

Article

A double-sampling extension of the German National Forest Inventory for design-based small area estimation on forest district levels

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Academic Editor: name

Version 13th May 2018 submitted to *Remote Sens.*

Abstract: The German National Forest Inventory consists of a systematic grid of permanent sample plots and provides a reliable evidence-based assessment of the state and the development of Germany's forests on national and federal state level in a 10 year interval. However, the data have yet been scarcely used for estimation on smaller management levels such as forest districts due to insufficient sample sizes within the area of interests and the implied large estimation errors. In this study, we present a double-sampling extension to the existing German National Forest Inventory (NFI) that allows for the application of recently developed design-based small area regression estimators. We illustrate the implementation of the estimation procedure and evaluate its potential for future large scale operational application by the example of timber volume estimation on two small scale management levels (45 and 405 forest district units respectively) over the entire area of the federal German state of Rhineland-Palatinate. An airborne laserscanning (ALS) derived canopy height model and a tree species classification map based on satellite data were used as auxiliary data in an ordinary least square regression model to produce the timber volume predictions. The results support that the suggested double-sampling procedure can substantially increase estimation precision on both management levels: the two-phase estimators were able to reduce the variance of the SRS estimator by 43% and 25% on average for the two management levels respectively.

Keywords: National forest inventory, small area estimation, forest districts, double sampling for regression within strata, cluster sampling, canopy height model, tree species classification

1. Introduction

The German National Forest Inventory (NFI) provides reliable evidence-based and accurate information of the current state and the development of Germany's forest over time. The NFI thereby has the responsibility to satisfy various information needs including reporting to public and state forestry administrations, wood-based industries and the public on the national level, as well as to the Food and Agriculture Organization of the United Nations (FAO) and to the United Nations Framework Convention on Climate Change (UNFCCC) on the international level [1]. The current design of the German NFI rests solely upon a terrestrial cluster inventory that is carried out at sample locations systematically distributed over the entire forest state area of Germany. In order to cover a large area of 114'191 ha [2], the sample size has been specifically chosen to satisfy high estimation accuracies for forest attributes on the national and federal state levels. However, sample sizes often drop dramatically when entering spatial units below the federal state level. This is particularly true for

31 forest management levels such as forest districts for which the estimation uncertainties turn out to be
32 unacceptably large due to the very limited number of sample plots within these units. For this reason,
33 the German NFI data have not yet been extensively incorporated into operational planning on forest
34 district management levels. In most German federal states, management strategies are thus still based
35 on expert judgments from time-consuming standwise inventories (SFI), which are prone to systematic
36 deviations [3] and do not provide any measure of uncertainty.

37 Some German federal states, such as Lower Saxony, have approached this problem by establishing
38 a regional Forest District Inventory (FDI) with a much higher sampling density than used by the NFI
39 in order to scientifically base their regional management strategies on quantitative and accurate
40 information [4]. However, such FDIs are cost-intensive and, facing increasing restrictions in budget
41 and staff resources, there has been a need for more cost-efficient inventory methods [5]. One method
42 which has proven to be efficient is double- or two-phase sampling [6–9]. Double-sampling incorporates
43 less expensive auxiliary information and can be used to either increase estimation precision under a
44 fixed terrestrial sample size, or maintain estimation precision under reduced terrestrial sample size.
45 Double-sampling procedures have already been used for stratification in the FDI of Lower Saxony
46 [10], and Grafström *et al.* [11] illustrated how to use the auxiliary information to determine optimised
47 balanced terrestrial sample designs. Recent studies have extended double-sampling to triple-sampling
48 estimation methods using auxiliary information derived at two different sampling intensities. An
49 example can be found in von Lüpke *et al.* [12] who illustrated an extension of the existing two-phase
50 FDI of Lower Saxony to a three-phase design that uses updates of past inventory data as additional
51 auxiliary information and allows for a significant reduction of the terrestrial sample size in intermediate
52 inventories. Another example is Massey *et al.* [13] who developed a triple-sampling extension based
53 on the ideas of Mandallaz [14] for the Swiss NFI that can significantly reduce the increase in estimation
54 uncertainty caused by the new annual inventory design.

55 Two-phase and three-phase samplings techniques have also been applied to small area estimation
56 (SAE). SAE techniques address the situation where the number of samples within a subunit, or small
57 area (SA), of the entire sampling frame is too small to provide reliable estimates for that unit. A broad
58 range of SA estimators used in forest inventories [8] originally comes from official statistics. One
59 such method that is commonly applied is known as indirect estimation [15], where statistical models
60 are used to convert auxiliary information into predictions of the target variable that is rarely or not
61 observed in the small area. These models are trained using data from outside the small area in order
62 to "borrow strength" from areas where information is available. Of numerous applications of SAE in
63 forestry [16–19], most use unit-level models, i.e. the inventory plot is the unit of the response variable
64 in the training data used for the model fit. Such unit-level models have been intensively investigated
65 for timber volume estimation using various remote sensing auxiliary data [20,21]. Other studies have
66 investigated area-level models, where the auxiliary information is only provided on the SA level [22].
67 Some studies have illustrated that even NFI data derived under low sampling densities can still be
68 used to provide acceptable precision of small area estimates on much smaller management levels. One
69 example is Breidenbach and Astrup [16] who used data from the Norwegian NFI to make small area
70 estimation for standing timber volume for 14 municipalities where the number of NFI samples within
71 these areas were between 1 and 35. The estimation errors under the applied model-dependent and
72 design-based small area estimators turned to be markedly smaller than under the standard one-phase
73 estimator. Another example is Magnussen *et al.* [23] who recently used the Swiss NFI data to estimate
74 timber volume within 108 Swiss forest districts with sample sizes between 9 and 206. Similar studies
75 using German NFI data for small area estimation have been lacking.

76 The objective of this study was to investigate whether the application of latest design-based small
77 area estimation methods allow to use the German NFI data to produce estimates of acceptable precision
78 on two forest district levels. The methods were tested in the German federal state Rhineland-Palatinate.
79 Three types of model-assisted design-based small area regression estimators were used to derive point
80 and variance estimates of mean standing timber volume for 45 and 405 forest management units on

81 the two respective district levels. The SA-estimators we considered were the *pseudo-small*, *extended*
 82 *pseudo-synthetic* and the *pseudo-synthetic* design-based small area estimator suggested by Mandallaz
 83 [24] and Mandallaz *et al.* [19]. Auxiliary data consisted of a canopy height model (CHM) obtained
 84 from a countrywide airborne laser scanning (ALS) and a tree species classification map to be used for
 85 regression within tree species strata. The estimation precisions were compared to those obtained by
 86 the standard one-phase estimator for cluster sampling under simple random sampling. The chosen
 87 double-sampling estimators were selected for several reasons: (i) the design-based framework relaxes
 88 dependencies on the regression model assumptions which seemed appropriate facing severe quality
 89 restrictions in the ALS data; (ii) the estimators can be used with *non-exhaustive*, i.e. non wall-to-wall,
 90 auxiliary information; (iii) all estimators are explicitly formulated for cluster sampling which has
 91 not yet been the case for frequently used model-dependent estimators; and (iv) the asymptotically
 92 unbiased g-weight variance partially accounts for estimating the regression coefficients on the same
 93 sample used for estimation (*internal model approach*) and is also robust under heteroscedasticity of
 94 the model residuals. The results from this study were considered to provide valuable information
 95 about the potential of the suggested small area estimation procedure and the incorporated auxiliary
 96 information for future operational large scale application.

97 2. Terrestrial sampling design of the German NFI

98 The German NFI is a periodic inventory that is carried out every 10 years over the entire forest area
 99 of Germany. The most recent inventory (BWI3) was conducted in 2011 and 2012. While information
 100 was originally gathered on a systematic 4x4 km grid, some federal states such as Rhineland-Palatinate
 101 have switched to a densified 2x2 km grid. The German NFI uses a cluster sampling design, which
 102 means that a sample unit consists of at most four sample locations (also referred to as *sample plots*)
 103 that are arranged in a square, called *cluster*, with a side length of 150 metres. The number of plots per
 104 cluster can vary between 1 and 4 depending on forest/non-forest decisions by the field crews on the
 105 individual plot level [25]. In the field survey of the BWI3, sample trees for timber volume estimation
 106 are selected according to the angle count sampling technique [26], using a basal area factor (BAF) of 4
 107 that is respectively adjusted for sample trees at the forest boundary by a geometric intersection of the
 108 boundary transect with the individual tree's inclusion circle [25]. A further inventory threshold for a
 109 tree to be recorded is a diameter at breast height (DBH) of at least 7 cm. For each sample tree that is
 110 selected by this procedure, the DBH, the absolute tree height, the tree diameter at 7 m (D7) and the
 111 tree species is measured and used to estimate the volume at the tree level. These volume estimates are
 112 based on the application of tree species specific taper curves that are adjusted to the set of diameters
 113 and corresponding height measurements taken from the respective sample tree [27].

114 3. Double sampling in the infinite population approach

115 3.0.1. One- and Two-Phase Sampling in the Infinite Population Approach

116 The estimators used in this study have been proposed by [19,24] and derive their mathematical
 117 properties under the so-called infinite population approach. Therefore, we shall first provide a short
 118 introduction into this general estimation framework. We start by assuming that the population P of
 119 trees $i \in 1, 2, \dots, N$ within a forest of interest F is exactly defined, and each tree i has a response variable
 120 Y_i (e.g. its timber volume) that can be used to define the population mean Y (e.g., the average timber
 121 volume per unit area) over F . Since a full census of all tree population individuals is almost never
 122 feasible, Y has to be estimated based on a sample. In the infinite population approach this sample is a
 123 set of points or locations x distributed independently and uniformly over the set of all possible points
 124 in F . Each point x has an associated local density $Y(x)$ (e.g., the timber volume per unit area) whose
 125 spatial distribution is given by a fixed (i.e. non stochastic) piecewise constant function. The population
 126 mean Y is mathematically equivalent to the integral of the local density function surface divided by
 127 the surface area of F , $\lambda(F)$, i.e. $Y = \frac{1}{N} \sum_{i=1}^N Y_i = \frac{1}{\lambda(F)} \int_F Y(x) dx$, and thus the population mean Y

corresponds to a spatial mean. Since the actual local density function is unobserved in its entirety, one estimates Y by taking a sample s_2 consisting of n_2 points and measuring each of their respective local densities. This sampling procedure is often referred to as *one-phase sampling* (OPS) and s_2 is referred to as the terrestrial inventory. In contrast to the one-phase approach, *two-phase* or *double-sampling* procedures use information from two nested samples (phases). Practically speaking, the terrestrial inventory s_2 is embedded in a large phase s_1 comprising n_1 sample locations that each provide a set of explanatory variables described by the column vector $\mathbf{Z}(x) = (z(x)_1, z(x)_2, \dots, z(x)_p)^\top$ at each point $x \in s_1$. These explanatory variables are derived from auxiliary information that is available in high quantity within the forest F . For every $x \in s_1$, $\mathbf{Z}(x)$ is transformed into a prediction $\hat{Y}(x)$ of $Y(x)$ using the choice of some prediction model. The basic idea of this method is to boost the sample size by providing a large sample of less precise but cheaper predictions of $Y(x)$ in s_1 and to correct any possible model bias, i.e., $\mathbb{E}(Y(x) - \hat{Y}(x))$, using the subsample of terrestrial inventory units where the value of $Y(x)$ is observed. In this context, it is also important to note that the response and auxiliary variables are assumed to be error-free and the resulting errors for the point estimates reflect only the uncertainty due to sampling.

