

# Ecosystem carbon balance in the Hawaiian Islands under different scenarios of future climate and land use change

Paul C. Selmants<sup>1,6</sup>, Benjamin M. Sleeter<sup>2</sup>, Jinxun Liu<sup>1</sup>, Tamara S. Wilson<sup>1</sup>,

Clay Trauernicht<sup>3</sup>, Abby G. Frazier<sup>4</sup>, Gregory P. Asner<sup>5</sup>

## **Affiliations:**

<sup>1</sup>U.S. Geological Survey, Moffett Field, CA, USA

<sup>2</sup>U.S. Geological Survey, Seattle, WA, USA

<sup>3</sup>University of Hawai‘i at Mānoa, Honolulu, HI, USA

<sup>4</sup>The East-West Center, Honolulu, HI, USA

<sup>5</sup>Arizona State University, Tempe, AZ, USA

<sup>6</sup>Author to whom correspondence should be addressed

**Email:** [pselmants@usgs.gov](mailto:pselmants@usgs.gov)

**Running title:** Hawai‘i carbon balance

**Keywords:** land use, climate change, carbon balance, Hawai‘i, scenarios, disturbance, ecosystem model

**Date:** March 15, 2021

## Abstract

The State of Hawai‘i passed legislation to be carbon neutral by 2045, a goal that will partly depend on carbon sequestration by terrestrial ecosystems. However, there is considerable uncertainty surrounding the future direction and magnitude of the land carbon sink in the Hawaiian Islands. We used simulation modeling to assess how projected future changes in climate and land use will influence ecosystem carbon balance in the Hawaiian Islands under all combinations of two radiative forcing scenarios (RCPs 4.5 and 8.5) and two land use scenarios (low and high) over a 90-year timespan from 2010-2100. Collectively, terrestrial ecosystems of the Hawaiian Islands acted as a net carbon sink under low radiative forcing (RCP 4.5) for the entire 90-year simulation period, with low land use change further enhancing carbon sink strength. In contrast, Hawaiian terrestrial ecosystems transitioned from a net sink to a net source of CO<sub>2</sub> to the atmosphere under high radiative forcing (RCP 8.5), with high land use accelerating this transition and exacerbating net carbon loss. A sensitivity test of the CO<sub>2</sub> fertilization effect on plant productivity revealed it to be a major source of uncertainty in projections of ecosystem carbon balance. Reconciling this uncertainty in how net photosynthesis will respond to rising atmospheric CO<sub>2</sub> will be essential to realistically constrain simulation models used to evaluate the effectiveness of ecosystem-based climate mitigation strategies.

## 1. Introduction

Terrestrial ecosystems are a major sink for atmospheric carbon dioxide (CO<sub>2</sub>), removing ~30% of human emissions on an annual basis and reducing the rate of increase in atmospheric CO<sub>2</sub> (Keenan and Williams 2018, Friedlingstein *et al* 2019). There is increasing recognition among policymakers that natural and agricultural ecosystems can contribute to climate mitigation, which has given rise to the popularity of “natural carbon solutions” (Cameron *et al* 2017). Defined as conservation and land management efforts aimed at enhancing ecosystem carbon storage (Griscom *et al* 2017), natural climate solutions are appealing because they are seen as cost-efficient and readily available (Galarraga *et al* 2017, Cameron *et al* 2017, Fargione *et al* 2018). However, effective implementation

is complicated by the uncertainty surrounding the future direction and magnitude of the land carbon sink, especially at the regional scale. Despite this uncertainty, evidence indicates that both interannual and long-term variability in carbon uptake by land ecosystems is driven primarily by fluctuations in climate, land use, and land cover change (Ahlström *et al* 2015, Prestele *et al* 2017, Friedlingstein *et al* 2019). Incorporating the interactive effects of land use and climate into spatially explicit future projections of ecosystem carbon balance could therefore provide a reference point to evaluate the effectiveness of land-based mitigation. Although a complex challenge, the growing number of sub-national jurisdictions that plan to incorporate land-based mitigation strategies into their emissions reduction efforts would benefit from understanding how future land use and climate-biosphere feedbacks will affect ecosystem carbon balance in their respective regions (Sleeter *et al* 2019).

The State of Hawai‘i exemplifies the challenges associated with projecting the interactive effects of future climate and land use change on ecosystem carbon balance at a regional scale. Hawai‘i was the first U.S. state to enact legislation committing to full carbon neutrality, requiring the state to account for and offset all of its greenhouse gas emissions by 2045 (State of Hawai‘i Acts 15 and 16). This legislation emphasizes the mitigation potential of natural ecosystems as a key component to emissions reduction, necessitating baseline estimates and future projections of land carbon sink strength. However, Hawai‘i’s challenging terrain complicates these assessment efforts. The main Hawaiian Islands are a complex mosaic of natural and human-dominated landscapes overlain by steep climate gradients across relatively short distances (supplmental figure 3), with mean annual temperature ranging from ~4-24° C (Giambelluca *et al* 2014) and mean annual rainfall ranging from ~180-9500 mm (Giambelluca *et al* 2013). Temperatures have risen rapidly in the Hawaiian Islands since the mid 1970s (Giambelluca *et al* 2008) and a long-term drying trend has persisted since the early 1920s (Frazier and Giambelluca 2017), resulting in reduced forest biomass and productivity (Barbosa and Asner 2017). These same drying and warming trends have increased the frequency and intensity of wildland fire (Trauernicht *et al* 2015, Trauernicht 2019) with predictable negative effects on ecosystem carbon balance (Selmants *et al* 2017). Ecosystem carbon stocks across the main Hawaiian Islands have also been strongly influenced by the legacy of past land use change (Osher *et*

70 *al* 2003, Asner *et al* 2011). Thousands of hectares of land were deforested beginning in the late 19th  
71 century to clear land for sugar plantations and cattle pasture (Cuddihy and Stone 1990). Since the  
72 mid-20th century, much of this agricultural land has been steadily converted to urban areas,  
73 commercial forestry plantations, or simply abandoned and colonized by non-native grass species  
74 (Suryanata 2009, Perroy *et al* 2016). Although these past trends surely inform the future impact of  
75 climate and land use change on ecosystem carbon balance, high spatial and temporal heterogeneity  
76 complicates realistic projection efforts. To date only one study has attempted to integrate land use,  
77 climate, and natural disturbances into future projections of Hawaiian ecosystem carbon balance, with  
78 projections limited to the mid-21st century under a single land use change scenario and moderate  
79 radiative forcing (SRES A1B, equivalent to RCP 6; Selman *et al* 2017).

80 We used a stochastic, spatially explicit simulation model to estimate ecosystem carbon balance for  
81 Hawai‘i’s natural and agricultural lands on an annual basis for the period 2010–2100 under a range of  
82 assumptions about future climate, land use, land cover, disturbance, and global CO<sub>2</sub> emissions  
83 (Daniel *et al* 2016, 2018, Sleeter *et al* 2019). We explored four unique scenarios that represent all  
84 combinations of two land use change pathways (low and high) and two radiative forcing pathways  
85 (representative concentration pathway [RCP] 4.5 and RCP 8.5). In addition to these four scenarios,  
86 we conducted a separate series of simulations to examine how ecosystem carbon balance estimates  
87 vary in response to different levels of a CO<sub>2</sub> fertilization effect (CFE) on net primary productivity  
88 (NPP; Sleeter *et al* 2019). our goals were to estimate changes in Hawaiian ecosystem carbon balance  
89 and their uncertainties under a range of plausible future scenarios, quantify the relative impact of  
90 major controlling processes such as land use change, disturbance, and climate change, and assess the  
91 sensitivity of model estimates to the introduction of a CFE on NPP.

## 92 **2. Methods**

93 We used the Land Use and Carbon Scenario Simulator (LUCAS), an integrated landscape change and  
94 carbon gain-loss model, to project changes in ecosystem carbon balance for the seven main Hawaiian

Islands under all combinations of two land use scenarios (low and high) and two radiative forcing scenarios (RCP 4.5 and RCP 8.5). We also developed a separate set of scenarios to test model sensitivity to different levels of a CFE on NPP. The landscape change portion of LUCAS is a state-and-transition model that applies a Monte Carlo approach to track the state type and age of each simulation cell in response to a pre-determined set of transitions (Daniel *et al* 2016). The carbon gain-loss portion tracks carbon stocks within each simulation cell over time as continuous state variables, along with a pre-defined set of continuous flows specifying rates of change in stock levels over time (Daniel *et al* 2018, Sleeter *et al* 2019). We parameterized the Hawai‘i LUCAS model to estimate annual changes in carbon stocks and fluxes in response to land use, land use change, wildland fire, and long-term climate variability for the time period 2010-2100.

