

1 An investigation of factors affecting inferences from and
2 reliability of state-space age-structured assessment
3 models

4 Timothy J. Miller^{1,2} ~~Greg~~Gregory L. Britten³

5 Elizabeth N. Brooks² Gavin Fay⁴ ~~Alex~~Alexander C. Hansell²

6 Christopher M. Legault² Chengxue Li² Brandon Muffley⁵

7 Brian C. Stock⁶ John Wiedenmann⁷

8 ~~02 May, 2025~~04 January, 2026

9 ¹corresponding author: timothy.j.miller@noaa.gov

10 ²Northeast Fisheries Science Center, Woods Hole Laboratory, 166 Water Street, Woods
11 Hole, MA 02543 USA

12 ³Biology Department, Woods Hole Oceanographic Institution, 266 Woods Hole Rd. Woods
13 Hole, MA, USA

14 ⁴Department of Fisheries Oceanography, School for Marine Science and Technology,
15 University of Massachusetts Dartmouth, 836 S Rodney French Boulevard, New Bedford,
16 MA 02740, USA

¹⁷ ⁵Mid-Atlantic Fishery Management Council, 800 North State Street, Suite 201, Dover, DE
¹⁸ 19901 USA

¹⁹ ⁶Institute of Marine Research, Nye Flødevigveien 20, 4817 His, Norway

²⁰ ⁷Department of Ecology, Evolution, and Natural Resources. Rutgers University

²¹

keywords: stock assessment, state-space, model selection, bias, convergence, retrospective patterns

Abstract

State-space models ~~are increasingly used for stock assessment , and evaluations of their statistical reliability and best practices for selecting among process error configurations are~~ have been promoted as the next-generation of fisheries stock assessment and evaluation of their reliability is needed. We simulated ~~72~~ operating models that varied fishing pressure ~~and observation error across process errors in recruitment, survival, selectivity, catchability, and /or natural mortality. We fit estimating models with different assumptions on the process error source and whether ,~~ magnitude of observation error, and sources of process error. For each operating model, we fit a range of estimating models with correct and incorrect configurations. We measured reliability of estimating models by convergence rate, accuracy of AIC-based model selection, estimation bias, and magnitude of retrospective patterns. All reliability measures were generally better with lower observation error, contrast in fishing pressure over time, and when median natural mortality ~~or a stock-recruit relationship were estimated. Estimating models without a stock-recruit relationship that assumed the correct rate is known. The magnitude of the log-likelihood gradients was not a reliable indicator of convergence. AIC can generally distinguish~~ process error source ~~and median natural mortality had high convergence rates and low bias. Bias was also low under many incorrect process error assumptions when there was contrast in fishing pressure and low observation error. Marginal AIC most accurately distinguished process errors on recruitment , survival, and selectivity, and other process error sources when variability was greater~~ with lower observation error and higher true process error variability. Distinguishing the stock recruit relationship with AIC required large contrast in spawning biomass and low recruitment variation, but bias in stock-recruit parameter estimation was prevalent. Retrospective pat-

47 terns were ~~generally small but were sizable for recruitment when observation error was high.~~
48 ~~These results help establish the statistical reliability of state space assessment models and~~
49 ~~pave the way for the next generation of fisheries stock assessment~~not large for mis-specified
50 models. These findings improve our understanding of when results from state space models
51 will be reliable.

Introduction

Application of state-space models in fisheries stock assessment and management has expanded dramatically within the International Council for the Exploration of the Sea (ICES), Canada, and the Northeast US (Nielsen and Berg 2014; Cadigan 2016; Pedersen and Berg 2017; Stock and Miller 2021). State-space models treat latent population characteristics as statistical time series with periodic observations that also may have error due to sampling or other ~~sources of measurement error~~ measurement properties. Traditional assessment models may use state-space approaches to account for temporal variability in population characteristics (Legault and Restrepo 1999; Methot and Wetzel 2013), but these models treat the annual parameters as penalized fixed ~~effects~~ effect parameters where the variance parameters controlling the penalties are assumed known (Thorson and Minto 2015). Modern state-space models can estimate the annually varying parameters as random effects with variance parameters estimated using maximum marginal likelihood or corresponding Bayesian approaches. These ~~latter random-effects~~ approaches are considered best practice and ~~a~~ are recommended for the next generation of stock assessment models (Hoyle et al. 2022; Punt 2023).

State-space stock assessment models, with nonlinear functions of latent parameters and multiple types of observations with varying distributional assumptions, are one of the most complex examples of this analytical approach. Statistical aspects of state-space models and their application within fisheries have been studied extensively, but previous work has focused primarily on linear and Gaussian state-space models (Aeberhard et al. 2018; Auger-Méthé et al. 2021). Therefore, current understanding of the reliability of state-space models does not extend to usage for stock assessment.

As state-space models provide greater flexibility by allowing multiple processes to vary as random effects (Nielsen and Berg 2014; Aeberhard et al. 2018; Stock et al. 2021), one of the most immediate questions regards the implications of mis-specification among alternative sources of process error. Incorrect treatment of population attributes as temporally varying

(Trijoulet et al. 2020; Liljestrand et al. 2024) could lead to misidentification of stock status and biased population estimates, ultimately impacting fisheries management decisions (Legault and Palmer 2016; Szuwalski et al. 2018; Cronin-Fine and Punt 2021). Furthermore, biological, fishery, and observational processes are often confounded in catch-at-age data, which may adversely affect the ability to distinguish between true process variability and observational error (~~Li et al. In review;~~ Punt et al. 2014; Stewart and Monnahan 2017; Cronin-Fine and Punt 2021; Fisch et al. 2023; Li et al. 2025a).

Li et al. (2024) conducted a full-factorial simulation-estimation study to assess model reliability when confounding random-effects processes (numbers-at-age, fishery selectivity, and natural mortality) were included. Their results suggest that while state-space models can generally identify sources of process error, overly complex models, even when misspecified (i.e., incorporating process error that did not exist in reality), often performed similarly to correctly specified models, with little to no bias in key management quantities. Similarly, Liljestrand et al. (2024) found little downside in assuming process error in recruitment or selectivity, even when it was absent.

Despite ~~mounting efforts, several limitations~~ increasing research on state space assessment models, several uncertainties in state space assessment modeling remain. First, confounding processes that can be treated as random effects in the model ~~were not~~ have not been thoroughly examined or tested within a simulation-estimation framework. Second, previous studies relied on operating models conditioned on specific fisheries, limiting their generalizability (~~Li Liljestrand et al. In review; Liljestrand 2024; Li et al. 2024~~ 2025a). In particular, the effects of observation error and underlying fishing history have not been fully isolated in simulation study designs, making it challenging to disentangle the interplay between process and observation error magnitudes, as demonstrated in Fisch et al. (2023). Third, explicitly modeling stock-recruit relationships (SRRs) as mechanistic drivers of population dynamics is promising (Fleischman et al. 2013; Pontavice et al. 2022), but reliability of inferences within integrated state-space age-structured models has not been evaluated. Evidence from

other studies suggests that when both process and observation errors are unknown, estimating density dependence parameters becomes highly uncertain (Knappe 2008; Polansky et al. 2009). In particular, Knappe (2008) demonstrated that stronger density dependence becomes increasingly difficult to estimate in the presence of observation error. Therefore, it is crucial to assess whether ~~density-dependence~~ density-dependent mechanisms can be estimated with sufficient precision for use in fisheries management (Auger-Méthé et al. 2016). Finally, although the importance of autocorrelation in process errors is recognized, investigations of the ability to distinguish state-space assessment models with and without autocorrelation and whether such misspecification is detrimental to estimation of important population metrics are lacking (Johnson et al. 2016; Xu et al. 2019).

In the present study, we conduct a simulation study with operating models (OMs) varying by degree of observation error, source and variability of process error, and fishing history. The simulations from these OMs are fitted with estimation models (EMs) that make alternative assumptions for sources of process error, whether ~~a~~ an SRR was estimated, and whether natural mortality is estimated. Given the confounding nature of process errors, developing diagnostic tools to detect model misspecification is of great scientific interest and could aid the next generation of stock assessments (Auger-Méthé et al. 2021). We evaluate whether ~~convergence-and~~ OM and EM attributes affect rates of convergence and the ability of Akaike Information Criterion (AIC) ~~can-to~~ correctly determine the source of process error ~~and-or~~ the existence of ~~a~~ an SRR. We also evaluate ~~when retrospective patterns occur and the degree of bias in the outputs of the assessment model that are effects of OM and EM attributes on~~ magnitude of retrospective patterns and bias in estimation of parameters and other model outputs important for management.

Methods

We used the Woods Hole Assessment Model (WHAM) to configure OM_s and EM_s in our simulation study (~~Miller and Stock 2020; Stock~~ [Stock](#) and Miller 2021; [Miller et al. 2025](#)). WHAM is an R package freely available via a ~~github~~ [Github](#) repository and is built on the Template Model Builder package (Kristensen et al. 2016). For this study we used version 1.0.6.9000, commit 77bbd94. WHAM has also been used to configure OM_s and EM_s for closed loop simulations evaluating index-based assessment methods (Legault et al. 2023) and is currently used or accepted for use in management of numerous ~~NEUS~~ [Northeast United States \(NEUS\)](#) fish stocks (e.g., NEFSC 2022a, 2022b; NEFSC 2024).

We completed a simulation study with a number of OM_s that can be categorized based on where process error random effects were assumed: ~~recruitment (R, assumed present in all models);~~ [R OM_s assume process error for recruitment only](#). Other OM categories assume [recruitment process errors along with process errors for](#) apparent survival (~~denoted~~ $R+S$), natural mortality ($R+M$), fleet selectivity ($R+Sel$), or index catchability ($R+q$). We refer to the ~~($R+S$)~~ OM_s as modeling apparent survival because on ~~logscale~~ [log-scale](#) the random effects (~~$\epsilon_{a,y}$~~) are additive to the total mortality (~~$F+M$~~ [fishing and natural mortality](#)) between numbers at age, thus they modify the survival term. ~~However, as Stock and Miller (2021) note, these random effects can be due to events other than mortality, such as immigration, emigration, misreported catch, and other sources of misspecification.~~ For each OM, assumptions about the magnitude of the variance of process errors and observations are required and the values we used were based on a review of the range of estimates from ~~Northeast United States (NEUS)~~ [NEUS](#) assessments using WHAM.

In total, we configured 72 OM_s with alternative assumptions about the source and magnitude of process errors, magnitude of observation error in indices and age composition data, and contrast in fishing pressure over time. ~~We fitted 20 EM_s to observations generated from each of~~ [For each OM, we simulated](#) 100 ~~simulations where process errors were also simulated~~

~~Each EM~~ time series of abundance at age with process errors, and for each realized time series, we simulated observation data sets. For each data set, we fitted a number of EMs that differed in assumptions about the source of process errors, whether natural mortality (or the median for models with process error in natural mortality) was estimated, and whether a Beverton-Holt SRR was estimated within the EM. Details of each of the OMs and EMs are described below.

We did not use the log-normal bias-correction feature for process errors or observations described by (Stock and Miller (2021) for OMs and EMs to simplify interpretation of the study results (Li et al. ~~In review~~2025b). All code we used to perform the simulation study and summarize results can be found at https://github.com/timjmiller/SSRTWG/tree/main/Project_0/code.

Operating models

Population

~~We intended the population demographics and observation types to represent a general NEUS groundfish stock.~~ The population consists of 10 age classes, ages 1 to 10+, with the last being a plus group that accumulates ages 10 and older. ~~We assume spawning occurs annually 1/4 of the way through the year.~~ The maturity at age was a logistic curve with $a_{50} = 2.89$ and slope = 0.88 (Figure S1, top left).

Weight at age (W_a) was generated with a von Bertalanffy growth function defining length at age:

$$L_a = L_\infty \left(1 - e^{-k(a-t_0)}\right),$$

where $t_0 = 0$, $L_\infty = 85$, and $k = 0.3$, and a ~~L-W~~length-weight relationship such that

$$W_a = \theta_1 L_a^{\theta_2},$$

where $\theta_1 = e^{-12.1}$ and $\theta_2 = 3.2$ (Figure S1, top right).

We assumed a Beverton-Holt SRR with constant pre-recruit mortality parameters for all OMs. We assume spawning occurs annually 0.25 of each year and recruitment at age 1 ($N_{1,y}$). All biological inputs to calculations of spawning ~~biomass-stock biomass (SSB)~~ per recruit (i.e., weight, maturity, and natural mortality at age) are constant in the ~~apparent survival ($R+S$) selectivity ($R+Sel$), and survey catchability ($R+q$)~~ process error OMs. Therefore, steepness and unfished recruitment are also constant over the time period for those OMs (Miller and Brooks 2021). We assumed a value of 0.2 for the natural mortality rate in OMs without process errors on natural mortality. We specified unfished recruitment equal to e^{10} and $F_{MSY} = F_{40\%} = 0.348$, which equates to a steepness of 0.69 and $a = 0.60$ and $b = 2.4 \times 10^{-5}$ for the Beverton-Holt parameterization

$$N_{1,y} = \frac{aSSB_{y-1}}{1 + bSSB_{y-1}}$$

(Figure S1, bottom right). ~~We assumed a value of 0.2 for the natural mortality rate in OMs without process errors on natural mortality and for the median rate for OMs with process errors on~~ For OMs with time-varying random effects for natural mortality, steepness is not constant. However, we used the same a and b parameters as other OMs, which equates to a steepness and R_0 at the median of the time series process for natural mortality. Similarly, for OMs with time-varying random effects for fishery selectivity, F_{MSY} also varies temporally, so equilibrium conditions for these OMs are defined for mean selectivity parameters.

We used two fishing scenarios for OMs. In the first scenario, the stock experiences overfishing at $2.5F_{MSY}$ for the first 20 years followed by fishing at F_{MSY} for the last 20 years (denoted $2.5F_{MSY} \rightarrow F_{MSY}$). In the second scenario, the stock is fished at F_{MSY} for the entire time period (40 years). The magnitude of the overfishing assumptions is based on average estimates of overfishing for NEUS groundfish stocks from Wiedenmann et al. (2019) and similar to the approach in Legault et al. (2023).

The second scenario represents the ideal situation where the stock is fished at an optimal level, but provides less contrast in stock sizes over time. We specified initial population

abundance at age at the equilibrium distribution that corresponds to fishing at either ~~$F = 2.5 \times F_{\text{MSY}}$~~ $F = 2.5 F_{\text{MSY}}$ or $F = F_{\text{MSY}}$. This implies that, for a deterministic model, the abundance at age would not change from ~~the first year to the next.~~

~~For OMs with time-varying random effects for M , steepness is not constant. However, we used the same a and b parameters as other OMs, which equates to a steepness and R_0 at the median year to year at the beginning of the time series process for M . For OMs with time-varying random effects for fishery selectivity, is also not constant, but since we use the same F history as other OMs, this corresponds to at the mean selectivity parameters.~~

Fleets

We assumed a single fleet operating year round for catch observations with logistic selectivity ~~for the fleet~~ ($a_{50} = 5$ and slope = 1; Figure S1, bottom left). This selectivity was used to define F_{MSY} for the Beverton-Holt SRR parameters above. We assumed a logistic-normal distribution with no correlation on the multivariate normal scale for the corresponding annual age-composition observations~~for the fleet~~.

Indices

Two time series of fishery-independent surveys measured in numbers are generated for the entire 40 year period with one occurring in the spring (0.25 of each year) and one in the fall (0.75 of each year), representing current bottom trawl surveys conducted in the NEUS. Catchability of both surveys are assumed to be 0.1. Like the fishing fleet, we assumed logistic selectivity for both indices ($a_{50} = 5$ and slope = 1) and a logistic-normal distribution with no correlation on the multivariate normal scale for the annual age-composition observations.

222 Observation Uncertainty

223 The standard deviation for log-aggregate catch was 0.1 for all OMs, a common assumption
224 for commercial removals in NEUS stock assessments. Two levels of observation error variance
225 (high and low) were specified for indices and all age composition observations (both indices
226 and catch). The low uncertainty specification assumed a standard deviation of 0.1 for both
227 series of log-aggregate index observations, and the standard deviation of the logistic-normal
228 for age composition observations was 0.3. In the high uncertainty specification, the standard
229 deviation for log-aggregate indices was 0.4 and that for the age composition observations
230 was 1.5. The low standard deviation for index observations is typical for fish stocks that
231 are consistently sampled across survey stations whereas the high value is typical for more
232 sporadically sampled stocks. The standard deviations for the age composition observations
233 were determined from the range of values estimated from WHAM fits to NEUS stocks that
234 assumed the logistic-normal model. For all EMs, the standard deviation for log-aggregate
235 observations was assumed known whereas that for the logistic-normal age composition ob-
236 servations was estimated.

237 Operating models with random effects on numbers at age

238 For operating models with random effects on recruitment ~~and(or)~~ only and also on ap-
239 parent survival (R, R+S), we assumed marginal standard deviations for recruitment of
240 $\sigma_R \in \{0.5, 1.5\}$ ~~and~~ the marginal standard deviations for apparent survival random effects
241 at older age classes ~~of~~ were $\sigma_{2+} \in \{0, 0.25, 0.5\}$. The full factorial combination of these
242 process error assumptions (~~2x3~~ 2 x 3 levels) and scenarios for fishing history (2 levels) and
243 observation error (2 levels) scenarios described above results in 24 different R ($\sigma_{2+} = 0$) and
244 R+S operating models (Table S1).

Operating models with random effects on natural mortality

All R+M OMs treat natural mortality as constant across age, but with annually varying random effects. WHAM treats natural mortality as a log-transformed parameter

$$\log M_{y,a} = \mu_M + \epsilon_{M,y}$$

that is a linear combination of a mean log-natural mortality parameter that is constant across ages ($\mu_M = \log(0.2)$) and any annual random effects are marginally distributed as $\epsilon_{M,y} \sim N(0, \sigma_M^2)$. The marginal standard deviations we assumed for log natural mortality random effects were $\sigma_M \in \{0.1, 0.5\}$ and the random effects were either uncorrelated or first-order autoregressive (AR1, $\rho_M \in \{0, 0.9\}$). Uncorrelated random effects were also included on recruitment with $\sigma_R = 0.5$ (hence, we denote these OMs as R+M). The full factorial combination of these process error assumptions and fishing history (2 levels) and observation error (2 levels) scenarios described above results in 16 different R+M OMs (Table S2).

Operating models with random effects on fleet selectivity

WHAM treats each selectivity parameter s as a logit-transformed parameter

$$\log \left(\frac{p_{s,y} - l_s}{u_s - p_{s,y}} \right) = \mu_s + \epsilon_{s,y}$$

that is a linear combination of a mean μ_s and any annual random effects marginally distributed as $\epsilon_{s,y} \sim N(0, \sigma_s^2)$, where the lower and upper bounds of the parameter (l_s and u_s) can be specified by the user. All selectivity parameters (a_{50} and slope parameters) were bounded by ~~0 and 10~~ $s_l = 0$ and $s_u = 10$ for all OMs and EMs. The marginal standard deviations we assumed for logit scale random effects were $\sigma_s \in \{0.1, 0.5\}$ and AR1 autocorrelation parameters of $\rho_s \in \{0, 0.9\}$. Like R+M OMs, the full factorial combination of these process error assumptions (2x2 levels) and scenarios described above for fishing history (2

levels) and observation error (2 levels) results in 16 different R+Sel OMs (Table S3).

Operating models with random effects on index catchability

Like selectivity parameters, WHAM treats catchability for an index i as a logit-transformed parameter

$$\log \left(\frac{q_{i,y} - l_i}{u_i - q_{i,y}} \right) = \mu_i + \epsilon_{i,y}$$

that is a linear combination of a mean μ_i and any annual random effects marginally distributed as $\epsilon_{i,y} \sim N(0, \sigma_i^2)$ where the lower and upper bounds of the catchability (l_i and u_i) can be specified by the user. We assumed bounds of 0 and 1000 for all OMs and EMs. For all OMs and EMs with process errors on catchability, the temporal variation only applies to the first index, which could be interpreted as capturing some unmeasured seasonal process that affects availability to the survey. The marginal standard deviations we assumed for logit scale random effects were $\sigma_i \in \{0.1, 0.5\}$ and AR1 autocorrelation parameters of $\rho_i \in \{0, 0.9\}$. Like R+M and R+Sel OMs, the full factorial combination of these process error assumptions and fishing history (2 levels) and observation error (2 levels) scenarios described above results in 16 different R+q OMs (Table S4).

Estimation models

For each of the data sets simulated from an OM, 20 EMs were fit. A total of 32 different EMs were fit across OMs where the subset of 20 depended on the source of process error in the OM (Table S5). The EMs have different assumptions about the source of process error (R+S, R+M, R+Sel, R+q) and whether or not 1) there is temporal autocorrelation, 2) a Beverton-Holt SRR is estimated, and 3) the natural mortality rate (μ_M , the constant or mean on log scale for R+M EMs) is estimated. For simplicity we refer to the derived estimate e^{μ_M} as the median natural mortality rate regardless of whether natural mortality random effects are estimated in the EM.

Subsets of 20 EMs in Table S5 were fit to ~~simulate~~simulated data sets from each of the OM process error ~~categories~~sources. For R and R+S OM, fitted EMs had matching process error assumptions as well as R+Sel, R+M, and R+q assumptions without autocorrelation. ~~Similarly,~~ For other OM process error ~~categories~~sources, we fit EMs with ~~matching~~correct process error assumptions~~as well as other process error types~~, the correct process error source but incorrect correlation assumption, and the incorrect process error source without autocorrelation. As such, EMs were configured correctly for the OM, or they had mis-specification in assumptions of process error autocorrelation, the source of process error, and(or) the SRR (Beverton-Holt or none).

The maturity at age, weight at age for catch and SSB, and observation error ~~variance of~~standard deviations for aggregate catch and indices were all assumed known at the true values. However, the variance parameters for the logistic-normal distributions for age composition observations were estimated in the EMs. ~~As such, EMs would either be configured completely correctly for the OM, or there could be mis-specification in assumptions of process error autocorrelation, the type of process error, or the SRR (Beverton-Holt or none).~~

Measures of reliability

Convergence

The first measure of reliability we investigated was frequency of convergence when fitting each EM to the simulated data sets. There are various ways to assess convergence of the fit (e.g., Carvalho et al. 2021; Kapur et al. 2025), but given the importance of estimates of uncertainty when using assessment models in management, we estimated ~~probablity~~probability of convergence as measured by occurrence of a positive-definite ~~hessian~~Hessian matrix at the optimized negative log-likelihood that could be inverted (i.e., providing Hessian-based standard error estimates). We also provide results in the Supplementary Materials for ~~the maximum~~convergence defined by the maximum absolute gradient $< 1^{-6}$ and the maximum

of the absolute ~~values among all gradients~~ gradient values for all fits of a given EM to all simulated data sets from a given OM that produced ~~hessian-based~~ Hessian-based standard errors for all estimated fixed effects. This provides an indication of how poor the calculated gradients can be, but still presumably converged adequately enough for parameter inferences. ~~We used the Clopper-Pearson exact method for constructing 95% confidence intervals of the probabilities of convergence (Clopper and Pearson 1934; Thulin 2014).~~

AIC for model selection

We ~~estimated the probability of selection~~ investigated the reliability of AIC-based model selection for two purposes. First, we analyzed selection of each process error model ~~structure~~ source (R, R+S, R+M, R+Sel, R+q) using marginal AIC. For a given ~~operating model~~ OM simulated data set, we compared AIC for EMs ~~that were all configured the same for with~~ different process error assumption conditional on whether median natural mortality (~~known or estimated~~) and the SRR (rate and the Beverton-Holt ~~or none~~).

~~We also estimated the probability of correctly selecting SRR were estimated.~~ Second, we analyzed AIC-based selection between EMs with ~~Beverton-Holt SRR assumed and models without the SRR (null model).~~ We made these comparisons between models that otherwise ~~assumed the same process error structure as the operating model and both of the compared models either estimate median natural mortality or assume it is known~~ and without the Beverton-Holt SRR assumed. Contrast in fishing pressure and time series with recruitment at low stock size ~~has~~ have been shown to improve estimation of SRR parameters (Magnusson and Hilborn 2007; Conn et al. 2010). Our preliminary inspections ~~of the proportions of simulations where the correct recruitment model was chosen~~ indicated generally poor performance of AIC in determining the Beverton-Holt SRR model for a given set of OM factors (including contrast in fishing pressure) ~~indicated generally poor performance of AIC,~~ even when the EM was configured with the correct process error source. Therefore, we ~~fit~~

~~logistic regression models to the indicator of Beverton-Holt models having lower AIC as~~
~~a function conditioned on the EMs having the correct process error assumption and also~~
~~considered the effect~~ of the log-standard deviation of the true $\log(\text{SSB})$ ($\log \text{SD}_{\text{SSB}}$; similar
to the log of the coefficient of variation for SSB) ~~on model selection~~ since simulations with
realized SSB producing low and high ~~recruitments~~ recruitment would have larger variation
in realized SSB.

All model selection results condition ~~on whether all of the compared estimating models~~
~~completed the~~ only on completion of the optimization process without failure for all of the
compared EMs. We did not condition on convergence as defined ~~by a gradient threshold~~
~~or invertibility of the hessian above~~ because optimization could correctly determine an inap-
propriate process error assumption by estimating variance parameters at the lower bound
of zero. Such an optimization could indicate poor convergence but the likelihood would be
equivalent to that without the mis-specified random effects and the AIC would be appro-
priately higher because more (variance) parameters were estimated. All other measures of
reliability described below (bias and Mohn's ρ) use these same criteria for inclusion of EM
fits in the summarized results.

Bias

We also investigated bias in estimation of various model attributes as a measure of reliability.

For a given model attribute we calculated the relative error

$$\text{RE}(\theta_i) = \frac{\hat{\theta}_i - \theta_i}{\theta_i}$$

$$\text{RE}(\theta_j) = \frac{\hat{\theta}_j - \theta_j}{\theta_j} \quad (1)$$

from fitting a given ~~estimating model~~ EM to simulated data set ~~i~~ j configured for a given
OM where ~~$\hat{\theta}_i$ and θ_i~~ $\hat{\theta}_j$ and θ_j are the estimated and true values for simulation ~~i~~ . ~~We~~

~~estimated bias as the median of the relative errors across all simulations for a given OM and EM combination. We constructed 95% confidence intervals for the median relative bias using the binomial distribution approach as in Miller and Hyun (2018) and Stock and Miller (2021). We present results for bias in terminal year estimates of j . We analyzed simulation results for estimates of terminal year SSB and recruitment, Beverton-Holt stock-recruit-SRR parameters (a and b), and median natural mortality rate. Results for terminal year fishing mortality were strongly negatively correlated with those for SSB and are provided in the Supplementary Materials.~~

Mohn's ρ

~~Finally, we investigated presence of retrospective patterns in fitted models as a measure of reliability. We calculated Mohn's ρ for SSB, fully-selected fishing mortality-fishing mortality (averaged over all age classes), and recruitment for each EM fit to each OM simulated data set (Mohn 1999). We fit 7-peels for each EM and calculated median 95% confidence intervals for $P = 7$ peels to each simulated data set and calculated Mohn's ρ using the same methods as that for relative bias.~~ for a given attribute θ as

$$\rho(\theta) = \frac{1}{P} \sum_{p=1}^P \frac{\hat{\theta}_{Y-p, Y-p} - \hat{\theta}_{Y-p, Y}}{\hat{\theta}_{Y-p, Y}} \quad (2)$$

~~where Y is last year of the full set of observations and $\hat{\theta}_{y, y'}$ is the estimate for attribute θ in year y from a model fit using data up to year $y' \geq y$. Thus, θ terms where $y' = Y$ refer to estimates from the fit to all years of data.~~

Results

Summarizing results across OM and EM attributes

~~For many R and R+S OMs, convergence rate declined when either the median natural~~

~~mortality rate or the Beverton-Holt SRR was estimated even when the process error~~
~~assumptions of the EMs and OMs matched (Figure S17, A). When there was high~~
~~observation error and constant fishing pressure ($F = F_{\text{MSY}}$ for all 40 years), convergence~~
~~was poor for of all~~ Because the OM and EM attributes that we investigated are numerous,
 we used two methods to summarize the most important factors for differences in results
 within a given OM process error source. The first method was fitting regression models
 with the response being each of the measures of reliability described above and predictor
 variables were defined based on OM and EM characteristics (e.g., MacKinnon et al.
 1995; Wang et al. 2017; Harwell et al. 2018). For the binary indicators of convergence
 and AIC-based selection of an SRR, we performed logistic regressions. For indicators of
 AIC-based selection of EM process error ~~configurations other than R EMs when fitted to R~~
~~OMs ($\sigma_{2+} = 0$) regardless of whether median natural mortality and SRRs were estimated.~~
~~Convergence of R EMs was high for all R and R+S OMs except when there was high~~
~~observation error and constant fishing pressure,~~ source (multiple categories) we performed
 multinomial regressions. For other measures of reliability we fit linear regression models to
 transformed responses. Because relative errors (Eq. 1) and ~~when median natural mortality~~
~~and SRRs were estimated. R+S EMs fit to R OMs exhibited poor convergence regardless of~~
~~whether natural mortality or a SRR was estimated. R+S EMs fit to R+S OMs had highest~~
~~convergence rates when there was contrast in fishing pressure and low observation error~~
 Mohn's ρ for the various parameters are bounded below at -1, we used a transformation of
 these values

$$y_j = \log \left[f \left(\hat{\theta}_j, \theta_j \right) + 1 \right] \quad (3)$$

where f is either the relative error (Eq. 1) or Mohn's ρ (Eq. 2) for simulation j , so that
 values are unbounded. For relative errors, y_j is the log-scale error. We omitted simulations
 where estimated attributes equal to zero ($\text{RE} = -1$). For all regressions we fit separate models
 with just individual OM and EM factors included, with all factors included, with all second
 order interactions, and with all third order interactions. For the multinomial regression,

we used the `vglm` function from the VGAM package (Yee 2008; Yee 2015). We tabulated percent reduction in residual deviance for each of the regression fits. We did not perform formal statistical analyses of effects of OM and EM attributes on results (e.g., ANOVA) because of the lack of independence of the “observations” that results from fitting multiple EMs to each simulated data set.

