**Potential impacts of climate change on the phenological synchrony between a consumer and its resource: the example of the spruce budworm and its tree host.**

**Abstract**

The ongoing anthropogenic climate change increasingly affects species phenology. Especially, the seasonal resting period, when organisms remain inactive during part of the year, is often driven by temperature. In the case of a consumer and a resource that both exhibit a seasonal resting period, the synchrony of the end of their respective resting period is fundamental for the persistence of their interaction. Since the consumer and its resource may react differently to a change in temperature regime, the synchrony between them will likely be altered. In this study, we propose a general theoretical model that determines the duration of the resting period according to temperature, and its effects on synchrony or mismatch between phenological stages of two interacting species. We then illustrate our approach using the spruce budworm – balsam fir system in Eastern Canada as a case study. We found that an increase in temperature would usually advance the end of the resting period. However, the effects of a warm or cold spell during the resting period would strongly vary according to the time and the duration of the spell. Depending on the way each species reacts to the same temperature shift, the mismatch between the consumer and its resource may increase or decrease. For the spruce budworm – balsam fir system, our model predicts that an increase in temperature may increase the mismatch between the insect and the tree in Southern sites, but may increase the synchrony in Northern sites. This kind of modelling approach is of prime importance to investigate potential effects of climate change on consumer – resource systems. The study of synchrony between interacting species is fundamental to predict future species distribution.

**Keywords:**

Consumer, resource, phenology, global warming, spruce budworm, balsam fir

**1. Introduction**

Anthropogenic climate change has increasingly disrupted ecological interactions for the past century (IPCC 2014 and references within). It is expected that this trend will continue and amplify as interacting species are likely to respond differently to similar environmental changes and selective pressures (Parmesan, 2006). Among ecological interactions, consumer-resource relationships are fundamental to the functioning of terrestrial and marine ecosystems. Trophic interactions between consumers and resources may be affected by climate change through: (1) direct changes in the life history traits (e.g., fecundity, mortality) of the consumer and/or the resource (Bale et al., 2002), (2) changes in the abundance of the consumer and/or the resource due to cascading effects from higher or lower trophic levels (e.g., changes in the consumer’s predators or competitors) (Both, van Asch, Bijlsma, Van Den Burg, & Visser, 2009), and (3) differential shifts in the phenology of the consumer and/or the resource leading to phenological mismatch (Kharouba et al., 2018).

Phenological mismatch between a consumer and a resource occurs when the timing of the species’ life cycle phases relevant to the interaction changes at different rates. The impact of a phenological mismatch on interacting species depends on the nature of the interaction and the direction of the phenological change (Renner & Zohner, 2018). In antagonistic interactions such as between a consumer and a resource, an increase in synchrony will likely favour the consumer while a decrease will favour the resource.

While there is a growing body of literature stating evidence for changes in species phenology, examples of phenological mismatch are still relatively scarce (Renner & Zohner, 2018), partly because they are expected to persist only for a short period of time over small spatial scales. Among studies, there is no clear pattern in the direction of the change: emergence dates of some interacting species are getting closer together, while others are getting further apart (Kharouba et al., 2018). The clearest cases of climate-driven mismatch have been reported at high latitudes (Reference ?) where many organisms rely on a seasonal resting period because temperatures affecting physiological processes tend to be below species optima for most of the year.

After a seasonal arrest in development that can last for several months, synchrony in springtime phenology, particularly between the emergence of phytophagous insects from diapause and the end of their host plants dormancy, is critical to the consumer’s fitness (Van Asch & Visser, 2007). Spring defoliators have evolved to exploit foliage at its annual optimal nutritional qualities, i.e., high concentration in nutrient and water and low concentrations in fibre and secondary metabolites (Hunter & Lechowicz, 1992)(Mattson & Scriber, 1987). This strategy requires a precise phenological match between the consumer and its resource because the quality of foliage declines quickly with foliar expansion. If the consumer emerges too early, it may encounter a long initial period with no food or poor food quality. If it emerges too late, the speed at which the nutritional quality of the foliage degrades may outpace the consumer’s capacity to adjust physiologically, thus negatively impacting its fitness.

    The seasonal resting phase of many organisms can be divided in two successive stages called endodormancy and ecodormancy in perennial woody plants, and diapause and quiescence in insects (Chuine & Régnière, 2017). Development and metabolism are generally inhibited by internal factors (e.g., depletion of energy reserves, hormones) in the first stage and by external factors (e.g., temperature, photoperiod) in the second one. Our study focuses on a system where both a consumer and its resource show a seasonal resting period during part of the year (e.g., during winter). For both, we assume that the first stage of the seasonal resting period is accomplished early in the winter before the return of favourable conditions, a common case in temperate and colder climates, and that temperature is the main driver that triggers the end of the second stage and therefore the resting period. In the remainder of this paper, the expression “resting period” will refer to the second stage of the seasonal resting period.

Observed patterns of change in phenological synchrony as a result of climate change have been difficult to explain mechanistically. The aim of this study is to investigate potential effects of global warming on the phenological synchrony between a consumer and its resource in the likely case where each species reacts independently of the other to temperatures. We begin with the general theoretical aspects that determine the duration of the resting period according to temperature, and its effects on synchrony / mismatch between phenological stages of the two species. Then we use a major insect pest of the Canadian boreal forest, the spruce budworm, and its main host, balsam fir, as a case study. We investigate the phenological mismatch across a gradient of latitudes, and the way this synchrony / mismatch is expected to vary with climate change in the future.

