

Propagule redirection: Habitat availability reduces colonization and increases recruitment in reef fishes

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Abstract. Increased habitat availability or quality can alter production of habitat-dependent organisms in two contrasting ways: (1) by enhancing input of new colonists to the new sites (the Field-of-Dreams Hypothesis); and (2) by drawing colonists away from existing sites (the Propagule Redirection Hypothesis), and thus reducing the deleterious effects of density. We conducted a field experiment on coral reef fishes in Moorea, French Polynesia, to quantify how differing levels of habitat availability (controlling for quality) increased and/or redirected colonizing larval fish. Focal reefs without neighboring reefs received two to four times more settlers than reefs with adjacent habitat, demonstrating that increased habitat redirected larval fish. At the scale of the entire reef array, total colonization increased 1.3-fold in response to a sixfold increase in reef area (and a 2.75-fold increase in adjusted habitat availability). Thus, propagules were both increased and redirected, a result midway between the Field-of-Dreams and Propagule Redirection Hypotheses. A recruitment model using our data and field estimates of density-dependent recruitment predicts that habitat addition increases recruitment primarily by ameliorating the negative effects of competition at existing sites rather than increasing colonization at the new sites *per se*. Understanding long-term implications of these effects depends upon the interplay among habitat dynamics, population connectivity, colonization dynamics, and density dependence.

Key words: attraction–production; coral reefs; density dependence; habitat selection; reef fish; settlement; settlement shadow.

INTRODUCTION

The spatial configuration, quality, and size of habitat patches can determine the distribution and abundance of organisms by affecting colonization, movement, and the strength of species interactions (MacArthur and Wilson 1967, Hanski 1998). Understanding the role of habitat is particularly important in light of tremendous variation in habitat availability, driven by both natural (Porter 1972, Gardner et al. 2005, Silliman et al. 2005, Bruno et al. 2007) and anthropogenic (e.g., Barel et al. 1985, Skole and Tucker 1993, Bellwood et al. 2004) causes. Environmental change often involves the loss (or gain) of a biogenic habitat critical to the persistence of other organisms. Thus, many restoration techniques focus on reestablishing the structural attributes of a system (e.g., trees or corals) based on the Field-of-Dreams Hypothesis: “if you build it, they will come” (a reference to the 1989 movie [Palmer et al. 1997]). This concept may be useful in understanding the response of species to restoration of degraded habitat, but it also may provide a valuable foundation for understanding the dynamics of species more generally, especially in response to natural fluctuations in habitat availability.

The Field-of-Dreams Hypothesis proposes that increased habitat availability will lead to a proportionate increase in colonists, with no resulting decrease in density at previously existing sites. Alternatively, new habitat may simply redirect colonists away from other suitable sites (Carr and Hixon 1997, Osenberg et al. 2002b): the Propagule Redirection Hypothesis. Propagule redirection may create settlement shadows (*sensu* Jones 1997) and unintentionally contribute to the further degradation of existing sites by removing potential colonists (Osenberg et al. 2002b). However, if postcolonization processes are density dependent, then redirection may still enhance the population by ameliorating deleterious effects of density in existing habitat. The net effect of redirection will depend on the strength of density dependence, the overall level of colonization, and the magnitude of propagule redirection (Osenberg et al. 2002b). Effects of propagule redirection on the response of populations to altered habitat availability also will depend upon connectivity, temporal scale, and habitat quality (Resetarits 2005, Resetarits and Binckley 2009) (see *Discussion* for more detail).

Studies of larval depletion by predators in which a “wall of mouths” (predators) reduce settlement to downstream habitats (Gaines and Roughgarden 1987, Hamner et al. 1988, Peterson and Black 1991) provide compelling evidence that biological processes can modify patterns of larval settlement. In a different

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context, foraging studies of pollinators suggest that the addition of flowers can lead to reductions in local density of the pollinators (reviewed in Mitchell et al. 2009). There are, however, few experimental studies of the effects of habitat on colonization per se. In one of the best examples, Resetarits (2005) examined oviposition by a tree frog (*Hyla chrysoscelis*) in relation to sets of pools that varied in the proportion of sites with predators. Oviposition was greater in predator-free ponds that had the fewest number of neighboring ponds with predators, demonstrating selective oviposition at both local and landscape scales. However, this work did not directly address redirection of ovipositing females (and their eggs), because the fates of the females (and their eggs) that could have oviposited at sites with some predator-containing ponds, but chose not to, were unknown. Did they perish, or were they redirected to other pools? In contrast, Resetarits and Binckley (2009), working with aquatic beetles, also varied the number of pools, keeping composition fixed. They found a proportionate increase in colonization rates as the number of pools increased, corroborating the Field-of-Dreams Hypothesis. However, this seminal (and to date, unique) study did not examine or extrapolate their results to postcolonization dynamics.

Propagule redirection is particularly relevant in species with demographically open local populations in which organisms must find and colonize new sites during each generation, and where colonists have the potential to colonize many possible sites. This phenomenon is particularly relevant to many benthic marine organisms, whose propagules are capable of dispersing long distances over short periods of time. For example, the majority of reef fishes have a bipartite life history, with pelagic larvae that disperse tens to hundreds of kilometers from their natal site.

Here, we (1) report a field test of the Propagule Redirection and Field-of-Dreams Hypotheses for coral reef fishes, (2) parameterize a model that integrates our observed levels of propagule redirection and previous experimental estimates of density dependence to evaluate the implications of propagule redirection on fish recruitment (i.e., postcolonization dynamics), and (3) use the model to assess the relative importance of increased colonization, propagule redirection, and reduced density dependence on fish recruitment.

MATERIALS AND METHODS

Study site and species

We conducted fieldwork on the northern shore of Moorea, French Polynesia (17°30' S, 149°50' W), a shallow lagoon with interspersed sand and reef inside a barrier reef (Galzin and Pointer 1985). Our experiment had two treatments (high vs. low habitat availability) arranged in 12 experimental blocks. To assess the effects of habitat availability on redirection we used a “competition” design, in which we monitored settlement to focal habitat that either did or did not have

neighboring sites “competing” for larvae. Each location was an open sand flat (at a depth of ~4 m) with no other hard-bottom habitat, so that each treatment would receive approximately equal larval input within a block without influences from neighboring natural reefs.

