Species-specific effects of the Urban Heat Island on tree growth across Berlin

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Abstract

This document outlines the rationale for an analysis of tree growth (potential) and its relationship with the Urban Heat Island (UHI) effect in Berlin using an extensive, publicly available data set. It introduces preliminary results and provides an outlook for up-coming and potential work.

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1 Introduction

Berlin features the most intense Urban Heat Island (UHI) in Germany due to its large extent and development intensity (Kuttler et al., 2015), with temperature increases of up to 12~K during day-time and 6~K on average for night-times (2001-2010, Fenner et al., 2014) in urban vs. rural areas. Consequently, urban green (infrastructure) systems are subjected to increased heat more frequently, potentially affecting their process dynamics - either positively or adversely. Their performance and health, however, is closely tied to local energy budgets (Grimmond et al., 1996; Hertel and Schlink, 2019), which in turn are decisive for controlling human wellbeing (e.g. Maras et al., 2016), amongst other factors. Assessing the effect of increased temperatures on green infrastructure, as part of the urban landscape, is therefore instrumental for understanding, and ultimately mitigating, the potential impact of future warming on increasingly urban societies (Norton et al., 2015).

Trees, in particular, provide shading as well as transpirative cooling in their vicinity (Endlicher et al., 2016; Gillner et al., 2015; Oke, 1982), and therefore can reduce ambient temperatures, infrastructure power-consumption and (human) thermal discomfort (Akbari et al., 2001; e.g. Gulyás et al., 2006; Hoyano, 1988; Mayer and Höppe, 1987); simultaneously, they provide numerous other environmental, cultural and

psychological services and/or benefits (see Tzoulas et al., 2007 for review). Further, recent tree growth dynamics as a proxy for on-going and future warming may provide an additional line of evidence to support the growing knowledge base on future climate-vegetation dynamics (Zhao et al., 2016) and may aid in mitigation and adaptation efforts (Brune, 2016; Pretzsch et al., 2017).

Trees and green infrastructure in urban areas show a tendency for enhanced growth rates and/or productivity compared to rural counterparts (Jia et al., 2018; Pretzsch et al., 2017), yet feature a broad range of effect size ranges and, in some cases, signs specific to species and locality. Zhao et al. (2016) showed that growth rates increased within urban clusters as urbanization intensifies using remotely sensed vegetation indices. Similarly, for Berlin, Dahlhausen et al. (2018), identified positive growth modulation in highly urbanized environments (using growth increments) for Tilia cordata Mill, the most abundant tree of the city, which they attributed to the UHI effect, while intermediate development intensity showed indications of being least favorable for tree growth. Further, Moser-Reischl et al. (2019) identified positive associations between air temperature and radial growth for two species commonly selected by urban planners (T. cordata, Rubinia pseudoacacia) in Munich. By contrast, Gillner et al. (2014) highlight decreased growth for Acer species (A. platanoides and pseudoplatanus), Platanus x hispanica and Quercus rubra with higher summer temperatures of the preceding year, especially when compounded with drought, in another German metropolis (Dresden). Differences in growth trends may result from contrasting species-specific characteristics, but are indeed affected by other processes and factors, such as water availability, pollution and road-salt loading, structural impedance through infrastructure or management, etc. (Pauleit et al., 2002; Quigley, 2004; Randrup et al., 2001; Rhoades and Stipes, 1999). Under climate change, atmospheric drought will likely be compounded with high temperatures - and intensified UHIs - more frequently, adding further stress to current urban disturbance regimes (Roloff et al., 2009).

Conditions affecting tree growth can vary greatly within urban areas or regions, and need to be accounted for when establishing relationships with pertinent drivers, such as the UHI effect. This typically complicates the extrapolation from individual sampling sites toward predicting effect sizes across entire urban areas and tree stocks. This is especially the case for studies reliant on labour-intensive methods which are limited logistically by sampling effort, reducing sample sizes, as well as species and spatial coverage.

To complement (existing) detailed dendroecological analyses of climate-growth relationships in Berlin for key species, we propose inferring growth modulation from a large data set in excess of 650000 individuals covering 94 genera and at least 600 species and/or cultivars provided by the Berlin Senate Administration (Senatsverwaltung). This data set contains information on location, species, trunk diameter (at breast height; DBH; see Tab. 1), and height, amongst other variables for the majority of street and park trees.