# 2. Methods

We begin with a unified description of the mechanism that determines the duration of the resting period of a consumer and its resource in terms of accumulation of ambient temperature. Then we list our data sources and explain the fitting methods for the spruce budworm - balsam fir system.

## **2.1 Theoretical development**

Throughout the resting period, an organism accumulates units of some quantity. The instantaneous rate of accumulation depends on the ambient temperature, and the resting period ends when a certain level of that quantity has been accumulated. For trees, this quantity can be heat, for example in degree-day models (Colombo, 1998) or more recent nonlinear models (Chuine, 2000)(Desbiens, 2007). For insects, the quantity can be the proportion of the corresponding life-cycle stage that they have completed (Cobbold & Powell, 2011)(Régnière, St-Amant, & Duval, 2012). Since the development rate is also temperature dependent, this quantity is ultimately also a measure of accumulated heat.

We denote time by *t* in days and temperature by *x = x(t)* in degrees Celsius. The instantaneous rate of accumulation is some nonnegative function of temperature, denoted by *R = R(x)*. For the range of temperatures that occur during the resting phase, *R(x)* is an increasing function. Indeed, while developmental rates typically decrease when temperatures exceed an upper threshold (Deutsch et al., 2008)(Amarasekare & Coutinho, 2014), such temperatures do not generally arise during the resting phase. The resting period begins at some time *t0* and ends at such time *t\** when the accumulated quantity reaches a certain threshold level *F*. As noted by Chuine and Régnière (2017), this concept of an accumulating quantity is “still the most important assumption in plant and animal phenology modelling”. The fundamental equation that connects all these quantities and determines the end of the resting period is

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|  |  | Eq. 1 |

A typical example for the accumulation rate function is the sigmoidal function

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|  |  | Eq. 2 |

where *b < 0* and *c* are two parameters to be estimated from data (Chuine, 2000)(Rebaudo & Rabhi, 2018). When the quantity of interest is the proportion of the life-cycle completed, it is natural to set the threshold level to be *F = 1*. If we divide Eq. 1 by *F* and include the term *1/F* into the function *R* in Eq. 2, we can standardize notation and compare different rate functions. We illustrate the rate function in Eq. 2 as well as the condition in Eq. 1 for two different species and two simplistic temperature time series in Figure 1. As temperature patterns during the resting period change because of global warming, the end time of the resting period of a species may shift. In particular, when temperatures increase, accumulation occurs faster and the phenology advances, i.e., the end time is earlier (Fig. 1D).

    Two species, such as a consumer and its resource, will likely have different forms of the rate accumulation function or the same form with different combinations of parameters (e.g., *b*, *c*, and *F*) even in the same temperature regime (compare solid and dashed curves in Fig. 1C), which typically leads to different end times of the resting period (bottom right panel). We denote these by *te\** (emergence time) for the insect (consumer) and by *tb\**(budburst time) for the host tree (resource). We call the difference in end times the *mismatch* between the two species (i.e., *mismatch=te\* - tb\**). When the end times of the resting period of two species respond differently to climate change, then the mismatch between the two species will change. This is the fundamental quantity that we study here (Fig. 1D).

One would expect that a consumer that crucially relies on a certain resource would have evolved a relatively small mismatch with that resource. However, this does not mean that its accumulation rate curve has to be the same as that of its resource. Even if the functional form is the same, many different parameter combinations in *R(x)* in Eq. 2 lead to the same end time of the resting period. Hence, as temperature patterns change, the phenologies of the two species may respond differently and the mismatch can increase or decrease. We use our theoretical model to predict general patterns of climate-change induced shifts in species phenologies and changes in the mismatch between interacting species. We use the spruce budworm and balsam fir system in eastern Canada to illustrate how estimated rate accumulation functions can be applied to different temperature scenarios to investigate how climate change may affect the phenological between these two species.

## **2.2 The spruce budworm – balsam fir system**

### **2.2.1 Study system**

The spruce budworm (SBW) (*Choristoneura fumiferana*) is the most destructive insect defoliator of the North American boreal forests (Fleming, 2000). It feeds primarily on balsam fir (*Abies balsamea*), white spruce (*Picea glauca)*, red spruce (*Picea rubens)*, and black spruce (*Picea mariana)*. This insect has a 1-year life cycle with a winter diapause at the second larval stage.  The first stage of the diapause ends in late winter and is followed by a period of quiescence during which development resumes at a rate dependent on outside temperatures. Once this post-diapause development is completed, second instar larvae emerge from their hibernacula (cocoon-like structure) in late April to late May in Eastern Canada and Northeastern USA.

