$ |
| K_SiOH4 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | | $K_{Si(OH)_4}^* = [H^+][SiO(OH)_3^-]/[Si(OH)_4]$ |
| K_SiOOH3 | mol/kg-soln, free pH scale | stoichiometric equilibrium constant |
| | | $K_{SiO(OH)_3^-}^* = [H^+][SiO_2(OH)_2^{2-}]/[SiO(OH)_3^-]$ |
| K_HNO2 | mol/kg-soln; mol/kg-H2O; mol/l | approximate value for equilibrium constant |
| | | $K_{HNO_2}^* = [H^+][NO_2^-]/[HNO_2]$ |
| K_HNO3 | mol/kg-soln; mol/kg-H2O; mol/l | approximate value for equilibrium constant |
| | | $K_{HNO_3}^* = [H^+][NO_3^-]/[HNO_3]$ |
| K_H2SO4 | mol/kg-soln; mol/kg-H2O; mol/l | approximate value for equilibrium constant |
| | | $K_{H_2SO_4}^* = [H^+][HSO_4^-]/[H_2SO_4]$ |
| K_HS | mol/kg-soln; mol/kg-H2O; mol/l | approximate value for equilibrium constant |
| | | $K_{HS^-}^* = [H^+][S^{2-}]/[HS^-]$ |
| Ksp_calcite | (mol/kg-soln) ² | stoichiometric equilibrium solubility product of calcite |
| | | $Ksp_{cal}^* = [Ca^{2+}][CO_3^{2-}]$ |
| Ksp_aragonite | $(\text{mol/kg-soln})^2$ | stoichiometric equilibrium solubility product of |
| | | aragonite |
| | | $Ksp_{ara}^* = [Ca^{2+}][CO_3^{2-}]$ |
| TA | mol/kg-soln | [TA], total alkalinity |
| рН | -, free scale | pH |
| pCO2 | atm, | partial pressure (fugacity) of CO ₂ in the water |
| CO2 | mol/kg-soln | $[CO_2]$ |
| HCO3 | mol/kg-soln | $[HCO_3^-]$ |
| CO3 | mol/kg-soln | $[CO_3^{2-}]$ |
| ВОН3 | mol/kg-soln | [B(OH) ₃] |
| BOH4 | mol/kg-soln | $\left[\mathrm{B}(\mathrm{OH})_{4}^{-}\right]$ |
| ОН | mol/kg-soln | [OH ⁻] |
| H3PO4 | mol/kg-soln | $[\mathrm{H_3PO_4}]$ |
| H2PO4 | mol/kg-soln | $[\text{H2PO}_4^-]$ |
| HPO4 | mol/kg-soln | $[HPO_4^{2-}]$ |
| PO4 | mol/kg-soln | $\left[\mathrm{PO}_4^{3-}\right]$ |
| SiOH4 | mol/kg-soln | $[Si(OH)_4]$ |
| SiOOH3 | mol/kg-soln | $[SiO(OH)_3^-]$ |
| SiO2OH2 | mol/kg-soln | $\left[\mathrm{SiO}_{2}(\mathrm{OH})_{2}^{2-}\right]$ |
| H2S | mol/kg-soln | $[H_2S]$ |
| HS | mol/kg-soln | [HS ⁻] |
| S2min | mol/kg-soln | $[S^{2-}]$ |
| NH4 | mol/kg-soln | [NH ₄ ⁺] |
| NH3 | mol/kg-soln | [NH ₃] |
| H2SO4 | mol/kg-soln | [H ₂ SO ₄] |
| HSO4 | mol/kg-soln | $[HSO_4^-]$ |
| SO4 | mol/kg-soln | $\left[\mathrm{SO}_4^{2-}\right]$ |

| HF | mol/kg-soln | [HF] |
|--------------------------|-------------|--|
| F | mol/kg-soln | [F-] |
| HNO3 | mol/kg-soln | [HNO ₃] |
| NO3 | mol/kg-soln | $[NO_3^-]$ |
| HNO2 | mol/kg-soln | [HNO ₂] |
| NO2 | mol/kg-soln | $[NO_2^-]$ |
| omega_calcite | - | saturation state Ω with respect to calcite |
| omega_aragonite | _ | saturation state Ω with respect to aragonite |
| revelle | _ | Revelle factor |
| c1 | _ | ionization fraction $c_1 = [CO_2]/[\sum CO_2]$ |
| c2 | _ | ionization fraction $c_2 = [HCO_3^-]/[\sum CO_2]$ |
| c3 | _ | ionization fraction $c_3 = [CO_3^{2-}]/[\sum CO_2]$ |
| dTAdSumCO2 | _ | $\frac{\partial [TA]}{[\partial \sum CO_2]}$ |
| | | with $[TA] = f([H^+], [\sum CO_2],)$ |
| b1 | _ | ionization fraction $b_1 = [B(OH)_3]/[\sum B(OH)_3]$ |
| b2 | _ | ionization fraction $b_2 = [B(OH)_4]/[\sum B(OH)_3]$ |
| dTAdSumBOH3 | _ | $\frac{\partial[TA]}{[\partial \sum B(OH)_3]}$ |
| | | with $[TA] = f([H^+], [\sum CO_2],)$ |
| so1 | _ | ionization fraction $so_1 = [H_2SO_4]/[\sum H_2SO_4]$ |
| so2 | _ | ionization fraction $so_2 = [HSO_4^-]/[\sum H_2SO_4]$ |
| so3 | _ | ionization fraction $so_3 = [SO_4^2]/[\sum H_2SO_4]$ |
| dTAdSumH2SO4 | _ | $\frac{\partial [TA]}{[\partial \sum H_2 SO_4]}$ |
| 41114541111 2 501 | | with $[TA] = f([H^+], [\sum CO_2],)$ |
| f1 | | with $[TA] = I([T], [\sum OO_2],)$ ionization fraction $f_1 = [HF]/[\sum HF]$ |
| f2 | _ | ionization fraction $f_1 = [\text{F}^-]/[\sum \text{HF}]$ |
| dTAdSumHF | _ | |
| dTAdSulliTr | - | |
| JUDA JII | | with $[TA] = f([H^+], [\sum CO_2],)$ |
| dTAdH | - | $\frac{\partial[TA]}{[\partial[H^+]]}$: buffer factor |
| 1771 1771 10 | | with $[TA] = f([H^+], [\sum CO_2],)$ |
| dTAdKdKdS | - | $\sum_{i} \frac{\partial [TA]}{\partial K_{i}^{*}} \frac{\partial K_{i}^{*}}{\partial S}$ |
| | | with $[TA] = f([H^+], [\sum CO_2],, K_i^*)$ |
| dTAdKdKdT | - | $\sum_{\mathrm{i}} rac{\partial [\mathrm{TA}]}{\partial \mathrm{K}_{\mathrm{i}}^{*}} rac{\partial \mathrm{K}_{\mathrm{i}}^{*}}{\partial \mathrm{T}}$ |
| | | with $[TA] = f([H^+], [\sum CO_2],, K_i^*)$ |
| dTAdKdKdd | - | $\sum_{\mathbf{i}} \frac{\partial [\mathrm{TA}]}{\partial \mathrm{K}^*_{\mathbf{i}}} \frac{\partial \mathrm{K}^*_{\mathbf{i}}}{\partial \mathrm{d}}$ |
| | | with $[TA] = f([H^+], [\sum CO_2],, K_i^*)$ |
| dTAdKdKdSumH2SO4 | - | $\sum_{i} \frac{\partial [TA]}{\partial K_{i}^{*}} \frac{\partial K_{i}^{*}}{\partial [\sum H_{2}SO_{4}]}$ |
| | | with $[TA] = f([H^+], [\sum CO_2],, K_i^*)$ |
| dTAdKdKdSumHF | _ | $\sum_{i} \frac{\partial [TA]}{\partial K_{i}^{*}} \frac{\partial K_{i}^{*}}{\partial [\Sigma HF]}$ |
| | | $ \begin{array}{c} \sum_{1} \partial K_{i}^{*} \partial [\sum H^{i}] \\ \text{with } [TA] = f([H^{+}], [\sum CO_{2}],, K_{i}^{*}) \end{array} $ |
| | I | with [111] = I([11], [_ 002],, II;) |

For elements that are calculated according to certain literature references, those references are given in appendix B.

3 Using AquaEnv

3.1 Basic features

3.1.1 calling the "K" functions directly

The elements K_CO2, K_HCO3, K_BOH3, K_W, K_HSO4, K_HF, K_NH4, K_H2S, K_H3PO4, K_H2PO4, K_HPO4, K_SiOH4, K_SiOOH3, KO_CO2, KO_O2, Ksp_aragonite, and Ksp_calcite can be calculated directly, witout creating an object of class *aquaenv*. This is done via functions that bear the same name as those elements

```
> K_CO2(15, 30)

[1] 9.089094e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"

> KO_CO2(15, 30)

[1] 0.03852158
attr(,"unit")
[1] "mol/(kg-soln*atm)"

> Ksp_calcite(15, 30, 100)

[1] 3.64455e-07
attr(,"unit")
[1] "(mol/kg-soln)^2"
```

3.1.2 Minimal aquaenv definition

Minimally, an object of class aquaenv can be defined with just a temperature and salinity value

```
> ae <- aquaenv(Tc = 15, S = 30)
> ae$K_CO2

[1] 9.089094e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"
```

Optionally, a mean depth can be given. As in the above case, the returned object of class aquaenv then contains a standard set of elements as shown by the names command.

```
> ae <- aquaenv(Tc = 15, S = 30, d = 10)
> ae$Ksp_calcite

[1] 3.588429e-07
attr(,"unit")
[1] "(mol/kg-soln)^2"
```

> names(ae)

| [1] | "Tc" | "Tk" | "S" | "C1" |
|------|------------|---------------|-----------------|------------|
| [5] | "I" | "d" | "hydroP" | "density" |
| [9] | "SumCO2" | "SumNH4" | "SumH2S" | "SumHNO3" |
| [13] | "SumHNO2" | "SumH3PO4" | "SumSiOH4" | "SumBOH3" |
| [17] | "SumH2SO4" | "SumHF" | "SumBr" | "ClConc" |
| [21] | "Na" | "Mg" | "Ca" | "K" |
| [25] | "Sr" | "molal2molin" | "free2tot" | "free2sws" |
| [29] | "tot2free" | "tot2sws" | "sws2free" | "sws2tot" |
| [33] | "K0_C02" | "K0_02" | "CO2_sat" | "02_sat" |
| [37] | "K_W" | "K_HSO4" | "K_HF" | "K_C02" |
| [41] | "K_HCO3" | "K_BOH3" | "K_NH4" | "K_H2S" |
| [45] | "K_H3P04" | "K_H2P04" | "K_HPO4" | "K_SiOH4" |
| [49] | "K_SiOOH3" | "K_HNO2" | "K_HNO3" | "K_H2SO4" |
| [53] | "K_HS" | "Ksp_calcite" | "Ksp_aragonite" | |

A minimal set of elements in an object of class *aquaenv* can be obtained by setting the flag skeleton to TRUE

```
> ae <- aquaenv(Tc = 15, S = 30, d = 10, skeleton = TRUE) > names(ae)
```

```
[1] "Tc"
                                                    "I"
                 "Tk"
                            "S"
                                        "C1"
                                                               "d"
 [7] "hydroP"
                 "density"
                                                               "SumHNO3"
                            "SumCO2"
                                        "SumNH4"
                                                    "SumH2S"
[13] "SumHNO2"
                 "SumH3PO4" "SumSiOH4" "SumBOH3"
                                                   "SumH2SO4" "SumHF"
[19] "K_W"
                 "K_HSO4"
                            "K_HF"
                                        "K_C02"
                                                    "K_HCO3"
                                                               "K_BOH3"
[25] "K_NH4"
                 "K_H2S"
                            "K_H3P04"
                                        "K_H2P04"
                                                   "K_HP04"
                                                               "K_SiOH4"
                            "K_HNO3"
[31] "K_SiOOH3" "K_HNO2"
                                        "K_H2SO4"
                                                    "K_HS"
```

3.1.3 Defining the complete aquaenv system in different ways

If enough information is given to define a complete speciation, i.e. either one of the pairs SumCO2 and pH, SumCO2 and TA, SumCO2 and CO2, or SumCO2 and pCO2, a full aquaenv system can be defined.

```
> Tc <- 15
> S <- 30
> d <- 10
```

```
> SumCO2 <- 0.002
> pH <- 8
> TA <- 0.002140323
> pCO2 <- 0.000533576
> CO2 <- 2.055419e-05
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, pH = pH)
> ae$TA
[1] 0.002140799
attr(,"unit")
[1] "mol/kg-soln"
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, TA = TA)
> ae$pH
[1] 7.998792
attr(,"pH scale")
[1] "free"
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, CO2 = CO2)
> ae$pH
[1] 7.999294
attr(,"pH scale")
[1] "free"
> names(ae)
                                           "S"
 [1] "Tc"
                        "Tk"
                                                              "C1"
                        "d"
 [5] "I"
                                           "hydroP"
                                                              "density"
 [9] "SumCO2"
                        "SumNH4"
                                           "SumH2S"
                                                              "SumHNO3"
                                                              "SumBOH3"
[13] "SumHNO2"
                        "SumH3P04"
                                           "SumSiOH4"
[17] "SumH2SO4"
                        "SumHF"
                                           "SumBr"
                                                             "ClConc"
                                                              "K"
[21] "Na"
                        "Mg"
                                           "Ca"
[25] "Sr"
                                                              "free2sws"
                        "molal2molin"
                                           "free2tot"
[29] "tot2free"
                        "tot2sws"
                                           "sws2free"
                                                              "sws2tot"
[33] "KO_CO2"
                        "K0_02"
                                           "CO2_sat"
                                                              "02_sat"
[37] "K_W"
                        "K_HS04"
                                           "K_HF"
                                                              "K_C02"
[41] "K_HCO3"
                        "K_BOH3"
                                           "K_NH4"
                                                              "K_H2S"
[45] "K_H3PO4"
                        "K_H2P04"
                                           "K_HP04"
                                                              "K_SiOH4"
[49] "K_SiOOH3"
                        "K_HNO2"
                                           "K_HNO3"
                                                              "K_H2SO4"
[53] "K_HS"
                        "Ksp_calcite"
                                           "Ksp_aragonite"
                                                              "TA"
[57] "pH"
                        "pC02"
                                           "C02"
                                                              "HC03"
[61] "CO3"
                        "BOH3"
                                           "BOH4"
                                                              "OH"
[65] "H3PO4"
                        "H2P04"
                                          "HP04"
                                                             "P04"
[69] "SiOH4"
                        "SiOOH3"
                                          "SiO2OH2"
                                                             "H2S"
```

```
[73] "HS" "S2min" "NH4" "NH3"
[77] "H2S04" "HS04" "S04" "HF"
[81] "F" "HN03" "N03" "HN02"
[85] "N02" "omega_calcite" "omega_aragonite"
```

As seen above, a full speciation is calculated along with the pH or total alkalinity respectively. If only pH or total alkalinity is needed, the calculation of the full speciation can be toggled off. Furthermore, the flag skeleton also works for a full system.

```
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, pH = pH, speciation = FALSE)
> names(ae)
```

```
"Tk"
                                        "S"
                                                          "C1"
 [1] "Tc"
 [5] "I"
                       "d"
                                        "hvdroP"
                                                          "density"
 [9] "SumCO2"
                       "SumNH4"
                                        "SumH2S"
                                                          "SumHNO3"
                       "SumH3P04"
                                        "SumSiOH4"
                                                          "SumBOH3"
[13]
     "SumHNO2"
                      "SumHF"
                                        "SumBr"
                                                          "ClConc"
[17] "SumH2SO4"
[21] "Na"
                      "Mg"
                                        "Ca"
                                                          "K"
[25] "Sr"
                       "molal2molin"
                                        "free2tot"
                                                          "free2sws"
                       "tot2sws"
                                        "sws2free"
                                                          "sws2tot"
[29] "tot2free"
[33] "KO_CO2"
                       "K0_02"
                                        "CO2_sat"
                                                          "02_sat"
[37] "K_W"
                      "K_HSO4"
                                        "K_HF"
                                                          "K_C02"
                                        "K_NH4"
                                                          "K_H2S"
[41] "K_HCO3"
                       "K_BOH3"
                                                          "K_SiOH4"
[45] "K_H3PO4"
                      "K_H2P04"
                                        "K_HP04"
[49] "K_SiOOH3"
                      "K_HN02"
                                        "K_HNO3"
                                                          "K_H2S04"
[53] "K_HS"
                       "Ksp_calcite"
                                        "Ksp_aragonite" "TA"
[57] "pH"
                      "pC02"
                                        "C02"
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, pH = pH, speciation = FALSE,
      skeleton = TRUE)
> names(ae)
 [1] "Tc"
                             "S"
                                                     "I"
                 "Tk"
                                         "Cl"
                                                                 "d"
                                                                 "SumHNO3"
 [7] "hydroP"
                 "density"
                             "SumCO2"
                                         "SumNH4"
                                                     "SumH2S"
                 "SumH3PO4"
                             "SumSiOH4"
                                                     "SumH2SO4"
                                                                 "SumHF"
[13] "SumHNO2"
                                         "SumBOH3"
```

Furtermore all the quantities needed for the explicit pH modelling approaches as given in Hofmann *et al.* (2008b) and Hofmann *et al.* (2008c) can be calculated by setting the flag dsa to TRUE. The revelle factor can be calculated using the flag revelle.

"K_C02"

"K_H2P04"

"K_H2SO4"

"K_HCO3"

"K_HP04"

"K_HS"

"K_BOH3"

"K_SiOH4" "TA"

```
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, pCO2 = pCO2, dsa = TRUE,
+ revelle = TRUE)
> ae$dTAdH
```

"K_HF"

"K_H3P04"

"K_HNO3"

"C02"

[19] "K_W"

[37] "pH"

[25] "K_NH4"

[31] "K_SiOOH3"

"K_HS04"

"K_H2S"

"K_HNO2"

"pC02"

```
[1] -17097.02
attr(,"unit")
[1] "(mol-TA/kg-soln)/(mol-H/kg-soln)"
attr(,"pH scale")
[1] "free"
```

> ae\$revelle

[1] 13.77498

If an ambivalent situation is created because the user enters too much information, an error message is displayed

```
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, CO2 = CO2, pCO2 = pCO2)
[1] "Error! Overdetermined system: entered pCO2: 0.000533576 , calculated pCO2: 0.000533575978437754"
[1] "Please enter only one of: pH, TA, CO2, or pCO2."
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, pH = pH, TA = TA)
[1] "Error! Overdetermined system: entered TA: 0.002140323 , calculated TA: 0.00214079928345428"
[1] "Please enter only one of: pH, TA, CO2, or pCO2."
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, pH = pH, CO2 = CO2)
[1] "Error! Overdetermined system: entered pH: 8 , calculated pH: 7.99929413967086"
[1] "Please enter only one of: pH, TA, CO2, or pCO2."
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, pH = pH, pCO2 = pCO2)
[1] "Error! Overdetermined system: entered pH: 8 , calculated pH: 7.99929412288978"
[1] "Please enter only one of: pH, TA, CO2, or pCO2."
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, TA = TA, CO2 = CO2)
[1] "Error! Overdetermined system: entered TA: 0.002140323 , calculated TA: 0.00214052081324705"
[1] "Please enter only one of: pH, TA, CO2, or pCO2."
> ae <- aquaenv(Tc, S, d, SumCO2 = SumCO2, TA = TA, pCO2 = pCO2)
[1] "Error! Overdetermined system: entered TA: 0.002140323 , calculated TA: 0.00214052080663004"
[1] "Please enter only one of: pH, TA, CO2, or pCO2."
```

3.1.4 Calculating $[\sum CO_2]$

 $[\sum CO_2]$ can be calculated by giving a constant pair of either pH and CO2, pH and TA, TA and CO2, or TA and pCO2

```
> pCO2 <- 0.0006952296
> CO2 <- 2.678134e-05
> pH <- 7.888569
> TA <- 0.0021
> Tc <- 15
> S <- 30
> d <- 10
> ae <- aquaenv(Tc, S, d, SumCO2 = NULL, pH = pH, CO2 = CO2)
> ae$SumCO2
[1] 0.001999994
attr(,"unit")
[1] "mol/kg-soln"
> ae <- aquaenv(Tc, S, d, SumCO2 = NULL, pH = pH, pCO2 = pCO2)
> ae$SumCO2
[1] 0.001999994
attr(,"unit")
[1] "mol/kg-soln"
> ae <- aquaenv(Tc, S, d, SumCO2 = NULL, pH = pH, TA = TA)
> ae$SumCO2
[1] 0.002000000
attr(,"unit")
[1] "mol/kg-soln"
> ae <- aquaenv(Tc, S, d, SumCO2 = NULL, TA = TA, CO2 = CO2)
> ae$SumCO2
[1] 0.002
attr(,"unit")
[1] "mol/kg-soln"
> ae <- aquaenv(Tc, S, d, SumCO2 = NULL, TA = TA, pCO2 = pCO2)
> ae$SumCO2
[1] 0.002
attr(,"unit")
[1] "mol/kg-soln"
```

3.1.5 Cloning an object of class aquaenv

It is possible to clone an obect of class aquaenv, either 1 to 1 or with different pH, TA, or K_CO2

```
> Tc <- 15
> S <- 30
> SumCO2 <- 0.002
> TA <- 0.00214
> ae <- aquaenv(Tc, S, SumCO2 = SumCO2, TA = TA)
> ae$pH
[1] 7.998381
attr(,"pH scale")
[1] "free"
> ae1 <- aquaenv(ae = ae)</pre>
> ae1$pH
[1] 7.998381
attr(,"pH scale")
[1] "free"
> pH <- 9
> ae2 <- aquaenv(ae = ae, pH = pH)
> ae2$TA
[1] 0.002982756
attr(,"unit")
[1] "mol/kg-soln"
> TA <- 0.002
> ae3 <- aquaenv(ae = ae, TA = TA)
> ae3$pH
[1] 7.548175
attr(,"pH scale")
[1] "free"
> K_CO2 <- 1e-06
> ae4 <- aquaenv(ae = ae, k_co2 = K_CO2)
> ae4$pH
[1] 7.998381
attr(,"pH scale")
[1] "free"
```

Note that K_CO2 as an input variable is in lower cases!

