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Effects of warming winter embryo incubation temperatures on larval cisco (Coregonus artedi) survival, growth, and critical thermal maximum --Manuscript Draft--

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Abstract:	Freshwater whitefishes, Salmonidae Coregoninae, are cold stenothermic fishes of ecological and socio-economic importance in northern hemisphere lakes that are warming in response to climate change. To address the effect of warming waters on coregonine reproduction we experimentally evaluated different embryo incubation temperatures on post-hatching survival, growth, and critical thermal maximum of larval cisco (Coregonus artedi) sampled from lakes Superior and Ontario. Embryos were incubated at water temperatures of 2.0, 4.4, 6.9, and 8.9°C to simulate present and increased winter temperatures, and hatched larvae were reared in a common environment. For both populations, larval survival and critical thermal maximum were negatively related to incubation temperature, and larval growth was positively related to incubation temperature. The magnitude of change across incubation temperatures was greater in the population sampled from Lake Superior than Lake Ontario for all traits examined. The more rapid decrease in survival and critical thermal maximum across incubation temperatures for larval cisco in Lake Superior, compared to those from Lake Ontario, suggests that Lake Superior larvae may possess a more limited ability to acclimate to and cope with increasing winter water temperatures. However, the rapid increase in growth rates across incubation temperatures in Lake Superior larvae suggests they could recover better from hatching at a small length induced by warm winters, as compared to Lake Ontario larvae. Our results suggest propagation and restoration programs may want to consider integrating natural habitat preferences and maximizing phenotypic variability to ensure offspring are set up for success upon stocking.		
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Dear Editor:

Please consider our manuscript entitled "Effects of warming winter embryo incubation temperatures on larval cisco (*Coregonus artedi*) survival, growth, and critical thermal maximum" for review as a full-length article.

We experimentally tested the effects warming winter incubation temperatures have on the survival, growth, and critical thermal limit of cisco larvae from lakes Superior and Ontario. Our key findings were that 1) cisco larvae survival and growth were negatively impacted from warming incubation temperatures and 2) larvae from Lake Superior had a greater magnitude of change from the coldest to warmest incubation temperatures than Lake Ontario cisco in all traits examined. Our results suggest that larval cisco survival from both populations could be greatly reduced from warming winter incubation conditions, and that early-life stage cisco from Lake Superior may possess a narrower ability to acclimate to and cope with increasing winter water temperatures from climate change than cisco from Lake Ontario.

Re-establishing native species, including cisco, in the Great Lakes is an active area of research. Key uncertainties associated with cisco restoration include understanding the role the environment plays in the development of phenotypes and the extent of plasticity within populations. Our results highlight the potential effects of climatic warming on cisco populations and the importance of integrating natural habitat preferences into hatchery propagation programs to ensure offspring are set up for success upon reintroduction. Currently, many coregonine hatchery facilities around the Great Lakes do not incubate embryos under natural lake thermal conditions (i.e., cold water temperatures, < 4.5°C). We expect this paper to be well-received based on extensive coregonine restoration and conservation efforts underway throughout North America and Eurasia.

All authors contributed to the study conception and funding acquisition. Material preparation, data collection, and data analysis were performed by Taylor Stewart (TS). The first draft of the manuscript was written by TS and all authors commented on subsequent versions of the manuscript. All authors read and approved the final manuscript.

This research has not been previously published and is not under consideration at any other journal. Our submission has been posted on a preprint server (bioRxiv.org) and we agree to update the preprint link with a link to our JGLR submission if it is accepted. Thank you for your time and consideration. Please do not hesitate to contact me with any further questions. We look forward to your response.

Sincerely,

Taylor R. Stewart (on behalf of all co-authors)

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Effects of warming winter embryo incubation temperatures on larval cisco (Coregonus artedi) 1 2 survival, growth, and critical thermal maximum 3 Taylor R. Stewart^{a,b*}, Mark R. Vinson^c, and Jason D. Stockwell^b 4 5 6 ^aDepartment of Biology, University of Vermont, 109 Carrigan Drive, Burlington, VT 05405, USA 7 8 ^bRubenstein Ecosystem Science Laboratory, University of Vermont, 3 College Street, 9 Burlington, VT 05401, USA 10 ^cU.S. Geological Survey, Great Lakes Science Center, Lake Superior Biological Station, 2800 Lake Shore Drive East, Ashland, WI 54806, USA 11 12

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Freshwater whitefishes, Salmonidae Coregoninae, are cold stenothermic fishes of ecological and socio-economic importance in northern hemisphere lakes that are warming in response to climate change. To address the effect of warming waters on coregonine reproduction we experimentally evaluated different embryo incubation temperatures on post-hatching survival, growth, and critical thermal maximum of larval cisco (Coregonus artedi) sampled from lakes Superior and Ontario. Embryos were incubated at water temperatures of 2.0, 4.4, 6.9, and 8.9°C to simulate present and increased winter temperatures, and hatched larvae were reared in a common environment. For both populations, larval survival and critical thermal maximum were negatively related to incubation temperature, and larval growth was positively related to incubation temperature. The magnitude of change across incubation temperatures was greater in the population sampled from Lake Superior than Lake Ontario for all traits examined. The more rapid decrease in survival and critical thermal maximum across incubation temperatures for larval cisco in Lake Superior, compared to those from Lake Ontario, suggests that Lake Superior larvae may possess a more limited ability to acclimate to and cope with increasing winter water temperatures. However, the rapid increase in growth rates across incubation temperatures in Lake Superior larvae suggests they could recover better from hatching at a small length induced by warm winters, as compared to Lake Ontario larvae. Our results suggest propagation and restoration programs may want to consider integrating natural habitat preferences and maximizing phenotypic variability to ensure offspring are set up for success upon stocking.

