- Factors affecting reliablity of state-space age-structured assessment models
- Timothy J. Miller<sup>1,2</sup> Greg Britten<sup>3</sup> Elizabeth N. Brooks<sup>2</sup>
- Gavin Fay<sup>4</sup> Alex Hansell<sup>2</sup> Christopher M. Legault<sup>2</sup>
- <sup>5</sup> Chengxue Li<sup>2</sup> Brandon Muffley<sup>5</sup> Brian C. Stock<sup>6</sup>
- John Wiedenmann<sup>7</sup>
- 12 March, 2025
- $_{8}$   $^{1} \rm{corresponding}$  author: timothy.j.miller@noaa.gov
- <sup>2</sup>Northeast Fisheries Science Center, Woods Hole Laboratory, 166 Water Street, Woods
- 10 Hole, MA 02543 USA
- <sup>3</sup>Woods Hole Oceanographic Institution
- $^{4}SMAST$
- <sup>13</sup> Institute of Marine Research
- <sup>6</sup>Mid-Atlantic Fisheries Management Council
- <sup>15</sup> Rutgers University

16

### $_{\scriptscriptstyle 17}$ Abstract

State-space assessment models, which are increasingly being used for management of com-18 mercially important fish stocks, can include process errors in many ways, and an evaluation 19 of the statistical reliability of the estimates of management output and best practices of selection among alternative process error configurations are needed. We simulated 72 oper-21 ating models with varying fishing pressure history, observation error error, and process error magnitude, correlation, and source (recruitment, survival, fishery selectivity, catchability, 23 and natural mortality). We fit estimating models with different assumptions on the process error source, whether median natural mortality was estimated, and whether a stock-recruit 25 relationship was estimated. Estimating models that assumed the correct process error source and median natural mortality rate was known, and did not estimate a stock recruit relationship generally had high probability of convergence and low bias in terminal year SSB estimation. Bias was also low under many incorrect process error assumptions when there was contrast in fishing pressure and low observation error. Stock-recruitment parameters were not reliably estimated in any scenarios, but the least accurate estimates occured under 31 ideal situations of low observation error, low recruitment variability, and temporal contrast in fishing pressure. However, estimation of median natural mortality rate was reliable for many OMs (other than those with survival process errors) with temporal contrast in fishing pressure regardless of the estimated process error type. Marginal AIC most accurately distin-35 guished process errors on recruitment, survival, and selectivity, as well as larger magnitude process errors of other types. Retrospective patterns for spawning stock biomass and fishing 37 mortality were generally weak, but those for recruitment were sizable when observation error was high, even with the correct process error assumptions.

# 40 Introduction

panded dramatically within International Council for the Exploration of the Sea (ICES), Canada, and the Northeast US (Nielsen and Berg 2014; Cadigan 2016; Pedersen and Berg 2017; Stock and Miller 2021). State-space models latent population characteristics as statistical time series with periodic observations that also may have error due to sampling or other sources of measurement error. Traditional assessment models may use state-space approaches to account for temporal variability in population characteristics (Legault and Restrepo 1999; Methot and Wetzel 2013), but these models treat the annual parameters as penalized fixed effects parameters where the variance parameters controlling the penalties are assumed known (Thorson and Minto 2015). Modern state-space models can estimate the annually varying parameters as random effects with variance parameters estimated using maximum marginal likelihood or corresponding Bayesian approaches. These latter approaches are considered best practice and a requirement for the next generation of stock assessment models (Hoyle et al. 2022; Punt 2023). State-space stock assessment models, with nonlinear functions of latent parameters and multiple types of observations with varying distributional assumptions, are one of the most complex examples of this analytical approach. Statistical aspects of state-space models and their application within fisheries have been studied extensively, but previous work has focused primarily on linear and Gaussian state-space models (Aeberhard et al. 2018; Auger-Méthé et al. 2021). Therefore, current understanding of the reliability of state-space models does not extend to usage for stock assessment. 61 As state-space models provide greater flexibility by allowing multiple processes to vary as 62 random effects (Nielsen and Berg 2014; Aeberhard et al. 2018; Stock et al. 2021), one of the most immediate questions regards the implications of mis-specification among alternative sources of process error. Incorrect treatment of population attributes as temporally varying

Application of state-space models in fisheries stock assessment and management has ex-

(Trijoulet et al. 2020; Liljestrand et al. 2024) could lead to misidentification of stock status and biased population estimates, ultimately impacting fisheries management decisions (Legault and Palmer 2016; Szuwalski et al. 2018; Cronin-Fine and Punt 2021). Furthermore, biological, fishery, and observational processes are often confounded in catch-at-age data, which may adversely affect ability to distinguish between true process variability and observational error (Li et al. In review; Punt et al. 2014; Stewart and Monnahan 2017; Cronin-Fine and Punt 2021; Fisch et al. 2023).

Li et al. (2024) conducted a full-factorial simulation-estimation study to assess model reliability when confounding random-effects processes (numbers-at-age, fishery selectivity, and
natural mortality) were included. Their results suggest that while state-space models can
generally identify sources of process error, overly complex models, even when misspecified
(i.e., incorporating process error that did not exist in reality), often performed similarly to
correctly specified models, with little to no bias in key management quantities. Similarly,
Liljestrand et al. (2024) found little downside in assuming process error in recruitment or
selectivity, even when it was absent.

Despite mounting efforts, several limitations remain. First, confounding processes that can be treated as random effects in the model were not thoroughly examined or tested within a simulation-estimation framework. Second, previous studies relied on operating models conditioned on specific fisheries, limiting their generalizability (Li et al. In review; Liljestrand et al. 2024). In particular, the effects of observation error and underlying fishing history were not fully isolated from their study design, making it challenging to disentangle the interplay between process and observation error magnitudes, as demonstrated in Fisch et al. (2023). Third, explicitly modeling stock-recruitment relationships as mechanistic drivers of population dynamics is promising (Fleischman et al. 2013; Du Pontavice et al. 2022) but can be challenging in state-space models. Evidence suggests that when both process and observation errors are unknown, estimating density dependence parameters becomes highly uncertain (Knape 2008; Polansky et al. 2009). In particular, Knape (2008) demonstrated

that stronger density dependence becomes increasingly difficult to estimate in the presence of observation error. Therefore, it is crucial to assess whether density dependence mechanisms can be estimated with sufficient precision before drawing ecological conclusions (Auger-Méthé et al. 2016).

In the present study, we conduct a simulation study with operating models (OMs) varying by degree of observation error, source and variability of process error, and fishing history. The simulations from these OMs are fitted with estimation models (EMs) that make alternative assumptions for sources of process error, whether a stock-recruit model was estimated, and 100 whether a constant, or, in some EMs, median, natural mortality is estimated. Given the 101 confounding nature of process errors, developing diagnostic tools to detect model misspeci-102 fication is of great scientific interest and could aid the next generation of stock assessments 103 (Auger-Méthé et al. 2021). We evaluate whether convergence and Akaike Information Cri-104 terion (AIC) can correctly determine the source of process error and the existence of a stock-105 recruit relationship. We also evaluate when retrospective patterns occur and the degree of 106 bias in the outputs of the assessment model that are important for management. 107

# $^{\circ}$ Methods

We used the Woods Hole Assessment Model (WHAM) to configure OMs and EMs in our simulation study (Miller and Stock 2020; Stock and Miller 2021). WHAM is an R package freely available as a github repository. For this study we used version 1.0.6.9000, commit 77bbd94. This package has also been used to configure OMs and EMs for closed loop simulations evaluating index-based assessment methods (Legault et al. 2023) and is used for management of haddock, butterfish, American plaice, bluefish, Atlantic cod, black sea bass, and yellowtail flounder in the Northeast US.

We completed a simulation study with a number of OMs that can be categorized based on where process error random effects are assumed: abundance at age (R, R+S), natural

mortality (R+M), fleet selectivity (R+Sel), or index catchability (R+q). For each OM assumption about variance of process errors and observations are required and the values we used were based on a review of the range of estimates from applications of WHAM in management of stocks of haddock, butterfish, and American plaice in the NE US.

In total, we configured 72 OMs with alternative assumptions about the source and variability 122 of process errors, level of observation error in indices and age composition data, and contrast 123 in fishing pressure over time. We fitted 20 EMs to observations from each of 100 simulations 124 where process errors were also simulated. For R+M, R+Sel, and R+q OMs, we used unique 125 seeds for each simulation, but we inadvertently used the same 100 seeds for all R and R+S 126 OMs. Each of the EMs made alternative assumptions about the source of process errors and 127 whether natural mortality (or the median for models with process error in natural mortality) 128 was estimated and whether a Beverton-Holt stock recruit relationship was estimated within the EM. Details of each of the operating and EMs are described below. 130

We did not use the log-normal bias-correction feature for process errors or observations
described by (Stock and Miller 2021) for operating and EMs (Li et al. In review). Simulations and model fitting were all carried out on the University of Massachusetts Green
High-Performance Computing Cluster. All code we used to perform the simulation study
and summarize results can be found at https://github.com/timjmiller/SSRTWG/tree/main/
Project\_0/code.

# Operating models

### 138 Population

The population consists of 10 age classes: ages 1 to 10+ and we assume spawning occurs each year 1/4 of the way through the year. The maturity at age was a logistic curve with  $a_{50} = 2.89$  and slope = 0.88 (Figure S1, top left). Weight at age was generated with a von Bertalanffy growth function

$$L_a = L_\infty \left( 1 - e^{-k(a - t_0)} \right)$$

where  $t_0 = 0$ ,  $L_{\infty} = 85$ , and k = 0.3, and a L-W relationship such that

$$W_a = \theta_1 L_a^{\theta_2}$$

where  $\theta_1 = e^{-12.1}$  and  $\theta_2 = 3.2$  (Figure S1, top right).

We assumed a Beverton-Holt stock recruit function with constant pre-recruit mortality parameters for all OMs. All post-recruit productivity components are constant in the NAA and survey catchability process error OMs. Therefore steepness and unfished recruitment are also constant over the time period for those OMs (Miller and Brooks 2021). We specified unfished recruitment =  $R_0 = e^{10}$  and  $F_{\rm MSY} = F_{40\%} = 0.348$  equated to a steepness of 0.69 and a = 0.60 and  $b = 2.4 \times 10^{-5}$  for the

$$N_{1,y} = \frac{aSSB_{y-1}}{1 + bSSB_{y-1}}$$

Beverton-Holt parameterization (Figure S1, bottom right). For OMs without process errors on natural mortality we assumed the rate was assumed 0.2. For OMs with process errors on natural mortality the median natural mortality rate was 0.2.

We used two fishing scenarios for OMs. In the first scenario, the stock experiences overfishing at  $2.5F_{\rm MSY}$  for the first 20 years and fishing at  $F_{\rm MSY}$  for the last 20 years (denoted  $2.5F_{\rm MSY} \rightarrow F_{\rm MSY}$ ). In the second scenario, the stock is fished at  $F_{\rm MSY}$  for the entire time period. The magnitude of the overfishing assumptions is based on average estimates of overfishing for NE groundfish stocks from (Wiedenmann et al. 2019). Legault et al. (2023) also used similar approaches to defining fishing mortality histories for OMs.

