

VARYING IMMIGRATION RATE

1.1 Assumptions of what has gone before

- TODO: link up these references! (and add recent papers into bib.)
- TODO: y-axis is wrong on relative abundance heat maps.
- TODO: rescale dynamics plots (reduce vertical dimension)
- Discussion of what we will now refer to as "default parameter values" (see table ??)
- Conclusion that there is something unrealistic about the immigration mechanism (rate too high)

[1]

1.2 Literature review for immigration

Possibly include here a summary of recent work on immigration, including IBT and meta-community theory. (Alternative is that this goes in introductory chapter.)

1.3 Motivation

One conclusion from chapter ?? is that the immigration rate, as given in the default parameter values (see table ??) is unrealistically high. The results presented in chapter ?? show that we are to expect no extinction of species, even at 90% HL, when running simulations with this immigration rate. This behaviour allowed us to explore community

responses to habitat loss that are not driven by changes in species richness. However for such a heavily impacted community to not exhibit local extinctions would be unusual in nature. This is may be considered an edge case - an open ecosystem with a strong influx of individuals from all species. Although the local habitat may be very close to total destruction the community is sustained by strong immigration from surrounding habitats. In reality such a strong and uniform rescue effect from immigration is unlikely due to spatial auto-correlation, differential dispersal rates and other effects (see discussion in section ??[references - York pollinator study]). (DISCUSS NUMBERS HERE R.E. RESCUE EFFECT?)

In this chapter we consider other scenarios by changing the immigration rate parameter. At one extreme we have the above scenario of high immigration, where extinction is impossible. At the other extreme we have closed communities with zero immigration. In this case there is no rescue effect from surrounding habitats and we may expect to see extinctions in response to habitat loss. Although a totally closed system does not exist in nature certain systems may come close. For example an island community that is a sufficiently distant from other land (see discussion on Island bio-geography theory) will have very low immigration rates, and systems that are effectively closed may also be artificially achieved in controlled situations (e.g. laboratory mesocosm).

For the default parameter values zero immigration results in the inevitable extinction of all non-plant species, in our simulated communities. Even for pristine habitats (0% HL) we do not see stable and persistent communities without some non-zero immigration. This result is demonstrated in chapter ??, where we explore factors contributing to stability and attempt to construct persistent closed-communities. For now we accept the general result that, under the default conditions, zero immigration results in community collapse. Therefore in this chapter we are interested in the regime between these two extremes of zero immigration, where we see many extinctions even at $HL = 0$, and high immigration where we see no extinctions even at $HL = 90$. In particular we look for some immigration value for which the communities are stable at low levels of habitat destruction, but where collapse is initiated as destruction is increased.

