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# Moisture content variation of ground vegetation fuels in boreal mesic and sub-xeric mineral soil forests in Finland

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1	Moisture content variation of ground vegetation fuels in boreal mesic and sub-xeric
2	mineral soil forests in Finland
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**Abstract:** Forest fire risk in Finland is estimated by the Finnish Forest Fire Index (FFI), which predicts the fuel moisture content (FMC) of the forest floor. We studied the FMC variation of four typical ground vegetation fuels, Pleurozium schreberi, Hylocomium splendens, Dicranum spp., and Cladonia spp., and raw humus in mature and recently clear-cut stands. Of these, six were sub-xeric Pinus sylvestris stands, and six mesic Picea abies stands. We analyzed FFI's ability to predict FMC and compared it with the widely applied Canadian Fire Weather Index (FWI). We found that in addition to stand characteristics ground layer FMC was highly dependent on the species so that *Dicranum* was the moistest, and *Cladonia* the driest. In the humus layer, the differences among species were small. Overall, the FWI was a slightly better predictor of FMC than the FFI. While the FFI predicted ground layer FMC generally well, the shape of the relationship varied among the four species. The use of auxiliary variables thus has potential in improving predictions of ignitions and forest fire risk. Knowledge of FMC variation could also benefit planning and timing of prescribed burnings. **Brief summary**: The studied four moss and lichen species were found to dry at different rates, thus having different ignition potential and fire risk. Stand type, and particularly developmental stage also affected the drying rates. The fire risk indices could be improved by using these variables, which could benefit fire prevention. **Keywords:** fire risk, forest fire index, forest type, prescribed burning, Norway spruce, Scots pine, stand structure Running head: Variation in moisture content of ground vegetation fuels

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#### Introduction

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In Finland, forest fires declined during the last century. This decline was particularly steep during the latter half of the century. The average annual burned area in 1950s was about 5,700 ha and in the 1970s it had declined to approximately 700 ha (Yearbook of Forest Statistics 1990-1991 (1992). In recent decades, the average annual burned area has varied between 200 and 800 ha, only occasionally exceeding 1,000 ha. The average size of an individual fire is currently about 0.4 ha (Finnish Statistical Yearbook of Forestry 2014). The climatological fire risk in Finland was relatively stable during the last century (Mäkelä et al. 2012), so the decline in fire occurrence is explained by other factors, such as efficiency in fire detection and suppression, and changes in ignition sources, stand structure, forest fragmentation, and vegetation (Päätalo 1998; Wallenius 2011). This is also supported by the difference between the fire regimes of Finland and neighbouring Sweden, where the annual burned area has been higher and large fires frequent (Lindberg et al. 2020). Although forest fires do not currently form a major risk to society or property in Finland, they still employ rescue services leading to a need to improve forest fire risk assessment methods. This is partially due to the fact, that although the burned area has been low, the annual number of fires has been about 1,300 in the 21st century (Finnish Statistical Yearbook of Forestry 2014). Thus, the smallsized but frequent forest fires burden regional rescue services and local fire brigades during the forest fire season. Several studies have also predicted that the general forest fire risk in Finland (Kilpeläinen et al. 2010; Lehtonen et al. 2014; Mäkelä et al. 2014) and the risk for large fires (Lehtonen et al. 2016) will increase in the 21st century. One way to improve the preparedness of rescue services is to improve the ability to predict potential fire hazard days. The fuel moisture content (FMC) of different fuels is one of the key factors when estimating fire risk. FMC is used to predict flammability, and it is also a factor in models predicting fire intensity and fire spread rate. Most forest fire indices are meteorological and use various weather data to compute indices for assessing fire risk (San-Miguel-Ayanz et al. 2003).

73 Currently, the most widely used fire index system is the Canadian Forest Fire Weather Index System 74 (CFFWIS), which was initially designed for the Canadian boreal forest. Since being published in 1970 75 (Van Wagner 1987), it has gradually been adopted in many parts of the world, including different 76 vegetation zones and fuel types (Dimitrakopoulos et al. 2011). The FMC estimation in CFFWIS is 77 divided into three moisture codes: Fine Fuel Moisture Code (FFMC), Duff Moisture Code (DMC) and 78 Drought Code (DC) (Van Wagner 1987). These moisture codes are calculated daily based on air 79 temperature, relative humidity (not in DC), wind speed (only FFMC), and rainfall. Two spread 80 indices are then estimated: initial spread index using wind and FFMC and build-up index combining 81 DMC and DC. The spread indices are then combined to determine the Fire Weather Index (FWI) (Van 82 Wagner 1987). 83 CFFWIS has proven suitable in forests with a flammable duff layer typically consisting of a humus 84 layer and moss cover like, for instance, the black spruce (Picea mariana) (Mill.) Britton, Sterns & 85 Poggenburg) forests in boreal Northern America (e.g. Ziel et al. 2020). Fennoscandian coniferous 86 forests have a similar type of duff structure, and CFFWIS has generally been found to work well there 87 (Granström and Schimmel 1998; Tanskanen et al. 2005). 88 Despite the increasing use of CFFWIS, national fire indices are still commonly used in many 89 countries. In Finland, the forest fire risk is estimated and predicted by the Finnish Forest Fire Index 90 (FFI). FFI was constructed in 1996 to replace the former fire index, which was based merely on 91 statistical correlations between weather variables and the occurrence of fires (Heikinheimo et al. 92 1998). In 1996, Sweden started to use CFFWIS as a national forest fire index system (Sjöström et al. 93 2019), but the Finnish Meteorological Institute (FMI) decided to develop its own index, partly 94 because CFFWIS was considered unnecessarily complicated with its hierarchical structure, and 95 because it was lacking solar radiation as an explaining variable (Heikinheimo et al. 1998). 96 FFI is based on empirical relationships between weather data and the volumetric moisture content of a 97 6-cm thick layer of forest floor. In short (see Supplement 1 and Vajda et al. (2014) for details), air 98 temperature values are obtained from the ground weather station network and spatially interpolated to 99 a 10 km ×10 km grid using the kriging method (Venäläinen and Heikinheimo 2003). Evaporation is

estimated based on this interpolated data and weather prediction models, and the precipitation is received from weather radars (Venäläinen and Heikinheimo 2003; Vajda *et al.* 2014). The index is a continuous variable calibrated to vary from 1.0 to 6.0, 6.0 being the driest. The index has been assigned a threshold value of 4.0, at which point it predicts a volumetric moisture content under 20%. When the index exceeds this threshold, a forest fire warning is announced in public media, which forbids the lighting of open fires. It must be noted that the FFI uses volumetric moisture content values based on non-destructive monitoring of fuels and thus they are not directly comparable with gravimetric moisture content values.