4. Estimators

4.1. Design-based one-phase estimator for cluster sampling (SRS)

The one-phase estimator for cluster sampling (SRS) constitutes the *status quo* that is currently applied under the existing one-phase sampling design of the German NFI in order to obtain point and variance estimates for the mean timber volume of a given estimation unit. In order to provide all estimators in the infinite population framework and ensure a consistent terminology with the two-phase estimators in Section 4.2, we will introduce the SRS estimator that is applied in the BWI3 algorithms [28] in the form given in Mandallaz [9] and Mandallaz *et al.* [29].

In order to calculate the local density $Y_c(x)$ at the cluster level, a cluster is defined as consisting of M sample locations (in the BWI3, we have $M = 4$) where $M - 1$ sample locations x_2, \dots, x_M are created close to the cluster origin x_1 by adding a fixed set of spatial vectors e_2, \dots, e_M to x_1 . The actual number of plots per cluster, $M(x)$, is a random variable due to the uniform distribution of x_l ($l = 1, \dots, M$) in the forest F and to the forest/non-forest decision for each sample location x_l :

$$M(x) = \sum_{l=1}^M I_F(x_l) \quad \text{where} \quad I_F(x_l) = \begin{cases} 1 & \text{if } x_l \in F \\ 0 & \text{if } x_l \notin F \end{cases} \quad (1)$$

The local density on cluster level $Y_c(x)$, which is in our case the timber volume per hectare, is then defined as the average of the individual sample plot densities $Y(x_l)$:

$$Y_c(x) = \frac{\sum_{l=1}^M I_F(x_l) Y(x_l)}{M(x)} \quad (2)$$

The local density $Y(x_l)$ on individual sample plot level was calculated according to the description in Mandallaz [9], which can be rewritten for angle-count sampling technique applied in the BWI3. The general form of $Y(x)$ in Mandallaz [9] is given as the Horwitz-Thompson estimator

$$Y(x_l) = \sum_{i \in s_2(x_l)} \frac{Y_i}{\pi_i \lambda(F)} \quad (3)$$

where Y_i is in our case the timber volume of the tree i recorded at sample location x in m^3 . Each tree has an inclusion probability π_i that is well defined as the proportion of its inclusion circle area $\lambda(K_i)$ within the forest area $\lambda(F)$, i.e. via their geometric intersection:

$$\pi_i = \frac{\lambda(K_i \cap F)}{\lambda(F)} \quad (4)$$

The radius R_i of the tree's inclusion circle K_i is given by $R_i = DBH_i/cf_{i,corr}$ (also referred to as *limiting distance*), where $cf_{i,corr}$ is the original counting factor cf corrected for potential boundary effects at the forest border. In case of angle-count sampling, we can rewrite π_i as

$$\pi_i = \frac{G_i}{cf_{i,corr}\lambda(F)} \quad (5)$$

since the intersection area $\lambda(K_i \cap F)/\lambda(F)$ can be expressed using the trees basal area G_i (in m^2) and the corrected counting factor:

$$\lambda(K_i \cap F) = \frac{G_i}{cf_{i,corr}} \quad \text{where} \quad cf_{i,corr} = cf \frac{\lambda(K_i)}{\lambda(K_i \cap F)} \quad (6)$$

Eq. 5 in Eq. 3 yields the rewritten form of $Y(x_l)$ for angle count sampling that conforms to the definition used in the BWI3 algorithms [28]:

$$Y(x_l) = \sum_{i \in s_2(x_l)} \frac{cf_{i,corr} Y_i}{G_i} = \sum_{i \in s_2(x_l)} nha_i Y_i \quad (7)$$

where nha_i is the number of trees per hectare represented by tree i . The local densities on cluster level can then be used to derive the estimated spatial mean \hat{Y}_c and its estimated variance $\hat{\mathbb{V}}(\hat{Y}_c)$ for any given spatial unit for which $n_2 \geq 2$ (n_2 denoting the number of clusters):

$$\hat{Y}_c = \frac{\sum_{x \in s_2} M(x) Y_c(x)}{\sum_{x \in s_2} M(x)} \quad (8a)$$

$$\hat{\mathbb{V}}(\hat{Y}_c) = \frac{1}{n_2(n_2 - 1)} \sum_{x \in s_2} \left(\frac{M(x)}{\bar{M}_2} \right)^2 (Y_c(x) - \hat{Y}_c)^2 \quad (8b)$$

with $\bar{M}_2 = \frac{\sum_{x \in s_2} M(x)}{n_2}$.

4.2. Design-based small area regression estimators for cluster sampling

All three considered small area estimators use ordinary least square (OLS) regression models to produce predictions of the local density $Y_c(x)$ directly on the cluster level c . We consider the internal model approach, where the estimators take into account that the regression coefficients on the cluster level were fitted using the same sample used for estimation. To apply this to small area estimation, the vector of estimated regression coefficients on the cluster level is found by "borrowing strength" from the entire terrestrial sample s_2 of the current inventory:

$$\hat{\beta}_{c,s_2} = \mathbf{A}_{c,s_2}^{-1} \left(\frac{1}{n_2} \sum_{x \in s_2} M(x) Y_c(x) \mathbf{Z}_c(x) \right) \quad (9a)$$

$$\mathbf{A}_{c,s_2} = \frac{1}{n_2} \sum_{x \in s_2} M(x) \mathbf{Z}_c(x) \mathbf{Z}_c^\top(x) \quad (9b)$$

$\mathbf{Z}_c(x)$ is the vector of explanatory variables on the cluster level, which is calculated as the weighted average of the explanatory variables $\mathbf{Z}(x_l)$ on the individual plot levels x_1, \dots, x_l (Eq. 10). The weight $w(x_l)$ is the proportion of the extraction area (support) within the forest F used to derive the explanatory variables from the raw auxiliary information.

$$\mathbf{Z}_c(x) = \frac{\sum_{l=1}^M I_F(x_l) w(x_l) \mathbf{Z}(x_l)}{\sum_{l=1}^M I_F(x_l) w(x_l)} \quad (10)$$

¹⁸⁶ The estimated design-based variance-covariance matrix $\hat{\Sigma}_{\hat{\beta}_{c,s_2}}$ accounts for the fact that the regression
¹⁸⁷ model is internal and reflects the sampling variability that occurs when estimating the regression
¹⁸⁸ coefficients on the realized sample s_2 . It is defined as

$$\hat{\Sigma}_{\hat{\beta}_{c,s_2}} = \mathbf{A}_{c,s_2}^{-1} \left(\frac{1}{n_2^2} \sum_{x \in s_2} M^2(x) \hat{R}_c^2(x) \mathbf{Z}_c(x) \mathbf{Z}_c^\top(x) \right) \mathbf{A}_{c,s_2}^{-1} \quad (11)$$

¹⁸⁹ with

$$\hat{R}_c = Y_c(x) - \mathbf{Z}_c^\top(x) \hat{\beta}_{c,s_2} = Y_c(x) - \hat{Y}_c(x) \quad (12)$$

¹⁹⁰ being the empirical model residuals at the cluster level, which by construction of OLS satisfy the
¹⁹¹ important zero mean residual property, i.e. $\frac{\sum_{x \in s_2} M(x) \hat{R}_c(x)}{\sum_{x \in s_2} M(x)} = 0$.

¹⁹²
¹⁹³ In the following, we will give a short description of each small area estimator and refer to
¹⁹⁴ Mandallaz *et al.* [19], Mandallaz [24], Mandallaz *et al.* [29] if the reader requires additional details or
¹⁹⁵ proofs. The estimators have also been implemented in the R-package *forestinventory* [30] which was
¹⁹⁶ used to compute all estimates in this study.

¹⁹⁷

¹⁹⁸ 4.2.1. Pseudo Small Area Estimator (PSMALL)

¹⁹⁹ All point information used for small area estimation is now restricted to that available at the
²⁰⁰ sample locations $s_{1,G}$ or $s_{2,G}$ in the small area G , with exception of $\hat{\beta}_{c,s_2}$ and $\hat{\Sigma}_{\hat{\beta}_{c,s_2}}$ which are always
²⁰¹ based on the entire sample s_2 . We thus first define the following quantities on the small area level:

$$\hat{\mathbf{Z}}_{c,G} = \frac{\sum_{x \in s_{1,G}} M_G(x) \mathbf{Z}_{c,G}(x)}{\sum_{x \in s_{1,G}} M_G(x)} \quad \text{where} \quad \mathbf{Z}_{c,G}(x) = \frac{\sum_{l=1}^M I_G(x_l) \mathbf{Z}(x_l)}{M_G(x)} \quad (13a)$$

$$Y_{c,G}(x) = \frac{\sum_{l=1}^M I_G(x_l) Y(x_l)}{M_G(x)} \quad \text{and} \quad \hat{Y}_{c,G}(x) = \hat{\mathbf{Z}}_{c,G}^\top \hat{\beta}_{c,s_2} \quad (13b)$$

$$\bar{\hat{R}}_{2,G} = \frac{\sum_{x \in s_{2,G}} M_G(x) \hat{R}_{c,G}(x)}{\sum_{x \in s_{2,G}} M_G(x)} \quad \text{where} \quad \hat{R}_{c,G}(x) = Y_{c,G}(x) - \hat{Y}_{c,G}(x) \quad (13c)$$

²⁰² Note that the restriction to G , i.e. $I_G(x_l) = \{0, 1\}$, is made on the individual sample plot level x_l ,
²⁰³ and $M_G(x) = \sum_{l=1}^M I_G(x_l)$ thus is the number of sample plots per cluster within the small area. The
²⁰⁴ asymptotically design-unbiased point estimate of PSMALL is then defined according to Eq. 14a. The
²⁰⁵ first term estimates the small area population mean of G by applying the globally derived regression
²⁰⁶ coefficients to the small area cluster means of the explanatory variables $\hat{\mathbf{Z}}_{c,G}$. The second term then
²⁰⁷ corrects for a potential bias of the regression model predictions in the small area G by adding the
²⁰⁸ mean of the empirical residuals $\bar{\hat{R}}_{2,G}$ in G . This correction is necessary because the zero mean residual
²⁰⁹ property that holds in F is not guaranteed to hold in small area G under this construction.

$$\hat{Y}_{c,G,PSMALL} = \hat{\mathbf{Z}}_{c,G}^\top \hat{\boldsymbol{\beta}}_{c,s_2} + \bar{R}_{2,G} \quad (14a)$$

$$\begin{aligned} \hat{\mathbb{V}}(\hat{Y}_{c,G,PSMALL}) &= \hat{\mathbf{Z}}_{c,G}^\top \hat{\boldsymbol{\Sigma}}_{\hat{\boldsymbol{\beta}}_{c,s_2}} \hat{\mathbf{Z}}_{c,G} + \hat{\boldsymbol{\beta}}_{c,s_2}^\top \hat{\boldsymbol{\Sigma}}_{\hat{\mathbf{Z}}_{c,G}} \hat{\boldsymbol{\beta}}_{c,s_2} \\ &+ \frac{1}{n_{2,G}(n_{2,G}-1)} \sum_{x \in s_{2,G}} \left(\frac{M_G(x)}{\bar{M}_{2,G}} \right)^2 (\hat{R}_{c,G}(x) - \bar{R}_{2,G})^2 \end{aligned} \quad (14b)$$

²¹⁰ with $\bar{M}_{2,G} = \frac{\sum_{x \in s_{2,G}} M_G(x)}{n_{2,G}}$.