To make our experimental arrays as uniform as possible, we constructed 168 Standardized Habitat Units (e.g., Forrester 1990) by attaching colonies of *Pocillopora verrucosa* (~15 cm diameter) to cinder blocks (one coral colony per cinder block) using Z-spar marine epoxy (Splash Zone Compound, Kopcoat, Pittsburg, Pennsylvania, USA). Each of 12 sites (i.e., statistical blocks) contained one replicate of each treatment and consisted of 14 SHUs. Pairs of focal SHUs for each treatment were placed 16 m apart (and perpendicular to the dominant current direction, with at least 15 m to the nearest natural reef). We then surrounded one pair of focal SHUs by 10 single SHUs arranged in a 2 m diameter circle (see Appendix A for a diagram). This constituted the “high habitat” treatment. In the “low habitat” treatment, the two focal SHUs lacked neighbors.

The Field-of-Dreams and Propagule Redirection Hypotheses differ in their predictions about the input of colonists (settlement) as well as the longer-term effects on recruitment (i.e., establishment of older life stages). Because the presence of older life stages (or heterospecifics) can modify settlement patterns (e.g., Schmitt and Holbrook 1996), we quantified and removed settlers on a daily basis (as described in the next paragraph). This allowed us to unambiguously evaluate the key prediction that distinguishes the Field-of-Dreams and Propagule Redirection Hypotheses: i.e., does added habitat proportionally increase settlement or simply redistribute settlers? However, this approach precluded us from quantifying subsequent postsettlement recruitment patterns. We therefore used existing data on the strength of postsettlement density-dependent survival to extrapolate the observed settlement effects to their longer-term recruitment consequences.

We monitored settlement daily for 28 days from 15 June to 12 July 2006. Each day between the hours of 0730 and 1130, all fishes that had settled the previous night were removed using hand nets and the anesthetic Eugenol (Munday and Wilson 1997). By removing fish daily, we (1) minimized mortality prior to sampling (most fish experience high levels of mortality during the first 1–2 weeks after settlement; thus weekly sampling would be too coarse), (2) reduced the effects of predator aggregation to sites with high prey density, and (3) eliminated potential confounding effects such as density-independent or density-dependent competition or predation by older conspecifics and heterospecifics. Because recently settled fish are highly attached to a given coral site, especially during their first 24 hours on the reef (Holbrook and Schmitt 2002), we assumed that new settlers had not moved between experimental arrays.

Data analysis

We collected settlers from 16 species and six families (Appendix E), but only 4 species, which together comprised >88% of all settlers, were sufficiently abundant to permit statistical analysis: *Dascyllus aruanus* (Pomacentridae), *Dascyllus flavicaudus* (Pomacentridae), *Halichoeres trimaculatus* (Labridae), and *Paragobiodon lacunicolus* (Gobiidae). We quantified settlement for these four species as the cumulative number of individuals that settled to the SHUs during the 28 days of sampling. To test for propagule redirection, we compared settlement of each fish species to the two focal SHUs either with or without neighbors (i.e., at high vs. low habitat availability). That is, for each block, we calculated $\ln(S_{-N}/S_{+N})$, where S is the cumulative settlement over 28 days to the focal SHUs that were isolated (no other corals within at least 15 m: $-N$, where N stands for neighbors) or to the two focal SHUs surrounded by 10 neighboring SHUs ($+N$). In the absence of propagule redirection, the numbers of settlers should be the same in these two treatments: i.e., $\ln(S_{-N}/S_{+N}) = 0$. If additional habitat depletes larvae and therefore reduces settlement to nearby habitat, then settlement to the two isolated SHUs should be greater than settlement to the two SHUs surrounded by neighbors: i.e., $\ln(S_{-N}/S_{+N}) > 0$.

To determine whether increased habitat availability affected total settlement to the entire experimental array, we compared the cumulative number of individuals that settled to the 12 SHUs at high habitat availability (i.e., to the two focal SHUs plus the 10 neighboring SHUs) vs. low habitat availability (i.e., to the two isolated SHUs): $\ln(S_{\text{high}}/S_{\text{low}})$. Note that $S_{\text{low}} = S_{-N}$, but we use different terminology to make our notation more intuitive for each comparison. The Propagule Redirection Hypothesis (in its extreme) predicts that settlement to the high habitat arrays should be the same as settlement to low habitat arrays ($S_{\text{high}}/S_{\text{low}} = 1$), whereas the Field-of-Dreams hypothesis predicts that the relative settlement should be proportional to habitat availability (e.g., $S_{\text{high}}/S_{\text{low}} = 12/2 = 6$; but see *Results* for a modification to this simple expectation).

Projected recruitment success

To assess how effects of habitat availability on fish settlement translate to changes in fish recruitment, we modified the Beverton-Holt recruitment function (Osenberg et al. 2002a) to incorporate habitat availability, and parameterized the model using existing field data on density dependence in *D. flavicaudus*, one of our focal species and a species known to compete for predator-free space within coral heads (Schmitt and Holbrook 2007)

$$N_t = \frac{e^{-\alpha} N_0}{1 + \frac{\beta(1 - e^{-\alpha}) \frac{N_0}{h}}{\alpha}} \quad (1)$$

where N_t is the number of recruits to a site (i.e., settlers

who survive t days); N_0 is the initial number of settlers to that site, α is the per capita density-independent mortality rate, β is the per capita density-dependent mortality rate, and h is the amount of habitat (i.e., N_0/h is the density of settlers). See Appendix B for details.

Habitat addition can increase recruitment and local density in two ways: (1) increased settlement (because there is more habitat; i.e., Field-of-Dreams); and (2) reduced density dependence (because settlers are spread out among more habitat; i.e., Propagule Redirection). The latter (reduced density dependence) only occurs if the increase in total settlement is less than proportional to the availability of habitat (i.e., if there is propagule redirection). We compared the relative importance of these two pathways by partitioning the potential change in recruitment due to these components (see Appendix D).

