



Centre Oceanogràfic
de Balears (COB-IEO)



Spatio-temporal fishing risk of large pelagic fish in the Mediterranean Sea

Master Thesis

Tom Leven

02312706

At the Faculty of Sciences

Ghent University

Main supervisor: Dr. Pilar Tugores

Co-supervisors: Dr. Diego Alvarez-Berastegui & Dr. Miguel Cabanellas-Reboreda

Executive Summary

The executive summary goes here.

Abstract

Abstract goes here.

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1 Introduction

Large pelagic fishes (LPF) including tunas, swordfish, and pelagic sharks, hold profound ecological and economic importance globally. They are characterized by their highly migratory nature, and their roles as apex predators. Their large-scale migrations are driven by the fact that food resources and suitable breeding grounds are far apart from each other (Fromentin and Powers, 2005; Tracey et al., 2023; Relano and Pauly, 2022a). Through these migrations, LPF form an important biological link between distant areas (trans-ocean migrations), and between coastal and pelagic production (Beamish et al., 2005). As apex predators, they also strongly influence the structure and function of ecosystems and intensive fishing can lead to cascading top-down effects (Baum and Worm, 2009; Young et al., 2015). They also have a high social and global importance for food security. Global catches of LPF, i.e., tuna and tuna-like species (mainly Thunnini, Xiphidae, and Istiophoridae), amounted to a record 8.3 million t reported landings in 2022, and tuna fisheries alone are worth 40 billion dollars annually (FAO, 2024; McKinney et al., 2020).

In the Mediterranean Sea (Fig. 1), bluefin tuna (BFT; *Thunnus thynnus*), albacore tuna (*Thunnus alalunga*), and swordfish (SWO; *Xiphias gladius*) are the most important LPF (Papaconstantinou and Farrugio, 2000). In 2023, these species alone accounted for approximately 63% of reported catches for all tuna and tuna-like species (in total around 60 thousand tonnes; ICCAT, 2025). Apart from their economic importance, these three species also form an important part of the culture and identity of numerous coastal communities in the Mediterranean, as they have been exploited there since ancient times (Addis et al., 2012; Usai, 2022; Di Natale et al., 2005; ?). The bluefin tuna fishery for example, is the oldest known industrial fishery in the world (Di Natale, 2012).

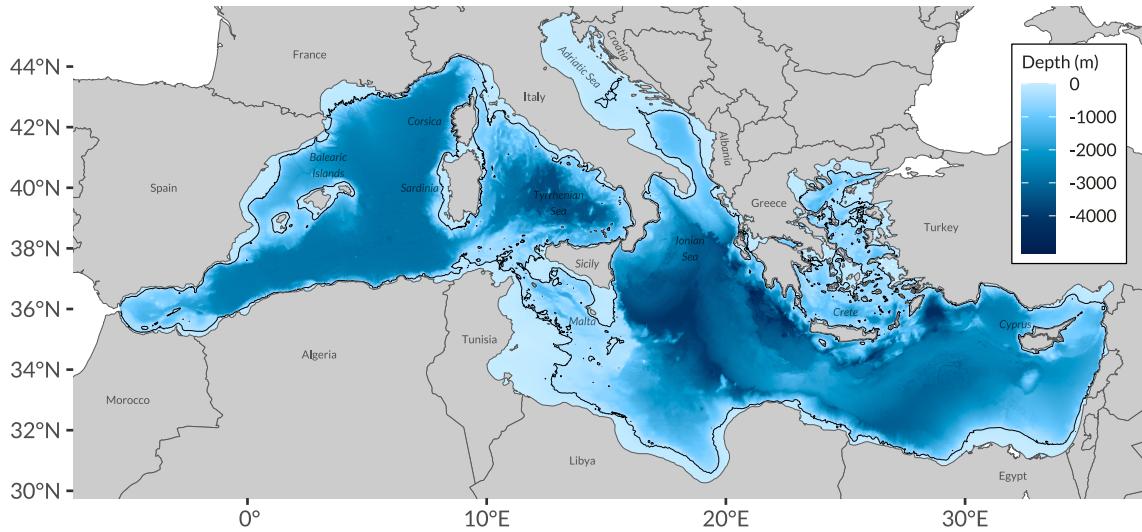


Figure 1: The Mediterranean Sea, study area for this analysis. The black line indicates the 200m isobar.

To sustainably manage these fish stocks and the ecosystems they occur in, it is essential to accurately estimate the location and intensity of fishing pressure (Piet et al., 2006; Russo et al., 2013; Maina et al., 2016). This information is increasingly needed for marine spatial planning, marine protected area designation, conservation of target and non-target species, and ecosystem-based fisheries management (Vespe et al., 2016; Sales Henriques et al., 2024; Tidd et al., 2015; Gerritsen and Lordan, 2010; James et al., 2018; Shea et al., 2023; Bell et al., 2025). Most fisheries management systems for example aim to regulate one or more of the following: catches, gear type, fishing location, seasonality, and effort. To achieve this, information on who, when, where and how much fishing is occurring is necessary (Orofino et al., 2023).

Traditionally, mostly logbooks are used to monitor where and when a vessel is fishing (Hintzen et al., 2012). Compiling logbook data is the responsibility of each skipper and is mandatory for vessels fishing in EU waters (European Commission, 1993). Information of catches is recorded daily at a spatial resolution of 1° longitude and 0.5° latitude (size of ICES areas). Since the 2000s however, there

has been an increase in the usage of automatic monitoring tools. Since 2005, Vessel Monitoring Systems (VMS) are mandatory for fishing vessels >15 m in the EU, to meet the increasing need for fine-scale spatial and temporal information from commercial fleets for compliance with the ecosystem approach to fisheries management (European Commission, 2003; Gerritsen and Lordan, 2010). In the EU, VMS usually have a ping rate between 1-2 hours, aiming at a compromise between appropriate resolution and costs (Shepperson et al., 2017). This data has however traditionally been confined to government regulators and other fisheries authorities, limiting its usage, particularly in the case of migratory species that travel between different countries' jurisdictions (Orofino et al., 2023; Hinz et al., 2013). More recently, the usage of Automatic Identification Systems (AIS) has emerged as a promising tool to monitor the spatio-temporal extent of fisheries and for fisheries Monitoring, Control and Surveillance (MCS; Zhang et al., 2022). Globally, AIS is compulsory for large vessels (over 300 tons), while in the EU it is mandatory for fishing vessels \geq 15m in length (European Commission, 2011). Originally designed as a collision avoidance tool, it allows analysis at higher spatial and temporal resolution due to the high ping rate of up to 2 seconds (Taconet et al., 2019; Kontas, n.d.). There are however some handicaps that can limit the coverage of this technology. This includes for example intentional turning-off, reception quality, and difficulty of discerning individual messages in areas with very high vessel density (Taconet et al., 2019). The increasing coverage and data availability however, enable the analysis of more complex fisheries behaviours (Natale et al., 2015) and AIS data is also publicly accessible, allowing for a fine-scale spatio-temporal assessment of fisheries between different countries' jurisdictions.

Most studies investigating fishing effort in the Mediterranean Sea have looked at the spatial footprint of bottom trawlers as shown by Ferrà et al. (2018) and Marsaglia et al. (2024). However, monitoring the fisheries on LPF in this area will also provide important information for their management. The International

Commission for the Conservation of Atlantic Tunas (ICCAT) is responsible for the sustainable management of LPF in the Mediterranean. Spatio-temporal data on fishing effort are incorporated into stock assessments for bluefin tuna, swordfish, and albacore tuna (ICCAT, 2024b,c,a). However, the data reported by contracting parties are often coarse in resolution (aggregated at $5^\circ \times 5^\circ$ or more rarely $1^\circ \times 1^\circ$) which limits the capacity to detect fine-scale fishing patterns. Incorporating fishing pressure derived from AIS, could provide an alternative for cross-validation of reported effort data and analysis on a finer spatial scale.

Monitoring fisheries of LPF is also highly relevant in terms of climate change, of which the effects have recently been proposed to be integrated in the species' management (ICCAT, 2022). For this, members of the ICCAT '*Subcommittee of Ecosystems and Bycatch*' have created the '*TunaMed Observatory*' with the aim to '*identify and monitor the variability of environmental processes in the Mediterranean Sea that affects the ecology of large pelagic fishes - with special attention on tunas-, and to investigate the potential role of climate change on this variability*' (ICCAT, 2023). As the spatial footprint of fishing is closely linked to the distribution of target species, changes in the distribution of fishing could thus be indicators of habitat change and climate-driven shifts in species range (?).

One of the most important gear types used in the Mediterranean for the capture of LPF are drifting longlines (FAO, 2025; ICCAT, 2024a,b,c). This gear type is however also associated with high levels of bycatch which includes species of sea turtles, pelagic sharks, and seabirds (Carpentieri et al., 2021). Thus, demonstrating the utility of fine scale fishing data for the conservation of non-target species for example through combining both species distribution models with this data (Welch et al., 2024). Apart from longlines, purse seine nets are used in the Mediterranean mainly for the capture of BFT, targetting schools of these fish during spawning. In the Mediterranean, this method is not associated to large amounts of bycatch but can have big impacts on the targetted fish as it is extracting whole schools (Car-

pentieri et al., 2021; ICCAT, 2024b)

To address current gaps in the monitoring of large pelagic fisheries in the Mediterranean, this thesis uses fine-scale data on fishing activities derived from AIS and provided by Global Fishing Watch (GFW) to examine spatio-temporal patterns in fishing over the past decade (2015-2024). This approach enables analysis of the whole Mediterranean Sea area, that would not be possible when relying on other effort estimates like VMS and logbooks. The study focuses on the two main gear types targeting LPF, namely drifting longlines and tuna purse seines, to determine '*hotspots*' of fishing activity corresponding to high risk areas for LPF, analyse temporal trends, and the relationship of fishing with depth and distance to port. More specifically, we aim to answer:

1. Where and when is fishing activity for large pelagic species most intense and have spatial patterns shifted between 2015 and 2024?
2. What are the seasonal patterns of fishing activity by gear type and do temporal patterns differ interannually?
3. To what extent does AIS capture the full scope of fishing activity in the Mediterranean?
4. What is the spatial relationship between fishing hours and environmental features and how could we use these as indicators of habitat changes?

2 Material and Methods

2.1 Data sources

2.1.1 Fishing hours

Data on fishing hours was obtained from Global Fishing Watch (GFW). This non-profit organization provides a global dataset of estimated fishing activity derived from AIS data (Global Fishing Watch, 2025). They process data from over 190,000 unique AIS transmitters, each assigned a unique Maritime Mobile Service Identity (MMSI). These AIS devices broadcast a vessel's location as frequently as every 2 seconds (Kontas, n.d.; Taconet et al., 2019). Along with the exact location, each AIS transmission includes a timestamp, speed, and heading of the vessel. GFW then analyses these positional data points, to infer fishing activity, via two different Convolutional Neural Networks (CNN's), which are described in detail in Kroodsma et al. (2018).

A first CNN classifies fishing vessels into one of sixteen fishing gear categories (Fig. 2) and predicts vessel characteristics such as length, tonnage, and engine power. This deep-learning model is trained on a dataset of vessels matched to official vessel registries which are *known* fishing vessels. Vessels without gear information in the registries used are assigned a gear if their movement patterns resemble those of a known vessel class. A second CNN classifies every obtained AIS position as either fishing or non-fishing based on characteristic fishing movement (Kroodsma et al., 2018; Fig. 3). These individual fishing events are then aggregated into grid cells spanning either 0.1° or 0.01° on a side. For the present analysis, 0.1° resolution was chosen, covering the study area (Mediterranean Sea; Fig 1) in sufficient detail.

Unique vessels are identified based on their MMSI number and the dataset con-

tains information on the vessel's registration and flag country. The gear class estimated by the CNN is also compared with the information from different vessel registries, such as the EU fleet register or the ICCAT record of vessels. In case the derived vessel class does not match to the one in the registry, GFW assigns the broadest gear type that allows for agreement. If for example, a vessel is registered as a purse seiner but is inferred to be a tuna purse seiner, it would ultimately be assigned to the purse seiner class (Fig. 2).



Figure 2: Hierarchy of fishing gears recognised by GFW

2.1.2 ICCAT catch data

Data on catches of large-pelagic species in the Mediterranean is openly available from the International Commission for the Conservation of Atlantic Tunas (ICCAT). [Nominal catch data](#) was obtained for the period 2015-2023 and filtered for the major large pelagic species like tunas and billfishes (ICCAT, 2025).

2.1.3 Bluefin Tuna vessels

Individual purse seine vessels that were assigned a bluefin tuna quota for 2024 were obtained from the corresponding national notices. In France, quotas are allocated and published by the *Ministère de la Transition écologique et de la Cohésion des territoires* (2024). In Spain, by the *Secretaría General de Pesca* (2024) and in Italy by the *Ministero dell'agricoltura, della sovranità alimentare e delle foreste* (2024). These countries were chosen exemplary because in the period from 2015 to 2023, they made up more than half of the total landings of BFT for all contracting parties to ICCAT, of which close to 90% are fished with purse seine nets (ICCAT, 2024b). Since the national notices did not contain the vessels MMSI number (which is the identifier used by GFW), the vessel names and registration numbers were cross-referenced with the European Fleet Register to obtain it (European Commission, Directorate-General for Maritime Affairs and Fisheries, 2025).