In addition to its role in wildfire, FMC plays an important role in prescribed burnings, used in Finland as a silvicultural tool and nowadays also for ecological restoration and management for biodiversity. Because of this, the scope of prescribed burnings in Finland has widened in recent years to a more diverse set of burnings with different ecological aims such as burnings of retention trees, restoration burnings in nature conservation areas and management burnings of sun-exposed and xeric habitats (for details see Lindberg *et al* 2020). The various aims also set diverse targets for fire impact and depth. However, despite the recognized importance of fire for restoration, the overall area of prescribed burns has declined in recent decades (Lindberg *et al*. 2020).

FMC is one of the most significant factors determining the potential days of prescribed burnings and intended burning depth (Sandberg 1980; Ferguson *et al.* 2002; Hille and den Ouden 2005; Hille and Stephens 2005). Because of different ecological aims, understanding how FMC develops in various fuels and their effect on fire impact and burning result is necessary. As an example, in silvicultural burnings and burnings on barren habitats, the aim is to decrease the organic layer, which requires a sufficiently low FMC. If the moisture of the ground layer and in some cases raw humus is too high, the burning effects are not fully achieved. In restoration burnings, more various moisture conditions are possible, since more diverse burning results are accepted (Lindberg *et al.* 2020).

Boreal ground layer species differ in their structure and growth form which affects their water-holding capacity (Peterson and Mayo 1975; Busby and Whitfield 1978; Pech 1989). The aim of this study was to determine the FMC variation of dominant forest floor mosses and lichens and raw humus in different stands of the two most common forest types in Southern Finland. We analyzed how the moisture content of selected species varied as a function of FFI, and we compared the ability of FFI and FWI to predict the FMC of selected fuel materials.

We hypothesize that as clear-cut areas and pine-dominated sub-xeric stands receive more radiation

and are more exposed to the drying effect of wind: i) ground vegetation fuels dry faster in clear-cut areas as compared to closed-canopy forests, ii) fuels in pine-dominated forests dry faster than in spruce-dominated forests, iii) varying water holding capacity of studied materials explains the possible differences in their FMC behavior and potential days of ignition.

# Materials and methods

Study area

The study area is located in Southern Finland in the Evo State Forest (Fig. 1) belonging to the southern boreal vegetation zone (Ahti *et al.* 1968). The elevation of the study area varies between 100-190 meters a.s.l., mean annual temperature in the region is +3.1°C, the average annual precipitation is 670 mm, and the growing season 160 days (Juvakka *et al.* 1995). The bedrock is mostly orogenic granitoid covered by a thick, stony morainic layer, but glacier sedimented areas such as deltas, sandur deltas and eskers with sand or gravel are also common (Okko 1972). Of the sampled stands, the sub-xeric stands were mostly located in sedimented, sandy soils and mesic stands on sandy or fine sandy moraines (Fig. 1).

# Figure 1

Experimental design and sampling

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Nearly 90% of Finnish forests are managed commercially (Finnish Statistical Yearbook of Forestry 2014). The management is typically done relatively uniformly, including artificial regeneration, 2-4 low thinnings, and clear-cutting with less than 3% retention of tree volume (Finnish Forestry, Practice and Management 2011, Kuuluvainen et al. 2019). The stands are thus evenly aged, relatively sparsely stocked and most often dominated by Norway spruce (Picea abies L.) H. Karst and Scots pine (Pinus sylvestris L.) The most common forest site types on mineral soils in Finland are mesic forests (Myrtillus-type), which cover 52% and sub-xeric forests (Vaccinium-type), which cover 26% of forests (Finnish Statistical Yearbook of Forestry 2014). Both forest types in their later successional stages are characterized by dwarf shrubs bilberry (Vaccinium myrtillus L.), lingonberry (Vaccinium vitis-idaea L.) and common heather (Calluna vulgaris L. (Hull)). In sub-xeric forests V. vitis-idaea and Calluna are dominant, and in mesic forests V. myrtillus is dominant and Calluna practically absent. Managed conifer-dominated mesic and sub-xeric forests on mineral soils typically have an easily distinguishable raw humus layer with a typical thickness of 3-5 cm in Southern Finland (Tamminen 1991). In these forests, moss and lichen dominated ground vegetation is the most common and the most important flammable fuel bed, where the majority of forest fires ignite and spread (Schimmel and Granström 1997; Tanskanen et al. 2005). A continuous moss carpet is typical in later successional stages of coniferous forests whereas in young successional stages it is less abundant, thus decreasing fire risk (Schimmel and Granström 1997). Yet, recent clear-cuts where the moss carpet still exists and herbs and graminoids have not yet colonized the areas are flammable similar to the mature forests. A recent study showed that a significant number of forest fires in Sweden are started in clear-cuts as the sparks produced by forest machines are an important source of ignitions (Sjöström et al. 2019). The raw humus layer is also potentially flammable, and the targets and success of prescribed burnings are often estimated by burning depth, which indicates the decrease of moss and raw humus layer.

The feather moss (*Pleurozium schreberi*) (Brid) Mitt. is the most abundant moss species with a coverage of approximately 30% in mesic and 35% in sub-xeric forests. (Mäkipää 2000a). Fork mosses (*Dicranum* spp., *D.polysetum Sw.* and *D.scoparium* Hedw. being the most dominant) cover about 10% in both mesic and sub-xeric types (Mäkipää 2000b), whereas stairstep moss (*Hylocomium splendens*) (Hedw.) is clearly more abundant in mesic types with a share over 10% but in sub-xeric types only 3% (Mäkipää 2000c). Reindeer lichens (*Cladonia* spp) are practically absent in mesic forests but patchy with an average share of 5% in sub-xeric forests (Nousiainen 2000). *Cladonias* abundance increases significantly in xeric and barren forests, which are less common (pooled share 4%) and are concentrated in Northern Finland (Finnish Statistical Yearbook of Forestry 2014).

Twelve forest stands from the study area were chosen, consisting of four different stand types and three replicates from each. The stand types were: 1. Sub-xeric, mature, *Pinus* dominated stand. 2. Sub-xeric, clear-cut area. 3. Mesic, mature, *Picea* dominated stand. 4. Mesic, open, clear-cut area (Fig. 1, Table 1). The age and standing stock of a stand is referred to as the developmental stage (either clear-cut or mature) and the combination of forest type and dominant tree species as stand type (either sub-xeric/*Pinus* or mesic/*Picea*) (Table 1).

#### Table 1

We selected individual stands from the forest planning databases of the study area, according to the following criteria: mature stands had to be over 70 years of age and be either *Pinus-* or *Picea-* dominated, with at least 70% dominance (Table 1). The clear-cut stands had to be harvested during the previous winter with no mechanical scarification. All stands had a distinctive raw-humus layer and a characteristic continuous moss layer with patches of *Cladonia* in sub-xeric stands. The growing stock and structure of the mature stands represented typical Finnish managed forest stands with an evenly aged structure and minor understory.