Keywords: Coregonus; cisco; climate change; embryo incubation; larvae; water temperature

Introduction:

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- Water temperatures are rising around the globe (Austin and Colman, 2008; Maberly et al., 2020;
- 40 O'Reilly et al., 2015; Woolway et al., 2020) and poses a threat to ectotherms, such as fish, that
- 41 have limited thermal tolerance ranges (Comte and Olden, 2017; Dahlke et al., 2020; Little et al.,
- 42 2020). Thermal tolerances vary with ontogenetic development (Dahlke et al., 2020; Sunday,
- 43 2020) and affect reproduction, metabolic rates, growth, and overall survival (Brett, 1979; Brown
- 44 et al., 2004; Busch et al., 2012; Gillooly et al., 2002; Little et al., 2020; Ohlberger et al., 2007).
- Vulnerability of fishes to climatic warming is highest for cold stenothermic species that lack the
- 46 ability to migrate to suitable temperatures. Specific vulnerability of local populations will likely
- depend on future climate regime shifts and the temperature requirements of spawners and
- embryos (Dahlke et al., 2020; Sunday, 2020). In the short-term, lacustrine spawners may cope
- 49 with warming waters by shifting spawning timing or using deeper and colder spawning habitat,
- or potentially in the long-term through thermal adaptation across generations. Adaptation,
- 51 however, may be too slow to keep pace with changing thermal conditions (Bruge et al., 2016).
- 52 For autumn spawners, spawning later in the season after waters have cooled sufficiently may still
- 53 impact embryo development due to warmer winter temperatures and earlier spring warming.

55 Freshwater whitefishes, Salmonidae Coregoninae (hereafter coregonines), are cold, stenothermic

- 56 fishes of ecological and socio-economic importance throughout the northern hemisphere (Elliott
- and Bell, 2011; Isaak, 2014; Jeppesen et al., 2012; Jonsson and Jonsson, 2014; Karjalainen et al.,
- 58 2015; Stockwell et al., 2009). In the Laurentian Great Lakes, cisco (*Coregonus artedi*) was
- 59 historically the most abundant ciscoe (sensu Eshenroder et al., 2016) species, a primary prey fish
- of lake trout (*Salvelinus namaycush*), and a commercial fishing target since the early 1800s
- 61 (Bogue, 2001; Chiarappa, 2005). Most cisco spawning stocks collapsed by the mid-1900s
- 62 (Baldwin et al., 2009; Koelz, 1929). Lake Superior stocks partially recovered by the early-1990s
- 63 (Stockwell et al., 2009), but contemporary abundance is considered to be below historical levels
- 64 (Rook et al., 2021). Present Lake Superior cisco population abundance is hypothesized to be
- limited by reduced and inconsistent survival of fish to age-1 due to climatic warming over the
- past two decades (Van Cleave et al., 2014) and lower overall ecosystem productivity due to
- 67 reduced phosphorus inputs as compared to 1900-1970 (Rook et al., 2021). Variable and weak
- 68 year-class strength of coregonines has been observed worldwide over the past several decades

69 and has been associated with annual variations in lake ice formation and winter-spring thermal 70 conditions (Anneville et al., 2015; Karjalainen et al., 2015; Marjomäki et al., 2004; Nyberg et al., 71 2001). 72 73 Most coregonines spawn nearshore in late-autumn, embryos incubate under ice, and hatch in 74 spring near ice-out, when rising spring water temperatures trigger hatching (Karjalainen et al., 75 2021, 2019, 2015; Stockwell et al., 2009). Increases in air temperature and water temperatures of 76 seasonally ice-covered lakes are projected to be greatest during the winter and spring, 77 respectively, in response to climate change (Christensen et al., 2007; Ozersky et al., 2021; 78 Schindler et al., 1990; Winslow et al., 2017). 79 80 The larval period of fishes is critical for year-class success (Cushing, 1990; Hjort, 1914), but the 81 physiological effects of thermal stress from non-optimal embryo incubation temperatures on 82 post-hatching survival are unclear. Additional physiological pressures as a result of warming 83 winter temperatures could be detrimental. The match-mismatch hypothesis postulates that larval 84 survival is dependent on a temporal and spatial match between larval feeding capabilities, such 85 as swimming ability and prey acquisition, and prey availability (Cushing, 1990). Warmer 86 incubation temperatures lead to earlier hatch dates and altered morphological developments, such 87 as smaller lengths and larger yolk sacs, that reduce larval feeding efficiency (Darowski et al., 88 1988), compared to colder incubated embryos (Karjalainen et al., 2015; Stewart et al., 2021a). 89 The selective pressures from elevated temperatures on embryonic and larval coregonine 90 development and survival may lead to adaptation, but the thermal trigger for the response and the 91 mechanism of the response are unknown. Consequently, quantifying the potential response and 92 adaptive capacity of cisco to warming winter and spring water temperatures is needed. 93 94 We experimentally evaluated how cisco embryo incubation temperatures influenced the survival 95 and performance of hatching larvae within and between two Great Lakes cisco populations. We 96 hypothesized that warmer, sub-optimal cisco embryo incubation temperatures decrease larval 97 survival, growth, and critical thermal limits compared to embryo incubation temperatures that 98 mimic cold, pre-climate change conditions. If our hypothesis is supported, we would expect a