## **2.1 Study area**

The spatial extent of this study was the terrestrial portion of the seven main Hawaiian Islands (figure 1), a total land area of 16,554 km<sup>2</sup>. We subdivided this landscape into a grid of 264,870 simulation cells, each of which was 250 x 250 m in size. Each simulation cell was assigned to one of 210 possible state types based on the unique combination of three moisture zones (dry, mesic, and wet; supplemental figure 1), seven islands, and ten discrete land cover classes (figure 1).

## **2.2 States and transitions**

We developed two land use scenarios (low and high) with transition pathways modified from Daniel *et al* (2016). Transitions between state types were pre-defined to represent urbanization, agricultural contraction, agricultural expansion, harvesting of tree plantations, and wildfire. Agriculture, forest, grassland, shrubland, and tree plantation state types each had multiple transition pathways, while the barren state type could only transition to developed (i.e., urbanization). Although most state types had an urbanization transition pathway, there was no transition pathway out of an urbanized (developed) state. Water and wetland state types remained static throughout the simulation period.

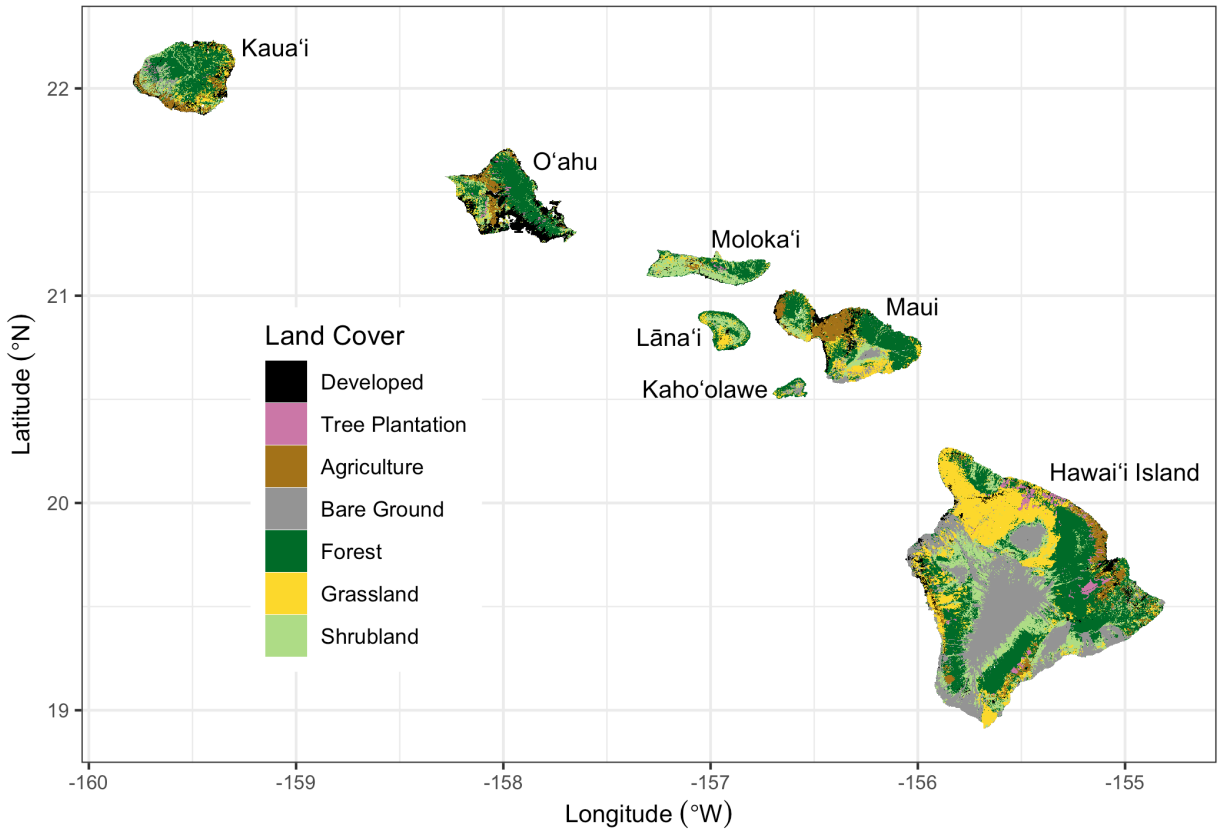


Figure 1: Land cover classification of the seven main Hawaiian Islands, adapted from Jacobi *et al.* (2017). Agriculture in this map combines herbaceous and woody crops, but these two crop types are treated as separate land cover classes in the simulation model. Water and Wetland land cover classes are not shown.

Transition targets were based on historical trends of land use change in the Hawaiian Islands from 1992-2011 (NOAA 2020) and on population projections for the State of Hawai'i (Kim and Bai 2018). For the high land use scenario, transition rates for each timestep and Monte Carlo realization were sampled from uniform distributions bounded by the median and maximum historical rates of agricultural contraction, agricultural expansion, and urbanization for each island. For the low land use scenario, rates of agricultural contraction and expansion were sampled from uniform distributions bounded by zero and the minimum historical rates for each island. Urbanization rates in the low land use scenario were based on island-level population estimates and projections at five year intervals from 2010-2045 (Kim and Bai 2018). We converted population projections into urbanization transition targets following Sleeter *et al* (2017) by calculating population density for each island and then projecting future developed area based on the five-year incremental change in island population. The spatial extent of agricultural contraction, agricultural expansion, and urbanization was constrained in both land use scenarios based on existing zoning maps (Daniel *et al* 2016). Transition targets for tree plantation harvest were set at ~75% of recent historical rates in the high land use scenario and ~40% of recent historical rates in the low land use scenario (Daniel *et al* 2016). In both land use scenarios, approximately 60% of tree plantation harvests were replacement harvests resulting in conversion to agriculture. The remaining 40% were rotation harvests replanted to *Eucalyptus* spp.

The wildfire transition sub-model was modified from Daniel *et al* (2016) by incorporating a new 21-year historical wildfire spatial database of the Hawaiian Islands (supplemental figure 2). We used this new spatial database to calculate historical wildfire size distribution and ignition probabilities for each unique combination of moisture zone (supplemental figure 1), island, and state type (figure 1) for the years 1999-2019. Starting in 2020, the number and size of fires was randomly drawn from one of these historical year-sets for each timestep and Monte Carlo realization, using burn severity probabilities from Selmants *et al* (2017). Wildfire in the low land use scenario was sampled from the subset of historical fire years at or below the median area burned statewide from 1999-2019. The high land use scenario sampled from historical fire years above the median area burned over the same 21-year period (supplemental figure 2a).

## 2.3 Carbon stocks and flows

The fate of carbon stocks was tracked for each simulation cell based on a suite of carbon flows (i.e., carbon fluxes) specifying the rates of change in these carbon stocks over time (Daniel *et al* 2018, Sleeter *et al* 2019). We defined carbon stocks as continuous state variables for each simulation cell, including live biomass, standing dead wood, down dead wood, litter, and soil organic carbon. We also included and tracked carbon in atmospheric, aquatic, and harvest product pools to enforce carbon mass balance (Daniel *et al* 2018). To transfer carbon between stocks, we defined baseline carbon flows as continuous variables resulting from growth, mortality, deadfall, woody decay, litter decomposition, and leaching (which includes runoff). We also defined carbon flows resulting from land use, land use change, and wildfire (Selmants *et al* 2017, Daniel *et al* 2018).

Initial carbon stocks and baseline carbon flows were estimated based on the moisture zone (supplmental figure 1), state type, and age of each simulation cell using a lookup table derived from the Integrated Biosphere Simulator (IBIS; Foley *et al* 1996, Liu *et al* 2020), a process-based dynamic global vegetation model. We initiated IBIS with minimal vegetation and simulated forward for 110 years using 30-year climate normals for the Hawaiian Islands (Giambelluca *et al* 2013, 2014). We calibrated IBIS carbon stocks with statewide gridded datasets of soil organic carbon (Soil Survey Staff 2016) and forest aboveground live biomass (Asner *et al* 2016). We also calibrated gross photosynthesis in IBIS using a Hawai‘i-specific gridded dataset derived from MODIS satellite imagery (Kimball *et al* 2017).