RESULTS

Settlement

Settlement was 2–4 times greater to the isolated focal SHUs than to the focal SHUs surrounded by neighboring SHUs, demonstrating that propagules were redirected: Fig. 1A, *D. flavicaudus* ($t_{11} = 11.67$, $P < 0.0001$), *D. aruanus* ($t_{11} = 12.17$, $P < 0.0001$), *H. trimaculatus* ($t_{11} = 6.22$, $P = 0.0003$), and *P. lacunicolus* ($t_{11} = 3.31$, $P = 0.007$). This pattern also was observed for the combined response of all other species that settled (Appendix E), although one uncommon species (*Acanthurus triostegus*) suggested the reverse trend.

Settlement to the entire experimental array (e.g., the focal SHUs as well as any neighboring SHUs) was 16–66% greater for high vs. low habitat availability (Fig. 1B), although this increase was significant for only two of the four species analyzed: *H. trimaculatus* ($t_{11} = 3.36$, $P = 0.0064$); *D. aruanus* ($t_{11} = 3.21$, $P = 0.0083$), *D. flavicaudus* ($t_{11} = 1.86$, $P = 0.088$), and *P. lacunicolus* ($t_{11} = 1.22$, $P = 0.25$; Fig. 1B). This ~36% increase in settlement (averaged across all four species) demonstrated that added habitat led to increased settlement. However, the response was far less than the predicted 500% (sixfold) increase expected under the simple Field-of-Dreams Hypothesis (Fig. 1B). Again, similar results were obtained for the combined response of all other species (with the exception of *A. triostegus*; Appendix E).

Our simple Field-of-Dreams prediction is based on the 6:1 disparity in number of SHUs in the two treatments. Complicating this prediction was the heterogeneous settlement patterns of fish on the array with neighbors: settlement was greater to “upstream” (vs. “downstream”) neighbors (Appendix E), but also greater to the focal SHUs (on a per SHU basis) relative to the neighbors (Appendix E). For the four focal species, settlement to a neighbor SHU was only 35% of that observed to a focal SHU. The magnitude of this effect did not differ among species. One explanation for this result is that fish settled preferentially to SHUs

placed side by side (i.e., focal SHUs; see Appendix A). We therefore adjusted the expectations under the Field-of-Dreams hypothesis to account for this possible difference in habitat quality. Instead of a sixfold difference in habitat, we assumed a 2.75-fold difference: i.e., $(10\text{SHUs}(0.35) + 2\text{SHUs}):2\text{SHUs}$. The observed settlement to the arrays with neighbors was still demonstrably lower than this expectation for three of the four fishes (Fig. 1B).

Projected recruitment success

Although total settlement of *D. flavicaudus* was only 16% greater at high habitat availability, projected recruitment using the Beverton-Holt equation was 125% greater (Fig. 2B). Had settlement increased in proportion to habitat, recruitment should have increased 175% (i.e., 2.75-fold). In contrast, if there was complete propagule redirection, so that habitat addition did not increase settlement but only reduced density-dependent mortality, recruitment would have increased by 114% (i.e., most of the observed 125% increase). Thus, the observed increase in recruitment due to habitat enhancement was primarily due to a reduction in density dependence via propagule redirection. The circular array provided additional information on effects of the spatial structure of habitat in the circular array, suggesting patterns of settlement may also be affected by the arrangement of habitat on a local scale (meters) (see Appendix C).

The increase in recruitment (and the partitioning to different components) was sensitive to the overall level of settlement. For example, if ambient settlement was only 10% of that observed, then the 16% increase in settlement observed for *D. flavicaudus* to the high habitat sites would have led to an approximately similar increase in recruitment (in this case 17%) because the effects of density would be small at such low input levels (Fig. 2A). Thus, the main effect of increased habitat availability would be via the increase in settlement (as slight as it was). On the other hand, if overall settlement were an order of magnitude greater than observed, then recruitment to the high habitat arrays would have increased by almost 93% despite the only modest (16%) increase in settlement to those arrays (Fig. 2C). If overall settlement increased even more (beyond the 10-fold change examined here), recruitment to the high habitat arrays would have increased in proportion to the change in habitat (i.e., by 175%). Thus, at high settlement, the main effect of habitat addition would be on the amelioration of density dependence rather than the increase in settlement, because the available habitat was already saturated.

DISCUSSION

Changes in habitat availability can affect population dynamics and species interactions by altering colonization rates and/or the strength of density-dependent interactions that arise following colonization. The life-

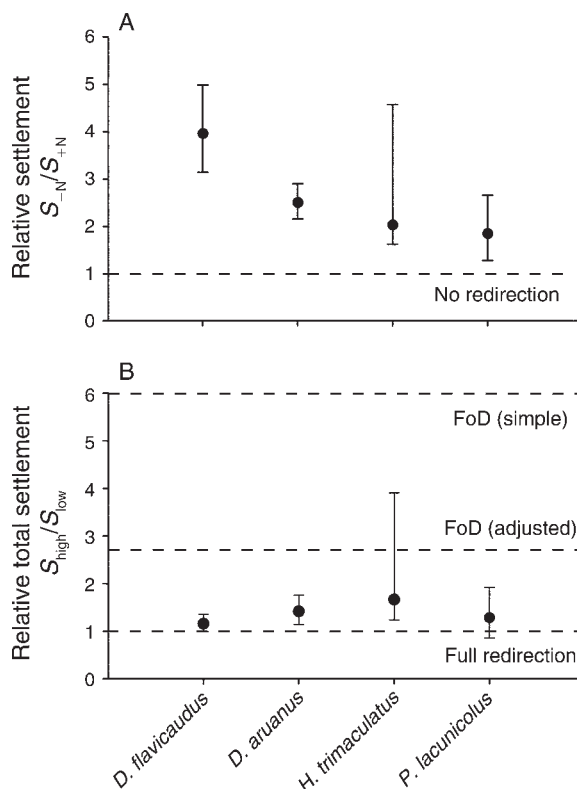


FIG. 1. Effects of habitat availability on patterns of relative settlement for four focal fish species (*Dascyllus ariuanus*, *D. flavicaudus*, *Halichoeres trimaculatus*, and *Paragobiodon lacunicolus*) on coral reefs in Moorea, French Polynesia. (A) The effect of habitat treatments on the average relative settlement of fish to focal Standardized Habitat Units (SHUs) with vs. without neighboring reefs: S_N/S_{+N} . The dashed line (Field-of-Dreams, FoD) represents the expected relationship if neighboring reefs have no effect on settlement to the focal reefs. (B) The effects of habitat on the relative total settlement of fish to the entire experimental array (i.e., to 12 SHUs vs. 2 SHUs): S_{high}/S_{low} . The lowest dashed line gives the expected result if larvae are fully redirected (i.e., total settlement is the same to both arrays, as in the extreme version of the Propagule Redirection Hypothesis). The two upper dashed lines give the expected results if there is no redirection (i.e., Field-of-Dreams Hypothesis, FoD) based on the sixfold increase in habitat area or 2.75-fold increase in habitat availability (adjusted for quality). Data points and error bars represent back-transformed means and 95% CI.