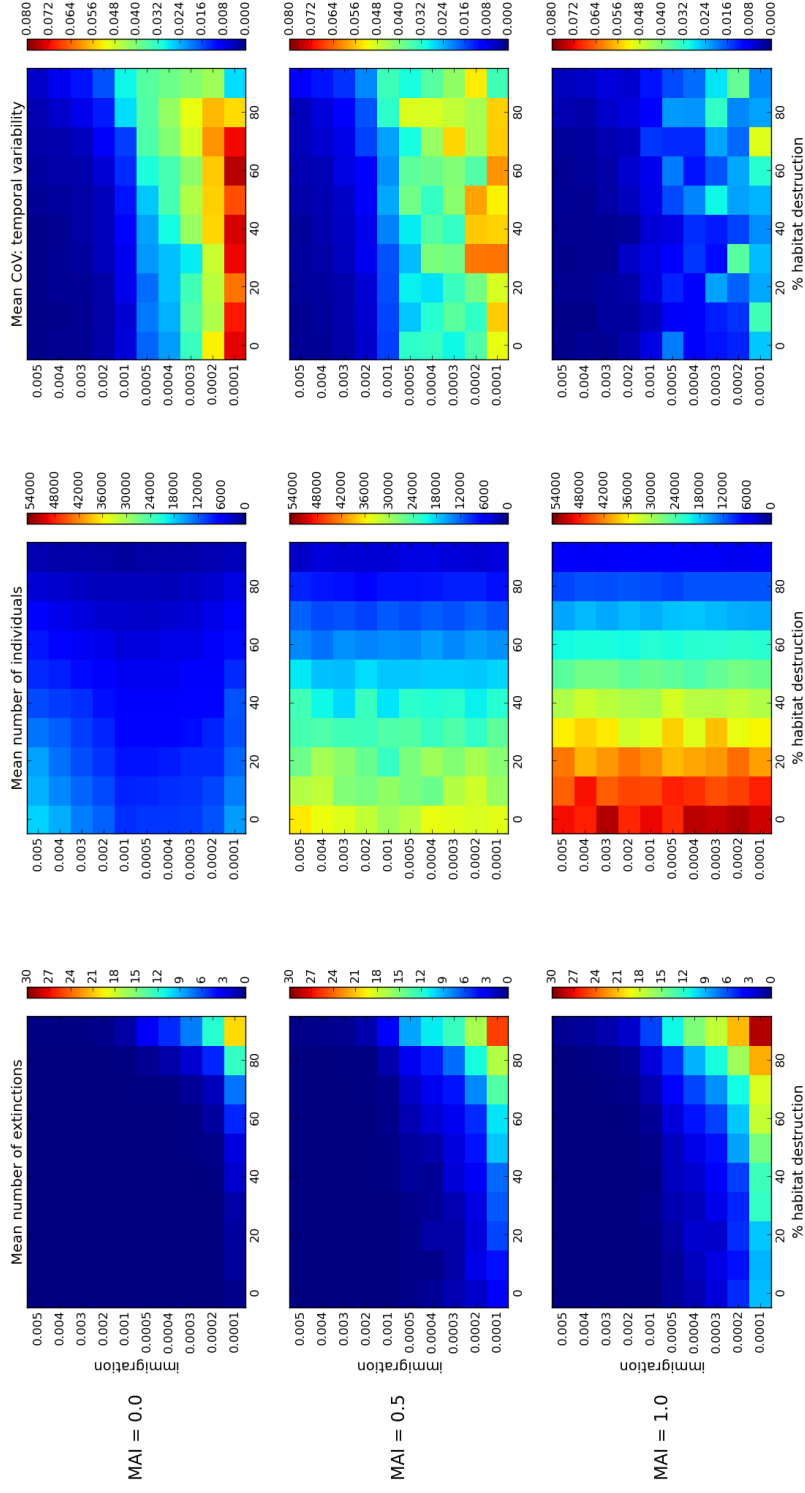


Figure 1.1: **Summary heat maps:** Each heat map shows the value of a certain response metric across a 2-dimensional slice of parameter space. The parameters varied are immigration rate (y-axis) and percentage habitat destruction *HL* (x-axis). Each row of plots corresponds to a different MAI ratio as indicated. To construct the heatmaps one hundred repeat simulations were run for each combination of parameter values, each simulation using a different underlying network. The mean value of the response metrics is taken over the hundred repeats. Therefore each pixel shows an estimate of the expectation value of the metric at those parameter values. The left column shows the expected number of species that are extinct at the end of a simulation; the central column shows the expected biomass (total number of individuals) at the end of a simulation; and the right column shows the expected temporal variability (coefficient of variation of total biomass) of the dynamics during the final thousand iterations of a simulation. The latter is used as a proxy for stability (see text).

1.4 Exploration of parameter space: habitat loss and immigration

Motivated by the above we explore a two dimensional slice of parameter space. The immigration rate (IR) and the level of habitat destruction (HL) are varied and one hundred repeat simulations are conducted at each pair value. As previously each repeat uses a different underlying network topology, generated by the same procedure as described in section ???. This procedure is repeated for three different MAI ratios: 0.0, 0.5 and 1.0. All other parameters are held constant at their default values.

In order to speed up the simulations certain metrics used in the previous analysis (chapter ??) are not calculated. Only two pieces of information are saved as output from these simulations: the underlying network structure and the abundance time-series for each species at each simulation iteration. By limiting the simulation output the scope for analysis is restricted but the parameter space can be explored in more detail (higher resolution, greater number of repeats). This 'first pass' scan of the parameter space allows us to construct a general picture of how the model behaves. It may also be used to identify regions of parameter space on which to focus further computational effort for e.g. spatial analysis.

The entire range of habitat destruction is explored from pristine landscape ($HL = 0\%$) to near total destruction ($HL = 90\%$). Ranges for the immigration rate were chosen based on previous simulations. Since $IM = 0.005$ is sufficiently high to prevent any extinctions, this was taken as the maximum of the range. Simulations using $IM = 0$ have already determined that this leads to community collapse, therefore these were not repeated. A value of $IM = 0.0001$ was heuristically selected as lower bound.

The results of these simulations can be concisely represented as heat maps over the region of parameter space explored. Figure 1.1 shows how the expected value of three summary metrics varies across this space. As was discussed in section ??, no species extinctions are expected for sufficiently high levels of immigration. As seen in the left column of figure 1.1 the expected number of extinctions is zero at high IR, across the range of HL values. It is found that reducing the IR leads to an increasing number of extinctions. At low IR extinctions are possible, even in pristine landscape. This fits the previous observation that zero immigration leads to high numbers of extinctions.

Increasing HL generally increases the number of expected extinctions. In the most extreme case of low immigration and high habitat destruction ($MAI = 1.0, HL = 90, IR = 0.0001$) an average of close to thirty extinctions may be expected. Although this expected loss of half of the species is fairly catastrophic, it does not guarantee total collapse of the community. The trophic constraints imposed in the food-web generation procedure ensure that at least 25% of species belong in the first (basal) trophic level. In practice it very rarely (quantify) reaches

above 30%. Therefore a loss of thirty species suggests that least 40% of the remaining species are non-basal¹. In other words, despite significant loss of species, there is some persistence in higher trophic levels. It is important to note that even at low IR species which go extinct during the simulation may be rescued. Therefore it is interesting to ask how much of this persistence in higher trophic levels is due immigration rescue effects. These could lead to a situation with a number of very rare species 'bubbling along' close to zero (becoming extinct and then being rescued), with a small core of more abundant species meaningfully engaged in interaction dynamics and demographic processes².

For all three MAI ratios there exists an IR where the expected number of extinctions is zero in pristine landscape, but increases with HL. So although the immigration rescue effect prevents total community collapse, we do have a situation where HL can initiate species extinctions. The IR at which this occurs is greater for high MAI ratio. This effect of MAI ratio on extinctions is general. On average we expect a greater number of extinctions for high MAI (1.0) than for low MAI ratio (0.0), all else being equal. At the lowest IR and with pristine habitat we may expect about one extinction with a MAI ratio of 0.0, compared to about ten extinctions with a MAI ratio of 1.0. This can possibly be explained by looking at the second column in figure 1.1. On average a higher MAI ratio leader to a greater total number of individuals at the end of a simulation³. This means that there are fewer empty landscape cells into which an individual may immigrate at random. This reduces the effective immigration rate and so weakens the rescue effect. If it is the case that there were previously rare species, only made viable by immigration, it is these that will be hit by this⁴.

