

Local and Spatial Processes Shape the Collapse and Recovery of Mutualistic Networks

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1 Abstract

2 Ecosystems can undergo abrupt critical transitions even when environmental change is gradual,
3 resulting in hysteresis, where recovery requires conditions far more favorable than those that
4 triggered collapse. While previous research has mainly focused on local network dynamics, the
5 role of spatial heterogeneity and dispersal in mutualistic ecosystem resilience remains less ex-
6 plored. This study examines how spatial structures—grid, random, small-world, and scale-free
7 networks—interact with dispersal rates to influence ecosystem recovery and mutualistic network
8 persistence. We find that with low dispersal, restoration cost is similar across spatial struc-
9 tures, but at intermediate dispersal rates, scale-free networks show faster recovery and smaller
10 restoration costs. At high dispersal rates, increased connectivity initially reduces restoration cost;
11 however, over time, homogenization weakens spatial heterogeneity, causing restoration cost to
12 increase. Importantly, increasing nestedness can delay collapse but also extends the recovery dis-
13 tance, making ecosystems harder to restore. By adjusting dispersal rates and transitioning from
14 homogeneous to heterogeneous spatial structures, we can decrease restoration cost and improve
15 ecosystem stability, offering key insights for ecological management strategies.

16

Introduction

17 Ecological systems are increasingly destabilized by accelerating global change, driven by habitat
18 fragmentation (Haddad et al., 2015; Wilcox and Murphy, 1985), climate warming (Walther et al.,
19 2002), nitrogen deposition, and widespread deforestation (Laurance et al., 2014). These stressors
20 can push ecosystems past critical thresholds, resulting in sudden and often irreversible regime
21 shifts. Such critical transitions, also referred to as tipping points, are typically driven by nonlin-
22 ear feedback mechanisms, where the ecosystem abruptly collapses once a bifurcation threshold
23 (environmental stressor) is crossed (Dakos and Bascompte, 2014; Scheffer et al., 2001). Recovery
24 from such a state is often slow and may not occur under the same environmental conditions that
25 caused the collapse. Instead, the system typically requires substantially more favorable condi-
26 tions to return to its original state—a dynamic known as hysteresis, where collapse and recovery
27 follow different pathways (Kéfi et al., 2014; Scheffer and Carpenter, 2003). Understanding the
28 conditions that govern collapse and recovery, and identifying the factors that influence the cost
29 of restoration, has become central to the study of ecological resilience (Biggs et al., 2018; Suding
30 et al., 2004).

31 While tipping points have been explored in a variety of systems, including shallow lakes
32 (Carpenter et al., 1999; van de Leemput et al., 2015), coral reefs (McCook, 1999), and savannas
33 (Walker, 1995), mutualistic communities have only recently been examined through this lens.
34 These systems, such as plant–pollinator networks, are characterized by positive feedback loops,
35 where species benefit from the presence of their interaction partners (Bascompte and Jordano,
36 2007; Bastolla et al., 2009). Such feedbacks promote coexistence under stable conditions but
37 can also increase vulnerability to abrupt collapse when environmental stress intensifies (Apari-
38 cio et al., 2021; Bascompte and Scheffer, 2023; Dakos and Bascompte, 2014; Lever et al., 2014).
39 Globally, pollinator populations are declining due to a combination of stressors, including pesti-
40 cide use, habitat loss, and emerging pathogens (Henry et al., 2012; Potts et al., 2010; Whitehorn
41 et al., 2012). These pressures elevate species mortality and may push mutualistic systems toward

42 critical thresholds or collapse. Predicting community responses to such stress is difficult, as it
43 depends not only on species traits but also on the structure, strength, and organization of their
44 interactions (Bascompte et al., 2006; May, 1972; McCann, 2000). Such findings emphasize the
45 need to study how specific environmental stressors impact the organization of species interac-
46 tions, and how resulting changes in network structure affect community stability and recovery
47 after disturbance. Network properties such as nestedness and connectance have been shown to
48 influence the robustness and stability of mutualistic communities by determining how interac-
49 tion strength is distributed across species (Bascompte and Jordano, 2013; Thébault and Fontaine,
50 2010). However, most studies on tipping points in mutualistic systems have focused on non-
51 spatial, isolated networks, overlooking the spatial complexity inherent in real-world ecosystems
52 (Aparicio et al., 2021; Dakos and Bascompte, 2014; Lever et al., 2014; Panahi et al., 2023).

53 In nature, species and their interactions are embedded within spatially-structured landscapes,
54 where local communities occupy discrete habitat patches connected by dispersal. These land-
55 scapes form metacommunities, where local and regional processes jointly determine biodiversity
56 dynamics (Leibold et al., 2004; Loreau et al., 2003). Dispersal can buffer systems against collapse
57 by enabling recolonization and rescue effects, yet excessive connectivity can synchronize popula-
58 tion dynamics across patches, reducing spatial insurance and amplifying the risk of system-wide
59 failure (Urban and Keitt, 2001; Wang and Loreau, 2014). The configuration of the spatial net-
60 work—whether regular, random, or scale-free—can strongly influence how disturbances spread
61 or are contained within the system (Gilaranz et al., 2015; Holland and Hastings, 2008). Recent
62 work has shown that the spatial organization of landscapes can fundamentally alter ecological
63 tipping points, with heterogeneous structures often delaying collapse and enhancing recovery
64 by promoting localized buffering and spatial insurance (Saade et al., 2023). Despite the increas-
65 ing recognition of spatial processes in ecological resilience, how local interaction structures (e.g.,
66 nestedness) interact with global spatial organization and dispersal to govern collapse, recovery,
67 and hysteresis remains poorly understood. In particular, the degree to which spatial heterogene-
68 ity can reduce restoration costs and mediate resilience in mutualistic metacommunities is largely

69 unexplored, representing a key gap this study aims to address.

70 Recent advances in complex systems theory have shown that multi-layer or network-of-
71 networks frameworks can offer valuable insights into the behavior of interdependent systems
72 (Buldyrev et al., 2010; Kivelä et al., 2014). In such a frameworks, different network layers—whether
73 ecological, infrastructural, or social—are coupled through dynamic processes like information
74 flow or dispersal (Boccaletti et al., 2014; Gao et al., 2012). In ecological contexts, dispersal effec-
75 tively links local communities across space, forming a multilayer structure where the collapse
76 or recovery of one patch can cascade through the system (Baruah, 2022; Fronhofer et al., 2023).
77 Studies on interdependent systems suggest that structural heterogeneity and the formation of
78 new connections across layers can delay tipping points and enhance system-wide robustness (Al-
79 termatt and Fronhofer, 2018; Carrara et al., 2012). By analogy, spatial heterogeneity and moderate
80 dispersal in ecological metacommunities may buffer local collapses and lower restoration costs
81 by redistributing resilience across scales.