We specified initial population abundance at age at the equilibrium distribution fishing at

either  $F=2.5\times F_{\rm MSY}$  or  $F=F_{\rm MSY}$  for the two alternative fishing histories. This implies that, for a deterministic model, the abundance at age would not change from the first year to the next.

For OMs with time-varying random effects for M, steepness is not constant, but we used the same alpha and beta parameters as other OMs this equates to a steepness and R0 at the median of the time series process for M. For OMs with time-varying random effects for fishery selectivity,  $F_{MSY}$  is also not constant however we use the same F history as other OMs which corresponds to Fmsy at the mean selectivity parameters.

#### 169 Fleets

We assumed a single fleet operating year round for catch observations with logistic selectivity for the fleet with  $a_{50}=5$  and slope = 1 (Figure S1, bottom left). This selectivity is was used to define  $F_{\rm MSY}$  for the Beverton-Holt stock recruitment parameters above. We assumed a logistic-normal distribution with no correlation on the multivariate normal scale for the age-composition observations for the fleet.

#### 175 Indices

Two time series of surveys are assumed and observed in numbers rather than biomass for the entire 40 year period with one occurring in the spring (0.25 of each year) and one in the fall (0.75 of each year). Catchability of both surveys are assumed to be 0.1. Like the fishing fleet, we assumed logistic selectivity for both indices with  $a_{50} = 5$  and slope = 1 and a logistic-normal distribution with no correlation on the multivariate normal scale for the age-composition observations.

## $_{182}$ Observation Uncertainty

Standard deviation for log-aggregate catch was 0.1. There were two levels of observation 183 error variance for indices and age composition for both indices and fleet catch. A low 184 uncertainty specification assumed standard deviation of both series of log-aggregate index 185 observations was 0.1 and the standard deviation of the logistic-normal for age composition 186 observations was 0.3 In the high uncertainty specification the standard deviation for log-187 aggregate indices was 0.4 and that for the age composition observations was 1.5. For all 188 EMs, standard deviation for log-aggregate observations was assumed known whereas that 189 for the logistic-normal age composition observations was estimated. 190

### Operating models with random effects on numbers at age

For operating models with random effects on recruitment and(or) survival (R, R+S) we assumed marginal standard deviations for recruitment of  $\sigma_R \in \{0.5, 1.5\}$  and marginal standard deviations for older age classes of  $\sigma_{2+} \in \{0, 0.25, 0.5\}$ . The full factorial combination of these process error assumptions and fishing history (2 levels) and observation error (2 levels) scenarios described above results in 24 different R ( $\sigma_{2+} = 0$ ) and R+S operating models (Table S1).

#### Operating models with random effects on natural mortality

All R+M OMs treat natural mortality constant across age, but with annually varying random effects. WHAM treats natural mortality as a log-transformed parameter

$$\log M_{y,a} = \mu_M + \epsilon_{M,y}$$

that is a linear combination of a mean log-natural mortality parameter that is constant across ages  $\mu_M = \log(0.2)$  and any annual random effects marginally distributed as  $\epsilon_{M,y} \sim$ 

N  $(0, \sigma_M^2)$ . Uncorrelated random effects were also included on recruitment with  $\sigma_R = 0.5$  (hence, R+M). The marginal standard deviations we assumed for log natural mortality random effects were  $\sigma_M \in \{0.1, 0.5\}$  and AR1 autocorrelation parameters of  $\rho_M \in \{0, 0.9\}$ . The full factorial combination of these process error assumptions and fishing history (2 levels) and observation error (2 levels) scenarios described above results in 16 different R+M OMs (Table S2).

## Operating models with random effects on fleet selectivity

MORE SPECIFICS about correlation of random effects? Both selectivity pars? just correlated by year? WHAM treats selectivity parameter s as a logit-transformed parameter

$$\log\left(\frac{p_{s,y} - l_s}{u_s - p_{s,y}}\right) = \mu_s + \epsilon_{s,y}$$

that is a linear combination of a mean  $\mu_s$  and any annual random effects marginally distributed as  $\epsilon_{s,y} \sim \mathrm{N}\left(0,\sigma_{s}^{2}\right)$  where the lower and upper bounds of the parameter  $(l_{s})$  and 213  $u_s$ ) can be specified by the user. All selectivity parameters are either  $a_50$  or slope parame-214 ters and we assume bounds of 0 and 10 for all selectivity parameters for all operating and 215 EMs. The marginal standard deviations we assumed for logit scale random effects were 216  $\sigma_s \in \{0.1, 0.5\}$  and AR1 autocorrelation parameters of  $\rho_s \in \{0, 0.9\}$ . Like R+M OMs, the 217 full factorial combination of these process error assumptions and fishing history (2 levels) 218 and observation error (2 levels) scenarios described above results in 16 different R+Sel OMs 219 (Table S3). 220

## Operating models with random effects on index catchability

Like selectivity parameters, WHAM treats catchability for an index i as a logit-transformed parameter

$$\log\left(\frac{q_{i,y} - l_i}{u_i - q_{i,y}}\right) = \mu_i + \epsilon_{i,y}$$

that is a linear combination of a mean  $\mu_i$  and any annual random effects marginally distributed as  $\epsilon_{i,y} \sim N(0, \sigma_i^2)$  where the lower and upper bounds of the catchability ( $l_i$  and  $u_i$ ) can be specified by the user. Here we assume bounds of 0 and 1000 for all operating and EMs. For operating and EMs with process errors on catchability, the temporal variation is only assumed for the first index. The marginal standard deviations we assumed for logit scale random effects were  $\sigma_i \in \{0.1, 0.5\}$  and AR1 autocorrelation parameters of  $\rho_i \in \{0, 0.9\}$ . Like R+M and R+Sel OMs, the full factorial combination of these process error assumptions and fishing history (2 levels) and observation error (2 levels) scenarios described above results in 16 different R+q OMs (Table S4).

### 233 Estimation models

For each data set simulated from an OM 20 EMs were fit. A total of 32 different EMs were fit across OMs where the subset of 20 depended on the source of process error in the OM 235 (Table S5). The EMs make different assumptions about the source of process error (R+S, 236 R+M, R+Sel, R+q) and whether there is temporal autocorrelation or not, and whether a 237 Beverton-Holt stock recruit function is estimated and the natural mortality rate  $(\mu_M,$  the 238 constant or mean on log scale for R+M EMs) is estimated. For simplicity we refer to the 239 derived estimate  $e^{\mu_M}$  as the median natural mortality rate regardless of whether natural 240 mortality random effects are estimated in the EM. The first 20 EMs in Table S5 were fit to 241 simulate data sets from R and R+S OMs. EMs 5 to 24 in Table S5 were fit to simulate data 242 sets from R+M OMs. EMs 5 to 20 and 25-28 in Table S5 were fit to simulate data sets from R+Sel OMs. Finally, EMs 5 to 20 and 29-32 in Table S5 were fit to simulate data sets from R+q OMs. The maturity at age, weight at age for catch and SSB, and observation error variance of aggregate catch and indices were all assumed known at the true values.

## <sup>247</sup> Measures of reliability

The first measure of reliability we investigated was frequency of convergence when fitting each EM to the simulated data sets. There are various ways to assess convergence of the fit 249 (e.g., Carvalho et al. 2021), but given the importance of estimates of uncertainty when using 250 assessment models in management, we estimated probability of convergence as measured by 251 occurrence of a positive-definite hessian matrix at the optimized negative log-likelihood that 252 could be inverted. However, we also provided results in the Supplementary Materials for 253 the maximum of the absolute values among all gradients for all fits of a given EM to all 254 simulated data sets from a given OM that produced hessian-based standard errors for all 255 estimated fixed effects. This provides an indication of how poor the calculated gradients can 256 be, but still presumably converged adequately enough for parameter inferences. 257

### 258 AIC for model selection

We estimated the probability of selection of each process error model structure (R, R+S, R+M, R+Sel, R+q) using marginal AIC. For a given operating model, we compared AIC for EMs that all made the same assumptions about median natural mortality (known or estimated) and stock recruitment model (Beverton-Holt or none).

We also estimated the probability of correctly selecting between models with Beverton-Holt stock recruit function assumed and models without the stock-recruit function (null model). We made these comparisons between models that otherwise assume the same process error structure as the operating model and both of the compared models either estimate median natural mortality or assume it is known. Contrast in fishing pressure and time series with recruitment at low stock size has been shown to improve estimability of stock-recruitment parameters (Magnusson and Hilborn 2007; Conn et al. 2010). Our preliminary inspections of the proportions of simulations where the correct recruitment model was chosen for a given set of OM factors (including contrast in fishing pressure) indicated generally poor performance of AIC. Therefore, we fit logistic regression models to the indicator of Beverton-Holt models having lower AIC as a function of the log-standard deviation of the true log(SSB) (similar to the log of the coefficient of variation for SSB) since simulations with realized SSB producing low and high recruitments would have larger variation in realized SSB.

All results only condition on whether all of the compared estimating models completed the
optimization process without failure. We did not condition on convergence as defined by a
gradient threshold or invertibity of the hessian because optimization can correctly determine
the the correct likelihood that would indicate poor convergence because variance parameters
may be at the lower bound of zero correctly for models that assume the incorrect process
error structure.

#### 282 Bias

283 For a given model attribute we calculated the relative error

$$RE(\theta_i) = \frac{\widehat{\theta}_i - \theta_i}{\theta_i}$$

from fitting a given estimating model to simulated data set i configured for a given OM where  $\hat{\theta}_i$  and  $\theta_i$  are the estimated and true values for simulation i. We estimated bias as the median of the relative errors across all simulations for a given OM and EM combination. We constructed 95% confidence intervals for the median relative bias using the binomial distribution approach as in Miller and Hyun (2018) and Stock and Miller (2021). We present results for terminal years SSB and recruitment, Beverton-Holt stock recruit parameters (a and b), and median natural mortality rate. Results for terminal year fishing mortality are strongly

291 negatively correlated with those for SSB and provided in the supplementary materials.

Similar to the AIC results, bias results only condition on whether the estimating model completed the optimization process without failure. We did not condition on convergence as defined by a gradient threshold or invertibity of the hessian because the optimized model can provide reliable estimation of SSB, F, M, and stock recruit parameters whether or not the model was able to estimate non-zero random effects. In practice, the model would be reconfigured to remove unnecessary process errors and produce otherwise equivalent parameter estimates.

### Mohn's ho

We estimated Mohn's  $\rho$  for SSB, fully-selected fishing mortality, and recruitment for each EM (Mohn 1999). We estimated 7 peels for each EM. We calculated median 95% confidence intervals for Mohn's  $\rho$  using the same methods as that for relative bias. Similar to the other results, retrospective results only condition on whether all of the peels of a given estimating model completed the optimization process without failure. We did not condition on convergence as defined by a gradient threshold or invertibity of the hessian because the optimized model can provide reliable estimation of SSB, fishing mortality rates, and recruitment whether or not the model was able to estimate non-zero random effects.

