

A temperature-driven model of phenological mismatch provides insights into the potential impacts of climate change on consumer-resource interactions

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3 **consumer-resource interactions**

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Abstract

1. Anthropogenic climate change is increasingly affecting species phenology. Because trophic interactions often occur at specific phenological stages, changes in one species' phenology may affect others through phenological mismatch. In the case of a consumer and a resource that both exhibit a seasonal resting period, the synchrony of the end of their respective resting period is fundamental for the persistence of their interaction. Since the consumer and its resource may react differently to a change in temperature regime, the synchrony between them will likely be altered.
2. The goal of this study is to investigate the potential effects of climate change of species synchrony.
3. We propose a general theoretical model that determines the duration of the resting period according to temperature, and its effects on the mismatch between phenological stages of two interacting species. We illustrate our approach using the spruce budworm – balsam fir system in eastern Canada.
4. We found that an increase in temperature would usually advance the end of the resting period. However, the effects of a warm or cold spell during the resting period would strongly vary according to the time and the duration of the spell. Depending on the way each species reacts to the same temperature shift, the mismatch between the consumer and its resource may increase or decrease. The spruce budworm – balsam fir model predicts that an increase in temperature may increase the mismatch between the insect and the tree in southern sites, but may increase the synchrony in northern sites.
5. This type of modelling approach is of prime importance to investigate potential effects of climate change on consumer – resource systems.

Keywords:

Balsam fir, consumer, climate change, phenology, resource, spruce budworm

1. Introduction

Anthropogenic climate change has increasingly disrupted ecological interactions for the past century (Pachauri et al., 2014). This trend is expected to continue and amplify as interacting species are likely to respond differently to similar environmental changes and selective pressures (Parmesan, 2006). Trophic interactions between consumers and resources, which are fundamental to the functioning of ecosystems, may be affected by climate change through: (1) direct changes in life history traits (e.g., fecundity, mortality) of consumer and/or resource (Bale et al., 2002), (2) changes in abundance of consumer and/or resource due to cascading effects from other trophic levels (e.g., changes in the consumer’s predators) (Both, et al., 2009), and (3) differential shifts in the phenology of consumer and/or resource leading to phenological mismatch (Kharouba et al., 2018).

The concept of phenological mismatch has evolved since its inception in the early 1990s and is still debated today (e.g., Singer & Parmesan, 2020’s response to Kharouba & Wolkovich, 2020). In its stricter, original form, it states that the recruitment of a consumer is highest if the most energy expensive part of its life cycle is in synchrony with the peak availability of its resource (Cushing, 1990). The concept was later extended to the case where the phenologies of consumer and resource varied (Visser & Holleman, 2001), but still assumed that maximum fitness of the consumer is achieved at phenological synchrony. However, recent observations

suggest that, in some systems, phenological mismatch is the historical baseline (Singer & Parmesan, 2010). Historical mismatch can result from trade-offs between fecundity and mortality (Singer & Parmesan, 2010), mutualistic interactions (Forrest & Thomson, 2012) or intraspecific competition (Iwasa et al., 1983). Relaxing the assumption of baseline synchrony complicates predictions of potential climate change impacts. Indeed, if synchrony is the baseline, any differential change will be detrimental to the consumer's fitness. If the baseline is a mismatch, a change that decreases asynchrony will likely be beneficial to the consumer (but see Régnière & Nealis, 2018) while the opposite might exacerbate the detrimental effect of asynchrony to the point of extinction of the consumer (Singer & Parmesan, 2010).

Many organisms have advanced their phenology in recent decades (Post et al., 2001; Parmesan & Yohe, 2003), particularly the timing of spring events at mid-high latitudes (Parmesan, 2006; Cohen et al., 2018). In a meta-analysis of 27 pairs of interacting species, Kharouba et al. (2018) found that phenology advanced by an average of 4 days/decade across species since the early 1980s. Interacting species have advanced their phenology by similar magnitudes, resulting in relatively small (6.1 days/decade) but significant increases in mismatch. The observed mismatch had no consistent direction as 31 interactions shifted closer while 23 shifted further apart. The clearest examples of climate-driven mismatch have been reported for insect herbivores at high altitudes or latitudes (Renner & Zohner, 2018), where many organisms rely on a seasonal resting period because temperatures affecting physiological processes tend to be below species optima for most of the year. After a seasonal arrest in development that can last for several months, synchrony between the emergence of phytophagous insects from diapause and the end of their host plants dormancy, is critical to the consumer's fitness (Van Asch & Visser, 2007). Spring defoliators have evolved to exploit foliage at its annual optimal nutritional

82 qualities, i.e., high concentration in nutrient and water and low concentrations in fibre and
83 secondary metabolites (Mattson & Scriber, 1987). This strategy requires a precise phenological
84 match between the consumer and its resource because the quality of foliage declines quickly with
85 foliar expansion. If the consumer emerges too early, it may encounter a long initial period with
86 no food or poor food quality. If it emerges too late, the speed at which the nutritional quality of
87 the foliage degrades may outpace the consumer's capacity to adjust physiologically, thus
88 negatively impacting its fitness.

89 The seasonal resting phase of many organisms is divided into two successive stages called
90 endodormancy and ecodormancy in perennial woody plants, and diapause and quiescence in
91 insects (Chaine & Régnière, 2017). Development and metabolism are generally inhibited by
92 internal factors (e.g., depletion of energy reserves, hormones) in the first stage and by external
93 factors (e.g., temperature, photoperiod) in the second. In our study system, the consumer and its
94 resource show a seasonal resting period during part of the year (e.g., during winter). For both, we
95 assume that the first stage is accomplished early in the winter before the return of favourable
96 conditions, a common case in temperate and colder climates, and that temperature is the main
97 driver that triggers the end of the second stage and therefore the resting period. We will refer to
98 the second stage of the seasonal resting period simply as the "resting period".

99 Science is currently unable to predict the direction or the magnitude of phenological
100 mismatch between consumer and resource induced by climate change and the associated risks
101 that it poses to species (Samplonius et al., 2021). While patterns of change in phenological
102 synchrony are observed at an increasing rate because of climate change (Kharouba et al. 2018),
103 they have been difficult to explain mechanistically. Different mechanisms can lead to a
104 phenological mismatch between a consumer and its resource. Different species can respond to

different climatic factors or to the same factor but in a different way. Species can also face different constraints in phenological plasticity (Both & Visser, 2001) or have different costs associated with phenological response (Gienapp & Visser, 2006).

The aim of this study is to investigate potential effects of climate change on the phenological synchrony between a consumer and its resource when both species respond to the same climatic factor (i.e., temperature). The interaction occurs at the end of both species' resting periods, which is often the case for insect herbivores in mid-high latitudes. We begin with the general theoretical aspects that determine the duration of the resting period according to temperature, and their effects on synchrony / mismatch between phenological stages of the two species. Then we use a major insect pest of the Canadian boreal forest, the spruce budworm (SBW), and its main host, balsam fir, as a case study. We investigate the phenological synchrony between budworm's emergence from winter diapause and balsam fir's budburst across a gradient of latitudes and a number of different future climates.

2. The general model

We begin with a unified description of the mechanisms that determine the duration of the resting period of a consumer and its resource in terms of accumulation of ambient temperature. Then, we develop the main theoretical results at the single species level and for two interacting species.

2.1 Theoretical development

Throughout the resting period, an organism accumulates units of some quantity. The instantaneous rate of accumulation depends on ambient temperature. The resting period ends when a certain level of the quantity has accumulated. For trees, this quantity can be heat, for

example in degree-day models (Colombo, 1998) or more recent nonlinear models (Chaine, 2000; Desbiens, 2007). For insects, the quantity can be the proportion of the corresponding life-cycle stage that they have completed (also referred to as physiological stage) (Cobbold & Powell, 2011; Régnière, et al., 2012). Since the development rate is temperature dependent, this quantity ultimately measures accumulated heat. This concept of an accumulating quantity is “still the most important assumption in plant and animal phenology modelling” (Chaine & Régnière, 2017).

We denote time by t in days and temperature by $x = x(t)$ in degrees Celsius. The instantaneous rate of accumulation is some nonnegative, increasing function of temperature, denoted by $R = R(x)$. While developmental rates may decrease when temperatures exceed an upper threshold (Amarasekare & Coutinho, 2014), such temperatures do not generally arise during the winter resting period in mid-high latitudes. The resting period begins at time t_0 and ends at time t^* when the accumulated quantity reaches the threshold level F . The fundamental equation that connects all these quantities and determines the end of the resting period is

$$\int_{t_0}^{t^*} R(x(t))dt = F. \tag{Eq. 1}$$

A typical example for the accumulation rate function is

$$R(x) = \frac{1}{1 + \exp(b(x - c))}, \tag{Eq. 2}$$

with parameters $b < 0$ and c (Chaine, 2000; Rebaudo & Rabhi, 2018). We illustrate the rate function in Eq. 2 as well as the condition in Eq. 1 for two different species and two simplistic temperature time series in Figure 1. As temperature patterns during the resting period change because of climate change, the end time of the resting period of a species may shift. When

temperatures increase, accumulation occurs faster and the phenology advances, i.e., the end time is earlier (Fig. 1D).

We are particularly interested in the case where the two species are a consumer and its resource. Their rate accumulation functions generally differ even in the same temperature regime (compare solid and dashed curves in Fig. 1C), which typically leads to different end times of the resting period (Fig. 1D). In view of our intended application, we also refer to the consumer as the insect and to the resource as the tree. We denote the corresponding end times by t_e^* (emergence time) for the consumer (insect) and by t_b^* (budburst time) for the resource (its host tree). We call the difference in end times the *mismatch* between the two species (i.e., $mismatch = t_e^* - t_b^*$). When the end times of the resting period of two species respond differently to climate change, then the mismatch between the two species will change. This is the fundamental quantity that we study here (Fig. 1D).

2.2 Theoretical results

Model equation (Eq. 1) can in general not be solved explicitly for the end time, t^* , at least not for realistic temperature time series $x(t)$. Instead, we derive a general approximation formula for how t^* changes when future temperature time series ($x_2(t)$) deviate from historical expectation ($x_1(t)$) by a small amount (see Supplementary Material). We use this formula to evaluate and discuss the impact of climate change on the phenology of a single species and the phenological mismatch between two species in two particular cases: a constant temperature difference throughout the resting period or a warm or cold spell during a short time period. Our results here apply to any general rate accumulation function.

165 **2.2.1 Phenology shift of a single species**

166 When temperatures differ by a constant (Δx), we write $x_2(t) = x_1(t) + \Delta x$. Then the
167 corresponding end times t_2^* and t_1^* are related by

$$t_2^* = t_1^* - \frac{\Delta x}{R(x_1(t_1^*))} \int_{t_0}^{t_1^*} R'(x_1(t)) dt. \tag{Eq. 3}$$

168 When temperatures differ by warm or cold spell of short duration Δt at time t_s of temperature
169 difference Δx (positive for a warm spell, negative for a cold spell), then the corresponding ends
170 of the seasonal resting phases are related by

$$t_2^* = t_1^* - \Delta x \Delta t \frac{R'(x_1(t_s))}{R(x_1(t_1^*))}. \tag{Eq. 4}$$

171 Both formulas show the expected qualitative pattern that if time series x_2 is warmer than x_1 , i.e.,
172 $\Delta x > 0$, then the phenology advances, i.e., end of the seasonal resting period t_2^* is before the
173 corresponding t_1^* (since all the terms after the “ $-$ ” sign are positive). More importantly, the
174 formulas allow us to quantify the expected shift of the end time of the resting period. We observe
175 that the shift depends on the *derivative* of the rate accumulation function. In particular, the impact
176 of a short temperature spell is proportional to the derivative, $R'(x(t_s))$, at the time of the spell (t_s).
177 Hence, the phenology is most sensitive to warm or cold spells where $R(\cdot)$ is at its maximal
178 slope. For the rate function in Eq. 2, this occurs at temperature $x=c$.

