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Abundance distribution patterns of intertidal bivalves Macoma balthica and Cerastoderma edule at the Murman coast tidal flats (the Barents Sea)

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Keywords:	<i>Cerastoderma edule</i> , cockle, <i>Macoma balthica</i> , Barents sea, Geographic abundance distribution, latitudinal variation, species ranges
Abstract:	Density distribution of the common infaunal bivalves, <i>Macoma balthica</i> and <i>Cerastoderma edule</i> , was studied along the Murman Coast of the Barents Sea during 2002–2010. In both species, abundance was generally higher in West Murman in contrast to East Murman. Highest density of <i>Macoma balthica</i> reaching 1535 ind. m ⁻² was observed in the Kola Bay. <i>Cerastoderma edule</i> was less abundant; its density rarely exceeded 10 ind. m ⁻² in all but one site, where 282 ind. m ⁻² was registered. Reconstruction of abundance distribution across <i>Macoma balthica</i> European geographic range revealed that it does not match "abundant-centre" pattern, having features of ramped north. On the other hand, distribution of <i>Cerastoderma edule</i> abundance across the range generally follows "abundant-centre" pattern but southern edge populations show relatively higher abundances as compared to those at the north edge (the Barents Sea). Nazarova_et_al_abstract.doc

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INTRODUCTION

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Patterns of species' abundance at the biogeographic extremes may provide insights into such essential issues in ecology as identification of borders between adjacent biogeographic regions 3 and causes of species range limits. Large-scale distribution of species abundance is also a key 4 element of applied sciences such as planning of protected areas and designation of species into 5 regional Red Lists. A widespread paradigm, known as "abundant-centre hypothesis" (ACH) is 6 widely used to test whether the highest species abundance belongs to the range centre and 7 declines towards range edges (Sagarin et al., 2006). Intertidal fauna is an ideal test system for studying latitudinal gradients. In Western Europe, many intertidal and high subtidal marine 9 invertebrates have extended ranges with distinct north (or north-eastern) and south (or south-10 western) limits characterized by diametrically opposed combination of environmental conditions. 11 Recent studies on wide-spread marine coastal invertebrates have shown that abundance centre 12 pattern does not apply to most species equally, and even closely related species could show 13 different spatial patterns (Sagarin & Gaines, 2002; Rivadeneira et al., 2010). 14 Empirical approaches to quantify abundance in different parts of the range vary widely. One of 15 the main challenges for examining the abundant centre pattern is a logistical difficulty of 16 sampling throughout the whole species range. Particularly, it could be a problem to get adequate 17 samples from the range edge, where local populations are not stable in time and can 18 19 spontaneously become extinct and later reoccupy suitable habitats (Sagarin & Gaines, 2002; Sexton et al., 2009). 20 Species ranges are not stable in time, often shifting, expanding and contracting (Gaston, 2009). 21 Fluctuations of environmental variables, such as climate oscillations, may alter population fitness 22 at the distribution limits and range shifts can lag behind environmental changes (Pfenninger et 23 al., 2007; Svenning et al., 2008). Marginal populations exhibit greater temporal variability in 24 abundance being near the species' limit of environmental tolerance, especially in extreme years 25 (Gaston, 2009). Since temporal variation can introduce errors in estimation of spatial abundance 26

distribution, it is highly important to apply monitoring data series to improve the understanding 27 of differences in abundance between central and edge populations (Holt & Keitt, 2000; 28 Johnstone & Chapin, 2003). The spatial distribution and patchy structure of population density 29 are also likely to determine range limits. Edge populations usually considered to be smaller and 30 31 more spatially fragmented than populations in the centre of the range (Brown et al., 1995). They may exhibit reductions in diversity and in number of immigrants due to spatial arrangement only, 32 thus resulting in an isolation-by-distance effect (Sexton et al., 2009). 33 Here, we present the results of the study of abundance distribution of Macoma balthica and 34 Cerastoderma edule, two key bivalve species at the tidal flats of the Barents Sea. Macoma 35 balthica is currently regarded as a complex of Pacific Macoma balthica balthica and Atlantic 36 Macoma balthica rubra subspecies (Nikula et al., 2007). In Europe this complex spreads from 37 Bay of Biscay up north to the western Kara sea (Figure 1 A). Southern distribution range limit of 38 Macoma has shifted several hundreds of kilometres north during last decades (Beukema et al., 39 40 2009). Pure M. balthica balhica northern distribution limit belongs to the Varanger Peninsula, while populations in the White and Barents seas are proved to form a broad "hybrid swarm" -41 42 their genetic composition is intermediate between Atlantic and Pacific forms (Strelkov et al., 2007). Cerastoderma edule range is shaped into at least two genetically different groups of 43 populations along the North Atlantic coast, but separation between these groups does not have a 44 subspecies level (Krakau et al., 2012). Earlier we suggested that low density of individuals is 45 typical for C. edule in the Barents Sea, a North-Eastern distribution limit for the species (Genelt-46 Yanovskiy et al., 2010)(Figure 1A). To study latitudinal clines in geographical abundance 47 distributions, we put together published records of mean M. balthica and C. edule density across 48 species' European ranges with our own observations from the Barents Sea. 49