After emergence, young larvae mine 1-year old needles until budburst. The larvae will start feeding on expanding buds and developing needles as soon as they become available. Year-old needles are nutrient poor (Mattson & Scriber, 1987). In contrast, expanding needles from swelling buds have the maximum concentration of nitrogen and mineral elements thus providing nutrient-rich food to the developing larva. Hence, the success of SBW populations in establishing feeding sites in the spring depends on the synchrony of their development with that of their host trees (Régnière & Nealis, 2008)(Volney & Fleming, 2007). Indeed, the population consequences for late-emerging SBW are severe: larval survival is depressed (from about 60% normal survival down to 10%); development is delayed, and average pupal mass decreases by about 50% for both sexes (Lawrence, Mattson, & Haack, 1997). In the boreal forests of Eastern Canada, the higher vulnerability of balsam fir to spruce budworm defoliation compared to other host species has been attributed to its early budburst phenology (Blais, 1957). Indeed, under similar environmental conditions, balsam fir’s budburst occurs two weeks prior to black spruce, closer to spruce budworm emergence. Many environmental factors may affect budburst phenology including the accumulation of cold temperature during the dormancy (i.e., chilling temperatures), warm temperatures during the quiescent phase (i.e., forcing temperatures), photoperiod, or snowmelt. Balsam fir budburst phenology appears to be related to forcing temperatures but not photoperiod (Osawa, Shoemaker, & Stedinger, 1983). In Eastern Canada and Northeastern USA, emergence of SBW second instar larvae from their hibernacula generally precedes balsam fir budburst by several days.

We present accumulation rate functions for spruce budworm and balsam fir separately.

### **2.2.2 Phenological model of spruce budworm’s spring emergence**

Spruce budworms spend 8-9 months in a seasonal resting period. The factors involved in the termination of the first phase are unknown. For modelling purposes, the second stage of the seasonal resting period is often assumed to start on March 1st. The heat accumulation rate in this stage is defined as follows (Régnière et al., 2012)

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|  |  | Eq. 3 |

where

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|  |  | Eq. 4 |

Accumulation happens only when the temperature lies between a minimal value (*xb*) and a maximal value (*xm*). The resulting instantaneous rate of accumulation according to temperature shows a humped-shape profile. Accumulation is integrated through time (see Eq. 1), until *F = 1*. Parameter values for spruce budworm have been estimated from laboratory experiments as , , , , *xb* = 2.5 °C, and *xm* = 35 °C (Régnière et al., 2012).

### **2.2.3 Phenological model of Balsam fir’s budburst**

We use the Uniforc model of Chuine (2000), to model balsam fir’s budburst phenology. Accordingly, the heat accumulation rate is the same as in the theoretical example (see Eq. 2). Accumulation starts at a given time *t0*, which is posterior to January 1rt (Desbiens, 2007), when trees have accumulated enough cold to end bud dormancy. Accumulation is integrated through time (see Eq. 1). Budburst occurs when accumulation reaches a threshold *F\**.

We fit the model to data from Quebec and New Brunswick (Desbiens, 2007), which resulted in the following parameter values: *b* = -0.1936, *c* = 10.99 °C, *t0* = 84 (March 25th), and *F\** = 13.63.  These differ only slightly from the parameterization by Desbiens (2007).

We perform sensitivity analysis on both models using partial rank correlation coefficient (Wu, Dhingra, Gambhir, & Remais, 2013).

## **2.3 Temperature data**

We selected six locations on a latitudinal gradient across Eastern Canada (Fig. 3) to explore the effects of various historical and future temperature regimes on the synchrony between spruce budworm and balsam fir phenology. The temperature data for these locations were calculated using BioSIM (Régnière, Saint-Amant, Béchard, & Moutaoufik, 2014). BioSIM interpolates weather station historical data and climate change scenarios for any location in North America. We used data from 1996 to 2016 to explore the latitudinal trends of past emergence for the insect, budburst for the tree, and the mismatch between the two events over latitude and across years.

    Last, we used predicted temperatures under different warming scenarios at the same locations in order to gain insights in expected trends for emergence, budburst and mismatch in the future. We selected three different scenarios, known as RCP2.6, RCP4.5 and RCP8.5 (van Vuuren et al., 2011). For each scenario, we generated a total of 1200 stochastic temperature time series over the 2001-2100 period.

    Both models used temperature data with a four-hour time interval. It allows for the capture of warm events within a day (i.e., a few hours of warm temperatures, while the average daily temperature stays low).

# 3. Results

## **3.1 Theoretical results**

Model equation (Eq. 1) can in general not be solved explicitly for the end time, *t\*,* at least not for realistic temperature time series *x(t)*. Instead, we derive a general approximation formula for how *t\** changes when a future temperature (*x2(t)*) time series deviates from historical expectation (*x1(t)*) by a small amount (see appendix). We use this formula to evaluate and discuss the impact of a climate change on the phenology of a single species and the phenological mismatch between two species in two particular cases: a constant temperature difference throughout the resting period or a warm or cold spell during a short time period.

### **3.1.1 Phenology shift of a single species**

In the case of a constant temperature difference, we write where is the difference. Then the corresponding end times *t2\** and *t1\** are related by

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|  |  | Eq. |

In the second case, if the difference in temperature between two years is a warm or cold spell of short duration at time *ts* of temperature difference (positive for a warm spell, negative for a cold spell), then the corresponding ends of the seasonal resting phases are related by

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|  |  | Eq. |

     Both formulas show the expected qualitative pattern that if time series *x2* is warmer than *x1*, i.e., , then the phenology advances, i.e., end of the seasonal resting period *t2\** is before the corresponding *t1\** (since all the terms after the “-“ sign are positive). More importantly, the formulas allow us to quantify the expected shift of the end time of the resting period. We observe that the shift depends on the *derivative* of the rate accumulation function. In particular, the impact of a short temperature spell is proportional to the derivative, *R’(x(ts))*, at the time of the spell (*ts*). Hence, the end time of the seasonal resting period is the most sensitive to warm or cold spells where has its maximal slope. For the rate function in Eq. 2, this occurs at temperature *x=c*.