82 In this study, we investigate how the interaction between local mutualistic structure, spatial
83 network topology, and dispersal rate governs collapse, recovery, and hysteresis in mutualistic
84 metacommunities. We model a system in which each habitat patch hosts an plant-pollinator
85 network embedded within spatial configurations of varying complexity (grid, random, small-
86 world, and scale-free). By combining synthetic simulations with 115 empirical networks from the
87 Web of Life database (Fortuna et al., 2014), we quantify how environmental stress influences the
88 cost of restoration and the conditions for recovery. We further evaluate the relative contributions
89 of local and spatial structures to resilience across dispersal regimes using quantitative model
90 comparisons. Our results reveal that spatial heterogeneity fundamentally alters the role of local
91 architecture, with scale-free landscapes under intermediate dispersal most effectively minimizing
92 hysteresis. Together, these insights highlight the need to integrate local interaction structure
93 with spatial network organization to better understand and manage the resilience of mutualistic
94 systems in changing environments.

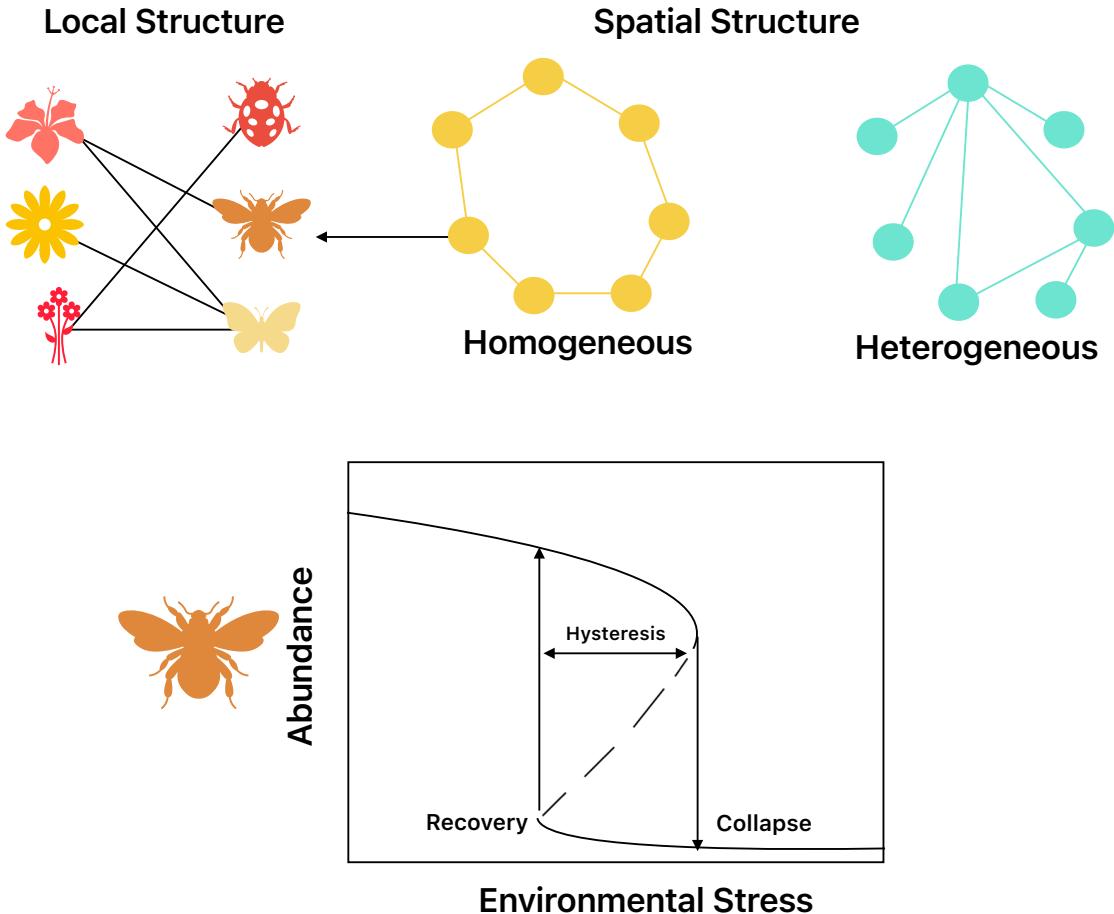


Figure 1: Schematic of local and spatial dynamics in mutualistic metacommunities. Local mutualistic networks of plants and pollinators are embedded within spatial networks that range from homogeneous (grid) to heterogeneous (scale-free) structures. The lower panel illustrates how increasing environmental stress drives a collapse in species abundance, followed by recovery as stress is reduced, forming a hysteresis loop. The distance between collapse and recovery thresholds represents the restoration cost, highlighting how the interplay between local interactions and spatial connectivity governs the recovery of mutualistic systems.

95

Methods

96

Spatial metacommunity framework

97 We modeled a spatially-explicit metacommunity in which each habitat patch harbors a mutu-
98 alistic community composed of plants and pollinators. These local communities are embedded
99 in a spatial network that governs the dispersal of individuals between patches. Each patch con-
100 tains a mutualistic interaction network, allowing us to isolate the effects of spatial structure and
101 dispersal. We systematically varied the underlying spatial configuration by generating networks
102 with four distinct topologies: regular (grid-like), small-world, random, and scale-free. This pro-
103 gression introduces increasing heterogeneity in the degree distribution of patches. All spatial
104 structure comprised $N = 101$ patches connected by 202 links, maintaining a constant average
105 degree of 4.

106

Local mutualistic dynamics

107 Within each patch, species dynamics were governed by a set of ordinary differential equations
108 describing population changes for S_P plant and S_A pollinator species (Lever et al., 2014). The
109 abundance of species i in patch l is denoted by P_i^l for plants and A_i^l for pollinators. The deter-
110 ministic dynamics are given by:

$$\begin{aligned}\frac{dP_i^l}{dt} &= P_i^l \left(\alpha_i^l - \sum_{k'=1}^{S_P} \beta_{ik'}^l P_{k'}^l + \frac{\sum_{k=1}^{S_A} \gamma_{ik}^l A_k^l}{1 + h_i^l \sum_{k=1}^{S_A} \gamma_{ik}^l A_k^l} \right) + \sigma \left(\sum_{m=1}^N a_{lm} \frac{P_i^m}{d_m} - P_i^l \right), \\ \frac{dA_i^l}{dt} &= A_i^l \left(\alpha_i^l - k_i^l - \sum_{k'=1}^{S_A} \beta_{ik'}^l A_{k'}^l + \frac{\sum_{k=1}^{S_P} \gamma_{ik}^l P_k^l}{1 + h_i^l \sum_{k=1}^{S_P} \gamma_{ik}^l P_k^l} \right) + \sigma \left(\sum_{m=1}^N a_{lm} \frac{A_i^m}{d_m} - A_i^l \right).\end{aligned}$$

111 Here, α_i represents the intrinsic growth rate ($\alpha_i = 0.15$), and k_i denotes an externally imposed
112 mortality affecting only pollinators. Intraspecific and interspecific competition coefficients are
113 set to $\beta_{ii} = 1$ and $\beta_{ij} = 0.05$, respectively. Mutualistic benefits saturate as partner abundance
114 increases, modeled through a type II functional response with handling time $h_i^l = 0.3$.