# Results

# 309 Convergence performance

For R and R+S OMs, convergence generally declines for most EMs when the median natural mortality rate is estimated and/or the Beverton-Holt stock recruit relationship is estimated even when the process error assumptions of the EMs and OMs match (Figure 1, A). When

there was high observation error and constant fishing pressure (change from  $2.5F_{\rm MSY}$  to  $F_{\rm MSY}$ ), convergence of all EM process error configurations other than R EMs was poor 314 for R OMs ( $\sigma_{2+}=0$ ) regardless of whether median natural mortality and stock-recruit 315 relationships were estimated. Convergence of R EMs was high for all R and R+S OMs except 316 when there was high observation error and constant fishing pressure and median natural 317 mortality and stock-recruit relationships were estimated. R+S EMs fit to R OMs exhibited 318 poor convergence regardless of whether natural mortality or a stock-recruit relationship was 319 estimated. R+S EMs fit to R+S OMs had highest convergence rates when there was contrast 320 in fishing pressure and low observation error. Convergence rates were high for all EMs when 321 fit to data from R+S OMs with lower observation error except those where median natural 322 mortality and/or stock-recruit relationships were estimated. 323

Convergence of all EMs fitted to R+M OMs was best when the OMs had higher natural mortality process error variability, low observation error, and contrast in fishing pressure (Figure 1, B). R+M EMs that estimated autocorrelation of process errors had poor convergence for R+M OMs when there was low natural mortality process error variability whether or not there was autocorrelation of the simulated process errors. R+S EMs fitted to data generated from R+M OMs always converged poorly whether median natural mortality and the Beverton-Holt stock-recruit relationship were estimated.

R+S EMs, in particular, converged poorly when fit to data generated from R+Sel OMs with lower selectivity process error variability or higher observation error (Figure 1, C). R+Sel EMs generally converged better than other EMs for R+Sel OMs with higher process error variability, lower observation error, and contrast in fishing pressure regardless of whether median natural mortality or a stock recruit relationship was estimated.

For R+q OMs, convergence of R+q EMs is generally better than that of other EMs when there is contrast in fishing history (Figure (1, D). Convergence of R+S EMs is generally worse than that of all other EMs across all OMs whether or not median natural mortality or a stock recruit relationship is estimated. Again, convergence probability generally declines for all EMs when median natural mortality or a stock recruit relationship is estimated.

We found a wide range of maximum absolute values of gradients for models that converged as defined here (Figure S2). The largest value observed for a given EM and OM combination was typically  $< 10^{-3}$ , but many converged models had values greater than 1. For many OMs, EMs that assumed the correct process error type and did not estimate median natural mortality or the Beverton-Holt stock-recruit relationship produced the lowest gradient values.

## AIC performance for process error structure

Marginal AIC accurately determines the correct process error assumptions in EMs when data are generated from R and R+S OMs, regardless of whether median natural mortality or a stock recruit relationship is estimated (Figure 2, A). Adding estimation of median natural mortality or a stock recruit relationship separately has a negligible effect on the accuracy of determining the correct process error assumption. When both are estimated, there is a noticeable reduction in accuracy when OMs have a constant fishing history, observation error is low and larger variability in recruitment process errors.

For R+M OMs, marginal AIC only accurately determined the correct process error model and correlation structure when observation error was low and variability in natural mortality process errors was high (Figures 2, B). Estimating the median natural mortality rate reduced the accuracy of AIC for OMs that assumed natural mortality process errors were independent. For OMs with poor accuracy, AIC most frequently selected EMs with process errors in catchability (R+q) or selectivity (R+Sel). Selection of R+S EMs was typically unlikely.

Marginal AIC most accurately determined the correct source of process error and correlation structure for R+Sel OMs with low observation error (Figures 2, C). When there was low variability in selectivity process errors and high observation error, R+q or R+S EMs were more likely to have the best AIC. Whether median natural morality or stock recruit relationships were estimated appeared to have little effect on the performance of AIC.

Marginal AIC most accurately determined the correct source of process error and correlation structure for R+q OMs with high variability in catchability process errors (Figures 2,D). The worst accuracy occurred for OMs with low variability in catchability process errors and high observation error. However, in these OMs, the marginal AIC accurately determined the correct source of process error (but not correlation structure) except when EMs estimated both median natural morality and the stock recruit relationships and OMs assumed a constant fishing pressure.

Our comparisons of model performance condition on assuming the true process error config-

# AIC performance for the stock-recruit relationship

373

uration is known (EM and OM process error types match). Broadly, we found generally poor accuracy of AIC in selecting models assuming a Beverton-Holt stock recruit function over the null model without an assumed stock-recruit relationship for all OMs. However, we also 376 found increased accuracy of AIC in determining the Beverton-Holt stock-recruit relationship with greater variation in spawning biomass generated in the simulated populations for nearly 378 every OM (Figure 3). 379 With R and R+S process error assumptions, probability of AIC accuracy for the B-H stock-380 recruit relationship as a function of SSB variability were greatest for OMs where there was 381 greater process variability in survival, lower process variability in recruitment and contrast in 382 fishing pressure (Figure 3, A). However, the largest variation in SSB (Figure 3, rug on x-axis) 383 occurred in OMs with larger recruitment variability ( $\sigma_R = 1.5$ ; Figure 3, A, right column 384 group). The weakest overall accuracy of AIC for the Beverton-Holt stock-recruit function 385 was when there was constant fishing pressure (Figure 3, A, second and fourth rows). The largest effect of simultaneously estimating median natural mortality was when there was

OMs had high observation error and constant fishing pressure (Figure 3, A, second row).

For R+M OMs, probability of AIC accuracy for the Beverton-Holt stock-recruit relationship was greatest for OMs where there was contrast in fishing pressure (Figure 3, B). The largest variation in SSB occurred in OMs with larger natural mortality variability and temporal AR1 correlation of the annual random effects ( $\sigma_M = 1.5$ ,  $\rho_M = 0.9$ ; Figure 3, B, right-most column) where AIC accuracy was also high even when there was constant fishing pressure. Interestingly, there seemed to be little difference in AIC accuracy whether there was lower or higher observation error or whether the EM simultaneously estimated median natural mortality.

For R+Sel OMs, contrast in fishing pressure over time is the primary mechanism creating variation in SSB and these are the OMs where AIC accuracy for the Beverton-Holt stock recruit relationship was greatest (Figure 3, C, first and third rows). There appears to be little effect of variability or correlation of selectivity process errors or whether median natural mortality was estimated on AIC accuracy.

Like the R+Sel OMs, the greatest accuracy for AIC in selecting the Beverton-Holt stockrecruit relationship occurred for R+q OMs where there was contrast in fishing pressure over time which is also where there was the greatest variation in SSB (Figure 3, D, first and third rows). There were some differences in effects of SSB variation on AIC accuracy when OMs had high observation error and constant fishing pressure. Across all of the OMs in Figure 3, the only estimated decline in AIC accuracy with SSB variation occurred when R+q OMs had high observation error and constant fishing pressure and greater variation in catchability process errors that also had AR1 autocorrelation.

### 410 Bias

### Spawning stock biomass and recruitment

For R OMs ( $\sigma_{2+} = 0$ ), there was no indication of bias (95% confidence intervals included 0) in terminal year SSB for any of the estimating models regardless of process error assumptions,

except when no SR assumption was made, recruitment variability was low, and there was contrast in fishing mortality and high observation error (Figure 4, A). However, errors in 415 terminal SSB estimates were highly variable when median natural mortality was estimated 416 and there was constant fishing pressure and high observation error (Figure 4, A, second row). 417 For R+S OMs, the EMs with matching process error assumption generally produced un-418 baised estimation of terminal SSB except when median natural mortality was estimated 419 and there was high observation error. In R+S OMs with low observation error, EMs with 420 incorrect process error assumptions typically provided biased estimation of terminal year 421 SSB. Estimating the Beverton-Holt stock-recruit relationship had little discernible effect on 422 bias of terminal year SSB estimation whereas estimating median M tended to produce more 423 variability in errors in terminal SSB estimation similar to R OMs.

For R+ M OMs with low variability in natural mortality process errors, low observation error 425 and contrast in fishing motality over time all EMs produced low variability in SSB estimation 426 error that indicated unbiasedness (Figure 4, B, third row). However, larger variability in 427 natural mortality process errors increased bias of EMs without the correct process error 428 type. Estimating median natural mortality increased variability of SSB estimation error 429 particularly for OMs with high observation error and constant fishing pressure over time and 430 increased bias in SSB estimation for many R+M OMs. Like R and R+S OMs, estimating a 431 stock-recruit relationship had little discernible effect on SSB bias. 432

For R+Sel OMs, there was no evidence of bias for any EMs when variability in selectivity process error and observation error was low, and with contrast in fishing mortality (Figure 4, C). The largest bias occurred for any EMs that estimated median natural mortality when the OMs had high observation error, constant fishing pressure, and greater variability in selectivity process errors ( $\sigma_{\text{Sel}} = 0.5$ ) or low selectivity process errors ( $\sigma_{\text{Sel}} = 0.1$ ) and low observation error. However, there was no evidence of bias of matching R+Sel EMs when observation error was low and variation in selectivity process errors was larger, whether

median natural mortality was estimated or not (Figure 4, C, third row). We only observed an effect of estimating the Beverton-Holt Stock recruit relationship for R+Sel OMs that had high observation error and contrast in fishing pressure where estimating the relationship produced less biased SSB estimation for many EMs (Figure 4, C, top row).

All EMs fit to data from R+q OMs with low observation error and contrast in fishing 444 pressure exhibited little evidence of bias in terminal SSB estimation except for R+M EMs 445 when there was no AR1 correlation in catchability process errors (Figure 4, D). Many EMs 446 also performed well in R+q OMs with low observation error, but no contrast in fishing 447 pressure. For R+q OMs with high observation error and contrast in fishing pressure, EMs 448 that estimated the Beverton-Holt stock recruit function exhibited less SSB bias than those 449 that did not. Estimating median natural mortality in the EMs only resulted in much more 450 variable SSB estimation errors when there was no contrast in fishing pressure (Figure 4, D, first and third rows).

For all OM process error types, relative errors in terminal year recruitment were generally more variable than SSB, but effects of R and R+S OM and EM attributes on bias (i.e, negative or positive or none) were similar (Figure S5, A). Furthermore, for EM configurations where bias in terminal bias in SSB was evident, median relative errors in recruitment often indicated stronger bias in recruitment of the same sign.

#### 458 Beverton-Holt parameters

In R and R+S OMs, EMs with the correct assumptions about process errors, provided the least biased estimation of Beverton-Holt stock-recruit relationship parameters when there was low observation error, a change in fishing pressure over time, and lower variability of recruitment process errors, and higher variability survival process errors, and there was little effect of estimating natural mortality (Figure S6, A). For other R and R+S OMs, estimating natural mortality often resulted in less biased estimation of stock-recruit parameters. There

was generally large variability in relative errors of the stock-recruit parameter estimates, but the lowest variability occurred with low variability in recruitment and little or no variability in survival process errors ( $\sigma_{2+} \in \{0, 0.25\}$ ), and contrast in fishing pressure. Across al R and R+S OMs, relative errors for the a parameter were often less variable than those for b.

