Quantifying the manageability of pollination networks in an invasion context

Supporting Information

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Contents

1   Finding a complex network's matchings

Our approach to find the minumum number of driver nodes relies on finding maximum matchings and maximal cardinality matchings. We start with a directed network in which the direction of the link represents the direction of control (Figure S1a). We then construct an alternative representation of the directed network in which each node of the directed network is represented by two nodes that indicate their outgoing and incoming links respectively (Figure S1b). Finding a maximum matching in this alternative representation is equivalent to finding the largest possible set of edges in which one node on the left-hand side is connected to at most one node on the right-hand side. To find the maximum matching we use the push-relabel algorithm implemented in `max_bipartite_matching` in the R package igraph 1.0.1 (Csardi & Nepusz 2006). Once we have the matching (shown in the Figure S1b) it is then easy to identify the roles of each node in this representation: nodes on the left-hand side that are connected to a matched (purple) link are superior while those connected to a matched link on the right-hand side are matched. This information can then be mapped back to the original representation to identify the control paths and the driver nodes in the network (Figure S1c). Figure S1d–f illustrate this approach for a network with bidirectional links.

The algorithm implemented in `max_bipartite_matching`, however, is only able to find **one** of possibly many maximum matchings in a network. Though one maximum matching is enough to calculate n_D and
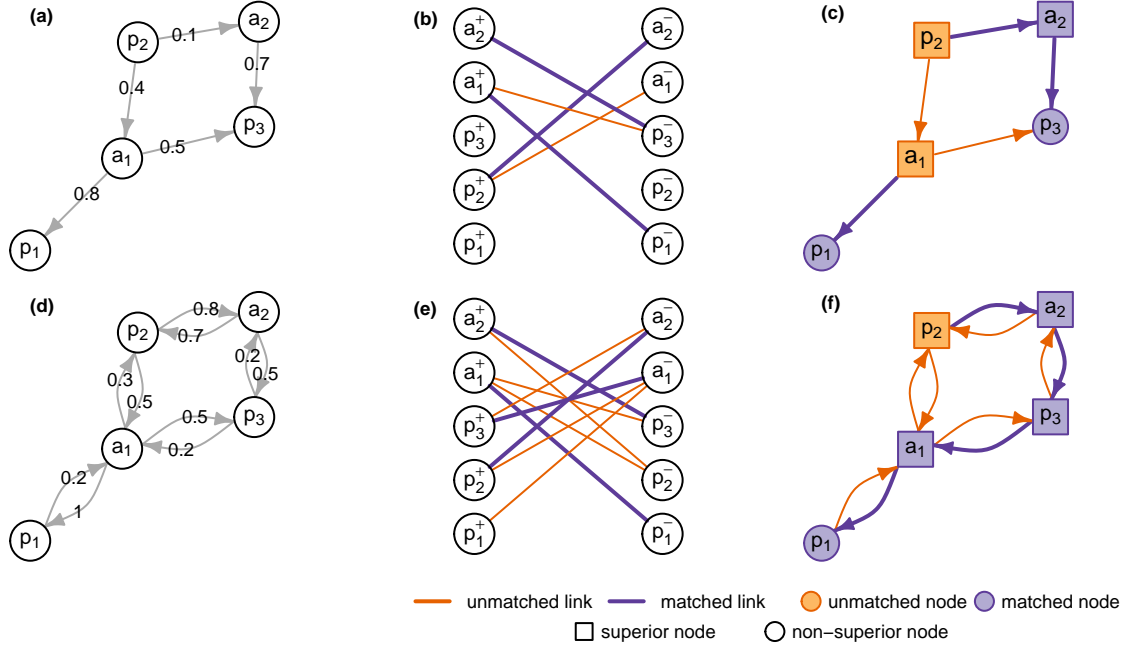
Figure S1: **Finding a maximum matching in a complex network**. (a & d) Directed networks that indicate the direction of control between species. (b & e) Alternative bipartite representations of the directed networks. (c & f) The matchings in the bipartite representation mapped back to the original network.

hence to provide indication of the manageability of a community, it is not sufficient to estimate the role of individual species. To do that, we need to calculate all possible maximum matchings (or, equivalently, all maximal cardinality matchings in weighted networks like ours). To do this, we again start from the alternative bipartite representation in Figure S1b and assign an identity to each of the links in the network (shown as numbers in Figure S2a). We will call this bipartite representation P. We then construct the line graph of the alternative bipartite representation $L(P)$ (Figure S2b). Each node in $L(P)$ represents a link in P and these are connected to each other only if and only if they share a common node in P. We then calculate H, the complement graph of $L(P)$ and identify all of its maximal cliques (Figure S2c). Here some extra definitions are necessary. First, H is a graph with the same nodes as $L(P)$ but that has a link between two nodes if and only if there is not a link in $L(P)$. Second, a clique is a subset of nodes such that all pairs of them are linked. Lastly, a maximal clique is a clique such that there are no cliques composed of more nodes (Gutin 2013). In this example, there are two maximal cliques: the one formed by 1, 3 and 5, and the one formed by 2, 3 and 5. The final step is then to map these cliques onto the original network to obtain all possible maximal cardinality matchings as shown in Figure 1 in the main text.

In the main text, we show all maximal cardinality matchings for a simple example network. To further illustrate our methodology here, we also show the approach for the smallest of our empirical networks, the uninvaded network at site 10 (Table S1; Figure S3). The largest component of this network is composed of 16 species of which two are non-invasive plants and the other 14 are pollinators. The one-to-one relationship between matched and superior nodes implies that in order to achieve full network controllability, most
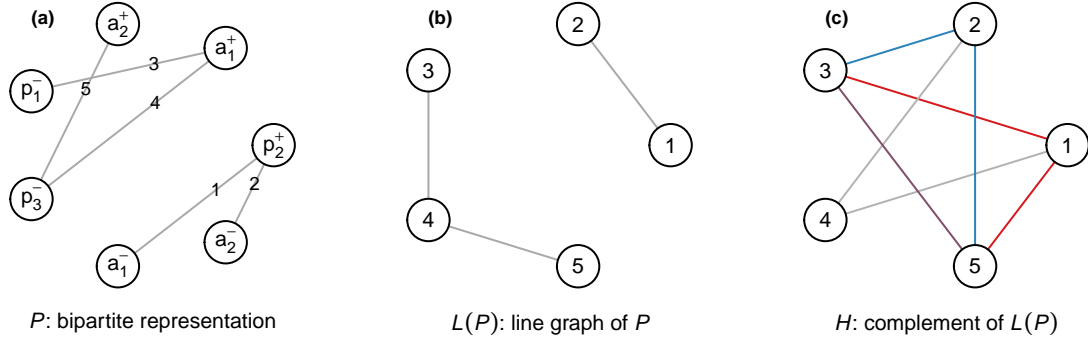
Figure S2: **Finding all possible maximal cardinality matchings**. (a) Alternative bipartite representation of the directed network in Figure S1a. (b) Line graph of the network in a. (c) Complement of the network in b. The two maximal cliques are shown in red and blue.

pollinators would be unmatched, and hence are classified as driver nodes that require external intervention. At the same time, both plants in the community, *Heracleum sphondylum* and *Rubus fructicosus*, and one of the pollinators, *Orthotylus/Lygocorus*, tend to be classified as superior nodes.