50 MATERIALS AND METHODS

51 Study area

including numerous bays and fjords. The longest among them, Kola Bay, divide Murman Coast into two regions: Western Murman and Eastern Murman. Murman environmental conditions are determined by westerly warm Murman Coastal Current (Denisov & Dzhenyuk, 1995). Interaction between Murman Coastal Current and Arctic water backflow causes remarkable

Murman Coast is a northern border of the Kola Peninsula with a coastal line of about 700 km,

- 57 cooling gradient from West to East Murman. Owing to gradient changes in environment this
- 58 whole area has transitional biogeographical status from boreal to arctic zone (Zenkevich,
- 59 1963). After E.F. Gourjanova, I.G. Zaks and P.V. Ushakov (Gourjanova & Ushakov, 1929;
- 60 Gourjanova et al., 1929; Gourjanova et al., 1930) we consider Western Murman, Kola inlet and
- Eastern Murman as three distinct regions of Murman Coast.

62 Sampling

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- Samples were collected at 18 sites across the Murman Coast (all for *C. edule* and 12 of them for *M. balthica*) (Figure 1, table 1). The sampling took place in July–August, from 2002 to
- 65 2010. At sites inhabited by *Macoma* alone, 5 cm deep cores with a surface area 0.03 m⁻² were
- 66 taken; for sites where *Macoma* and *Cerastoderma* occurred together, 0.1 m⁻² corers were used.
- 67 All samples were sieved through 1 mm mesh. Each tidal level was characterized with five
- samples at all soft-bottom sites studied.
- 69 All collected bivalves were sorted, identified and counted. Shell length of each individual
- was measured using callipers to the nearest 0.5 mm. Age of individuals was determined by
- 71 counting annual growth marks on outer side of the shell. This method of age estimation was used
- 72 basing on previous studies of Macoma and Cerastoderma in the region, confirming that
- structural marks on shells of the molluses in the subarctic seas reflect the annual growth patterns,
- due to the strong seasonal variation in sea water temperature (Agarova, 1979; Maximovich et al.,
- 75 1992; Maximovich & Gerasimova, 2004, Nazarova et al, 2010).

Statistical analysis

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Similar statistical procedures were employed for both species. We started with describing the abundances from each site. We used both medians and means as measures of central tendency and created Tukey boxplots (Tukey, 1976) to illustrate abundance variability. Data on abundance from individual samples from one site was used for plotting. Also year-to-year variability of mean abundance and size-frequency distributions of the species were analysed. Kruskal-Wallis test (Hollander, Wolfe, 1973) was used to compare regional average abundances. The data from published records of mean C. edule and M. balthica abundance are presented together with our own observations in two maps (Figure 5). For the comparisons we used only mean densities of bigger than spat-sized individuals of Macoma and Cerastoderma. Full details on all papers used in assessment of global distribution patterns are listed in electronic supplementary (appendix A). All calculations were done with R (R-core, 2014). In all cases, statistical significances

RESULTS

were tested at 5% probability level.

The abundance of Macoma balthica at Western Murman and Kola Bay sites varied from 90 hundreds to thousands individuals per square meter. M. balthica mean density was significantly 91 different at three regions of the Murman coast (Kruskal-Wallis test: $\chi^2 = 17.6$, p < 0.0001). 92 Highest mean M. balthica density (3350±520 ind. m⁻²) was observed in 2005 at Cape Abram 93 (Kola Bay). M. balthica density at other Kola bay sites (Pala, Retinskoe and Nagornoe) was 94 95 much alike Western Murman sites (figures 2A and 3A). The Eastern Murman M. balthica mean density was lower than in Western Murman and Kola bay and rarely exceeded 100 ind. m⁻² 96 (Figure 3A). However, two major exceptions from general pattern were found in 2008 97 (1208±73 ind. m⁻² in Ivanovskaya bay and 387±77 ind. m⁻² in Yarnyshnaya bay respectively). In 98 Dalne-Zelenestkaya bay (Eastern Murman), mean long-term (from 2002 to 2008) abundance was 99 47 ± 3 ind. m⁻² (Figure 2A). 100

The most typical type of M. balthica size structure was bimodal, but both cases of

prevalence of larger and smaller molluscs could occur in same or subsequent year. So, in 2008

Yarnyshnaya was dominated by 4 mm molluscs, while 17-18 mm-long *M. balthica* prevailed in nearby Dalne-Zelenetskaya and Gavrilovo (Figure 4A).