3.1.6 Preparing input variables

Input variables for the function aquaenv need to be in mol/kg-solution and on the free pH scale. Data in other concentration units or pH scales can be converted using the function convert.

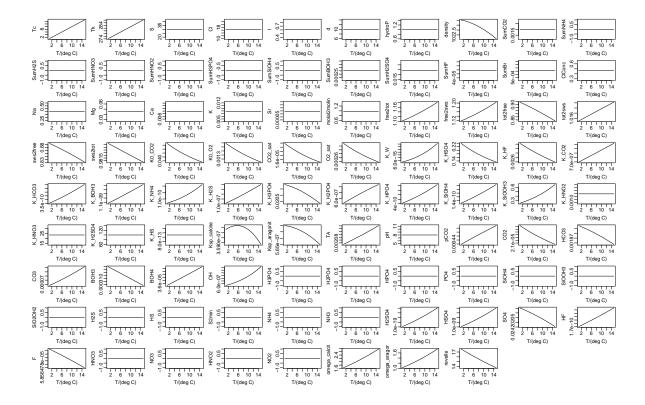
3.1.7 Vectors as input variables

One of the input variables for the function aquaenv may be a vector. All the other input variables are then assumed to be constant. The elements of the resulting two dimensional object of class aquaenv are then vectors containing the elements as a function of the input variable for which a vector is given.

```
> SumC02 <- 0.002
> pH <- 8
> Tc <- 1:15
> S <- 30
> d <- 10
> ae <- aquaenv(Tc, S, d, SumC02 = SumC02, pH = pH, revelle = TRUE)
> ae$revelle

[1] 17.60496 17.30007 16.99847 16.70067 16.40712 16.11820 15.83423 15.55550
[9] 15.28223 15.01461 14.75276 14.49681 14.24681 14.00282 13.76484
```

A two dimensional object of class aquaenv can be visualized using the plot function. For coppletions of the usef,cthexdefault savingsocille pleatheristic for for the class aquaenv results in a new plotting device being opened. Setting the flag newdevice to FALSE prevents that.



The plot function plots all elements of the respective object of class aquaenv. This, however, might not be what the user wants, especially if a larger plotting device cannot properly displayed like in the case above. In this case the parameter what can be used. Note, however, that the default setting for calling plot with the parameter what is that mfrow=c(1,1). So if one wants to plot several elements, mfrow needs to be set to a suitable value.

```
> plot(ae, xval = Tc, xlab = "T/(deg C)", what = c("pH", "C02", + "HC03", "C03"), newdevice = FALSE, mfrow = c(1, 4))

\frac{1}{2} \frac{1}{2} \frac{1}{4} \frac{1}{6} \frac{1}{8} \frac{10}{12} \frac{12}{14} \frac{14}{14} \frac{10}{12} \frac{10}{12} \frac{10}{14} \frac{10}{12} \frac{10
```

The following chunks of example code show other possible definitions of objects of class aquaenv with vectors as input variables.

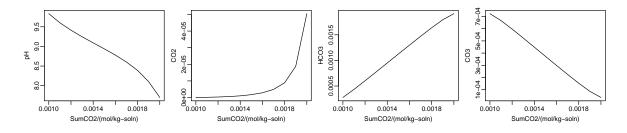
```
> ae <- aquaenv(Tc = 1:15, S = 30, d = 10, SumCO2 = SumCO2, TA = 0.0023) > plot(ae, xval = 1:15, xlab = "T/(deg C)")
```

$$>$$
 ae $<$ - aquaenv($Tc = 15$, $S = 20:30$, $d = 10$, $SumCO2 = SumCO2$, $TA = TA$) $>$ plot(ae, xval = 20:30, xlab = "S")

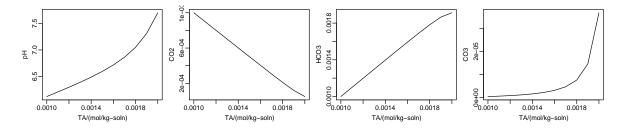
- > ae <- aquaenv(Tc = 15, S = 30, d = seq(1, 1000, 200), SumCO2 = SumCO2,
- + TA = TA, revelle = TRUE, dsa = TRUE)
- > plot(ae, xval = seq(1, 1000, 200), xlab = "depth/m")

Interesting to note is that also, e.g., SumCO2, TA, pH and SumNH4 can be vectors

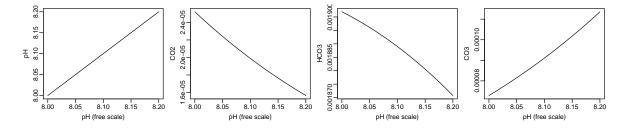
> ae <- aquaenv(10, 20, SumCO2 = seq(0.001, 0.002, 1e-04), TA = 0.002) > plot(ae, xval = ae\$SumCO2, xlab = "SumCO2/(mol/kg-soln)", what = c("pH", "CO2", "HCO3", "CO3"), newdevice = FALSE, mfrow = c(1, 4))



- > ae <- aquaenv(10, 20, SumCO2 = 0.002, TA = seq(0.001, 0.002,
- + 1e-04))
- > plot(ae, xval = aeTA, xlab = "TA/(mol/kg-soln)", what = c("pH",
- + "CO2", "HCO3", "CO3"), newdevice = FALSE, mfrow = c(1, 4))



> ae <- aquaenv(10, 20, SumCO2 = 0.002, pH = seq(8, 8.2, 0.001)) > plot(ae, xval = ae\$pH, xlab = "pH (free scale)", what = c("pH", + "CO2", "HCO3", "CO3"), newdevice = FALSE, mfrow = c(1, 4))



```
> ae <- aquaenv(10, 20, SumCO2 = 0.002, SumNH4 = seq(1e-04, 2e-04,
+ 1e-05), pH = 8)
> ae$NH3

[1] 1.534559e-06 1.688014e-06 1.841470e-06 1.994926e-06 2.148382e-06
[6] 2.301838e-06 2.455294e-06 2.608749e-06 2.762205e-06 2.915661e-06
[11] 3.069117e-06
attr(,"unit")
[1] "mol/kg-soln"
```

3.1.8 Calculating $[\sum CO_2]$ from input vectors

The functionality of calculating SumCO2 can also be used together with vectors as input variables.

```
> ae <- aquaenv(Tc = 11:15, S = 30, SumCO2 = NULL, pH = pH, CO2 = CO2,
+ revelle = TRUE, dsa = TRUE)
> ae$SumCO2

[1] 0.002415976 0.002463447 0.002511183 0.002559165 0.002607376
attr(,"unit")
[1] "mol/kg-soln"

Two further examples

> ae <- aquaenv(Tc = 15, S = 20:30, SumCO2 = NULL, pH = pH, pCO2 = pCO2)
> plot(ae, xval = 20:30, xlab = "S")

> ae <- aquaenv(Tc = 15, S = 30, d = seq(1, 1000, 100), SumCO2 = NULL,
+ pH = pH, TA = TA)
> plot(ae, xval = seq(1, 1000, 100), xlab = "depth/m")
```

3.1.9 Conversion from and to a dataframe

Objects of class aquaenv can be converted to an R data.frame to further post-process them with standard R means. Similarly, R data.frames can be converted to objects of class aquaenv to use the plotting facilities that exist for objects of class aquaenv. This can be helpful for plotting output of a dynamic model run, e.g. from R package **deSolve**, and will be shown later in this document.

```
> aedataframe <- as.data.frame(ae)
> aetest <- aquaenv(ae = aedataframe, from.data.frame = TRUE)</pre>
```

3.1.10 Converting elements in an obect of class aquaenv

Elements of an object of class *aquaenv* are calculated in, e.g., the concentration unit mol/kg-solution (molinity). The function **convert** can be used to convert all elements in an object of class *aquaenv* that share a common attribute, e.g. the unit.

3.1.11 Quantities needed for explicit pH modelling

As already mentioned above, the quantities needed for the explicit pH modelling approach (direct substitution approach - DSA) as presented by Hofmann *et al.* (2008b) can be calculated with the function aquaenv by setting the flag dsa.

```
> ae <- aquaenv(Tc = 15, S = 30, d = 10, SumCO2 = 0.002, pH = 8, + dsa = TRUE, revelle = TRUE)
```

This command calculated the buffer factor and the partial derivatives of [TA] with respect to other summed quantities referred to in Hofmann *et al.* (2008b)

> ae\$dTAdH

```
[1] -17142.03
attr(,"unit")
[1] "(mol-TA/kg-soln)/(mol-H/kg-soln)"
attr(,"pH scale")
[1] "free"
> ae$dTAdSumCO2
[1] 1.045937
attr(,"unit")
[1] "(mol-TA/kg-soln)/(mol-SumCO2/kg-soln)"
```

as well the sums partial derivatives of [TA] with respect to the equilibrium constants (K^* 's) multiplied with the partial derivatives of the respective equilibrium constant with one of their variables (i.e., S, T, d, SumH2SO4, od SumHF) as introduced in Hofmann *et al.* (2008a).

> ae\$dTAdKdKdS

```
[1] 3.986499e-06
attr(,"unit")
[1] "(mol-TA/kg-soln)/\"psu\""
```

> ae\$dTAdKdKdSumH2SO4

```
[1] -0.001086849
attr(,"unit")
[1] "(mol-TA/kg-soln)/(mol-SumH2S04/kg-soln)"
```

Furthermore, the ionization fractions used for the pH dependent fractional stoichiometric pH modelling approach described in Hofmann *et al.* (2008c) are calculated as well

> ae\$c1

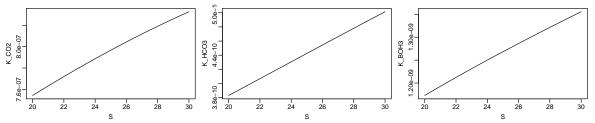
[1] 0.01025964

3.2 The plot.aquaenv function

In the previous sections, the plot function has been introduced. What actually is called if the first element of the arguments list of plot is an object pf class aquaenv is the function plot.aquaenv. This is a multifunctional tool to visualize information contained in an object of class aquaenv. For the convenience of the users, plot.aquaenv combines the call of standard R plotting functions and the previous call of the function par to set parameters like mfrow, mar, etc. as well as the opening of a plotting device with a certain size. As already shown above, setting the flag newdevice to FALSE suppresses the opening of a new plotting device (this feature is needed here to create a plot that will be woven into the LATEX document by Sweave).

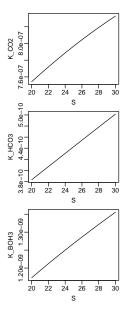
For example

```
> ae <- aquaenv(10, 20:30)
> plot(ae, xval = 20:30, xlab = "S", what = c("K_C02", "K_HC03", "K_B0H3"), size = c(10, 2), mfrow = c(1, 3), newdevice = FALSE)
```



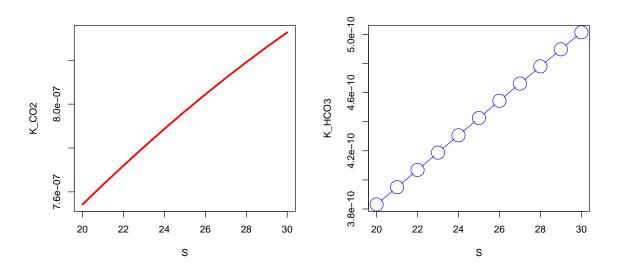
and

```
> plot(ae, xval = 20:30, xlab = "S", what = c("K_C02", "K_HC03", "K_B0H3"), size = c(2, 10), mfrow = c(3, 1), newdevice = FALSE)
```



Furthermore the parameter device can be specified which allows the user to write the plots to .eps and .pdf files. The parameter filename can be used to specify a filename other than the default filename "aquaenv".

These features make the function plot.aquaenv different from standard R plotting functions. However, if the flags newdevice and setpar are set to FALSE, plot.aquaenv behaves like a "normal" R plotting function



Furthermore, the function plot.aquaenv can be used to create "cumulative" plots and "Bjerrum" plots. This will be explained in some of the following sections.

3.3 Using objects of class aquaenv in dynamic models

3.3.1 Ordinary dynamic models

It is convenient to use objects of class *aquaenv* in a dynamic model, e.g. solved using the R package **deSolve**. This can be illustrated with an example. (For information about how to set up a dynamic model with **deSolve**, consult the documentation of **deSolve**).

A list of parameters is specified

```
> parameters <- list(</pre>
                     = 25
         S
                             , # psu
                             , #
         Tc\_min
                    = 5
                                 degrees C
                             , # degrees C
                     = 25
         Tc_max
                     = 10
                             , # m
         d
                                 , # 1/d
                                                         proportionality factor for air-water exchange
         k
                     = 0.4
         r0x
                    = 0.0000003 , \# mol-N/(kg*d)
                                                   maximal rate of oxic mineralisation
         rNitri
                    = 0.0000002 , # mol-N/(kg*d)
                                                   maximal rate of nitrification
                    = 0.000006 , # mol-N/(kg*d)
         rPP
                                                   maximal rate of primary production
         ksDINPP
                     = 0.000001
                                 , # mol-N/kg
         ksNH4PP
                     = 0.000001
                                 , # mol-N/kg
                     = 0.1
                                       1/d
                                                    (dispersive) transport coefficient
                                 , # mol/kg-soln
         02_io
                    = 0.000296
                                 , # mol/kg-soln
         NO3 io
                     = 0.000035
         SumNH4_io
                    = 0.000008
                                   # mol/kg-soln
                                 , # mol/kg-soln
         SumCO2_io
                    = 0.002320
                                 , # mol/kg-soln
                     = 0.002435
         TA\_io
         C_Nratio
                    = 8
                                 , # mol C/mol N
                                                      C:N ratio of organic matter
                     = 30
                                   # timestep from which PP begins
```

```
+ b = 50 , # timestep where PP shuts off again

+ modeltime = 100 # duration of the model

+ )
```

A model function is defined which will be executed every timestep of the numerical integration. An object of class *aquaenv* is created in each timestep, some of its elements are used to calculated kinetic rate expressions and the whole object is returned as output.

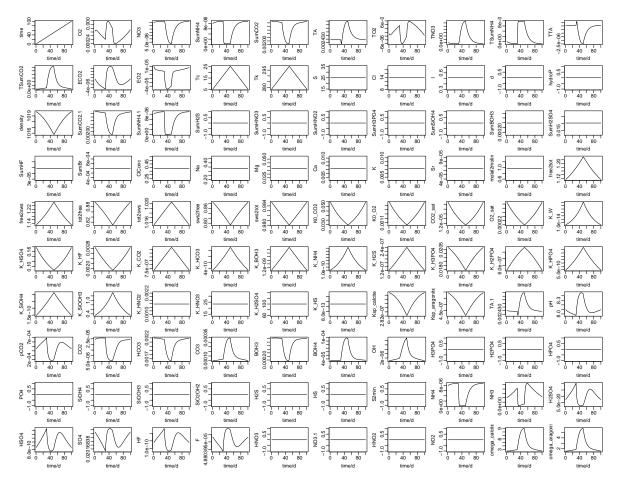
```
> Waddenzeebox <- function(timestep, currentstate, parameters)</pre>
   with (
          as.list(c(currentstate,parameters)),
           Tc \leftarrow c(seq(Tc\_min, Tc\_max, (Tc\_max-Tc\_min)/(modeltime/2)),
                    seq(Tc_max, Tc_min, -(Tc_max-Tc_min)/(modeltime/2)))[[round(timestep)+1]]
            ae <- aquaenv(Tc=Tc, S=S, SumCO2=SumCO2, SumNH4=SumNH4, TA=TA)
            ECO2
                    <- k * (ae$CO2_sat - ae$CO2)
                    <- k * (ae$02_sat - 02)
           E02
                    <- D*(02 io
            TN2
                                    - 02)
                                  - NO3)
            TNO3
                   <- D*(NO3_io
            TSumNH4 <- D*(SumNH4_io - SumNH4)
                                 - TA)
            TTA <- D*(TA_io
            TSumCO2 \leftarrow D*(SumCO2\_io - SumCO2)
           RNit
                      <- rNitri
                     <- r0x
            R\Omega x
           ROxCarbon <- ROx * C_Nratio
           pNH4PP <- 0
           RPP <- 0
            if ((timestep > a) && (timestep < b))</pre>
                      <- rPP * ((SumNH4+NO3)/(ksDINPP + (SumNH4+NO3)))
               pNH4PP <- 1 - (ksNH4PP/(ksNH4PP + SumNH4))
            else
               RPP <- 0
              7
            RPPCarbon <- RPP * C_Nratio
                               + EO2 - ROxCarbon - 2*RNit + (2-2*pNH4PP)*RPP + RPPCarbon
            d02
                    <- T02
                    <- TNO3
                              + RNit -(1-pNH4PP)*RPP
            dNO3
            dSumCO2 <- TSumCO2 + ECO2 + ROxCarbon - RPPCarbon
            dSumNH4 <- TSumNH4 + ROx - RNit - pNH4PP*RPP
                              + ROx - 2*RNit -(2*pNH4PP-1)*RPP
            ratesofchanges <- c(d02, dN03, dSumNH4, dSumCO2, dTA)
           transport <- c(T02=T02, TN03=TN03, TSumNH4=TSumNH4, TTA=TTA, TSumC02=TSumC02)
           airseaexchange <- c(ECO2=ECO2, EO2=EO2)</pre>
           return(list(ratesofchanges, list(transport, airseaexchange, ae)))
         }
```

The model is solved

```
> with (as.list(parameters),
+ {
+ initialstate <- c(02=02_io, N03=N03_io, SumNH4=SumNH4_io, SumC02=SumC02_io, TA=TA_io)
+ times <- c(0:modeltime)
+ output <- as.data.frame(vode(initialstate,times,Waddenzeebox,parameters, hmax=1))[-1,]
+ })</pre>
```

and the output can be plotted in the same way as a two dimensional object of class aquaenv by converting it to an object of class aquaenv using the from.data.frame flag of the function aquaenv

```
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = outputtime,
+ xlab = "time/d", mfrow = c(10, 10), newdevice = FALSE)
```



3.3.2 Models using the explicit pH modelling approach

3.3.2.1 In one single model

Since an object of class *aquaenv* can contain all quantities necessary to employ the explicit pH modelling approaches as introduced by Hofmann *et al.* (2008b,a,c), they can be readily used in an explicit pH model.

As an example, we give a model that calculates the pH in the "classical" way in every timestep using aquaenv, also employs the explicit pH modelling approach (direct substitution approach - DSA) given in Hofmann et al. (2008b) and additionally employs fractional stoichiometry as given in Hofmann et al. (2008c). The pH evolution is thus calculated in three different ways which allows comparing the three values for consistency.

Again, a list of parameters is defined

```
> parameters <- list(</pre>
       \mathcal{S}
                    = 25
                             , # psu
       Tc
                    = 15
                             , # degrees C
                             , # m
                    = 10
       d
                                  , # 1/d
                                                       proportionality factor for air-water exchange
                    = 0.4
       r0x
                    = 0.0000003 , # mol-N/(kg*d) maximal rate of oxic mineralisation
                    = 0.0000002 , # mol-N/(kg*d) maximal rate of nitrification = 0.0000006 , # mol-N/(kg*d) maximal rate of primary production
       rNitri
       rPP
                    = 0.000001 , # mol-N/kg
       ksSumNH4
                                                          (dispersive) transport coefficient
                                 , # mol/kg-soln
       02_io
                    = 0.000296
                                 , # mol/kg-soln
       NO3_io
                    = 0.000035
                                 , # mol/kg-soln
       SumNH4_io = 0.000008
                                 , # mol/kg-soln
, # mol/kg-soln
       SumCO2_io
                    = 0.002320
                    = 0.002435
       TA_{-}io
       C Nratio
                                , # mol C/mol\ N
                                                      C:N ratio of organic matter
                    = .30
                                 , # timestep from which PP begins
       b
                    = 50
                                 , # timestep where PP shuts off again
       modeltime
                   = 100
                                   # duration of the model
```

And a model function is defined. Again, an object of class *aquaenv* is created in each timestep and respective elements are used.