stage responsible for colonizing new habitat varies among species. For example, in many freshwater and terrestrial systems adults colonize habitats, often as sites for oviposition, e.g., frogs (Resetarits 2005), aquatic insects (Stav et al. 1999, Resetarits 2001), and butterflies (Rauscher 1979, Renwick and Chew 1994). In other systems, earlier life stages are dispersive and adults are more sedentary (e.g., most marine organisms and plants). Although our study focused on marine fishes, these insights likely apply to a variety of systems in which local dynamics are relatively open.

The simplest expectation is that colonization will be proportional to habitat availability (i.e., the Field-of-

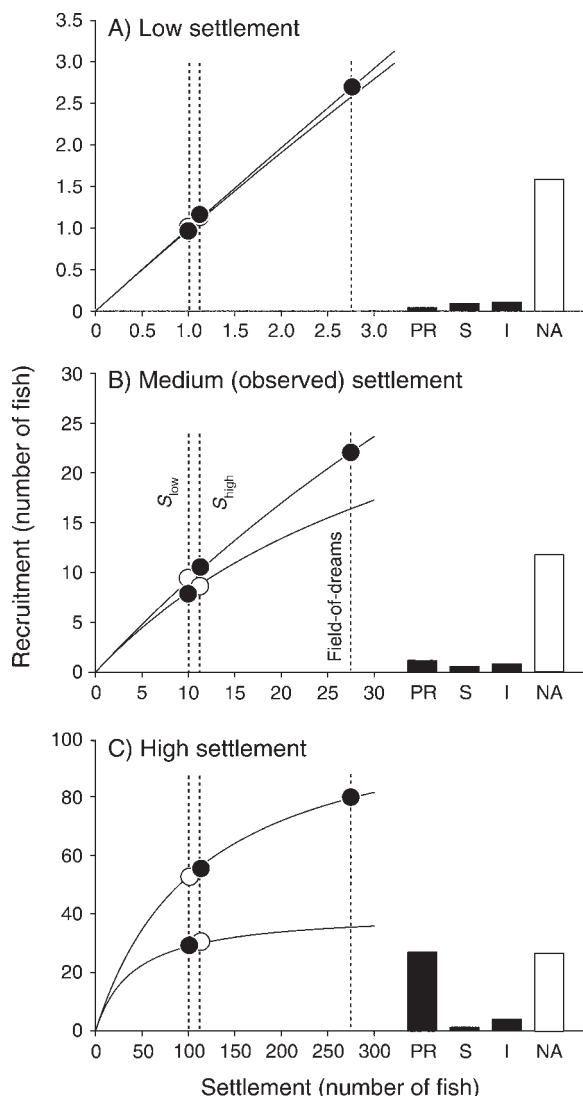


FIG. 2. Extrapolated effects of adjusted habitat availability on settlement and recruitment of *Dascyllus flavicaudus*, under: (A) low (10% of observed), (B) medium (observed), and (C) high (10 times observed) levels of settlement. For each panel, recruitment functions are given for low (lower curve) and high (upper curve) adjusted habitat availability (where $h_{\text{high}}/h_{\text{low}} = 2.75$). Three settlement intensities (total number of fish to an array) are indicated with vertical dashed lines: settlement at low habitat availability (S_{low}), settlement at high habitat availability (S_{high}), and expected settlement in the absence of redirection (i.e., if settlement is proportional to habitat availability: the Field-of-Dreams Hypothesis). In all panels, $S_{\text{high}} = 1.17S_{\text{low}}$ (i.e., the observed response of *D. flavicaudus* to the increase in habitat: Fig. 1B). Solid circles give the expected recruitment and settlement under low and high habitat availability with observed settlement, and under high habitat assuming proportional settlement. Open circles give the expected recruitment and settlement if only density dependence (or only settlement) is affected by habitat addition. The difference in recruitment between the two most extreme recruitment levels gives the potential effect on recruitment under the Field-of-Dreams Hypothesis (i.e., the expected increase in recruitment if a 2.75 increase in habitat led to a 2.75 \times increase in settlement). That potential can be divided into four components (histograms on

Dreams Hypothesis). In our study, increased habitat availability led to a significant increase in the total number of settlers, but the magnitude of this increase was small (36% averaged across focal species) relative to the increase in habitat (175% or 500%). This small increase in settlement led to a greater increase in projected recruitment (based on field data and the habitat-modified Beverton-Holt model), because propagule redirection reduced the intensity of density dependence. Thus, increased habitat increased production in two ways: first, by increasing the colonization rate (i.e., more habitat led to more colonists), and second, by decreasing density at other sites and thus reducing the effect of density dependence (because of propagule redirection). This effect is likely general, although the overall effect and relative importance of increased settlement (Field-of-Dreams effect) vs. relaxed density dependence (via the Propagule Redirection effect) will likely depend upon system properties such as propagule supply, connectivity, temporal scale of reproduction (and overlap among cohorts), and the nature of interactions between colonists.

Propagule redirection: connectivity and temporal and spatial scales

Variation in habitat availability, and predictions from different hypotheses (such as Field-of-Dreams and Propagule Redirection), will influence three processes that manifest over different time scales. The first step is initial colonization, which in our system was represented by settlement. The Field-of-Dreams and Propagule Redirection hypotheses differ in their predictions about colonization, but the implications of these hypotheses depend upon subsequent dynamics that play out over longer time scales, both within and between generations.