### **3.1.2 Effects on the mismatch between two species**

Since there is no interaction during the resting phase, heat accumulation occurs independently in each species. The mismatch between two species then changes when the two species respond differently to changes in temperature.  It is obvious that the mismatch between two species increases if the phenology of the earlier emerging species advances more quickly, and it decreases if the phenology of the earlier emerging species advances more slowly. Equations 5 and 6 indicate that information about the current phenology (*R(x(t\*))*) and the sensitivity (*R’(x(t))*) are required to quantify the change in phenology in each species and therefore the change in mismatch.

For a theoretical example, let us consider the scenario that the resource emerges before the consumer and is more sensitive at lower temperature. Then, if temperatures are higher than historical averages early in the season, the phenology of the resource is likely to advance more than that of the consumer, so that the mismatch would increase.  Vice versa, if temperatures exceed historical averages only late in the season, when the consumer is more sensitive, the phenology of the consumer is expected to advance more, so that the mismatch would decrease. Hence, a change in mismatch is most likely if the two species are most sensitive at different temperatures, i.e., when the maximal slopes of their respective rate accumulation functions occur at very different temperatures.

In reality, the periods of high sensitivity of the two species may overlap and the rate functions at emergence time (the terms in the denominators in Eqs. 5 and 6) could differ significantly. As a result, the effect of temperature increases depends on details of each scenario. We illustrate this dependence using the rate function for balsam fir (Eq. 2) and for spruce budworm (Eq. 3) together with a simplified time series of daily mean temperatures as modeled by

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|  |  | Eq. 7 |

where the mean, amplitude and offset have been chosen to match historical averages in the city of Fredericton (NB, Canada).  We denote the end times of the consumer for the time series *xi(t)* by *te,i\** (emergence time) and of the resource by *tb,i\** (budburst time).

We begin with the scenario that the future temperature time series differs from historical expectation by a constant, i.e., we use Eq 5 to approximate the advance in phenology. In this simplified model, the end of the seasonal resting period for the resource (consumer) occurs on day 127 (131), the rate curve has its highest slope at 7.14 degrees (19.2 degrees), and the end of the resting period advances by about 3.75 (3.88) days per degree increase in mean temperature. The linear approximation in Eq. 5 captures the actual end of the resting period very well (Fig. 2A).  Since the resting period of the resource at historical temperature regimes ends earlier, but advances more slowly with increasing mean temperature than that for the consumer, the mismatch decreases over time.  However, the difference is small (0.13 days per degree increase). We note that with this simplified temperature time series, the balsam fir emerges before the spruce budworm, whereas in reality the budworm usually emerges earlier. More realistic time series give us the empirically observed pattern (see Section 3.2).

In the second scenario, the future time series differs from historical expectation by a short spell of duration and temperature difference . We apply the corresponding formula (Eq. 6) to each species. Then the mismatch changes according to

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|  |  | Eq. 8 |

Since the resource emerges roughly four days prior to the consumer under historical expectations in our simplified time series (see preceding scenario), mismatch 1 is positive. As in the theoretical example above, the resource is more sensitive to lower temperatures than the consumer (Fig. 2B). In fact, the sensitivity of the resource is a relatively narrow peak around the maximum at *x=c1*. The sensitivity of the consumer peaks at much higher temperatures (higher, in fact, than usually arise before emergence), but the peak is much broader than for the resource. Hence, the phenology of the consumer will advance noticeably no matter when a warm spell occurs, whereas the phenology of the resource will advance noticeably only when it occurs near the temperature *x=c1*. In addition, the rate function at emergence (the denominators in Eq. 8) differs vastly between the two species. For our particular time series, the values are and , respectively. Consequently, even if the resource is more sensitive than the consumer at the time of the spell (), the difference in parentheses in Eq. 8 can still be negative because the denominator in the first term is much larger than in the second. Indeed, this is what happens with the simplified time series and our study system: the phenology of both species advances, and that of the consumer advances more, so that the mismatch decreases, no matter when a warm spell happens. The situation with realistic time series that vary across latitude is more nuanced (see below).

Despite their simplicity, these scenarios give us some important insights into the evolution of the mismatch. The advance of the phenology of each species is a nonlinear function of temperature increase and each species has a temperature regime during which its phenology is most sensitive to change, namely when the slope of the rate function is large. The mismatch between two species increases if the phenology of the earlier emerging species advances more quickly, and it decreases if the phenology of the earlier emerging species advances more slowly. A change in mismatch is most likely if the rate accumulation functions of the two species have their maximal slopes far apart.

## **3.2 Case study: spruce budworm - balsam fir system**

### **3.2.1 Fitting and sensitivity**

We analyzed the residuals of the fitting of the balsam fir model in order to check its quality. The residuals follow a Normal distribution centred on 0 (Fig. 4A). Even more importantly, no obvious pattern can be observed for the residuals across latitude (Fig. 4B) in the range that we used throughout the study.