The second method involved fitting classification and regression trees (Breiman et al. 1984) to show how the OM and EM attributes, and their interactions, partition the values for each measure of reliability (e.g., Gonzalez et al. 2018; Collier et al. 2022). We used classification trees for categorical measures (convergence and AIC) and regression trees for the other measures with continuous scales (relative error and Mohn’s ρ). The response variables were the same as the regressions for the deviance reduction analyses. We used the `rpart` function in the `rpart` package (Therneau and Atkinson 2025) to fit trees. Full trees were determined using default settings except that we increased the number of cross-validations to 100. For clarity, we manually pruned the full trees to show just the primary branches. ~~Convergence rates were high for all EMs when fit to data from R+S OMs with lower observation error except those where median natural mortality and/or SRRs were estimated.~~

~~Convergence of all EMs fitted to R+M OMs was highest when the OMs had higher natural mortality process error variability, low observation error, and contrast in fishing pressure (Figure S17, B).~~ We also provide detailed results for all measures of reliability at each combination of OM and EM attributes in the Supplementary Materials. For confidence intervals of probability of convergence, we used the Clopper-Pearson exact method (Clopper and Pearson 1934; Thulin 2014). For AIC selection of process error source we provide estimates of the proportions of simulations where each EM type was selected. For AIC selection of the SRR (a binary indicator for each simulated data set), we fit logistic regressions and present resulting predicted probabilities of correctly selecting the SRR as a function of SSB variability ($\log \text{SD}_{\text{SSB}}$ described above). We estimated bias as the median of the relative errors across all simulations for a given OM and EM combination. We constructed

95% confidence intervals for the median relative bias, and Mohn's ρ using the binomial distribution approach (Thompson 1936) as in Miller and Hyun (2018) and Stock and Miller (2021).

Results

Convergence performance

For probability of convergence, the EM process error assumption was the single attribute that resulted in the largest percent reduction in deviance (14-28%) for all OM process error sources other than R+M EMs that estimated autocorrelation of process errors had poor convergence for R+M OMs when there was low natural mortality process error variability regardless of autocorrelation of the simulated process errors. R+S EMs fitted to data generated from R+M OMs always converged poorly whether or not OMs where the EM median natural mortality and the Beverton-Holt SRR were estimated. rate assumption (estimated or known) explained the most residual deviance (>11%; Table 1). However, including interactions of OM and EM factors also provided large reductions in residual deviance (35-47%), suggesting successful convergence depended on a combination OM and EM attributes.

The R Classification trees for each OM process error source all had the primary branch defined using the same attribute that provided the largest reduction in deviance (Figure 1). EMs that assumed R+S EMs, in particular, had poor convergence when fit to data generated from RS process errors converged poorly for all OMs that were simulated with the alternative process error assumptions (R, R+Sel OMs with lower selectivity process error variability or higher observation error (Figure S17, C). M, R+Sel EMs generally converged better than other EMs for Sel, and R+Sel OMs with higher process error variability, lower observation error, and contrast q OMs). For all trees, branches based on the OM fishing mortality history showed better convergence when the OM included a change in fishing pressure regardless of

~~whether~~. Branches based on whether the Beverton-Holt SRR was assumed or not, showed better convergence when it was not estimated and branches based on the median natural mortality ~~or a SRR was estimated~~.

~~For~~ rate assumption showed better convergence when it was treated as known. For some R+q OMs, convergence of R+M and R+q EMs was generally better than that of other EMs. Sel OMs, better convergence was also observed when there was contrast in fishing pressure (Figure (S17, D)). Convergence of R+S EMs was generally worse than that of all other EMs across all OMs whether or not median natural mortality or a SRR was estimated. Again, convergence probability generally declined for all EMs when median natural mortality or a SRR was estimated. lower observation uncertainty.

When convergence is defined by a gradient threshold, the primary factor explaining deviance reduction is the same that for Hessian-based convergence for all OM process error sources, but there are some differences in deviance reduction for secondary factors (Table S6), and probability of convergence, overall, was lower (Figure S2). We found a wide range of maximum absolute values of gradients for models that ~~converged~~ had invertible Hessians (Figure S3). The largest value observed for a given EM and OM combination was typically $< 10^{-3}$, but many converged models had values greater than 1. For many OMs, EMs that assumed the correct process error ~~type~~ source and did not estimate median natural mortality or the Beverton-Holt SRR produced the lowest gradient values.

AIC performance

~~Marginal AIC accurately determined~~

Process error source

For AIC selection of the correct process error ~~assumptions in EMs when data were generated from R and~~ configuration, the magnitude of observation and process error variation were the

attributes that resulted in the largest percent reductions in deviance across OM process error sources other than R OMs (Table 2). Both sources of variation explained large reductions in deviance for R+S OMs, regardless of whether median natural mortality or (17-22%) and R+Sel (8-26%) OMs, whereas variance of process errors provided the major reductions for R+M (>9%) and R+q (>13%) OMs. Comparatively, none of the OM or EM attributes explained particularly large reductions in deviance for R OMs, but fishing history, whether a SRR was estimated (Figure S4, A). Attempting to estimate, and whether median natural mortality or a SRR separately had a negligible effect on the accuracy of determining the correct process error assumption. When both were estimated, there was a noticeable reduction in accuracy when OMs had a constant fishing pressure, low observation error, and larger variability in recruitment process errors. was known or estimated provided similar and the largest reductions (approximately 5-6%). Inclusion of second and third order interactions, did not provide large reductions in deviance for any of the OM process error sources.

For all OM process error sources other than R OMs, the attributes defining the primary branches of classification trees matched those that provided the largest reductions in deviance (Figure 2). Across all OMs, AIC was more accurate for the process error source when process error variability was greater and when observation error was lower. For R+M OMs, marginal AIC only accurately determined the correct process error model and correlation structure when observation error was low and variability in natural mortality process errors was high (Figures S4, B). Of these OMs, estimating the median natural mortality rate only reduced the accuracy of AIC when natural mortality process errors were independent and fishing pressure was constant. For OMs with poor model selection accuracy, AIC most frequently selected EMs with process errors in catchability (R+q) or selectivity (R+Sel). Selection of R+S EMs was generally unlikely.

Marginal AIC most accurately determined the correct source of process error and correlation structure for R+Sel OMs with low observation error (Figures S4, C). When there was low variability in selectivity process errors and high observation error, R+q or R+S EMs were

~~more likely to have the best AIC. Whether median natural mortality or SRRs were estimated appeared to have little effect on the performance of AIC.~~

~~Marginal AIC most accurately determined the correct source of process error and S OMs, there was a tendency to select R OMs when observation error was higher and apparent survival variation was lower ($\sigma_{2+} = 0.25$), but accuracy for the process error source was otherwise highly accurate. Larger variability of process error relative to observation error was also required for accurate identification of the correct correlation structure for R+q OMs with high variability in catchability process errors (Figures S4,D). The R-M, R+q, and R+q OMs with low variability in catchability process errors and high observation error had the least model selection accuracy. However, for these OMs, the marginal AIC accurately determined the correct source of process error (but not correlation structure) except when Sel OMs (Figure S4). No branches were estimated for classification trees fit to the R OMs assumed a-, likely because accuracy was high across all simulations (0.94), although inspection of the fine-scale results shows there is some degradation in AIC selection when an SRR and median natural mortality rate are estimated for R OMs with constant fishing pressure and EMs estimated both median natural mortality and the SRR high observation error (Figure S4, top left).~~

~~AIC performance for the stock-recruit-~~

Stock-recruit relationship

~~Our comparisons of model performance conditioned on assuming the true process error configuration is known (EM and OM process error types match) and we focus on results where the EMs assume median natural mortality is known because there was little difference in results when the EMs estimated this parameter. Broadly, we found generally poor accuracy of AIC in selecting models assuming a Logistic regressions for AIC selection of the Beverton-Holt SRR over the null model without an SRR for all OMs SRR, showed OM fishing history~~

and $\log SD_{SSB}$ provided substantial reductions in deviance for R+M ($>13\%$), R+Sel ($>26\%$), and R+q ($>24\%$) OM (Table 3). For R OM, fishing history provided the largest reduction in deviance ($>9\%$), whereas none of the attributes individually provided large reductions in deviance for R+S OM (all $<5\%$). However, ~~we also found increased accuracy of AIC in determining the Beverton-Holt SRR when the simulated population exhibited greater variation in spawning biomass for nearly every OM (Figure S19).~~

~~With inclusion of all attributes provided larger reductions in deviance than the sum of individual contributions for both R ($>30\%$) and R+S ($\sim 19\%$) OM. Further fits for R and R+S process error assumptions, probability of lowest AIC for the B-H SRR as a function of SSB variability were greatest for OM with contrast in fishing pressure and lower process variability in recruitment (Figure S19, A). The largest variation in SSB occurred in OM with larger recruitment variability ($\sigma_R = 1.5$; Figure S19, A, right column group), but the same high AIC accuracy was achieved for OM with lower recruitment variability at lower levels of SSB variation. The level of observation error had little effect on AIC accuracy.~~ OM that including different combinations of two factors additively showed fits that included $\log SD_{SSB}$ and recruitment variation only provided essentially the same reduction in deviance as the models with all factors. For all OM process error sources, inclusion of interaction terms provided relatively little reduction in residual deviance.

~~For R+M OM, probability of lowest AIC for the Beverton-Holt SRR increased steeply with variation in SSB whether it was induced by~~ Attributes defining the primary branches of classification trees for AIC selection of the SRR assumption were the same as those explaining the largest reductions in deviance for the logistic regression models (Figure 3). All branches based on $\log SD_{SSB}$ showed better accuracy with larger variability in SSB and all branches based on fishing history showed better accuracy when there was contrast in fishing ~~or variation in natural mortality process error.~~ (Figure S19, B). There was little difference in AIC accuracy whether the natural mortality process errors were correlated and, similar to pressure. Branches based on OM observation error or recruitment variability (R

~~and R+S OMs, there was also little effect due to level of observation error.~~

~~For R+Sel OMs,) showed better accuracy when they were lower. For R OMs, a combination of lower recruitment variability, contrast in fishing pressure over time was the primary source of variation in SSB and these are the OMs where AIC accuracy for the Beverton-Holt SRR was greatest (Figure S19, C). There was little effect of variability or correlation of selectivity process errors or the level of observation error on AIC accuracy.~~

~~Like the, and higher SSB variability produced AIC accuracy over 0.8. For R+S OMs, lower recruitment variability and observation error and higher SSB variability produced AIC accuracy of 0.79. For R+Sel OMs, the greatest accuracy for AIC in selecting the Beverton-Holt SRR occurred for M, R+Sel, and R+q OMs where there was contrast in fishing pressure over time which is also where there was the greatest variation in SSB (Figure S19, D). There was also little effect of variability or correlation of catchability process errors or the level of observation error on AIC accuracy, accuracy of 0.87 to 0.94 was observed with just increased SSB variability.~~

Bias

Terminal year spawning stock biomass, fishing mortality, and recruitment

~~For R OMs ($\sigma_{2+} = 0$), there was no indication of bias (95% confidence intervals included 0) in terminal year SSB. Regression models for log-scale errors in SSB that included the various OM and EM factors showed little reduction in deviance (<5%) for any of the estimating models regardless of process error assumptions, except when no SR assumption was made, recruitment variability was low, and there was contrast in fishing mortality and high observation error (Figure S10, A). However, errors in terminal SSB estimates were highly variable when factors across all OM process error sources (Table 4). The attributes producing the largest reductions were the EM assumption for median natural mortality was~~

estimated and there was constant fishing pressure and high observation error (Figure S10, A, second row).

For (known or estimated) for R, R+S OMs, the EMs with matching process error assumptions generally produced unbiased estimation of terminal SSB except when median natural mortality was estimated and there was high observation error. In M, R+S OMs with low observation error, EMs with incorrect process error assumptions typically provided biased estimation of terminal year SSB. Estimating the Beverton-Holt SRR had little discernible effect on bias of terminal year SSB estimation whereas estimating median M tended to produce more variability in errors in terminal SSB estimation similar to ROMs.

For RSel, and R+M OMs with low variability in natural mortality process errors, low observation error and contrast in fishing mortality over time all EMs produced low variability in SSB estimation error that indicated unbiasedness (Figure S10, B, third row). However, larger variability in natural mortality process errors increased bias of EMs without the correct process error type. Estimating median natural mortality increased variability of SSB estimation error particularly for OMs with high observation error and constant fishing pressure over time. It also increased bias in SSB estimation for many Rq OMs (1-3%). EM process error assumption for R+M OMs. Like Rand RS OMs (4%) and fishing history for all OM process error sources (1-5%). Including second order interactions provided the largest reductions in residual deviance (10- 26%). Including third order interactions also provided further reductions for R, R+S OMs, estimating a SRR had little discernible effect on SSB bias.

For, and R+Sel OMs, there was no evidence of bias for any EMs when variability in selectivity process error and observation error was low, and q OMs between 5 and 11%.

In all regression trees, branches based on fishing history and level of observation error generally showed less bias in SSB with contrast in fishing mortality (Figure S10, C). The largest bias occurred for any EMs that estimated median natural mortality when the OMs

had high observation error, constant fishing pressure, and greater variability in selectivity process errors ($\sigma_{\text{Sel}} = 0.5$) or low selectivity process errors ($\sigma_{\text{Sel}} = 0.1$) and low observation error. However, there was no evidence of SSB bias for correctly specified R+Sel EMs when observation error was low and variation in selectivity process errors was larger, whether median natural mortality was estimated or not (Figure S10, C, third row). We only observed an effect of estimating the Beverton-Holt SRR for R+Sel OMs that had high observation error and contrast in fishing pressure where estimating the SRR produced less biased SSB estimation for many EMs (Figure S10, C, top row).

All EMs fit to data from R+q OMs with low observation error and contrast in fishing pressure exhibited little evidence of bias in terminal SSB estimation except for R+M EMs when there was no AR1 correlation in catchability process errors (Figure S10, D). Many EMs also performed well in and lower observation error (Figure 4). For scenarios where there was bias, it was generally positive (over-estimation). For branches based on treatment of median natural mortality rate, bias was generally less when it was known rather than estimated. For some R+q OMs with low observation error, but no contrast in fishing pressure. For Sel and R+q OMs with high observation error and contrast in fishing pressure, EMs that estimated the Beverton-Holt SRR exhibited less SSB bias than those that did not. Estimating median natural mortality in the EMs only resulted in much more variable SSB estimation errors when there was no contrast in fishing pressure (Figure S10, D, first and third rows), less bias in SSB was shown when the EM process error assumption was correct.

For all OM process error types, relative errors in terminal year recruitment were generally more variable than SSB, but effects of R and R+S Results for bias in fishing mortality and recruitment generally matched those for SSB, except that directions of bias for fishing mortality were opposite to those for SSB and recruitment. Effects of individual OM and EM attributes on bias (i.e., negative or positive or none) were similar (Figure S12, A). Furthermore, for EM configurations where bias in terminal SSB was evident, median relative errors in recruitment often indicated stronger bias in recruitment of the same sign. factors on

regression models were similarly small as measured by reduction in deviance (Tables S7 and S8). Factors defining the primary branches of regression trees were in most cases identical to those for SSB (Figures S5 and S6).

Stock-recruit parameters

~~Across all OMs, there was generally less bias and (or) lower variability in estimation of the Beverton-Holt a parameter than the~~ Regression models for log-scale errors of estimates of both the Beverton-Holt a and b parameter. ~~In Rand R~~ parameters showed none of the factors explained large percent reductions in deviance (Table 5). The OM fishing history provided the largest deviance reduction for most OM process error sources for both parameters, but reductions were generally less than 6%. Exceptions were the R+S OMs, EMs with the correct assumptions about process errors provided the least biased estimation of Beverton-Holt SRR parameters when there was a change in fishing pressure over time and lower variability of recruitment process errors, but there was little effect of estimating median natural mortality and a small increase in bias for those OMs that had high observation error (Figure S7, A). For other R and Sel OMs where a and b were reduced by approximately 11% and 8%, respectively, and the R+S OMs, estimating natural mortality often resulted in less biased estimation of SRR parameters. There was generally large variability in relative errors of the SRR-q OMs where b was reduced by 10%. The EM process error assumption provided similar reductions in deviance for both parameters for R OMs. Including interactions also did not produce important reductions in deviance.

For regression trees of log-scale errors in Beverton-Holt a and b parameter estimates, ~~but the lowest variability occurred with low variability in recruitment and little or no variability in survival process errors ($\sigma_{2+} \in \{0, 0.25\}$), and contrast in fishing pressure.~~

~~In R+M OMs, the most accurate estimation of SRR parameters for all EM process error assumptions occurred when there was a change in fishing pressure, greater variability in~~

~~natural mortality process errors, and lower observation error (Figure S7, B). Relative to the R, less bias was indicated with contrast in OM fishing pressure for all branches in trees for each OM process error source (Figures 5 and 6). For all branches based on recruitment variability in trees for R and R+S OMs, there was even less effect of estimating median natural mortality on estimation bias for the SRR parameters.~~

~~Bias for SRR parameters was large and variability in relative errors was greatest for most EMs fit to R+Sel OMs with constant fishing pressure (Figure S7, C). Less bias in parameter estimation occurred for OMs with a change less bias in both a and b was observed with less recruitment variability. For R OMs with contrast in fishing pressure and the best accuracy occurred for those OMs that had low observation error and more variable and uncorrelated selectivity process errors, and when the EMs had with the correct process error assumption. There was little effect of estimating natural mortality on relative errors for SRR parameters.~~

~~Like R greater recruitment variability EMs that assumed the incorrect R+Sel OMs, relative errors in SRR parameters for R+q OMs were more accurate for most EM process error types when OMs had contrast in fishing pressure and lower observation error (Figure S7, D). However, the best accuracy occurred for those OMs that had M process errors produced less bias in both a and b than other process error assumptions. Across all combinations of OM and EM attributes, some bias was observed for both parameters, but there was generally less bias and(or) lower variability in catchability process errors. The worst accuracy of SRR parameter estimation regardless of EM type occurred when R+q OMs had low observation error and constant fishing pressure estimation of the a parameter than the b parameter (Figure S7, D, fourth row).~~

Median natural mortality rate

~~Across all OMs and EMs there was little effect of estimating SRRs on the bias in estimation~~

of Fitted regression models for log-scale errors in median natural mortality (Figure S13).
Median natural mortality rate was estimated accurately by all rate showed largest percent
reductions in residual deviance for R+S and R+M models (Table 6). The largest reductions
for a single attribute was the EM process error types for all R OMs except those with
high observation error and constant fishing pressure, in which case relative errors were high
(Figure S13, A, $\sigma_{2+} = 0$). For assumption ($>20\%$) and fishing history ($>15\%$) for R+S
OMs estimation of median natural mortality rate was most accurate when observation error
was low and there was contrast in fishing pressure and the EM process error type was correct.
For, Fishing history also provided $>10\%$ reduction for R+M OMs, median natural mortality
was estimated most accurately, regardless of EM process error type, when OMs had a change
in fishing pressure and low observation error (Figure S13, B). However, those but reductions
for all factors in R, R+M OMs that also had greatest variability in AR1 correlated natural
mortality process errors only had unbiased estimation when the EM process error type was
correct. Sel, and R+q OMs were relatively low ($<6\%$). Interactions of OM and EM factors
also provided substantial further reductions for R+S and R+M OMs (between 8 and 15%
for second order interactions).
All EM process error types accurately estimated Regression trees with branches based on
fishing history showed less bias in median natural mortality rate for R+Sel OMs that had
with contrast in fishing pressure, low observation error, and low selectivity process error
variability (Figure S13, C). When selectivity process error variability increased, the incorrect
EM process errors produce more biased estimation of median natural mortality rate. The
least accurate estimation occurred for all EM process error types when observation error was
high and fishing pressure was constant.
Like and branches based on level of observation error showed less bias with more precise
observations (Figure 7). For R OMs, branches based on EM process error assumption showed

less bias with EMs assuming the correct R and the incorrect R+S assumption. For R+Sel OMs, all EM process error types produced accurate estimation of median natural mortality rate when fit to S and R+q OMs with contrast in fishing pressure, low observation error and low catchability process error variability (Figure S13, D). Most M OMs, branches based on EM process error showed only the correct EM process error types produced biased estimation of median natural mortality when R+q OMs had high observaiton error and constant fishing pressure assumption with less bias.

Mohn's ρ

Regression models for Mohn's ρ for SSB was small in absolute value for all R and R+S OMs, regardless of EM process error types, and whether median natural mortality rate or SRRs were estimated (Figure S14, A) of SSB showed little reduction in deviance for any of the OM an EM attributes (<2%; Table 7). The strongest retrospective patterns (highest absolute lack of explanatory power is also reflected in the regression trees where median Mohn's ρ values) occurred in OMs with the largest apparent survival process error variability are near zero unless a large combinations of OM and EM conditions occur (Figure 8). For example, in R+S OMs, with constant fishing pressure, high observation error, and contrast in fishing pressure, but only for EMs with the incorrect process error type and where median natural mortality rate was assumed known (median ρ was approximately -0.15). For higher apparent survival process error, EMs that assume R+M , R+Sel, and R+q OMs, process errors have a median Mohn's $\rho = -0.068$.

Similarly, poor explanatory power of the OM and EM attributes occurred when we fit regression models for Mohn's ρ was also small in absolute value, but median values were all closer to 0 than the largest values in the R and R+S OMs (Figure S14, B-D). For these OMs, there was no noticeable effect of estimation of median natural mortality rate or SRRs on of fishing mortality and recruitment (Tables S9 and S10). Regression trees for Mohn's ρ

for any EM process error types.

of fishing mortality were similar to those for SSB in that median values of Mohn's ρ for recruitment was small in absolute value for all R OMs with low variability in recruitment process errors, regardless of EM process error type, and whether median natural mortality rate or SRRs were estimated (Figure S16, A) were close to zero for most combinations of OM and EM attributes (Figure S8). However, R and R+S OMs with greater recruitment process variability and higher observation error had we observed median Mohn's ρ for recruitment greater than zero for most EMs even when the EM process error type was correct. In R+S OMs with lower 0.1 at branches much closer to the base of the trees with fewer interactions of the OM and EM attributes (Figure S9). These branches with consistently large retrospective patterns were typically defined by larger OM observation error, EMs with the correct process error type exhibited better median Mohn's ρ close to 0 than EMs with the incorrect process error type. For R+M, R+Sel, and R+q OMs, results for OM constant fishing pressure, or incorrect EM process error configuration. Comparing regression model and regression tree fits, attributes defining the primary branches for all regression trees of all Mohn's ρ for recruitment are similar to those for SSB, but the range in median values and variation in Mohn's ρ values for a given OM are generally larger for recruitment (Figure S16, B-D) values (SSB, fishing mortality, and recruitment) generally matched those that explained the largest reductions in deviance.

Discussion

Assessing convergence

Analyses of model convergence across simulations can be useful for understanding the utility of alternative convergence criteria used in applications to real data for directing the practitioner to more appropriate random effects configurations. It is common during the

~~assessment model fitting process to check that the maximum absolute gradient component is less than some threshold prior to inspecting the Hessian of the optimized likelihood for invertibility (Carvalho et al. 2021). However, there is no accepted standard for the gradient threshold (e.g., Lee et al. 2011; Hurtado-Ferro et al. 2014; Rudd and Thorson 2018) and some thresholds would exclude models that in fact have an invertible Hessian. We found the Hessian at the optimized log-likelihood can often be invertible when the maximum absolute gradient was much larger than what would be perceived to be a sensible threshold.~~

Poor convergence was common in our results when the incorrect process error source was assumed. Li et al. (2024) found that convergence ~~rate~~ could be a useful diagnostic especially for separating the correct ~~model~~ simpler process error assumption from overly complex models. ~~However, the criteria for convergence used in their study may also lead to limited ability to distinguish the correct model from overly simplistic models, a pattern that was also noted by Liljestrand et al. (2024) in which one process error may absorb all sources of process error when the magnitude of other process errors are low.~~

~~Often poor convergence result when~~ Poor convergence often occurs when parameter estimates are at their bounds (Carvalho et al. 2021), ~~and this also applies to variance parameters for random effects with state-space assessment models. Even.~~ However, even when the Hessian is invertible for a converged model, parameters that are poorly informed will have extremely large variance estimates. This further inspection can lead to a more appropriate and often more parsimonious model configuration where the problematic parameters are not estimated. For example, process error variance parameters in state-space models that are estimated close to 0 indicates that the random effects are estimated to have little or no variability and removing these process errors is warranted. ~~Generally, our results suggest we can expect lower probability of convergence of state-space assessment models when estimating natural mortality or SRRs because of the difficulty distinguishing these parameters from others being estimated in assessment model with data that are typically available.~~ Our experiments did not aim to emulate the practitioner decision process in

~~developing model configurations (e.g. removing a source of process error and refitting the~~
~~model when process error variance parameters were estimated close to 0).~~ Evaluating the
determining an appropriate model configurations, but evaluating the efficacy of such a deci-
sion process when applying EMs might be important in closed loop simulations ~~(e.g. MSE)~~
aimed at quantifying management performance (e.g., a management strategy evaluation).

It is common during the assessment model fitting process to check that the maximum
absolute gradient component is less than some threshold prior to inspecting the Hessian
of the optimized likelihood for invertibility (Carvalho et al. 2021), but we found reliance on
magnitude of the gradient values for fitted models as a convergence criterion questionable.
There is no accepted standard for the gradient threshold (e.g., Lee et al. 2011; Hurtado-Ferro
et al. 2014; Rudd and Thorson 2018), but the Hessian at the optimized log-likelihood was
often invertible when the maximum absolute gradient was much larger than what might be
perceived to be a sensible threshold in some of our simulations. Therefore, the gradient
criterion could exclude models that in fact have an invertible Hessian.