2.2.2 Effects on the mismatch between two species

When heat accumulation occurs independently in each species, the mismatch between the two changes if they respond differently to changes in temperature. The mismatch increases if the phenology of the earlier emerging species advances more quickly and decreases if it advances more slowly. Equations 3 and 4 require information about the current phenology ($R(x_1(t_1^*))$) and the sensitivity ($R'(x_1(t_1))$) to quantify the change in phenology in each species and therefore the change in mismatch.

For a theoretical example, we consider the scenario that the resource emerges before the consumer and is more sensitive at lower temperatures. Then, if temperatures are higher than historical averages early in the season, the phenology of the resource is likely to advance more than that of the consumer, so that the mismatch increases. Vice versa, if temperatures exceed historical averages only late in the season, when the consumer is more sensitive, the phenology of the consumer is expected to advance more, so that the mismatch decreases. Hence, a change in mismatch is most likely if the two species are most sensitive at different temperatures, i.e., when the maximal slopes of their respective rate accumulation functions occur at very different temperatures.

In reality, the periods of high sensitivity of the two species may overlap more or less, and the rate functions at emergence time (the terms in the denominators in Eqs 3 and 4) could differ significantly. As a result, the effect of temperature increases depends on details of each scenario. We illustrate this dependence using a simplified time series of daily mean temperatures as modelled by

$$x_i(t) = 6.9 + 15\cos\left(\frac{2\pi(t - 200)}{365}\right) \quad \text{Eq. 5}$$

where the mean, amplitude and offset have been chosen to match historical averages in Fredericton (NB, Canada). We denote the end times of the consumer for the time series $x_i(t)$ by $t_{e,i}^*$ (emergence time) and of the resource by $t_{b,i}^*$ (budburst time).

When future temperatures differ from historical expectation by a constant (Δx), we apply Eq. 3 to each species (R_e is for the insect, and R_b is for the tree). The resulting mismatch changes according to

$$\underbrace{t_{e,2}^* - t_{b,2}^*}_{\text{mismatch2}} = \underbrace{t_{e,1}^* - t_{b,1}^*}_{\text{mismatch1}} - \Delta x \left(\frac{\int_{t_0}^{t^*} R_e'(x_1(t)) dt}{R_e(x_1(t_{e,1}^*))} - \frac{\int_{t_0}^{t^*} R_b'(x_1(t)) dt}{R_b(x_1(t_{b,1}^*))} \right). \quad \text{Eq. 6}$$

The resulting linear approximation captures the actual end of the resting period very well (see supplementary Fig. S1). The derivative has to be integrated over the whole development period. Thus, it is difficult to estimate the resulting trend at first glance because it mostly depends on the amount of time that temperatures stay in the sensitive range of each species (that drives the numerator values), but it also depends on the value of the function at emergence time (denominator).

When future time series differ from historical expectation by a short spell of duration Δt and temperature difference Δx , we apply the corresponding formula (Eq. 4) to each species. Then the mismatch changes according to

$$\underbrace{t_{e,2}^* - t_{b,2}^*}_{\text{mismatch2}} = \underbrace{t_{e,1}^* - t_{b,1}^*}_{\text{mismatch1}} - \Delta x \Delta t \left(\frac{R'_e(x_1(t_s))}{R_e(x_1(t_{e,1}^*))} - \frac{R'_b(x_1(t_s))}{R_b(x_1(t_{b,1}^*))} \right). \quad \text{Eq. 7}$$

215 In case of non-overlapping sensitivity periods, each species exhibits a sensitivity that is a
 216 relatively narrow peak around a maximum at $x=c_i$, and they peak at different temperatures. The
 217 phenology of one species will advance noticeably only when the spell occurs near its peak
 218 temperature. The other species is not affected (Fig. 2A).

219 In the case of two overlapping sensitivity periods, one species (e.g., the consumer) may
 220 exhibit a broader sensitivity that peaks at a different temperature than the other species (e.g., the
 221 resource). Hence, in this example, the phenology of the consumer will advance noticeably no
 222 matter when a warm spell occurs, whereas the phenology of the resource will advance noticeably
 223 only when it occurs near the temperature $x=c_I$ (Fig. 2B).

224 However, among all scenarios, the rate function at emergence (the denominators in Eq. 7)
 225 may differ vastly between the two species. Consequently, even if the resource is more sensitive
 226 than the consumer at the time of the spell ($R'_e > R'_b$), the difference in parentheses in Eq. 7 can
 227 still be negative if the denominator in the first term is much larger than in the second.

228 Despite their simplicity, these scenarios give us some important insights into the evolution
 229 of the mismatch. The advance of the phenology of each species is a nonlinear function of
 230 temperature increase and each species has a temperature regime during which its phenology is
 231 most sensitive to change, namely when the slope of the rate function is large. In this section, we
 232 used Eq. 1 to develop general theoretical insights. The model is agnostic about the specific R
 233 functions used. The model does not make any assumption on the consequences of a change in
 234 synchrony between the two interacting species, in terms of survival, reproduction, or any life

235 history trait. These aspects are system-specific. In the following section, we apply the model
236 using realistic temperature time series in order to make predictions for a specific case study, at a
237 given place and time.

238 **3. A case study: the spruce budworm – balsam fir system**

239 We describe the spruce budworm – balsam fir system and provide the accumulation function for
240 both species. Then, we list our data sources and explain the fitting methods for the spruce
241 budworm - balsam fir system. Finally, we give the results for this case study. We use R (R core
242 team, 2020) to implement the model, analyze the results, and generate the figures. Past and future
243 temperature data were obtained using BioSIM ((Régnière, et al., 2014).

244 **3.1 Study system**

245 The spruce budworm (SBW) (*Choristoneura fumiferana*) is the most destructive defoliator of
246 North American boreal forests (Fleming, 2000). It feeds primarily on balsam fir (*Abies*
247 *balsamea*), black spruce (*Picea mariana*), white spruce (*Picea glauca*), and red spruce (*Picea*
248 *rubens*). This univoltine insect has an 8-9 months winter resting period (diapause). The first
249 stage ends in late winter and is followed by a quiescent period during which development
250 resumes at a rate dependent on outside temperatures. Diapause ends when second instar larvae
251 emerge from their hibernacula (cocoon-like structure) in late April to late May. Larvae feed on
252 expanding buds and developing needles as soon as they become available. Expanding needles
253 from swelling buds are nutrient rich with a maximum concentration of nitrogen and mineral
254 elements (Mattson & Scriber, 1987). Hence, the success of SBW populations in establishing
255 feeding sites in the spring depends on the synchrony of their development with that of their host
256 trees (Régnière & Nealis, 2008). Consequences for late-emerging SBW are severe: larval

survival is depressed (from about 60% normal survival down to 10%), development is delayed, and average pupal mass decreases by about 50% (Lawrence et al., 1997). In the boreal forests of eastern Canada, the higher vulnerability of balsam fir to SBW defoliation compared to other host species has been attributed to its early budburst (Blais, 1957). The emergence of SBW generally precedes balsam fir budburst by several days. Balsam fir phenology appears to be related to forcing temperatures but not photoperiod (Osawa et al., 1983).

We assessed historical patterns and climate change predictions of synchrony between SBW's emergence from diapause and balsam fir's budburst using an existing model of budworm phenology and a general budburst model fitted to observations of budburst phenology in Canada. Each species has its own accumulation function which we used to illustrate the general theoretical case (see section 2.2.2 and Fig. 2).

3.2 Phenological models

3.2.1 Phenological model of spruce budworm's spring emergence

The first stage of SBW diapause is assumed to end in late winter, but the physiological processes and environmental factors that control this are still unknown (Bean, 1961). For modelling purposes, March 1st is generally taken as the start of the quiescent stage (Régnière, et al., 2012). Taking an earlier starting date would not affect the predicted date of emergence since temperatures are generally well below the 2.5°C development threshold prior to March 1st across the vast majority of SBW's geographical range. The heat accumulation rate of the overwintering stage (L_{20}) of SBW is defined as (Régnière, et al., 2012)

$$R(x(t)) = \begin{cases} \beta_1 \left[\frac{1}{1 + \exp(\beta_2 - \beta_3)} - \exp((\tau - 1)/\beta_4) \right], & \text{if } x_b \leq x(t) \leq x_m \\ 0, & \text{otherwise} \end{cases} \quad \text{Eq. 8}$$

277 where

$$\tau = \frac{x(t) - x_b}{x_m - x_b} \quad \text{Eq. 9}$$

278 Accumulation occurs only when the temperature lies between a minimal (x_b) and a maximal
279 value (x_m). Parameter values for SBW, estimated by Régnière et al. (2012) from laboratory
280 experiments, are $\beta_1 = 0.194$, $\beta_2 = 3.0$, $\beta_3 = 5.94$, $\beta_4 = 0.034$, $x_b = 2.5$ °C, $x_m = 35$ °C and $F=1$.
281 The accumulation rate function is increasing for temperatures up to 31°C, which is well above the
282 temperatures observed during the quiescent stage.

283 **3.2.2 Phenological model of balsam fir’s budburst**

284 Process-based tree phenology models describe the individual or interactive effects of
285 environmental conditions (e.g., temperature or photoperiod) on bud development (Chuine &
286 Régnière, 2017). We use the *Uniforc* model of Chuine (2000) to model balsam fir’s budburst
287 phenology. *Uniforc* predicts bud development as a function of temperature in the second stage of
288 seasonal resting (i.e., ecodormancy). The heat accumulation rate is the same as in the theoretical
289 example (see Eq. 2). Accumulation starts some time after January 1st (Desbiens, 2007), when
290 trees have accumulated enough cold to end bud dormancy. Budburst occurs when accumulation
291 reaches a threshold F^* .

292 We fitted the Uniforc model to budburst phenology data collected in the 1980s and 1990s
293 in Quebec and New Brunswick (Desbiens, 2007; Régnière pers. comm. 2020). Each year, bud

development was observed in different sites during the growing season at time intervals ranging from two days to two weeks. Budburst occurs when buds develop from class I to II according to the class scheme developed by Dorais & Kettela (1982). The budburst date was defined as the date when 50% of the buds in the site have reached stage II. We obtained temperature data at each site for each year using BioSIM. We estimated parameter values of the Uniforc model using simulated annealing in order to predict budburst date according to temperatures during development period.

We performed sensitivity analysis on both models using partial rank correlation coefficients (Wu et al., 2013).

3.3 Historical and future temperature regimes in eastern Canada

We selected six locations on a latitudinal gradient across eastern Canada (Fig. 4) to explore the effects of historical and future temperature regimes on the synchrony between SBW and balsam fir phenology. The temperature data for these locations were calculated using BioSIM (Régnière, et al., 2014). BioSIM interpolates weather station historical data and climate change scenarios across North America. We used data from 1996 to 2016 to explore latitudinal trends of past insect emergence, tree budburst and their mismatch across years.

We used predicted temperatures under different warming scenarios to reveal expected trends for emergence, budburst and mismatch. We selected three different scenarios, known as RCP2.6 (moderate warming), RCP4.5 (intermediate) and RCP8.5 (strong) (van Vuuren et al., 2011). For each scenario, we generated 1200 stochastic temperature time series over the 2001-2100 period using BioSIM.

Both models used temperature data with a four-hour time interval, which allows for the capture of warm events within a day.

3.4. Mismatch in the spruce budworm - balsam fir system

3.4.1 Fitting and sensitivity

Fitting the *Uniforc* model to phenological data from Quebec and New Brunswick resulted in the parameter values: $b = -1.32$, $c = 7.14$ °C, $t_0 = 87$ (March 28th), and $F^* = 18.6$ (RMSE = 12.6). The residuals of this fitting follow a Normal distribution centred on 0 (Fig. 5A). There is no obvious pattern for the residuals across latitude in the range of our study (Fig. 5B).