From each stand, samples of three dominant moss and/or lichen species were collected on 17 days during summer 2003. The days were chosen using FFI values received from the Finnish Meteorological Institute, so that they would cover different weather and drying conditions (Fig. 2). Sampling was focused especially on dry and drying periods whereas, during constant wet periods (which covered the most part of the sampling period), it was not carried out.

We sampled each stand in the afternoons of the sampling days. On each occasion, five randomly chosen samples consisting of moss or lichen and raw humus were taken with humus auger with a diameter of 5.8 cm, height of 10 cm and volume of 264 cm<sup>3.</sup> The samples were taken from a 300 m<sup>2</sup> circular sample plot and were located at least 30 m from the stand edge. In mesic stands, the sampled species were: *Pleurozium.schreberi*, *Dicranum spp* (*D. polysetum* being the most abundant) and *Hylocomium splendens.*, and on sub-xeric stands *Pleurozium*, *Dicranum* and *Cladonia*. (C. *rangiferina* (L.) Weber ex F.H. Wigg. being the most abundant). The third replication of mesic clearcut area had an insufficient cover of *Hylocomium*, so only *Pleurozium* and *Dicranum* were sampled.

Each sample was then divided into two layers: surface and raw humus. Five subsamples of each layer were pooled into one sample representing the average from that stand. Thus, each sampled stand had six combined samples: a combined sample of each of the three surface species, and three combined samples from raw humus under each species. The collective samples were preserved during transportation in air-tight plastic bags. The fresh-weighing and drying was done directly after transportation with a minimum of 18 hours of oven-drying at 105 °C. Sufficient drying time was ensured by experimental dryings before actual sampling. After drying, the samples were weighed and the dry-weight FMC was determined.

Data analysis

The noon values of FFI and FWI were used in analysis. The FWI values were received from FMI and calculated according to Van Wagner and Pickett (1985) using weather data from the nearest

meteorological station located approximately 4 km south-west of the center of the study area. The wind values came from the nearest available station, about 25 km north-east of the study area. We modeled FMC separately for each species, and the surface and raw humus layers, as a function of FFI, stand type, and the development class. Preliminary analyses showed that the shape of the relationship between FMC and FFI varied among the species and was often non-linear. We thus used generalized additive modeling (e.g. Zuur et al. 2009), in which FMC was predicted as a smooth function of FFI. For the strictly positive data (FMC), we used a Gaussian error distribution and log-link function, and the smoothers were allowed to vary as a function of developmental stage. To avoid problems with overfitting and to ensure biologically realistic model behavior, we used monotonically decreasing Psplines as smoothers and limited their flexibility (number of knots in the splines k = 4). To compare the performance of FFI to the more widely used FWI, we then repeated the analyses, using FWI as the continuous predictor in place of FFI. The models were compared using pseudo-R<sup>2</sup> values for both (models with FFI and FWI). For model validation (sensu Zuur et al. 2009), we visually inspected the residuals as a function of FMC and each predictor, as well as day of year to ensure there were no temporal patterns in the residuals (Supplement 2). All models were fitted using R (R Core Team 2019) and the package scam (Pva 2018).

The observed and predicted days of ignition of surface fuels in different stands were analyzed by calculating a probability using FMC frequencies. In Fennoscandia, the FMC values for moisture content of extinction have been estimated to range from 25 to 35 % (Granström and Schimmel 1998; Tanskanen *et al.* 2005). We used the lower limit since it was considered a more suitable estimate for the timing of prescribed burnings, which was justified because in prescribed burnings one aim is to decrease organic material and ensure a sufficient ecological impact (Lindberg *et al.* 2020). The frequencies over threshold value were compared to all the values of the examined variables or their combinations. Thus, if for instance *Pleurozium* in sub-xeric clear-cuts had 21 observations under a 25% threshold value of FMC, these 21 were compared to all 51 observations in sub-xeric clear-cuts resulting in a probability ratio of 41% (21/51) X 100=41%).

#### Results

During the measurement period, the FMC of surface layer varied between 3% and 300% (Fig. 2). The overall patterns in how the moisture conditions changed during the summer were similar among the species, sites and site types, but the levels differed greatly among species and sites (Fig. 2). It should be noted that the weather conditions during summer 2003 were relatively variable with no long dry periods. This is visible in the distribution of the FFI values, where the highest values (4-6) are missing, which means that the driest circumstances did not occur during sampling (Fig. 2).

### Figure 2

Of the species, *Dicranum* was generally the moistest and *Cladonia* the driest, whereas *Pleurozium* and *Hylocomium* were between the two. When modeling the FMC as a function of FFI, stand type and developmental stage, several patterns were visible in the surface layer. First, there were clear differences between species in the shape of the relationship between FMC and FFI. *Pleurozium*, *Hylocomium* and *Cladonia* had a tendency for a steadier decline compared to *Dicranum*, which retained moisture up to a higher FFI before declining more rapidly in moisture content (Fig. 3). It is noteworthy that, despite the quick decline at higher FFI values for *Dicranum*, the predicted moisture content in mature stands stayed above the 25-35% level, considered a threshold of ignition (Fig. 3). Stand type was not a significant predictor for any of the species in the surface layer (Table 2). The effect of the developmental stage was significant in the smoother terms only (Table 3, Fig. 4). Plot-level random effects were significant only for *Pleurozium*.

For the raw humus layer, the relationship between FFI and fuel moisture content were close to linear in most cases, and the differences in the smoothers were clearly smaller compared to the surface layer (Table 2). Similarly, the effect of stand type was different from the surface layer so that, for both *Pleurozium* and *Dicranum*, the sub-xeric sites were drier than the mesic sites (Table 3). Plot-level

289 random effects were significant only for *Cladonia*. The raw humus variation among the stand types 290 was lower but clear among the developmental stages and, in all stands, well above the 25-35% level. 291 292 Table 2 293 Table 3 294 Figure 3 295 Figure 4 296 297 FWI predicted the FMC of surface layers slightly better than FFI (Table 4). Both models predicted the 298 FMCs of Pleurozium and Hylocomium better than Dicranum and Cladonia. In raw humus, the 299 prediction ability was clearly lower, and FWI and FFI performed practically equally (Table 4). The 300 predicted moisture variation curves as a function of FWI are shown in Supplement 3. 301 302 Table 4 303 304 The potential fire hazard days (i.e., days during which the FMC values were under 25%) were highest 305 in Cladonia and lowest in Dicranum (Table 5). Clear-cut areas and sub-xeric pine stands had more 306 fire hazard days than mature stands and mesic spruce-stands. The predicted fire hazard days by FFI 307 formed 6% of sampled days, whereas the observed FMCs of > 25% during the same sampled days 308 was 28%. 309 310 Table 5 311 312 **Discussion** 313 314 Our results showed that the composition of ground floor vegetation has an effect on the flammability 315 of the surface layer in Fennoscandian boreal forests, and how it varies during the fire season. This 316 flammability was further modulated by the effect of stand growing stock along the lines shown in

earlier studies (Granström and Schimmel 1998; Tanskanen *et al.* 2005; Tanskanen *et al.* 2006). The differences among species and developmental stages in how the surface layer moisture varied were prominent. As an example, *Dicranum* in mature stands retained a moisture content well above the 25-35% threshold of the FFI value of 4 (the threshold for public warning), whereas *Cladonia* was close to the flammability threshold throughout the range of FFI values included in the sample here.