A factor affecting the convergence criteria, particularly for maximum likelihood estimation
of models with random effects, is numerical accuracy. All optimizations performed in these
simulations are of the Laplace approximation of the marginal likelihood and, therefore, gra-
dients and Hessians are also with respect to this approximation (see TMB::sdreport in the
Template Model Builder package). Functionality within the Template Model Builder pack-
age exists (i.e., TMB::checkConsistency) to check the validity of the Laplace approximation
and the utility of this as a diagnostic for state-space assessment models should be explored
further. Furthermore, numerical methods are used to calculate and invert the Hessian for
variance estimation for models with random effects. ~~Along with our results,~~ Our results,
along with the potential lack of accuracy imposed by these approximations, ~~suggests~~ suggest
at least investigating whether the Hessian is positive definite when the calculated absolute
gradients are not terribly large (e.g, < 1).

816 Configuring process error

817 ~~Of the OM process error configurations we considered, we found AIC to be accurate for~~
818 ~~selecting models with process errors on recruitment and apparent survival (Rand~~ We found
819 accuracy of marginal AIC-based selection for the correct process error source required only
820 low observation error for R, R+S). ~~Fitting models to other OMs rarely preferred, R+S~~
821 ~~EMsSel, and Rand +q OMs. R+S EMs were nearly always selected for the matching OMs;~~
822 ~~a similar result was reported by Liljestrand~~ M OMs further required higher process error
823 variability, but this also improved accuracy for the other OM process errors sources when
824 there was higher observation error. These results seem consistent with Li et al. (2024). ~~For~~
825 ~~other sources of process error, Their simulation studies investigated models with multiple~~
826 process error sources and found good ~~accuracy of AIC was improved when there was larger~~
827 ~~variability in the process errors and/or lower observation error in detecting correct process~~
828 error assumptions for simulations based on two stocks (Gulf of Maine cod and Southern
829 New England-Mid-Atlantic yellowtail founder) that are well sampled by NEUS bottom trawl
830 surveys used as indices in the respective assessments and poor accuracy for Atlantic mackerel,
831 a semi-pelagic species that is observed relatively poorly.

832 ~~Across all OM process error configurations, AIC performed poorly in identifying that the~~
833 ~~presence of~~

834 Stock recruitment relationships

835 Variation in SSB was the most important factor for using marginal AIC to correctly
836 distinguish the Beverton-Holt SRR ~~in the OM unless there was contrast in fishing pressure~~
837 ~~possibly in combination with other factors such as lower variability in recruitment process~~
838 ~~errors (in R from the null model without an SRR. For R+M, R+Sel, and R+S models) or~~
839 ~~greater variation in natural mortality process errors (for R+M OMs, Fig. S19).~~ As such,
840 properly accounting for process error in natural mortality could be important (Li et al.

~~2024)~~ when evaluating SRRs in state-space models. Curiously, ~~we did not find a marked~~
~~effect of the level of observation error on ability to~~ OM, the SRR was accurately detected
when the CV of SSB over the time series was at least 40 to ~~detect the SRR, but it is~~
possible that AIC would perform better if observations have even lower uncertainty than
~~we considered~~ 50% ($\log \text{SD}_{\text{SSB}} = -0.9$ to -0.7) regardless of any other OM or EM attributes.
Detection of the SRR for R and R+S OM, the SRR was accurately detected, but
this lower level ($\sigma_R = 0.5$) was assumed for all of the other OM process error source and
represents the lower range of estimates from recent NEUS stock assessments. Our results
assumed that the EM process error configuration was correct, but this may not be a strong
limitation given the ability of AIC to distinguish the process error source in many scenarios.
Although we did not compare models with alternative SRRs (e.g., Ricker ~~and vs.~~ Beverton-
Holt), we do not expect AIC to perform any better distinguishing between relationships and
may be more difficult than distinguishing from the null model even with larger variability in
SSB. Our finding that AIC tended to choose simpler recruitment models in ~~most~~ many cases
contrasts with the noted bias in AIC for more complex models (Shibata 1976; Katz 1981;
Kass and Raftery 1995); ~~but, whereas those~~. However, these earlier findings apply to the
much more common comparison of models that are fit to raw and independent observations,
~~here we are comparing whereas our comparisons of~~ state-space models ~~which~~ account for
observation error and separately estimate process errors in latent variables.
Our results comport with those of de Valpine and Hastings (2002) who found AIC could not
distinguish among state-space SRRs that were fit just to SSB and recruitment observations
(i.e., not within an assessment model). Similarly, Britten et al. (In review) found AIC
could not reliably distinguish the Beverton-Holt SRR from no SRR, nor identify alternative
environmental effects on SRR parameters. However, Miller et al. (2016) did find AIC to
prefer ~~a~~ an SRR with environmental effects when applied to data for the ~~SNEMA~~ Southern
New England-Mid-Atlantic yellowtail flounder stock and AIC also selected an environmental
covariate on ~~a~~ an SRR for the most recent stock assessment of Georges Bank yellowtail

flounder (NEFSC 2025). Both of these yellowtail flounder stocks have large changes in stock size and the values of environmental covariates over time. Additionally, this species is well-observed by the bottom trawl survey that is used for an index in assessment models.

~~As expected, bias in all parameters and assessment output was generally improved with lower observation error.~~ Estimation of SRR parameters was only moderately reliable in ideal scenarios of low observation error and contrast in fishing for ~~some~~ R+Sel and R+M OMs ~~; but generally with large temporal variability in process errors.~~ Otherwise, SRR parameter estimation was biased and(or) highly variable. We found substantial bias in estimated SRR parameters in R and R+S OMs particularly with high variability in recruitment and apparent survival process errors, suggesting that practitioners should be cautious with SRR inferences when fitted assessment models have these properties. We only evaluated effects of SSB variability on accuracy of AIC in identifying the SRR, but those results suggests we might find less bias for the SRR parameters in such cases as well. Another condition that could improve perception of bias in our simulation studies is restricting results to fits that converged with Hessian-based standard errors for all parameters, but Britten et al. (In review) did not find less SRR parameter bias when restricting estimates using a gradient-based criterion. A simulation study by Stock and Miller (2021) examining configurations of environmental covariate effects on a Beverton-Holt SRR for the previously mentioned, well-observed, Southern New England-Mid-Atlantic yellowtail flounder stock found little or no bias for the density-independent mortality parameter a , but still biased estimation of the density-dependent parameter b .

~~On the other hand, estimation of~~

Estimating assessment model quantities

As expected, bias in parameters, SSB, and other assessment output was generally improved with lower observation error. Estimation of median natural mortality was reliable in many

OM scenarios with contrast in fishing pressure, consistent with Hoenig et al. (2025). ~~In some OMs, when EMs estimated the SRR parameters and median natural mortality, bias for those parameters was improved. Conversely, for some R+Sel and R+q OMs where there was bias in natural mortality due to high observation error, estimating the SRR reduced the bias in median natural mortality rate. However, estimating median natural mortality did cause~~ However, we found poor accuracy in ~~SSB estimation~~ terminal SSB estimation when estimating median natural mortality in many OMs ~~without~~ when there was no contrast in fishing pressure over time and ~~with~~ higher observation error. ~~Thus~~ Therefore, estimating median natural mortality should be approached with caution in state-space assessment models, particularly given its significant impact on determination of reference point and stock status (Li et al. 2024).

Negligible retrospective patterns

Incorrect EM process error assumptions did not produce strong retrospective patterns for SSB for any OMs regardless of whether median natural mortality or ~~a~~ an SRR was estimated, ~~but some weak retrospective patterns occur~~ although some weak patterns occurred when observation error was high and there was contrast in fishing pressure. However, retrospective patterns tended to be more variable for recruitment and were sometimes large even when the EM was correct. Therefore, we recommend ~~emphasis~~ de-emphasis on inspection of ~~retrospective patterns primarily for SSB and F~~ patterns for recruitment, but further research on retrospective patterns in other assessment model parameters, management quantities such as biological reference points, and projections may be beneficial (Brooks and Legault 2016). The general lack of retrospective patterns with mis-specified process errors is perhaps to be expected. Retrospective patterns are often induced in simulation studies by rapid changes in a quantity such as index catchability, natural mortality, or perceived catch during years toward the end of the time series (Legault 2009; Miller and Legault 2017; Huynh et al. 2022;

Breivik et al. 2023). In our simulations, the process errors changing over time may have trends in ~~particular~~certain simulations, particularly when strong autocorrelation is imposed, but the random effects have no trend on average across simulations. Szuwalski et al. (2018) and ~~lietal24~~Li et al. (2024) also found relatively small retrospective patterns when the source of mis-specification was temporal variation in ~~demography~~demographic attributes. Indeed, it is common for the flexibility provided by temporal random effects to reduce retrospective patterns (Miller et al. 2018; Stock et al. 2021; Stock and Miller 2021), though it does not necessarily indicate a more accurate assessment model (Perretti et al. 2020; Li et al. 2024; Liljestrand et al. 2024). Our results together with the existing literature seem to suggest that when a strong retrospective pattern is observed in an assessment it is more likely to be due to a mis-specification of a rapid shift in some model attribute rather than whether a particular process is assumed to be randomly varying temporally.

Summarization approach

~~Our simulation study examined the importance of several factors for reliable inferences from state-space age-structured assessment models. Contrast in fishing pressure was consistently an important factor across all~~We found the use of regression models and classification and regression trees extremely useful in understanding the most important OM and EM attributes explaining variation in the measures of reliability we examined ~~. AIC accurately distinguished models with process errors on recruitment only (R) or on recruitment and apparent survival (R+S). Accuracy for other process error types required a strong signal (high process variability) with low noise (low observation uncertainty). Therefore, we expect practitioners will find R+S configurations to provide satisfactory diagnostics across a range of life history and data quality scenarios. AIC generally performed poorly for selecting the SRR, but performance was improved with~~across all simulations. The classification and regression trees are generally a good tool for determining the OM and EM attributes that produce better

or worse measures of reliability. However, determining the combination of attributes that produce the best or worst measures of reliability can be challenging using the trees alone. For example, in the regression tree for median natural mortality rate estimates in R OMs (Figure 7), both of the first branches imply bias is low regardless of OM fishing history, but when OM fishing pressure is constant, results are much better when OM observation error is low (median RE about -6%) than when OM observation error is high (median RE about 40%). The default pruning of the trees can exclude these lower branches. However, inspection of deviance explained by various regression models shows the ~9% reduction in residual deviance by including second order interaction of all OM and EM factors (Table 6), indicating that the interaction of factors may be important, thereby complimenting the regression tree analysis. Higher order interactions of some factors could also provide reductions in deviance and, therefore, inspection of results for each combinations of OM and EM factors, as provided in the Supplementary Materials, can also be important.

Recommendations and conclusions

Our findings regarding model convergence suggests practitioners using state-space models and maximum marginal likelihood for estimation should not heavily weight the magnitude of the gradient values in determining convergence as long as the maximum absolute value is around 1 or lower. Instead, positive-definiteness of the Hessian of the minimized negative log-likelihood should be evaluated.

Unfortunately, whether the practitioner includes a Beverton-Holt SRR will often depend on biological plausibility of this particular SRR because using AIC to determine its validity required a combination of low recruitment variability and, contrast in fishing pressure. Some large variation in SSB over time, and lower observation error, which applies to a limited number of managed stocks. Furthermore, some bias in estimation in at least one of the SRR parameters existed in nearly all OM-EM combinations (and MSY-based reference points

968 should be expected. Because bias in terminal SSB and retrospective patterns were indifferent
969 to whether or not the SRR was estimated, ~~and convergence was slightly better~~ the prevalence
970 of bias in SRR parameter estimation, and often better convergence without the SRR, we
971 recommend a sensible default ~~would be to fit models without an assumed SRR~~ is to exclude
972 an SRR when fitting assessment models, as also suggested by Brooks (2024).

973 We found marginal AIC can, in many cases, accurately distinguished models with process
974 errors. We saw the best accuracy for models with process errors on recruitment only
975 (R), recruitment and apparent survival (R+S), and recruitment and selectivity (R+Sel),
976 especially with lower observation error. However, AIC could also distinguish R+M and R+q
977 process errors when variability of those processes was greater. The R+S assumption for
978 process errors is common in applications of WHAM in the NEUS and the SAM assessment
979 framework (Nielsen and Berg 2014) in ICES, and we can have some confidence that
980 practitioners are correctly arriving at this assumption over other sources of process error
981 using marginal AIC.

982 Acknowledgements

983 This work was funded by NOAA Fisheries Northeast Fisheries Science Center. We thank
984 Jon Deroba, two anonymous reviewers, and the associate editor for helpful comments on ~~an~~
985 ~~earlier version~~ earlier versions of this manuscript that markedly improved its clarity.

References

- Aeberhard, W.H., Flemming, J.M., and Nielsen, A. 2018. Review of State-Space Models for Fisheries Science. *Annual Review of Statistics and Its Application* **5**(1): 215–235. doi:10.1146/annurev-statistics-031017-100427.
- Auger-Méthé, M., Field, C., Albertsen, C.M., Derocher, A.E., Lewis, M.A., Jonsen, I.D., and Mills Flemming, J. 2016. State-space models’ dirty little secrets: Even simple linear Gaussian models can have estimation problems. *Scientific reports* **6**(1): 26677. doi:10.1038/srep26677.
- Auger-Méthé, M., Newman, K., Cole, D., Empacher, F., Gryba, R., King, A.A., Leos-Barajas, V., Mills Flemming, J., Nielsen, A., Petris, G., and others. 2021. A guide to state-space modeling of ecological time series. *Ecological Monographs* **91**(4): e01470. doi:10.1002/ecm.1470.
- Breiman, L., Friedman, J.H., Olshen, R.A., and Stone, C.J. 1984. Classification and regression trees. Chapman; Hall/CRC, New York, NY USA. doi:10.1201/9781315139470.
- Breivik, O.N., Aldrin, M., Fuglebakk, E., and Nielsen, A. 2023. Detecting significant retrospective patterns in state space fish stock assessment. *Canadian Journal of Fisheries and Aquatic Sciences* **80**(9): 1509–1518. doi:10.1139/cjfas-2022-0250.
- Britten, G., Brooks, E.N., and Miller, T.J. In review. Identification and performance of environmentally-driven stock-recruitment relationships in state space assessment models. *Canadian Journal of Fisheries and Aquatic Sciences*.
- Brooks, E.N. 2024. Pragmatic approaches to modeling recruitment in fisheries stock assessment: A perspective. *Fisheries Research* **270**: 106896. doi:10.1016/j.fishres.2023.106896.
- Brooks, E.N., and Legault, C.M. 2016. Retrospective forecasting – evaluating performance of stock projections for New England groundfish stocks. *Canadian Journal of Fisheries and Aquatic Sciences* **73**(6): 935–950. doi:10.1139/cjfas-2015-0163.
- Cadigan, N.G. 2016. A state-space stock assessment model for northern cod, including under-

reported catches and variable natural mortality rates. Canadian Journal of Fisheries and Aquatic Sciences **73**(2): 296–308. doi:10.1139/cjfas-2015-0047.

Carvalho, F., Winker, H., Courtney, D., Kapur, M., Kell, L., Cardinale, M., Schirripa, M., Kitakado, T., Yemane, D., Piner, K.R., Maunder, M.N., Taylor, I., Wetzel, C.R., Doering, K., Johnson, K.F., and Methot, R.D. 2021. A cookbook for using model diagnostics in integrated stock assessments. Fisheries Research **240**: 105959. doi:https://doi.org/10.1016/j.fishres.2021.105959.

Clopper, C.J., and Pearson, E.S. 1934. The use of confidence or fiducial limits illustrated in the case of the binomial. Biometrika **26**(4): 404–413. doi:10.1093/biomet/26.4.404.

Collier, Z.K., Zhang, H., and Soyoye, O. 2022. Alternative methods for interpreting Monte Carlo experiments. Communications in Statistics - Simulation and Computation: 1–16. doi:10.1080/03610918.2022.2082474.

Conn, P.B., Williams, E.H., and Shertzer, K.W. 2010. When can we reliably estimate the productivity of fish stocks? Canadian Journal of Fisheries and Aquatic Sciences **67**(3): 511–523. doi:10.1139/F09-194.

Cronin-Fine, L., and Punt, A.E. 2021. Modeling time-varying selectivity in size-structured assessment models. Fisheries Research **239**: 105927. Elsevier.

de Valpine, P., and Hastings, A. 2002. Fitting population models incorporating process noise and observation error. Ecological Monographs **72**(1): 57–76.

Fisch, N., Shertzer, K., Camp, E., Maunder, M., and Ahrens, R. 2023. Process and sampling variance within fisheries stock assessment models: Estimability, likelihood choice, and the consequences of incorrect specification. ICES Journal of Marine Science **80**(8): 2125–2149. doi:10.1093/icesjms/fsad138.

Fleischman, S.J., Catalano, M.J., Clark, R.A., and Bernard, D.R. 2013. An age-structured state-space stock–recruit model for Pacific salmon (*Oncorhynchus spp.*). Canadian Journal of Fisheries and Aquatic Sciences **70**(3): 401–414. doi:10.1139/cjfas-2012-0112.

Gonzalez, O., O’Rourke, H.P., Wurpts, I.C., and Grimm, K.J. 2018. Analyzing Monte Carlo

- simulation studies with classification and regression trees. *Structural Equation Modeling: A Multidisciplinary Journal* **25**(3): 403–413. doi:10.1080/10705511.2017.1369353.
- Harwell, M., Kohli, N., and Peralta-Torres, Y. 2018. A survey of reporting practices of computer simulation studies in statistical research. *The American Statistician* **72**(4): 321–327. doi:10.1080/00031305.2017.1342692.
- Hoenig, J.M., Hearn, W.S., Leigh, G.M., and Latour, R.J. 2025. Principles for estimating natural mortality rate. *Fisheries Research* **281**: 107195. doi:10.1016/j.fishres.2024.107195.
- Hoyle, S.D., Maunder, M.N., Punt, A.E., Mace, P.M., Devine, J.A., and A’mar, Z.T. 2022. Preface: Developing the next generation of stock assessment software. *Fisheries Research* **246**: 106176. doi:10.1016/j.fishres.2021.106176.
- Hurtado-Ferro, F., Szuwalski, C.S., Valero, J.L., Anderson, S.C., Cunningham, C.J., Johnson, K.F., Licandeo, R., McGilliard, C.R., Monnahan, C.C., Muradian, M.L., Ono, K., Vert-Pre, K.A., Whitten, A.R., and Punt, A.E. 2014. Looking in the rear-view mirror: Bias and retrospective patterns in integrated, age-structured stock assessment models. *ICES Journal of Marine Science* **72**(1): 99–110. doi:10.1093/icesjms/fsu198.
- Huynh, Q.C., Legault, C.M., Hordyk, A.R., and Carruthers, T.R. 2022. A closed-loop simulation framework and indicator approach for evaluating impacts of retrospective patterns in stock assessments. *ICES Journal of Marine Science* **79**(7): 2003–2016. doi:10.1093/icesjms/fsac066.
- Johnson, K.F., Councill, E., Thorson, J.T., Brooks, E., Methot, R.D., and Punt, A.E. 2016. Can autocorrelated recruitment be estimated using integrated assessment models and how does it affect population forecasts? *Fisheries Research* **183**: 222–232. doi:10.1016/j.fishres.2016.06.004.
- Kapur, M.S., Ducharme-Barth, N., Oshima, M., and Carvalho, F. 2025. Good practices, trade-offs, and precautions for model diagnostics in integrated stock assessments. *Fisheries Research* **281**: 107206. doi:10.1016/j.fishres.2024.107206.
- Kass, R.E., and Raftery, A.E. 1995. Bayes factors. *Journal of the American Statistical*

Association **90**(430): 773–795. doi:10.1080/01621459.1995.10476572.

Katz, R.W. 1981. On some criteria for estimating the order of a Markov chain. *Technometrics* **23**(3): 243–249. doi:10.1080/00401706.1981.10486293.

Knape, J. 2008. Estimability of density dependence in models of time series data. *Ecology* **89**(11): 2994–3000. doi:10.1890/08-0071.1.

Kristensen, K., Nielsen, A., Berg, C.W., Skaug, H., and Bell, B.M. 2016. TMB: Automatic differentiation and Laplace approximation. *Journal of Statistical Software* **70**(5): 1–21. doi:10.18637/jss.v070.i05.

Lee, H.-H., Maunder, M.N., Piner, K.R., and Methot, R.D. 2011. Estimating natural mortality within a fisheries stock assessment model: An evaluation using simulation analysis based on twelve stock assessments. *Fisheries Research* **109**(1): 89–94. doi:10.1016/j.fishres.2011.01.021.

Legault, C.M. 2009. Report of the retrospective working group, 14-16 january 2008. US Department of Commerce Northeast Fisheries Science Center Reference Document 09-01. US Department of Commerce Northeast Fisheries Science Center. Woods Hole, MA.

Legault, C.M., and Palmer, M.C. 2016. In what direction should the fishing mortality target change when natural mortality increases within an assessment? *Canadian Journal of Fisheries and Aquatic Sciences* **73**(3): 349–357. doi:10.1139/cjfas-2015-0232.

Legault, C.M., and Restrepo, V.R. 1999. A flexible forward age-structured assessment program. *Col. Vol. Sci. Pap. ICCAT* **49**(2): 246–253.

Legault, C.M., Wiedenmann, J., Deroba, J.J., Fay, G., Miller, T.J., Brooks, E.N., Bell, R.J., Langan, J.A., Cournane, J.M., Jones, A.W., and Muffley, B. 2023. Data-rich but model-resistant: An evaluation of data-limited methods to manage fisheries with failed age-based stock assessments. *Canadian Journal of Fisheries and Aquatic Sciences* **80**(1): 27–42. doi:10.1139/cjfas-2022-0045.

Li, C., Deroba, J.J., Berger, A.M., Goethel, D.R., Langseth, B.J., Schueller, A.M., and

- Miller, T.J. 2025a. Random effects on numbers-at-age transitions implicitly account for movement dynamics and improve performance within a state-space stock assessment. *Canadian Journal of Fisheries and Aquatic Sciences* **82**. doi:10.1139/cjfas-2025-0092.
- Li, C., Deroba, J.J., Miller, T.J., Legault, C.M., and Perretti, C. 2025b. Guidance on bias-correction of log-normal random effects and observations in state-space assessment models. *Canadian Journal of Fisheries and Aquatic Sciences*. doi:10.1139/cjfas-2025-0093.
- Li, C., Deroba, J.J., Miller, T.J., Legault, C.M., and Perretti, C.T. 2024. An evaluation of common stock assessment diagnostic tools for choosing among state-space models with multiple random effects processes. *Fisheries Research* **273**: 106968. doi:10.1016/j.fishres.2024.106968.
- Liljestrand, E.M., Bence, J.R., and Deroba, J.J. 2024. The effect of process variability and data quality on performance of a state-space stock assessment model. *Fisheries Research* **275**: 107023. doi:10.1016/j.fishres.2024.107023.
- MacKinnon, D.P., Warsi, G., and Dwyer, J.H. 1995. A simulation study of mediated effect measures. *Multivariate Behavioral Research* **30**(1): 41–62. doi:10.1207/s15327906mbr3001__3.
- Magnusson, A., and Hilborn, R. 2007. What makes fisheries data informative? *Fish and Fisheries* **8**(4): 337–358. doi:10.1111/j.1467-2979.2007.00258.x.
- Methot, R.D., and Wetzel, C.R. 2013. Stock synthesis: A biological and statistical framework for fish stock assessment and fishery management. *Fisheries Research* **142**: 86–99. doi:10.1016/j.fishres.2012.10.012.
- Miller, T.J., and Brooks, E.N. 2021. Steepness is a slippery slope. *Fish and Fisheries* **22**(3): 634–645. doi:10.1111/faf.12534.
- Miller, T.J., Curti, K.L., and Hansell, A.C. 2025. Space for WHAM: A multi-region, multi-stock generalization of the woods hole assessment model with an application to black sea bass. *Canadian Journal of Fisheries and Aquatic Sciences* **82**: 1–26. doi:10.1139/cjfas-2025-0097.

1120 Miller, T.J., Hare, J.A., and Alade, L. 2016. A state-space approach to incorporating envi-
 1121 ronmental effects on recruitment in an age-structured assessment model with an appli-
 1122 cation to Southern New England yellowtail flounder. *Canadian Journal of Fisheries and*
 1123 *Aquatic Sciences* **73**(8): 1261–1270. doi:10.1139/cjfas-2015-0339.

1124 Miller, T.J., and Hyun, S.-Y. 2018. Evaluating evidence for alternative natural mortality
 1125 and process error assumptions using a state-space, age-structured assessment model.
 1126 *Canadian Journal of Fisheries and Aquatic Sciences* **75**(5): 691–703. doi:10.1139/cjfas-
 1127 2017-0035.

1128 Miller, T.J., and Legault, C.M. 2017. Statistical behavior of retrospective patterns and
 1129 their effects on estimation of stock and harvest status. *Fisheries Research* **186**: 109–120.
 1130 doi:10.1016/j.fishres.2016.08.002.

1131 Miller, T.J., O’Brien, L., and Fratantoni, P.S. 2018. Temporal and environmental variation
 1132 in growth and maturity and effects on management reference points of Georges Bank
 1133 Atlantic cod. *Canadian Journal of Fisheries and Aquatic Sciences* **75**(12): 2159–2171.
 1134 doi:10.1139/cjfas-2017-0124.

1135 Mohn, R. 1999. The retrospective problem in sequential population analysis: An investi-
 1136 gation using cod fishery and simulated data. *ICES Journal of Marine Science* **56**(4):
 1137 473–488. doi:10.1006/jmsc.1999.0481.

1138 NEFSC. 2022a. Final report of the haddock research track assessment working group. Avail-
 1139 able at https://s3.us-east-1.amazonaws.com/nefmc.org/14b_EGB_Research_Track_Haddock_WG_Report.pdf.

1140 NEFSC. 2022b. Report of the American plaice research track working group. Available at
 1141 https://s3.us-east-1.amazonaws.com/nefmc.org/2_American-Plaice-WG-Report.pdf.

1142 NEFSC. 2024. Butterfish research track assessment report. US Dept Commer Northeast
 1143 Fish Sci Cent Ref Doc. 24-03; 191 p.

1144 NEFSC. 2025. Yellowtail flounder research track working group report. Available at
 1145 <https://d23h0vhsm26o6d.cloudfront.net/10c.-Yellowtail-Flounder-RT-WG-Report.pdf>.

1146 Nielsen, A., and Berg, C.W. 2014. Estimation of time-varying selectivity in stock assessments

- using state-space models. *Fisheries Research* **158**: 96–101. doi:10.1016/j.fishres.2014.01.014.
- Pedersen, M.W., and Berg, C.W. 2017. A stochastic surplus production model in continuous time. *Fish and Fisheries* **18**(2): 226–243. doi:10.1111/faf.12174.
- Perretti, C.T., Deroba, J.J., and Legault, C.M. 2020. Simulation testing methods for estimating misreported catch in a state-space stock assessment model. *ICES Journal of Marine Science* **77**(3): 911–920. doi:10.1093/icesjms/fsaa034.
- Polansky, L., De Valpine, P., Lloyd-Smith, J.O., and Getz, W.M. 2009. Likelihood ridges and multimodality in population growth rate models. *Ecology* **90**(8): 2313–2320. doi:10.1890/08-1461.1.
- Pontavice, H. du, Miller, T.J., Stock, B.C., Chen, Z., and Saba, V.S. 2022. Ocean model-based covariates improve a marine fish stock assessment when observations are limited. *ICES Journal of Marine Science* **79**(4): 1259–1273. doi:10.1093/icesjms/fsac050.
- Punt, A.E. 2023. Those who fail to learn from history are condemned to repeat it: A perspective on current stock assessment good practices and the consequences of not following them. *Fisheries Research* **261**: 106642. doi:10.1016/j.fishres.2023.106642.
- Punt, A.E., Hurtado-Ferro, F., and Whitten, A.R. 2014. Model selection for selectivity in fisheries stock assessments. *Fisheries Research* **158**: 124–134. doi:10.1016/j.fishres.2013.06.003.
- Rudd, M.B., and Thorson, J.T. 2018. Accounting for variable recruitment and fishing mortality in length-based stock assessments for data-limited fisheries. *Canadian Journal of Fisheries and Aquatic Sciences* **75**(7): 1019–1035. doi:10.1139/cjfas-2017-0143.
- Shibata, R. 1976. Selection of the order of an autoregressive model by Akaike’s information criterion. *Biometrika* **63**(1): 117–126. doi:10.1093/biomet/63.1.117.
- Stewart, I.J., and Monnahan, C.C. 2017. Implications of process error in selectivity for approaches to weighting compositional data in fisheries stock assessments. *Fisheries Research* **192**: 126–134. doi:10.1016/j.fishres.2016.06.018.
- Stock, B.C., and Miller, T.J. 2021. The Woods Hole Assessment Model (WHAM): A general state-space assessment framework that incorporates time- and age-varying processes via

random effects and links to environmental covariates. *Fisheries Research* **240**: 105967.
doi:10.1016/j.fishres.2021.105967.