Postcolonization survival translates settlement into recruitment. Our recruitment model incorporated effects of density dependence within cohorts, but over longer terms, as multiple cohorts potentially build up at a site, effects among cohorts may also become important. Thus, recruitment dynamics will be affected by how many cohorts occupy a local site, and the patterns of density dependence among those groups. For example, the Field-of-Dreams hypothesis works well even in the presence of propagule redirection if within-cohort density dependence is sufficiently strong (e.g., Fig. 2C). Similarly, if multiple cohorts co-occur and interact with one another, then Field-of-Dreams may also be realized even if there is strong propagule redirection. Accumulation of cohorts will intensify density dependence in a

← the right side), the first three of which comprise the realized effect: Propagule Redirection (PR), which ameliorates density-dependent mortality, increased settlement (S), the interaction between density dependence and settlement (I), and the remainder (NA), which represents the potential recruitment that is not achieved.

way analogous to our result for high settlement levels (Fig. 2C): if the combined effects of settlers and prior residents saturate the system, then there will be little effect of redirection. Conversely, weak density dependence (within and among cohorts) will cause Field-of-Dreams to fail in the presence of propagule redirection (e.g., Fig. 2A).

Finally, dynamics across generations will define how recruitment at one time step affects future changes in local population sizes. In relatively closed systems, the dynamics of a newly colonized patch will be largely internally driven (i.e., future “colonists” will be internally produced). In such a case, local reproduction will increase local abundance, and this will eventually lead to a local abundance that is proportional to local habitat availability. Thus, although the initial short-term response may be affected by propagule redirection, the long-term response will not. In open systems (with no self-recruitment) the within-patch dynamics will always depend on the external supply of propagules. Thus, if local abundance is reduced due to propagule redirection it cannot be regained by future within-patch dynamics.

Propagule redirection may be important even if its effects are transient, because it can influence the time scale for recovery of ecosystems following a disturbance or in response to seasonally regenerating habitats. Populations that experience frequent disturbance may never reach their full production potential if the projected time to a fully recovered patch through slow natural colonization is long (relative to timing of disturbances: e.g., Robertson 1996). Alternatively, if colonization of newly available habitat is high and does not deplete colonization of other habitats, then the overall system may recover from disturbances quickly. Thus, it may not be as important to know whether or not increased habitat availability will *eventually* reach its full potential, but at what time scale new habitats will “fill up.”

Propagule redirection: unbridled conjecture

Propagule redirection can increase recruitment if it ameliorates negative effects of density (e.g., via predator attraction or competition). In contrast, if density has beneficial effects (e.g., predator dilution, mate action [Sweatman 1985, White et al. 2010]), then propagule redirection could slow (or even negate) positive responses to increased habitat availability. This may be pronounced in systems with biogenic habitat if the habitat benefits from animals that occupy the habitat. As in ant-plant or flower-pollinator mutualisms (Mitchell et al. 2009), some invertebrates and fishes (including *D. flavicaudus*) use coral for structure and/or food and also provide positive benefits to the coral (Goldshmid et al. 2004, Stewart et al. 2006, Holbrook et al. 2008). If there is propagule redirection (Fig. 1A), then corals effectively “compete” for colonists of their mutualists. This may have led to the evolution of signaling systems (between corals and their symbionts) that have not yet

been fully appreciated by marine scientists. Indeed, recent studies demonstrate that many marine organisms can locate and orient to cues produced by their habitat (Lecchini 2004, Simpson et al. 2005). Thus, some marine systems may be more analogous to co-evolved terrestrial plant pollinators than previously thought. Indeed, because many habitats are biogenic, future studies should incorporate propagule redirection and the effects of changing densities on the dynamics of the colonists as well as their habitat.

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LITERATURE CITED

- Barel, C. D. N., et al. 1985. Destruction of fisheries in Africa's lakes. *Nature* 315:19–20.
- Bellwood, D. R., T. P. Hughes, C. Folke, and M. Nystrom. 2004. Confronting the coral reef crisis. *Nature* 429:827–833.
- Bruno, J. F., E. R. Selig, K. S. Casey, C. A. Page, B. L. Willis, C. D. Harvell, H. Sweatman, and A. M. Melendy. 2007. Thermal stress and coral cover as drivers of coral disease outbreaks. *PLoS Biology* 5:1220–1227.
- Carr, M. H., and M. A. Hixon. 1997. Artificial reefs: the importance of comparisons with natural reefs. *Fisheries* 22: 28–33.
- Forrester, G. E. 1990. Factors influencing the juvenile demography of a coral reef fish. *Ecology* 71:1666–1681.
- Gaines, S. D., and J. Roughgarden. 1987. Fish in offshore kelp forests affect recruitment to intertidal barnacle populations. *Science* 235:479–481.
- Galzin, R., and J. P. Pointer. 1985. Moorea Island, Society Archipelago. Pages 73–102 in *Fifth International Coral Reef Symposium*, Papeete, Tahiti.
- Gardner, T. A., I. M. Cote, J. A. Gill, A. Grant, and A. R. Watkinson. 2005. Hurricanes and Caribbean coral reefs: impacts, recovery patterns, and role in long-term decline. *Ecology* 86:174–184.
- Goldshmid, R., R. Holzman, D. Weihs, and A. Genin. 2004. Aeration of corals by sleep-swimming fish. *Limnology and Oceanography* 49:1832–1839.
- Hamner, W. M., M. S. Jones, J. H. Carleton, I. R. Hauri, and D. M. Williams. 1988. Zooplankton, planktivorous fish, and water currents on a windward reef face—Great Barrier Reef, Australia. *Bulletin of Marine Science* 42:459–479.
- Hanski, I. 1998. Metapopulation dynamics. *Nature* 396:41–49.
- Holbrook, S. J., A. J. Brooks, R. J. Schmitt, and H. L. Stewart. 2008. Effects of sheltering fish on growth of their host corals. *Marine Biology* 155:521–530.
- Holbrook, S. J., and R. J. Schmitt. 2002. Competition for shelter space causes density-dependent predation mortality in damselfishes. *Ecology* 83:2855–2868.
- Jones, G. 1997. Relationships between recruitment and post-recruitment processes in lagoonal populations of two coral reef fishes. *Journal of Experimental Marine Biology and Ecology* 213:321–346.
- Lecchini, D. 2004. Experimental assessment of sensory abilities of coral reef fish larvae in the detection of their settlement location. *CR Biologies (Comptes Rendus Biologies)* 327:159–171.