²¹¹

²¹² The variance-covariance matrix of the auxiliary vector $\hat{\boldsymbol{\Sigma}}_{\hat{\mathbf{Z}}_{c,G}}$ is thereby defined as

$$\hat{\boldsymbol{\Sigma}}_{\hat{\mathbf{Z}}_{c,G}} = \frac{1}{n_{1,G}(n_{1,G}-1)} \sum_{x \in s_{1,G}} \left(\frac{M_G(x)}{\bar{M}_{1,G}} \right)^2 (\mathbf{Z}_{c,G}(x) - \hat{\mathbf{Z}}_{c,G})(\mathbf{Z}_{c,G}(x) - \hat{\mathbf{Z}}_{c,G})^\top \quad (15)$$

²¹³ with $\bar{M}_{1,G} = \frac{\sum_{x \in s_{1,G}} M_G(x)}{n_{1,G}}$.

²¹⁴

²¹⁵ The estimated design-based variance of $\hat{Y}_{c,G,PSMALL}$ is given by Eq. 14b. Basically, the first
²¹⁶ term constitutes the variance introduced by the uncertainty in the regression coefficients, whereas
²¹⁷ the second term expresses the variance caused by estimating the exact auxiliary mean in G using a
²¹⁸ non-exhaustive sample $s_{1,G}$. The third term is the variance of the model residuals and thus accounts for
²¹⁹ the inaccuracies of the model predictions. Note that the first term can also be rewritten using g-weights
²²⁰ [29, pg.14] which ensures some beneficial calibration of the auxiliary variables to the first-phase sample.

²²¹

²²² 4.2.2. Pseudo Synthetic Estimator (PSYNTH)

²²³ The PSYNTH estimator is commonly applied when no terrestrial sample is available within
²²⁴ the small area G (i.e. $n_{2,G} = 0$). The point estimate (Eq. 16a) is thus only based on the predictions
²²⁵ generated by applying the globally derived regression coefficients to the small area cluster means of
²²⁶ the explanatory variables $\hat{\mathbf{Z}}_{c,G}$. Note that the bias correction term using the empirical residuals (Eq.
²²⁷ 14a) can no longer be applied. The PSYNTH estimator thus has a potential unobservable design-based
²²⁸ bias.

$$\hat{Y}_{c,G,PSYNTH} = \hat{\mathbf{Z}}_{c,G}^\top \hat{\boldsymbol{\beta}}_{c,s_2} \quad (16a)$$

$$\hat{\mathbb{V}}(\hat{Y}_{c,G,PSYNTH}) = \hat{\mathbf{Z}}_{c,G}^\top \hat{\boldsymbol{\Sigma}}_{\hat{\boldsymbol{\beta}}_{c,s_2}} \hat{\mathbf{Z}}_{c,G} + \hat{\boldsymbol{\beta}}_{c,s_2}^\top \hat{\boldsymbol{\Sigma}}_{\hat{\mathbf{Z}}_{c,G}} \hat{\boldsymbol{\beta}}_{c,s_2} \quad (16b)$$

²²⁹ The contribution to the variance by the model residuals in small area G can also no longer be
²³⁰ considered (Eq. 16b). As a result, the synthetic estimator will usually have a smaller variance than
²³¹ estimators that consider the model residuals, but at the cost of a potential bias. Note that the PSYNTH
²³² estimator is still design-based, but one purely has to rely on the validity of the regression model within
²³³ the small area as it is the case in the model-dependent framework.

²³⁴

²³⁵ 4.2.3. Extended Pseudo Synthetic Estimator (EXTPSYNTH)

²³⁶ The EXTPSYNTH estimator (Eq. 17) has been proposed by Mandallaz [24] as a transformed
²³⁷ version of the PSMALL estimator that has the form of the PSYNTH estimator but remains
²³⁸ asymptotically design unbiased. It has the advantage that the mean of the empirical model residuals
²³⁹ of the OLS regression model for the entire area F and the small area G are by construction both
²⁴⁰ zero at the same time, i.e. $\bar{R}_c = \bar{R}_{c,G} = 0$. This is realized by extending the auxiliary vector $\mathbf{Z}_c(x)$

241 by the indicator variable $I_{c,G}$ which takes the value 1 if the entire cluster lies within the small area
 242 G and 0 if the entire cluster is outside G , i.e. $I_{c,G}(x) = \frac{M_G(x)}{M(x)}$. The extended auxiliary vector thus
 243 becomes $\hat{\mathbf{Z}}_c^\top(x) = (\mathbf{Z}_c^\top(x), I_{c,G}(x))$ and the new regression coefficient using $\hat{\mathbf{Z}}_c(x)$ instead of $\mathbf{Z}_c(x)$
 244 in Eq. 9 is denoted as $\hat{\boldsymbol{\theta}}_{s_2}$. All remaining components are calculated by plugging in $\hat{\mathbf{Z}}_c(x)$ in Eq. 13.
 245 A decomposition of $\hat{\boldsymbol{\theta}}_{s_2}$ reveals that the residual correction term is now included in the regression
 246 coefficient $\hat{\boldsymbol{\theta}}_{s_2}$ [29].

$$\hat{Y}_{c,G,EXTPSYNTH} = \hat{\mathbf{Z}}_{c,G}^\top \hat{\boldsymbol{\theta}}_{c,s_2} \quad (17a)$$

$$\hat{\mathbb{V}}(\hat{Y}_{c,G,EXTPSYNTH}) = \hat{\mathbf{Z}}_{c,G}^\top \hat{\boldsymbol{\Sigma}}_{\hat{\boldsymbol{\theta}}_{c,s_2}} \hat{\mathbf{Z}}_{c,G} + \hat{\boldsymbol{\theta}}_{c,s_2}^\top \hat{\boldsymbol{\Sigma}}_{\hat{\mathbf{Z}}_{c,G}} \hat{\boldsymbol{\theta}}_{c,s_2} \quad (17b)$$

247 However, it is important to note that $\hat{R}_{c,G} = 0$ under the extended regression model only holds if
 248 the sample plots x_1, \dots, x_l of a cluster are *all* either inside or outside the small area, i.e. $M_G(x) \equiv M(x)$,
 249 and thus $I_{c,G}(x) = \frac{M_G(x)}{M(x)}$ can only take the values 1 or 0. Mandallaz *et al.* [29] assumed that the
 250 effects on the estimates should be negligible as the number of occasions where $M_G(x) < M(x)$ was
 251 considered to be small in practical implementations. It was thus a further objective of this study to
 252 investigate the actual number of occurrences as well as effects of this phenomenon by comparing the
 253 estimates of EXTPSYNTH to those of PSMALL.

254 4.3. Measures of estimation accuracy

255 The estimation precision was quantified by the estimation error, which is the ratio of the standard
 256 error and the point estimate (here \hat{Y} stands for the point estimate produced under the various
 257 estimators):

$$\text{error} [\%] = \frac{\sqrt{\hat{\mathbb{V}}(\hat{Y})}}{\hat{Y}} * 100 \quad (18)$$

258 We further calculated the 95% confidence interval for each estimate. The confidence intervals
 259 were used heuristically for hypothesis testing to determine whether the point estimates of the three
 260 estimators for a given small area were statistically different. The confidence intervals for the SRS
 261 estimator can be obtained as:

$$CI_{1-\alpha}(\hat{Y}_c) = \hat{Y}_c \pm t_{n_2-1,1-\frac{\alpha}{2}} \sqrt{\hat{\mathbb{V}}(\hat{Y}_c)} \quad (19)$$

262 The confidence intervals for the PSMALL and EXTPSYNTH estimates are calculated as:

$$CI_{1-\alpha}(\hat{Y}_{c,G,EXTPSYNTH}) = \hat{Y}_{c,G,EXTPSYNTH} \pm t_{n_2,G-1,1-\frac{\alpha}{2}} \sqrt{\hat{\mathbb{V}}(\hat{Y}_{c,G,EXTPSYNTH})} \quad (20a)$$

$$CI_{1-\alpha}(\hat{Y}_{c,G,PSMALL}) = \hat{Y}_{c,G,PSMALL} \pm t_{n_2,G-1,1-\frac{\alpha}{2}} \sqrt{\hat{\mathbb{V}}(\hat{Y}_{c,G,PSMALL})} \quad (20b)$$

263 For the PSYNTH estimates, the confidence intervals are

$$CI_{1-\alpha}(\hat{Y}_{c,G,PSYNTH}) = \hat{Y}_{c,G,PSYNTH} \pm t_{n_2-p,1-\frac{\alpha}{2}} \sqrt{\hat{\mathbb{V}}(\hat{Y}_{c,G,PSYNTH})} \quad (21)$$

264 with p being the number of parameters used in the regression model including the intercept term.

265 In order to address the potential benefits of the small area estimators compared with the SRS
 266 approach, we calculated the *relative efficiency* (*RE*, Eq. 22) which can be interpreted as the relative
 267 sample size under SRS needed to achieve the variance under the double-sampling (DS) estimators.

$$RE = \frac{\hat{V}(\hat{Y}_{SRS})}{\hat{V}(\hat{Y}_{DS})} \quad (22)$$

where \hat{Y} stands for the point estimate produced under the respective estimator.