Alive individuals of *C. edule* were revealed only at 11 sites out of 18. We have not found any alive cockles in Klimkovka, Dolgaya, Tyuva, Gavrilovo, Drozdovka, Ivanovskaya and Gremikha. Density of *C. edule* in Pechenga, Retinskoe, Bolshaya Volokovaya was so low that there was no possibility to estimate it using our sampling scheme, only solitary individuals were present at the tidal flats during our observations. Only empty shells of *C. edule* were wound in Gavrilovo and Drozdovka, while Shelpino population completely become extinct after 2009.

In Ura bay only solitary cockles were present in 2005, but in 2010 abundance increased up to 282 ± 68 ind. m⁻². The other site with rather high abundance was Pala bay (45 ± 9 ind. m⁻² in 2006). In subsequent 2007 cockle abundance at this site had dropped, and only solitary cockles were present (Figure 2B). Cockle abundance was notably lower at all other sites (less than 10 ind. m⁻² at majority of sites studied) (Figure 3B). Mean abundance at the Western Murman was significantly higher than at other two regions (Kruskal-Wallis test: $\chi 2 = 21.4$, p < 0.001).

Unimodal age structures with prevalence of single *C. edule* generation was typical for most sites (22-24 mm-long cockles in Ura bay, 2005; 26-28 mm in Shelpino, 2005 and 20-24 mm in Pala, 2007) (Figure 4B).

120 DISCUSSION

Our results generally match with the expectation that amphi-boreal species *Macoma balthica* will show higher densities in the Barents Sea in comparison to the more temperate-zone *Cerastoderma edule*. Cockle is supposed to be the most sensitive bivalve to low air temperatures on European shores - severe winters were shown to cause mass mortalities among adult *Cerastoderma edule* in the Wadden Sea (Strasser et al., 2001). What effects of severe winters are more critical for cockle survival is still unclear, but probably the sudden rise of the air temperature in subsequent spring may be even more crucial than winter freezing of upper bottom level during low tides (Strasser et al., 2001). The long-term monitoring data from the Wadden

129 Sea provided evidences that long-term variability in average winter temperatures affects bivalve 130 population dynamics in numerous ways (Beukema et al., 2009; Beukema & Dekker, 2014). 131 Community-level analyses revealed that mild warm winters in the Wadden Sea give advantage to 132 recruitment in decapods Crangon crangon and Carcinus maenas, which in turn essentially feed 133 on postlarvae of bivalves Macoma balthica, Cerastoderma edule, Mya arenaria and Mytilus edulis (Beukema & Dekker, 2014). The Barents Sea is lacking Carcinus maenas, and distribution 134 edge of this species belongs to North Norway. Crangon crangon also cannot be regarded as 135 important predator on the Barents Sea intertidal because it does not seem to form permanent 136 137 populations in the Murman Coast (Kuznetsov, 1960). Regional variability in mean abundance, found in both M. balthica and C. edule could be 138 139 interpreted by difference in general environmental conditions between Western and Eastern Murman. West Murman is an area of numerous fjord-like bays, among which Kola Bay is the 140 longest (57 km) and Pechenga and Ura bay are more typical – 16 and 19 km long respectively. 141 142 Eastern Murman coast contrary is comparatively straight. Also, westerly Murman Coastal Current dominates the nearshore hydrodynamics in Eastern Murman, while longer bays of 143 144 Western Murman are more isolated and thus provide more stable conditions (Denisov & Dzhenyuk, 1995). 145 According to our data, mean *Macoma* abundance was significantly lower in Eastern 146 Murman sites in comparison to Western Murman and Kola Bay sites. In 1973 mean density of 147 M. balthica at Dalne-Zelenetskaya bay (23±9 ind. m⁻²) was much alike the present-day values 148 (Agarova et al., 1976). At Western Murman and in Kola Bay clam densities were comparable 149 150 with their densities in the White Sea, in the Baltic Sea and in the nearest waters – the northern part of the Norwegian Sea (for example Semenova, 1974; Aschan, 1988; Bostrom & Bonsdorf, 151 2000; Oug, 2001; Varfolomeeva & Naumov, 2013; Rousi et al., 2013; Gerasimova & 152 Maximovich, 2013). Similar to the Eastern Murman values of mean *Macoma* abundances were 153 154 typical to the southern part of the Norwegian Sea and the Wadden Sea (Sneli, 1968; Beukema,