In order to test the accuracy of both insect and tree models, we compared predicted budburst and emergence date with available data for two years (2013, 2014) in two sites in Quebec (Pureswaran, et al., 2019). Measured budburst occurred between May 16th and June 17th, in 2013, and between May 29th and June 15th, in 2014. The model predicts a median budburst on May 27th in 2013, and May 28th in 2014. For the insect, the emergence peak occurred on May 9th in 2013, and on May 23th in 2014. The model predicts a median emergence on May 20th in 2013, and May 21th in 2014. Hence, the model gives satisfactory results.

The budworm model is sensitive to most parameters (Fig. 5C). The only exception is x_m (the maximal temperature) since very high temperatures are rare during late winter and spring, and to a certain extent β_1 . Increasing parameters β_2 , β_4 , and x_b (minimal temperature) delays emergence, while increasing β_3 strongly advances phenology. The tree model is most sensitive to parameters t_0 (when the tree starts accumulating heat) and b (which drives the speed of

accumulation). An increase in t_0 postpones phenology, while an increase in b advances it (Fig. 5D).

3.4.2 Spruce budworm – balsam fir system across latitude

The model predicts median dates of emergence and budburst for each year. Results for past and future data were analyzed using a one-way analysis of variance (see supplementary material for full results). Both emergence (Fig. 6A) and budburst (Fig. 6B) occur later at higher latitudes: $p < 0.05$ between sites 1, 2, 3 (South) and 4, 5, 6 (North), although some discrepancies may occur due to altitude (sites 2 and 4). However, insects and trees are not affected by temperatures in the same way. Hence, the mismatch between both species varies across latitude ($p < 0.05$ between sites 1, 3 and 5, 6) (Fig. 6C). At lower latitudes, emergence is expected to occur 5 to 10 days before budburst. Hence, the majority of the larval population may have time to emerge before budburst occurs, which allows larvae to benefit from an important source of nutrients. Fuentealba, et al. (2017) evaluated SBW performance on balsam fir according to their phenological mismatch. According to their study, we can infer SBW survival to be around 45% and reproductive output (fecundity * survival) to be 45%. At higher latitudes, emergence may sometimes occur before budburst and sometimes after, which would lead to 30% survival and 20% reproductive output. In brief, insects seem more adapted to their host phenology at lower than at higher latitudes.

3.4.3 Predicted trends according to warming scenarios

Across all scenarios, emergence and budburst are expected to occur earlier when temperatures increase. The differences in emergence, budburst and mismatch across latitude is highly significant ($p < 10^{-16}$ for almost all sites and for all scenarios). Warmer scenarios lead to an increase of variance, more than a real shift in date compared to less warm scenarios. Insects and trees react differently to temperature increase. Balsam fir shows an historical (1996-2016)

358 difference between northern and southern sites of 10 days on average. This difference stays
359 approximately the same in case of warmer temperatures (i.e., budburst date is shifted similarly
360 across latitude, see Fig. 6B). By contrast, emergence of SBW is expected to shift differently
361 across latitude (about 15 days difference compared to 10 historically, see Fig. 6A).

362 Therefore, the mismatch is affected. Southern sites are expected to show emergence occurring
363 before budburst most of the time. In case of a moderate warming scenario (RCP2.6), the waiting
364 time is expected to be short, which would allow the larvae to have access to a rich source of
365 nutrients. Thus, a moderate warming could be beneficial for the insect at low latitudes. In case of
366 greater warming (RCP4.5, RCP8.5), variance may lead to emergence occurring too early some
367 years, leading to larvae dying from starvation. Thus, a greater temperature increase may lead to
368 stronger population variances among years.

369 In northern sites, all warming scenarios lead to an increased synchrony between the insect and
370 its host. The overall pattern appears similar to what is currently expected in southern sites.
371 However, the variance is expected to be greater with warmer scenarios, which may increase
372 insect mortality from time to time, when emergence occurs too late.

373 **4. Discussion**

374 We propose a temperature-driven mechanistic model to determine the end of the seasonal resting
375 period of species. We investigate the effects of temperature on species phenology, and its
376 consequences on consumer-resource synchrony, across latitude and under future climate
377 scenarios.

4.1 Expected impacts of climate change on phenological mismatch between consumer and resource

A growing body of literature shows phenological shifts of species due to climate change with different species shifting at different rates (Visser & Both, 2005; Donnelly et al., 2011; Kharouba et al., 2018). By linking development with temperature, our modelling approach predicts the expected phenological shift under various future climate scenarios for a species whose resting period is determined by temperature. More specifically, the change in phenology is determined by the interaction of the new temperature pattern and the slope of the rate accumulation function (R'); see Eq. 8 and Supplementary Material. For example, a warm spell occurring when the slope of development rate is very low may have little effect compared to the same warm spell occurring when development is more sensitive to temperature (i.e., when R' is high); see Fig. 2B. Consecutive spells will have additive effects: several warm spells will advance phenology several times, while a cold spell may cancel effects of a warm spell.

For a consumer to efficiently exploit a resource, some form of temporal synchrony is often required. Observations of consumer-resource systems show that, as an effect of climate change, their degree of synchrony can increase or decrease, but the mechanisms behind this difference are unclear (Kharouba et al., 2018). We focused on the duration of the seasonal resting period, which constitutes a significant stage in many species' life cycle. As different species react differently to temperature changes, we applied our model to each species separately to predict whether their degree of synchrony will increase or decrease as temperature patterns change.

A phenological shift of the consumer and its resource may affect their population dynamics and subsequently the larger food web in which they are embedded. For example, an

increase of the phenological mismatch may limit the consumer's food intake, which affects its biomass, its life cycle, and potentially even its ability to persist in the considered geographic zone (Simmonds et al., 2020). If a subsidiary resource is available, the consumer may switch resources (Stålhandske et al., 2016). A decrease of the phenological mismatch may lead to a stronger depletion of the resource. Both of these outcomes may cause ripple effects through the food web. Thus, a shift in relative phenology between the consumer and its resource can affect species demography in many ways (Miller-Rushing et al., 2010), as well as their geographic distribution. Our work, which is based on heat accumulation, improves our ability to predict the direction and the magnitude of the change in phenological synchrony, a research challenge identified by earlier studies (Kharouba et al. 2018).

4.2 The spruce budworm - balsam fir system and its general implications

We apply our general theory to predict the future mismatch between spruce budworm and balsam fir across latitude under different warming scenarios. According to the general theory, a warm spell occurring at any time during the heat accumulation period is likely to affect the insect because its $R'(\cdot)$ has a wide profile (see Fig. 2B, and section 2.2.2). The tree will react only if the warm spell occurs within its narrow sensitivity period. Therefore, under a change in temperature regime that would lead to several warm spells, the insect will advance its phenology more than the tree. Simulations with realistic temperature time series (showing several warm spells) confirm the theoretical results. Under an increase of temperatures, insect phenology would advance by 15 days on average, while the tree phenology would advance by 7 days on average. The expected overall mismatch increases by approximately 7 days on average (see Fig. 6 and section 3.4.2).

Budworms have to emerge within two weeks of budburst to have access to a rich source of nutrients. A shift in phenology increasing the likelihood of this pattern would in turn increase larval survival and population persistence in previously non-favourable regions. Under historical temperature regimes, budworm emergence date in northern sites is close to budburst date, which causes the budworm to emerge too late in some years, while emergence in southern sites occurs within 15 days before budburst. An increase of the mismatch benefits the insect in northern sites since it would emerge before budburst most of the time, while in southern sites, the budworm would emerge too early some years. A shift in phenology that leads to emergence occurring too early (or too late) would cause severe budworm mortality and prevent its persistence in some parts of its current distribution. Conversely, an increase in the persistence of the insect in other regions will increase damages to host trees. Since insect pests (especially SBW) are one of the main drivers of forest composition in the black spruce, white spruce and balsam fir zones (Harper et al., 2003), a change in phenological synchrony between these species may affect their future geographic distributions.

A change in mismatch between budworm and balsam fir would also affect other potential host species. It is known that black spruce is a suitable host, but its phenological mismatch with budworm protects the tree from a long-term budworm establishment (Pureswaran, et al., 2015). Moreover, in northern sites, black spruce seems to be somehow protected because the budworm cannot establish for a long period due to a high frequency of cold years (Pureswaran, et al., 2015). However, our model predicts that warming should lead to a better synchrony between the insect and balsam fir in these sites, therefore the budworm might be able to establish and generate outbreaks, which in turn would affect black spruce populations. To test this hypothesis, we

444 require *in situ* phenological data for a parameterization of the tree model (Eq. 2) for black
445 spruce.

446 Our model assumes that neither the consumer nor the resource affects the phenology of
447 the other species. Only temperature drives phenology. Recent observations of interactions
448 between defoliation and host phenology suggest that severe defoliation by budworm may
449 advance budburst phenology (Deslauriers et al., 2019). According to our model, an increase in
450 temperature should increase the mismatch by advancing the budworm phenology more than that
451 of the tree. If defoliation leads to an advance in phenology for the tree, then the resulting
452 mismatch would stay close to the historical mismatch, which would reduce the benefit of this
453 physiological response for the tree.

454 **4.3 Future extensions**

455 The present study provides novel insights towards a better understanding of the effects of
456 an increase of temperatures on species phenology. Further studies may consider specific features
457 that we do not include in the model. First, our study is limited to cases where the phenology is
458 directly constrained by environmental variables (more specifically temperature). Organisms can
459 also use temperature (or any other environmental variables) indirectly as cues for predicting an
460 optimal time window to have access to resources or convenient environmental conditions (e.g.,
461 photoperiod, rainfall) (McNamara et al., 2011).

462 Second, we measure mismatch as the time lag between peaks of consumer demand and
463 resource availability. Some authors have argued that more precise measurements should take into
464 account the shape and location of the phenological distributions around these peaks (e.g., Lindén,
465 2018) while others showed that, in some cases, time lag between phenological peak dates was a

better predictor of resource availability than the overlap between phenological distributions (Ramakers et al., 2020).

Third, a change in phenological synchrony will have an effect on the whole life cycle, which would be beyond the scope of the present study. Rising temperatures are likely to affect physiological processes at later life stages, by affecting metabolism for example (Brown, et al., 2004). It is also clear that a change in the mismatch is likely to affect the ecological and evolutionary dynamics of the consumer-resource system. These aspects provide avenue for further studies.

The theoretical framework presented here can be applied to other consumer-resource systems. The concrete example of the spruce budworm - balsam fir system informs the required methodology. To estimate the mismatch, the two rate accumulation functions must be calibrated accurately. It appears that the model is sensitive to the temporal resolution of the temperature time series. There are two timescales at which temperatures affect the outcome of the model. The first one occurs at the daily scale. Average daily temperatures lead to different results than a four-hour time interval for temperatures (used throughout the present study) since variations within a day may allow for development during a few hours, even when the average temperature over the whole day may not.

The second one occurs at a broader timescale (e.g., weekly). For the study system, temperature data modelled by BioSim (Régnière, et al., 2014) lead to realistic emergence and budburst dates when compared to real data (see section 3.2) and to a phenological mismatch, i.e., budworm usually emerging before budburst, similar to field observations (Pureswaran, et al., 2019). More simplistic distributions that do not contain any warm spell can lead to the opposite

488 pattern (i.e., budburst occurs first). Realistic temperatures show warm periods during days with
489 low average temperatures (i.e., late winter - early spring), which advance the insect phenology
490 compared to simplified (cosine) distributions. The tree is less sensitive to these short warm
491 events.

492 The modelling approach presented here allows for investigation of potential effects of
493 climate change on consumer-resource systems. Synchrony / mismatch between a consumer and
494 its resource is fundamental to predict future species distribution. Future studies may go further by
495 including more complex mechanistic approaches (e.g., energy budget models) in order to model
496 the whole life cycle of the consumer, which could lead to new insights on the dynamics of the
497 systems. Our model could be linked to models that explore the consequences of phenological
498 mismatch on the population dynamics of consumer – resource systems (e.g., Bewick et al., 2016).