- MacArthur, R. H., and E. O. Wilson. 1967. The theory of island biogeography. Princeton University Press, Princeton, New Jersey, USA.
- Mitchell, R. J., R. J. Flanagan, B. J. Brown, N. M. Waser, and J. D. Karron. 2009. New frontiers in competition for pollination. *Annals of Botany* 103:1403–1413.
- Munday, P. L., and S. K. Wilson. 1997. Comparative efficacy of clove oil and other chemicals in anaesthetization of *Pomacentrus amboinensis*, a coral reef fish. *Journal of Fish Biology* 51:931–938.
- Osenberg, C. W., C. M. St. Mary, R. J. Schmitt, S. J. Holbrook, P. Chesson, and B. Byrne. 2002a. Rethinking ecological inference: density dependence in reef fishes. *Ecology Letters* 5:715–721.
- Osenberg, C. W., C. M. St. Mary, J. A. Wilson, and W. J. Lindberg. 2002b. A quantitative framework to evaluate the attraction–production controversy. *ICES Journal of Marine Science* 59:S214–S221.
- Palmer, M. A., R. F. Ambrose, and N. L. Poff. 1997. Ecological theory and community restoration ecology. *Restoration Ecology* 5:291–300.
- Peterson, C. H., and R. Black. 1991. Preliminary evidence for progressive sestonic food depletion in incoming tide over a broad tidal sand flat. *Estuarine Coastal and Shelf Science* 32: 405–413.
- Porter, J. W. 1972. Predation by *Acanthaster* and its effect on coral species diversity. *American Naturalist* 106:487–492.
- Rausher, M. D. 1979. Larval habitat suitability and oviposition preference in three related butterflies. *Ecology* 60:503–511.
- Renwick, J. A. A., and F. S. Chew. 1994. Oviposition behavior in Lepidoptera. *Annual Review of Entomology* 39:377–400.
- Resetarits, W. J. 2001. Colonization under threat of predation: avoidance of fish by an aquatic beetle, *Tropisternus lateralis* (Coleoptera: Hydrophilidae). *Oecologia* 129:155–160.
- Resetarits, W. J. 2005. Habitat selection behaviour links local and regional scales in aquatic systems. *Ecology Letters* 8: 480–486.
- Resetarits, W. J., and C. A. Binckley. 2009. Spatial contagion of predation risk affects colonization dynamics in experimental aquatic landscapes. *Ecology* 90:869–876.
- Robertson, D. R. 1996. Interspecific competition controls abundance and habitat use of territorial Caribbean damselfishes. *Ecology* 77:885–899.
- Schmitt, R. J., and S. J. Holbrook. 1996. Local-scale patterns of larval settlement in a planktivorous damselfish: Do they predict recruitment? *Australian Journal of Marine and Freshwater Research* 47:449–463.
- Schmitt, R. J., and S. J. Holbrook. 2007. The scale and cause of spatial heterogeneity in strength of temporal density dependence. *Ecology* 88:1241–1249.
- Silliman, B. R., J. van de Koppel, M. D. Bertness, L. E. Stanton, and I. A. Mendelsohn. 2005. Drought, snails, and large-scale die-off of southern U.S. salt marshes. *Science* 310: 1803–1806.
- Simpson, S. D., M. Meekan, J. Montgomery, R. McCauley, and A. Jeffs. 2005. Homeward Sound. *Science* 308:221.
- Skole, D., and C. Tucker. 1993. Tropical deforestation and habitat fragmentation in the Amazon: satellite data from 1978 to 1988. *Science* 260:1905–1910.
- Stav, G., L. Blaustein, and J. Margalith. 1999. Experimental evidence for predation risk sensitive oviposition by a mosquito, *Culiseta longiareolata*. *Ecological Entomology* 24:202–207.
- Stewart, H. L., S. J. Holbrook, R. J. Schmitt, and A. J. Brooks. 2006. Symbiotic crabs maintain coral health by clearing sediments. *Coral Reefs* 25:609–615.
- Sweatman, H. P. A. 1985. The influence of adults of some coral reef fishes on larval recruitment. *Ecological Monographs* 55: 469–485.
- White, J. W., J. F. Samhouri, A. C. Stier, C. L. Womald, S. L. Hamilton, and S. A. Sandin. 2010. Synthesizing mechanisms of density dependence in reef fishes: behavior, habitat configuration, and observation scale. *Ecology* 91:1949–1961.

APPENDIX A

Experimental design and map of study site (*Ecological Archives* E091-202-A1).

APPENDIX B

Recruitment model, parameterization, and projections (*Ecological Archives* E091-202-A2).

APPENDIX C

Local scale variation in settlement relative to prevailing currents (*Ecological Archives* E091-202-A3).

APPENDIX D

Partitioning the effect on recruitment (*Ecological Archives* E091-202-A4).

APPENDIX E

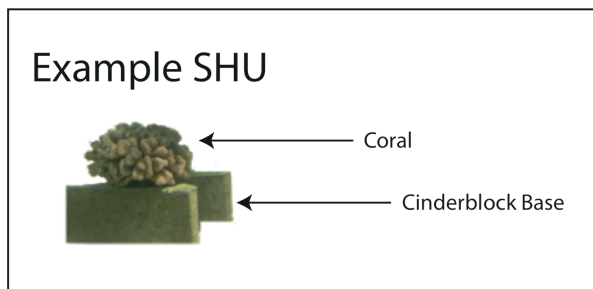
Results for each species (*Ecological Archives* E091-202-A5).

APPENDICES

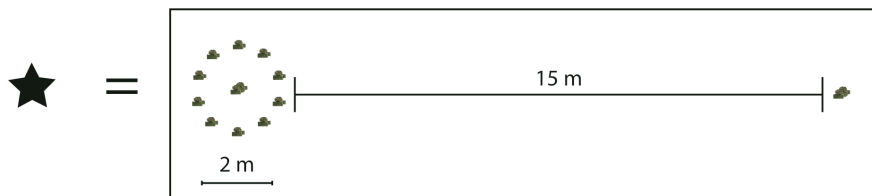
Appendix A: Experimental Design and Map of Study Site

i) A standardized habitat unit (SHU); ii) the spatial arrangement of a single replicate block (consisting of 14 SHUs: two alone, two in the center of the circle, ten on the edge of the circle), showing the treatments with (left hand side) and without (right hand side) neighboring SHUs; and iii) a map of the north shore of Moorea, showing the approximate locations of the 12 experimental blocks.

