**Title: Exact** integer linear programming solvers outperform simulated annealing for solving conservation planning problems

**Authors:** Richard Schustera,b,\*, Jeffrey O. Hansonc, Matt Strimas-Mackeyd, Joseph R. Bennetta

a Department of Biology, 1125 Colonel By Drive, Carleton University, Ottawa ON, K1S 5B6 Canada.

b Ecosystem Science and Management Program, 3333 University Way, University of Northern British Columbia, Prince George BC, V2N 4Z9 Canada.

c School of Biological Sciences, The University of Queensland, Brisbane, QLD 4072, Australia

d Cornell Lab of Ornithology, Cornell University, Ithaca, NY 14850 USA.

\*Corresponding author: Department of Biology, 1125 Colonel By Drive, Carleton University, Ottawa ON, K1S 5B6 Canada. Email: [richard.schuster@glel.carleton.ca](mailto:richard.schuster@glel.carleton.ca), Phone: +1 250 631 8324, ORCID: 0000-0003-3191-7869

**Abstract**

The resources available for conserving biodiversity are limited, and so protected areas need to be established in places that will achieve objectives for minimal cost. Two of the main algorithms for solving systematic conservation planning problems are Simulated Annealing (SA) and exact integer linear programming (EILP) solvers. Using a case study in British Columbia, Canada, we compare the cost-effectiveness and processing times of SA used in Marxan versus EILP using both commercial and open-source algorithms. Plans for expanding protected area systems based on EILP algorithms were 12 to 30% cheaper than plans of Marxan using SA, due to EILP’s ability to find optimal solutions as opposed to approximations. The best EILP solver we examined was on average 1071 times faster than the Marxan SA algorithm tested. The performance advantages of EILP solvers were also observed when we aimed for spatially compact solutions by including a boundary penalty. One practical advantage of using EILP over SA is that the analysis does not require calibration, saving even more time. Given the performance of EILP solvers, they can be used to generate conservation plans in real-time during stakeholder meetings and can facilitate rapid sensitivity analysis, and contribute to a more transparent, inclusive, and defensible decision-making process.

**Introduction**

Area-based systematic conservation planning aims to provide a rigorous, repeatable, and structured approach for designing new protected areas that efficiently meet conservation objectives (Margules and Pressey 2000). Historically, spatial conservation decision-making often evaluated parcels opportunistically as they became available for purchase, donation, or under threat (Pressey et al. 1993, Pressey and Bottrill 2008). Although purchasing such areas may improve the status quo, such decisions may not substantially and cost-effectively enhance the long-term persistence of species or communities (Joppa and Pfaff 2009, Venter et al. 2014). Systematic conservation planning, on the other hand, is a multi-step process that involves framing conservation planning problems as optimization problems with clearly defined objectives (e.g. minimize acquisition cost) and constraints (Margules and Pressey 2000). These optimization problems are then solved to obtain candidate reserve designs (termed solutions), which are used to guide protected area acquisitions and land policy (Schwartz et al. 2018). Due to the systematic, evidence-based nature of these tools, they can help contribute to a transparent, inclusive, and more defensible decision-making process (Margules and Pressey 2000).

Today, Marxan is the most widely used systematic conservation planning software, having been used in 184 countries to design marine and terrestrial reserve systems (Ball et al. 2009). Although Marxan supports several algorithms for solving conservation planning problems, most conservation planning exercises use its implementation of simulated annealing (SA), an iterative, stochastic metaheuristic algorithm for approximating global optima of complex functions (Kirkpatrick et al. 1983). By conducting thousands of simulations to determine the impact of different candidate solutions, Marxan aims to generate solutions that are near-optimal. One of the reasons why Marxan uses SA instead of exact integer linear programming (EILP) solvers, is that EILP solvers were historically not well suited to solve problems with nonlinear constraints and penalties, such as problems trying to create spatially compact or connected solutions (i.e. compactness and connectivity goals) and generally took a lot longer than SA to solve problems (Sarkar et al. 2006, Haight and Snyder 2009). However, the SA approach provides no guarantee on solution quality. As a consequence, conservation scientists and practitioners have no way of knowing how suboptimal their solutions are. Optimal being the configuration of protected areas that delivers the desired benefits and the lowest cost. The discussion about the pros and cons of linear programming versus heurists, such as SA, in conservation planning spans more than two decades (Cocks and Baird 1989, Underhill 1994, Church et al. 1996, Rodrigues and Gaston 2002, Önal 2004), but the EILP shortcomings mentioned above have largely been overcome in recent years (Beyer et al. 2016).