499

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Authors' contribution

All authors conceptualized the study. FL and SP built the theoretical model. SP calibrated the tree model, gathered the temperature data, and performed the simulations. SP, JNC and FL analyzed the results, wrote and edited the manuscript.

Data Availability

Code will be stored on Zenodo public depository. Data on balsam fir phenology can be made available on demand to Dr. J. Régnière (Natural Resources Canada).

511 **References**

512 Amarasekare, P., & Coutinho, R. M. (2014). Effects of temperature on intraspecific competition
513 in ectotherms. *The American Naturalist*, 184(3), E50-65. doi: 10.1086/677386

514 Bale, J. S., Masters, G. J., Hodkinson, I. D., Awmack, C., Bezemer, T. M., Brown, V. K.,
515 Butterfield, J., Buse, A., Coulson, J. C., Farrar, J., Good, J. E. G., Harrington, R., Hartley,
516 S., Jones, T. H., Lindroth, R. L., Press, M. C., Symrnioudis, I., Watt, A. D., & Whittaker, J.
517 B. (2002). Herbivory in global climate change research: direct effects of rising temperature
518 on insect herbivores. *Global Change Biology*, 8(1), 1–16. doi: 10.1046/j.1365-
519 2486.2002.00451.x

520 Bean, J. L. (1961). Predicting emergence of second-instar spruce budworm larvae from
521 hibernation under field conditions in minnesota. *Annals of the Entomological Society of*
522 *America*, 54(2), 175–177. doi: 10.1093/aesa/54.2.175

523 Bewick, S., Cantrell, R. S., Cosner, C., & Fagan, W. F. (2016). How resource phenology affects
524 consumer population dynamics. *The American Naturalist*, 187(2), 151–166.

525 Blais, J. R. (1957). Some relationships of the spruce budworm, *Choristoneura fumiferana* (Clem.)
526 to black spruce, *Picea mariana* (Moench) Voss. *The Forestry Chronicle*, 33(4), 364–372.
527 doi: 10.5558/tfc33364-4

528 Both, C., van Asch, M., Bijlsma, R. G., Van Den Burg, A. B., & Visser, M. E. (2009). Climate
529 change and unequal phenological changes across four trophic levels: constraints or
530 adaptations? *Journal of Animal Ecology*, 78(1), 73–83. doi: 10.1111/j.1365-
531 2656.2008.01458.x

- Both, C., & Visser, M. E. (2001). Adjustment to climate change is constrained by arrival date in a long-distance migrant bird. *Nature*, 411(6835), 296–298. doi: 10.1038/35077063
- Brown, J. H., Gillooly, J. F., Allen, A. P., Savage, V. M., & West, G. B. (2004). Toward A Metabolic Theory Of Ecology. *Ecology*, 85(7), 1771–1789. doi: 10.1890/03-9000
- Chaine, I. (2000). A united model for budburst of trees. *Journal of Theoretical Biology*, 207, 337–347. doi: 10.1006/jtbi.2000.2178
- Chaine, I., & Régnière, J. (2017). Process-based models of phenology for plants and animals. *Annual Review of Ecology, Evolution, and Systematics*, 48(1), 159–182. doi: 10.1146/annurev-ecolsys-110316-022706
- Cobbold, C. A., & Powell, J. A. (2011). Evolution stabilises the synchronising dynamics of poikilotherm life cycles. *Bulletin of Mathematical Biology*, 73(5), 1052–1081. doi: 10.1007/s11538-010-9552-1
- Cohen, J. M., Lajeunesse, M. J., & Rohr, J. R. (2018). A global synthesis of animal phenological responses to climate change. *Nature Climate Change*, 8(3), 224–228. doi: 10.1038/s41558-018-0067-3
- Colombo, S. J. (1998). Climatic warming and its effect on bud burst and risk of frost damage to white spruce in Canada. *The Forestry Chronicle*, 74(4), 567–577. doi: 10.5558/tfc74567-4
- Cushing, D. H. (1990). Plankton production and year-class strength in fish populations: An update of the match/mismatch hypothesis. *Advances in Marine Biology*, 26(C), 249–293. doi: 10.1016/S0065-2881(08)60202-3
- Desbiens, M. (2007). *Relation phénologique entre le débourrement des bourgeons chez le sapin*

- 553 *baumier et l'émergence des larves de deuxième stade de la tordeuse des bourgeons de*
554 *l'épinette, Choristoneura fumiferana (Lepidoptera: Tortricidae)*. ProQuest.
- 555 Deslauriers, A., Fournier, M.-P., Carteni, F., & Mackay, J. (2019). Phenological shifts in conifer
556 species stressed by spruce budworm defoliation. *Tree Physiology*, 39(4), 590–605. doi:
557 10.1093/treephys/tpy135
- 558 Donnelly, A., Caffarra, A., & O'Neill, B. F. (2011). A review of climate-driven mismatches
559 between interdependent phenophases in terrestrial and aquatic ecosystems. *International*
560 *Journal of Biometeorology*, 55(6), 805–817. doi: 10.1007/s00484-011-0426-5
- 561 Dorais, L., & Kettela, E. G. (1982). A review of entomological survey and assessment techniques
562 used in regional spruce budworm. *Choristoneura Fumiferana*.
- 563 Fleming, R. A. (2000). Climate change and insect disturbance regimes in Canada's boreal forests.
564 *World Resource Review*, 12(3), 521–548.
- 565 Forrest, J., & Thomson, J. D. (2012). Pollinator experience, neophobia and the evolution of
566 flowering time. *Proceedings of the Royal Society B: Biological Sciences*, 276(1658), 935–
567 943. doi: 10.1098/rspb.2008.1434
- 568 Fuentealba, A., Pureswaran, D., Bause, É., & Despland, E. (2017). How does synchrony with
569 host plant affect the performance of an outbreaking insect defoliator? *Oecologia*, 184(4),
570 847–857. doi: 10.1007/S00442-017-3914-4
- 571 Gienapp, P., & Visser, M. E. (2006). Possible fitness consequences of experimentally advanced
572 laying dates in Great Tits: differences between populations in different habitats. *Functional*
573 *Ecology*, 20(1), 180–185. doi: 10.1111/j.1365-2435.2006.01079.x

- 574 Harper, K., Boudreault, C., DeGrandpré, L., Drapeau, P., Gauthier, S., & Bergeron, Y. (2003).
575 Structure, composition, and diversity of old-growth black spruce boreal forest of the Clay
576 Belt region in Quebec and Ontario. *Environmental Reviews*, 11(1 SUPPL.), 2003. doi:
577 10.1139/a03-013
- 578 Iwasa, Y., Odendaal, F. J., Murphy, D. D., Ehrlich, P. R., & Launer, A. E. (1983). Emergence
579 patterns in male butterflies: A hypothesis and a test. *Theoretical Population Biology*, 23(3),
580 363–379. doi: 10.1016/0040-5809(83)90024-2
- 581 Kharouba, H. M., Ehrlén, J., Gelman, A., Bolmgren, K., Allen, J. M., Travers, S. E., &
582 Wolkovich, E. M. (2018). Global shifts in the phenological synchrony of species
583 interactions over recent decades. *Proceedings of the National Academy of Sciences*, 115(20),
584 5211–5216. doi: 10.1073/pnas.1714511115
- 585 Kharouba, H. M., & Wolkovich, E. M. (2020). Disconnects between ecological theory and data
586 in phenological mismatch research. *Nature Climate Change*, 10(5), 406–415. doi:
587 10.1038/s41558-020-0752-x
- 588 Lawrence, R. K., Mattson, W. J., & Haack, R. A. (1997). White spruce and the spruce budworm:
589 Defining the phenological window of susceptibility. *Canadian Entomologist*, 129(2), 291–
590 318. doi: 10.4039/Ent129291-2
- 591 Lindén, A. (2018). Adaptive and nonadaptive changes in phenological synchrony. *Proceedings of*
592 *the National Academy of Sciences of the United States of America*, 115(20), 5057–5059. doi:
593 10.1073/pnas.1805698115
- 594 Mattson, W., & Scriber, M. (1987). Feeding ecology of insect folivores of woody plants:
595 nitrogen, water, fiber, and mineral considerations. In F. Slansky & J. Rodriguez (Eds.), *The*

- 596 *Nutritional Ecology of Insects, Mites, and Spiders* (pp. 105–146). John Wiley & Sons, New
597 York.
- 598 McNamara, J. M., Barta, Z., Klaassen, M., & Bauer, S. (2011). Cues and the optimal timing of
599 activities under environmental changes. *Ecology Letters*, *14*(12), 1183–1190. doi:
600 10.1111/j.1461-0248.2011.01686.x
- 601 Miller-Rushing, A. J., Høye, T. T., Inouye, D. W., & Post, E. (2010). The effects of phenological
602 mismatches on demography. *Philosophical Transactions of the Royal Society B: Biological*
603 *Sciences*, *365*(1555), 3177–3186. doi: 10.1098/rstb.2010.0148
- 604 Osawa, A., Shoemaker, C. A., & Stedinger, J. R. (1983). A stochastic model of balsam fir bud
605 phenology utilizing maximum likelihood parameter estimation (*Abies balsamea*, Quebec).
606 *Forest Science*, *29*(3), 478–490. doi: 10.1093/forestscience/29.3.478
- 607 Pachauri, R. K., Allen, M. R., Barros, V. R., Broome, J., Cramer, W., Christ, R., Church, J. A.,
608 Clarke, L., Dahe, Q., & Dasgupta, P. (2014). *Climate change 2014: synthesis report.*
609 *Contribution of Working Groups I, II and III to the fifth assessment report of the*
610 *Intergovernmental Panel on Climate Change*. Ipcc.
- 611 Parmesan, C. (2006). Ecological and evolutionary responses to recent climate change. *Annual*
612 *Review of Ecology, Evolution, and Systematics*, *37*(1), 637–669. doi:
613 10.1146/annurev.ecolsys.37.091305.110100
- 614 Parmesan, C., & Yohe, G. (2003). A globally coherent fingerprint of climate change impacts
615 across natural systems. *Nature*, *421*(6918), 37–42. doi: 10.1038/nature01286
- 616 Post, E., Forchhammer, M. C., Stenseth, N. C., & Callaghan, T. V. (2001). The timing of life-
617 history events in a changing climate. *Proceedings of the Royal Society B: Biological*

618 *Sciences*, 268(1462), 15–23. doi: 10.1098/rspb.2000.1324

619 Pureswaran, D. S., De Grandpré, L., Paré, D., Taylor, A., Barrette, M., Morin, H., Régnière, J.,
620 Kneeshaw, D. D. (2015). Climate-induced changes in host tree–insect phenology may drive
621 ecological state-shift in boreal forests. *Ecology*, 96(6), 1480–1491. doi: 10.1890/13-2366.1

622 Pureswaran, D. S., Neau, M., Marchand, M., De Grandpré, L., & Kneeshaw, D. (2019).
623 Phenological synchrony between eastern spruce budworm and its host trees increases with
624 warmer temperatures in the boreal forest. *Ecology and Evolution*, 9(1), 576–586. doi:
625 10.1002/ece3.4779

626 Ramakers, J. J. C., Gienapp, P., & Visser, M. E. (2020). Comparing two measures of
627 phenological synchrony in a predator–prey interaction: Simpler works better. *Journal of*
628 *Animal Ecology*, 89(3), 745–756. doi: 10.1111/1365-2656.13143

629 Rebaudo, F., & Rabhi, V.-B. (2018). Modeling temperature-dependent development rate and
630 phenology in insects: review of major developments, challenges, and future directions.
631 *Entomologia Experimentalis et Applicata*, 166(8), 607–617. doi: 10.1111/eea.12693

632 Régnière, J., & Nealis, V. G. (2008). The fine-scale population dynamics of spruce budworm:
633 survival of early instars related to forest condition. *Ecological Entomology*, 33(3), 362–373.
634 doi: 10.1111/j.1365-2311.2007.00977.x

635 Régnière, J., & Nealis, V. G. (2018). Two sides of a coin: host–insect synchrony fitness trade-offs
636 in the population dynamics of the western spruce budworm. *Insect Science*, 25, 117–126.
637 doi: 10.1111/1744-7917.12407

638 Régnière, J., Saint-Amant, R., Béchard, A., & Moutaoufik, A. (2014). *BioSIM 10: User's*
639 *manual*, A. Nat. Resour. Can., Can. For. Serv. Laurentian Forestry Centre, Québec

(Quebec). Inf. Rep. LAU-X-137E.