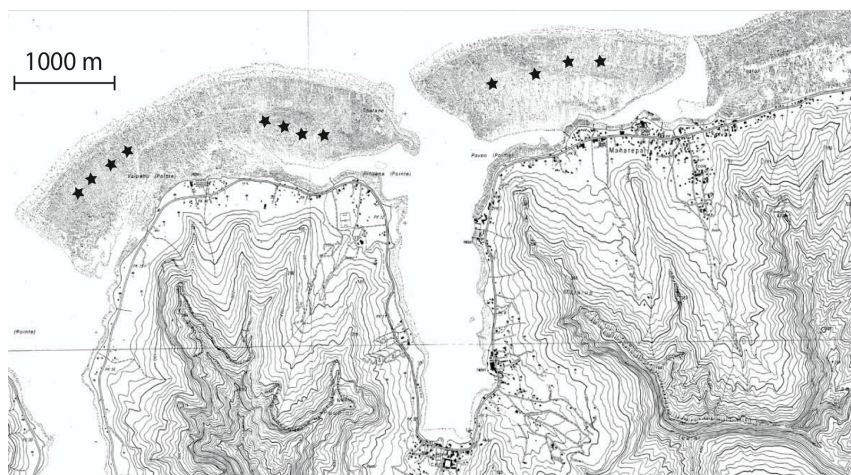
i)



ii)



iii)



Map Courtesy of Moorea. Berkeley.edu

Appendix B: Recruitment Model, Parameterization and Projections

We started with the standard formulation of the Beverton-Holt recruitment function (e.g., Osenberg et al. 2002):

$$D_t = \frac{e^{-\alpha t} D_0}{1 + \frac{\beta(1 - e^{-\alpha t}) D_0}{\alpha}}, \quad (\text{A1})$$

where D_t is the density of recruits to a site (i.e., settlers who survive t days); D_0 is the initial density of settlers to that site, α is the per capita density-independent mortality rate, and β is the per capita density-dependent mortality rate. Because we were interested in the separate effects of habitat (h) and numbers (N), per se, we re-expressed density ($D = N/h$) to obtain:

$$N_t = \frac{e^{-\alpha t} N_0}{1 + \frac{\beta(1 - e^{-\alpha t}) \frac{N_0}{h}}{\alpha}}, \quad (\text{A2})$$

Notice that the parameters have the same meanings and units in A2 as in A1 (i.e., α : 1/day, t : day; β : corals / fish / day). We then parameterized the Beverton-Holt model (fitting α and β) using data from previous studies of density dependence conducted on one of the focal species, *D. flavicaudus* (Holbrook and Schmitt 2002, Schmitt and Holbrook 2007). The corals used by Schmitt and Holbrook (2007) were approximately twice the areal dimension of ours, so we parameterized the model by doubling their coral numbers to reflect the difference in areal cover of corals in the two studies (i.e., h is expressed as the number of corals of the size used in our study).

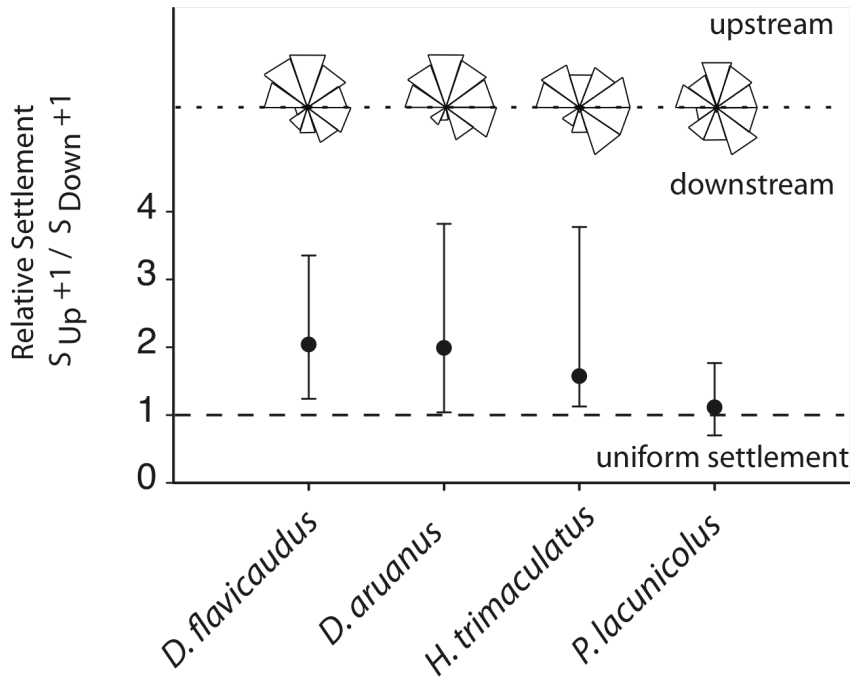
Recruitment patterns of newly settled reef fishes, including *D. flavicaudus*, are driven by high mortality within the first few days following their arrival to the reef (Almany and Webster 2006, Schmitt and Holbrook 2007). We therefore calculated the expected number of survivors after two days on the reef (i.e., “recruitment”). This time scale is the same as the one used to generate the data with which we parameterized the model (Schmitt and Holbrook 2007). We assumed that density dependence was most strong among fish that settled within 2 weeks of one another, and therefore used the total number of settlers per coral observed in our 28 day (i.e., + 2 weeks) field study. We then compared the levels of settlement and recruitment in a system with low habitat availability ($h = 1$) to one with high habitat availability ($h = 2.75$); this 2.75 fold difference in habitat availability is within the range of habitat losses observed in coral reef systems (e.g. Connell et al. 2004, Bruno and Selig 2007). Expected settlement intensities were taken from our field experiment (low vs. high habitat availability) for *D. flavicaudus*, but we also examined the results under much lower and higher settlement intensities to better generalize the results.