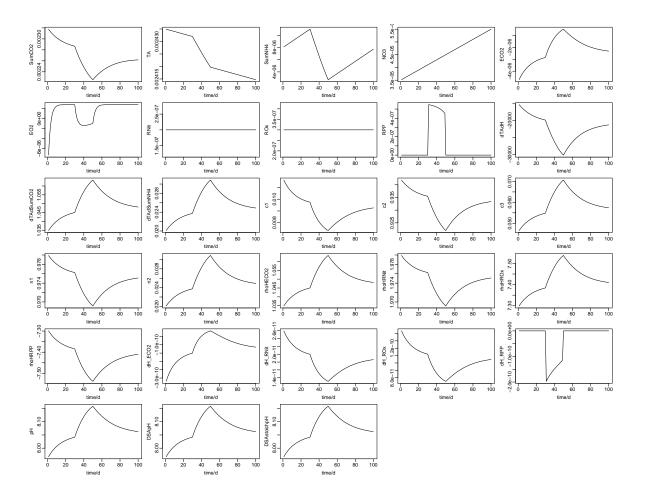
```
> boxmodel <- function(timestep, currentstate, parameters)</pre>
+ {
    with (
          as.list(c(currentstate,parameters)),
            # the "classical" implicit pH calculation method is applied in aquaenv
            ae <- aquaenv(Tc=Tc, S=S, SumCO2=SumCO2, SumNH4=SumNH4, TA=TA, dsa=TRUE)
            EC02
                     <- k * (ae$CO2_sat - ae$CO2)
                     <- k * (ae$02_sat - 02)
            E02
            RNit
                       <- rNitri
            ROx
                      <- r0x
            if ((timestep > a) && (timestep < b))
                RPP <- rPP * (SumNH4/(ksSumNH4 + SumNH4))</pre>
              }
            else
                RPP <- 0
```

```
d02
        <- EO2 - C_Nratio*ROx - 2*RNit + C_Nratio*RPP
dNO3
        <- RNit
dSumCO2 <- ECO2 + C_Nratio*ROx - C_Nratio*RPP
dSumNH4 <- ROx - RNit - RPP
        <- ROx - 2*RNit - RPP
# The DSA pH
dH <- (dTA - (dSumCO2*ae$dTAdSumCO2 + dSumNH4*ae$dTAdSumNH4))/ae$dTAdH
DSApH <- -log10(H)
# The DSA pH using pH dependent fractional stoichiometry (= using partitioning coefficients)
rhoHECO2 <- ae$c2 + 2*ae$c3
rhoHRNit <- 1 + ae$n1
rhoHROx <- C_Nratio * (ae$c2 + 2*ae$c3) - ae$n1
rhoHRPP \leftarrow (C_Nratio * (ae$c2 + 2*ae$c3)) + ae$n1
dH_ECO2 <- rhoHECO2*ECO2/(-ae$dTAdH)</pre>
dH_RNit <- rhoHRNit*RNit/(-ae$dTAdH)</pre>
dH_ROx <- rhoHROx*ROx /(-ae$dTAdH)</pre>
dH_RPP <- rhoHRPP*RPP /(-ae$dTAdH)
dH_stoich <- dH_ECO2 + dH_RNit + dH_ROx + dH_RPP
DSAstoichpH <- -log10(H_stoich)
ratesofchanges \leftarrow c(d02, dN03, dSumNH4, dSumC02, dTA, dH, dH\_stoich)
processrates <- c(ECO2=ECO2, EO2=EO2, RNit=RNit, ROx=ROx, RPP=RPP)</pre>
               <- c(DSApH=DSApH, rhoHECO2=rhoHECO2, rhoHRNit=rhoHRNit, rhoHROx=rhoHROx,</pre>
DSA
                    rhoHRPP=rhoHRPP, dH_ECO2=dH_ECO2, dH_RNit=dH_RNit, dH_ROx=dH_ROx,
                     dH_RPP=dH_RPP, DSAstoichpH=DSAstoichpH)
return(list(ratesofchanges, list(processrates, DSA, ae)))
```

The model is solved

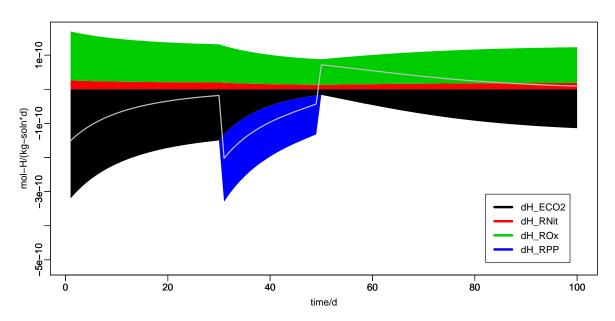
```
> with (as.list(parameters),
+ {
+ H_init <<- 10^(-(aquaenv(Tc=Tc, S=S, SumCO2=SumCO2_io, SumNH4=SumNH4_io, TA=TA_io, speciation=FALSE)$pH))
+ initialstate <<- c(O2=O2_io, NO3=NO3_io, SumNH4=SumNH4_io, SumCO2=SumCO2_io, TA=TA_io, H=H_init, H_stoich=H_init)
+ times <<- c(O:modeltime)
+ output <<- as.data.frame(vode(initialstate, times, boxmodel, parameters, hmax=1))[-1,]
+ })</pre>
```

and output can be plotted. Again using plot.aquaenv. Note that here the parameter what is used.



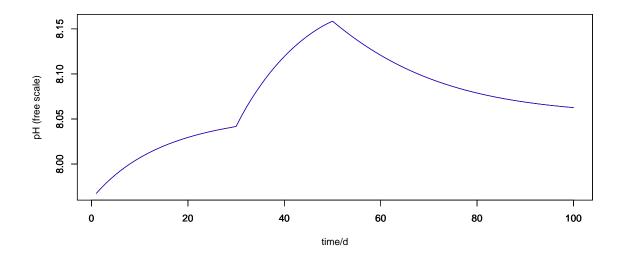
Here, the cumulative plotting functionality of plot.aquaenv can be employed as well to visualize the influences of the different kinetically modelled processes on $[H^+]$.

```
> what <- c("dH_ECO2", "dH_RNit", "dH_ROx", "dH_RPP")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+ what = what, xlab = "time/d", size = c(7, 5), ylab = "mol-H/(kg-soln*d)",
+ legendposition = "bottomright", cumulative = TRUE, newdevice = FALSE)
```



Finally, the pH values calculated with the three different methods can be plotted in one single graph to see that they are identical, i.e. the three methods of pH calculation are consistent with each other

```
> ylim <- range(output$DSApH, output$DSAstoichpH, output$pH)
> plot(output$pH, ylim = ylim, type = "l", xlab = "time/d", ylab = "pH (free scale)")
> par(new = TRUE)
> plot(output$DSApH, ylim = ylim, type = "l", col = "red", xlab = "",
+ ylab = "")
> par(new = TRUE)
> plot(output$DSAstoichpH, ylim = ylim, type = "l", col = "blue",
+ xlab = "", ylab = "")
```

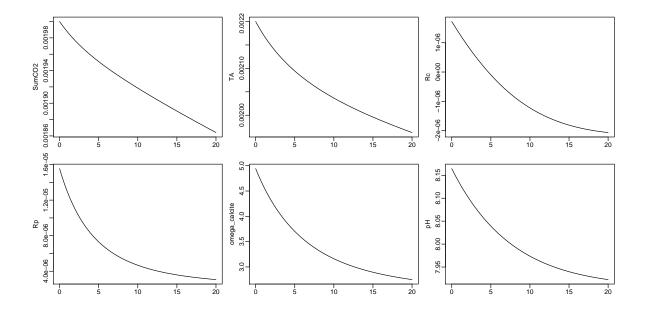


3.3.2.2 In three separate models

3.3.2.2.1 The implicit pH modelling approach

```
A list of parameters
```

```
> parameters <- list(</pre>
                = .35
                            , # psu
                = 15
                            , # degrees C
                            , # mol/kg-soln (comparable to Wang2005)
     SumCO2\_tO = 0.002
                = 0.0022
                            , # mol/kg-soln (comparable to Millero1998)
                            , # 1/d
                = 0.5
                                                  proportionality factor for air-water exchange
                = 0.000001 , \# mol/(kg-soln*d)
                                                    max rate of calcium carbonate precipitation
     kp
                            , # -
                = 2.0
                                                exponent for kinetic rate law of precipitation
     n
     modeltime = 20
                            , # d
                                             duration of the model
     outputsteps = 100
                             #
                                             number of outputsteps
The model function
> boxmodel <- function(timestep, currentstate, parameters)</pre>
   with (
         as.list(c(currentstate,parameters)),
                <- aquaenv(Tc=Tc, S=S, SumCO2=SumCO2, TA=TA, SumSiOH4=0,</pre>
           ae
                           SumBOH3=0, SumH2SO4=0, SumHF=0)
                <- kc * ((ae$CO2_sat) - (ae$CO2))
           Rc
                <- kp * (1-ae$omega_calcite)^n
           dSumCO2 <- Rc - Rp
                  <- -2*Rp
           ratesofchanges <- c(dSumCO2, dTA)
          processrates <- c(Rc=Rc, Rp=Rp)
           return(list(ratesofchanges, list(processrates, ae)))
+ }
Solving the model
> with (as.list(parameters),
         initialstate <<- c(SumCO2=SumCO2_t0, TA=TA_t0)</pre>
                 <<- seq(0,modeltime,(modeltime/outputsteps))</pre>
         times
                     <-- as.data.frame(vode(initialstate,times,boxmodel,parameters, hmax=1))
         output
       })
Visualizing the output
> what <- c("SumCO2", "TA", "Rc", "Rp", "omega_calcite", "pH")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
       xlab = "", mfrow = c(2, 3), size = c(12, 5), what = what,
       newdevice = FALSE)
```



3.3.2.2.2 The explicit pH modelling approach

A list of parameters

```
> parameters <- list(</pre>
                               , # psu
                  = 35
                  = 15
                               , # degrees C
      Tc
                               , # mol/kg-soln (comparable to Wang2005)
      SumCO2_tO
                  = 0.002
      TA\_t0
                  = 0.0022
                               , # mol/kg-soln
                                                 (comparable to Millero1998)
                               , # 1/d
                  = 0.5
                                                        proportionality factor for air-water exchange
      kc
                                                            \hbox{\it max rate of calcium carbonate precipitation}\\
                  = 0.000001
                               , # mol/(kg-soln*d)
      kр
                  = 2.0
                                                      exponent for kinetic rate law of precipitation
                               , # d
      modeltime
                  = 20
                                                   duration of the model
      outputsteps = 100
                                                   number of outputsteps
```

The model function

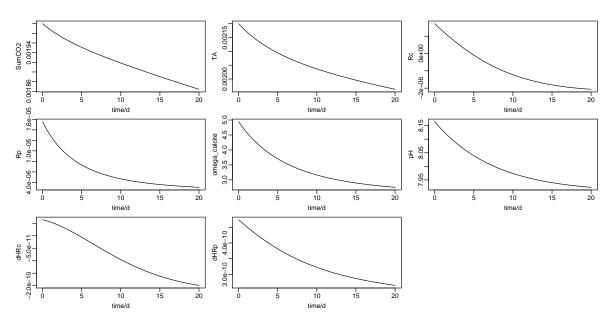
```
> boxmodel <- function(timestep, currentstate, parameters)</pre>
    with (
          as.list(c(currentstate,parameters)),
                  <- aquaenv(Tc=Tc, S=S, SumCO2=SumCO2, pH=-log10(H), SumSiOH4=0,</pre>
            ae
                              SumBOH3=0, SumH2SO4=0, SumHF=0, dsa=TRUE)
                  <- kc * ((ae$CO2_sat) - (ae$CO2))
            Rc
                  <- kp * (1-ae$omega_calcite)^n
            Rр
            dSumCO2 <- Rc - Rp
                               -(ae$dTAdSumCO2*Rc ))/ae$dTAdH
            dHRp
                    <- (-2*Rp -(ae$dTAdSumCO2*(-Rp)))/ae$dTAdH
            dН
                    <- dHRc + dHRp
            ratesofchanges <- c(dSumCO2, dH)
```

```
+ processrates <- c(Rc=Rc, Rp=Rp)
+ outputvars <- c(dHRc=dHRc, dHRp=dHRp)
+
+ return(list(ratesofchanges, list(processrates, outputvars, ae)))
+ }
+ )
+ )</pre>
```

Solving the model

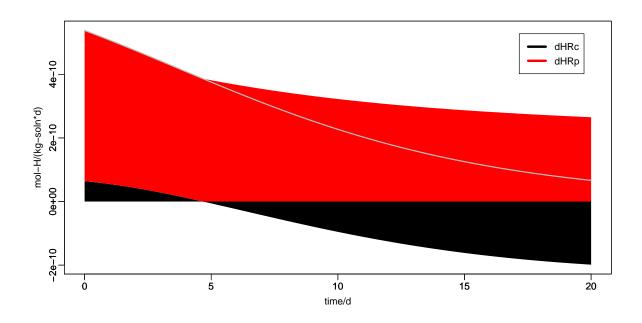
Visualizing the output

```
> what <- c("SumCO2", "TA", "Rc", "Rp", "omega_calcite", "pH",
+    "dHRc", "dHRp")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+    xlab = "time/d", mfrow = c(3, 3), size = c(15, 10), what = what,
+    newdevice = FALSE)
```



Cumulatively plotting the influences of the two processes on the pH

```
> what <- c("dHRc", "dHRp")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+ xlab = "time/d", what = what, ylab = "mol-H/(kg-soln*d)",
+ legendposition = "topright", cumulative = TRUE, newdevice = FALSE)
```



3.3.2.2.3 The fractional stoichiometric approach

A list of parameters

```
> parameters <- list(</pre>
                                    , # psu
       \mathcal{S}
                     = 35
                      = 15
                                    , # degrees C
       Tc
                                   , # mol/kg-soln (comparable to Wang2005)
, # mol/kg-soln (comparable to Millero1998)
       SumCO2_tO
                     = 0.002
       TA\_tO
                     = 0.0022
                     = 0.5
                                   , # 1/d
, # mol/(kg-soln*d)
                                                                proportionality factor for air-water exchange
       kc
                                                                    max rate of calcium carbonate precipitation
                      = 0.000001
       kр
                      = 2.0
                                                             exponent for kinetic rate law of precipitation
       {\tt modeltime}
                    = 20
                                    , # d
                                                          duration of the model
        outputsteps = 100
                                                          number of outputsteps
```

The model function

```
> boxmodel <- function(timestep, currentstate, parameters)</pre>
    with (
          as.list(c(currentstate,parameters)),
                  <- aquaenv(Tc=Tc, S=S, SumCO2=SumCO2, pH=-log10(H), SumSiOH4=0,</pre>
            ae
                              SumBOH3=0, SumH2SO4=0, SumHF=0, dsa=TRUE)
            Rc
                  <- kc * ((ae$CO2_sat) - (ae$CO2))
                  <- kp * (1-ae$omega\_calcite)^n
            Rр
            dSumCO2 <- Rc - Rp
            rhoc
                     <- ae$c2 + 2*ae$c3
            rhop
                    <- 2*ae$c1 + ae$c2
                    <- rhoc*Rc/(-ae$dTAdH)
            dHRc
            dHRp
                    <- rhop*Rp/(-ae$dTAdH)
            dН
                     <- dHRc + dHRp
```

```
ratesofchanges <- c(dSumCO2, dH)
             processrates
                              <- c(Rc=Rc, Rp=Rp)
                              <- c(dHRc=dHRc, dHRp=dHRp, rhoc=rhoc, rhop=rhop)
             outputvars
             return(list(ratesofchanges, list(processrates, outputvars, ae)))
Solving the model
> with (as.list(parameters),
           aetmp <- aquaenv(Tc=Tc, S=S, SumCO2=SumCO2_t0, TA=TA_t0, SumSiOH4=0,</pre>
                              SumBOH3=0, SumH2SO4=0, SumHF=0)
           H_t0 <- 10^(-aetmp$pH)</pre>
           initialstate <<- c(SumCO2=SumCO2_t0, H=H_t0)</pre>
                          <-- seq(0,modeltime,(modeltime/outputsteps))
           times
           output
                          <\!\!<\!\!-\text{ as.data.frame}(\texttt{vode}(\texttt{initialstate},\texttt{times},\texttt{boxmodel},\texttt{parameters},~\texttt{hmax=1}))
Visualizing the output
> what <- c("SumCO2", "TA", "Rc", "Rp", "omega_calcite", "pH",
         "dHRc", "dHRp", "rhoc", "rhop")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
         xlab = "time/d", mfrow = c(3, 4), size = c(15, 10), what = what,
         newdevice = FALSE)
                                0.00215
  0.00194
                                                                                            1.0e-05
                                                            Rc
0e+00
                              Ϋ́
                                                                                          Вp
                                                              -2e-06
                     15
                                                                                       20
                10
                                             10
  5.0
                                8.15
  4.0 4.5
omega_calcite
                                                            dHRc
-5.0e-11
                                                                                            4.0e-
                                                                                          dHRp
                              PH
8.05
  3.5
                                                                                            3.0e-10
  3.0
                                7.95
                     15
                                                   15
                                                                                 15
                                                                                                               15
                10
                                             10
                                                                           10
                                                                                                         10
                                                                                                        time/d
                                0.94
  1.08
rhoc
                              rhop
                                0.92
  1.06
```

3.4 Titration simulation: the function titration

10

With the function **titration AquaEnv** provides a powerful tool to simulate titrations. A two dimensional object of class *aquaenv* will be created where the second dimension is the

amount of titrant added. For this purpose, three extra elements are added to the *aquaenv* object that will be created:

| elemer | \mathbf{nt} | unit | explanation |
|----------|---------------|-----------------|--|
| delta_co | onc_titrant | mol/kg-solution | the offset in concentration of the titrant that is caused by adding the titrant to the sample |
| delta_m | nass_titrant | kg | the amount of mass of titrant solution added |
| delta_m | noles_titrant | mol | the amount of moles of titrant added |

Each one of this elements is a suitable xval for plotting an aquaenv object generated by titration.

3.4.1 Titration with HCl

The standard titration type is titration with hydrochloric acid (HCl). A simple example will illustrate this best.

An object of class *aquaenv* needs to be created to define the initial conditions of the titration. That is temperature, salinity, depth, the concentrations of all summed quantities and the initial pH (or [TA]).

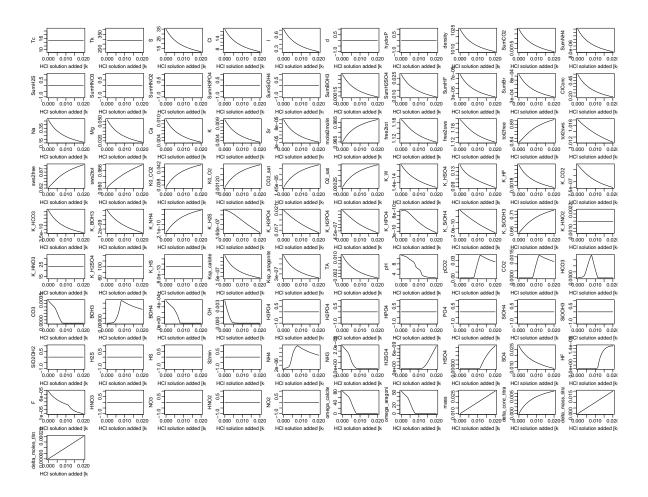
```
> ae_init <- aquaenv(Tc = 15, S = 35, SumCO2 = 0.0035, SumNH4 = 2e-05, pH = 11.3)
```

Then titration can be run to create the object describing the simulated titration. In this example the titrant is HCl of the relatively low concentration of 0.01 mol/kg-solution. The sample solution amounts to 10 g. To sweep a considerable pH range quite a lot of sample needs to be added: 20 g. This means the salinity of the solution in the titration vessel will change due to dilution with the titrant solution. For this reason, the salinity of the titrant solution needs to be given via the parameter S_titrant. However, we assume the titrant does not contain borate, sulfate or fluoride, that is why we do not set the flag seawater_titrant to TRUE.

```
> ae <- titration(ae_init, mass_sample = 0.01, mass_titrant = 0.02,
+ conc_titrant = 0.01, S_titrant = 0.5, steps = 100)</pre>
```

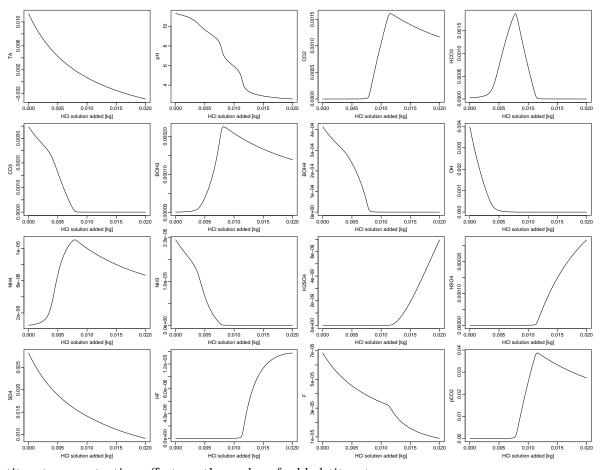
To get a quick overview, all elements of the obtained aquaenv object can be plotted

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]", + mfrow = c(10, 10), newdevice = FALSE)
```



Then, a selection of elements can be plotted as a function of the added titrant mass,

```
> what <- c("TA", "pH", "CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", + "NH4", "NH3", "H2SO4", "HSO4", "SO4", "HF", "F", "pCO2") > plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]", + what = what, size = <math>c(12, 8), mfrow = c(4, 4), newdevice = FALSE)
```

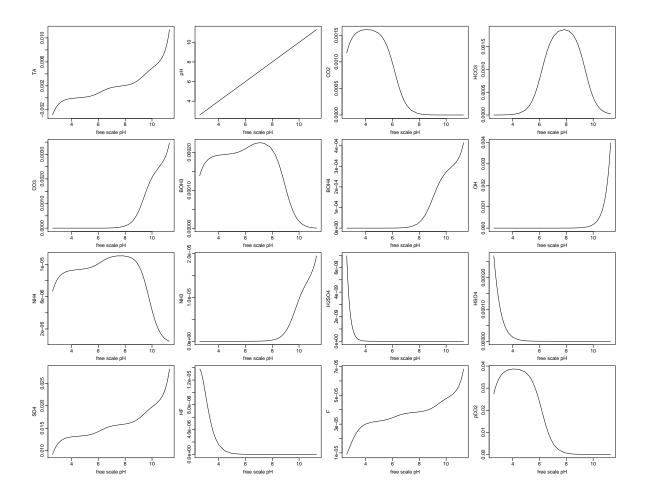


titrant concentration offset, or the moles of added titrant

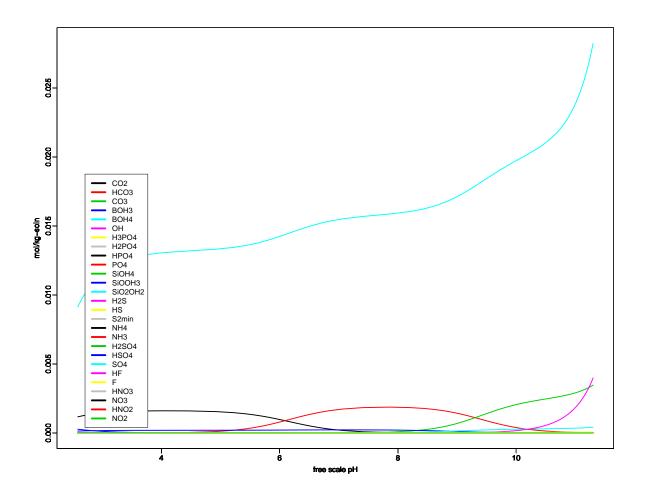
```
> plot(ae, xval = ae\$delta_conc_titrant, xlab = "[HCl] offset added [mol/kg-soln]",
+ what = what, size = c(14, 10), mfrow = c(12, 8))
> plot(ae, xval = ae\$delta_moles_titrant, xlab = "HCl added [mol]",
+ what = what, size = c(14, 10), mfrow = c(12, 8))
```

However, it is also possible to plot this selection of elements against othe variables, e.g., the calculated free scale pH