5. Case study

5.1. Study area and small area units

The German federal state Rhineland-Palatinate (RLP) is located in the western part of Germany and borders Luxembourg, France and Belgium. With 42.3% (appr. 8400 km²) of the entire state area (19850 km²) covered by forest, RLP is one of the two states with the highest forest coverage among all federal states of Germany [2]. The forests of RLP are further characterised by a pronounced diversity in bioclimatic growing conditions that have strong influence on the local growth dynamics as well as tree species composition [31] and are further characterised by large variety of forest structures ranging from characteristic oak coppices (Moselle valley), pure spruce, beech and scots pine forests (i.a. Hunsrück and Palatinate forest) up to mixed forests comprising variable proportions of oak, larch, spruce, Scots pine and beech. Around 82% of the forest area in RLP are mixed forest stands and 69% of the forest area exhibit a multi-layered vertical structure. The forest area of RLP are divided into 3 ownership classes, i.e. state forest (27%), communal forest (46%) and privately owned forest (27%). The forest service of RLP has the legal mandate to sustainably manage the state and communal forest area (73% of the entire forest area), including forest planning, harvesting and the sale of wood [32]. For this reason, the entire forest area has been spatially organised in 3 main hierarchical management units (Figure 1). On the upper level, RLP has been divided into 45 Forstämter (FA), which are further divided into a total number of 405 Forstreviere (FR). The next level are the forest stands (104'184 in total) for which expert judgements are conducted by SFIs in a 5 to 10 year period in order to set up management strategies for the upcoming 10 years. The FAs and FRs constituted the SA units for which design-based small area estimations of the mean standing timber volume were calculated by incorporating the available terrestrial inventory data of the BWI3 in the estimators described in Section 4. The average area of the SA units was 43'777 ha on the FA-level, and 4624 ha on the FR level.

5.2. Terrestrial sample

Rhineland-Palatinate (RLP) is covered by a 2x2 km inventory grid of the German NFI. In the last inventory (BWI3) conducted in the year 2011 and 2012, timber volume information was derived for 2810 clusters (8092 plots) in the field survey. The local timber volume density on the plot and cluster level for this sample was consequently calculated according to Section 4.1. In the framework of this survey, the plot center coordinates were re-measured with the differential global satellite navigation system (DGPS) technique. Knowledge about the exact plot positions were considered crucial to provide optimal comparability between the terrestrial observations and the information derived from the auxiliary information. A comparison of the DGPS coordinates with the so-far used target coordinates revealed that 90% of all horizontal deviations lay in the range of 25 meters. A detailed analysis of horizontal DGPS errors in RLP by Lamprecht *et al.* [33] indicated that 80% of the plots should not exceed horizontal DGPS errors of 8 meters. For 162 plots, the DGPS coordinates were replaced by their target coordinates due to missingness or implausible values. The terrestrial sample size $n_{2,G}$ within the FA units was 46 clusters on average and ranged between 11 and 64. Within the FR units, $n_{2,G}$ was considerably smaller with an average of 5 clusters and a range between 0 and 13.

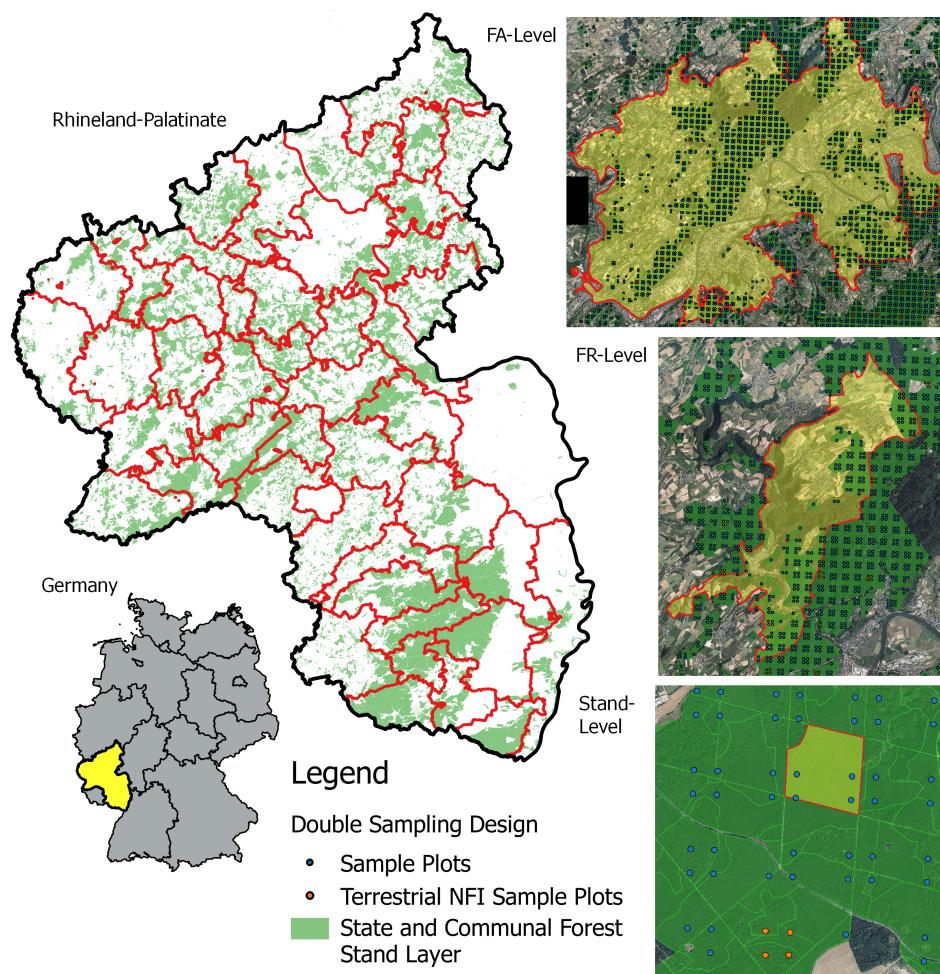


Figure 1. Left: Study area with delineated FA forest management units. Right: Example for each of the three management units (from top to bottom): FA, FR and forest stand unit overlaid with the extended double-sampling cluster design. Green: Forest stand polygon layer defining the state and communal forest area of this study.

308 5.3. Extension to double-sampling design

309 In order to apply the small area estimators (Section 4.2), the existing NFI design was extended
 310 to a double-sampling cluster design by densifying the existing systematic 2x2 km grid to a grid size
 311 of 500x500 m that constituted the large first phase s_1 (Figure 1, right). The existing terrestrial phase
 312 s_2 was integrated by replacing the target coordinates of the respective s_1 clusters by the terrestrially
 313 measured DGPS coordinates. The sampling frame was further restricted to the communal and state
 314 forest area. The forest/non-forest decision for each plot was thereby made by a spatial intersection of
 315 the plot center coordinates with a polygon layer of the communal and state forest stand layer provided
 316 by the forest service. Using this stand layer provided the advantage to consistently apply the same
 317 forest/non-forest definition to the entire sample s_1 in order to decide about excluding or including a
 318 plot in the sampling frame. The terrestrial sample size n_2 was thus reduced to 2055 clusters (5791 plots).
 319 Table 1 provides a short descriptive summary about the volume densities and the main attributes of
 320 the NFI plots located in the state and communal forest sampling frame. The densification led to an
 321 average sample size $n_{1,G}$ of 759 clusters (range: 246 – 1022) in the FA units, and 88 clusters (range: 1 –
 322 194) in the FR units.

Table 1. Descriptive statistics of the forest observed on NFI sample plots located within communal and state forest area ($n_2=5791$).

Variable	Mean	SD	Maximum
Timber Volume (m ³ /ha)	300.86	195.55	1375.31
Mean DBH (mm)	354.90	137.22	1123.20
Mean height (dm)	239.60	72.43	497.43
Mean stem density per hectare	101.00	114.01	1010.31

323 5.4. Auxiliary data

324 5.4.1. LiDAR canopy height model

325 A prerequisite for the application of the suggested two-phase small area estimators is the
 326 identification of suitable auxiliary data available over the entire study area. From 2003 to 2013,
 327 the topographic survey institution of RLP conducted an airborne laserscanning acquisition over the
 328 entire federal state during leaf-off conditions in order to derive a countrywide digital terrain model
 329 (DTM) as well as a digital surface model (DSM). For this study, the recorded ALS data was used to
 330 create a canopy height model (CHM) in raster format, providing discrete information about the canopy
 331 surface height of the forest area in a spatial resolution of 5 meters (Fig. 2, top). The CHM was calculated
 332 as the difference between the digital terrain model and the digital surface model that were derived by
 333 a Delauney interpolation of the ground and first ALS pulses respectively. A more detailed description
 334 of the procedure can be found in Hill *et al.* [34]. The CHM provided the most valuable information to
 335 be used in the OLS regression model for predicting the timber volume on the plot and cluster level.
 336 However, it should be noted that the prolonged acquisition period of the ALS campaign led to the
 337 possibility of poor temporal alignment with the BWI3 survey, sometimes up to 10 years. In addition,
 338 the quality of the CHM varied substantially as ALS technology evolved over the years. For example,
 339 the ALS acquisitions recorded in 2002 and 2003 exhibited particularly poor quality with about only 0.04
 340 point per m², whereas more recent datasets contained more than 5 points per m². Furthermore, CHM
 341 information was not available at 16 sample locations due to sensor failures. These plots were deleted
 342 from the sampling frame and treated as missing at random. This assumption was considered to be
 343 reasonable as the respective sample locations did not systematically exclude specific forest structures.

344 5.4.2. Tree species map

345 Additional auxiliary data was derived from a countrywide satellite-based classification map
 346 predicting the five main tree species [35], i.e. European beech, Sessile and Pedunculate oak, Norway
 347 spruce, Douglas fir and Scots pine (Fig. 2, bottom). The tree species map has a grid size of 5x5 m
 348 and was calculated from 22 bi-temporal satellite images (SPOT5 and RapidEye) using a spatially
 349 adaptive classification algorithm [36]. As timber volume estimation on the tree level is often based
 350 on species-specific biomass and volume equations, the use of tree species information has often been
 351 stated as a key factor for improving the precision of timber volume estimates [37]. In this respect,
 352 incorporating the tree species map was particularly attractive as it predicts five of the seven tree species
 353 that are used in the BWI3 taper functions [27] to calculate the timber volume of a sample tree. However,
 354 due to unavailable satellite data, the tree species map excluded one large patch with an area of 415
 355 km² in the south-west part of RLP covering an entire FA unit consisting of 10 FR units. In 9 additional
 356 FR units, the tree species information was also missing for a subset of the sample locations due to two
 357 additional patches with areas of 76 km² and 100 km² respectively in the northern part of RLP. For these
 358 19 FR units, small area estimation was thus restricted to using only the available CHM information in
 359 the regression model. Thus, 411 of 5791 sample locations (approximately 7%) used to fit the regression

³⁶⁰ model were affected by missing tree species information. A summary of the sample sizes and missing
³⁶¹ auxiliary data for both the CHM and the tree species map is provided in Table 2.

Table 2. Sample size for each phase in entire study area. $n_{\{1,2\},plot}$: number of plots. $n_{\{1,2\}}$: number of clusters. TSPEC: tree species map information.