The development of moisture content between the surface layer and raw humus was clear. Rain usually affects the surface layer saturating it rapidly. The raw humus layer receives some moisture, especially in heavier rains, but dries slowly. However, during longer dry periods, the surface layer and raw humus dry more thoroughly. Long drought periods did not occur during the sampling period so the FMCs in such circumstances could not be compared.

The FMC variation of surface and raw humus layers was great, especially in higher FMCs, which can be due to several reasons. The same FFI values estimated for a  $10 \text{ km} \times 10 \text{ km}$  square were used for all stands, so differences in rainfall between stands may have occurred due to local showers. The FMCs were determined layer by layer, which overlooks moisture variation within layers. It is known that the moisture gradient within layers is steep (Vasander and Lindholm 1985), so the upper parts of the surface layer could be clearly drier than the FMCs observed in this study.

When considering differences among the species in the surface layer, *Dicranum* was consistently the moistest, and *Cladonia* the driest. *Pleurozium* and *Hylocomium* were between these two and showed a relatively similar moisture behavior as presented by Busby and Whitefield (1978). The higher FMCs and slower drying curve of *Dicranum* is probably due to its dense tomentum-covered structure (Peterson and Mayo 1975), which leads to a higher moisture retaining capacity. As reported previously (Mutch and Gastineau 1970; Granström and Schimmel 1998), *Cladonia* was the driest surface fuel. This is explained by its gelatinous thallus, loose structure and high surface-to-volume ratio resulting in extreme moisture behavior (Heatwole 1966; Pech 1989, 1991).

FMC varied among stand types. The results of the FMC variation of the surface layer are in accordance with previous studies in which the differences between stands correlate with their ground vegetation flammability (Tanskanen et al. 2006). Using 30% threshold values for the FMC of moss layer, Tanskanen et al. (2006) reported two times more potential days of ignition in open than in mature areas, and in *Pinus*-dominated stands two to three times higher than in *Picea*-dominated stands. In our study, the differences between clear-cut and mature developmental stages were clear, but the impact of site type and the associated dominant tree species was smaller. Comparison between the Finnish FFI and Canadian FWI showed that FWI was consistently a better predictor for the moisture content of the surface layer fuels, irrespective of the species. For the raw humus layer, the two indices performed almost identically. The better performance of FWI for surface fuels was similar to what Tanskanen et al. (2005) reported. Thus the CFFWIS could well be used in Finland. Our results support the conclusions of Tanskanen et al. (2005) and Vajda et al. (2014) suggesting that FFI could be improved by using forest stand variables. Such parameters as developmental stage and dominant tree species could likely improve the FFIs prediction ability significantly, which could eventually help practical fire suppression activities by better anticipation and preparation. Fire history studies in Fennoscandia have reported great variation in fire cycles. The shorter cycles have been typical in *Pinus*-dominated forests, especially in south- and middle boreal forests (e.g., Lehtonen and Kolström 2000), whereas in more northern and *Picea*-dominated forests, the cycle has been longer (e.g. Wallenius 2004). The differences have been explained by meteorological factors, dominant tree species, vegetation, fire suppression and general human influence (Wallenius 2004, 2011). According to our results, the differences in reported fire cycles could be partially explained by dominant tree species and changes in ground floor vegetation, especially in lichen-bryophyte ratio. For example, the abundance of *Cladonia* has substantially decreased in recent decades in Finland (Nousiainen 2000; Mäkipää and Heikkinen 2003; Tonteri et al. 2013). At the same time, a notable increase in the abundance of *Dicranum* has been documented especially in Northern Finland (Mäkipää 2000b). It is possible that reduction in the cover of fast-drying *Cladonia* and increase in the

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cover of slowly-drying *Dicranum* has partially reduced forest fire risk particularly in Northern Finland.

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In our study, the large variation of FMC in different stands and ground floor fuel materials show that potential days for prescribed burnings also have a large variation, especially when the variable ecological targets of burnings are taken into account. An often presented rule of thumb in guidelines for prescribed burnings is that the forest fire warning in Finland (FFI value 4) could be considered as a general threshold for successful burnings (Lemberg and Puttonen 2002). According to our results, this assumption is too simplistic, since suitable days for prescribed burning also seem to occur with lower FFI values. Yet it should be noted that the selected level of FMC 25% should be interpreted as a level where burning of studied surface layer fuels is possible. Thus, the various goals of prescribed burnings should be taken into account when suitable burning conditions are determined. For instance, in most restoration burnings no special burning depth is targeted as it is in silvicultural burnings. On the other hand, denser stands where restoration burnings are performed dry slower than regeneration areas. Also, if the aim is also to burn the humus layer, long drought periods are needed since the FMC values of raw humus did not reach the ignition threshold limits within the range of the FFI values we analyzed. Thus, a stand-specific monitoring of surface fuel and raw humus layer is recommended so that all potential burning days – whose small number often functions as a limiting factor – could be utilized more effectively, and the targeted impacts of burnings could be ensured.