- 155 1976; Jensen & Jensen, 1985; Reise et al., 1994) (Figure 5A).
- Populations of *Macoma balthica* in the South-Western Barents Sea (Murman coast) are
- 157 certainly not marginal within the species' European part of the range (figures 1A and 5A) -
- 158 Macoma also occur in the Kara sea on the North-East and quite typical for the White Sea
- intertidal communities, where it reaches high densities (Semenova, 1974; Denisenko et al., 2003;
- 160 Strelkov et al., 2007; Gerasimova & Maximovich, 2013).
- Densities of *M. balthica* in Ivanovskaya, the easternmost site for this species in our dataset,
- were dissimilar with other Eastern Murman localities, because in this site *Macoma* appeared to
- be more subtidal than intertidal species it occurs just below the low water level. Subtidal
- 164 Macoma balthica population was previously reported in the Pechora Sea (southeastern part of
- the Barents Sea) with mean densities of 654 ind. m⁻², which is twice lower than in Ivanovskaya
- 166 (Denisenko et al., 2003).
- Our data suggest that *Macoma balthica* abundance distribution within the species range does not
- clearly follow the "abundant centre" (Figure 5A) and is more likely a "ramped north" (for names
- of distribution shapes see Sagarin & Gaines, 2002). Surprisingly, mean individual densities of
- bigger than spat-sized *Macoma* in Western Murman sites were more similar to those in the White
- and Baltic Seas, and were higher than long-term mean values available for North Sea and Bay of
- 172 Biscay (see appendix A).
- Mean densities of adult *Cerastoderma edule* on the Murman Coast of the Barents Sea,
- varying in most sites around 10 ind. m⁻² were generally lower than in other parts of the species
- range, presumably in the North and Irish Seas (Beukema, 1989; Ivell, 1989; André & Rosenberg,
- 176 1991; Bachelet et al., 1992; Ramon, 2003; Dare et al., 2004) (Figure 5B). Long-term mean
- density of cockles in the Dalne-Zelenetskaya bay was equal to previously reported values
- 178 (Agarova et al., 1976). Very similar to the Barents Sea cockle density was previously reported
- 179 from Tromsø, North Norway (Oug, 2001). That data also supports hypothesis that low abundance
- 180 is typical feature of the northern part of the species range. Taking into account that at

southernmost distribution limit, Merja Zerga (Morocco), cockle density may reach 1200 ind. m⁻² 181 with long-term mean of 638 ind. m⁻² (Gam et al., 2010), abundance pattern could be probably 182 characterized as an "abundance center", but with a higher abundance at the southern range edge, 183 as compared to the Barents Sea. (Figure 5B). Current easternmost Cerastoderma local population 184 is located in Tryashina (~68° 59' N, ~36° 30' E), we did not find any alive cockles further east. 185 However, in Drozdovka (~100 km east from Tryashina) we have found out two non-fossilized 186 cockle shells, which mean that temporary outbreak settlement of this species occurred here some 187 years ago. Murman coast clearly is a dynamic range edge for Cerastoderma, and thus in five 188 other sites (Pechenga, Retinskoe, Tyuva, Bolshaya Volokovaya, Gavrilovo) there was no 189 possibility to estimate cockle density using standard sampling methods and only shells were 190 found. No cockles were also found in the inner part of the Kola Bay, Cape Abram and Nagornoe 191 (Figure 3B). We suppose that it could be explained by known sensitivity of Cerastoderma to 192 pollution load (Savari et al., 1989), which should be significant near the city and port of 193 194 Murmansk, largest coastal industrial area in the Barents Sea region. Cerastoderma edule population dynamics has a very typical feature of periodically occurring 195 mass mortality events, when adult densities sharply reduces from thousands of individuals to 196 single cockles per square meter due to a variety of factors. After that slow population recovery 197 takes place (Strasser at al., 2001; Malham et al., 2012; Callaway et al., 2013). In the Wadden Sea 198 during periods, called after Ducrotoy et al. (1991) as "recovery", cockle density may reach 199 54000 ind. m⁻² including individuals of all age groups (Malham et al., 2012). According to our 200 study, Cerastoderma local population in Ura bay on Western Murman unambiguously had 201 features of recovery phase of population dynamics in 2010 after "crisis" phase observed in 2005 202 203 (Figure 2B). During our first visit, only solitary individuals were found across the tidal flat, while in 2010 mean cockle density was 282 ind. m⁻², which is highest density value for the 204 Barents Sea. Our results on cockle local population recovery in Ura bay support the dependence 205 206 (at least, partial) of Barents sea populations of invertebrates and pelagic fish on larval drift from

207	west (Gjøsaeter, 1995; Mikkelsen & Pedersen, 2004).		
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215	REFERENCES		
216	Agarova I.Ya. (1979) Some aspects of the linear growth of bivalve mollusks in the population		
217	of Cerastoderma edule L. Molluscs: main results of studies: abstracts of the annual all-		
218	union conference. Leningrad 6, 80-82. (in Russian)		
219	Agarova I.Ya., Voronova M.N., Galtsova V.V., Ioffe B.I., Letova N.V., Streltsov V.E. and		
220	Streltsova S.I. (1976) Distribution and ecology of benthic fauna on intertidal sandbank		
221	Dalniy Plaj. Ecological studies of soft-bottom littoral. Apatity, 95–186. (in Russian)		
222	André C. and Rosenberg R. (1991) Adult-larval interactions in the suspension-feeding bivalves		
223	Cerastoderma edule and Mya arenaria. Marine ecology progress series. Oldendorf 71(3),		
224	227–234.		
225	Aschan M. (1988) Soft bottom macrobenthos in a Baltic archipelago: Spatial variation and		
226	optimal sampling strategy. In Annales Zoologici Fennici 25 (2), 153–164.		
227	Bachelet G, Guillou J., and Labourg P. J. (1992) Adult-larval and juvenile interactions in the		
228	suspension-feeding bivalve, Cerastoderma edule (L.): field observations and experiments.		