99 negative relationship between embryo incubation temperature and the larval traits examined for 100 wild cisco populations when reared artificially. 101 102 **Methods:** 103 Ethics 104 All work described here was approved for ethical animal care under University of Vermont's Institutional Animal Care and Use Committee (Protocol # PROTO202000021). 105 106 107 Crossing Design and Fertilization 108 Cisco were collected from the Apostle Islands, Lake Superior (46.85°, -90.55°) and Chaumont 109 Bay, Lake Ontario (44.05°, -76.20°) in December 2019. Eggs and milt were stripped from 12 110 females and 16 males from each population and artificially fertilized under a blocked, nested 111 full-sib, half-sib fertilization design to create a maximum of 48 families. A single fertilization 112 block consisted of four males each paired to three unrelated females, where all offspring of a 113 given female were full siblings (Stewart et al., 2021a). 114 115 For clarity, our operational use of a population represents a single species sampled from a single 116 location within a single lake (e.g., cisco from the Apostle Islands region in Lake Superior). Our 117 sampling efforts represent a single location within large lakes and does not likely capture the 118 possible genetic variation within a species or population. 119 120 Rearing Conditions 121 Full embryo incubation methods are described in Stewart et al. (2021). Embryos were incubated 122 in 24-well cell culture microplates placed in climate-controlled chambers (Memmert[®] 123 IPP260Plus) at mean (SD) constant temperatures of 2.0 (0.5), 4.4 (0.2), 6.9 (0.2), and 8.9 124 (0.3)°C. These temperatures were selected to mimic present and potentially warmer winter temperatures (Titze and Austin, 2014) at typical cisco spawning depths (<100 m, Goodyear, 125 126 1982). Reconstituted freshwater medium was used during fertilizations and incubations 127 (International Organization For Standardization 6341, 2012) to standardize the chemical 128 properties of the water among all treatments and between populations. After hatching, larvae

were photographed alive and ventrally (Nikon® D5600 and Nikon® AF-S DX 18-55mm lens). 130 Total length was measured from images using Olympus[®] LCmicro. 131 132 Newly-hatched larvae were transferred to rearing tanks segregated by population and incubation 133 temperature. Larvae from Lake Superior were reared in four (4 incubation treatments x 1 replicate) 150-liter oval tanks. Larvae from Lake Ontario were reared in eight (4 incubation 134 135 treatments x 2 replicates) 150-liter oval tanks. Lake Ontario larvae were divided equally by 136 families (i.e., up to 24 of 48 total larvae per family per replicate tank) into replicate tanks per 137 incubation temperature treatment. Lake Superior larvae were unreplicated as a result of low 138 fertilization success and embryo survival - insufficient numbers were available for multiple 139 rearing tanks. All rearing tanks were supplied with chilled, recirculating water maintained at 140 6.5° C (mean (SD) = 6.36 (1.17)). Water temperatures ($\pm 0.2^{\circ}$ C) were recorded hourly. Larvae in 141 all rearing tanks were exposed to the same photoperiod cycle (i.e., 12-hr light, 12-hr dark) with 142 gradual sunrise and sunset transitions. Full spectrum (i.e., 380-780 nm), white LED lights 143 (AquaShift® MLA-WH) were used to simulate daylight. Dead larvae were removed and counted 144 each day. Larvae were fed Artemia and transitioned to Otohime A dry feed one-week post-hatch. 145 Food was provided ad libitum. 146 147 Thermal Challenge 148 After 60 days, larvae from each population, incubation treatment, and replicate rearing tank were 149 thermally challenged. Because larvae within and among rearing tanks did not hatch on the same 150 day, 60-days post-hatch was calculated from the date of 50% hatching for each rearing tank. 151 Larvae were transferred to 5.4-liter clear, rectangular tanks, with two replicate tanks per rearing 152 tank and approximately 50 larvae, or as many available, were used in each replicate tank. Water 153 temperature was 10°C and larvae were allowed to acclimate to this temperature for 12 hours 154 prior to the thermal challenge. The water in the thermal challenge system was recirculated among all replicate tanks and aerated. During the thermal challenge, water temperatures were 155 156 raised from 10.0°C at a constant rate of 0.5°C per 30 minutes until all larvae were deceased. 157 Larvae were considered terminated when loss of equilibrium was achieved and were motionless 158 for at least 5 seconds. Once endpoint criteria were met, larvae were euthanized, photographed,