```
> plot(ae, xval = aepH, xlab = "free scale pH", what = what, size = c(12, + 8), mfrow = c(4, 4), newdevice = FALSE)
```

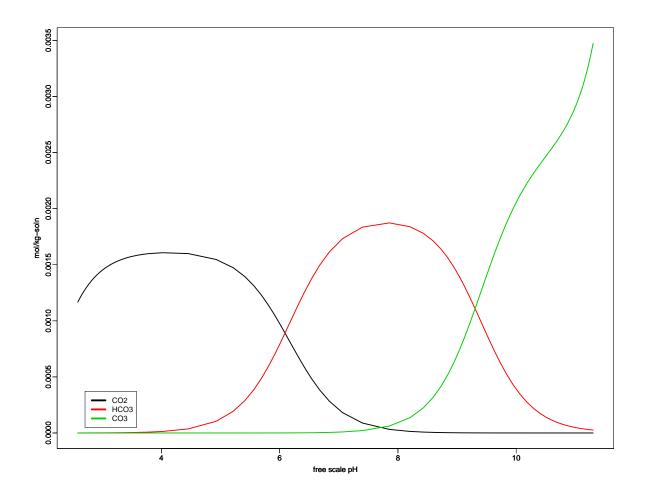


As mentioned earlier, the function plot.aquaenv offers the possibility of creating bjerrum plots from objects obtained with titration. The simplest way to do that is (remember the newdevice=FALSE is just needed to produce plots that are nicely woven into this vignette)

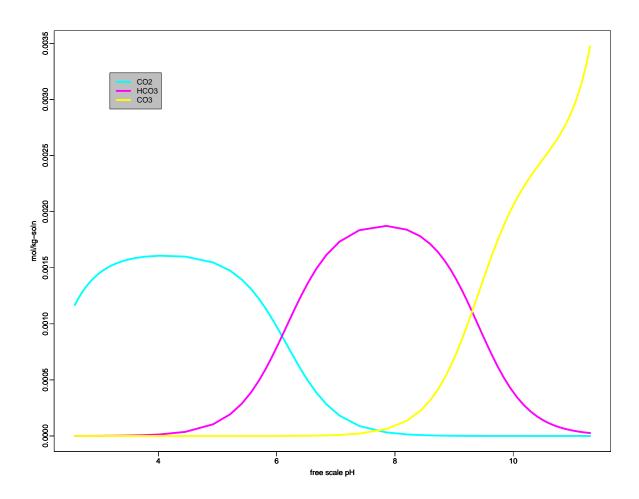


Or just select a few elements

```
> what <- c("CO2", "HCO3", "CO3")
> plot(ae, what = what, bjerrum = TRUE, newdevice = FALSE)
```

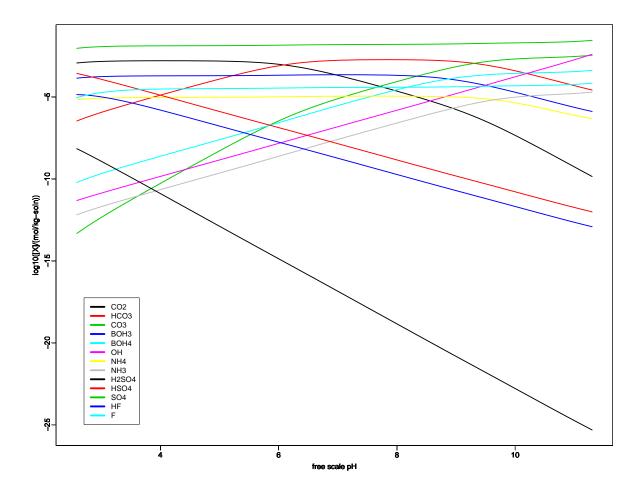


Again, the plotting style can be customized



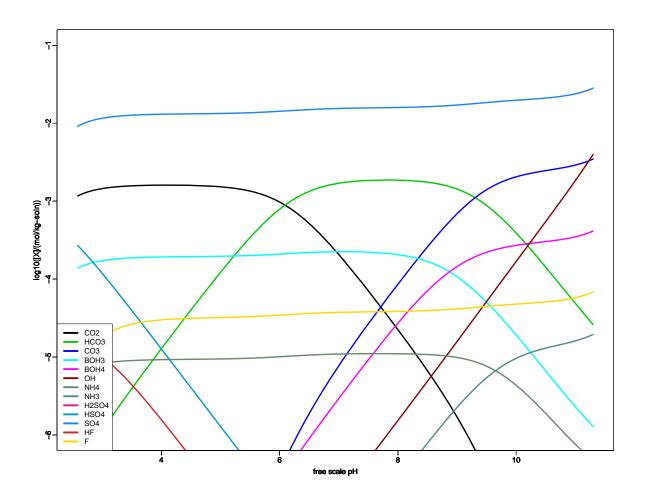
However, generally Bjerrum plots are done on the log scale. This can be accomplished using the flag log

```
> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",
+ "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, newdevice = FALSE)
```



Furthermore, we can zoom in to the region of most interest to marine scientists

```
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6, + -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6, colors()[seq(100, 250, 6)]), newdevice = FALSE)
```



3.4.2 Titration with NaOH

Similar to the titration with HCl, also a titration with NaOH can be simulated

```
> ae <- titration(aquaenv(Tc = 15, S = 35, SumCO2 = 0.0035, SumNH4 = 2e-05, + pH = 2), mass_sample = 0.01, mass_titrant = 0.02, conc_titrant = 0.01, + S_{titrant} = 0.5, steps = 50, type = "NaOH")
```

Plottinge everything

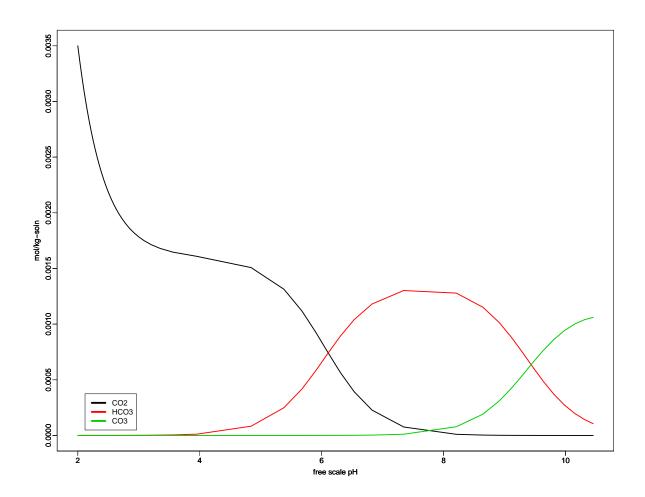
```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "NaOH solution added [kg]", + mfrow = c(10, 10))
```

Plotting selectively

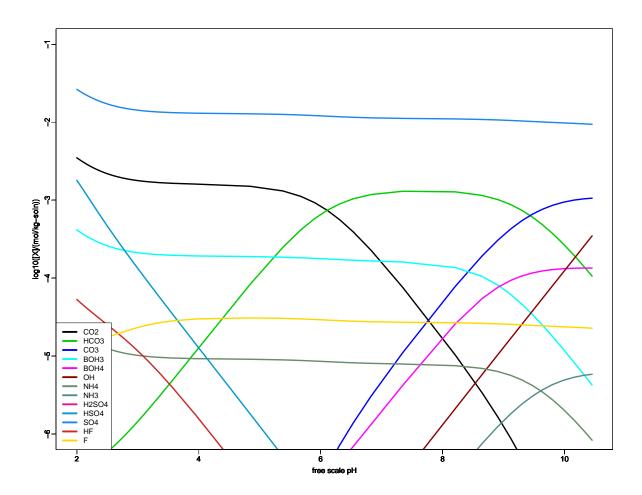
```
> what <- c("TA", "pH", "CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4", "NH3", "H2SO4", "HSO4", "SO4", "HF", "F", "pCO2") > plot(ae, xval = ae$delta_mass_titrant, xlab = "NaOH solution added [kg]", what = what, size = <math>c(12, 8), mfrow = c(4, 4)) > plot(ae, xval = ae$pH, xlab = "free scale pH", what = what, size = c(12, 8), mfrow = c(4, 4))
```

Bjerrum plots

```
> what <- c("CO2", "HCO3", "CO3")
> plot(ae, what = what, bjerrum = TRUE, newdevice = FALSE)
```



```
> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",
+ "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6,
+ -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6,
+ colors()[seq(100, 250, 6)]), newdevice = FALSE)
```



3.4.3 Titration with a titrant with high concentrations and a large sample volume - classical Bjerrum plots

The Bjerrum plots created in the previous two sections do not really look like the classical textbook ones. This is because we simulated a titration with a small sample volume and a titrant with low concentrations. As a result the total concentrations like, e.g., total carbonate decreased due to dilution. In simulating a titration with a rather large volume and a titrant with high concentrations the volume and salinity corrections do not matter any more and graphs known from textbooks (e.g. Zeebe and Wolf-Gladrow 2001) are produced.

```
> ae <- titration(aquaenv(Tc = 15, S = 35, SumCO2 = 0.0035, SumNH4 = 2e-05, + pH = 11.3), mass_sample = 100, mass_titrant = 0.5, conc_titrant = 3, + S_{titrant} = 0.5, steps = 100)
```

Plotting everything

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]", + mfrow = c(10, 10))
```

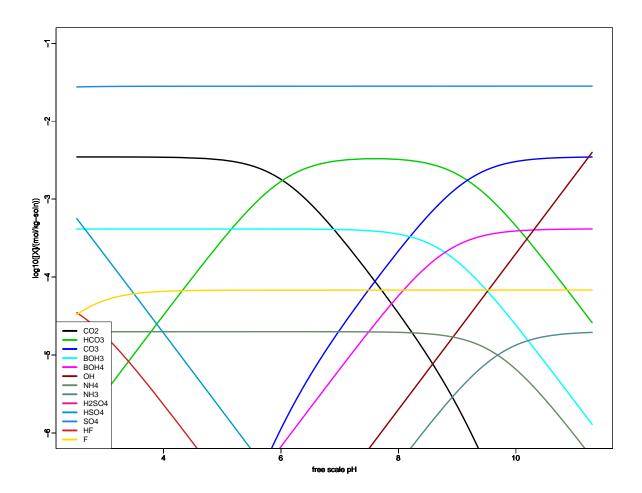
Plotting selectively and with different elements for xval

```
> what <- c("TA", "pH", "CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", + "NH4", "NH3", "H2SO4", "HSO4", "SO4", "HF", "F", "pCO2")
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]", + what = what, size = c(12, 8), mfrow = c(4, 4))
> plot(ae, xval = ae$pH, xlab = "free scale pH", what = what, size = c(12, + 8), mfrow = c(4, 4))
> plot(ae, xval = ae$delta_conc_titrant, xlab = "[HCl] offset added [mol/kg-soln]", + what = what, size = c(12, 8), mfrow = c(4, 4))
> plot(ae, xval = ae$delta_moles_titrant, xlab = "HCl added [mol]", + what = what, size = c(12, 8), mfrow = c(4, 4))
```

Creating different kinds of Bjerrum plots

and the classical textbook one

```
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6, + -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6, + colors()[seq(100, 250, 6)]), newdevice = FALSE)
```



3.5 Calculating information from titration curves: the function TAfit

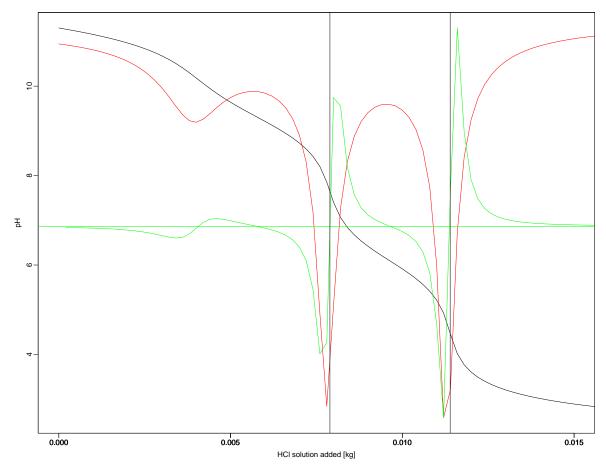
3.5.1 A little theory

While titrating a sample of natural seawater with HCl there one sees two clear equivalence points (Dickson 1981) The second equivalence point is the equivalence point of total alkalinity and the difference between the second and the first equivalence point signifies the total amount of $\sum CO_2$ of the sample Hansson and Jagner (1973).

This can be illustrated with <code>AquaEnv</code>. The respective titration curve can be plotted, together with its first and second derivative. Furthermore, the equivalence points can be marked with vertical lines (Please note that for a titrant concentration of 0.01 mol/kg-solution and 0.01 kg of sample, the value of the concentration (in mol/kg-solution) of total alkalinity and total carbonate equals the value of the total amount (in mol)).

```
> ae_init <- aquaenv(Tc = 15, S = 35, SumCO2 = 0.0035, SumNH4 = 2e-05, 
+ pH = 11.3)
> ae <- titration(ae_init, mass_sample = 0.01, mass_titrant = 0.02, 
+ conc_titrant = 0.01, S_titrant = 0.5, steps = 100)
> plot(ae, xval = ae\$delta_mass_titrant, xlab = "HCl solution added [kg]",
```

```
+ what = "pH", xlim = c(0, 0.015), newdevice = FALSE)
> par(new = TRUE)
> plot(ae$delta_mass_titrant[1:100], diff(ae$pH), type = "l", col = "red",
+ xlim = c(0, 0.015), ylab = "", xlab = "", yaxt = "n")
> par(new = TRUE)
> plot(ae$delta_mass_titrant[2:100], diff(diff(ae$pH)), type = "l",
+ col = "green", xlim = c(0, 0.015), ylab = "", xlab = "",
+ yaxt = "n")
> abline(h = 0, col = "green")
> abline(v = ae$TA[[1]]) - ae$SumCO2[[1]])
```

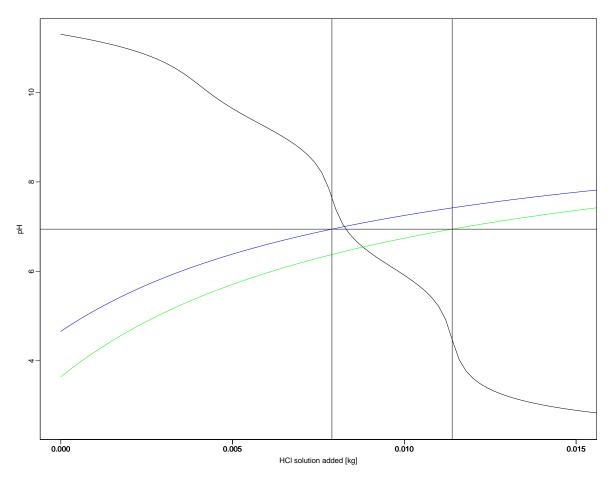


Following classical chemical textbooks (e.g. Skoog and West 1982), one can determine [TA] and $[\sum CO_2]$ of a sample by graphically determining those equivalence points. However, there is no mechanistic understanding of the contents of the solution involved in doing so.

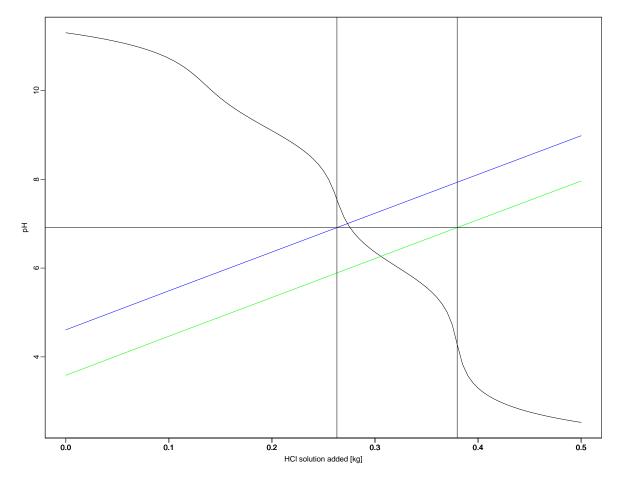
Other methods, called "Gran evaluations" (Gran 1952; Hansson and Jagner 1973; Dickson 1981; Haraldsson, Anderson, Hassellov, Hulth, and Olsson 1997; Anderson, Turner, Wedborg, and Dyrssen 1999), try to linearize the mechanistic model of what is going on in the solution during titration. They define the so called linear "Gran functions" and try to find their roots to determine the equivalence points. We will illustrate that by plotting the Gran functions

F0 (blue) and F2 (-F1, green) and again mark the equivalence points with vertical lines. The y=zero line for the Gran functions is indicated by a horizontal line

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",
      what = "pH", xlim = c(0, 0.015), newdevice = FALSE)
> prot1 <- c()
> for (i in 1:length(ae$pH)) {
     prot1 <- c(prot1, (10^-(ae$pH[[i]]) + ae$HSO4[[i]] + ae$HF[[i]] +</pre>
          ae$CO2[[i]] - ae$CO3[[i]] - ae$BOH4[[i]] - ae$OH[[i]]))
+ }
> par(new = TRUE)
> plot(ae$delta_mass_titrant, prot1, type = "1", col = "blue",
      xlim = c(0, 0.015), ylab = "", xlab = "", yaxt = "n", ylim = c(-0.015,
          0.015))
> prot2 <- c()
> for (i in 1:length(ae$pH)) {
     prot2 <- c(prot2, (10^-(ae$pH[[i]]) + ae$HSO4[[i]] + ae$HF[[i]] -</pre>
          ae$HCO3[[i]] - 2 * ae$CO3[[i]] - ae$BOH4[[i]] - ae$OH[[i]]))
+
+ }
> par(new = TRUE)
> plot(ae$delta_mass_titrant, prot2, type = "1", col = "green",
     xlim = c(0, 0.015), ylab = "", xlab = "", yaxt = "n", ylim = c(-0.015)
         0.015))
> abline(v = ae$TA[[1]])
> abline(v = ae$TA[[1]] - ae$SumCO2[[1]])
> abline(h = 0)
```



One can see that the Gran functions actually are not linear. This is due to volume and salinity change effects during the titration. This can be overcome by either employing "modified Gran functions" (see Haraldsson $et\ al.\ 1997$) that correct for the volume changes or by using a titration with a titrant with high concentrations and a large sample volume (Please not that here the value of the concentration of total alkalinity and total carbonate does not equal their total amount and need to be converted with the factor 100/3)



Another proposed method of determining [TA] and $[\sum CO_2]$ is to not only determine the two equivalence points, but to fit the whole titration curve with a theoretical titration curve based on a mechanistic model of what is going on in the solution during the titration (Dickson 1981; DOE 1994; Anderson *et al.* 1999). The function **titration** of **AquaEnv** provides exactly such a theoretical titration curve and the function **TAfit** makes use of this fact to determine [TA] and $[\sum CO_2]$ of a sample by non linear curve fitting.

3.5.2 Determining [TA] and $[\sum CO_2]$ by non linear curve fitting

3.5.2.1 Proof of concept

First, a proof of concept will show that the function TAfit is implemented consistently. Some "data" can be generated with the titration function.

```
> initial_ae <- aquaenv(Tc = 15, S = 35, SumCO2 = 0.002, TA = 0.0022)
> ae <- titration(initial_ae, mass_sample = 0.01, mass_titrant = 0.003, conc_titrant = 0.01, S_titrant = 0.5, steps = 20)
```

Now, the input data for the TAfit routine can be generated: a table with the added mass of the titrant and the resulting free scale pH

```
> titcurve <- cbind(ae$delta_mass_titrant, ae$pH)</pre>
```

Note that For the TAfit all total quantities except SumCO2 (SumNH4, SumH2S, SumH3PO4, SumSiOH4, SumHNO3, SumHNO2, SumBOH3, SumH2SO4, SumHF) need to be known. However, the latter three can be calculated from salinity as it is done in this example.

Thus, we see that TAfit calculates the correct SumCO2 and TA values.

Trying the Lueker, Dickson, and Keeling (2000) (K_CO2 and K_HCO3) and Perez and Fraga (1987b) (K_HF) values

TAfit can also take E (V) values as input variables, so we generate E values using E0=0.4 V and the Nernst equation. (But before that, we calculate a titration with a titrant with the same salinity as seawater such that S does not change during the titration. otherwise we would need to calculate the S profile for the titration extra to use it to convert to the total scale in the following step.) However, to do so we first need to convert our pH curve to the seawater pH scale. According to (DOE 1994, p.7, ch.4, sop.3) and Dickson, Sabine, and Christian (2007), the Nernst equation relates E to the total proton concentration.

Again, TAfit can be executed, this time also calculating E0. Note that the flag verbose=TRUE causes TAfit to show the progress of the fitting procedure in a plot window.