2.2 Data filtering

All data filtering was conducted using R (v4.4.1; R Core Team, 2024) within the RStudio environment (Posit team, 2024). The GFW dataset was first cropped and adjusted to a shapefile of the Mediterranean Sea (Fig. 1), obtained from the General Fisheries Commission for the Mediterranean (GFCM). Subsequent filtering was based on the assigned gear type and GFW registry information. Only entries with the gear type drifting longlines or tuna purse seines, and that were registered with ICCAT were retained (Fig. 3). These gear types were chosen as they are the main *metiers* involved in the exploitation of large-pelagic species in the Mediterranean. Since 2015, on average, 95% of BFT's, and 99% of albacore tunas are caught using either longlines or purse seine nets and close to 100% of swordfish catches use longlines (ICCAT, 2024b,a,c) The GFW dataset is available from the year 2012 until 2024. However, only entries from 2015 were retained, in order to

avoid masking *real* fishing dynamics with the increase in adoption of AIS devices, which only became mandatory in the EU in 2014 for all vessels > 15 m in length (European Commission, 2011). It is estimated however, that in the Mediterranean the EU fishing fleet >15 m is 100% equipped with AIS since 2018 (Taconet et al., 2019).

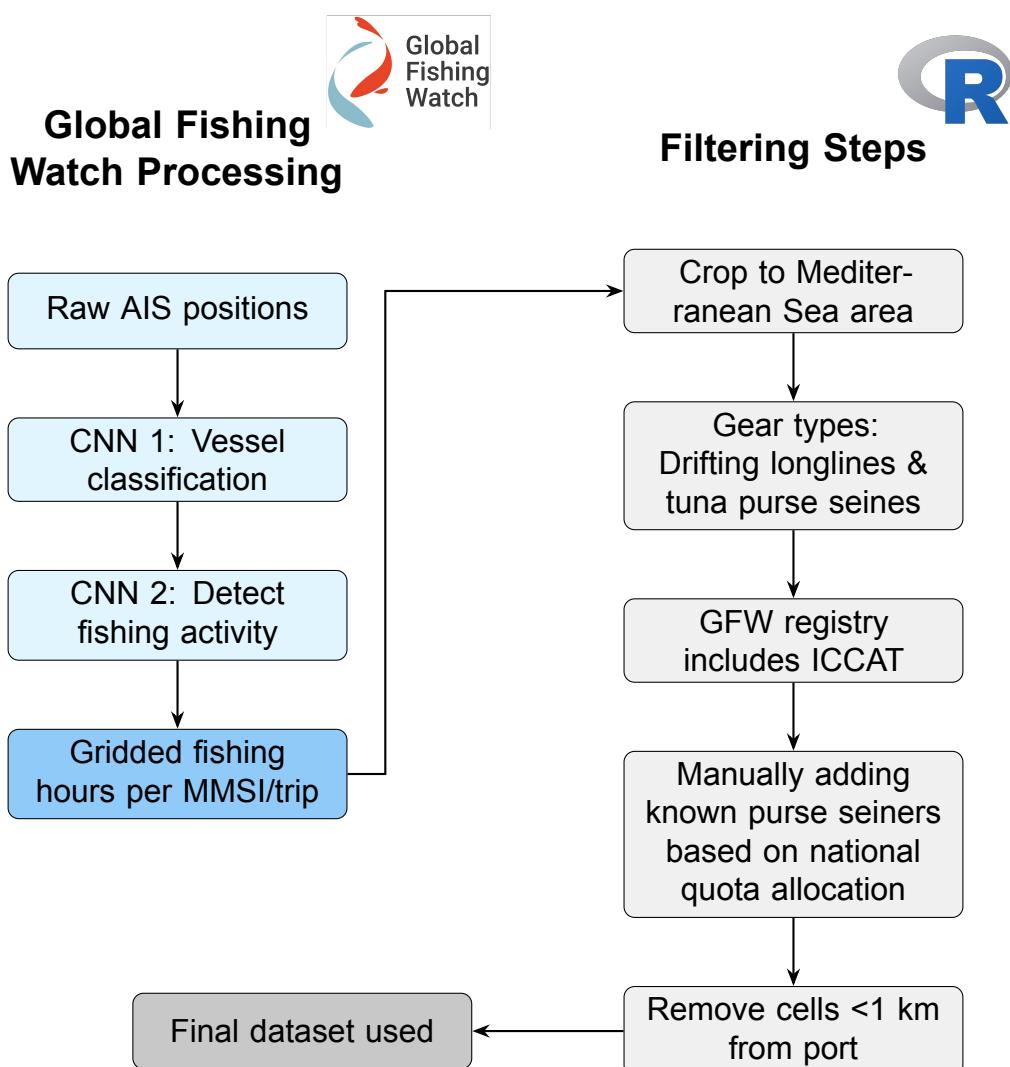


Figure 3: Global Fishing Watch data processing pipeline and filtering steps to obtain the final dataset used in this study. CNN = Convolutional Neural Network

In case of the tuna purse seiners, these filtering steps also removed some vessels which are known to be fishing for bluefin tuna based on national quota allocations

(obtained for Spain, France, and Italy). These vessels were removed by the filtering steps either because GFW assigns them a different gear type, or they are not present in the registry information from ICCAT that GFW uses. For the present analysis, these vessels were thus, manually included after comparison with the information from the national vessel registries.

Irregular vessel movement patterns occurring in or near ports can falsely resemble fishing activity and should therefore be excluded when estimating fishing effort. These movements include cruising towards the harbour to land catches or vessel maintenance (de Souza et al., 2016). Thus, to remove these areas from the analysis, the points inside the 1 km boundary around ports were removed. Distance from port was determined based on a dataset from GFW, which contains anchorages that are either known ports, or that contained at least 20 unique stationary vessels since 2012 (Global Fishing Watch, 2020).

To enable seasonal analysis, each date was assigned to a season. These were defined as winter (Dec. 21 - Mar. 19), spring (Mar. 20 - Jun. 20), summer (Jun. 21 - Sep. 21), and fall (Sep. 22 - Dec. 20) based on the approximate timing of solstices and equinoxes following common definitions.

2.3 Data analysis

To identify persistent hotspot areas throughout the one decade study time and analyse spatio-temporal trends, the Emerging Hotspot Analysis (EHA) tool was used in ArcGIS Pro (Esri Inc., 2024). For this, multidimensional netCDF (network common data form) files of the fishing hours data were generated after curation and filtering in R, using the *terra* package (v1.8.18; Hijmans, 2025). Subsequently, this data was read in as multidimensional raster files in ArcGIS Pro, with the dimensions corresponding to longitude, latitude and time. Data was aggregated annually, using the sum of fishing hours per year for each cell. From this, a space-time

cube, which is the input required for the EHA tool, was created (Fig. 4A).

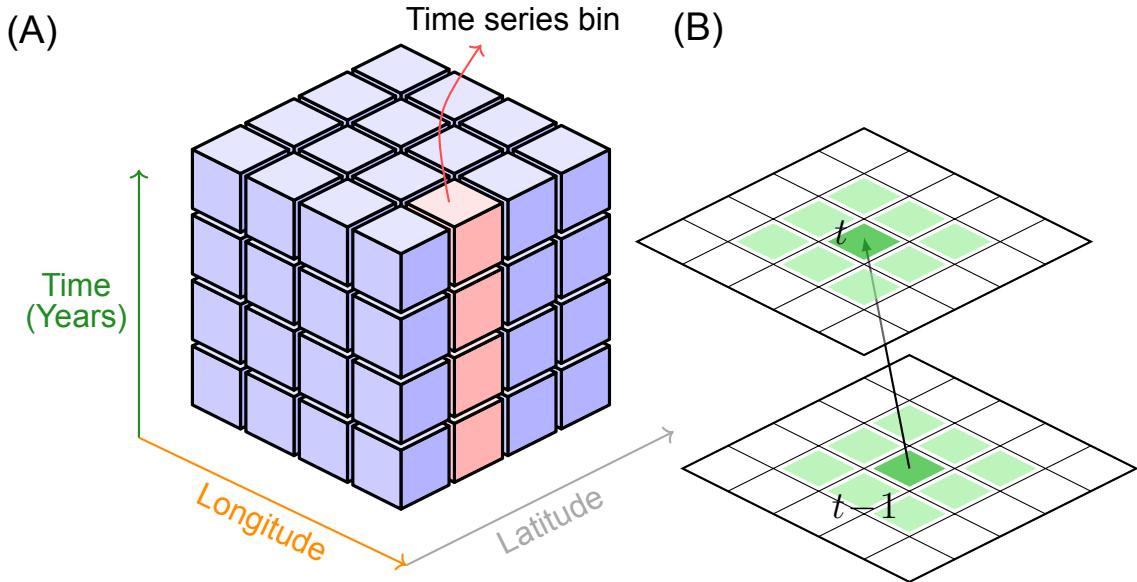


Figure 4: A) Structure of the space-time cube in ArcGIS. Each individual cube corresponds to one *neighbourhood bin*, which is the sum of all coloured cells in B. One *time series bin* corresponds to the same location over time (red). B) Conceptualization of space-time dependency as implemented in the Emerging Hotspot Analysis tool. One *neighbourhood bin* is defined as the cell itself (darkgreen) plus the cells surrounding it (lightgreen), as well as those cells in the previous time step ($t - 1$).

EHA uses a combination of two statistical methods. First, the Getis-Ord G_i^* statistic to identify areas where low/high values are spatially clustered (Getis and Ord, 1992). The null hypothesis states that the sum of values of location i and its neighbours, is not significantly different from what would be expected by chance, based on all observations (neighbours are defined as shown in Fig. 4B). Thus, each *neighbourhood* contains the cell itself, plus all cells contiguous with it via edges and corners at time t and $t - 1$. Each neighbourhood is compared to all global observations at the current and preceding time step. Based on the neighbourhood definition, a binary spatial weight matrix is constructed, where each entry $w_{i,j}$, is either 1 (if features i and j are neighbours) or 0 otherwise. The G_i^* statistic is then calculated as:

$$G_i^* = \frac{\sum_{j=1}^n w_{i,j}x_j - \bar{X} \sum_{j=1}^n w_{i,j}}{S \sqrt{\frac{\sum_{j=1}^n w_{i,j}^2 - \left(\sum_{j=1}^n w_{i,j}\right)^2}{n-1}}} \quad (1)$$

where x_j is the value for feature j , $w_{i,j}$ is the spatial weight between feature i and j , and n is the total number of features. The terms \bar{X} and S represent the global mean and standard deviation of the attribute values, respectively, and are given by:

$$\bar{X} = \frac{\sum_{j=1}^n x_j}{n} \quad (2)$$

$$S = \sqrt{\frac{\sum_{j=1}^n x_j^2}{n} - (\bar{X})^2} \quad (3)$$

The implementation of G_i^* in the EHA tool also applies a False Discovery Rate (FDR) correction to account for multiple testing and spatial dependency in the data. This approach is preferred over methods like Bonferroni correction, which only accounts for multiple testing, as FDR is less conservative and less likely to miss true positives (Caldas de Castro and Singer, 2006). EHA is thus, a spatio-temporal extension of the G_i^* statistic, as it extends each cell not only to its spatial but also to the temporal neighbours.

Second, EHA applies the Mann-Kendall trend test to evaluate whether there is a monotonic upward or downward trend in each time series bin (Mann, 1945; Kendall and Gibbons, 1990). The non-parametric Mann-Kendall statistic S analy-

ses each time series bin. It ranks and compares each point x_i (for $i = 1, 2, \dots, n - 1$) to all subsequent points x_j (for $j = i + 1, i + 2, \dots, n$) and is given by (Kendall and Gibbons, 1990, Section 1.9)

$$S = \sum_{i=1}^{n-1} \sum_{j=i+1}^n \text{sign}(x_j - x_i) \quad (4)$$

Where the sign function is defined as:

$$\text{sign}(x_j - x_i) = \text{sign}(R_j - R_i) \begin{cases} 1 & x_i < x_j \\ 0 & x_i = x_j \\ -1 & x_i > x_j \end{cases} \quad (5)$$

and R_i and R_j are the ranks of observations x_i and x_j of each time series. Thus, for every time point, it assigns a 1 if the value is higher than the previous one, a 0 if the value is the same, and a -1 if the value is lower. These scores are then summed for each time series bin and under the null hypothesis of no trend, the value of S is zero. To assess the significance of S , the variance V_0^* can be calculated as (Kendall and Gibbons, 1990, Section 4.9)

$$V_0^*(S) = n(n - 1)(2n + 5)/18 - \sum_{j=1}^m t_j(t_j - 1)(2t_j + 5)/18 \quad (6)$$

where n is the total number of observations, and m the number of groups with tied ranks, each with t_j tied observations.

3 Results

3.1 Cumulative apparent fishing hours

A total of 213 unique longliners and purse seiners were recorded for 2015-2024. These vessels together accounted for an average of 122 760 hours per year, with an average of 576 hours per vessel per year. Areas that show the highest effort throughout the study time include the Mediterranean coast of Spain, around Sardinia and Sicily, south of Malta, the Adriatic, and south of Cyprus (Fig. 5A). Areas with high effort generally also show the lowest coefficient of variation (Fig. 5B). There does not appear to be any fishing activity based on AIS around the African Mediterranean coast (Fig. 5).