The are interesting trends in the expected total biomass of a community. As noted communities with higher MAI ratio support more individuals on average. Also a high MAI ratio seems to make the total biomass of a community insensitive to IR. In the case of purely antagonistic communities (MAI=0.0) the greatest number of individuals is supported by those with either very low or very high IR. The latter is an intuitive result of births due to immigration supplementing births due to reproduction in the local community. This supplementary effect reduces with IR, leading to fewer individuals. The increase in total individuals at very low IR is harder to explain. We know that at zero immigration all non-plant species go extinct. So we may expect that, in this region of low immigration, non-plant species become increasingly rare. In an antagonistic community the plant species would benefit from the rarity of their animal consumers. Therefore we may expect that this increase in the biomass is accounted for by an increased abundance of plants⁵.

¹If all thirty species lost are non-basal we are left with 3/5 basal species to 2/5 non-basal.

²If this is the case it leads intuitively to the concept of 'viable' species, possibly more useful here than 'non-extinct'.

³Mechanism behind this?

⁴To determine if this is what is happening need to look at total abundances?

⁵This proposed mechanism may be working in reverse in the MAI=1.0 communities.

..This trend extends to the response to HL - as HL increases antagonistic communities with low and high levels of immigration maintain higher total biomasses than those with intermediate immigration rates. However we may suspect that the compositions of these communities may differ..INVESTIGATE.

- Importantly it should be noted that communities with higher MAI ratios support a greater number of individuals, on average, than those with low MAI ratio, regardless of immigration. This is especially visible at low levels of HL - considering the case of immigration = 0.005 : the expected number of individuals for a community with $MAI = 0.0$ is around 20,000, compared to around 50,000 for a community with $MAI = 1.0$. It is not as clear (from the heatmaps) whether this difference remains as significant at high levels of habitat destruction - INVESTIGATE THIS. We may conclude from this that, in some sense mutualism is better for the overall community than 'non-mutualism'. Is it that it is a better form of reproduction for the plants. Does it only benefit those species involved in the mutualistic interactions, such that we end up with communities dominated by mutualists (which out compete non-mutualists)? Or does mutualism benefit the wider community also?
- For higher MAI ratio the response of total community biomass to habitat loss does not appear to be mediated by immigration. In fact immigration rate appears to make little difference to the expected total biomass, for these hybrid antagonistic-mutualistic communities. (In the case of $MAI = 0.5$ there may be some remnant of the effect discussed above.) Again we may expect that community composition is changing, without noticeable effect on total biomass. It makes sense that mutualistic communities do not respond in the same way to a reduction in the number of non-plant species (if this is what is driving the previous pattern). It may be that the benefit due to the loss in animal consumers is balanced by the harm due to the loss in mutualistic partners. If this is the case we may expect a shift away from mutualistic dominance towards antagonistic dominance in the lower (basal?) trophic levels (or if not 'dominance' at least a shift in relative abundances.)
- The coefficient of variation (CoV) of the overall biomass is taken as a crude measure for the stability of the dynamics. In general we see a superposition of trends in the response of CoV to immigration and habitat loss. In general increasing habitat destruction, decreases the stability of the dynamics (for contiguous destruction - c.f. conflicting result for random destruction). Also decreasing the immigration values, decreases the stability of the dynamics in general. This second trend makes sense - since we know that there are stable co-existence equilibria for this model at high levels of immigration, and that there are not for zero immigration, it is expected that the system should pass through a region of instability in switching between these regimes (bifurcation?).

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- There is one exception to the above - which is that at very low immigration, or very high habitat destruction, the trend appears to be disrupted. More on this?
- Both trends in stability are expressed most strongly in antagonistic communities (case of $MIA = 0$), and are only weakly visible in pure mutualistic communities ($MAI = 0.0$). This suggests that mutualism is a stabilising influence, or at least dampens the destabilising effects of varying immigration and of habitat destruction. (This is also suggested by the above notes trends in total biomass.)

1.4.1 Example dynamics

Figure 1.2 illustrates the loss of stability in passing from a high to a low IR regime. The dynamics of three example antagonistic communities are depicted for each regime. The increase in temporal variability is greater for antagonistic communities than for those with mutualism (again we see mutualism to be a stabilising force), as is clear from the right-hand column of figure 1.1. In the left-hand panel of figure 1.2 we see that the total biomass of each community is fairly constant, after an initial transience. It appears that the system undergoes stochastic fluctuations about its stable equilibrium⁶. In the right-hand panel, which has very low IR, we see that this stability is reduced. It may be that the reduction in IR increases the time it takes for the system to reach equilibrium, or it may be that the effect of the stochastic fluctuations is amplified because of a reduction in the stability of the equilibrium.

Is the increased abundance, recorded in the heat maps, due to a shift in the equilibrium population or the oscillations (transient or otherwise?).