```
> fit2 <- TAfit(initial_ae, Etitcurve, conc_titrant = 0.01, mass_sample = 0.01,
+ Evals = TRUE, verbose = TRUE, seawater_titrant = TRUE)
> fit2

$TA
[1] 0.0022
attr(,"unit")
[1] "mol/kg-soln"

$SumCO2
[1] 0.002
```

```
attr(,"unit")
[1] "mol/kg-soln"
$E0
[1] 0.4
attr(,"unit")
[1] "V"
$sumofsquares
[1] 7.51883e-31
```

attr(,"pH scale")

[1] "free"

Furthermore, TAfit can fit K_CO2 as well, however, one single value for the whole titration curve is fitted, i.e. there is no correction for K_CO2 changes due to changing S due to mixing

```
with the titrant
> fit3 <- TAfit(initial_ae, titcurve, conc_titrant = 0.01, mass_sample = 0.01,
      S_{titrant} = 0.5, K_{CO2fit} = TRUE)
> fit3
$TA
[1] 0.002201253
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.002002897
attr(,"unit")
[1] "mol/kg-soln"
$K_C02
[1] 9.169194e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"
$sumofsquares
[1] 0.0001173601
> initial_ae$K_CO2
[1] 9.385853e-07
attr(,"unit")
[1] "mol/kg-soln"
```

One can see that the fitted value for K_CO2 is not the same as the value in the initial aquaenv object, which is the "correct" value. That is, because during data creation K_CO2 changed along the course of the titration due to changes in salinity. Assuming that the titrant has the same salinity as the sample (and is made up of natural seawater, i.e. containing SumBOH4, SumH2SO4 and SumHF as functions of S), then the "correct" K_CO2 should be fitted. This can be accomplished in TAfit by not giving the argument S_titrant (i.e. assuming the titrant has the same salinity as the sample) and setting the flag seawater_titrant to TRUE

> ae <- titration(initial_ae, mass_sample = 0.01, mass_titrant = 0.003,
+ conc_titrant = 0.01, steps = 20, seawater_titrant = TRUE)</pre>

```
> titcurve <- cbind(ae$delta_mass_titrant, ae$pH)</pre>
> fit4 <- TAfit(initial_ae, titcurve, conc_titrant = 0.01, mass_sample = 0.01,</pre>
      K_CO2fit = TRUE, seawater_titrant = TRUE)
> fit4
$TA
[1] 0.0022
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.002
attr(,"unit")
[1] "mol/kg-soln"
$K_C02
[1] 9.385853e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"
$sumofsquares
[1] 8.16471e-29
Furthermore, TA, SumCO2, K_CO2 and E0 can be fitted at the same time.
> Etitcurve <- cbind(titcurve[, 1], (0.4 - ((Constants$R/10) *
      initial_ae$Tk/Constants$F) * log(10^-titcurve[, 2])))
> fit5 <- TAfit(initial_ae, Etitcurve, conc_titrant = 0.01, mass_sample = 0.01,
      K_CO2fit = TRUE, seawater_titrant = TRUE, Evals = TRUE)
> fit5
$TA
[1] 0.0022
attr(,"unit")
[1] "mol/kg-soln"
```

\$SumCO2

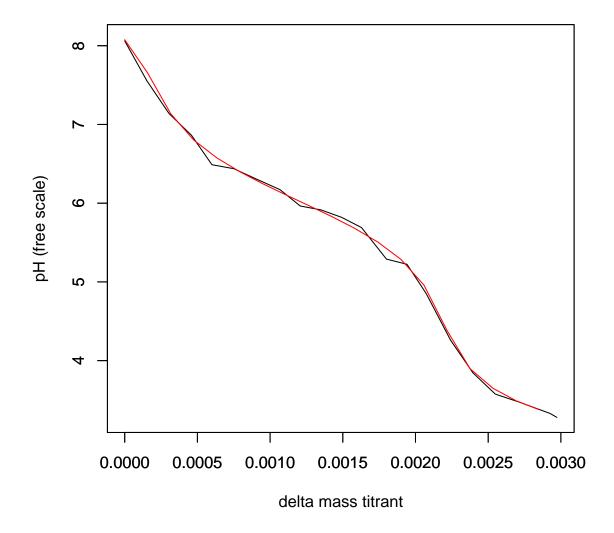
```
[1] 0.002
attr(,"unit")
[1] "mol/kg-soln"
$E0
[1] 0.4043664
attr(,"unit")
[1] "V"
$K_C02
[1] 9.385853e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"
$sumofsquares
[1] 1.700981e-30
Sometimes, the obtained titration curve is not equally spaced on the x axis. TAfit can deal
with such curves if the flag equalspaced is set to FALSE
> neqsptitcurve <- rbind(titcurve[1:9, ], titcurve[11:20, ])</pre>
> fit6 <- TAfit(initial_ae, negsptitcurve, conc_titrant = 0.01,
      mass_sample = 0.01, seawater_titrant = TRUE, equalspaced = FALSE,
      verbose = TRUE, debug = TRUE)
> fit6
$TA
[1] 0.002199682
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.002000279
attr(,"unit")
[1] "mol/kg-soln"
$sumofsquares
[1] 0.002378033
Finally, some "noise" is added to the generated data
> noisetitcurve <- titcurve * rnorm(length(titcurve), mean = 1,
    sd = 0.01)
```

```
> fit7 <- TAfit(initial_ae, noisetitcurve, conc_titrant = 0.01,
+ mass_sample = 0.01, seawater_titrant = TRUE, verbose = TRUE)
> fit7
```

```
$TA
[1] 0.002178266
attr(,"unit")
[1] "mol/kg-soln"
$SumC02
[1] 0.002003503
attr(,"unit")
[1] "mol/kg-soln"
$sumofsquares
[1] 0.06177392
```

The flag verbose=TRUE prompts to show the traject of the fitting procedure in a plot window. However, each new fit is plotted over the first one and Sweave includes only the first plot in each code chunk in the resulting LATEX file. Therefore, we use the flag debug=TRUE to visualize the final fit

```
> ylim = range(noisetitcurve[, 2], calc)
> xlim = range(tit$delta_mass_titrant, noisetitcurve[, 1])
> plot(noisetitcurve[, 1], noisetitcurve[, 2], xlim = xlim, ylim = ylim,
+ type = "l", xlab = "delta mass titrant", ylab = "pH (free scale)")
> par(new = TRUE)
> plot(tit$delta_mass_titrant, calc, xlim = xlim, ylim = ylim,
+ type = "l", col = "red", xlab = "", ylab = "")
```



3.5.2.2 Test with generated data from literature

Dickson (1981) provided a synthetic dataset to test total alkalinity fitting programs. This dataset is included in **AquaEnv** as sample_dickson. Following quantities are given

```
> conc_titrant <- 0.3
```

> mass_sample <- 0.2

> S_titrant <- 14.835

> SumBOH3 <- 0.00042

> SumH2SO4 <- 0.02824

> SumHF <- 7e-05

Note that all concentrations are in mol/kg-solution and the mass of the sample is in kg. Note further that the salinity of the titrant has been calculated from its ionic strength of $0.3 \,$ mol/kg-soln.

In the original dataset as represented in sample_dickson, the mass of titrant is given in g which needs to be converted to kg

```
> sam <- cbind(sample_dickson1981[, 1]/1000, sample_dickson1981[, + 2])  
Then an attempt to recalculate the [TA] and [\sum CO<sub>2</sub>] values given in Dickson (1981) ([TA]=0.00245 mol/kg-soln and [\sum CO<sub>2</sub>] 0.00220 mol/kg-soln) can be done
```

```
> dicksonfit <- TAfit(aquaenv(Tc = 25, S = 35, SumBOH3 = SumBOH3,
+ SumH2SO4 = SumH2SO4, SumHF = SumHF), sam, conc_titrant, mass_sample,
+ S_titrant = S_titrant, debug = TRUE)
> dicksonfit
```

\$TA

```
[1] 0.002464256
attr(,"unit")
[1] "mol/kg-soln"
```

\$SumCO2

```
[1] 0.002191958
attr(,"unit")
[1] "mol/kg-soln"
```

\$sumofsquares

[1] 0.01287255

This shows the fit is not accurate. Why is that so?

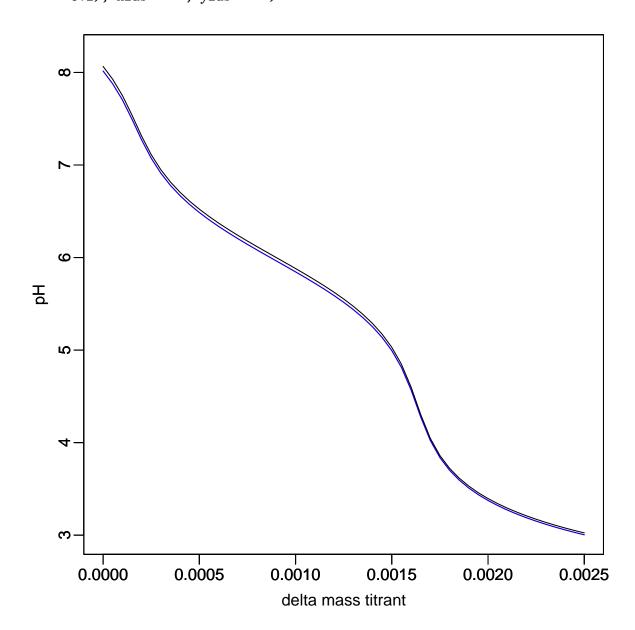
3.5.2.2.1 Does the salinity correction (S_titrant) matter?

Let us calculate a theoretical titration without salinity correction

```
> dicksontitration1 <- titration(aquaenv(Tc = 25, S = 35, SumCO2 = 0.0022,
+ SumBOH3 = SumBOH3, SumH2SO4 = SumH2SO4, SumHF = SumHF, TA = 0.00245),
+ mass_sample = mass_sample, mass_titrant = 0.0025, conc_titrant = conc_titrant,
+ steps = 50, type = "HC1")
and one with salinity correction
> dicksontitration2 <- titration(aquaenv(Tc = 25, S = 35, SumCO2 = 0.0022,
+ SumBOH3 = SumBOH3, SumH2SO4 = SumH2SO4, SumHF = SumHF, TA = 0.00245),
+ mass_sample = mass_sample, mass_titrant = 0.0025, conc_titrant = conc_titrant,
+ S_titrant = S_titrant, steps = 50, type = "HC1")</pre>
```

Now the difference between both curves (in red and blue) and the "Dickson" curve (in black) can be visualized

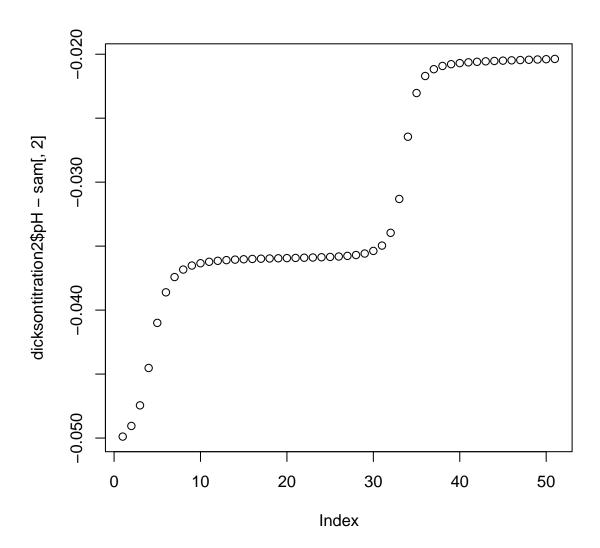
```
> plot(dicksontitration1, xval = dicksontitration1$delta_mass_titrant,
+ what = "pH", xlim = c(0, 0.0025), ylim = c(3, 8.2), newdevice = FALSE,
+ col = "red", xlab = "delta mass titrant")
> par(new = TRUE)
> plot(dicksontitration2, xval = dicksontitration2$delta_mass_titrant,
+ what = "pH", xlim = c(0, 0.0025), ylim = c(3, 8.2), newdevice = FALSE,
+ col = "blue", xlab = "")
> par(new = TRUE)
> plot(sam[, 1], sam[, 2], type = "l", xlim = c(0, 0.0025), ylim = c(3,
+ 8.2), xlab = "", ylab = "")
```



That means, the salinity correction makes no significant difference (the red and the blue curve cannot be discerned), because the relation between the total amount of sample and the added amount of titrant is very large: salinity only drops from 35 to 34.75105.

But there is an offset between the "Dickson" curve and our curve

> plot(dicksontitration2\$pH - sam[, 2])



3.5.2.2.2 Does fitting K_CO2 as well improve the fit?

```
> dicksonfit2 <- TAfit(aquaenv(Tc = 25, S = 35, SumBOH3 = SumBOH3,
+ SumH2SO4 = SumH2SO4, SumHF = SumHF), sam, conc_titrant, mass_sample,</pre>
```

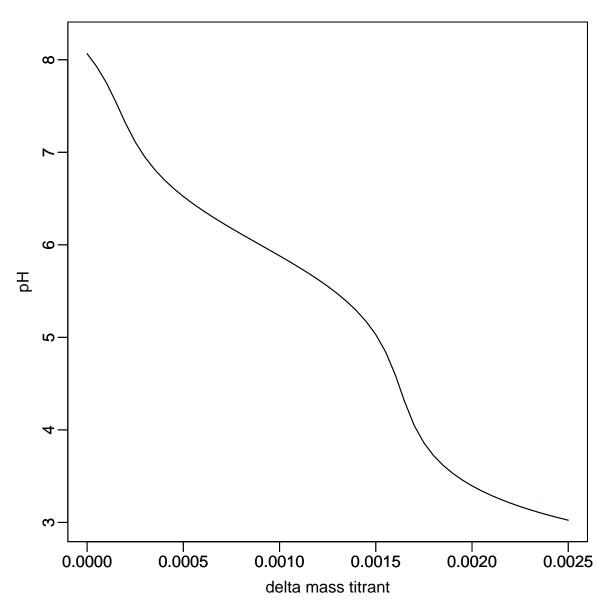
```
S_titrant = S_titrant, debug = TRUE, K_CO2fit = TRUE)
> dicksonfit2
$TA
[1] 0.002458081
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.002194006
attr(,"unit")
[1] "mol/kg-soln"
$K_C02
[1] 1.030960e-06
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"
$sumofsquares
```

Yes it does, but it is not optimal yet.

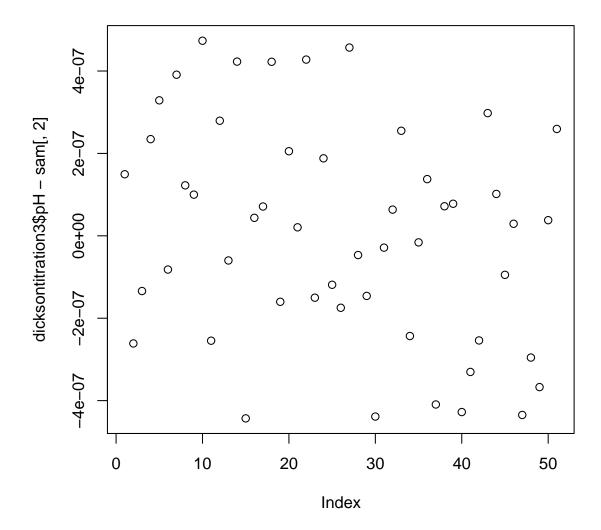
[1] 0.005724457

There still remains one major difference between the calculations carried out in Dickson (1981) and the calculations in **AquaEnv**: Dickson (1981) uses fixed values for the equilibrium constants and does not calculate them as functions of temperature and salinity. Furthermore, the values that are used in Dickson (1981) are not exactly the same as are obtained in **AquaEnv** for the same salinity and temperature.

Let us calculate a theoretical titration curve employing exactly the same equilibrium constant values as used in Dickson (1981) and plot the result together with the "Dickson" curve



Plotting the differences between both curves reveals that they are the same down to 1 umol/kg-soln.



Calculating [TA] and $[\sum CO_2]$ using TAfit and exactly the same equilibrium constant values as used in Dickson (1981)

```
> dicksonfit3 <- TAfit(aquaenv(Tc = 25, S = 35, SumBOH3 = SumBOH3, 
+ SumH2SO4 = SumH2SO4, SumHF = SumHF, k_w = 4.32e-14, k_co2 = 1e-06, 
+ k_co3 = 8.2e-10, k_co3 = 1.78e-09, k_co4 = (1/12.3), k_co4 = (1/408), 
+ sam, conc_titrant, mass_sample, S_titrant = S_titrant, debug = TRUE, 
+ k_w = 4.32e-14, k_co2 = 1e-06, k_co3 = 8.2e-10, k_co3 = 1.78e-09, 
+ k_co4 = (1/12.3), k_co4 = (1/408)) 
> dicksonfit3
```

\$1A [1] 0.00245 attr(,"unit")

[1] 1.017149e-08

```
[1] "mol/kg-soln"
$SumCO2
[1] 0.0022
attr(,"unit")
[1] "mol/kg-soln"
$sumofsquares
[1] 3.279302e-12
reveals that now exactly the same values are calculated as are given in Dickson (1981).
3.5.2.3 Test with data from literature Fitting test data from Dickson et al. (2007)
(SOP3b, page 11).
> mass_sample
                   <- 140.32 / 1000 #kg
> density_titrant <- 1.02393/1000 #kg/ml
> conc_titrant
                   <- 0.10046
                                     #mol/kg-soln
> sample <- cbind(sample_dickson2007[,1] * density_titrant, sample_dickson2007[,2])
> fit <- TAfit(aquaenv(Tc=24.25, S=33.923), sample,
               mass_sample=mass_sample, conc_titrant=conc_titrant,
               datxbegin=sample[,1][[1]], verbose=TRUE, debug=TRUE,
               Evals=TRUE, electrode_polarity="neg", k1k2="lueker",
               khf="perez", seawater_titrant=TRUE,
               SumCO2Zero=TRUE)
> fit
$TA
[1] 0.002259093
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0
attr(,"unit")
[1] "mol/kg-soln"
$E0
[1] -0.3946336
attr(,"unit")
[1] "V"
$sumofsquares
```

One can see that with the values given in Dickson et al. (2007) being 2260.06 umol/kg-soln and 394.401 mV that there is a small difference between the Dickson et al. (2007) values and

the values obtained here. This is in spite of the fact that **AquaEnv** has been checked for consistancy against Dickson *et al.* (2007) in terms of every constant, formula and K value.

We can try to recreate the curve given in Dickson $et\ al.\ (2007)$ with a theoretical titration in **AquaEnv**

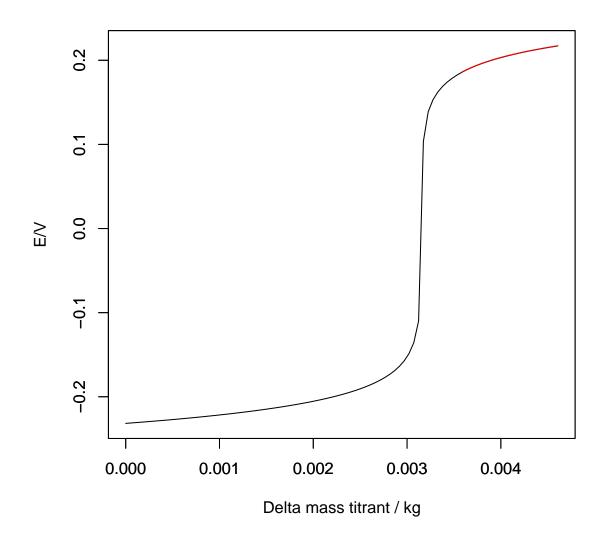
```
> testtit <- titration(aquaenv(Tc = 24.25, S = 33.923, SumCO2 = 0,
+ TA = 0.00226006, SumBOH3 = 0), mass_sample = mass_sample,
+ conc_titrant = conc_titrant, k1k2 = "lueker", khf = "perez",
+ mass_titrant = sample[, 1][[length(sample[, 1])]], steps = 90,
+ seawater_titrant = TRUE)</pre>
```

E is related to total pH (Dickson et al. 2007) via the Nerst equation

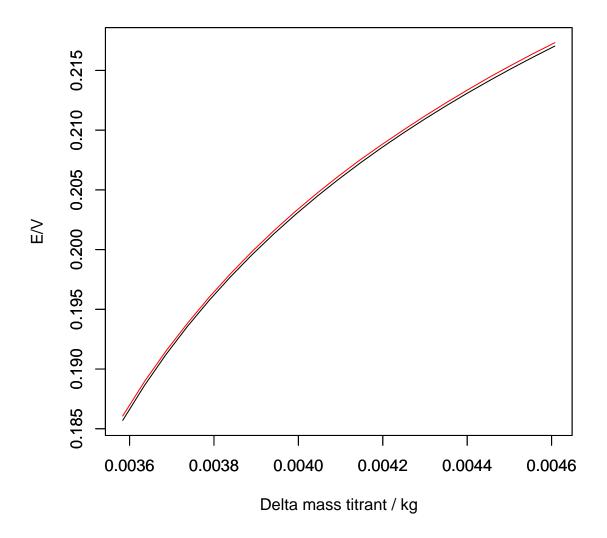
```
> testtotpH <- convert(testtit$pH, "pHscale", "free2tot", Tc = 24.25, 
+ S = 33.923) 
> testE <- -(-0.394401 - (((Constants$R/10) * testtit$Tk[[1]])/Constants$F) * + log(10^{(-testtotpH)}))
```

That lets us visualize the whole curve

```
> ylim <- range(testE, sample[, 2])
> xlim <- range(testtit$delta_mass_titrant, sample[, 1])
> plot(testtit$delta_mass_titrant, testE, type = "l", xlim = xlim,
+ ylim = ylim, xlab = "Delta mass titrant / kg", ylab = "E/V")
> par(new = TRUE)
> plot(sample[, 1], sample[, 2], type = "l", col = "red", xlim = xlim,
+ ylim = ylim, xlab = "", ylab = "")
```

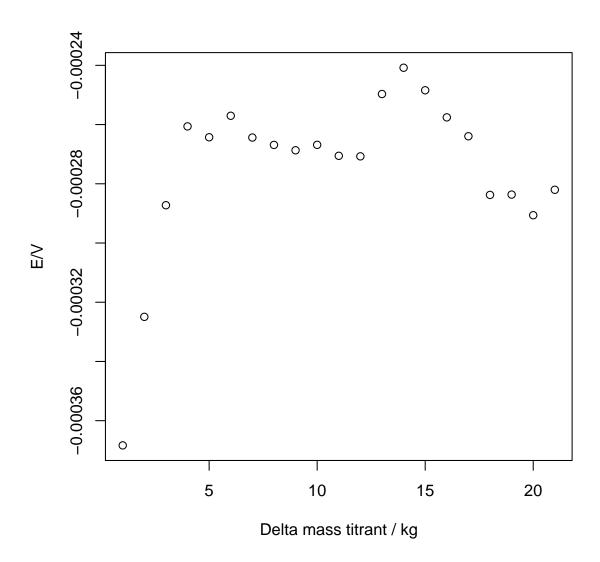


just the part to be fitted



and the difference between the curves

```
> plot(testE[71:91] - sample[, 2], xlab = "Delta mass titrant / kg", + ylab = "E/V")
```



That shows that the curves are slightly different. However, with an offset for E0 of 0.26 mV the fit would be rather good. It is unclear where this offset comes from.

3.5.2.4 Real data

3.5.2.4.1 sample1

The dataset sample1 is a titration curve for TA determination, obtained in the lab of the NIOO-CEME, Yerseke, The Netherlands.

Metadata of the sample which is:

| x-value | volume of titrant added (in ml) with a Metrohm Omega 665 Dosimat |
|-----------------------|---|
| y-value | pH measured on the NBS scale with a Metrohm Omega 714 pH Meter (Buffer pH 4: citric acid, sodium hydroxide, chlor hydrogen; Buffer pH 7: di-sodium hydrogen phosphate, potassium di-hydrogen phosphate) |
| Tc | 23.5 °C |
| S | 34 |
| volume of the sample | 10 ml |
| titrant type | HCl |
| titrant concrntration | 0.01 N = 0.01 M |

We assume that the salinity of the titrant solution is 0.5 (the salinity of a solution with a ionic strength of 0.01 mol/kg-solution (0.01N) according to: I=(19.924~S)/(1000~-1.005~S)). Furthermore, the the x values have to be converted to kg titrant, assuming the density of the titrant solution is aequivalent to the density of seawater with S=0.5, and last but not least 1 ml = 0.000001 m3.