2 Control profiles

Ruths & Ruths (2014) proposed a heuristic-based method to provide insight into what might be required to control a network. The method is based on the idea that the reasons behind a node being classified as a driver node can be precisely identified, and that the relative contribution of these reasons can be used to characterise the control profile of the network. Specifically, nodes can be deemed to be driver nodes for any three reasons:

1. **Because it is a source node**. Source nodes are nodes that exclusively have outgoing links. Therefore, other nodes cannot control it and it must be externally controlled instead.

2. **Because of an external dilation**. Dilations occur whenever a control signal needs to branch out in order to reach all nodes in a network. One kind of dilation, which arises due to a surplus of sink nodes (those that only have incoming links) is called external dilation. This concept might be easier to understand by looking at Figure S3c. The source node *a* branches out into nine paths, but it can only control one of them. We say then that the sink nodes *d* to *k* become driver nodes because of an external dilation.

3. **Because of an internal dilation**. The remaining dilations (those that do not arise from a surplus of sink nodes) are called internal dilations. Node *j* in Figure S3c and the driver nodes in Figure S7 are examples of this category.

Mathematically, the number of driver nodes D is the sum of the number of source nodes, the number of external dilation points, and the number of internal dilation points, $D = N_s + N_e + N_i$. A network can
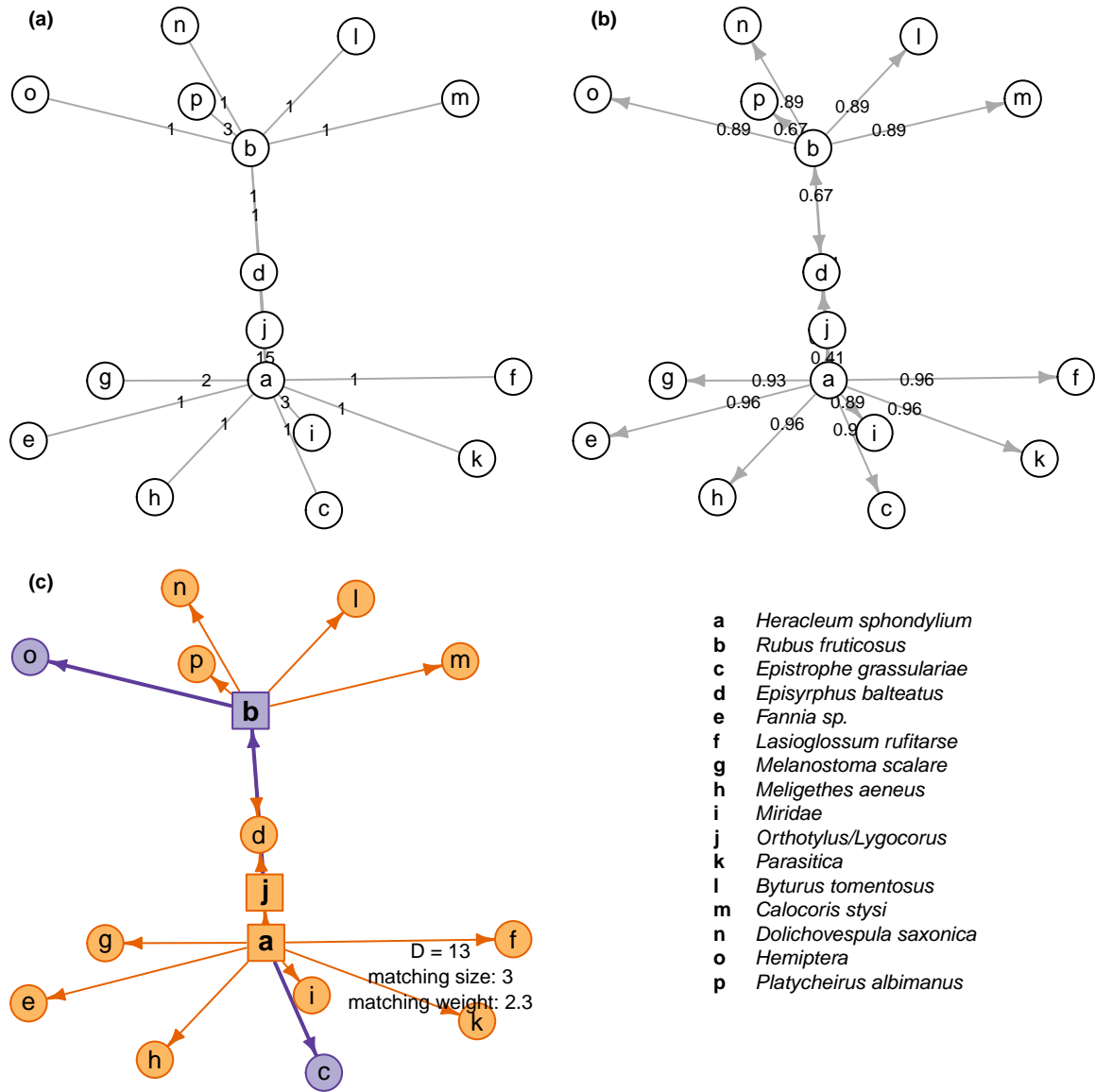
Figure S3: **Illustration of the procedure with an empirical network**. (a) The initial visitation network. (b) The directed network in which the direction of control is determined based on the mutual dependences and asymmetries. (c) One of the possible maximum matchings calculated using the procedures illustrated in the Figures S1 and S2. Legend as in Figure S1

then be characterised by the relative proportion of the three kind of controls

$$\eta_s + \eta_e + \eta_i = 1,$$

where $\eta_s = N_s/D$, $\eta_e = N_e/D$, and $\eta_i = N_i/D$. These terms can be calculated with relative ease. D can be calculated using a maximum matching algorithm. N_s can be calculated by inspecting the network's degree distribution. $N_e = max(0, N_t - N_s)$, where N_t is the number of sink nodes in the network. And we can solve for N_i once we have all other information.