eIn a recent simulation study, Beyer et al. (2016) found that Marxan with simulated annealing can deliver solutions that are orders of magnitude below optimality. They compared Marxan to exact integer linear programming (EILP) (Wolsey and Nemhauser 1999), which minimizes or maximizes an objective function (a mathematical equation describing the relationship between actions and outcomes) subject to a set of constraints and conditional on the decision variables (the variables corresponding to the selection of actions to implement) being integers (Beyer et al. 2016). Unlike metaheuristic methods such as SA, prioritization using EILP will find the optimal solution or can be instructed to return solutions within a defined level of suboptimality. Some have argued that EILP algorithms are well-suited for solving conservation planning problems (Cocks and Baird 1989, Underhill 1994, Rodrigues and Gaston 2002), but until recent advances in computational capacity and algorithms, it has been impossible to solve the Marxan-like systematic conservation planning problems with EILP for large problems (Haight and Snyder 2009, Beyer et al. 2016). (Beyer et al. 2016) recently introduced a linearization solution to the nonlinear constraint problem to find efficient solutions in an EILP framework, which greatly improved the utility of EILP for solving conservation planning problems.

Here we compare exact integer linear programming solvers with simulated annealing as used in Marxan, for solving minimum set systematic conservation planning problems (Rodrigues et al. 2000) using real-world data from Western North America. The goal of solving the minimum set problem is to find the places that maximize biodiversity, while minimizing reserve cost. We found that EILP generated high quality solutions 1,000 times faster than simulated annealing that could save over $100 million (or 13%) for realistic conservation scenarios when compared to solutions obtained from simulated annealing. These results also hold true for problems aiming for spatially compact solutions. Our findings open up new possibilities for scenario generation to quickly explore and compare different conservation prioritization scenarios in real-time.

**Material and Methods**

*Study area*

We focused on a 27,250 km2 portion of the Georgia Basin, Puget Trough and Willamette Valley of the Pacific Northwest region spanning the US and Canada, corresponding to the climate envelope indicative of the Coastal Douglas-fir (CDF) Biogeoclimatic zone in southwestern British Columbia (Meidinger and Pojar 1991) (Supplementary Information Figure S1). Land cover in the region is diverse, with approximately 57% of the land in forest, 8% as savanna or grassland, 5% in cropland, 10% being urban or built and the rest in wetland, water or barren.

*Biodiversity data.*

We used species distribution models for 72 bird species as our conservation features at a 1-ha grid cell resolution (Supplementary Table 1).The distribution models were based on data from eBird, a citizen-science effort that has produced the largest and most rapidly growing biodiversity database in the world (Hochachka et al. 2012, Sullivan et al. 2014). From the 2013 eBird Reference Dataset (<http://ebird.org/ebird/data/download>) we used a total of 12,081 checklists in our study area, then filtered these checklists to retain only those from March – June to capture the breeding season, <1.5 hours in duration, <5 km travelled, and a maximum of 10 visits to a given location to improve model fit. Sampling locations <100 m apart were collapsed to one location, yielding 5,470 checklists from 2,160 locations, visited from 1-10 times and 2.53 times on average. The R package unmarked (version 0.9-9; Fiske and Chandler 2011) provided the framework for all species distribution models, which necessarily include two parts: occupancy and detection (Mackenzie et al. 2002). This form of distribution modelling, also known as occupancy modelling uses the information from repeat visits to a site to infer estimates of detectability of a species as well as estimates of probability of occurrence. For further details on biodiversity data see (Rodewald et al. 2019).

*Property layer and land cost*.