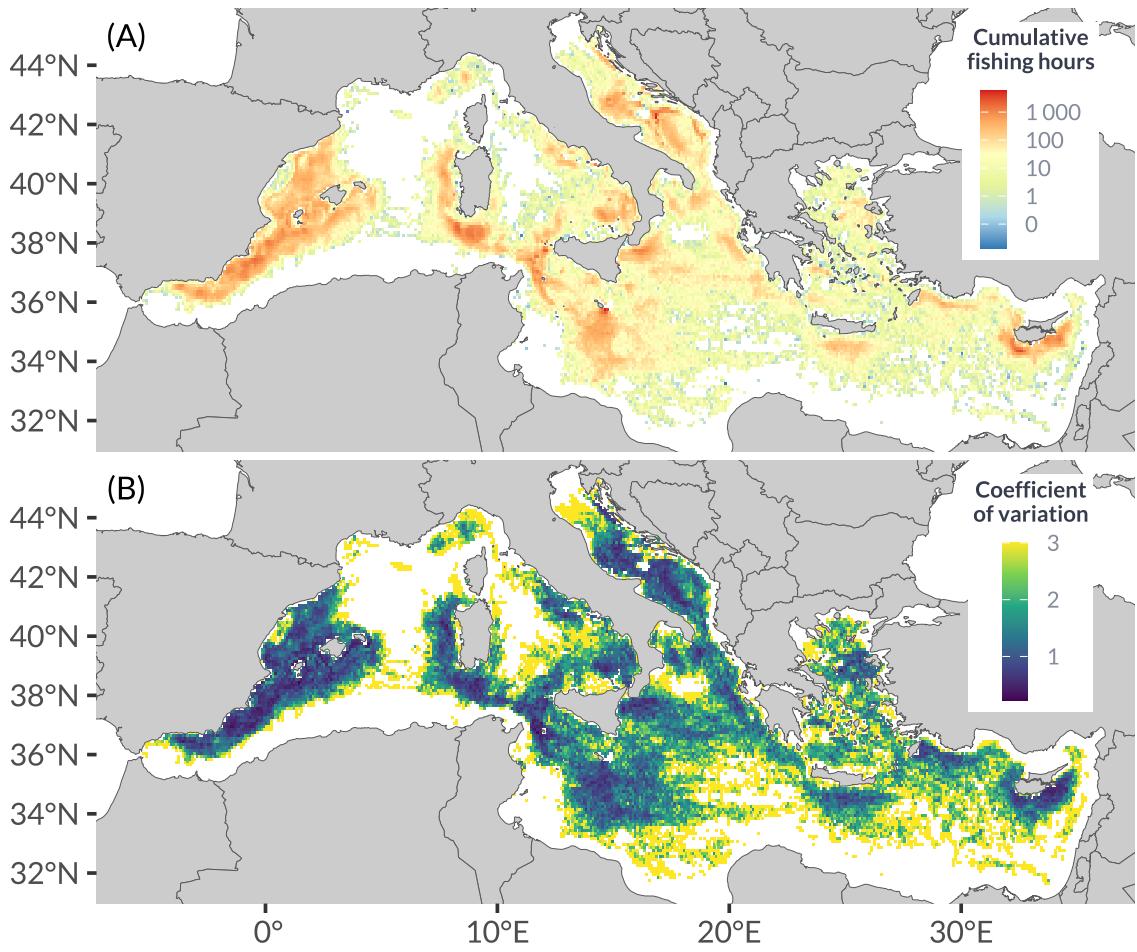


Figure 5: Summary statistics for both longliners and purse-seiners. A) Cumulative fishing hours in the Mediterranean (2015-2024). Colors are on a log-scale. B) Coefficient of variation between years (standard deviation divided by mean) for each cell.

3.2 Fishing hotspots and high risk areas

Numerous persistent longline hotspot areas were identified throughout the Mediterranean including the southern coast of Spain, south of Sardinia, around Sicily and Malta, as well as south of Cyprus (Fig. 6A). These hotspots show however, differing trends. Fishing hours in the south of Malta appear to be increasing throughout, with a similar trend of increase in the Adriatic Sea. Other areas that appear to be consistent hotspots, like the east of Sicily, show a decreasing trend (Fig. 6B).

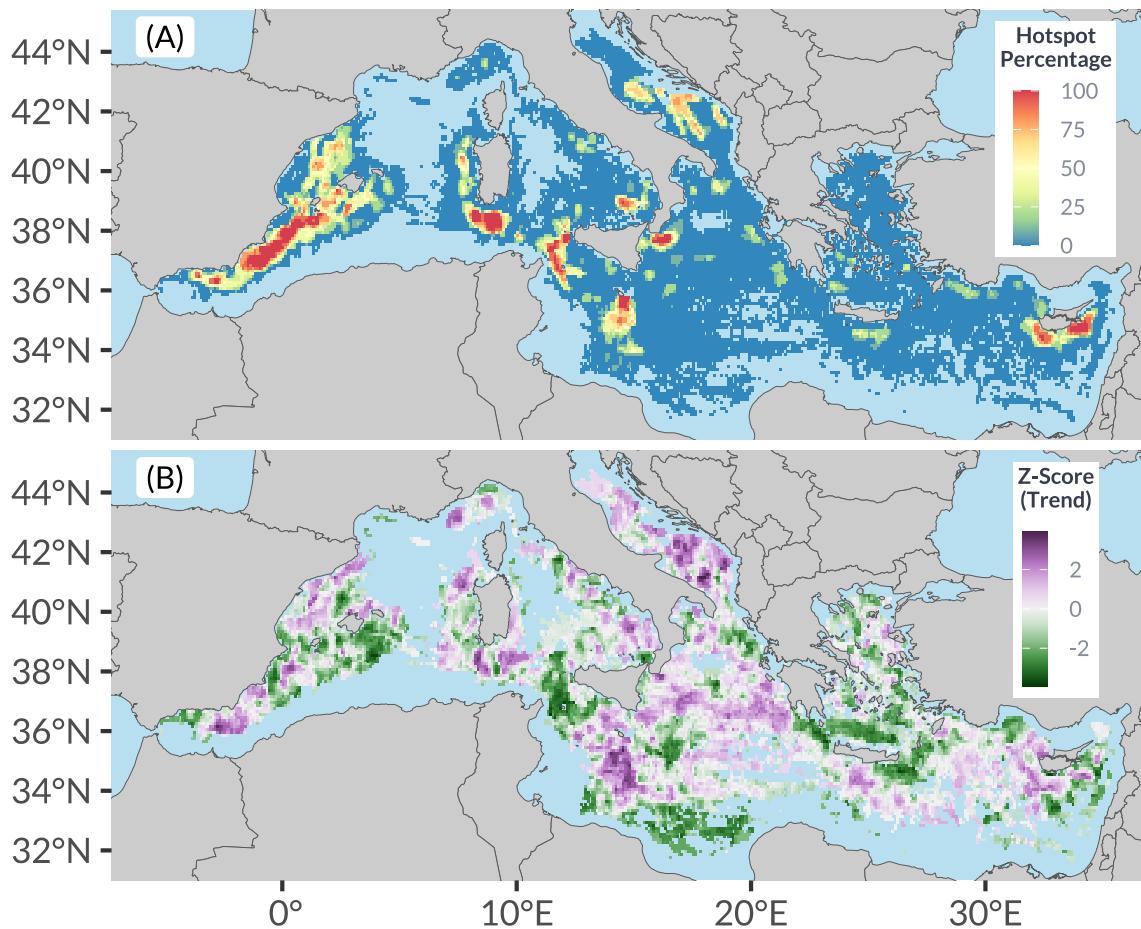


Figure 6: Hotspot percentage and trend scores for longliners in the Mediterranean. The percentage reflects the years in which a given cell was a hotspot based on the Getis-Ord Gi^* statistic. The Z-score is derived from the Mann-Kendall statistic.

Purse seine hotspots appear persistent around the Balearic Islands, in the Adriatic Sea, and along the Calabrian coast in Italy (Fig. 7A). Trends in these areas show a clear increase around Ibiza and in the central Adriatic, while areas around the coast in the Adriatic show a decrease (Fig. 7B). The hotspot area along Calabria shows no clear trend.

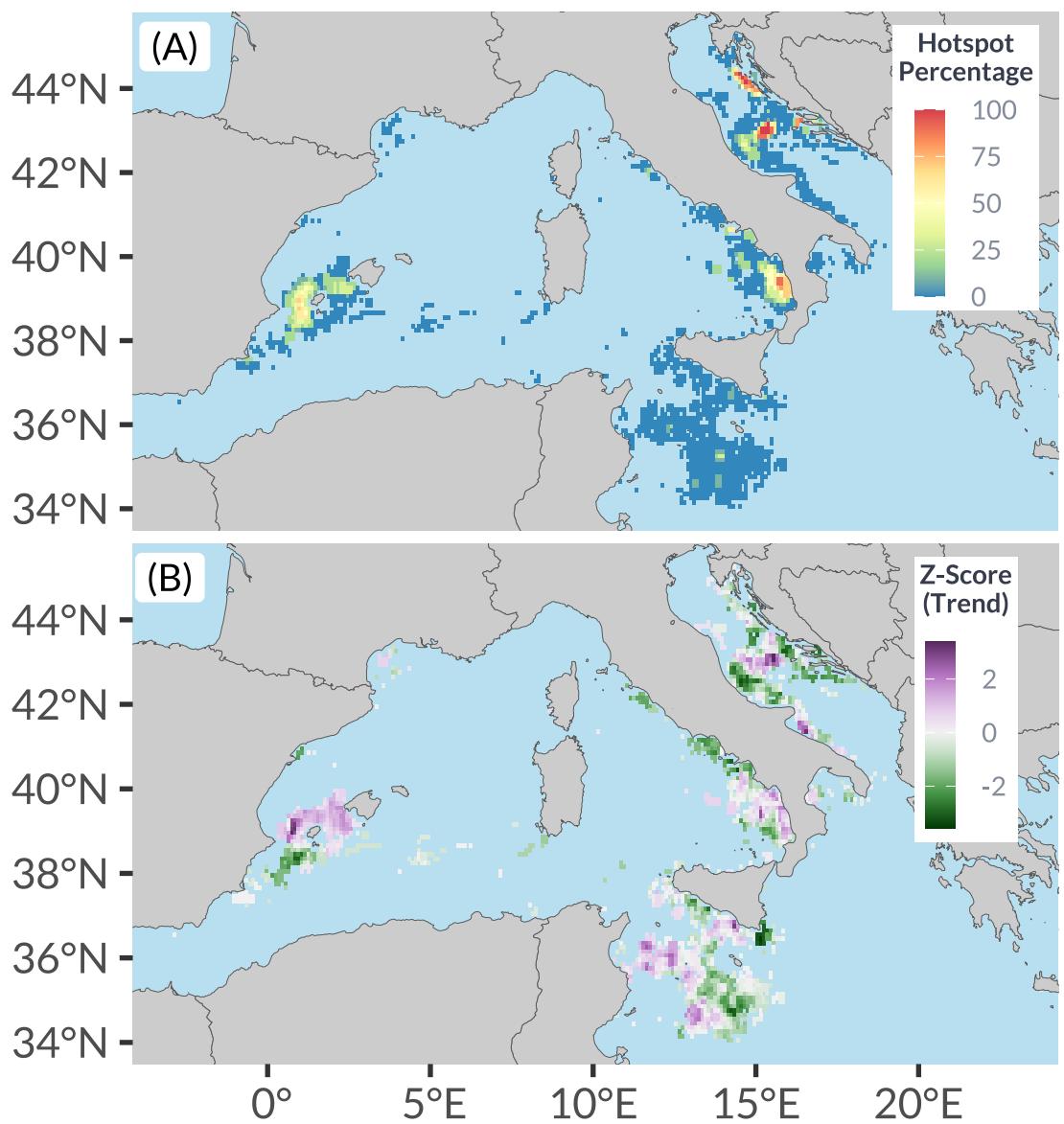


Figure 7: Hotspot percentage and trend scores for purse seiners in the Mediterranean. The percentage reflects the years in which a given cell was a hotspot based on the Getis-Ord Gi^* statistic. The Z-score is derived from the Mann-Kendall statistic.

3.3 Temporal changes

Longline fishing hours show a clear seasonal trend where activity is highest during warmer months (mainly in the summer) and lowest in the colder months around winter (Fig. 8A). Some areas show high effort earlier in the season in spring (for example south of Malta) and others are more persistent later in the season in fall (for instance around Ibiza). The time series show a similar seasonal trend, although the intensity varies between years, and it appears that the longline season is expanding between years (Fig. 8B). The highest annual longline fishing hours throughout the study time were recorded for 2022 and the lowest for 2015 (Tab. 1).

Table 1: Annual sum of fishing hours for purse seiners and longliners.