The mutualistic strength γ_{ik} is modulated by the species' connectivity as:

$$\gamma_{ik} = \gamma_0 \frac{\epsilon_{ik}}{t_i^\delta}$$

where ϵ_{ik} represents the entries of the plant pollinator interaction matrix (obtained from empirical networks), γ_0 sets the baseline mutualistic strength ($\gamma_0 = 1$), t_i is the degree of species i , and $\delta = 0.5$ introduces a trade-off between specialization and per-interaction benefit. For simplicity, the main simulations assumed identical parameter values across species and patches; however, we verified that the qualitative patterns are robust to variation in mutualistic strength (Figure S2), local-network heterogeneity (Figure S3), stochastic perturbations (Figure S4), and heterogeneous species-level parameters (Figure S5).

Dispersal within spatial networks

Dispersal is modeled as a random-walk diffusion process, affecting both plant and pollinator species. The network of spatial connections is encoded by an adjacency matrix a_{lm} , where $a_{lm} = 1$ indicates a direct link between patches l and m , and d_m is the degree of patch m . The net movement of species i into patch l is given by:

$$\sigma \left(\sum_{m=1}^N a_{lm} \frac{X_i^m}{d_m} - X_i^l \right),$$

where X_i^l corresponds to either P_i^l or A_i^l , and σ is the dispersal rate, varied between 0 (no dispersal) and 1 (high dispersal). This formulation assumes uniform diffusion across neighboring patches, such that individuals spread equally from a given patch to all its neighbors. In regular networks, the diffusion term cancels out under uniform abundance, whereas in heterogeneous networks, local differences in connectivity generate directional flow, influencing spatial persistence and recovery dynamics.

133 *Simulation protocol and hysteresis quantification*

134 To investigate how the system responds to environmental stress, we incrementally increased the
135 mortality parameter k_i in steps of 0.01, integrating the system numerically using a fourth-order
136 Runge–Kutta solver until a steady state was reached. Extinction was defined as the point where
137 all pollinator abundances dropped below a threshold of 0.01. Following collapse, we reversed
138 the process by gradually decreasing k_i to determine the conditions under which the system
139 could recover. Recovery was defined as the point where at least one pollinator species exceeded
140 the same abundance threshold. The distance between the collapse and recovery points defines
141 the hysteresis width, which we interpret as a measure of the restoration cost. Each simulation
142 involved solving a system of $N \times (S_P + S_A)$ coupled differential equations.

143 *Empirical networks*

144 To examine the generality of our findings, we analyzed 115 empirical plant–pollinator networks
145 obtained from the Web of Life database (www.web-of-life.es). We restricted our dataset to
146 networks containing no more than 100 species in total (plants plus pollinators) to maintain com-
147 putational feasibility while preserving a broad spectrum of structural diversity. These empirical
148 systems encompass a variety of interaction patterns—from highly specialized to strongly gen-
149 eralized—providing a representative gradient of local network organization through which to
150 assess spatial and dispersal effects on system resilience.

151 We quantified the internal organization of each empirical network using nestedness, a key
152 descriptor of mutualistic network that captures the tendency of specialist species to interact with
153 subsets of generalist partners. To quantify this pattern, we used the metric proposed by (Fortuna
154 et al., 2019), which is mathematically equivalent to the NODF measure (Almeida-Neto et al.,
155 2008) but avoids penalizing species that share the same number of interaction partners. For each
156 pair of plant or pollinator species, we calculated the proportion of shared interaction partners
157 relative to the smaller of their degrees and then averaged these values across all pairs to obtain

158 the observed nestedness of each network. To allow comparisons among networks differing in
159 size, connectance, or sampling intensity, we standardized nestedness using a null model-based
160 z -score. Following the probabilistic null model (Bascompte et al., 2003) randomized matrices were
161 generated by drawing each cell independently with probability $\pi_{ij} = (p_i + q_j)/2$ where p_i and q_j
162 denote the empirical row and column marginal frequencies of plant and pollinator respectively.
163 This null model preserves the expected generalization level of both plants and pollinators, while
164 allowing variation around the empirical structure. For each empirical network, we generated
165 100 randomized matrices computed their mean (μ) and standard deviation (σ) of nestedness
166 and computed the standardized nestedness value as $z = \frac{x - \mu}{\sigma}$, where x denotes the observed
167 nestedness of the empirical network. This normalization allows direct comparison of nestedness
168 across systems differing in size, connectance, and sampling intensity. The resulting nestedness
169 z -score thus reflects how strongly a network deviates from random expectations—positive values
170 indicating more nested organization than expected by chance.

171 To evaluate the influence of local structural variability on resilience under a fixed spatial
172 configuration, we systematically embedded each one of the 115 empirical networks—each char-
173 acterized by a distinct nestedness z -score—within the same spatial topologies. This approach
174 allowed us to isolate how differences in local mutualistic architecture modify system behavior
175 when the spatial framework is held constant. In practice, for a given spatial structure (e.g., grid,
176 random, small-world, or scale-free), each empirical network served as the local interaction matrix
177 for all patches, while dispersal dynamics and global connectivity patterns remained unchanged.
178 This design ensured that any observed variation in collapse thresholds, recovery trajectories, or
179 restoration costs arose specifically from differences in local network organization, rather than
180 from alterations in the spatial structure or dispersal regime. By repeating this process across
181 multiple spatial configurations, we systematically disentangled the relative contributions of local
182 architecture and spatial topology to metacommunity resilience.

183

Results

184 *Dispersal Rates and Spatial Structures in Collapse and Recovery Dynamics*

185 We observed that the impact of dispersal on collapse and recovery dynamics varied across spa-
186 tial structures (Figure 2). In the absence of dispersal, collapse points were nearly identical across
187 all networks, as spatial configuration could not influence dynamics without connectivity among
188 patches. At intermediate dispersal rates, scale-free networks exhibited a delayed collapse and
189 faster recovery compared with grid and random networks, reflecting the stabilizing role of het-
190 erogeneous connectivity (Figure 2, S1). At high dispersal, scale-free systems collapsed and recov-
191 ered earlier, indicating that strong coupling accelerates both loss and recovery through homoge-
192 nization. The hysteresis width, representing restoration cost, declined with increasing dispersal
193 but rose again beyond a critical threshold, suggesting diminishing stabilizing effects at very high
194 dispersal. These results highlight intermediate dispersal as the regime that most effectively bal-
195 ances stability and recovery potential across spatial configurations.