In R+M OMs, the most accurate estimation of stock-recruit parameters for all EM process error assumptions occurred when there was a change in fishing pressure combined with either low variability in natural mortality process errors and high observation error or vice versa (Figure S6, B). Relative to the R, and R+S OMs, there was even less effect of estimating median natural mortality on estimation bias for the stock-recruit relationship parameters, but similarly there was generally large variability of the parameter estimates.

Bias for stock-recruit parameters was very strong and variability in relative errors was greatest for most EMs fit to R+Sel OMs with constant fishing pressure (Figure S6, C). Less bias
in parameter estimation occurred for OMs with a change in fishing pressure over time and
of those OMs and the most accurate estimation occurred for EMs with process error types
that matched the OMs and the OMs had low observation error, contrast in fishing mortality and more variable and uncorrelated selectivity process errors. There was little effect of
estimating natural mortality on relative errors for stock recruit parameters.

Like R+Sel OMs, relative errors in stock-recruit parameters were less variable for most EM process error types when OMs had contrast in fishing pressure and those OMs that also had lower observation error generally had the least variability (Figure S6, D). The worst accuracy of stock-recruit parameter estimation regardless of EM type occured when R+q OMs had low observation error and constant fishing pressure (Figure S6, D, fourth row).

### 487 Median natural mortality rate

Median natural mortality rate is estimated accurately by all EM process error types for all R OMs except those where observation error is high and fishing pressure is constant where

variability in relative errors is high (Figure S7, A,  $\sigma_{2+} = 0$ ). For R+S OMs estimation of median natural mortality rate is most accurate when observation error is low and there is contrast in fishing pressure and the EM process error type is correct. There is little effect of estimating stock-recruit relationships on the bias for median natural mortality.

For R+M OMs, median natural mortality was estimated most accurately, regardless of EM process error type, when OMs had a change in fishing pressure and low observation error (Figure S7, B). However, those R+M OMs that also had greatest variability in AR1 correlated natural mortality process errors only had unbiased estimation when the EM process error type was correct. Like R, and R+S OMs, there was little effect of estimating stock-recruit relationships on the bias for median natural mortality.

All EM process error types produced accurate estimation of median natural mortality rate for 500 R+Sel OMs that had contrast in fishing pressure, low observation error, and low selectivity 501 process error variability (Figure S7, C). When selectivity process error variability increased 502 the incorrect EM process errors produce more biased estimation of median natural mortality 503 rate. The least accurate estimation occurred for all EM process error types when observation 504 error was high and there was constant fishing pressure. Again, like other OM process error 505 types, there was little effect of estimating stock-recruit relationships on the bias for median 506 natural mortality. 507

Like R+Sel OMs, all EM process error types produced accurate estimation of median natural mortality rate when fit to R+q OMs with contrast in fishing pressure, low observation error and low catchability process error variability (Figure S7, D). Most EM process error types produced biased estimation of median natural mortality when R+q OMs had high observation error and constant fishing pressure. Again, there was no discernible effect of estimating stock-recruit relationships on bias.

## Mohn's $\rho$

Mohn's  $\rho$  for SSB was small in absolute value for all R and R+S OMs, regardless of EM 515 process error types, and whether median natural mortality rate or stock-recruit relationships 516 were estimated (Figure 5, A). The strongest retrospective patterns occurred in OMs with 517 the largest survival process error variability, high observation error, and contrast in fishing 518 pressure, but only for EMs without the correct process error type and where median natural 519 mortality rate was assumed known. For R+M, R+Sel, and R+q OMs, Mohn's  $\rho$  was also 520 small in absolute value, but median values were all closer to 0 than the largest values in the R 521 and R+S OMs (Figure 5,B-D). For these OMs, there was no noticeable effect of estimation of 522 median natural mortality rate or stock recruit relationship on Mohn's  $\rho$  for any EM process 523 error types. 524 Mohn's  $\rho$  for recruitment was small in absolute value for all R OMs with low variability in

525 recruitment process errors, regardless of EM process error type, and whether median natural 526 mortality rate or stock-recruit relationships were estimated (Figure S9, A). However, R and 527 R+S OMs with greater recruitment process variability and higher observation error had 528 median Mohn's  $\rho$  for recruitment greater than zero for most EMs even when the EM 520 process error type was correct. In R+S OMs with lower observation error, EMs with the 530 correct process error type exhibited better median Mohn's  $\rho$  close to 0 than EMs with the 531 incorrect process error type. For R+M, R+Sel, and R+q OMs, results for Mohn's  $\rho$  for 532 recruitment are similar to those for SSB, but the range in median values and variation in 533 Mohn's  $\rho$  values for a given OM are generally larger for recruitment (Figure S9, B-D). 534

# Discussion

## 536 Convergence

Convergence results can be useful for understanding how lack of convergence in applications 537 to real data might direct the practitioner to which alternative random effects configurations 538 is more appropriate. Therefore, the type of convergence that we might use as a diagnostic 539 is important. It is common during assessment model fitting to check that the maximum 540 absolute gradient component is less than some threshold, but there is no standard. However, 541 if the Hessian were checked for positive definiteness conditional on a satisfactory gradient, 542 some models may be excluded that in fact have an invertible hessian. We found the hessian 543 at the optimized log-likelihood can often be invertible when the gradient approach with a 544 sensible criterion (e.g.,  $10^{-3}$ ) would indicate lack of convergence. Another factor affecting the convergence criteria is numerical accuracy. With the process er-546 rors modeled as random effects we are optimizing an (Laplace) approximation of the marginal log-likelihood and therefore hessians and gradients are also with regard to the approximation. Furthermore, numerical methods are used to calculate the hessian (Optimhess?), and also 549 to invert it for variance estimation. All of these approximations are affecting the gradients 550 and invertibility of the hessian (although likely worse for some models than others), which, 551 along with our results, suggests at least investigating whether the hessian is positive definite 552 when the calculated gradients are not terribly large. Often high gradients and/or lack of 553 positive definite hessian result when parameter estimates are at their bounds (Carvalho et 554 al. 2021), and this also applies to variance or autocorrelation parameters for random effects 555 with state-space assessment models. 556 Even when hessian-based variance can be accomplished, parameters that are poorly estimated will have extremely large variances. This further inspection can lead to a more appropriate and often more parsimonious model configuration where the problematic parameters are not estimated. For example, process error variance parameters that are estimated close to 0 indicates that the random effects are estimated to have little or no variability and removing these process errors is warranted. Generally, our results suggest we can expect lower probability of convergence of state-space assessment models when estimating natural mortality or stock-recruit relationships because of the difficulty distinguishing these parameters from others being estimated in assessment model with data typically available.

## 66 AIC

567

process errors on recruitment and survival (R and R+S). Fitting models to other OMs rarely 568 preferred R+S EMs and R and R+S EMs were nearly always selected for the matching OMs. 560 For other sources of process error, accuracy of AIC was improved to useful levels when there 570 was larger variability in the process errors and/or lower observation error. 571 We found AIC to performs weakly in determining stock recruit relationships unless there 572 is large contrast in SSB and low variation in recruitment process errors. Although we did not compare models with alternative stock-recruit relationships (e.g., Ricker and Beverton-574 Holt), we do not expect AIC to perform any better distinguishing between relationships. 575 Our finding that AIC tended to choose simpler recruitment models in most cases contrasts 576

Among the process error configurations we used in OMs, we found AIC to be accurate for

with the noted bias in AIC for more complex models (Shibata 1976; Katz 1981; Kass and Raftery 1995), but whereas those findings apply to more common fitting of models to raw data, state-space models account for observation error and estimate process errors in latent variables.

Our results comport well with those of de Valpine and Hastings (2002) which found AIC could not distinguish among state-space stock-recruit models that were fit just to SSB and recruitment "observations" and Britten et al. which found AIC could not reliably distinguish alternative environmental effects on stock-recruit parameters. Miller et al. (2016) did find

AIC to prefer a stock-recruit relationship with environmental effects for SNEMA yellowtail flounder, but there was a large change in stock size estimated over time and flatfish are wellobserved by the NEFSC bottom trawl survey used for an index in the assessment model.

### $_{588}$ Bias

As we might have expected, bias in all parameters and assessment output was generally better 580 with lower observation error. Estimation of stock-recruit relationship parameters was not 590 reliable in any of the OM-EM combinations, but estimation of median natural mortality was 591 feasible in many OM scenarios with temporal contrast in fishing pressure. For OMs where 592 there was bias in natural mortality due to high observation error (R+Sel and R+q OMS), 593 estimating the stock-recruit relationship seemed to alleviate the bias. However, estimation 594 of median natural mortality can cause large differences between the true and estimated SSB 595 (that may or may not be unbiased on average) when there is less contrast in fishing pressure 596 over time and higher observation error. 597

# 598 Retrospective patterns

Incorrect process error assumptions for EMs did not produce strong retrospective patterns for SSB for any OMs whether median natural mortality and a stock-recruit relationship were estimated or not, but some weak retrospective patterns occur when observation error was high and there was contrast in fishing pressure. Retrospective patterns tended to be more variable for recruitment and can be large even when the EM is correct. Therefore, we recommend emphasis on inspection of retrospective patterns primarily for SSB and F, but further research on retrospective patterns in other assessment model parameters, management quantitities, and projections may be beneficial (Brooks and Legault 2016).

## 607 Conclusions

- All together what do results mean? E.g., when do you get good convergence and retrospective
- patterns and high/none bias? When do you get good convergence and no retros and no/big
- 610 bias?
- 611 For R+S OMs with low observation error, temporal contrast in fishing pressure, and large
- variation in survival process errors (3rd row), some EMs with the incorrect process error
- assumptions, estimate median M to be 0, and SSB is also underestimated, but Mohn's  $\rho$  for
- SSB is better than when M is known. Fortunately, AIC would not choose the wrong EM
- process error type in these situations.
- 616 When R+S is preferred AMONG THE TYPES CONSIDERED HERE, it probably means
- its the correct process error .

# 618 Acknowledgements

This work was funded by NOAA Fisheries Northeast Fisheries Science Center.