Year	Fishing hours	
	Purse seiners	Longliners
2015	4,193	71,846
2016	5,313	89,136
2017	5,901	107,345
2018	6,128	106,130
2019	5,931	113,261
2020	5,636	124,927
2021	5,706	133,361
2022	5,675	149,059
2023	5,085	139,068
2024	6,661	137,406

Regarding the purse seine fleet, fishing hours show a very pronounced seasonal trend with a peak in spring (Fig. 9). The core areas of the fishery during spring are the Balearic Islands, along the coast of Calabria (south-west Italy), and the central Adriatic. In the Adriatic, there appears to be purse seine activity throughout the whole year (Fig. 9A). The purse seine season for large-pelagic species is limited to the months of May and June and is consistent between years (Fig. 9B). Highest annual fishing hours for purse seiners were recorded in 2024 and the lowest in

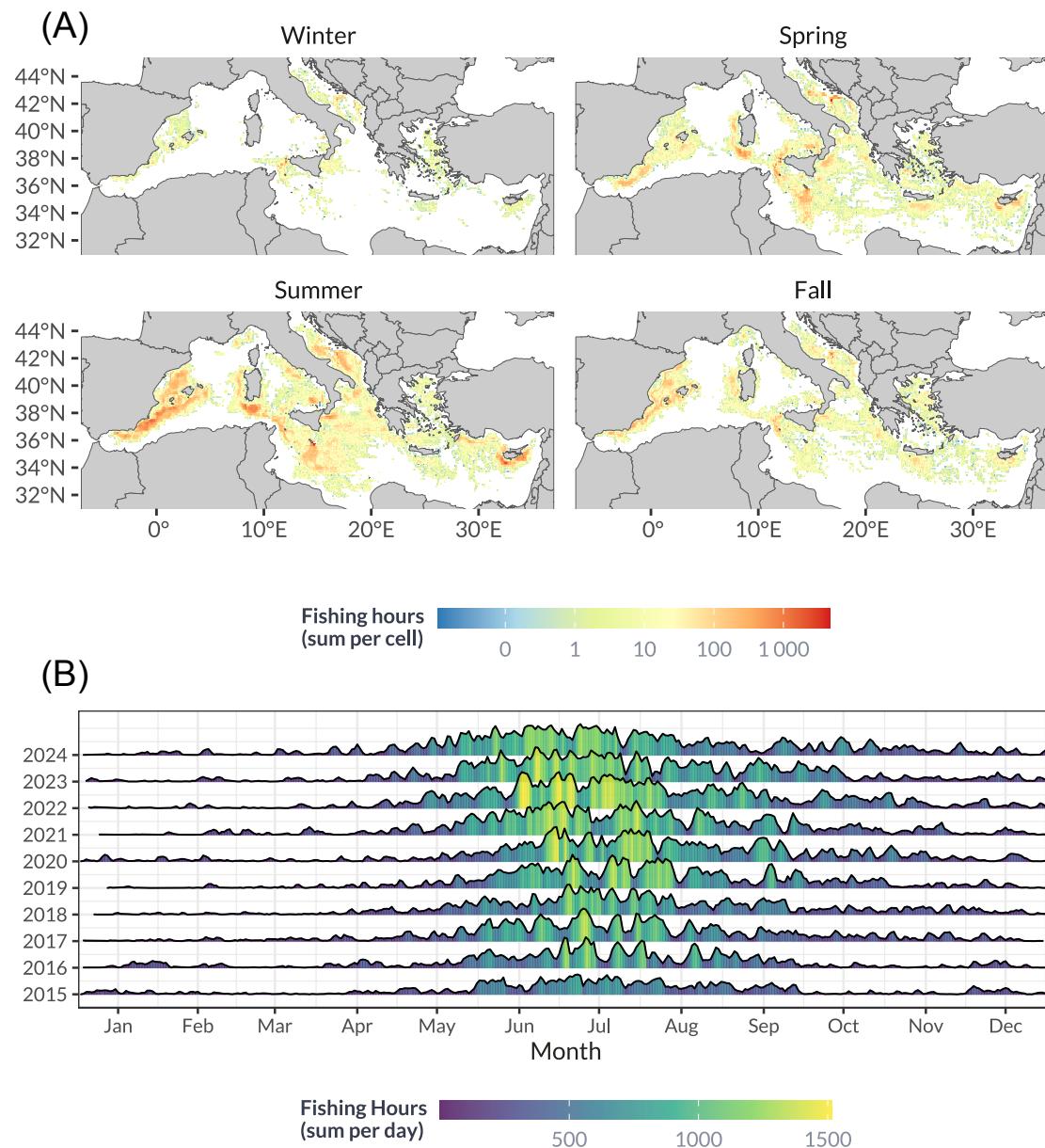


Figure 8: Temporal changes in longline fishing hours. A) Spatial differences between seasons. Hours are summed per season and cell and colour is on a log-scale. Seasons are defined based on calendar dates: winter (Dec 21 - Mar 19), spring (Mar 20 - Jun 20), summer (Jun 21 - Sep 21), and fall (Sep 22 - Dec 20). B) Time series of fishing hours, summed per year and day.

2015 (Tab. 1).

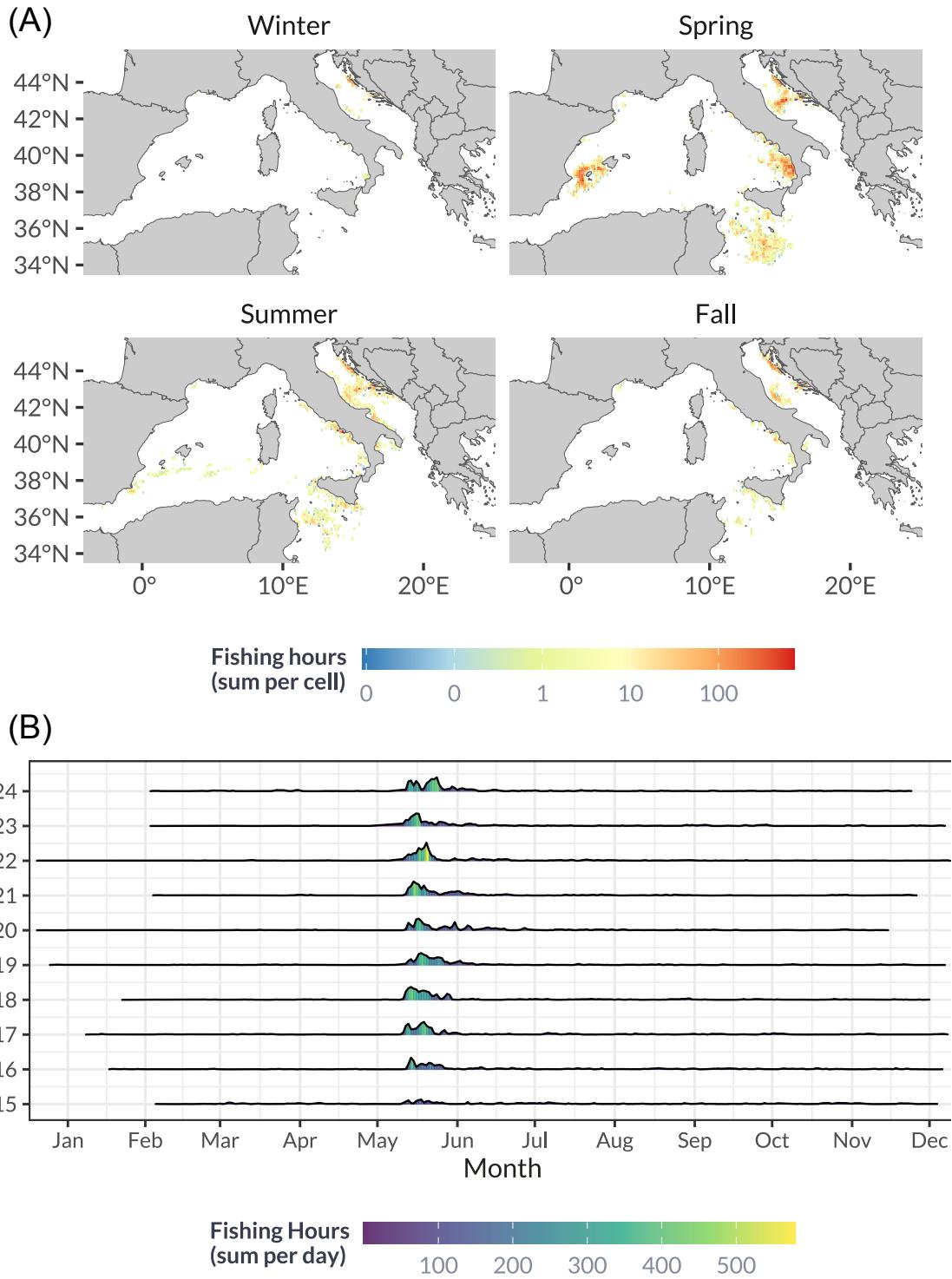


Figure 9: Temporal changes in purse seine fishing hours. A) Spatial differences between seasons. Seasons are defined based on calendar dates: winter (Dec 21 - Mar 19), spring (Mar 20 - Jun 20), summer (Jun 21 - Sep 21), and fall (Sep 22 - Dec 20). Hours are summed per season and cell and colour is on a log-scale. B) Time series of fishing hours, summed per year and day.

3.4 Flag countries

Vessels were flagged to a total of 10 countries and the majority of vessels analysed were longliners (Tab. 2; see Fig. S2 and S3 for an overview by country). Italy shows the highest amount of both purse seiners and longliners identified in the GFW data.

Table 2: Number of vessels by country and gear type in 2024 from our detections compared to vessels with drifting longlines currently registered with the GFCM and tuna purse seine vessels currently registered with ICCAT. ‘–’ indicates no recorded vessels for that gear. PS = purse seiners, LL = longliners

Country	Detected PS	Reg. PS	% PS	Detected LL	Reg. LL	% LL
Albania	–	2	0.0%	–	–	–
Algeria	3	39	7.7%	–	–	–
EU-Croatia	2	3	66.7%	–	–	–
EU-Cyprus	–	1	0.0%	14	15	93.3%
Egypt	–	2	0.0%	–	–	–
EU-France	13	21	61.9%	–	4	0.0%
EU-Greece	–	–	–	8	24	33.3%
EU-Italy	16	19	84.2%	60	83	72.3%
EU-Malta	–	2	0.0%	19	17	111.8%
Libya	–	15	0.0%	–	–	–
Morocco	1	5	20.0%	–	–	–
EU-Spain	5	7	71.4%	22	30	73.3%
Tunisia	–	59	0.0%	–	–	–
Turkey	–	36	0.0%	–	–	–
Total / Mean	40	213	30.1%	123	173	58.6%

Most regions with high fishing activity are fishing grounds shared by multiple countries (Fig. 10). Regions with high overlap between flag countries for longliners include the Balearic Islands, south of Crete, and south of Malta, which are also areas with high fishing hours (Fig. 5A). For purse seiners, fishing generally is more concentrated and thus, overlap is also higher, as seen in the core fishing areas of the Balearic Islands and south of Malta (Fig. 10).

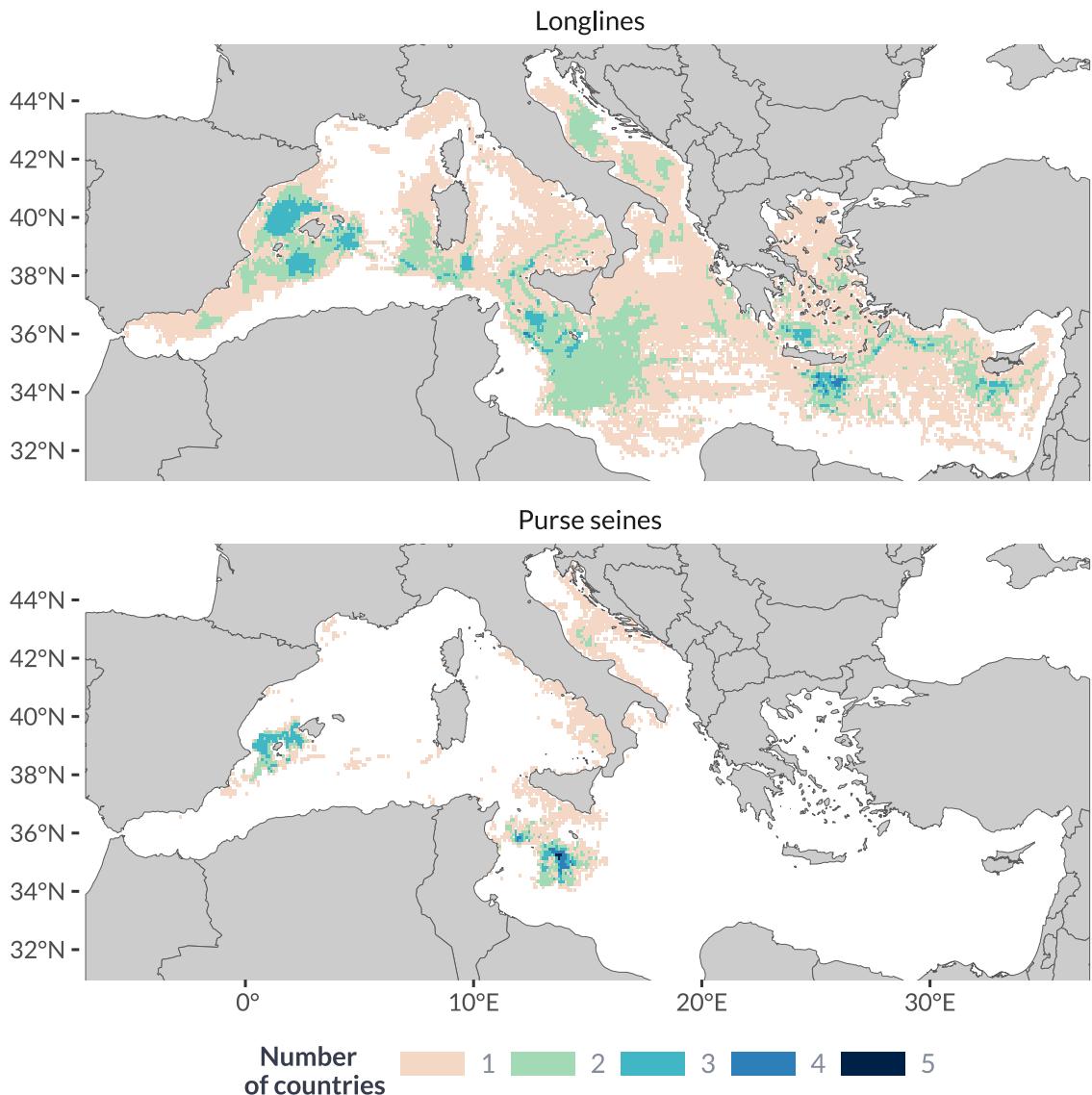


Figure 10: Number of countries fishing per cell for longliners and purse seiners between 2015-2024.