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159 and preserved in 95% ethanol. The elapsed time and temperature at termination of each 160 individual larvae was recorded and total length was measured from the images. 161 162 All larvae from the 8.9°C treatment died during the acclimation period from an unknown cause, 163 thus, only thermal challenge data from 2.0, 4.4, and 6.9°C are presented. 164 165 Statistical Analyses 166 All statistical analyses were performed in R version 4.0.5 (R Core Team, 2021). 167 168 Larval survival was estimated for each rearing tank, across all families, as the percent of larvae 169 surviving between hatching and 60 days after the date of 50% hatching. Our estimates of larval 170 survival from Lake Superior are unreplicated. However, useful information can still be gleaned 171 without strict statistical testing (e.g., Davies and Gray, 2015). Observations of single estimates of 172 larval survival across incubation temperatures could foster further hypotheses and lead to more 173 focused studies. 174 175 Similar to larval survival estimates, larval growth rate estimates for Lake Superior were 176 unreplicated. To this end, we qualitatively compared absolute growth rates between populations 177 and across incubation temperatures by generating bootstrapped confidence intervals for the 178 observed absolute growth rate estimates. For each population, incubation temperature treatment, 179 and replicate rearing tank, a bootstrapped mean length-at-hatch was calculated from random 180 sampling with replacement from the observed lengths-at-hatch, and a bootstrapped mean final 181 length was calculated from random sampling with replacement from the observed final lengths. 182 The difference between the bootstrapped mean final length and bootstrapped mean length-at-183 hatch was calculated and divided by the duration of the larval experiment (i.e., absolute growth 184 rate). The bootstrap procedure was repeated 10,000 times. The bootstrapped absolute growth rate 185 distributions were used to calculate the 2.5 and 97.5 percentile values (i.e., 95% confidence 186 interval) as a measure of variation around the observed absolute growth rate, and to qualitatively 187 assess the likelihood of differences in growth among populations and incubation temperature 188 treatments, in absence of replication. For Lake Ontario, the 95% confidence intervals were 189 calculated as the mean 2.5 and 97.5% percentiles across replicate tanks. Comparisons were made

by examining the overlap of the observed mean absolute growth rate to the bootstrapped 95% confidence intervals of all pairwise comparisons.

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The critical thermal maxima (CTMax) of larval cisco from each population and incubation temperature treatment was expressed as the arithmetic mean of the temperature at which endpoint criteria were reached (Mora and Ospina, 2001). Although we have estimates for each individual larva within replicate thermal challenge tanks, the larvae from Lake Superior were reared in a single rearing tank and thus the estimates are not independent and cannot be treated as true replicates. Therefore, a similar bootstrap approach as described for larval growth was used to qualitatively compare CTMax among populations and incubation temperature treatments. For each population, incubation temperature treatment, and replicate rearing tank, we generated a bootstrap sample by randomly selecting, with replacement, a termination temperature n times, where n equals the number of observations in the experiment. The CTMax was calculated for each bootstrapped sample and the distribution of bootstrapped CTMax was used to calculate the 95% confidence interval as a measure of variation around the observed CTMax. The bootstrap procedure was repeated 10,000 times. For Lake Ontario, the 95% confidence intervals were calculated as the mean 2.5 and 97.5% percentiles across replicate tanks. Comparisons were made by examining the overlap of the observed mean CTMax to the bootstrapped 95% confidence intervals of all pairwise comparisons.

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Results:

- 211 Larval Survival
- A total of 9,605 larvae hatched and were reared from lakes Superior (2,332 larvae) and Ontario
- 213 (7,273 larvae) across all incubation temperatures. Larval survival was highest at the 2.0°C
- incubation temperature and decreased with warming incubation temperatures for both
- populations (Figure 1). Survival rates were 38.7% at 2.0°C, 17.7% at 4.4°C, 1.1% at 6.9°C, and
- 216 5.4% at 8.9°C for Lake Superior larvae and 43.3% at 2.0°C, 35.3% at 4.4°C, 12.4% at 6.9°C, and
- 2.6% at 8.9°C for Lake Ontario larvae. Larval survival was higher for Lake Ontario larvae than
- Lake Superior larvae across all incubation temperature treatments, except 8.9°C. Lake Ontario
- 219 larvae had similar survival rates (< 9% difference) at the 2.0 and 4.4°C incubation temperatures,

220 whereas Lake Superior larval survival decreased 21% from the 2.0° to 4.4°C incubation 221 temperatures (Figure 1). 222 223 Larval Growth 224 Larval cisco absolute growth rates increased with warming incubation temperatures in both 225 populations (Figure 2). Larvae from Lake Superior had lower absolute growth rates at 2.0 and 226 4.4°C (0.049 and 0.044 mm day⁻¹, respectively) compared to Lake Ontario (0.056 and 0.061 mm 227 day⁻¹, respectively). Absolute growth rates increased at 6.9°C for Lake Superior (0.057 mm day⁻¹ 228 1) and 8.9°C for Lake Ontario (0.078 mm day-1), and both populations had similar absolute growth rates at 6.9 and 8.9°C (mean difference <0.001 and 0.012 mm day⁻¹, respectively; Figure 229 230 2). 231 232 Thermal Challenge 233 Critical thermal limit in larval cisco decreased with warming incubation temperatures in Lake 234 Superior and Lake Ontario (Figure 3). Larvae from Lake Superior incubated at 2.0°C had the highest CTMax (25.81°C). However, CTMax in Lake Superior decreased by 0.83 and 0.77°C 235 236 between the 2.0 to 4.4°C and the 4.4 and 6.9°C incubation temperature treatments, respectively. 237 CTMax was similar for Lake Ontario larvae incubated at 2.0 and 4.4°C (24.99 and 24.96°C, 238 respectively) and decreased at 6.9°C (24.67°C). 239 240 **Discussion:** 241 Survival, growth rates, and critical thermal limits of cisco larvae from lakes Superior and Ontario 242 were influenced by embryo incubation temperatures that were warmer than current natural winter 243 water temperatures in these lakes. First, we found a negative relationship between larval survival 244 and embryo incubation temperature. Second, warmer embryo incubation temperatures increased 245 larval growth rates. Third, critical thermal limits decreased with warming incubation temperatures. Lastly, the magnitude of change across incubation temperature treatments was 246 247 greater in cisco from the historically colder Lake Superior than Lake Ontario for all traits 248 examined. These results suggest increased risk to Laurentian Great Lakes cisco populations in 249 response to projected climatic warming.