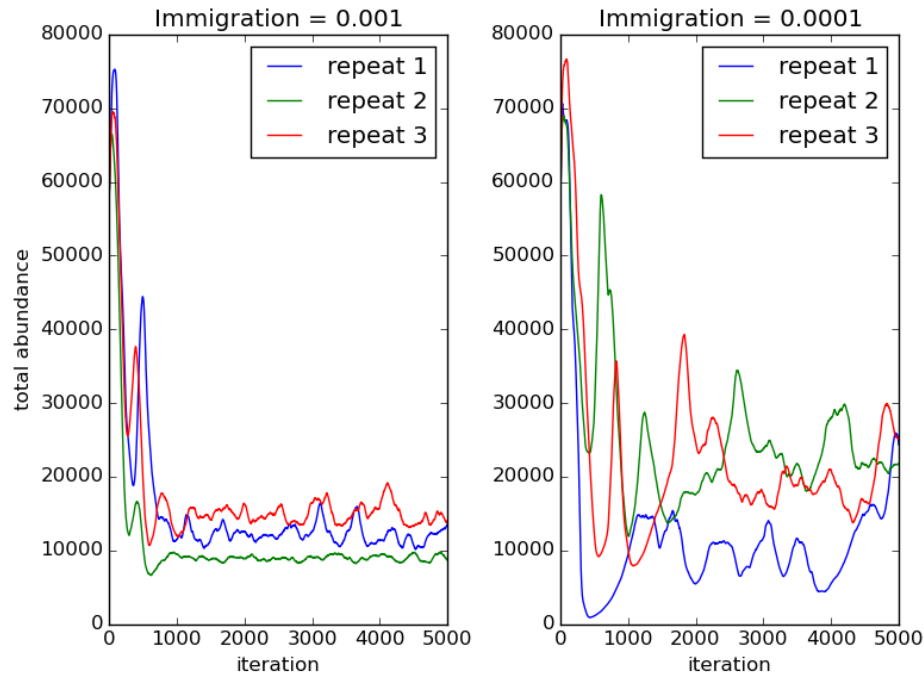
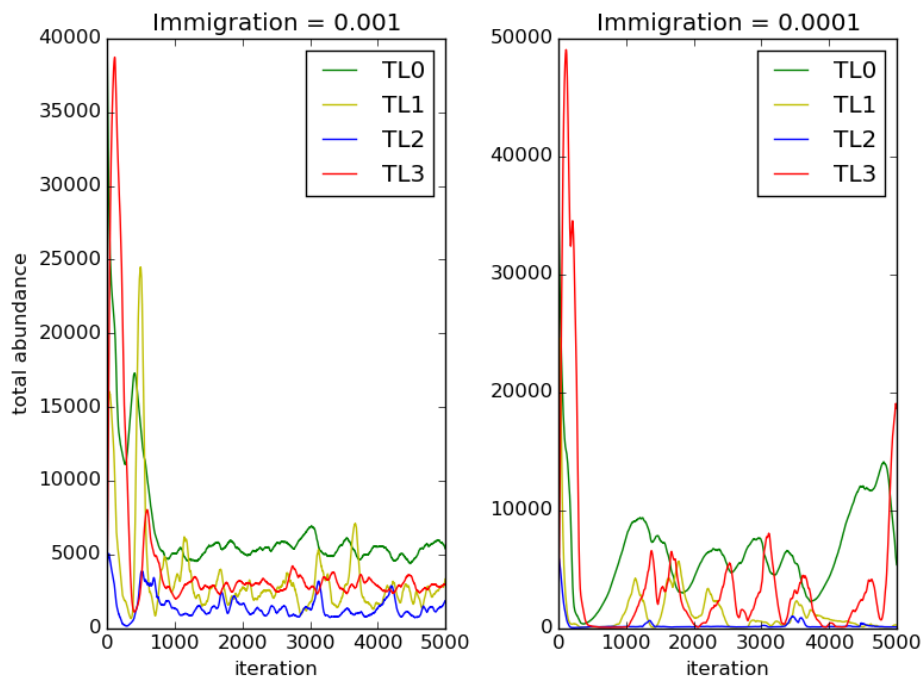


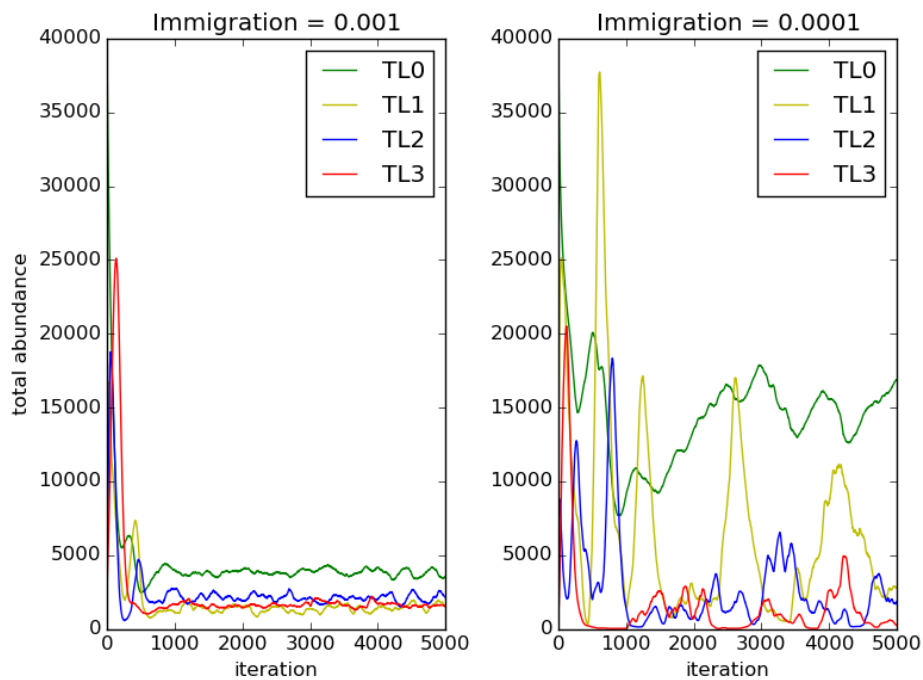
Figure 1.2: Temporal dynamics of the total biomass of communities over the course of six simulations. Each panels shows the dyanmics for three distinct siulations, each in a different colour. The left panels shows communities with a high immigration rate, and the right paenl for a low immigration rate. In all cases there is no habitat destruction $HL = 0$.

⁶Test for this?