A comparison of fishing hours from the GFW data with catch data from ICCAT shows that AIS under-represents fishing activity by non-EU countries relative to EU countries (Fig. 11). Even though, many non-EU countries account for a substantial share of the total reported catches. Notably, AIS also does not capture any French longline vessels.

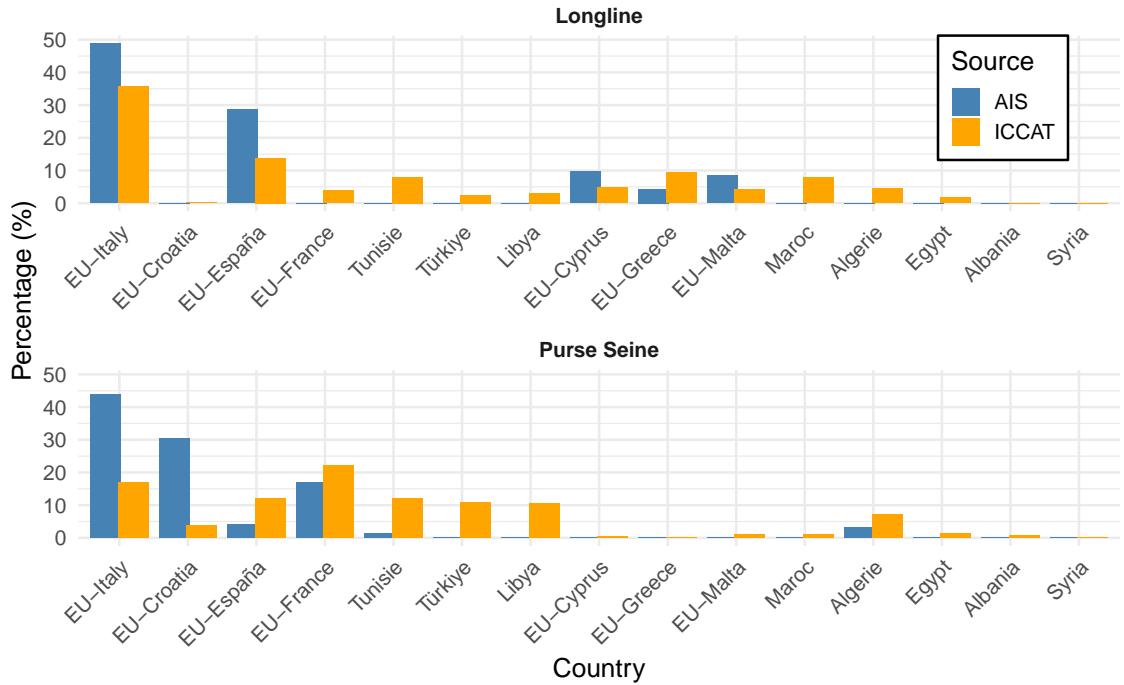


Figure 11: Comparison of relative percentages between GFW AIS data and ICCAT catch data. AIS percentages are relative to the total fishing hours between all countries. ICCAT percentages are relative to the total weight of catches between all countries.

3.5 Depth and distance to port

The relationship between the cumulative proportion of fishing hours and the distance to port reveals that most fishing activity of both gear types is concentrated less than 100 km from the closest port (Fig. 12A). The trend for the depth is different between gear types, where most purse seine fishing occurs at shallower depths (> 1000 m; 50% above 500 m depth) and longline fishing takes place over much greater depth ranges (Fig. 12B; 50% above 1600 m depth).

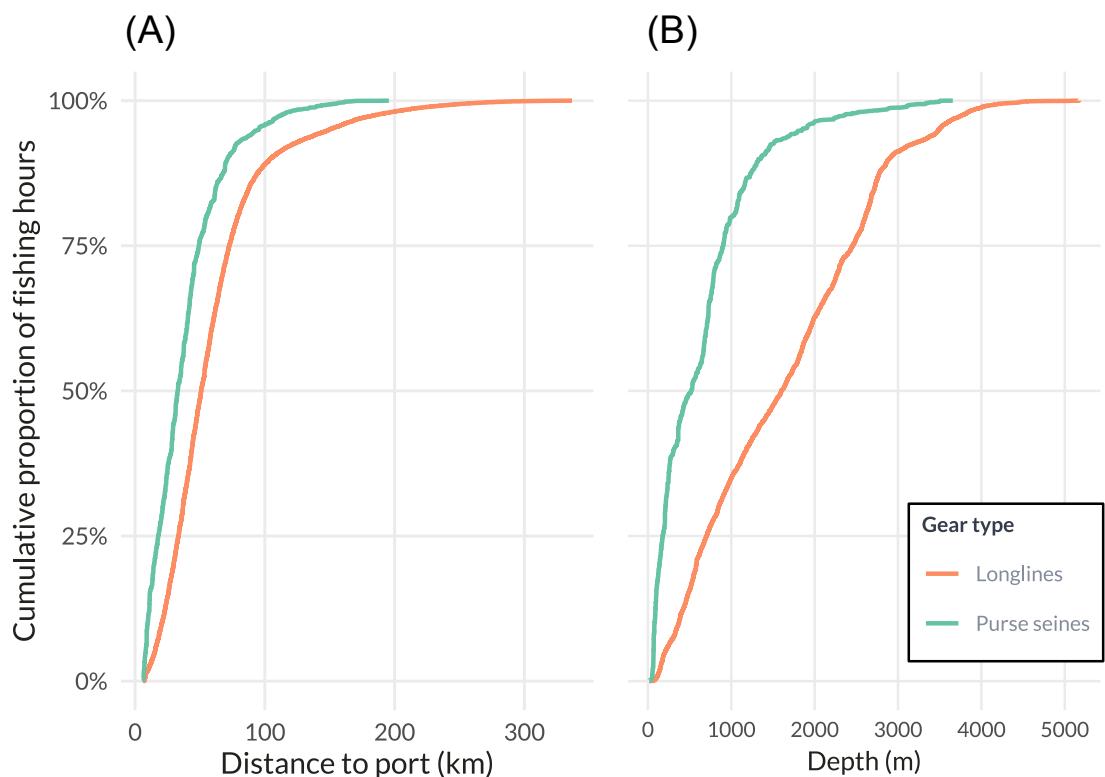


Figure 12: Cumulative proportion of fishing hours with A) Distance to port and B) Depth.

4 Discussion

Our analysis of fishing activity targetting LPF in the Mediterranean revealed that a total of 213 unique longline and purse seine vessels that accounted for an average of over 120 000 fishing hours annually, with concentrated effort along the Spanish Mediterranean coast, Sardinia, Sicily, Malta, the Adriatic, and around Cyprus. These areas correspond to some of the main known spawning and feeding grounds of LPF in the Mediterranean. Supporting this, both fleets also exhibited strong seasonal patterns, with longline fishing peaking in summer and purse seine fishing concentrated in spring driven by the migration towards spawning grounds. Most fishing activity occurred within 100 km of ports, with purse seiners operating in shallower waters compared to the broader depth range of longliners. Monitoring the depth or distance to port of individual fleets could provide insights into potential shifts in distribution related to climate change for example. Notably, the majority of vessels detected by AIS were flagged to EU countries, especially Italy, and AIS data suggested an under-representation of non-EU fishing activity when compared to official catch and vessel records from ICCAT and GFCM.

4.1 Main risk areas (hotspots)

Considering that fisheries of highly migratory LPF occur around aggregations of these animals they should generally follow patterns in the distribution of the target species which aggregate mainly for spawning but also in feeding areas. The fishing hotspots identified generally correspond to spawning and feeding grounds of LPF.

In the Mediterranean, most research on these important areas has focused on BFT. Large BFT > 25 kg come to the Mediterranean Sea for spawning and stay from mid-May until mid-July while some individuals have been shown to reside in the Mediterranean the whole year (Fromentin and Lopuszanski, 2013). They

are caught by both longlines and purse seines (ICCAT, 2024b). Purse seine hotspots around the Balearic Islands and in the Tyrrhenian Sea correspond to known spawning grounds of BFT (Medina, 2020). Notably however, the spawning ground in the strait of Sicily was not consistently identified as a hotspot, probably due to a lack of purse seiners equipped with AIS from southern Mediterranean countries. The most consistent hotspot in our study was the Adriatic Sea where purse seine activity is not limited to spring unlike the other hotspots. One explanation for this could be that a resident BFT population frequents this area throughout the year and catches demonstrate mainly individuals below reproductive age in this area (?). This population stays in the Adriatic Sea and only migrates out of the Mediterranean Sea after reaching maturity around the age of 8 years. Aggregations of BFT in the Adriatic appear to not be related to spawning but to feeding and Croatia is currently the only country that is allowed to catch BFT below > 30 kg and use them in fattening farms. This explains the year-round fishing activity by the Croatian purse seine fleet which is also reflected in the catches reported to ICCAT (Fig. S4).

Swordfish migratory behaviour is highly complex but local increases in density appear to be dictated by seasonal migration to spawning areas when the 24°C isotherm occurs (Arocha, 2007; Palko et al., 1981). In the Mediterranean, swordfish are known to spawn from June to August between the Balearic Islands and east of the Strait of Gibraltar, in the Tyrrhenian Sea and the Strait of Messina, the Levantine Sea (between Crete and Cyprus), and the Gulf of Taranto in the Ionian Sea (see Fig. 5 in Arocha, 2007). Juveniles and adults occupy different habitats with juveniles being more associated to coastal areas and adults more to pelagic waters (Damalas and and, 2014). Swordfish are one of the main targets of the Mediterranean longline fleets. The largest persistent fishing hotspot that was identified spans from west of the Balearic Islands until the east of the Strait of Gibraltar, overlapping with the spawning ground identified there. Other persistent

hotspots overlapping with known spawning areas are located in the Tyrrhenian Sea close to Sicily and to the east of Sicily in the Ionian Sea. Catches of juveniles are generally quite high in the Mediterranean and one of the management regulations currently in place for swordfish is a minimum size limit (100 cm [ICCAT Rec 16-05], Mauricio, 2020). This minimum size limit has however led to a higher discard rate and higher effort, as since then, more individuals are discarded as they fall under the minimum size, which in turn increases the effort necessary to reach the quota (García-Barcelona et al., 2020). The swordfish stock in the Mediterranean is currently still considered overfished but a seasonal closure has been implemented from January to the end of March that also extends to the albacore fleet during October and November [ICCAT Rec 16-05].

Capture as non-target species in Mediterranean longline fisheries is a concern for seabirds, pelagic elasmobranchs, and sea turtles. Drifting longlines are estimated to be responsible for the bycatch of about 27 000 individual sea turtles annually in the Mediterranean alone (Carpentieri et al., 2021). The same authors also identified the most important parameters when investigating capture rates of loggerhead turtles to be fishing effort, including detailed information on the métier and the time spent fishing. The longline hotspots that were identified in our analysis, show overlap with estimates of high relative abundance of loggerhead turtles *Caretta caretta* most notably in the Tyrrhenian Sea, the Balearic Islands, and south of Malta (DiMatteo et al., 2022; Burgess et al., 2010). Burgess et al. (2010) analysed the longline fishery around Malta and reported that *Caretta caretta* makes up 40% of the total catch in terms of individual animals. Considering that this area sees some of the highest longline activity identified in this study and shows an increasing trend, closer monitoring on bycatch rates is necessary. Although there appear to be recent advances in reducing bycatch of turtles for example in Spain, these are driven by changes in métier (mainly due to deeper sets that sink faster) and not through targetted management. Future economically driven changes in

métier could thus, easily reverse this decrease. The results and methodology employed here, could be a starting point to identify high risk areas overlapping with important turtle habitat, to better target the deployment of onboard observers. Further studies could look into fine-scale spatial overlap between non-target species and longlines, to better estimate the risk for specific species groups.

4.2 Trends

The spawning of BFT in the Mediterranean has been shown to be highly reliant on the position of a frontal region where newly arrived, less saline Atlantic water meets more saline, resident Mediterranean waters around the Balearic Islands (Alvarez-Berastegui et al., 2012). The position of this front is however, variable between years (Balbín et al., 2014). The position of the purse seine hotspots in which the individuals migrating from the Atlantic to spawn are targetted, should thus follow the position of this front between years. In the present analysis we were able to show that the location of fishing in this hotspot changes between years and following the variable front is a possible explanation for this.

The clear seasonality of the purse seine fleet can be attributed to the timing of BFT migration into the Mediterranean and the implemented TAC's which are reached in a very short amount of time. The longline season appears to be expanding over time which could be related to temperatures being higher for longer due to climate change or an increase in fishing activity is necessary to reach the TAC's. In general however, there are no large-scale shifts in fishing grounds and they stay mostly in similar areas from 2015 until 2024. This finding shows that large variation in spawning grounds is rare for the LPF targetted by the gears analysed here.

Some longline fleets analysed here have had changes in the gear that is used recently. Italian and Spanish longline fleets targetting swordfish are now using mostly mid-water longlines instead of traditional surface longlines (Bertolino, F.

and Camolese, C. and Dell'Aquila, M. and Mariani, A. and Valastro, M., 2015; Barcelona, Salvador García et al., 2010). As this gear occurs in deeper waters, a potential way of monitoring changes would be the depth in which a fleet operates.