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251 Our hypothesis that larval survival is highest at the coldest incubation temperature, which 252 mimicked the natural lake environment, was supported. Between the two lakes, Lake Superior 253 cisco had a sharper decline in larval survival from 2.0 to 4.4°C compared to Lake Ontario cisco. 254 Though both populations are cold adapted, the result suggests Lake Superior cisco were more 255 cold-water adapted than those from Lake Ontario. Lake Superior is colder than Lake Ontario 256 (Mason et al., 2016; Millar, 1952) and spawning cisco from Lake Superior were collected at an 257 open lake location, whereas spawning cisco from Lake Ontario cisco were collected in a shallow, 258 protected bay. Water temperatures in shallower protected habitats increase more rapidly after 259 ice-out and have higher maximum spring and summer temperatures (i.e., Lake Ontario sampling 260 location; Minns et al., 2011) compared to deeper, open-water locations (i.e., Lake Superior 261 sampling location; Titze and Austin, 2014) because the heat capacity of water is positively 262 related to depth and water is mixed less in protected bays (Assel et al., 2003; Gan and Liu, 2020; 263 Verburg and Antenucci, 2010). Interactions among winter and spring temperatures, hatching 264 dates, zooplankton availability and larval size-dependent predation mortality influence year-class 265 strength of vendace (C. albula) and European whitefish (C. lavaretus) in Europe (Anneville et 266 al., 2009; Marjomäki et al., 2004; Mehner et al., 2011; Miller et al., 1988). Spring warming rates 267 in particular appear to play a critical role in prey availability and larval growth and survival of 268 autumn-spawning coregonines (Karjalainen et al., 2015; Myers et al., 2014). 269 270 The transition from endogenous to exogenous feeding is critical to larval fish survival (Cushing, 271 1990; Hjort, 1914). Higher winter temperatures induce earlier coregonine embryo hatching and 272 cause larvae to have smaller lengths-at-hatch and larger yolk-sac volumes (Karjalainen et al., 273 2015; Stewart et al., 2021a). Larvae hatching with larger yolk sacs may have more time to 274 transition to exogenous feeding (Hjort, 1914; Lucke et al., 2020; Miller et al., 1988), but at a cost 275 to swimming efficiency and predator avoidance (Darowski et al., 1988; Myers et al., 2014). In 276 wild populations, earlier hatching may also increase the mismatch between the onset of spring 277 plankton blooms and larval prey, increasing the risk for starvation and higher larval mortality 278 (Cushing, 1990; Myers et al., 2014). Interactions among winter and spring temperatures, 279 hatching dates, zooplankton availability and larval size-dependent predation mortality influence 280 year-class strength of vendace (C. albula) and European whitefish (C. lavaretus) in Europe 281 (Anneville et al., 2009; Marjomäki et al., 2004; Mehner et al., 2011; Miller et al., 1988). Spring

warming rates in particular appear to play a critical role in prey availability and larval growth and survival of autumn-spawning coregonines (Karjalainen et al., 2015; Myers et al., 2014). Our experiment provided cisco larvae a predator-free environment with *ad libitum* food immediately after hatching, yet we still observed sharp declines in larval survival for those incubated at increased temperatures. We suggest an additional or alternative hypothesis for a survival bottleneck under climate change scenarios is that larval cisco survival may not be as limited by prey availability but instead by reduced physiological condition caused by warmer embryo incubations.

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Rapid larval growth is associated with high survival (Blaxter, 1986; Houde, 1989; Miller et al., 1988; Myers et al., 2014; Ware, 1975). In our experiment, larval cisco exhibited low survival despite higher absolute growth rates when incubated at warmer temperatures. These results did not support our hypothesis that warmer, sub-optimal cisco incubation temperatures decrease larval growth rates. Coregonine embryos incubated at high temperatures (i.e., > 6°C) often hatch prematurely and are underdeveloped (Colby and Brooke, 1970; Price, 1940). Warmer incubations may require free-floating embryos to rapidly convert yolk for development. In this sense, higher absolute growth rates gained from warming incubation temperatures may not be optimal for survival. Larval cisco from Lake Superior use a mixed-feeding strategy with endogenous energy reserves (i.e., yolk) and exogenous feeding overlapping at lengths between 10.0-12.0 mm (Lucke et al., 2020). Embryos incubated at colder temperatures (i.e., 2.0 and 4.4°C) from the sampled populations of lakes Superior and Ontario cisco had mean lengths-athatch from 9.9-11.3 mm, whereas mean lengths-at-hatch ranged from 8.7-9.7 mm at the warmest incubation temperature (8.9°C; Stewart et al., 2021). A combination of field and experimental data suggests that cold, long incubations with prolonged development results in larger length-athatch with less endogenous energy reserves, the ability to immediately use a mixed endogenous and exogenous feeding strategy, and lower growth rates could be the best scenario to maximize larval survival (Lucke et al., 2020; Stewart et al., 2021a). However, this 'goldilocks scenario' may only work if all biotic and abiotic conditions (e.g., water temperature, appropriately sized prey, etc.) match cisco phenotypes. This hypothesis remains to be tested, as our experiments provided a stable and optimal temperature and feeding environment, conditions that cannot be assumed to occur in the wild.