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Dynamics from four individual simulation runs, with biomasses aggregated by trophic level. Each panel represents the dynamics of a single simulation run. The coloured lines represent the temporal dynamics of the biomass of each trophic level, as indicated in the legends. In all cases there is **no habitat destruction** ($HL = 0$). Two immigration scenarios are presented. **Left column: high immigration. Right column: low immigration.**



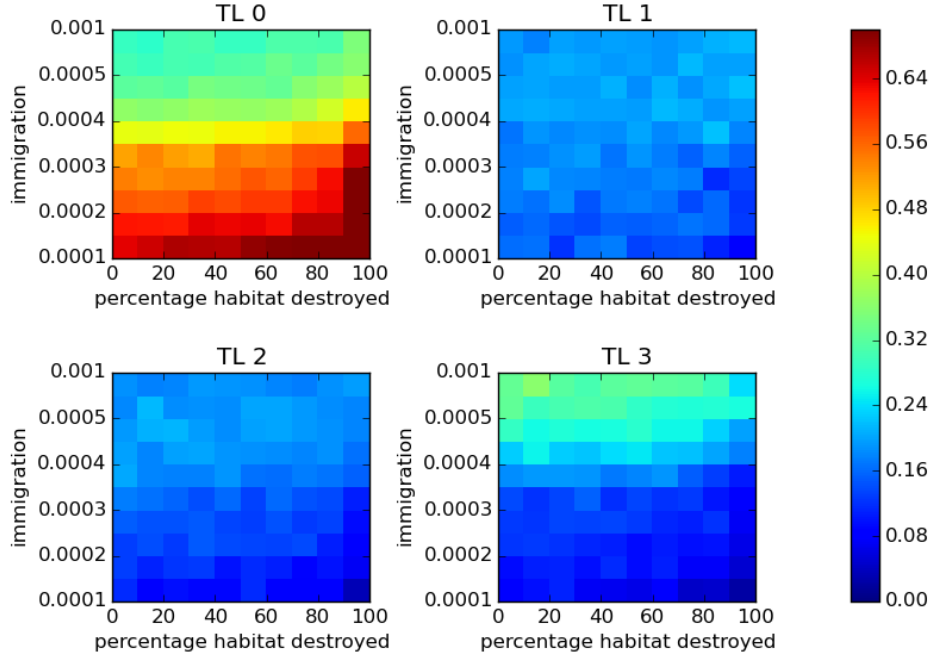
Figures 1.4.1 show example dynamics by trophic level, for both antagonistic and mutualistic communities, in the high and low IR regimes.

It is important to note the large amplitude fluctuations in biomass that are induced by the reduction in IR. **Dynamics such as these bring into question the validity of the results, which take a snapshot of species abundances at the end of the simulation, under the assumption of steady state.**

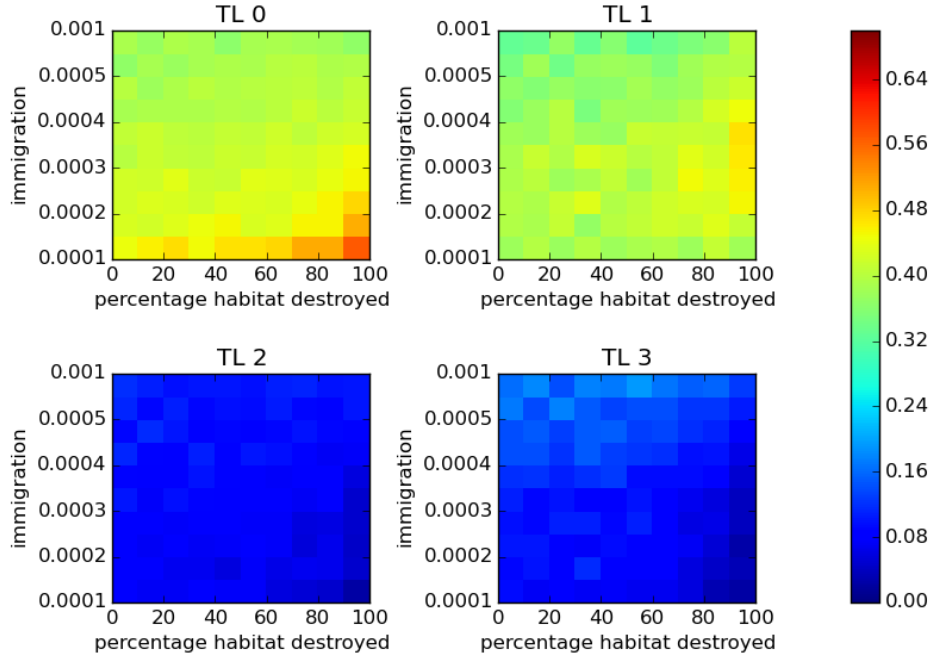
1.4.2 Relative abundances

Figure 1.4.2 shows the relative abundance of each trophic level for antagonistic and mutualistic communities, across the parameter space. For purely antagonistic communities the proportion of individuals in each trophic level varies strongly with IR and weakly with HL. The mutualistic communities (MAI=1.0) show much less variation in their trophic composition. Of particular interest is that, with very little disturbance, antagonistic communities become dominated by plants. This does not happen for mutualistic communities, except at very high levels of disturbance. This again suggests that mutualism is a stabilising force, and that it benefits all species in the community not just those that engage in mutualistic interactions - or maybe not judging by the following figure, and the fact that TL2 and TL3 are consistently low.