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# Conclusions

Our results show that the different ground vegetation fuels differ in their moisture variation and ignition potential. Developmental stage and stand type of the forest affect the moisture variation of the studied fuels. Canadian FWI predicted the FMC of surface layer better than Finnish FFI, so it could be used in Finland. We conclude that, by using additional predictor variables, the ability of forest fire indices to predict fuel moisture could be improved. This could benefit forest fire prevention by enhancing early warning systems and by developing a GIS-based system providing online stand-wise

399	FMC estimates of surface fuels, which could be utilized in practical firefighting as well as in
400	prescribed burning.
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402	Abbreviations
403	CFFWIS Canadian Forest Fire Weather Index System
404	DC Drought Code
405	DMC Duff Moisture Code
406	FFI Finnish Forest Fire Index
407	FFMC Fine Fuel Moisture Code
408	FMC Fuel moisture content
409	FMI Finnish Meteorological Institute
410	FWI Canadian Fire Weather Index
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21	References
28	
29	Ahti T, Hamet-Ahti L, Jalas J (1968) Vegetation zones and their sections in
30	northwestern Europe. Annales Botanici Fennici 5, 169–211.
31	Busby JR, Whitfield, DWA (1978) Water potential, water content, and net assimilation of some
32	boreal forest mosses. Canadian Journal of Botany 56, 1551–58. https://doi.org/10.1139/b78-
33	184
34	Dimitrakopoulos AP, Bemmerzouk AM, Mitsopoulos ID (2011) Evaluation of the Canadian fire
35	weather index system in an eastern Mediterranean environment. Meteorological. Applications
36	18, 83–93. https://doi.org/10.1002/met.214
37	Ferguson SA, Ruthford JE, McKay SJ, Wright D, Wright C, Ottmar R (2002) Measuring moisture
38	dynamics to predict fire severity in longleaf pine forests. International Journal of Wildland
39	Fire 11(4), 267–279. https://doi.org/10.1071/WF02010
40	Finnish Forestry - Practice and Management (2011) Metsäkustannus, Helsinki. 271 p.
41	Finnish statistical yearbook of forestry 2014 (2014) Finnish Forest Research Institute.
42	http://urn.fi/URN:ISBN:978-951-40-2506-8
43	Granström A, Schimmel J (1998) Utvärdering av det kanadensiska brandrisksystemet –
44	testbränningar och uttorkningsanalyser. (In Swedish with English abstract: Assessment of the
45	Canadian forest fire danger rating system for Swedish fuel conditions.) P21-244/98. (Rescue
46	Service: Karlstad, Sweden)
47	Heatwole H (1966) Moisture exchange between the atmosphere and some lichens of the genus
48	Cladonia. Mycologia 58, 148-156. Available at https://www.jstor.org/stable/3756996
49	[Verified 28 November 2020].
50	Heikinheimo M, Venäläinen A, Tourula T (1998) A soil moisture index for the assessment of
51	forest fire potential in the boreal zone. In Proceedings of the International Symposium on
52	Applied Agrometeorology and Agroclimatology (Volos, Greece), Office for Official
53	Publication of the European Commission (Luxembourg), NR Dalezios (ed), EUR 18328-
54	COST 77, 79, 711, 549– 555.

455	Hille MG, den Ouden J (2005) Fuel load, humus consumption and humus moisture dynamics in
456	Central European Scots pine stands. International Journal of Wildland Fire 14, 153-159
457	https://doi.org/10.1071/WF04026
458	Hille MG, Stephens S (2005) Mixed Conifer Forest Duff Consumption during Prescribed Fires:
459	Tree Crown Impacts. Forest Science 51(5), 417-424. Available at
460	https://nature.berkeley.edu/stephenslab/wp-content/uploads/2015/04/Hille-Stephens-duff-FS
461	[Verified 28 November 2020].
462	Juvakka M, Viinikainen J, Puputti I, Kuupakko S (1995) Vesijaon tutkimusalue, hoito- ja
463	käyttösuunnitelma 1994–2003. [Plan for the management and use of forests in Vesijako
464	research area 1994–2003]. Metlan tutkimusmetsien julkaisusarja 5. Vantaa. 228 p. ISSN
465	1238-0830. (In Finnish).
466	Kilpeläinen A, Kellomäki S, Strandman H, Venäläinen A (2010) Climate change impacts on
467	forest fire potential in boreal conditions in Finland. Climatic Change 103, 383-398
468	https://doi.org/10.1007/s10584-009-9788-7
469	Kuuluvainen T, Lindberg H, Vanha-Majamaa I, Keto-Tokoi P, Punttila P (2019) Low-level
470	retention forestry, certification and biodiversity: case Finland. Ecological Processes 8, 47.
471	https://doi.org/10.1186/s13717-019-0198-0
472	Lehtonen H, Kolström T (2000) Forest fire history in Viena Karelia, Russia. Scandinavian
473	Journal of Forest Research 15, 585-590. https://doi.org/10.1080/02827580050216833
474	Lehtonen I, Ruosteenoja K, Venäläinen A, Gregow H (2014) The projected 21st century forest
475	fire risk in Finland under different greenhouse gas scenarios. Boreal Environment Research
476	<b>19</b> , 127-139. Available at:
477	https://www.researchgate.net/publication/285955800_The_projected_21st_century_forest-
478	fire_risk_in_Finland_under_different_greenhouse_gas_scenarios [Verified 28 November
479	2020].
480	Lehtonen I, Venäläinen A, Kämäräinen M, Peltola H, Gregow H (2016) Risk of large-scale fires
481	in boreal forests of Finland under changing climate. Natural Hazards and Earth System
482	Sciences 16, 239253. https://doi.org/10.5194/nhess-16-239-2016

483	Lemberg T, Puttonen P (2002) Kulottajan käsikirja. Metsälehti kustannus. Vammalan kirjapaino.
484	(Guide for prescribed burning. Textbook. In Finnish)
485	Lindberg H, Punttila P, Vanha-Majamaa I (2020) The challenge of combining variable retention
486	and prescribed burning in Finland. Ecological Processes 9, 4 (2020).
487	https://doi.org/10.1186/s13717-019-0207-3
488	Mäkelä HM, Laapas M, Venäläinen, A (2012) Long-term temporal changes in the occurrence of a
489	high forest fire danger in Finland. Natural Hazards and Earth System Sciences 12, 2591-2601
490	https://doi.org/10.5194/nhess-12-2591-2012
491	Mäkelä HM, Venäläinen A, Jylhä K, Lehtonen I, Gregow H (2014) Probabilistic
492	projections of climatological forest fire danger in Finland. Climate Research 60, 73-85.
493	Available at https://www.jstor.org/stable/24896175 [Verified 28 November 2020].
494	Mäkipää R (2000a) Pleurozium schreberi. In: Reinikainen A, Mäkipää R, Vanha-Majamaa I,
495	Hotanen J-P. (eds.) 2000. Kasvit muuttuvassa metsäluonnossa. [Summary in English:
496	Changes in the frequency and abundance of forest and mire plants in Finland since 1950].
497	Tammi, Jyväskylä. 384 p.
498	Mäkipää R (2000b) Dicranum. In: Reinikainen A, Mäkipää R, Vanha-Majamaa I, Hotanen J-P.
499	(eds.) 2000. Kasvit muuttuvassa metsäluonnossa. [Summary in English: Changes in the
500	frequency and abundance of forest and mire plants in Finland since 1950]. Tammi, Jyväskylä.
501	384 p.
502	Mäkipää R (2000c) Hylocomium splendensi. In: Reinikainen A, Mäkipää R, Vanha-Majamaa I,
503	Hotanen J-P. (eds.) 2000. Kasvit muuttuvassa metsäluonnossa. [Summary in English:
504	Changes in the frequency and abundance of forest and mire plants in Finland since 1950].
505	Tammi, Jyväskylä. 384 p.
506	Mäkipää R, Heikkinen J (2003) Large-scale changes in abundance of terricolous bryophytes
507	and macrolichens in Finland. Journal of Vegetation Science 14, 497–508.
508	https://doi.org/10.1111/j.1654-1103.2003.tb02176.x