Régnière, J., St-Amant, R., & Duval, P. (2012). Predicting insect distributions under climate change from physiological responses: spruce budworm as an example. *Biological Invasions*, 14(8), 1571–1586. doi: 10.1007/s10530-010-9918-1

Renner, S. S., & Zohner, C. M. (2018). Climate change and phenological mismatch in trophic interactions among plants, insects, and vertebrates. *Annual Review of Ecology, Evolution, and Systematics*, 49(1), 165–182. doi: 10.1146/annurev-ecolsys-110617-062535

Samplonius, J. M., Atkinson, A., Hassall, C., Keogan, K., Thackeray, S. J., Assmann, J. J., Burgess, M. D., Johansson, J., Macphie, K. H., Pearce-Higgins, J. W., Simmonds, E. G., Varpe, Ø., Weir, J. C., Childs, D. Z., Cole, E. F., Daunt, F., Hart, T., Lewis, O. T., Pettorelli, N., Sheldon, B. C., & Phillimore, A. B. (2021). Strengthening the evidence base for temperature-mediated phenological asynchrony and its impacts. *Nature Ecology and Evolution*, 5(2), 155–164. doi: 10.1038/s41559-020-01357-0

Simmonds, E. G., Cole, E. F., Sheldon, B. C., & Coulson, T. (2020). Phenological asynchrony: a ticking time-bomb for seemingly stable populations? *Ecology Letters*, 23(12), 1766–1775. doi: 10.1111/ele.13603

Singer, M. C., & Parmesan, C. (2010). Phenological asynchrony between herbivorous insects and their hosts: Signal of climate change or pre-existing adaptive strategy? *Philosophical Transactions of the Royal Society B: Biological Sciences*, 365(1555), 3161–3176. doi: 10.1098/rstb.2010.0144

Singer, M., & Parmesan, C. (2020). *Misunderstanding mismatch*. doi: 10.32942/osf.io/rxmct

Stålhandske, S., Olofsson, M., Gotthard, K., Ehrlén, J., Wiklund, C., & Leimar, O. (2016).

Phenological matching rather than genetic variation in host preference underlies geographical variation in host plants used by orange tip butterflies. *Biological Journal of the Linnean Society*, 119(4), 1060–1067. doi: 10.1111/bij.12838

Van Asch, M., & Visser, M. E. (2007). Phenology of forest caterpillars and their host trees: The importance of synchrony. *Annual Review of Entomology*, 52, 37–55. doi: 10.1146/annurev.ento.52.110405.091418

van Vuuren, D. P., Edmonds, J., Kainuma, M., Riahi, K., Thomson, A., Hibbard, K., Hurtt, G. C., Kram, T., Krey, V., Lamarque, J. F., Masui, T., Meinshausen, M., Nakicenovic, N., Smith, S. J., Rose, S. K. (2011). The representative concentration pathways: an overview. *Climatic Change*, 109(1), 5–31. doi: 10.1007/s10584-011-0148-z

Visser, M. E., & Holleman, L. J. M. (2001). Warmer springs disrupt the synchrony of oak and winter moth phenology. *Proceedings of the Royal Society B: Biological Sciences*, 268(1464), 289–294. doi: 10.1098/rspb.2000.1363

Visser, M. E., & Both, C. (2005). Shifts in phenology due to global climate change: the need for a yardstick. *Proceedings of the Royal Society B: Biological Sciences*, 272(1581), 2561–2569. doi: 10.1098/rspb.2005.3356

Wu, J., Dhingra, R., Gambhir, M., & Remais, J. V. (2013). Sensitivity analysis of infectious disease models: methods, advances and their application. *Journal of The Royal Society Interface*, 10(86), 20121018. doi: 10.1098/rsif.2012.1018

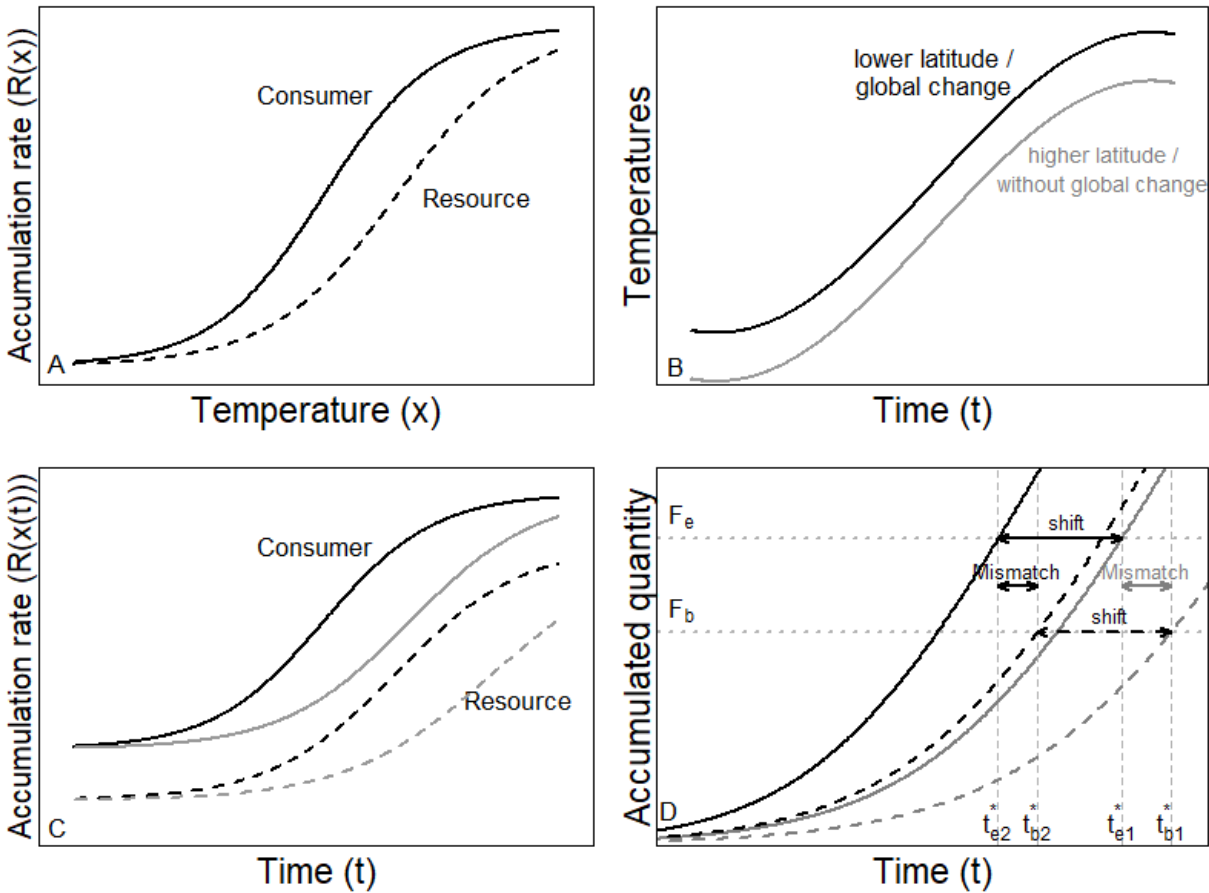
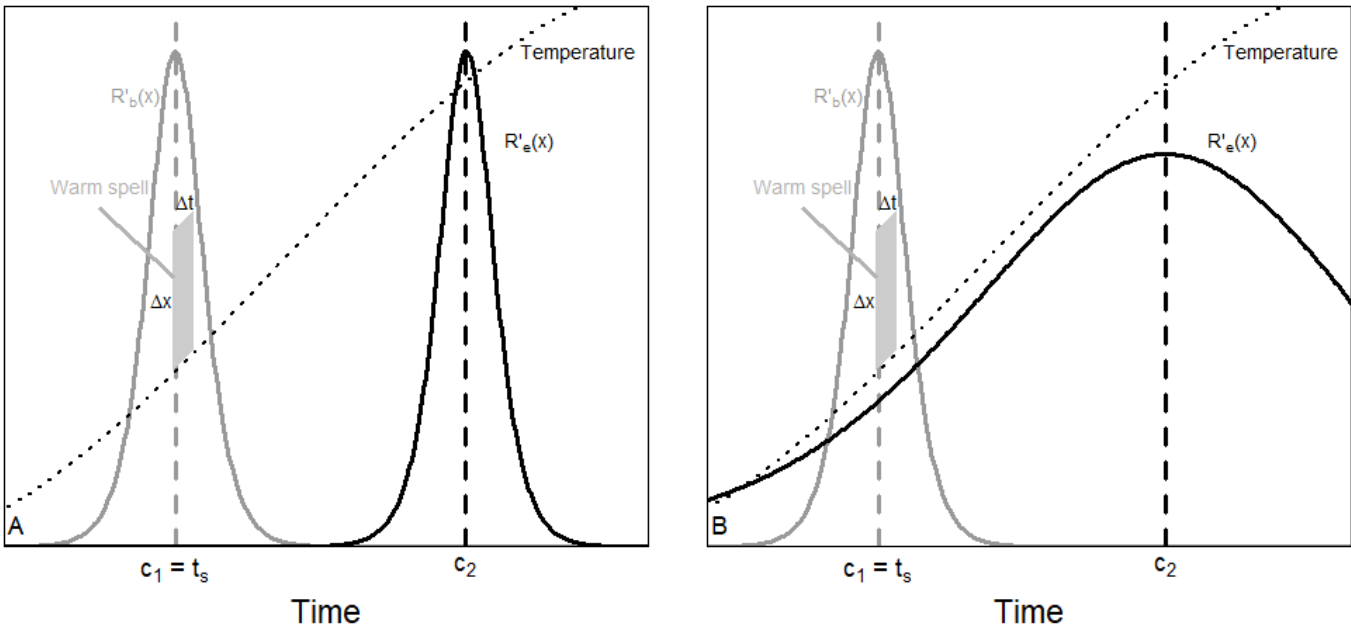


Figure 1: Illustration of theoretical development. (A) The rate accumulation function for two different species (solid line is consumer and dashed line is resource). (B) Two simplified temperature time series (black line is warmer and grey line is cooler). (C) Four combinations of rate accumulation; each species with two different temperature time series. (D) The resulting end of the resting phase for the consumer in cooler (grey solid line) and warmer (black solid line) temperatures and for the resource in cooler temperatures (black dashed line). The difference within species (grey vs. black line) indicates the shift in emergence in space (due to latitude or altitude) or time (due to global change). The difference between species (solid vs. dashed line) indicates the mismatch in the end of the seasonal resting phase for a fixed temperature regime

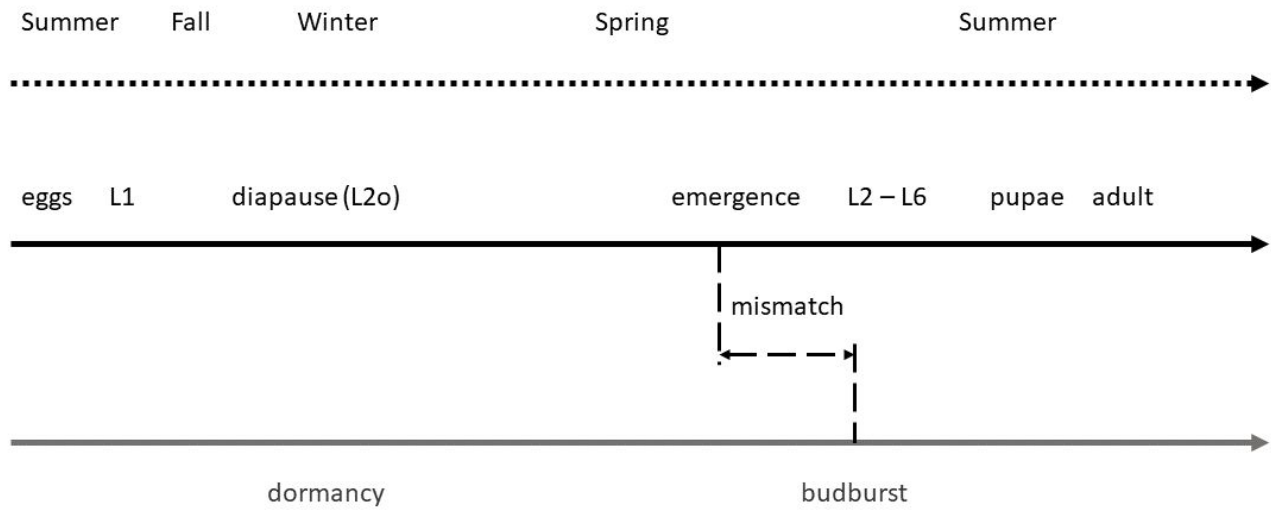
692 (same location and same time). Accumulation functions for both the consumer and the resource
693 use Eq. 2 with the following parameter values: $b_c = -1.32$, $c_c = 7.14$, $b_r = -1.16$, $c_r = 8.14$.