Upon applying this framework to our networks with directions determined by asymmetry we, identify that they are all external dilation dominated (Figure S4). Ruths & Ruths (2014) explains that "external-dilation-dominated networks exhibited strong aspects of top-down control". Because of the surplus of sink nodes, an external intervention that is applied to a source node will induce a correlated response among their subordinate species. For example, an increase in the abundance of *Heracleum sphondylum* (Figure S3) is likely to induce an increase in the abundance of the pollinators it interacts with. Therefore it can be expected that in order to fully control the network it is not sufficient to just apply interventions to the source nodes.

The control profile of our pollination networks provides additional justification for our decision to determine the importance of a species not only by its f_D but also with f_S. Many species that act like control sinks in our network will be classified as a driver node because they are necessary to *fully* control the dynamics of the community as opposed to playing a central role in influencing the abundance of other species in the community.

Using th Ruths & Ruths (2014) approach also highlights the importance of appropiately selecting the direction of control (see Section 6). A network with bidirectional links weighted by the mutual dependences would have zero source nodes and zero sink nodes. Therefore, the control profile of such a network would have been completely uninformative, as it would just highlight that the proportion of internal dilations is $\eta_i = 1$.

3   Properties of empirical networks

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The networks studied had species richness ranging between 19 and 87 (16–86 when considering only the largest component in each network). As shown by the network asymmetry *AS* (Blüthgen *et al.* 2007), the
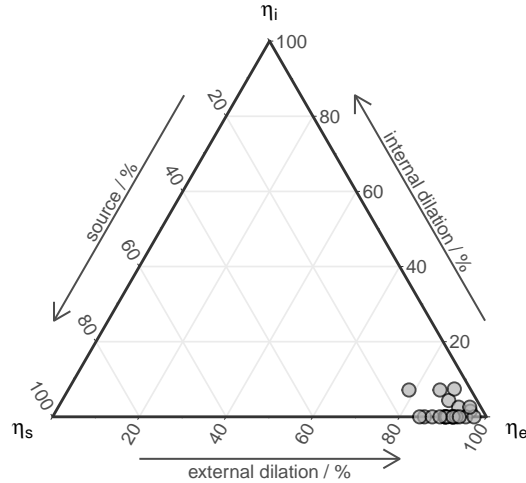
Figure S4: **Control profiles of the studied networks**. Following Ruths & Ruths (2014), profiles are shown in a ternary plot in which the corners correspond to points (1, 0, 0), (0, 1, 0), and (0, 0, 1) in the three-dimensional (η_s, η_e, η_i) space such that $\eta_s + \eta_e + \eta_i = 1$, and $\eta_s, \eta_e,$ and $\eta_i > 0$.

networks had a low ratio of plants to pollinators overall. Furthermore, the networks had relatively low levels of nestedness (when measured using the quantitative version of the NODF index; Almeida-Neto & Ulrich 2011). Details for each network can be found in Table S1.

4 Visitation as a proxy for species interdependence

Visitation frequency has been shown to be an appropriate surrogate for inter-specific effects in pollination networks (Vázquez *et al.* 2005; Bascompte *et al.* 2006). Nevertheless visitation is not equivalent to pollen deposition and might be insufficient to reflect the dependences of plants on animals and vice versa (Alarcón 2010; King *et al.* 2013). We therefore investigated the effect of calculating the dependences using visitation or pollination effectiveness and importance—two metrics more proximate to plant reproductive success (Figure S5). We did this by comparing *(i)* the manageability of the community and *(ii)* the percentage of interactions that maintained the direction of dependence. To do this, we used data collected by Ballantyne *et al.* (2015) from a low diversity pollination community at a dry lowland heathland in Dorset, UK (50° 43.7'N 2° 07.2'W). First, deposition networks were quantified using the mean single visit deposition—the number of conspecific pollen grains effectively deposited on a virgin stigma during a single visit by a particular animal (Ne'Eman *et al.* 2010; King *et al.* 2013; Ballantyne *et al.* 2015). Second, visitation networks were constructed by counting the visits to flowers during Single Visit Depositions. Finally, pollinator importance networks were constructed as the product of pollinator efficiency and visit frequency.

At a network scale (Figure S5), the driver-node density n_D was consistent among the three weighting

Table S1: **Properties of the analysed plant-pollinator communities**. Invasive plants were *Carpobrotus affine acinaciformis* (car), *Opuntia stricta* (op), and *Impatients grandulifera* (imp). All properties, with the exception of the networks' total species richness (R), correspond to the network's largest component. We show the number of species (n_s), the number of plants (n_p), the number of pollinators (n_a), the network connectance (c), the network asymmetry (AS), and the network nestedness (NODF index). British networks were assembled by Lopezaraiza-Mikel et al. (2007), Spanish were networks assembled by Bartomeus et al. (2008).

site	invader	R	n_s	n_p	n_a	c	AS	NODF	location
1	—	35	35	9	26	0.17	-0.49	8.68	Cap de Creus, Spain
1	car	57	57	10	47	0.17	-0.65	13.27	Cap de Creus, Spain
2	—	40	38	9	29	0.18	-0.53	12.66	Cap de Creus, Spain
2	car	38	38	11	27	0.21	-0.42	15.04	Cap de Creus, Spain
3	—	31	29	6	23	0.22	-0.59	14.30	Cap de Creus, Spain
3	op	33	28	6	22	0.24	-0.57	13.29	Cap de Creus, Spain
4	—	35	35	10	25	0.17	-0.43	12.43	Cap de Creus, Spain
4	car	57	57	14	43	0.14	-0.51	13.70	Cap de Creus, Spain
5	—	35	33	7	26	0.23	-0.58	13.05	Cap de Creus, Spain
5	op	32	32	8	24	0.19	-0.50	10.96	Cap de Creus, Spain
6	—	30	25	7	18	0.23	-0.44	9.77	Cap de Creus, Spain
6	op	37	37	9	28	0.17	-0.51	12.45	Cap de Creus, Spain
7	—	37	30	3	27	0.38	-0.80	24.86	Bristol, United Kingdom
7	imp	57	57	8	49	0.20	-0.72	14.36	Bristol, United Kingdom
8	—	48	43	3	40	0.36	-0.86	6.84	Bristol, United Kingdom
8	imp	87	83	13	70	0.12	-0.69	8.67	Bristol, United Kingdom
9	—	55	53	11	42	0.14	-0.58	13.77	Bristol, United Kingdom
9	imp	86	86	11	75	0.13	-0.74	13.40	Bristol, United Kingdom
10	—	19	16	2	14	0.57	-0.75	13.35	Bristol, United Kingdom
10	imp	54	49	5	44	0.26	-0.80	9.04	Bristol, United Kingdom