229

230

pp.175-182.

In Marine eutrophication and population dynamics. Olsen and Olsen, Fredensborg,

- 231 Beukema J. J. (1976) Biomass and species richness of the macro-benthic animals living on the 232 tidal flats of the Dutch Wadden Sea. Netherlands Journal of Sea Research, 10(2), 236–261. Beukema J. J. (1989) Long-term changes in macrozoobenthic abundance on the tidal flats of 233 234 the western part of the Dutch Wadden Sea. Helgoländer Meeresuntersuchungen 43(3-4), 235 405–415. **Beukema J. J., and Dekker R.** (2014) Variability in predator abundance links winter 236 temperatures and bivalve recruitment: correlative evidence from long-term data in a tidal 237 flat. Marine Ecology Progress Series 513, 1–15. 238 Beukema J. J., Dekker R., and Jansen J. M. (2009) Some like it cold: populations of the 239 240 tellinid bivalve *Macoma balthica* (L.) suffer in various ways from a warming climate. Marine Ecology Progress Series 384, 135–145. 241 242 Boström C. and Bonsdorff E. (2000) Zoobenthic community establishment and habitat complexity-the importance of seagrass shoot-density, morphology and physical disturbance 243 244 for faunal recruitment. Marine Ecology Progress Series 205, 123–138. Brown J. H., Mehlman D. W., and Stevens G. C. (1995) Spatial variation in 245 246 abundance. *Ecology* 76(7), 2028–2043.
- Callaway R., Burdon D., Deasey A., Mazik K. and Elliott M. (2013) The riddle of the sands:
- 248 how population dynamics explains causes of high bivalve mortality. *Journal of Applied*
- 249 Ecology 50(4), 1050–1059.
- Dare P.J., Bell M.C., Walker P. and Bannister R.C.A. (2004) Historical and current status of
- 251 cockle and mussel stocks in The Wash. CEFAS Lowestoft. 85 pp.
- Denisenko S. G., Denisenko N. V., Lehtonen K. K., Andersin A. B., and Laine A. O. (2003)
- Macrozoobenthos of the Pechora Sea (SE Barents Sea): community structure and spatial
- distribution in relation to environmental conditions. Marine Ecology Progress Series 258,
- 255 109–123.
- 256 **Denisov V. V., Dzhenyuk S. L.** (1995) Chapter 2. Abiotic conditions. Abiotic conditions of Kola

257	Peninsula. In: The Biological Resources of the Coastal zone of the Kola Peninsula. Modern		
258	State and Rational Usage. Russian Academy of Sciences, Apatity, Murmansk, Russia. pp.		
259	10–25. (In Russian)		
260	Ducrotoy J. P., Rybarczyk H., Souprayen J., Bachelet G., Beukema J., Desprez M., and		
261	Ibanez F. (1991) A comparison of the population dynamics of the cockle (<i>Cerastoderma</i>		
262	edule, L.) in North-Western Europe. In: M. Elliott and JP. Ducrotoy. (eds.) Estuaries and		
263	coasts: spatial and temporal intercomparisons. International Symposium Series, ECSA-19		
264	Symposium, Olsen and Olsen, Fredensborg. pp. 173-184		
265	Gam M., de Montaudouin X., and Bazairi H. (2010) Population dynamics and secondary		
266	production of the cockle Cerastoderma edule: A comparison between Merja Zerga		
267	(Moroccan Atlantic Coast) and Arcachon Bay (French Atlantic Coast). Journal of Sea		
268	Research, 63(3), 191–201.		
269	Gaston K. J. (2009) Geographic range limits of species. Proceedings of the Royal Society B:		
270	Biological Sciences, 276(1661), 1391–1393.		
271	Genelt-Yanovskiy E., Poloskin A., Granovitch A., Nazarova S., and Strelkov P. (2010)		
272	Population structure and growth rates at biogeographic extremes: A case study of the		
273	common cockle, Cerastoderma edule (L.) in the Barents Sea. Marine pollution bulletin,		
274	61(4), 247–253.		
275	Gerasimova A. V. and Maximovich N. V. (2013) Age-size structure of common bivalve		
276	mollusc populations in the White Sea: the causes of instability. Hydrobiologia, 706(1),		
277	119–137.		
278	Gjøsæter H. (1995) Pelagic fish and the ecological impact of the modern fishing industry in the		
279	Barents Sea. Arctic 48(3), 267–278.		
280	Gourjanova E.F., Zaks I.G. and Ushakov P.V. (1930) A littoral of western Murman. Studies of		
281	the Seas of USSR 2, 47–52. (In Russian)		