694

695 **Figure 2:** Sensitivity of resource accumulation functions, $R'(x(t))$, for two species when there is
696 no (A) or some (B) overlap. In (A), the warm spell indicated will advance the phenology of
697 species 1 (grey curve) but not alter that of species 2 (black curve). In (B), the same warm spell
698 will advance both phenologies. For this illustration, we used the accumulation function in Eq 2
699 for species 1 with parameter values for balsam fir (see section 3.4.1). We used the same function
700 for species 2 in (A), but with $c = 15.2$. In (B), we used the accumulation function for spruce
701 budworm in Eq. 8 with corresponding parameter values (see section 3.2.1). The temperature is
702 chosen according to Eq. 5 (from January to June).

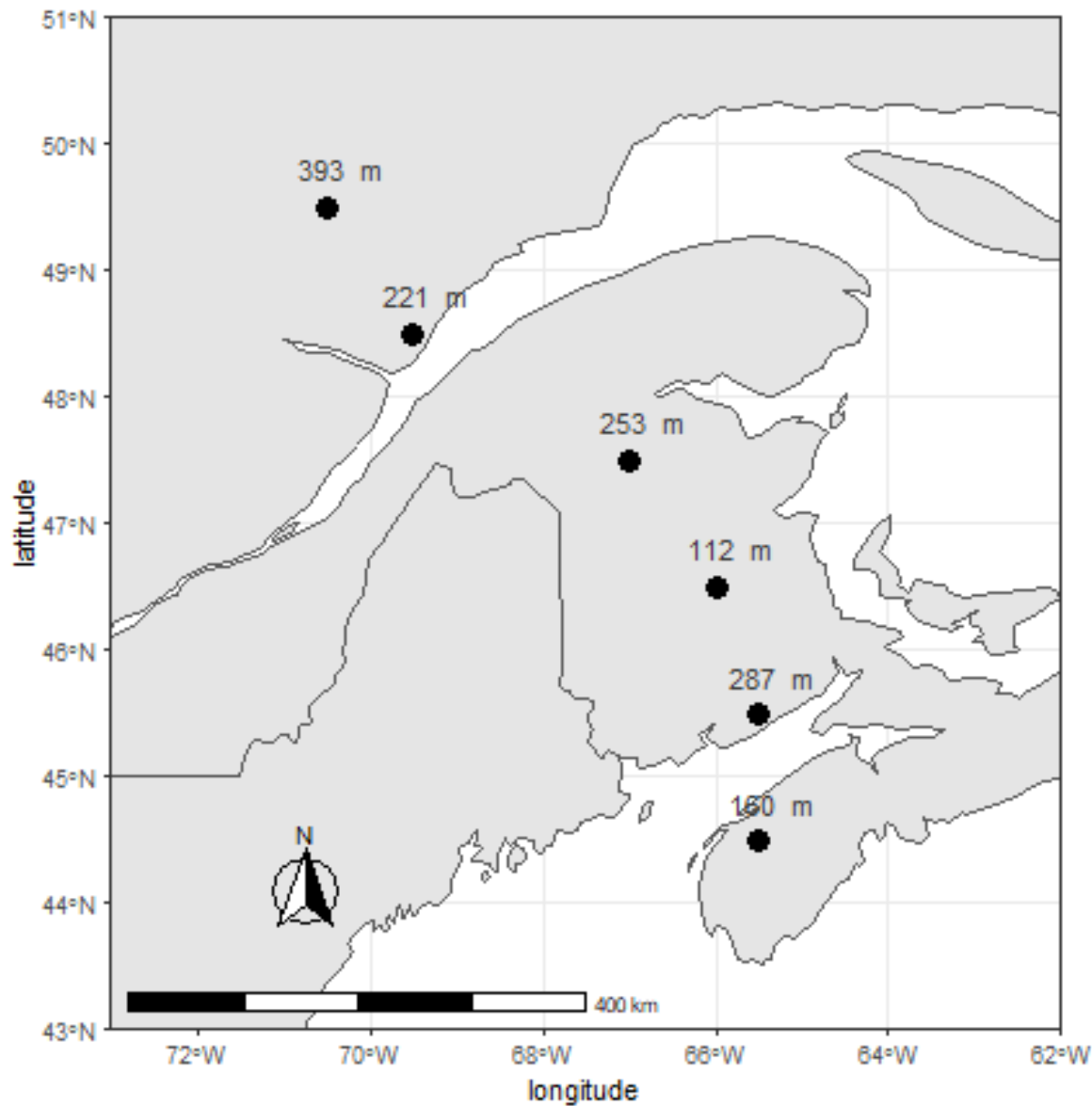
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704

705 Figure 3: life cycles of balsam fir and spruce budworm. Black line represents SBW life cycle, and grey

706 line is balsam fir life cycle.



707

708 **Figure 4:** Location of the sample sites where temperature data were collected for past and future
709 trends. Points are located across a gradient of latitude in Nova-Scotia, New Brunswick, and
710 Quebec. Altitude is indicated above each point.

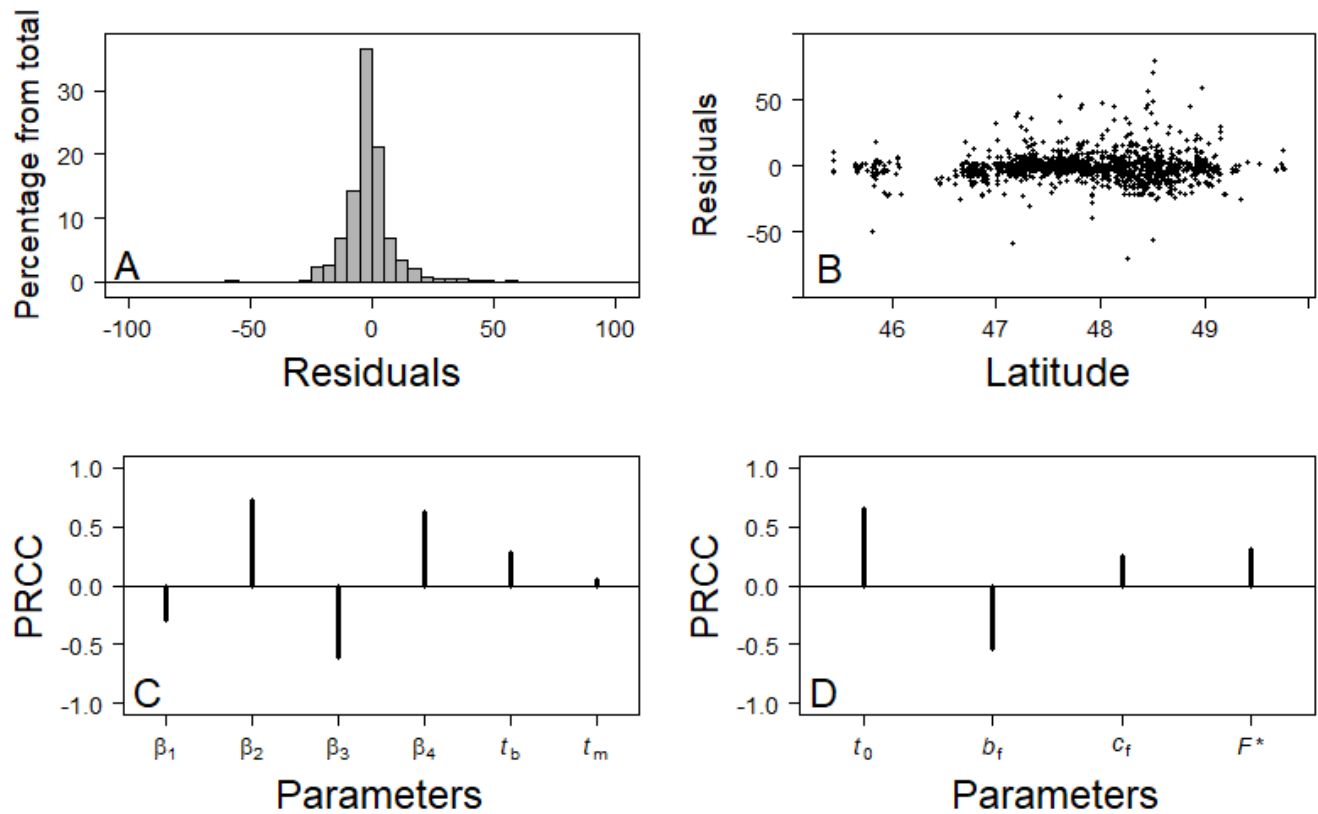
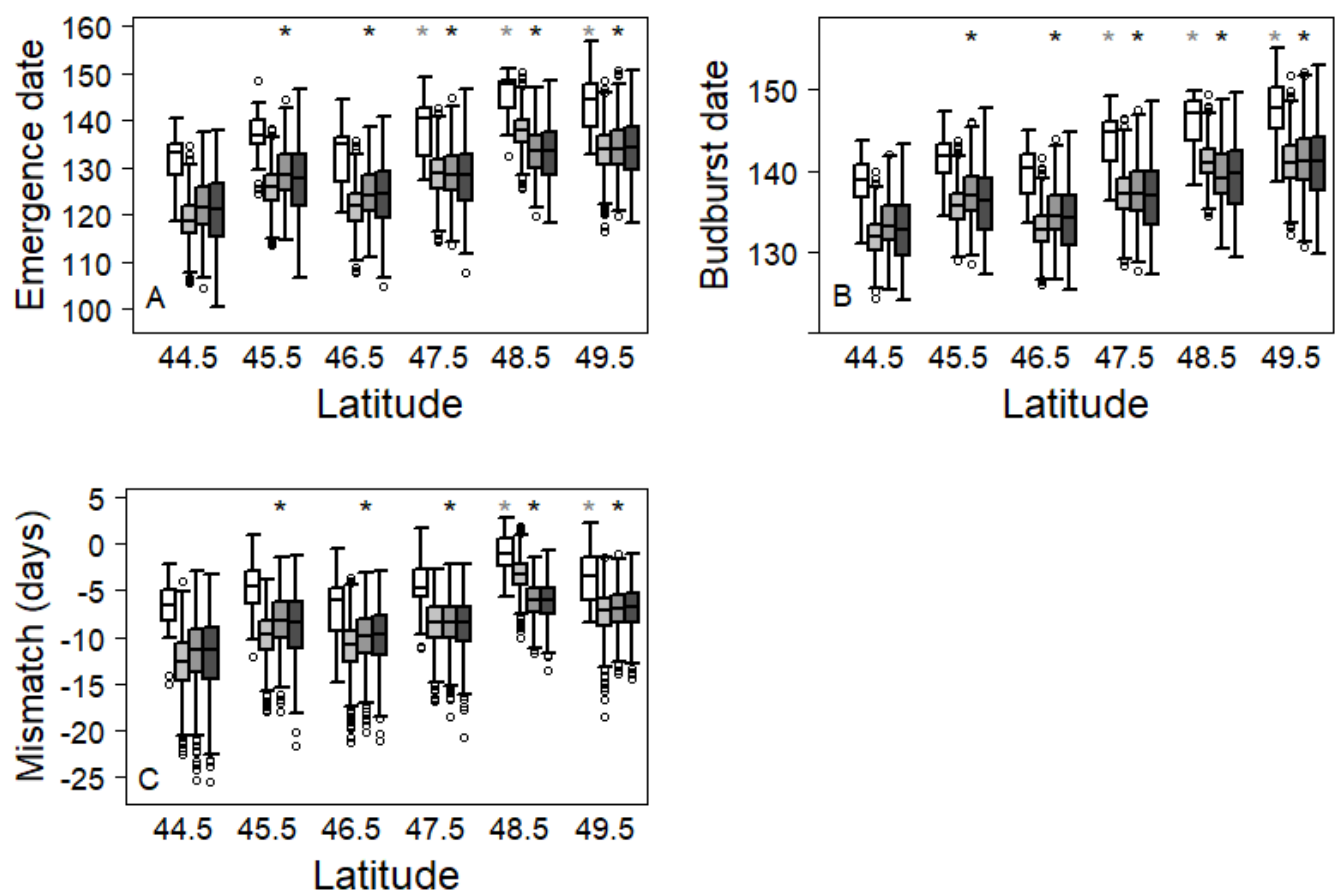