Table 1: Available records by category in entire data set (n), and those with age and DBH entries (n_{full}).

Category	n	\mathbf{n}_{full}
Park	257985	151527
Street	363905	361381
Riparian	46364	0

In a space-for-time substitution, growth of individual species can be assessed across the entire city of Berlin, and related to effects of the UHI, while accounting for other location-specific factors, such as street characteristics, development intensity, available soil volume, etc. Comparable applications are found, for example, in Quigley (2004) and Pretzsch et al. (2017). The former inferred absolute growth potential for species across successional groups (early, mid, late stage), and between rural and urban conspecifics, yet lacked spatially-explicit effect size estimates across the urban-rural space and was limited to comparatively small sample sizes per group ($n_{total} = 230$ divided in 15 species, 3 groups and 2 locations). Pretzsch et al. (2017) applied linear hierarchical models to infer growth modulation on annual basis for different time periods and urban vs. rural locations while accounting for stand-level variability; however, for Berlin only 145 individuals of one species (T. cordata) were assessed. As mentioned previously, climate-growth relationships can vary substantially between species, and in fact, Quigley (2004) and Pretzsch et al. (2017) report contrasting results regarding average tree diameter, i.e. smaller or larger for urban vs. rural trees of same age. Consequently,

this variability of effect sizes and directions calls for a more comprehensive assessment across species and with greater spatial coverage throughout Berlin.

We therefore propose applying a statistical model that is fully spatially explicit, while also allowing to account for the nested nature of the data set (e.g. streets and districts) as well as other pertinent factors using hierarchical, generalized additive models (see Section 2). As a result, the absolute growth potential of a species can be inferred given, for example, a specific location, age or UHI magnitude. Further, the impact of UHI loading can be predicted for a single species across all of Berlin as a continuous surface. The inclusion of independent, tree-level growth data, however, is paramount as it allows:

- 1) validating the publicly-available Senate dataset (DBH) while providing a more reliable estimate of tree age,
- 2) applying the model with annual basal area as a response (enabling incorporating effects of varying climate/UHI intensities over time), and
- 3) comparing DBH (Senate data) and basal area-derived models for a subset of locations and species, increasing confidence in Berlin-wide predictions.

2 Methods

2.1 Study area

Berlin is one of the largest metropolitan areas in Central Europe (892 km^2) with a population of approximately 3.6 million, and a maximum extent of 38 km in North-South and 45 km in East-West directions. It is located in North-Eastern Germany, and lies in the temperate zone with warm-humid climate (Dfb) according to the updated Köppen-Geiger classification (Beck et al., 2018), with mean annual temperature of approximately $10^{\circ}C$ and precipitation of 575 mm (Tempelhof weather station, DWD). Berlin features low relief (approximately 30 m to 60 m with 120 m at solitary peaks), and is centered around a glacial outwash valley (sands, gravel), bordered by two plateaus consisting of glacial till and clay in the North-East and South, as well as sands in the South-West. The city provides extensive public green space covering around 30 % of its area (suvkberlin2019?), with an extensive urban forest of nearly 700000 publically-managed trees along streets, in parks and in riparian areas. It has a pronounced urban heat island, owing to its high degree of built-up and sealed surfaces, with air temperatures up to $12^{\circ}C$ higher in urban compared to adjacent rural areas (Fenner et al., 2014).