Net primary production for each simulation cell was calculated as the mean IBIS-derived value for each combination of moisture zone and state type adjusted with a spatially explicit stationary growth multiplier to reflect local variation driven by microclimate (Sleeter *et al* 2019). We calculated this spatial growth multiplier as the NPP anomaly for each simulation cell relative to mean NPP values for each combination of moisture zone and state type based on empirical relationships between total annual NPP and mean annual rainfall or temperature (Schuur 2003, Del Grosso *et al* 2008) using Hawai‘i-specific climate data (Giambelluca *et al* 2013, 2014). Soil carbon flux to the atmosphere



( $R_h$ ) and aquatic soil carbon losses (leaching and overland flow) were estimated as the ratio of the IBIS-derived flux for each combination of moisture zone and state type to the microclimate-adjusted NPP value for each simulation cell. All other carbon flow rates were estimated as the ratio of the mean IBIS-derived flux for each combination of moisture zone and state type to the size of the originating carbon stock at each age (Sleeter *et al* 2018, Daniel *et al* 2018). Climate change impacts on carbon flows were represented by temporal growth and decay multipliers applied to each simulation cell based on statistically downscaled CMIP5 climate projections for the Hawaiian Islands under each of two radiative forcing scenarios (RCP 4.5 and RCP 8.5; Timm *et al* 2015, Timm 2017). The impact of future changes in rainfall and temperature on NPP were represented by annual growth multipliers calculated using empirical NPP models (Schuur 2003, Del Grosso *et al* 2008) and climate model projections of temperature and rainfall for each radiative forcing scenario. The effect of future warming on turnover rates of dead organic matter were represented by temporal decay multipliers calculated using Q10 functions and climate model temperature projections for each radiative forcing scenario. We applied a Q10 of 2.0 for wood and soil organic matter decay flows (Kurz *et al* 2009, Sleeter *et al* 2019) and a Q10 of 2.17 for litter decay flows (Bothwell *et al* 2014). Transition-triggered carbon flows resulting from disturbances associated with land use change, timber harvesting, and wildfire were based on values from Don *et al* (2011), Selmants *et al* (2017), and Daniel *et al* (2018).

#### **2.4 CO<sub>2</sub> fertilization effect**

Increasing atmospheric CO<sub>2</sub> concentrations stimulate leaf-level photosynthesis, potentially increasing NPP as well (Franks *et al* 2013). However, the magnitude and persistence of this effect is highly uncertain, particularly across a range of climatic conditions and over long time spans (Walker *et al* 2020). Following Sleeter *et al* (2019), we developed a separate set of scenarios designed to test the sensitivity of LUCAS model projections of ecosystem carbon balance to different rates of a CO<sub>2</sub> fertilization effect (CFE). We incorporated a CFE multiplier for NPP that represented the percent increase in NPP for every 100 ppm increase in atmospheric CO<sub>2</sub> concentration under the high land use and high radiative forcing (RCP 8.5) scenario. We tested five CFE levels ranging from 5% to

15%, which is within the range of CFEs observed in free air CO<sub>2</sub> enrichment (FACE) experiments. For all levels, we assumed CFEs reached saturation at an atmospheric CO<sub>2</sub> concentration of 600 ppm, with no further stimulation of NPP despite a continued increase in CO<sub>2</sub> concentration to 930 ppm by 2100. This 600ppm threshold generally coincides with the upper limit from FACE experiments and is reached by the year 2060 under RCP 8.5.

## 2.5 Scenario simulations and analysis

Each of the four unique scenarios were run for 90 years at an annual timestep and repeated for 30 Monte Carlo realizations, using initial conditions corresponding to the year 2010. All simulations were performed within the SyncroSim (version 2.2.4) software framework with ST-Sim (version 3.2.13) and SF (version 3.2.10) add-on modules (Daniel *et al* 2016, 2018). Model inputs and outputs were prepared with the R statistical computing platform (R Core Team 2019) using the tidyverse (Wickham *et al* 2019), raster (Hijmans 2020), and rsyncrosim (Daniel *et al* 2020) packages. Carbon stocks and fluxes for the seven main Hawaiian Islands were calculated for each scenario by summing within each Monte Carlo realization on an annual basis and then calculating annual means as well as the annual upper and lower limits of the 30 Monte Carlo realizations. Carbon balance for the seven main Hawaiian Islands was calculated on annual basis for each scenario and Monte Carlo realization as net biome productivity (NBP), which was equal to annual carbon input in the form of NPP minus the annual sum of all carbon losses from terrestrial ecosystems, including heterotrophic respiration ( $R_h$ ) from litter and soil, carbon fluxes to the atmosphere triggered by land use and land use change, wildfire emissions, and aquatic carbon losses through leaching and overland flow. Positive NBP values indicated ecosystems of the seven main Hawaiian Islands were acting as a net sink for atmospheric CO<sub>2</sub>, while negative NBP values indicated that these ecosystems were acting as a net carbon source to the atmosphere (Chapin *et al* 2006).

### 3. Results

#### 3.1 Carbon stocks and fluxes

Terrestrial ecosystems of the seven main Hawaiian Islands stored an estimated 316 Tg of carbon at the beginning of the simulation period in 2010 (figure 2a), with 58% in soil organic matter, 22% in living biomass, and 20% in surface dead organic matter (litter and dead wood; figure 2b). Ecosystems accumulated carbon in all scenarios but at different rates, with trajectories shaped primarily by climate change and to a lesser extent by land use change. The highest and most consistent projected accumulation of ecosystem carbon occurred under the combination of low radiative forcing and low land use change, yielding a ~15% increase in ecosystem carbon to an average of 363 Tg by 2100 (figure 2a). In contrast, high radiative forcing and high land use change resulted in the lowest ecosystem carbon gain, reaching a peak of ~332 Tg in 2063 and a decline to 327 Tg in 2100, resulting in a net increase of only 3% by the end of the simulation period (figure 2a). Ecosystem carbon accumulation was driven exclusively by increasing soil organic carbon across all four scenarios, all other stocks declined over time (figure 2b).

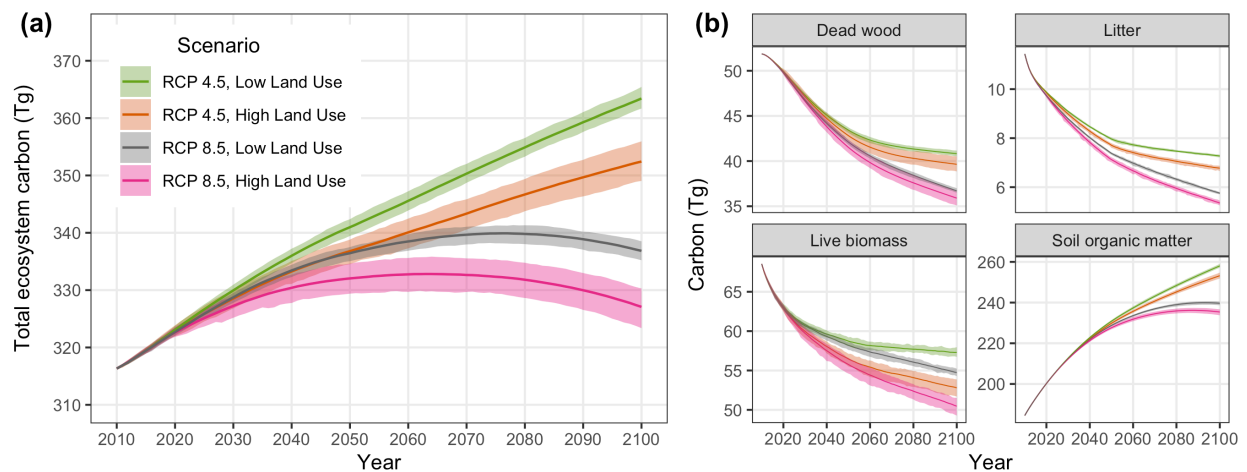


Figure 2: Projected changes in total ecosystem carbon storage (a) and individual carbon stocks (b) for the seven main Hawaiian Islands. Solid lines indicate the mean of 30 Monte Carlo realizations for each scenario, with shaded areas indicating the minimum and maximum range of Monte Carlo realizations.