509	Mutch, RW, Gastineau OW (1970) Timelag and equilibrium moisture content of reindeer lichen.
510	Res. Pap. INT-76. Ogden, UT: U.S. Department of Agriculture, Forest Service, Intermountain
511	Forest and Range Experiment Station. 8 p. https://doi.org/10.5962/bhl.title.68840
512	Nousiainen H (2000) Cladina In: Reinikainen A, Mäkipää R, Vanha-Majamaa I, Hotanen, J-P.
513	(eds.) 2000. Kasvit muuttuvassa metsäluonnossa. [Summary in English: Changes in the
514	frequency and abundance of forest and mire plants in Finland since 1950]. Tammi, Jyväskylä.
515	384 p.
516	Okko M (1972) Jäätikön häviämistapa Toisen Salpausselän vyöhykkeessä Lammilla.
517	Summary: Deglaciation in the Second Salpausselka ice-marginal belt at Lammi, South
518	Finland. Terra <b>84</b> (3), 115-123.
519	Päätalo ML (1998) Factors influencing occurrence and impacts of fires in Northern European
520	forests. Silva Fennica 32(2), 185-202. https://doi.org/10.14214/sf.695
521	Péch G (1989) A model to predict the moisture content of reindeer lichen. Forest Science 35,
522	1014-1028. https://doi.org/10.1093/forestscience/35.4.1137
523	Péch G (1991) Dew on reindeer lichen. Canadian Journal of Forest Research 21, 1415–1418.
524	https://doi.org/10.1139/x91-198
525	Peterson W, Mayo J (1975) Moisture stress and its effect on photosynthesis in <i>Dicranum</i>
526	polysetum. Canadian Journal of Botany 53, 2897–2900. Available at
527	https://fdocuments.in/document/moisture-stress-and-its-effect-on-photosynthesis-in-
528	dicranum-polysetum.html [Verified 28 November 2020].
529	Pya N (2018) scam: Shape Constrained Additive Models. R package version 1.2-3.
530	R Core Team (2019) R: A language and environment for statistical computing. Version 3.5. R
531	Foundation for Statistical Computing, Vienna, Austria.
532	Sandberg DV (1980) Duff reduction by prescribed burning in Douglas-fir. USDA Forest Service
533	Research Paper PNW-272, 18 p. https://doi.org/10.2737/PNW-RP-272
534	San-Miguel-Ayanz J, Carlson JD, Alexander M, Tolhurst K, Morgan G, Sneeuwjagt R, Dudley M
535	(2003) Current methods to assess fire danger potential. In Wildland Fire Danger Estimation
536	and Mapping. The Role of Remote Sensing Data, Chuvieco E (ed), World Scientific

537	Publishing: Singapore; 21–61. Available at
538	https://pdfs.semanticscholar.org/da16/f999aff0083cfee5820cfad43a8d6d1e4c41.pdf [Verified
539	28 November 2020].
540	Schimmel J, Granström A (1997) Fuel succession and fire behavior in the Swedish boreal forest.
541	Canadian Journal of Forest Research 27, 1207–1216. https://doi.org/10.1139/x97-072
542	Sjöström J, Plathner FV, Granström A (2019) Wildfire ignition from forestry machines in boreal
543	Sweden. International Journal of Wildland Fire 28(9), 666-677.
544	https://doi.org/10.1071/WF18229
545	Tamminen P (1991) Kangasmaan ravinnetunnusten ilmaiseminen ja viljavuuden alueellinen
546	vaihtelu. Summary: Expression of soil nutrient status and regional variation in soil fertility of
547	forested sites in southern Finland. Folia Forestalia 777. 40 p. http://urn.fi/URN:ISBN:951-
548	40-1170-8
549	Tanskanen H, Venäläinen A, Puttonen P, Granström A (2005) Impact of stand structure on
550	surface fire ignition potential in Picea abies and Pinus sylvestris forests in southern Finland.
551	Canadian Journal of Forest Research 35, 410-420. https://doi.org/10.1139/X04-188
552	Tanskanen H, Granström A, Venäläinen A, Puttonen P (2006) Moisture dynamics of moss
553	dominated surface fuel in relation to the structure of Picea abies and Pinus sylvestris stands.
554	Forest Ecology and Management. 226, 189-198. doi:10.1016/j.foreco.2006.01.048
555	Tonteri T, Salemaa M, Rautio P (2013) Changes of understorey vegetation in Finland in 1985-
556	2006. In: Merilä, P. & Jortikka, S. (eds.). Forest Condition Monitoring in Finland – National
557	report. The Finnish Forest Research Institute. [Online report].
558	http://urn.fi/URN:NBN:fi:metla-201305087583.
559	Vajda A, Venäläinen A, Suomi I, Junila P, Mäkelä HM (2014) Assessment of forest fire danger in
560	a boreal forest environment: description and evaluation of the operational system applied in
561	Finland. Meteorological Applications 21, 879-887. https://doi.org/10.1002/met.1425
562	Van Wagner CE (1987) Development and Structure of the Canadian Forest Fire Weather Index
563	System; Forestry Technical Report 35; Canadian Forestry Service: Ottawa, ON, Canada,