```
> sample <- cbind((sample1[, 1] * 1e-06 * aquaenv(Tc = 23.5, S = 0.5)$density),
+ sample1[, 2])</pre>
```

We reduce the amount of datapoints: for runtime reasons

```
> sample <- sample[seq(1, length(sample[, 1]), 3), ]</pre>
```

We convert the concentration of the titrant from molarity (mol/l) to mol/kg-soln

```
> conc_titrant <- convert(0.01, "conc", "molar2molin", Tc = 23.5, S = 0.5)
```

We calculate the mass of the sample in kg (and 10 ml = 0.00001 m3)

```
> mass_sample <- 1e-05 * aquaenv(Tc = 23.5, S = 34)$density
```

No the fitting procedure can be performed. Note that the datapoints are not equally spaced and the pH scale of the measured pH values is the NBS scale.

```
> sample1_fit <- TAfit(aquaenv(Tc = 23.5, S = 34), sample, conc_titrant,
+ mass_sample, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
+ debug = TRUE, verbose = TRUE)
> sample1_fit

$TA
[1] 0.005729476
attr(,"unit")
```

```
[1] "mol/kg-soln"

$SumCO2
[1] 0.0042921
attr(,"unit")
[1] "mol/kg-soln"

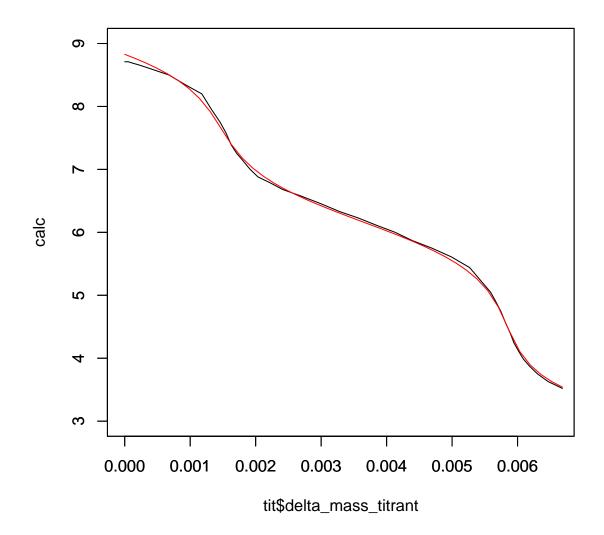
$sumofsquares
[1] 0.1238645
```

Note that AquaEnv uses the Davies equation to calculate the activity coefficient for the proton upon conversion from/to NBS scale. The Davies equation is only valid up to a ionic strength (I) of 0.5 (Zeebe and Wolf-Gladrow 2001), that is around S=24.48. To accurately determine [TA] for samples with higher salinity using the function TAfit, one needs to provide either pH values on the free, total or seawater scale or E (V) values.

```
The [TA] value obtained with the Peene-Method was 0.0058740 mol/l, i.e.
> convert(0.005874, "conc", "molar2molin", Tc = 23.5, S = 34)
[1] 0.005741753
attr(,"unit")
[1] "mol/kg-soln"
the difference in umol/kg-soln is
> (0.005724596 - sample1_fit$TA) * 1e+06
[1] -4.880155
attr(,"unit")
[1] "mol/kg-soln"
and the SumCO2 value was 0.0522 mgC/l, i.e.
> convert((0.0522/12), "conc", "molar2molin", Tc = 23.5, S = 34)
[1] 0.004252064
attr(,"unit")
[1] "mol/kg-soln"
the difference in umol/kg-soln is
> (0.004239359 - sample1_fit$SumCO2) * 1e+06
[1] -52.74089
attr(,"unit")
[1] "mol/kg-soln"
```

Note the deviation of the measured curve from the simulated curve, especially at high pH's (and high salinities) as the titration proceeds, salinity drops due to dilution with non saline titrant (HCl solution) and the agreement between the curves gets better

```
> plot(sample[, 1], sample[, 2], xlim = c(0, 0.0066), ylim = c(3,
+ 9), type = "l", xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit$delta_mass_titrant, calc, xlim = c(0, 0.0066), ylim = c(3,
+ 9), type = "l", col = "red")
```



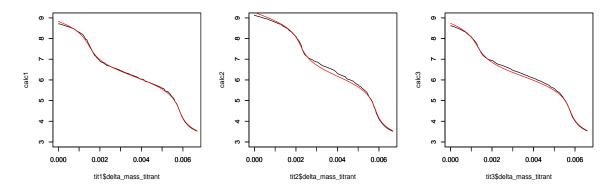
this shows that NBS pH values with initial salinities of more than 25 are not excellent input values for TAfit.

3.5.2.4.2 sample 1 to sample 3

The datasets sample2 and sample3 are similar to sample1 and their metadata can be found in their helpfiles.

Preparing the input for TAfit

```
> sam1 < -cbind((sample1[, 1] * 1e-06 * aquaenv(Tc = 23.5, S = 0.5) density),
     sample1[, 2])
> sam2 <- cbind((sample2[, 1] * 1e-06 * aquaenv(Tc = 22.9, S = 0.5)$density),
      sample2[, 2])
> sam3 < - cbind((sample3[, 1] * 1e-06 * aquaenv(Tc = 23.3, S = 0.5)$density),
      sample3[, 2])
> conc_titrant1 <- convert(0.01, "conc", "molar2molin", Tc = 23.5,</pre>
      S = 0.5)
> conc_titrant2 <- convert(0.01, "conc", "molar2molin", Tc = 22.9,</pre>
      S = 0.5)
> conc_titrant3 <- convert(0.01, "conc", "molar2molin", Tc = 23.3,</pre>
      S = 0.5
> mass_sample1 <- 1e-05 * aquaenv(Tc = 23.5, S = 34)$density
> mass_sample2 <- 1e-05 * aquaenv(Tc = 22.9, S = 33.4)$density
> mass_sample3 <- 1e-05 * aquaenv(Tc = 23.3, S = 33.8)$density
Performing the fitting
> sample1_fit <- TAfit(aquaenv(Tc = 23.5, S = 34), sam1, conc_titrant1,
      mass_sample1, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, verbose = TRUE)
> tit1 <- tit
> calc1 <- calc
> sample2_fit <- TAfit(aquaenv(Tc = 22.9, S = 33.4), sam2, conc_titrant2,
      mass_sample2, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, verbose = TRUE)
> tit2 <- tit
> calc2 <- calc
> sample3_fit <- TAfit(aquaenv(Tc = 23.3, S = 33.8), sam3, conc_titrant3,
      mass_sample3, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, verbose = TRUE)
> tit3 <- tit
> calc3 <- calc
Visualizing the final fits
> par(mfrow = c(1, 3))
> plot(sam1[, 1], sam1[, 2], xlim = c(0, 0.0066), ylim = c(3, 9),
     type = "1", xlab = "", ylab = "")
```



Comparing the resulting values to the values obtained with the graphical method (J. Peene)

[1] 0.005725735 0.005723115

```
> c(TA3peene, sample3_fit$TA)

[1] 0.005664072 0.005658898

> c(SumCO21peene, sample1_fit$SumCO2)

[1] 0.004252064 0.004287438

> c(SumCO22peene, sample2_fit$SumCO2)

[1] 0.003543385 0.003551793

> c(SumCO23peene, sample3_fit$SumCO2)

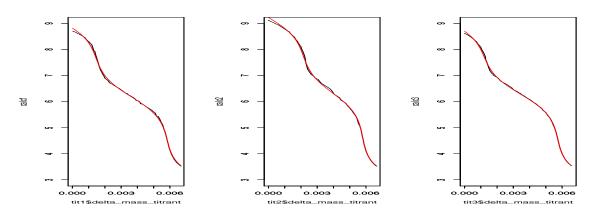
[1] 0.004379538 0.004410422
```

3.5.2.4.3 sample1 to sample3: fitting K_CO2 as well

Performing the fitting. Note that, with fitting K_CO2 as well, it might be necessary to choose a small initial stepwidth-factor and/or bring the parameter TASumCO2guess close to expected values for [TA] and $[\sum CO_2]$, otherwise the optimization routine might create an error, e.g. by "trying" negative K_CO2 values.

```
> sample1_fit <- TAfit(aquaenv(Tc = 23.5, S = 34), sam1, conc_titrant1,
      mass_sample1, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, K_CO2fit = TRUE, nlscontrol = nls.lm.control(factor = 0.1),
      verbose = TRUE)
> tit1 <- tit
> calc1 <- calc
> sample2_fit <- TAfit(aquaenv(Tc = 22.9, S = 33.4), sam2, conc_titrant2,
      mass_sample2, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, K_CO2fit = TRUE, TASumCO2guess = 0.006, nlscontrol = nls.lm.control(fa
      verbose = TRUE)
> tit2 <- tit
> calc2 <- calc
> sample3_fit <- TAfit(aquaenv(Tc = 23.3, S = 33.8), sam3, conc_titrant3,
      mass_sample3, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, K_CO2fit = TRUE, nlscontrol = nls.lm.control(factor = 0.1),
      verbose = TRUE)
> tit3 <- tit
> calc3 <- calc
Visualizing the resulting fits
> par(mfrow = c(1, 3))
> plot(sam1[, 1], sam1[, 2], xlim = c(0, 0.0066), ylim = c(3, 9),
```

```
+ type = "l", xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit1$delta_mass_titrant, calc1, xlim = c(0, 0.0066), ylim = c(3, 4), type = "l", col = "red")
> plot(sam2[, 1], sam2[, 2], xlim = c(0, 0.0066), ylim = c(3, 9), type = "l", xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit2$delta_mass_titrant, calc2, xlim = c(0, 0.0066), ylim = c(3, 4), type = "l", col = "red")
> plot(sam3[, 1], sam3[, 2], xlim = c(0, 0.0066), ylim = c(3, 9), type = "l", xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit3$delta_mass_titrant, calc3, xlim = c(0, 0.0066), ylim = c(3, 4), type = "l", col = "red")
```



Comparing the resulting values to the values obtained with the graphical method (J. Peene)

```
> TA1peene <- convert(0.005874, "conc", "molar2molin", Tc = 23.5,
+ S = 34)
> TA2peene <- convert(0.005856, "conc", "molar2molin", Tc = 22.9,
+ S = 33.4)
> TA3peene <- convert(0.005794, "conc", "molar2molin", Tc = 23.3,
+ S = 33.8)
> SumC021peene <- convert(0.0522/12, "conc", "molar2molin", Tc = 23.5,
+ S = 34)
> SumC022peene <- convert(0.043488/12, "conc", "molar2molin", Tc = 22.9,
+ S = 33.4)
> SumC023peene <- convert(0.05376/12, "conc", "molar2molin", Tc = 23.3,
+ S = 33.8)
> c(TA1peene, sample1_fit$TA)
```

[1] 0.005741753 0.005725985

> c(TA2peene, sample2_fit\$TA)

```
[1] 0.005725735 0.005682684
> c(TA3peene, sample3_fit$TA)
[1] 0.005664072 0.005626975
> c(SumCO21peene, sample1_fit$SumCO2)
[1] 0.004252064 0.004298992
> c(SumCO22peene, sample2_fit$SumCO2)
[1] 0.003543385 0.003577105
> c(SumCO23peene, sample3_fit$SumCO2)
[1] 0.004379538 0.004445158
```

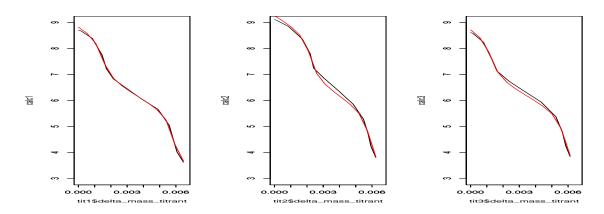
3.5.2.4.4 sample1 to sample3: fitting ≈ 10 datapoints

To test if the fit wil be significantly worse if the amount of datapoints in the titration curve is reduced, we reduce the amount of datapoints in the curve to around 10.

```
> sam1 \leftarrow cbind((sample1[, 1] * 1e-06 * aquaenv(Tc = 23.5, S = 0.5)$density),
      sample1[, 2])[seq(1, length(sample1[, 1]), 10), ]
> sam2 < -cbind((sample2[, 1] * 1e-06 * aquaenv(Tc = 22.9, S = 0.5)$density),
      sample2[, 2])[seq(1, length(sample2[, 1]), 10), ]
> sam3 <- cbind((sample3[, 1] * 1e-06 * aquaenv(Tc = 23.3, S = 0.5)$density),
      sample3[, 2])[seq(1, length(sample3[, 1]), 10), ]
Performing the fitting
> sample1_fit <- TAfit(aquaenv(Tc = 23.5, S = 34), sam1, conc_titrant1,
      mass_sample1, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, verbose = TRUE)
> tit1 <- tit
> calc1 <- calc
> sample2_fit <- TAfit(aquaenv(Tc = 22.9, S = 33.4), sam2, conc_titrant2,
      mass_sample2, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, verbose = TRUE)
> tit2 <- tit
> calc2 <- calc
> sample3_fit <- TAfit(aquaenv(Tc = 23.3, S = 33.8), sam3, conc_titrant3,
      mass_sample3, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, verbose = TRUE)
> tit3 <- tit
> calc3 <- calc
```

Visualizing the resulting fit

```
> par(mfrow = c(1, 3))
> plot(sam1[, 1], sam1[, 2], xlim = c(0, 0.0066), ylim = c(3, 9),
      type = "1", xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit1$delta_mass_titrant, calc1, xlim = c(0, 0.0066), ylim = c(3,
      9), type = "1", col = "red")
> plot(sam2[, 1], sam2[, 2], xlim = c(0, 0.0066), ylim = c(3, 9),
      type = "1", xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit2$delta_mass_titrant, calc2, xlim = c(0, 0.0066), ylim = c(3,
      9), type = "1", col = "red")
> plot(sam3[, 1], sam3[, 2], xlim = c(0, 0.0066), ylim = c(3, 9),
      type = "1", xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit3$delta_mass_titrant, calc3, xlim = c(0, 0.0066), ylim = c(3, 0.0066))
      9), type = "1", col = "red")
```



Comparing the resulting values to the values obtained with the graphical method (J. Peene)

```
> TA1peene <- convert(0.005874, "conc", "molar2molin", Tc = 23.5,
+ S = 34)
> TA2peene <- convert(0.005856, "conc", "molar2molin", Tc = 22.9,
+ S = 33.4)
> TA3peene <- convert(0.005794, "conc", "molar2molin", Tc = 23.3,
+ S = 33.8)
> SumC021peene <- convert(0.0522/12, "conc", "molar2molin", Tc = 23.5,
+ S = 34)
> SumC022peene <- convert(0.043488/12, "conc", "molar2molin", Tc = 22.9,
+ S = 33.4)
> SumC023peene <- convert(0.05376/12, "conc", "molar2molin", Tc = 23.3,
+ S = 33.8)
> c(TA1peene, sample1_fit$TA)
```

```
[1] 0.005741753 0.005739229
> c(TA2peene, sample2_fit$TA)
[1] 0.005725735 0.005719017
> c(TA3peene, sample3_fit$TA)
[1] 0.005664072 0.005665640
> c(SumCO21peene, sample1_fit$SumCO2)
[1] 0.004252064 0.004310279
> c(SumCO22peene, sample2_fit$SumCO2)
[1] 0.003543385 0.003563105
> c(SumCO23peene, sample3_fit$SumCO2)
```

[1] 0.004379538 0.004443528

That means using just ≈ 10 datapoints is feasible, as the fits do not get worse significantly

3.5.2.4.5 sample4_1 to sample4_2: standard seawater

sample4_1 to sample4_4 are titration curves of reference material (standard seawater) "Batch 82" (A. Dickson) for TA determination, obtained in the lab of the NIOO-CEME, Yerseke, The Netherlands. Metadata can again be obtained from the helpfiles of the datasets.

We assume the salinity of the titrant solution is 0.5 (the salinity of a solution with a ionic strength of 0.01 mol/kg-solution (0.01N) according to: I = (19.924 S) / (1000 - 1.005 S). We convert the x values to kg titrant, and assume the density of the titrant solution is aequivalent to the density of seawater with S=0.5. Finally, 1 ml = 0.000001 m³.

```
> sam4_1 \leftarrow cbind((sample4_1[, 1] * 1e-06 * aquaenv(Tc = 21, S = 0.5)$density), + sample4_1[, 2])
```

We convert the concentration of the titrant from molarity (mol/l) to mol/kg-soln

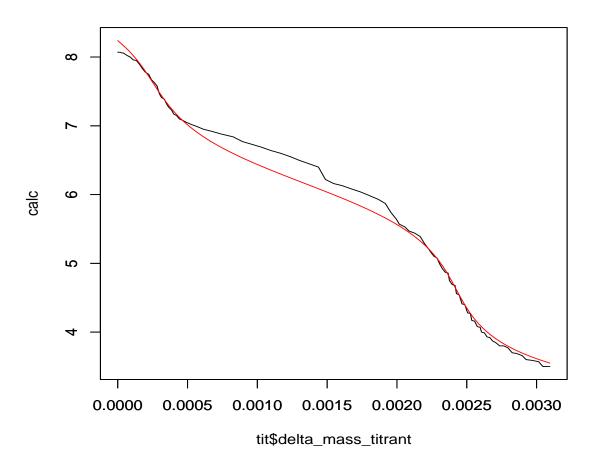
```
> conc_titrant <- convert(0.01, "conc", "molar2molin", Tc = 21, + S = 0.5)
```

We calculate the mass of the sample in kg (10 ml = 0.00001 m3)

```
> mass_sample <- 1e-05 * aquaenv(Tc = 21, S = 35.356)$density
```

We perform the fitting procedure: note that the datapoints are NOT equally spaced and the pH scale of the measured pH values is the NBS scale, (debug=TRUE toggles the debug mode: the last simulated titration tit and the converted pH profile calc are made global variables for investigation and plotting)

The fit can be visualized



(Note that the fit is very good on the steep slopes and not so good inbetween. This is because the sample titration curve is not equally spaced. There are smaller titration steps in the steep parts, resulting in more points there. Since **AquaEnv** minimizes the sum of squares of all points without weighting the points according to their density in a certain region of the curve, the regions with higher point density have implicitly a higher weight during the fitting procedure.)

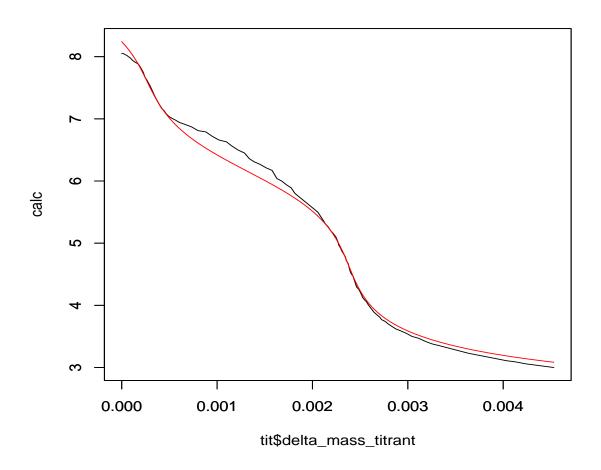
and the results can be compared to results obtained with the graphical method (J.Peene)

```
> peeneTA <- convert(0.002418, "conc", "molar2molin", Tc = 21,
+ S = 35.356)
> peeneSumC02 <- convert(0.025872/12, "conc", "molar2molin", Tc = 21,
+ S = 35.356)
> c(peeneTA, sample4_1_fit$TA)

[1] 0.002359562 0.002374952
> c(peeneSumC02, sample4_1_fit$SumC02)

[1] 0.002103894 0.002128026
```

```
sample4_2 is a replica titration of sample4_1.
Preparations
> sam4_2 <- cbind((sample4_2[, 1] * 1e-06 * aquaenv(Tc = 22.8,
      S = 0.5)$density), sample4_2[, 2])
> conc_titrant <- convert(0.01, "conc", "molar2molin", Tc = 22.8,</pre>
     S = 0.5
> mass_sample <- 1e-05 * aquaenv(Tc = 22.8, S = 35.356)$density
Fitting
> sample4_2_fit \leftarrow TAfit(aquaenv(Tc = 22.8, S = 34.356, SumH3P04 = 3e-07,
      SumSiOH4 = 2e-06, SumHNO2 = 1e-08, SumHNO3 = 8.3e-07), sam4_2,
      conc_titrant, mass_sample, S_titrant = 0.5, equalspaced = FALSE,
      pHscale = "nbs", debug = TRUE)
> sample4_2_fit
$TA
[1] 0.002338329
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.002079495
attr(,"unit")
[1] "mol/kg-soln"
$sumofsquares
[1] 1.311792
Visualisation
> xlim <- range(sam4_2[, 1], tit$delta_mass_titrant)</pre>
> ylim <- range(sam4_2[, 2], calc)
> plot(sam4_2[, 1], sam4_2[, 2], xlim = xlim, ylim = ylim, type = "l",
      xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit$delta_mass_titrant, calc, xlim = xlim, ylim = ylim,
      type = "1", col = "red")
```



Comparison to graphical method