Sampling frame	$n_{1,plot}$	n_1	$n_{2,plot}$	n_2
communal and state forest	96'854	33'365	5791	2055
missing CHM	18	10	0	0
missing TSPEC	7060	3587	414	385
missing CHM and TSPEC	3	2	0	0
missing CHM or TSPEC	7075	3595	414	385

³⁶² 5.5. Calculation of the explanatory variables

³⁶³ 5.5.1. Canopy height model

³⁶⁴ The continuous explanatory variables derived from the CHM were the mean canopy height
³⁶⁵ (*meanheight*) and the standard deviation (*stddev*). The quantities were calculated by evaluating the
³⁶⁶ raster values around each sample location within a circle with a predefined radius of 12 meters, i.e.
³⁶⁷ the support. In order to correct for edge effects at the forest border, the intersection of each support
³⁶⁸ area to the state and communal forest area was determined using a polygon mask provided by the
³⁶⁹ state forest service. The percentage of the support within the forest layer was used as the weight
³⁷⁰ $w(x_l)$ introduced in Eq. 10 in order to derive the weighted mean of the explanatory variables on the
³⁷¹ cluster level. Neglecting the support adjustment would deteriorate the coherence between explanatory
³⁷² variables computed at the forest boundary and the corresponding local density that already includes
³⁷³ a potential boundary adjustment, thus introducing unnecessary noise to the model. The boundary
³⁷⁴ adjustment to the support also makes the sampling frame more consistent for the different data sources
³⁷⁵ (Section 5.3).

³⁷⁶ The ALS acquisition year (*ALSpyear*) was added as a categorical variable in order to account for the
³⁷⁷ time lag with the terrestrial survey as well as to help explain the heterogeneity in the data introduced
³⁷⁸ by the varying ALS quality. In 2008, a sensor error produced particularly poor ALS quality so the year
³⁷⁹ was divided accordingly into two factor levels, denoted 2008_1 and 2008. Furthermore, in order to
³⁸⁰ increase the number of observations per factor level the years 2006 and 2007 were pooled together and
³⁸¹ the same was done for 2012 and 2013. The result was nine factor levels denoted as 2002, 2003, 2007,
³⁸² 2008_1, 2008, 2009, 2010, 2011 and 2012.

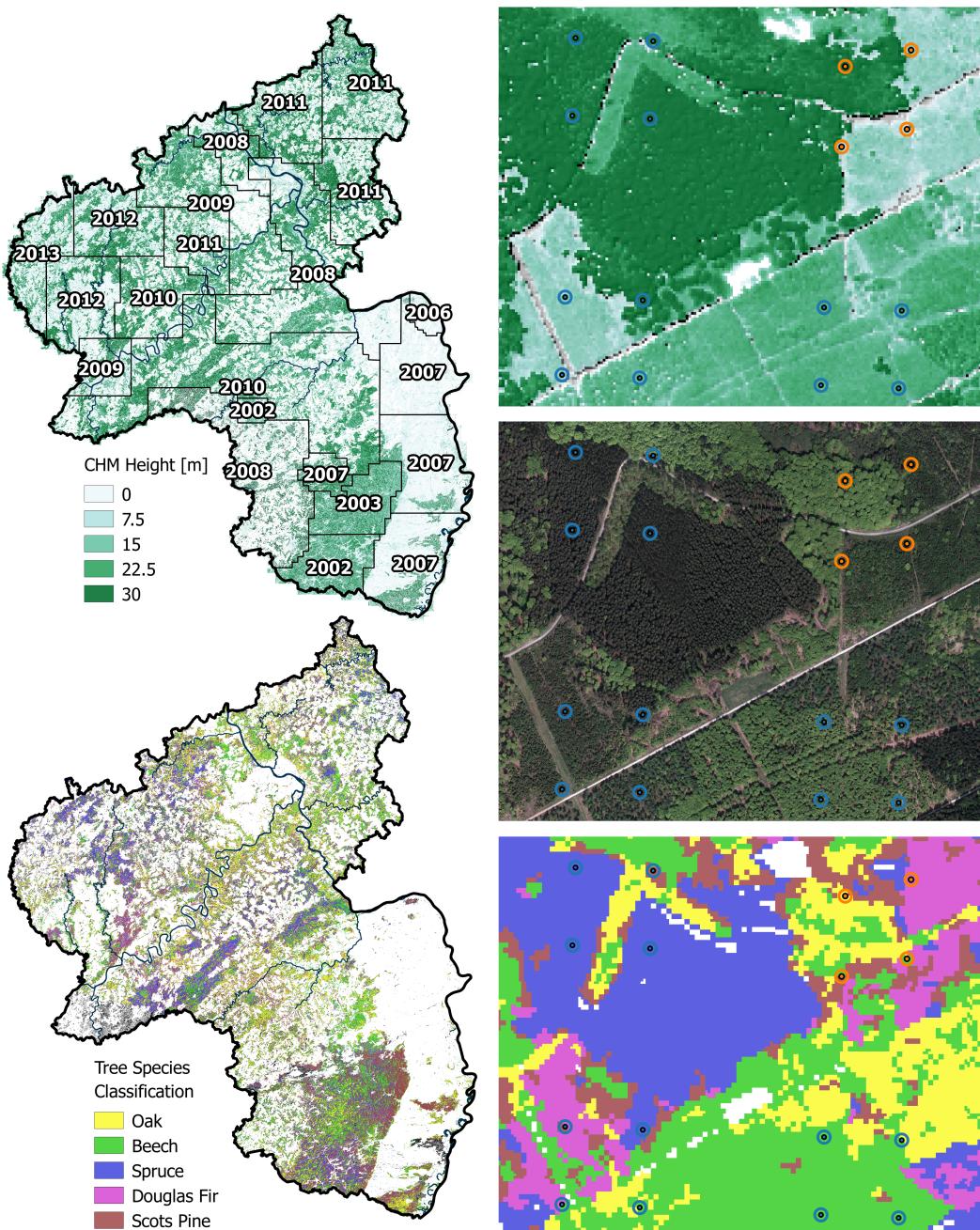


Figure 2. Left: CHM (top) and tree species classification map (bottom) available on the federal state level. Right: Magnified illustration of the supports used to derive the explanatory variables from the auxiliary data. From top to bottom: CHM, aerial image, tree species classification.

383 5.5.2. Tree species map

384 The tree species map was used to predict the main tree species at each sample plot which served
 385 as an additional categorical variable *treespecies* in the regression model. In the first step, one of the five
 386 tree species was assigned to a sample location if 100% of the raster values within the edge-corrected
 387 support were classified as that species. Otherwise, the sample location was assigned the value 'mixed'.
 388 Likewise for the CHM variables, the support radius was 12 meters although the use of different
 389 support sizes for each explanatory variable would be in agreement with the two-phase estimators
 390 presented in Section 4.2. The specific setting for the support size and the percentage threshold was
 391 found to be optimal in order to yield the best possible regression model precision when incorporating

the *treespecies* variable as an additional predictor. In a second step, the *treespecies* variable was also passed through a calibration model in order to reduce the effects of misclassification errors on the regression model coefficients and to increase model accuracy. The calibration model consisted of a decision tree from a random forest algorithm [38] that was trained to predict the actual main plot tree species (known for all terrestrial plots) based on available auxiliary variables. These variables were the predicted *treespecies* variable, the mean canopy height and standard deviation of the CHM, as well as the proportion of coniferous trees estimated from the classification map and the growing region derived from a polygon map. The algorithm was grown with 2000 trees considering 3 of the predictors for each split. We thus applied this calibration model to the *treespecies* variable derived at all sample locations s_1 . Table 3 gives the classification accuracies [39] of the *treespecies* variable after calibration. More details on the processing of the explanatory variables and identification of optimal parameter settings for their calculation are described in Hill *et al.* [34].

Table 3. Classification accuracies of the *treespecies* variable before and after calibration. n_{ref} : number of terrestrial reference plots. n_{class} : number of classified plots.

Main plot species	Producer's accuracy[%]	User's accuracy[%]	n_{ref}	n_{class}
Beech	22.31	47.02	883	419
Douglas Fir	24.78	48.72	230	117
Oak	11.07	48.48	289	66
Spruce	53.15	61.13	651	566
Scots Pine	22.91	46.07	179	89
Mixed	84.49	64.53	3152	4127
Overall accuracy: 61.96%			5384	5384

5.6. Regression Model

The model selection process for this study required a substantial time commitment due to sophisticated challenges such as: a) the heterogeneity of the remote sensing data, b) the identification of the optimal support sizes under angle count sampling, and c) the incorporation of tree species information. Here, only a summary of the extensive analysis that was performed is provided but the reader can refer to Hill *et al.* [34] if more details are desired.

The model with highest adjusted R^2 and lowest RMSE was achieved using *meanheight*, *meanheight*², *stddev*, *ALSyear* and *treespecies* as main effects, and including interaction terms between *meanheight* and *ALSyear*, *stddev* and *ALSyear*, *meanheight* and *stddev*, and *meanheight* and *treespecies*. Summary information about the adjusted R^2 , RMSE and RMSE% of the selected models is provided in Table 4. As the two-phase estimators described in Section 4.2 derive and apply the regression coefficients and the residuals on the aggregated cluster level, we re-evaluated the model as used in the estimators on the cluster level (formulas given in Appendix) and found improved model fits compared to the plot level (adjusted R^2 of 0.59 and RMSE of 101.61 m³/ha and 33.6%). The stratification by the ALS acquisition year substantially improved the model fit, indicating that it is an effective means in accounting for the noise in the data caused by ALS quality variations and time-gaps between the ALS and the terrestrial survey. However, the stratification led to a highly unbalanced data set when a further *treespecies* stratification was included. For this reason, a individual species modeling within each *ALSyear* stratum remained infeasible, but might have further improved the model fit. An additional evaluation of the model's performance within each ALS acquisition year stratum revealed that the quality of the model fit substantially varied between the strata (Table 5). In particular, values above the overall adjusted R^2 were higher in ALS acquisition years close to the terrestrial survey date compared to years with larger time gaps.

As described in Section 5.4.2, the information of the tree species classification map was missing within 1 FA and 19 FR units. For these small area units, we applied the regression model without

the *treespecies* variable (Table 4, reduced model). However, the adjusted R^2 s of the full and reduced model were found to be very similar on both the plot and cluster level. This implied that the variance reduction of the reduced model when applied to the two-phase estimators would likely be comparable to that of the full model. For this reason, a joint evaluation of the estimation results is performed in Section 6.

Table 4. Model fit specifications for the two OLS regression models on the cluster level. Interaction terms are indicated by ':'. () give the respective values on the plot level.

model terms	model	R^2_{adj}	RMSE	RMSE%
meanheight + stddev + meanheight ² + treespecies + ALSyear + meanheight:treespecies + meanheight:ALSyear + meanheight:stddev + stddev:ALSyear	full model	0.58 (0.48)	90.11 (139.22)	29.76 (45.98)
meanheight + stddev + meanheight ² + ALSyear + meanheight:ALSyear + meanheight:stddev + stddev:ALSyear	reduced model	0.55 (0.45)	95.23 (144.13)	31.65 (47.60)

Table 5. R^2 , RMSE and RMSE% on the cluster level of the full regression model within ALS acquisition year strata (*ALSyear*). *Area_{ALSyear}*: Area covered by ALS acquisition given in km². *n*: sample size of validation data. () give the respective values on the plot level.