schemes (0.33 for deposition, 0.33 for the visitation, and 0.38 for the pollinator-importance network, respectively). The choice of weighting can also have an impact on at the species level. Therefore we calculated f_D and f_S (the frequency at which each species is classified as a driver or superior node, respectively, within the set of all possible maximal cardinality matchings in a network) and calculated it's correlation among all three weighting schemes. Although visitation and deposition produce moderately different results, we found a very strong agreement between the order produced by visitation and importance (Table S2). Finally, we also investigated whether the asymmetry of mutual dependence, which defines the direction of control, was consistent among the three possible weighting schemes. We found again that the direction of the dominant dependence was maintained was consistent for 95% of the interactions weighted by visitation or importance, the two most appropriate metrics for pollinator and plant dependence (Table S2).

Altogether, the evidence supports the idea that visitation is a suitable metric to estimate the mutual dependence of species pairs. First, it is directly related to pollinator foraging. Second, it produces results within our controllability framework that are consistent with plant reproductive success (as estimated by the importance metric).
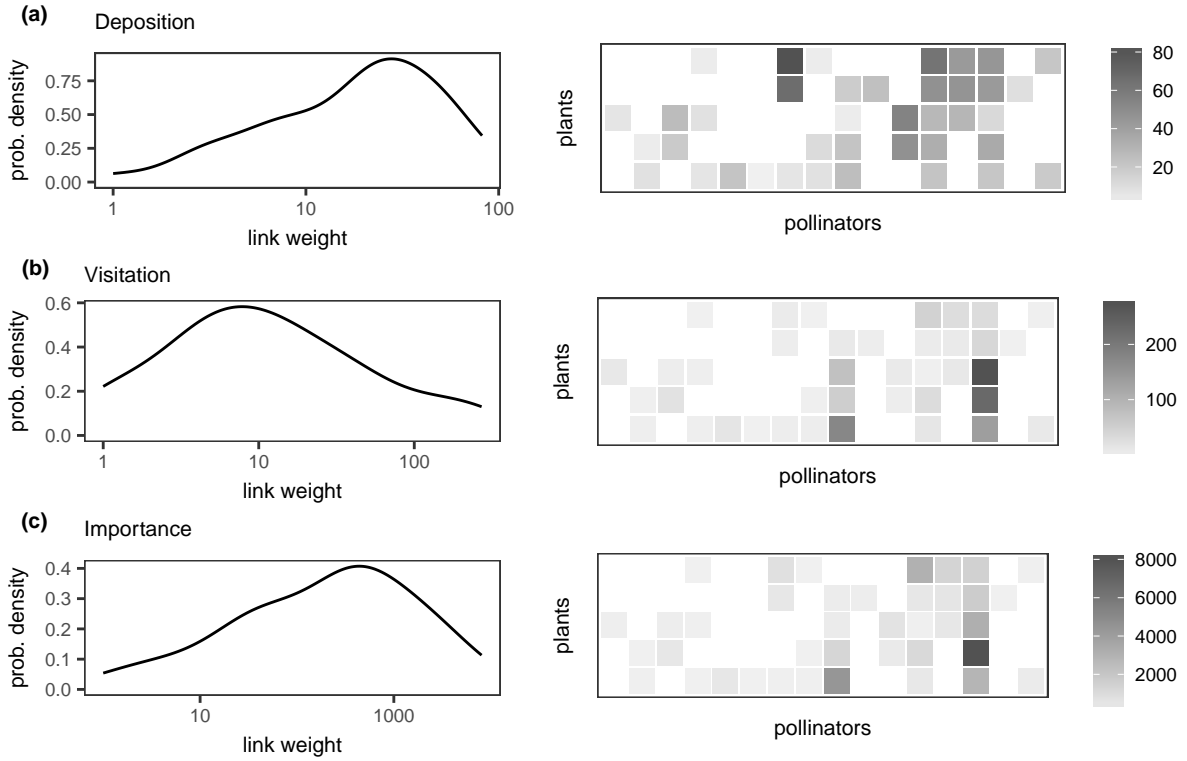
Figure S5: **Distribution of interaction weights for the pollen deposition, visitation and pollinator importance networks**. Note that the x axes in the density plots have been log-transformed.

Table S2: **Agreement among network weighting schemes**. Spearman correlation coeffcients (with p-value) of species' f_D and the percentage of interactions that share the direction of dependence obtained using the three weighting schemes and an unweighted scheme.

	unweighted	deposition	importance	visitation
unweighted	-	0.93 (< 0.001)	0.85 (< 0.001)	0.85 (< 0.001)
deposition	87%	-	0.86 (< 0.001)	0.87 (< 0.001)
importance	77%	74%	-	1 (< 0.001)
visitation	82%	74%	95%	-

5   Sensitivity analysis

Our approach is fundamentaly based on the network structure. Often, the majority of the interactions that make up this structure are weak; in our networks this means that most interactions are formed by a small number of observed pollination visits, and therefore those weak intercations are less conspicuous in the field than strong ones. To strengthen the case of our approach, we therefore evaluated the robustness of our results to simulated sampling limitations.

To do so we removed a portion of the links, for each network, and calculated how three control metrics of the subsampled network compare to those of the full network. Specifically, we calculated the difference between the n_D of the subsampled and the full network, the Spearman correlation between the vectors f_D—which contain the frequency with which each species is classified as a driver species, and the Spearman

correlation between f_S—which contain the frequency with which each species is classified as a superior node. We removed links from the full network with a probability inversely proportional to the interaction weight, and varied the portion of links removed between 50% and 95% at 5% increments. We repeated the procedure ten times for each removed level and each network to obtain a total sample size of $n = 2000$.