# References

- <sup>621</sup> Aeberhard, W.H., Flemming, J.M., and Nielsen, A. 2018. Review of State-Space Models
- for Fisheries Science. Annual Review of Statistics and Its Application 5(1): 215–235.
- doi:10.1146/annurev-statistics-031017-100427.
- Auger-Méthé, M., Field, C., Albertsen, C.M., Derocher, A.E., Lewis, M.A., Jonsen, I.D.,
- and Mills Flemming, J. 2016. State-space models' dirty little secrets: Even simple
- linear gaussian models can have estimation problems. Scientific reports 6(1): 26677.
- doi:10.1038/srep26677.
- Auger-Méthé, M., Newman, K., Cole, D., Empacher, F., Gryba, R., King, A.A., Leos-
- Barajas, V., Mills Flemming, J., Nielsen, A., Petris, G., and others. 2021. A guide to
- state-space modeling of ecological time series. Ecological Monographs 91(4): e01470.
- doi:10.1002/ecm.1470.
- Britten, G., Brooks, E.N., and Miller, T.J. Reliability of stock-recruitment function estima-
- tion in state space assessment models.
- <sup>634</sup> Brooks, E.N., and Legault, C.M. 2016. Retrospective forecasting evaluating performance
- of stock projections for new england groundfish stocks. Canadian Journal of Fisheries
- and Aquatic Sciences **73**(6): 935–950. doi:10.1139/cjfas-2015-0163.
- 637 Cadigan, N.G. 2016. A state-space stock assessment model for northern cod, including under-
- reported catches and variable natural mortality rates. Canadian Journal of Fisheries and
- Aquatic Sciences **73**(2): 296–308. doi:10.1139/cjfas-2015-0047.
- 640 Carvalho, F., Winker, H., Courtney, D., Kapur, M., Kell, L., Cardinale, M., Schirripa, M.,
- Kitakado, T., Yemane, D., Piner, K.R., Maunder, M.N., Taylor, I., Wetzel, C.R., Doering,
- K., Johnson, K.F., and Methot, R.D. 2021. A cookbook for using model diagnostics in
- integrated stock assessments. Fisheries Research 240: 105959. doi:https://doi.org/10.
- 1016/j.fishres.2021.105959.
- <sup>645</sup> Conn, P.B., Williams, E.H., and Shertzer, K.W. 2010. When can we reliably estimate the
- productivity of fish stocks? Canadian Journal of Fisheries and Aquatic Sciences **67**(3):

- 511–523. doi:10.1139/F09-194.
- <sup>648</sup> Cronin-Fine, L., and Punt, A.E. 2021. Modeling time-varying selectivity in size-structured
- assessment models. Fisheries Research 239: 105927. Elsevier.
- de Valpine, P., and Hastings, A. 2002. Fitting population models incorporating process noise
- and observation error. Ecological Monographs **72**(1): 57–76.
- Du Pontavice, H., Miller, T.J., Stock, B.C., Chen, Z., and Saba, V.S. 2022. Ocean model-
- based covariates improve a marine fish stock assessment when observations are limited.
- 654 ICES Journal of Marine Science **79**(4): 1259–1273. doi:10.1093/icesjms/fsac050.
- <sup>655</sup> Fisch, N., Shertzer, K., Camp, E., Maunder, M., and Ahrens, R. 2023. Process and sampling
- variance within fisheries stock assessment models: Estimability, likelihood choice, and the
- consequences of incorrect specification. ICES Journal of Marine Science 80(8): 2125–
- 658 2149. Oxford University Press.
- <sup>659</sup> Fleischman, S.J., Catalano, M.J., Clark, R.A., and Bernard, D.R. 2013. An age-structured
- state-space stock-recruit model for pacific salmon (oncorhynchus spp.). Canadian Jour-
- nal of Fisheries and Aquatic Sciences **70**(3): 401–414. NRC Research Press.
- 662 Hoyle, S.D., Maunder, M.N., Punt, A.E., Mace, P.M., Devine, J.A., and A'mar, Z.T. 2022.
- Preface: Developing the next generation of stock assessment software. Fisheries Research
- **246**: 106176.
- 665 Kass, R.E., and Raftery, A.E. 1995. Bayes factors. Journal of the American Statistical
- Association **90**(430): 773–795. doi:10.1080/01621459.1995.10476572.
- Katz, R.W. 1981. On some criteria for estimating the order of a Markov chain. Technomet-
- rics **23**(3): 243-249. doi:10.1080/00401706.1981.10486293.
- 669 Knape, J. 2008. Estimability of density dependence in models of time series data. Ecology
- 89(11): 2994–3000. doi:10.1890/08-0071.1.
- Legault, C.M., and Palmer, M.C. 2016. In what direction should the fishing mortality target
- change when natural mortality increases within an assessment? Canadian Journal of
- Fisheries and Aquatic Sciences **73**(3): 349–357. NRC Research Press.

- Legault, C.M., and Restrepo, V.R. 1999. A flexible forward age-structured assessment program. Col. Vol. Sci. Pap. ICCAT **49**(2): 246–253.
- 676 Legault, C.M., Wiedenmann, J., Deroba, J.J., Fay, G., Miller, T.J., Brooks, E.N., Bell,
- R.J., Langan, J.A., Cournane, J.M., Jones, A.W., and Muffley, B. 2023. Data-rich but
- model-resistant: An evaluation of data-limited methods to manage fisheries with failed
- age-based stock assessments. Canadian Journal of Fisheries and Aquatic Sciences 80(1):
- 680 27-42. doi:10.1139/cjfas-2022-0045.
- 681 Li, C., Deroba, J.J., Berger, A.M., Goethel, D.R., Langseth, B.J., Schueller, A.M., and
- Miller, T.J. In review. Random effects on numbers-at-age transitions implictly 2 account
- for movement dynamics and improve 3 performance within a state-space stock assessment.
- 684 Canadian Journal of Fisheries and Aquatic Sciences.
- 685 Li, C., Deroba, J.J., Miller, T.J., Legault, C.M., and Perretti, C. In reviewIn review. Guid-
- ance on bias-correction of log-normal random effects and observations in state-space
- assessment models. For Canadian Journal of Fisheries and Aquatic Sciences, State-space
- special collection.
- 689 Li, C., Deroba, J.J., Miller, T.J., Legault, C.M., and Perretti, C.T. 2024. An evalu-
- ation of common stock assessment diagnostic tools for choosing among state-space
- models with multiple random effects processes. Fisheries Research 273: 106968.
- doi:10.1016/j.fishres.2024.106968.
- 693 Liljestrand, E.M., Bence, J.R., and Deroba, J.J. 2024. The effect of process variability and
- data quality on performance of a state-space stock assessment model. Fisheries Research
- **275**: 107023. Elsevier.
- Magnusson, A., and Hilborn, R. 2007. What makes fisheries data informative? Fish and
- Fisheries 8(4): 337–358. doi:https://doi.org/10.1111/j.1467-2979.2007.00258.x.
- 698 Methot, R.D., and Wetzel, C.R. 2013. Stock synthesis: A biological and statistical frame-
- work for fish stock assessment and fishery management. Fisheries Research 142: 86–99.
- doi:10.1016/j.fishres.2012.10.012.

- Miller, T.J., and Brooks, E.N. 2021. Steepness is a slippery slope. Fish and Fisheries 22(3):
- 634–645. doi:10.1111/faf.12534.
- Miller, T.J., Hare, J.A., and Alade, L. 2016. A state-space approach to incorporating envi-
- ronmental effects on recruitment in an age-structured assessment model with an appli-
- cation to Southern New England yellowtail flounder. Canadian Journal of Fisheries and
- 706 Aquatic Sciences **73**(8): 1261–1270. doi:10.1139/cjfas-2015-0339.
- Miller, T.J., and Hyun, S.-Y. 2018. Evaluating evidence for alternative natural mortality
- and process error assumptions using a state-space, age-structured assessment model.
- Canadian Journal of Fisheries and Aquatic Sciences **75**(5): 691–703. doi:10.1139/cjfas-
- 710 2017-0035.
- Miller, T.J., and Stock, B.C. 2020. The Woods Hole Assessment Model (WHAM). Available
- from https://timjmiller.github.io/wham/.
- Mohn, R. 1999. The retrospective problem in sequential population analysis: An investi-
- gation using cod fishery and simulated data. ICES Journal of Marine Science **56**(4):
- 473–488. doi:10.1006/jmsc.1999.0481.
- Nielsen, A., and Berg, C.W. 2014. Estimation of time-varying selectivity in stock assessments
- using state-space models. Fisheries Research 158: 96–101. doi:10.1016/j.fishres.2014.01.014.
- Pedersen, M.W., and Berg, C.W. 2017. A stochastic surplus production model in continuous
- time. Fish and Fisheries **18**(2): 226–243. doi:10.1111/faf.12174.
- Polansky, L., De Valpine, P., Lloyd-Smith, J.O., and Getz, W.M. 2009. Likelihood ridges
- and multimodality in population growth rate models. Ecology **90**(8): 2313–2320.
- doi:10.1890/08-1461.1.
- Punt, A.E. 2023. Those who fail to learn from history are condemned to repeat it: A per-
- spective on current stock assessment good practices and the consequences of not following
- them. Fisheries Research **261**: 106642.
- Punt, A.E., Hurtado-Ferro, F., and Whitten, A.R. 2014. Model selection for selectivity in
- fisheries stock assessments. Fisheries Research 158: 124–134. Elsevier.

- Shibata, R. 1976. Selection of the order of an autoregressive model by akaike's information criterion. Biometrika **63**(1): 117–126. doi:10.1093/biomet/63.1.117.
- Stewart, I.J., and Monnahan, C.C. 2017. Implications of process error in selectivity for approaches to weighting compositional data in fisheries stock assessments. Fisheries Research 192: 126–134. Elsevier.
- Stock, B.C., and Miller, T.J. 2021. The Woods Hole Assessment Model (WHAM): A general state-space assessment framework that incorporates time- and age-varying processes via random effects and links to environmental covariates. Fisheries Research **240**: 105967. doi:https://doi.org/10.1016/j.fishres.2021.105967.
- Stock, B.C., Xu, H., Miller, T.J., Thorson, J.T., and Nye, J.A. 2021. Implementing twodimensional autocorrelation in either survival or natural mortality improves a state-space assessment model for Southern New England-Mid Atlantic yellowtail flounder. Fisheries Research 237: 105873.
- Szuwalski, C.S., Ianelli, J.N., and Punt, A.E. 2018. Reducing retrospective patterns in stock
   assessment and impacts on management performance. ICES Journal of Marine Science
   75(2): 596–609. Oxford University Press.
- Thorson, J.T., and Minto, C. 2015. Mixed effects: A unifying framework for statistical modelling in fisheries biology. ICES Journal of Marine Science **72**(5): 1245–1256. doi:10.1093/icesjms/fsu213.
- Trijoulet, V., Fay, G., and Miller, T.J. 2020. Performance of a state-space multispecies model: What are the consequences of ignoring predation and process errors in stock assessments? Journal of Applied Ecology **57**(1): 121–135. Wiley Online Library.
- Wiedenmann, J., Free, C.M., and Jensen, O.P. 2019. Evaluating the performance of datalimited methods for setting catch targets through application to data-rich stocks: A case study using northeast U.S. Fish stocks. Fisheries Research **209**(1): 129–142. doi:https: //doi.org/10.1016/j.fishres.2018.09.018.