```
> peeneTA <- convert(0.00239, "conc", "molar2molin", Tc = 22.8,
+    S = 35.356)
> peeneSumC02 <- convert(0.02532/12, "conc", "molar2molin", Tc = 22.8,
+    S = 35.356)
> peeneCalcSumC02 <- convert(0.02625722/12, "conc", "molar2molin",
+    Tc = 22.8, S = 35.356)
> c(peeneTA, sample4_2_fit$TA)
```

[1] 0.002333386 0.002338329

> c(peeneSumCO2, peeneCalcSumCO2, sample4_2_fit\$SumCO2)

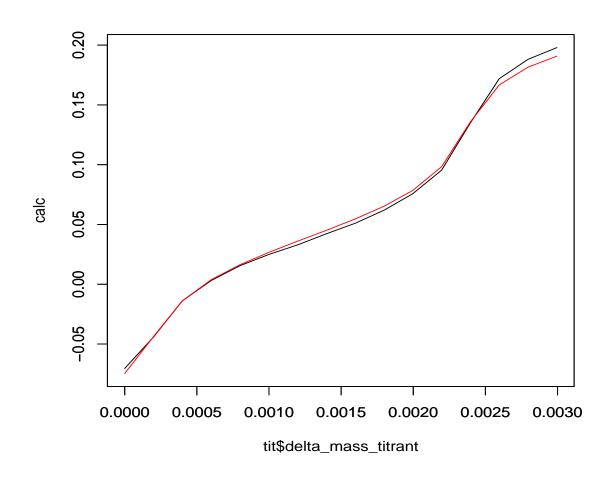
[1] 0.002060018 0.002136270 0.002079495

3.5.2.4.6 sample4_3 to sample4_4: E curves

The datasets sample4_3 and sample4_4 have been obtained by manually titrating samples of the standard seawater mentioned above and recording the "raw" potential (E) values coming from the electrode.

Preparations

```
> sam4_3 < - cbind((sample4_3[, 1] * 1e-06 * aquaenv(Tc = 21, S = 0.5)$density),
      sample4_3[, 2]/1000)
> conc_titrant <- convert(0.01, "conc", "molar2molin", Tc = 21,</pre>
      S = 0.5)
> mass_sample <- 1e-05 * aquaenv(Tc = 21, S = 0)$density
Fitting
> sample4_3_fit \leftarrow TAfit(aquaenv(Tc = 21, S = 34.356, SumH3P04 = 3e-07,
      SumSiOH4 = 2e-06, SumHNO2 = 1e-08, SumHNO3 = 8.3e-07), sam4_3,
      conc_titrant, mass_sample, S_titrant = 0.5, Evals = TRUE,
      verbose = TRUE, electrode_polarity = "neg", debug = TRUE)
> sample4_3_fit
$TA
[1] 0.002383378
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.002192699
attr(,"unit")
[1] "mol/kg-soln"
$E0
[1] -0.3847173
attr(,"unit")
[1] "V"
$sumofsquares
[1] 0.0002022453
Visualizing
> xlim <- range(sam4_3[, 1], tit$delta_mass_titrant)</pre>
> ylim <- range(sam4_3[, 2], calc)
> plot(sam4_3[, 1], sam4_3[, 2], xlim = xlim, ylim = ylim, type = "1",
      xlab = "", ylab = "")
> par(new = TRUE)
> plot(tit$delta_mass_titrant, calc, xlim = xlim, ylim = ylim,
      type = "1", col = "red")
```



Fitting $K_{-}CO2$ as well

```
$E0

[1] -0.3921561

attr(,"unit")

[1] "V"

$K_C02

[1] 7.042922e-07

attr(,"unit")

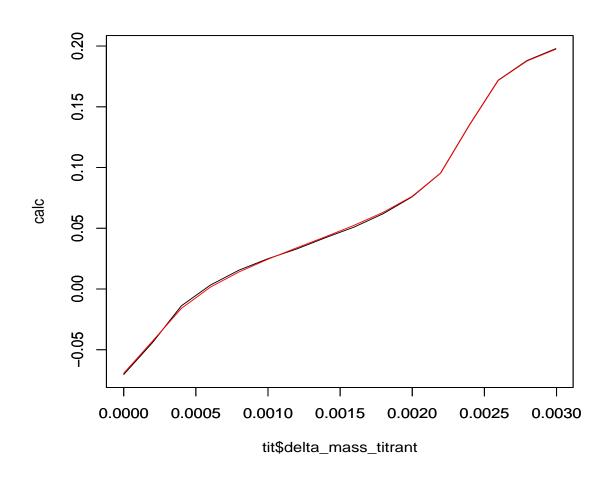
[1] "mol/kg-soln"

attr(,"pH scale")

[1] "free"

$sumofsquares

[1] 1.766105e-05
```



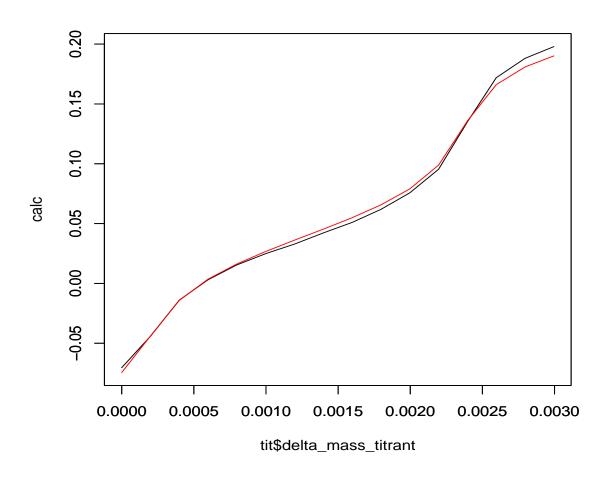
Fitting with the K values from Dickson (1981)

[1] 0.00218946

```
attr(,"unit")
[1] "mol/kg-soln"

$E0
[1] -0.3850324
attr(,"unit")
[1] "V"

$sumofsquares
[1] 0.0002403367
```



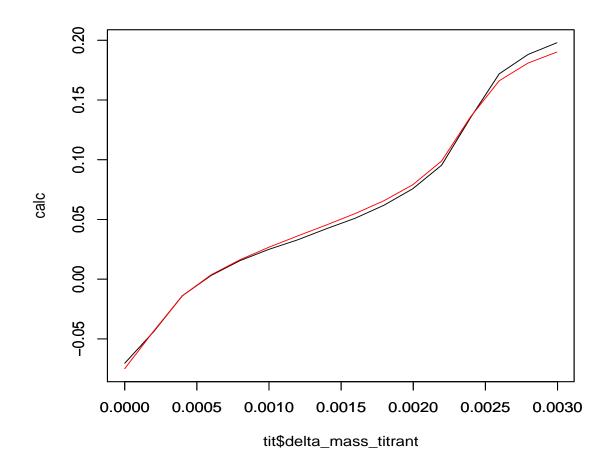
Fitting with titrant salinity assumed equal to sample salinity

```
$E0
[1] -0.3840105
attr(,"unit")
[1] "V"
```

\$sumofsquares

[1] 0.000245281

Visualizing



Fitting with titrant concentration assumed to be exactly 0.01 mol/kg-soln

```
> sample4_3_{fit_4} \leftarrow TAfit(aquaenv(Tc = 21, S = 34.356, SumH3P04 = 3e-07, SumSiOH4 = 2e-06, SumHNO2 = 1e-08, SumHNO3 = 8.3e-07), sam4_3, conc_titrant = 0.01, mass_sample, S_titrant = 0.5, Evals = TRUE, verbose = TRUE, electrode_polarity = "neg", debug = TRUE) > <math>sample4_3_{fit_4}
```

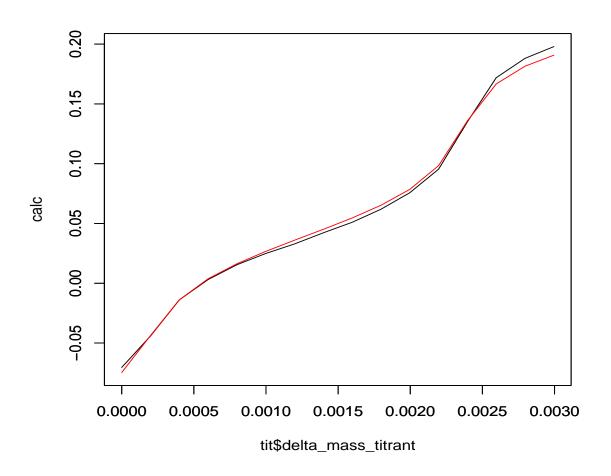
```
$TA
[1] 0.002379434
attr(,"unit")
[1] "mol/kg-soln"

$SumCO2
[1] 0.002188906
attr(,"unit")
[1] "mol/kg-soln"

$EO
[1] -0.3847282
attr(,"unit")
[1] "V"

$sumofsquares
```

[1] 0.0002036067



Preparations for sample4_4

```
> sam4_4 \leftarrow cbind((sample4_4[, 1] * 1e-06 * aquaenv(Tc = 21, S = 0.5)$density), + sample4_4[, 2]/1000) > conc_titrant \leftarrow convert(0.01, "conc", "molar2molin", Tc = 21, + S = 0.5) > mass_sample \leftarrow 1e-05 * aquaenv(Tc = 21, S = 0)$density
```

Fitting

```
> sample4_4_fit <- TAfit(aquaenv(Tc = 22.4, S = 34.356, SumH3P04 = 3e-07,
+ SumSiOH4 = 2e-06, SumHNO2 = 1e-08, SumHNO3 = 8.3e-07), sam4_4,
+ conc_titrant, mass_sample, S_titrant = 0.5, Evals = TRUE,
+ verbose = TRUE, electrode_polarity = "neg", debug = TRUE)
> sample4_4_fit
```

\$TA

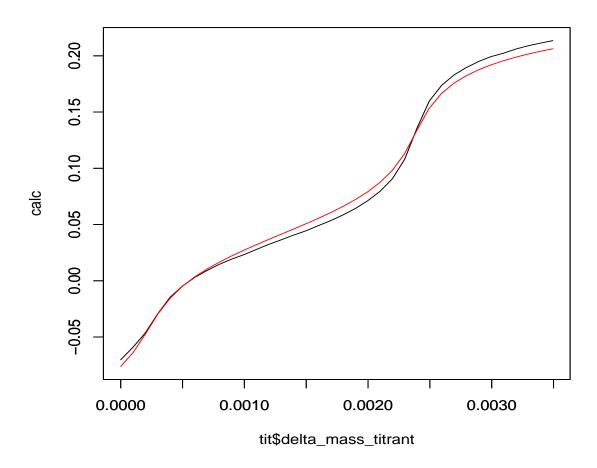
[1] 0.002400287

```
attr(,"unit")
[1] "mol/kg-soln"

$SumCO2
[1] 0.002179599
attr(,"unit")
[1] "mol/kg-soln"

$EO
[1] -0.3875244
attr(,"unit")
[1] "V"

$sumofsquares
[1] 0.001209710
```



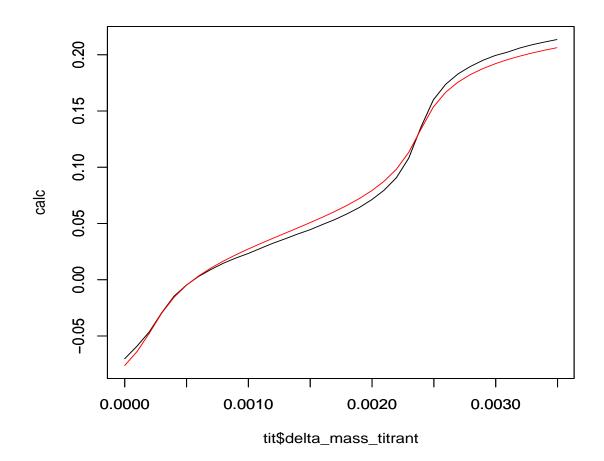
Fitting with concentrations of summed quantities calculated from salinity

```
attr(,"unit")
[1] "V"
```

\$sumofsquares

[1] 0.001209842

Visualizing



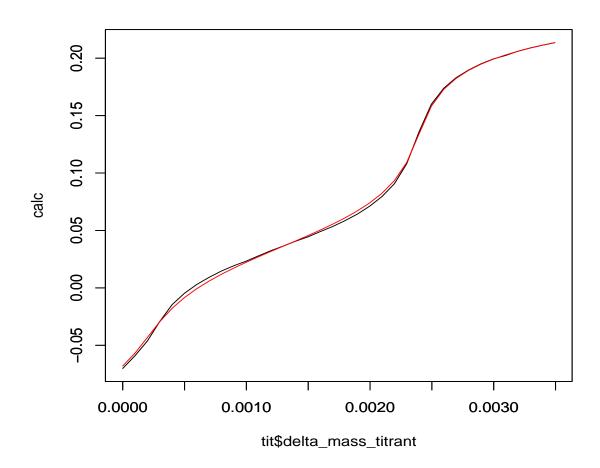
Fitting K_CO2 as well

```
> sample4_4_fit_2 <- TAfit(aquaenv(Tc = 22.4, S = 34.356, SumH3P04 = 3e-07, SumSiOH4 = 2e-06, SumHNO2 = 1e-08, SumHNO3 = 8.3e-07), sam4_4,
```

```
conc_titrant, mass_sample, S_titrant = 0.5, Evals = TRUE,
      verbose = TRUE, electrode_polarity = "neg", K_CO2fit = TRUE,
      debug = TRUE)
> sample4_4_fit_2
$TA
[1] 0.002406076
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.002204483
attr(,"unit")
[1] "mol/kg-soln"
$E0
[1] -0.3949774
attr(,"unit")
[1] "V"
$K_C02
[1] 6.148002e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"
$sumofsquares
```

Visualizing

[1] 0.0001361478



 ${\bf 3.5.2.4.7} \quad {\bf Some \ rudimentary \ statistics \ on \ sample 4_1 \ to \ sample 4_4}$

(Beware, these statistics are just "back of the envelope" calculations, since not all the measurements are independent and for different quantities not the same number of measurements have been carried out.)

The graphical method

```
> peeneTA <- c(2353.904, 2326.871)
> peeneDIC <- c(2098.849, 2054.267, 2130.305)
> paste(mean(peeneTA), "+/-", sd(peeneTA))

[1] "2340.3875 +/- 19.1152176158159"

> paste(mean(peeneDIC), "+/-", sd(peeneDIC))

[1] "2094.473666666667 +/- 38.2073555396515"
```

```
pH curve fitting
```

```
> pHTA <- c(sample4_1_fit$TA, sample4_2_fit$TA) * 1e+06
> pHDIC <- c(sample4_1_fit$SumCO2, sample4_2_fit$SumCO2) * 1e+06
> paste(mean(pHTA), "+/-", sd(pHTA))
[1] "2356.64068645046 +/- 25.8961384594687"
> paste(mean(pHDIC), "+/-", sd(pHDIC))
[1] "2103.76061832307 +/- 34.3162691977244"
E curve fitting
> ETA <- c(sample4_3_fit$TA, sample4_3_fit_1$TA, sample4_3_fit_2$TA,
      sample 4\_3\_fit\_3\$TA, \ sample 4\_3\_fit\_4\$TA, \ sample 4\_4\_fit\$TA,
      sample4_4_fit_1$TA, sample4_4_fit_2$TA) * 1e+06
> EDIC <- c(sample4_3_fit$SumCO2, sample4_3_fit_1$SumCO2, sample4_3_fit_2$SumCO2,
      sample4_3_fit_3$SumCO2, sample4_3_fit_4$SumCO2, sample4_4_fit$SumCO2,
      sample 4\_4\_fit\_1\$SumCO2, \ sample 4\_4\_fit\_2\$SumCO2) \ * \ 1e+06
> paste(mean(ETA), "+/-", sd(ETA))
[1] "2392.25389187223 +/- 10.8961998604983"
> paste(mean(EDIC), "+/-", sd(EDIC))
[1] "2191.55246039690 +/- 9.03451422906339"
pH and E curve fitting combined
> totalTA <- c(pHTA, ETA)
> totalDIC <- c(pHDIC, EDIC)</pre>
> paste(mean(totalTA), "+/-", sd(totalTA))
[1] "2385.13125078787 +/- 19.8073522476785"
> paste(mean(totalDIC), "+/-", sd(totalDIC))
[1] "2173.99409198213 +/- 39.5542040449721"
```

3.5.2.4.8 sample5_1 to sample5_2: freshwater samples

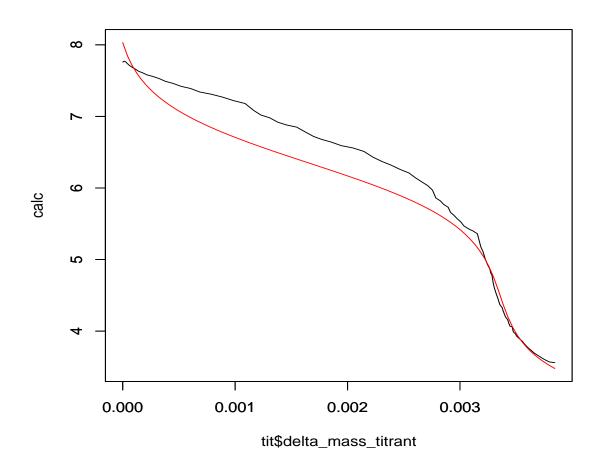
sample5_1 to sample5_3 are titration curves of freshwater samples obtained in the lab of the NIOO-CEME, Yerseke, The Netherlands. sample5_1 and sample5_2 are pH curves.

Preparations

```
> sam5_1 < -cbind((sample5_1[, 1] * 1e-06 * aquaenv(Tc = 21, S = 0.5)$density),
      sample5_1[, 2])
> conc_titrant <- convert(0.01, "conc", "molar2molin", Tc = 21,</pre>
      S = 0.5)
> mass_sample <- 1e-05 * aquaenv(Tc = 17.5, S = 0)$density
Fitting
> sample5_1_fit <- TAfit(aquaenv(Tc = 17.5, S = 0), sam5_1, conc_titrant,
      mass_sample, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
      debug = TRUE, verbose = TRUE)
> sample5_1_fit
$TA
[1] 0.00337084
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.003434296
attr(,"unit")
[1] "mol/kg-soln"
$sumofsquares
[1] 6.007394
Visualizing
> xlim <- range(sam5_1[, 1], tit$delta_mass_titrant)</pre>
> ylim <- range(sam5_1[, 2], calc)</pre>
> plot(sam5_1[, 1], sam5_1[, 2], xlim = xlim, ylim = ylim, type = "l",
      xlab = "", ylab = "")
> par(new = TRUE)
```

> plot(tit\$delta_mass_titrant, calc, xlim = xlim, ylim = ylim,

type = "1", col = "red")



Comparing to graphical method

```
> peeneTA <- convert(0.003314, "conc", "molar2molin", Tc = 17.5, + S = 0) > c(peeneTA, sample5_1_fit$TA)
```

[1] 0.003318352 0.003370840

Preparations for sample5_2

```
> sam5_2 < -cbind((sample5_2[, 1] * 1e-06 * aquaenv(Tc = 21, S = 0.5)$density), + sample5_2[, 2])
> conc_titrant < -convert(0.01, "conc", "molar2molin", Tc = 21, + S = 0.5)
> mass_sample < -1e-05 * aquaenv(Tc = 18.3, S = 0)$density
Fitting
```

```
> sample5_2_fit <- TAfit(aquaenv(Tc = 18.3, S = 0), sam5_2, conc_titrant,
+ mass_sample, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",</pre>
```

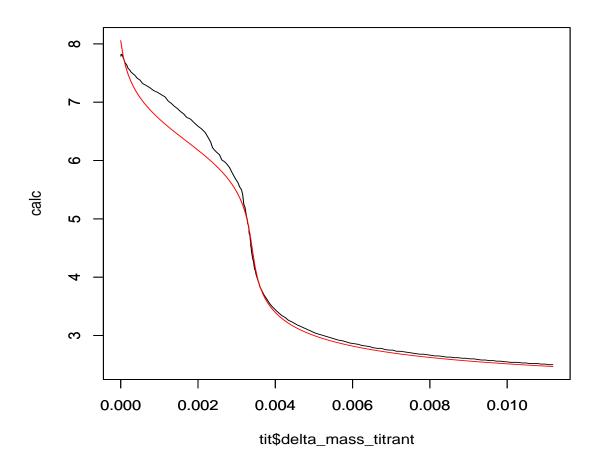
```
+ debug = TRUE, verbose = TRUE)
> sample5_2_fit
```

```
$TA
[1] 0.003417041
attr(,"unit")
[1] "mol/kg-soln"

$SumCO2
[1] 0.003474673
attr(,"unit")
[1] "mol/kg-soln"

$sumofsquares
```

[1] 5.060395



Comparing to the graphical method

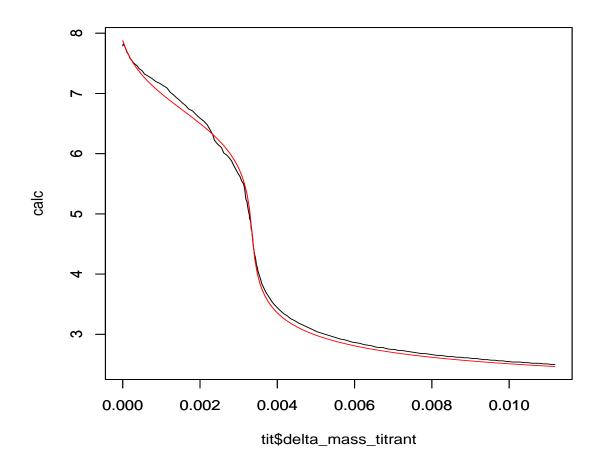
```
> peeneTA <- convert(0.00329, "conc", "molar2molin", Tc = 18.3,
+     S = 0)
> peeneSumCO2 <- convert(0.039647/12, "conc", "molar2molin", Tc = 18.3,
+     S = 0)
> c(peeneTA, sample5_2_fit$TA)