Overall, the results of the sensitivity analysis indicate that our approach might still be useful for management even in the absence of complete sampling. Specifically, as the proportion of sampled interaction decreased, the variability of n_D increased, but was overall very similar to that obtained when using all available interactions (Figure S6a). Similarly, the correlation between the relative frequency with which species are classified as a driver (f_D) or as a superior node (f_S) was highly correlated even for *extreme* levels of undersampling (Figure S6b). For instance, we found that with a sampling level of 50%, the mean Spearman correlation coefficient was 0.66 for f_D and 0.76 for f_S. Finally, the results suggest that the relative ranking of species is better conserved in f_S than for f_D throughout all subsampling levels.
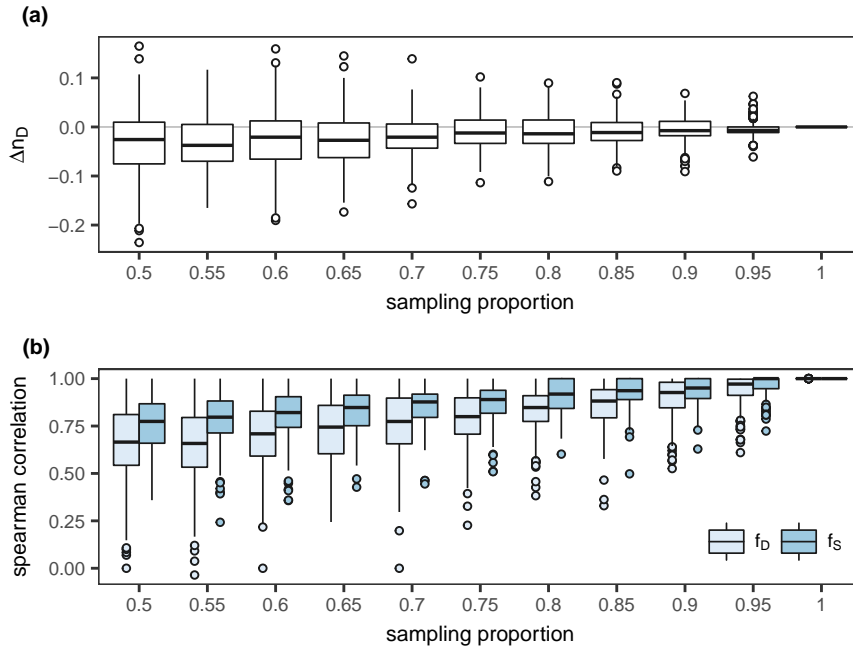

Figure S6: **Sensitivity to subsampling**. (a) The difference between the n_D of the subsampled and the full networks. (b) The spearman correlation of the relative importance of species between the subsampled and the full networks.

6   Chosing the direction of control

Our decision to adopt the direction of asymmetries as the direction of control is coherent with ecological processes and also confers several advantages. First, it is consistent with previous studies, which facilitates

the comparability between our findings and those in other systems explored using structural controllability. Second, it reduces the number of control cycles in the network—which require special treatment in our approach. Third, the reduced number of links facilitates the computation of all possible maximum matchings.

For ilustration, we show all maximal cardinality matchings for an example network in which links are weighted based on the mutual dependences and not asymmetries (Figure S7). We note, however, that using mutual dependences, or indeed any set up that can lead to bidirectional links, is problematic for a couple of reasons. First, having two links between interacting species and using our maximum matching approach to calculate the control configurations, can result in matchings that include cycle inducing links (Figure S8). Strictly speaking these configurations are maximal cardinality matchings in the alternative bipartite configuration (Figure S1e) but not in the directed network (Figure S1f). While bidirectional links are the norm when using mutual dependences, they only appear in the asymmetry-weighted networks whenever the mutual dependences between two species are strictly symmetric. In this special case, we include two links weighted with 0.5 to reflect the fact that the species affect each other to the same extent.
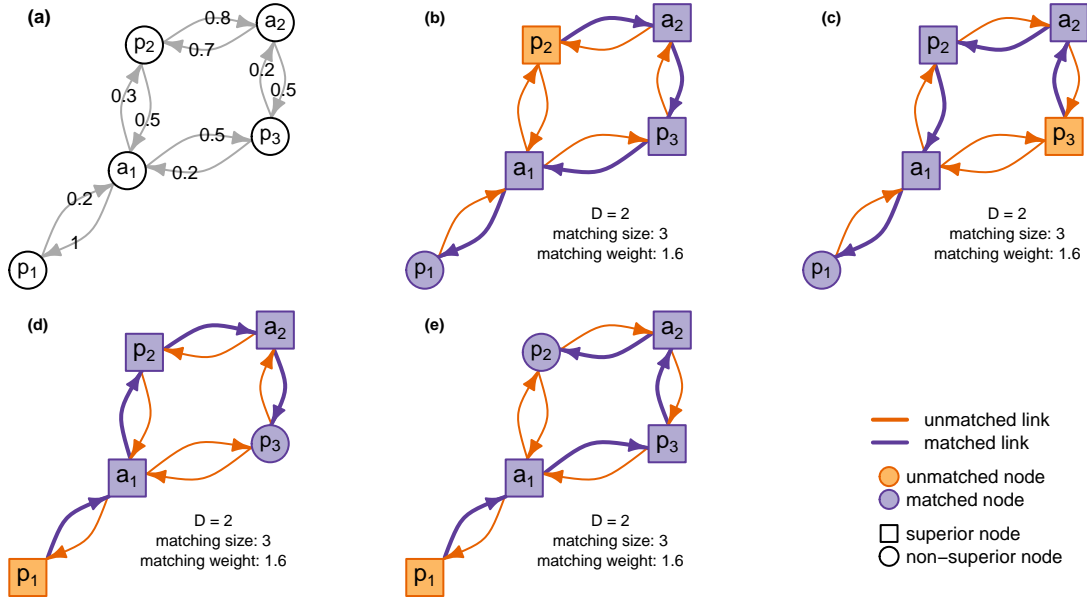

Figure S7: **Matchings of a simple network with bidirectional links**. (a) A network with bidirectional links. (b-e) All possible control configurations.