196

197 Building on the observed collapse and recovery dynamics, we next examined how local net-
198 work architecture—specifically nestedness—modulates restoration cost (hysteresis width) across
199 spatial structures and dispersal regimes (Figure 3). The first row of Figure 3 shows how restora-
200 tion cost varies with nestedness for grid, random, and scale-free networks under low ($\sigma \approx 0$),
201 intermediate ($\sigma = 0.125$), and high dispersal ($\sigma = 1$), while the second row presents the corre-
202 sponding regression slopes that quantify the direction and strength of these relationships.

203 In grid and random networks, restoration cost increased consistently with nestedness across
204 all dispersal levels, reflecting a persistent positive slope. In grids, this relationship remained
205 equally strong regardless of dispersal, whereas in random networks the positive slope weakened
206 at intermediate dispersal before strengthening again at high dispersal. In contrast, scale-free
207 networks showed a clear dispersal-dependent reversal: the slope was positive at low dispersal,

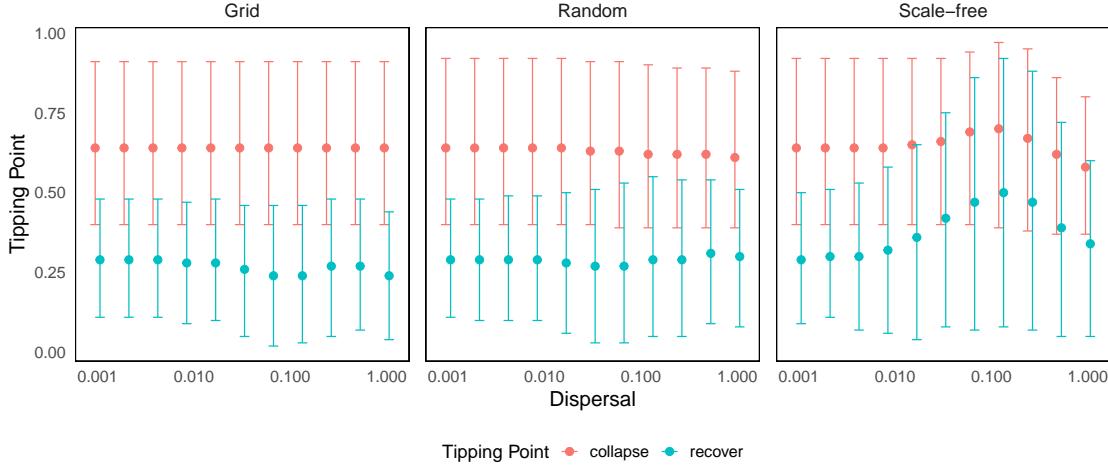


Figure 2: Collapse and recovery thresholds across spatial and dispersal regimes. Each subplot presents results from 115 mutualistic network realizations across three spatial configurations (grid, random, and scale-free) and varying dispersal rates. Filled circles indicate the mean collapse and recovery points, while error bars represent the 95% confidence intervals. The figure highlights that at intermediate dispersal, scale-free networks experience a delayed collapse and a faster recovery relative to grid and random structures.

208 became negative at intermediate levels—indicating that moderate movement through heterogeneous
 209 connectivity reduced restoration cost—and approached zero at high dispersal as spatial
 210 mixing homogenized the system. Among all structures, grid networks exhibited the steepest
 211 positive slopes, followed by random and then scale-free networks. The lowest restoration cost
 212 occurred in highly nested scale-free systems at intermediate dispersal, marking the regime where
 213 spatial heterogeneity and moderate connectivity most effectively enhance recovery efficiency.

214 *Local and Spatial Contributions to Restoration Cost*

215 To evaluate how local and spatial structures jointly influence restoration cost, we mapped hys-
 216 teresis width against nestedness (local property) and spatial configuration across dispersal regimes
 217 (Figure 4). At low dispersal, variation in restoration cost was mainly associated with nestedness,

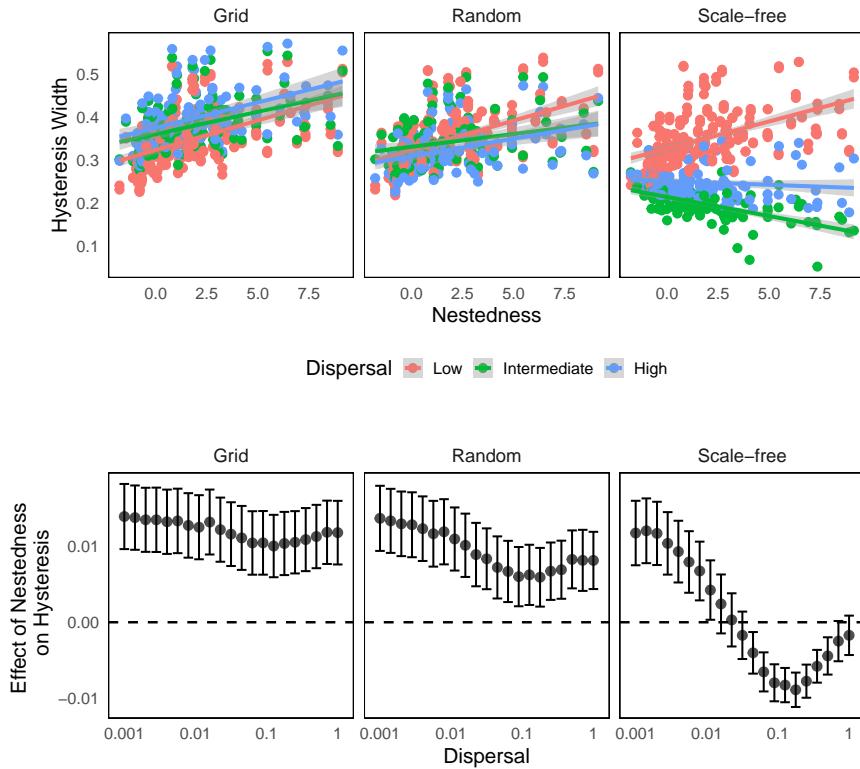


Figure 3: Variation in restoration cost with nestedness across spatial structures and dispersal regimes. The first row shows how restoration cost (measured as hysteresis width) varies with nestedness across three spatial structures—grid, random, and scale-free—under low (red), intermediate (green), and high (blue) dispersal rates. Linear model fits illustrate the direction and strength of these relationships. The second row depicts the estimated slopes from these models, highlighting whether nestedness increases or decreases restoration cost; dashed lines mark zero to distinguish positive and negative effects. Grid and random networks generally exhibit a positive association between nestedness and restoration cost, whereas scale-free networks show a dispersal-dependent shift—nestedness lowers restoration cost at intermediate dispersal, indicating enhanced recovery efficiency under heterogeneous spatial connectivity.