Net primary production (NPP) for the seven main Hawaiian Islands declined across all four scenarios,

driven primarily by climate change and to a lesser extent by land use change (Fig. 3). The combination of high radiative forcing (RCP 8.5) and high land use change led to the steepest decline in NPP over time, driven by intense long-term drying on the leeward sides of islands under RCP 8.5 (supplemental figure 4) and sustained losses of forest and shrubland land area in the high land use scenario (supplmental figure 5). In contrast, climate change led to increased heterotrophic respiration ( $R_h$ ) over time, such that more intense warming under RCP 8.5 (supplmental figure 4) resulted in  $R_h$  being ~ 3% higher by 2100 than under RCP 4.5 (figure 3). Heterotrophic respiration declined substantially over time in the high land use scenario (figure 3) because of long-term reductions in forest and shrubland land area (supplemental figure 5), similar to trends in NPP. Transition-triggered carbon fluxes to the atmosphere from land use, land use change, and wildfire were largely independent of changes in climate, stabilizing by mid-century at an average of ~0.4 Tg y<sup>-1</sup> in the high land use scenario and ~0.2 Tg y<sup>-1</sup> in the low land use scenario (figure 3). Uncertainty around transition-triggered carbon fluxes were higher in the high land use scenario, driven primarily by greater variability in wildland fire probabilities.

### ***3.2 Ecosystem carbon balance***

Net biome productivity (NBP) averaged approximately 0.6 Tg C y<sup>-1</sup> at the start of the simulation period and declined over time in all four scenarios (figure 4). On average, terrestrial ecosystems of the seven main Hawaiian Islands collectively acted as a net carbon sink throughout the simulation period under the RCP 4.5 radiative forcing scenario, but carbon sink strength was ~40% lower in the high land use scenario compared to the low land use scenario by the end of the simulation period (figure 4). In contrast, ecosystems of the Hawaiian Islands acted as a net carbon source to the atmosphere toward the latter half of the simulation period under RCP 8.5, with the transition from sink to source occurring 15 years earlier on average in the high land use scenario than in the low land use scenario (figure 4). The high land use scenario under RCP 8.5 represented a ~40% larger net source of carbon to the atmosphere by the year 2100 than the low-land use scenario under the same radiative forcing. Over the entire simulation period, both global emissions reductions and local avoided land conversion



Figure 3: Projected changes in net primary production (NPP), heterotrophic respiration (Rh) and carbon fluxes induced by land use and land use change for the seven main Hawaiian Islands. Solid lines indicate the mean of 30 Monte Carlo realizations for each scenario, with shaded areas indicating the minimum and maximum range of Monte Carlo realizations.

resulted in substantial increases in cumulative NBP (figure 5). However, switching from RCP 8.5 to RCP 4.5 increased cumulative NBP in the Hawaiian Islands more than twice as much as reducing emissions from local land use change and wildfire disturbance (figure 5). Switching from RCP 8.5 to RCP 4.5 under the low land use scenario yielded the greatest cumulative increase in NBP, resulting in a median gain of 26.5 Tg of carbon over the entire 90-year simulation period.

### ***3.3 CO<sub>2</sub> fertilization effect***

Projected estimates of both total ecosystem carbon storage and ecosystem carbon balance were highly sensitive to differing rates of a CFE on plant productivity. Under the high radiative forcing (RCP 8.5) and high land use scenario, the inclusion of a CFE ranging from 5-15% led to ~33-98 Tg of additional carbon storage in ecosystems by the end of the century, a ~10-30% increase (figure 6a). Compared to the reference scenario (0% CFE), a 5% CFE was sufficient to transform Hawaiian Island ecosystems from a net carbon source to the atmosphere during the latter half of the 21st century (figure 4b) to a net carbon sink for the entire simulation period (figure 6b), completely offsetting all other carbon losses induced by high radiative forcing and high land use. Net carbon sink strength was further enhanced at higher CFE rates, with NBP increasing by an average of 0.07 Tg C y<sup>-1</sup> for each 1% increase in CFE (figure 6b). When compared to other scenarios, applying a 5% CFE to the high radiative forcing and high land use scenario resulted in a mean annual NBP of  $0.46 \pm 0.3$  Tg C y<sup>-1</sup>, roughly equivalent to mean annual NBP in the low radiative forcing and low land use scenario with no CFE ( $0.52 \pm 0.12$ ). A 15% CFE applied to the high radiative forcing and high land use scenario resulted in a mean annual NBP of  $1.18 \pm 0.29$  Tg C y<sup>-1</sup>, more than double that of the low radiative forcing and low land use scenario with no CFE.



Figure 4: Projected changes in net biome productivity (NBP) for the seven main Hawaiian Islands. Values above zero indicate terrestrial ecosystems are acting as a net carbon sink for atmospheric carbon and values below zero indicate ecosystems are acting as a net carbon source to the atmosphere. Solid lines indicate the mean of 30 Monte Carlo realizations for each scenario, with shaded areas indicating the minimum and maximum range of Monte Carlo realizations. The dashed horizontal line in each panel represents the boundary between ecosystems acting as a net carbon sink (positive NBP values) and a net carbon source (negative NBP values).



Figure 5: Projected changes in cumulative net biome productivity (NBP) for the seven main Hawaiian Islands when switching from the high to low land use change scenario under each radiative forcing scenario (top panel) and when switching from the high (RCP 8.5) to low (RCP 4.5) radiative forcing scenario under each land use scenario (bottom panel). Box plots indicate the median (vertical black line), 25th and 75th percentiles (colored boxes), 10th and 90th percentiles (thin horizontal lines), and values outside of this range (black circles). Note the different x-axis scales in each panel.



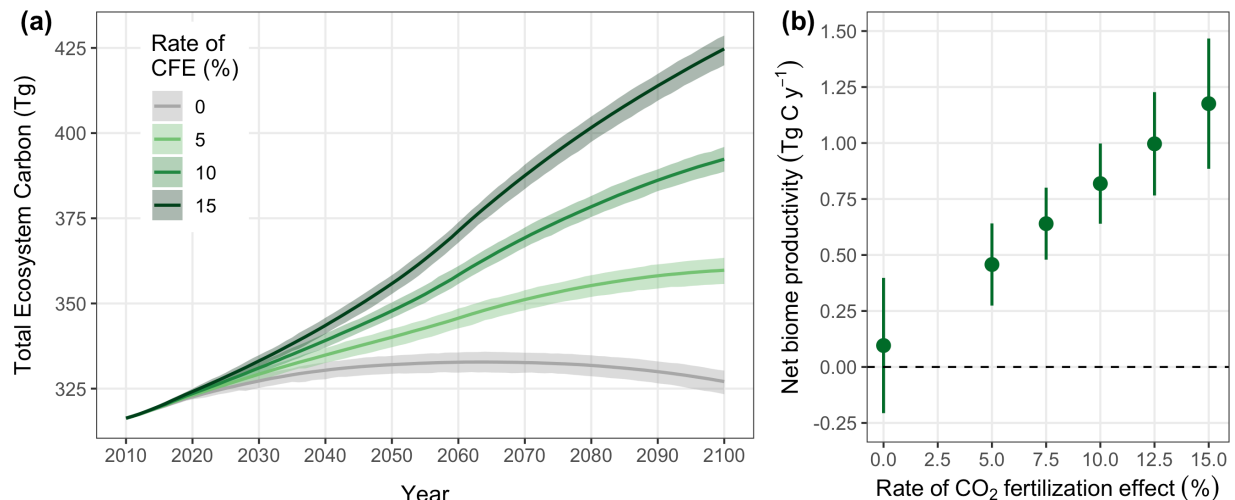


Figure 6: Sensitivity of projected changes in total ecosystem carbon storage (a) and mean annual net biome productivity (b) to different rates of carbon dioxide fertilization in the seven main Hawaiian Islands under the RCP 8.5 radiative forcing and high land use scenario. The carbon dioxide fertilization effect (CFE) is the percent change in net primary productivity (NPP) for every 100 ppm increase in atmospheric carbon dioxide. The CFE for all rates is capped at 600 ppm, which is achieved around the year 2060. Solid lines in (a) indicate the mean total ecosystem carbon storage across 30 Monte Carlo realizations for each CFE rate, with shaded areas indicating the minimum and maximum range of Monte Carlo realizations. Solid circles in (b) represent mean annual net biome productivity averaged across all years and Monte Carlo realizations for each CFE rate, with vertical lines indicating the standard deviation of the mean. The dashed horizontal line in (b) represents the boundary between ecosystems acting as a net carbon sink (positive NBP values) and a net carbon source (negative NBP values).