Stock, B.C., Xu, H., Miller, T.J., Thorson, J.T., and Nye, J.A. 2021. Implementing two-dimensional autocorrelation in either survival or natural mortality improves a state-space assessment model for Southern New England-Mid Atlantic yellowtail flounder. *Fisheries Research* **237**: 105873. doi:10.1016/j.fishres.2021.105873.

Szuwalski, C.S., Ianelli, J.N., and Punt, A.E. 2018. Reducing retrospective patterns in stock assessment and impacts on management performance. *ICES Journal of Marine Science* **75**(2): 596–609. doi:10.1093/icesjms/fsx159.

Therneau, T., and Atkinson, B. 2025. *Rpart*: Recursive Partitioning and Regression Trees. Available from <https://github.com/bethatkinson/rpart>.

Thompson, W.R. 1936. On confidence ranges for the median and other expectation distributions for populations of unknown distribution form. *Annals of Mathematical Statistics* **7**(3): 122–128. doi:10.1214/aoms/1177732502.

Thorson, J.T., and Minto, C. 2015. Mixed effects: A unifying framework for statistical modelling in fisheries biology. *ICES Journal of Marine Science* **72**(5): 1245–1256. doi:10.1093/icesjms/fsu213.

Thulin, M. 2014. The cost of using exact confidence intervals for a binomial proportion. *Electronic Journal of Statistics* **8**(1): 817–840. doi:10.1214/14-EJS909.

Trijoulet, V., Fay, G., and Miller, T.J. 2020. Performance of a state-space multispecies model: What are the consequences of ignoring predation and process errors in stock assessments? *Journal of Applied Ecology* **57**(1): 121–135. doi:10.1111/1365-2664.13515.

Wang, S., Cadigan, N.G., and Benoît, H.P. 2017. Inference about regression parameters using highly stratified survey count data with over-dispersion and repeated measurements. *Journal of Applied Statistics* **44**(6): 1013–1030. doi:10.1080/02664763.2016.1191622.

Wiedenmann, J., Free, C.M., and Jensen, O.P. 2019. Evaluating the performance of data-limited methods for setting catch targets through application to data-rich stocks:

1201 A case study using northeast U.S. Fish stocks. *Fisheries Research* **209**(1): 129–142.
 1202 doi:10.1016/j.fishres.2018.09.018.

1203 Xu, H., Thorson, J.T., Methot, R.D., and Taylor, I.G. 2019. A new semi-parametric method
 1204 for autocorrelated age- and time-varying selectivity in age-structured assessment models.
 1205 *Canadian Journal of Fisheries and Aquatic Sciences* **76**(2): 268–285. doi:10.1139/cjfas-
 1206 2017-0446.

1207 Yee, T.W. 2008. The VGAM package. *R News* **8**(2): 28–39. Available from [https://journal.](https://journal.r-project.org/articles/RN-2008-014/)
 1208 [r-project.org/articles/RN-2008-014/](https://journal.r-project.org/articles/RN-2008-014/).

1209 Yee, T.W. 2015. *Vector generalized linear and additive models: With an implementation in*
 1210 *R*. Springer, New York, NY USA. doi:10.1007/978-1-4939-2818-7.

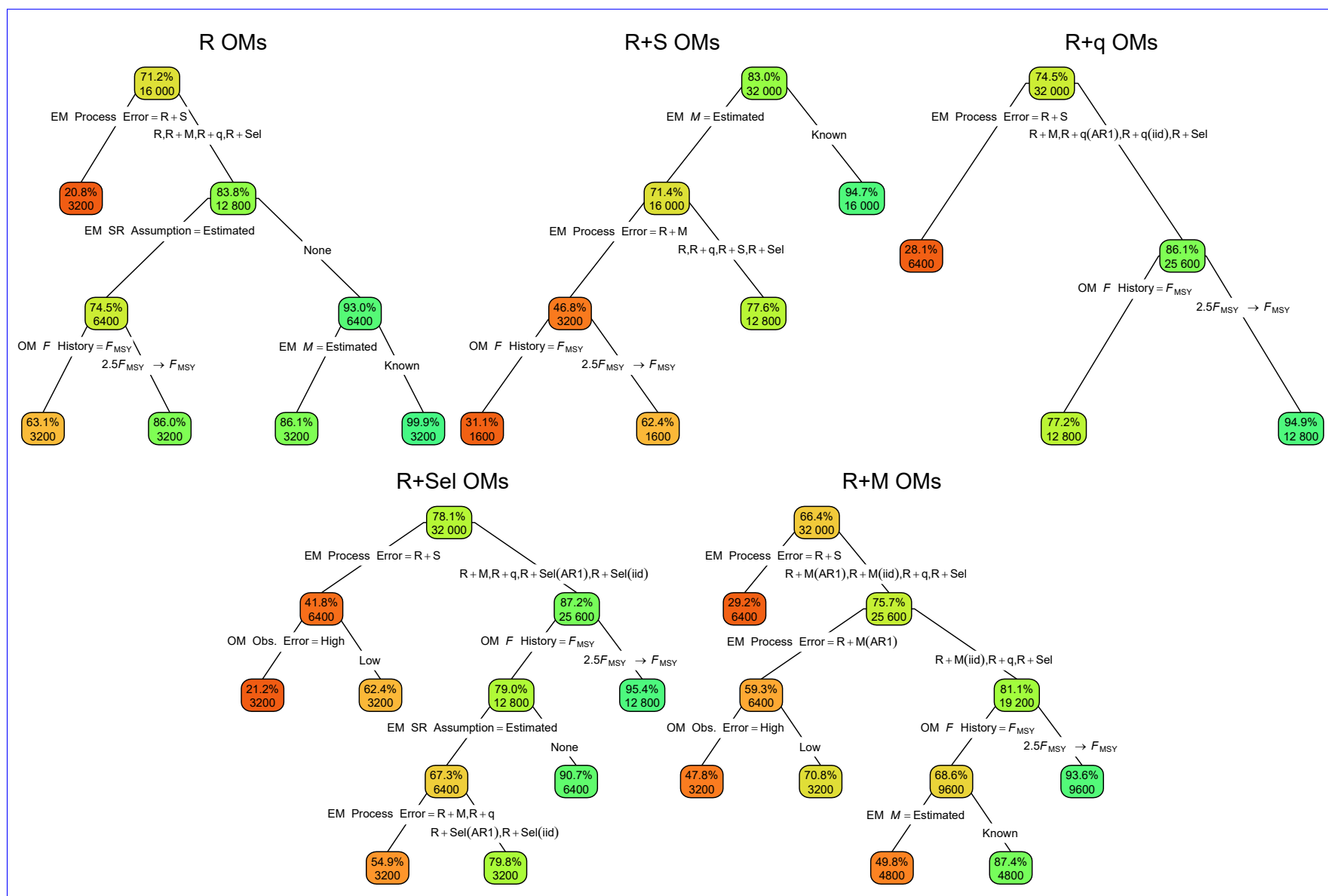


Fig. 1. Estimated probability of fits Classification trees indicating primary factors determining convergence as defined by providing hessian-based Hessian-based standard errors for EMs assuming alternative process error (colored points and lines) R, and median natural mortality (estimated or known) and Beverton-Holt stock-recruit relationships (estimated or not; along x-axis) when fitted to operating models that have R and R+S (A), R+Sel (B) M, R+M (C), or Sel and R+q OMs. Nodes denote percent convergence (Dtop) process error structures. Circled values indicate results where the EM process error structure matches that and number of fits (bottom) for the operating model and vertical lines represent 95% confidence intervals corresponding subset.

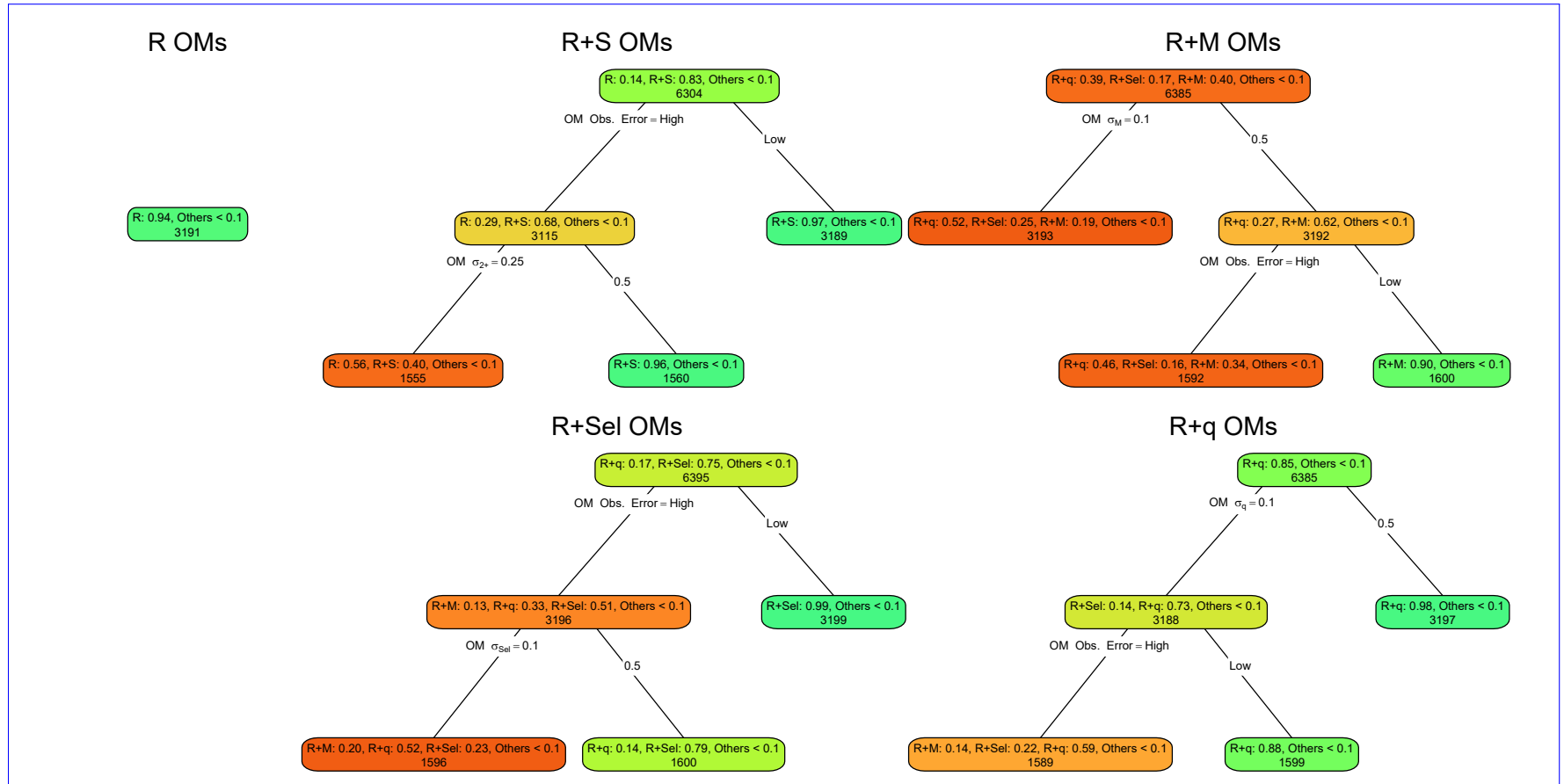


Fig. 2. Classification trees indicating primary factors determining which EM process error assumption provides the lowest AIC for R+S, R+M, R+Sel and R+q OMs. Each node shows the proportion of EM process error models with lowest AIC (top) and number of observations (bottom) for the corresponding subset. Lower or higher accuracy of the process error assumption are indicated by more red or green polygons, respectively.

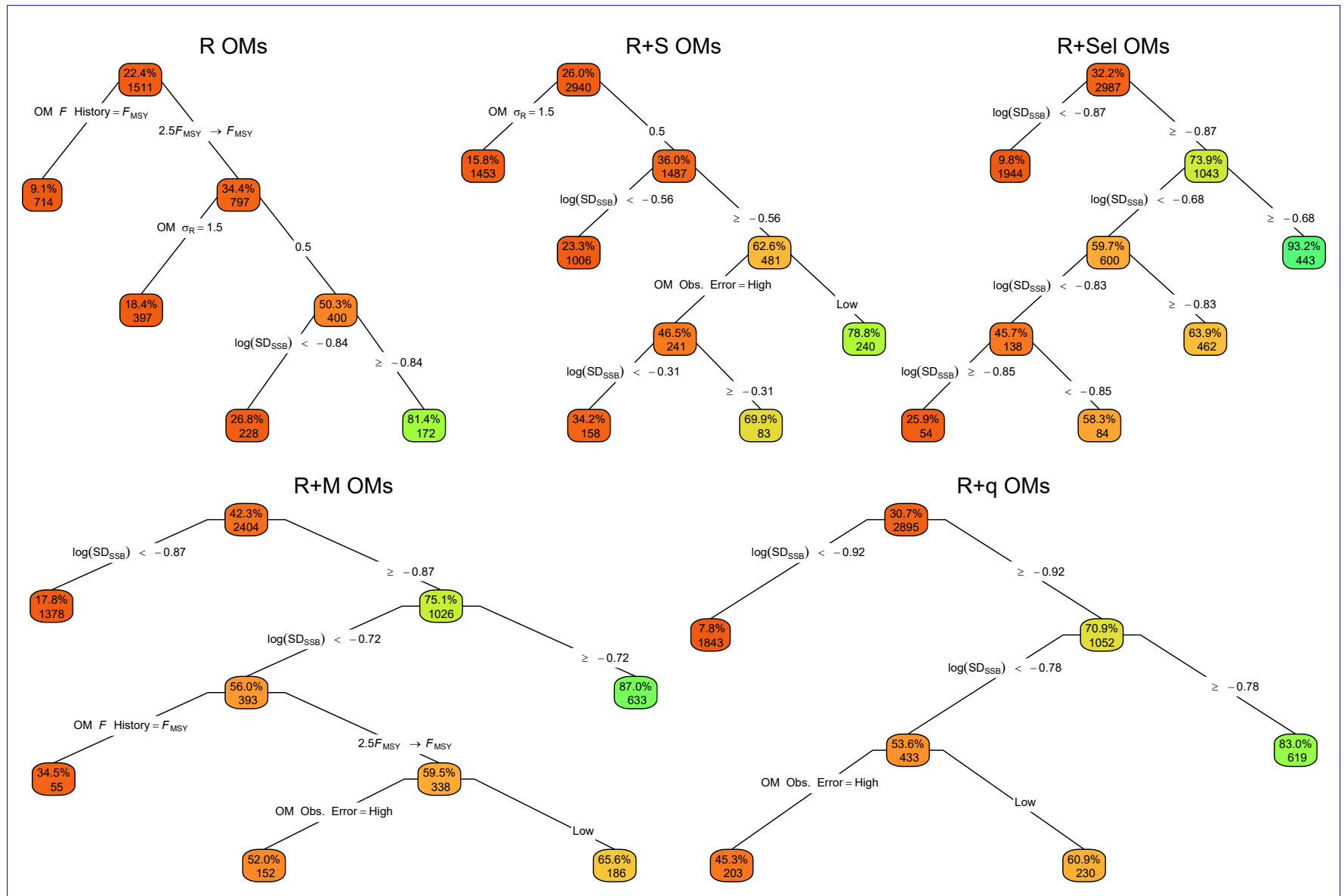


Fig. 3. Estimated probability of lowest AIC for EMs assuming alternative process error structures. Classification trees indicating primary factors determining which EM SRR assumption (colored bars) conditional on alternative assumptions for median natural mortality (estimated none or known) and Beverton-Holt stock-recruit relationships (estimated or not; along x-axis) when fitted to operating models that have provides the lowest AIC for Rand, R+S(A), R+Sel(B)M, R+M(C), or Sel and R+q (D) process error structuresOMs. Striped bars indicate results where All EMs assume the EM correct process error structure matches source. Nodes denote the percentage of EMs that assume the SRR with lowest AIC (top) and number of observations

Fig. 4. Regression trees indicating primary factors determining reductions in sums of squares of errors in estimation measured by Eq. 3 for terminal year SSB for R+S, R+M, R+Sel and R+q OMs. Each node shows the median error (top) and number of observations (bottom) for the corresponding subset. Median errors closer to or further from zero are indicated by more green or red polygons, respectively.

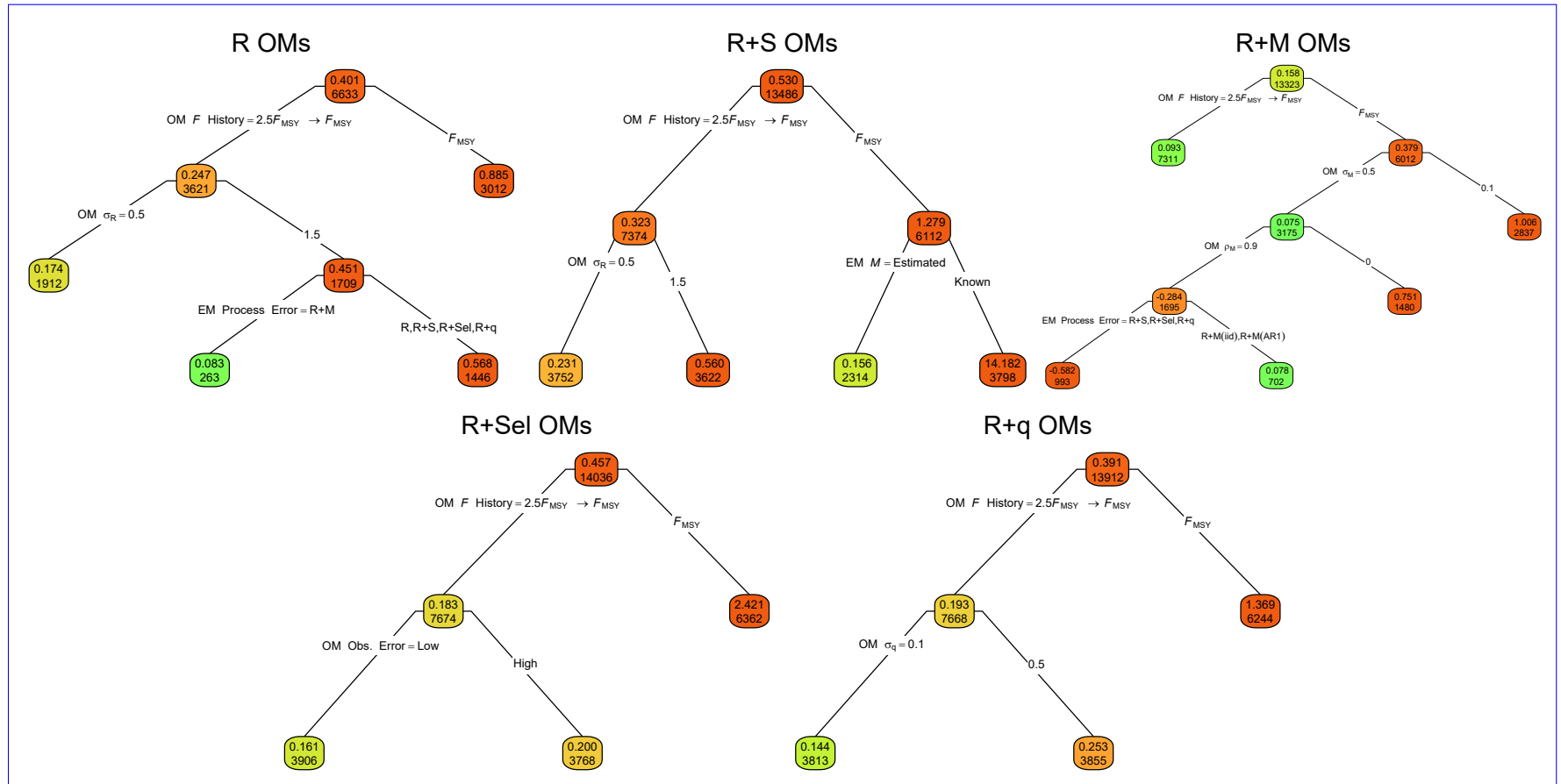


Fig. 5. Estimated probability-Regression trees indicating primary factors determining reductions in sums of lowest AIC from logistic regression on the log-standard deviation squares of the true log(SSB) errors in each simulation estimation measured by Eq. 3 for estimating model with Beverton-Holt stock-recruit relationships, rather than the otherwise equivalent EM without the stock-recruit relationship. Results are conditional on median M is known in the EM and alternative assumptions EMs having the correct process error structure: Beverton-Holt SRR parameter a for R and R+S (A), R+Sel (B)M, R+M (C), or Sel and R+q OMs. Each node shows the median error (Dtop) -, and median M is assumed known in the EM. Solid and dashed lines are for OMs with and without temporal contrast in fishing pressure, respectively, and polygons represent 95% confidence intervals. Range number of results indicates the range of log-standard deviation of log observations (SSBbottom) for simulations the corresponding subset. Lower or higher median absolute errors of the particular OM process error assumption are indicated by more green or red polygons, respectively.

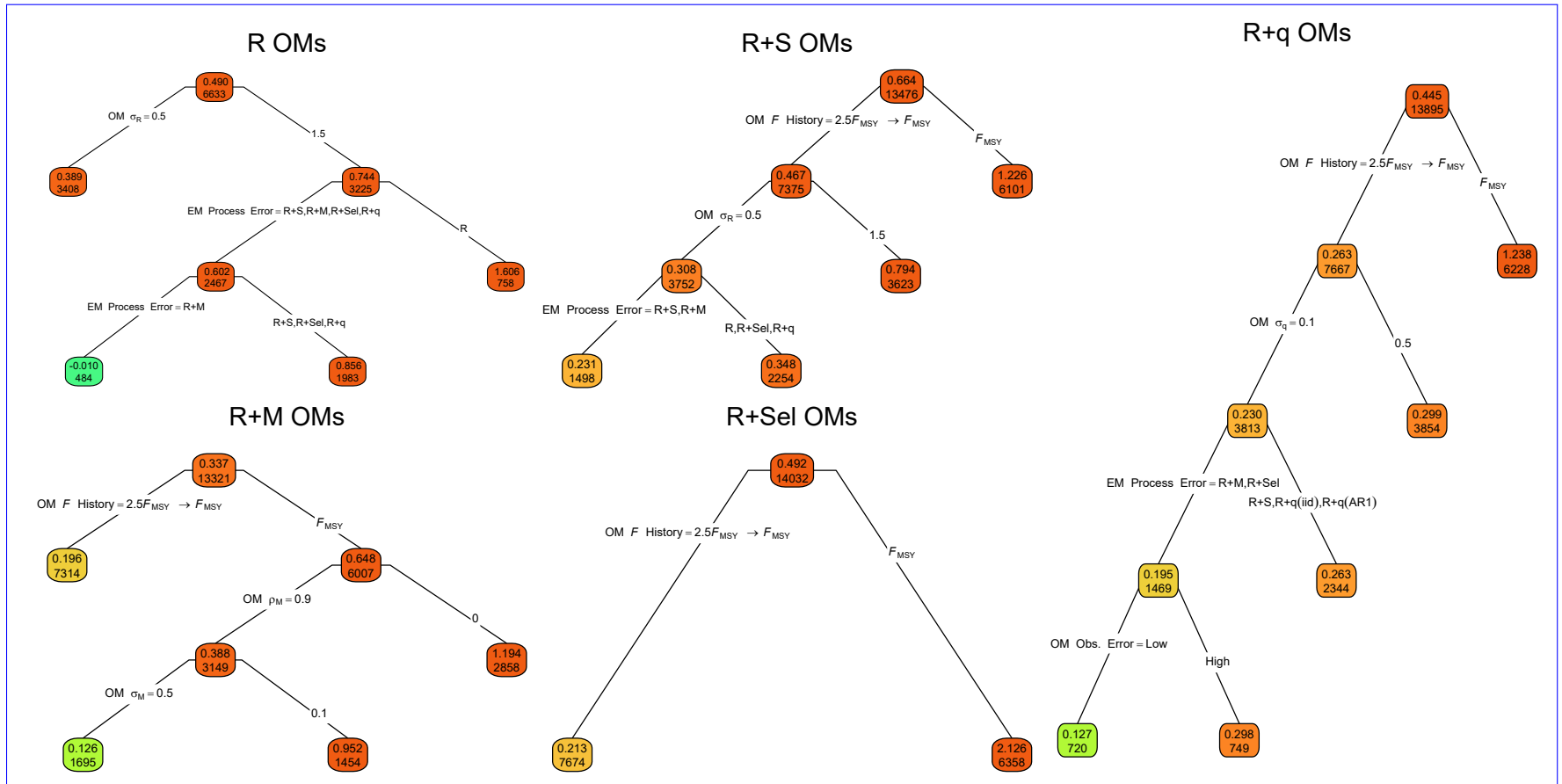


Fig. 6. Median relative error Regression trees indicating primary factors determining reductions in sums of terminal year SSB squares of errors in estimation measured by Eq. 3 for estimating models fitted to data sets simulated with alternative process error structures: the Beverton-Holt SRR parameter b for R and R+S (A), R+Sel (B) M, R+M (C), or Sel and R+q OMs. Each node shows the median error (top) and number of observations (bottom) for the corresponding subset. Circled values indicate results where Lower or higher median absolute errors of the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals assumption are indicated by more green or red polygons, respectively.