We incorporated spatial heterogeneity in land cost (Ando et al. 1998, Polasky et al. 2001, Ferraro 2003, Naidoo et al. 2006) in our plans by using property data and 2012 land value assessments from the Integrated Cadastral Information Society of British Columbia (BC). This process resulted in 193,623 properties for BC which were subsequently used as planning units (Schuster et al. 2014). Property data, including tax assessment land values from Washington State came from the University of Washington’s Washington State Parcel Database (<https://depts.washington.edu/wagis/projects/parcels/>; Version: StatewideParcels\_v2012n\_e9.2\_r1.3; Date accessed: 2015/04/30), as well as San Juan County Parcel Data with separate signed user agreement. The combined property layer included 1.92 million polygons. Property data, including tax assessment land values from Oregon State had to be sourced from individual counties, which included Benton, Clackamas, Columbia, Douglas, Lane, Linn, Marion, Multnomah, Polk, Washington and Yamhill. The combined property layer for Oregon included 605,425 polygons. We converted the polygon cost values to 1-ha raster cells for consistency with the biodiversity data by calculating area weighted mean values of cost per raster cell. Using tax assessment values as an estimate of conservation cost is an underestimate because tax assessment values are often lower than market value, but estimates of market values over larger areas are rarely available and tax assessments do provide a good general approximation.

*Spatial prioritization*

We compared EILP and SA for solving the minimum set spatial prioritization problem (Ball et al. 2009). In this formulation, the landscape is divided into a set of discrete planning units. Each planning unit is assigned a financial cost (here we use the assessed land value) and a conservation value for a set of features that we wish to protect (here the occupancy probability for a set of species). We also define representation targets for each species as the amount of habitat we hope to protect for that species. The goal of this prioritization problem is to optimize the trade-off between conservation benefit and financial cost (McIntosh et al. 2017). Achieving this goal involves finding the set of planning units that meets the conservation targets for the minimum possible cost (i.e. min cost: such that conservation value ≥ target). Details on the Marxan problem formulation can be found in Ball et al. (2009) and the EILP formulation in Beyer et al. (2016). Three key parameters that are important for Marxan analysis, which we also use here are: species penalty factor, number of iterations, and number of restarts (Ardron et al. 2010). Briefly, the species penalty factor is the penalty given to a reserve system for not adequately representing a feature, the number of iterations determines how long the annealing algorithms will run, and the number of restarts determines how many different solutions Marxan will generate (for more details see SI Appendix S1). For all scenarios, we used 1 km2 planning units, generated by aggregating the species and cost data to this coarser resolution from the original 1-ha cells. Aggregation was accomplished by taking the sum of cost data and the mean of species data for all 1-ha cells within the larger 1 km2 cells.

*EILP solvers (commercial vs open source)*

A variety of EILP solvers currently exist, and both commercial and open source solvers are available. All solvers yield optimal solutions to EILP problems, but there are substantial differences in performance (i.e. time taken to solve a problem) and in the size of problems that can be solved (Lin et al. 2017). For the purposes of performance testing we opted for one of the best commercial solvers currently available, Gurobi (Gurobi Optimization Inc. 2017). In a recent benchmark study, Gurobi outperformed other solver packages for more complex formulations and a practical use-case (Luppold et al. 2018). To investigate solver performance of packages that are freely available to everyone, we also tested the open source solver SYMPHONY (Ralphs et al. 2019). Both Gurobi and SYMPHONY can be used from R. For Gurobi we used the R package provided with the software (Gurobi version 8.1-0) and for SYMPHONY the Rsymphony package (version 0.1-28; Harter et al. 2017). We used the prioritizr R package to solve EILP problems for both Gurobi and SYMPHONY solvers (Hanson et al. 2019).