282

Gourjanova E.F., Zaks I.G. and Ushakov P.V. (1929) A littoral of Kola bay. Proceedings of the

- Leningrad Society of Naturalist, 60(2), 17–107. (In Russian) 283 284 Gourjanova E.F. and Ushakov P.V. (1929) A littoral of eastern Murman. Studies of the Seas of 285 *USSR* 10, 5–40. (In Russian) Hollander M. L. and Wolfe D.A. (1973) Nonparametric Statistical Methods. New York: John 286 287 Wiley and Sons Perry Press. Holt R. D. and Keitt T. H. (2000) Alternative causes for range limits: a metapopulation 288 289 perspective. Ecology Letters 3(1), 41–47. **Ivell R.** (1981) A quantitative study of a Cerastoderma – Nephthys community in the Limfjord, 290 Denmark, with the special reference to production of Cerastoderma edule. Journal of 291 292 Molluscan Studies 47, 147–170. Jensen K. T. and Jensen J. N. (1985) The importance of some epibenthic predators on the 293 density of juvenile benthic macrofauna in the Danish Wadden Sea. Journal of 294 Experimental Marine Biology and Ecology 89(2), 157–174. 295 296 Johnstone J. F. and Chapin F. S. (2003). Non-equilibrium succession dynamics indicate continued northern migration of lodgepole pine. Global Change Biology 9(10), 1401-297 1409. 298 Krakau, M., Jacobsen, S., Jensen, K. T., and Reise, K. (2012). The cockle Cerastoderma 299 edule at Northeast Atlantic shores: genetic signatures of glacial refugia. Marine biology, 300 301 *159*(1), 221-230. 302 **Kuznetsov V.V.** (1960) White Sea and biological features of its flora and fauna. Moscow – 303 Leningrad: Publishing House of the Academy of Sciences of the USSR. (In Russian) 304 Malham S. K., Hutchinson T. H. and Longshaw M. (2012) A review of the biology of European cockles (Cerastoderma spp.). Journal of the Marine Biological Association of 305 306 the United Kingdom 92(07), 1563–1577.
- Maximovich N.V., Gerasimova A. V. and Kunina T. A. (1992) Production caracteristics of

 Macoma balthica population in Chupa bay (White Sea). I. Linear growth. Vestnik Sankt-

309 Peterburgskogo Universiteta. Series 3: Biology 4 (24), 12-19. (In Russian) 310 Maximovich N. V. and Gerasimova A. V. (2004) Age determination of the White Sea bivalves by the shell morphology. In Proceedings of the V Scientific Session of the Marine 311 312 Biological Station of St. Petersburg State University, St. Petersburg, Russia: 29–30. (in 313 Russian) Mikkelsen N. and Pedersen T. (2004) How can the stock recruitment relationship of the 314 Barents Sea capelin (Mallotus villosus) be improved by incorporating biotic and abiotic 315 316 factors? Polar Research 23(1), 19-26. Nazarova S. A., Genelt-Yanovsky E. A. and Maximovich N. V. (2010) Linear growth of 317 318 Macoma balthica in the Murmansk tidal zone (the Barents Sea) Vestnik Sankt-319 Peterburgskogo Universiteta. Series 3: Biology 4, 35-43. (In Russian) 320 Nikula R., Strelkov P. and Väinölä R. (2007) Diversity and trans-arctic invasion history of mitochondrial lineages in the north atlantic Macoma balthica complex (Bivalvia: 321 322 Tellinidae). Evolution 61(4), 928–941. Oug E. (2001) Polychaetes in intertidal rocky and sedimentary habitats in the region of Tromsø, 323 324 northern Norway. Sarsia 86(1), 75–83. **Pfenninger M., Nowak C., and Magnin F.** (2007) Intraspecific range dynamics and niche 325 evolution in Candidula land snail species. Biological Journal of the Linnean Society 90(2), 326 303-317. 327 Ramón M. (2003) Population dynamics and secondary production of the cockle Cerastoderma 328 edule (L.) in a backbarrier tidal flat in the Wadden Sea. Scientia Marina 67(4), 429–443. 329 330 R Core Team (2014) R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL http://www.R-project.org/. 331 Reise K., Herre E. and Sturm M. (1994) Biomass and abundance of macrofauna in intertidal 332 sediments of Königshafen the northern Wadden Sea. Helgoländer 333 in 334 Meeresuntersuchungen 48(2-3), 201–215.