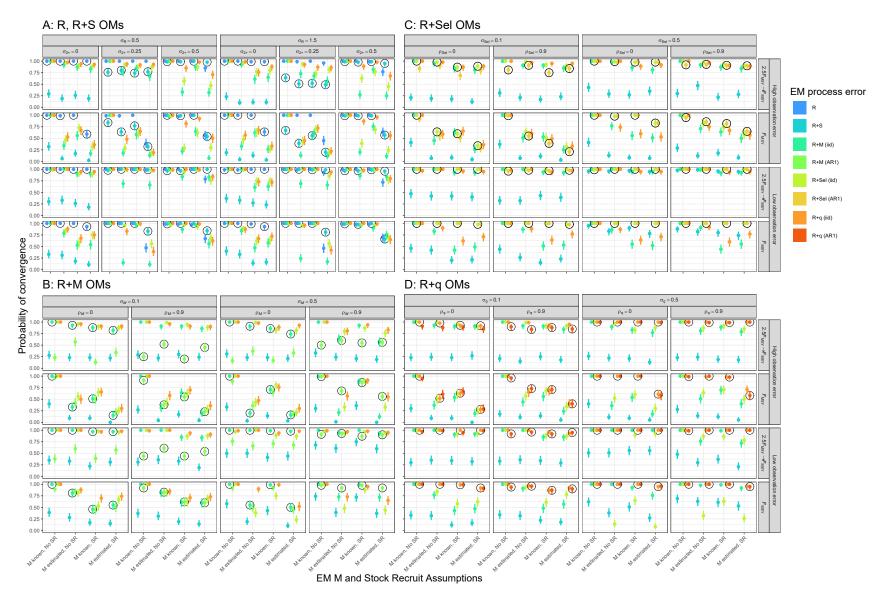


Fig. 1. Estimated probability of fits providing hessian-based standard errors for EMs assuming alternative process error (colored points and lines), and median natural mortality (estimated or known) and Beverton-Holt stock recruit functions (estimated or not; along x-axis) when fitted to operating models that have R and R+S (A), R+Sel (B), R+M (C), or R+q (D) process error structures. Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

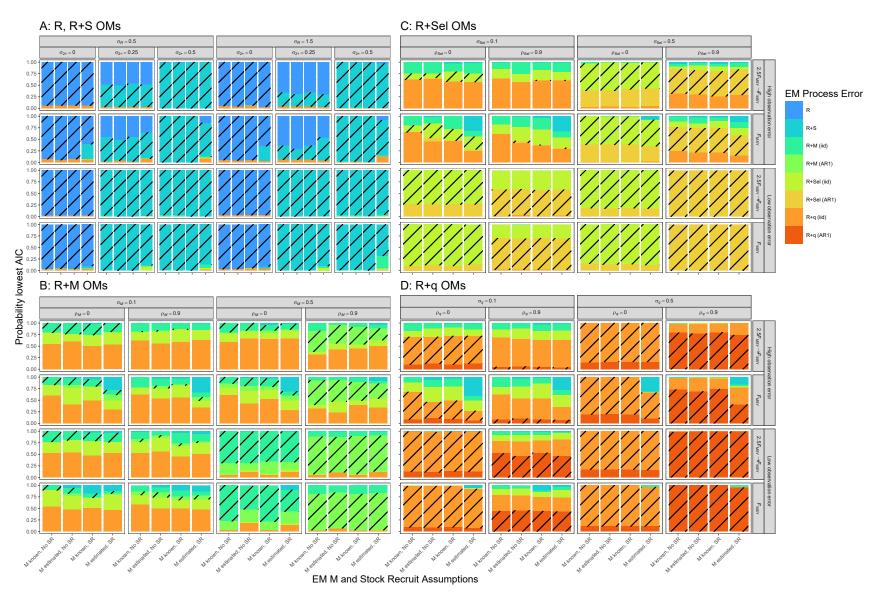


Fig. 2. Estimated probability of lowest AIC for EMs assuming alternative process error structures (colored bars) conditional on alternative assumptions for median natural mortality (estimated or known) and Beverton-Holt stock recruit functions (estimated or not; along x-axis) when fitted to operating models that have R and R+S (A), R+Sel (B), R+M (C), or R+q (D) process error structures. Striped bars indicate results where the EM process error structure matches that of the operating model.

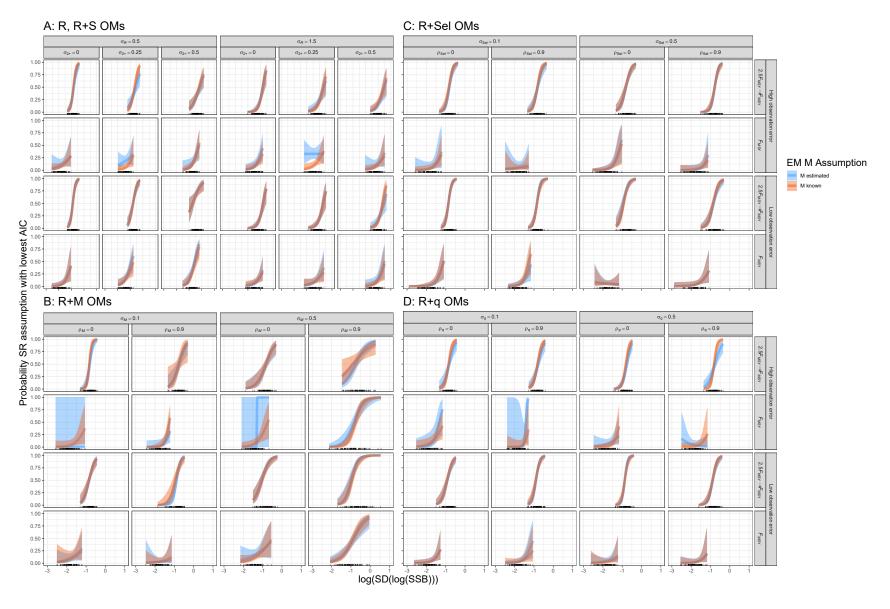


Fig. 3. Estimated probability of lowest AIC from logistic regression on the log-standard deviation of the true log(SSB) in each simulation for estimating model with Beverton-Holt stock recruit functions, rather than the otherwise equivalent EM without the stock recruit function. Results are conditional on alternative assumptions for median natural mortality (estimated or known) and on EMs having the correct process error structure: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Rug along x-axis denotes SD(log(SSB)) values for each simulation amd polygons represent 95% confidence intervals.

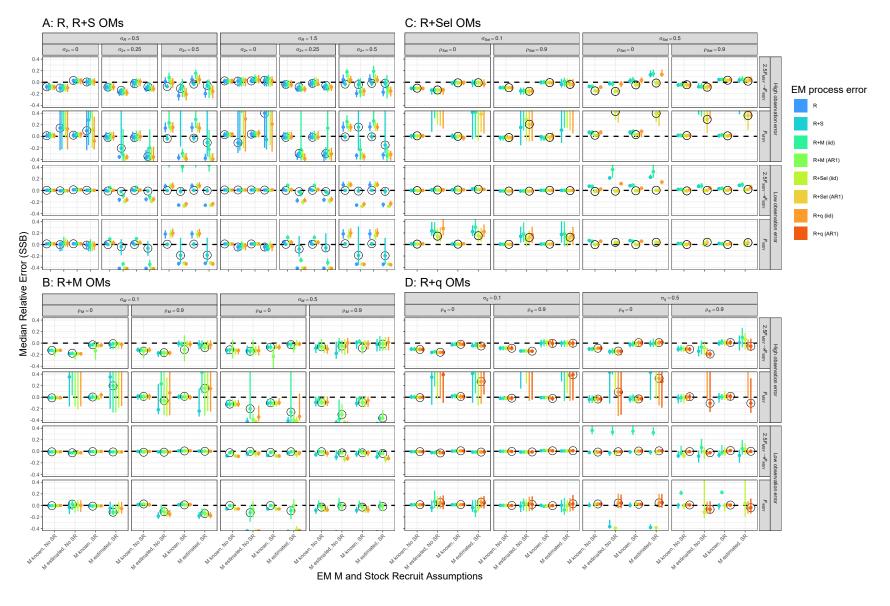


Fig. 4. Median relative error of terminal year SSB for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Diamond shaped points denote results where the EM process error assumption matches that of the operating model. Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

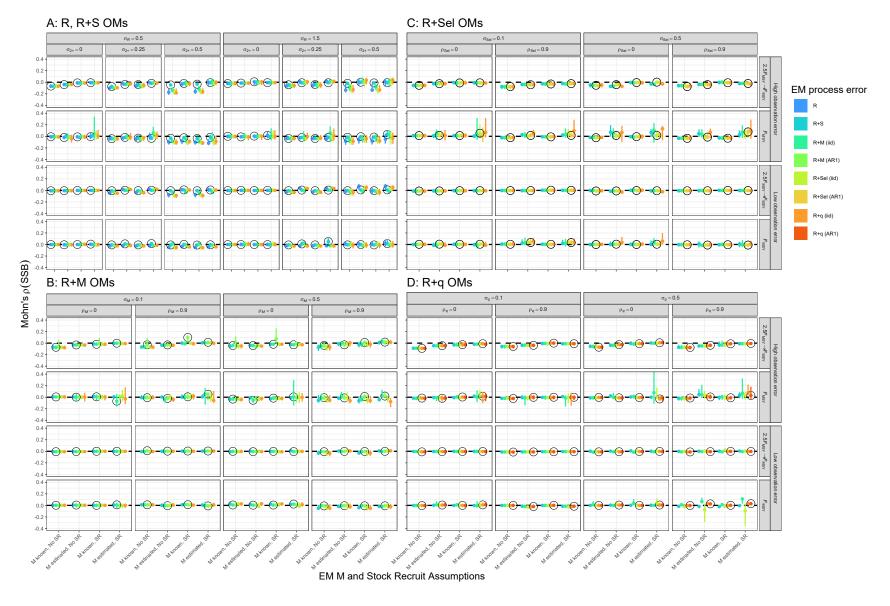


Fig. 5. Median Mohn's rho for SSB for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

## <sup>754</sup> Supplementary Materials

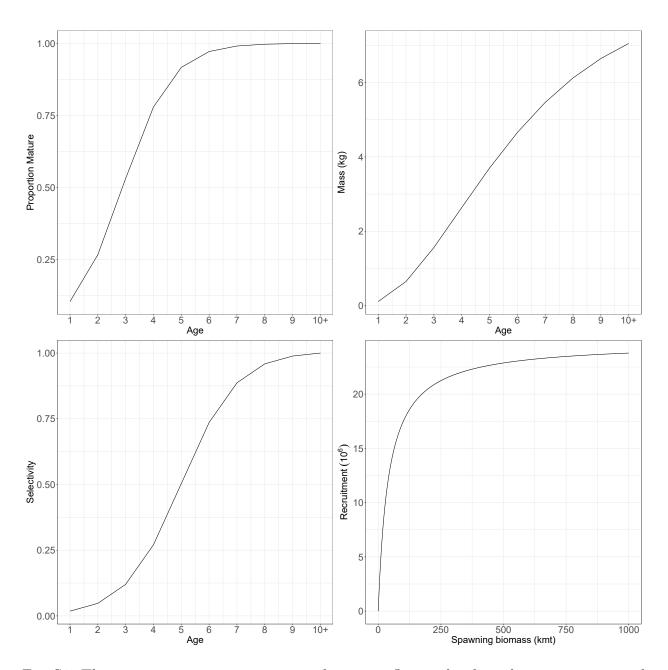


Fig. S1. The proportion mature at age, weight at age, fleet and index selectivity at age, and Beverton-Holt stock recruit relationship assumed for the population in all operating models. For operating models with random effects on fleet selectivity, this represents the selectivity at the mean of the random effects.