We consider two ways to tackle this limitation. The first, is to calculate the matching(s) as usual and then, before calculating f_S and f_D, filter out the cycle containing matchings (Figure S8). In the special case that the network has no maximal cardinality matchings that correspond to feasible control configurations because they all include cycles, we reduce the size of the cliques we enumerate in H (Figure S2) and repeat. Instead of calculating the matching(s) for the network with bidirectional cycles, in the second approach, we calculate the matching(s) of 2^n unidirectional versions of the network, where n is the number
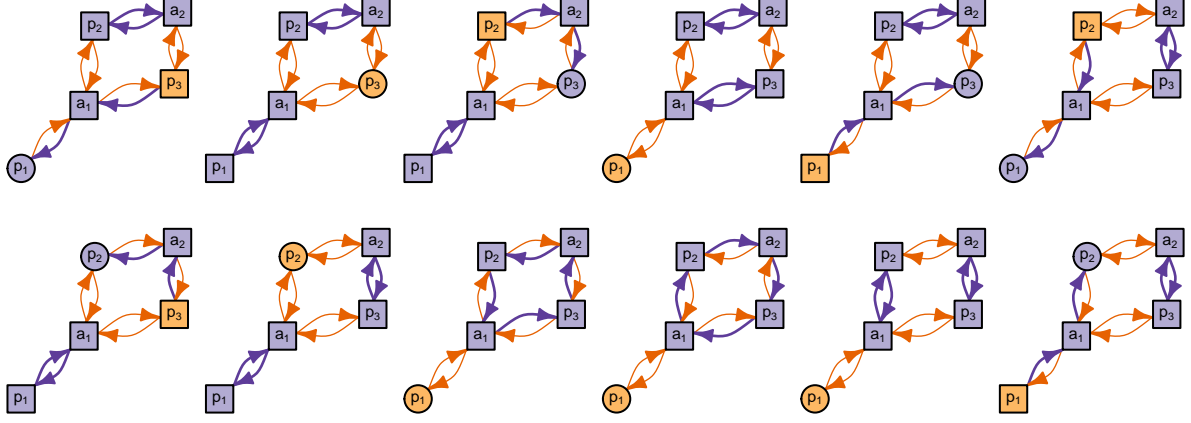
Figure S8: **Cycles in a bidirectional network**. Our approach finds twelve configurations for Figure S7a that are matchings in the alternative bipartite representation but not in the directed network.

of bidirectional cycles in the network (Figure S9). We then average the results of these unidirectional networks to find the overall f_S and f_D for a given species.
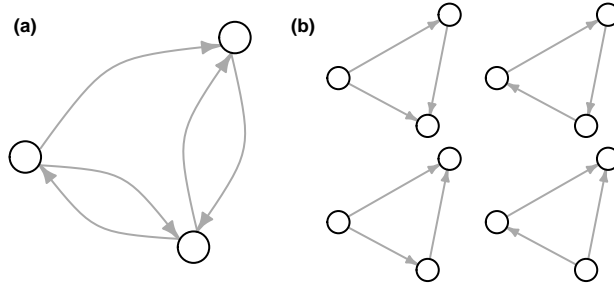

Figure S9: **Dealing with cycles**. The network on the left, which has $n = 2$ bidirectional cycles, can be decomposed into 2^n unidirectional networks.

The number of maximal cardinality matchings increases disproportionally fast with the number of links in a network. For instance, our example network in Figure 1, which is composed of unidirectional links weighted by the asymmetry, has two different maximal cardinality matchings. In contrast, the equivalent network with bidirectional links has 16 different maximal cardinality matchings (Figure S7 and S8). To further illustrate the implications of this growth, some of our empirical networks had approximately ten thousand different maximal cardinality matchings using asymmetries but one hundred million when using mutual dependences. Consequently, the computational cost of either filtering matchings with cycles or calculating the matchings for 2^n networks surpassed the capabilities of our available computational resources and rendered us unable to calculate all possible maximal cardinality matchings when using mutual dependences in most of our bidirectional networks. Nevertheless, as most of our networks weighted by asymmetry had between zero and four cycles, we successfully used the second approach explained in the previous paragraph to calculate the all maximal cardinality matchings for these networks.

We highlight that although the direction of control might affect the f_S and f_D, it doesn't strongly affect

the network's relative driver-node density n_D. We reached this conclusion by calculating the n_D obtained assuming that (a) the direction of control is governed by the asymmetry, (b) bidirectional dependences, (c) that plants depend on pollinators, or that (d) pollinators depend on plants. Although the n_D values were quantitatively different between the approaches (Figure S10), the Spearman correlation among these four options suggested a very high agreement (Table S3).
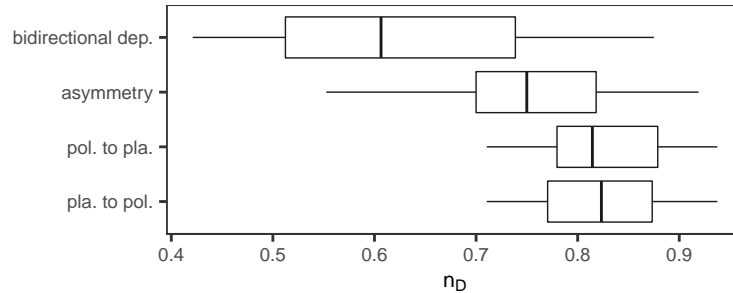

Figure S10: **Absolute differences among asumptions on the direction of control.**

Table S3: **Agreement among assumptions on the direction of control**. Spearman correlation coefficients of the driver-nodes density n_D four different assumptions of the direction of control..

	pol. to pla.	asymmetry	bidirectional dep.
pla. to pol.	1	0.97	0.96
pol. to pla.		0.98	0.96
asymmetry			0.98

7 Statistical methods

7.1 Manageability

7.1.1 Genrealised linear models for n_D

Candidate models were compared using AICc and the relative importance of the explanatory variables was evaluated using the sum of Akaike weights over candidate models that accounted for 95% of the evidence (Burnham & Anderson 2003; Bates *et al.* 2015; Bartoń 2016). Coefficient estimates were averaged following Buckland *et al.* (1997) while confidence intervals were calculated following Lukacs *et al.* (2010). Only models that accounted for 95% of the evidence weight were considered to quantify the relative importance of the fixed effects.

7.1.2 Network randomisations

Randomisations that maintained the degree of the visitation networks were generated using the quasiswap-count algorithm and the function `commsim` in vegan 2.3-3 (Oksanen *et al.* 2016). After generating the

randomised networks, we then calculated the mutual dependences and interaction asymmetries of each and determined n_D using our maximum-matching framework. Finally, we calculated the average rank (akin to a p-value) of n_D for each empirical network compared to the corresponding values of each set of 999 randomisations (Veech 2012).

Similar as in the previous null model, to evaluate the effect of the direction of the asymmetries, we calculated the average rank of the empirical n_D when compared to that of the randomisations.