[1] 0.003294805 0.003417041
> c(peeneSumCO2, sample5_2_fit$SumCO2)

[1] 0.003308742 0.003474673

Fitting K_CO2 as well
> sample5_2_fit_1 <- TAfit(aquaenv(Tc = 18.3, S = 0), sam5_2, conc_titrant,
+     mass_sample, S_titrant = 0.5, equalspaced = FALSE, pHscale = "nbs",
+     verbose = TRUE, K_CO2fit = TRUE, debug = TRUE)
> sample5_2_fit_1
```

```
$TA
[1] 0.00336031
attr(,"unit")
[1] "mol/kg-soln"
$SumCO2
[1] 0.003568864
attr(,"unit")
[1] "mol/kg-soln"
$K_C02
[1] 2.023563e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"
$sumofsquares
[1] 0.9279765
```



3.5.2.4.9 sample5_3: E curve

The dataset sample5_3 has been obtained by manually titrating samples of the freshwater sample mentioned above and recording the "raw" potential (E) values coming from the electrode.

Preparations

mass_sample, S_titrant = 0.5, Evals = TRUE, verbose = TRUE,

```
+ electrode_polarity = "neg", debug = TRUE)
> sample5_3_fit
```

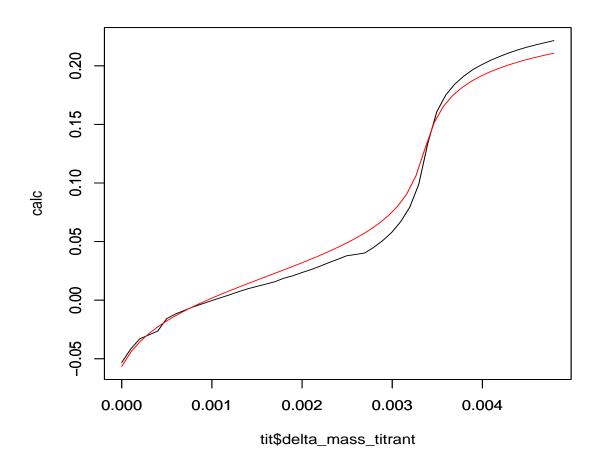
```
$TA
[1] 0.003371395
attr(,"unit")
[1] "mol/kg-soln"

$SumCO2
[1] 0.00353039
attr(,"unit")
[1] "mol/kg-soln"

$EO
[1] -0.3865431
attr(,"unit")
[1] "V"

$sumofsquares
```

[1] 0.003445975

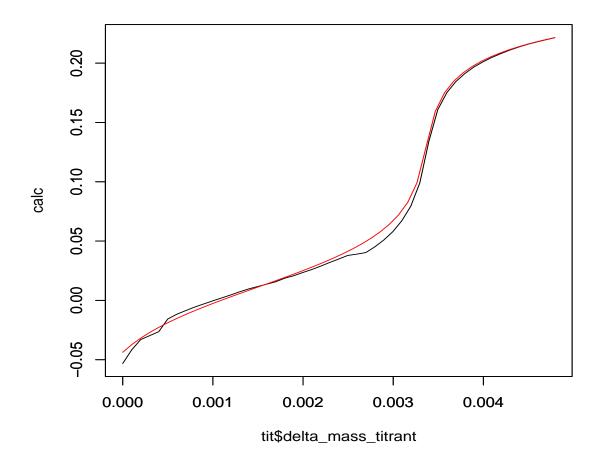


Fitting $K_{-}CO2$ as well

```
attr(,"unit")
[1] "V"

$K_C02
[1] 2.392563e-07
attr(,"unit")
[1] "mol/kg-soln"
attr(,"pH scale")
[1] "free"

$sumofsquares
[1] 0.0006593513
```



3.5.2.4.10 Some rudimentary statistics on sample5_1 to sample5_4 (Beware, these statistics are just "back of the envelope" calculations, since not all the measurements are independent and for different quantities not the same number of measurements

have been carried out.)

```
The graphical method
```

```
> peeneTA <- c(3310.61, 3286.456)
> peeneDIC <- c(3300.358)
> paste(mean(peeneTA), "+/-", sd(peeneTA))

[1] "3298.533 +/- 17.0794571927799"
> paste(mean(peeneDIC), "+/-", sd(peeneDIC))

[1] "3300.358 +/- NA"
```

```
pH curve fitting
> pHTA <- c(sample5_1_fit$TA, sample5_2_fit$TA, sample5_2_fit_1$TA) *</pre>
> pHDIC <- c(sample5_1_fit$SumCO2, sample5_2_fit$SumCO2, sample5_2_fit_1$SumCO2) *
      1e+06
> paste(mean(pHTA), "+/-", sd(pHTA))
[1] "3382.7305953731 +/- 30.1769214263983"
> paste(mean(pHDIC), "+/-", sd(pHDIC))
[1] "3492.61078871658 +/- 69.0542374221847"
E curve fitting
> ETA <- c(sample5_3_fit$TA, sample5_3_fit_1$TA) * 1e+06
> EDIC <- c(sample5_3_fit$SumCO2, sample5_3_fit_1$SumCO2) * 1e+06
> paste(mean(ETA), "+/-", sd(ETA))
[1] "3374.56336350352 +/- 4.48120487234782"
> paste(mean(EDIC), "+/-", sd(EDIC))
[1] "3642.01356486105 +/- 157.859653501884"
pH and E curve fitting combined
> totalTA <- c(pHTA, ETA)</pre>
> totalDIC <- c(pHDIC, EDIC)</pre>
> paste(mean(totalTA), "+/-", sd(totalTA))
[1] "3379.46370262527 +/- 21.9169955876238"
> paste(mean(totalDIC), "+/-", sd(totalDIC))
[1] "3552.37189917437 +/- 123.735679008991"
```

4 Extending AquaEnv

It is very simple for the user to create own functions that use **AquaEnv** and extend its functionality. We will demonstrate that by creating simple analogons for the **AquaEnv** functions titration and TAfit.

The function simpletitration will take the following arguments

```
aquaenv an object of class aquaenv: minimal definition, contains all information about the system: T, S, d, total concentrations of nutrients etc

volume the volume of the (theoretical) titration vessel in l

amount the amount of titrant added in mol

steps the amount of titrant is added in

type the type of titrant: either "HCl" or "NaOH"
```

The function is defined as

```
> simpletitration <- function(aquaenv,
                                                         # an object of class aquaenv: minimal definition,
                                                         # contains all information about the system:
                                                         # T, S, d, total concentrations of nutrients etc
                                volume.
                                                         \mbox{\tt\#} the volume of the (theoretical) titration vessel in 1
                                                         # the amount of titrant added in mol
                                amount,
                                                         # the amount of steps the amount of titrant is added in
                                steps,
                                                         # the type of titrant: either "HC1" or "NaOH"
                                type)
    {
      directionTAchange <- switch(type, HCl = -1, NaOH = +1)</pre>
      TAconcchangeperstep <- convert(((amount/steps)/volume), "conc", "molar2molin", aquaenv$Tc, aquaenv$S)
      aquaenvtemp <- aquaenv
      for (i in 1:steps)
                       <- aquaenvtemp$TA + (directionTAchange * TAconcchangeperstep)</pre>
          aquaenvtemp <- aquaenv(ae=aquaenvtemp, TA=TA)</pre>
                       <- merge(aquaenv, aquaenvtemp)</pre>
      aquaenv[["DeltaCTitrant"]] <- convert((amount/volume)/steps*(1:(steps+1)),
                                               "conc", "molar2molin", aquaenv$Tc, aquaenv$S)
      {\tt return(aquaenv)} \quad {\tt\#\ object\ of\ class\ aquaenv\ which\ contains\ a\ titration\ simulation}
```

and can be used to create a bjerrum plot

```
> ae <- simpletitration(aquaenv(Tc = 15, S = 35, SumCO2 = 0.0035,

+ SumNH4 = 2e-05, pH = 11.3), volume = 100, amount = 1.5, steps = 100,

+ type = "HCl")

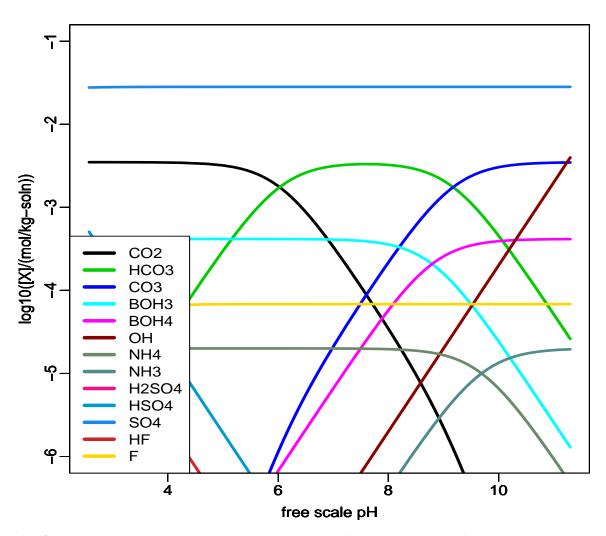
> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",

+ "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")

> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6,

+ -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6,

+ colors()[seq(100, 250, 6)]), newdevice = FALSE)
```



The function simpletitration in turn can be used to create a simple analogon to TAfit with the arguments

ae an object of class aquaenv: minimal definition, contains all infor-

mation about the system: T, S, d, total concentrations of nutri-

ents etc

pHmeasurements a table containing the titration curve: basically a series of pH

values (pH on free proton scale)

volume the volume of the titration vessel
amount the total amount of the titrant added

TAguess=0.0025 a first guess for [TA] and [SumCO2] to be used as initial values

for the optimization procedure

type="HCl" the type of titrant: either "HCl" or "NaOH"

defined as

```
# basically a series of pH values (pH on free proton scale)
                    volume,
                                           # the volume of the titration vessel
                                           # the total amount of the titrant added
                    amount.
                    TAguess=0.0025,
                                           # a first guess for [TA] and [SumCO2] to be used as
                                           # initial values for the optimization procedure
                    type="HC1")
                                           # the type of titrant: either "HCl" or "NaOH"
ae\$Na <- NULL  # make sure ae gets cloned as "skeleton": cloneaquaenv determines "skeleton"
                # TRUE or FALSE from the presence of a value for Na
residuals <- function(state)</pre>
    ae$SumCO2 <- state[[1]]</pre>
              <- simpletitration(aquaenv(ae=ae, TA=state[[2]]), volume=volume,</pre>
                                   {\tt amount=amount, steps=(length(pHmeasurements)-1), type=type)\$pH}
    residuals <- pHmeasurements-pHcalc
    return(residuals)
require(minpack.lm)
out <- nls.lm(fn=residuals, par=c(TAguess, TAguess)) #guess for TA is also used as guess for SumCO2
                          <- list(out$par[[2]], out$par[[1]], out$deviance)</pre>
attr(result[[1]], "unit") <- "mol/kg-soln"
attr(result[[2]], "unit") <- "mol/kg-soln"
names(result)
                          <- c("TA", "SumCO2", "sumofsquares")
return(result) # a list of three values
                # ([TA] in mol/kg-solution, [SumCO2] in mol/kg-solution, sum of the squared residuals)
```

The function simpleTAfit can be used to calculate TA and SumCO2

A Abbreviations for references used throughout the code and in the helpfiles

Atkins1996 Atkins (1996) Boudreau1996 Boudreau (1996) DOE1994 DOE (1994) Dickson1979a Dickson and Riley (1979a) Dickson1981 Dickson (1981) Dickson1984 Dickson (1984) Dickson1987 Dickson and Millero (1987) Dickson1990 Dickson (1990a) Emerson and Hedges (2008) Emerson2008 Follows2006 Follows, Ito, and Dutkiewicz (2006) Hofmann et al. (2008b) Hofmann2008 Lewis1998 Lewis and Wallace (1998) Lueker2000 Lueker *et al.* (2000) Millero1981 Millero and Poisson (1981) Millero1988 Millero et al. (1988)Millero1995 Millero (1995) Millero et al. (1995) Millero1995a Mucci1983 Mucci (1983) Perez and Fraga (1987a) Perez1987a ${\rm Riordan} 2005$ Riordan et al. (2005) Roy1993b Roy et al. (1993b) Sundquist1979 Sundquist, Plummer, and Wigley (1979) Weiss1970 Weiss (1970) Weiss1974 Weiss (1974) Wischmeyer et al. (2003) Wischmeyer 2003 Zeebe2001 Zeebe and Wolf-Gladrow (2001)

B References for the elements of an object of class aquaenv

| element | references |
|---------|--|
| Cl | (DOE 1994, chapter 5, p. 11), and (Zeebe and Wolf-Gladrow 2001, p. 100, footnote 3) |
| I | (DOE 1994, chapter 5, p. 13, 15), (Zeebe and Wolf-Gladrow 2001, p.12), and (Roy et al. 1993b, p.257). Note that the approximation $I/(mol/kg\text{-solution}) \approx 0.0199201$ S is given in (Millero 1982, p. 428.). This relationship converted into mol/kg-H ₂ O and the last digits adjusted (from 0.0199201 to 0.019924) results in the formula used here. |
| density | Millero and Poisson (1981) and (DOE 1994, chapter 5, p. 6f). |

SumBr, ClConc, Na, Mg, Ca, K, Sr molal2molin free2tot,

tot2free free2sws,

tot2sws, sws2free, sws2tot

 $K0_CO2$

 $K0_{-}O2$

 $K_{-}W$

 $K_{-}HSO4$

K_HF

K_CO2, K_HCO3

K_BOH3

K_NH4

(DOE 1994, chapter 5, p.11)

(Roy et al. 1993b, p.257), and (DOE 1994, chapter 5, p. 15)

(Dickson 1984, p.2302), (DOE 1994, chapter 4, p.16), (Zeebe and Wolf-Gladrow 2001, p.57, 261)

(Dickson 1984, p.2303), (Zeebe and Wolf-Gladrow 2001, p.57)

Weiss (1974), (DOE 1994, chapter 5, p. 13) (here it is stated that the unit is mol/(kg-solution*atm)), (Millero 1995, p.663), (Zeebe and Wolf-Gladrow 2001, p.257)

derived from a formula for the oxygen saturation concentration in $ml-O_2/kg$ -solution by Weiss (1970) using the first virial coefficient of oxygen (Atkins 1996, p. 41, 1029) and the atmospheric oxygen fugacity (Williams 2004)

(Millero 1995, p.670) (**original reference**, but slightly different formula for seawater pH), (DOE 1994, chapter 5, p. 18) (NOT the original reference! DOE (1994) cites in an update from 1997 Millero (1995)! However the version of the formula used here is the one converted to total pH scale given in DOE (1994)), and (Zeebe and Wolf-Gladrow 2001, p. 258). Constant type (stoichiometric), pH scale (total, converted to free here), and concentration unit (mol/kg-solution squared): (DOE 1994, chapter 5, p. 12,18), pH scale also in (Zeebe and Wolf-Gladrow 2001, p. 258).

(DOE 1994, chapter 5 page 13), (Zeebe and Wolf-Gladrow 2001, p. 260), Dickson (1990b) (original reference). Constant type (stoichiometric), pH scale (free) , and concentration unit (mol/kg-H₂O converted to mol/kg-solution here): (DOE 1994, chapter 5, p. 13).

(Dickson and Riley 1979b, p. 91) (original reference), (DOE 1994, c. 5, p. 15), (Roy et al. 1993b, p. 257), (Dickson and Millero 1987, p. 1783), (Millero 1995, p. 664), (Zeebe and Wolf-Gladrow 2001, p. 260) (converted to molinty and total scale). Constant type (stoichiometric), pH scale (free), and concentration unit ($mol/kg-H_2O$ converted to mol/kg-solution here): (DOE 1994, chapter 5, p. 15, 16).

(Roy et al. 1993b, p. 254) (original reference), (DOE 1994, chapter 5, p.14) (in a version converted to mol/kg-H₂O), (Millero 1995, p. 664), (Zeebe and Wolf-Gladrow 2001, p. 255). Constant type (stoichiometric) and concentration unit (mol/kg-H₂O converted to mol/kg-solution here): (DOE 1994, chapter 5, p. 14, 15), pH scale (total, converted to free here): In (DOE 1994, chapter 5, p. 12) the total scale is stated for the formula for high salinities and thus can be inferred for the formula for low salinities. The scale is also indirectly stated for both formulations in the original reference Roy et al. (1993b). Note that in Roy et al. (1993b) a function for pure water (Millero 1979) is cited and a function for seawater is derived. In Millero (1995) it is stated that for S<5 the fresh water formula of (Millero 1979) should be used and for S>=5 the seawater formula derived in Roy et al. (1993b). However, both formulations do not always intersect at S=5. The true intersection with respect to salinity S is a function of temperature Tk. Here, we first calculate this intersection by numerical root finding and then decide which formulation to use. This practise results in a continuous function with respect to S. (Note that Millero (1979) is restated wrongly in Roy et al. (1993b): one of the numerical values for the function for $K_{CO_2}^*$ is given as 310.48919, but correct is 2310.48919. However, in Millero (1995) this value is stated correctly.)

(Dickson 1990a, p. 763) (original, but mol/kg-H₂O version), (DOE 1994, ch. 5, p. 14), (Zeebe and Wolf-Gladrow 2001, p. 262), (Millero 1995, p.669) (mol/kg-H₂O version) , agrees with data in Roy, Roy, Lawson, Vogel, Moore, Davis, and Millero (1993a). Constant type (stoichiometric) and concentration unit (mol/kg-solution): (DOE 1994, chapter 5, p. 14), pH scale (total): (DOE 1994, chapter 5, p. 12) and (Zeebe and Wolf-Gladrow 2001, p.263).

Millero et al. (1995) (original reference), (Millero 1995, p.671). Constant type (stoichiometric) and concentration unit (mol/kg-solution): (Millero 1995, p.671), pH scale (seawater, converted to free here): Lewis and Wallace (1998) (in corrections of Millero (1995)).

Millero et al. (1988) (original reference), (Millero 1995, p.671). Constant type (stoichiometric) and

 $K_{-}H2S$

| 11.25 | concentration unit (mol/kg-solution): (Millero 1995, p.671), PH scale (seawater, converted to free here): Lewis and Wallace (1998) (in corrections of Millero (1995)). |
|---|--|
| K_H3PO4, K_H2PO4, K_HPO4 | (Millero 1995, p.670) (original reference, but formula for seawater scale pH), (DOE 1994, ch. 5, p 16,17), agrees with data in Dickson and Riley (1979a). Constant type (stoichiometric), concentration unit (mol/kg-solution), and pH scale (total, converted to free here): (DOE 1994, chapter 5, p. 12, 16, 17). |
| K_SiOH4 | Millero et al. (1988) (original reference), (DOE 1994, chapter 5, p 17), (Millero 1995, p.671) (formula for seawater scale pH) Constant type (stoichiometric), concentration unit (mol/kg-H ₂ O converted to mol/kg-solution here by omitting the conversion summand ln(1-0.001005 S)), and pH scale (total, converted to free here): (DOE 1994, chapter 5, p. 12, 17). |
| K_SiOOH3 | Wischmeyer et al. (2003) (original reference), corrected due to personal communication with Dieter Wolf-Gladrow (one of the authors). The corrected version can be obtained from either Dieter Wolf-Gladrow or Andreas F Hofmann (a.hofmann@nioo.knaw.nl). Constant type (stoichiometric), concentration unit (mol/kg-solution), and pH scale (total, converted to free here): Wischmeyer et al. (2003). |
| K_HNO2 | Constant value, not a function of temperature and salinity! Obtained as a hybrid pk value (featuring the activity of the proton but the concentration of other species (see Zeebe and Wolf-Gladrow (2001) for a treatment of different types of equilibrium constants) in molar concentration (mol/l) on the NBS pH scale (Durst 1975) from Riordan <i>et al.</i> (2005). Used as an approximation for the stoichiometric $K_{HNO_2}^*$ in mol/kg-solution on the free proton pH scale here. |
| K_H2SO4 | Constant value, not a function of temperature and salinity! Obtained as a standard pK value from (Atkins 1996, p. 1045). Used as an approximation for the stoichiometric $K_{H_2SO_4}^*$ in mol/kg-solution on the free proton pH scale here. |
| K_HS | Constant value, not a function of temperature and salinity! Obtained as a standard pK value from (Atkins 1996, p. 1045). Used as an approximation for the stoichiometric $K_{Hs^-}^*$ in mol/kg-solution on the free proton pH scale here. |
| Ksp_calcite, Ksp_aragonite | Mucci (1983) (original reference), Boudreau (1996). Note that in there are errors in Boudreau (1996): b_0 for calcite is not 0.7712 but 0.77712 and b_1 for aragonite is not 0.001727 but 0.0017276. |
| рН | As given in Dickson (1984), p. 2303 (use of "m") and Dickson and Riley (1979a), p. 91f all concentrations appearing in the definition of the total and the seawater pH scale are molal (mol/kg-H ₂ O) concentrations. But in Roy <i>et al.</i> (1993b), p. 257 and in DOE (1994), chapter 4, SOP 6, p. 1 it is stated, that concentrations for the seawater and total pH scale are in mol/kg-solution. To be consistent with DOE (1994) molin concentrations (mol/kg-solution) are chosen for calculating the pH. |
| revelle | (Zeebe and Wolf-Gladrow 2001, p.73) |
| dTAdKdKdS, dTAdKdKdT, dTAdKdKddd, dTAdKdKd- SumH2SO4, dTAdKdKd- SumHF | Hofmann <i>et al.</i> (2008a) |
| | |

The values for K_W, K_HSO4, K_HF, K_CO2, K_HCO3, K_BOH3, K_NH4, K_H2S, K_H3PO4, K_H2PO4, K_HPO4, K_SiOH4, K_SiOOH3, Ksp_calcite, Ksp_aragonite obtained as functions of salinity S and temperature Tc from the above references are pressure corrected using the given depth d and the calculated hydrostatic pressure hydroP according to Millero (1995) with corrections by Lewis and Wallace (1998).

In general it is to be said that all corrections from Lewis and Wallace (1998) have been applied.

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