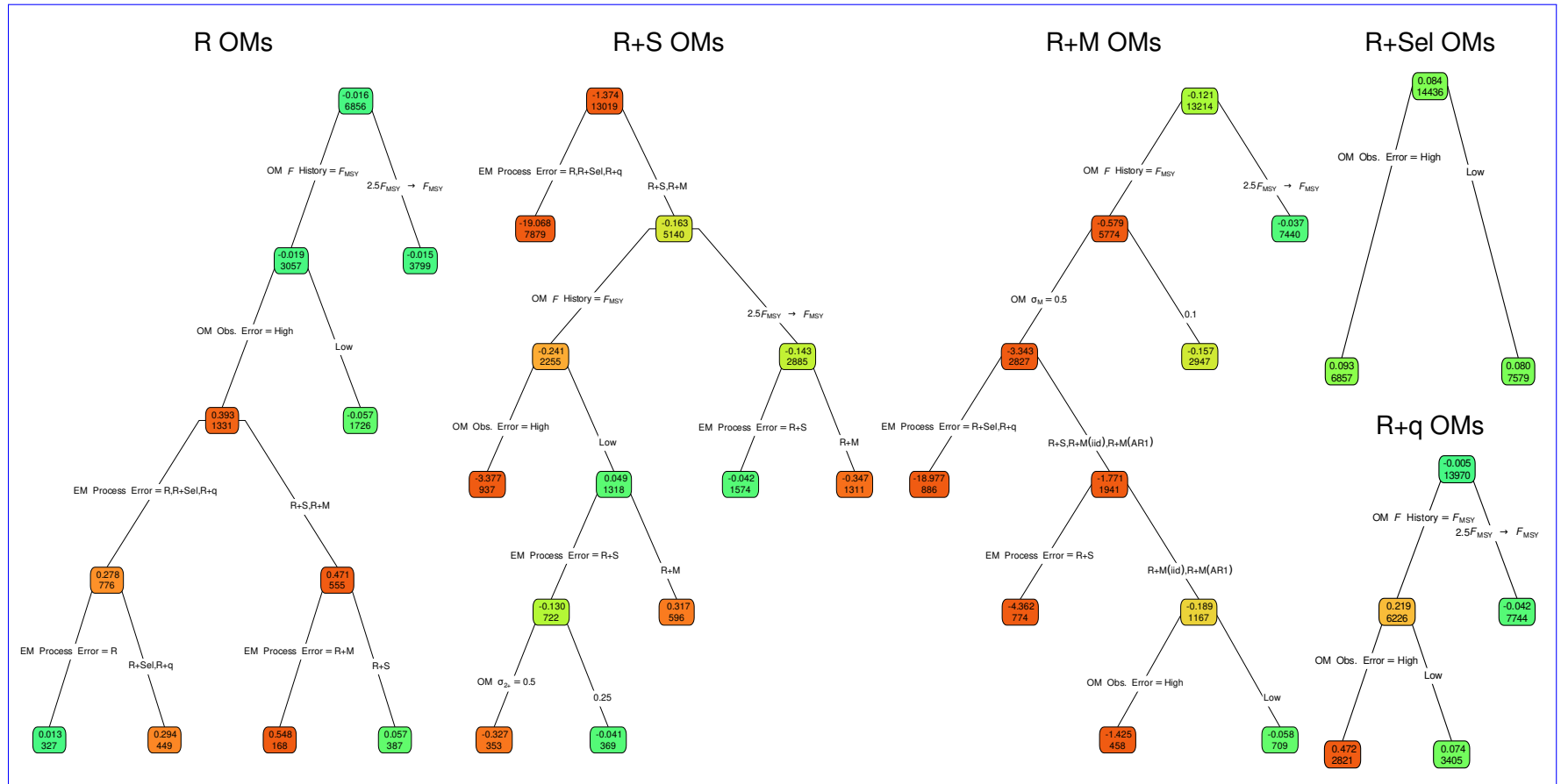


Fig. 7. Median Mohn's rho Regression trees indicating primary factors determining reductions in sums of squares of errors in estimation measured by Eq. 3 for SSB—the median natural mortality rate for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or Sel and R+q OMs. Each node shows the median error (Top) and number of observations (bottom) for the corresponding subset. Circled values indicate results where Lower or higher median absolute errors of the EM-process error structure matches that of the operating model and vertical lines represent 95% confidence intervals assumption are indicated by more green or red polygons, respectively.

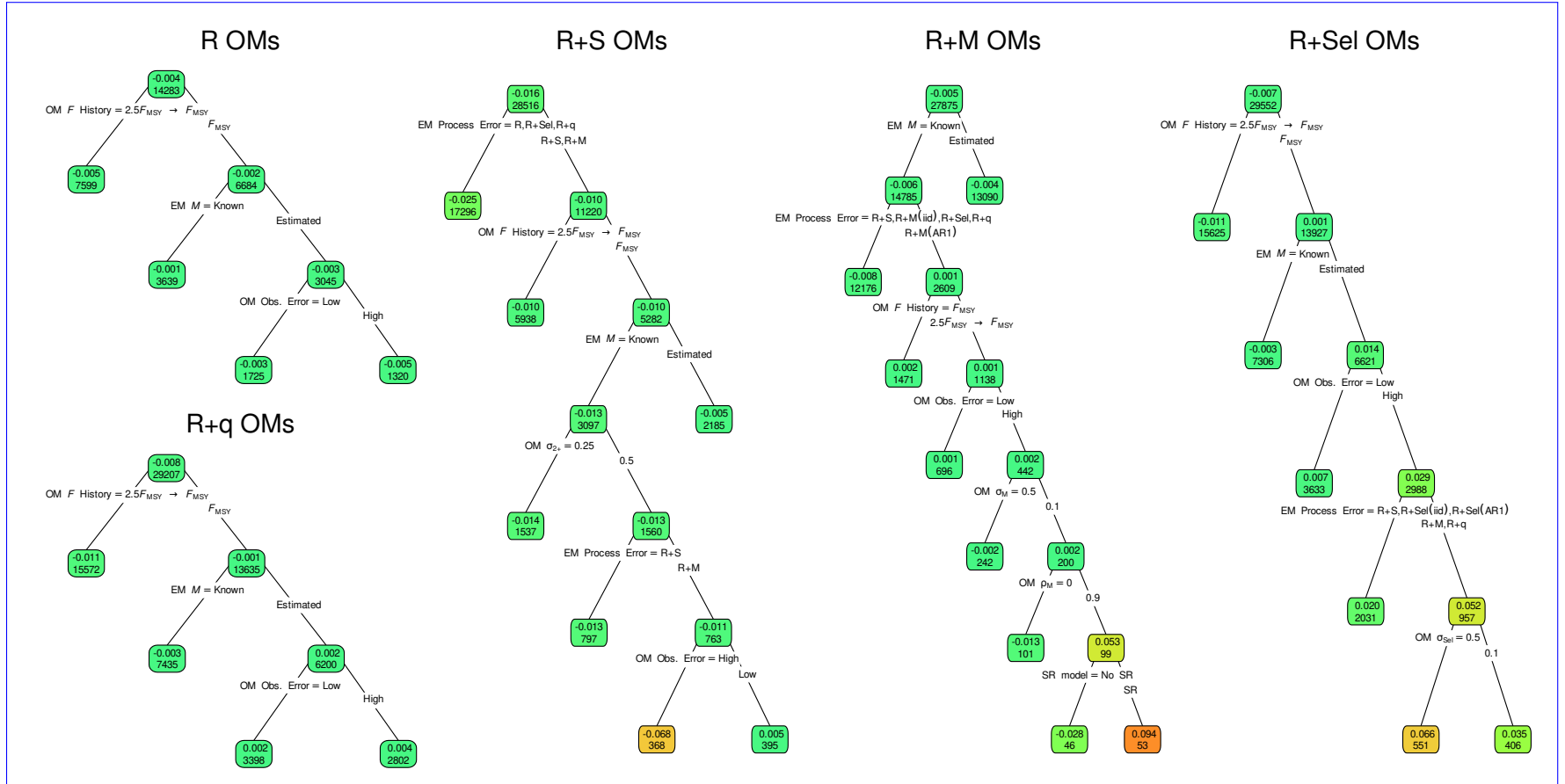


Fig. 8. Regression trees indicating primary factors determining reductions in sums of squares of errors in transformed Mohn's ρ (Eq. 3) for SSB for R+S, R+M, R+Sel and R+q OM. Each node shows the median Mohn's ρ (top) and number of observations (bottom) for the corresponding subset. Median Mohn's ρ closer to or further from zero are indicated by more green or red polygons, respectively.

Table 1. For each OM process error source (columns), percent reduction in deviance for logistic regression models fit to indicators of convergence (providing Hessian-based standard errors) with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM Process Error	27.95	4.58	14.68	17.24	24.66
EM M Assumption	1.07	11.43	2.45	0.56	1.46
EM SR Assumption	2.88	3.30	1.24	2.47	1.59
OM Obs. Error	0.75	4.64	2.06	4.54	1.60
OM F History	2.32	3.37	1.63	3.30	2.59
OM σ_R	0.10	0.02	—	—	—
OM σ_{2+}	—	0.40	—	—	—
OM σ_M	—	—	0.22	—	—
OM ρ_M	—	—	0.17	—	—
OM σ_{Sel}	—	—	—	1.81	—
OM ρ_{Sel}	—	—	—	0.02	—
OM σ_q	—	—	—	—	0.34
OM ρ_q	—	—	—	—	<0.01
All factors	39.54	31.46	24.85	34.83	36.31
+ All Two Way	45.03	39.89	35.20	42.81	43.70
+ All Three Way	47.02	44.57	37.88	45.51	46.87

Table 2. For each OM process error source (columns), percent reduction in deviance for multinomial logistic regression models fit to indicators of EM process error assumption with lowest AIC with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	5.52	1.05	0.52	0.61	1.32
EM SR Assumption	5.60	0.75	1.13	0.93	1.95
OM Obs. Error	2.96	22.46	3.42	25.67	5.03
OM F History	5.77	0.62	0.94	0.91	2.05
OM σ_R	0.10	0.66	—	—	—
OM σ_{2+}	—	16.86	—	—	—
OM σ_M	—	—	9.06	—	—
OM ρ_M	—	—	0.38	—	—
OM σ_{Sel}	—	—	—	7.59	—
OM ρ_{Sel}	—	—	—	0.60	—
OM σ_q	—	—	—	—	13.50
OM ρ_q	—	—	—	—	0.75
All factors	20.98	46.12	16.58	40.83	25.99
+ All Two Way	22.02	48.94	21.63	44.08	30.17
+ All Three Way	22.05	49.98	22.36	44.54	31.38

Table 3. For each OM process error source (columns), percent reduction in deviance for logistic regression models fit to indicators of EM SRR assumption (none or Beverton-Holt) with lowest AIC with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	0.04	0.21	0.18	0.02	0.01
OM Obs. Error	<0.01	0.65	0.14	0.04	0.02
OM F History	9.17	3.79	13.08	26.56	24.60
OM σ_R	3.54	4.74	—	—	—
OM σ_{2+}	—	0.14	—	—	—
OM σ_M	—	—	1.14	—	—
OM ρ_M	—	—	0.05	—	—
OM σ_{Sel}	—	—	—	0.02	—
OM ρ_{Sel}	—	—	—	0.17	—
OM σ_q	—	—	—	—	0.36
OM ρ_q	—	—	—	—	0.02
$\log(SD_{SSB})$	4.11	1.59	33.39	41.36	39.23
All factors	31.52	18.99	34.23	43.77	42.31
+ All Two Way	34.79	22.24	35.99	45.84	44.04
+ All Three Way	35.41	23.09	37.57	46.39	44.63

Table 4. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to errors in estimation measured by Eq. 3 for the terminal year SSB with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	2.28	1.15	1.04	2.92	3.26
EM SR assumption	0.10	0.06	0.08	0.06	0.08
EM Process Error	0.43	4.28	0.40	0.11	1.05
OM Obs. Error	1.63	0.07	0.78	0.32	<0.01
OM F History	2.62	3.15	1.28	3.22	4.72
OM σ_R	0.03	0.01	—	—	—
OM σ_{2+}	—	0.93	—	—	—
OM σ_M	—	—	0.18	—	—
OM ρ_M	—	—	0.01	—	—
OM σ_{Sel}	—	—	—	0.16	—
OM ρ_{Sel}	—	—	—	0.04	—
OM σ_q	—	—	—	—	1.02
OM ρ_q	—	—	—	—	0.06
All factors	7.59	9.86	3.93	7.04	10.64
+ All Two Way	17.99	25.56	10.06	13.44	22.43
+ All Three Way	23.39	36.74	13.76	16.55	31.11

Table 5. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to errors in estimation measured by Eq. 3 for the Beverton-Holt SRR parameters with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

<u>Factor</u>	<u>Beverton-Holt a</u>					<u>Beverton-Holt b</u>				
	<u>R</u>	<u>R+S</u>	<u>R+M</u>	<u>R+Sel</u>	<u>R+q</u>	<u>R</u>	<u>R+S</u>	<u>R+M</u>	<u>R+Sel</u>	<u>R+q</u>
<u>EM M Assumption</u>	<u>0.02</u>	<u>1.05</u>	<u>0.02</u>	<u>0.11</u>	<u>0.02</u>	<u>0.05</u>	<u>1.06</u>	<u>0.03</u>	<u>0.01</u>	<u>0.40</u>
<u>EM Process Error</u>	<u>2.74</u>	<u>0.18</u>	<u>0.20</u>	<u>1.25</u>	<u>1.90</u>	<u>2.29</u>	<u>1.21</u>	<u>0.12</u>	<u>1.40</u>	<u>3.06</u>
<u>OM Obs. Error</u>	<u>0.16</u>	<u><0.01</u>	<u>0.01</u>	<u>0.04</u>	<u><0.01</u>	<u><0.01</u>	<u>0.01</u>	<u>0.05</u>	<u>0.01</u>	<u>0.01</u>
<u>OM F History</u>	<u>3.15</u>	<u>3.34</u>	<u>5.60</u>	<u>11.37</u>	<u>10.00</u>	<u>1.16</u>	<u>1.17</u>	<u>2.01</u>	<u>7.97</u>	<u>3.87</u>
<u>OM σ_R</u>	<u>2.31</u>	<u>0.74</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>1.67</u>	<u>0.52</u>	<u>~</u>	<u>~</u>	<u>~</u>
<u>OM σ_{2+}</u>	<u>~</u>	<u>0.29</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.01</u>	<u>~</u>	<u>~</u>	<u>~</u>
<u>OM σ_M</u>	<u>~</u>	<u>~</u>	<u>0.30</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.13</u>	<u>~</u>	<u>~</u>
<u>OM ρ_M</u>	<u>~</u>	<u>~</u>	<u>0.51</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.22</u>	<u>~</u>	<u>~</u>
<u>OM σ_{Sel}</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.13</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.05</u>	<u>~</u>
<u>OM ρ_{Sel}</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.07</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.04</u>	<u>~</u>
<u>OM σ_q</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.04</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>0.10</u>
<u>OM ρ_q</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u><0.01</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u>~</u>	<u><0.01</u>
<u>All factors</u>	<u>8.07</u>	<u>5.15</u>	<u>6.73</u>	<u>12.64</u>	<u>11.79</u>	<u>4.91</u>	<u>3.75</u>	<u>2.55</u>	<u>9.12</u>	<u>7.22</u>
<u>+ All Two Way</u>	<u>9.96</u>	<u>7.37</u>	<u>9.76</u>	<u>13.59</u>	<u>13.65</u>	<u>7.55</u>	<u>7.15</u>	<u>4.32</u>	<u>10.08</u>	<u>12.16</u>
<u>+ All Three Way</u>	<u>11.22</u>	<u>8.15</u>	<u>11.13</u>	<u>14.48</u>	<u>14.87</u>	<u>9.78</u>	<u>9.02</u>	<u>5.26</u>	<u>11.08</u>	<u>14.73</u>

Table 6. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to errors in estimation measured by Eq. 3 for the median natural mortality rate parameter with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM SR assumption	0.21	0.38	0.11	0.26	0.43
EM Process Error	1.98	20.36	3.16	0.94	1.31
OM Obs. Error	4.74	0.79	0.40	2.23	1.88
OM F History	5.07	15.11	10.65	0.24	2.38
OM σ_R	≤ 0.01	0.01	—	—	—
OM σ_{2+}	—	5.04	—	—	—
OM σ_M	—	—	5.32	—	—
OM ρ_M	—	—	0.85	—	—
OM σ_{Sel}	—	—	—	1.30	—
OM ρ_{Sel}	—	—	—	0.37	—
OM σ_q	—	—	—	—	0.46
OM ρ_q	—	—	—	—	0.06
All factors	12.64	40.10	21.29	5.54	6.52
+ All Two Way	21.17	48.12	36.19	9.87	11.71
+ All Three Way	23.03	50.38	42.82	11.58	14.64

Table 7. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to transformed Mohn's ρ values for each simulation (Eq. 3) for SSB with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	0.79	0.18	0.15	0.95	1.24
EM SR assumption	<0.01	0.01	<0.01	<0.01	<0.01
EM Process Error	<0.01	0.22	0.14	0.08	0.04
OM Obs. Error	0.12	0.03	0.05	0.18	0.21
OM F History	0.84	0.14	0.07	1.08	1.56
OM σ_R	0.01	0.01	~	~	~
OM σ_{2+}	~	0.02	~	~	~
OM σ_M	~	~	0.01	~	~
OM ρ_M	~	~	<0.01	~	~
OM σ_{Sel}	~	~	~	0.01	~
OM ρ_{Sel}	~	~	~	0.02	~
OM σ_q	~	~	~	~	0.01
OM ρ_q	~	~	~	~	0.01
All factors	1.89	0.63	0.43	2.43	3.29
+ All Two Way	3.63	1.10	0.91	4.75	6.22
+ All Three Way	4.27	1.65	1.50	5.73	7.53

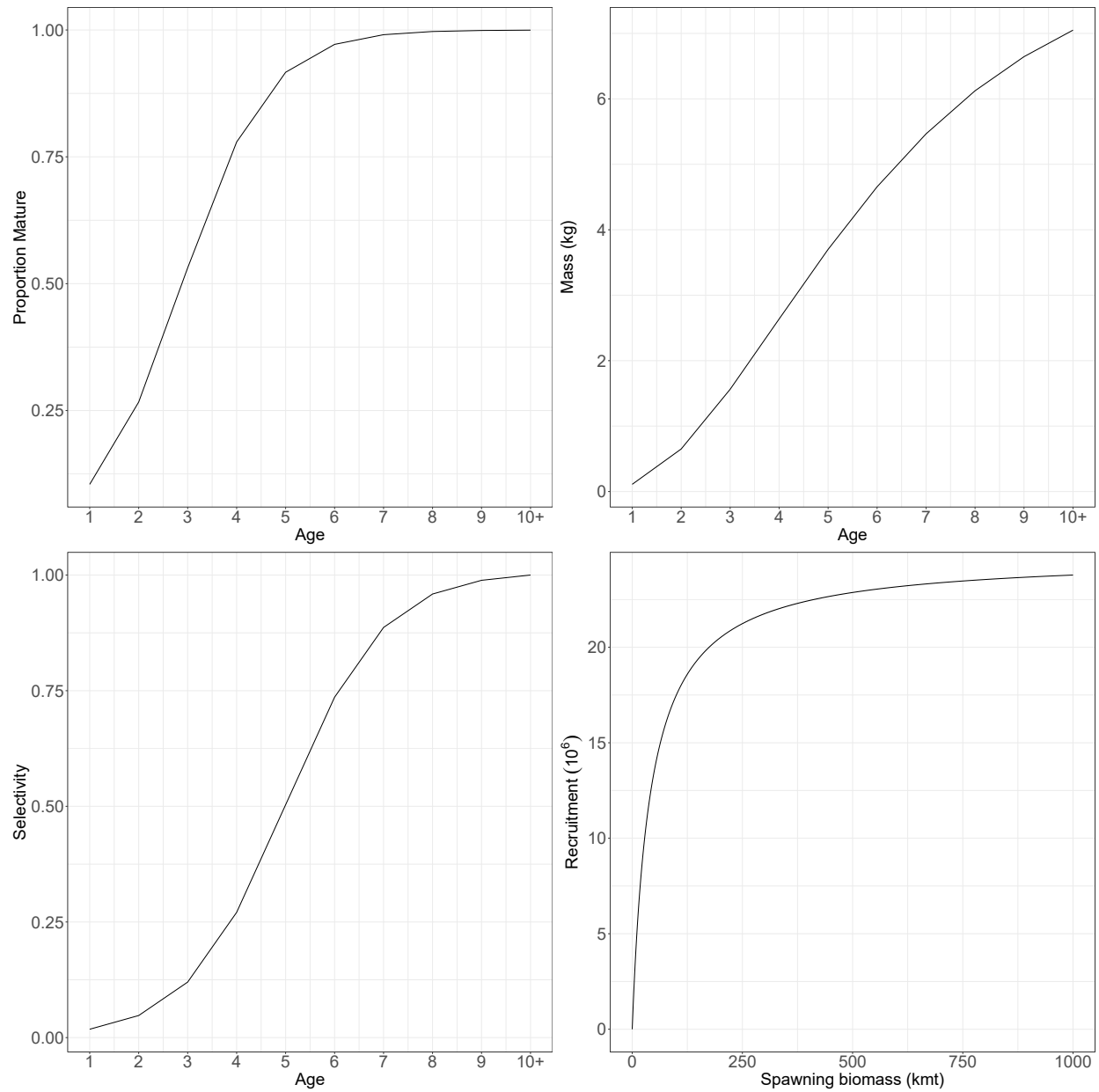


Fig. S1. The proportion mature at age, weight at age, fleet and index selectivity at age, and Beverton-Holt stock-recruit relationship-SRR assumed for the population in all operating models-OMs. For operating models-OMs with random effects on fleet selectivity, this represents the selectivity at the mean of the random effects.

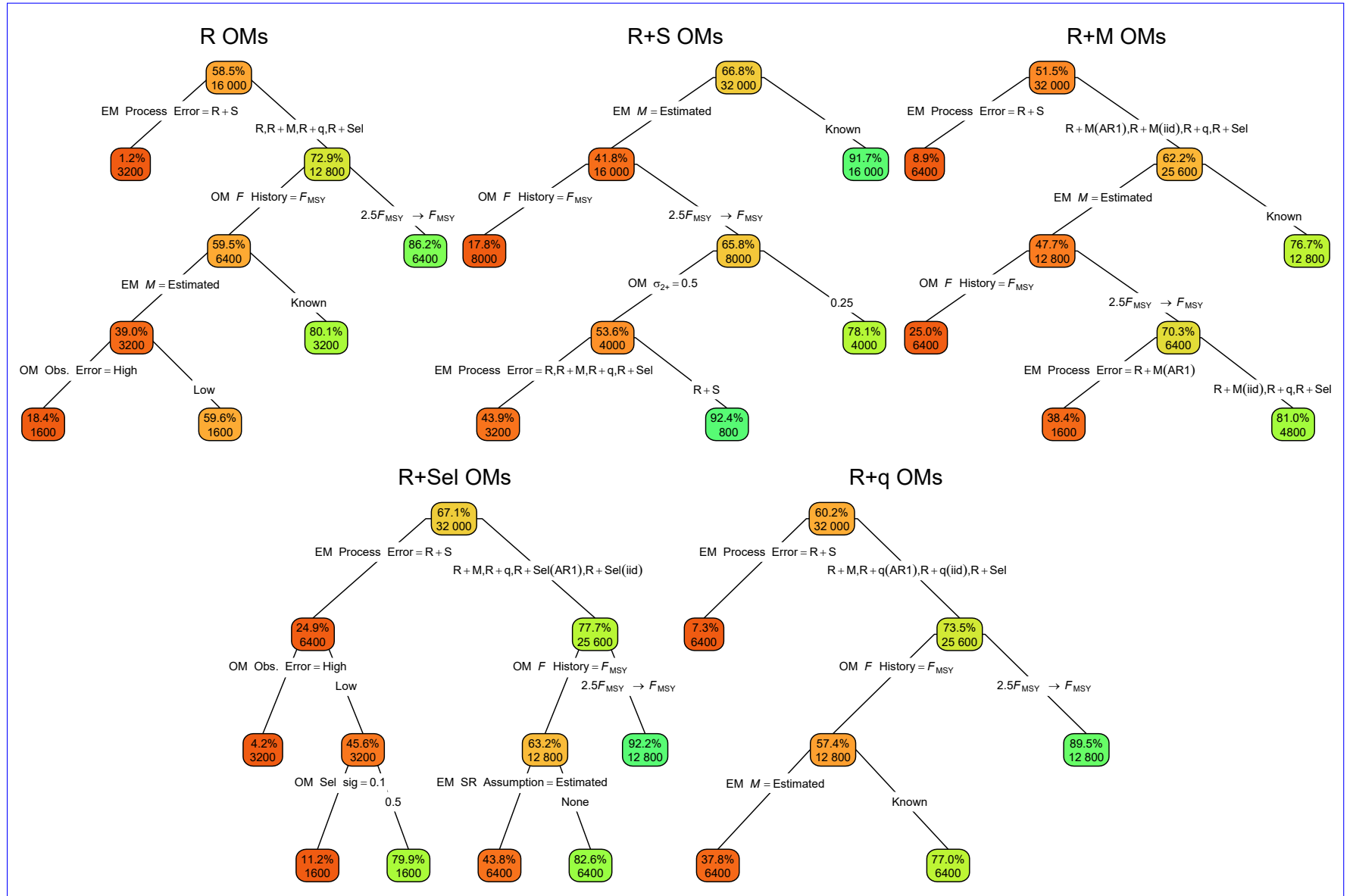


Fig. S2. Classification trees indicating primary factors determining convergence as defined by a maximum absolute gradient $< 10^{-6}$ for R, R+S, R+M, R+Sel and R+q OMs. Nodes denote percent convergence (top) and number of fits (bottom) for the corresponding subset. Lower or higher convergence rates are indicated by more red or green polygons, respectively

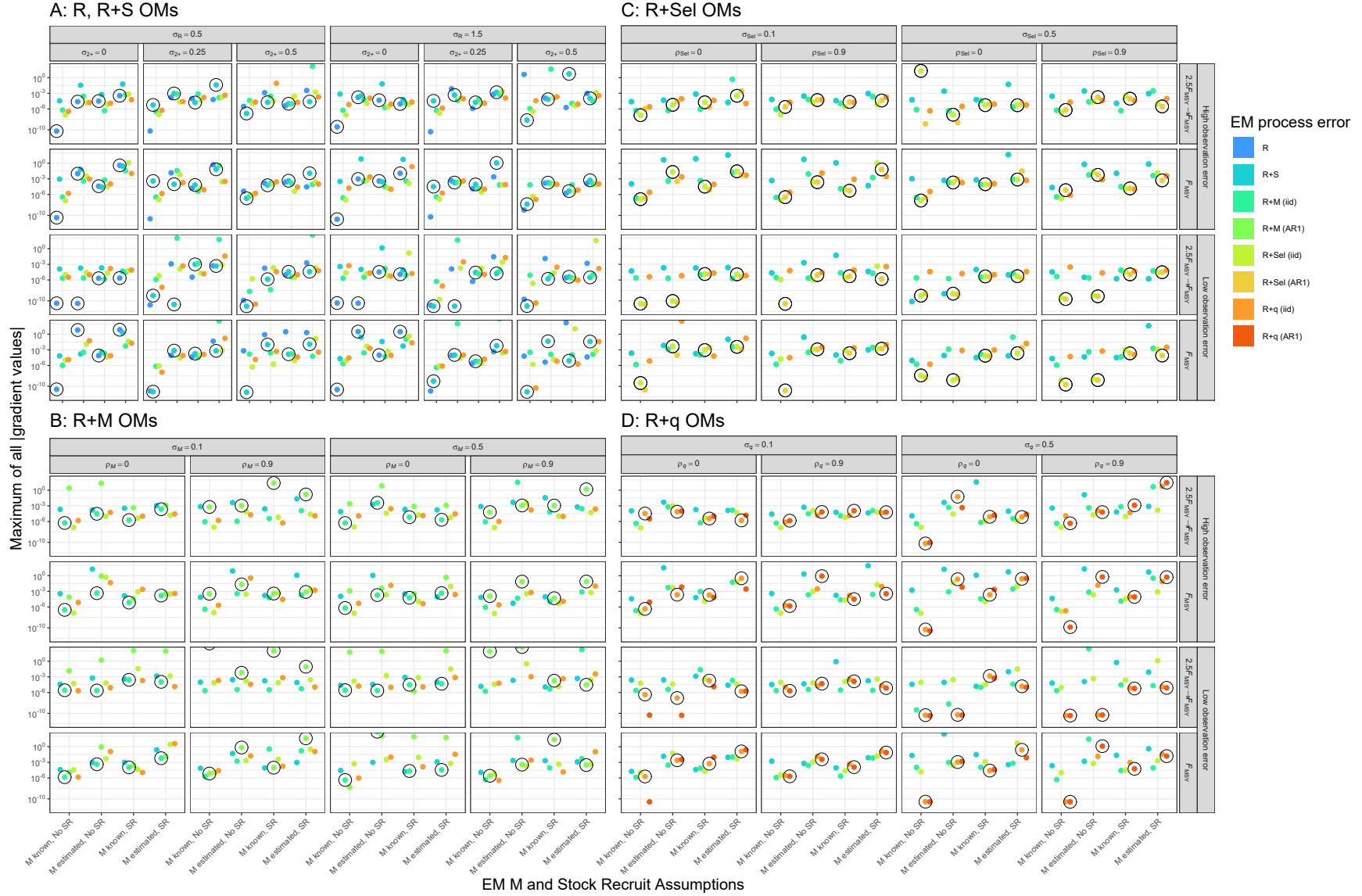


Fig. S3. The maximum of the absolute values of all gradient values for all fits that provided Hessian-based standard errors across all simulated data sets of a given OM configuration (A: R and R+S, B: R+M, C: R+Sel, or D: R+q). Results are conditional on EM fits with alternative process error assumptions (colored points and lines), median natural mortality (estimated or known) and recruitment assumptions (Beverton-Holt SRR or not). Circled values indicate results where the EM process error structure matches that of the OM and vertical lines represent 95% confidence intervals.