*Scenarios investigated*

We investigated a range of scenarios that were computationally feasible for this study. For both Marxan and prioritzr we created the following range of scenarios: i) vary conservation targets between 10 and 90% protection of features in 10% increments (9 variations), using ii) 10 – 72 features (5 variations) as targets, and iii) with spatial extents of 9,282 planning units, 37,128 planning units, and 148,510 planning units (3 variations), resulting in a total of 135 scenarios created (Table 1). For Marxan, we also varied two additional parameters, i) the number of iterations ranged from 104 to 108 (5 variations) and ii) species penalty factors (SPF) of 1, 5, 25, and 125 were explored (4 variations, roughly spanning two orders of magnitude) for a total of 2,700 scenarios investigated in Marxan (Table 1). Exploring ranges of values for number of iterations and SPF are recommended for calibration of Marxan to increase Marxan’s ability to approximate the optimal solution (Ardron et al. 2010). As the processing time for the most complex problem in Marxan (90% target, 72 features, 148,510 planning units, 108 iterations) was >8 hours, we restricted the full range of scenarios to those mentioned above. The maximum number of planning units we used is within the range of previous studies using Marxan (e.g. Venter et al. 2014; Runge et al. 2016), although using more than 50,000 planning units with SA is discouraged without extensive parameter calibration, as near optimal solutions will be hard to find for problems of that size (Ardron et al. 2010). To allow for a fair contrast between SA and EILP, that focuses on algorithmic comparisons and not within SA variation, we focused our results and discussion on the best solution achieved with Marxan across 10 repeat runs.

As systematic conservation planners often aim for spatially compact solutions to their problems, we also investigated a range of scenarios using a term called boundary length modified (BLM), which is used to improve the clustering and compactness of a solution (McDonnell et al. 2002). We randomly selected a 225 x 225 pixel region of the study area to generate a problem with 50, 625 planning units, the maximum recommended for Marxan. After initial calibration we set the number of features/species to 72, SPF to 25 and number of iterations for Marxan to 108. We varied targets between 10 and 90% protection of features in 10% increments, and used the following BLM values: 0.1; 1; 10; 100; 1,000 for a total of 45 scenarios. Both Marxan and prioritzr allow a user to specify BLM values as presented here. For the details on the mathematical formulation of the spatial compactness constraint in ILP, please see SI Appendix S2 and (Beyer et al. 2016).

All analyses were conducted on a desktop computer with an Intel Core i7-7820X Processor and 128 GB RAM running Ubuntu 18.04 and R v 3.5.3. All data, scripts and full results are available online (<https://osf.io/my8pc/>) and will be archived in a persistent repository with a DOI pending acceptance of the manuscript.

**Results**

EILP algorithms (Gurobi, Symphony) outperformed SA (Marxan) in terms of their ability to find minimal cost solutions across all scenarios that met conservation targets. Summarizing across calibrated Marxan scenarios (number of iterations > 100,000 and species penalty factor 5 or 25), the range of savings ranged from 0.8% to 52.5% (median 12.6%, SI Figure S2) when comparing EILP results to the best (cheapest) solution for a Marxan scenario. For example, at the 30% protection target EILP solvers resulted in solutions that were $55 million cheaper than SA (Figure 1a), because it selected cheaper and less parcels in the optimal solution. With these savings an additional 961 ha could be protected (13,897 ha vs 12,936 ha) using an EILP algorithm by raising the representation targets until the cost of the resulting solution matched that of the Marxan solution using SA. In general, SA performed reasonably well at smaller problem sizes, fewer planning units and features and low targets, but as the problem size and complexity increased SA was less consistent in finding good solutions (SI Figure S2). Cost profiles across targets, number of features and number of planning units are shown in SI Figures S3-5.

The shortest processing times were achieved using the prioritizr package and the commercial solver Gurobi, followed by prioritizr and the open source solver Symphony, and lastly Marxan (Figure 1b). Gurobi had the shortest processing times across all scenarios investigated, Symphony tied with Gurobi in some scenarios and took up to 78 times longer than Gurobi in other scenarios (mean = 14 times, SI Figure S6), and Marxan took between 1.8 and 1995 times longer than Gurobi (mean = 281 times, SI Figure S7). The longest processing times for Gurobi, SYMPHONY and Marxan for a single scenario were 40 seconds, 31 minutes, and 8 hours respectively. For the most complex problem (i.e. targets = 90%, 72 features; 148,510 planning units), Marxan calibration across the 5 number of iterations and 4 species penalty factor values took a total of 5 days 7 hours, compared to 30 seconds using Gurobi and 28 minutes using SYMPHONY. Time profiles across targets, number of features and number of planning units are shown in SI Figures S8-10.