    In order to test the accuracy of the models, we compared predicted budburst and emergence date with available data for budburst and emergence dates for two years (2013, 2014) in two sites in Quebec (Pureswaran, Neau, Marchand, De Grandpré, & Kneeshaw, 2019). Measured budburst occurred between day 136 (May 16th) and day 168 (June 17th), in 2013, and between day 149 (May 29th) and day 166 (June 15th), in 2014. The model predicts a median budburst on day 147 (May 27th) in 2013, and day 148 (May 28th) in 2014. Hence, the tree model shows good accuracy.    A similar test was done on the insect model. According to Pureswaran, Neau, Marchand, De Grandpré, & Kneeshaw (2019), the emergence peak occurred on day 129 (May 9th) in 2013, and on day 143 (May 23rd) in 2014. The model predicts a median emergence on day 140 (May 20th) in 2013, and day 141 (May 21st) in 2014. The accuracy of the insect model is also satisfactory.

    Our analysis shows that the budworm model is sensitive to most parameters (Fig. 4C). The only exception is *xm* (the maximal temperature) since very high temperatures are rare during late winter and spring, and to a certain extent . Increasing parameters , , and *xb* (minimal temperature) would delay emergence, while increasing would strongly advance phenology. The tree model is most sensitive to parameters *t0* (when the tree starts accumulating heat) and *b*  (which drives the speed of accumulation). An increase in *t0* would postpone budburst date, while an increase in *b* would advance phenology (Fig. 4D).

### **3.2.2 Spruce budworm – Balsam fir system across latitude**

Both emergence (Fig. 5A) and budburst (Fig. 5B) occur later at higher latitudes, although some discrepancies may occur due to altitude. However, insects and trees are not affected by temperatures in the same way. Hence, the mismatch between both events varies across latitude (Fig. 5C). On the one hand, at lower latitudes, emergence is expected to occur 5 to 10 days before budburst. Hence, the whole larvae population may have time to emerge before budburst occurs, which allows larvae to benefit from an important source of nutrients. On the other hand, at higher latitudes, emergence may sometimes occur before budburst and sometimes after. In some years, emergence may occur a few days before the budburst, and larvae may have to wait a few days (up to 10 days), while in other years, emergence may occur a few days after budburst. In brief, insects should be more adapted to their host phenology at lower latitude than at higher latitudes, where larvae may experience some years with harsher conditions than others.

### **3.2.3 Predicted trends according to warming scenarios**

Across all scenarios, emergence and budburst are expected to occur earlier when temperatures increase. There is still a latitudinal pattern. Warmer scenarios lead to an increase of variance, more than a real shift in date compared to less warm scenarios. Insects and trees react differently to temperature increase. Balsam fir shows a difference between northern and southern sites of 10 days on average. This difference stays roughly the same in case of warmer temperatures (i.e., budburst date is shifted similarly across latitude, see Fig. 5A). On the other hand, emergence of spruce budworm is expected to shift differently across latitude (about 15 days compared to 10 nowadays, see Fig. 5B).

    Therefore, the mismatch is affected. Southern sites are expected to show emergence occurring before budburst most of the time. In case of a moderate warming scenario (RCP2.6), the waiting time is expected to be short, which would allow the larvae to have access to the nutrient-rich swelling buds and developing needles. Thus, a moderate warming would be beneficial for the insect at low latitudes. In case of greater warming (RCP4.5, RCP8.5), variance may lead to emergence occurring either too early or too late. Larvae may or may not have access to nutrients depending on years. Thus, a greater temperature increase may lead to stronger population variances among years.

    On the other hand, in northern sites, all warming scenarios lead to a better synchrony between the insect and its host tree (i.e., a reduced mismatch). The overall pattern looks similar to what is expected nowadays in southern sites. But the variance is expected to be greater with warmer scenarios, which may provoke an increase of insect mortality from time to time, when emergence occurs too late.

# 4. Discussion

We investigated the effects of temperature on the end of the seasonal resting period, and its consequences on consumer-resource synchrony.

## **4.1 Expected impacts of climate change on phenological mismatch between consumer and resource**

There is a growing body of literature showing phenological shift among species due to climate change. However, different species do so at different rates (Visser & Both, 2005)(Donnelly, Caffarra, & O’Neill, 2011)(Kharouba et al., 2018). Our model provides a mechanism that links development and temperature. Thus, it specifically formulates the relationship between a change in temperature and the resulting phenological shift. In the likely case where the accumulation function is specific to each species, the effects of a given temperature regime will also be species-specific. Our model also shows that even for a given species, a change in temperature may have different effects depending on the base temperature at which the change is applied. Hence, a warm spell occurring when the slope of development rate is very low may have little effect compared to the same warm spell occurring when development is more sensitive to temperature (i.e., when *R’* is high) (see Fig. 2B, and 4.2). Moreover, several consecutive spells will have additive effects. Several warm spells will advance phenology several times, while a cold spell may cancel effects of a warm spell.

In the case of two interacting species, if a consumer shows a response to temperature that differs from that of its resource, a change in climatic patterns may lead to either an increase or a decrease of the mismatch between the two species. This pattern has been already observed in real systems and seems to change at an accelerating rate (Kharouba et al., 2018). Increasing deviation of climate conditions from normals could be an explanation. This increase could be compounded by increasing rates of spring development as suggested in our general model.

The present study investigates the case of two species whose respective life cycles are driven by the same external factor (i.e., temperature), but in potentially different ways. The resulting synchrony between the two species is an indirect synchrony driven by temperature. This kind of modelling approach is of prime importance to investigate potential effects of global warming on consumer-resource systems (Chuine and Régnière, 2017). Synchrony / mismatch between a consumer and its resource is fundamental to predict future species distribution. Our work improves our ability to predict the direction and the magnitude of the change in phenological synchrony, a research challenge identified by earlier studies (Kharouba et al. 2018).