## 4. Discussion

We estimated that terrestrial ecosystems of the Hawaiian Islands have been a consistent net sink for atmospheric carbon over the last decade (figure 4). For the time period 2011-2019, net biome productivity (NBP) averaged  $0.64 \text{ TgC y}^{-1}$  and ranged from  $0.46$  to  $0.88 \text{ TgC y}^{-1}$  across all scenarios. Based on this mean annual NBP estimate, Hawaiian terrestrial ecosystems offset approximately 13% of 2015 statewide CO<sub>2</sub> emissions from energy production and transportation ( $5.04 \text{ TgC}$ ), the State of Hawai‘i’s largest source of greenhouse gas emissions (State of Hawai‘i 2019). Future projections indicate Hawaiian terrestrial ecosystems will continue to be a net sink for atmospheric carbon if global CO<sub>2</sub> emissions peak around 2040 and then decline (RCP 4.5), and that carbon sink strength can be further enhanced by reducing the intensity and extent of future land use change. If, however,

global CO<sub>2</sub> emissions continue to rise throughout the 21st century (RCP 8.5), our projections indicate Hawaiian ecosystems will transition from a net sink to a net source of CO<sub>2</sub> to the atmosphere, with high levels of land use change accelerating this transition and exacerbating net carbon loss. Our model results also indicate that projections of ecosystem carbon balance are highly sensitive to the introduction of a CFE. Even a 5% increase in NPP for every 100ppm increase in atmospheric CO<sub>2</sub> was sufficient to completely offset all other carbon losses induced by the high radiative forcing and high land use scenario, maintaining Hawaiian Island ecosystems as a net carbon sink for the entire simulation period instead of transitioning to a net carbon source by mid-century. Reconciling the high uncertainty surrounding the response of net photosynthesis to rising atmospheric CO<sub>2</sub> is essential to more realistically constrain model projections of ecosystem carbon balance.

#### *4.1 Impact of different climate and land use pathways*

By comparing ecosystem carbon balance estimates under different scenario combinations, we were able to assess the relative impact of both global emissions reductions and regional actions to reduce emissions from land use, land use change, and wildland fire (figure 5). Global adherence to a lower emissions trajectory (i.e., switching from RCP 8.5 to RCP 4.5) had the largest impact, resulting in a median cumulative increase of 26 Tg C sequestered by Hawaiian ecosystems over the 90-year simulation period. Long-term reductions in the intensity of land use change also consistently led to an increase in ecosystem carbon sequestration, but to a lesser degree than global emissions reductions. Switching from the high to the low land use scenario resulted in a median cumulative retention of an additional 11.6 Tg C in Hawaiian ecosystems by 2100. The combination of global climate mitigation and local reductions in land use conversion had the largest potential benefit to ecosystem carbon sequestration, reducing cumulative net losses by over 400% (37.7 Tg C). Notably, the relative impact of reducing emissions from land use change was much greater under the high radiative forcing pathway (RCP 8.5). Cumulative NBP increased by 130% when switching from the high to low land use scenario under RCP 8.5, as opposed to a 37% cumulative increase in NBP when switching from high to low land use under RCP 4.5. These results demonstrate that reducing ecosystem carbon losses

from land use change, harvest, and wildland fire can be an important component of greenhouse gas reduction efforts by sub-national jurisdictions like the State of Hawai‘i, regardless of the global emissions trajectory. These results also highlight the utility of Hawai‘i’s multi-pronged approach of participating in global climate mitigation efforts by reducing emissions from the energy and transportation sectors while also reducing land use emissions to minimize positive feedbacks to the climate system.

## **4.2 Comparison to other studies**

There are few estimates of contemporary ecosystem carbon balance for the main Hawaiian Islands, and even fewer model projections of future ecosystem carbon balance in response to climate and land use change. Our mean annual NBP estimate of  $0.64 \text{ TgC y}^{-1}$  for the period 2011-2020 agrees well with a recent State of Hawai‘i Greenhouse Gas Inventory report, which estimated an annual net carbon sink of  $0.66 \text{ Tg C}$  in 2015 from agriculture, forestry, and other land uses (State of Hawai‘i 2019). In contrast, our NBP estimate for the past decade was  $\sim 88\%$  higher than a previous statewide LUCAS model estimate covering the same time period ( $0.341 \text{ Tg C y}^{-1}$ ; Selmants *et al* 2017). This discrepancy was likely driven by modifications in how we calculated NPP, soil  $R_h$ , and soil aquatic carbon loss compared to previous versions of the LUCAS model, as well our model’s finer spatial resolution (Selmants *et al* 2017, Daniel *et al* 2018). Previous versions of a Hawai‘i LUCAS model were run at 1-km spatial resolution and simulation cells within each unique combination of moisture zone and state type all had the same mean IBIS-derived NPP value applied to them at the beginning of the simulation period. In contrast, our NPP estimates at 250-m spatial resolution were adjusted on a cell-by-cell basis using Hawai‘i-specific climate data as described in section 2.3. As a result, our statewide NPP estimates from 2011-2020 were 9.5% lower on average than previous LUCAS model estimates for Hawai‘i during the same time period (Selmants *et al* 2017), likely because of the greater influence of more arid simulation cells. Soil carbon losses via  $R_h$ , leaching, and overland flow in previous versions of the LUCAS model were calculated as the ratio of the IBIS-derived flux to the size of the originating carbon stock, in this case soil organic carbon to 1-m depth (Daniel *et al* 2018).

Here we calculated soil  $R_h$  and aquatic carbon losses as the ratio of the mean IBIS-derived flux to the microclimate-adjusted NPP value of each simulation cell, which is a more realistic driver than stock size (Jackson *et al* 2017). Compared to previous Hawai'i LUCAS model estimates (Selmants *et al* 2017), soil  $R_h$  and aquatic carbon losses from 2011-2020 were reduced by an average of 15% and 21%, respectively, which widened the gap between carbon gain (NPP) and carbon losses and accounted for the overall increase in NBP estimates for this time period.

### 4.3 Limitations of this study

There is ample evidence that increasing atmospheric  $CO_2$  concentrations can stimulate NPP (Norby *et al* 2005, Zhu *et al* 2016), but the magnitude and persistence of this effect remains highly uncertain (Franks *et al* 2013, Walker *et al* 2020). Our results demonstrate that long-term projections of ecosystem carbon balance are highly sensitive to uncertainty in CFE strength. With no CFE, Hawaiian ecosystems transitioned into a net source of  $CO_2$  to the atmosphere beginning in the latter half of the 21st century under the high land use and high radiative forcing scenario. However, a 5% increase in NPP for every 100 ppm increase in atmospheric  $CO_2$  applied to the same scenario resulted in Hawaiian ecosystems remaining a net carbon sink throughout the entire simulation period. A 15% CFE applied to the high land use and high radiative forcing scenario resulted in a nearly 5-fold increase in mean annual NBP averaged across all years and Monte Carlo realizations. Despite this demonstrated sensitivity to a CFE, the combined effect of several potentially attenuating factors makes it difficult to determine the appropriate magnitude of a CFE with any degree of confidence (Walker *et al* 2020). Nitrogen and phosphorus limitation can reduce or eliminate a CFE (Reich *et al* 2006, Norby *et al* 2010, He *et al* 2017, Terrer *et al* 2019), as can water limitation and heat stress (Obermeier *et al* 2017, Birami *et al* 2020). Forest age may also be a factor, with young aggrading forests showing a strong positive growth response to  $CO_2$  fertilization (Walker *et al* 2019), while old-growth forests show little to no response (Jiang *et al* 2020, Yang *et al* 2020). This evidence indicates that a CFE may be highly variable across space and time, suggesting that it may be unrealistic to apply a single CFE rate value across an entire region over several decades. Until

mechanistic understanding is improved, the most conservative approach when projecting future ecosystem balance in the context of climate mitigation planning may be to assume no CFE, with the knowledge that any realized CFE will only enhance ecosystem carbon sequestration.

## 5. Conclusion

## Acknowledgements

This study was funded by the U.S. Geological Survey Biological Carbon Sequestration Program. Thanks to Leonardo Frid and Colin Daniel of ApexRMS for assistance with SyncroSim software, and to Nicholas Koch of Forest Solutions, Inc. for information on *Eucalyptus* spp. harvesting in Hawai‘i. Thanks also to Christian Giardina and Zhiliang Zhu for providing the initial impetus for this research. Any use of trade, firm, or product names is for descriptive purposes only and does not imply endorsement by the U.S. Government.