The ability of larval cisco to use favorable nursery habitat near the lake surface is directly related to their ability to tolerate spring-summer surface water temperatures. The increase in CTMax with decreased incubation temperature supported our hypothesis that cold, pre-climate change conditions would maximize thermal performance. The different magnitudes of change between cisco from lakes Superior and Ontario could be explained from evolutionary adaptations to local conditions. Fish populations from high-latitude, low-temperature locales often compensate for slower metabolism and lower growth rates by having more efficient physiological performance than low-latitude populations (i.e., countergradient variation; (Conover and Present, 1990; Jonassen, 2000; Reist et al., 2006). Lake Superior experiences colder and less seasonal variation in water temperature than Lake Ontario (Calamita et al., 2021; Zhang et al., 2018), and larval cisco from Lake Superior may have more efficient physiological adaptations (e.g., cardiac and respiratory performance) which could explain the high thermal tolerance at cold incubation temperatures and sensitivity to increased temperatures. Our results suggest research on mechanisms driving the observed differences in CTMax between populations (e.g., cardiac failure, oxidative damage to tissue, body mass, stress biomarkers, protein denaturation, etc.) may prove insightful.

Our results have implications for current and proposed hatchery-based restoration efforts of coregonines in the Laurentian Great Lakes (Bronte et al., 2017). We found that cisco offspring from two of the Great Lakes raised at warm incubation temperatures (*i.e.*, > 4.5°C) had lower overall performance than individuals incubated at cold temperatures (*i.e.*, < 4.5°C). Many coregonine hatchery facilities around the Great Lakes do not or cannot incubate embryos under natural lake thermal conditions (*i.e.*, cold water temperatures, < 4.5°C; Bronte et al., 2017). Hatchery-produced fish can have lower fitness in natural environments than wild fish (Araki et al., 2008; Bailey et al., 2010; Christie et al., 2014). Offspring from parents haphazardly selected for artificial breeding and reared in captivity before release have the potential to induce strong directional selection and harm naturally recruiting populations (Araki and Schmid, 2010; Tingley III et al., 2019). Transgenerational effect of lower larval performance and its potential effect on the response to selection are unknown but warrants investigation (Araki et al., 2008; Araki and Schmid, 2010; Christie et al., 2014; Ford, 2002). The consequences an artificial

344 environment may have on the genetic diversity within a population and fitness of post-stocking 345 individuals needs to be considered in ongoing restoration and conservation efforts (Tingley III et 346 al., 2019). 347 348 Identifying the genetic mechanisms (i.e., SNPs and gene expression) involved in the thermal 349 adaptation and acclimation of coregonine populations is an important next step. Variation in 350 certain genetic markers and survival under thermal stress may allow managers to determine the 351 genotypes associated with increased survival at variable or increasing temperatures (Narum et 352 al., 2013). Examining gene expression across populations and temperature treatments will help 353 identify and evaluate the function of differentially expressed genes and potential physiological 354 pathways that may be disproportionately under- or over-represented with thermal stress 355 (Rougeux et al., 2018). Furthermore, the combination of genomic tools (e.g., genome-wide 356 association study and RNA-seq) in thermal ecology experiments can provide valuable insight 357 into the functional significance of markers associated with thermal tolerance (Rougeux et al., 358 2018). Considerable progress has been recently made in advancing our genomic knowledge of 359 Laurentian Great Lakes coregonines that will provide a foundation for this work (Ackiss et al., 360 2020; Blumstein et al., 2020; Eaton et al., 2021; Lachance et al., 2021). 361 362 **Conclusion:** 363 The rapidity at which winter environments are changing has revealed our 'blind spot' for winter 364 biology (Ozersky et al., 2021). The results presented here and elsewhere (Karjalainen et al., 365 2016, 2015; Stewart et al., 2021a, 2021b) focus on how coregonine reproduction may be 366 impacted by a warming climate and suggest that while we have much to learn, the effects of 367 warming winters will vary among populations and with the magnitude of warming. These results 368 highlight the importance of integrating natural habitat preferences into stock propagation 369 programs to ensure offspring are set up for success upon reintroduction. A challenge for 370 managers and propagation facilities is to consider the impact embryo incubation conditions may 371 have on larval survival and performance in relation to production targets. Additionally, 372 propagation and stocking may accomplish the short-term restoration objective of supplementing

wild populations, but other limiting factors (e.g., habitat loss, anthropogenic disturbances, water

quality, invasive species) also need to be addressed to achieve long-term population conservation