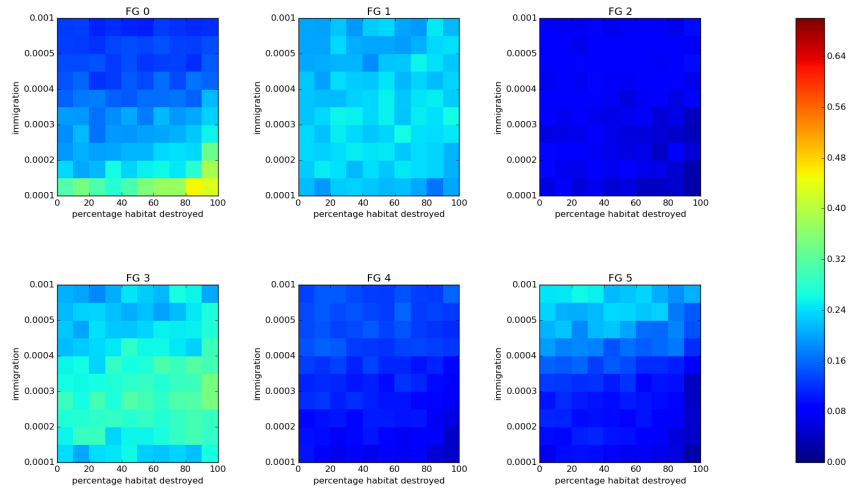
1.4. EXPLORATION OF PARAMETER SPACE: HABITAT LOSS AND IMMIGRATION



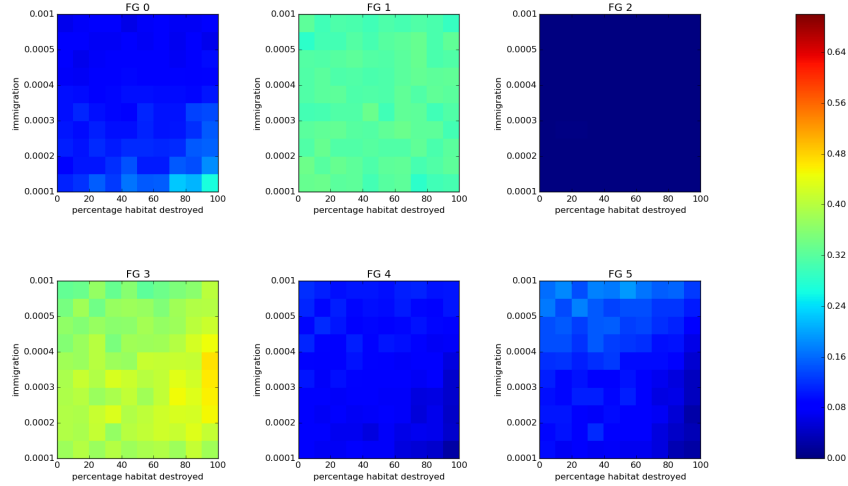
The relative abundance of species belonging to each of the four trophic levels. **Above: MAI = 0.0. Below: MAI = 1.0.** Each pixel on the heat maps corresponds to an average over one hundred repeat simulations at those parameter values. The abundances are measured at the end of each simulation.



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The relative abundance of species belonging to each of the six functional groups. **Above: MAI = 0.5. Below: MAI = 1.0.** Each pixel on the heat maps corresponds to an average over one hundred repeat simulations at those parameter values. The abundances are measured at the end of each simulation.



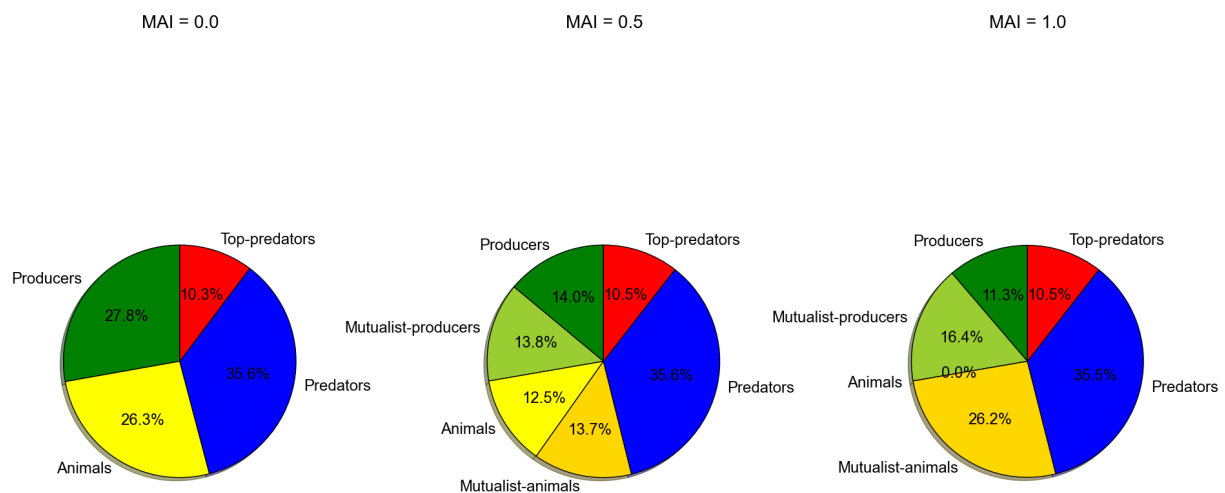


Figure 1.3: These diagrams indicate the mean number of species belonging to each functional group for the three MAI ratios in consideration. The results are averaged over one thousand simulations with the given MAI ratio, selected from the total ensemble of simulations (used to generate e.g. fig 1.1). The numbers are independent all simulations parameters other than those used in the construction of the initial network (as outlined in section ??) i.e. connectance, number of species, MAI ratio, and trophic constraints.

1.5 Further investigation of these trends - RADS, dynamics

TODO:

- plot RADS (single run with TL colours, average with error bars) and RAS (relative abundances of trophic levels) to demonstrate some of the above arguments - are they correct in their interpretation?
- more heatmaps - relative abundances by trophic level/functional group
- plot some example dynamics to demonstrate changes in stability (use total biomass and by trophic level. Maybe one with all species.)