<i>ALSyear</i>	<i>Area_{ALSyear}</i>	R^2	RMSE	RMSE%	<i>n</i>
2012	2807	0.65 (0.61)	98.52 (135.84)	29.62 (44.87)	156 (408)
2011	4361	0.60 (0.57)	96.89 (146.21)	29.66 (48.29)	354 (883)
2010	4182	0.64 (0.51)	76.38 (120.90)	27.57 (39.93)	420 (1171)
2009	2100	0.53 (0.42)	92.22 (133.42)	33.31 (44.07)	218 (559)
2008	2968	0.61 (0.48)	87.10 (130.38)	32.20 (43.06)	247 (701)
2008_1	2116	0.43 (0.33)	117.99 (175.43)	33.64 (57.94)	157 (394)
2007	3498	0.56 (0.46)	82.43 (136.47)	26.57 (45.08)	135 (418)
2003	602	0.34 (0.27)	85.92 (154.48)	27.31 (51.02)	145 (529)
2002	775	0.52 (0.44)	87.25 (141.55)	27.22 (46.75)	97 (314)

Concerning the existence of outliers or leverage points in the training set for the model, it should be noted that it is more problematic for PSMALL, PSYNTH and EXTPSYNTH to simply remove them as one might be inclined to do in a model-dependent context. Strictly speaking, outlier removal in the design-based context essentially means that those plots, and implicitly any potentially similar plots that were not realized in the selected sample, have been removed from the sampling frame and are no longer considered part of the forest area of interest. While this may be valid for some obvious typos or measurement errors, it is generally not advisable to manipulate the sampling frame after observing data collected from it, especially when the observation in question lies within the small area of interest. However, for sake of completeness, we conducted an analysis of influential observations

[443] [40, pp. 160–167] on the plot level for the full regression model. We calculated the leverage values and
 [444] found that 10% of all observations exceeding a predefined critical threshold, i.e. twice the average of
 [445] the hat matrix diagonal entries. Further investigation revealed that several leverage points showed
 [446] unusually large *meanheight* values compared to their respective timber volume densities. They tended
 [447] to occur in ALS acquisition years with longer time gaps to the terrestrial survey date and were thus
 [448] more likely caused by harvesting activities in the sample plot area. Although these areas likely affected
 [449] by harvest should clearly not be removed from the sampling frame, it does provide more justification
 [450] for the inclusion of the *ALSpyear* variable to mitigate the implied effects.

451 6. Results

452 6.1. General estimation results

453 An application of the SRS, PSMALL and EXTPSYNTH estimator was not feasible for 17 of all 405
 454 FR-units due to an insufficient terrestrial sample size of $n_{2,G} < 2$. We further restricted the calculation
 455 of the PSMALL and EXTPSYNTH estimator to small area units with a minimum terrestrial sample
 456 size of $n_{2,G} \geq 4$ to avoid unstable estimates. This affected 65 additional FR units and limited unbiased
 457 two-phase estimations to 321 (79%) of the 405 FR units. It should be noted that also the PSYNTH
 458 estimator could not be applied for 2 FR-units since $n_{1,G} < 2$. Due to substantially larger sample sizes,
 459 all estimators could however be applied to all 45 FA units. The average value and the range of the
 460 mean timber volume estimates over the evaluated FA and FR units turned out to be very similar
 461 between all estimators (Table 6). An additional pairwise comparison of the 95% confidence intervals
 462 revealed that the four estimators did in fact not produce statistically different point estimates for all
 463 FA and FR units. This confirmed that the differences between the estimators are solely found in the
 464 precision which they provide for the point estimates.

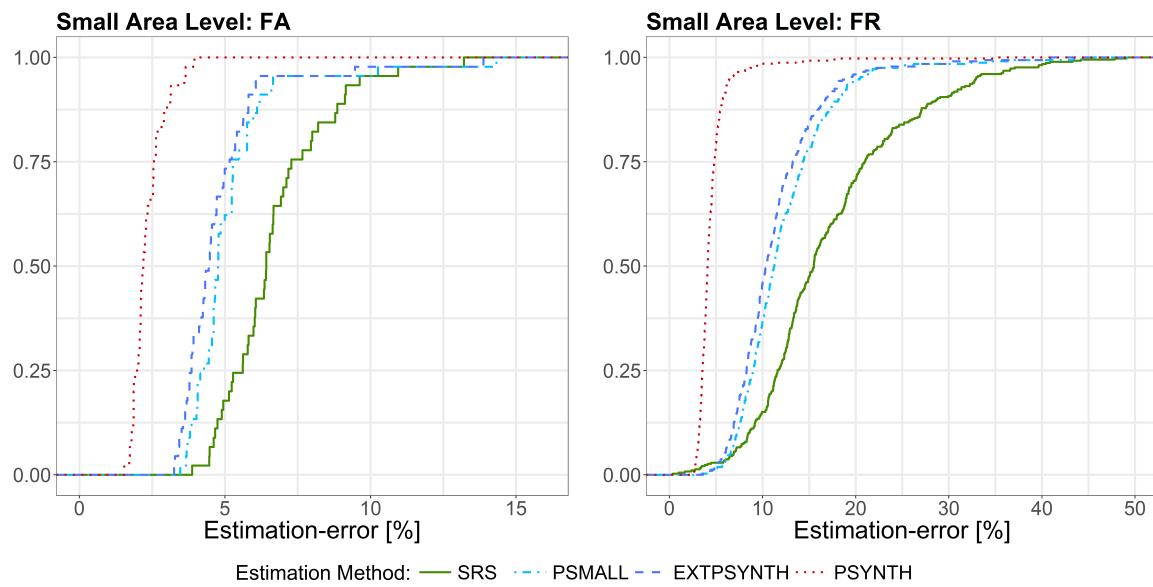
Table 6. Descriptive summary of point estimates and estimation errors on the two forest district levels.
 N_u : number of evaluated small area units.

District level	Estimator	Point estimates			error[%]		
		mean	min	max	mean	min	max
FA	SRS ($N_u=45$)	300.16	215.91	392.84	6.69	3.87	13.21
	PSMALL ($N_u=45$)	307.29	209.26	417.10	5.16	3.46	14.33
	EXTPSYNTH ($N_u=45$)	307.27	209.01	415.02	4.78	3.25	13.88
	PSYNTH ($N_u=45$)	306.90	223.51	409.92	2.34	1.54	3.95
FR	SRS ($N_u=388$)	301.83	99.89	612.13	18.32	0.34	104.97
	PSMALL ($N_u=321$)	308.15	159.64	568.67	12.24	3.48	44.94
	EXTPSYNTH ($N_u=321$)	308.38	154.07	544.34	11.34	3.60	40.91
	PSYNTH ($N_u=403$)	307.82	166.01	444.29	4.65	2.56	62.51

465 6.2. Estimation errors

466 On both small area levels, the design-unbiased estimators PSMALL and EXTPSYNTH led to a
 467 substantial reduction in the estimation error compared to the SRS estimator (Fig. 3). On the FA level,
 468 the SRS estimator yielded an estimation error of 6.7% on average compared to 5.2% and 4.8% under
 469 EXTPSYNTH and PSMALL respectively (Table 6). The cumulative error distribution (Fig. 3, left)
 470 reveals that under the SRS estimator, errors less than 5% were achieved for 17% of the FA units (8 of
 471 45). This proportion could be increased to 62% (28 FA units) and 73% (33 FA units) by application of
 472 the PSMALL and EXTPSYNTH estimator. 95% of all estimates exhibited errors less than 9.5% under
 473 the SRS estimator and less than 6.6% when using PSMALL or EXTPSYNTH. Estimation errors higher
 474 than 10% only appeared twice for each of the three estimators.

475 Although the estimation errors were substantially larger overall on the FR level compared to the
 476 FA level due to smaller sample sizes, the error reduction from SRS by PSMALL and EXTPSYNTH were
 477 even more pronounced (Fig. 3, right). The average error under the SRS estimator was 18.3%, while
 478 it was 11.3% and 12.2% under PSMALL and EXTPSYNTH (Table 6). Errors smaller than 10% were
 479 achieved for 15% of the FR units by the SRS estimator, and for 46% by the PSMALL and EXTPSYNTH
 480 estimator. 95% of the 321 FR units where PSMALL and EXTPSYNTH could be applied exhibited errors
 481 less than 20%. In comparison, the SRS estimates resulted in errors less than 36.6% for 95% of the 388 FR
 482 units.



483 **Figure 3.** Cumulative distribution of estimation errors under SRS, PSMALL, EXTPSYNTH and the
 484 PSYNTH estimator. *Left:* Results for the 45 FA units. *Right:* Results for the 388 (SRS), 321 (PSMALL,
 485 EXTPSYNTH) and 403 (PSYNTH) FR units.

486 On both small area levels, the PSYNTH estimator resulted in much smaller estimation errors
 487 compared to PSMALL and EXTPSYNTH. This was as expected, since the PSYNTH variance estimate
 488 does not take the residual variation in each small area unit into account (Section 4.2.2). Compared
 489 to the asymptotically design-unbiased estimators PSMALL and EXTPSYNTH, the estimation errors
 490 produced by PSYNTH thus seem to be too optimistic. One should also recall that the estimates of the
 491 PSYNTH estimator are potentially design-biased.

492 6.3. Comparison of PSMALL and EXTPSYNTH

493 Figure 3 reveals that the error distribution of PSMALL and EXTPSYNTH are very similar, with
 494 PSMALL showing marginally higher estimation errors. In order to investigate the differences between
 495 PSMALL and EXTPSYNTH, we compared the g-weight variances of both estimators for all 321 FR
 496 units (Fig. 4, left). As obvious, PSMALL yielded slightly larger variances for the vast majority of
 497 the estimates. As addressed in Section 4.2.3, one possible explanation for differences was the effect
 498 of one or more clusters not entirely being included in a small area unit, as this would constitute an
 499 assumption violation of the EXTPSYNTH estimator. This violation was actually observed in 155 of
 500 the 321 FR units (48%). We compared the variances of PSMALL and EXTPSYNTH for all small areas
 501 that did not have the violations using a Wilcoxon Signed-Rank Test [41] on a 5% significance level.
 502 This test was also performed pairwise for groups $n_{2,G} \leq 6$, $n_{2,G} > 6$ and $n_{2,G} > 10$. The distribution
 503 of variances from EXTPSYNTH was found to be highly significantly lower than that of PSMALL
 504 except for the group of $n_{2,G} > 10$. The latter was expected since the variances of both estimators are
 505 asymptotically equivalent under large terrestrial sample sizes $n_{2,G}$ within the small area [29, pp.17–18].