Model	$\sigma_R$	$\sigma_{2+}$	Fishing History	Observation Uncertainty
$NAA_1$	0.5		$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$NAA_2$	1.5		$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_3$	0.5	0.25	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_4$	1.5	0.25	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_5$	0.5	0.50	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_6$	1.5	0.50	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_7$	0.5		$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_8$	1.5		$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_9$	0.5	0.25	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_{10}$	1.5	0.25	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_{11}$	0.5	0.50	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_{12}$	1.5	0.50	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$NAA_{13}$	0.5		$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{14}$	1.5		$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{15}$	0.5	0.25	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{16}$	1.5	0.25	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{17}$	0.5	0.50	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{18}$	1.5	0.50	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{19}$	0.5		$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{20}$	1.5		$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$NAA_{21}$	0.5	0.25	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$\mathrm{NAA}_{22}$	1.5	0.25	$F_{ m MSY}$	Index SD = 0.4, Age composition SD = $1.5$
$NAA_{23}$	0.5	0.50	$F_{ m MSY}$	Index SD = 0.4, Age composition SD = $1.5$
$NAA_{24}$	1.5	0.50	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$

Table S2. Distinguishing characteristics of the operating models with random effects on natural mortality. Standard deviations (SD) are for log-normal distributed indices and logistic normal distributed age composition observations (fleet and indices). Fishing mortality changes after year 20 (of 40) for fishing histories where fishing mortality is not constant. For AR1 process errors,  $\sigma$  is defined for the marginal distribution of the processes.

Model	$\sigma_R$	$\sigma_{M}$	$ ho_M$	Fishing History	Observation Uncertainty
$M_1$	0.5	0.1	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_2$	0.5	0.5	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_3$	0.5	0.1	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_4$	0.5	0.5	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_5$	0.5	0.1	0.0	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_6$	0.5	0.5	0.0	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_7$	0.5	0.1	0.9	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_8$	0.5	0.5	0.9	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$M_9$	0.5	0.1	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$M_{10}$	0.5	0.5	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$M_{11}$	0.5	0.1	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$M_{12}$	0.5	0.5	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$M_{13}$	0.5	0.1	0.0	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$M_{14}$	0.5	0.5	0.0	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$M_{15}$	0.5	0.1	0.9	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$M_{16}$	0.5	0.5	0.9	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$

Table S3. Distinguishing characteristics of the operating models with random effects on selectivity. Standard deviations (SD) are for log-normal distributed indices and logistic normal distributed age composition observations (fleet and indices). Fishing mortality changes after year 20 (of 40) for fishing histories where fishing mortality is not constant. For AR1 process errors,  $\sigma$  is defined for the marginal distribution of the processes.

Model	$\sigma_R$	$\sigma_{ m Sel}$	$ ho_{ m Sel}$	Fishing History	Observation Uncertainty
$\mathrm{Sel}_1$	0.5	0.1	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_2$	0.5	0.5	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_3$	0.5	0.1	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_4$	0.5	0.5	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_5$	0.5	0.1	0.0	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_6$	0.5	0.5	0.0	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_{7}$	0.5	0.1	0.9	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_8$	0.5	0.5	0.9	$F_{ m MSY}$	Index SD = $0.1$ , Age composition SD = $0.3$
$\mathrm{Sel}_9$	0.5	0.1	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$\mathrm{Sel}_{10}$	0.5	0.5	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$\mathrm{Sel}_{11}$	0.5	0.1	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$\mathrm{Sel}_{12}$	0.5	0.5	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$\mathrm{Sel}_{13}$	0.5	0.1	0.0	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$\mathrm{Sel}_{14}$	0.5	0.5	0.0	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$\mathrm{Sel}_{15}$	0.5	0.1	0.9	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$Sel_{16}$	0.5	0.5	0.9	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$

Table S4. Distinguishing characteristics of the operating models with random effects on catchability. Standard deviations (SD) are for log-normal distributed indices and logistic normal distributed age composition observations (fleet and indices). Fishing mortality changes after year 20 (of 40) for fishing histories where fishing mortality is not constant. For AR1 process errors,  $\sigma$  is defined for the marginal distribution of the processes.

Model	$\sigma_R$	$\sigma_q$	$\rho_q$	Fishing History	Observation Uncertainty
$q_1$	0.5	0.1	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$q_2$	0.5	0.5	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$q_3$	0.5	0.1	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.1$ , Age composition SD = $0.3$
$q_4$	0.5	0.5	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = 0.1, Age composition SD = $0.3$
$q_5$	0.5	0.1	0.0	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$q_6$	0.5	0.5	0.0	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$q_7$	0.5	0.1	0.9	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$q_8$	0.5	0.5	0.9	$F_{ m MSY}$	Index SD = 0.1, Age composition SD = $0.3$
$q_9$	0.5	0.1	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$q_{10}$	0.5	0.5	0.0	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$q_{11}$	0.5	0.1	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$q_{12}$	0.5	0.5	0.9	$2.5F_{\mathrm{MSY}} \to F_{\mathrm{MSY}}$	Index SD = $0.4$ , Age composition SD = $1.5$
$q_{13}$	0.5	0.1	0.0	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$q_{14}$	0.5	0.5	0.0	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$q_{15}$	0.5	0.1	0.9	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$
$q_{16}$	0.5	0.5	0.9	$F_{ m MSY}$	Index SD = $0.4$ , Age composition SD = $1.5$

Table S5. Distinguishing characteristics of the estimating models.

Model	Recruitment model	Mean $M$	Process error assumption
$\mathrm{EM}_1$	Mean recruitment	0.2	Recruitment ( $\sigma_R$ estimated)
$\mathrm{EM}_2$	Beverton-Holt	0.2	Recruitment ( $\sigma_R$ estimated)
$EM_3$	Mean recruitment	Estimated	Recruitment ( $\sigma_R$ estimated)
$\mathrm{EM}_4$	Beverton-Holt	Estimated	Recruitment ( $\sigma_R$ estimated)
$\mathrm{EM}_5$	Mean recruitment	0.2	Recruitment and survival $(\sigma_R, \sigma_{2+} \text{ estimated})$
$\mathrm{EM}_6$	Beverton-Holt	0.2	Recruitment and survival $(\sigma_R, \sigma_{2+} \text{ estimated})$
$\mathrm{EM}_7$	Mean recruitment	Estimated	Recruitment and survival ( $\sigma_R$ , $\sigma_{2+}$ estimated)
$\mathrm{EM}_8$	Beverton-Holt	Estimated	Recruitment and survival $(\sigma_R, \sigma_{2+} \text{ estimated})$
$EM_9$	Mean recruitment	0.2	Recruitment and uncorrelated natural mortality ( $\sigma_R,  \sigma_M$ estimated, $\rho_M = 0$ )
$\mathrm{EM}_{10}$	Beverton-Holt	0.2	Recruitment and uncorrelated natural mortality ( $\sigma_R,  \sigma_M$ estimated, $\rho_M = 0$ )
$\mathrm{EM}_{11}$	Mean recruitment	Estimated	Recruitment and uncorrelated natural mortality ( $\sigma_R$ , $\sigma_M$ estimated, $\rho_M=0$ )
$\mathrm{EM}_{12}$	Beverton-Holt	Estimated	Recruitment and uncorrelated natural mortality ( $\sigma_R,\sigma_M$ estimated, $\rho_M=0$ )
$\mathrm{EM}_{13}$	Mean recruitment	0.2	Recruitment and uncorrelated fleet selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ estimated, $\rho_{\rm Sel}=0$ )
$\mathrm{EM}_{14}$	Beverton-Holt	0.2	Recruitment and uncorrelated fleet selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ estimated, $\rho_{\rm Sel}=0$ )
$EM_{15}$	Mean recruitment	Estimated	Recruitment and uncorrelated fleet selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ estimated, $\rho_{\rm Sel}=0$ )
$\mathrm{EM}_{16}$	Beverton-Holt	Estimated	Recruitment and uncorrelated fleet selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ estimated, $\rho_{\rm Sel}=0$ )
$\mathrm{EM}_{17}$	Mean recruitment	0.2	Recruitment and uncorrelated catchability (spring index) $(\sigma_R, \sigma_q)$ estimated, $\rho_q = 0$
$\mathrm{EM}_{18}$	Beverton-Holt	0.2	Recruitment and uncorrelated catchability (spring index) $(\sigma_R, \sigma_q)$ estimated, $\rho_q = 0$
$EM_{19}$	Mean recruitment	Estimated	Recruitment and uncorrelated catchability (spring index) $(\sigma_R,  \sigma_q  \text{estimated},  \rho_q = 0)$
$\mathrm{EM}_{20}$	Beverton-Holt	Estimated	Recruitment and uncorrelated catchability (spring index) $(\sigma_R,  \sigma_q  \text{estimated},  \rho_q = 0)$
$\mathrm{EM}_{21}$	Mean recruitment	0.2	Recruitment and AR1 natural mortality ( $\sigma_R$ , $\sigma_M$ , $\rho_M$ estimated)
$\mathrm{EM}_{22}$	Beverton-Holt	0.2	Recruitment and AR1 natural mortality ( $\sigma_R$ , $\sigma_M$ , $\rho_M$ estimated)
$\mathrm{EM}_{23}$	Mean recruitment	Estimated	Recruitment and AR1 natural mortality $(\sigma_R, \sigma_M, \rho_M \text{ estimated})$
$\mathrm{EM}_{24}$	Beverton-Holt	Estimated	Recruitment and AR1 natural mortality ( $\sigma_R$ , $\sigma_M$ , $\rho_M$ estimated)
$\mathrm{EM}_{25}$	Mean recruitment	0.2	Recruitment and AR1 selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ , $\rho_{\rm Sel}$ estimated)
$\mathrm{EM}_{26}$	Beverton-Holt	0.2	Recruitment and AR1 selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ , $\rho_{\rm Sel}$ estimated)
$\mathrm{EM}_{27}$	Mean recruitment	Estimated	Recruitment and AR1 selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ , $\rho_{\rm Sel}$ estimated)
$\mathrm{EM}_{28}$	Beverton-Holt	Estimated	Recruitment and AR1 selectivity ( $\sigma_R$ , $\sigma_{\rm Sel}$ , $\rho_{\rm Sel}$ estimated)
$EM_{29}$	Mean recruitment	0.2	Recruitment and AR1 catchability (spring index) $(\sigma_R,  \sigma_q,  \rho_q  \text{estimated})$
${\rm EM}_{30}$	Beverton-Holt	0.2	Recruitment and AR1 catchability (spring index) ( $\sigma_R$ , $\sigma_q$ , $\rho_q$ estimated)
$\mathrm{EM}_{31}$	Mean recruitment	Estimated	Recruitment and AR1 catchability (spring index) ( $\sigma_R$ , $\sigma_q$ , $\rho_q$ estimated)
$EM_{32}$	Beverton-Holt	Estimated	Recruitment and AR1 catchability (spring index) $(\sigma_R, \sigma_q, \rho_q)$ estimated)



Fig. S2. The maximum of the absolute values of all gradient values for all fits that provided hessian-based standard errors across all simuated data sets of a given OM configuration (A: R and R+S, B: R+M, C: R+Sel, or D: R+q). Results are conditional on EM fits with alternative process error type (colored points and lines), median natural mortality (estimated or known) and recruitment assumptions (Beverton-Holt stock recruit or not). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.