1.6 Points for discussion

A comparison of the relative merits of being a mutualist versus a non-mutualist is worthwhile. Importantly it must be remembered that mutualistic interactions are also trophic interactions. In our case, energy is transferred from producer to animal. So there is some loss/detriment to the producer. However this loss is balanced or made up for by the benefit of reproduction. An assumption, in our model at least, is that the mutualistic method of reproduction must be 'better' in some sense than the non-mutualistic alternative in order to justify the cost of the trophic interaction involved. In the context of our model this means..

Table ?? shows the default parameter values used for most simulations. Lets consider the potential benefit of switching a single herbivorous link for a mutualistic link, for either party. If the plant is a non-mutualist it must impart 20% of its energy to the offspring when reproducing (this happens with a probability of 0.01 on each iteration). It is also subject to lose 70% when it is encountered by this herbivore. If it were to switch this herbivorous link for a mutualistic link it would only lose 25% of its energy in the interaction, and it would pass on a seed that is almost guaranteed (really?) to generate an offspring. Therefore the cost of reproducing is slightly increased for a mutualist, but the cost of interacting with an individual from the trophic level above is dramatically reduced. There is an additional benefit that the mutualistic reproduction can occur over a greater distance. The net gain loss of this change depends on the probability/rate of interaction. We should investigate this, however the results suggest that being a mutualist is of significant benefit to plants - mutualistic plants are relatively more abundant than non-mutualistic ones, except in the case of high habitat loss or low immigration (when there are few enough mutualistic partners that interactions become infrequent?)

For animals there is no cost to carrying and spawning the seed of their mutualistic partners. The only change in the switching of mutualistic links is the amount of energy that they receive from the interaction. During a herbivorous interaction, the herbivore takes 70%

of the plant's energy, and assimilates it with an efficiency of 80%. Therefore it obtains 60% of the plant's energy. During a mutualistic interaction the animal-mutualist takes and assimilates 25% of the plant's energy. Therefore on an interaction by interaction basis there is a negative trade off for an animal in switching its link to mutualistic. However there may be an emergent benefit in that this type of interaction is much better for plants, therefore increasing the plant biomass and therefore indirectly benefiting animal (mutualists and non-mutualists?) due to the increased frequency of interactions (density of plants).

- Do mutualistic plants only reproduce mutualistically? (almost certain yes)
- Does MATINGRESOURCE apply to mutualistic interactions?
- Why are top predators able to eat plants?!! Does omnivory trade off apply?? - too many individuals belong to TL3 in general.

Mutualism is general stabilises dynamics, and leads to communities with more realistic biomass pyramids - i.e. dominated by the first two trophic levels, with fewer individuals in TL2/3.

1.7 Habitat loss with low immigration

1.8 Questions for Alan

- Please not to look at correcting this document word by word, but rather to look at the general flow and structure of the discussion.
- Do you know about sideways whole page figures?!
- Since the dynamics do not necessarily reach steady state should I re-do analysis with average over a time window?
- Can I use "we"??
- Tense?
- Figure 1.1 summary heatmaps: too much information in one figure (feels that way from discussion).
- TODO: linear interactions? (should it be that freq/predabun should be linear in preyabun? Same gradient across species?) If so we should be able to fit a GLV, if not via Timme, then via repeated simulations and some numerical optimisation.