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375 and viability (Tingley III et al., 2019). Maximizing phenotypic variation and adaptability to 376 changing conditions (i.e., portfolio effect; Schindler et al., 2015, 2010) is a strong consideration 377 in restoration and conservation efforts. Embracing management strategies that foster increased 378 early-life stage fitness could improve the ability of coregonines to cope with environmental 379 change in the wild and aid in addressing recruitment bottlenecks. 380 381 **Acknowledgments:** 382 We thank staff at the Wisconsin Department of Natural Resources Bayfield Fisheries Field 383 Station, U. S. Geological Survey (USGS) Tunison Laboratory of Aquatic Science, and New 384 York State Department of Environmental Conservation Cape Vincent Fisheries Station for field 385 collections of spawning adults. Rachel Taylor, Dan Yule, and Caroline Rosinski helped with 386 fertilizations and incubation experiment maintenance. [insert name] provided the USGS solicited 387 review that strengthened the manuscript, as did anonymous peer reviewers and Stockwell and 388 Dr. Ellen Marsden laboratory members. This work was funded by the USGS [grant/cooperative 389 agreement number G16AP00087 and G17AC00042] to the Vermont Water Resources and Lakes 390 Studies Center and the University of Vermont. Additionally, this work was made possible with 391 funds made available to Lake Champlain by Senator Patrick Leahy through the Great Lakes 392 Fishery Commission. Any use of trade, product, or firm names is for descriptive purposes only 393 and does not imply endorsement by the U.S. Government. 394 395 **References:** 396 Ackiss, A.S., Larson, W.A., Stott, W., 2020. Genotyping- by- sequencing illuminates high 397 levels of divergence among sympatric forms of coregonines in the Laurentian Great Lakes. 398 Evol. Appl. 13, 1037–1054. https://doi.org/10.1111/eva.12919 399 Anneville, O., Lasne, E., Guillard, J., Eckmann, R., Stockwell, J.D., Gillet, C., Yule, D.L., 2015. 400 Impact of Fishing and Stocking Practices on Coregonid Diversity. Food Nutr. Sci. 06, 401 1045–1055. https://doi.org/10.4236/fns.2015.611108 402 Anneville, O., Souissi, S., Molinero, J.C., Gerdeaux, D., 2009. Influences of human activity and 403 climate on the stock-recruitment dynamics of whitefish, Coregonus lavaretus, in Lake 404 Geneva. Fish. Manag. Ecol. 16, 492–500. https://doi.org/10.1111/j.1365-2400.2009.00703.x 405 Araki, H., Berejikian, B.A., Ford, M.J., Blouin, M.S., 2008. Fitness of hatchery-reared