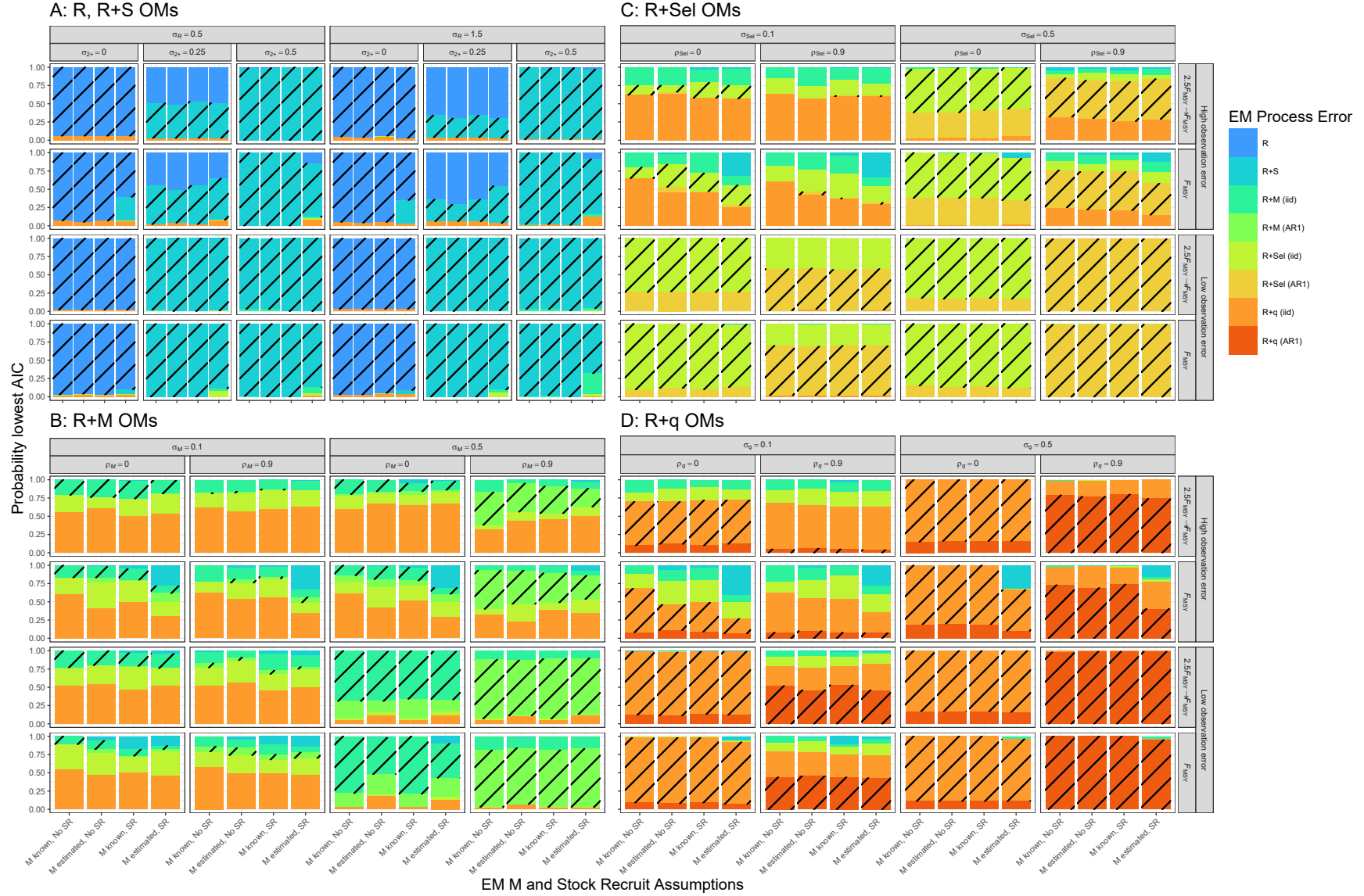


Fig. S4. Estimated probability of lowest AIC for EMs assuming alternative process error assumptions (colored bars) conditional on alternative assumptions for median natural mortality (estimated or known) and Beverton-Holt SRR (estimated or not; along x-axis) when fitted to OMs that have R and R+S (A), R+Sel (B), R+M (C), or R+q (D) process error sources. Striped bars indicate results where the EM process error structure matches that of the OM.

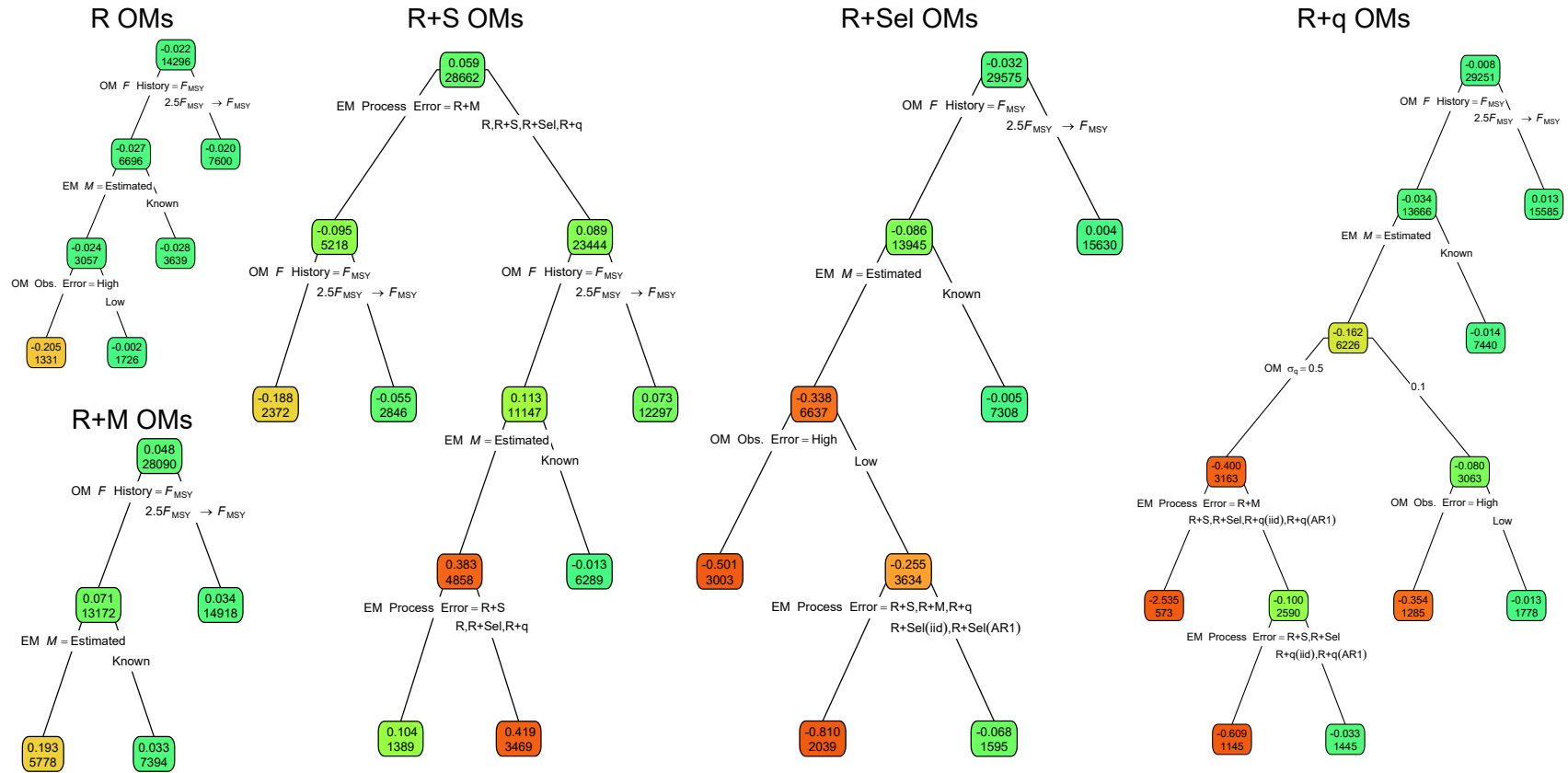


Fig. S5. Regression trees indicating primary factors determining reductions in sums of squares of errors in estimation measured by Eq. 3 for terminal year fully-selected fishing mortality for R+S, R+M, R+Sel and R+q OM. Each node shows the median error (top) and number of observations (bottom) for the corresponding subset. Median errors closer to or further from zero are indicated by more green or red polygons, respectively.

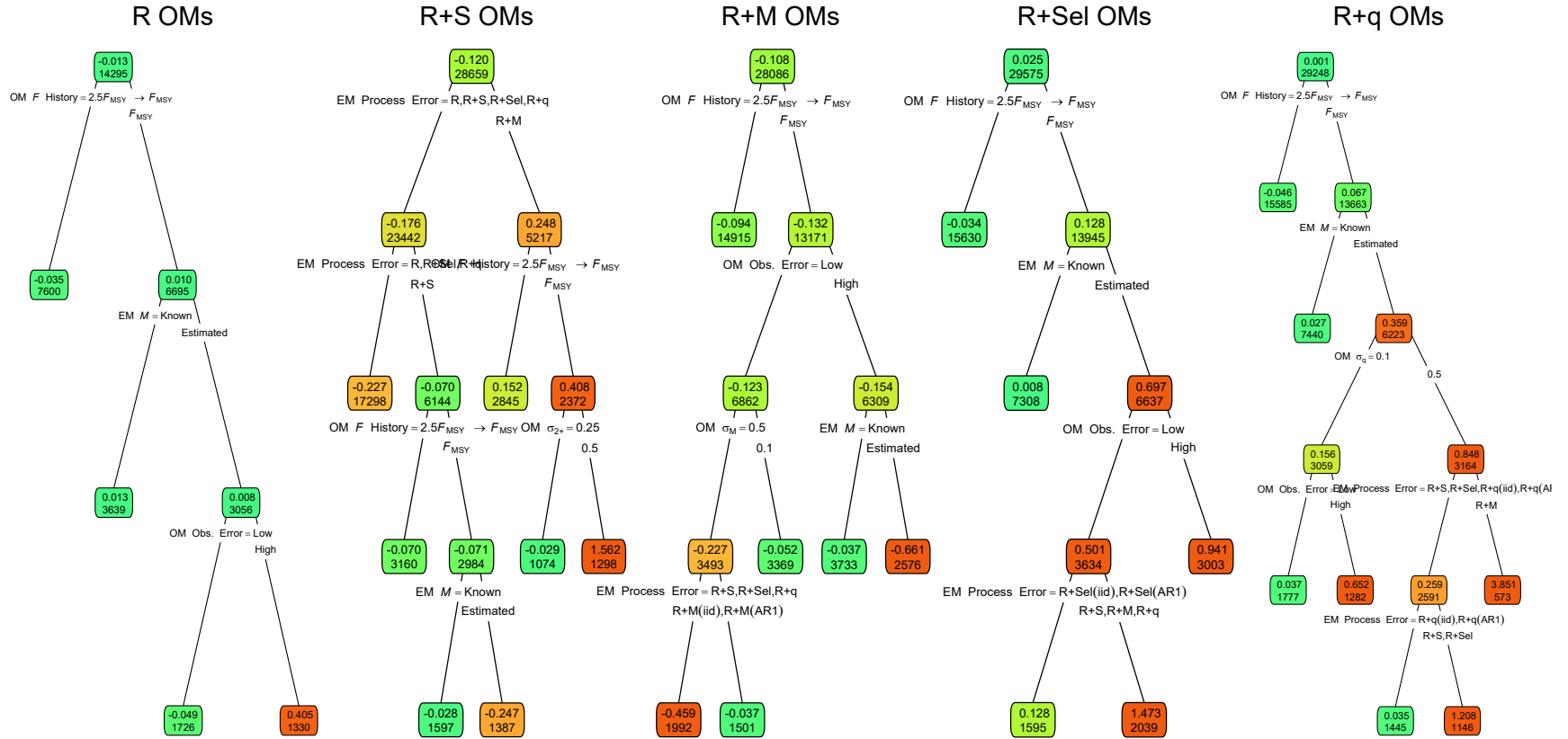


Fig. S6. Regression trees indicating primary factors determining reductions in sums of squares of errors in estimation measured by Eq. 3 for terminal year recruitment for R+S, R+M, R+Sel and R+q OM. Each node shows the median error (top) and number of observations (bottom) for the corresponding subset. Median errors closer to or further from zero are indicated by more green or red polygons, respectively.

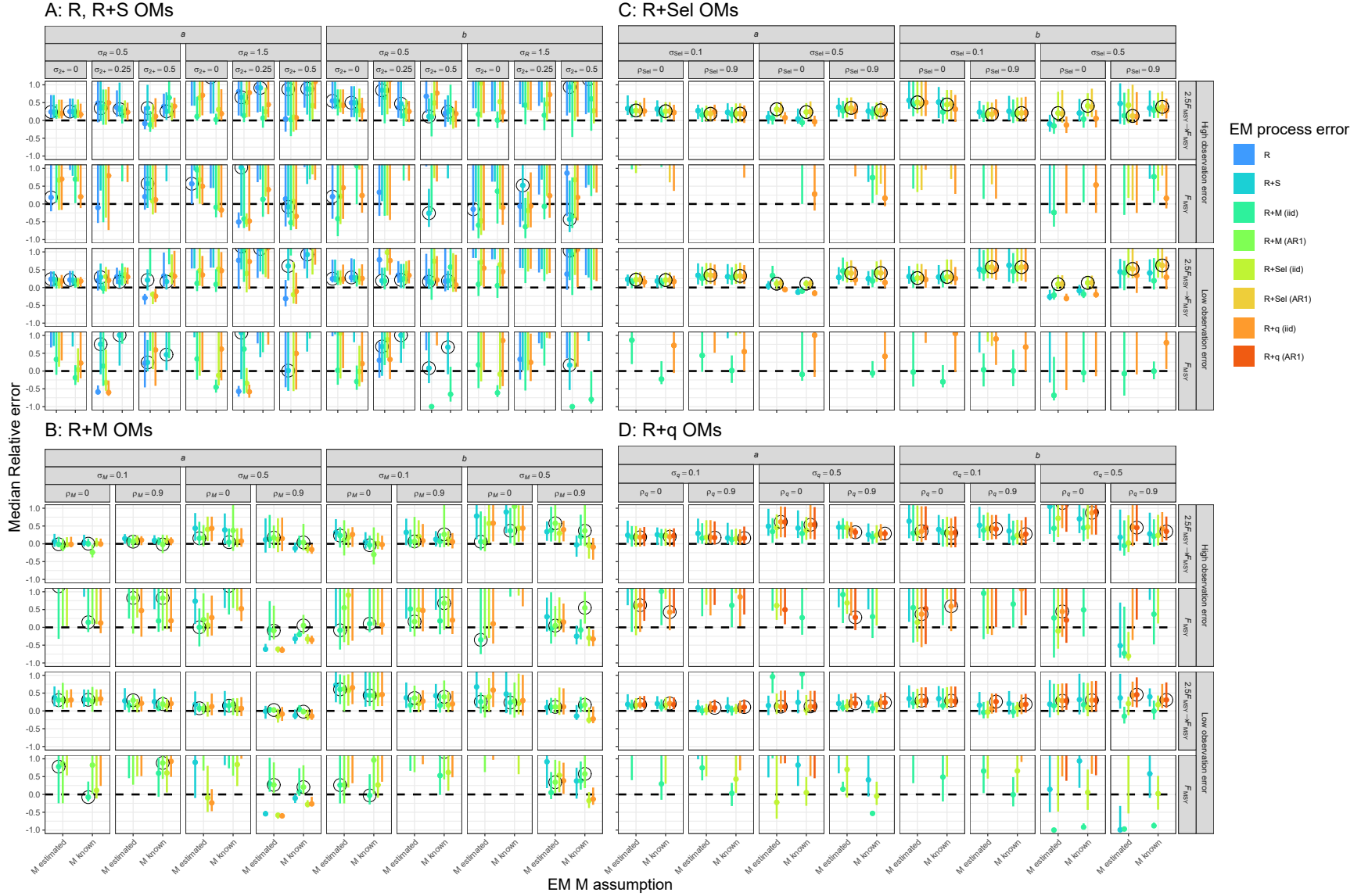


Fig. S7. Median relative error of Beverton-Holt SRR parameters (a and b) for EMs fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the OM and vertical lines represent 95% confidence intervals.

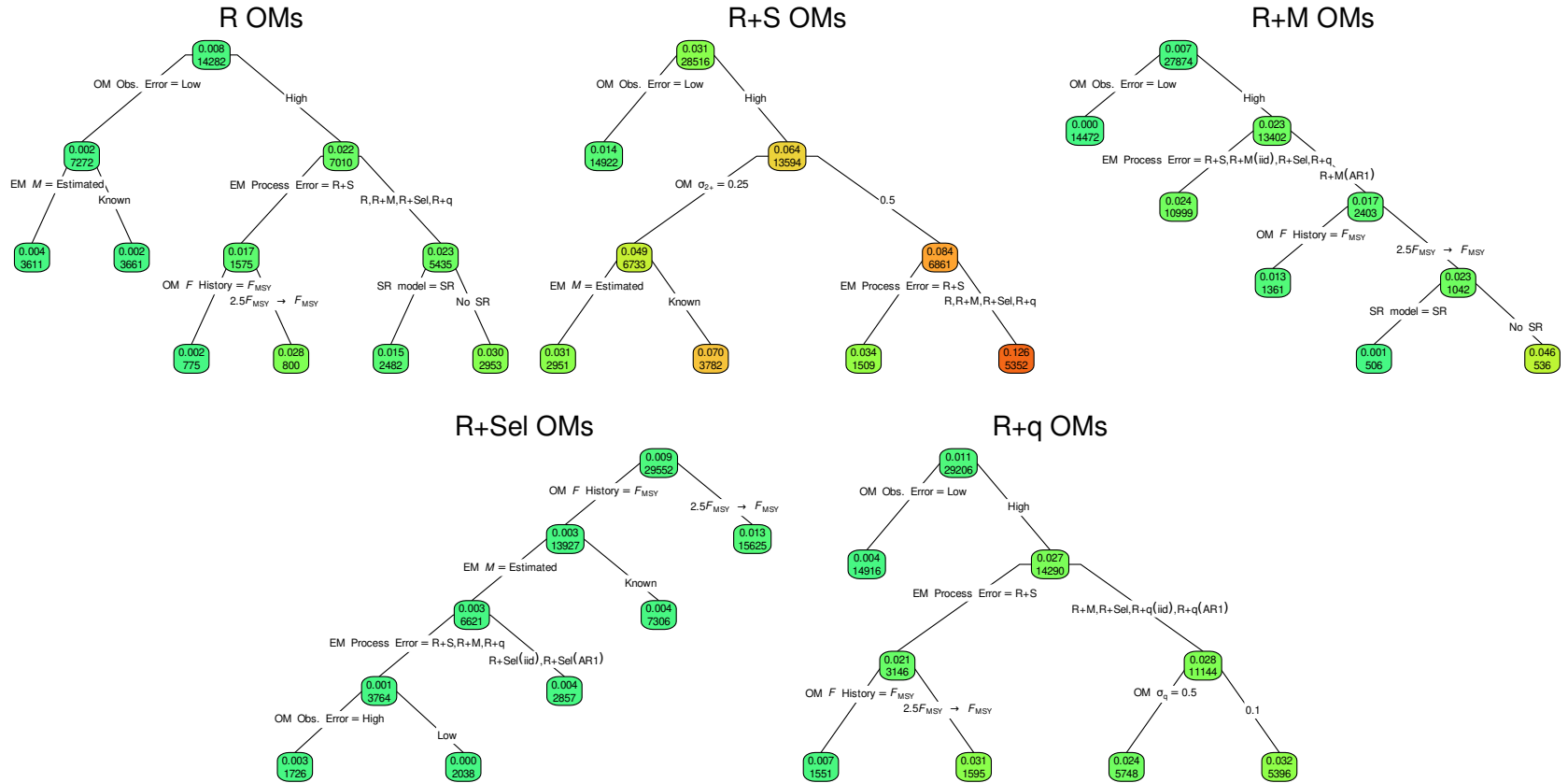


Fig. S8. Regression trees indicating primary factors determining reductions in sums of squares of errors in transformed Mohn's ρ (Eq. 3) for fishing mortality averaged over all age classes for R+S, R+M, R+Sel and R+q OMs. Each node shows the median Mohn's ρ (top) and number of observations (bottom) for the corresponding subset. Median Mohn's ρ closer to or further from zero are indicated by more green or red polygons, respectively.

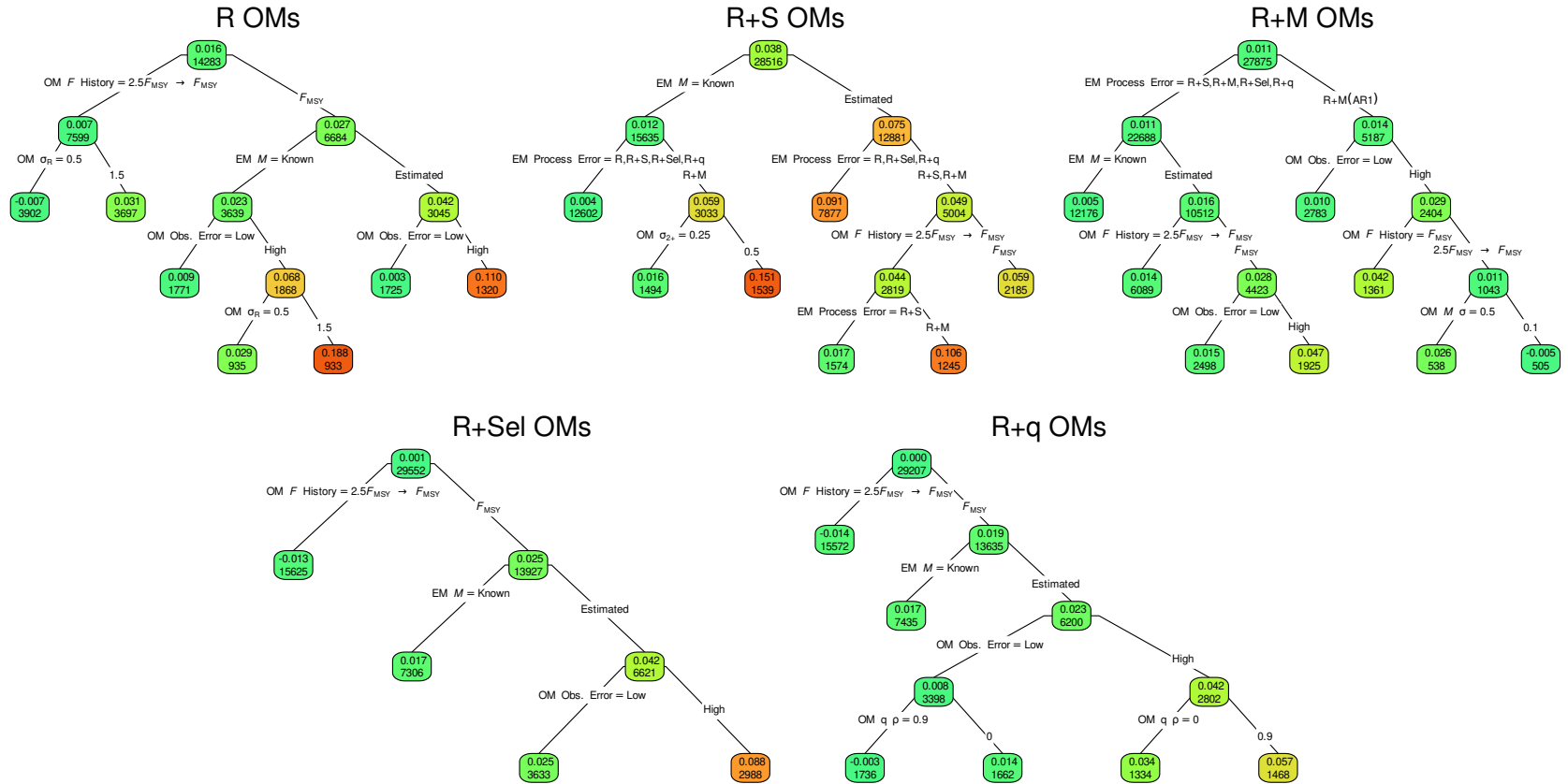


Fig. S9. Regression trees indicating primary factors determining reductions in sums of squares of errors in transformed Mohn's ρ (Eq. 3) for recruitment for R+S, R+M, R+Sel and R+q OMs. Each node shows the median Mohn's ρ (top) and number of observations (bottom) for the corresponding subset. Median Mohn's ρ closer to or further from zero are indicated by more green or red polygons, respectively.

Table S1. Distinguishing characteristics of the ~~operating models~~ OMs with random effects on recruitment and apparent survival (R_{\sim} , $R+S$). ~~Standard~~ When observation uncertainty is low, standard deviations ~~(SD) are~~ for log-normal distributed indices and logistic normal distributed age composition observations ~~(fleet are 0.1 and indices) 0.3, respectively, and when it is high, standard deviations are 0.4 and 1.5, respectively.~~ Fishing mortality either changes from $2.5F_{MSY}$ to F_{MSY} after year 20 (of 40) ~~for fishing histories where fishing mortality or~~ is ~~not~~ constant at F_{MSY} over all years.

Model	σ_R	σ_{2+}	Fishing History	Observation Uncertainty
1	0.5		$2.5F_{MSY} \rightarrow F_{MSY}$	Low
2	1.5		$2.5F_{MSY} \rightarrow F_{MSY}$	Low
3	0.5	0.25	$2.5F_{MSY} \rightarrow F_{MSY}$	Low
4	1.5	0.25	$2.5F_{MSY} \rightarrow F_{MSY}$	Low
5	0.5	0.50	$2.5F_{MSY} \rightarrow F_{MSY}$	Low
6	1.5	0.50	$2.5F_{MSY} \rightarrow F_{MSY}$	Low
7	0.5		F_{MSY}	Low
8	1.5		F_{MSY}	Low
9	0.5	0.25	F_{MSY}	Low
10	1.5	0.25	F_{MSY}	Low
11	0.5	0.50	F_{MSY}	Low
12	1.5	0.50	F_{MSY}	Low
13	0.5		$2.5F_{MSY} \rightarrow F_{MSY}$	High
14	1.5		$2.5F_{MSY} \rightarrow F_{MSY}$	High
15	0.5	0.25	$2.5F_{MSY} \rightarrow F_{MSY}$	High
16	1.5	0.25	$2.5F_{MSY} \rightarrow F_{MSY}$	High
17	0.5	0.50	$2.5F_{MSY} \rightarrow F_{MSY}$	High
18	1.5	0.50	$2.5F_{MSY} \rightarrow F_{MSY}$	High
19	0.5		F_{MSY}	High
20	1.5		F_{MSY}	High
21	0.5	0.25	F_{MSY}	High
22	1.5	0.25	F_{MSY}	High
23	0.5	0.50	F_{MSY}	High
24	1.5	0.50	F_{MSY}	High

Table S2. Distinguishing characteristics of the ~~operating models~~ OMs with random effects on recruitment and natural mortality (R+M). ~~Standard~~ When observation uncertainty is low, standard deviations ~~(SD) are~~ for log-normal distributed indices and logistic normal distributed age composition observations ~~(fleet are 0.1 and indices) 0.3, respectively, and when it is high, standard deviations are 0.4 and 1.5, respectively.~~ Fishing mortality either changes from $2.5F_{\text{MSY}}$ to F_{MSY} after year 20 (of 40) for fishing histories where fishing mortality or is not constant at F_{MSY} over all years. For AR1 process errors, ~~σ~~ σ_M is defined for the marginal distribution of the processes.