2.2 Data sources

An overview of all data used in this study, including sources, types, and application, is provided in Table X. Most notably, characteristics of street trees and their respective environments were collated from the city's tree inventory, as well as various elements of the Berlin Environment Atlas, which are openly available via daten.berlin.de and curated by the city's government. The data used from the inventory includes species, location, age, and diameter at breast height (dbh). The environment atlas holds information on several variables relevant to tree growth, such as soil type and nutrient exchange capacity, distance to groundwater, surface area of tree planting beds ("Baumscheibe"), as well as general characteristics of the urban environment like building height. It also includes spatially-explicit model outputs for local climate simulations representative of typical conditions, such as air temperature (2 m) on a 2015 summer's day at different hours (0400, 1400, 2200). These outputs are available at city block or planning unit level, and were used as a proxy for the general temperature distribution influenced by urbanization across the investigated growth period across the previous 80 years (see section X for details). This was deemed reasonable as Berlin's built-up area has not changed markedly over the past 50 years (i.e., from about 52 to 61 %) (Mohamed, 2017). The utility of using this high-resolution data set of air temperatures over a derived urban heat island product was assessed by... Further insight into the role of the urban environment was assessed by including local climate zone classes derived by WUDAPT. These describe the proportion of urban area with a specific cover, such as dense high rises, sealed surfaces, etc. For this study, the LCZ 2, representing dense mid-rise built up cover was used as a proxy for the degree of urbanization, and is the most frequent LCZ for Berlin (Fenner et al., 2017) (found ref in . Use percent sealed surface instead?

Data description used for generating maps and route optimization. Resolution and radius are provided in m, the latter referring to the buffer size across which data was averaged to estimate trees' environmental conditions. A zero-radius refers to a point extraction from categorical and location specific data. Planting Bed Area, and Soil type are not currently used.