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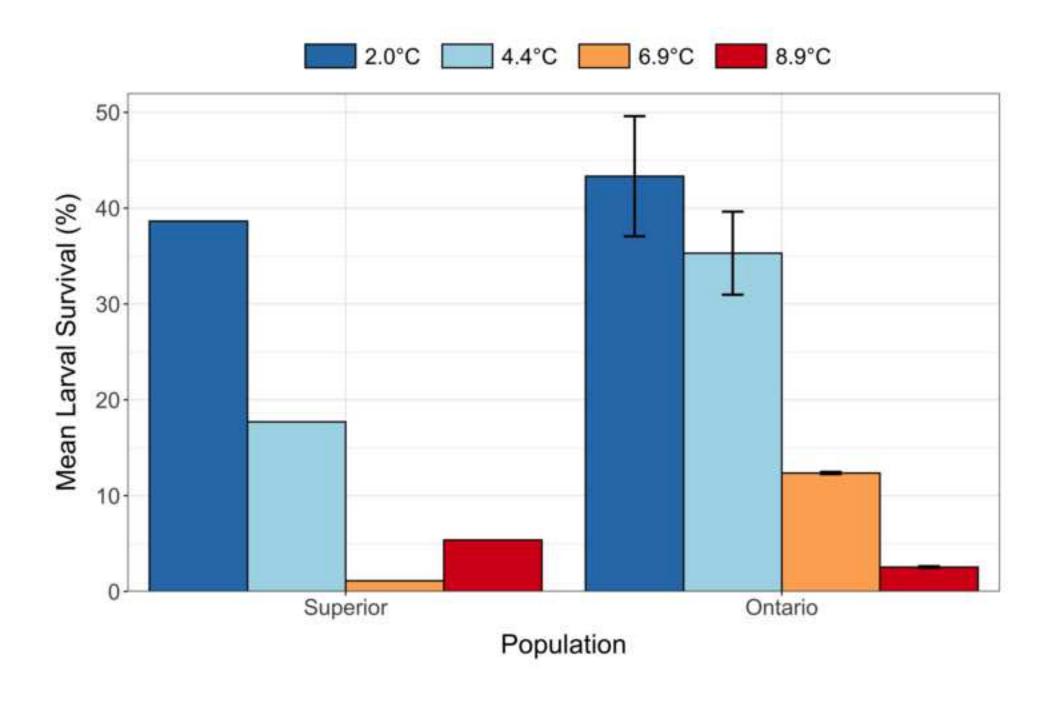
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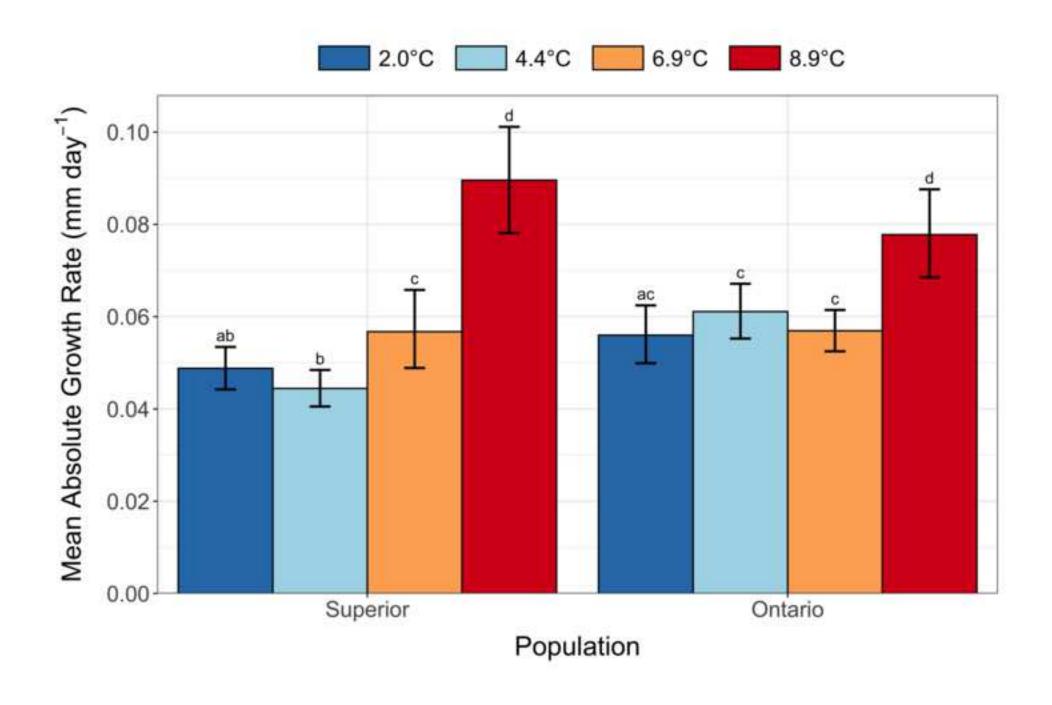
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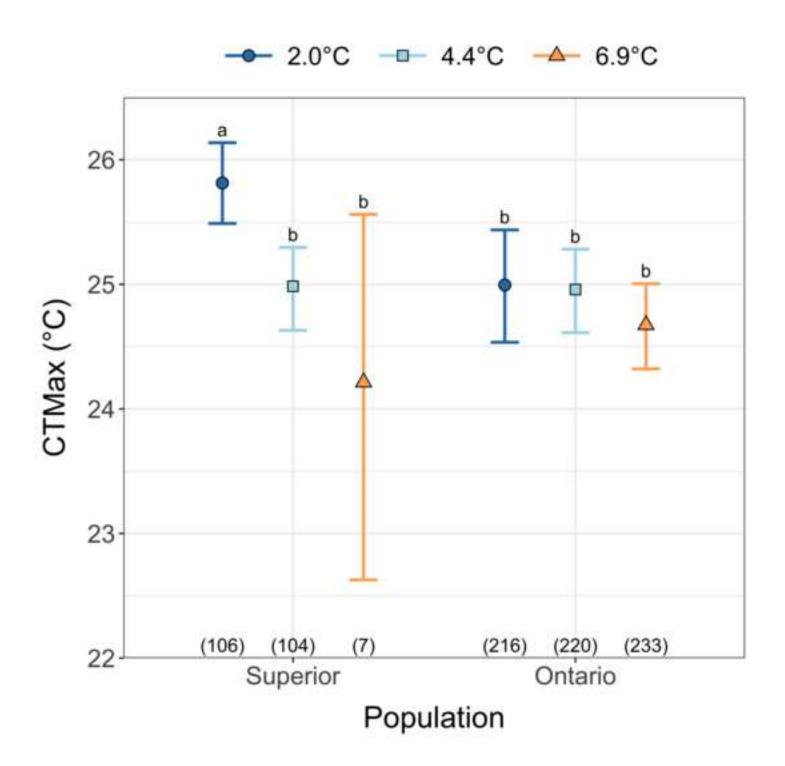
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646 Figure captions: 647 648 Figure 1. Mean larval survival (%) for larval cisco (Coregonus artedi) from Lakes Superior and 649 Ontario incubated at 2.0, 4.4, 6.9, and 8.9°C across replicate rearing tanks. Error bars indicate 650 standard error. Lake Superior mean survival estimates are unreplicated and thus do not have 651 error estimates. 652 653 Figure 2. Mean absolute growth rates (mm day⁻¹) for larval cisco (*Coregonus artedi*) from lakes 654 Superior and Ontario incubated at 2.0, 4.4, 6.9, and 8.9°C. Error bars indicate 95% bootstrapped 655 confidence intervals. Letters indicate overlap of the observed mean absolute growth rate to the 656 bootstrapped 95% confidence intervals of all pairwise comparisons. 657 Figure 3. Critical thermal maxima (CTMax; °C) for larval cisco (Coregonus artedi) from lakes 658 659 Superior and Ontario incubated at 2.0, 4.4, and 6.9°C. Error bars indicate 95% bootstrapped 660 confidence intervals. Letters indicate overlap of the observed CTMax to the bootstrapped 95% 661 confidence intervals of all pairwise comparisons. Sample sizes are indicated in parentheses.







Declaration of Interest Statement

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☑ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.	
□The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:	
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Author Credit Statement

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Taylor R. Stewart: Writing- Original draft preparation, Conceptualization, Methodology, Investigation, Formal analysis, Visualization, Data curation, Funding acquisition. **Mark R. Vinson**: Writing- Reviewing and Editing, Conceptualization, Funding acquisition. **Jason D. Stockwell**: Writing- Reviewing and Editing, Conceptualization, Funding acquisition, Supervision