Model	σ_R	σ_M	ρ_M	Fishing History	Observation Uncertainty
1	0.5	0.1	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
2	0.5	0.5	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
3	0.5	0.1	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
4	0.5	0.5	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
5	0.5	0.1	0.0	F_{MSY}	Low
6	0.5	0.5	0.0	F_{MSY}	Low
7	0.5	0.1	0.9	F_{MSY}	Low
8	0.5	0.5	0.9	F_{MSY}	Low
9	0.5	0.1	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
10	0.5	0.5	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
11	0.5	0.1	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
12	0.5	0.5	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
13	0.5	0.1	0.0	F_{MSY}	High
14	0.5	0.5	0.0	F_{MSY}	High
15	0.5	0.1	0.9	F_{MSY}	High
16	0.5	0.5	0.9	F_{MSY}	High

Table S3. Distinguishing characteristics of the ~~operating models~~ OMs with random effects on recruitment and selectivity (R+Sel). ~~Standard~~ When observation uncertainty is low, ~~standard deviations (SD) are~~ standard deviations ~~for log-normal distributed indices and logistic normal distributed age composition observations (fleet are 0.1 and indices)0.3, respectively, and when it is high, standard deviations are 0.4 and 1.5, respectively.~~ Fishing mortality either ~~changes from $2.5F_{\text{MSY}}$ to F_{MSY} after year 20 (of 40) for fishing histories where fishing mortality or~~ is ~~not~~ constant at F_{MSY} over all years. For AR1 process errors, ~~σ~~ σ_{Sel} is defined for the marginal distribution of the processes.

Model	σ_R	σ_{Sel}	ρ_{Sel}	Fishing History	Observation Uncertainty
1	0.5	0.1	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
2	0.5	0.5	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
3	0.5	0.1	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
4	0.5	0.5	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
5	0.5	0.1	0.0	F_{MSY}	Low
6	0.5	0.5	0.0	F_{MSY}	Low
7	0.5	0.1	0.9	F_{MSY}	Low
8	0.5	0.5	0.9	F_{MSY}	Low
9	0.5	0.1	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
10	0.5	0.5	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
11	0.5	0.1	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
12	0.5	0.5	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
13	0.5	0.1	0.0	F_{MSY}	High
14	0.5	0.5	0.0	F_{MSY}	High
15	0.5	0.1	0.9	F_{MSY}	High
16	0.5	0.5	0.9	F_{MSY}	High

Table S4. Distinguishing characteristics of the ~~operating models~~ OMs with random effects on recruitment and catchability (R+q). ~~Standard~~ When observation uncertainty is low, ~~standard deviations (SD) are~~ standard deviations ~~for log-normal distributed indices and logistic normal distributed age composition observations (fleet are 0.1 and indices)0.3, respectively, and when it is high, standard deviations are 0.4 and 1.5, respectively.~~ Fishing mortality either ~~changes from $2.5F_{\text{MSY}}$ to F_{MSY} after year 20 (of 40) for fishing histories where fishing mortality or is not constant at F_{MSY} over all years.~~ For AR1 process errors, ~~σ~~ σ_q is defined for the marginal distribution of the processes.

Model	σ_R	σ_q	ρ_q	Fishing History	Observation Uncertainty
1	0.5	0.1	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
2	0.5	0.5	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
3	0.5	0.1	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
4	0.5	0.5	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	Low
5	0.5	0.1	0.0	F_{MSY}	Low
6	0.5	0.5	0.0	F_{MSY}	Low
7	0.5	0.1	0.9	F_{MSY}	Low
8	0.5	0.5	0.9	F_{MSY}	Low
9	0.5	0.1	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
10	0.5	0.5	0.0	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
11	0.5	0.1	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
12	0.5	0.5	0.9	$2.5F_{\text{MSY}} \rightarrow F_{\text{MSY}}$	High
13	0.5	0.1	0.0	F_{MSY}	High
14	0.5	0.5	0.0	F_{MSY}	High
15	0.5	0.1	0.9	F_{MSY}	High
16	0.5	0.5	0.9	F_{MSY}	High

Table S5. Distinguishing characteristics of the ~~estimating models~~ EMs and ~~operating model~~ indication (+) of which OM process error categories sources (R, R+S, R+M, R+Sel, R+q) ~~where used each EM configuration was fit.~~

Model	Recruitment model	Median M	Process error	R,R+S OMs	R+M OMs	R+Sel OMs	R+q OMs
1	Mean recruitment	0.2	R ($\sigma_{2+} = 0$)	+			
2	Beverton-Holt	0.2	R ($\sigma_{2+} = 0$)	+			
3	Mean recruitment	Estimated	R ($\sigma_{2+} = 0$)	+			
4	Beverton-Holt	Estimated	R ($\sigma_{2+} = 0$)	+			
5	Mean recruitment	0.2	R+S (σ_{2+} estimated)	+	+	+	+
6	Beverton-Holt	0.2	R+S (σ_{2+} estimated)	+	+	+	+
7	Mean recruitment	Estimated	R+S (σ_{2+} estimated)	+	+	+	+
8	Beverton-Holt	Estimated	R+S (σ_{2+} estimated)	+	+	+	+
9	Mean recruitment	0.2	R+M ($\rho_M = 0$)	+	+	+	+
10	Beverton-Holt	0.2	R+M ($\rho_M = 0$)	+	+	+	+
11	Mean recruitment	Estimated	R+M ($\rho_M = 0$)	+	+	+	+
12	Beverton-Holt	Estimated	R+M ($\rho_M = 0$)	+	+	+	+
13	Mean recruitment	0.2	R+Sel ($\rho_{Sel} = 0$)	+	+	+	+
14	Beverton-Holt	0.2	R+Sel ($\rho_{Sel} = 0$)	+	+	+	+
15	Mean recruitment	Estimated	R+Sel ($\rho_{Sel} = 0$)	+	+	+	+
16	Beverton-Holt	Estimated	R+Sel ($\rho_{Sel} = 0$)	+	+	+	+
17	Mean recruitment	0.2	R+q ($\rho_q = 0$)	+	+	+	+
18	Beverton-Holt	0.2	R+q ($\rho_q = 0$)	+	+	+	+
19	Mean recruitment	Estimated	R+q ($\rho_q = 0$)	+	+	+	+
20	Beverton-Holt	Estimated	R+q ($\rho_q = 0$)	+	+	+	+
21	Mean recruitment	0.2	R+M (ρ_M estimated)		+		
22	Beverton-Holt	0.2	R+M (ρ_M estimated)		+		
23	Mean recruitment	Estimated	R+M (ρ_M estimated)		+		
24	Beverton-Holt	Estimated	R+M (ρ_M estimated)		+		
25	Mean recruitment	0.2	R+Sel (ρ_{Sel} estimated)			+	
26	Beverton-Holt	0.2	R+Sel (ρ_{Sel} estimated)			+	
27	Mean recruitment	Estimated	R+Sel (ρ_{Sel} estimated)			+	
28	Beverton-Holt	Estimated	R+Sel (ρ_{Sel} estimated)			+	
29	Mean recruitment	0.2	R+q (ρ_q estimated)				+
30	Beverton-Holt	0.2	R+q (ρ_q estimated)				+
31	Mean recruitment	Estimated	R+q (ρ_q estimated)				+
32	Beverton-Holt	Estimated	R+q (ρ_q estimated)				+

Table S6. For each OM process error source (columns), percent reduction in deviance for logistic regression models fit to indicators of convergence (maximum absolute gradient $< 10^{-6}$) with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM Process Error	30.40	0.45	17.57	16.04	24.03
EM M Assumption	2.38	24.11	4.42	1.02	2.66
EM SR Assumption	1.80	0.32	0.96	3.38	2.13
OM Obs. Error	0.12	0.77	0.33	1.76	0.28
OM F History	3.51	6.33	2.36	5.86	5.30
OM σ_R	<0.01	<0.01	~	~	~
OM σ_{2+}	~	<0.01	~	~	~
OM σ_M	~	~	0.39	~	~
OM ρ_M	~	~	0.09	~	~
OM σ_{Sel}	~	~	~	1.08	~
OM ρ_{Sel}	~	~	~	0.01	~
OM σ_q	~	~	~	~	0.06
OM ρ_q	~	~	~	~	<0.01
All factors	43.69	35.72	29.33	34.57	40.69
+ All Two Way	50.53	42.99	43.91	45.93	48.62
+ All Three Way	52.30	48.41	46.81	47.71	50.40

Table S7. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to errors in estimation measured by Eq. 3 for the terminal year fully-selected fishing mortality with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	2.26	1.33	1.26	2.93	3.26
EM SR assumption	0.11	0.07	0.08	0.07	0.09
EM Process Error	0.46	4.18	0.38	0.13	1.02
OM Obs. Error	1.61	0.06	0.86	0.41	<0.01
OM F History	2.49	3.23	1.42	3.22	4.55
OM σ_R	0.02	0.02	—	—	—
OM σ_{2+}	—	0.87	—	—	—
OM σ_M	—	—	0.16	—	—
OM ρ_M	—	—	0.01	—	—
OM σ_{Sel}	—	—	—	0.24	—
OM ρ_{Sel}	—	—	—	0.05	—
OM σ_q	—	—	—	—	1.03
OM ρ_q	—	—	—	—	0.05
All factors	7.42	9.96	4.37	7.26	10.43
+ All Two Way	17.63	25.76	10.94	13.88	22.07
+ All Three Way	22.97	37.03	14.74	17.32	30.74

Table S8. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to errors in estimation measured by Eq. 3 for the terminal year recruitment with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	1.96	0.40	0.69	3.52	3.03
EM SR assumption	0.06	0.02	0.05	0.02	0.05
EM Process Error	0.39	4.74	0.41	0.12	1.16
OM Obs. Error	1.47	0.08	0.64	0.18	<0.01
OM F History	2.54	2.66	1.11	4.18	5.06
OM σ_R	0.03	0.01	—	—	—
OM σ_{2+}	—	1.05	—	—	—
OM σ_M	—	—	0.36	—	—
OM ρ_M	—	—	0.02	—	—
OM σ_{Sel}	—	—	—	0.23	—
OM ρ_{Sel}	—	—	—	0.06	—
OM σ_q	—	—	—	—	1.09
OM ρ_q	—	—	—	—	0.06
All factors	6.90	9.01	3.43	8.58	10.90
+ All Two Way	16.48	24.64	9.73	15.76	22.75
+ All Three Way	21.46	35.60	13.56	19.07	31.15

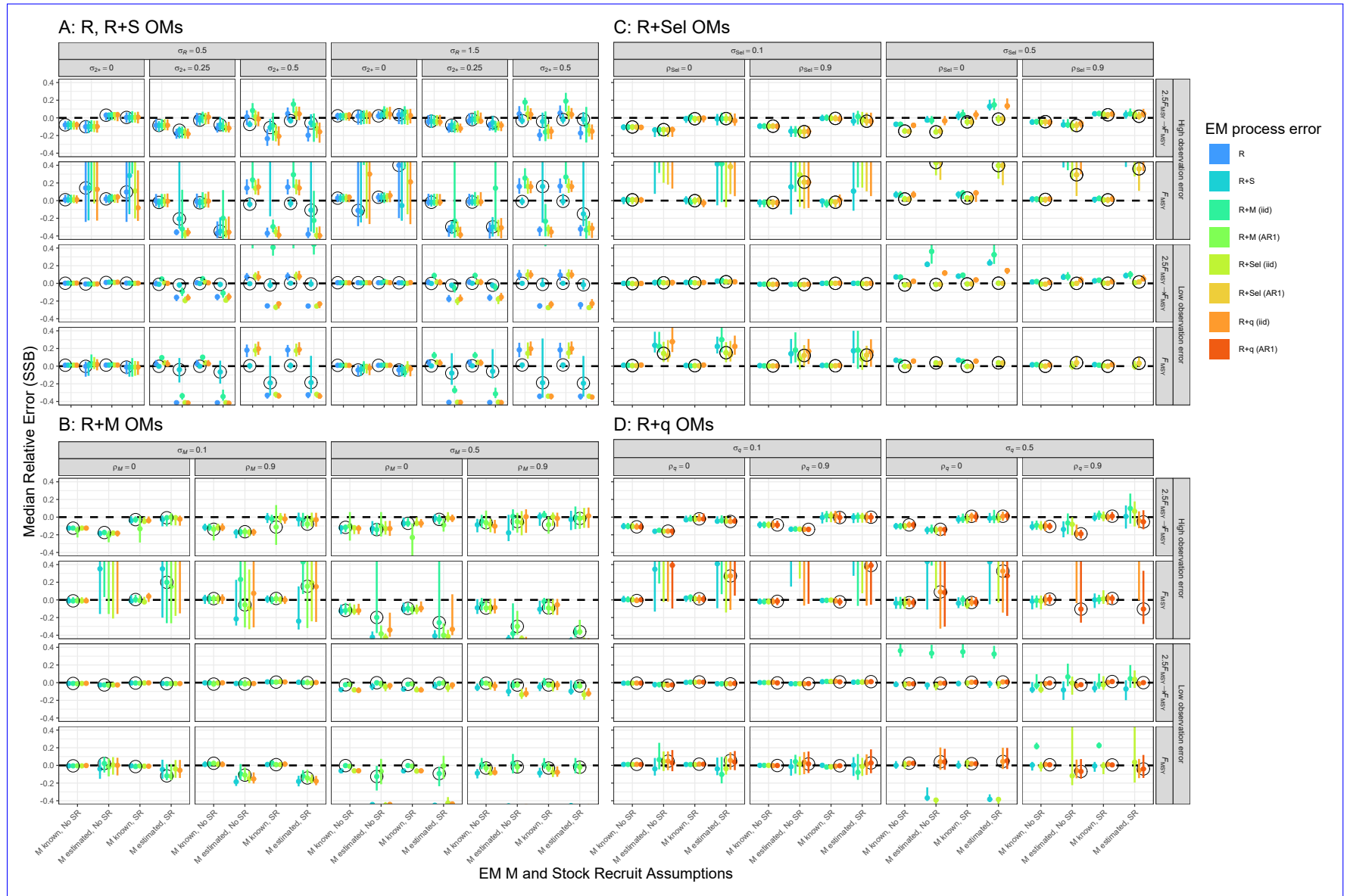


Fig. S10. The maximum Median relative error of the absolute values of all gradient values terminal year SSB for all fits that provided hessian-based standard errors across all simulated EMs fitted to data sets of a given OM configuration (A: simulated with alternative process error sources: R and R+S (A), B: R+M (B), C: R+Sel (C), or D: R+q (D)). Results are conditional on EM fits with alternative process error type (colored points and lines), median natural mortality (estimated or known) and recruitment assumptions (Beverton-Holt stock-recruit relationship or not). Circled values indicate results where the EM process error structure matches that of the operating model OM and vertical lines represent 95% confidence intervals.

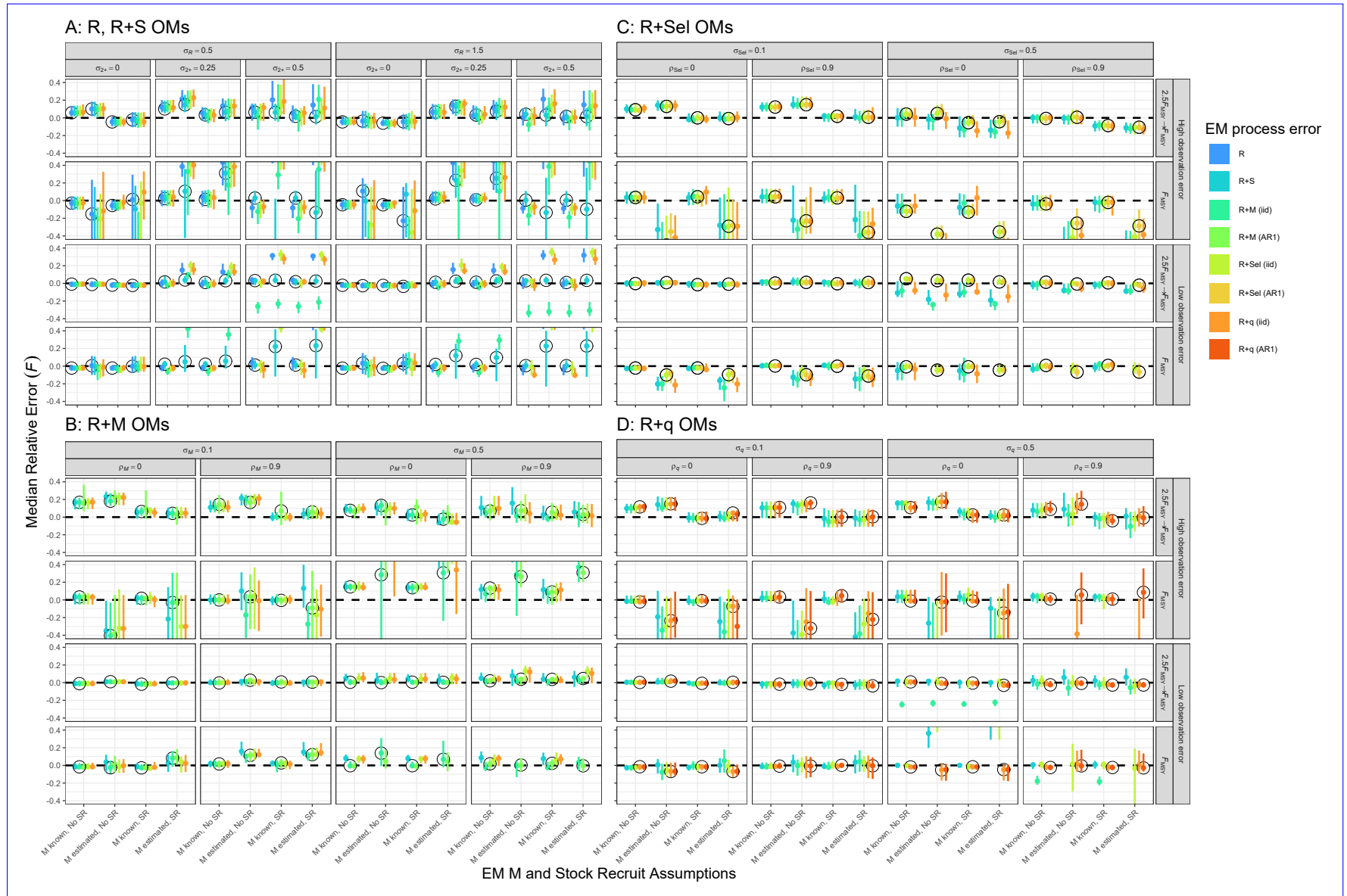


Fig. S11. Probability of estimating models providing maximum absolute values of gradients less than 10^{-6} assuming alternative process error structures (colored points and lines), and median natural of terminal year fully-selected fishing mortality (estimated or known) and Beverton-Holt stock-recruit relationships (estimated or not; along x-axis) when for EMs fitted to operating models that have data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model OM and vertical lines represent 95% confidence intervals.

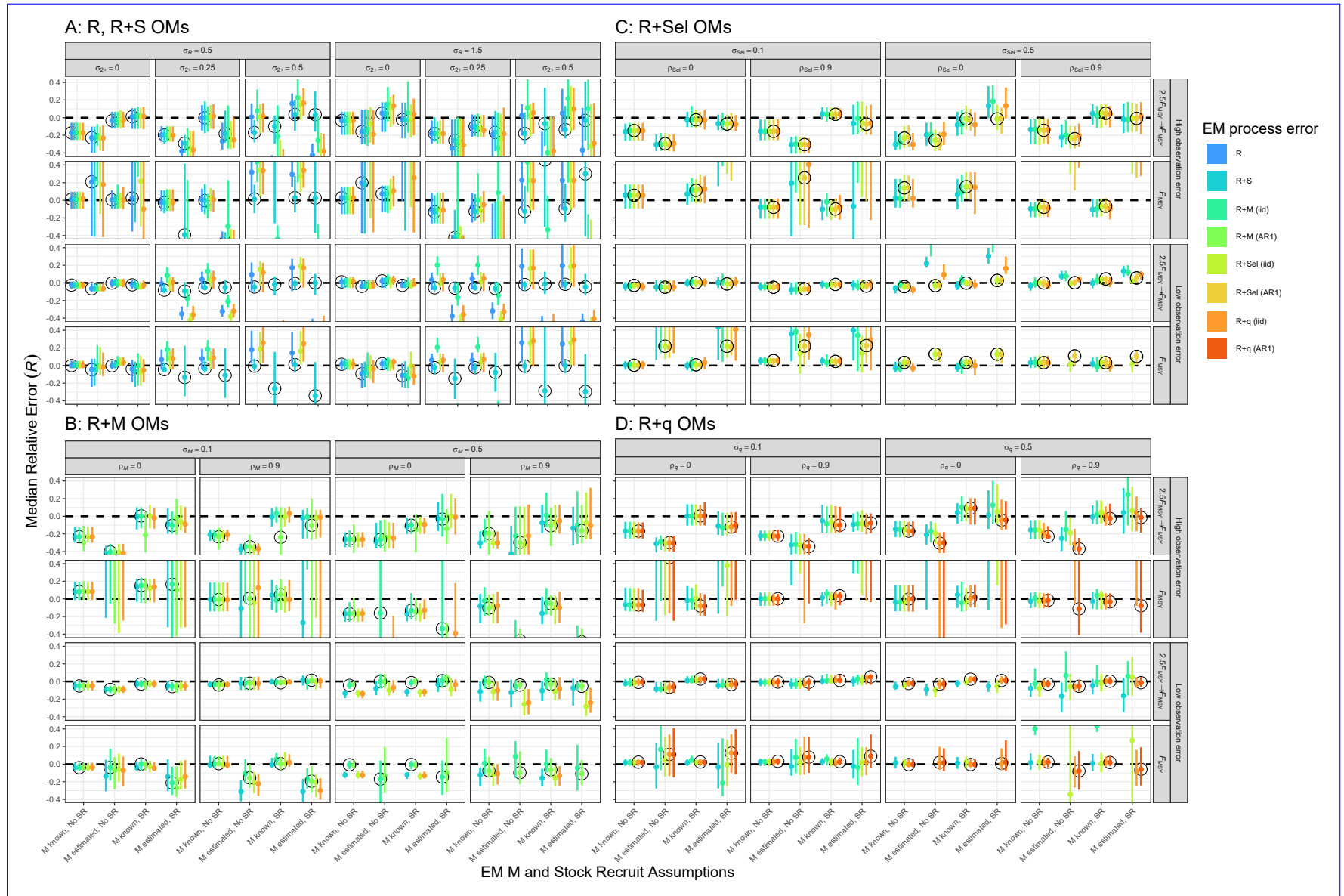


Fig. S12. Estimated probability-Median relative error of lowest AIC from logistic regression on the log-standard deviation of the true log(SSB) in each simulation-terminal year recruitment for estimating model EMs fitted to data sets simulated with Beverton-Holt stock-recruit relationships, rather than the otherwise equivalent EM without the stock-recruit relationship. Results are conditional on alternative assumptions for median natural mortality (estimated or known) and on EMs having the correct process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Rug along x-axis denotes $SD(\log(SSB))$. Circled values for each simulation indicate results where the EM process error structure matches that of the

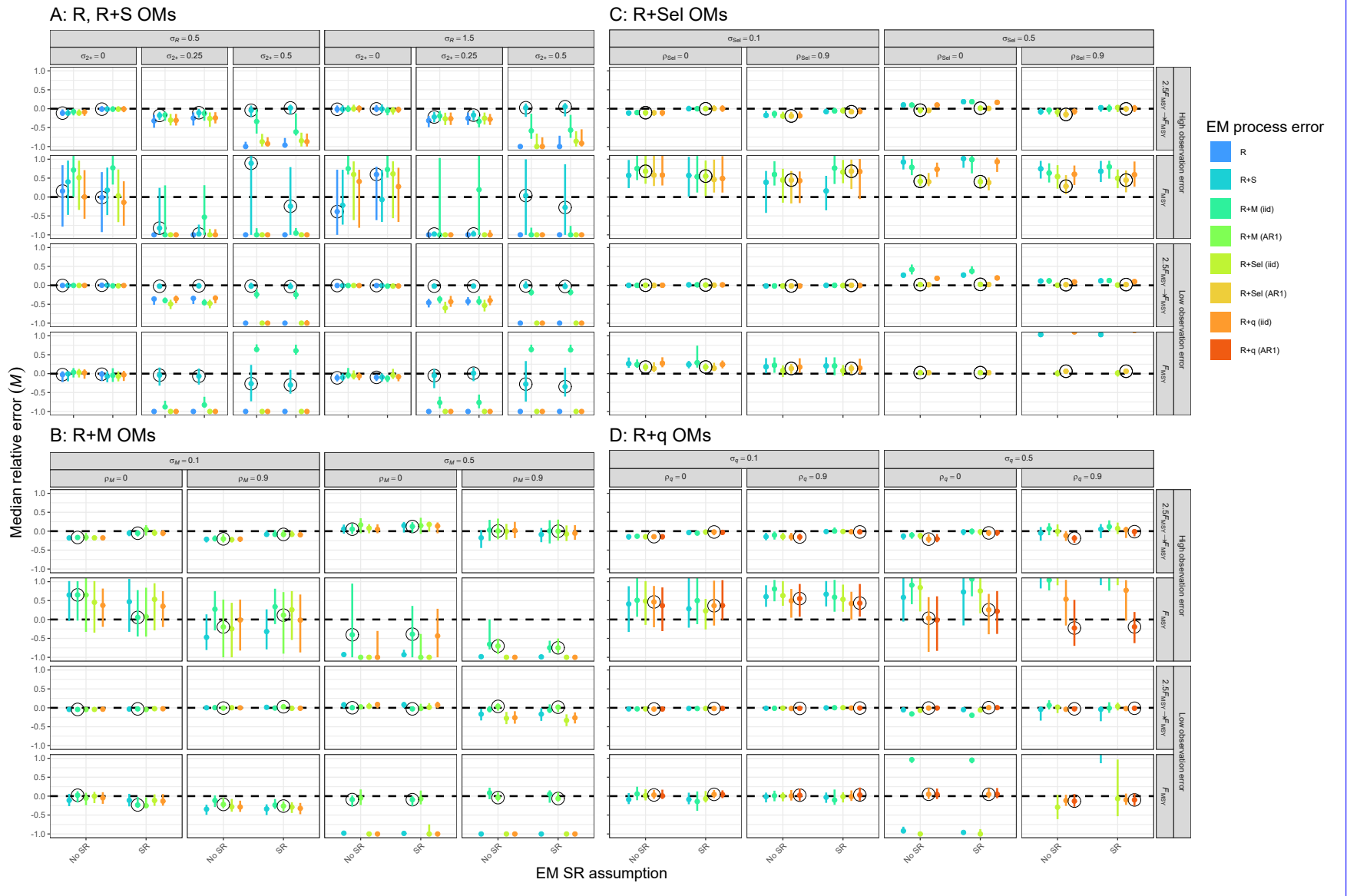


Fig. S13. Median relative error of median natural mortality for EMs fitted to data sets simulated with alternative process error sources: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the OM and vertical lines represent 95% confidence intervals.