Name Accessed

```
100
150
https://www.wudapt.org/continental-lcz-maps/
Demuzere et al. (2019)
Berlin Veg/Building Height
Oct '20
Polygon
m
150
https://daten.berlin.de/
Berlin Soil Nutrients, Bodenkundliche Kennwerte 2015 (Umweltatlas)
Nov '20
Polygon
mol\ m^{-2}
0
https://daten.berlin.de/
Berlin Climate Model, Air temperature 2015 (Umweltatlas)
Feb '21
Polygon
^{\circ}C
20
https://daten.berlin.de/
Planting Bed Area
Oct '20
Polygon
m^2
https://daten.berlin.de/
Berlin Soils
Oct '20
Polygon
0
https://daten.berlin.de/
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2.3 General analyses

The proposed statistical method is from the class of hierarchical, generalized additive models (GAM, or GAMM for mixed models/hierarchical models). In these models combinations of continuous and categorical predictor variables can be summed to estimate a response. In particular, continuous variables that are linearly, as well as non-linearly related to the response can be represented by applying a transfer function, typically termed "smoothing function" (Wood, 2017); these are constructed using a number of base functions of varying complexity and form, which provides a high degree of flexibility, ideal for fitting ecosystem dynamics which are rarely linear (Pedersen et al., 2019), or correctly represented with deterministic functional forms (e.g. quadratic equations). In general, a GAM can be written as:

$$E(Y) = g^{-1} \left(\beta_0 + \sum_{i=1}^n f_i(x_i) \right), \tag{1}$$

and

$$y = E(Y) + \epsilon, \tag{2}$$

where Y is taken from an appropriate distribution and corresponding link function g, β_0 is the intercept and f_i represents a smooth function of a predictor (Pedersen et al., 2019), and $\epsilon \sim \mathcal{N}(0, \sigma^2)$. Note, that f_i consists of a smooth (e.g. spline) constructed via basis functions of different form and complexity, multiplied by a coefficient:

$$f_i(x_i) = \sum_{k=1}^K \beta_{i,k} b_{i,k}(x_i).$$
 (3)

Nested data structures (e.g. due to similar road [type]) can be accounted for by introducing random effects (Wood, 2017), while spatial dependence between observations can be included by constructing smoothing functions with e.g. northings and eastings, as for example done in (Augustin et al., 2009). Ultimately, the implementation of a such a GAMM will allow for establishing continuous prediction surfaces of growth potential (approximated via DBH) for individual species across urban areas (including parks) of Berlin.

Currently, DBH has been modeled using a hierarchical linear model (linear mixed effects model) with lme4 (Bates et al., 2015) in R Core Team (2020) (see Section 3). The general form of this model is:

$$Y_{i,j} = (\beta_0 + b_{0,i,j}) + (\beta_1 + b_{1,i,j}) \cdot x_i + \epsilon_{i,j}, \tag{4}$$

where β_0 is the intercept with its random component b_0 , and β_1 the slope with its random component b_1 . The random errors are assumed i.i.d. and distributed as $b \sim \mathcal{N}(0, \tau^2)$. The model for which results are presented in Figure 5 estimates DBH from tree age and the local UHI intensity as continuous covariates with random slopes and intercepts for each species; note, that for computational efficiency each genera was modeled separately. Further, models were established for the three most abundant species per genera with at least 1000 individuals.

2.4 Available and required data

Additional remote sensing data for local context, higher resolution UHI data, and most important, temporally-resolved tree growth data are still required and/or would greatly improve the confidence in effect size estimates (see Section 1). See Tab. 2 for details on data sets.

Table 2: Data requirements for analysis, including currently available/accessed and required/desired data.

Subject /	Desc.	Type	Obsv. (n)	Resolution (m)	Source
Relevance					
available or access	ed				
urban trees	riparian, street,	tabular, spatial	668254		Senatsverwaltung
	park; basic	points			Berlin
	mensuration data				
	(not for riparian)				
UHI Effect	raster data set with	raster		200	UHI Explorer,
	summer day/night				[@chakraborty2019]
	time (global				
	coverage; Berlin				
•1	included)	. 1 1 1	150550		C 1
soil coverage	available soil area	tabular, spatial	178576		Senatsverwaltung
(Baumscheibe)	and bounding	polygons			Berlin
	infrastructure of				
vegetation and	trees data specific for	tabular, spatial			Senatsverwaltung
building height	indidivual building	polygons			Berlin
0 0	(complexes) and	1 00			
	vegetation				
required and/or de	esired				
UHI Effect	raster data set with	raster		< 50	? Landsat / Sentinel
	summer day/night				,
	time (global				
	coverage; Berlin				
	included)				
road / street	orientation and	tabular			? Senatsverwaltung,
characteristics	width of streets			~ 0	urban planning
street tree density	planting density of	tabular or raster		< 50	? Senatsverwaltung,
	trees as proxy for				urban planning
	potential				
	density-dependent inhibition				
landcover data	providing local	tabular or raster		100 - 200	? Landsat /
	development				Sentinel,
	intensity				Senatsverwaltung
tree growth series	$tree\ specific,$	tabular			? Field work,
	incremental				partners

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Table 3: Binned age-distribution	tor trees in	Berlin data set	and entries missing	age intormation
Table 9. Diffied age distribution	101 01005 111	Dermi data set	, and churics impoing	age milormanon.

Genera	(0,30]	(30,60]	(60,90]	(90,120]	$(120,\!150]$	150+	Total	Missing
							(n)	(n)
Tilia	43060	66929	37197	5403	205	113	163127	10220
Acer	33027	52167	14941	3074	203	139	140815	37264
Quercus	12085	20155	8808	4478	1012	1842	64464	16084
Betula	6302	11026	2135	138	4	1	29013	9407
Platanus	5242	12739	5105	1736	859	119	26714	914
Aesculus	5757	8651	6589	1808	201	45	25909	2858
Robinia	4587	9351	2270	376	30	9	24238	7615
Populus	1603	7819	3223	1108	207	100	19973	5913
Carpinus	6411	5125	934	303	15	15	17951	5148
Prunus	6837	3929	326	39	5	0	17904	6768
Fraxinus	6235	4862	1160	252	21	5	16835	4300
Pinus	2118	3730	2039	345	18	4	14915	6661
Other	33246	21960	5755	3832	285	863	106396	40455
Marg. Totals	166510	228443	90482	22892	3065	3255	668254	153607

3 Preliminary results

Tree locations are clustered and structured based on their category, i.e. riparian, street and park trees (Fig. 1). Planting in space and time shows species-specific patterns (by districts), often related to major events, such as the start and end of armed and/or political conflict. Table 3 shows the binned distribution of genera across age classes.

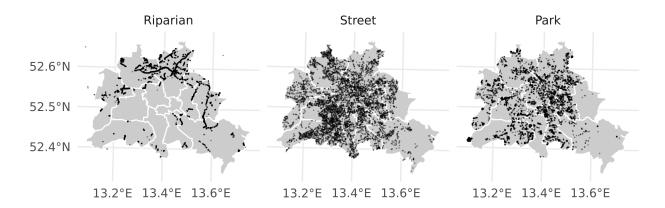
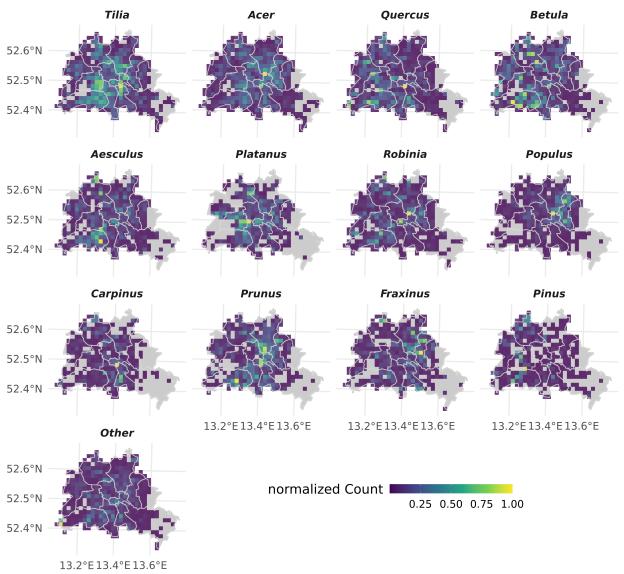


Figure 1: Individual tree locations for three categories available in Berlin Senate urban tree data set. Note, that for each category 7000 observations were subsampled from the available pool to facilitate visualization.

The distribution of the UHI effect is highly irregular and clustered in space (Fig. 3), and also shows variability through time (data not shown, refer to the urban heat island explorer).

The exposure to increased heat-loading of individual genera (and consequently species) is highly uneven throughout the city (Fig. 4). Street and park trees of most genera are clustered in urban areas with intermediate to high UHI loading, while riparian trees, and some street and park trees of other genera tend to be spread more evenly across Berlin's UHI range.

Note, that results below are preliminary and should be considered as a template for future



Data source: daten.berlin.de; WFS Service, accessed: 2019-12-15

Figure 2: Gridded counts for the 11 most frequent genera, as well as *Pinus* and remaining genera. Note, that counts are standardized to unity for individual genera.

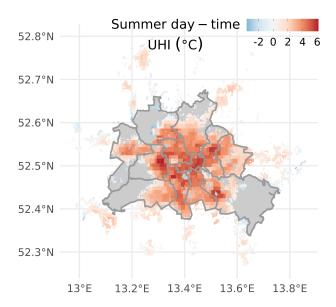


Figure 3: Estimate of UHI intensity based on the algorithm in (Chakraborty and Lee, 2019), comparing urban with rural pixels within the greater metropolitan cluster. Presented values are averaged over the summer of 2007.

outputs, rather than used for inference. The effect of UHI loading on absolute growth potential varies between genera and species (Fig. 5). Most notably, *Quercus*, the 3rd-most frequent genera, shows decreased absolute growth with increasing UHI loading, while the most frequent genera, *Tilia*, features contrasting relationships between species. The estimated effect sizes presented here are linear. However, temperature may exert a non-linear control on absolute growth and, hence, applying a method able to capture such dynamics may result in somewhat different effect sizes / behavior. Additionally, if temperatures increase in the future under climate warming, any non-linear effects may become more enhanced, stressing the need for a more flexible model fit and structure (i.e. using GAMM over linear models-).

4 Outlook

We seek to build upon and improve the current analysis by:

- validating the database with independent observations
- incorporating more pertinent covariates as dependent variables in the linear mixed model
- testing multiple model structures with formal model selection procedures
- checking model residuals for spatial auto-correlation and accounting for it where necessary to ensure unbiased estimates of effect sizes
- repeating the above with a hierarchical GAM (i.e. GAMM) to allow for:
 - estimating continuous prediction surfaces for UHI impacts on individual species' growth (similar to results in Figure 5) under recent conditions
 - estimating absolute, species-specific growth potential under increased temperatures and UHI loading under climate change, ideally based on climate simulations (otherwise step-wise increases based on RCP scenarios) for the key species. Note, that complications presented by 'out-of-sample' predictions will be addressed.
 - assess potential age-dependent UHI impacts on individual species.
 - repeating the above (GAMM) with total and incremental basal area as responses

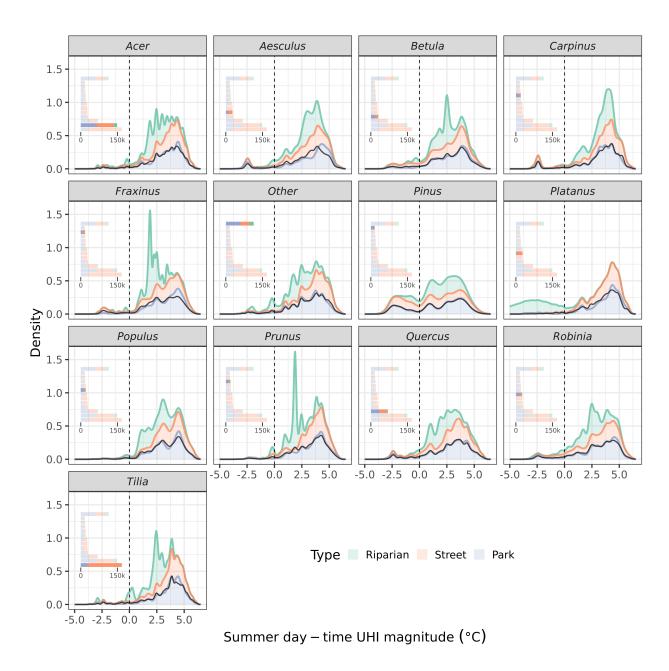


Figure 4: Empirical density distribution of all individuals within the presented genera along the UHI continuum. UHI intensities were extracted for each tree location, and the distribution hence represents the first detailed overview of the exposure of Berlin's trees to urban heat loading. The black line is the density across all three categories. Insets show corresponding tree totals.

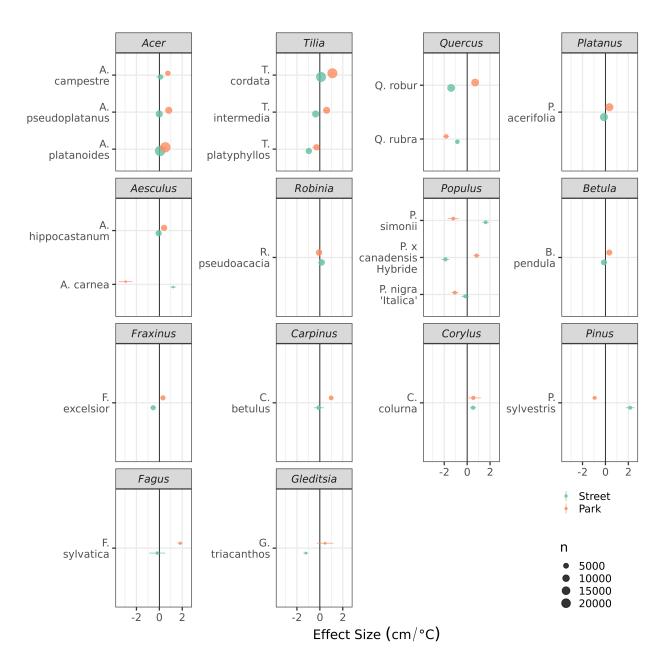


Figure 5: Impact of UHI loading on tree diameter (DBH), accounting for age and inter-specific differences from the linear mixed model (via random slopes and intercepts). Line-ranges are standard errors of predicted effect sizes (i.e. slopes). Differences between street and park trees are considerable for some species, and may be due to local clustering and/or spatial under-representation across the UHI continuum. Further investigations need to address the degree of spatial autocorrelation and account for it where required in linear mixed models, and with smoothing interactions in a GAMM implementation.

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5.0.1 Colophon

This report was generated on 2021-05-18 14:49:17 using the following computational environment and dependencies:

The current Git commit details are:

#> Local: master C:/Users/ahurl/Documents/_work/p024_gfz_berlin-trees/berlin.trees

#> Remote: master @ origin (https://github.com/the-Hull/berlin.trees.git)

#> Head: [43beb0c] 2021-05-18: add refs