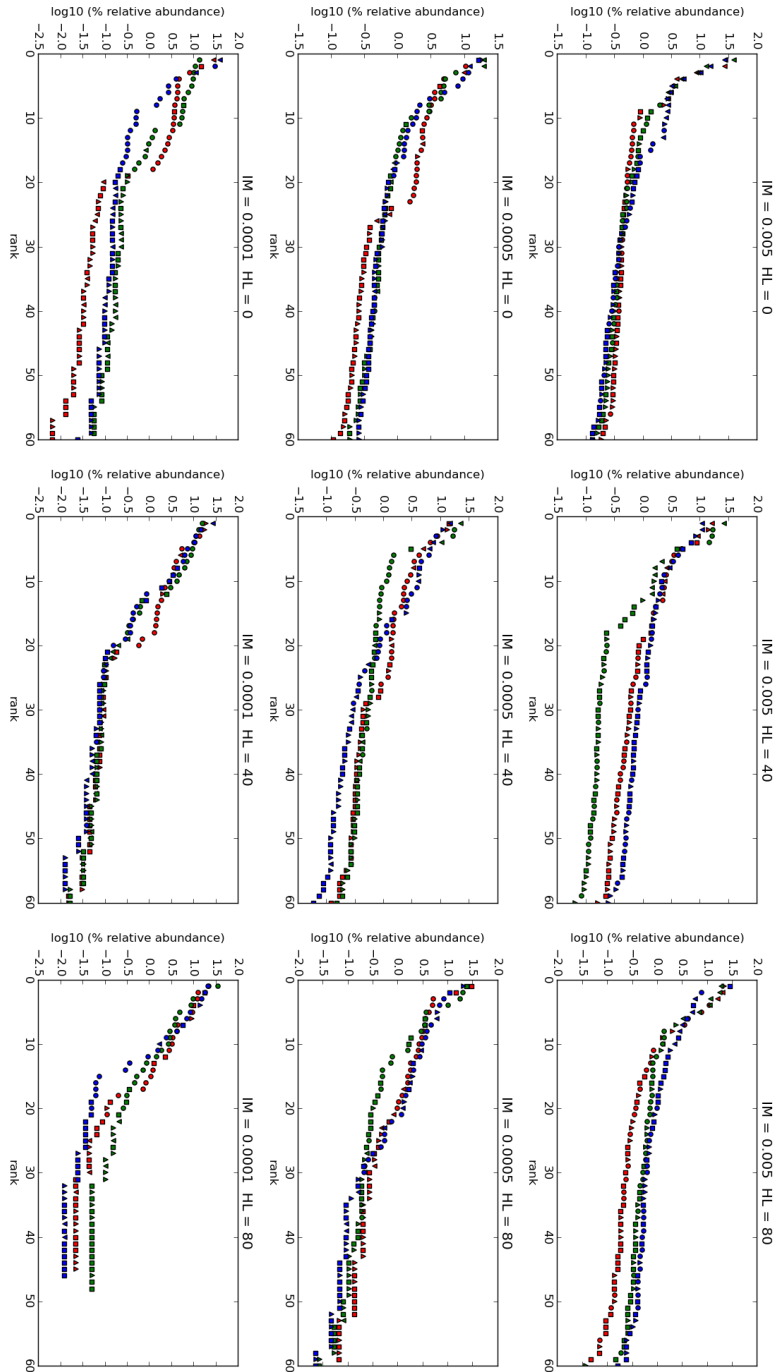


Figure 1.4: **Rank abundance distributions** for individual simulation runs, for nine different pair values of immigration rate and habitat destruction. Each distribution is for a single community at the end of an individual simulation run. The different colours correspond to different MAI ratios: red = 0.0; green = 0.5; blue = 1.0. And the different symbols correspond to different trophic levels: circle = 0; upwards triangle = 1; square = 2; downwards triangle = 3.

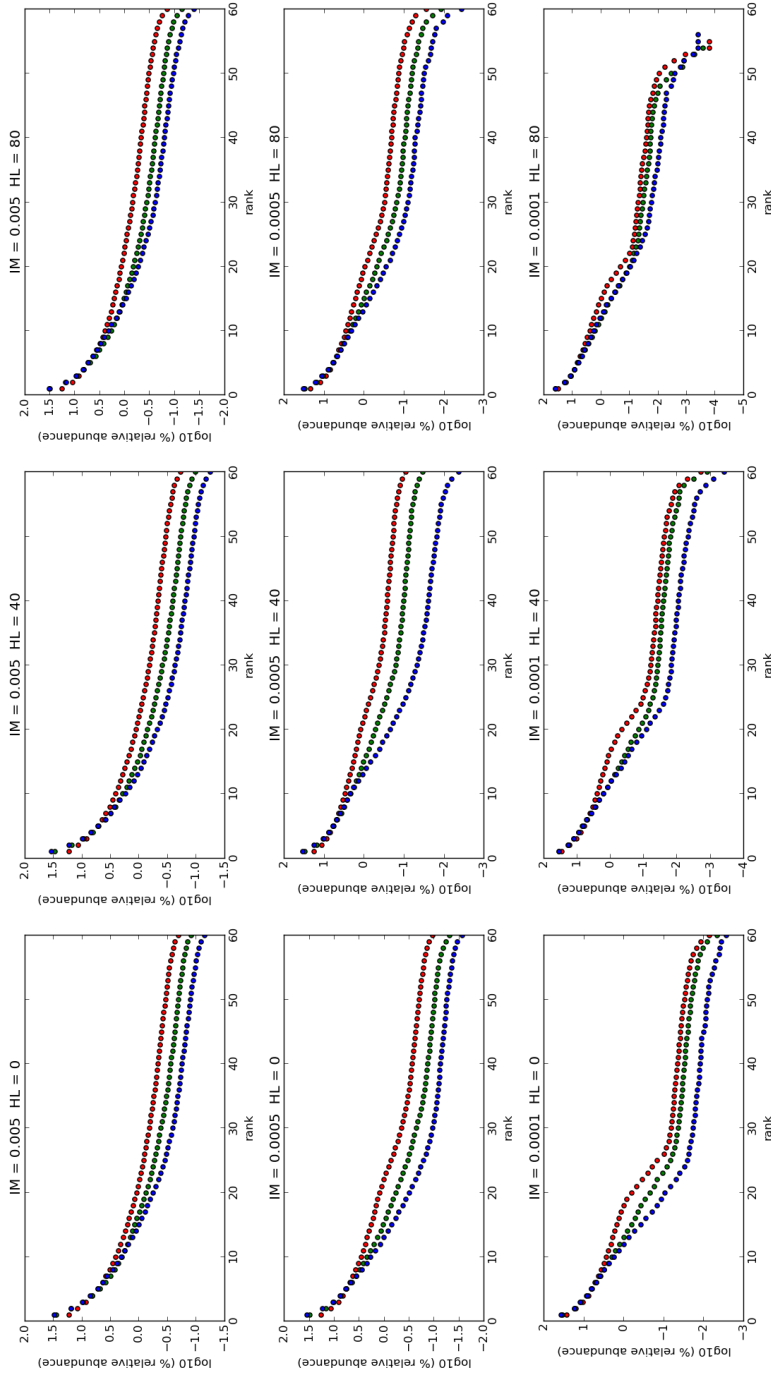


Figure 1.5: **Average rank abundance distributions** over one hundred simulation runs, for nine different pair values of immigration rate and habitat destruction. Each distribution is calculated using the mean relative abundance of the ranked species, averaged over the final abundances of one hundred repeat simulations. The different colours correspond to different MAI ratios: red = 0.0; green = 0.5; blue = 1.0.

BIBLIOGRAPHY

- [1] W. J. RIPPLE AND R. L. BESCHTA, *Trophic cascades in yellowstone: The first 15 years after wolf reintroduction*, Biological Conservation, 145 (2012), pp. 205–213.