EILP algorithms (Gurobi, Symphony) also outperformed SA (Marxan) when using a BLM to achieve compacter solutions. This was true for objective function values (Figure 2a) as well as for processing times (Figure 2b). Through finding optimal solutions, using EILP resulted in objective function values 5.65 to 149% (mean 22.7%) lower than SA values. Gurobi was the fastest solver to find solutions to problems including BLM in 44 of 45 scenarios, in one case SYMPHONY was faster. SYMPHONY outperformed Marxan in 44 of 45 scenarios, and took on average 13.7 times as long as Gurobi to find a solution (range -0.31 to 42.6). Marxan was never faster than Gurobi and took on average 104.6 times as long as Gurobi to find a solution (range 3.09 to 190.8). An example of the spatial representation of the solutions for a 10% target is shown in SI Figure S11.

**Discussion**

We found that EILP algorithms outperformed SA both in terms of cost-effectiveness and processing times, even when including linearized non-linear problem formulations, when planning for spatially compact solutions. There have been calls for using EILP in solving conservation planning problems in the past (Underhill 1994, Rodrigues and Gaston 2002), but we are now at a point where making this switch is both advisable and computationally feasible, where technical capacity exists. Our study provides a systematic test case, using real world data to build on the findings of (Beyer et al. 2016) and show that their results hold for a realistic case study. We further expanded the scope of testing to include assessed land values in order to give estimates of how much better optimal solution can perform in terms of cost savings, compared to SA solutions. Finally, we showcase that even open source EILP solvers are much faster than SA algorithms as implemented in Marxan, which is very encouraging for non-academic user that would otherwise have to buy Gurobi licenses (Gurobi is free for academic use). The combination of the superior performance findings by both (Beyer et al. 2016) and this study indicates that EILP approaches should be strongly considered as improvements for minimum set conservation planning problems, currently solved using SA. This improvement is especially important in real world applications as the speed of generating solutions can be advantageous in iterative and dynamic planning processes that usually occur when planning for conservation (Sarkar et al. 2006). Given Marxan’s flexibility to use optimization methods other than SA, we hope that a future version of Marxan will include EILP solvers.

One practical advantage of using EILP over SA is that the analysis does not require parameter calibration. Unlike EILP, parameter calibration is a crucial task in every Marxan/SA project and the species penalty factors, number of SA iterations, and number of SA restarts must be calibrated to improve solution quality (Ardron et al. 2010). This task can be very time consuming, especially for larger problems (e.g. 50,000 planning units). Ideally all possible combinations of parameters should be explored, but this further increases processing time. For instance, exploring three different parameter values would result in 27 different scenarios to explore (i.e. 3 × 3 × 3). Although we omitted calibration runs prior to finalizing and presenting results in this study, the parameter calibration step took several days for the most complex problem we investigated in this study. Yet none of this calibration time is necessary using EILP. An added benefit is that the somewhat subjective process of setting values for these three parameters can be eliminated using EILP as well.

Recommended practices for Marxan analyses caution against using SA for conservation planning exercises with more than 50,000 planning units (Ardron et al. 2010). Such large-sized problems have occurred in the past and, as increasingly high resolution data become available, may become more common in the future (e.g. Venter et al. 2014; Runge et al. 2016). Unlike SA, EILP/prioritizr can solve problem sizes with more than one million planning units (Hanson 2018, Schuster et al. 2019). Realistically, as problem sizes grow beyond what was intended for Marxan/SA projects, EILP will run into problems solving very large problems (>1 million planning units) that include non-linear constraints, such as optimizing compactness or connectivity, as those problem formulations need to be linearized for EILP to work. A potential future solution to this issue could be the use of nonlinear integer programming for more problems including non-linear constraints (Grossmann 2002, Lee and Leyffer 2011). Whether EILP would also outperform SA for more complex problem formulations, such as dynamic problems or problems with multiple objectives, still needs to be explored. Potential solutions would be to linearize the problem, or incorporate algorithms like Mixed Integer Quadratically Constrained Programming (Franco et al. 2014).