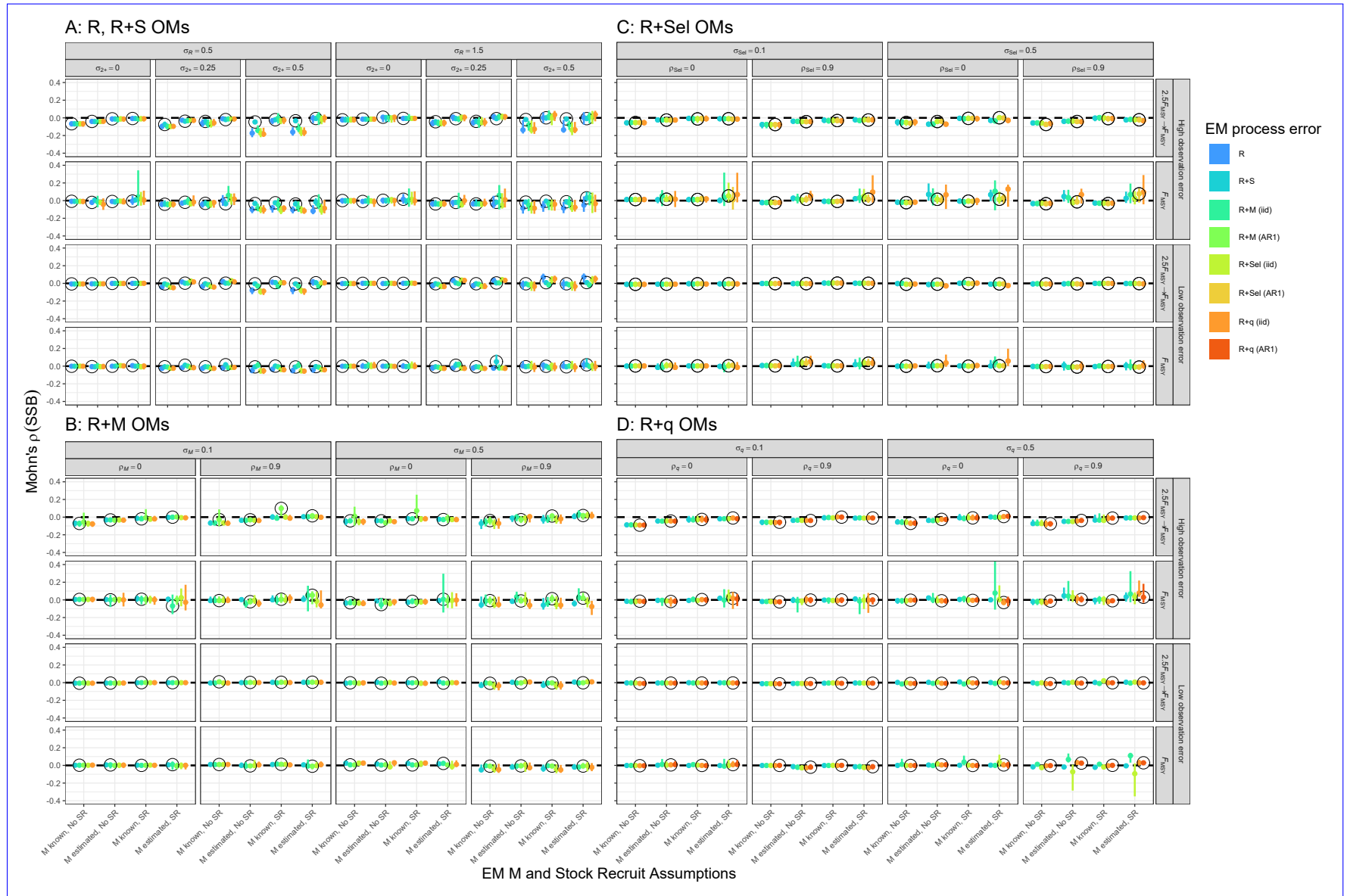
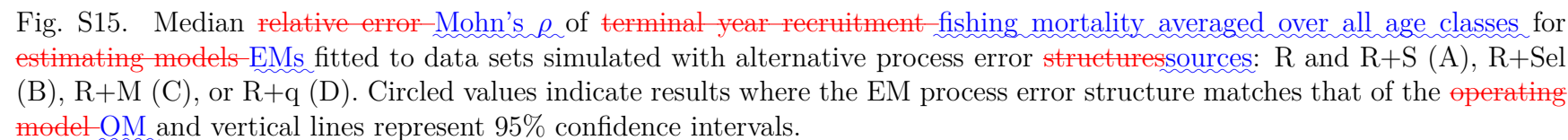


Fig. S14. Median ~~relative error of terminal-year fully-selected-fishing-mortality~~ Mohn's ρ for ~~estimating-models~~ SSB for EMs fitted to data sets simulated with alternative process error ~~structures~~ sources: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the ~~operating-model-OM~~ OM and vertical lines represent 95% confidence intervals.



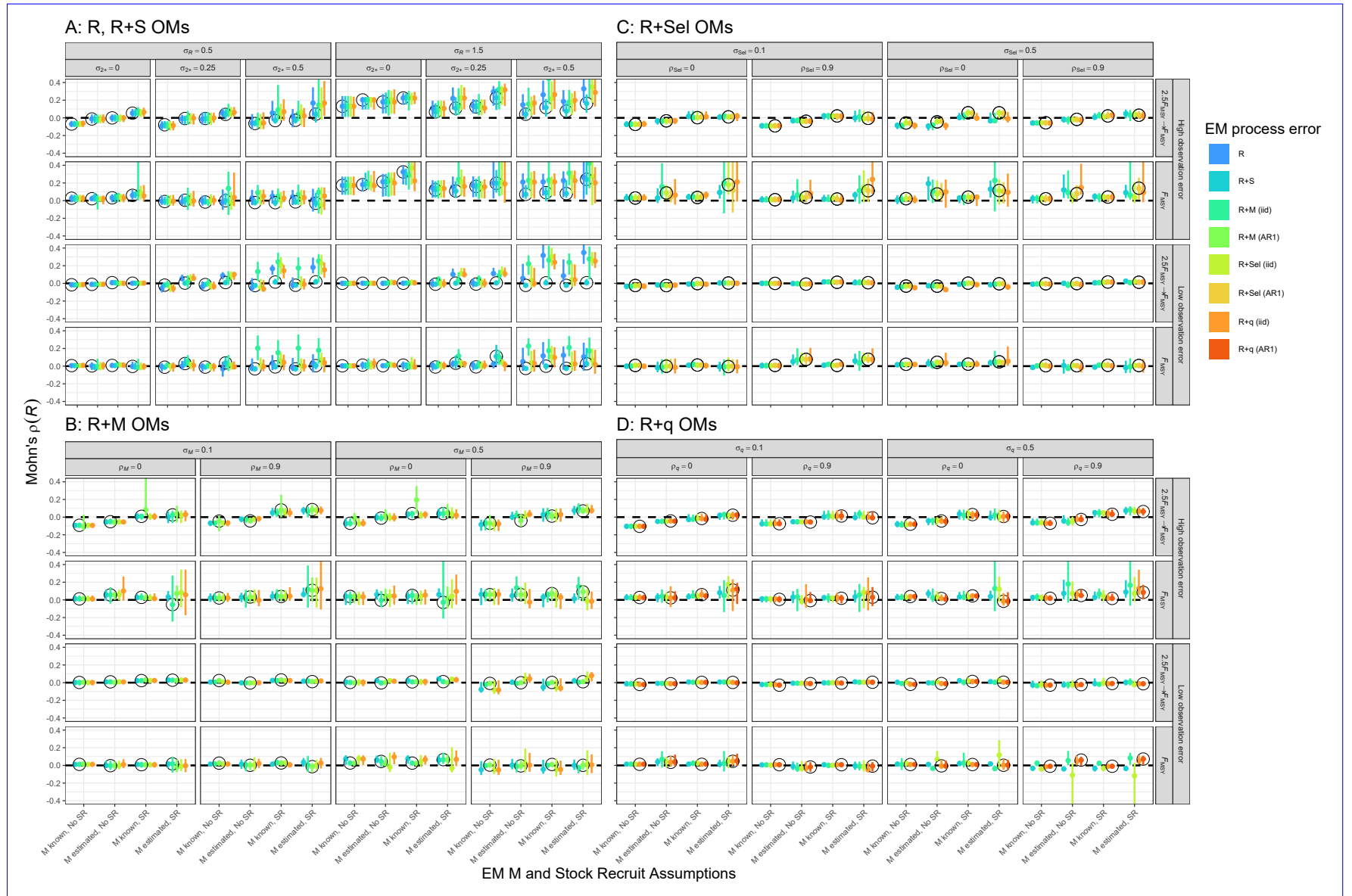


Fig. S16. Median relative error Mohn's ρ of Beverton-Holt stock-recruit parameters (a and b) recruitment for estimating models EMs fitted to data sets simulated with alternative process error structures sources: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model OM and vertical lines represent 95% confidence intervals.

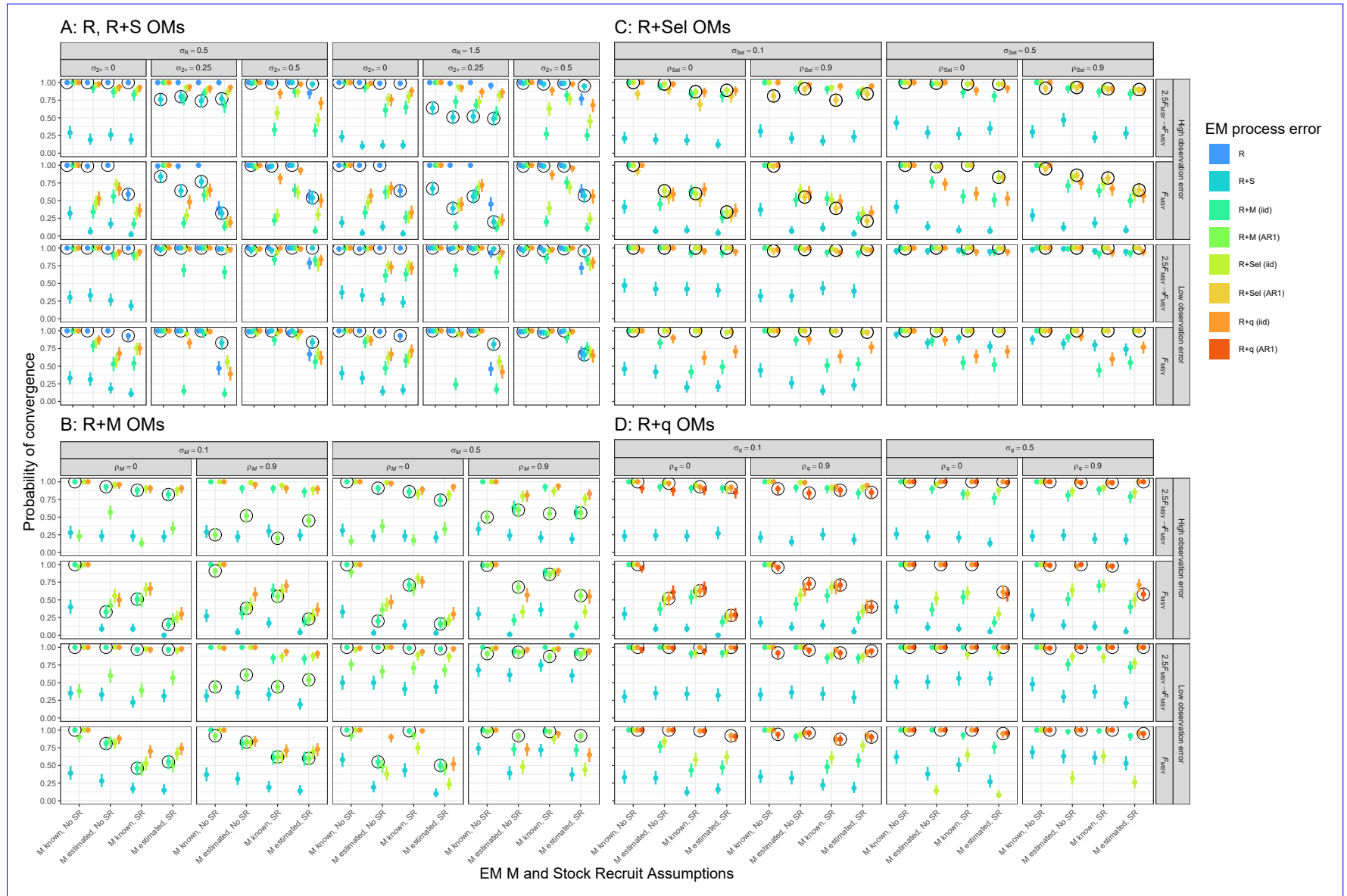


Fig. S17. Median relative error Probability of EMs providing Hessian-based standard errors with alternative process error (colored points and lines), and median natural mortality for estimating models (estimated or known) and Beverton-Holt SRR (estimated or not; along x-axis) assumptions when fitted to data sets simulated with alternative process error structures: OMs that have R and R+S (A), R+Sel (B), R+M (C), or R+q (D) process error sources. Circled values indicate results where the EM process error structure matches that of the operating model OM and vertical lines represent 95% confidence intervals.

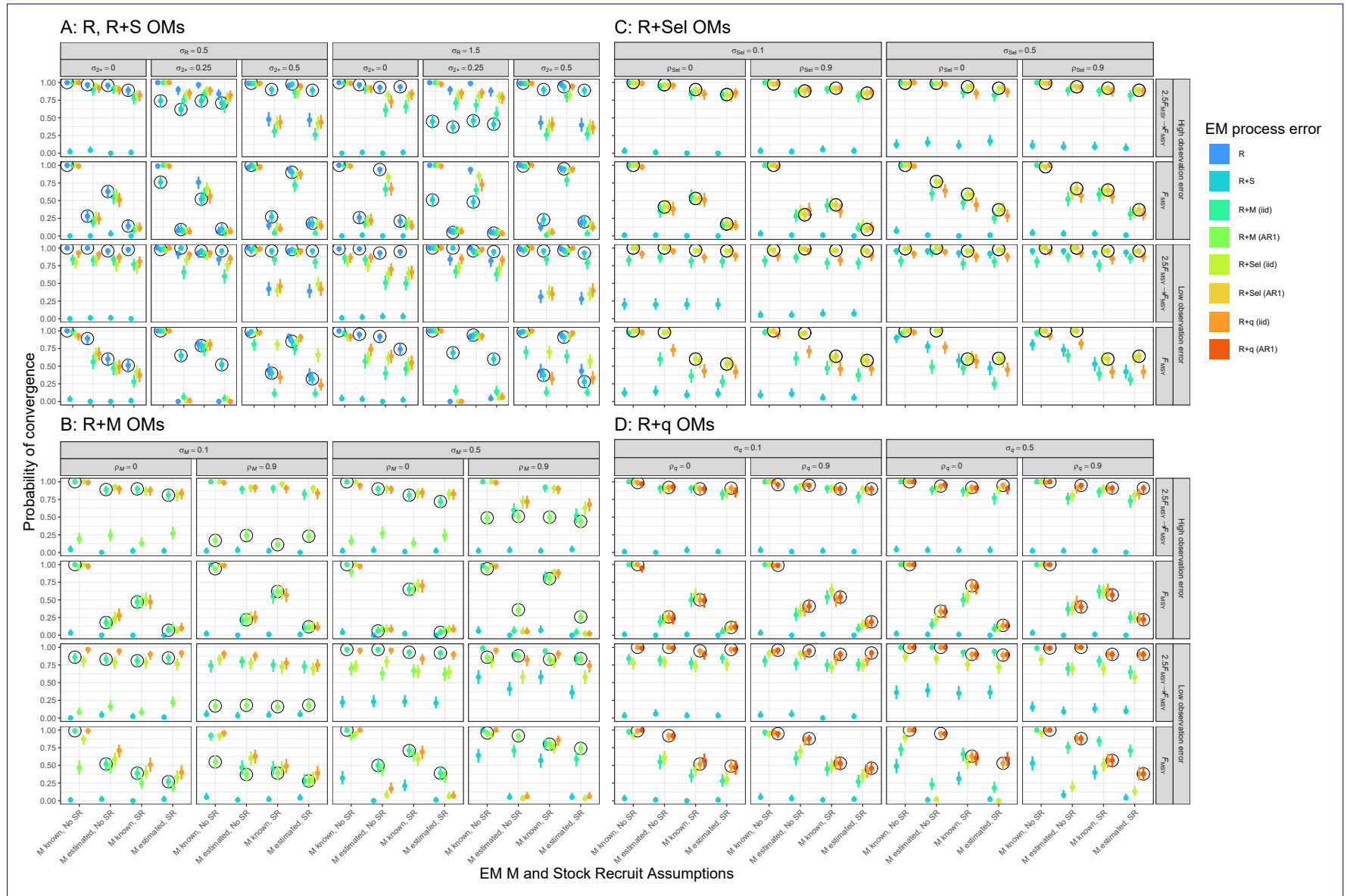


Fig. S18. Median Mohn's ρ Probability of fishing-mortality averaged over all age-classes for estimating models fitted to data sets simulated EMs providing maximum absolute values of gradients less than 10^{-6} with alternative process error structures: (colored points and lines), and median natural mortality (estimated or known) and Beverton-Holt SRR (estimated or not; along x-axis) assumptions when fitted to OMs that have R and R+S (A), R+Sel (B), R+M (C), or R+q (D) process error sources. Circled values indicate results where the EM process error structure matches that of the operating model OM and vertical lines represent 95% confidence intervals.

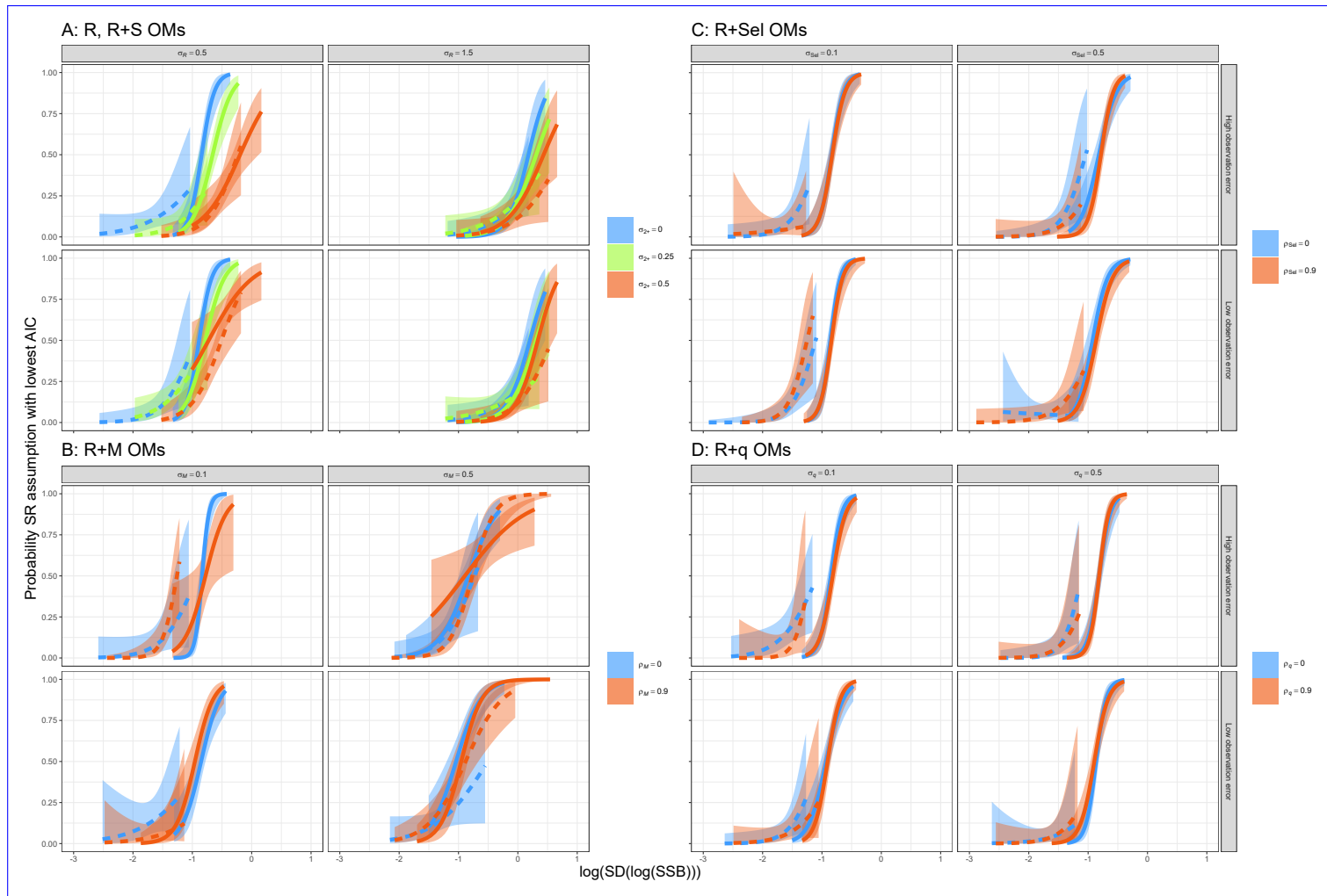


Fig. S19. Median Mohn's ρ Probability of recruitment lowest AIC from logistic regression on the log-standard deviation of the true $\log(\text{SSB})$ in each simulation for estimating models fitted to data sets simulated EM with Beverton-Holt SRRs, rather than the otherwise equivalent EM without the SRR. Results are conditional on median M is known in the EM and alternative assumptions EMs having the correct process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where, and median M is assumed known in the EM process error structure matches that of the operating model. Solid and vertical dashed lines are for OMs with and without temporal contrast in fishing pressure, respectively, and polygons represent 95% confidence intervals. Range of results indicates the range of log-standard deviation of $\log(\text{SSB})$ for simulations of the particular OM.

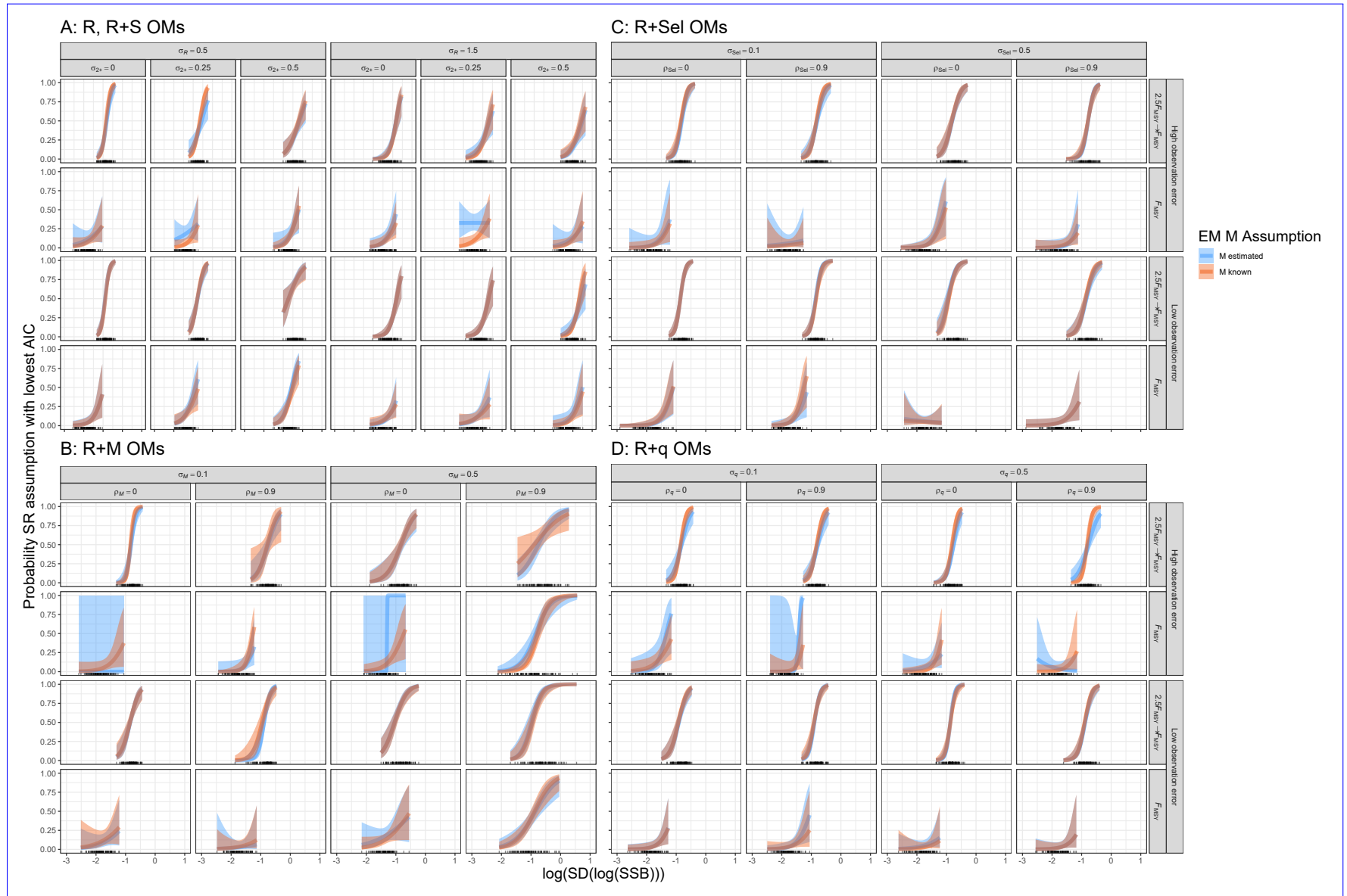


Fig. S20. Estimated probability of lowest AIC from logistic regression on the log-standard deviation of the true log(SSB) in each simulation for EM with Beverton-Holt SRRs, rather than the otherwise equivalent EM without the SRR. Results are conditional on alternative assumptions for median natural mortality (estimated or known) and on EMs having the correct process error structure: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Rug along x-axis denotes $SD(\log(SSB))$ values for each simulation and polygons represent 95% confidence intervals.

Table S9. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to transformed Mohn's ρ values for each simulation (Eq. 3) for fishing mortality averaged over all age classes with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	0.06	0.09	0.01	0.12	0.01
EM SR assumption	0.01	<0.01	0.01	0.02	0.01
EM Process Error	0.03	0.07	0.02	0.06	0.03
OM Obs. Error	0.16	0.10	0.05	0.02	0.07
OM F History	0.07	0.02	0.03	0.24	0.03
OM σ_R	<0.01	0.01	~	~	~
OM σ_{2+}	~	0.09	~	~	~
OM σ_M	~	~	<0.01	~	~
OM ρ_M	~	~	<0.01	~	~
OM σ_{Sel}	~	~	~	0.01	~
OM ρ_{Sel}	~	~	~	<0.01	~
OM σ_q	~	~	~	~	<0.01
OM ρ_q	~	~	~	~	0.01
All factors	0.32	0.38	0.12	0.48	0.15
+ All Two Way	0.65	0.67	0.30	0.95	0.43
+ All Three Way	1.18	1.11	0.63	1.34	0.90

Table S10. For each OM process error source (columns), percent reduction in deviance for linear regression models fit to transformed Mohn's ρ values for each simulation (Eq. 3) for recruitment with each OM and EM factor (rows) included individually, combined, and with second and third order interactions.

Factor	R	R+S	R+M	R+Sel	R+q
EM M Assumption	0.86	0.56	0.16	1.00	1.27
EM SR assumption	<0.01	0.02	0.01	0.01	0.01
EM Process Error	0.01	0.59	0.18	0.07	0.04
OM Obs. Error	0.34	0.01	0.08	0.24	0.27
OM F History	0.91	0.22	0.06	1.20	1.67
OM σ_R	<0.01	0.14	~	~	~
OM σ_{2+}	~	0.11	~	~	~
OM σ_M	~	~	0.01	~	~
OM ρ_M	~	~	<0.01	~	~
OM σ_{Sel}	~	~	~	0.01	~
OM ρ_{Sel}	~	~	~	0.01	~
OM σ_q	~	~	~	~	0.01
OM ρ_q	~	~	~	~	0.01
All factors	2.28	1.74	0.51	2.66	3.51
+ All Two Way	4.20	2.74	1.08	5.08	6.51
+ All Three Way	4.83	3.79	1.79	6.03	7.82