4.3 Data caveats

AIS technology was not designed for the estimation of fishing effort and thus, there are some persistent challenges in deriving fishing effort from these tracking devices. First, there are differences in the coverage of different AIS receivers. Global Fishing Watch receives raw AIS signals from terrestrial receivers, satellite receivers, and *Dynamic AIS*, which are receivers onboard thousands of large vessels that travel along major shipping routes and relay AIS signals they receive to satellites. Reception of these signals is however, not homogenous in all places, especially in areas with high vessel density like the Mediterranean. Additionally, there have been changes in the AIS sources that underlie the calculation of fishing hours from GFW throughout our study period. This could potentially mask real variation in fishing hours and lead to an apparent increase in activity while in reality only coverage increased. This is the reason why time series derived from this data is difficult to analyse, however GFW is currently working on an additional dataset that quantifies AIS coverage for each vessels' trip which would enable a more detailed, quantitative analysis of differences in fishing activity over time. Second, AIS is biased towards larger vessels (> 15 m) and towards countries in which this technology is mandatory as shown in this study. This is especially relevant in the Mediterranean, as the majority of fishing vessels belong to the small-scale fleet and an important part of Total Allowable Catches are held by non-EU countries. French longliners for example operate in the Gulf of Lions but we were not able to detect any French longline vessel through AIS, likely because they are < 15 m (Farrugio, 2013). Third, The fishing detection model used by GFW is a general model that analyses different métiers and even different gear types under the

same *effort* metric which is fishing hours. These hours include multiple fishing-related activities. This can be for example, setting gear, soak time, gear hauling but also searching for fish. The two main gear types in our study usually use different measures of effort. For pelagic longlines for example the number of hooks or hauls and for tuna purse seines for instance the number of sets. This can result in overestimation of fishing hours as recently demonstrated by Hintzen et al. (2025). The authors showed that GFW data can overestimate the real time spent fishing by between 30% and 380% but also that the geographic position of hauls was identical based on self-sampling data from two pelagic trawling fleets. This illustrates that the results of our studies should not be interpreted as an absolute measure of fishing effort exerted but rather as a relative quantitative measure that is indicative of broad-scale trends especially considering the large time frame for which data was analysed. Fourth, AIS is not able to detect different métiers as small changes in gear type for example can mean that a different species is being targetted. This is one of the limitations here, as all three LPF are targetted by different types of longlines, but we are not able to differentiate between the fishing hours of each métier based on AIS. Further research could improve this for example through combining logbook data on catches with AIS data to estimate catch per unit effort (CPUE), allowing a more accurate evaluation of fishing operations and thus, resource abundance (Niu et al., 2024).

The results of this study demonstrate the broad distribution of fishing in almost the whole Mediterranean Sea and can contribute to the sustainable exploitation of migratory LPF. The methods used here could be combined with other data like log books to get a more detailed information on changes in relative abundance of target species. This information would provide more detailed information to management which is necessary for the implementation of ecosystem-based management. The areas we identified as fishing hotspots and potentially also the migratory paths between them could be prime targets for management for

example to implement seasonal or spatial closures. Relano and Pauly (2022b) proposed to protect migratory LPF through *Blue Corridors*, essentially protected areas along the species' migratory pathways. Our findings could help in identifying these areas. Mapping of the intensity of fishing does also have direct implications for conservation, as bycatch rates can be quite high for pelagic longliners in the Mediterranean.

5 Conclusion

Conclusion goes here.

6 Acknowledgements

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Bibliography

- Addis, P., R. Cannas, A. Cau, J. M. Dean, G. Greco, I. Locci, and M. Secci. (2012). Social, cultural and basic economic analysis of the trap fishery of Sardinia: first step towards parameterization. *Collective Volume of Scientific Papers, ICCAT*, 67, 380–389. https://www.iccat.int/Documents/CVSP/CV067_2012/CV067010380.pdf accessed 2025-05-29.
- Alvarez-Berastegui, D., P. Reglero, F. Alemany, L. Ciannelli, R. Balbín, and J. L. López-Jurado. (2012). Geographically and environmentally driven spawning distributions of tuna species in the western mediterranean sea. *Marine Ecology Progress Series*, 463, 273–284. <https://doi.org/10.3354/meps09800>.
- Arocha, F. (2007). Swordfish reproduction in the atlantic ocean: An overview. *Gulf and Caribbean Research*, 19, 21–36.
- Balbín, R., J.L. López-Jurado, M.M. Flexas, P. Reglero, P. Vélez-Velchí, C. González-Pola, J.M. Rodríguez, A. García, and F. Alemany. (2014). Interannual variability of the early summer circulation around the balearic islands: Driving factors and potential effects on the marine ecosystem. *Journal of Marine Systems*, 138, 70–81. <https://doi.org/10.1016/j.jmarsys.2013.07.004>. The wrapping up of the IDEADOS project: International Workshop on Environment, Ecosystems and Demersal Resources, and Fisheries.
- Barcelona, Salvador García, Ortiz de Urbina, José M., de la Serna, José M., Alot, Enrique, and Macías, David. (2010). Seabird bycatch in spanish mediterranean large pelagic longline fisheries, 2000-2008. *Aquatic Living Resources*, 23, 363–371. <https://doi.org/10.1051/alr/2010022>.
- Baum, Julia K. and Boris Worm. (2009). Cascading top-down effects of changing oceanic predator abundances. *Journal of Animal Ecology*, 78, 699–714. <https://doi.org/10.1111/j.1365-2656.2009.01531.x>.

Beamish, R.J., G.A. McFarlane, and J.R. King. (2005). Migratory patterns of pelagic fishes and possible linkages between open ocean and coastal ecosystems off the pacific coast of north america. *Deep Sea Research Part II: Topical Studies in Oceanography*, 52, 739–755. <https://doi.org/10.1016/j.dsr2.2004.12.016>.

Bell, James B., Johannes H. Fischer, Ana P.B. Carneiro, Shane Griffiths, Alessandra Bielli, Sebastián Jiménez, Steffen Oppel, Richard A. Phillips, Helen M. Wade, Oliver Yates, and Stuart A. Reeves. (2025). Evaluating the effectiveness of seabird bycatch mitigation measures for pelagic longlines in the south atlantic. *Biological Conservation*, 302. <https://doi.org/10.1016/j.biocon.2025.110981>.

Bertolino, F. and Camolese, C. and Dell'Aquila, M. and Mariani, A. and Valastro, M. (2015). Swordfish *Xiphias gladius L.* fisheries using mesopelagic longline in the mediterranean sea by italian fishing fleet. *Collective Volume of Scientific Papers, ICCAT*, 71, 2073–2078. https://www.iccat.int/Documents/CVSP/CV071_2015/n_5/CV071052073.pdf accessed 2025-06-02.

Burgess, E., M. Dimech, R. Caruana, M. Darmanin, H. Raine, A. Harrison, and P. J. Schembri. (2010). Social, cultural and basic economic analysis of the trap fishery of Sardinia: first step towards parameterization. *Collective Volume of Scientific Papers, ICCAT*, 65, 2262–2269. https://www.iccat.int/Documents/CVSP/CV065_2010/n_6/CV065062262.pdf accessed 2025-05-29.

Caldas de Castro, Marcia and Burton H. Singer. (2006). Controlling the false discovery rate: A new application to account for multiple and dependent tests in local statistics of spatial association. *Geographical Analysis*, 38, 180–208. <https://doi.org/10.1111/j.0016-7363.2006.00682.x>.

Carpentieri, P., A. Nastasi, M. Sessa, and A. Srour, eds. (2021). *Incidental Catch*

of Vulnerable Species in Mediterranean and Black Sea Fisheries – A Review
volume 101 of *Studies and Reviews*. General Fisheries Commission for the
Mediterranean, FAO, Rome <https://doi.org/10.4060/cb5405en>.

Damalas, Dimitrios and Persefoni Megalofonou and. (2014). Age-dependent habitat identification of mediterranean swordfish: Application on commercial fishery data and potential use in fisheries management. *Reviews in Fisheries Science & Aquaculture*, 22, 257–273. [10.1080/23308249.2014.951987](https://doi.org/10.1080/23308249.2014.951987).

de Souza, Erico N., Kristina Boerder, Stan Matwin, and Boris Worm. (2016). Improving fishing pattern detection from satellite ais using data mining and machine learning. *PLOS ONE*, 11, 1–20. <https://doi.org/10.1371/journal.pone.0158248>.

Di Natale, A. (2012). Literature on the Eastern Atlantic and Mediterranean Tuna Trap Fishery. *Collective Volume of Scientific Papers, ICCAT*, 67, 175–220. https://www.iccat.int/Documents/CVSP/CV067_2012/CV067010175.pdf accessed 2025-05-29.

Di Natale, A., A. Celona, and A. Mangano. (2005). A series of catch records by the harpoon fishery in the Strait of Messina from 1976 to 2003. *Collective Volume of Scientific Papers, ICCAT*, 58, 1348–1359. https://www.iccat.int/Documents/CVSP/CV058_2005/n_4/CV058041348.pdf accessed 2025-05-29.

DiMatteo, Andrew, Ana Cañadas, Jason Roberts, Laura Sparks, Simone Pani-gada, Olivier Boisseau, Anna Moscrop, Caterina Maria Fortuna, Giancarlo Lauriano, Draško Holcer, Hélène Peltier, Vincent Ridoux, Juan Antonia Raga, Jesús Tomás, Annette C. Broderick, Brendan J. Godley, Julia Haywood, David March, Robin Snape, Ricardo Saggarminaga, and Sandra Hochscheid. (2022). Basin-wide estimates of loggerhead turtle abundance in the mediterranean

sea derived from line transect surveys. *Frontiers in Marine Science*, 9. <https://doi.org/10.3389/fmars.2022.930412>.

Esri Inc. (2024). ArcGIS Pro (Version 3.4). <https://www.esri.com/en-us/arcgis/products/arcgis-pro/overview>. Esri Inc.

European Commission. (1993). Council Regulation No. 2847/93 of 12 October 1993 establishing a control system applicable to the common fisheries policy. *Official Journal of the European Union*, 261, 93. <https://eur-lex.europa.eu/eli/reg/1993/2847/oj/eng>.

European Commission. (2003). Commission Regulation No. 2244/2003 of 18 December 2003 laying down detailed provisions regarding satellite-based vessel monitoring systems. *Official Journal of the European Union*, L333, 17–27. <https://eur-lex.europa.eu/eli/reg/2003/2244/oj/eng>.

European Commission. (2011). Commission Directive 2011/15/EU of 23 February 2011 amending Directive 2002/59/EC of the European Parliament and of the Council establishing a Community vessel traffic monitoring and information system Text with EEA relevance. *Official Journal of the European Union*, L49/33. <https://eur-lex.europa.eu/eli/dir/2011/15/oj/eng>.

European Commission, Directorate-General for Maritime Affairs and Fisheries. (2025). EU Fleet Register, Version 1.2.0.1. https://webgate.ec.europa.eu/fleet-europa/index_en accessed 2025-05-17.

FAO. (2024). *The State of World Fisheries and Aquaculture 2024 – Blue Transformation in Action*. Food and Agriculture Organization, Rome <https://doi.org/10.4060/cd0683en>.

FAO. (2025). Fishing gear types: Drifting longlines. technology fact sheets. <https://www.fao.org/fishery/en/geartype/233/en> accessed 2025-05-26.

Farrugio, H. (2013). *Fisheries in the Gulf of Lions*. RAC/SPA, UNEP-MAP, Tunis https://www.rac-spa.org/sites/default/files/doc_open_seas/fisheries_gulf_lions.pdf. Ed. RAC/SPA.

Ferrà, Carmen, Anna Nora Tassetti, Fabio Grati, Giulio Pellini, Piero Polidori, Giuseppe Scarcella, and Gianna Fabi. (2018). Mapping change in bottom trawling activity in the mediterranean sea through ais data. *Marine Policy*, 94, 275–281. <https://doi.org/10.1016/j.marpol.2017.12.013>.

Fromentin, Jean-Marc and Daniel Lopuszanski. (2013). Migration, residency, and homing of bluefin tuna in the western mediterranean sea. *ICES Journal of Marine Science*, 71, 510–518. <https://doi.org/10.1093/icesjms/fst157>.

Fromentin, Jean-Marc and Joseph E Powers. (2005). Atlantic bluefin tuna: population dynamics, ecology, fisheries and management. *Fish and Fisheries*, 6, 281–306. <https://doi.org/10.1111/j.1467-2979.2005.00197.x>.

García-Barcelona, S., D. Macías, S. Saber, M.-J. José Gómez-Vives, P. Rioja, and J. Ortiz de Urbina. (2020). Effects of ICCAT Rec [2016-05] size regulation on the discards of the Spanish longline fishery targeting swordfish (*Xiphias gladius* Linnaeus, 1758) in the western Mediterranean. *Collective Volume of Scientific Papers, ICCAT*, 76, 172–179. https://www.iccat.int/Documents/CVSP/CV076_2019/n_3/CV076003172.pdf accessed 2025-06-02.

Gerritsen, Hans and Colm Lordan. (2010). Integrating vessel monitoring systems (VMS) data with daily catch data from logbooks to explore the spatial distribution of catch and effort at high resolution. *ICES Journal of Marine Science*, 68, 245–252. <https://doi.org/10.1093/icesjms/fsq137>.

Getis, Arthur and J. K. Ord. (1992). The analysis of spatial association by use of distance statistics. *Geographical Analysis*, 24, 189–206. <https://doi.org/10.1111/j.1538-4632.1992.tb00261.x>.

Global Fishing Watch. (2020). Distance from port in meters, Version 20201104.

<https://globalfishingwatch.org/datasets-and-code/>.