## Data and code availability

Tabular model output data and metadata are available in machine readable format from the USGS ScienceBase data repository at <https://doi.org/10.5066/P9AWLFKZ>. Model input data and R code used to format input data, summarize output data, and compile this manuscript are available from a GitHub repository at [https://github.com/selmants/HI\\_Model](https://github.com/selmants/HI_Model).

## ORCID

Paul C. Selman <https://orcid.org/0000-0001-6211-3957>

Benjamin M. Sleeter <https://orcid.org/0000-0003-2371-9571>

Jinxun Liu <https://orcid.org/0000-0003-0561-8988>

Tamara S. Wilson <https://orcid.org/0000-0001-7399-7532>

Abby G. Frazier <https://orcid.org/0000-0003-4076-4577>

Gregory P. Asner <https://orcid.org/0000-0001-7893-6421>

## References

- Ahlström A, Raupach M R, Schurgers G, Smith B, Arneth A, Jung M, Reichstein M, Canadell J G, Friedlingstein P, Jain A K, Kato E, Poulter B, Sitch S, Stocker B D, Viovy N, Wang Y P, Wiltshire A, Zaehle S and Zeng N 2015 The dominant role of semi-arid ecosystems in the trend and variability of the land CO<sub>2</sub> sink *Science* **348** 895–9 Online: <https://science.sciencemag.org/content/348/6237/895>
- Asner G P, Hughes R F, Mascaro J, Uowolo A L, Knapp D E, Jacobson J, Kennedy-Bowdoin T and Clark J K 2011 High-resolution carbon mapping on the million-hectare Island of Hawaii *Frontiers in Ecology and the Environment* **9** 434–9 Online: <http://doi.wiley.com/10.1890/100179>
- Asner G P, Sousan S, Knapp D E, Selmants P C, Martin R E, Hughes R F and Giardina C P 2016 Rapid forest carbon assessments of oceanic islands: A case study of the Hawaiian archipelago *Carbon Balance and Management* **11** Online: <http://www.cbmjournal.com/content/11/1/1>
- Barbosa J M and Asner G P 2017 Effects of long-term rainfall decline on the structure and functioning of Hawaiian forests *Environmental Research Letters* **12** 094002 Online: <https://doi.org/10.1088%2F1748-9326%2Faa7ee4>
- Birami B, Nägele T, Gattmann M, Preisler Y, Gast A, Arneth A and Ruehr N K 2020 Hot drought reduces the effects of elevated CO<sub>2</sub> on tree water-use efficiency and carbon metabolism *New Phytologist* **226** 1607–21 Online: <https://nph.onlinelibrary.wiley.com/doi/abs/10.1111/nph.16471>
- Bothwell L D, Selmants P C, Giardina C P and Litton C M 2014 Leaf litter decomposition rates increase with rising mean annual temperature in Hawaiian tropical montane wet forests *PeerJ* **2** e685 Online: <https://peerj.com/articles/685>

414 Cameron D R, Marvin D C, Remucal J M and Passero M C 2017 Ecosystem management and land  
 415 conservation can substantially contribute to California's climate mitigation goals *Proceedings of the*  
 416 *National Academy of Sciences* **114** 12833–8 Online: <http://www.pnas.org/content/114/48/12833>

417 Chapin F S, Woodwell G M, Randerson J T, Rastetter E B, Lovett G M, Baldocchi D D, Clark D A,  
 418 Harmon M E, Schimel D S, Valentini R, Wirth C, Aber J D, Cole J J, Goulden M L, Harden J W,  
 419 Heimann M, Howarth R W, Matson P A, McGuire A D, Melillo J M, Mooney H A, Neff J C,  
 420 Houghton R A, Pace M L, Ryan M G, Running S W, Sala O E, Schlesinger W H and Schulze E-D  
 421 2006 Reconciling carbon-cycle concepts, terminology, and methods *Ecosystems* **9** 1041–50

422 Cuddihy L W and Stone C P 1990 *Alteration of Hawaiian vegetation: Effects of humans, their*  
 423 *activities and introductions* (Honolulu, Hawaii: University of Hawaii Press)

424 Daniel C, Hughes J, Embrey A, Frid L and Lucet V 2020 *Rsyncrosim: The r interface to syncrosim*  
 425 Online: <https://github.com/rsyncrosim/rsyncrosim>

426 Daniel C J, Frid L, Sleeter B M and Fortin M-J 2016 State-and-transition simulation models: A  
 427 framework for forecasting landscape change *Methods in Ecology and Evolution* **7** 1413–23

428 Daniel C J, Sleeter B M, Frid L and Fortin M-J 2018 Integrating continuous stocks and flows into  
 429 state-and-transition simulation models of landscape change *Methods in Ecology and Evolution* **9**  
 430 1133–43

431 Del Grosso S, Parton W, Stohlgren T, Zheng D, Bachelet D, Prince S, Hibbard K and Olson R 2008  
 432 Global potential net primary production predicted from vegetation class, precipitation, and  
 433 temperature *Ecology* **89** 2117–26

434 Don A, Schumacher J and Freibauer A 2011 Impact of tropical land-use change on soil organic  
 435 carbon stocks – a meta-analysis *Global Change Biology* **17** 1658–70

436 Fargione J E, Bassett S, Boucher T, Bridgman S D, Conant R T, Cook-Patton S C, Ellis P W, Falcucci  
 437 A, Fourqurean J W, Gopalakrishna T, Gu H, Henderson B, Hurteau M D, Kroeger K D, Kroeger T,  
 438 Lark T J, Leavitt S M, Lomax G, McDonald R I, Megonigal J P, Miteva D A, Richardson C J,

439 Sanderman J, Shoch D, Spawn S A, Veldman J W, Williams C A, Woodbury P B, Zganjar C,  
 440 Baranski M, Elias P, Houghton R A, Landis E, McGlynn E, Schlesinger W H, Siikamaki J V,  
 441 Sutton-Grier A E and Griscom B W 2018 Natural climate solutions for the United States *Science*  
 442 *Advances* **4** eaat1869 Online: <https://advances.sciencemag.org/content/4/11/eaat1869>

443 Foley J A, Prentice I C, Ramankutty N, Levis S, Pollard D, Sitch S and Haxeltine A 1996 An  
 444 integrated biosphere model of land surface processes, terrestrial carbon balance, and vegetation  
 445 dynamics *Global Biogeochemical Cycles* **10** 603–28

446 Franks P J, Adams M A, Amthor J S, Barbour M M, Berry J A, Ellsworth D S, Farquhar G D,  
 447 Ghannoum O, Lloyd J, McDowell N, Norby R J, Tissue D T and Caemmerer S von 2013 Sensitivity  
 448 of plants to changing atmospheric CO<sub>2</sub> concentration: From the geological past to the next century  
 449 *New Phytologist* **197** 1077–94 Online: <http://doi.wiley.com/10.1111/nph.12104>

450 Frazier A G and Giambelluca T W 2017 Spatial trend analysis of Hawaiian rainfall from 1920 to  
 451 2012 *International Journal of Climatology* **37** 2522–31 Online:  
 452 <https://rmets.onlinelibrary.wiley.com/doi/abs/10.1002/joc.4862>

453 Friedlingstein P, Jones M W, O’Sullivan M, Andrew R M, Hauck J, Peters G P, Peters W, Pongratz J,  
 454 Sitch S, Quéré C L, Bakker D C E, Canadell J G, Ciais P, Jackson R B, Anthoni P, Barbero L, Bastos  
 455 A, Bastrikov V, Becker M, Bopp L, Buitenhuis E, Chandra N, Chevallier F, Chini L P, Currie K I,  
 456 Feely R A, Gehlen M, Gilfillan D, Gkritzalis T, Goll D S, Gruber N, Gutekunst S, Harris I, Haverd V,  
 457 Houghton R A, Hurtt G, Ilyina T, Jain A K, Joetzjer E, Kaplan J O, Kato E, Klein Goldewijk K,  
 458 Korsbakken J I, Landschützer P, Lauvset S K, Lefèvre N, Lenton A, Lienert S, Lombardozzi D,  
 459 Marland G, McGuire P C, Melton J R, Metzl N, Munro D R, Nabel J E M S, Nakaoka S-I, Neill C,  
 460 Omar A M, Ono T, Peregon A, Pierrot D, Poulter B, Rehder G, Resplandy L, Robertson E,  
 461 Rödenbeck C, Séférian R, Schwinger J, Smith N, Tans P P, Tian H, Tilbrook B, Tubiello F N, Werf G  
 462 R van der, Wiltshire A J and Zaehle S 2019 Global Carbon Budget 2019 *Earth System Science Data*  
 463 **11** 1783–838 Online: <https://essd.copernicus.org/articles/11/1783/2019/>