564	Volume 1, 48 p. Available at https://cfs.nrcan.gc.ca/publications?id=19927 [Verified 28
565	November 2020].
566	Van Wagner CE, Pickett TL (1985) Equations and FORTRAN program for the Canadian Forest
567	Fire Weather Index System. Canadian Forestry Service, Petawawa National Forestry Institute,
568	Chalk River, Ontario. Forestry Technical Report 33. 18 p. Available at
569	https://cfs.nrcan.gc.ca/publications?id=19973 [Verified 28 November 2020].
570	Vasander H, Lindholm T (1985) Fire intensities and surface temperatures during prescribed
571	burning. Silva Fennica 19(1), 1-15. https://doi.org/10.14214/sf.a15406
572	Venäläinen A, Heikinheimo M (2003) The Finnish forest fire index calculation system. In:
573	Zschau, J. & Kuppers, A. (ed.). Early Warning Systems for Natural Disaster Reduction.
574	Springer, 467 p.
575	Wallenius T (2004) Fire histories and tree ages in unmanaged boreal forests in Eastern
576	Fennoscandia and Onega peninsula. Academic Dissertation, June 2004. University of Helsinki,
577	Faculty of Biosciences, Department of Biological and Environmental Sciences and Faculty of
578	Agriculture and Forestry, Department of Forest Ecology. http://urn.fi/URN:ISBN:952-10-
579	1893-3
580	Wallenius T (2011) Major decline in fires in coniferous forests - reconstructing the phenomenon
581	and seeking for the cause. Silva Fennica 45, 139-155. https://doi.org/10.14214/sf.36
582	Yearbook of forest statistics 1990-1991 (1992) Finnish Forest Resource Institute.
583	http://urn.fi/URN:ISBN:951-40-1205-4
584	Ziel RH, Bieniek PA, Bhatt US, Strader H, Rupp, TS, York AA (2020) Comparison of Fire
585	Weather Indices with MODIS Fire Days for the Natural Regions of Alaska. Forests 11(5),
586	516; https://doi.org/10.3390/f11050516
587	Zuur A, Ieno EN, Walker N, Saveliev AA, Smith, GM (2009) Mixed Effects Models and
588	Extensions in Ecology with R. New York, NY. Springer.
589	
590	

# Tables and figure captions

Table 1. The sampled stands. In clear-cut areas the dominant tree species refers to species of the precut stand. Pine: *Pinus sylvestris*, spruce: *Picea abies*, birch: *Betula* spp.

Stand	Developmental stage	Stand type	Age, years	Average height, meters	Standing stem volume: cubic meters/hectare	Standing tree species percentages by volume (pine/spruce/birch)
SXC1	clear-cut	sub-xeric/pine	0	0	0	-
SXC2	clear-cut	sub-xeric/pine	0	0	0	-
SXC3	clear-cut	sub-xeric/pine	0	0	0	-
SXM1	mature	sub-xeric/pine	90	24	210	90/10/0
SXM2	mature	sub-xeric/pine	120	26	250	100
SXM3	mature	sub-xeric/pine	120	25	240	100
MC1	clear-cut	mesic/spruce	0	0	0	-
MC2	clear-cut	mesic/spruce	0	0	0	-
MC3	clear-cut	mesic/spruce	0	0	0	-
MM1	mature	mesic/spruce	75	26	260	10/80/10
MM2	mature	mesic/spruce	90	28	310	10/90/0
MM3	mature	mesic/spruce	90	27	290	10/90/10

Layer	Species	Variable	Estimate	Std. Error	t	р	
Surface	Pleurozium	Intercept	4.75	1.72	2.76	0.006	**
		Dev. stage mature forest	2.24	2.42	0.92	0.356	
		Site type sub-xeric	-0.23	0.16	-1.47	0.144	
Surface	Dicranum	Intercept	5.00	0.90	5.56	< 0.001	***
		Dev. stage mature forest	0.58	0.91	0.64	0.523	
		Site type sub-xeric	-0.10	0.10	-0.98	0.327	
Surface	Hylocomium	Intercept	3.98	0.23	17.36	< 0.001	***
		Dev. stage mature forest	2.85	2.54	1.12	0.266	
Surface	Cladonia	Intercept	3.63	0.17	21.11	< 0.001	***
		Dev. stage mature forest	2.13	2.58	0.82	0.412	
Raw humus	Pleurozium	Intercept	5.39	0.21	26.01	< 0.001	***
		Dev. stage mature forest	0.13	0.35	0.36	0.719	
		Site type sub-xeric	-0.20	0.05	-4.24	< 0.001	***
Raw humus	Dicranum	Intercept	4.96	0.41	12.16	< 0.001	***
		Dev. stage mature forest	0.73	0.51	1.43	0.156	
		Site type sub-xeric	-0.16	0.06	-2.44	0.016	*
Raw humus	Hylocomium	Intercept	5.30	0.37	14.50	< 0.001	***
		Dev. stage mature forest	0.43	0.47	0.92	0.363	
Raw humus	Cladonia	Intercept	4.76	0.09	54.53	< 0.001	***
		Dev. stage mature forest	0.62	0.28	2.24	0.027	*

Significant variables (p < 0.05) are in bold

Layer	Species	Smoother term	F p	)	
Surface	Pleurozium	s(FFI) x Dev. stage clearcut	32.18	< 0.001 ***	
		s(FFI) x Dev. stage mature forest	27.33	< 0.001 ***	
		plot (random effect)	3.09	< 0.001 ***	
	Dicranum	s(FFI) x Dev. stage clearcut	27.37	< 0.001 ***	
		s(FFI) x Dev. stage mature forest	29.96	< 0.001 ***	
		plot (random effect)	0.04	0.393	
	Hylocomium	s(FFI) x Dev. stage clearcut	15.18	< 0.001 ***	
		s(FFI) x Dev. stage mature forest	12.75	< 0.001 ***	
		plot (random effect)	0.31	0.326	
	Cladonia	s(FFI) x Dev. stage clearcut	28.54	< 0.001 ***	
		s(FFI) x Dev. stage mature forest	11.76	< 0.001 ***	
		plot (random effect)	0.00	0.841	
Raw humus	Pleurozium	s(FFI) x Dev. stage clearcut	11.07	< 0.001 **	
		s(FFI) x Dev. stage mature forest	2.49	0.111	
		plot (random effect)	0.19	0.366	
	Dicranum	s(FFI) x Dev. stage clearcut	5.93	0.004 **	
		s(FFI) x Dev. stage mature forest	2.36	0.118	
		plot (random effect)	1.73	0.023 *	
	Hylocomium	s(FFI) x Dev. stage clearcut	3.66	0.060	
		s(FFI) x Dev. stage mature forest	6.77	0.011 *	
		plot (random effect)	0.00	0.815	
	Cladonia	s(FFI) x Dev. stage clearcut	30.09	< 0.001 ***	
		s(FFI) x Dev. stage mature forest	5.18	0.026 *	
		plot (random effect)	2.84	0.017 *	

Significant variables (p < 0.05) are in bold

Table 4. Performance of the Finnish Forest Fire Index (FFI) compared to the Canadian Fire Weather Index (FWI) as a predictor of FMC in different layers, measured as pseudo-R2.