Global Fishing Watch. (2025). Global AIS-based Apparent Fishing Effort Dataset,

Version 3.0. <https://doi.org/10.5281/zenodo.14982712>.

Hijmans, Robert J. (2025). *terra: Spatial Data Analysis*. <https://rspatial.org/>.

R package version 1.8-18, <https://rspatial.github.io/terra/>.

Hintzen, Niels T., Francois Bastardie, Doug Beare, Gerjan J. Piet, Clara Ulrich,

Nicolas Deporte, Josefina Egekvist, and Henrik Degel. (2012). VMStools: Open-source software for the processing, analysis and visualisation of fisheries log-book and VMS data. *Fisheries Research*, 115-116, 31–43. <https://doi.org/10.1016/j.fishres.2011.11.007>.

Hintzen, Niels T, Katie Brigden, Hendrik-Jan Kaastra, Steven Mackinson, Martin A Pastoors, and Lennert van de Pol. (2025). Bias in global fishing watch ais

data analyses results in overestimate of northeast atlantic pelagic fishing impact.

ICES Journal of Marine Science, 82, fsaf033. [10.1093/icesjms/fsaf033](https://doi.org/10.1093/icesjms/fsaf033).

Hinz, H, L G Murray, G I Lambert, J G Hiddink, and M J Kaiser. (2013).

Confidentiality over fishing effort data threatens science and management progress. *Fish and Fisheries*, 14, 110–117. <https://doi.org/10.1111/j.1467-2979.2012.00475.x>.

ICCAT. (2022). Resolution by ICCAT on Climate Change. <https://www.iccat.int/Documents/Recs/compendiopdf-e/2022-13-e.pdf> accessed 2025-05-25.

ICCAT. (2023). Terms of reference for the mediterranean tuna habitat observatory

initiative. *Collective Volume of Scientific Papers, ICCAT*, 80, 155–161. https://www.iccat.int/Documents/CVSP/CV080_2023/n_7/CV080070155.pdf accessed 2025-05-29.

ICCAT. (2024a). Executive Summary ALB-MD. https://www.iccat.int/Documents/SCRS/ExecSum/ALB_MED_ENG.pdf accessed 2025-05-21.

ICCAT. (2024b). Executive Summary BFT-E. https://www.iccat.int/Documents/SCRS/ExecSum/BFT_E_ENG.pdf accessed 2025-03-16.

ICCAT. (2024c). Executive Summary SWO-MD. https://www.iccat.int/Documents/SCRS/ExecSum/BFT_E_ENG.pdf accessed 2025-04-20.

ICCAT. (2025). Nominal catches of Atlantic tunas and tuna-like fish (including sharks), by gear, region and flag (Excel table). <https://www.iccat.int/en/accesingdb.html> accessed 2025-05-07.

James, Mark, Tania Mendo, Esther L. Jones, Kyla Orr, Ali McKnight, and John Thompson. (2018). Ais data to inform small scale fisheries management and marine spatial planning. *Marine Policy*, 91, 113–121. <https://doi.org/10.1016/j.marpol.2018.02.012>.

Kendall, M. G. and J. D. Gibbons. (1990). *Rank Correlation Methods* 5th edition. Oxford University Press, New York.

Kontas, Vasilis. (n.d.). How often do the positions of the vessels get updated on Marine Traffic? <https://support.marinetraffic.com/en/articles/9552905-how-often-do-the-positions-of-the-vessels-get-updated-on-marinetraffic> accessed 2025-01-05.

Kroodsma, David A., Juan Mayorga, Timothy Hochberg, Nathan A. Miller, Kristina Boerder, Francesco Ferretti, Alex Wilson, Bjorn Bergman, Timothy D. White, Barbara A. Block, Paul Woods, Brian Sullivan, Christopher Costello, and Boris Worm. (2018). Tracking the global footprint of fisheries. *Science*, 359, 904–908. <https://doi.org/10.1126/science.aoa5646>.

Maina, I., S. Kavadas, S. Katsanevakis, S. Somarakis, G. Tserpes, and S. Georgakarakos. (2016). A methodological approach to identify fishing grounds: A case study on greek trawlers. *Fisheries Research*, 183, 326–339. <https://doi.org/10.1016/j.fishres.2016.06.021>.

Mann, H. B. (1945). Nonparametric tests against trend. *Econometrica*, 13, 245–259. <https://doi.org/10.2307/1907187>.

Marsaglia, Luca, Antonio Parisi, Simone Libralato, Nathan A Miller, Pete Davis, Fernando S Paolo, Fabio Fiorentino, Germana Garofalo, Marco Costantini, and Tommaso Russo. (2024). Shedding light on trawl fishing activity in the mediterranean sea with remote sensing data. *ICES Journal of Marine Science*, 82. <https://doi.org/10.1093/icesjms/fsae153>.

Mauricio, O. (2020). Estimation of undersize mediterranean swordfish (*Xiphias gladius*) catches between 2008 - 2018 for the longline main fleets. *Collective Volume of Scientific Papers, ICCAT*, 77, 317–329. https://www.iccat.int/Documents/CVSP/CV077_2020/n_3/CV077030317.pdf accessed 2025-06-02.

McKinney, Raiana, James Gibbon, Esther Wozniak, and Grantly Galland. (2020). Netting billions 2020: A global tuna valuation. <https://www.pewtrusts.org/-/media/assets/2020/10/nettingbillions2020.pdf> accessed 2025-05-29.

Medina, Antonio. (2020). Reproduction of atlantic bluefin tuna. *Fish and Fisheries*, 21, 1109–1119. <https://doi.org/10.1111/faf.12489>.

Ministero dell'agricoltura, della sovranità alimentare e delle foreste. (2024). Allegato 1, TONNO ROSSO - CAMPAGNA DI PESCA 2024. *Gazzetta Ufficiale della Repubblica Italiana*. https://www.gazzettaufficiale.it/atto/serie_generale/caricaDettaglioAtto/originario?atto.dataPubblicazioneGazzetta=

2024-05-21&atto.codiceRedazionale=24A02500&elenco30giorni=false

accessed 2025-05-29.

Ministère de la Transition écologique et de la Cohésion des territoires. (2024). Arrêté du 13 février 2024 établissant les modalités de répartition du quota de thon rouge (*Thunnus thynnus*) accordé à la France pour la zone «océan Atlantique à l'est de la longitude 45° O et Méditerranée» pour l'année 2024. *Journal officiel de la République française*. https://www.legifrance.gouv.fr/download/pdf?id=wSCtx11Gzpq9uW0cYXc7s-_AvWkqbw3aGTWSBldcbDg=.

Natale, Fabrizio, Maurizio Gibin, Alfredo Alessandrini, Michele Vespe, and Anton Paulrud. (2015). Mapping Fishing Effort through AIS Data. *PLOS ONE*, 10, 1–16. <https://doi.org/10.1371/journal.pone.0130746>.

Niu, Zixuan, Zhaojun Chen, Wei Yu, Haihong Guo, Ruichen Zhu, and Xiaoyi Zhang. (2024). Temporal and spatial variations in squid jigging catch efficiency in the oyashio extension region. *Fisheries Oceanography*, 33, e12692. <https://doi.org/10.1111/fog.12692>.

Orofino, Sara, Gavin McDonald, Juan Mayorga, Christopher Costello, and Darcy Bradley. (2023). Opportunities and challenges for improving fisheries management through greater transparency in vessel tracking. *ICES Journal of Marine Science*, 80, 675–689. <https://doi.org/10.1093/icesjms/fsad008>.

Palko, B. J., G. L. Beardsley, and W. J. Richards. (1981). Synopsis of the Biology of the Swordfish, *Xiphias gladius*, Linnaeus. Technical Report NMFS Circular 441 United States Department of Commerce, National Oceanic and Atmospheric Administration (NOAA) <https://www.fao.org/4/ap932e/ap932e.pdf>.

Papaconstantinou, C. and H. Farrugio. (2000). Fisheries in the Mediterranean. *Mediterranean Marine Science*, 1, 5–18. <https://doi.org/10.12681/mms.2>.

Piet, G. J., F. J. Quirijns, L. Robinson, and S. P. R. Greenstreet. (2006). Potential pressure indicators for fishing, and their data requirements. *ICES Journal of Marine Science*, 64, 110–121. <https://doi.org/10.1093/icesjms/fs1006>.

Posit team. (2024). *RStudio: Integrated Development Environment for R*. Posit Software, PBC Boston, MA <http://www.posit.co/>.

R Core Team. (2024). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing Vienna, Austria <https://www.R-project.org/>.

Relano, Veronica and Daniel Pauly. (2022a). Philopatry as a tool to define tentative closed migration cycles and conservation areas for large pelagic fishes in the pacific. *Sustainability*, 14. <https://doi.org/10.3390/su14095577>.

Relano, Veronica and Daniel Pauly. (2022b). Philopatry as a tool to define tentative closed migration cycles and conservation areas for large pelagic fishes in the pacific. *Sustainability*, 14. <https://doi.org/10.3390/su14095577>.

Russo, T., A. Parisi, and S. Cataudella. (2013). Spatial indicators of fishing pressure: Preliminary analyses and possible developments. *Ecological Indicators*, 26, 141–153. <https://doi.org/10.1016/j.ecolind.2012.11.002>.

Sales Henriques, Nuno, Karim Erzini, Jorge M.S. Gonçalves, and Tommaso Russo. (2024). Let's measure it: An approach of high-resolution estimates of bottom fixed net fishing effort at national level. *Fisheries Research*, 278. <https://doi.org/10.1016/j.fishres.2024.107118>.

Secretaría General de Pesca. (2024). Resolución de 8 de marzo de 2024, de la secretaría general de pesca, por la que se dispone la asignación de cuotas de atún rojo y publicación del censo específico de la flota autorizada para el ejercicio de la pesca del atún rojo creado por el real decreto 46/2019, de 8 de

febrero, por el que se regula la pesquería de atún rojo en el atlántico oriental y mediterráneo. *Boletín Oficial del Estado*, 32846–32871. <https://www.boe.es/boe/dias/2024/03/20/pdfs/BOE-A-2024-5591.pdf> accessed 2025-05-29.

Shea, Brendan D., Austin J. Gallagher, Lindsay K. Bomgardner, and Francesco Ferretti. (2023). Quantifying longline bycatch mortality for pelagic sharks in western pacific shark sanctuaries. *Science Advances*, 9. <https://doi.org/10.1126/sciadv.adg3527>.

Shepperson, Jennifer L, Niels T Hintzen, Claire L Szostek, Ewen Bell, Lee G Murray, and Michel J Kaiser. (2017). A comparison of VMS and AIS data: the effect of data coverage and vessel position recording frequency on estimates of fishing footprints. *ICES Journal of Marine Science*, 75, 988–998. <https://doi.org/10.1093/icesjms/fsx230>.

Taconet, M., D. Kroodsma, J. Gee, and J. A. Fernandes. (2019). Introduction of Global Atlas of AIS-based fishing activity. In *Global Atlas of AIS-based fishing activity - Challenges and opportunities* (M. Taconet, D. Kroodsma, and J. A. Fernandes, eds.) FAO, Rome <http://www.fao.org/3/ca7012en/ca7012en.pdf> accessed 2025-05-29.

Tidd, Alex N., Youen Vermand, Paul Marchal, John Pinnegar, Julia L. Blanchard, and E. J. Milner-Gulland. (2015). Fishing for Space: Fine-Scale Multi-Sector Maritime Activities Influence Fisher Location Choice. *PLOS ONE*, 10, 1–14. <https://doi.org/10.1371/journal.pone.0116335>.

Tracey, Sean R., Barrett W. Wolfe, Klaas Hartmann, Julian Pepperell, and Sam M. Williams. (2023). Movement behavior of swordfish provisions connectivity between the temperate and tropical southwest pacific ocean. *Scientific Reports*, 13. <https://doi.org/10.1038/s41598-023-38744-z>.

Usai, Maria Pina. (2022). *Tuna: underwater natural and cultural heritage. The*

Tunèa case study, a project for the re-connection between coastal community and marine ecosystem 887–895. Firenze University Press <http://doi.org/10.36253/979-12-215-0030-1.84>.

Vespe, Michele, Maurizio Gibin, Alfredo Alessandrini, Fabrizio Natale, Fabio Mazzarella, and Giacomo C. Osio and. (2016). Mapping eu fishing activities using ship tracking data. *Journal of Maps*, 12, 520–525. <https://doi.org/10.1080/17445647.2016.1195299>.

Welch, Heather, Tyler Clavelle, Timothy D. White, Megan A. Cimino, David Kroodsma, and Elliott L. Hazen. (2024). Unseen overlap between fishing vessels and top predators in the northeast pacific. *Science Advances*, 10. <https://doi.org/10.1126/sciadv.ad15528>.

Young, Jock W., Brian P.V. Hunt, Timothée R. Cook, Joel K. Llopiz, Elliott L. Hazen, Heidi R. Pethybridge, Daniela Ceccarelli, Anne Lorrain, Robert J. Olson, Valerie Allain, Christophe Menkes, Toby Patterson, Simon Nicol, Patrick Lehodey, Rudy J. Kloster, Haritz Arrizabalaga, and C. Anela Choy. (2015). The trophodynamics of marine top predators: Current knowledge, recent advances and challenges. *Deep Sea Research Part II: Topical Studies in Oceanography*, 113, 170–187. <https://doi.org/10.1016/j.dsr2.2014.05.015>. Impacts of climate on marine top predators.