Fig. S3. Probability of estimating models providing maximum absolute values of gradients less than  $10^{-6}$  assuming alternative process error (colored points and lines), and median natural mortality (estimated or known) and Beverton-Holt stock recruit functions (estimated or not; along x-axis) when fitted to operating models that have R and R+S (A), R+Sel (B), R+M (C), or R+q (D) process error structures. Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

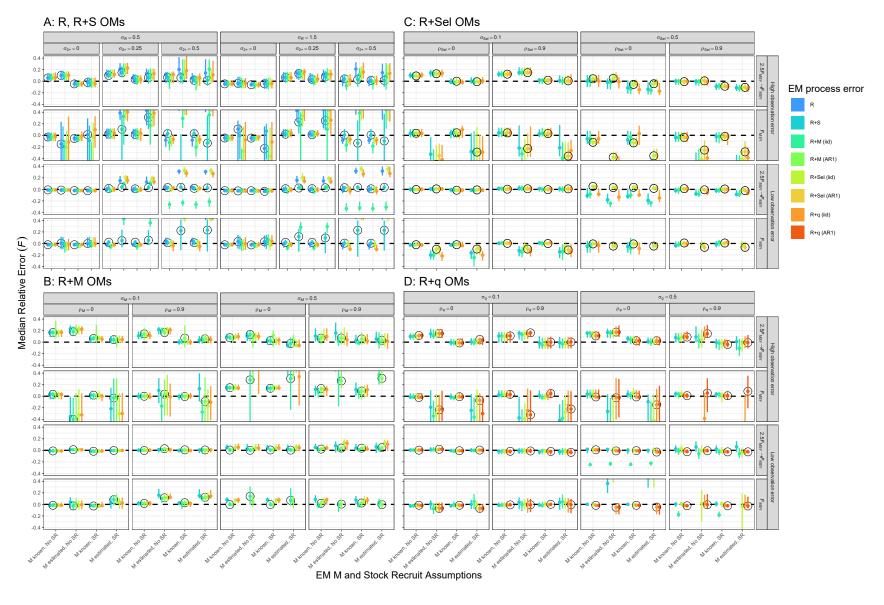


Fig. S4. Median relative error of terminal year fully-selected fishing mortality for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

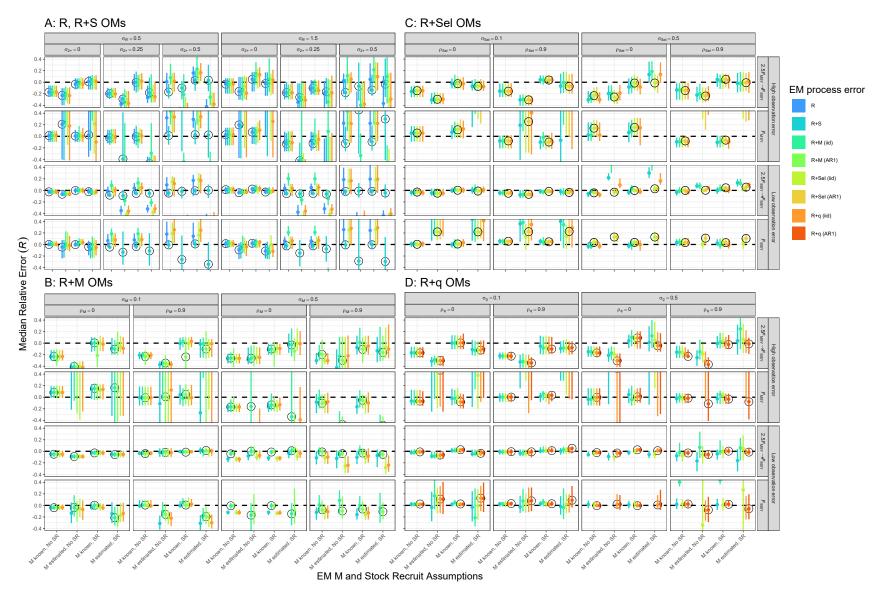


Fig. S5. Median relative error of terminal year recruitment for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

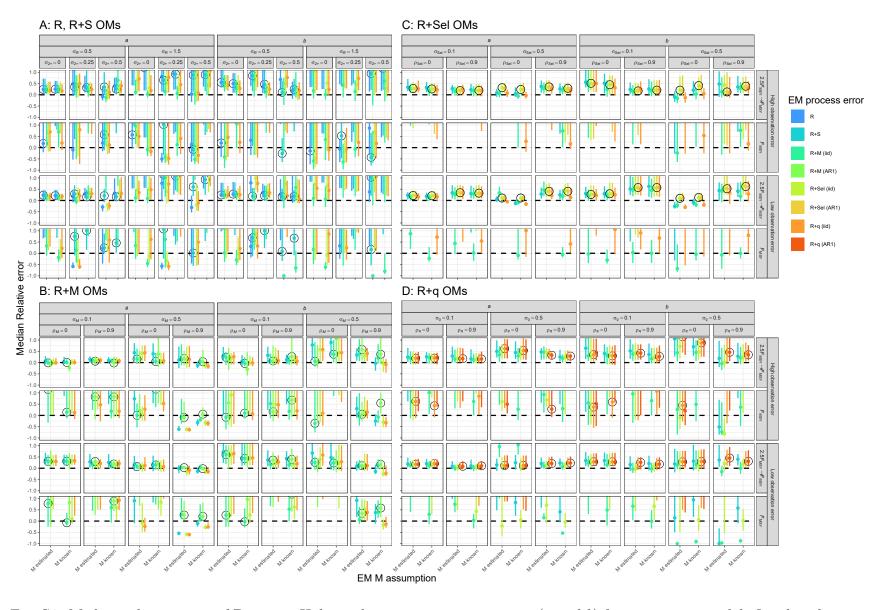


Fig. S6. Median relative error of Beverton-Holt stock recruitment parameters (a and b) for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

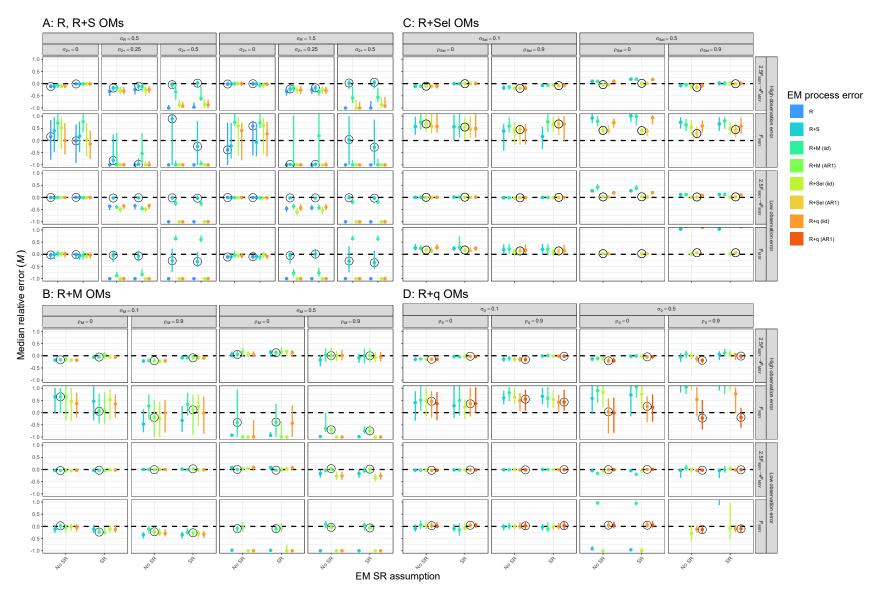


Fig. S7. Median relative error of median natural mortality for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

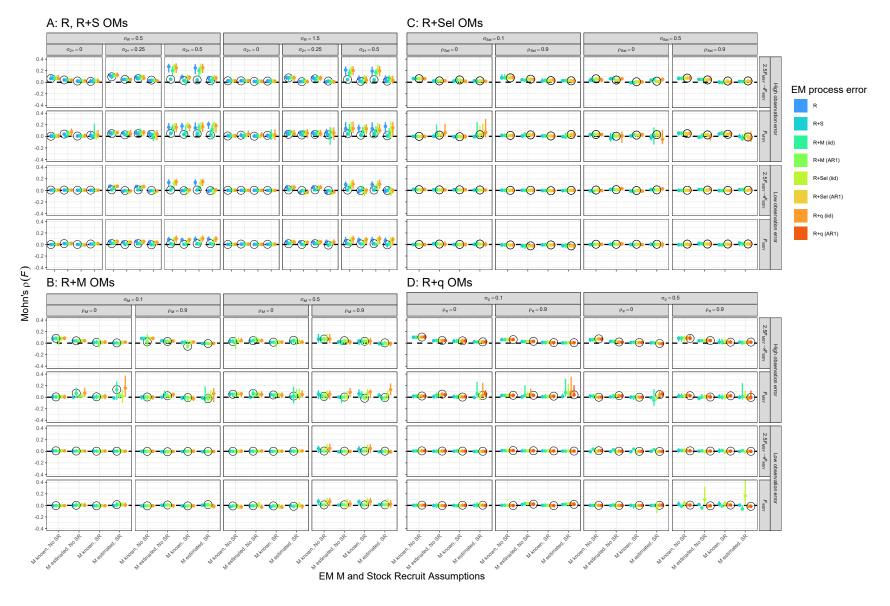


Fig. S8. Median Mohn's  $\rho$  of fishing mortality averaged over all age classes for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.

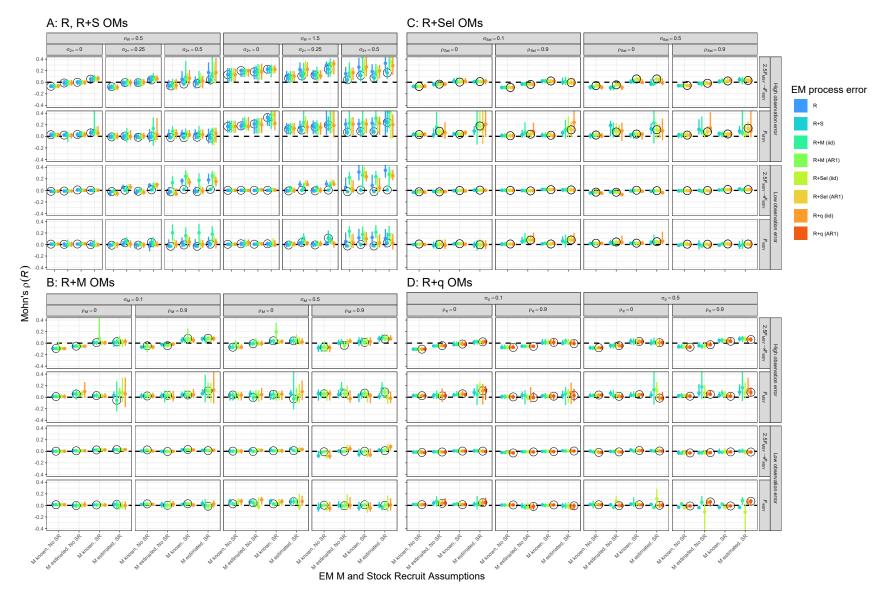


Fig. S9. Median Mohn's  $\rho$  of recruitment for estimating models fitted to data sets simulated with alternative process error structures: R and R+S (A), R+Sel (B), R+M (C), or R+q (D). Circled values indicate results where the EM process error structure matches that of the operating model and vertical lines represent 95% confidence intervals.