7.2 Relative importance

7.2.1 Genrealised mixed effects models for f_S and f_D

To facilitate comparison among the continuous variables, we scaled them so that they all had a mean of zero and a standard deviation of one. Although the importance of plants and pollinators or invasive and non-invasive species could respond differently to our structural metrics, our data set did not contain enough variation to include the corresponding interactions terms for these latter two predictors. All network metrics were calculated using the R package bipartite 2.06 (Dormann *et al.* 2008). Candidate models and estimates were assessed using the same procedure as in the models for n_D.

8 Matching's weight threshold

We use the frequency with which a species is classified as a driver node f_D and the frequency with which a species is classified as a superior node f_S in a set of accepted maximal cardinality matchings to infer their relative importance. Maximal cardinality matchings were accepted if the matching weight was over a certain threshold. This threshold was defined as a proportion of the maximum matching's weight. Here we evaluate the impact that a particular choice of this threshold has on both f_D and f_S. The number of accepted maximal cardinality matchings increased rapidly as the threshold at which they are accepted decreases. Nevertheless, this number stabilises below approximately 0.6-0.7 (Figure S11).

We then examined how species relative importance changes with the chosen threshold. We observed, that regardless of the matching weight's threshold, species tended to be superior nodes in all or none of the matchings, which equates to f_S values of one or zero, and the differences between important and unimportant species were maintained (Figure S12). In contrast, species' f_D reached intermediate levels of importance and showed higher variability across thresholds (Figure S13). We note here that we were unable to compute all the maximal cardinality matchings for one of our networks (uninvaded community
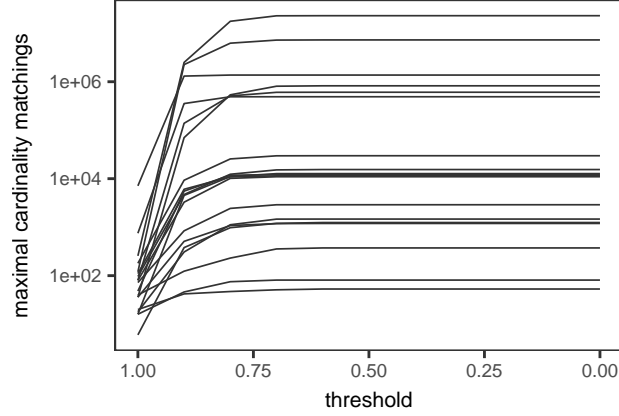
Figure S11: **Number of maximal cardinality matchings as a function of the weight threshold**. Each line corresponds to one network.

in site 8). This is because, we estimated the number of matchings to be more than one billion, and this surpassed our available computational resources.

Finally, we examined the agreement across thresholds by calculating the Spearman correlation matrix of species' f_S and f_D for each network. The mean of those matrices is shown in Table S4. Among all thresholds, we found a very high agreement for f_S, and high to very high for f_D. With the exception of using a threshold of 1 (only maximum matchings are accepted), the choice of the threshold has limited impact on the relative importance of species. For convenience, we use an intermediate level of 0.8. Nevertheless, the differences are marginal and any threshold $<= 0.9$ is likely to produce very similar results.

Table S4: **Spearman correlation coefficient among different matching weight thresholds**. The upper triangle shows values for f_D and the lower triangle shows values for f_S