Zhang, Chongliang, Yong Chen, Binduo Xu, Ying Xue, and Yiping Ren. (2022). The dynamics of the fishing fleet in china seas: A glimpse through ais monitoring. *Science of The Total Environment*, 819. <https://doi.org/10.1016/j.scitotenv.2022.153150>.

A Supplementary Material

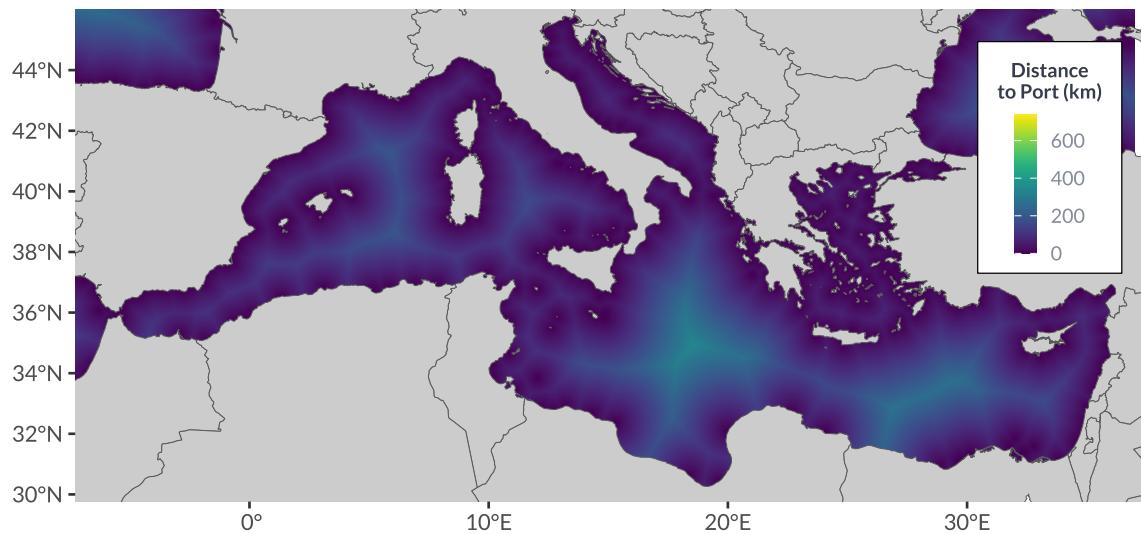


Figure S1: Distance to port in km throughout the Mediterranean.

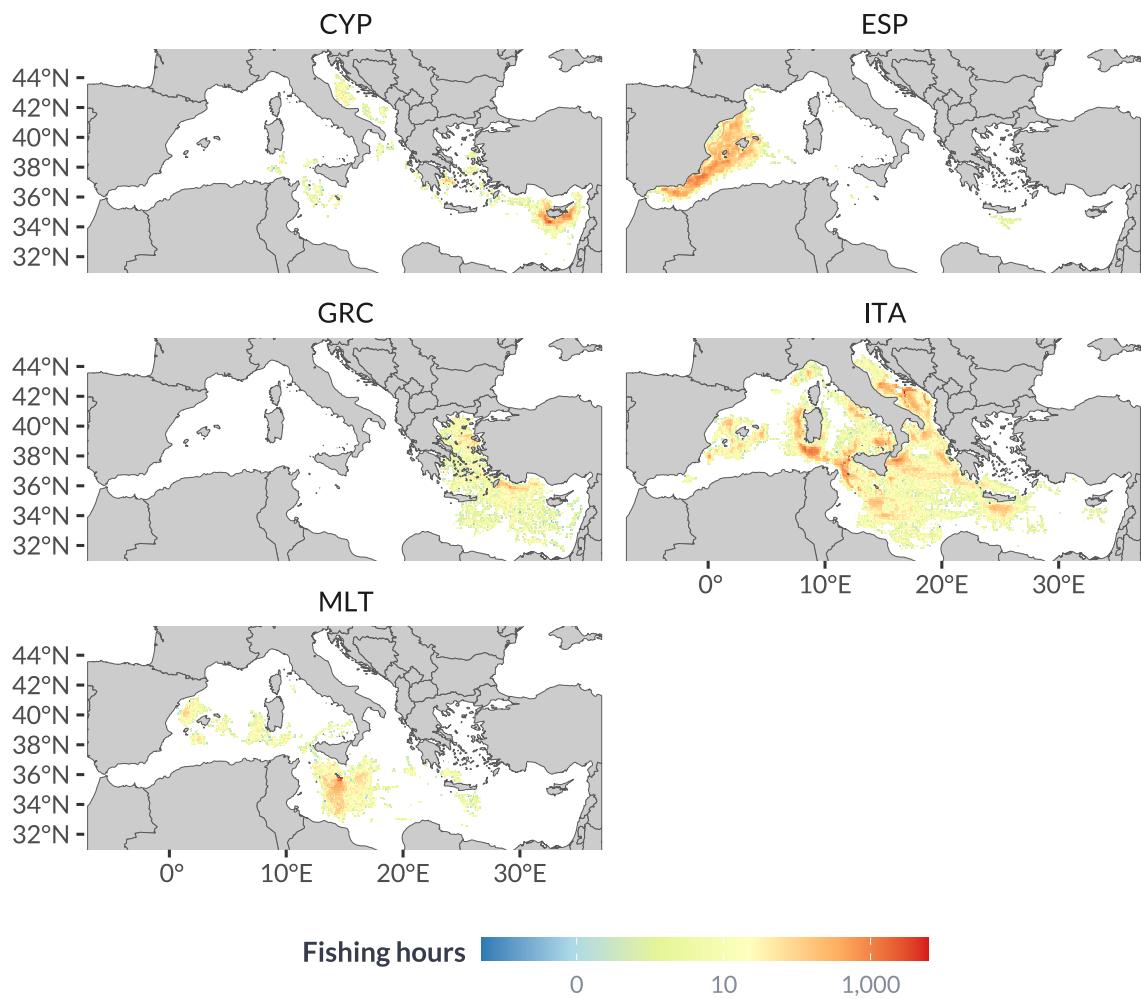


Figure S2: Sum of longline fishing hours per flag country (2015-2024). CYP = Cyprus; ESP = Spain, GRC = Greece, ITA = Italy, MLT = Malta

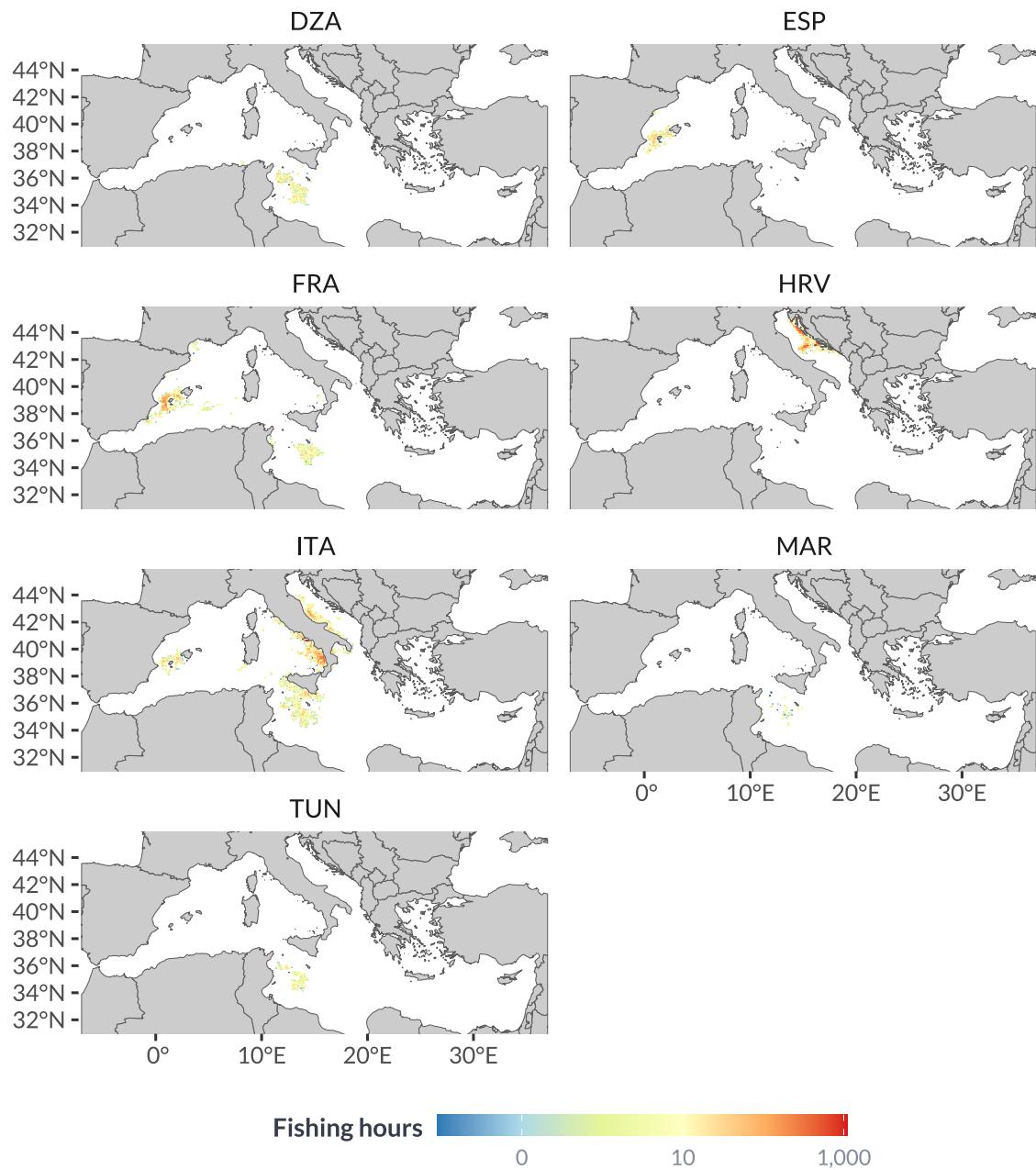


Figure S3: Sum of purse seine fishing hours per flag country (2015-2024). DZA = Algeria, ESP = Spain, FRA = France, HRV = Croatia, ITA = Italy, MAR = Morocco, TUN = Tunisia.

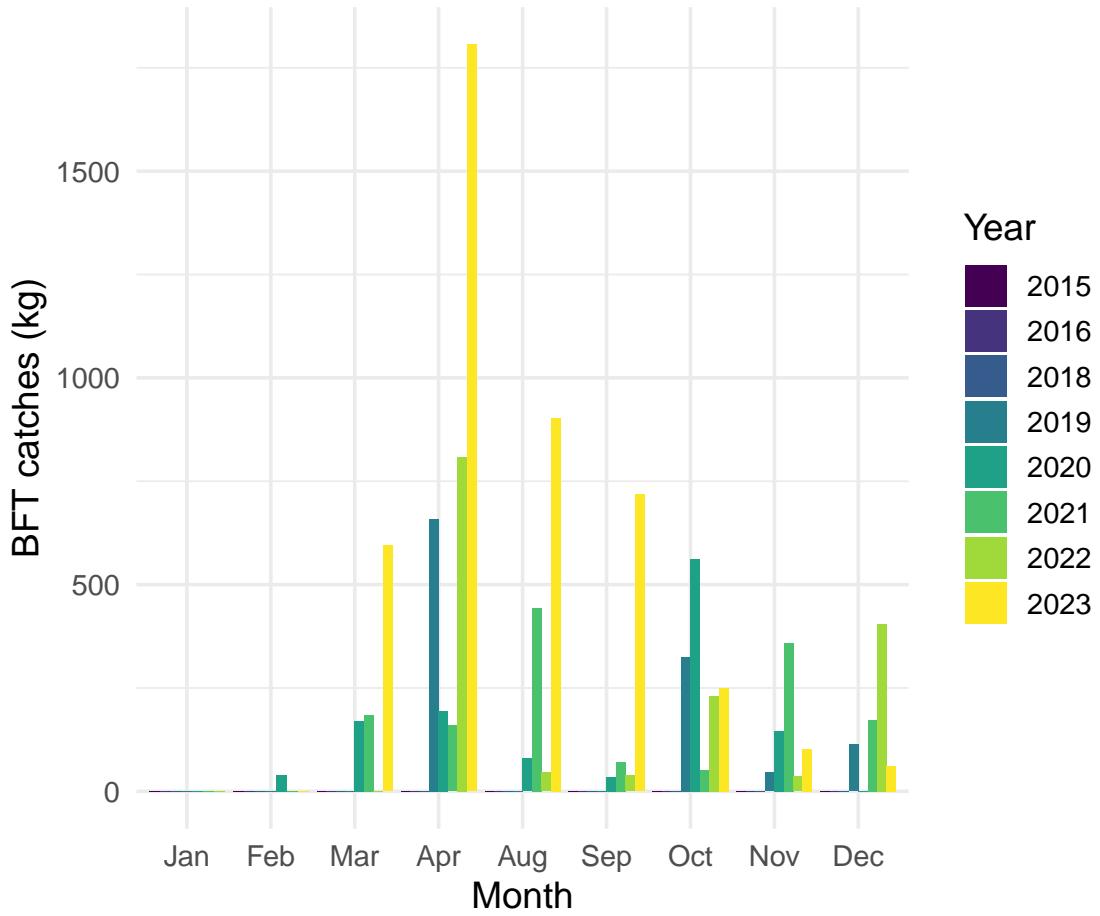


Figure S4: Reported bluefin tuna (BFT) purse seine catches of Croatia. May, June, and July were excluded as they are orders of magnitude higher and did not allow for graphic comparison between months with lower activity.