Finally, we argue that another strength of EILP solvers, especially Gurobi, is that they can be used to quickly explore and compare different conservation prioritization scenarios in real-time. This ability could be used to great advantage during stakeholder meetings, to explore various scenarios and undertake rapid sensitivity analysis.

**Conclusion**

EILP algorithms substantially outperform SA as used in minimum set systematic conservation planning, both in terms of solution cost, as well as in terms of time required to find near optimal or optimal solutions. Using an EILP algorithm, as implemented in the R package prioritizr, has the added benefit that users do not need to worry about or set parameters such as species penalty factors or number of iterations, which significantly reduces the time a user spends on finding suitable values for these parameters. Given the potential EILP is showing for conservation planning, we recommend users consider adding this modified approach to solving systematic conservation planning problems.

**Acknowledgements**

RS is supported by a Liber Ero Fellowship and Environment and Climate Change Canada (ECCC), JOH by ECCC, MSM by endowments at the Cornell Lab of Ornithology, and JRB by Natural Sciences and Engineering Research Council of Canada and ECCC. We thank W. Hochachka for providing code fore processing eBird data. All data, scripts and full results are available online (<https://osf.io/my8pc/>) and will be archived in a persistent repository with a DOI pending acceptance of the manuscript.

**References**

Ando, A. et al. 1998. Species Distributions, Land Values, and Efficient Conservation. - Science 279: 2126–2128.

2010. Marxan Good Practices Handbook, Version 2 (JA Ardron, HP Possingham, and CJ Klein, Eds.). - Pacific Marine Analysis and Research Association.

Ball, I. R. R. et al. 2009. Marxan and relatives: Software for spatial conservation prioritisation. - In: Moilanen, A. et al. (eds), Spatial conservation prioritisation: Quantitative methods and computational tools. Oxford University Press, pp. 185–195.

Beyer, H. L. et al. 2016. Solving conservation planning problems with integer linear programming. - Ecological Modelling 328: 14–22.

Church, R. L. et al. 1996. Reserve selection as a maximal covering location problem. - Biological Conservation 76: 105–112.

Cocks, K. D. and Baird, I. A. 1989. Using mathematical programming to address the multiple reserve selection problem: An example from the Eyre Peninsula, South Australia. - Biological Conservation 49: 113–130.

Ferraro, P. J. 2003. Assigning priority to environmental policy interventions in a heterogeneous world. - Journal of Policy Analysis and Management 22: 27–43.

Fiske, I. J. and Chandler, R. B. 2011. unmarked : An R Package for Fitting Hierarchical Models of Wildlife Occurrence and Abundance. - Journal Of Statistical Software 43: 128–129.

Franco, J. F. et al. 2014. A mixed-integer quadratically-constrained programming model for the distribution system expansion planning. - International Journal of Electrical Power & Energy Systems 62: 265–272.

Grossmann, I. E. 2002. Review of Nonlinear Mixed-Integer and Disjunctive Programming Techniques. - Optimization and Engineering 3: 227–252.

Gurobi Optimization Inc. 2017. Gurobi Optimizer Reference Manual, Version 7.5.1.

Haight, R. G. and Snyder, S. A. 2009. Integer programming methods for reserve selection and design. - In: Moilanen, Atte; Wilson, Kerrie A.; Possingham, Hugh, eds. Spatial conservation prioritization. Quantitative methods and computational tools. Oxford, UK: Oxford University Press: 43-57. Chapter 4.: 43–57.

Hanson, J. 2018. Conserving evolutionary processes.

Hanson, J. et al. 2019. prioritizr: Systematic Conservation Prioritization in R, Version 4.0.2.

Harter, R. et al. 2017. Rsymphony: SYMPHONY in R.

Hochachka, W. M. et al. 2012. Data-intensive science applied to broad-scale citizen science. - Trends in ecology & evolution 27: 130–137.