Figure 5: Fitting residuals of the balsam fir model and sensitivity analysis. (A) Residuals follow a Normal distribution centered on 0. (B) No obvious latitudinal patterns can be found on the residuals within the range of latitudes that is used throughout the rest of the study. (C) Partial Rank Correlation Coefficient (PRCC) shows that the SBW model is sensitive to most parameters especially β_2 , β_4 and x_b that delay emergence, and β_3 that advances phenology. (D) The tree model is mostly sensitive to b_f that hastens budburst, and t_0 that delays budburst.



718

719 **Figure 6:** Latitudinal distribution of (A) median emergence date of SBW (Julian days), (B)
720 median budburst date, and (C) mismatch between emergence and budburst date. For each
721 latitude, the white box (left one) represents the 1996-2016 period. Grey boxes represent expected
722 outcomes according to RCP 2.6 (light grey), RCP 4.5 (dark grey), and RCP 8.5 (black) scenarios
723 over 2021 to 2100. A black (grey) star means that the corresponding site shows a significant
724 difference with the most southern site for present (future) predicted trait. Both emergence and
725 budburst are expected to occur later at higher latitudes. Overall warming scenarios, both events
726 are expected to occur earlier in the year. Nowadays, emergence is expected to occur 5 to 10 days
727 before budburst at low latitudes, while at higher latitudes, emergence may sometimes occur
728 before budburst and sometimes after. For all warming scenarios, an increase in mismatch is

729 expected. At low latitudes, emergence may occur too early some years, which may lead to low
730 survival of SBW. At higher latitudes, emergence is expected to systematically occur a few days
731 before budburst, which would increase survival of SBW.

Supplementary information

A temperature-driven model of phenological mismatch provides insights into the potential impacts of climate change on consumer-resource interactions

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1 Theoretical developments

In this supplementary material, we give the details for the mathematical derivation of the two sensitivity formulas for the end time of the seasonal resting period of a species. The general equation that connects the start time t_0 , the rate curve $R(x)$ and the threshold F to the end time t^* of the resting period is

$$\int_{t_0}^{t^*} R(x(t))dt = F. \quad \text{Eq. S1}$$

General features

We want to determine how t^* changes when the temperature $x = x(t)$ changes by a small amount.

More formally, we will derive a formula for the linear approximation

$$t^*(\epsilon) = t^*(0) + \epsilon \frac{dt^*}{d\epsilon} \quad \text{Eq. S2}$$

where ϵ measures the magnitude of the small change, $t^*(0)$ is the end time when there is no change in the temperature time series from historical data, and the derivative is the sensitivity of the end time with respect to small changes.

We write the change in temperature as $x(t) + \epsilon z(t)$, where $z(t)$ is the pattern in which the temperature differs from the expectation and ϵ is small. Since the end time now depends on ϵ , we write $t^* = t^*(\epsilon)$. The sensitivity of the end time with respect to ϵ is given by the derivative

$$\frac{dt^*}{d\epsilon} \text{ for } \epsilon = 0. \quad \text{Eq. S3}$$

This expression will depend on the pattern of temperature difference, $z(t)$. We will discuss two specific patterns below.

When we substitute these expressions into the defining equation for t^* above, ϵ appears twice: once in the upper limit of integration and once in the integrand. To emphasize these two occurrences, we write the left-hand side of the equation as a function of two variables, namely

$$I(t^*(\epsilon), R(x + \epsilon x)) = \int_{t_0}^{t^*(\epsilon)} R(x(t) + \epsilon z(t)) dt. \quad \text{Eq. S4}$$

When we differentiate the equation that defines the end time, $I(t^*, R) = F$, with respect to ϵ , we use the chain rule repeatedly and obtain

$$\frac{d}{d\epsilon} I(t^*(\epsilon), R(x + \epsilon x)) = \frac{\partial I}{\partial t^*} \frac{dt^*}{d\epsilon} + \frac{\partial I}{\partial R} \frac{dR}{dx} \frac{dx}{d\epsilon} = 0. \quad \text{Eq. S5}$$

The derivative of the integral in Eq. S4 with respect to the end time is simply the integrand evaluated at the end time. The derivative of the integral with respect to the integrand is the integral itself since this is linear. The derivative of the rate function with respect to x is the usual derivative and the derivative of x with respect to ϵ is z , by our definition above. Then we can solve the above equation for the quantity we are looking for and find

$$\frac{dt^*}{d\epsilon} = \frac{-\int_{t_0}^{t^*} R'(x(t))z(t)dt}{R(x(t^*))}. \quad \text{Eq. S6}$$

Hence, the end time has the linear approximation

$$t^*(\epsilon) \approx t^*(0) + \epsilon \frac{dt^*}{d\epsilon} = t^*(0) + \epsilon \frac{-\int_{t_0}^{t^*} R'(x(t))z(t)dt}{R(x(t^*))}. \quad \text{Eq. S7}$$

As expected, the pattern by which the temperature deviates, $z(t)$, appears in this formula. We look at two interesting special cases for this pattern.

Specific patterns

The first case is that the temperature change is constant throughout the period, independent of time.

In that case, we can set $\epsilon z(t) = \Delta x$ to be the constant temperature difference. Then the function

$z(t)$ drops out of the above integral and the end time is given by

$$t^*(\epsilon) \approx t^*(0) - \Delta x \frac{\int_{t_0}^{t^*} R'(x(t)) dt}{R(x(t^*))}. \quad \text{Eq. S8}$$

Since $R'(x) > 0$ and $R(x) > 0$, the end time decreases if the temperature increases, i.e., the phenology advances. We knew this already from general consideration, but now we have an explicit expression for how much the advance is per degree increase.

The second case in which we can simplify the general formula is that there is a warm or cold spell of relatively short duration at a particular time during the resting phase. Then $\epsilon z(t) = \Delta x$ during the spell of duration Δt , starting at time t_s , and $z(t) = 0$ otherwise. The integral in the numerator of Eq. S6 can be approximated by

$$\epsilon \int_{t_0}^{t^*} R'(x(t)) z(t) dt = \Delta x \int_{t_s}^{t_s + \Delta t} R'(x(t)) dt \approx \Delta x \Delta t R'(x(t_s)) \quad \text{Eq. S9}$$

Hence, the expression for the end time is approximately

$$t^*(\epsilon) \approx t^*(0) - \Delta x \frac{\Delta t R'(x(t_s))}{R(x(t^*))} \quad \text{Eq. S10}$$

This means that the end time is most sensitive to a warm or cold spell when the derivative of the rate function is the highest, all other things being equal.

The two formulas (Eq. S8 and Eq. S10) may seem different, but they express the same idea. One has to integrate R' for all times where the two time series differ. When the two time series differ by a constant for all times, the one has to integrate over the entire time series. When the two

time series differ only on an interval of length Δt , then one has to integrate over only that interval. If the interval is short, then the value of $R(x(t))$ does not change much and therefore the integral is approximated by the product of the length of the interval (Δt) and the value of the integrand ($R(x(t_s))$).

Derivative of the rate function

$$R(x) = \frac{1}{1 + \exp(b(x - c))}, \quad \text{Eq. S11}$$

we can explicitly calculate the derivative as

$$R'(x) = \frac{-b \exp(b(x - c))}{(1 + \exp(b(x - c)))^2}, \quad \text{Eq. S12}$$

which is positive since b is negative. To find the maximum of the derivative, we differentiate again and find

$$R''(x) = \frac{-b^2 \exp(b(x - c))(1 - \exp(b(x - c)))}{(1 + \exp(b(x - c)))^3} \quad \text{Eq. S13}$$

The maximum of R occurs where $R = 0$, which happens when $x = c$ (see Fig. 2).

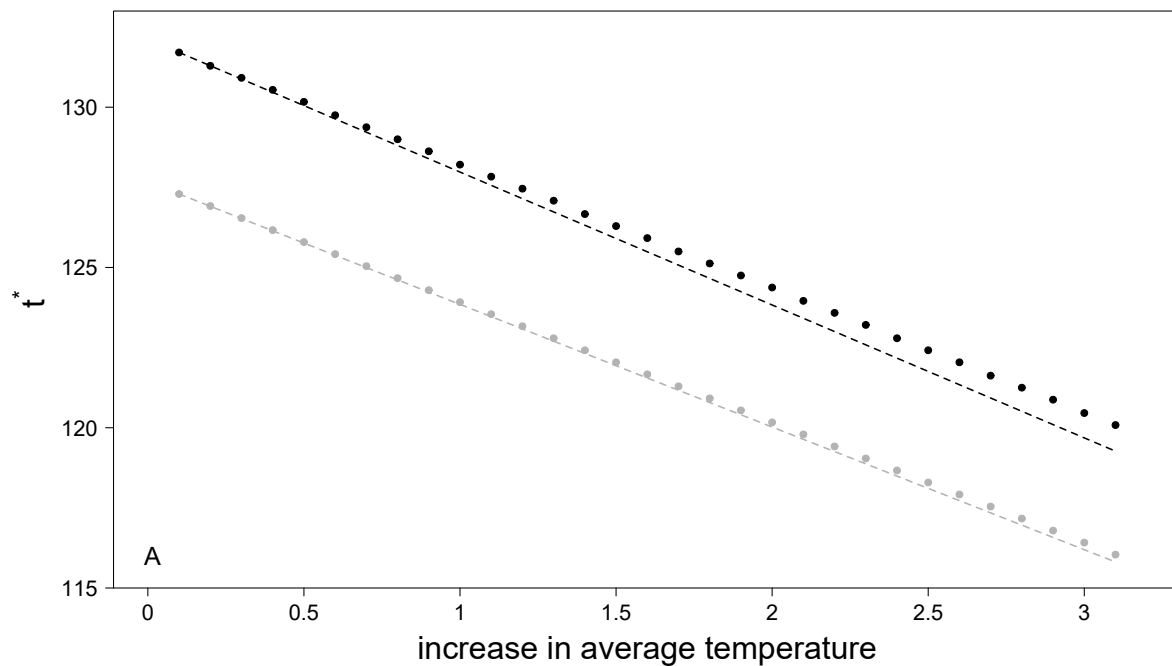


Figure S1: Effects of a constant temperature difference on species phenology. Black is the consumer (SBW), and grey is the resource (balsam fir). A constant temperature difference advances species phenology. Dotted is the predicted value (Eq. 3 used with the *R* functions of SBW and balsam fir), dashed is the linear approximation from the model with simple time series.