464 Galarraga I, Murieta E S de and França J 2017 Climate policy at the sub-national level *Trends in*



465 *Climate Change Legislation* ed A Averchenkova, S Fankhauser and M Nachmany (Edward Elgar  
 466 Publishing) pp 143–74 Online: <https://www.elgaronline.com/view/9781786435774.00018.xml>

467 Giambelluca T W, Chen Q, Frazier A G, Price J P, Chen Y-L, Chu P-S, Eischeid J K and Delporte D  
 468 M 2013 Online Rainfall Atlas of Hawai‘i *Bull. Amer. Meteor. Soc.* **94** 313–6

469 Giambelluca T W, Diaz H F and Luke M S A 2008 Secular temperature changes in Hawai‘i  
 470 *Geophysical Research Letters* **35** Online:  
 471 <https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1029/2008GL034377>

472 Giambelluca T W, Shuai X, Barnes M L, Alliss R J, Longman R J, Miura T, Chen Q, Frazier A G,  
 473 Mudd R G, Cuo L and Businger A D 2014 *Evapotranspiration of Hawai‘i* Online:  
 474 <http://evapotranspiration.geography.hawaii.edu/downloads.html>

475 Griscom B W, Adams J, Ellis P W, Houghton R A, Lomax G, Miteva D A, Schlesinger W H, Shoch  
 476 D, Siikamäki J V, Smith P, Woodbury P, Zganjar C, Blackman A, Campari J, Conant R T, Delgado C,  
 477 Elias P, Gopalakrishna T, Hamsik M R, Herrero M, Kiesecker J, Landis E, Laestadius L, Leavitt S M,  
 478 Minnemeyer S, Polasky S, Potapov P, Putz F E, Sanderman J, Silvius M, Wollenberg E and Fargione  
 479 J 2017 Natural climate solutions *Proceedings of the National Academy of Sciences* **114** 11645–50  
 480 Online: <https://www.pnas.org/content/114/44/11645>

481 He L, Chen J M, Croft H, Gonsamo A, Luo X, Liu J, Zheng T, Liu R and Liu Y 2017 Nitrogen  
 482 Availability Dampens the Positive Impacts of CO<sub>2</sub> Fertilization on Terrestrial Ecosystem Carbon and  
 483 Water Cycles *Geophysical Research Letters* **44** 11, 590–11, 600 Online:  
 484 <https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1002/2017GL075981>

485 Hijmans R J 2020 *Raster: Geographic data analysis and modeling* Online:  
 486 <https://CRAN.R-project.org/package=raster>

487 Jackson R B, Lajtha K, Crow S E, Hugelius G, Kramer M G and Piñeiro G 2017 The Ecology of Soil  
 488 Carbon: Pools, Vulnerabilities, and Biotic and Abiotic Controls *Annual Review of Ecology,*  
 489 *Evolution, and Systematics* **48** 419–45 Online:

490 <http://www.annualreviews.org/doi/10.1146/annurev-ecolsys-112414-054234>

491 Jacobi J, Price J, Gon III S and Berkowitz P 2017 Hawaii Land Cover and Habitat Status: U.S.  
 492 Geological Survey data release Online: <https://doi.org/10.5066/F7DB80B9>

493 Jiang M, Medlyn B E, Drake J E, Duursma R A, Anderson I C, Barton C V M, Boer M M, Carrillo Y,  
 494 Castañeda-Gómez L, Collins L, Crous K Y, De Kauwe M G, Santos B M dos, Emmerson K M, Facey  
 495 S L, Gherlenda A N, Gimeno T E, Hasegawa S, Johnson S N, Kännaste A, Macdonald C A, Mahmud  
 496 K, Moore B D, Nazaries L, Neilson E H J, Nielsen U N, Niinemets Ü, Noh N J, Ochoa-Hueso R,  
 497 Pathare V S, Pendall E, Pihlblad J, Piñeiro J, Powell J R, Power S A, Reich P B, Renchon A A,  
 498 Riegler M, Rinnan R, Rymer P D, Salomón R L, Singh B K, Smith B, Tjoelker M G, Walker J K M,  
 499 Wujeska-Klaus A, Yang J, Zaehle S and Ellsworth D S 2020 The fate of carbon in a mature forest  
 500 under carbon dioxide enrichment *Nature* **580** 227–31 Online:  
 501 <https://www.nature.com/articles/s41586-020-2128-9>

502 Keenan T and Williams C 2018 The Terrestrial Carbon Sink *Annual Review of Environment and*  
 503 *Resources* **43** 219–43 Online: <https://doi.org/10.1146/annurev-environ-102017-030204>

504 Kim Y-S and Bai J 2018 *Population and economic projections for the State of Hawaii to 2045*  
 505 (Hawaii Department of Business, Economic Development & Tourism) Online:  
 506 <https://dbedt.hawaii.gov/economic/economic-forecast/2045-long-range-forecast/>

507 Kimball H L, Selman P C, Moreno A, Running S W and Giardina C P 2017 Evaluating the role of  
 508 land cover and climate uncertainties in computing gross primary production in Hawaiian Island  
 509 ecosystems *PLOS ONE* **12** e0184466 Online:  
 510 <http://journals.plos.org/plosone/article?id=10.1371/journal.pone.0184466>

511 Kurz W A, Dymond C C, White T M, Stinson G, Shaw C H, Rampley G J, Smyth C, Simpson B N,  
 512 Neilson E T, Trofymow J A, Metsaranta J and Apps M J 2009 CBM-CFS3: A model of  
 513 carbon-dynamics in forestry and land-use change implementing IPCC standards *Ecological*  
 514 *Modelling* **220** 480–504 Online:

515 <http://www.sciencedirect.com/science/article/pii/S0304380008005012>

516 Liu J, Sleeter B M, Zhu Z, Loveland T R, Sohl T, Howard S M, Key C H, Hawbaker T, Liu S, Reed B,  
517 Cochrane M A, Heath L S, Jiang H, Price D T, Chen J M, Zhou D, Bliss N B, Wilson T, Sherba J, Zhu  
518 Q, Luo Y and Poulter B 2020 Critical land change information enhances the understanding of carbon  
519 balance in the United States *Global Change Biology* **26** 3920–9 Online:  
520 <https://onlinelibrary.wiley.com/doi/abs/10.1111/gcb.15079>

521 NOAA 2020 *Coastal Change Analysis Program (C-CAP) Regional Land Cover: Hawaii* (NOAA  
522 Office of Coastal Management) Online: <https://coast.noaa.gov/digitalcoast/data/>

523 Norby R J, DeLucia E H, Gielen B, Calfapietra C, Giardina C P, King J S, Ledford J, McCarthy H R,  
524 Moore D J P, Ceulemans R, Angelis P D, Finzi A C, Karnosky D F, Kubiske M E, Lukac M, Pregitzer  
525 K S, Scarascia-Mugnozza G E, Schlesinger W H and Oren R 2005 Forest response to elevated CO<sub>2</sub> is  
526 conserved across a broad range of productivity *Proceedings of the National Academy of Sciences*  
527 **102** 18052–6 Online: <https://www.pnas.org/content/102/50/18052>

528 Norby R J, Warren J M, Iversen C M, Medlyn B E and McMurtrie R E 2010 CO<sub>2</sub> enhancement of  
529 forest productivity constrained by limited nitrogen availability *Proceedings of the National*  
530 *Academy of Sciences* **107** 19368–73 Online: <https://www.pnas.org/content/107/45/19368>

531 Obermeier W A, Lehnert L W, Kammann C I, Müller C, Grünhage L, Luterbacher J, Erbs M, Moser  
532 G, Seibert R, Yuan N and Bendix J 2017 Reduced CO<sub>2</sub> fertilization effect in temperate C<sub>3</sub> grasslands  
533 under more extreme weather conditions *Nature Climate Change* **7** 137–41 Online:  
534 <https://www.nature.com/articles/nclimate3191>

535 Osher L J, Matson P A and Amundson R 2003 Effect of land use change on soil carbon in Hawaii  
536 *Biogeochemistry* **65** 213–32 Online: <https://doi.org/10.1023/A:1026048612540>

537 Perroy R L, Melrose J and Cares S 2016 The evolving agricultural landscape of post-plantation  
538 Hawai‘i *Applied Geography* **76** 154–62 Online:  
539 <http://linkinghub.elsevier.com/retrieve/pii/S0143622816304490>