335	Rivadeneira M. M., Hernaez P., Antonio Baeza J., Boltana S., Cituentes M., Correa C.,		
336	and Thiel M. (2010) Testing the abundant-centre hypothesis using intertidal porcelain		
337	crabs along the Chilean coast: linking abundance and life-history variation. Journal of		
338	biogeography, 37(3), 486–498.		
339	Sagarin R. D., and Gaines S. D. (2002) The 'abundant centre' distribution: to what extent is it a		
340	biogeographical rule? Ecology letters 5(1), 137–147.		
341	Sagarin R. D., Gaines S. D. and Gaylord B. (2006) Moving beyond assumptions to understan		
342	abundance distributions across the ranges of species. Trends in Ecology and Evolution		
343	21(9), 524–530.		
344	Savari A., Sylvestre C., Sheader M., Le Gal Y. and Lockwood A.P.M. (1989) Stress studies		
345	on the common cockle (Cerastoderma edule L.) in Southampton Water. Topics in Marine		
346	biology. Scientia Marina 53(2-3), 729-735.		
347	Semenova N.L. (1974) The distribution of the bivalve Macoma balthica (L.) in some inlets of		
348	Kandalaksha Bay of the White Sea. Proceedings of the White Sea Biological Station of the		
349	Moscow State University 4, 87–102. (In Russian)		
350	Sexton J. P., McIntyre P. J., Angert A. L. and Rice K. J. (2009) Evolution and ecology of		
351	species range limits. Annual Review of Ecology, Evolution, and Systematics 40, 415–436		
352	Sneli J. A. (1968) The intertidal distribution of polychaetes and molluscs on a muddy shore in		
353	Nord-Möre, Norway. Sarsia 31(1), 63–68.		
354	Strasser M., Reinwald T. and Reise K. (2001) Differential effects of the severe winter of		
355	1995/96 on the intertidal bivalves Mytilus edulis, Cerastoderma edule and Mya arenaria in		
356	the Northern Wadden Sea. Helgoland Marine Research 55(3), 190–197.		
357	Strelkov P., Nikula R. and Väinölä R. (2007) Macoma balthica in the White and Barents Seas:		
358	properties of a widespread marine hybrid swarm (Mollusca: Bivalvia). Molecular ecology		
359	16(19), 4110–4127.		
360	Svenning J. C., Normand S. and Kagevama, M. (2008) Glacial refugia of temperate trees in		

361	Europe: insights from species distribution modelling. Journal of Ecology 96(6), 1117-			
362	1127.			
363	Tukey J. W. (1976) Exploratory data analysis. Massachusetts: Addison-Wesley.			
364	Varfolomeeva M. and Naumov A. (2013) Long-term temporal and spatial variation of			
365	macrobenthos in the intertidal soft-bottom flats of two small bights (Chupa Inlet,			
366	Kandalaksha Bay, White Sea). <i>Hydrobiologia</i> 706(1), 175–189.			
367	Zenkevich L.A. (1963) Biology of the Seas of the U.S.S.R. Moscow. Translated from Russian t			
368	English by S. Botsharskaya. New York: Interscience Publishers.			

Table 1. Investigation sites and sampling details

Region	Site (code)	Latitude, longitude	Observation years
Western Murman	Pechenga (PG)	69.58, 31.27	2005
	Ura bay (UR)	68.99, 36.51	2010
Kola bay	Pala bay (PL)	69.19, 33.37	2006-2007
	Retinskoe (RT)	69.11, 33.38	2005
	Cape Abram (AB)	68.98, 33.03	2008
	Severnoe Nagornoe (NG)	68.90, 33.06	2005
	Tyuva (TU)	69.17, 33.63	2005
	Bolshaya Volokovaya (BV)	69.27, 33.62	2005
Eastern Murman	Klimkovka (KL)	69.23, 34.64	2009
	Dolgaya (DO)	69.17, 33.63	2009
	Gavrilovo (GV)	69.17, 35.86	2008
	Yarnyshnaya (YA)	69.09, 36.05	2004 - 2008
	Dalne-Zelenetskaya (DZ)	69.11, 36.10	2002 - 2009
	Shelpino (SH)	69.10, 36.21	2005, 2008
	Porchnikha (PR)	69.08, 36.25	2006 - 2007
	Tryashina (TR)	68.99, 36.51	2008
	Drozdovka (DR)	68.30, 38.44	2008
	Ivanovskaya subtidal (IV)	68.29, 38.71	2008
	Gremikha (GR)	68.03, 39.85	2009

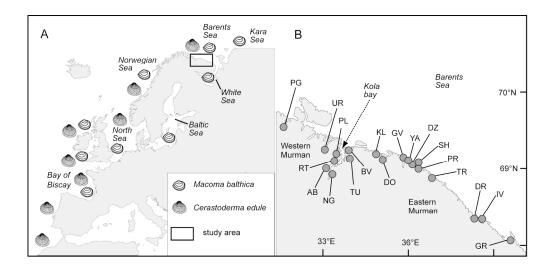
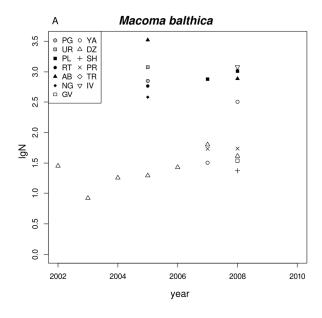


Fig 1.> Distribution range map of *Macoma balthica* (European part) and *Cerastoderma edule* (A) and study area map (B). See table 1 for sample sites details.