2 Analysis of variance

Here, we present the full results of the analysis of variance done on emergence date, budburst date and mismatch across latitude, for past/present temperatures, and for the three RCP scenarios. There are 6 sites ranged from site 1 (southern site: 44.5° N) to site 6 (northern site: 49.5° N) (see Fig. 4 and 6, and main text for details). The analysis was performed with R.

2.1 Historical data

Emergence date

```
##              Df Sum Sq Mean Sq F value    Pr(>F)
## Site              5    3374    674.8    17.89 3.2e-13 ***
## Residuals    120    4527     37.7
##
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
##
## Pairwise comparisons using t tests with pooled SD
##
##      Site1  Site2  Site3  Site4  Site5
## Site2 0.20440    -      -      -      -
## Site3 1.00000 0.64190    -      -      -
## Site4 0.00574 1.00000 0.02647    -      -
## Site5 2.0e-09 0.00021 2.0e-08 0.01493    -
## Site6 1.5e-08 0.00102 1.4e-07 0.05283 1.00000
##
## P value adjustment method: bonferroni
```

Budburst date

```
##              Df Sum Sq Mean Sq F value    Pr(>F)
## Site              5    1332    266.43    21.72 1.94e-15 ***
```

```
## Residuals    120    1472    12.26
```

```
##
```

```
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

```
##
```

```
## Pairwise comparisons using t tests with pooled SD
```

```
##
```

```
##           Site1    Site2    Site3    Site4    Site5
```

```
## Site2 0.21247    -        -        -        -
```

```
## Site3 1.00000 1.00000    -        -        -
```

```
## Site4 0.00062 1.00000 0.01185    -        -
```

```
## Site5 6.5e-09 0.00050 3.5e-07 0.18502    -
```

```
## Site6 3.1e-12 8.8e-07 2.2e-10 0.00155 1.00000
```

```
##
```

```
## P value adjustment method: bonferroni
```

Mismatch

```
##           Df Sum Sq Mean Sq F value    Pr(>F)
```

```
## Site           5   545.7   109.13    11.08 8.7e-09 ***
```

```
## Residuals    120  1182.1     9.85
```

```
##
```

```
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

```
##
```

```
## Pairwise comparisons using t tests with pooled SD

##

##      Site1  Site2  Site3  Site4  Site5
## Site2 0.5382    -      -      -      -
## Site3 1.0000  0.5175    -      -      -
## Site4 0.2684  1.0000  0.2572    -      -
## Site5 1.4e-07  0.0014  1.3e-07  0.0038    -
## Site6 0.0042  1.0000  0.0039  1.0000  0.2517
##

## P value adjustment method: bonferroni
```

2.2 RCP 2.6 data

Emergence date

```
##              Df Sum Sq Mean Sq F value Pr(>F)
## Site         5 305237   61047    3334 <2e-16 ***
## Residuals    7194 131707         18
##

## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

##

## Pairwise comparisons using t tests with pooled SD

##

##      Site1  Site2  Site3  Site4  Site5
```

```
## Site2 <2e-16      -      -      -      -
## Site3 <2e-16 <2e-16      -      -      -
## Site4 <2e-16 <2e-16 <2e-16      -      -
## Site5 <2e-16 <2e-16 <2e-16 <2e-16      -
## Site6 <2e-16 <2e-16 <2e-16 <2e-16 <2e-16
```

```
##
```

```
## P value adjustment method: bonferroni
```

Budburst date

```
##              Df Sum Sq Mean Sq F value Pr(>F)
```

```
## Site          5  94260    18852    2896 <2e-16 ***
```

```
## Residuals    7194   46827         7
```

```
##
```

```
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

```
##
```

```
## Pairwise comparisons using t tests with pooled SD
```

```
##
```

```
##      Site1 Site2 Site3 Site4 Site5
```

```
## Site2 <2e-16      -      -      -      -
```

```
## Site3 <2e-16 <2e-16      -      -      -
```

```
## Site4 <2e-16 <2e-16 <2e-16      -      -
```

```
## Site5 <2e-16 <2e-16 <2e-16 <2e-16      -
```

Site6 <2e-16 <2e-16 <2e-16 <2e-16 1

##

P value adjustment method: bonferroni

Mismatch

Df Sum Sq Mean Sq F value Pr(>F)

Site 5 66174 13235 2316 <2e-16 ***

Residuals 7194 41116 6

##

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

##

Pairwise comparisons using t tests with pooled SD

##

Site1 Site2 Site3 Site4 Site5

Site2 <2e-16 - - -

Site3 <2e-16 <2e-16 - -

Site4 <2e-16 <2e-16 <2e-16 -

Site5 <2e-16 <2e-16 <2e-16 <2e-16 -

Site6 <2e-16 <2e-16 <2e-16 <2e-16 <2e-16

##

P value adjustment method: bonferroni

2.3 RCP 4.5 data

Emergence date

```
##                Df Sum Sq Mean Sq F value Pr(>F)
## Site          5 139304    27861    1045 <2e-16 ***
## Residuals     7194 191768         27
##
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
##
## Pairwise comparisons using t tests with pooled SD
##
##      Site1  Site2  Site3  Site4  Site5
## Site2 <2e-16   -       -       -       -
## Site3 <2e-16 <2e-16   -       -       -
## Site4 <2e-16 1.0000 <2e-16   -       -
## Site5 <2e-16 <2e-16 <2e-16 <2e-16   -
## Site6 <2e-16 <2e-16 <2e-16 <2e-16 0.006
##
## P value adjustment method: bonferroni
```

Budburst date

```
##                Df Sum Sq Mean Sq F value Pr(>F)
## Site          5  51477    10295    1046 <2e-16 ***
```

```
## Residuals    7194    70813        10

##

## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

##

## Pairwise comparisons using t tests with pooled SD

##

##      Site1 Site2 Site3 Site4 Site5
## Site2 <2e-16  -      -      -      -
## Site3 <2e-16 <2e-16  -      -      -
## Site4 <2e-16 0.62   <2e-16  -      -
## Site5 <2e-16 <2e-16 <2e-16 <2e-16  -
## Site6 <2e-16 <2e-16 <2e-16 <2e-16 <2e-16

##

## P value adjustment method: bonferroni
```

Mismatch

```
##              Df Sum Sq Mean Sq F value Pr(>F)
## Site         5  23457    4691   651.4 <2e-16 ***
## Residuals    7194   51815         7

##

## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

##
```



```
## Pairwise comparisons using t tests with pooled SD
```

```
##
```

```
##      Site1 Site2 Site3 Site4 Site5
```

```
## Site2 <2e-16 - - -
```

```
## Site3 <2e-16 <2e-16 - -
```

```
## Site4 <2e-16 0.22 <2e-16 -
```

```
## Site5 <2e-16 <2e-16 <2e-16 <2e-16 -
```

```
## Site6 <2e-16 <2e-16 <2e-16 <2e-16 <2e-16
```

```
##
```

```
## P value adjustment method: bonferroni
```

2.4 RCP 8.5 data

Emergence date

```
##              Df Sum Sq Mean Sq F value Pr(>F)
```

```
## Site      5 155376   31075    727.7 <2e-16 ***
```

```
## Residuals 7194 307197      43
```

```
##
```

```
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

```
##
```

```
## Pairwise comparisons using t tests with pooled SD
```

```
##
```

```
##      Site1 Site2 Site3 Site4 Site5
```

```
## Site2 <2e-16      -      -      -      -
## Site3 <2e-16 <2e-16      -      -      -
## Site4 <2e-16 0.202  <2e-16      -      -
## Site5 <2e-16 <2e-16 <2e-16 <2e-16      -
## Site6 <2e-16 <2e-16 <2e-16 <2e-16 0.014
##
## P value adjustment method: bonferroni
```

Budburst date

```
##              Df Sum Sq Mean Sq F value Pr(>F)
## Site          5  58046    11609    690.5 <2e-16 ***
## Residuals    7194 120951         17
##
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
##
## Pairwise comparisons using t tests with pooled SD
##
##      Site1   Site2   Site3   Site4   Site5
## Site2 < 2e-16      -      -      -      -
## Site3 8.2e-16 < 2e-16      -      -      -
## Site4 < 2e-16 0.0097  < 2e-16      -      -
## Site5 < 2e-16 < 2e-16 < 2e-16 < 2e-16      -
```

```
## Site6 < 2e-16 < 2e-16 < 2e-16 < 2e-16 < 2e-16
```

```
##
```

```
## P value adjustment method: bonferroni
```

Mismatch

```
##           Df Sum Sq Mean Sq F value Pr(>F)
```

```
## Site      5  25363    5073    572.3 <2e-16 ***
```

```
## Residuals 7194  63763         9
```

```
##
```

```
## Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

```
##
```

```
## Pairwise comparisons using t tests with pooled SD
```

```
##
```

```
##      Site1  Site2  Site3  Site4  Site5
```

```
## Site2 < 2e-16  -      -      -      -
```

```
## Site3 < 2e-16 < 2e-16  -      -      -
```

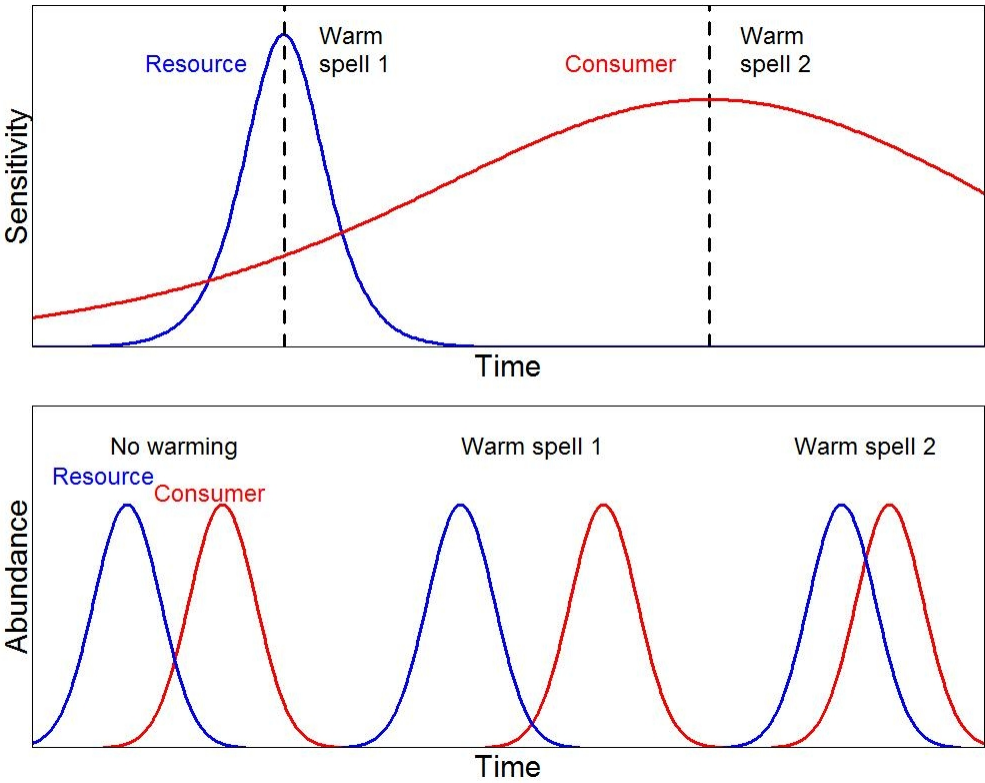
```
## Site4 < 2e-16 1      < 2e-16  -      -
```

```
## Site5 < 2e-16 < 2e-16 < 2e-16 < 2e-16  -
```

```
## Site6 < 2e-16 < 2e-16 < 2e-16 < 2e-16 1.1e-11
```

```
##
```

```
## P value adjustment method: bonferroni
```



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