Prestele R, Arneth A, Bondeau A, Noblet-Ducoudré N de, Pugh T A M, Sitch S, Stehfest E and Verburg P H 2017 Current challenges of implementing anthropogenic land-use and land-cover change in models contributing to climate change assessments *Earth System Dynamics* **8** 369–86 Online: <https://esd.copernicus.org/articles/8/369/2017/>

R Core Team 2019 *R: A language and environment for statistical computing* (Vienna, Austria: R Foundation for Statistical Computing) Online: <https://www.R-project.org/>

Reich P B, Hobbie S E, Lee T, Ellsworth D S, West J B, Tilman D, Knops J M H, Naeem S and Trost J 2006 Nitrogen limitation constrains sustainability of ecosystem response to CO<sub>2</sub> *Nature* **440** 922–5 Online: <http://www.nature.com/articles/nature04486>

Schuur E A 2003 Productivity and global climate revisited: The sensitivity of tropical forest growth to precipitation *Ecology* **84** 1165–70

Selmants P C, Giardina C P, Jacobi J D and Zhu Z 2017 *Baseline and projected future carbon storage and carbon fluxes in ecosystems of Hawai‘i* (U.S. Geological Survey) Online: <https://doi.org/10.3133/pp1834>

Sleeter B M, Liu J, Daniel C, Rayfield B, Sherba J, Hawbaker T J, Zhiliang Zhu, Selmants P C and Loveland T R 2018 Effects of contemporary land-use and land-cover change on the carbon balance of terrestrial ecosystems in the United States *Environ. Res. Lett.* **13** 045006 Online: <http://stacks.iop.org/1748-9326/13/i=4/a=045006>

Sleeter B M, Marvin D C, Cameron D R, Selmants P C, Westerling A L, Kreitler J, Daniel C J, Liu J and Wilson T S 2019 Effects of 21st-century climate, land use, and disturbances on ecosystem carbon balance in California *Global Change Biology* **25** 3334–53

Sleeter B M, Wilson T S, Sharygin E and Sherba J T 2017 Future scenarios of land change based on empirical data and demographic trends *Earth’s Future* **5** 1068–83 Online: <http://doi.wiley.com/10.1002/2017EF000560>

Soil Survey Staff 2016 *Soil Survey Geographic (SSURGO) Database, Natural Resources*

565 *Conservation Service, United States Department of Agriculture* (Natural Resources Conservation  
 566 Service, USDA) Online: <https://sdmdataaccess.sc.egov.usda.gov>.

567 State of Hawai‘i 2019 *Hawaii Greenhouse Gas Emissions Report for 2015* (Hawaii State  
 568 Department of Health, Clean Air Branch) Online:  
 569 [https://health.hawaii.gov/cab/files/2019/02/2015-Inventory\\_Final-Report\\_January-2019-004-1.pdf](https://health.hawaii.gov/cab/files/2019/02/2015-Inventory_Final-Report_January-2019-004-1.pdf)

570 Suryanata K 2009 Diversified Agriculture, Land Use, and Agrofood Networks in Hawaii *Economic*  
 571 *Geography* **78** 71–86 Online: <http://doi.wiley.com/10.1111/j.1944-8287.2002.tb00176.x>

572 Terrer C, Jackson R B, Prentice I C, Keenan T F, Kaiser C, Vicca S, Fisher J B, Reich P B, Stocker B  
 573 D, Hungate B A, Peñuelas J, McCallum I, Soudzilovskaia N A, Cernusak L A, Talhelm A F, Sundert  
 574 K V, Piao S, Newton P C D, Hovenden M J, Blumenthal D M, Liu Y Y, Müller C, Winter K, Field C  
 575 B, Viechtbauer W, Lissa C J V, Hoosbeek M R, Watanabe M, Koike T, Leshyk V O, Polley H W and  
 576 Franklin O 2019 Nitrogen and phosphorus constrain the CO<sub>2</sub> fertilization of global plant biomass  
 577 *Nature Climate Change* 1–6 Online: <https://www.nature.com/articles/s41558-019-0545-2>

578 Timm O E 2017 Future warming rates over the Hawaiian Islands based on elevation-dependent  
 579 scaling factors *International Journal of Climatology* **37** 1093–104 Online:  
 580 <https://rmets.onlinelibrary.wiley.com/doi/abs/10.1002/joc.5065>

581 Timm O E, Giambelluca T W and Diaz H F 2015 Statistical downscaling of rainfall changes in  
 582 Hawai‘i based on the CMIP5 global model projections *Journal of Geophysical Research:*  
 583 *Atmospheres* **120** 92–112

584 Trauernicht C 2019 Vegetation—Rainfall interactions reveal how climate variability and climate  
 585 change alter spatial patterns of wildland fire probability on Big Island, Hawaii *Science of The Total*  
 586 *Environment* **650** 459–69 Online:  
 587 <http://www.sciencedirect.com/science/article/pii/S0048969718333187>

588 Trauernicht C, Pickett E, Giardina C P, Litton C M, Cordell S and Beavers A 2015 The contemporary  
 589 scale and context of wildfire in Hawai‘i *Pacific Science* **69** 427–44

Walker A P, De Kauwe M G, Medlyn B E, Zaehle S, Iversen C M, Asao S, Guenet B, Harper A,  
Hickler T, Hungate B A, Jain A K, Luo Y, Lu X, Lu M, Luus K, Megonigal J P, Oren R, Ryan E, Shu  
S, Talhelm A, Wang Y-P, Warren J M, Werner C, Xia J, Yang B, Zak D R and Norby R J 2019 Decadal  
biomass increment in early secondary succession woody ecosystems is increased by CO<sub>2</sub> enrichment  
*Nature Communications* **10** Online: <http://www.nature.com/articles/s41467-019-08348-1>

Walker A P, Kauwe M G D, Bastos A, Belmecheri S, Georgiou K, Keeling R, McMahon S M, Medlyn  
B E, Moore D J, Norby R J, Zaehle S, Anderson-Teixeira K J, Battipaglia G, Brien R J, Cabugao K  
G, Cailleret M, Campbell E, Canadell J, Ciais P, Craig M E, Ellsworth D, Farquhar G, Fatichi S,  
Fisher J B, Frank D, Graven H, Gu L, Haverd V, Heilman K, Heimann M, Hungate B A, Iversen C M,  
Joos F, Jiang M, Keenan T F, Knauer J, Körner C, Leshyk V O, Leuzinger S, Liu Y, MacBean N,  
Malhi Y, McVicar T, Penuelas J, Pongratz J, Powell A S, Riutta T, Sabot M E, Schleucher J, Sitch S,  
Smith W K, Sulman B, Taylor B, Terrer C, Torn M S, Treseder K, Trugman A T, Trumbore S E,  
Mantgem P J van, Voelker S L, Whelan M E and Zuidema P A 2020 Integrating the evidence for a  
terrestrial carbon sink caused by increasing atmospheric CO<sub>2</sub> *New Phytologist* **229** 2413–45

Wickham H, Averick M, Bryan J, Chang W, McGowan L D, François R, Golemund G, Hayes A,  
Henry L, Hester J, Kuhn M, Pedersen T L, Miller E, Bache S M, Müller K, Ooms J, Robinson D,  
Seidel D P, Spinu V, Takahashi K, Vaughan D, Wilke C, Woo K and Yutani H 2019 Welcome to the  
tidyverse *Journal of Open Source Software* **4** 1686

Yang J, Medlyn B E, Kauwe M G D, Duursma R A, Jiang M, Kumarathunge D, Crous K Y, Gimeno  
T E, Wujeska-Klaus A and Ellsworth D S 2020 Low sensitivity of gross primary production to  
elevated CO<sub>2</sub> in a mature eucalypt woodland *Biogeosciences* **17** 265–79 Online:  
<https://www.biogeosciences.net/17/265/2020/>

Zhu Z, Piao S, Myneni R B, Huang M, Zeng Z, Canadell J G, Ciais P, Sitch S, Friedlingstein P,  
Arnell A, Cao C, Cheng L, Kato E, Koven C, Li Y, Lian X, Liu Y, Liu R, Mao J, Pan Y, Peng S,  
Peñuelas J, Poulter B, Pugh T A M, Stocker B D, Viovy N, Wang X, Wang Y, Xiao Z, Yang H, Zaehle  
S and Zeng N 2016 Greening of the Earth and its drivers *Nature Climate Change* **6** 791–5 Online:

