**Running head:** Biotic factors and succession along tropical elevational gradient

**Title: The effects of pathogenic fungi, herbivores, and predators on pioneer vegetation of tropical forests change with elevation.**

Piotr Szefer, Austin Sau, Kenneth Molem, Jonah Philip, Martin Volf, and Vojtech Novotny

**Abstract**:

In the tropics pathogenic fungi, herbivores and their predators have a significant impact on plant community assembly processes. These plant-enemy interactions can be notably influenced by environmental gradients. In the context of regenerating tropical forest, this phenomenon can potentially affect the trajectory of succession at different elevations, and cause variation inflation in succession outcomes. Moreover, the relative significance of individual biotic factors can vary depending on elevation and may be differentially influenced by key plant traits. In this paper we present the results of a manipulative experiment replicated at three elevations to study the effect of focal predators (birds, bats, and ants), herbivorous insets and pathogenic fungi on plant communities on vegetation at the pioneer stage of tropical forest regeneration.

In the experiments responses of biotic factors to elevation depended on a specific characteristic of the community (biomass, diversity, richmess or density). We defined top-down effects as variation in the LRR that cannot be explained by the quadratic function of an experimental garden richness and biomass. Top-down and bottom up effect were both present and the same factor can have top-down effect on one of the community characteristic and at the same time some other characteristic can be bottom-up controlled. Some linear trends in top-down control were observed for diversity and fungi (increasing in top-down strength), richness and insects and biomass and predators (decreasing in strength). Trophic cascades caused by the focal predators were weak, but got stronger and more evident with increasing elevation. Insects shape community composition of pioneer stages of regeneration resulting in dominance of a few early succession woody plant species. Fungi influence composition and increase density of woody plants only at the highest elevation. Fungi affected plant community only at the high elevation. At the low elevation insects increase dissimilarity by decreasing abundance differences between sites. At mid and high elevations insects reduce and increase the species turnover rate respectively. [???] Predictive power is reduced at mid-elevation at Numba and plats obtain lower biomass due to the lack of insects. The strongest predictor [of what?] was biomass, which means that plants that would be established at the site are also the ones that benefit most from the biotic factor removal. All these results point to a strong bottom-up control.

**Including species richness as a co variate in ordination removes the effect of treatments. Similar conclusion can be drawn from model comparisons, where in most cases the most complicated models are the best, these are models comparing treatment and site effects with models. The effects that were better predicted by the bottom-up forcing (higher adjusted R2 values an dlowes AIC values for models without the treatment effect) were richness and density. Treatment was a significant factor in case of diversity and biomass effects. But generally the predictive power of the models was relatively low (looking at the adjusted R square).**

When abundance, species richness, locations, species’ ability to change their SLA value, and phylogeny SLA significantly improved model for predator exclusion only. Water content of leaves was significantly in all models and significant for fungicide treatment. Percent herbivory improved models for predator and insecticide treatments.

**Key words**: tropical forest, pioneer succession, fungicide, insecticide herbivory, tri-trophic interactions, Papua New Guinea.

**Introduction**

Identifying the deterministic predictors of successional dynamics is important for better predicting successional trajectories (Chazdon 2008, Norden et al. 2015). The pioneer stages of succession play a crucial role in determining the pace, progression, and endpoint of succession.  Natural enemies play a significant role in shaping competitive dynamics within real-world plant communities (Comita and Stump 2020). Herbivorous insects and pathogenic fungi can influence competitive outcomes thorough many mechanisms, including apparent competition (ref), density dependent (ref) and enemy-spillover effects. Multiple studies conducted in tropical forests have demonstrated the importance of fungi and herbivorous insects as mortality factors (Gripenberg et al. 2014, Bagchi et al. 2014). Furthermore, the effects of natural enemies on plant communitites can vary depending on specific plant traits (Jia et al. 2020).

Biotic interactions, including interactions with natural enemies, can have a profound impact on the survival and dynamics of plant species. High competitive strategies may confer resilience against natural enemies, but plants must strike a balance between resistance and competitiveness (Radny and Meyer 2018). The expression and variation of traits during the assembly process of tropical forests are often influenced by biotic factor filters. For instance, resource-acquisitive strategies may be disadvantageous when natural enemy densities are high and heterospecifics densities are low (Adler et al. 2013, Lasky et al. 2014). This can hinder facilitation process and slower down the succession (Fichtner et al. 2017).

Natural enemies have been found to enhance the predictability of community composition and increase diversity without causing a significant negative effect on plant biomass and their actions seem to be trait mediated (higher SLA and lower LDMC when insects are present) (Szefer et al. 2020). Additionally birds, bats, ants and other predators put constant pressure on herbivorous insects. The dynamics of tri-trophic interactions may are determined by traits. Predator feeding mode affects the behaviour of the pray, which in turn may result in dominance of specific traits, and change the plant community (Schmitz 2017). These affects seem to be stronger in more mature ecosystems of primary tropical forests (Letourneau et al. 2004, Barber and Marquis 2011, Sam et al. 2015, Roslin et al. 2017, Houska Tahadlova et al. 2023), and may be also important for successional proceesses. In contrast, forest gaps food webs, at least in lowland PNG, are bottom-up controlled (Szefer et al. 2022), but the variability of these effects haven't been yet evaluated along ecological gradients.

Impact of individual biotic factors may differ along ecological gradients (Prach and Walker 2011). For example, gradient of increasing distance from the wet tropical forest interior, weakened the density-dependent effect of pathogenic fungi and herbivorous insects on seedling survival (Krishnadas et al. 2018). These effects however may not be as simple alnog natural gradients. The loss of crucial biotic interactions may be sometimes compensated by others in being the main sources of mortality, like it may be in the case of competing key mortality factors like *vertebrate seed-eaters vs* pathogenic fungi and insects (Williams et al. 2021). It is therefore important to understand how plant-enemy interactions shape regeneration in contexts of changing diversity and abiotic conditions. Enemies along the gradient (Krishnadas et al. 2020) [*good citations from the introduction there*]. Including the local natural plant-enemy and tri-trophic interactions may shed some light onto unpredictability of succession (ref, how good are these predictions?, examples where this worked? Stronger tropic cascades are predicted for vertebrate predator and invertebrate herbivore (Borer et al. 2005). This would suggest that bird predation may be relatively more important at higher elevations for plant richness and diversity. “Diversity within trophic levels may limit tropic cascade strengths through compensatory dynamics within tropic levels” according to Community Regulation Hypothesis (Holt and Polis 1997, Pace et al. 1999).

**Elevation gradients are well established models for studying potential climate changes and diversity impacts (Becker et al. 2007), and the impact of biotic factors and the strength of tri-trophic interactions may change alnog it**. For example, predation rates in the tropics are predicted to be stronger (that theory in Roslin et al 2017) at lower altitudes (Roslin et al. 2017). Similarly, the diversity and abundance of another key predator in tropics – ants, decreases and is nearly absent above 1500 m a.s.l. (ref).With reduced pressure from top preedators, the top-down control may get weaker. In fact our study suggests, that in lowland tropical forest gaps bottom-up effects may dominate, and that fungi may be involved in more complex not only antagonistic, but also mutualistic interactions (Szefer et al. 2022). But the elevational gradient includes also shift in trait values (*“plant defences, particularly leaf toughness and flavonoid compounds, tend to be higher at high elevations”* (Rasmann et al. 2014).) of slow growing, more resource conservative plants. This coupled with decreasing abundances at the consumer levels (herbivores and their predators) at higher elevations **fungi may be the most important density mortality factor at higher elevations (ref)**. Thus, succession at higher elevation may be more affected by recruitment limitation and functional traits, particularly growth rate, of individual plant species and how this relates to natural enemies. At the higher elevations plant traits related to growth rate and characteristic for the acquisitive resource use strategy (SLA, Water content) should be better at predicting better species’ success in the absence of natural enemies.Plant recruitment limitation (species richness high or low when recruitment limitation is low or high respectively) and inter-specific competition (high competition when diversity is low and biomass high), makes species’ growth rates, approximated by SLA (Osone et al. 2008, Kleinschmidt et al. 2020), an increasingly better predictors of their success in secondary succession with higher elevation.

Pioneer plants posses specific morphological and physiological characteristics that enable them to establish and thrive in newly created environments (ref, Chazdon, Lohbeck 2013 z Impact of .). For a pioneer woody plant, achieving canopy development despite competition from grasses and herbaceous plants is considered a success (Charles 2020). ***Successfull species should be able to have good gap colo****nization capabilities and strong competitive ability (Pywell et al. 2003). Wood density and proximity to the intact forest predicted survival of canopy trees, but it may be different for secondary successional trees (Charles et al. 2018). Therefore,* pioneer plants in tropical forests exhibit numerous acquisitive traits, such as high photosynthetic activity, rapid growth rate, high specific leaf area (SLA), **low wood and tissue density**, and low herbivore resistance. Although these traits make them vulnearble to higher mortality rate, they can compensate for herbivore damage (Whitfeld et al. 2012, Chazdon 2014).

When studying regeneration, trait-based approaches are more practical and informative compared to species-level analysis (ref). Traits have proven to be effective in the restoration of specific targets defined by trait composition (Laughlin 2014) and have the potential to predict long-term survival and growth rates (Werden et al. 2018). What is notable is that species' long-term growth and survival rates in a regenerating community can be accurately predicted based on their traits and short-term dynamics **(Mart**ínez-Garza et al. 2013)**.** However, it is important not to overlook the influence of biotic interactions, which can significantly alter competitive outcomes between species (Kim et al. 2013) and affect trait expression in plants (Radny and Meyer 2018, Koprivnikar et al. 2021).

Here we present results of an experimental exclusion of fungal pathogens, insects and spiders and their predators from pioneer (0-1.5 year) stages of regenerating gaps in tropical forest at three elevations. We hypothesize that the strength of the plant-natural-enemy interactions **increases with elevation and drive changes in regenerating plant community assembly and dynamics. Specifically we predict, that natural enemies** (i.e. fungi, insect herbivores) will **increase** richness and diversity of the vegetation consistently along the elevation as more conservative and slow growing species are present. The biotic factors’ will have different effects on the community assembly processes, i.e. are filters that increase prediction of the future community composition. With a reduced control of predators, sronger effects of herbivorous insects and fungi, as well as stronger abiotic filters at higher elevations should in turn increase the predictability of the community composition. Finally, we attempt to evaluate how biotic factors filter palnts based on their traits (leaf water content, SLA) in regenerating community, and evaluate variation in this relationship along elevation.