175x85mm (300 x 300 DPI)



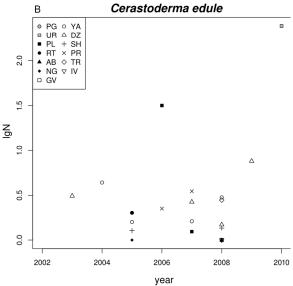
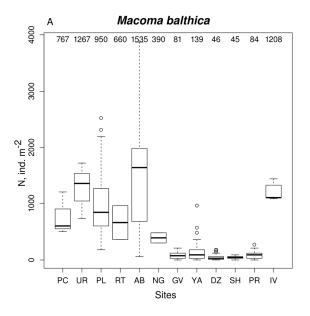


Fig. 2. Year-to-year variability of *Macoma balthica* (A) and *Cerastoderma edule* (B) mean abundance (IgN). Symbol colour corresponds to the region: Western Murman – solid grey, Kola pay – solid black and Eastern Murman – open symbols.

101x208mm (300 x 300 DPI)



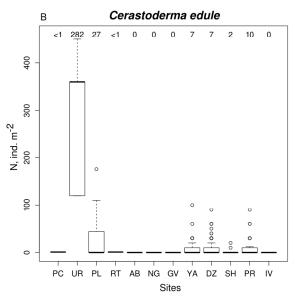


Fig. 3. Mean abundance (N, ind.m-2) of *Macoma balthica* (A) and *Cerastoderma edule* (B) in the Barents sea. Samples are arranged from West to East. At the box-plots box marked Q1, median and Q3, whiskers marked 1.5 IQR. Mean abundance indicated at the top. 101x210mm (300 x 300 DPI)

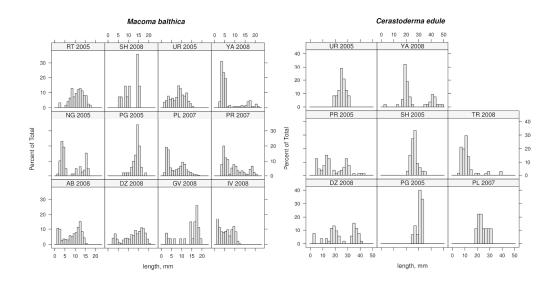


Fig. 4. Size frequency distributions of *Macoma balthica* (A) into 1 mm classes and *Cerastoderma edule* (B) into 2 mm classes.

209x105mm (300 x 300 DPI)

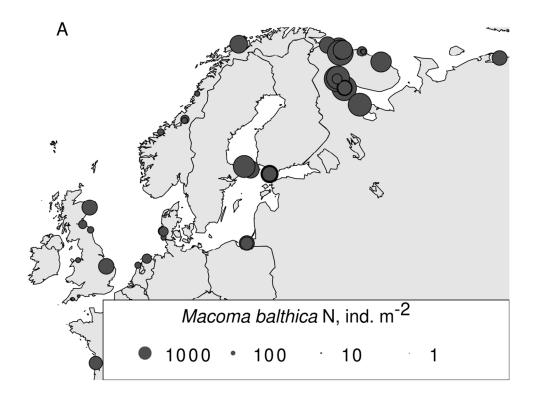


Fig. 5. Geographic abundance distributions of *Macoma balthica* (A) and *Cerastoderma edule* (B). Circles squares are proportional to mean abundance. For sources of data see Appendix A. $109 \times 84 \text{mm}$ (300 x 300 DPI)

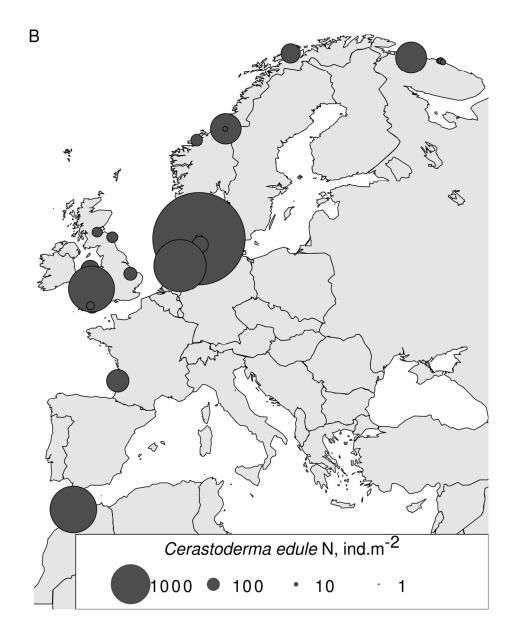


Fig. 5. Geographic abundance distributions of *Macoma balthica* (A) and *Cerastoderma edule* (B). Circles squares are proportional to mean abundance. For sources of data see Appendix A. $108 \times 139 \, \text{mm}$ (300 x 300 DPI)