    This phenological shift may affect persistence of the consumer and/or its resource. For example, an increase of the mismatch between the two species may prevent the consumer from interacting with the resource, which may prevent the consumer from completing its life cycle, and may lead to a decrease in consumer biomass, or even its extirpation from the ecosystem. In some cases, if a subsidiary resource is available, the consumer may switch resources, which has some cascading effects throughout the food chain. On the other hand, a decrease in mismatch may lead to a stronger depletion of the resource, which may cause ripple effects through the food web. Thus, this phenological shift between the consumer and its resource is likely to affect species demography (Miller-Rushing, Høye, Inouye, & Post, 2010). It may even affect the geographic range of species distribution (for the consumer and/or the resource).

## **4.2 The spruce budworm - balsam fir system and its general implications**

We apply our general theory to the spruce budworm - balsam fir system.  The rate accumulation functions of the insect and its host tree have different shapes. Thus, the two species should react differently to global warming. We predict the future mismatch across latitude under different warming scenarios. Our conclusions are of prime importance for the future of the boreal forest in Eastern Canada. First, the mismatch between emergence and budburst is critical for the geographic distribution of the spruce budworm. If emergence occurs within two weeks before budburst, budworm populations may have access to a rich source of nutrients when buds open, which in turn increases the likelihood of larvae survival, and population persistence. On the other hand, if emergence occurs too early (or too late), budworm may run out of energy before budburst (miss the budburst and the access to a rich source of nutrients). In both cases, budworm populations may face population decline, which could lead to the budworm population not persisting at that location. Second, the model also gives information for the distribution of the host tree. The fact that the spruce budworm may (may not) be in better synchrony with its host tree would lead to a potential increase (decrease) of the threat on the tree, which in turn may affect the potential latitudinal distribution of the tree. (I don’t like this last sentence. **Can we find a more affirmative formulation for it?)**

Last, an increase in the mismatch between the insect and its host tree may affect other species. Hence, if the spruce budworm is not locally in synchrony with balsam fir anymore, it may establish on another host (e.g., black spruce) that may show a better synchrony between budburst date and emergence date at that location (Pureswaran et al., 2015). It is known that defoliator insects (especially spruce budworm), along with fire cycles and soil composition are the main drivers of the forest composition in the black spruce, white spruce and balsam fir zones (Harper et al., 2003). A switch of host would potentially disrupt this equilibrium, since other tree species would experience an increase of the threat due to spruce budworm, which in turn may affect their geographic distribution. It therefore has implications for the boreal forest in Eastern Canada. To test this hypothesis, one would require *in situ* phenological data that would allow for a parameterization of the tree model (Eq. 2) for white spruce and black spruce.

Nonetheless, this concrete example has implications that go beyond the spruce budworm - balsam fir system. Hence, a good estimate of the mismatch between two interacting species relies on accuracy of both models. The analysis of this case study leads to four general conclusions. First, the difference in model robustness according to the source of forcing variables (temperature) is a critical point and may be explained by the way the models have been calibrated on real data. The calibration of the tree model was done on phenology data collected in Quebec and New-Brunswick. Data comes from trees that experienced real seasonal variations. On the other hand, the calibration of the insect model was done on experimental data (Régnière, 1990) coming from budworms that grew under different temperature regimes. We acknowledge that sampling L2 instar on site is very challenging due to its size (1- or 2-mm length). However, despite these limitations in data collection, it seems that experimental data may not be as accurate as sampling data to calibrate phenological models, which is likely to play a fundamental role in the generalization of this kind of approach.

Second, this study system also emphasizes the need of using real temperature data. For the system shown, simulations done with real temperature data lead to different results than simulations done with theoretical temperatures (using a cosine function). All things being equal, the theoretical temperatures lead to budburst always occurring before emergence (Fig. 2A), while real temperatures lead to emergence mostly occurring before budburst (Fig. 5). The later pattern is consistent with real observations (Pureswaran, Neau, Marchand, De Grandpré, & Kneeshaw, 2019). The difference between the two outputs can be explained by the lack of warm spell in the theoretical distribution (i.e., temperatures increase monotonically from January 1st to Summer). Real temperatures show warm spells during the same period of time, which advances the phenology of the budworm that is more sensitive than the tree to these short warm events.

Third, even different sources of real temperature data can lead to different results. Results shown throughout this article were obtained using temperature data from BioSim (Régnière, Saint-Amant, Béchard, & Moutaoufik, 2014). Using this interpolation method for temperature data, both models show a good level of accuracy when compared to real data (see 3.2). However, if another interpolation method is used (Thornton et al., 2020), the two models may show different behaviour. The BioSim temperature data shows some warm spells throughout late Winter- early Spring period compared to the other interpolation methods. Hence, emergence of the insect is expected to occur two weeks earlier when the BioSim temperature data is used. The source of temperature data is therefore another critical point.

    Last, the way temperature is applied on the model is also a major factor. If one considers an average temperature per day, results are different from using a four-hours time interval for temperatures (used throughout the present study). Hence, variations within a day may allow for development during a few hours, while the average temperature over the whole day may be too low to allow for development. A four-hours time interval advances phenology. The insect model shows a greater difference (about 10 days earlier using the four-hours interval) than the tree model (less than a week earlier). This sensitivity is therefore species-dependent. These four general conclusions are fundamental to develop future applied models.