Joppa, L. N. and Pfaff, A. 2009. High and far: biases in the location of protected areas. - PloS one 4: e8273.

Kirkpatrick, S. et al. 1983. Optimization by Simulated Annealing. - Science 220: 671–680.

Lee, J. and Leyffer, S. 2011. Mixed Integer Nonlinear Programming. - Springer Science & Business Media.

Lin, C. Y. et al. 2017. Participant Selection Problem: Relative Performance of Five Optimization Solvers. - Proceedings of the 8th International Conference on Computer Modeling and Simulation: 24–31.

Luppold, A. et al. 2018. Evaluating the performance of solvers for integer-linear programming. in press.

Mackenzie, D. I. et al. 2002. Estimating site occupancy rates when detection probabilities are less than one. - Ecology 83: 2248–2255.

Margules, C. R. and Pressey, R. L. 2000. Systematic conservation planning. - Nature 405: 243–53.

McDonnell, M. D. et al. 2002. Mathematical Methods for Spatially Cohesive Reserve Design. - Environmental Modeling & Assessment 7: 107–114.

McIntosh, E. J. et al. 2017. The Impact of Systematic Conservation Planning. - Annual Review of Environment and Resources 42: annurev-environ-102016-060902.

Meidinger, D. and Pojar, J. 1991. Ecosystems of British Columbia. - British Columbia Ministry of Forests.

Naidoo, R. et al. 2006. Integrating economic costs into conservation planning. - Trends in ecology & evolution 21: 681–7.

Önal, H. 2004. First-best, second-best, and heuristic solutions in conservation reserve site selection. - Biological Conservation 115: 55–62.

Polasky, S. et al. 2001. Selecting Biological Reserves Cost-Effectively: An Application to Terrestrial Vertebrate Conservation in Oregon. - Land Economics 77: 68–78.

Pressey, R. L. and Bottrill, M. C. 2008. Opportunism, Threats, and the Evolution of Systematic Conservation Planning. - Conservation Biology 22: 1340–1345.

Pressey, R. et al. 1993. Beyond opportunism: key principles for systematic reserve selection. - Trends in ecology & evolution 8: 124–128.

Ralphs, T. et al. 2019. coin-or/SYMPHONY: Version 5.6.17. - Zenodo.

Rodewald, A. D. et al. 2019. Tradeoffs in the value of biodiversity feature and cost data in conservation prioritization. - Sci Rep 9: 1–8.

Rodrigues, A. S. L. and Gaston, K. J. 2002. Optimisation in reserve selection procedures—why not? - Biological Conservation 107: 123–129.

Rodrigues, A. S. et al. 2000. Flexibility, efficiency, and accountability: adapting reserve selection algorithms to more complex conservation problems. - Ecography 23: 565–574.

Runge, C. A. et al. 2016. Incorporating dynamic distributions into spatial prioritization (N Roura-Pascual, Ed.). - Diversity and Distributions 22: 332–343.

Sarkar, S. et al. 2006. Biodiversity Conservation Planning Tools: Present Status and Challenges for the Future. - Annual Review of Environment and Resources 31: 123–159.

Schuster, R. et al. 2014. Bird Community Conservation and Carbon Offsets in Western North America. - Plos One in press.

Schuster, R. et al. 2019. Optimizing the conservation of migratory species over their full annual cycle. - Nature Communications 10: 1754.

Schwartz, M. W. et al. 2018. Decision Support Frameworks and Tools for Conservation. - Conservation Letters 11: e12385.

Sullivan, B. L. et al. 2014. The eBird enterprise: an integrated approach to development and application of citizen science. - Biological Conservation 169: 31–40.

Underhill, L. G. 1994. Optimal and suboptimal reserve selection algorithms. - Biological Conservation 70: 85–87.

Venter, O. et al. 2014. Targeting Global Protected Area Expansion for Imperiled Biodiversity. - PLOS Biology 12: e1001891.

Wolsey, L. A. and Nemhauser, G. L. 1999. Integer and combinatorial optimization. - John Wiley & Sons.