This was also confirmed by a visual comparison of the absolute differences in the variances (Fig. 4, right) which decreased with increasing terrestrial sample size. Performing the same comparison for small areas with violations also revealed the EXTPSYNTH variances to be significantly smaller than the respective PSMALL variances until sample sizes $n_{2,G} > 10$. Based on these investigations, it was not possible to determine whether the differences for sample sizes smaller than 10 were caused by the violations or just reflect the general tendency of EXTPSYNTH to produce smaller variances than PSMALL under small sample sizes. However, a visual inspection provided some evidence that the violations created a statistically significant influence on the EXTPSYNTH variance (Fig. 4, left, red diamonds) that makes it appear to be slightly over-optimistic. For sample sizes of $n_{2,G} < 6$, a weakly significant difference between the EXTPSYNTH variances of those small areas with violations and the EXTPSYNTH variances without violation was also indicated by an unpaired Wilcoxon Rank-Sum Test. However, the differences were still marginal and a comparison of the confidence intervals of PSMALL and EXTPSYNTH revealed that the variance differences did not lead to statistically significant point estimates.

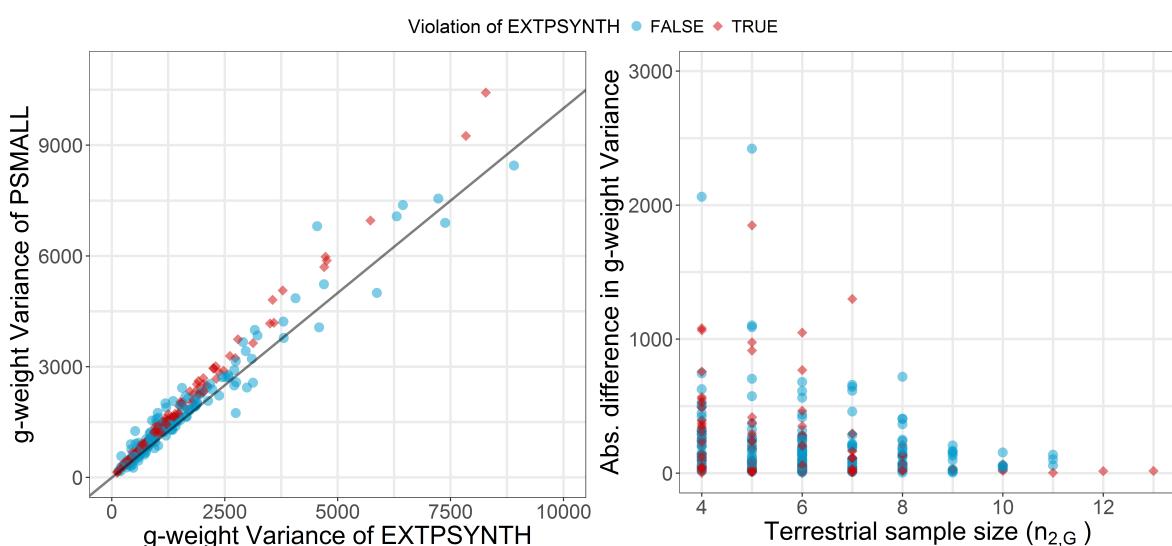


Figure 4. Left: Comparison of the g-weight variance between the PSMALL and the EXTPSYNTH estimator for the 321 FR units. Right: Difference in g-weight variance between the PSMALL and the EXTPSYNTH estimator in dependence of the terrestrial data ($n_{2,G}$) in the FR unit.

6.4. Variance reduction compared to SRS

The variance reduction relative to SRS for PSMALL and EXTPSYNTH are described in Figure 5 and Table 7. A direct comparison of the variances within the small area units revealed that the application of the design-unbiased estimators (PSMALL, EXTPSYNTH) led to a variance reduction compared to SRS in all FA units. In 75% of the FA units, the EXTPSYNTH estimator was able to reduce the variance by up to 54.1%. The reduction in variance can also be expressed in the relative efficiency values, which were 2.02 on average and ranged between 1.18 and 4.13 on the FA level. On FR level, the reduction in variance even reached values of 90% and relative efficiencies of 30 (Table 7 and Fig. 5). The PSMALL estimator again yielded slightly lower variance reductions and relative efficiencies due to the generally smaller variances of the EXTPSYNTH estimator (Section 6.3).

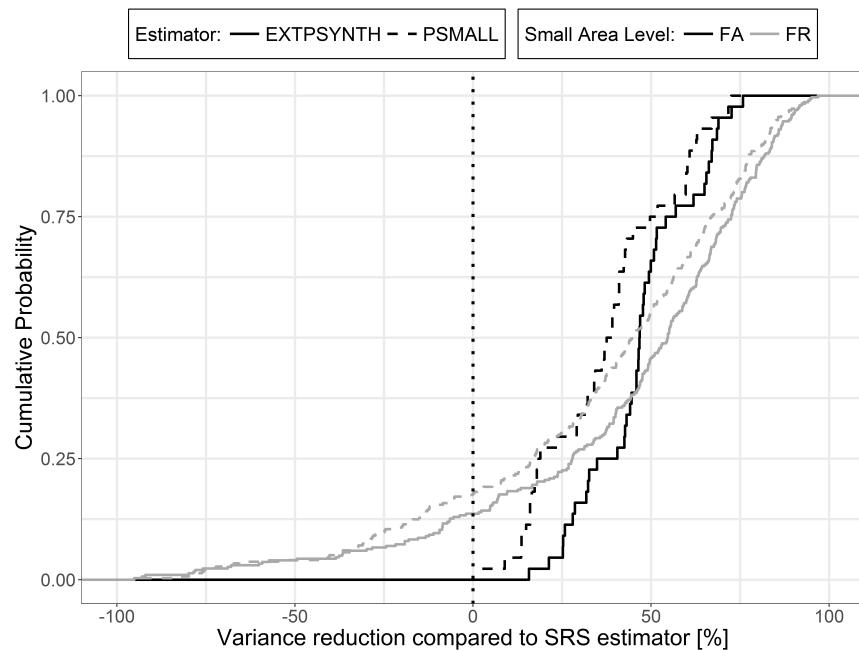


Figure 5. Cumulative distribution of variance reduction by the PSMALL and EXTPSYNTH compared to the SRS estimator for the 45 FA and 321 FR units.

Table 7. Descriptive summary of variance reduction compared to SRS and relative efficiencies on the two forest district levels. N_u : number of evaluated small area units.

District level	Estimator	Variance reduction [%]			relative efficiency		
		mean	min	max	mean	min	max
FA	PSMALL ($N_u=45$)	33.51	2.6	72.5	1.74	1.03	3.64
	EXTPSYNTH ($N_u=45$)	43.30	15.7	75.8	2.03	1.18	4.13
FR	PSMALL ($N_u=321$)	12.48	-1203.9	96.8	2.54	0.08	31.61
	EXTPSYNTH ($N_u=321$)	24.75	-892.7	97.0	2.95	0.10	33.70

527 Cases also occurred on the FR level where one or both two-phase estimators produced larger
 528 variance values than under the SRS estimator. This happened in 19% of the FR units under the
 529 EXTPSYNTH, and in 24% of the FR units under the PSMALL estimator. One possible reason for this
 530 was supposed to be a large residual variance due to a poor performance of the regression model
 531 within the small area unit. In order to investigate this hypothesis, we analyzed the three variance
 532 terms of the PSMALL estimator (Eq. 14b), i.e. the variance introduced by the uncertainty of the
 533 regression coefficients (term 1), the variance caused by estimating the auxiliary means (term 2), and
 534 the variance of the model residuals (term 3). In general, the residual term is expected to make the
 535 largest contribution to the overall variance since it's sample size is based on $n_{2,G}$ whereas the auxiliary
 536 term and the coefficient term are based on larger sample sizes, i.e. $n_{1,G}$ and n_2 respectively. Figure 6
 537 illustrates the share of the overall variance by the residual term of the PSMALL estimator scaled by
 538 the overall percentage reduction or increase of the variance compared to SRS for various small area
 539 sample sizes $n_{2,G}$. Not surprisingly, the residual term generally constitutes the dominating part of the
 540 PSMALL variance (around 84% on average). It has to be noted that such high residual term dominance
 541 does not necessarily indicate that the PSMALL variance will be disproportionately large (Figure 6, right).
 542 However, the vast majority of cases where the PSMALL variance was considerably larger than the
 543 SRS variance occurred where the residual term contributed over 75% to the overall PSMALL variance
 544 (Figure 6, left). Among those cases, the most pronounced were observed under small sample sizes

*n*_{2,G} < 5. Here, the average increase in variance compared to SRS of those FR units with *n*_{2,G} = 4 was 272%, compared to 62% for FR units with *n*_{2,G} > 4. In contrast, the decreases in variance compared to SRS (Figure 6, right) were much more homogeneous in magnitude and also independent of the terrestrial sample size. Since *n*_{2,G} is the same for PSMALL and SRS, these observations imply that in the problematic small areas, the sum of square residuals for the regression model are likely larger than the sum of square local densities for the clusters in *s*_{2,G}. This indicates the presence of outliers with large residuals, which likely arise when there was forest loss after the ALS scanning but before the terrestrial survey year.

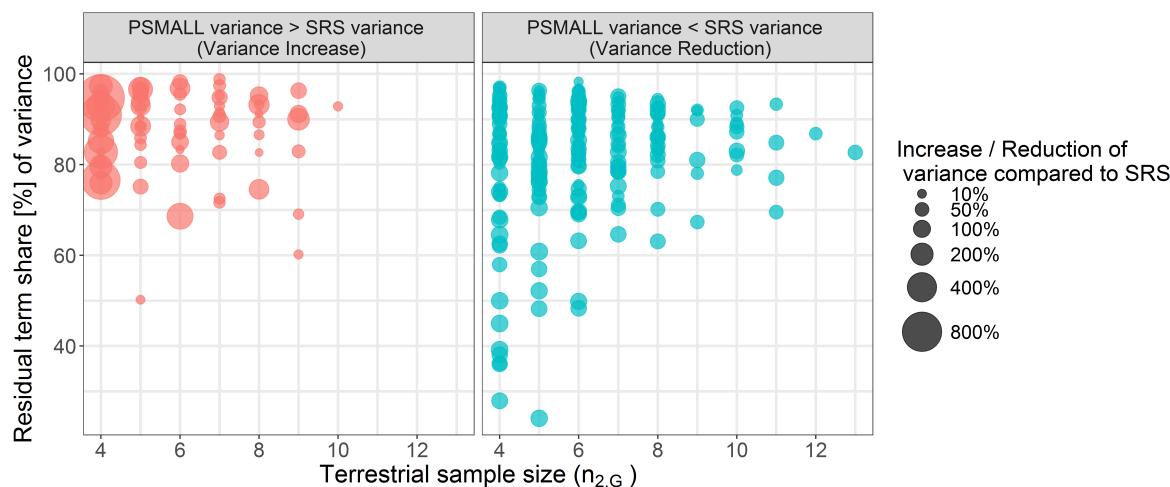


Figure 6. Share of the overall variance by the residual term of the PSMALL estimator for various small area sample sizes. Points are scaled by the overall percentage reduction/increase of the variance compared to SRS.