## **4.3 Limitations of the approach**

Although the present study provides novel insights towards a better understanding of temperature of species phenology, we should also point out some limitations. First, our study is limited to cases where the phenology is directly constrained by environmental variables (more specifically temperature). It is well known that temperature (or any other environmental variable) can also be used differently. For example, organisms can use environmental variables indirectly as cues for predicting optimal time window to have access to resources or convenient environmental conditions (e.g., photoperiod, rainfall) (McNamara, Barta, Klaassen, & Bauer, 2011).

    Second, we measure mismatch as the time lag between peaks of consumer demand and resource availability. Some authors have argued that more precise measurements should take into account the shape and location of the phenological distributions around these peaks (e.g., (Lindén, 2018)) while others showed that, in some cases, time lag between phenological peak dates was a better predictor of resource availability than the overlap between phenological distributions (Ramakers, Gienapp, & Visser, 2020).

## **4.4 Conclusion**

The modelling approach presented here is of prime importance to investigate potential effects of global warming on consumer-resource systems. Synchrony / mismatch between a consumer and its resource is fundamental to predict future species distribution. Future studies may go further by including more complex mechanistic approaches (e.g., energy budget models) in order to model the whole life cycle of the consumer, which would lead to new insights on the dynamics of the systems. Our model could be linked to models that explore the consequences of phenological mismatch on the population dynamics of a consumer and its resource (e.g., (Bewick, Cantrell, Cosner, & Fagan, 2016)).

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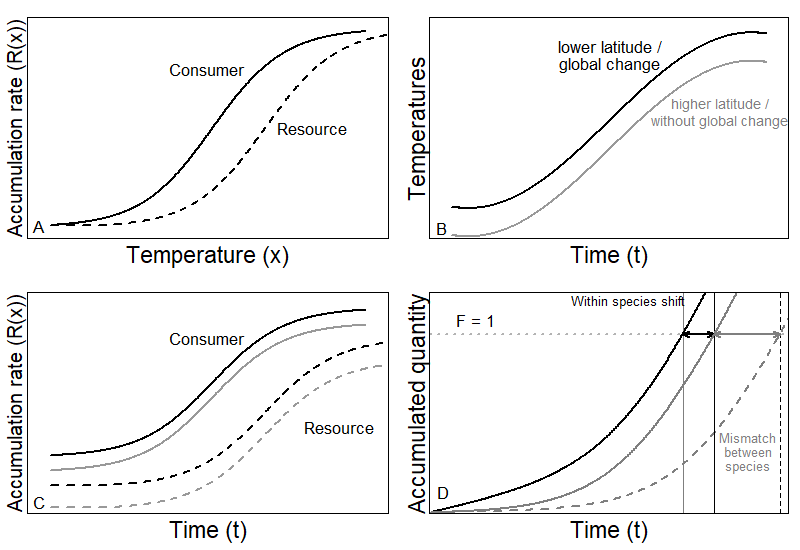
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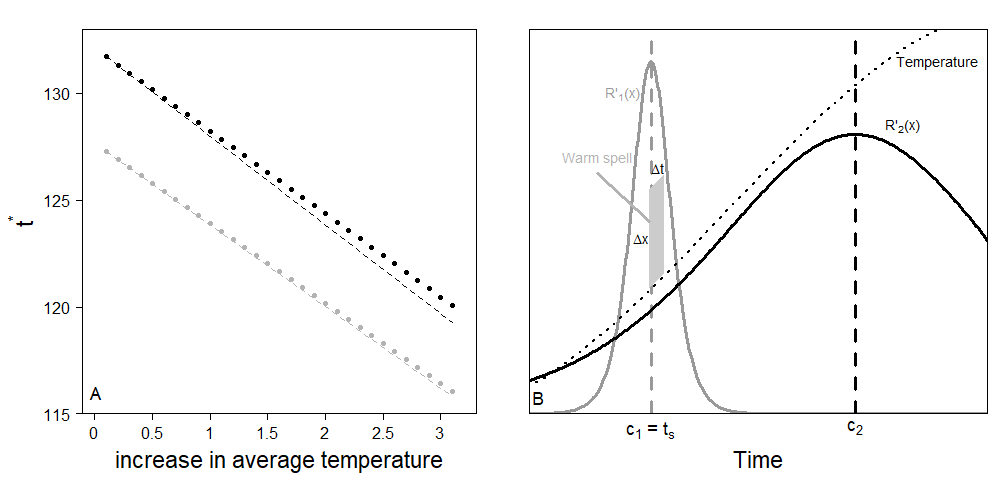
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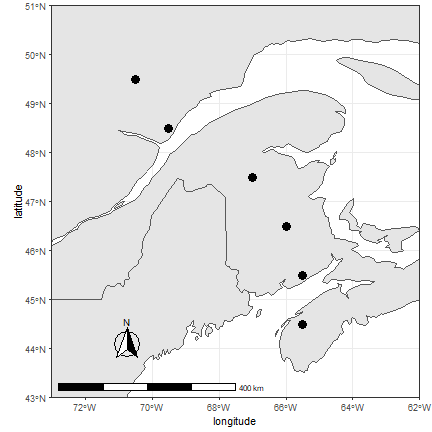
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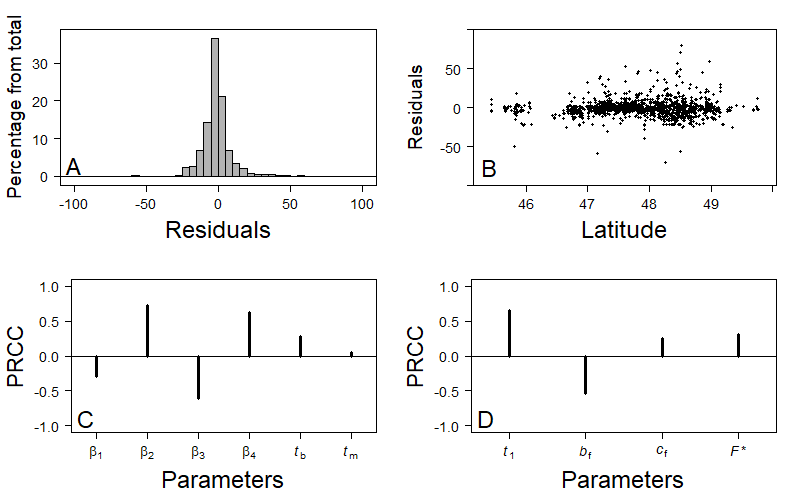
**Figure 1**: Illustration of theoretical development. (A) The rate accumulation function for two different species (solid line is consumer and dashed line is resource). (B) two simplified temperature time series (black line is warmer and grey line is cooler). (C) Four combinations of rate accumulation; each species with two different temperature time series. (D) The resulting end of the resting phase for the consumer in cooler (grey solid) and warmer (black solid) temperatures and for the resource in cooler temperatures (black dashed). The difference within species (grey vs black) indicates the shift in emergence in space (due to latitude or altitude) or time (due to global change). The difference between species (solid vs dashed) indicates the mismatch in the end of the seasonal resting phase for a fixed temperature regime (same location and same time).



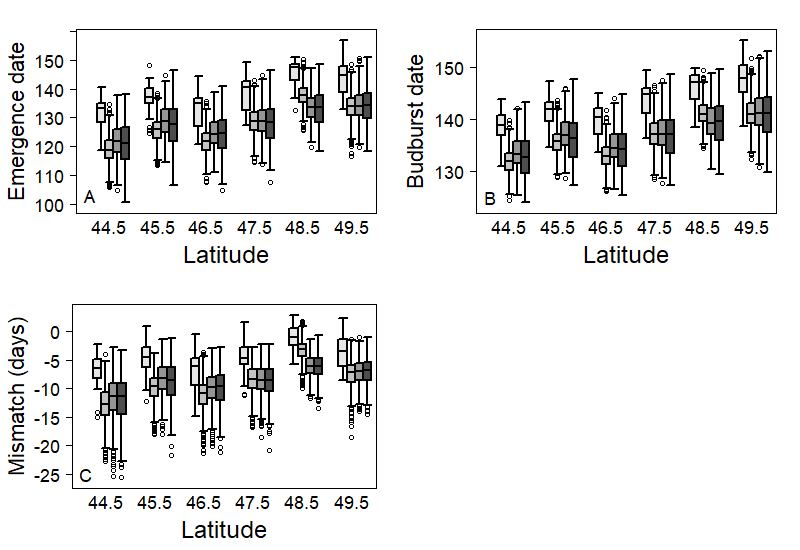
**Figure 2**: Effects of (A) a constant temperature difference, and (B) a short warm spell, on species phenology. For both panels, black is the consumer (spruce budworm), and grey is the resource (balsam fir). (A) A constant temperature difference advances species phenology. Solid is the predicted value, dashed is the linear approximation from the model with simple time series. (B) The two species have their R’ that peaks at different temperatures. A short warm spell will mostly affect the species which R’ is highest at that time (in this example, the tree is more sensitive than the insect).



**Figure 3**: Location of the sample sites where temperature data were collected for past and future trends. Points are located across a gradient of latitude in Nova-Scotia, New Brunswick, and Quebec.



**Figure 4**: Fitting residuals and sensitivity analysis of the Balsam fir model. (A) Residuals follow a Normal distribution centered on 0. (B) No obvious latitudinal patterns can be found on the residuals within the range of latitudes that is used throughout the rest of the study. (C) Partial Rank Correlation Coefficient (PRCC) shows that the budworm model is sensitive to most parameters especially , and *xb* that delay emergence, and that advance phenology. (D) The tree model is mostly sensitive to *b* that advances phenology, and *t0* that delays budburst.



**Figure 5**: Latitudinal distribution of (A) emergence date of L2 instar (Julian days), (B) budburst date (Julian days), and (C) mismatch between emergence and budburst date. For each latitude, the white box (left one) represents the 1996-2016 period. Grey boxes represent expected outcome according to RCP 2.6 (light grey), RCP 4.5 (dark grey), and RCP 8.5 (black) scenarios over 2021 to 2100.  Both emergence and budburst are expected to occur later at higher latitudes. Over all warming scenarios, both events are expected to occur earlier in the year. Warmer scenarios generate more variance. Nowadays, emergence is expected to occur 5 to 10 days before budburst at low latitudes, while at higher latitudes, emergence may sometimes occur before budburst and sometimes afterwards.  For all warming scenarios, an increase in mismatch is expected. At low latitudes, emergence may occur too early some years, which may lead to low survival of L2. At higher latitudes, emergence is expected to systematically occur a few days before budburst, which would increase survival of L2.