218 indicating that local network architecture dominated resilience when movement between patches
 219 was limited. With increasing dispersal, patterns progressively shifted toward the spatial axis,

220 showing that spatial connectivity—particularly in scale-free landscapes—became the principal
221 driver of reduced restoration cost. This transition marked a threshold where spatial organiza-
222 tion began to govern recovery dynamics. At high dispersal, the effects of both local and spatial
223 structure converged, reflecting the homogenizing influence of strong connectivity. Overall, inter-
224 mediate dispersal emerged as the regime where spatial heterogeneity most effectively reduced
225 restoration cost and enhanced system resilience.

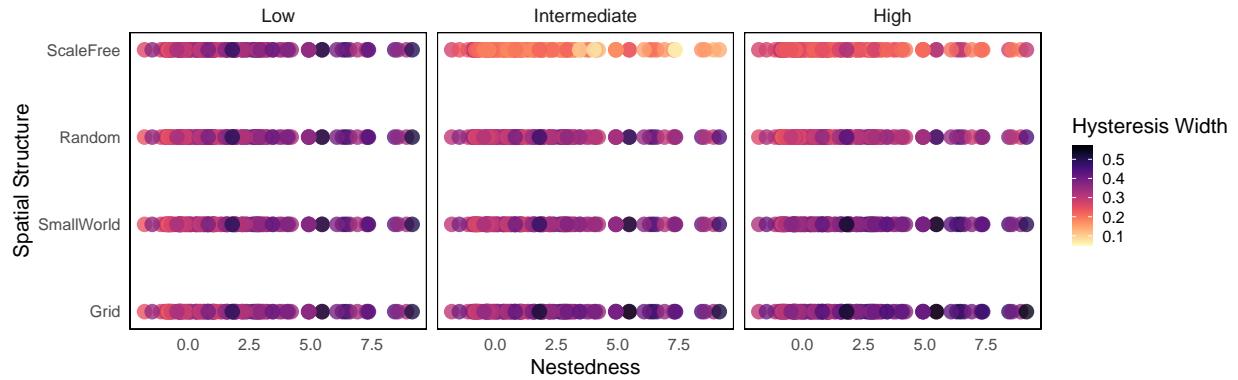


Figure 4: Heatmap illustrating the combined effects of local and spatial structures on restoration cost across dispersal regimes. Restoration cost (hysteresis width) is shown as a function of local network nestedness (x-axis) and spatial structure type (grid, random, scale-free; y-axis) for varying dispersal rates. At low dispersal, variation in restoration cost is largely governed by local structure, reflecting patch-level control of resilience. With increasing dispersal, restoration cost patterns shift toward the spatial axis, indicating that network connectivity—particularly in scale-free structures—enhances recovery and lowers restoration costs. At high dispersal, the influence of both axes converges, consistent with landscape homogenization. Overall, the figure highlights that intermediate dispersal optimally balances local and spatial processes to enhance ecosystem resilience.

226 We quantitatively evaluated how local and spatial network structures shape restoration cost
227 (measured as hysteresis width) across varying dispersal rates. Using generalized linear models
228 (GLMs) and comparing their performance with Akaike Information Criterion (AIC), we found
229 that models combining both local (nestedness) and spatial (network topology) predictors consis-
230 tently outperformed single-factor models, underscoring the interdependence of local and global
231 processes (Figure 5). At low dispersal rates, local structural properties exerted stronger positive
232 effects on restoration cost, indicating that limited connectivity amplifies local feedbacks and de-
233 lays recovery. In contrast, at intermediate dispersal rates, spatial structure became the dominant
234 driver, where heterogeneous connectivity—particularly in scale-free networks—significantly re-
235 duced restoration cost. At high dispersal, both effects converged, reflecting homogenization
236 across the landscape that diminished the distinct influence of either structure. Overall, these
237 results demonstrate that local architecture governs resilience under isolation, whereas spatial
238 heterogeneity mitigates restoration cost under moderate dispersal, promoting efficient recovery
239 at the metacommunity scale. Statistical comparisons confirmed that models integrating both local
240 and spatial predictors best explained restoration cost, emphasizing the synergistic-not additive-
241 nature of their effects.

242 Discussion

243 Our work shows that resilience in mutualistic metacommunities arises from the dynamic in-
244 teraction between local mutualistic structure and the architecture of spatial connectivity. By
245 combining 115 empirical plant–pollinator networks with multiple spatial configurations and dis-
246 persal regimes, we reveal how these two scales jointly control collapse, recovery, and the cost of
247 restoration. Dispersal mediates a transition in control—from recovery governed by local interac-
248 tions at low connectivity to spatially-coordinated stabilization at intermediate dispersal, where
249 heterogeneous, scale-free landscapes most effectively reduce hysteresis. Unlike previous studies
250 that analyzed local mutualistic networks in isolation, our framework integrates spatial hetero-