**Materials and methods**

*Experimental design*

Experimental treatments (fungicide MANCOZEB …, Imidacloprid + Lambda cychalothrin, Exclosures, Herbivore addition, locally dominant generalist Curc001 in Wanang and CurcXXX in Numba and Yawan, Undisturbed growth and repeated treatments. P**lant species from three tropical forest sites: Wanang (200 m a.s.l.), Numba (700 m a.s.l.), and Yawan (1900 m a.s.l.). We refer to the sites as low, mid- and high elevations throughout the text.**

*Statistical analyses*

We used log-response ratios (LRRs) to compare the effect strengths between sites. We used Bray-Curtis dissimilarities between control plot and each of the four treatments (fungicide, herbivore, insecticide and exclusion of key predators) to test our prediction that some treatments may affect community composition more strongly at higher elevations.

*SLA analysis*

We modeled natural logarithm of total biomass at a given plot predicted with logarithm of SLA [m2/kg dry leaf mass] as a linear function using a phylogenetic generalised least square (assumption, both high and low values of SLA may not be the best option for plants **[explain in the intro why it is like that]**). We tested for the best modle of the trait evolution using the function fitContinuous. And we chose based on likelihood the OU model. For which we estimated the alpha parameter, that was later used to provide correlation structure for the model.

*LRR correlations*

If fungi can compensate for herbivorous insect mortality, then reduction of one of the effects in a manipulative experiment would cause an increase of the other. This can be partially evaluated using correlation of the effect ratios. If the compensation is complete, or in a sense affects the same ‘fitness components’ [functional redundancy] of a plant community then we would expect no correlation between positive LRR effects of the two factors. Additionally, all individual treatment effects should be around zero: because removal of any given natural enemies would be immediately and completely replaced by another. But in case where there is a correlation, this would exclude this type of compensation. We assume that the effects of the following factors: all herbivorous insects, Correlation of predators and herbivore exclusion and herbivore addition.

We calculated log-response ratios (LRR) for plant biomass, species richness, species diversity and stem density to be able to compare the effect sizes at individual elevations. LRR is generally calculated as the logarithm of ratio of values for present to not-present sites. Three treatments (netting, insecticide, and fungicide spraying) were exclusion treatments and for these LRR was calculated as *log(xc/xt)*. This gives the effect of a given biotic factor. For the addition of herbivores however the formula was modified to *log(xt/xc)*.

Bertellet’s and Levene’s homogeneity tests were performed for LRR of all descriptors comparisons between sites. If at least one test showed lack of homogeneity we accounted this in statistical model by providing *varIdent(form = ~1|site)* option as weights in the *gls* function.

Community composition … and Bray-Curtis analysis, Random vs deterministic changes (Raup-Crick).

We approached species success’ depandency on traits in two ways. First we used random forest method to evaluate importance of total sppecies biomass at a given plot, the plot richness, species SLA, water content and herbivory on their presence-absence dynamics when comparing control plot with a treeatment plot. I.e., how well traits predicted wether species werer lost, gained or remained on the treatment plot compared to the control. Algorithm…, hyperparameters?

CWM for different treatments we also evaluated whether treatments affected trait community composition, by calcualting community weighted means (CWM) for SLA, water content and herbivory [**i dont think I have performed these analyses yet**].

Phylogeny …

Response variable for the SLA effect ...

We built multiple, competitive statistical models explaining relationship between species’ traits and their abilities to change biomass in response to a treatment. Those included linear and quadratic relationships, and site interactions while accounting for phylogenetical relationships between species. We used chi-square test of the log-likelihood ratio to compare models. If two models were not significantly different from each other then we would always go with a model with lower AIC. We compared model without traits (null model) and with traits, allowing for nonlinear (quadratic) trait vs Δbio relationship. If the model with the traits was better, we tested whether removal of a single trait would improve the quadratic model. With a two trait model we would try to reduce it further to only one trait. Final model would then be tested for the necessity of the quadratic terms, which will be removed in all combinations for two/three trait model. Test for the interaction. The final model co-variates were always plant species richness at the control plot, logarithm of abundance (number of stems >1 cm DBH in the plot); log-response of SLA for a species (to account for SLA plasticity for a given species and the location, All GLS models accountd for phylogeny. This ensures that we look only at the trait effect on the biomass change.

**Results**

We identified **157** (total of 8594 kg) **woody plant species**, with an average range from xxx to xxx per plot (Tab. SXX). Our data revealed that diversity at the 750 m location was significantly higher compared to other locations (Supplementary Information: Appendix I: Fig. S1). Additionally, based on a quadratic Poisson regression model applied to our data, we observed that richness peaked around 1250 m a.s.l. Elevations were similar in their baseline productivity (total biomass, **Fig. S1**). At the high elevation the plant abundance, richness and diversity was significantly higher than at the low elevation. Mid elevation had many outliers. However, no significant differences in variance between the sites were detected (**Tab. S1**).

Fungi exhibited a positive effect on woody plant richness at low and high elevations (Fig. 1), but this effect was not significant at the mid-elevation. At the high elevation the removal of fungi also had a positive effect on woody plant density. Insects only reduced biomass at mid-elevation. However, at the high elevation insects increased diversity and richness. Removal of insects caused increase in richness at low and high elevations. Predator removal resulted in an increase in diversity and richness, but only at the high elevation. The addition of herbivores reduced biomass and woody plant density at low elevation, as well as richness at both low and high elevations. In the GLS function) showed significant difference of LRR in Yawan, that was not observed for LM model (**Tab. S1**).

Correlation between LRR of the insect effects (i.e. insects vs additional insects, and insects vs predators) were observed in 37.5 % of cases. There were no correlation at the lowest elevation. Close number of correlations in mid and high elevations (four and five out of eight respectively) coincided only in the case of biomass for the insect vs predator effects comparison (**Fig. S2**). If we also include correlations between predator exclusion (P) and P+A this ratio gets significantly higher for Yawan (9 out of 12 correlations). All observed correlations of the **insects** with **top-predators** effects were positively correlated. In contrast, all insect effects and effects of herbivore addition were negatively correlated.

Insect exclusion had a significant impact on the community composition across all elevations. The application of fungicide treatment had a similar effect, but only at the highest elevation (Fig, 3; Appendix I, Tab. S1).

At the low elevation, the control group exhibited higher baseline dissimilarity compared to mid-elevation. Only the addition of herbivores resulted in a consistent decrease in dissimilarity with increasing elevation (Fig. 4).

At the low elevation, the species composition primarily changed through the turnover of species (high balanced component, low gradient component of dissimilarity). However, at the two other locations, the community underwent changed through shifts in the abundance structure were more pronounced than species turnover. Treatments affected this partition. At the low elevation insects reduced dissimilarity and by mainly reducing the gradient component (**Appendix I, Fig S4**). At the mid elevation reduction of the balanced component by insect and predator presence and increased levels of herbivory were leveled out by increase in the gradient components, but not in the herbivore addition treatment. This resulted in no changes to general dissimilarity. At the highest elevation increase in balanced component due to insects and presence of predators was reflected in the increase in dissimilarity. No change in the gradient component were observed.

At the highest studied elevation, all treatments caused a shift from deterministic in their divergence (predictably unpredictable to random) towards random community assembly (Fig. 3). Similar shift, but weaker in cases of insecticide and fungicide were observed at the lowest elevation. Opposite patterns were observed at the mid-elevation, where communities showed rather random assembly and only addition of herbivores significantly increased divergence of the community composition.

Plants with higher water content had more biomass per species in the absence of herbivores and fungi in mid and high elevation (**Tab. SXXX, Appendix I**). Species’ faith (i.e. wether they were lost, gained, or stayed compared to the control) was best predicted by individual species’ total biomass. In all cases water content was the second most important trait, while SLA and herbivory as the third most important trait only at the fungicide treatment.

Three studied traits varied in their importance in predicting the change in biomass (Δ biomass) in response to a treatment (Tab. XXX, Fig. XXX). SLA was significant, nonlinear only in the predator exclusion treatment. Species with higher water content tend to increase biomass in response to the fungicide treatment. Water content was a significant co-variate in all models. Most variation in the biomass change was explained by abundance, site identity, intra-specific shifts in SLA values.

**Discussion**

Our results show that biotic factors affect pioneer vegetation, but the direction and magnitude of their effects are elusive and vary depending on the properties of the studied plant community and the elevations. The biomass remained robust to experimental treatments, indicating a strong bottom-up biomass control. However, we observed that the effects of fungi, insects, and predators on richness, diversity, and density were generally positive at high and low elevations, while they were absent at the mid elevation. This pattern suggests a clear top-down forcing influence. At least for richness, density and diversity measures, patterns seem to be correlated with the local plant species richness at each sites: low species richness plots have stronger positive effects and high species richness site (Numba, mid elevation) has weak to no effects. However, species richness was a poor predictor of a LRR at the level of individual gardens. In case of strong bottom-up control, response of plant community to a treatment should not be correlated with its biomass or richness at the control (benchmark) plot. If under a treatment we see significant correlation this indicates that the effect of the treatment is correlated to initial community. This is a convoluted effect, meaning that it is top-down gets stronger with more biomass – which is also a definition of bottom-up… complicated.

where changes in community are driven by the initial community we should be able to explain much of the LRR effect based on the initial community composition (control plots). The direction and strength of the effects seems to be richness-related. **Turns out that the strength of the effect is directly related to the baseline biomass at the plot, regardless of the elevation… (and treatment?) [CHECK IT OUT]. Is high richness of a plot reduces the effect of top-down factors**. This is another evidence for bottom-up control.

Fungi and insects show similar effects at different elevations. High and low elevations exhibit increased richness and density, possibly due to negative density dependence and niche space expansion for plant species. At the mid elevation, fungi and insects can compensate for each other's effects, resulting in a nullification of individual treatments. [***COMPENSATION***] If compensation is present then this also affects the results of insects effects. Lets say that we remove insects from a plot, and there is a strong compensation by either fungi or other factors, like a fast recolonization, then the effect would not be noticeable for both of these treatements, and there will be no correlation between their effects. To be certain about the presence of these effects additional treatment combining insecticide and fungicide treatment should be added. Additionally, if top-down effects are strong, herbivores are strongly suppressed. In this situation, insecticide treatment may have negligible effect on vegetation (no effect of insects, but effect of predators sign? - density in Y). Of course we cannot exclude also other factors, like rhodents and soil communities. This calls for more thoughtful experimental designs. Lack of correlation does not necessarily imply lack of effects, and effects are often present without correlation. Mid elevation is the only exception where no effects or correlation are observed. The presence of effects without correlation suggests different underlying mechanisms.

Significant positive effects of top predators on richness and diversity of top-predators occurred more often at the highest elevation. In a classic, mechanistic model of the top-down control, when the effects of insects on vegetation biomass are negative/positive, then we would expect the effects of predators to be the opposite. The lack of a general effect does not necessarily mean that the **cascading effects** are not present. But at least, *negative correlation of the effects is expected*. Significant correlations of herbivororus insects and their predators effects on vegeation, *were either not significant or positive for biomass, diversity, richness, and density*. It means that the exclosure treatment has the same effect on vegatation, as the insecticide treatment. It is possible, that in the exclosure spiders are released from the predatory pressure and exert stronger pressure on herbivores. Limitations of surveying the plots for insects to see exactly what had happened. *Exclusion intensifies spider predation either by* ***creating conditions for web building spiders****, or by reducing predatory spider mortality, which intensifies predation on herbivores. At least in Wanang no cosistent increase in spider abundance was observed. But we concluded weak top-down control there (basically no effect of the exclosure). This might be different for N and Y (****evaluate using preliminary data****).*In conclusion: Presence of predators increases the total abundance of insects and decreases (although not always) the proportion of Arachnids. So there is some supprot for our claim that predatory birds in pioneer tropical forest vegetation may be feeding predominantly on Arachnids, which than cascades onto the insect herbivore community.]

In the case of a strong bottom-up control there should be no correlation between LRR effects. These positive correlations of insects and predator effects on biomass were absent at the lowest elevation studied. This could suggest weak effects of birds at these elevations. These exact patterns plays out in N and Y for richness, diversity and density, but more often at the highest elevation. This may suggest that top-down control and cascading effects of birds, bats and ants may be stronger at the highest elevations. [Exclusion of top predators has the same effect as insect exclusion on species turnover in Numba and Yawan. While in Wanang this may suggest compensation by fungi, at the high elevation the correlation of effects indicates some additional mechanins within the tritrophic interaction. Addition of generalist herbivores also correlated with the predator effects. If the effects of insects are mainly caused by specialists, then generalists interfering with their feeding could weaken their effects.

Only insects affected the plant species composition at all elevations, and fungi also at the highest elevation. The analysis of beta diversity components show that all insect related treatments (P, P+H, I) affect species turnover and dominance structure in very similar way. Insect herbivores acted as filters for species at the mid elevation, but they increased species turnover at the highest elevation. At the lowest elevation, only insects had an effect on dominance structure changes. These effects again may relate to humped-shaped patterns in regional species richness with a peak at the mid-elevation. In Numba species rich regional pool may increase the competitive landscape for plants, and it can be more difficult for species-specific natural enemies to accumulate, because the community composition is unpredictable. Assembly analysis confirms this idea. In Numba, plant community has a strong evidence that it assembles randomly. On the other hand, when species pool of pioneers is smaller, like at the high and low elevations, insects may increase richness most probably through apparent competition, and open up space for new species. What could be the reason for such stark differences? At the lowest elevation species composition changed mainly through turnover of species (balanced component), in contrast to two other sites, where community changed through a combination of both species turnover and the abundance structure (gradient component). *At the highest elevation, there was a shift in the proportions of beta diversity components. Balanced component was reduced and gradient component increased. This means that in the absence of insects, species turnover plays bigger role in the community processes at the highest elevation.* Reduction of dissimilarity means higher predictability of the plant community assembly. **At the high species turnover site, the effects of biotic factors may be weaker than at more stable sites**. If community is highly variable, contingent site effects can make detection of a biotic factor effect challenging. **Insect herbivores (P, I, H combined) can suppress plants in communities with low variability in species composition (Numba)**, which allows new species to establish. Moreover, insects reduce dissimilarity for highly variable communities.

We tried to investigate whether **SLA will become better predictor of plant success (fitness) with elevation. This would be the case if traits would become better at predicting the faith of an individual species.** CWM was negatively correlated with species richness, which means that dominant species in more rich communities had lower SLA and LDMC. This pattern was also observed by Zang et al. (2019) (CWM Among the eight measured traits, the SLA, LNC, and LPC were negatively associated with local species richness Zang et al 2013) for the old growth forest and was attributed to shifts from resource acquisitive to conservative strategies with accumulating species. Our result shows that this relationship is always present and it is not a result of trait value shifts by individual species. *One of interesting results from our study was a g*eneral negative correlation of CWMs of SLA and the leaf water content was **higher at low diversity plots**. This means that Higher SLA was a better strategy or was a result of a lack of competiton between plants at species poor sites. This relationship of CWM and species richnes was independent of elevation and treatment. C**WM change was due to having species with lower SLA to achieve higher abundances in the high richness plots (Supplement Figure SX), rather than species adjusting their SLA.** Plants showed **limited ability to change their SLA in response to treatments** and richness of the surrounding community [figure?] (Figure SX. SLA~richness).

At least in the case of **SLA richness does not correlate with variation in SLA** [what???], as it would be predicted by the limited similarity hypothesis. But it is difficult to make such broad statements based only on two traits. Species optima are better evaluated in a multi-variate trait space (Muscarella and Uriarte 2016). At species poor plots, plant species with higher SLA were able to dominate (and increase the overall weighted average of the SLA). Relying on higher SLA is not a good strategy in species rich communities. This might be caused by an increased competition between plants for some limited resource or resources. In Wanang none of the treatment had the ability to change the shape of this relationship.

If traits were better at predicting species success at higher elevations as predicted in case of weak top-down control we should see significant correlation between SLA and Log(TotBio) at the control plot and the R2 should increase with the elevation. This trend is not present at all. If a factor affects community composition SLA should be **even better predictor of success when the factor is removed**. This is not the case. What else can we get from the phylogenetic regression analysis? That the correlation seem to be present and stronger at some treatments. Insects select based on SLA. Plants with lower SLA tend to have higher biomass irrespective of the elevation. Numba is the most peculiar site: Lower intercept for both insecticide and fungicide treatment. In Numba the slope for the insecticide treatment is less steep.

**When we controlled for baseline abundance, richness, site identity, trait plasticity (LRR SLA), and phylogeny we found rather limited effect of studied traits. Differential effects of elevation found only for the SLA and water content, at predator an P+H sites.** It seems that the effects of insects are not trait mediated. At some circumstances however, like our hypothesized intermediate predation there seem to be a fair amount of variation explained by it. **SLA is a moderately good predictor of insect herbivory** (Whitfeld et al. 2012, Zhang et al. 2023). There are different optima for species at each elevations. In Wanang, plants with average SLA will respond positively, or at least non-negatively. When we add herbivores to the exclosure, their additional effect will affect water content, and not SLA, but it will promote high water content plants in the exclosure with herbivores, this was found also for fungi. This is different in Yawan and numba. For exclosure numba has the opposite effect to that found in Wanang for SLA and Yawan has a linearly decreasing effect with SLA.

Fungi, insects and predators show differential effect with elevation on biomass, richness, diversity and density of pioneer succession forest. Surprisingly, elevated abundances of herbivores had no effect on plant biomass in Numba when there was a clear effect of insects at this elevation. This lack of consistency because the ability to reduce plant biomass is attributed mainly to highly specialized herbivores. That would, however result in some effects of insects on richness and diversity. These were not observed. Numba is the most species rich from our sites. These species have lowest SLA [SITE COMPARISON, SPECIES BASED SLA] from all sites. These species may have limited compensation abilities. When there are no insects plants with higher SLA tend to occupy the community [there should be no difference between ]. This cause the increase in biomass. For the natural communities lower SLA reduces ability of additional herbivores to impact the biomass. It can be tested in the second manuscript. But without morpho-species I would not be able to evaluate specialization.

Insects affect processes more. They change community composition but possibly not through density dependent effects. Insects have weaker effects at the mid elevation at richness, biomass and diversity but different mode of action [explain]. Selecting for higher SLA plants can be simply one of the mechanisms of the insect action. If biotic factors are not so important to the community processes then SLA relationship should be positive at least for the Numba site.

Keep in mind that we are probably are affecting mutualistic species more than necrotrophs (In our case Glomus AM, but it can be also parasitic). So first observation is that fungi increase species richness. The effect is not huge but consistently higher than zero at all elevations. Fungi increase density of woody plants at highest elevation. In addition, decrease diversity at the lowest. Mutualist should help plants to endure competition (“AM fungi significantly contribute to the uptake of soil nutrients, increase plant biomass and confer on the plant improved resistance to stress and pathogens” [Smith, S. E. & Read, D. J. *Mycorrhizal Symbiosis* (Academic Press, 2008)]).

**IN summary, the e**ffects of insects and other biotic factors are weak at the mid-elevation. The effects of herbivores and pathogenic fungi seem to affect trait community composition, but are absent at Numba (no relationship between SLA and tota biomass), which is surprising. Most complex was the situation in Yawan, where we got effects of fungi, insec herbivores (both addtion and removal) as well as tropic cascade of top-predators. High diversity of plants at the 750 reduces the effects of fungi and herbivores and cascades Herbivores may be more specialized there also because of the mid-range effects and thus reduce their feeding range. Also variation in species composition between gardens may be higher here [I can test that]. This definitely does something to the biotic factors. Only there insects had negative effect on biomass, but increase in generalist abundance did not. Could generalists induce compensation? Would that result in the change in SLA response it is definitely correlatied as only Numba had some significant relationship for the SLA ~ totbio. But that seems to be bottom up control of the effects? And possibly also the plant community is more strongly affected by the competitive processes at 750. We concluded that at 200 effects of birds and bats are weak. Cascades seem to increase in strength at the highest elevation (and affect richness and diversity). Birds bats and ants somehow can affect arthropod predators, that releases insects from the pressure of their arthropod predators and this may have positive effect on plants.

Only insect removal was able to change the plant community composition. Only at highest elevation fungi were also responsible for the change as well. Predictability of a species success based on their traits was limited. Some further questions: (1) Is stronger inter-species competition reduces the importance of the tropic cascades? (2) What are possible consequences of the **bottom-up regulated richness vs top-down regulated richness [SIMULATIONS?]**. What about the community composition? Insects, reduce conspecific densities near shade-intolerant species: “which natural enemies are responsible depends on the **mycorrhizal association** and **shade tolerance** of tree species” (Jia et al. 2020).

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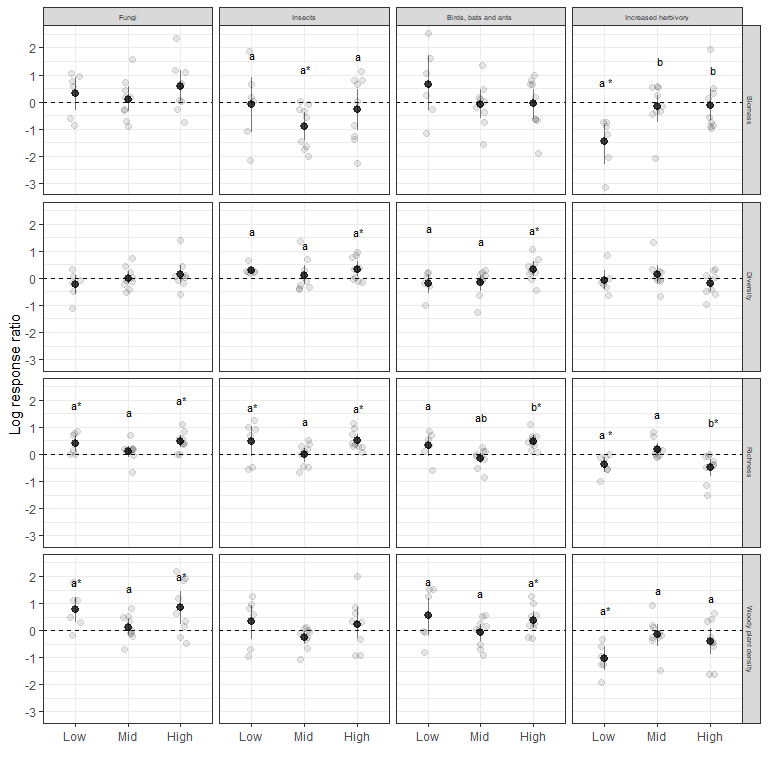
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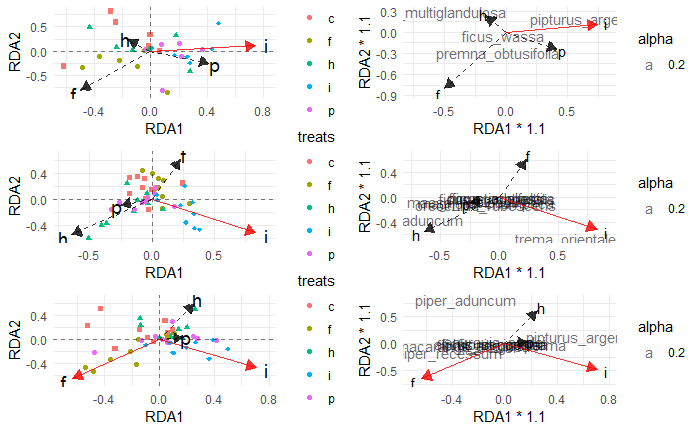
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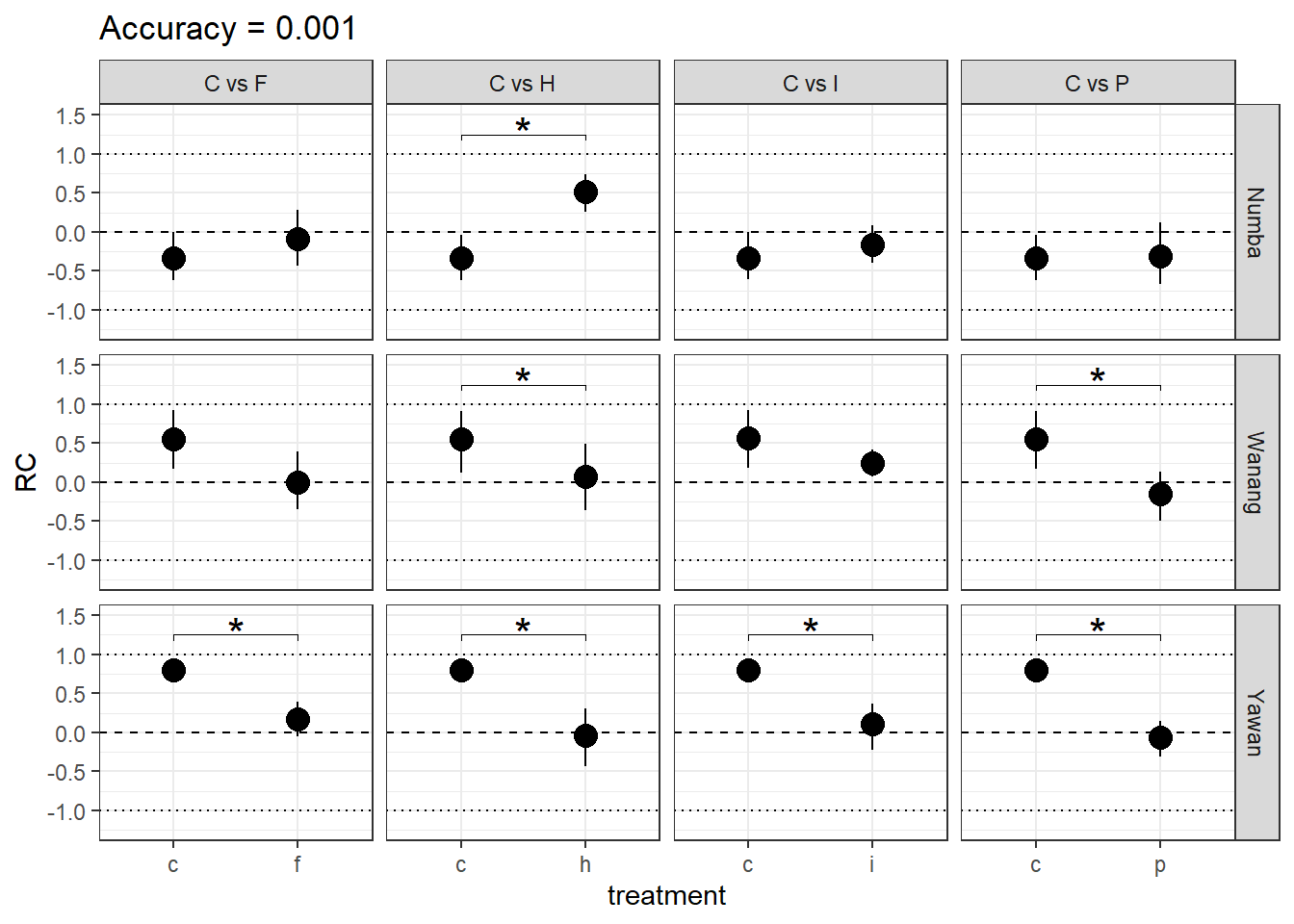
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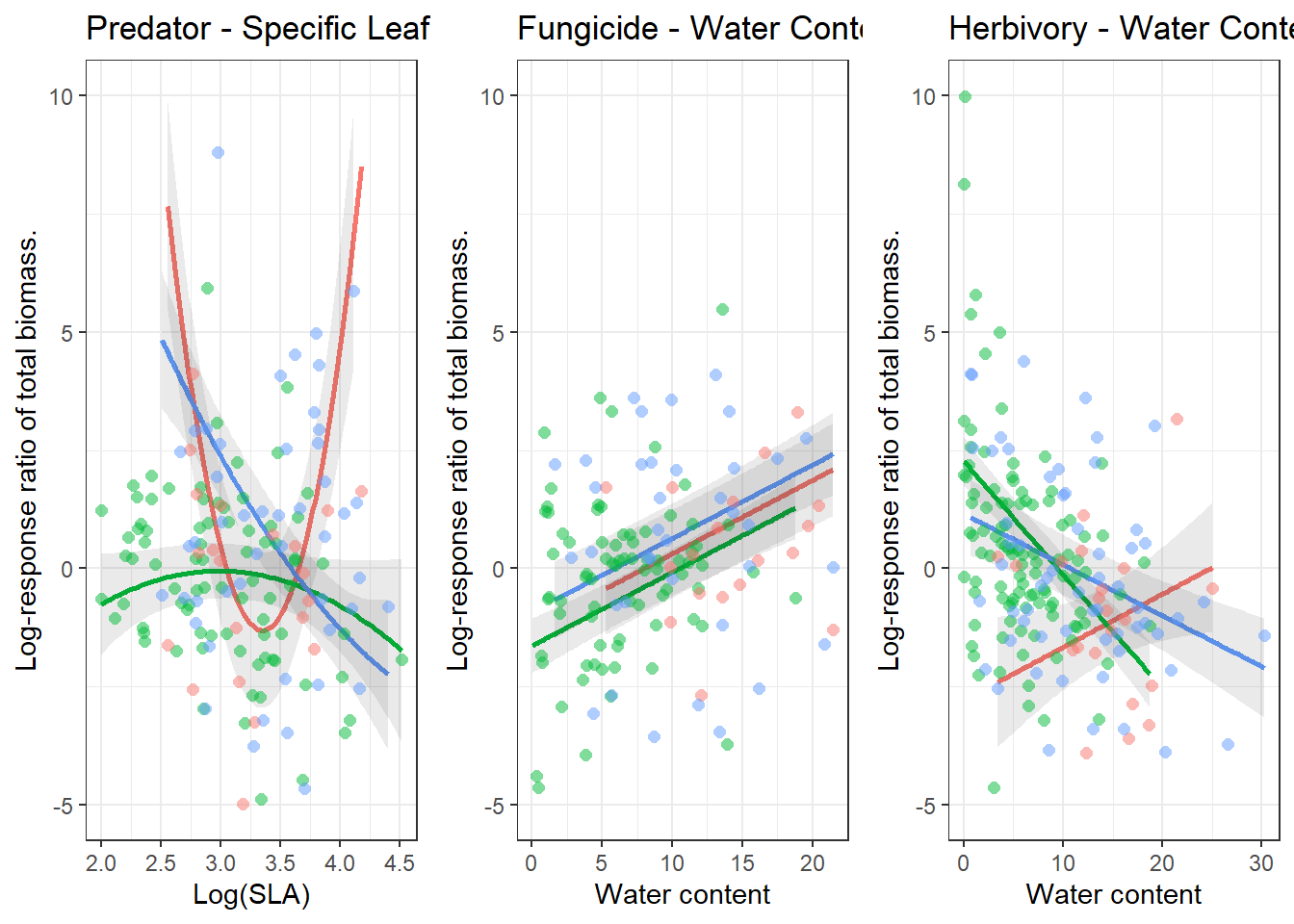
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Figure 1. Magnitude and direction of the effects of fungi, insects and their predators on biomass, richness, diversity and density of woody plants at low (Wanang, 200 m a.s.l.) mid (Numba, 750 m a.s.l.), and high (Yawan 1900 m a.s.l.) elevations. Log response ratio is a log-ratio of a given descriptor value from plots where a biotic factor was present to where it was absent. Grey points represent empirical values. Letters indicate statistically significant ($\\alpha$ = 0.05) differences for pairwise comparisons between elevations within a treatment-descriptor combination with Tukey correction for multiplicity. Letters are added only in cases where at least one mean-to-mean difference or mean-to-zero difference. Zero effect are indicated with a dashed line. Stars in the upper index indicate significant (ɑ = 0.05) differences of the effect mean from zero (significance of the effect at a given site)

Figure 2. Ordination diagrams

Figure 3. Raup-Crick

Figure 3: Traits