553 7. Discussion

554 7.1. Performance of estimators

555 With the objective of extending the use of the German NFI data to additional estimation on
 556 small-scale management levels, we evaluated the performance of design-based small area regression
 557 estimators with respect to their suitability for future operational large scale application. For this reason,
 558 we conducted a case study in the German federal state of Rhineland-Palatinate where we applied the
 559 SRS, the PSMALL and the EXTPSYNTH estimators to produce estimates of the mean timber volume on
 560 two forest management levels over the entire federal state area, comprising 45 and 405 small area units
 561 respectively. In order to assess and compare the performance of the estimators, it was of particular
 562 interest to gather information about the magnitudes of estimation precision they can provide.

563 Our study showed that on both small area levels, the PSMALL and the EXTPSYNTH estimators
 564 generally led to a substantial reduction in estimation error compared to the standard one-phase
 565 SRS estimator. On the upper management level (FA districts), PSMALL and EXTPSYNTH produced
 566 estimation errors smaller than 5% for 73% of the small areas compared to only 17% under the one-phase
 567 SRS estimator. The same level of precision could not be achieved on the lower management level
 568 (FR districts) primarily due to substantially smaller terrestrial sample sizes. However, in 95% of the
 569 FR units, the estimation errors could be limited to 20% compared to 40% under SRS. A pairwise
 570 comparison of the confidence intervals revealed that the estimators did not produce significantly
 571 different point estimates. The much smaller estimation errors of the PSYNTH estimator reflected the
 572 fact that it does not try to correct for potential bias in the point estimate which can lead to overly
 573 optimistic estimation errors and confidence intervals. One should thus prefer the unbiased estimates
 574 of PSMALL or EXTPSYNTH whenever their calculation is possible.

575 For several FR units, it was observed that the PSMALL and the EXTPSYNTH estimator can
576 occasionally produce larger variances than the SRS estimator. It is important to note that this is in
577 perfect agreement with the theory of both two-phase estimators and can theoretically appear if the
578 residual variance in the small area, which generally constitutes the dominating part of the two-phase
579 variance, turns out to be much higher than the variance of the terrestrial data in the small area. The
580 empirical findings of our study suggest that such cases can particularly occur if moderate or poor
581 model fits within a small area are combined with small terrestrial sample sizes (≤ 5) in the small area.
582 A closer look on these small areas thus might reveal the reason for the poor prediction performance
583 and help to improve the model fit. Nonetheless, it should be kept in mind that small terrestrial sample
584 sizes can also cause the SRS estimator to not reflect the actual variation of the local density within a
585 small area. In this case, the two-phase variance estimate might be larger but more realistic. Whereas a
586 visual analysis of aerial images, remote sensing data or stand maps might give some further evidence
587 for or against this hypothesis, a definite proof is practically infeasible.

588 We were also able to empirically confirm that the EXTPSYNTH estimator generally produces
589 slightly smaller variances and estimation errors than the PSMALL estimator. This is most probably
590 caused by marginally smaller model residuals due to the intercept adjustment to the terrestrial data
591 in the small area unit, which is primarily a means to ensure the zero mean residual property of the
592 EXTPSYNTH estimator. However, our analysis indicated that the difference between the two estimators
593 is negligible for sample sizes ≥ 10 due to their asymptotic equivalency. We further investigated a
594 potential impact on the EXTPSYNTH variance caused by the assumption violation that one or more
595 clusters are not entirely included in the small area unit and found a slight but statistically significant
596 tendency to be over-optimistic for sample sizes smaller than 6. More empirical evidence must be
597 gathered before generalizing this as a rule of thumb for the application of the EXTPSYNTH under
598 cluster sampling. It thus seems recommendable to prefer the EXTPSYNTH to the PSMALL estimator
599 if its assumptions are not violated since it yields slightly smaller variances under mathematically
600 soundness. Even if the differences between both estimators were marginal and did not lead to
601 significantly different point estimates, PSMALL can serve as a safe alternative if the EXTPSYNTH
602 assumption is violated. Aside from this, calculating both PSMALL and EXTPSYNTH and subsequently
603 compare their results is always recommended to reveal suspicious deviations.

604 7.2. Auxiliary data

605 The auxiliary data used in our study were derived from two remote sensing sources, i.e. an ALS
606 canopy height model and a tree species classification map. Likewise in many similar studies, the ALS
607 mean canopy height proved to be the explanatory variable with highest predictive power. However,
608 the large time-gaps of up to 10 years between the ALS acquisition and the terrestrial survey date caused
609 the substantial introduction of artificial noise in the data. Whereas a post-stratification to the ALS
610 acquisition years was an effective means to counteract the implied residual inflation, several leverage
611 points were unambiguously caused by the temporal asynchronicity. Undetectable forest loss during
612 the gap between the ALS acquisition and the NFI was also likely a cause for high residual variance
613 in some small area units compared to the terrestrial data variance, which subsequently led to higher
614 variances than the SRS estimator. As opposed to the ALS data, the availability of a country-wide tree
615 species classification map has yet been unique among all German federal states. Whereas the study of
616 Hill *et al.* [34] already showed that the tree species information was able to improve the model fit, it has
617 yet not been used to its full potential. One reason for this was the impossibility of modeling individual
618 tree species within each ALS acquisition year, which would add further explanatory power. Another
619 reason was the lack of available satellite data for classification in some parts of the country, which
620 led to missing values in the inventory data and restricted 19 FR units to a simpler regression model.
621 Promising steps with respect to more up-to-date canopy height information have already been made, as
622 the topographic survey institution of RLP will from this year on provide a country-wide canopy height
623 model derived from aerial imagery acquisitions. These campaigns will in the future be conducted in a

624 two-year period and allow to derive canopy height information matching the dates of terrestrial forest
 625 inventories. A study of Kirchhoefer *et al.* [42] recently indicated that similar model performance for
 626 German NFI data can be achieved using such imagery-based canopy height models. Additionally, the
 627 improved coverage and repetition rate of the Sentinel-2 satellite [43] will allow to produce annually
 628 updated tree species classification maps. We consider these alternative auxiliary data sources to also
 629 solve the problem of missing explanatory variables at inventory plots. One could also make use of
 630 the exhaustive information within the two-phase estimators by using the true auxiliary means [19,24],
 631 which could further decrease estimation errors. Previous studies of Mandallaz *et al.* [19] however
 632 showed that given a reasonable large sample size of the first phase, the differences in the estimation
 633 error are usually small. With respect to the substantial improvements in the temporal synchronicity
 634 between auxiliary and terrestrial inventory data, we consider the demonstrated double-sampling
 635 approach also to be very efficient for the estimation of change [44].

636 8. Conclusion

637 The study led to two major conclusions: (1) the EXTPSYNTH and PSMALL estimator generally
 638 achieved substantially smaller estimation errors on the two investigated forest district levels compared
 639 to the SRS estimator. Thus, the demonstrated small area estimation procedure constitutes a major
 640 contribution to an additional use of the German NFI data for estimation below the federal state
 641 level. Further close cooperation with the forest authorities is crucial to evaluate whether the achieved
 642 error levels are already sufficient enough in order to support forest planning decisions. A first
 643 study will concentrate on testing the EXTPSYNTH and PSMALL confidence intervals as a validation
 644 source for the stand-wise inventories. (2) Despite the quality restrictions, the ALS data and the tree
 645 species map were found to be well suited to model the mean timber volume on the plot and cluster
 646 level. With the prospect of more frequently updated aerial canopy height models and tree species
 647 maps, the two data sources will become even more attractive to be used as an integral part of future
 648 operational applications. The improving availability of remote sensing data will also allow to extent
 649 the demonstrated estimation procedure to the estimation of change. We consider this to be one of the
 650 next milestones towards a future operational use of the demonstrated small area estimation procedure.

651 **Acknowledgments:** We want to express our gratitude to Prof. H. Heinemann (Chair of Land Use Engineering,
 652 ETH Zurich) for supporting this study. We also want to explicitly thank Dr. Johannes Stoffels and Dr. Henning
 653 Buddenbaum from the Environmental Sensing and Geoinformatic Group of University of Trier for providing the
 654 ALS data and tree species classification map, and Dr. Kai Husmann and Dr. Christoph Fischer from the Northwest
 655 German Forest Research Institution Göttingen for their advice in processing the terrestrial inventory data. Special
 656 gratitude is also owed to Dr. Thomas Riedel from the Thünen Institute for providing the densified NFI sample
 657 grid, and Dr. Alexander Massey for proofreading.

658 **Author Contributions:** Andreas Hill conducted the study and wrote the manuscript. Daniel Mandallaz developed
 659 the design-based estimators and supported the statistical analysis. Joachim Langshausen supported the study on
 660 the part of the State Forest Service Rhineland-Palatinate and cross-checked the analysis and the manuscript.

661 **Conflicts of Interest:** The authors declare no conflict of interest.

662 Appendix

663 R-squared on cluster level

664 The R^2 on the cluster level is calculated using the number of plots $M(x)$ of each cluster in order to
 665 weight for the varying number of plots on which $Y_c(x)$ and $\hat{Y}_c(x)$ are based on.

$$R^2 = \frac{\sum_{x \in s_2} \left(\frac{M(x)}{M_2} \right)^2 \left(\hat{Y}_c(x) - \bar{Y}_c \right)^2}{\sum_{x \in s_2} \left(\frac{M(x)}{M_2} \right)^2 \left(Y_c(x) - \bar{Y}_c \right)^2}$$

666 $Y_c(x)$ and $\hat{Y}_c(x)$ are the predicted and observed local densities on the cluster level calculated according
 667 to Equations 2 and 12. \hat{Y}_c is the estimated sample mean corresponding to the weighted mean over all
 668 observed local densities on the cluster level (Eq. 8).

669 *RMSE on cluster level*

670 The same weights $M(x)$ are also applied to calculate the RMSE on the cluster level. n_2 is the
 671 number of clusters used in the modeling frame.

$$\text{RMSE} = \sqrt{\frac{1}{n_2} \sum_{x \in s_2} \left(\frac{M(x)}{\bar{M}_2} \right)^2 \left(\hat{Y}_c(x) - Y_c(x) \right)^2}$$

672 The *relative* or *normalized* RMSE is calculated by dividing the RMSE by the estimated sample mean \hat{Y}_c :

$$\text{RMSE}[\%] = \frac{\text{RMSE}}{\hat{Y}_c}$$

673 Note that the weights $\frac{M(x)}{\bar{M}_2} \equiv 1$ if the number of plots per cluster is constant.

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