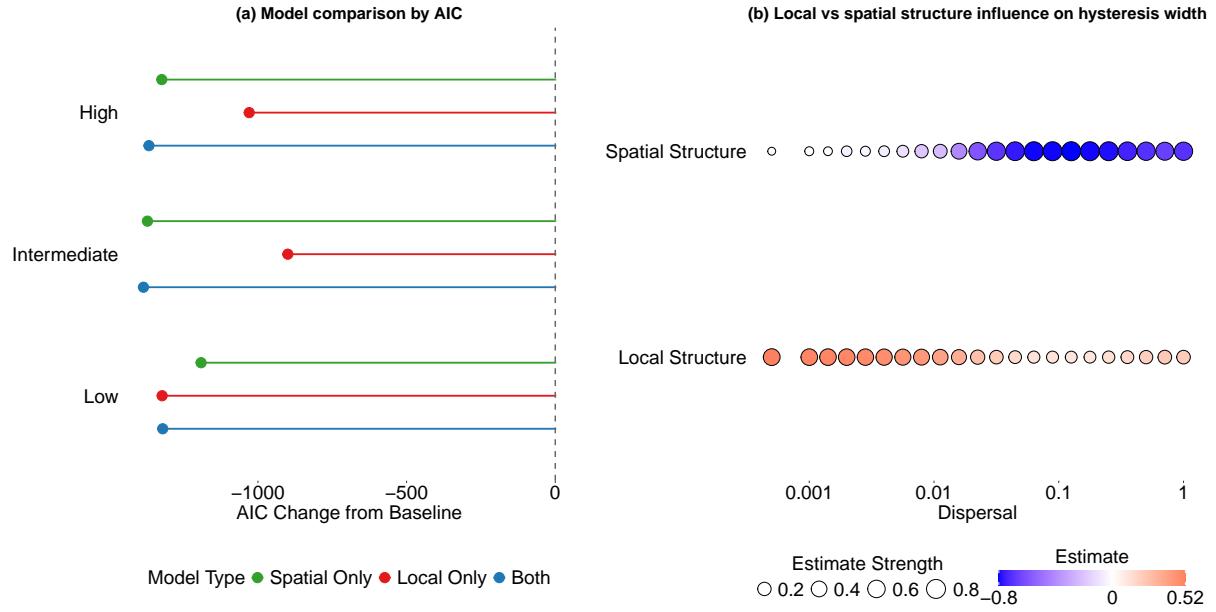


Figure 5: Quantitative evaluation of local and spatial contributions to restoration cost across dispersal regimes. (a) Model comparison using Akaike Information Criterion (AIC) across low, intermediate, and high dispersal rates. Models incorporating both local (nestedness) and spatial (network topology) predictors consistently yielded lower AIC values than single-factor models, indicating that restoration cost (hysteresis width) is jointly governed by local and spatial factors. (b) Standardized regression coefficients (β) for local and spatial predictors across dispersal regimes. Point color represents the direction and magnitude of effects, while size denotes effect strength. At low dispersal, local structure strongly increases restoration cost, reflecting dominant local feedbacks under limited connectivity. At intermediate dispersal, spatial structure becomes the primary driver, where heterogeneous connectivity markedly reduces restoration cost. At high dispersal, both effects converge, reflecting landscape homogenization. Overall, while local structure amplifies restoration cost in isolated conditions, spatial heterogeneity mitigates it under moderate dispersal, promoting more efficient recovery at the landscape scale.

251 geneity and dispersal processes, demonstrating that stability emerges not from either local or
 252 spatial structure alone, but from their coupling. This integration identifies the dispersal regime

253 where spatial heterogeneity most strongly promotes recovery and minimizes restoration cost.

254 Abrupt transitions in ecosystems typically emerge from reinforcing feedbacks between species
255 interactions and environmental stress (Dakos and Bascompte, 2014; May, 1972; Scheffer et al.,
256 2001). Although these dynamics are well documented in isolated communities, real landscapes
257 consist of spatially-distributed patches linked through dispersal (Leibold et al., 2004; Loreau et al.,
258 2003). Spatial connectivity can either buffer collapse by facilitating recolonization or propagate
259 disturbances that synchronize declines (Dakos et al., 2019; Gilarranz et al., 2015). Our results
260 extend this theoretical foundation by quantifying how spatial topology and dispersal together
261 determine restoration cost—a metric that reflects how much effort is required to return a system
262 to its original state after collapse. By explicitly coupling local nestedness with spatial structure,
263 we identify how landscape heterogeneity governs the balance between local persistence and re-
264 gional synchronization (Holland and Hastings, 2008; Kéfi et al., 2014; Saavedra et al., 2017).

265 At low dispersal, local architecture dominates system behavior: each patch recovers indepen-
266 dently, consistent with earlier theoretical predictions (Lever et al., 2014; Rohr et al., 2014). As
267 dispersal increases, however, spatial links allow recovery in one patch to spread to others, low-
268 ering restoration cost through a “spatial insurance effect” (Loreau et al., 2003). This effect was
269 strongest in heterogeneous (scale-free) spatial networks, where hub-like nodes channel recovery
270 potential and restrict the propagation of collapse. Excessive dispersal, however, erased spatial
271 differences, synchronized patch dynamics, and widened hysteresis (Dakos et al., 2019; Wang and
272 Loreau, 2014). This dual behavior highlights a general ecological trade-off: dispersal can promote
273 recovery but, when too strong, amplifies systemic vulnerability—a phenomenon also observed
274 in interdependent infrastructures and food webs (Buldyrev et al., 2010; Gao et al., 2012).

275 From a restoration perspective, our findings clarify how the influence of local and spatial
276 structure shifts with dispersal. While local properties such as nestedness regulate resilience,
277 their effects depend strongly on spatial organization. Highly nested networks embedded in
278 scale-free landscapes exhibited the lowest restoration costs, revealing that spatial heterogeneity
279 can compensate for the recovery inefficiency of densely connected local architectures. Crucially,

280 these patterns were not sensitive to simplifying assumptions: the same qualitative behaviour
281 emerged under environmental heterogeneity, stochastic perturbations, variation in local mutual-
282 istic structure, and species-level parameter differences drawn from uniform ranges (Figs. S2–S5).
283 Thus, optimizing restoration in mutualistic systems may require not only local rewiring but also
284 targeted modification of spatial connectivity. Moderate dispersal in heterogeneous landscapes
285 can create recovery corridors without inducing full synchronization,—a principle consistent with
286 metacommunity theory and recent work on spatial design for ecosystem resilience (Bascompte
287 et al., 2019; Keitt et al., 1997).

288 Our framework has limitations that open directions for further research. Although we in-
289 troduced heterogeneity in species parameters, all patches in our model experienced the same
290 underlying environmental conditions, whereas in real landscapes demographic rates and inter-
291 action strengths are shaped by local abiotic factors such as temperature, moisture, nutrient avail-
292 ability, and disturbance history. Such spatial variation in environmental drivers can generate
293 patch-specific collapse and recovery dynamics (Dakos et al., 2019) that our uniform environmen-
294 tal backdrop does not capture. Adaptive behaviors—such as trait evolution, partner switching
295 (Cai et al., 2020; Kaiser-Bunbury et al., 2010; Kondoh, 2003; Valdovinos et al., 2013) were also
296 not considered, yet these can enhance persistence and modify spatial tipping points (Valdovi-
297 nos et al., 2016). Furthermore, we focused on purely mutualistic layers, while antagonistic links
298 (e.g., herbivory, parasitism) often coexist and may redistribute extinction risks or stabilize certain
299 modules (Glaum and Kessler, 2017; Pilosof et al., 2017). Finally, dispersal was assumed static and
300 uniform; in nature, movement is adaptive and species-specific, potentially feeding back on spatial
301 structure and recovery trajectories. Extending the model to account for these processes would
302 increase ecological realism and improve predictions of collapse and restoration in fragmented
303 landscapes.

304 Together, these findings establish that spatial structure has a stronger qualitative role than
305 local architecture in determining restoration efficiency (Figure 5(b)). While increasing nestedness
306 can delay collapse, it may also widen hysteresis by slowing recovery once disruption occurs.

307 Spatial heterogeneity can counterbalance this cost by redistributing recovery potential across the
308 network, effectively reducing the restoration burden. This mechanism mirrors observations in
309 fragmented pollination networks, coral reef systems, and stepping-stone habitats, where mod-
310 erate connectivity enhances recolonization and stabilizes mutualistic interactions (Keitt et al.,
311 1997; Kormann et al., 2016). By identifying when and how landscape structure alters the re-
312 silience of mutualistic communities, this study bridges theoretical ecology with conservation
313 practice. It suggests that maintaining spatial heterogeneity—rather than altering local network
314 design alone—may represent a more efficient and cost-effective path to restoring and sustaining
315 ecosystem function in a changing world.

316 Data Availability

317 Codes and data are available in a Github repository (<https://github.com/subhendu-math/Local->
318 [spatial-effect.git](#)).

319 Acknowledgments

320 Subhendu thank the members of Bascompte Lab for discussions. Funding was provided by SNSF
321 (grant number 310030_197201 to JB).

322 Literature Cited

- 323 Almeida-Neto, M., Guimaraes, P., Guimaraes Jr, P. R., Loyola, R. D., and Ulrich, W. (2008). A
324 consistent metric for nestedness analysis in ecological systems: reconciling concept and mea-
325 surement. *Oikos*, 117(8):1227–1239.
- 326 Altermatt, F. and Fronhofer, E. A. (2018). Dispersal in dendritic networks: Ecological conse-
327 quences on the spatial distribution of population densities. *Freshwater Biology*, 63(1):22–32.

- 328 Aparicio, A., Velasco-Hernández, J. X., Moog, C. H., Liu, Y.-Y., and Angulo, M. T. (2021).
329 Structure-based identification of sensor species for anticipating critical transitions. *Proceedings*
330 *of the National Academy of Sciences*, 118(51):e2104732118.
- 331 Baruah, G. (2022). The impact of individual variation on abrupt collapses in mutualistic networks.
332 *Ecology Letters*, 25(1):26–37.
- 333 Bascompte, J., García, M. B., Ortega, R., Rezende, E. L., and Pironon, S. (2019). Mutualistic
334 interactions reshuffle the effects of climate change on plants across the tree of life. *Science*
335 *Advances*, 5(5):eaav2539.
- 336 Bascompte, J. and Jordano, P. (2007). Plant-animal mutualistic networks: the architecture of
337 biodiversity. *Annu. Rev. Ecol. Evol. Syst.*, 38(1):567–593.
- 338 Bascompte, J. and Jordano, P. (2013). *Mutualistic networks*, volume 53. Princeton University Press.
- 339 Bascompte, J., Jordano, P., Melián, C. J., and Olesen, J. M. (2003). The nested assembly of plant–
340 animal mutualistic networks. *Proceedings of the National Academy of Sciences*, 100(16):9383–9387.
- 341 Bascompte, J., Jordano, P., and Olesen, J. M. (2006). Asymmetric coevolutionary networks facili-
342 tate biodiversity maintenance. *Science*, 312(5772):431–433.
- 343 Bascompte, J. and Scheffer, M. (2023). The resilience of plant–pollinator networks. *Annual Review*
344 *of Entomology*, 68(1):363–380.
- 345 Bastolla, U., Fortuna, M. A., Pascual-García, A., Ferrera, A., Luque, B., and Bascompte, J. (2009).
346 The architecture of mutualistic networks minimizes competition and increases biodiversity.
347 *Nature*, 458(7241):1018–1020.
- 348 Biggs, R., Peterson, G. D., and Rocha, J. C. (2018). The regime shifts database. *Ecology and Society*,
349 23(3).

- 350 Boccaletti, S., Bianconi, G., Criado, R., Del Genio, C. I., Gómez-Gardenes, J., Romance, M.,
351 Sendina-Nadal, I., Wang, Z., and Zanin, M. (2014). The structure and dynamics of multilayer
352 networks. *Physics reports*, 544(1):1–122.
- 353 Buldyrev, S. V., Parshani, R., Paul, G., Stanley, H. E., and Havlin, S. (2010). Catastrophic cascade
354 of failures in interdependent networks. *Nature*, 464(7291):1025–1028.
- 355 Cai, W., Snyder, J., Hastings, A., and D’Souza, R. M. (2020). Mutualistic networks emerging from
356 adaptive niche-based interactions. *Nature Communications*, 11(1):5470.
- 357 Carpenter, S., Brock, W., and Hanson, P. (1999). Ecological and social dynamics in simple models
358 of ecosystem management. *Conservation Ecology*, 3(2).
- 359 Carrara, F., Altermatt, F., Rodriguez-Iturbe, I., and Rinaldo, A. (2012). Dendritic connectivity
360 controls biodiversity patterns in experimental metacommunities. *Proceedings of the National
361 Academy of Sciences*, 109(15):5761–5766.
- 362 Dakos, V. and Bascompte, J. (2014). Critical slowing down as early warning for the onset of col-
363 lapse in mutualistic communities. *Proceedings of the National Academy of Sciences*, 111(49):17546–
364 17551.
- 365 Dakos, V., Matthews, B., Hendry, A. P., Levine, J., Loeuille, N., Norberg, J., Nosil, P., Scheffer, M.,
366 and De Meester, L. (2019). Ecosystem tipping points in an evolving world. *Nature Ecology &
367 Evolution*, 3(3):355–362.
- 368 Fortuna, M. A., Barbour, M. A., Zaman, L., Hall, A. R., Buckling, A., and Bascompte, J. (2019).
369 Coevolutionary dynamics shape the structure of bacteria-phage infection networks. *Evolution*,
370 73(5):1001–1011.
- 371 Fortuna, M. A., Ortega, R., and Bascompte, J. (2014). The web of life. *arXiv preprint
372 arXiv:1403.2575*.

- 373 Fronhofer, E. A., Bonte, D., Bestion, E., Cote, J., Deshpande, J. N., Duncan, A. B., Hovestadt, T.,
374 Kaltz, O., Keith, S., Kokko, H., et al. (2023). Causes and consequences of dispersal in biodiverse
375 spatially structured systems: what is old and what is new? *arXiv preprint arXiv:2312.00166*.
- 376 Gao, J., Buldyrev, S. V., Stanley, H. E., and Havlin, S. (2012). Networks formed from interdepen-
377 dent networks. *Nature Physics*, 8(1):40–48.
- 378 Gilarranz, L. J., Sabatino, M., Aizen, M. A., and Bascompte, J. (2015). Hot spots of mutualistic
379 networks. *Journal of Animal Ecology*, pages 407–413.
- 380 Glaum, P. and Kessler, A. (2017). Functional reduction in pollination through herbivore-induced
381 pollinator limitation and its potential in mutualist communities. *Nature Communications*,
382 8(1):2031.
- 383 Haddad, N. M., Brudvig, L. A., Clobert, J., Davies, K. F., Gonzalez, A., Holt, R. D., Lovejoy, T. E.,
384 Sexton, J. O., Austin, M. P., Collins, C. D., et al. (2015). Habitat fragmentation and its lasting
385 impact on earth's ecosystems. *Science Advances*, 1(2):e1500052.
- 386 Henry, M., Beguin, M., Requier, F., Rollin, O., Odoux, J.-F., Aupinel, P., Aptel, J., Tchamitchian,
387 S., and Decourtey, A. (2012). A common pesticide decreases foraging success and survival in
388 honey bees. *Science*, 336(6079):348–350.
- 389 Holland, M. D. and Hastings, A. (2008). Strong effect of dispersal network structure on ecological
390 dynamics. *Nature*, 456(7223):792–794.
- 391 Kaiser-Bunbury, C. N., Muff, S., Memmott, J., Müller, C. B., and Caflisch, A. (2010). The robust-
392 ness of pollination networks to the loss of species and interactions: a quantitative approach
393 incorporating pollinator behaviour. *Ecology Letters*, 13(4):442–452.
- 394 Kéfi, S., Guttal, V., Brock, W. A., Carpenter, S. R., Ellison, A. M., Livina, V. N., Seekell, D. A.,
395 Scheffer, M., Van Nes, E. H., and Dakos, V. (2014). Early warning signals of ecological transi-
396 tions: methods for spatial patterns. *PLoS One*, 9(3):e92097.

- 397 Keitt, T. H., Urban, D. L., and Milne, B. T. (1997). Detecting critical scales in fragmented land-
398 scapes. *Conservation Ecology*, 1(1).
- 399 Kivelä, M., Arenas, A., Barthelemy, M., Gleeson, J. P., Moreno, Y., and Porter, M. A. (2014).
400 Multilayer networks. *Journal of Complex Networks*, 2(3):203–271.
- 401 Kondoh, M. (2003). Foraging adaptation and the relationship between food-web complexity and
402 stability. *Science*, 299(5611):1388–1391.
- 403 Kormann, U., Scherber, C., Tscharntke, T., Klein, N., Larbig, M., Valente, J. J., Hadley, A. S., and
404 Betts, M. G. (2016). Corridors restore animal-mediated pollination in fragmented tropical forest
405 landscapes. *Proceedings of the Royal Society B: Biological Sciences*, 283(1823):20152347.
- 406 Laurance, W. F., Clements, G. R., Sloan, S., O'connell, C. S., Mueller, N. D., Goosem, M., Venter,
407 O., Edwards, D. P., Phalan, B., Balmford, A., et al. (2014). A global strategy for road building.
408 *Nature*, 513(7517):229–232.
- 409 Leibold, M. A., Holyoak, M., Mouquet, N., Amarasekare, P., Chase, J. M., Hoopes, M. F., Holt,
410 R. D., Shurin, J. B., Law, R., Tilman, D., et al. (2004). The metacommunity concept: a framework
411 for multi-scale community ecology. *Ecology Letters*, 7(7):601–613.
- 412 Lever, J. J., van Nes, E. H., Scheffer, M., and Bascompte, J. (2014). The sudden collapse of
413 pollinator communities. *Ecology Letters*, 17(3):350–359.
- 414 Loreau, M., Mouquet, N., and Gonzalez, A. (2003). Biodiversity as spatial insurance in heteroge-
415 neous landscapes. *Proceedings of the National Academy of Sciences*, 100(22):12765–12770.
- 416 May, R. M. (1972). Will a large complex system be stable? *Nature*, 238(5364):413–414.
- 417 McCann, K. S. (2000). The diversity–stability debate. *Nature*, 405(6783):228–233.
- 418 McCook, L. J. (1999). Macroalgae, nutrients and phase shifts on coral reefs: scientific issues and
419 management consequences for the great barrier reef. *Coral Reefs*, 18:357–367.

- 420 Panahi, S., Do, Y., Hastings, A., and Lai, Y.-C. (2023). Rate-induced tipping in com-
421 plex high-dimensional ecological networks. *Proceedings of the National Academy of Sciences*,
422 120(51):e2308820120.
- 423 Pilosof, S., Porter, M. A., Pascual, M., and Kéfi, S. (2017). The multilayer nature of ecological
424 networks. *Nature Ecology & Evolution*, 1(4):0101.
- 425 Potts, S. G., Biesmeijer, J. C., Kremen, C., Neumann, P., Schweiger, O., and Kunin, W. E. (2010).
426 Global pollinator declines: trends, impacts and drivers. *Trends in Ecology & Evolution*, 25(6):345–
427 353.
- 428 Rohr, R. P., Saavedra, S., and Bascompte, J. (2014). On the structural stability of mutualistic
429 systems. *science*, 345(6195):1253497.
- 430 Saade, C., Fronhofer, E. A., Pichon, B., and Kéfi, S. (2023). Landscape structure affects
431 metapopulation-scale tipping points. *The American Naturalist*, 202(1):E17–E30.
- 432 Saavedra, S., Rohr, R. P., Bascompte, J., Godoy, O., Kraft, N. J., and Levine, J. M. (2017). A struc-
433 tural approach for understanding multispecies coexistence. *Ecological Monographs*, 87(3):470–
434 486.
- 435 Scheffer, M., Carpenter, S., Foley, J. A., Folke, C., and Walker, B. (2001). Catastrophic shifts in
436 ecosystems. *Nature*, 413(6856):591–596.
- 437 Scheffer, M. and Carpenter, S. R. (2003). Catastrophic regime shifts in ecosystems: linking theory
438 to observation. *Trends in Ecology & Evolution*, 18(12):648–656.
- 439 Suding, K. N., Gross, K. L., and Houseman, G. R. (2004). Alternative states and positive feedbacks
440 in restoration ecology. *Trends in Ecology & Evolution*, 19(1):46–53.
- 441 Thébault, E. and Fontaine, C. (2010). Stability of ecological communities and the architecture of
442 mutualistic and trophic networks. *Science*, 329(5993):853–856.

- 443 Urban, D. and Keitt, T. (2001). Landscape connectivity: a graph-theoretic perspective. *Ecology*,
444 82(5):1205–1218.
- 445 Valdovinos, F. S., Brosi, B. J., Briggs, H. M., Moisset de Espanes, P., Ramos-Jiliberto, R., and Mar-
446 tinez, N. D. (2016). Niche partitioning due to adaptive foraging reverses effects of nestedness
447 and connectance on pollination network stability. *Ecology Letters*, 19(10):1277–1286.
- 448 Valdovinos, F. S., Moisset de Espanés, P., Flores, J. D., and Ramos-Jiliberto, R. (2013). Adaptive
449 foraging allows the maintenance of biodiversity of pollination networks. *Oikos*, 122(6):907–917.
- 450 van de Leemput, I. A., van Nes, E. H., and Scheffer, M. (2015). Resilience of alternative states in
451 spatially extended ecosystems. *PLoS One*, 10(2):e0116859.
- 452 Walker, B. (1995). Conserving biological diversity through ecosystem resilience. *Conservation
453 Biology*, 9(4):747–752.
- 454 Walther, G.-R., Post, E., Convey, P., Menzel, A., Parmesan, C., Beebee, T. J., Fromentin, J.-M.,
455 Hoegh-Guldberg, O., and Bairlein, F. (2002). Ecological responses to recent climate change.
456 *Nature*, 416(6879):389–395.
- 457 Wang, S. and Loreau, M. (2014). Ecosystem stability in space: α , β and γ variability. *Ecology
458 Letters*, 17(8):891–901.
- 459 Whitehorn, P. R., O’Connor, S., Wackers, F. L., and Goulson, D. (2012). Neonicotinoid pesticide
460 reduces bumble bee colony growth and queen production. *Science*, 336(6079):351–352.
- 461 Wilcox, B. A. and Murphy, D. D. (1985). Conservation strategy: the effects of fragmentation on
462 extinction. *The American Naturalist*, 125(6):879–887.