|	0	0.1	0.2	0.3	0.4	0.5	0.6	0.7	0.8	0.9	1
0	—	1	1	1	1	1	1	0.995	0.947	0.83	0.402
0.1	1	—	1	1	1	1	1	0.995	0.947	0.83	0.402
0.2	1	1	—	1	1	1	1	0.995	0.947	0.83	0.402
0.3	1	1	1	—	1	1	1	0.995	0.947	0.83	0.402
0.4	1	1	1	1	—	1	1	0.995	0.947	0.83	0.402
0.5	1	1	1	1	1	—	1	0.995	0.947	0.83	0.402
0.6	1	1	1	1	1	1	—	0.995	0.947	0.832	0.404
0.7	1	1	1	1	1	1	1	—	0.953	0.845	0.426
0.8	1	1	1	1	1	1	1	1	—	0.903	0.493
0.9	0.994	0.994	0.994	0.994	0.994	0.994	0.994	0.994	0.995	—	0.642
1	0.982	0.982	0.982	0.982	0.982	0.982	0.982	0.982	0.983	0.987	—
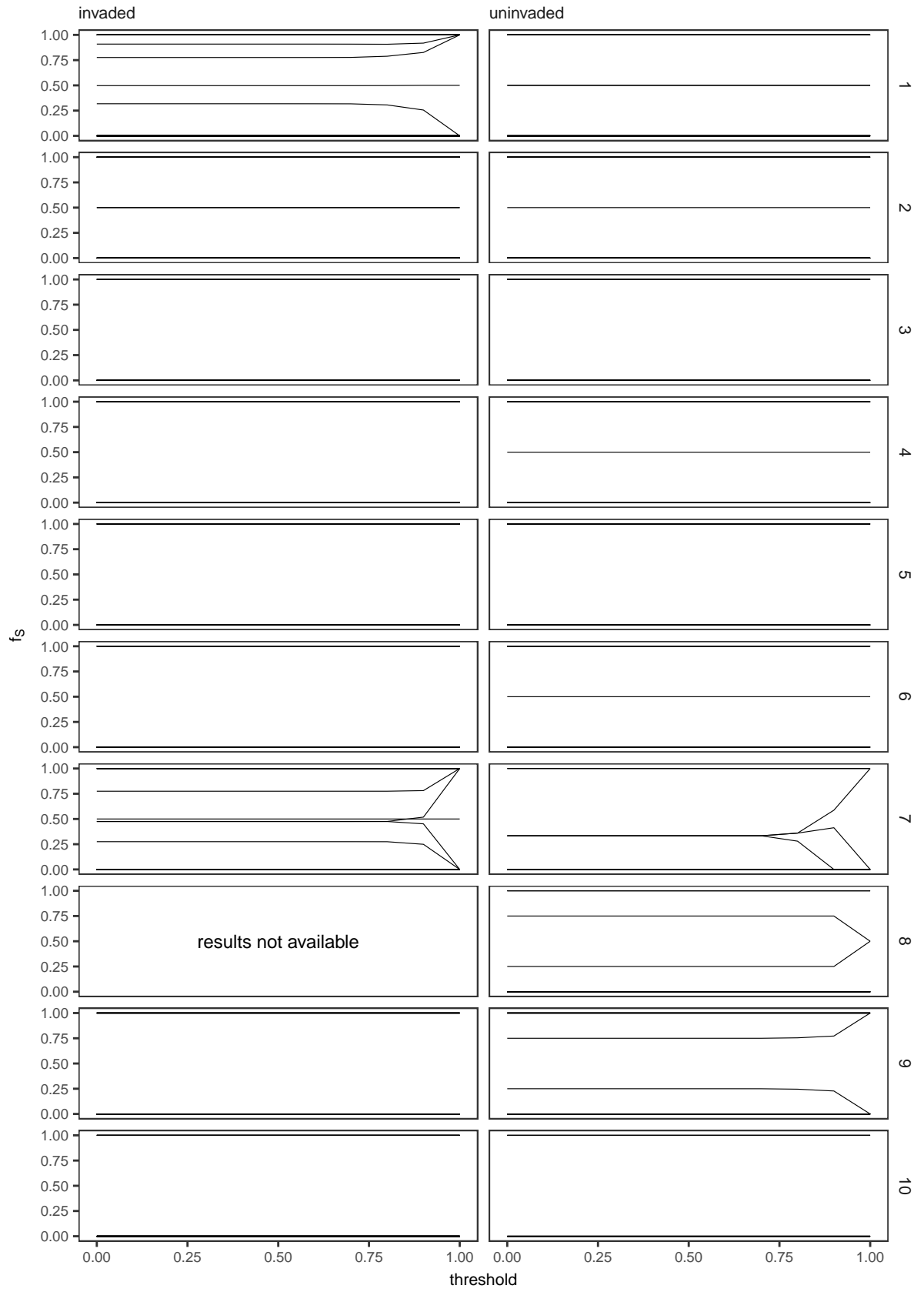
Figure S12: **Relative requency with which a species was classified as a superior node (f_S) as a function of the matching weight threshold**. Rows correspond to network pairs.
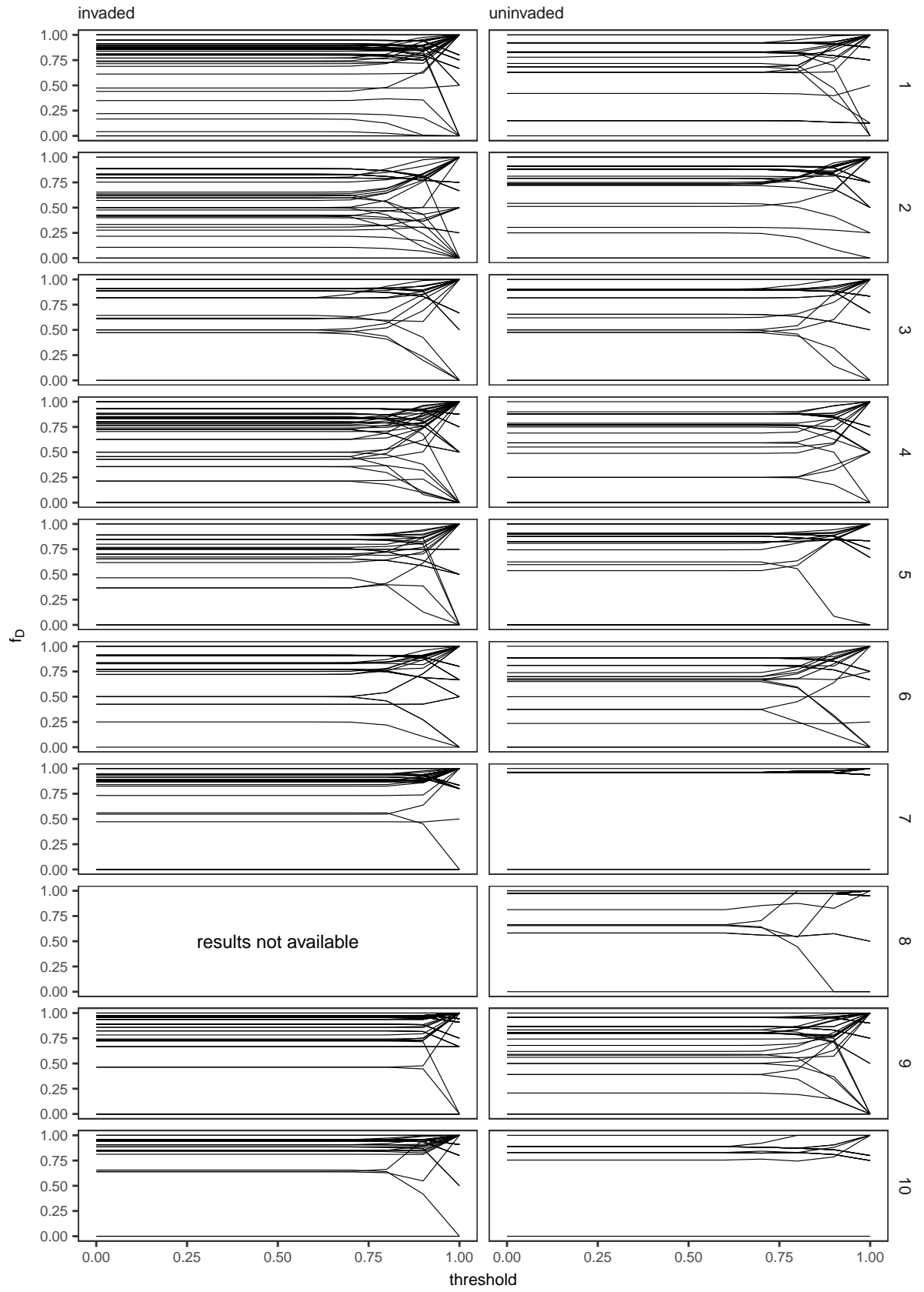
Figure S13: **Relative frequency with which a species was classified as a driver node (f_D) as a function of the matching weight threshold**. Rows correspond to network pairs.

9 n_D models results

Table S5: **Coefficients of the n_D models**. Estimates are averaged over the models that accounted for 95% of the evidence weight.

	est.	imp.	C.I.
(Intercept)	3.75	1.00	3.1
ratio plant/pollinator	-5.96	1.00	3.6
spp. richness	-0.01	0.66	0.024
connectance	-2.54	0.48	7.9
invasion status	-0.04	0.32	0.21
study site	0.08	0.30	0.33

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