Surface layer	FFI	FWI	
	R <sup>2</sup>	$R^2$	
Pleurozium	0.55	0.64	
Dicranum	0.46	0.54	
Hylocomium	0.6	0.69	
Cladonia	0.45	0.52	
Raw humus	FFI	FWI	
	R <sup>2</sup>	R <sup>2</sup>	
Pleurozium	0.26	0.25	
Dicranum	0.36	0.34	
Hylocomium	0.35	0.36	
Cladonia	0.42	0.34	

Table 5. The potential fire hazard days (defined as fuel moisture content values under 25%) of studied surface layer materials, stand types and developmental stages. (MT= mesic stand, SX= sub-xeric stand, C=clear-cut area, M=mature stand, FFI pred = the potential days of ignition predicted by Finnish Forest Fire Index (FFI), index values> 4)

6	7	5
6	7	6

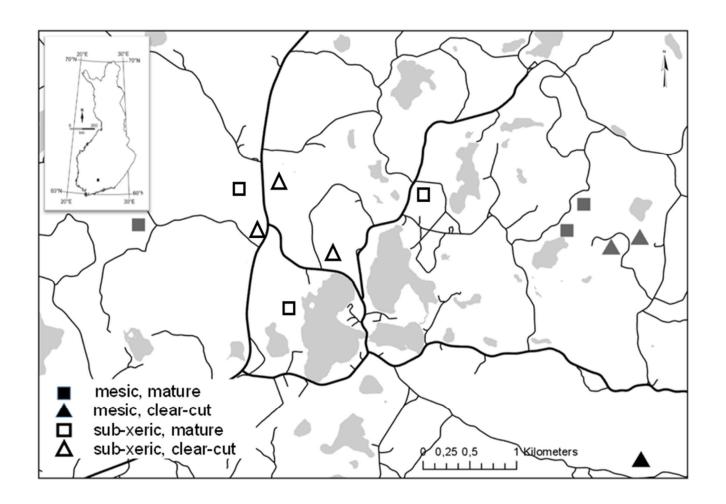
	MTC	MTM	SXC	SXM	MT	SX	С	М	FFFI pred	Total
Pleurozium	54 %	8 %	41 %	31 %	28 %	36 %	47 %	20 %	6 %	32 %
Dicranum	32 %	0 %	27 %	8 %	14 %	18 %	29 %	4 %	6 %	16 %
Hylocomium	54 %	4 %			22 %		54 %	4 %	6 %	22 %
Cladonia			71 %	20 %		45 %	71 %	20 %	6 %	45 %
Total	45 %	4 %	46 %	20 %	21 %	33 %	46 %	12 %	6 %	28 %
FFI > 4									6 %	
FFI < 4									94 %	

Figure 1. Location of sampled stands

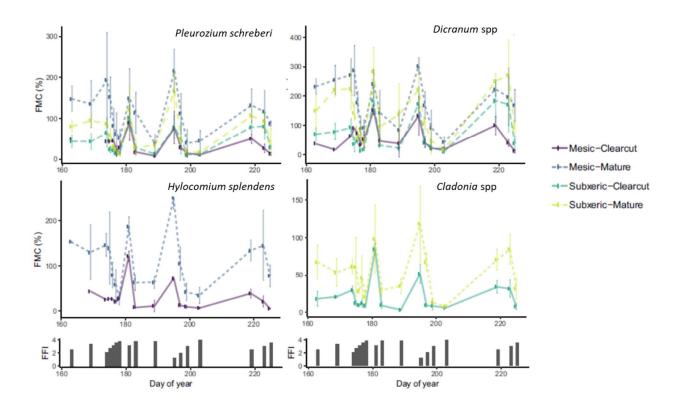
Figure 2. The observed fuel moisture contents (FMC) and Finnish Forest Fire Index (FFI) values on sampling days. Note the different y-axes.

Figure 3. The predicted fuel moisture content (%) of each studied species, by stand type and developmental stage, as a function of Finnish Forest Fire Index (FFI). Dotted lines show the 25-35% moisture content.

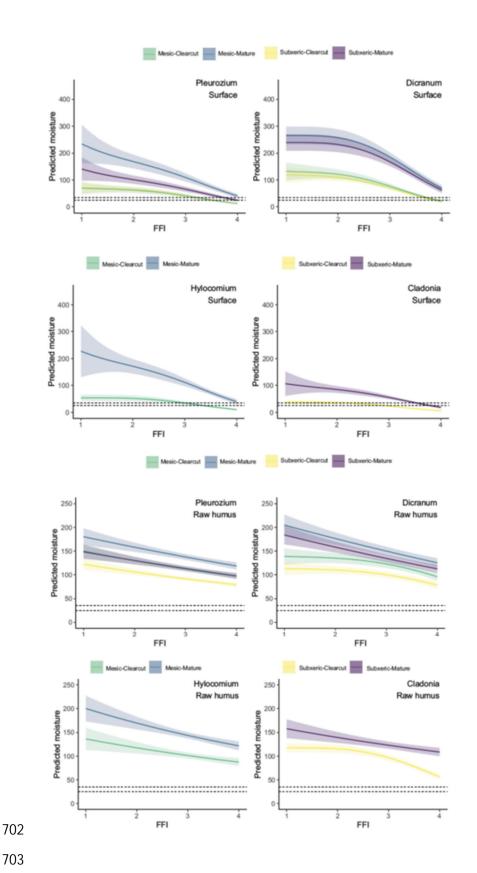
Figure 4. The predicted fuel moisture content (%) by studied species, as a function of Finnish Forest Fire Index (FFI) on different stand types and developmental stages. Dotted lines show the 25-35% moisture content.



**Figure 1**.



**Figure 2.** 



**Figure 3.** 

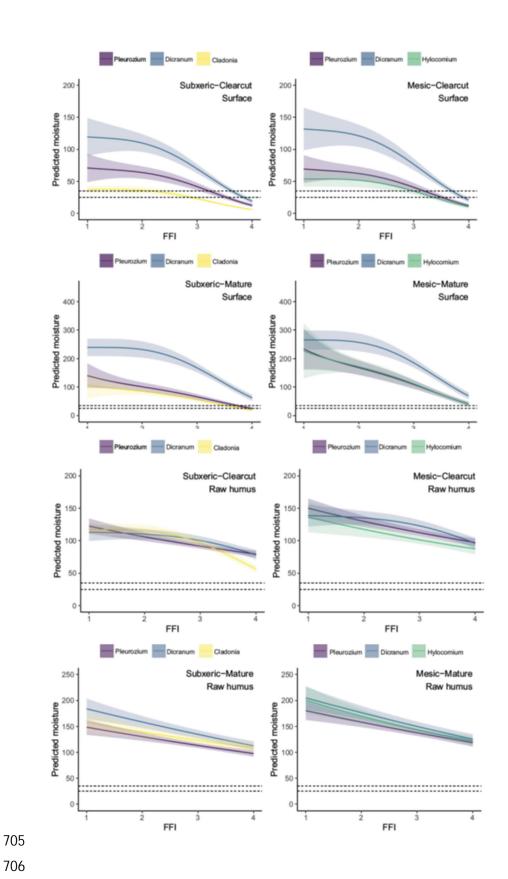


Figure 4.