Appendix C: Local Scale Variation in Settlement Relative to Prevailing Currents

The circular array provided additional information on propagule redirection because the current at our sites was predominantly unidirectional, coming in over the reef crest to the north of the study sites and flowing over the sites before exiting the lagoon through deeper channels (Hench et al. 2008). Thus, we compared cumulative settlement to the SHUs in the “upstream” (i.e., most northerly) half of the circle to the “downstream” half. This statistical approach was similar to the previous two, except that we added 1 to the number of settlers due to the presence of zeros at some downstream sectors: $\ln[(S_{\text{up}}+1)/(S_{\text{down}}+1)]$. For all analyses, we used paired t-tests, comparing $\ln(S)$ (or $\ln(S+1)$) for the two treatments ($n = 12$ blocks).

Within the circular grid of neighboring SHUs, there was a spatial pattern that further suggested propagule redirection (i.e., settlement shadows, sensu Jones 1997). Settlement to the five upcurrent SHUs was 29-165% greater than to the five downcurrent SHUs for all four focal species, although this pattern was significant for only two of the four species (Fig. C1): *D. flavicaudus* ($t_{11} = 2.80$; $P = 0.017$), *H. trimaculatus* ($t_{11} = 2.65$; $P = 0.023$), *D. aruanus* ($t_{11} = 2.06$; $P = 0.062$), and *P. lacunicolus* ($t_{11} = 0.44$; $P = 0.67$).

Figure C1 The spatial distribution of fish on the circular array of ten SHUs in the high habitat treatment. Wedge size represents the relative magnitude of settlement averaged across the twelve sites. Points give the back-transformed mean (\pm 95%CI) relative settlement to the five upcurrent SHUs vs. the five downcurrent SHUs: $(S_{\text{up}}+1)/(S_{\text{down}}+1)$, $n=12$ replicate arrays.



Appendix D: Partitioning the Effect on Recruitment

Habitat addition can increase recruitment and local density via two mechanisms: 1) increased settlement (because there is more habitat to attract larvae: i.e., Field-of-Dreams); and 2) reduced density dependence (because settlers are spread out among more habitat: i.e., Propagule Redirection). To partition the potential change in recruitment to these two components, we explored six recruitment levels defined by the two recruitment functions (low vs. high habitat availability: i.e., $h = 1$ or $h = 2.75$ in Equation 1 or A2) and three settlement levels (observed under low habitat availability, observed under high habitat availability, and the expected settlement if settlement was proportional to habitat availability). We then used these recruitment levels to define:

1) the *potential* recruitment effect predicted by the Field-of-Dreams Hypothesis (i.e., the difference in recruitment predicted under high habitat vs. low habitat availability assuming settlement was proportional to habitat availability), and

2) the *realized* recruitment effect (i.e., the difference in predicted recruitment under high vs. low habitat availability given observed settlement responses). We partitioned the realized effect into that attributable to:

- a) propagule redirection combined with reduced density dependence (i.e., the change in recruitment when habitat is increased but settlement remains at ambient levels);
- b) increased settlement only (i.e., the change in recruitment when settlement is increased but density dependence is unaffected); and
- c) the interaction between settlement augmentation and density dependence (arising from non-linearities in the recruitment function).

The difference between the potential and the realized effects is the amount of recruitment under the Field-of-Dreams Hypothesis that was not achieved. The results of this partitioning are summarized in Figure 2.

Appendix E. Results for Each Species.

Table E1. Settlement of all recorded species of fish summed over 12 sites and 28 days. ‘Circle’ refers to 10 neighboring SHUs on the outer rim of the “high habitat availability” treatments; Focal refers to the two focal SHUs in each treatment, either surrounded by neighboring corals (high habitat availability) or not (low habitat availability treatment).

Family (common name) <i>Genus species</i>	Total Settlement		
	High Habitat Availability		Low Habitat Availability
	Circle	Focal	Focal
Holocentridae (Squirrelfishes)			
<i>Neoniphon sammara</i>	0	3	2
Apogonidae (Cardinalfishes)			
<i>Apogonichthys ocellatus</i>	1	3	0
Pomacentridae (Damsel-fishes)			
<i>Chromis viridis</i>	2	0	15
<i>Dascyllus aruanus</i>	94	35	87
<i>Dascyllus flavicaudus</i>	111	29	120
<i>Pomacentrus pavo</i>	4	1	5
Labridae (Wrasses and Parrotfishes)			
<i>Halichoeres trimaculatus</i>	72	28	61
<i>Pseudocheilinus hexataenia</i>	1	0	0
<i>Scarus oviceps</i>	4	1	3
<i>Chlorurus sordidus</i>	2	1	0
Gobiidae (Gobies)			
<i>Eviota sp.</i>	5	0	3
<i>Paragobiodon lacunicolus</i>	39	28	53
<i>Paragobiodon modestus</i>	7	0	2
<i>Priolepis squamogena</i>	4	0	3
Acanthuridae (Surgeonfishes)			
<i>Acanthurus olivaceus</i>	0	2	0
<i>Acanthurus triostegus</i>	5	16	6
Total	351	147	359

Literature Cited

- Almany, G. R. and M. S. Webster. 2006. The predation gauntlet: early post-settlement mortality in reef fishes. *Coral Reefs* **25**:19-22.
- Bruno, J. F. and E. R. Selig. 2007. Regional decline of coral cover in the Indo-Pacific: timing, extent, and subregional comparisons. *PLoS ONE* **2**:e711, 711-718.
- Connell, J. H., T. E. Hughes, C. C. Wallace, J. E. Tanner, K. E. Harms, and A. M. Kerr. 2004. A long-term study of competition and diversity of corals. *Ecological Monographs* **74**:179-210.
- Hench, J. L., J. J. Leichter, and S. G. Monismith. 2008. Episodic circulation and exchange in a wave-driven coral reef and lagoon system. *Limnology and Oceanography* **53**:2681-2694.
- Holbrook, S. J. and R. J. Schmitt. 2002. Competition for shelter space causes density-dependent predation mortality in damselfishes. *Ecology* **83**:2855-2868.
- Jones, G. 1997. Relationships between recruitment and postrecruitment processes in lagoonal populations of two coral reef fishes. *Journal of Experimental Marine Biology and Ecology* **213**:321-346.
- Schmitt, R. J. and S. J. Holbrook. 2007. The scale and cause of spatial heterogeneity in strength of temporal density dependence. *Ecology* **88**:1241-1249.